



1 Tropospheric aerosol hygroscopicity measurements in China

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26 **Abstract**

27 Hygroscopicity largely determines phase state, chemical reactivity, optical properties and
28 cloud nucleation activities of aerosol particles, thus significantly affecting their impacts on
29 visibility, atmospheric chemistry and climate. In the last twenty years a large number of field
30 studies have investigated hygroscopicity of tropospheric aerosols in China under sub- and super-
31 saturated conditions. Aerosol hygroscopicity measurements in China are reviewed in this paper: 1)
32 a comprehensive summary and critical discussion of aerosol hygroscopicity measurements in
33 China is provided; 2) available measurement data are compiled and presented under a consistent
34 framework to enhance their accessibility and usability; 3) current knowledge gaps are identified,
35 and an outlook which could serve as guidelines for planning future research is also proposed.

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37



38 **1 Introduction**

39 In the last few decades, rapid industrial, economic and social developments in China have
40 caused large emissions of gaseous and particulate pollutants into the troposphere (Li et al., 2017a),
41 where they are mixed with gases and aerosols from natural sources. Under unfavourable
42 meteorological conditions (i.e. when air is stagnant and stable), severe air pollution occurs, due to
43 accumulation of primary pollutants and more importantly, formation of secondary pollutants (Zhu
44 et al., 2011; He et al., 2014; Zhang et al., 2015; An et al., 2019; Lu et al., 2019; Zhang et al., 2019c).
45 During severe air pollution events, PM_{2.5} could exceed a few hundred $\mu\text{g m}^{-3}$ (Guo et al., 2014;
46 Huang et al., 2014) and O₃ could reach up to >200 ppbv (Wang et al., 2017a). The concept of air
47 pollution complex has been proposed to describe the complexity of air pollution in China,
48 characterized by complex sources and complex interactions of a myriad of gaseous and particulate
49 pollutants (Zhu et al., 2011; Lu et al., 2019; Chu et al., 2020). Thanks to the implementation of
50 effective air pollution control measures, substantial decrease in PM_{2.5} has occurred nationwide in
51 the last several years (Zhang et al., 2019b); however, slight but significant increase in O₃ has been
52 observed in many regions during the same period (Li et al., 2019a; Lu et al., 2020), revealing the
53 complexity and difficulty in synergistic control of PM_{2.5} and O₃.

54 Hygroscopicity, one of the most important physicochemical properties of aerosols, determines
55 the amount of water associated with aerosol particles under ambient conditions (mainly relative
56 humidity, and temperature to a less extent) and significantly affects their environmental and
57 climatic impacts (Kreidenweis and Asa-Awuku, 2014; Tang et al., 2019). Hygroscopicity is
58 referred to hygroscopic properties under subsaturated conditions from a specific view, while from
59 a general view, it is referred to both hygroscopic properties under subsaturated conditions and
60 cloud condensation nucleation (CCN) activities under supersaturated conditions. Due to their



61 hygroscopicity, aerosol particles will take up water (i.e. hygroscopic growth) and lead to increase
62 in particle mass and size (Kreidenweis and Asa-Awuku, 2014; Tang et al., 2016; Wu et al., 2018b;
63 Tang et al., 2019). Therefore, hygroscopicity largely determines optical properties of aerosols and
64 as a result their impacts on visibility and direct radiative forcing under subsaturated conditions
65 (Titos et al., 2016; Zhao et al., 2019); on the other hand, hygroscopicity is also closely linked to
66 CCN activities of aerosols and thus their abilities to form cloud droplets under supersaturated
67 conditions (Kreidenweis and Asa-Awuku, 2014; Farmer et al., 2015; Tang et al., 2016), thereby
68 having important implications for their indirect radiative forcing (Dusek et al., 2006; McFiggans
69 et al., 2006; Farmer et al., 2015). Furthermore, hygroscopicity determines aerosol liquid water
70 content (ALWC) and thus phase state, acidity and chemical reactivities of aerosols (Bertram and
71 Thornton, 2009; Liu et al., 2017; Tang et al., 2017; Wu et al., 2018b), playing critical roles in
72 secondary aerosol formation as well as removal and production of trace gases. In addition,
73 hygroscopic growth measurements can provide valuable insights into mixing states of aerosols
74 (Swietlicki et al., 2008; Riemer et al., 2019). Due to its importance, tropospheric aerosol
75 hygroscopicity has been investigated in China by a number of field studies in the last 10-20 years,
76 as reviewed in this paper.

77 Swietlicki et al. (Swietlicki et al., 2008) summarized and analyzed hygroscopic properties of
78 ambient aerosols measured using H-TDMA (Hygroscopic Tandem Differential Mobility Analyser)
79 prior to September 2007, when ambient aerosol hygroscopicity was seldom explored in China. The
80 effects of hygroscopicity on aerosol light scattering have been reviewed and summarized on the
81 global scale (Titos et al., 2016; Burgos et al., 2019), and a very recent paper also briefly
82 summarizes aerosol light scattering enhancement studies in China (Zhao et al., 2019). A book
83 chapter (Kreidenweis and Asa-Awuku, 2014) discussed in brief hygroscopic growth and light



84 scattering enhancement of ambient aerosols, but only a few measurements conducted in China
85 were included. In addition, a recent paper (Tang et al., 2019) has reviewed aerosol hygroscopicity
86 measurement techniques. However, aerosol hygroscopicity measurements in China have not been
87 reviewed yet.

88 In this paper we provide a comprehensive review of hygroscopic properties of ambient aerosols
89 measured using H-TDMA in China; in addition, CCN activities of tropospheric aerosols measured
90 in China are also reviewed and discussed. Via using the single hygroscopicity parameter (κ), we
91 attempt to reconcile hygroscopic properties examined at $<100\%$ RH (relative humidity) with CCN
92 activities measured at $>100\%$ RH. A number of studies measured light scattering enhancement
93 factors, $f(\text{RH})$, of ambient aerosols in China (Zhao et al., 2019), but most of these studies are not
94 included herein for two reasons: 1) $f(\text{RH})$ measurements in China have been reviewed in brief very
95 recently (Zhao et al., 2019); 2) it is not trivial to convert measured $f(\text{RH})$ to growth factors or κ
96 values (Kreidenweis and Asa-Awuku, 2014). Nevertheless, we note that some methods have been
97 proposed to convert measured $f(\text{RH})$ to κ (Kuang et al., 2017; Kuang et al., 2018). Single particle
98 techniques were employed to investigate hygroscopic properties of tropospheric aerosols (Li et al.,
99 2016); however, as numbers of particles examined in single particle studies are usually too limited
100 to provide enough information for the overall aerosol hygroscopicity, these studies are not
101 discussed herein. Although not covered in this review, remote sensing techniques can also be used
102 to retrieve aerosol hygroscopicity in the troposphere (Lv et al., 2017; Bedoya-Velásquez et al.,
103 2018; Tang et al., 2019; Dawson et al., 2020).

104 The first goal of this paper is to provide a comprehensive overview of hygroscopic properties
105 and CCN activities of tropospheric aerosols in China via reviewing previous field studies. The
106 second goal is to compile and present measurement data (as compiled in Tables S1-S5) reported



107 by previous work using a consistent framework (i.e. via using the single hygroscopicity parameter)
108 to enhance their accessibility and usability. The third goal, perhaps more importantly, is to identify
109 knowledge gaps in this field and then to provide an outlook which can serve as practical guidelines
110 for planning future research. In this paper, Section 2 describes the methodology adopted in this
111 paper to analyse and review previous studies, and previous measurements of hygroscopic
112 properties and CCN activities of tropospheric aerosols in China are reviewed and discussed in
113 Sections 3 and 4. In the end, Section 5 outlines knowledge gaps and research perspectives.

114 **2 Methodology**

115 **2.1 Hygroscopic properties**

116 H-TDMA instruments, initially developed ~40 years ago (Liu et al., 1978; McMurry et al.,
117 1983; Rader and McMurry, 1986; McMurry and Stolzenburg, 1989), have been widely used in
118 field and laboratory studies (Kreidenweis et al., 2005; Svenningsson et al., 2006; Gysel et al., 2007;
119 Sjogren et al., 2008; Swietlicki et al., 2008; Duplissy et al., 2009; Asmi et al., 2010; Liu et al.,
120 2011; Wu et al., 2011; Kreidenweis and Asa-Awuku, 2014; Zieger et al., 2017; Tang et al., 2019).
121 Technical details of H-TDMA measurements, including operation principles, data analysis and etc.,
122 have been detailed in a review paper (Swietlicki et al., 2008). In brief, an aerosol flow, dried to
123 <20% RH, is passed through an aerosol neutralizer and the first DMA (Differential Mobility
124 Analyzer) to produce quasi-monodisperse aerosols with a given mobility diameter; after that, the
125 aerosol flow is delivered through a humidifier to be humidified to a given RH, and subsequently
126 aerosol size distributions are measured using the second DMA coupled with a CPC (Condensation
127 Particle Counter). The hygroscopic growth factor, GF, is defined as the ratio of the aerosol mobility
128 diameter at a given RH to that at dry conditions. As aerosol particles at a given size may have
129 different hygroscopic properties and thus display different GF values at a given RH, probability



130 distribution functions of GF (i.e. number fractions of aerosol particles at each GF) have also been
131 reported in some studies.

132 The measured distribution functions of GF are usually smoothed and skewed due to several
133 reasons, e.g., the finite width of the DMA's transfer function, and several TDMA inversion
134 algorithms have been proposed to convert the H-TDMA raw data to the probability density
135 function of GF (Stolzenburg and McMurry, 1988; Stratmann et al., 1997; Voutilainen et al., 2000;
136 Cocker et al., 2001; Cubison et al., 2005; Gysel et al., 2009). The algorithm developed by Gysel
137 et al., TDMAinv, is currently the most widely used one. Errors and uncertainties of H-TDMA data
138 can come from several sources, including RH and temperature variability, electrical mobility
139 classification, particle non-equilibrium in the second DMA, and etc. Swietlicki et al. (Swietlicki
140 et al., 2008) comprehensively discussed the sources and magnitudes of these errors and how they
141 can be reduced or minimized. In addition, guidelines used for H-TDMA measurements, including
142 instrumental design, calibration, validation and operation as well as data analysis, have been
143 recommended in literature (Duplissy et al., 2009; Massling et al., 2011).

144 H-TDMA measurements of ambient aerosols were typically conducted for a few different
145 particles diameters at a given relative humidity (RH); most measurements were carried out at 90%
146 RH, though some studies also reported growth factors (GF) at other RH. To facilitate comparison
147 of GF reported at different RH, we convert GF measured at a given RH to κ using Eqs. (1-2)
148 (Petters and Kreidenweis, 2007; Tang et al., 2016):

$$149 \quad \kappa = (\text{GF}^3 - 1) \left(\frac{B}{\text{RH}} - 1 \right) \quad (1)$$

$$150 \quad B = \exp \left(\frac{A}{d_0 \cdot \text{GF}} \right) \quad (2)$$



151 where d_0 is the dry particle diameter; A , which describes the Kelvin effects, is equal to 2.1 nm at
152 298.15 K if the surface tension is assumed to be the same as water (0.072 J m^{-2}) (Petters and
153 Kreidenweis, 2007; Tang et al., 2016). Converting GF to κ also facilitates comparison between
154 hygroscopic properties and CCN activities. For a few studies which reported GF at different RH,
155 we focus GF measured at 90% RH; if the data at 90% are not available, we then choose
156 measurements at the RH closest to 90%.

157 To further facilitate comparison between different measurements, Swietlicki et al. (Swietlicki
158 et al., 2008) classified aerosol hygroscopicity into four groups according to their GF at 90% RH.
159 This methodology was adopted by Ye et al. (Ye et al., 2013) who reported aerosol hygroscopic
160 growth measurements in Shanghai. Nevertheless, Ye et al. (Ye et al., 2013) classified aerosol
161 particles into three modes (instead of four), and the criteria used are slightly different from
162 Swietlicki et al. (Swietlicki et al., 2008). Here we adopt the method proposed by Ye et al. (Ye et
163 al., 2013), who classified aerosol hygroscopicity into three modes, including the nearly-
164 hydrophobic (NH, $\kappa < 0.1$), the less-hygroscopic (LH, $0.1 < \kappa < 0.25$) and the more-hygroscopic
165 (MH, $\kappa > 0.25$) modes. However, here a few further statements are necessary. First, terminologies
166 used differ in previous studies for aerosol hygroscopicity modes. For example, bimodal aerosol
167 hygroscopicity was frequently observed in China (as discussed in Section 3), and the nearly-
168 hydrophobic mode defined by Ye et al. (Ye et al., 2013) was called the less-hygroscopic mode or
169 the low-hygroscopic mode in several studies. Second, actual aerosol hygroscopicity in the
170 troposphere may not perfectly fit into one of the three modes defined by Ye et al. (Ye et al., 2013).

