



1 **Changing characteristics of atmospheric CH₄ on the Tibetan Plateau,**
2 **records from 1994 to 2017 at Mount Waliguan station**

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18 **Abstract.** A 24-year long-term observation of atmospheric CH₄ was presented at
19 Mt. Waliguan (WLG) station, the only WMO/GAW global station in inland of Eurasia.
20 Overall, during 1994-2017, continuously increase of atmospheric CH₄ was observed at
21 WLG with yearly growth rate of 5.1 ± 0.1 ppb yr⁻¹, although near-zero and even
22 negative growth appeared in some particular periods, e.g., 1999-2000, and 2004-2006.
23 The average CH₄ mole fraction was only 1805.8 ± 0.1 ppb in 1995, but unprecedented
24 elevated ~ 100 ppb and reached a historic high of 1903.8 ± 0.1 ppb in 2016. The seasonal
25 averages of atmospheric CH₄ at WLG were ordered by summer, winter, autumn and
26 spring, and the correlation slopes of $\Delta\text{CO}/\Delta\text{CH}_4$ showed a maximum in summer and
27 minimum in winter, which was almost opposite to other sites in the northern
28 hemisphere, e.g., Mauna Loa, Jungfraujoch, and was caused by regional transport.
29 Strong potential sources at WLG were predominately identified in northeast (cities, e.g.,
30 Xining, Lanzhou) and southwest (the Northern India), and air masses from west and
31 northwest regions were accompanied with higher CH₄ mole fractions than that from
32 city regions.

33

34 What is interesting is that obviously changes appeared in different observing periods.
35 Generally, i) the amplitudes of diurnal or seasonal cycles were continuously increasing
36 over time, ii) the wind sectors with elevated CH₄ moved from ENE-...-SSE sectors in
37 early periods to NNE-...-E sectors (city regions) in later years, iii) the area of source
38 regions was increasing along with the years, and strong sources gradually shifted from
39 northeast to southwest, iv) the annual growth rates in recent years (e.g., 2013-2016)
40 were significantly larger than that in early periods (e.g., 1998-2012). We conclude that
41 the site was more and more affected by regional sources along with the time. Northern
42 India was possibly becoming the strongest source area to WLG rather than city regions
43 before. The case study in the Tibetan Plateau showed that the atmospheric CH₄
44 observed in Qinghai-Tibetan Plateau changed not as expected, the annual growth rate
45 was even larger than that in city regions in some period (e.g., 7.3 ± 0.1 ppb yr⁻¹ in 2013-



46 2016). It is unambiguous that the anomalously fluctuations of atmospheric CH₄ in this
47 region are a warning to the world, its increasingly annual growth rate may be a
48 dangerous signal to global climate change.



49 1 Introduction

50 Since the pre-industrial era, the emissions of greenhouse gases (GHGs) have increased
51 continuously, and larger absolute increases were found in recent years with the
52 concentration higher than ever now (WMO, 2019). The GHGs could perturb the
53 infrared radiation balance, trap the heat in the atmosphere, which contributes to global
54 warming, melting glaciers, extreme weather events and many other global climate
55 changes (IPCC, 2014). The recent 30-years from 1983 to 2012 were the warmest of the
56 last 800-years in the Northern Hemisphere, and half of the rising surface temperature
57 was due to increased GHGs emissions (IPCC, 2014). As one of the most important
58 GHGs, the global warming effect of methane (CH_4) is just after carbon dioxide (CO_2)
59 (Etminan et al., 2016). It has an 8-12 years atmospheric lifetime (Battle et al., 1996),
60 with the global warming potential of ~ 23 times greater than CO_2 over a 100 year
61 horizon (Weber et al., 2019). About 17% of radiative forcing by long-lived greenhouse
62 gases was contributed by CH_4 during 1750-2016 (Etminan et al., 2016). Since the
63 beginning of the industrial era, the concentration of CH_4 is rapidly increased because
64 of the influence of anthropogenic activities (Saunio et al., 2016). The result by the
65 analyses of ice cores in Antarctica showed that the atmospheric concentration of CH_4
66 has reached unprecedented over the last 0.8 million years (IPCC, 2014).

67 In the beginning of 1990s, CH_4 concentration appeared a decreasing trend in global
68 scale. However, high growth rates were found in 1998, which was possibly due to the
69 higher global mean temperature (Dlugokencky et al., 1998; Nisbet et al., 2014).
70 Subsequently, a low growth rate sustained over 1999 to 2006, except for special years
71 (2002/2003) with El Niño events (Dlugokencky et al., 1998). The annual growth rates
72 dropped from $\sim 12 \text{ ppb yr}^{-1}$ to near 0 from late 1980s to 1999-2006 (Nisbet et al., 2019).
73 But thereafter, the atmospheric CH_4 concentration keeps rising from 2007. During
74 2007-2013, the annual growth rate of methane was $5.7 \pm 1.2 \text{ ppb yr}^{-1}$. After 2013, the
75 atmospheric CH_4 grew even at rates not observed after 1980s, such as $12.7 \pm 0.5 \text{ ppb}$
76 yr^{-1} in 2014 and $10.1 \pm 0.7 \text{ ppb yr}^{-1}$ in 2015 (Nisbet et al., 2016, 2019). The overall



77 global growth rate was 7.1 ppb yr^{-1} in recent 10 years (WMO, 2019). And the not
78 expected increase since 2007 would make it difficult to meet the targets of carbon
79 emission reduction in the future. The World Meteorological Organization/Global
80 Atmospheric Watch programme (WMO/GAW) annual greenhouse gas bulletin revealed
81 that globally averaged CH_4 mole fraction reached a new high with $1869 \pm 2 \text{ ppb}$ in 2018
82 (Rubino et al., 2019), which was $\sim 259\%$ of pre-industrial levels ($\sim 722 \text{ ppb}$ around 1750
83 C.E.) (Etheridge et al., 1998; WMO, 2019).

84 The atmospheric CH_4 is mainly emitted from natural sources (about 40%, e.g.,
85 ruminants and wetlands) and anthropogenic sources (about 60%, e.g., paddies, cattle
86 ranch, coal mine, fossil fuel and biomass burning) (Hausmann et al., 2016; Saunois et
87 al., 2016). Studies from GAW observations indicated that the causes of recent increase
88 were likely attributed to anthropogenic emissions at mid-latitudes in the northern
89 hemisphere and the wetlands in the tropics (WMO, 2019). The rapid development of
90 population growth, economic expansion and countries urbanization has led to more and
91 more fossil fuel production and consumption (e.g., the large-scale exploitation of
92 natural gas, oil and coal) and biomass burning, consequently large amounts of
93 anthropogenic CH_4 were emitted around the world in recent years (Galloway, 1989;
94 Streets and Waldhoff, 2000; Wang et al., 2002; Lin et al., 2014; Hausmann et al., 2016).
95 The recent carbon isotope study revealed that biogenic emissions might also have
96 driven CH_4 increase, including microbial sources whether from rice, paddies,
97 ruminants, termites, enteric fermentation or all of these (Nisbet et al., 2016; Schaefer et
98 al., 2016; Wolf et al., 2017). 90% of CH_4 destruction in the atmosphere are mainly from
99 the reaction with hydroxyl radicals (OH) (Vaghjiani and Ravishankara, 1991; Bousquet
100 et al., 2011), an important oxidant in the troposphere (Logan et al., 1981). Therefore,
101 the interannual variability of OH or the decline oxidative capacity of the atmosphere
102 may also cause the recently increased CH_4 growth rates (Rigby et al., 2017; Turner et
103 al., 2017).

104 To get accurate understanding of atmospheric CH_4 , a systematic observations



105 network would perform the best. Hundreds of CH₄ observation stations worldwide are
106 running under the framework of WMO/GAW. Since 1978, systematic measurements of
107 atmospheric CH₄ began around the world (Blake et al., 1982; Rasmussen and Khalil,
108 1984; Dlugokencky et al., 1994). On the northern slope of the Mauna Loa volcano,
109 Hawaii, there exists the first global station Mauna Loa (MLO), which was performed
110 about 3397m above sea level (a.s.l.) and far away from local sources and sinks. It has
111 the longest records of continuous atmospheric CH₄ observation (Keeling et al., 1976).
112 Later, many types of the sites were installed CH₄ observation system, such as the
113 Barrow (BRW), South Pole (polar site, SPO) (Dlugokencky et al., 1995), Cape Grim
114 (CGO) in Australia (coastal/island sites) (Pearman and Beardsmore, 1984),
115 Minamitorishima (MNM) in Japan (coastal/island sites) (Wada et al., 2007),
116 Jungfrauoch (JFJ) in Switzerland and Mount Waliguan in China (continental mountain
117 site) (Zhou et al., 2004; Loov et al., 2008). Even though, the exact causes of
118 significantly increased CH₄ emissions in past years are still remained unclear and
119 debated, especially for the anomalous periods with suddenly large growth, due to the
120 time and space sparsity of measurements and the crude model approaches, which
121 limited our understanding of the global variation of atmospheric CH₄ (Saunio et al.,
122 2019; Weber et al., 2019). As long as the reasons for rising CH₄ emissions contributed
123 by natural sources (e.g. wetlands), anthropogenic sources (e.g. fossil fuels), or climate
124 change feedbacks remain uncertainties, it will be impractical to predict CH₄ trends in
125 the future, and then to develop realistic management (Nisbet et al., 2019). Therefore, it
126 is essential to establish typical observing regions and perform long observations.