171 **2.2 CCN activities**

172 A variety of instruments have been developed to measure CCN number concentrations
173 (Twomey, 1963; Sinnarwalla and Alofs, 1973; Fukuta and Saxena, 1979; Hudson, 1989; Ji et al.,



174 1998; Chuang et al., 2000; McMurry, 2000; Nenes et al., 2001; Otto et al., 2002; VanReken et al.,
175 2004; Roberts and Nenes, 2005; Frank et al., 2007; Kreidenweis and Asa-Awuku, 2014). Currently
176 the most widely used one is the continuous-flow streamwise thermal gradient CCN counter based
177 on the design of Roberts and Nenes (Roberts and Nenes, 2005; Lance et al., 2006) and
178 commercialized by Droplet Measurement Technologies, and mode details of this instrument can
179 be found elsewhere (Roberts and Nenes, 2005; Lance et al., 2006).

180 Measurements of size-resolved CCN activities have been discussed in a number of previous
181 studies (Lance et al., 2006; Frank et al., 2007; Petters et al., 2007; Rose et al., 2008; Good et al.,
182 2010; Moore et al., 2010; Rose et al., 2010; Bougiatioti et al., 2011). In many studies, an aerosol
183 flow sampled from the ambient air, after dried to <20% RH, is passed through an aerosol
184 neutralizer and then a DMA to produce quasi-monodisperse aerosols. The aerosol flow is
185 subsequently split into two flows; one flow is sampled into a CCN counter to measure number
186 concentrations of cloud condensation nuclei ([CCN]), and the other one is sampled into a CPC to
187 measure number concentrations of condensation nuclei ([CN]). At a given supersaturation,
188 activation fractions ([CCN]/[CN]) are measured as a function of particle diameter (selected using
189 the DMA) and then fitted by an activation curve to determine the activation diameter at which the
190 activation fraction is equal to 0.5 (Snider et al., 2006; Rose et al., 2008; Sullivan et al., 2009;
191 Bougiatioti et al., 2011; Cerully et al., 2011), and activation fractions can be measured at one or
192 more supersaturation as a function of particle diameter. Methods used for instrument calibration
193 and data correction, which can be found in literature (Frank et al., 2007; Petters et al., 2007; Rose
194 et al., 2008; King et al., 2009; Petters et al., 2009; Moore et al., 2010), are not discussed herein.
195 Furthermore, κ can be derived from the determined activation diameter at a given supersaturation
196 (Petters and Kreidenweis, 2007).



197 Maximum activation fractions may not approach one for ambient aerosols, and generally two
198 methods have been used to fit the data. If the maximum activation fraction of the fitted activation
199 curve is not fixed (three-parameter fit), the derived activation diameter (d_a) and single
200 hygroscopicity parameter (κ_a) describe the average properties of activated particles; if it is forced
201 to be 1 (two-parameter fit), the derived activation diameter (d_t) and single hygroscopicity
202 parameter (κ_t) describe the overall aerosol properties (Rose et al., 2010). For aerosols with bimodal
203 hygroscopicity distribution, κ_a is comparable to the κ determined using H-TDMA for the more-
204 hygroscopic mode, while κ_t is comparable to the average κ for the two modes. In addition to d_a and
205 d_t , the apparent cut-off diameter (above which [CN] is equal to [CCN] at a given supersaturation.),
206 d_{cut} (and thus κ_{cut}), can be determined if it is assumed that particles at each size are internally mixed
207 and that larger particles are activated first (Rose et al., 2010; Hung et al., 2014). The determination
208 of d_{cut} does not required size-resolved activation fractions, but needs the overall activation fractions
209 and aerosol number size distribution (Burkart et al., 2011; Hung et al., 2014). Our review paper is
210 focused on κ_a and to a less extent κ_t , and only discusses κ_{cut} when neither κ_a nor κ_t was reported.

211 In addition, [CCN] and [CCN]/[CN] were also measured at one or more supersaturation in
212 Tianjin (Deng et al., 2011; Yang et al., 2012; Zhang et al., 2012), Zhangjiakou (Hebei) (Lu and
213 Guo, 2012), Shijiazhuang (Hebei) (Lu and Guo, 2012), Xingtai (Hebei) (Wang et al., 2018b),
214 Qingdao (Li et al., 2015a), Shanghai (Leng et al., 2013; Leng et al., 2014), Guangzhou (Duan et
215 al., 2017; Duan et al., 2018) and Mt. Huang (Fang et al., 2016), as well as over marginal seas of
216 China (Zhu et al., 2019; Gao et al., 2020) and northwestern Pacific (Wang et al., 2019a; Zhu et al.,
217 2019). As these studies did not carry out size-resolved measurements and thus did not report
218 critical diameters or κ , they are not further discussed herein.



219 **3 Hygroscopic growth**

220 A number of aerosol hygroscopic growth measurements have been carried out in China since
221 2001 using H-TDMA (or very similar instruments). Most of these measurements were performed
222 in three regions with severe air pollution, including the North China Plain (NCP), Yangtze River
223 Delta (YRD) and Pearl River Delta (PRD), and these studies are discussed in Sections 3.1-3.3. In
224 addition, as discussed in Section 3.4, several measurements were also conducted at other locations
225 in the east or south China.

226 **3.1 North China plain (NCP)**

227 The North China Plain is a heavily polluted region where many aerosol hygroscopic growth
228 measurements were conducted, and as summarized in Table S1. In this section we review the
229 measurements carried out at urban sites in Beijing (Section 3.1.1), rural sites in Beijing (Section
230 3.1.2), other urban/suburban sites (Section 3.1.3) and other rural sites (Section 3.1.4).

231 **3.1.1 Urban sites in Beijing**

232 Aerosol hygroscopic growth has been measured at three urban sites in Beijing, including the
233 PKU site, the IAP site, and the CAMS site.

234 **PKU site:** The PKU site is located on the campus of Peking University (39°59'20"N,
235 116°18'26"E), which is between the fourth and fifth ring road in the northwest of Beijing. All the
236 measurements (Massling et al., 2009; Meier et al., 2009; Wu et al., 2016; Wu et al., 2017; Wang et
237 al., 2018c) took place on the roof of a six-floor building (~30 m above ground), which is ~100 m
238 away from a major road.

239 Aerosol hygroscopic growth was first measured at the PKU site during 2004-2005 (Massling
240 et al., 2009; Meier et al., 2009). Massling et al. (Massling et al., 2009) measured aerosol
241 hygroscopic growth (at 90% RH) in June-July 2004 and January-February 2005. Aerosol



242 hygroscopicity exhibited trimodal distribution, and κ were found to be in the range of 0-0.028,
243 0.036-0.176 and 0.175-0.386 for the low-, medium- and high-hygroscopic modes (Massling et al.,
244 2009). In addition, no obvious difference in aerosol hygroscopicity was found between summer
245 and winter. Ammonium sulfate was the major inorganic species for the high-hygroscopic mode,
246 while fresh carbonaceous materials (e.g., soot) dominated the low-hygroscopic mode (Massling et
247 al., 2009). Aerosol hygroscopicity was found to increase with particle size and pollution levels
248 (Massling et al., 2009), as more secondary inorganic species were formed.

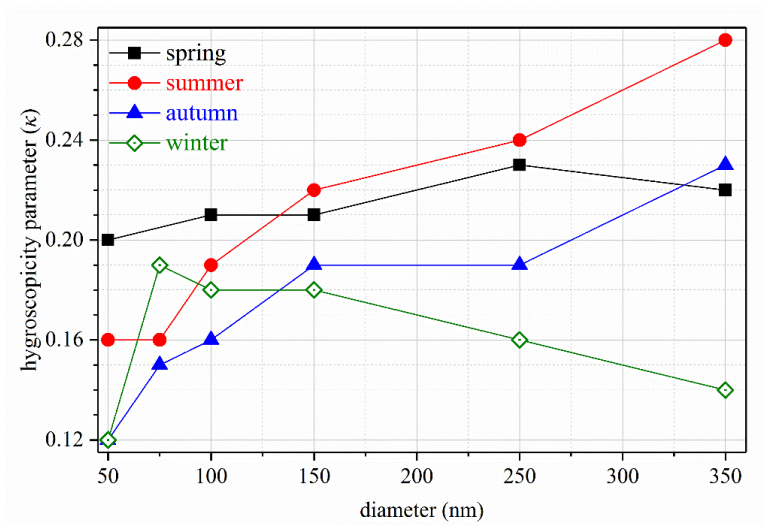
249 Meier et al. (Meier et al., 2009) further explore aerosol hygroscopic growth (at 90% RH) at
250 the PKU site in January 2005. Similar to the work by Massling et al. (Massling et al., 2009), three
251 aerosol hygroscopicity modes were identified, with the κ values being 0-0.027, 0.036-0.154 and
252 0.152-0.366 for low-, medium- and high-hygroscopic modes (Meier et al., 2009); however, no
253 obvious dependence of aerosol hygroscopicity on air pollution levels was found. The average κ
254 were found to first increase (30-80 nm) and then decrease with particle size (80-350 nm). Measured
255 GF at 90% RH were compared with these calculated from size-resolved inorganic compositions
256 measured offline, and discrepancies between measured and calculated GF were attributed to the
257 effects of organics contained (Meier et al., 2009). In addition, hygroscopic growth at 55% and 70%
258 RH was also explored for 30-400 nm aerosol particles (Meier et al., 2009), and GF at 55% and 70%
259 RH, compared to 90% RH, displayed similar dependence on particle size.

260 Wu and co-workers (Wu et al., 2016; Wu et al., 2017; Wang et al., 2018c) carried out
261 extensive aerosol hygroscopic growth measurements (at 90% RH) at the PKU site during 2014-
262 2015. Bimodal aerosol hygroscopicity distribution was observed in May-June 2014 (Wu et al.,
263 2016), dominated by the hydrophilic mode, and the average κ appeared to increase with particle
264 size, from 0.160 at 50 nm to 0.280 at 250 nm. In addition, number fractions of aerosol particles in



265 the hydrophilic mode first increased with particle size up to 150 nm, and then did not show
266 significant change with further increase in particle size (Wu et al., 2016); to be more specific,
267 average number fractions of aerosol particles in the hydrophilic mode were ~ 0.6 at 50 nm and
268 increased to ~ 0.8 above 150 nm. For each particle size, aerosol hygroscopicity was found to be
269 larger during new particle formation (NPF) periods, compared to non-NPF periods (Wu et al.,
270 2016), because more secondary species were found during NPF periods typically associated with
271 strong photochemical processes. Aerosol mass spectrometry (AMS) measurements suggested that
272 both aerosol hygroscopicity was dominated by inorganics, the contribution of which increased
273 with particle size and pollution levels (Wu et al., 2016). It was further found that the measured κ
274 could be well predicted using the AMS data, and the derived κ of organics depended linearly on
275 their O:C ratios (Wu et al., 2016).

276 The PKU site was affected by a series of biomass burning events in May-June 2014, and the
277 effect of biomass burning on aerosol composition and hygroscopicity was examined (Wu et al.,
278 2017). During biomass burning events, biomass burning contributed significantly to the production
279 and growth of aerosols in the Aitken mode, and the contribution of organics and black carbon to
280 mass concentrations of submicrometer aerosols reached 60% and 18% (Wu et al., 2017).
281 Hygroscopicity and number fractions of aerosols in the hydrophobic mode were relatively
282 invariable during biomass burning events, and the average κ , which showed no variation with
283 particles size (50-250 nm), were determined to be ~ 0.1 (Wu et al., 2017), substantially smaller than
284 those in the same period without significant impacts by biomass burning (Wu et al., 2016).



285

286 **Figure 1.** Change in average κ with aerosol diameter at the PKU site in four different seasons
287 between May 2014 to January 2015 (Wang et al., 2018c).