127 China has the largest anthropogenic CH₄ emissions in the world (Janssens-
128 Maenhout et al., 2019). Qinghai-Tibetan Plateau has an average altitude over 4000m
129 a.s.l., which has long been recognized as the roof of the world. By coincidence, the two
130 largest CH₄ source regions in the world (i.e. Eastern China and Northern India) trapped
131 the Tibetan Plateau in the middle (Zhang et al., 2011; Fu et al., 2012; Wilson and Smith,
132 2015). Under the characteristics of special geographical conditions, lower population



133 density, rarely industrial activities and high sensitivity to external disturbances, the
134 Tibetan Plateau is undoubtedly one of ideal regions to observe continual CH₄ signal
135 (Zhou et al., 2005; Fu et al., 2012; Zhang et al., 2013). Most of the previous studies
136 reported the short-term CH₄ variations in China and concluded the importance of long-
137 term observation (Cai et al., 2000; Zou et al., 2005; Wang et al., 2009; Fang et al., 2013),
138 which is of great value to enhance the understanding of the global carbon cycle (Yuan
139 et al., 2019). As the rapid development of China and India, the year to year difference
140 as well as the sources and sinks of CH₄ on the Tibetan Plateau might change
141 significantly over time. Since 1994, in-situ measurements of atmospheric CH₄ have
142 been launched at Mt.Waliguan (WLG) station. To study the long-term variations of
143 atmospheric CH₄ at the WLG and get a new insight of its characteristics in the inland
144 of the Eurasia, in this study, we evaluated the performance of a 24 year long-term in
145 situ observations of CH₄ at Mt.Waliguan baseline observatory, which is the longest time
146 observing records in China. Temporal patterns, annual variations, long-term trends, air
147 mass transports, spatial distribution of potential sources were analyzed. In addition, the
148 case studies combining atmospheric CO measurements and a separate analysis between
149 the Tibetan Plateau and the city regions were performed to constrain the contribution
150 of anthropogenic emissions.

151

152 **2 Methodology**

153 **2.1 Measurement site**

154 The Mt.Waliguan (WLG, 36.28° N, 100.09° E, 3816m a.s.l.) station is situated at the
155 edge of northeastern of the Tibetan (Qinghai-Xizang) Plateau, which was in remote
156 western China and isolated from populated and industrial regions (Fig. 1). WLG was
157 the only WMO/GAW global background station in Eurasia and running by the China
158 Meteorological Administration (CMA). The surrounding areas of the site are pristine
159 with sparse vegetation, naturally arid and semi-arid grasslands. Small farms with yak



160 and sheep are in the valley. Two adjacent large cities Xining (~2.2 million populations)
161 and Lanzhou are located about 90km northeast and 260km east of the station,
162 respectively. The Longyangxia hydroelectric station (~380 km²) is located
163 approximately 13km south to southwest of the WLG. The predominant winds at WLG
164 are mainly from southwest and east in winter and summer, respectively (Zhou et al.,
165 2004; Zhang et al., 2011), which is controlled by Tibet Plateau monsoon.
166 Simultaneously, diurnal variations of vertical winds at WLG is influenced by mountain-
167 valley breezes, where upslope flow brings heated air masses from the boundary layer
168 to the site in daytime and downslope flow results in cool air masses transport from
169 mountain peak to the site. Under this unique location, the observation at WLG could
170 obtain essential information on CH₄ sources and sinks from Eurasia (Zhou et al., 2005;
171 Zhang et al., 2013).

172

173 2.2 Instrumental setup

174 Atmospheric CH₄ has been measured quasi-continuously using a HP 5890 gas
175 chromatograph (GC) equipped with a flame ionization detector (FID) since July 1994,
176 and an Agilent 6890N GC equipped with a FID since June 2008. Both of the systems
177 used the same sampling procedures. A Cavity Ring Down Spectroscopy system (Picarro
178 G1301) began in January 2009 and the instrument was upgraded to Picarro G2401 in
179 2015. Ambient air is delivered to the above systems at about 5 L/min by a KNF
180 Neuberger N2202 vacuum pump via a dedicated 0.95 cm o.d. sample line from an 80m
181 intake line attached to an 89m steel triangular tower located approximately 15m from
182 the main observatory. The residence time of the ambient air from the top of the tower
183 to the instrument is 30 s. The ambient air is first passed through a 7 mm stainless steel
184 membrane filter located upstream of the pump and then (after the pump) passed through
185 a pressure relief valve set at 1 atm to release excess air and pressure. The ambient air is
186 then dried to a dew point of approximately -60°C by passing it through a glass trap
187 submerged in a -70°C methanol bath. All standard gases supplied to the instruments are



188 from pressurized 37.5 L treated aluminum alloy cylinders fitted with high-purity, two-
189 stage gas regulators. Stainless steel tubing (0.32 cm o.d., 0.22 cm i.d.) is used for the
190 standard gas sample line and the ambient sample line after the cold trap. The automated
191 sampling module equipped with a VICI 8 ports valve is designed to sample from
192 separate gas streams (standard tanks and ambient air). According to the comparability
193 target of WMO/GAW program (WMO, 2019), methane mole fractions are referenced
194 to a Working High standard (WH) and a Working Low standard (WL). Additionally, a
195 calibrated cylinder filled with compressed ambient air is used as a Target gas (T) to
196 check the precision and stability of the system routinely. Diagram of the observing
197 system during different periods could be seen at Zhou et al. (2004) and Fang et al.
198 (2013). Here, we focus on the longest continuous measurements of CH₄ from August
199 1994 to May 2017 at WLG. Data gaps in limited periods are because of the malfunction
200 of instrument and the maintenance of the sampling system.

201 The records of CO in this study was initially observed by an RGA-3 gas
202 chromatograph (GC) equipped with an HgO reduction detector (Trace Analytical Inc.)
203 since 1994. An automated sampling module was designed to sample from ambient air
204 and a series 9 standards. Detailed diagram of the system was described by Zhang et al.
205 (2011). Since 2010, the CO has been measured by the Cavity Ring Down Spectroscopy
206 instrument (Picarro G1302 and G2401 since 2015). The scale for all of the CO
207 measurement were further updated to WMO X2014A.

208

209 **2.3 Data processing**

210 Most on-site CH₄ observations were unavoidably influenced by local sources and other
211 complex conditions (e.g., traffic transportation, various topography). As a result, the
212 records cannot fully represent the regional atmospheric CH₄ in well-mixed conditions
213 (Liu et al., 2019). To precisely get representative regional records, we exclude CH₄
214 measurements influenced by local sources adjacent to the site (e.g., agricultural fields,
215 cities, traffic emissions). The hourly CH₄ data were classified as Local/Regional events



216 through the meteorological approach, which was based on essential meteorological
217 information, similar to previous studies by Zhou et al. (2004) and Liu et al. (2019). In
218 this study, the CH₄ records associated with surface wind from selected sectors (i.e.
219 NNE-...-ENE in spring, NE-...-SE in summer, NE-...-ESE in autumn, and NE-ENE in
220 winter) were flagged as local representative. Subsequently, we further rejected portion
221 of daytime records to minimize the effect of human activities (e.g., rush hours),
222 including 9:00-13:00 LT (local time) in spring and summer, 9:00-14:00 LT in autumn,
223 10:00-17:00 LT in winter. Finally, we filtered CH₄ data into local events when the
224 surface wind speed was less than 1.5 m s⁻¹ to minimize the very local accumulation.

225 To understand the influence of local surface wind, the hourly CH₄ data was
226 calculated versus 16 horizontal wind directions (Fang et al., 2013). In this study, we
227 used the ‘polarPlot’ function located in the ‘openair’ package of the statistical software
228 R (R Core Team, 2020). It shows the bivariate (i.e. wind speed and wind direction)
229 polar plot of CH₄ concentrations, and the concentrations are calculated as a continuous
230 surface by modelling using smoothing techniques (Carslaw et al., 2006; Diederich,
231 2007). Also, conditional probability function (CPF) was used to detect the probability
232 of which wind directions are dominated by high CH₄ mole fractions (Uria-Tellaetxe and
233 Carslaw, 2014). In order to study the pollution transport pathways of air masses at
234 WLG, the cluster analysis of 3 days back trajectories was applied using the Hybrid
235 Single-Particle Lagrangian Integrated Trajectory (HYSPPLIT) dispersion model
236 (Draxier and Hess, 1998; Rousseau et al., 2004) on the strength of gridded
237 meteorological data (2004-2017) from the National Oceanographic and Atmospheric
238 Administration’s Air Resources Laboratory (NOAA-ARL). The trajectories in four
239 months, including January, April, July and October, were calculated to represent the
240 seasons of winter, spring, summer and autumn, respectively. The spatial source
241 distributions of annual CH₄ were analyzed using the Potential Source Contribution
242 Function (PSCF) approach, which computed the conditional probability of the
243 residence times of air parcels with greater concentration than threshold transport to the



244 exactly receptor site (Ashbaugh et al., 1985). In this study, PSCF value was calculated
245 in 0.5×0.5 -degree grid cell (i, j) :

$$246 \quad \text{PSCF}_{ij} = m_{ij} / n_{ij} \quad (1)$$

247 n_{ij} represents the number of endpoints that terminate in the ij th grid cell, while the
248 number of trajectories with concentration exceed the threshold value was defined as m_{ij}
249 (Polissar et al., 1999). In order to reduce the abnormal influence of small n_{ij} values in
250 some grid cells, PSCF_{ij} was further computed by an arbitrary weighting function W_{ij} as
251 below.

$$252 \quad W_{ij} = \left\{ \begin{array}{ll} 1.00 & 3n_{ave} < n_{ij} \\ 0.70 & 1.5n_{ave} < n_{ij} \leq 3n_{ave} \\ 0.42 & n_{ave} < n_{ij} \leq 1.5n_{ave} \\ 0.05 & n_{ij} \leq n_{ave} \end{array} \right\} \quad (2)$$

253 W_{ij} represents the weight of cell (i, j) , n_{ij} is the number of trajectory endpoints that fall
254 in the ij th grid cell, while the n_{ave} shows the mean number of the endpoints in all grid
255 cells.

256 In order to fill the data gaps so as to evaluate the long-term CH_4 trend, we applied
257 the curve fitting approach by Thoning et al. (1989). We also calculated the trend curve
258 that excluded the influence of seasonal variation, and then got the annual growth rates
259 of the average of the first derivative of the trend curve. The function consists of the
260 polynomial part and the annual harmonics part:

$$261 \quad f(t) = a_0 + a_1 t + a_2 t^2 + \dots + a_{(k-1)} t^{(k-1)} + \sum_{n=1}^{nh} c_n [\sin(2n \pi t) + \varphi_n] \quad (3)$$

262 'k' represents the number of polynomial part. 'nh' is the number of harmonics part. We
263 applied $k = 3$ polynomial terms (a quadratic) for multi-year trends and $nh = 4$ yearly
264 harmonics for seasonal cycles in this study. The fast Fourier transform (FFT) was
265 utilized to smooth the fitting residuals (Press et al., 1992).

266 The significant difference test was applied by the 'scheirerRayHare' function in
267 the 'rcompanion' package of R software, which is a non-parametric test for two-way
268 ANOVA analysis. And the multiple comparison was used by Wilcoxon rank sum test



269 with R (R Core Team, 2020). For the correlation analysis between CH₄ and CO, we
270 obtained the detrended time series of CH₄ and CO from 2004-2017 based on the method
271 by Thoning et al. (1989). The detrended values are denoted as ΔCH_4 and ΔCO . To
272 obtain the correlation slopes of ΔCH_4 and ΔCO accurately, a rolling linear regression
273 was applied to ΔCH_4 and ΔCO time series by the 'roll_lm' function in 'roll' package of
274 R (R Core Team, 2020). We successively moved a 24-h time window by 1 h over the
275 whole time series. Similar to the study by Tohjima et al. (2014), we set 3 criteria to
276 achieve a better quality control of the slopes. When (i) the number of CH₄ record is less
277 than 5 in 24 hours, (ii) the coefficient variation of the correlation slope is more than
278 15%, (iii) the absolute value of the correlation slope is less than 0.8 ($|R| < 0.8$), the
279 correlation slopes were identified as statistically insignificant and inaccurately and were
280 rejected. In order to understand the year to year variations, we further analyzed the
281 different periods over 1994-2017. The entire CH₄ time series were divided according
282 to the significant stages or the critical time period of atmospheric CH₄ variations from
283 previously studies (Zhou et al., 2004; Fang et al., 2013; Zhang et al., 2013; Nisbet et
284 al., 2019; WMO, 2020), appropriately every five years a period. Unless special notes,
285 the average values in this study were presented with 95% confidence intervals (CIs).

286

287 **3 Results**

288 **3.1 Diurnal variations**

289 Generally, distinct diurnal cycles were observed in four seasons during 1994-2017. The
290 CH₄ mole fraction increased from early morning and reached the maximum at noon and
291 a trough in late afternoon (Fig. 2f). However, differences also existed from different
292 seasons. In spring, the atmospheric CH₄ apparently increased from 9:00 to 13:00 LT
293 with the daily amplitude of 5.7 ± 2.4 ppb. In summer, the elevated CH₄ also appeared
294 during 9:00-13:00 LT at noon, and the daily amplitude was 4.3 ± 2.6 ppb. In autumn,
295 the diurnal variation showed the mean amplitude of 4.5 ± 2.4 ppb, significantly elevated



296 CH₄ reached at 9:00-13:00 LT with one peak at noon. In winter, largely increasing
297 presented in the daytime at 9:00-17:00 LT, with the largest amplitude of 6.2 ± 2.4 ppb
298 among four seasons. For the diurnal variation over the whole monitoring period, the
299 highest CH₄ mole fraction was observed in winter and the minimum value was found
300 in spring (Fig. 2f).

301 Different patterns for diurnal CH₄ cycles were also found over different periods.
302 In 1994-1997 and 1998-2002, the CH₄ mole fractions in winter were apparently higher
303 than the other seasons (Fig. 2a-b). But its value was the highest in summer during the
304 period of 2003-2007, 2008-2012 and 2013-2017 (Fig. 2c-e). The atmospheric CH₄
305 values in winter were gradually falling behind the other seasons, and the gaps among
306 different seasons were increasing, especially for summer. Before 2002, diurnal cycles
307 in four seasons were ambiguous (Fig. 2a-b), but significant diurnal variations appeared
308 afterwards (Fig. 2c-e). The peak to trough amplitude almost increased along with the
309 time in almost all seasons. For example in spring, the amplitude was 6.5 ± 3.0 , $4.7 \pm$
310 1.8 , 5.6 ± 2.6 , 6.2 ± 2.4 and 6.8 ± 3.4 ppb over 1994-1997, 1998-2002, 2003-2007,
311 2008-2012 and 2013-2017, respectively (Table S1).

312

313 **3.2 The impact of local surface winds**

314 As observed by the previous short-term variations, the atmospheric CH₄ at WLJ was
315 significantly influenced by local surface wind from northeast to southeast sectors (Fig.
316 3f). Slight differences were also found among seasons. In spring, the atmospheric CH₄
317 was enhanced by 2.5-6.5 ppb compared to the seasonal average (1839.7 ± 1.4 ppb)
318 when the wind was originating from NNE-NE-ESE-E sectors. In summer and autumn,
319 the wind from NNE-NE-ESE-E-ESE induced higher CH₄ mole fractions, with
320 enhancement of about 3-9.5 ppb and 4 to 18 ppb, respectively. In winter, the CH₄ mole
321 fractions significantly elevated from the wind sectors that same as those found in spring,
322 with value of 7-21 ppb than seasonal average (1854.5 ± 4.8 ppb). Relatively, the
323 amplitude of enhancements in winter and autumn were larger than those in spring and



324 summer.

325 What interesting is that wind sectors elevating CH₄ mole fractions vary in different
326 periods. The early periods (i.e. 1994-1997 and 1998-2002) were different from the
327 recent periods (2003-2007, 2008-2012 and 2013-2017). The elevated CH₄ was
328 predominately from about ENE-E-ESE-SE-SSE sectors in early years (Fig. 3a-b), but
329 evolved to NNE-NE-ENE-E sectors in later years (Fig. 3c-e). Furthermore, the
330 amplitude of enhancements was almost increasing continuously along with the time.
331 For example, in autumn, the maximum CH₄ mole fractions were from E, ENE, ENE,
332 NE and ENE sectors in 1994-1997, 1998-2002, 2003-2007, 2008-2012 and 2013-2017,
333 with the successively increasing enhancements of 8.6, 12.1, 14.7, 16.8 and 19.7 ppb,
334 respectively.

335 We applied the CPF to hourly CH₄ and CO data by considering intervals of entire
336 data percentiles including 0-20, 20-40, 40-60, 60-80 and 80-100 to draw the CPF polar
337 plot. It is clear that different sources only affected CH₄ mole fractions on different
338 percentile range. For example, for most wind speed-directions the CPF probability of
339 CH₄ being greater than the 60th percentile was tending to zero (Fig. 4). And it is apparent
340 that most sources contributed to the less than 60th percentiles of CH₄ mole fraction (e.g.
341 40-60) (Fig. 4). The specific sources were prominent for specific percentile ranges. The
342 wind from the southwest and southeast was important on the cases of the higher
343 percentiles, resulted in the highest CH₄ mole fractions of 1849-1872 ppb for 60-80
344 percentile and 1872-2031 ppb for 80-100 percentile (Fig. 4). It's more obvious that the
345 CO showed gradually shifted sources with the increase of percentile ranges. The areas
346 where the CPF probabilities were higher is to the NW-SW sectors when the percentages
347 ranged from 0 to 40th. Nevertheless, when the percentages were larger than 60th, the
348 high probability areas completely moved to NE-SE sectors (Fig. 4).
349



350 3.3 Air mass pathways and potential source distributions

351 3.3.1 Air mass transports

352 Figure 5 illustrates the cluster analysis to the 3-day back trajectories to WLG during
353 2004-2017. In spring, the majority of the air masses were from west and northwest
354 regions, which accounted for about 24% (cluster 3) and 44% (cluster 5) of total
355 trajectories (Fig. 5a). These air masses were also accompanied with higher CH₄ mole
356 fractions than those from east and northeast regions, i.e. cluster 1 (13.3% of total) and
357 cluster 4 (11.69%) (Table 1). The largest enhancement was ~18 ppb (relative to spring
358 average) by cluster 3. In summer, 45% of air masses (cluster 1) were from eastern
359 regions. But the high CH₄ mole fractions were on cluster 2 and cluster 5 from northwest
360 and west regions, although low percentages were found (cluster 2: 26%; cluster 5: 7%)
361 (Fig. 5b). The highest CH₄ mole fraction was associated with cluster 2, with ~9 ppb
362 larger than the average in summer. In autumn, large proportion of air masses was
363 originating from west and southwest station, such as cluster 2 (49%) and cluster 3 (32%)
364 (Fig. 5c). The highest CH₄ was from cluster 3 with enhancement of ~4 ppb than the
365 seasonal average. Similar to autumn, the air masses were primarily from northwest in
366 winter, including northwest cluster 3 (59%) and southwest cluster 1 (34%) regions (Fig.
367 5d). The highest CH₄ mole fractions was on cluster 1 with the enhancement of ~7 ppb
368 than the average value.

369

370 3.3.2 Spatial distribution of potential source regions

371 In this study, the potential sources were analyzed over different periods, i.e. 2004-2007,
372 2008-2012 and 2013-2017. Generally, the strong potential sources were located at the
373 northeast to southeast of the station, especially in summer, but the source regions
374 differed in various seasons as well as years (Fig. 6). The regions of potential source in
375 spring (Fig. 6a-c) and winter (Fig. 6j-l) was obviously larger than that in summer (Fig.



376 6d-f) and autumn (Fig. 6g-i). There were also trends for the CH₄ source regions along
377 with years: i) the area of potential source regions was increasing with the years, and ii)
378 the location of strong potential sources changed along with the time. For example in
379 autumn and winter, the strength of CH₄ sources were very strong in the southeast to
380 northeast during 2004-2007 (Fig. 6g & i), and then weakened in 2008-2012 (Fig. 6h & k),
381 and finally (i.e. 2013-2017), almost vanished in eastern regions but moved to southwest
382 with very large distribution area (Fig. 6i & l).

383

384 **3.4 Extracting the well-mixed ambient methane**

385 To precisely understand characteristics of atmospheric CH₄, e.g., seasonal cycle or
386 long-term trend, it is vital to identify the CH₄ records that were influenced by local
387 sources and sinks. In this study, we analyzed hourly CH₄ measurements during 1994-
388 2017, and 47.3% of CH₄ data were classified as regional representative, with the
389 average CH₄ mole fraction of 1847.9 ± 0.3 ppb. The local representative data was
390 obviously larger than regional events, with an average value of 1858.2 ± 0.4 ppb (Table
391 2). The proportion of regional events increased slightly before 2012, but significantly
392 reduced in recent years (e.g., 2013-2017). The filtered regional/local time series was
393 shown in Figure 7. It can be seen that the CH₄ mole fractions obviously increased from
394 1994 to 2017. The atmospheric CH₄ showed strong growth and displayed large
395 fluctuation at WLG (Fig. 7). In 1995, the average CH₄ mole fraction was only 1805.8
396 ± 0.1 ppb, however, the average value increased 98 ppb by the year of 2016 ($1903.8 \pm$
397 0.1 ppb) (Table 3).