288

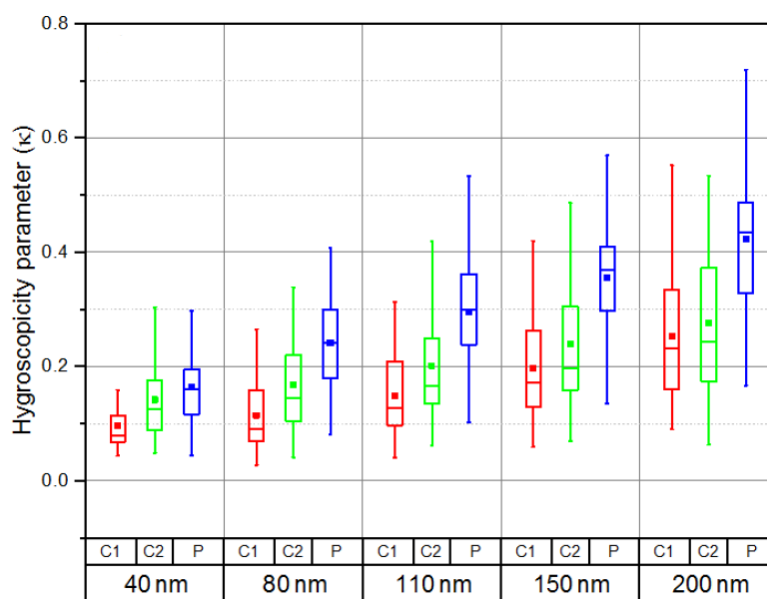
289 Seasonal variation of aerosol hygroscopic growth was investigated at the PKU site from May
290 2014 to January 2015 (Wang et al., 2018c), and the result is displayed in Figure 1. Average κ
291 increased significantly with particle size (50-350 nm) in summer and autumn, when strong
292 photochemical processes enhanced secondary aerosol formation and led to particle growth (Wang
293 et al., 2018c); in fact, number fractions of particles in the hydrophilic mode increased with
294 pollution levels, and they dominated the accumulation mode when $PM_{2.5}$ mass concentration
295 exceeded $100 \mu\text{g}/\text{m}^3$. In contrast, as shown in Figure 1, average κ only increased slightly with
296 particles size (50-350 nm) in spring while decreased substantially with particle size (75-350 nm)
297 in winter (Wang et al., 2018c), indicating significant contribution of primary species to aerosol
298 particles. Furthermore, being different to summer and autumn, substantial amounts of aerosol
299 particles in the hydrophobic mode were always observed in spring and winter (Wang et al., 2018c).
300 Another important feature revealed by Figure 1 is that for 150-350 nm aerosols, the hygroscopicity



301 was always highest in summer and lowest in winter (Wang et al., 2018c), and the difference
302 between the two seasons increased with particle size.

303 In addition, aerosol hygroscopic growth was investigated in March-April 2015 at the roof of
304 the Environmental Science Building (40°0'17"N, 116°19'34"E) on the campus of Tsinghua
305 University (Fajardo et al., 2016). This site, very close to the PKU site, is usually affected by the
306 same air masses. Number size distributions under dry and ambient conditions were measured for
307 10-500 nm particles to explore aerosol hygroscopicity under ambient RH (Fajardo et al., 2016).
308 No obvious aerosol growth was observed for RH below 50% (Fajardo et al., 2016); however, the
309 aerosol volume was increased by ~80% when RH reached 50%, and further increase in ambient
310 RH led to further hygroscopic growth.

311 **IAP site:** The IAP site is located at the Institute of Atmospheric Physics, Chinese Academy
312 of Science (39.97°N, 116.37°E) between the third and fourth ring roads in northern Beijing. All the
313 aerosol hygroscopic growth measurements (Wang et al., 2017c; Wang et al., 2019b; Fan et al.,
314 2020; Jin et al., 2020) were conducted at 90% RH at the ground level.



315

316 **Figure 2.** Size-resolved κ during the control clean (C1), the non-control clean (C2) and the non-
317 control polluted (P) periods. Solid squares represent the average κ , boxes represent the 25th, 50th,
318 and 75th percentiles, and extremities represent the 5th and 95th percentiles. Reprint with
319 permission by Wang et al. (Wang et al., 2017c). Copyright 2017 Copernicus Publications.

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321 Wang et al. (Wang et al., 2017c) investigated aerosol hygroscopic growth at the IAP site in
322 August-October 2015, when emission control measures were implemented for the 2015 China
323 Victory Day parade. Three periods with different pollution levels, including the control clean (C1),
324 the non-control clean (C2) and the non-control polluted (P) periods, were specifically examined to
325 evaluate the effect of emission control. Figure 2 shows that aerosol hygroscopicity increased with
326 particle size and pollution level (Wang et al., 2017c), due to enhanced contribution of secondary
327 species. For example, κ increased from 0.100 at 40 nm to 0.250 at 200 nm during C1, from 0.140
328 at 40 nm to 0.280 at 200 nm during C2, and from 0.160 at 40 nm to 0.420 at 200 nm during the



329 polluted period (Wang et al., 2017c). Furthermore, number fractions of particles in the more
330 hygroscopic mode increased in the polluted period, compared to C1 and C2. For 40 nm particles,
331 a quasi-unimodal hygroscopicity distribution was observed during C1, while bimodal or quasi-
332 trimodal distributions were observed during the other two periods; in contrast, bimodal patterns
333 were always observed for 150 nm particles (Wang et al., 2017c). It was also found that for all the
334 three periods, the average κ were always larger during the daytime than the nighttime (Wang et al.,
335 2017c).

336 A following study (Wang et al., 2019b) measured aerosol hygroscopic growth at the IAP site
337 in November-December 2016. Overall the average κ were found to increase with particle size,
338 from 0.164 at 40 nm to 0.230 at 200 nm during the clean period and from 0.155 at 40 nm to 0.290
339 at 200 nm during the polluted period (Wang et al., 2019b); compared to the clean period, the
340 average κ during the polluted period were smaller for 40 nm particles but larger for 80-200 nm
341 particles. In addition, bimodal distributions were always observed (Wang et al., 2019b). Number
342 fractions of particles in the less-hygroscopic mode was larger for 40 nm particles and smaller for
343 80-200 nm particles during the polluted period (Wang et al., 2019b), when compared to the clean
344 period, reflecting the compositional variation in 40 and 80-200 nm particles during the two periods.
345 Diurnal variation of aerosol hygroscopicity was also explored, displaying significant differences
346 between clean and polluted periods (Wang et al., 2019b).

347 Jin et al. (Jin et al., 2020) further analyzed size-resolved aerosol composition and
348 hygroscopicity measured at the IAP site in November-December 2016 (Wang et al., 2019b). The
349 size-dependent κ derived from measured GF at 90% RH was used to calculate ALWC at ambient
350 RH, assuming that a constant κ could be used to calculate GF at different RH (Jin et al., 2020); in
351 addition, size-resolved aerosol composition measured using AMS was used as input in



352 ISORROPIA-II to simulate ALWC at ambient RH. ALWC simulated using ISORROPIA-II were
353 found to be significantly smaller than calculated ALWC when RH was <60% (Jin et al., 2020),
354 because ISORROPIA-II failed to estimate water uptake by organics at low RH. Overall, organic
355 materials were estimated to contribute to (30±22)% of ALWC (Jin et al., 2020), highlighting the
356 importance of organics to aerosol hygroscopicity in urban Beijing.

357 Fan et al. (Fan et al., 2020) further conducted aerosol hygroscopic growth measurements at
358 the IAP site in May-June 2017, and bimodal hygroscopicity distributions were also observed for
359 40-200 nm aerosols. The summertime measurement in 2017 was compared with the wintertime
360 measurement at the same site in 2016 (Wang et al., 2019b), and the size dependence of aerosol
361 hygroscopicity was found to differ for the two seasons (Fan et al., 2020). The average κ increased
362 from 0.158 at 40 nm to 0.271 at 110 nm in winter, and further increase in particle size (to 200 nm)
363 led to slight decrease in κ (Fan et al., 2020); for comparison, the average κ increased with particles
364 size in summer, from 0.211 at 40 nm to 0.267 at 200 nm (Wang et al., 2019b). It was suggested
365 that the size dependence of aerosol hygroscopicity was mainly determined by the size-resolved
366 mass fractions of secondary inorganic species (Fan et al., 2020).

367 **CAMS site:** Wang et al. (Wang et al., 2018a) measured aerosol hygroscopic growth (30-90%
368 RH) of ambient aerosols on the campus of Chinese Academy of Meteorological Sciences, located
369 between the second and third ring roads in west Beijing. Measurements were conducted on a
370 building roof (~53 m above ground level) in December 2016, and the distance between the site
371 and a major road with heavy traffic was <200 m. Aerosol hygroscopic growth displayed unimodal
372 when RH did not exceed 60%, while bimodal distributions were usually observed at 70% and 80%
373 RH; in addition, aerosol hygroscopic growth occasionally exhibited trimodal distribution at 85%
374 and 90% RH (Wang et al., 2018a). Measured GF at 90% RH were used to calculate κ , which were



375 determined to be 0.010-0.015 and 0.286-0.358 for the hydrophobic and hydrophilic modes (Wang
376 et al., 2018a), both increasing with particle size (50-200 nm). Number fractions of hydrophobic
377 particles exceeded 50% at 50 and 100 nm, while hydrophilic particles frequently became dominant
378 in terms of number concentrations at 150 and 200 nm (Wang et al., 2018a). In addition,
379 hygroscopicity decreased at 50 nm but increased at 200 nm during heavily polluted periods (Wang
380 et al., 2018a), indicating their difference in compositions and sources.

381 **3.1.2 Rural sites in Beijing**

382 Aerosol hygroscopic growth were measured at two rural sites in Beijing, including Yufa
383 (Achtert et al., 2009) and Huairou (Wang et al., 2020b). The Yufa site (39.51°N, 116.31°E) is ~1.2
384 km away from a high-traffic expressway and ~50 km south to urban Beijing, and can be considered
385 as a representative rural and regional background site. Achtert et al. (Achtert et al., 2009) measured
386 aerosol hygroscopic growth as a function of RH (56, 76, 85 and 91%) on a four-floor building (22
387 m above the ground) at this site in August-September 2006. GF at 91% RH, ranging from 1.15 to
388 1.80 for 30-300 nm particles, were found to be larger in the accumulation mode than the Aitken
389 mode (Achtert et al., 2009); furthermore, increase in mass fractions of sulfate during polluted
390 periods led to increase in aerosol hygroscopicity with pollution level. Diurnal variation of aerosol
391 hygroscopicity was also explored (Achtert et al., 2009): hygroscopicity was found to be higher in
392 the daytime than the nighttime for the Aitken mode, whereas no significant difference in
393 hygroscopicity was observed between daytime and nighttime for the accumulation mode.

394 The Huairou site (40.42°N, 116.69°E) is located on the campus of the University of the
395 Chinese Academy of Sciences, ~60 km northeast from the center of Beijing. It was mainly
396 influenced by regional transport of pollutants from downtown Beijing (Tan et al., 2018) and small
397 local sources nearly (e.g., moderate traffic and small residential areas). Aerosol hygroscopic



398 growth (at 90% RH) was measured at this site in January-March 2016 (Wang et al., 2020b). The
399 average κ were determined to be 0.162-0.208 for 50-300 nm particles (Wang et al., 2020b), and
400 mass fractions of nitrate, which contributed significantly to aerosol hygroscopic growth, reached
401 44% during polluted episodes.

402 3.1.3 Other urban/suburban sites

403 Aerosol hygroscopic growth was measured at other four urban/suburban sites in NCP,
404 including two sites in Tianjin, one site in Hebei Province and one site in Shanxi Province.

405 **Tianjin:** The Wuqing site is located next to the Wuqing Meteorological Station (39°23'N,
406 117°0'E) in the west area of Wuqing (Tianjin), surrounded by mixed agricultural, residential and
407 industrial regions. This site is a good place to study regional air pollution in NCP, as it is ~30 km
408 northwest to the urban Tianjin, ~80 km southeast to the urban Beijing, ~130 km southwest to
409 Tangshan (Hebei), and ~160 km northeast to Baoding (Hebei). Aerosol hygroscopic growth was
410 measured at three RH (90%, 95% and 98.5%) at this site in July-August 2009 (Liu et al., 2011).
411 Bimodal hygroscopicity distribution was observed over the whole period, and the average κ ,
412 derived from GF measured at 90% RH, increased from 0.250 at 50 nm to 0.340 at 250 nm (Liu et
413 al., 2011). Compared to the nighttime, both the average κ and number fractions of particles in the
414 more-hygroscopic mode were larger during the daytime (Liu et al., 2011). The average κ were
415 found to increase with particle size for the more-hygroscopic mode, from 0.310 at 50 nm to 0.390
416 at 250 nm (Liu et al., 2011); in contrast, they decreased with particle size for the nearly
417 hydrophobic mode, from 0.054 at 50 nm to 0.025 at 250 nm. It was found that inorganics play an
418 important role for hygroscopic growth of the accumulation mode (Liu et al., 2014), while organics
419 were very importance for hygroscopic properties of the Aitken mode. In addition, κ calculated



420 from aerosol compositions measured offline were consistent with those derived from H-TDMA
421 measurements (Liu et al., 2014).