398

399 **3.5 Correlation analysis between CH₄ and CO**

400 Because part of CH₄ and CO in atmosphere are from the same anthropogenic sources
401 (e.g., fossil fuel combustion), we calculated the regression slopes of $\Delta\text{CO}/\Delta\text{CH}_4$ from
402 2004 to 2017 (Fig. S1). Figure 8 presents the average seasonal cycles of the $\Delta\text{CO}/\Delta\text{CH}_4$



403 slopes. Generally, the slopes were larger in summer and lower in winter during the
404 observing period, except for 2004-2007 with the high slope in autumn. Additionally,
405 the regression slopes increased along with the time, which showed the maximum in
406 2013-2017 and the minimum in 2004-2007. For the year to year variations, the
407 $\Delta\text{CO}/\Delta\text{CH}_4$ slopes showed large fluctuations from 2004 to 2017 at WLG (Fig. 9). The
408 slopes showed decreasing trend during 2004-2007 but then increased from 2007 to
409 2010, and again decreased after 2010. In spring and summer, increasing trends appeared
410 again after 2014. The slopes in summer were almost the largest but the lowest in winter.

411

412 3.6 Variation of long-term records

413 3.6.1 Seasonal cycles

414 In order to further investigate the characteristics of atmospheric CH_4 , we divided the
415 CH_4 observations into two main regions according to the above analysis, including
416 geographical conditions, the effect of surface winds, the long-range transports and the
417 potential source distributions. The first region was covered the northeast to southeast
418 (NNE-...SE) of WLG, which was denoted as City Regions (CR). The second region
419 was located south to west (S-...-W) of the station and was well known Tibet (Qinghai-
420 Xizang) Plateau (TP) (Fig. S2). Accordingly, the hourly CH_4 records when the surface
421 wind coming from these sectors were divided into two subsets (i.e. TP and CR). The
422 long-term variations between the two regions as well as the total regional time series
423 (Total) were compared and analyzed to explore new sight of atmospheric CH_4 variation
424 at WLG.

425 Overall, at WLG, the seasonal averages of CH_4 were ordered by summer (1850.0
426 ± 0.3 ppb), winter (1847.4 ± 0.3 ppb), autumn (1844.4 ± 0.3 ppb) and spring ($1841.2 \pm$
427 0.3 ppb), except during 1994-1997 with the maximum in winter and minimum in
428 autumn (Fig. 10). Seasonal averages in CR were significantly different to that in TP and
429 also the entire regional data (Total). The seasonal average in TP was mostly higher than



430 that in CR, except for wintertime (Table S2). The atmospheric CH₄ in August was
431 mostly the maximum and the April was the minimum for the total regional time series
432 (Total), with the seasonal amplitude of 13.4 ppb. The peak to trough amplitude in CR
433 (~15 ppb) was higher than that in TP (~13 ppb) during 1994-2016. Additionally,
434 seasonal amplitudes indicated different trends between CR and TP. For CR, the seasonal
435 amplitude was firstly dropped and then increased along with time, which were similar
436 to the variation of total regional events (Total). But for TP, the amplitude displayed a
437 continuously increasing trend, with values of about 15.9, 19.3, 21.6, 23.4 and 22.4 ppb
438 in 1994-1997, 1998-2002, 2003-2007, 2008-2012 and 2013-2016, respectively.

439

440 3.6.2 Long-term trend

441 In the 1990s, the CH₄ growth rates were very low and even negative at WLG.
442 Subsequently, during 2002-2006, a steady period was found with a near-zero growth
443 rates. After 2007, the atmospheric CH₄ was raised significantly (Fig. 11a). In the year
444 of 1997/1998, 2000/2001, 2007/2008 and 2011/2012, larger amplitude of the growth
445 rates was found and strong growth appeared (Fig. 11b). The growth rates fluctuated
446 evenly with both positive and negative values before 2009. But almost all of the growth
447 rates showed a positive value after 2009. From 1990s to 2010s, three apparently
448 developing stages (i.e. highlighted green, blue and red blocks) were presented, the CH₄
449 mole fraction slightly decreased in 1998-2000 (green color), and then go through a
450 relative steady period in 2003-2006 (blue color), finally increased steadily after 2007
451 (red color) (Fig. 11).

452 The overall annual growth rates were 5.1 ± 0.1 ppb yr⁻¹ over 1994-2016 at WLG
453 (Table 4). However, the periodic annual growth rates were 4.6 ± 0.1 , 2.6 ± 0.2 , 5.3 ± 0.2 ,
454 7.6 ± 0.2 and 5.7 ± 0.1 ppb yr⁻¹ in 1994-1997, 1998-2002, 2003-2007, 2008-2012 and
455 2013-2016, respectively. The CH₄ growth rate in CR was significantly different from
456 that in TP (Fig. S3). In 1994-1997, 2003-2007 and 2013-2016, the growth rates in TP
457 were obviously larger than that in CR (Table 4). But in 2003-2007 and 2008-2012, the



458 CR showed higher annual growth rates. In addition, for the entire observing period (i.e.
459 1994-2016), the growth rates in both TP (5.2 ± 0.1 ppb yr⁻¹) and CR (5.0 ± 0.1 ppb yr⁻
460 ¹) were similar to the overall annual growth rates (Fig. S3).

461

462 3.7 Case study for air mass transport

463 As described above, the northeast and southeast city regions might act as strong
464 regional sources influencing the atmospheric CH₄ at WLG. Therefore, to analyze the
465 effect of long-distance transport of emissions from cities, we further excluded the
466 regional data by air mass transport, and the rest of regional records were denoted as
467 'TR'. We applied the monthly cluster analysis to hourly trajectories over 2005-2007,
468 2008-2012 and 2013-2017. The cluster results from city regions were presented in detail
469 in Fig. S4 and Table S3.

470 The proportion of trajectories from cities was 40.3%, 32.5% and 6.8% in 2005-
471 2007, 2008-2012 and 2013-2017, respectively. And about 22.8%, 35.6% and 38.4% of
472 the regional records were associated with air masses transport from city regions in
473 2005-2007, 2008-2012 and 2013-2017, respectively (Table 5). The average CH₄ value
474 when air masses transport from city regions (1863.0 ± 0.3 ppb) was obviously higher
475 than other sectors (1850.6 ± 0.2 ppb) (Fig. S5). However, after excluding the CH₄
476 records when air trajectories transport from city regions, the growth rates of TR in 2008-
477 2012 (10.1 ± 0.1 ppb yr⁻¹) and 2013-2017 (6.3 ± 0.1 ppb yr⁻¹) (Table 5) were higher
478 than the original regional data series (i.e. 7.6 ± 0.2 and 5.7 ± 0.1 ppb yr⁻¹) (Table 4).
479 And the overall growth rate for TR in 2005-2016 or 1994-2016 was still similar to
480 original data series (Total), no significant difference was found (Fig. S5).

481



482 4 Discussion

483 4.1 Anthropogenic emission on temporal patterns

484 In the early years, the daily cycles of atmospheric CH₄ at WLG were not distinct and
485 the amplitude was small (Fig. 2a-b), which were similar to the previous studies by
486 Zhang et al. (2013) and Fang et al. (2013). The ambiguous diurnal patterns indicated
487 that the local sources were weak at WLG in the past. The apparent diurnal cycles after
488 2000 may be attributed to the intense activities by human (e.g., grazing, fuel burning),
489 which was aggravated in daytime and weak in nighttime (Fang et al., 2013). The
490 meteorological conditions (e.g., diffusion and transport) could contribute to the
491 increasing CH₄ amplitudes. The WLG was remote from the populate center, the good
492 diffusion condition in daytime may bring high CH₄ mole fractions to the site. The
493 increasing amplitude (Table S1) and CH₄ mole fractions over time suggested that the
494 WLG was affected by increasingly local sources (e.g., human activities) (Zhou et al.,
495 2004). The maximum was found in summer in recent years (Fig. 2e) could also be
496 ascribed to the transport of anthropogenic sources by the meteorological factors. In
497 summer, the intensely herd or graze activities around WLG might enhance the regional
498 CH₄ emissions and hence contribute to the higher CH₄ mole fractions. The higher CH₄
499 mole fraction in winter in past years (Fig. 2a) was probably because of the way of
500 heating (e.g., large biomass burning) as well as the adverse diffusion conditions in cold
501 weather.

502 The previous study by Zhou et al. (2004) found that higher CH₄ mole fractions
503 appeared when the winds come from the ENE-E-ESE-SE sectors at WLG during 1991-
504 2002, which was similar to our study in similar period (Fig. 3a-b). Causes of the
505 elevated CH₄ from these wind sectors could be attributed to the large plantation of
506 highland barley as well as high population density in those areas (Fang et al., 2013).
507 Two largest cities Xining and Lanzhou are situated in northeast and east of WLG,
508 respectively, which could emit large amount of CH₄ by human activities. Besides,



509 previous studies on black carbon (BC) and carbon monoxide (CO) indicated that the
510 emissions from the Yellow River Canyon industrial area, ~500km away from northeast
511 of WLG may also donate to the high CH₄ values originating from ENE and NE sectors
512 (Zhou et al., 2003). In summer, the prevailing wind directions were from NE-...-ESE
513 sectors (~46%) (Fig. S6), and the CH₄ mole fractions were also higher in the related
514 sectors. However, in the autumn and winter, although the prevailing wind and high wind
515 speed were from SSW-...-W sectors (~ 40-50%) (Fig. S6), the high CH₄ mole fractions
516 were from almost the opposite wind sectors of NNE-...-ESE (Fig. 3f), which indicated
517 that strong local sources were distributed from northeast to southeast (city regions), and
518 even covered the emissions of natural sources. As time goes on, the wind sectors with
519 high CH₄ mole fractions changed and concentrated on ESE to ENE sectors, and the
520 amplitudes of enhancements were increasing, which further implied the effect of
521 stronger emissions from anthropogenic sources in city regions in recent years.