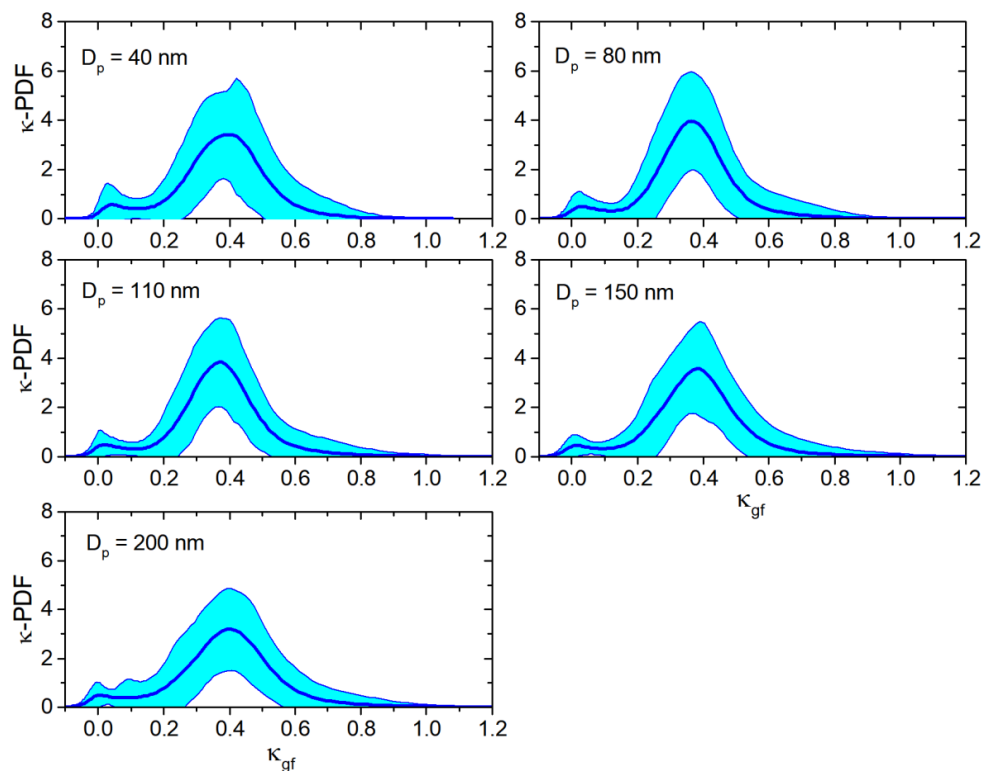
422 Two different methods were used to estimate ALWC at the Wuqing site in July-August 2009
423 (Bian et al., 2014). For the first method, κ derived from GF measurements at 90-98.5% RH were
424 assumed to be constant at different RH, and thus ALWC could be calculated from particle number
425 size distribution (Bian et al., 2014); for the second method, size-resolved aerosol composition,
426 only taking into account water soluble inorganic ions, was used as input in ISORROPIA-II to
427 predict ALWC. ALWC estimated using the first method agreed with those using the second method
428 for >60% RH, but was much larger compared to the second method when ambient RH was <60%
429 (Bian et al., 2014).

430 In March 2018, Ding et al. (Ding et al., 2019) carried out aerosol hygroscopic growth
431 measurements (70-85% RH) at the NKU site, an air quality research supersite at Nankai University
432 (38°59'N, 117°20'E), which was ~20 km away from downtown Tianjin. GF measured at 85% RH
433 were used to calculate average κ , being 0.301-0.477, 0.203-0.386 and 0.281-0.419 on 13th, 14th
434 and 15th March (Ding et al., 2019). In addition, the average κ were found to be larger during
435 polluted periods than clean periods, as the contribution of nitrate, sulfate and ammonium in the
436 accumulation mode increased during polluted periods (Ding et al., 2019). It was also found that
437 for the accumulation mode, κ were larger in the nighttime than the daytime (Ding et al., 2019).
438 Water-soluble inorganic ions measured offline were used as input in the ISORROPIA-II to predict
439 aerosol hygroscopicity, and measured and predicted κ showed good agreement (Ding et al., 2019),
440 implying that the contribution of organics to aerosol hygroscopic growth was quite limited.

441 **Hebei Province:** The Xingtai site is located at the National Meteorological Basic Station in
442 Xingtai (37.18°N, 114.37°E), a heavily polluted city in the center of NCP, and aerosol hygroscopic



443 growth (at 85% RH) was measured at this site in May-June 2016 (Wang et al., 2018b). As shown
444 in Figure 3, quasi-unimodal aerosol hygroscopicity distribution was observed and number
445 fractions of particles in the more-hygroscopic mode was $\sim 90\%$ for 40-200 nm particles (Wang et
446 al., 2018b), indicating that they were highly aged and internally mixed. The average κ were
447 determined to be 0.364-0.39 (Wang et al., 2018b), significantly larger than those reported for most
448 of other sites in NCP. No obvious dependence of average κ on particle size was observed, and the
449 average κ were found to be larger in daytime than nighttime, especially during new particle
450 formation events.



451

452 **Figure 3.** Mean probability density functions of κ and their standard deviations (shaded areas) for
453 40, 80, 110, 150 and 200 nm particles at the Xingtai site in May-June 2016, as derived from



454 measured GF at 85% RH. Reprint with permission by Wang et al. (Wang et al., 2018b). Copyright
455 2018 Copernicus Publication.

456

457 For the campaign at the Xingtai site in May-June 2016 (Wang et al., 2018b), aerosol
458 hygroscopicity on a clean day (21 May) was compared with that on a highly polluted day (23 May).
459 Aerosol hygroscopicity was higher on the polluted day (Chen et al., 2019), likely due to the
460 enhanced formation of nitrate as revealed by ACSM (aerosol chemical speciation monitor)
461 measurements. Furthermore, aerosol hygroscopicity increased with particles size (40-200 nm) on
462 both days, from 0.288 to 0.339 on 21 May and from 0.325-0.352 on 23 May (Chen et al., 2019).

463 **Shanxi Province:** The Xinzhou site (38.24°N, 112.43°E, 1500 m above sea level) was located
464 on the border between the NCP and the Loess Plateau. This suburban and regional site, surrounded
465 by agricultural land with limited local anthropogenic emissions, was located ~360 km southwest
466 to Beijing, ~78 km northwest to Taiyuan and ~10 km south to the city nearby. Aerosol hygroscopic
467 growth (85% RH) was investigated for 25-200 nm aerosols at this site in July-August 2014 (Zhang
468 et al., 2017). Quasi-unimodal aerosol hygroscopicity distribution was observed, indicating highly
469 aged and internally mixed particles. The average κ were determined to be 0.420-0.528,
470 significantly larger than those observed at other sites in the NCP; in addition, no obvious
471 dependence of κ on particle size was found (Zhang et al., 2017).

472 **3.1.4 Other rural sites**

473 Aerosol hygroscopic growth was measured at other two rural sites in NCP, i.e. the Xianghe
474 site and the Wangdu site (both in Hebei). The Xianghe site (39.75°N, 116.96°E), surrounded by
475 residential areas and farmlands, is considered as a typical rural site in NCP and is located ~5 km
476 west to the center of Xianghe town and ~70 km southeast to Beijing. At this site, aerosol



477 hygroscopic growth (at 87% RH) was measured in July-August 2013 (Zhang et al., 2016b).
478 Trimodal aerosol hygroscopicity distributions were observed for 50-350 nm particles (Zhang et al.,
479 2016b), and the average κ were determined to be 0.020-0.056, 0.170-0.211 and 0.365-0.455 for
480 nearly-hydrophobic, less-hygroscopic and more-hygroscopic modes. Aerosol hygroscopicity
481 showed some dependence on air masses (Zhang et al., 2016b): air masses which were transported
482 from the north with high speed winds typically contained larger number fractions of hydrophobic
483 species and exhibited lower hygroscopicity, whereas no obvious difference in aerosol
484 hygroscopicity and mixing state were observed for other air masses.

485 The Wangdu site (38.71°N, 115.16°E), a rural site located in the center area of NCP, was ~200
486 km southwest to Beijing, and aerosol hygroscopic growth (at 90% RH) was measured at this site
487 in June 2014 (Wang et al., 2017b). Bimodal aerosol hygroscopicity distribution was always
488 observed (Wang et al., 2017b), and the average κ were found to increase with particle size, from
489 0.240 at 30 nm to 0.320 at 250 nm.

490 **3.2 Yangtze River Delta (YRD)**

491 A number of aerosol hygroscopic growth measurements have been carried out since 2009 in
492 three large cities (Shanghai, Hangzhou and Nanjing) in the Yangtze River Delta.

493 **3.2.1 Shanghai**

494 Ambient aerosol hygroscopic growth was measured at two sites in Shanghai (Ye et al., 2011;
495 Ye et al., 2013; Wang et al., 2014; Xie et al., 2017; Li et al., 2018; Wang et al., 2020a). The FDU
496 site (31°18'N, 121°29'E) is located on the building roof of Department of Environmental Science
497 and Engineering, Fudan University; the Pudong site (31.22°N, 121.55°E) is located in Pudong
498 Meteorological Bureau. Both sites are considered as urban sites, surrounded by residential,
499 industrial and traffic areas, and their distance is <10 km.



500 **FDU site:** At the FDU site, Ye et al. (Ye et al., 2011) measured aerosol hygroscopic growth
501 (30-200 nm) at 20-85% RH in January-February 2009. Bimodal hygroscopic growth distribution
502 was always observed at 85% RH, and κ derived from measured GF at 85% RH were determined
503 to be 0.027-0.063 and 0.291-0.381 for the less- and more-hygroscopic modes (Ye et al., 2011). The
504 average κ decreased with particle size for the less hygroscopic mode while increased with particle
505 size for the more hygroscopic mode (Ye et al., 2011); in addition, number fractions of particles in
506 the less hygroscopic mode decreased with particle size. The change in GF with RH (20-85%) was
507 also discussed for particles with different sizes (Ye et al., 2011).

508 Compositional data provided by ATOFMS (Aerosol Time-of-Flight Mass Spectrometry) were
509 used to interpret GF measured at 85% RH for 250 nm particles on 18-19 January and 10 February
510 2009 (Ye et al., 2011). Bimodal aerosol hygroscopicity distribution was observed for 250 nm
511 particles, including a nearly-hydrophobic mode with κ of 0.029-0.061 and a more-hygroscopic
512 mode with κ of 0.387-0.399 (Wang et al., 2014). Aerosols in the more-hygroscopic mode consisted
513 predominantly of secondary species (e.g., OC-amine, sulfate and nitrate), while biomass burning
514 aerosols, uncoated EC, secondary organic compounds, and dust/ash were frequently identified in
515 the nearly-hydrophobic mode (Wang et al., 2014).

516 Aerosol hygroscopic growth (at 85% RH) was also measured at this site in February-March
517 2014 (Wang et al., 2020a). Aerosol hygroscopicity was found to exhibit bimodal distribution at
518 250 nm, and the average κ were determined to be 0.029 and 0.376 for nearly-hydrophobic and
519 more-hydrophilic modes (Wang et al., 2020a). Nearly-hydrophobic particles typically included
520 biomass burning aerosol, fresh EC and high molecular mass OC, while more-hydrophilic particles
521 included aged EC, amine-rich particles, and etc. (Wang et al., 2020a). Furthermore, a statistic



522 method was developed to estimate aerosol hygroscopicity from single particles mass spectra
523 (Wang et al., 2020a).