522

523 4.2 Pollutant sources regions

524 4.2.1 Sources regions

525 The air masses from east and northeast regions passed over the cities of Xining and
526 Lanzhou (capital of Gansu province), which is the populated center and industrial area
527 (Fig. 5). However, the highest CH₄ values was not observed when air mass was from
528 these sectors. Instead, high CH₄ mole fractions were frequently observed when air mass
529 from the northwest to southwest (Table 1). It was possibly due to that the air masses
530 from west and northwest had passed through the northwest of Qinghai province and the
531 central area of Xinjiang Uygur Autonomous Region (XUAR), where located Ge'ermu
532 urban area (the second largest city of Qinghai) with rapid industrial development,
533 natural gas and petroleum resources exploitation and large crops residue burning
534 (Zhang et al., 2013). Similar to the CPF percentile analysis (Fig. 4), the southwest or
535 northwest region away from the site may be also strong source regions.



536 Most potential source identified in northwestern regions (Fig. 6) was possibly due
537 to CH₄ emissions from the northwest Gansu province, the northwest Qinghai province
538 and the southeast of XUAR. The different source distribution by seasons could be
539 attributed to the effect of westerlies or the southeast monsoons (Zhou et al., 2004). The
540 obviously increasing source region was clear evidence for the strong effect of the
541 expansion of human activity. Moreover, the pattern of source region moved from the
542 east to the southwest, especially in autumn and winter, indicated that the southwest
543 away from the WLG, e.g., Northern India, were gradually becoming a strong CH₄
544 source region. India has abundant cattle as well as extensive large-scale coal mining,
545 large amount of CH₄ emissions may transport from northern India to the northeastern
546 Tibetan Plateau (Fig. 6i & l). The air mass transport result (Fig. 5d) also support the
547 result that the southwest air masses (cluster 1) contributed the highest CH₄ mole
548 fractions. The studies of atmospheric Hg at WLG by Fu et al. (2012) also revealed this
549 phenomenon.

550

551 4. 2. 2 Different sources between CH₄ and CO

552 The percentile polar plot clearly showed the specific distribution of different CH₄ mole
553 fractions (Fig. 4). The result revealed that most areas around WLG contributed to low
554 CH₄ mole fractions, the southeast and southwest of the site exist two strong source
555 regions. It is of great possible that the anthropogenic emission from cities (e.g.,
556 Lanzhou, Chengdu, etc.) was the only cause for high values in the southeast, and the
557 southwest region away from WLG was possibly due to sources from other countries,
558 such as India. Unlikely to the CH₄, the high CO mole fractions were consistent from
559 east regions (urbanized area) (Fig. 4), indicating strong anthropogenic sources in city
560 regions (i.e. Xining and Lanzhou) (Zhang et al., 2011).

561 The seasonal cycles of $\Delta\text{CO}/\Delta\text{CH}_4$ slopes (high in summer and low in winter) (Fig.
562 8) may primarily be due to the effect of monsoons and air mass transport. In summer,
563 air masses arriving at WLG were predominantly transported from the northeast to east



564 city regions (e.g., Xining, Lanzhou) with the largest CO mole fractions. In contrast, the
565 air masses were mainly from the southwest in winter, which carried strong CH₄
566 emissions but few CO emissions (Zhang et al., 2011) (Table 1). Hence the opposite two
567 air mass transport lead to a peak in summer and a trough in winter (Fig. 8). Moreover,
568 we could see apparently regional polarization in the concentration ratio of CH₄ and CO
569 (Fig. S7), implying the different strong source distribution between CH₄ and CO at
570 WLG. The cluster results (Fig. 5b & d) and the potential sources analysis (Fig. 6f & l)
571 also support this seasonal variation. Tohjima et al. (2014) found an opposite variation
572 at Hateruma Island, which showed low slope values in summer. It could be attributed
573 to different local sources and sinks, suggested the special topography condition and
574 local source distribution around WLG. The large $\Delta\text{CO}/\Delta\text{CH}_4$ fluctuations (Fig. 9) over
575 the study period was likely because of the anomaly years of different CH₄ or CO mole
576 fractions as well as source regions. In 2007, large increase of ΔCH_4 appeared, and from
577 2010 to 2013, the ΔCO decreased significantly (Fig. S1). Before 2010, large air masses
578 and potential source regions were identified in eastern regions (city regions) with the
579 highest CO emission (Fig. 5; Fig. 6). After 2010, the southwest regions showed high
580 contributions, with the highest CH₄ emission but relatively lower CO emission.
581 Therefore, obviously variation of the slopes presented with almost an increase in 2007-
582 2010 and a decrease after 2010 (Fig. 9).

583

584 4.3 Long-term variations

585 4.3.1 Seasonal cycles

586 The seasonal variations with maximum in summer and minimum in spring at WLG
587 (Fig. 10) were consistent with the previously short-term studies, e.g., Zhang et al.
588 (2013) in 2002-2006. However, it was almost opposite to observation in the adjacent
589 stations such as Lin'an, Shangdianzi and Longfengshan in China (Fang et al., 2013,
590 2016), as well as most global stations in the northern hemisphere, e.g., MLO, BRW and



591 JFJ (Dlugokencky et al., 1995; Loov et al., 2008). We further compared similar
592 WMO/GAW global stations in the north hemisphere, including MLO (19.54° N, -
593 155.58° E, 3397m a.s.l.) (Dlugokencky et al., 1995, 2019a), JFJ (46.55° N, 7.99° E,
594 3580m a.s.l.) (Zellweger et al., 2016), MNM (24.29° N, 153.98° E, 7.1m a.s.l.)
595 (Matsueda et al., 2004; Tsutsumi et al., 2006), as well as the marine boundary layer
596 (MBL) from NOAA/ESRL lab at similar latitude (Dlugokencky et al., 2019b). It can be
597 seen that the stations in the northern hemisphere (i.e. MLO, JFJ and MNM) and the
598 MBL showed an opposite trend to WLG with the minimum in summer and maximum
599 in winter or spring (Fig. 12). And the seasonal amplitude at WLG (~14 ppb) was lower
600 than many other sites in the northern hemisphere, about 35-70 ppb, e.g., MLO, BRW,
601 bialystok in Poland, ochsenkopf in Germany and beromunster in switzerland
602 (Dlugokencky et al., 1995; Thompson et al., 2009; Popa et al., 2010; Satar et al., 2016).
603 MBL also showed larger amplitude than that in WLG (Fig. 12). The low amplitude at
604 WLG was possibly because of high elevation of continental mountain sites (e.g., WLG,
605 JFJ), where were relatively less affected by local influences than the coastal/island sites
606 (Yuan et al., 2019). The peak in summer at WLG may be attributed to larger grazing
607 and human activities than other seasons. The CH₄ emissions from yaks and other
608 ruminants in Tibetan Plateau (alpine pasture) were very strong in summer, preceded
609 only by paddy emission (Fang et al., 2013; Zhang et al., 2013). Furthermore, the
610 photochemical capacities were very weak (high altitude) and the dynamic transport by
611 air flow from polluted northeast/southeast region was also strong in summer at WLG,
612 which all induced high CH₄ mole fractions in summer and consequently an opposite
613 trend with other sites (Ma et al., 2002; Xiong et al., 2009).

614

615 4.3.2 Long-term trends in different observing periods

616 The entirely fluctuant trend of atmospheric CH₄ in 1994-2016 at WLG (Fig. 11) was
617 similar to the global trend reported by quite a few studies (Bergamaschi et al., 2013;
618 Rigby et al., 2017; Nisbet et al., 2019). The previous study by Zhou et al. (2004) showed



619 the CH₄ annual increase of 4.5 ppb yr⁻¹ in 1992-2001, which was similar to that in 1994-
620 1997 (4.6 ± 0.1 ppb yr⁻¹) and 1997-2002 (2.6 ± 0.2 ppb yr⁻¹) in our study (Table 4).
621 Tohjima et al. (2002) found similar growth rates that the CH₄ at Cape Ochi-ishi and
622 Hateruma Island in 1995-2000 was increased about 4.5 and 4.7 ppb yr⁻¹, respectively. In
623 early 1990s, the CH₄ trend at WLG is very low, which was similar to the global growth
624 rates (Fig. 11b). The levels of OH radicals may have controlled the decrease or increase
625 of CH₄ in the atmosphere during this period (Dlugokencky et al., 1998; Rigby et al.,
626 2017; Turner et al., 2017). The growth rates were high in 1998 (Fig. 11b), which may
627 have been due to the high temperatures, large biomass burning and weak destruction
628 (Cunnold et al., 2002; Lelieveld et al., 2004; Simmonds et al., 2005).

629 The continuously larger CH₄ growth rates after 2007 at WLG (Fig. 11), e.g., 7.6 ±
630 0.2 ppb yr⁻¹ in 2008-2012, 5.7 ± 0.1 ppb yr⁻¹ in 2013-2016 (Table 4), were similar to the
631 recent study by Nisbet et al. (2016) and (2019), which showed the global CH₄ increased
632 by 5.7 ± 1.2 ppb yr⁻¹ in 2007-2014, and the much higher of 12.7 ± 0.5 ppb yr⁻¹ in 2014,
633 10.1 ± 0.7 ppb yr⁻¹ in 2015, 7.0 ± 0.7 ppb yr⁻¹ in 2016 and 7.7 ± 0.7 in 2017 ppb yr⁻¹.
634 The average growth rate in the northern hemisphere was 7.3 ± 1.3 ppb in 2007 and 8.1
635 ± 1.6 ppb in 2008 (Dlugokencky et al., 2009), which was also similar to the observation
636 at WLG (Table 4). After 2007, most sites in the northern hemisphere displayed large
637 CH₄ growth rates. Fang et al. (2013) showed that the annual growth rate of CH₄ was
638 9.4 ± 0.2 ppb yr⁻¹ in 2009-2011 at WLG, which were a little higher than our study in
639 similar period. The adjacent stations in China also revealed high CH₄ growth rates, e.g.,
640 8.0 ± 1.2 ppb yr⁻¹ at Lin'an in 2009-2011, 7.9 ± 0.9 ppb yr⁻¹ at Longfengshan in 2009-
641 2011, and 10 ± 0.1 ppb yr⁻¹ at Shangdianzi in 2009-2013 (Fang et al., 2013, 2016),
642 which was higher than the similar period in 2008-2012 or 2013-2016 at WLG (Table
643 4). The CH₄ measurements in other countries, e.g., Beromünster tall tower station also
644 showed high growth rate of 9.66 ppb yr⁻¹ in 2012-2014 (Satar et al., 2016). The very
645 warm temperatures, large biomass burning and the climatic anomaly e.g., El Niño, La
646 Niña event, were likely enhanced the CH₄ emissions after 2007 (Dlugokencky et al.,



647 2009). Additionally, the anomalously sharply increasing or decreasing years (e.g.,
648 2007/2008) may have a significant influence on the overall CH₄ trend (Fig. 11), and
649 these frequent anomalies appeared in most long-term observation stations, e.g., MLO
650 in USA (Dlugokencky et al., 2009), Mount Zugspitze in Germany (Yuan et al., 2019),
651 which were also possibly attributed to climatic forces, such as the exception during the
652 El Niño oscillation, forest fires, volcanic eruptions, and extreme weather events
653 (Keeling et al., 1995; Dlugokencky et al., 2009; Keenan et al., 2016; Nisbet et al., 2019).
654 The study by Satar et al. (2016) at Beromünster, Switzerland explained that the short-
655 term spikes were possibly related to emissions from agricultural activities, while the
656 longer lasting peaks were because of air mass transport and mixing.