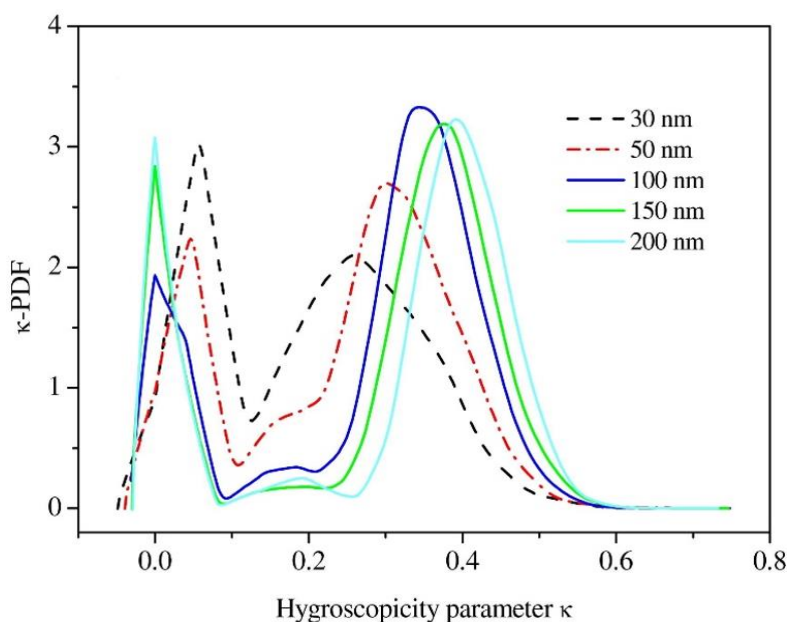
524 Xie et al. (Xie et al., 2017) further measured aerosol hygroscopic growth (83% RH) at the FDU
525 site in December 2014-January 2015. Bimodal aerosol hygroscopicity distribution (nearly
526 hydrophobic and more hygroscopic modes) was usually observed, and the average κ increased
527 from 0.161 at 40 nm to 0.345 at 400 nm (Xie et al., 2017). Number fractions of nearly hydrophobic
528 particles increased during polluted periods for all the sizes considered (40-400 nm), indicating
529 significant contribution of primary particles during haze events (Xie et al., 2017); however, the
530 increase in number fractions of nearly hydrophobic particles during pollution events were less
531 significant for larger particles, suggesting that primary emissions contributed more to smaller
532 particles.

533 Mixing state and hygroscopic growth (at 85% RH) were explored at the FDU site in July 2017
534 specifically for ambient black carbon (BC) aerosols (120, 240 and 260 nm) (Li et al., 2018).
535 Number fractions of BC particles decreased with particle size, from ~80% for 120 nm to ~60% for
536 360 nm. Hygroscopicity of BC particles displayed unimodal distribution, and their GF at 85% RH
537 peaked at ~1.0 (Li et al., 2018). Enhancement in hygroscopicity of BC particles, due to their aging
538 via condensation of secondary species, was frequently observed (Li et al., 2018): during the
539 nighttime nitrate contributed significantly to BC aging, while formation of secondary organic
540 materials played an important role during the daytime.

541 **Pudong site:** Aerosol hygroscopic growth (at 91% RH) was studied at the Pudong site in
542 September 2009 (Ye et al., 2013). As shown in Figure 4, aerosol hygroscopicity was found to be
543 trimodal, including a nearly-hydrophobic mode and a more-hygroscopic mode, as well as a less-
544 hygroscopic mode with much less abundance (Ye et al., 2013). The average κ increased from 0.270



545 at 30 nm to 0.390 at 200 nm for the more-hygroscopic mode (Ye et al., 2013), and decreased from
546 0.054 at 30 nm to 0.011 at 200 nm for the nearly-hydrophobic mode.



547
548 **Figure 4.** Probability distribution functions of the hygroscopicity parameter (κ) for 30, 50, 100,
549 150 and 200 nm aerosols at the Pudong site in September 2009. Reprint with permission by Ye et
550 al. (Ye et al., 2013). Copyright 2013 Elsevier Ltd.

551

552 3.2.2 Hangzhou

553 Up to now only one aerosol hygroscopic growth study was carried out in Hangzhou, at the ZJU
554 site located on the Huajiachi campus of Zhejiang University (30°16'N, 120°11'E). Aerosol
555 hygroscopic growth was measured at 70-90% RH (mainly at 82%) in December 2009-January
556 2010 (Zhang et al., 2011). Bimodal hygroscopicity distribution was observed for 50-200 nm
557 aerosols, while unimodal hygroscopicity distribution was observed for 30 nm aerosols (Zhang et
558 al., 2011). The average κ decreased from 0.121 at 30 nm to 0.065 at 80 nm for the low-hygroscopic



559 mode, and further increase in particle size (up to 200 nm) did not lead to significant change in κ
560 (Zhang et al., 2011). For comparison, the average κ increased from 0.303 at 30 nm to 0.343 at 80
561 nm for the more-hygroscopicity mode, and further increase in particle size only resulted in very
562 small increase in κ . In addition, number fractions of particles in the more-hygroscopic mode
563 increased from ~48% at 30 nm to ~70% at 100 nm, and remained nearly constant for 100-200 nm
564 (Zhang et al., 2011).

565 3.2.3 Nanjing

566 Aerosol hygroscopic growth was measured at three urban/suburban sites in Nanjing. The
567 NUIST site (32°207'N, 118°717'E) is a suburban site located on the 12th floor of the Meteorological
568 building at Nanjing University of Information Science and Technology, with several large
569 petrochemical factories and a busy expressway nearby. The NATC site (32.0°N, 118.7°E) is a
570 typical urban site at Nanjing Advanced Technical College, located in the centre business district
571 with heavy residential and traffic emissions. The JEMC site is an urban site on the 6th floor of the
572 building of Jiangsu Environmental Monitoring Centre (~18 m above the ground), located in the
573 urban area and surrounded by a variety of sources such as residence, restaurants, office blocks and
574 traffic.

575 **NUIST site:** Wu et al. (Wu et al., 2014) measured aerosol hygroscopic growth as a function of
576 RH (60-90%) at the NUIST site in May-July 2012, and bimodal hygroscopicity distributions were
577 frequently observed at 90% RH for 40-200 nm aerosols. For the more-hygroscopic mode, κ were
578 determined to be 0.294-0.349, increasing with particle size (except for 40 nm); while for the less-
579 hygroscopic mode, κ were found to decrease with particle size, from 0.079 at 40 nm to 0.040 at
580 200 nm (Wu et al., 2014). The average aerosol hygroscopicity measured at this site in Nanjing
581 seemed to be slightly lower than those reported in Beijing, Shanghai and Guangzhou.



582 Yang and co-workers further investigated aerosol (30-230 nm) hygroscopic growth (at 90%
583 RH) at this site in April-May 2014 (Xu et al., 2015; Yang et al., 2019), and bimodal hygroscopicity
584 distribution was observed. The κ values were found to be very low (close to 0) for the low-
585 hygroscopic mode, and decreased from 0.232 at 30 nm to 0.186 at 230 nm for the medium
586 hygroscopic mode. Aerosol hygroscopicity measured in April-May 2014 (Xu et al., 2015; Yang et
587 al., 2019) were significantly lower than that measured in May-June 2012 at the same site (Wu et
588 al., 2014). One possible reason was that in April-May 2014 organic species made a large
589 contribution to submicrometer aerosols (21-38% by mass) (Xu et al., 2015; Yang et al., 2019), thus
590 leading to substantial decrease in aerosol hygroscopicity.

591 **NATC site:** In August 2013, Li et al. (Li et al., 2015b) investigated hygroscopic growth at 90%
592 RH for 32-350 nm aerosols. A less-hydrophobic mode (κ : 0.017-0.031) and a more-hygroscopic
593 mode (κ : 0.178-0.229) were observed during the campaign (Li et al., 2015b). Aerosol
594 hygroscopicity reported at the NATC site in August 2013 (Li et al., 2015b) was lower than these
595 reported at the NUIST site in May-June 2012 (Wu et al., 2014) and in April-May 2014 (Xu et al.,
596 2015; Yang et al., 2019), perhaps because the contribution of low hygroscopic primary particles
597 (e.g., soot) from local emission was larger at the NATC site (an urban site), compared to the
598 NUIST site (a suburban site).

599 **JEMC site:** At the JEMS site, 40-200 nm aerosol hygroscopic growth was measured at 85%
600 RH in January-February 2015 (Zhang et al., 2018). The average κ were determined to be 0.200-
601 0.271 for 40-200 nm particles (Zhang et al., 2018), significantly larger than those (0.081-0.126 for
602 32-350 nm particles) reported for the NATC site in August 2013 (Li et al., 2015b), and the reason
603 was unclear. Bimodal hygroscopicity distribution was also observed (Zhang et al., 2018); similar
604 to two previous studies in Nanjing (Wu et al., 2014; Li et al., 2015b), number fractions of particles



605 in the low hygroscopic mode and their average κ both decreased with particle size, while the
606 average κ increased with particle size for the more hygroscopic mode (except for 40 nm).

607 **3.3 Pearl River Delta (PRD)**

608 A series of aerosol hygroscopic growth studies were conducted in PRD, to be more specific, at
609 two rural sites (Xinken and Wanqinsha) and one suburban site (Panyu) in Guangzhou and one
610 suburban site (HKUST) in Hong Kong.

611 **3.3.1 Rural sites in Guangzhou**

612 The Xinken site (22.6°N, 113.6°E), located near the Pearl River estuary, is ~50 km southeast
613 to urban Guangzhou, and the Wanqinsha site is located ~9 km northwest of Xinken. Both are
614 typical rural background sites with no major pollution sources nearby, and air quality at both sites
615 are affected by regional transport combined with limited local sources, such as traffic, ships,
616 biomass burning and cooking (Cheng et al., 2006; Eichler et al., 2008; Kim et al., 2011).

617 Eichler et al. (Eichler et al., 2008) measured aerosol hygroscopic growth (30-91% RH) at the
618 Xinken site in October-November 2004. The average GF at 91% RH were determined to 1.45,
619 1.53, 1.6 and 1.56 for 80, 140, 250 and 380 nm particles (Eichler et al., 2008), corresponding to κ
620 of 0.244, 0.283, 0.324 and 0.288, respectively. Inorganic aerosol compositions measured offline
621 were used to calculate GF, and the average difference between the measured and calculated GF
622 was found to be <8% (Eichler et al., 2008), suggesting that the contribution of organics to aerosol
623 hygroscopicity was rather small.

624 In a following study (Kim et al., 2011), aerosol hygroscopic growth (at 85% RH) of ultrafine
625 particles (40, 50, 60 and 80 nm) was investigated at the Wanqinsha site in October-November 2008.
626 During photochemical events, GF varied between 1.13 and 1.55, and particles consisted mainly of
627 ammonium sulfate and organic materials (Kim et al., 2011). For comparison, during combustion



628 events (i.e. affected by biomass burning and traffic emission), aerosol particles were mainly
629 composed of non-hygroscopic carbonaceous species and smaller amounts of potassium, and
630 correspondingly measured GF were reduced to 1.05-1.15 (Kim et al., 2011).

631 **3.3.2 Urban/suburban sites in Guangzhou**

632 The Panyu site, located at the top of Mt. Dazhengang (23°00'N, 113°21'E, 150 m above the sea
633 level), is surrounded by residential areas without major pollution sources nearby and can be
634 considered as a suburban site in Guangzhou (Tan et al., 2013). Several aerosol hygroscopic growth
635 measurements at 90% RH have been carried out at this site since 2011 (Tan et al., 2013; Jiang et
636 al., 2016; Cai et al., 2017; Tan et al., 2017; Cai et al., 2018; Hong et al., 2018; Liu et al., 2018).

637 Aerosol hygroscopic growth was first measured at this site in November-December 2011 (Tan
638 et al., 2013). Bimodal hygroscopicity distributions were observed for 40, 80, 110, 150 and 200 nm
639 particles, and κ were determined to be 0.045-0.091 and 0.290-0.323 for the less- and more-
640 hygroscopic modes (Tan et al., 2013). In general, both hygroscopicity and number fractions
641 increased with particle size for the more-hygroscopic mode, whereas they both decreased with
642 particle size for the less-hygroscopic mode. Average hygroscopicity was found to be larger during
643 the daytime than the nighttime for both modes (Tan et al., 2013), and hygroscopicity and number
644 fractions of particles in the more-hygroscopic mode increased during polluted periods, when
645 compared to clean periods.

646 Jiang et al. (Jiang et al., 2016) compared aerosol hygroscopicity measured at this site between
647 winter (December 2012-January 2013) and summer (July-September 2013), and no obvious
648 difference in average κ was found between the two seasons. Trimodal hygroscopicity distributions
649 were observed for 40-200 nm particles, and κ were determined to be 0.290-0.339, ~0.15 and ~0.015
650 for more-, less- and non-hygroscopic modes (Jiang et al., 2016). Similar to the work by Tan et al.