657 It is well established that the human activities are mainly responsible for the recent
658 rapid CH₄ growth rates and anomalies. The analysis from Emissions Database for
659 Global Atmospheric Research (EDGAR) showed the CH₄ emission by per sector in
660 China (Fig. S8) (Crippa et al., 2019). During the observing period, the waste, the oil
661 and natural gas and open burning continuously emitted large amount of CH₄ into the
662 air. After 2000, the CH₄ emission from solid fuel increased greatly in China. After 2003,
663 the CH₄ emitted from rice cultivation also increased continuously (Fig. S8). The
664 increased emissions from these sectors may greatly contribute to the CH₄ increase at
665 WLG, as well as the other regions in China. In addition, the recent studies revealed that
666 China's coal sector may have dominated the clearly positive trend in recent years, which
667 contributed the highest proportion of the anthropogenic CH₄ emissions (~33%)
668 (Janssens-Maenhout et al., 2019; Miller et al., 2019). In 2010-2015, China's coal
669 production increased obviously (from 3400 to 4000 million metric tons), but emissions
670 trends of CH₄ by rice, agriculture, ruminants, waste, and oil/gas have grown slightly
671 and even remained flat (EIA, USA). The isotopic evidence suggests that the significant
672 increase of biogenic emissions was the dominant factor of CH₄ rise, especially in the
673 tropical wetlands with strong rainfall anomalies, or the agricultural sources such as rice
674 paddies and ruminants, fossil fuel emissions have not been the main cause (Nisbet et



675 al., 2016). The study by Chen et al. (2013) illustrated that the warming (0.2 °C per
676 decade) in the Tibetan Plateau resulted in substantial emission of CH₄ due to the
677 permafrost thawing and glaciers melting. However, up to now, the specific causes of
678 such distinct variability around the years, e.g., the spikes or near-zero CH₄ growth rates,
679 are not yet determined.

680

681 4. 3. 3 Annual growth rate in Qinghai-Tibetan Plateau

682 Although similar annual growth rates were found among the City Regions (CR), the
683 Tibet Plateau (TP) and total regional records (Total) in the entire observing period
684 (1994-2016) (Fig. 11), significant differences were found in short-term periods (Fig.
685 S3). In 2013-2016 (Table 4), the TP showed larger growth rate than that in CR, implying
686 stronger CH₄ source in the Tibetan Plateau in recent years. The seasonal amplitude in
687 the Tibetan Plateau was continuously increasing, which also revealed that the Tibetan
688 Plateau was intensively affected by strong regional sources. Without air mass transport
689 from the city regions, the significantly increased annual growth rate (TR) in 2008-2012
690 (10.1 ± 0.1 ppb yr⁻¹) and 2013-2016 (6.3 ± 0.1 ppb yr⁻¹) (Fig. S5 and Table 5) suggested
691 that there were possibly other strong CH₄ sources around WLG not from cities.
692 Northern India and eastern China were obviously the largest two source regions of CH₄
693 at WLG (Fig. S9) (Crippa et al., 2019). Since the Tibetan Plateau was coincidentally
694 trapped in the middle of the largest increased source areas, the atmospheric CH₄ at
695 WLG was very likely dominated by long-distance transport from these two regions.
696 Although CH₄ emissions increased slowly during 1994-2002, and even negative trend
697 appeared in southeast China (Fig. S9), significantly increased emissions appeared in
698 both southeast and southwest Asia after 2007. The rapidly increased CH₄ would
699 probably make it difficult to meet the goals of carbon emission reduction in the future.
700 Especially on the scenario of quick increasing CH₄ on the Qinghai-Tibetan Plateau due
701 to the emission from two largest source regions. In view of the integrated eco-
702 environmental change processes and unique topography in the Qinghai-Tibetan



703 Plateau, it may provide us one of the last precious regions to study global climate
704 changes (Chen et al., 2013). The anomalously year to year fluctuations of atmospheric
705 CH₄ in Tibetan Plateau were unquestionably a warning or alarm to the world, and the
706 unprecedented annual growth rate might be a dangerous signal to global climate change.

707

708 **5 Conclusion**

709 The atmospheric CH₄ at Mt.Waliguan increased continuously during 1994-2017.
710 Although near-zero and even negative growth appeared in some particular periods, e.g.,
711 1999-2000, and 2004-2006, the overall trend of CH₄ was increased rapidly, especially
712 in recent decade. Obvious diurnal cycle was found with the peak at noon and a trough
713 at late afternoon. Due to the unique geophysical locations and transport pathway, the
714 seasonal averages of CH₄ at WLG displayed an opposite trend with sites in the northern
715 hemisphere, with summer maximum and spring minimum. Large amount of air masses
716 was from west and northwest regions of WLG, which accompanied with higher CH₄
717 mole fractions than that from city regions. The Northern India possibly became a strong
718 source of CH₄ to WLG rather than city regions before.

719 As time goes by, the temporal patterns (e.g., seasonal amplitude), the annual
720 variations, the long-term trends or potential source distribution of CH₄ at WLG are all
721 changing. Thus, the long-term verification is extremely important to accurately
722 understand CH₄ variations. The case study in Qinghai-Tibetan Plateau revealed
723 unprecedented annual growth rates of CH₄. In recent years, the Tibetan Plateau even
724 showed larger growth rate than that in city regions. Tibetan Plateau was with the highest
725 average altitude and was almost impervious to strong human activities. There is no
726 doubt that the anomalously variation and the unprecedented annual growth rate of
727 atmospheric CH₄ in this region might be a dangerous signal to global climate change.

728

729

730



731 *Data availability.* The gridded meteorological data (2004-2017) from NOAA-ARL was
732 available at <ftp://arlftp.arlhq.noaa.gov/pub/archives/gdas1/>. The data from MLO, JFJ,
733 MNM station was downloaded from World Data Centre for Greenhouse Gases
734 (WDCGG) at <https://gaw.kishou.go.jp/>. The MBL data was available at
735 ftp://aftp.cmdl.noaa.gov/data/trace_gases/CH4/flask/surface/. The geographical
736 distribution of annual emission data by Emissions Database for Global Atmospheric
737 Research (EDGAR) was from website
738 https://edgar.jrc.ec.europa.eu/overview.php?v=50_GHG.

739

740

741 *Author contributions.* SL, SF and ZF designed the research. SL performed the data
742 processing with assistance of SF and MG. The station were monitored, maintained ML
743 and PL, and they collected, preprocessed, and provided the hourly observational
744 dataset. SL and SF finished the manuscript with contributions from all the co-authors.

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1057 **Table 1.** The statistics for cluster analysis result for both CH₄ and CO at WLG station. The clusters
1058 from urban areas are highlighted with face bold.

	Cluster	Number	Average CH ₄ mole fraction
Spring	1	1243	1853.4 ± 2.7
	2	685	1852.6 ± 3.4
	3	2231	1877.6 ± 2.5
	4	1093	1850.5 ± 2.2
	5	4108	1860.8 ± 1.5
Summer	1	3981	1869.9 ± 1.2
	2	2244	1878.5 ± 2.6
	3	1040	1866.3 ± 2.6
	4	916	1857.8 ± 2.4
	5	578	1876.3 ± 5.1
Autumn	1	1133	1870.1 ± 3.7
	2	4235	1868.8 ± 1.3
	3	2745	1873.6 ± 1.6
	4	550	1865.2 ± 3.6
Winter	1	3066	1879.8 ± 2.1
	2	601	1872.6 ± 3.4
	3	5261	1865.8 ± 1.0

1059



1060 **Table 2.** The statistics of filtered CH₄ data series over different periods during 1994-2017 at WLG
1061 station.

Year	Regional representative			Local representative		
	Hours	Percentage (%)	Mean (ppb)	Hours	Percentage (%)	Mean (ppb)
1994-1997	6481	46.9	1801.7 ± 0.5	7346	53.1	1806.1 ± 0.5
1998-2002	5332	47.6	1832.6 ± 0.7	5877	52.4	1837.9 ± 0.6
2003-2007	12421	49.2	1832.2 ± 0.3	12850	50.8	1839.2 ± 0.3
2008-2012	11314	49.2	1856.4 ± 0.4	11664	50.8	1867.0 ± 0.4
2013-2017	10140	43.6	1894.9 ± 0.5	13121	56.4	1907.4 ± 0.5
1994-2017	45688	47.3	1847.9 ± 0.3	50858	52.7	1858.2 ± 0.4

1062



1063 **Table 3.** Yearly average CH₄ mole fractions at WLG station.

Year	Mean (ppb)	year	Mean (ppb)
1995	1805.8 ± 0.1	2006	1835.6 ± 0.2
1996	1804.6 ± 0.2	2007	1839.9 ± 0.5
1997	1806.8 ± 0.2	2008	1865.9 ± 0.3
1998	1827.0 ± 0.1	2009	1849.1 ± 0.1
1999	1820.2 ± 0.1	2010	1857.9 ± 0.2
2000	1820.0 ± 0.2	2011	1872.6 ± 0.4
2001	1849.2 ± 0.4	2012	1881.2 ± 0.3
2002	1835.5 ± 0.2	2013	1896.3 ± 0.2
2003	1842.5 ± 0.3	2014	1890.4 ± 0.1
2004	1836.7 ± 0.1	2015	1905.6 ± 0.3
2005	1837.4 ± 0.1	2016	1903.8 ± 0.1

1064



1065 **Table 4.** Annual growth rates of atmospheric CH₄ in the City Regions (CR), the Tibet Plateau (TP),
1066 and original total regional records (Total) during 1994-2016 at WLG station.

	1994-1997	1998-2002	2003-2007	2008-2012	2013-2016	1994-2016
CR	2.8 ± 0.1	3.6 ± 0.2	5.6 ± 0.2	7.1 ± 0.2	5.5 ± 0.1	5.0 ± 0.1
TP	3.4 ± 0.1	3.0 ± 0.2	6.8 ± 0.2	5.6 ± 0.2	7.3 ± 0.1	5.2 ± 0.1
Total	4.6 ± 0.1	2.6 ± 0.2	5.3 ± 0.2	7.6 ± 0.2	5.7 ± 0.1	5.1 ± 0.1

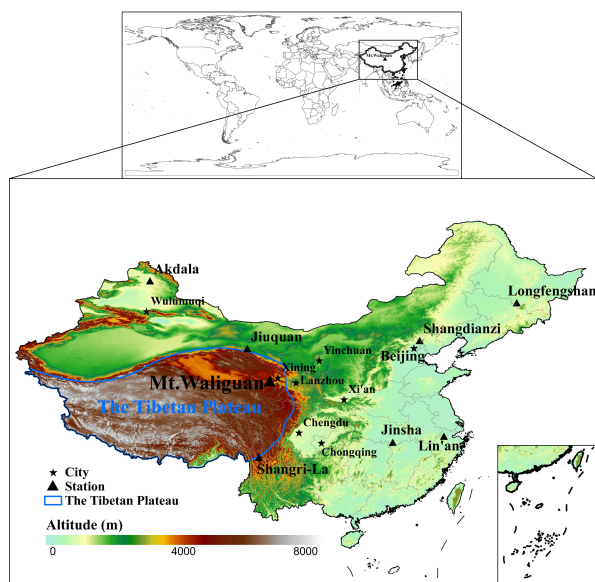
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1068 **Table 5.** The statistics of CH₄ data without air mass transport from city regions (TR) over different
1069 periods during 2005-2016 at WLG station.

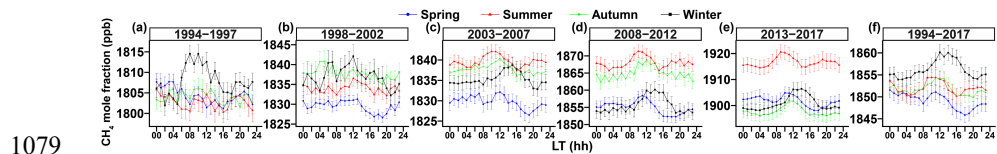
	Transport regions	Hours	Percentage (%)	Average (ppb)	Updated growth rate (ppb yr ⁻¹)
2005-2007	TR	6922	77.2	1824.9 ± 0.2	2.7 ± 0.2
	City	2041	22.8	1835.9 ± 0.5	-
2008-2012	TR	7060	64.4	1853.7 ± 0.2	10.1 ± 0.1
	City	4254	35.6	1861.0 ± 0.3	-
2013-2016	TR	4152	61.6	1888.2 ± 0.3	6.3 ± 0.1
	City	2591	38.4	1888.5 ± 0.5	-
2005-2016	TR	18134	67.1	1850.6 ± 0.2	7.0 ± 0.1
	City	8886	32.9	1863.2 ± 0.3	-

1070

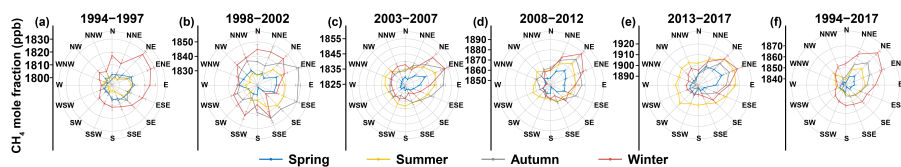


1071

1072 **Figure 1.** Location of Mt. Waliguan WMO/GAW global station as well as other regional stations in
1073 China. The gradient color indicates altitude. The digital elevation model (DEM) was downloaded
1074 from Geospatial Data Cloud site, Computer Network Information Center, Chinese Academy of
1075 Sciences (<http://www.gscloud.cn>), and then processed by ArcGIS software. The China map was
1076 derived from © National basic Geomatics Center of China (<http://www.ngcc.cn/ngcc/>). The world
1077 map was obtained from © OpenStreetMap (<https://www.openstreetmap.org/>). And the other shpfile
1078 file data and entire map were created by ArcGIS software.

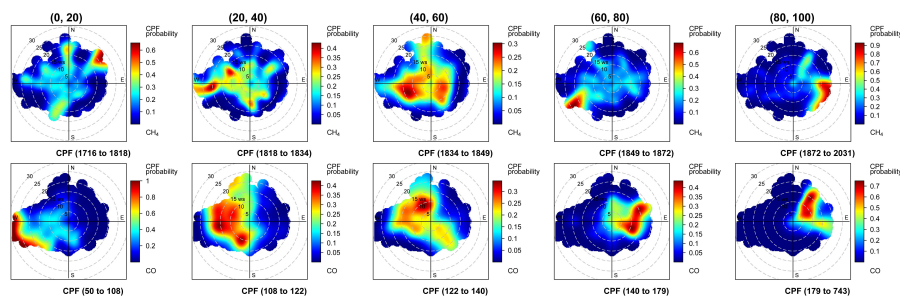


1080 **Figure 2.** Diurnal CH₄ cycles in different periods from 1994 to 2017 at WLG station. The lines with
1081 different colors represent various seasons. Error bars indicate the 95% confidence intervals.



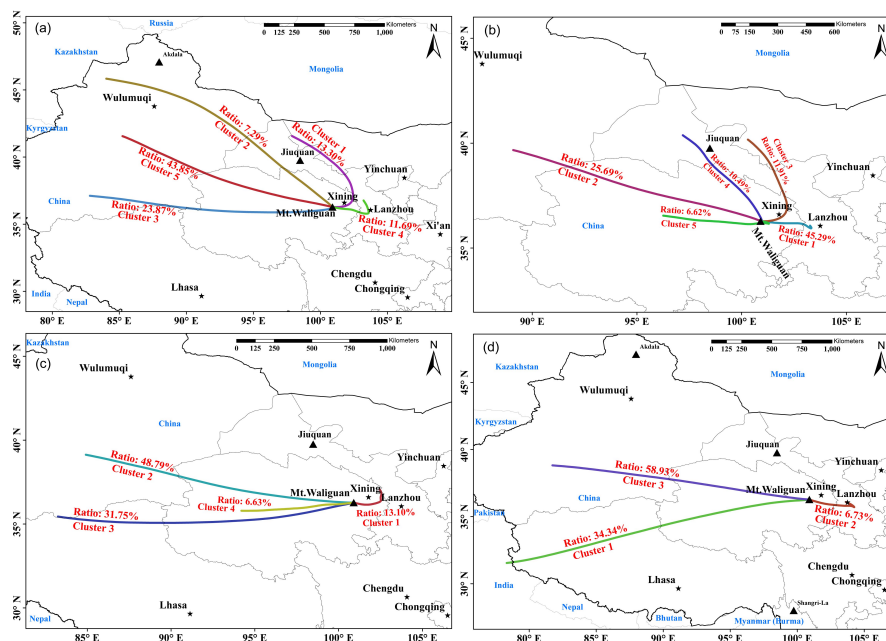
1082

1083 **Figure 3.** Wind-rose distribution of average hourly CH₄ records from 16 horizontal wind directions
1084 over different periods in 1994-2017 at WLG station. The different colors represent the CH₄ data in
1085 different seasons. Error bars in all directions indicate 95% confidence intervals.



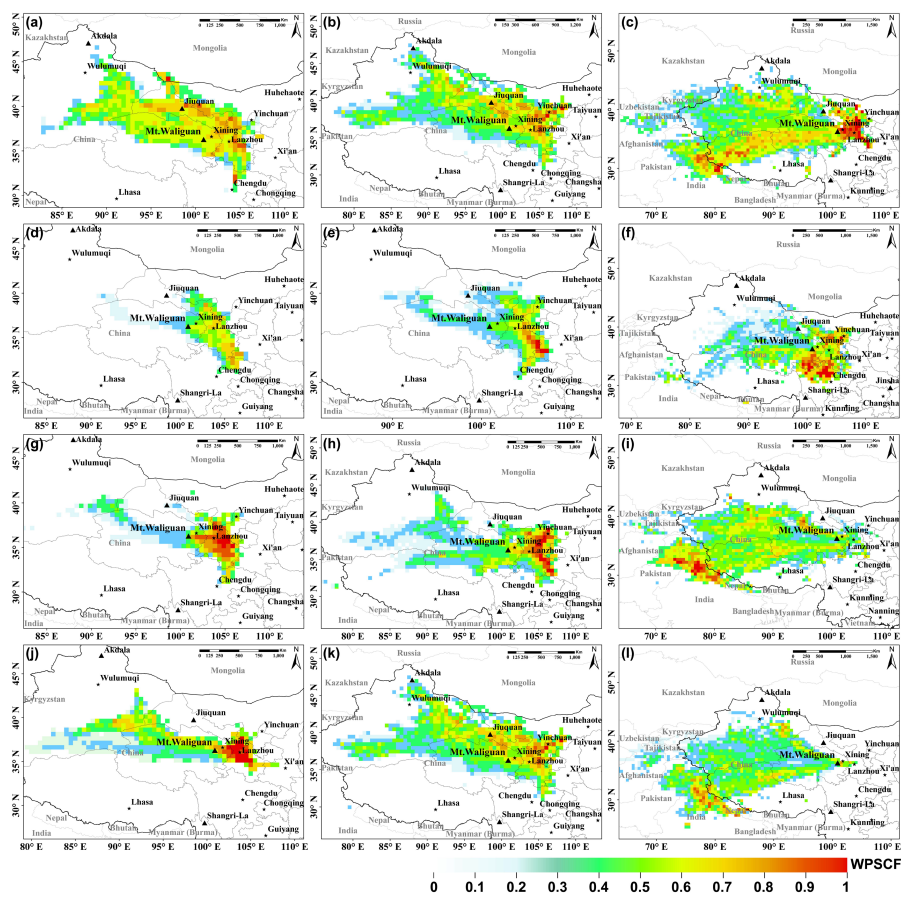
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1087 **Figure 4.** The polar plot of the distribution of CH₄ and CO concentration probability in different
1088 percentile ranges at WLG station. The analysis was based on the conditional probability functions
1089 (CPF) by Ashbaugh et al. (1985). The top plots show the measurements of CH₄ from 1994 to 2017.
1090 The bottom plots show the CO measurements in 2004-2017. ‘ws’ means the wind speed. The values
1091 in the bottom of each panel show the range of concentration in relevant percentile range. Gradient
1092 colors represent the levels of CPF probability in different percentile ranges.