651 (Tan et al., 2013), hygroscopicity and number fractions increased with particle size for the more-
652 hygroscopic mode, with no distinct difference between winter and summer (Jiang et al., 2016); for
653 the non-hygroscopic mode, hygroscopicity and number fractions both decreased with particle size,
654 and their number fractions were slightly lower in winter than in summer. Furthermore, the average
655 κ were larger during daytime than nighttime for both seasons, and the diurnal variation was more
656 profound in summer (Jiang et al., 2016).

657 Tan et al. (Tan et al., 2017) measured aerosol hygroscopic growth in January-March 2014, and
658 κ increased from 0.204 at 40 nm to 0.312 at 200 nm. The κ values derived from GF measured at
659 90% RH were used to calculate ALWC under ambient conditions, and meanwhile aerosol inorganic
660 species measured were used as input in ISORROPIA-II to predict ALWC. Good agreement
661 between calculated and predicted ALWC were found for RH >70%, but significant differences
662 were found at <70% RH (Tan et al., 2017). Liu et al. (Liu et al., 2018) further explored aerosol
663 hygroscopic growth measured in February-March 2014 at this site, and found that the average κ
664 values increased from 0.261 at 80 nm to 0.323 at 200 nm. In addition, bimodal hygroscopicity
665 distribution was observed, and κ increased from 0.382 at 80 nm to 0.432 at 200 nm for the more
666 hygroscopic mode (Liu et al., 2018).

667 Aerosol hygroscopic growth (at 90% RH) were further measured at this site in November-
668 December 2014 (Cai et al., 2017; Cai et al., 2018). Bimodal hygroscopicity distributions were
669 observed for 40-200 nm particles, and the average κ increased with particle size, from 0.213 at 40
670 nm to 0.312 at 200 nm. The κ values derived from size-resolved chemical compositions measured
671 using AMS were significantly lower than those derived from GF measurements (Cai et al., 2017;
672 Cai et al., 2018), probably because using a constant κ value (0.1) may underestimate
673 hygroscopicity of aerosol organics.



674 Aerosol composition and hygroscopic growth at 90% RH were investigated at this site in
675 September-October 2016 (Hong et al., 2018), using an ACSM and a H-TDMA. Bimodal
676 hygroscopicity distributions were observed; the more-hygroscopic mode was dominant at 100 and
677 145 nm, while less- and more-hygroscopic modes were of similar magnitude at 30 and 60 nm
678 (Hong et al., 2018). The average aerosol hygroscopicity increased with particle size, and no
679 obvious diurnal variation was observed (Hong et al., 2018); however, aerosol hygroscopicity was
680 higher during the daytime for the less-hygroscopic mode while slightly lower in the afternoon for
681 the more-hygroscopic mode. Hygroscopicity closure analysis suggested that taking into account
682 the dependence of GF on composition for organics led to better agreement between measured and
683 calculated GF (Hong et al., 2018). It was further found that GF increased linearly with O:C ratios
684 for organics, and the derived GF appeared to be less sensitive to the changes of O:C ratios during
685 polluted periods.

686 3.3.3 Hong Kong

687 Since 2011, H-TDMA and online mass spectrometry were employed by Chan and co-workers
688 (Lopez-Yglesias et al., 2014; Yeung et al., 2014; Cheung et al., 2015) to investigate aerosol
689 composition and hygroscopic growth at the HKUST supersite (22°20'N, 114°16'E) on the east coast
690 of Hong Kong. It is a typical suburban and coastal site with no major pollution sources nearby.

691 Aerosol hygroscopic growth at 90% RH was first investigated at this site in 2011 (Yeung et al.,
692 2014), and bimodal aerosol hygroscopicity distributions were observed with a dominant more-
693 hygroscopic mode and a weak less-hygroscopic mode at 75, 100, 150 and 200 nm. The average κ
694 were determined to be 0.330-0.360 during May, 0.370-0.390 during the first half of September,
695 0.210-0.250 during the second half of September and 0.290-0.320 during November (Yeung et al.,



696 2014), caused by compositional variations in different air masses; however, no obvious
697 dependence of average κ on particle size was found.

698 Number fractions of particles in the more-hygroscopic mode were always >0.8 (Yeung et al.,
699 2014), except for 75 nm particles in the second half of September (~ 0.45) which was dominantly
700 affected by continental air masses. When compared to maritime aerosols, hygroscopicity of
701 aerosols in the more-hygroscopic mode was substantially lower for continental aerosols which
702 contained larger proportions of organic materials (Yeung et al., 2014). Hygroscopicity closure
703 analysis suggested that using a constant GF (1.18) at 90% RH for organic materials, instead of
704 considering the dependence of GF on their oxidation degree, would lead to better agreement
705 between measured and calculated GF (Yeung et al., 2014), likely because inorganic species (such
706 as sulfate) contributed dominantly to the overall aerosol hygroscopicity during the entire campaign.

707 In addition, hygroscopic growth at the HKUST site was investigated as a function of RH (10-
708 90%) in 2011-2012 (Lopez-Yglesias et al., 2014; Cheung et al., 2015), and both hysteresis behavior
709 and continuous hygroscopic growth of ambient aerosols were observed.

710 **3.4 Other locations**

711 In addition to NCP, YRD and PRD, measurements of aerosol hygroscopic growth were also
712 conducted in other regions in China, as discussed below.

713 **Taipei:** Hygroscopic growth (15-90% RH) was investigated for 53, 82, 95 and 202 nm aerosols
714 at an urban site in Taipei (Taiwan Province) in October-December 2001 (Chen et al., 2003).
715 Bimodal hygroscopicity distribution was observed for all the particles at 90% RH: while κ (0.049-
716 0.068) showed no obvious dependence on particle size for the less hygroscopic mode, they
717 increased from 0.274 at 53 nm to 0.422 at 202 nm for the more hygroscopic mode (Chen et al.,
718 2003). No obvious hygroscopic growth was observed at $<45\%$ RH (Chen et al., 2003), and bimodal



719 hygroscopic growth behavior appeared at ~76% RH for all the sizes (53-202 nm), becoming more
720 noticeable with further increase in RH.

721 **Mt. Huang:** Mt. Huang (30°08'N, 118°09'E) is located in the mountainous area of east China
722 with large forest coverages and limited anthropogenic activities. Aerosol hygroscopic growth at
723 50-85% RH was examined in September-October 2012 at the mountain foot (~464 m above the
724 sea level) and the mountain top (~1860 m above the sea level) (Wu et al., 2018a). No significant
725 particle growth was observed below 60% RH at both sites, and bimodal growth behavior appeared
726 at ~75% RH except 40 nm particles and became more evident at higher RH (Wu et al., 2018a).
727 Hygroscopicity was higher in the daytime than the nighttime for both modes. In addition,
728 hygroscopicity was slightly higher at the mountain foot than the mountain top for both modes
729 (except 200 nm particles in the more-hygroscopic mode) (Wu et al., 2018a); the reason was that
730 more secondary inorganic species were formed at the mountain foot due to human activities, while
731 on the mountain top the contribution of organics increased. Compared to NCP, YRD and PRD sites,
732 the overall aerosol hygroscopicity was lower at Mt. Huang (Wu et al., 2018a), as it is located in a
733 clean region with smaller fractions of secondary inorganic aerosols.

734 In July 2014 aerosol hygroscopic growth (at 85% RH) was further studied at the top of Mt.
735 Huang (Xu, 2015; Chen et al., 2016; Wang et al., 2016). The average κ were determined to be
736 0.275, 0.266 and 0.290 at 70, 150 and 230 nm (Chen et al., 2016; Wang et al., 2016), in good
737 agreement with the previous study conducted at the same site in 2012 (Wu et al., 2018a). At a
738 given particle size, aerosol hygroscopicity was found to be higher in the daytime than the nighttime
739 (Chen et al., 2016; Wang et al., 2016); furthermore, aerosol hygroscopicity was higher for air
740 masses from northwest than those from southeast. The derived κ depended positively on mass
741 fractions of inorganics and negatively on organics (Chen et al., 2016; Wang et al., 2016). In



742 addition, unimodal aerosol hygroscopicity distribution occurred with high frequency (47.5%)
743 during the campaign, and it also appeared more frequently in the afternoon with GF (at 85% RH)
744 in the range of 1.25-1.45 (Chen et al., 2016; Wang et al., 2016).

745 **Shouxian:** In June-July 2016, Qian et al. (Qian et al., 2017) studied hygroscopic growth (at
746 90% RH) of 50-250 nm aerosols at Shouxian National Climate Observatory (32°26'N, 116°48'E)
747 in east China, a rural site surrounded by farmlands at Shouxian, Anhui Province. Bimodal aerosol
748 hygroscopicity distribution was observed, and the average κ increased with particle size, from
749 0.129 at 50 nm to 0.279 at 250 nm (Qian et al., 2017).

750 **East China Sea:** Total suspended particles were collected during a cruise over the East China
751 Sea (22-35°N and 119-126°E) in May-June 2014 and dissolved in deionized water. The resulting
752 solutions were atomized to generated aerosols, and their hygroscopic growth was then measured
753 at 5-90% RH (Yan et al., 2017). The average κ was determined to be 0.88 for the whole cruise, and
754 the daytime average (0.81) was smaller than the nighttime average (0.95) (Yan et al., 2017), due
755 to less chloride loss in the nighttime. It is to be assessed to which extent aerosols generated by Yan
756 et al. (Yan et al., 2017) can actually mimic ambient aerosols.

757 **3.5 Summary**

758 Geographically speaking, almost all the aerosol hygroscopic growth studies were conducted in
759 east China, especially in NCP, YRD and PRD. Aerosol hygroscopic growth in other regions in
760 China remains to be explored, and measurements at rural and remote areas with limited
761 anthropogenic impacts are very scarce. In addition, previous measurements were mainly
762 performed at or close to the ground level, except these carried out on the top of Mt. Huang (Chen
763 et al., 2016; Wang et al., 2016; Wu et al., 2018a).



764 It can be concluded that submicrometer aerosols in China usually exhibit bimodal
765 hygroscopicity distribution (i.e. nearly-hydrophobic and more-hygroscopic modes). However,
766 trimodal distributions, with a medium-hygroscopic mode with limited importance, were also
767 reported by several studies (Massling et al., 2009; Meier et al., 2009; Ye et al., 2013; Jiang et al.,
768 2016; Zhang et al., 2016b; Wang et al., 2017c; Wang et al., 2018a), and quasi-unimodal
769 hygroscopicity distributions existed but were quite sparse (Chen et al., 2016; Wang et al., 2016;
770 Wang et al., 2017c; Zhang et al., 2017; Wang et al., 2018b).

771 For the more-hygroscopic mode, κ usually increased with particle size, except for the
772 measurements carried out at HKUST site (Yeung et al., 2014) where no obvious dependence on
773 particle diameter was found. For the nearly-hydrophobic mode, κ usually decreased with particle
774 size (Liu et al., 2011; Ye et al., 2011; Zhang et al., 2011; Tan et al., 2013; Ye et al., 2013; Wu et al.,
775 2014; Jiang et al., 2016; Zhang et al., 2016b; Qian et al., 2017; Zhang et al., 2018), though different
776 results were also reported in several studies (Chen et al., 2003; Massling et al., 2009; Meier et al.,
777 2009; Li et al., 2015b; Wang et al., 2018a; Wu et al., 2018a).

778 Average aerosol hygroscopicity, especially for the more-hygroscopic mode, usually increased
779 with pollution levels (Massling et al., 2009; Wu et al., 2016; Wang et al., 2017c; Wang et al., 2018a;
780 Chen et al., 2019; Ding et al., 2019; Wang et al., 2019b), attributed to increased mass fractions of
781 secondary inorganic aerosols. However, different results were also reported (Meier et al., 2009),
782 especially for particles at or below 50 nm (Achtert et al., 2009; Wang et al., 2018a; Wang et al.,
783 2019b) for which primary emissions could play an important role.