1093

1094 **Figure 5.** Cluster analysis to the 72-h back trajectories in different seasons during 2004-2017 ending
1095 at WLG station. The (a), (b), (c) and (d) represents spring, summer, autumn and winter, respectively.
1096 The lines with different colors denote different cluster analysis results. The proportion of trajectories
1097 on each cluster is also marked.



1098

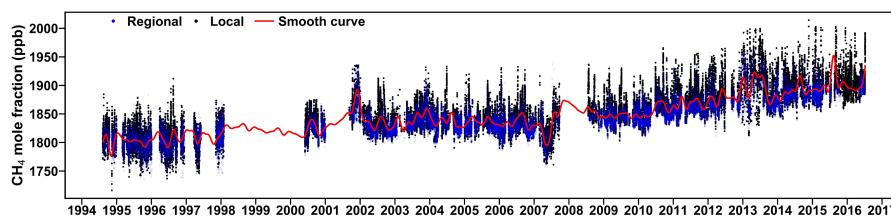
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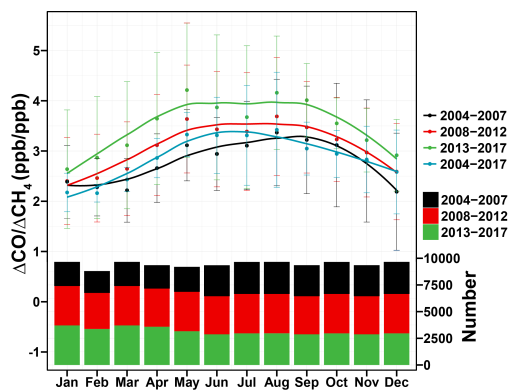
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Figure 6. Geographical distribution of weighted potential source of CH₄ in different periods over 1994-2017 at WLG station. The gradient color shows strong levels of potential source regions in different seasons, i.e. spring (a, b, c), summer (d, e, f), autumn (g, h, i) and winter (j, k, l), and different periods, i.e. 2004-2007 (a, d, g, j), 2008-2012 (b, e, h, k) and 2013-2017 (c, f, i, l).



1103
1104 **Figure 7.** Filtered hourly CH₄ data series from 1994 to 2017 at WLG station. The blue points with
1105 transparency represent regional events. The black points are the filtered local events. The red lines
1106 are calculated smooth values to the regional data by curve-fitting routine of Thoning et al. (1989).



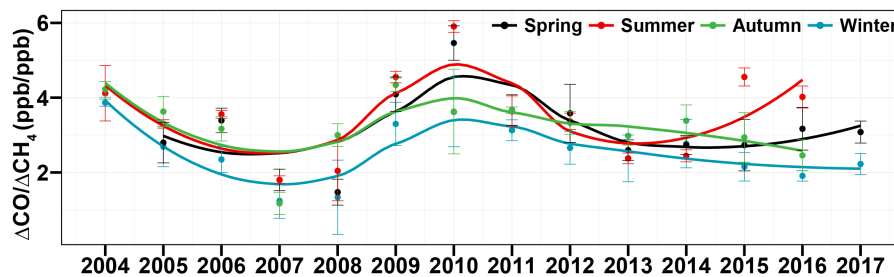
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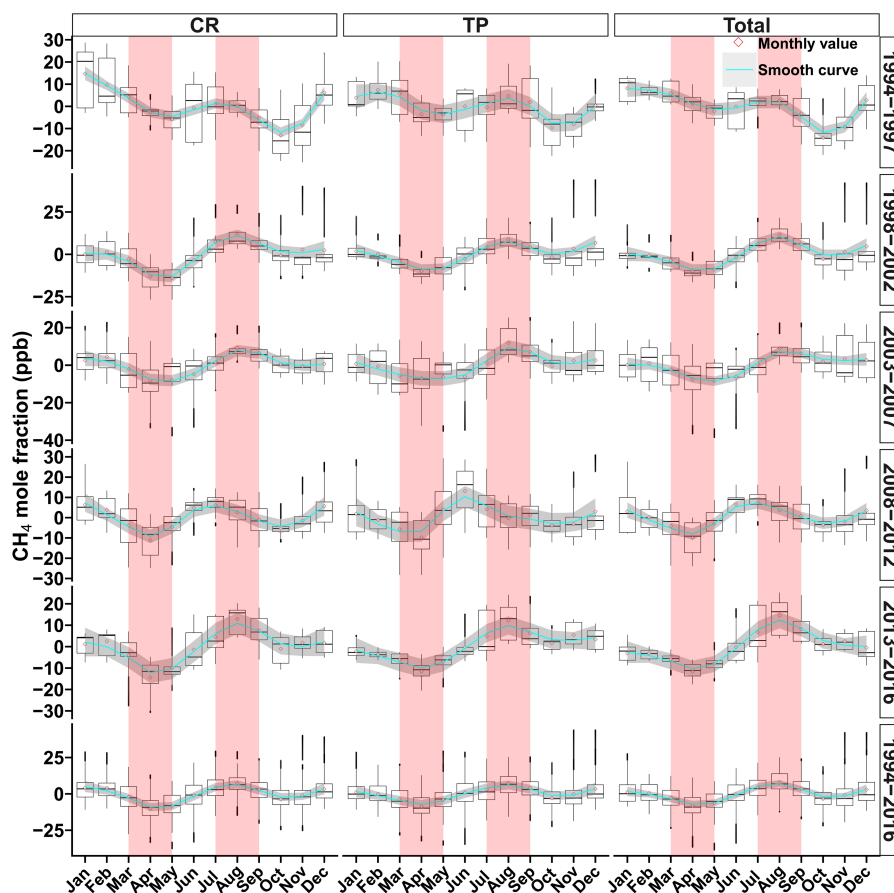
Figure 8. Average seasonal variation of $\Delta\text{CO}/\Delta\text{CH}_4$ slopes in different periods over 2004-2017 at WLG station. The error bars show the standard deviation of the monthly averages. The vertical bars are the monthly numbers of data in different periods.



1111

1112

Figure 9. Long-term trend of $\Delta\text{CO}/\Delta\text{CH}_4$ slopes over 2004-2017 at WLG station.



1113

1114 **Figure 10.** Monthly variations of regional CH₄ mole fractions from 1994 to 2016 at WLG station.

1115 The ‘CR’, ‘TP’ and ‘Total’ represents the measurements from the City Regions, the Tibet Plateau

1116 and the original total regional records, respectively. The box respectively shows the 25th percentile,

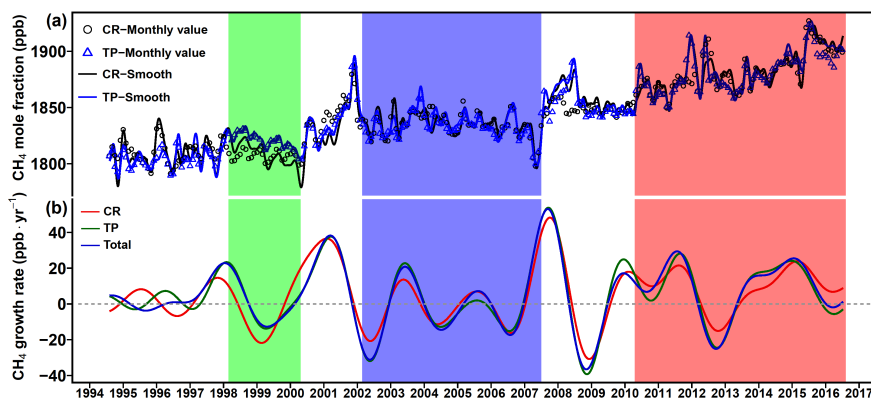
1117 the median and the 75th percentile from bottom to top. The bottom and the top whisker respectively

1118 reaches the minimum and 1.5 times the IQR (interquartile range). The black points are identified as

1119 outliers. The red squares are the averages. The cyan lines are the smooth curve of averages using

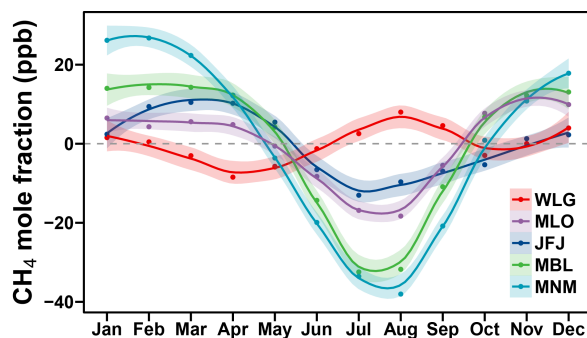
1120 the method of loess (Local Polynomial Regression Fitting). The gray bands are the 95% confidence

1121 interval of smooth curve.



1122

1123 **Figure 11.** The top panel (a) shows the smooth curve and monthly values of CH₄ mole fraction in
1124 the City Regions (CR) and the Tibet Plateau (TP) during 1994-2016 at WLG station. The bottom
1125 panel (b) is the annual growth rates of atmospheric CH₄ records from CR, TP as well as the total
1126 regional time series (Total). The growth rates are calculated from the first derivative of trend curves.
1127 The smooth curve and the trend is calculated by the method of Thoning et al. (1989).



1128

1129 **Figure 12.** The seasonal cycles of atmospheric CH₄ observed over 1994-2017 at WMO/GAW global
1130 stations of Mauna Loa (MLO), Jungfrauoch (JFJ), Minamitorishima (MNM) and Mt. Waliguan
1131 (WLG) in the northern hemisphere. The data of other sites (except WLG) are from WDCGG. The
1132 data in the marine boundary layer (MBL) are from NOAA / ESRL lab at the similar latitude to WLG.