784 A few studies examined aerosol hygroscopic growth at different seasons (Massling et al., 2009;
785 Jiang et al., 2016; Wang et al., 2018c; Fan et al., 2020). No obvious difference in the overall aerosol
786 hygroscopicity was observed between summer and winter at the PKU site (Beijing) (Massling et

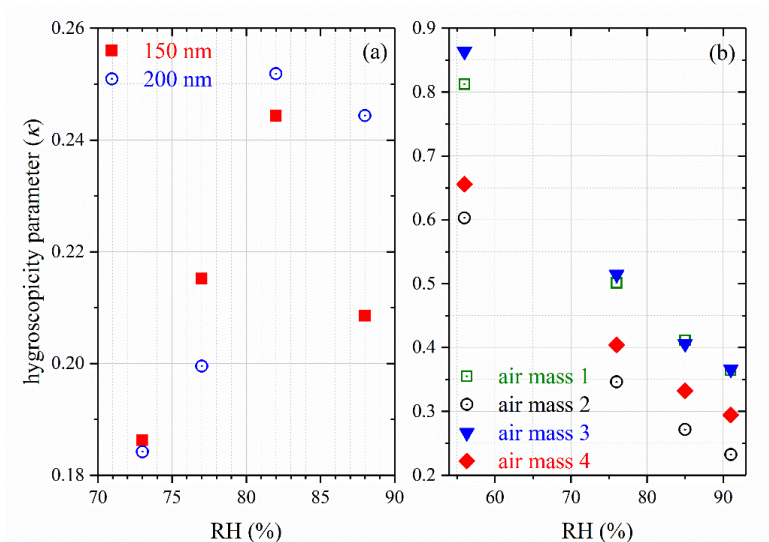


787 al., 2009), the IAP site (Beijing) (Fan et al., 2020) and the Panyu site (Guangzhou) (Jiang et al.,
788 2016). However, one study (Wang et al., 2018c) suggested that the overall hygroscopicity, and
789 especially hygroscopicity of 150-350 particles, was highest in summer and lowest in winter at the
790 PKU site (Beijing).

791 Diurnal variations of aerosol hygroscopic growth were also investigated. Most of these studies
792 suggested that aerosol hygroscopicity was generally higher in the daytime, compared to the
793 nighttime. For example, hygroscopicity was higher in the daytime than the nighttime for the Aitken
794 mode at the Yufa site (Beijing) in August-September 2006 (Achtert et al., 2009), while no
795 significant difference was found between daytime and nighttime for the accumulation mode. In
796 addition, aerosol hygroscopicity was found to be higher at the daytime than the nighttime at the
797 IAP site (Beijing) in August-October 2015 (Wang et al., 2017c), at the Wuqing site (Tianjin) in
798 July-August 2009 for the more hygroscopic mode (Liu et al., 2011), at the Xingtai site (Hebei) in
799 May-June 2016 (Wang et al., 2018b), at the Panyu site (Guangzhou) in November-December 2011
800 (Tan et al., 2013), December 2012-Janurary 2013 (Jiang et al., 2016) and July-September 2013
801 (Jiang et al., 2016), and at Mt. Huang in September-October 2012 (Wu et al., 2018a) and July 2014
802 (Chen et al., 2016; Wang et al., 2016). The underlying reason might be that photochemical
803 processes led to increased relative contribution of secondary aerosols. However, there are also
804 expectations. For example, κ was larger in the nighttime than the daytime for the accumulation
805 mode at the NKU site (Tianjin) in March 2017 (Ding et al., 2019). In addition, no obvious diurnal
806 variation in average aerosol hygroscopicity was observed at the Panyu site (Guangzhou) in
807 September-October 2016 (Hong et al., 2018), though aerosol hygroscopicity was higher during the
808 daytime for the less hygroscopic mode and slightly lower in the afternoon for the more-
809 hygroscopic mode.



810 While aerosol hygroscopic growth measurements were typically carried out at a single RH at
811 around 90%, several studies also investigated aerosol hygroscopic growth as different RH (Chen
812 et al., 2003; Eichler et al., 2008; Achtert et al., 2009; Meier et al., 2009; Liu et al., 2011; Ye et al.,
813 2011; Zhang et al., 2011; Cheung et al., 2015; Wang et al., 2018a; Wu et al., 2018a). As shown in
814 Figure 5, for the measurement carried out at ZJU site (Hangzhou) in December 2009-January 2010
815 (Zhang et al., 2011), κ derived from measured GF at different RH (73-88%) varied from 0.186 to
816 0.244 for 150 nm particles and from 0.184 to 0.252 for 200 nm particles. For the measurement
817 carried out at the Yufa site (Beijing) in August-September 2006 (Achtert et al., 2009), κ were found
818 to decrease with increasing RH (56-91%) for 250 nm particles, varying from ~ 0.3 to ~ 0.8 .
819 Considerable variations of κ with RH were also reported in other studies (Chen et al., 2003; Meier
820 et al., 2009; Ye et al., 2011; Cheung et al., 2015). Therefore, it can be concluded that using a
821 constant κ to describe aerosol hygroscopic growth at different RH may not always be proper. In
822 addition, during most H-TDMA measurements aerosols were first dried at low RH (typically $< 15\%$)
823 and then humidified to a given RH, and as a result these measurements could not simulate the
824 formation of supersaturated droplets which may exist even when RH was below the corresponding
825 deliquescence RH but above the efflorescence RH.



826

827 **Figure 5.** Single hygroscopicity parameters (κ) derived from GF measured as different RH. (a)
828 150 and 200 nm particles at the ZJU site (Hangzhou) in December 2009-January 2010 (Zhang et
829 al., 2011); (b) 250 nm particles at the Yufa site (Beijing) in August-September 2006 for four typical
830 air masses (Achtert et al., 2009).

831

832 4 CCN activities

833 As stated in Section 2.2, we only discuss CCN activity measurements which reported κ values
834 herein. Sections 4.1-4.3 review measurements conducted in NCP, YRD and PRD, and
835 measurements carried out in other regions in China are discussed in Section 4.4.

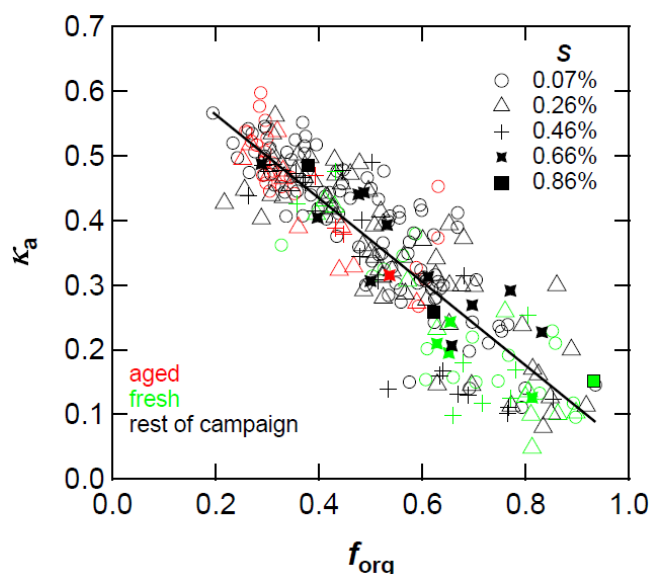
836 4.1 North China plain (NCP)

837 4.1.1 Beijing

838 In August-September 2006, size-resolved CCN activities were measured at the Yufa site
839 (Gunthe et al., 2011). Maximum activation fractions were around 1 for supersaturation in the range
840 of 0.26-0.86%; however, they only reached ~ 0.8 on average at 0.07% supersaturation, and these



841 inactive particles were mainly soot. For the entire measurement period, the average κ_a and κ_t were
842 both determined to be 0.3 ± 0.1 . CCN activities were found to increase with particle size due to
843 increased mass fractions of soluble inorganics (Gunthe et al., 2011), and κ_a was measured to be
844 ~ 0.2 at ~ 40 nm and ~ 0.5 at 200 nm. During periods affected by aged regional pollution, mass
845 fractions of soluble inorganics were enhanced, leading to increase in κ_a (0.35 ± 0.05) (Gunthe et al.,
846 2011); in contrast, mass fractions of organics increased during periods influenced by fresh city
847 pollution, resulting in decrease in κ_a (0.22 ± 0.07).



848
849 **Figure 6.** Dependence of κ_a on mass fractions of organics for three periods over the campaign (red:
850 the aged regional pollution period; green: the fresh city pollution period; black: the rest of the
851 campaign). Reprint with permission by Gunthe et al. (Gunthe et al., 2011). Copyright 2011
852 Copernicus Publications.

853

854 As shown in Figure 6, the measured CCN activities decreased as mass fractions of organics
855 increased (Gunthe et al., 2011); furthermore, the measured κ_a could be quantitatively described by



856 mass fractions of soluble inorganics and organics, and their κ were determined to be 0.7 and 0.1.
857 Aerosol CCN activities during a rapid particle growth event on 23 August were further examined
858 (Wiedensohler et al., 2009), during which CCN size distribution was dominated by the growing
859 nucleation mode instead of the accumulation mode in usual.

860 Measurements were carried out at the PKU site to investigate size-resolved CCN activities in
861 May-June 2014 (Wu et al., 2017). Similar to the concurrent H-TDMA measurements, average κ_a
862 was determined to be ~ 0.10 during biomass burning events, displaying no dependence on particles
863 size (Wu et al., 2017). CCN activities of submicrometer particles were significantly reduced during
864 biomass burning periods, due to increased mass fractions of organics and black carbon.
865 Furthermore, average κ calculated from aerosol compositions measured using AMS were
866 consistent with those derived from hygroscopic growth and CCN activity measurements (Wu et
867 al., 2017), if κ were assumed to be 0.53 and 0 for inorganics and organics.

868 Zhang et al. (Zhang et al., 2017) investigated size-resolved CCN activities at the IAP site in
869 November-December 2014 and August-September 2015, and maximum activation fractions were
870 found to be much smaller than one, indicating large fractions of CCN-inactive particles from local
871 primary emissions. The average κ_a , which ranged from 0.22 to 0.31 for 60-150 nm particles and
872 increased with particle size (Zhang et al., 2017), agreed well with those derived from the
873 concurrent H-TDMA measurements (Wang et al., 2017c). In addition, κ (0.32 ± 0.11) calculated
874 using ACSM-measured aerosol composition were significantly larger than those derived from
875 hygroscopic growth (0.25 ± 0.08) and CCN activities (0.26 ± 0.04) (Zhang et al., 2017). This was
876 because hygroscopicity estimated using ACSM-measured composition did not consider the
877 contribution of smaller and less-hygroscopic particles (aerosol hygroscopicity became lower for
878 smaller particles, but ACSM only detected >60 nm particles).



879 In November-December 2016, Zhang et al. (Zhang et al., 2019a) further investigated size-
880 resolved CCN activities at the IAP site and found that [CCN] was significantly increased during
881 nucleation-initiated haze episodes. It was suggested that increase in particle size contributed >80%
882 to the observed increase in [CCN] (Ren et al., 2018; Zhang et al., 2019a), while the effect of aerosol
883 hygroscopicity enhancement, due to change in aerosol composition, was much smaller.

884 **4.1.2 Other locations in NCP**

885 Zhang et al. (Zhang et al., 2014; Zhang et al., 2017) measured size-resolved CCN activities at
886 the Xianghe site (39.75°N, 116.96°E) in June-July 2013. Average κ_a were determined to be 0.24-
887 0.32 during polluted periods, showing no dependence on particle size; in contrast, κ_a increased
888 from ~0.22 at ~50 nm to ~0.38 at ~180 nm for background days (Zhang et al., 2014). Compared
889 to polluted periods, κ_a were ~20% larger under background conditions for the accumulation mode
890 (100-200 nm), as the contribution of aerosol organics from fresh biomass burning was significantly
891 increased during pollution events (Zhang et al., 2014); however, κ_a were very similar for the
892 nucleation/Aitken modes (40-100 nm) under background and polluted conditions.

893 Size-resolved CCN activities were further investigated at Xianghe site in July-August 2013
894 (Ma et al., 2016; Tao et al., 2020), and it was found that κ_a increased with particle size, from
895 0.22 ± 0.02 at 46 nm to 0.38 ± 0.02 at 179 nm. Compared to that derived from concurrent H-TDMA
896 measurements, aerosol hygroscopicity derived from CCN activities were slightly lower for <50
897 nm particles but higher for >100 nm particles (Ma et al., 2016; Zhang et al., 2016b).

898 Zhang and co-workers (Zhang et al., 2016a; Li et al., 2017b; Zhang et al., 2017) also
899 investigated size-resolved CCN activities at the Xinzhou site in July-August 2014. The average κ_a
900 were determined to be 0.42-0.51 for 37-150 nm particles, exhibiting no dependence on particle
901 size (Zhang et al., 2017); compared to other sites in the NCP, aerosols at the Xinzhou site displayed



902 significantly higher CCN activities, as aerosols observed at this site were highly aged after
903 undergoing regional transport for a long time. The average κ_a (0.48 ± 0.07) (Zhang et al., 2017)
904 agreed well with that (0.47 ± 0.03) determined from concurrent H-TDMA measurements (Zhang et
905 al., 2017), both much significantly larger than that (0.41 ± 0.06) calculated from ACSM-measured
906 aerosol composition, probably because such calculation may underestimate the hygroscopicity of
907 aerosol organics.

908 **4.2 Yangtze River Delta (YRD)**

909 Size-resolved CCN activity measurements were conducted in August 2013 at the NBM site
910 (32.04°N , 118.70°E) on the Jiangxi Island in the Yangtze River (Ma et al., 2017). This site, located
911 in a suburban area of Nanjing, did not have significant local emission at that time. The κ_a values
912 were found to range from ~ 0.1 to ~ 0.8 during the campaign, being 0.35 ± 0.13 on average (Ma et
913 al., 2017), and no significant variation in average κ_a was found for biomass burning, urban, marine
914 and industrial air masses. In addition, κ_a increased from 0.30 ± 0.08 at ~ 55 nm to 0.34 ± 0.08 at 67
915 nm, due to larger contribution of low-hygroscopic organics at 50 nm; however, further increase in
916 particle size up to ~ 149 nm did not lead to obvious increase in κ_a (Ma et al., 2017), likely because
917 aerosols arriving at this site were heavily aged and well internally mixed.

918 Long-term size-resolved CCN activities were studied in January-December 2013 at the Lin'an
919 site (Hangzhou, Zhejiang Province) (Che et al., 2016; Che et al., 2017), which is a WMO Global
920 Atmospheric Watch regional station (30.3°N , 119.73°E , 138 m above the sea level) located in the
921 center of YRD. Maximum activation fractions were close to one at high supersaturation but only
922 reached ~ 0.89 at 0.1% supersaturation. Values of κ_a and κ_t were almost identical (~ 0.25) at 40-50
923 nm and increased to ~ 0.42 (κ_a) and ~ 0.40 (κ_t) at 100-150 nm (Che et al., 2017), suggesting that
924 larger particles contained larger fractions of hygroscopic species (e.g., soluble inorganics).



925 Furthermore, CCN activities were also compared under nine different weather-pollution conditions
926 (Che et al., 2016), and κ were determined to be ~ 0.7 and ~ 0.1 for inorganics and organics during
927 haze episodes and ~ 0.6 and ~ 0.2 for other episodes.

928 **4.3 Pearl River Delta (PRD)**

929 Rose et al. (Rose et al., 2010; Rose et al., 2011) explored size-resolved CCN activities in July
930 2006 at the Backgarden site, which is a suburban site (23.55°N, 113.07°E) located ~ 60 km
931 northwest of Guangzhou. Maximum activation fractions were close to 1 at medium and high
932 supersaturation (0.47-1.27%) and well below 1 at low supersaturation (0.068-0.27%) (Rose et al.,
933 2010), and particles not activated were mainly externally mixed soot with an estimated median κ
934 of ~ 0.01 (Rose et al., 2011). The average κ_a and κ_t were determined to be 0.34 and 0.30 over the
935 entire campaign; to be more specific, κ_a and κ_t were almost identical (~ 0.3) for small particles and
936 increased to 0.4-0.5 and ~ 0.33 for large particles (Rose et al., 2010). Increase in average κ_a with
937 diameter was mainly due to enhanced mass fractions of inorganics for larger particles (Rose et al.,
938 2011). Compared to the rest of the campaign, κ_a and κ_t were reduced by $\sim 30\%$ on average during
939 biomass burning events (0.34 versus 0.24), when mass fractions of organics were substantially
940 increased; moreover, the decrease in κ_t during biomass burning events was very substantial for
941 < 100 nm particles but quite small for ~ 200 nm particles (Rose et al., 2010). It was further found
942 that assuming κ to be ~ 0.6 for inorganics and ~ 0.1 for organics could approximate the observed
943 CCN activities over the entire campaign (Rose et al., 2011).

944 Size-resolved CCN activities were investigated at the Panyu site in November-December 2014
945 (Cai et al., 2018), and the average κ_a were found to increase from 0.21 at 58 nm to 0.30 at 156 nm,
946 because mass fractions of organics, measured using AMS, decreased with particle size. The
947 average κ derived from H-TDMA measurements agreed well with those derived from CCN



948 measurements; however, they were larger than those calculated from size-resolved chemical
949 compositions, and the difference between measured and calculated κ increased with particle size
950 (Cai et al., 2018). This discrepancy was probably because assuming a constant κ (0.1) may
951 underestimate the hygroscopicity of aerosol organics.

952 Aerosol CCN properties were studied at the HKUST site in May 2011 (Meng et al., 2014), and
953 maximum activation fractions were found to exceed 0.9 for the entire campaign, implying that the
954 difference between κ_a and κ_t should be small. CCN activities were found to increase with particle
955 size, with average κ_a being determined to be 0.28 at 46 nm to 0.39 at 116 nm (Meng et al., 2014),
956 due to increase in volume fractions of inorganics as revealed by AMS measurements. It was further
957 found that the measured κ_a could be reasonably well predicted using volume fractions of inorganics
958 and organics (Meng et al., 2014), and their κ were determined to be 0.6 and 0.1.

959 **4.4 Other locations**

960 Hung et al. (Hung et al., 2014; Hung et al., 2016) measured [CCN], [CN] and aerosol number
961 size distribution in August 2011 at a rural site and in June 2012 at an urban site in Taiwan. The
962 rural site (25.89°N, 121.57°E) is ~15 km away from Taipei, while the urban site (25.01°N, 121.54°E)
963 is located on the campus of National Taiwan University in a metropolitan area of Taipei. At the
964 rural site, κ_{cut} increased from ~0.1 at ~50 nm to ~0.35 at ~165 nm during the first period which
965 was significantly affected by anthropogenic emissions, while increased from ~0.04 at ~70 nm to
966 ~0.28 at ~175 nm for the second period not significantly affected by anthropogenic emissions
967 (Hung et al., 2014). Overall, κ_{cut} was larger in the first period than the second period, probably due
968 to the impacts of aged air masses originating from cities nearby during the first period. Compared
969 to the rural site, κ_{cut} were much smaller at the urban site, increasing from ~0.021 at ~90 nm to 0.10



970 at ~250 nm (Hung et al., 2016), indicating that fresh anthropogenic aerosols tended to exhibit lower
971 hygroscopicity.

972 Shipborne size-resolved CCN activity measurements were carried out in September 2012 over
973 remote regions of the South China Sea and East China Sea (Atwood et al., 2017). Under marine
974 background conditions, the average κ_a were determined to be 0.65 ± 0.11 and 0.46 ± 0.17 for the
975 accumulation and Aitken modes (Atwood et al., 2017). Compared to marine background
976 conditions, CCN activities were reduced after extensive precipitation, with average κ_a determined
977 to be 0.54 ± 0.14 and 0.34 ± 0.11 for the accumulation and Aitken modes; whereas during periods
978 impacted by biomass burning, κ_a was reduced to 0.40 ± 0.03 for the accumulation mode but
979 increased instead to 0.56 ± 0.25 for the Aitken mode (Atwood et al., 2017).

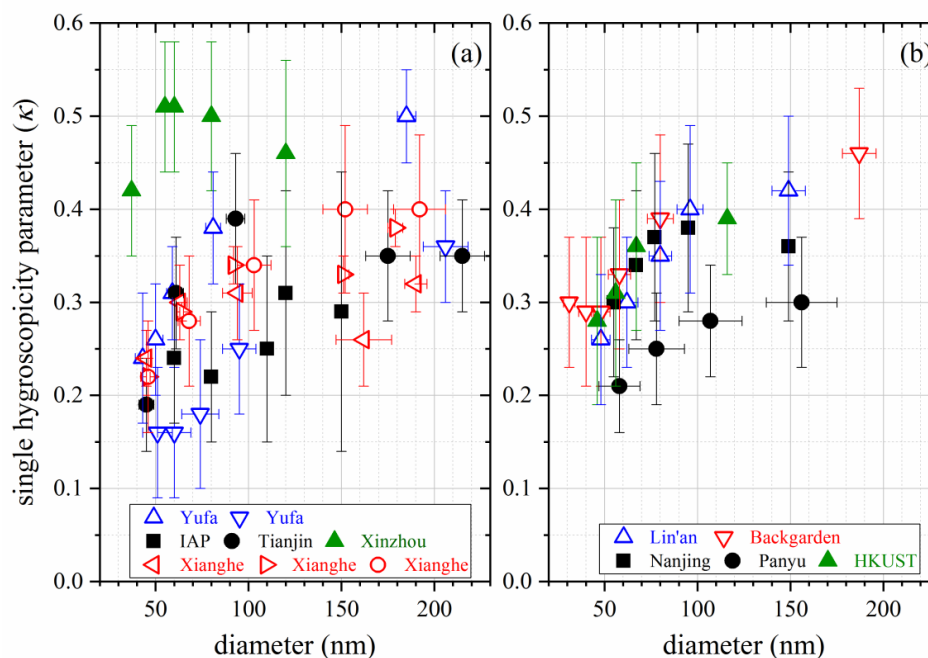
980 Size-resolved CCN activities were explored over north South China Sea ($19^{\circ}39'N$ to $22^{\circ}43'N$,
981 $113^{\circ}44'E$ to $118^{\circ}12'E$) in August 2018 (Cai et al., 2020), and no obvious dependence of κ_a on
982 particle size (50-100 nm) was observed. The campaign-averaged κ was determined to be ~ 0.40
983 (Cai et al., 2020), larger than these measured in the PRD region but smaller than those measured
984 over remote marine regions. This is because the air in north South China Sea was affected by both
985 continental air masses (low hygroscopicity) and marine background (high hygroscopicity).

986 **4.5 Summary**

987 Similar to H-TDMA measurements, CCN activity measurements in China were mainly carried
988 out in NCP, YRD and PRD, and almost all the measurements took place at or close to the ground
989 level. In addition, the number of CCN activity measurements is much smaller than H-TDMA
990 measurements. The limited number of field studies preclude any solid conclusions on diurnal and
991 seasonal variations of aerosol CCN activities being drawn.



992 Maximum activation fractions were typically found to be considerably smaller than 1 (Rose et
993 al., 2010; Gunthe et al., 2011; Che et al., 2017; Zhang et al., 2017), especially at low
994 supersaturation, and CCN-inactive particles were usually attributed to low hygroscopic primary
995 particles (e.g., soot) from local sources. The average κ , reported by previous studies, were
996 generally found to be in the range of 0.30-0.35; however, CCN activities could be significantly
997 reduced if measurement sites were affected by fresh urban pollution or biomass burning (Rose et
998 al., 2010; Gunthe et al., 2011; Zhang et al., 2014; Wu et al., 2017), due to enhanced contribution
999 of soot and organics. The average κ observed at the Xinzhou site appeared to be larger than those
1000 reported at other continental site (Zhang et al., 2017), probably because aerosols arriving at this
1001 site were heavily aged. In addition, two studies which investigated aerosol CCN activities in the
1002 marine boundary layer reported larger κ values (Atwood et al., 2017; Cai et al., 2020), compared
1003 to those at continental sites.



1004



1005 **Figure 7.** Measured κ_a as a function of particle diameter reported by previous studies (Rose et al.,
1006 2010; Deng et al., 2011; Gunthe et al., 2011; Deng et al., 2013; Meng et al., 2014; Zhang et al.,
1007 2014; Che et al., 2016; Ma et al., 2016; Che et al., 2017; Ma et al., 2017; Zhang et al., 2017; Cai
1008 et al., 2018; Tao et al., 2020) in the NCP (a) and other regions in China (b). Solid symbols represent
1009 urban/suburban sites, and open symbols represent rural sites.

1010

1011 Figure 7 summarizes size dependence of κ_a reported by CCN measurements at continental sites
1012 in China, and measurement data related to specific cases (e.g., biomass burning events) are not
1013 included (Rose et al., 2010; Wu et al., 2017). As shown in Figure 7, in general κ_a increased with
1014 particle size, as mass fractions increased with particle size for soluble inorganics and decreased for
1015 organics. Nevertheless, no obvious dependence of κ_a on particle size was also observed in Xinzhou
1016 (Zhang et al., 2017) and Nanjing (Ma et al., 2017), probably because aerosol particles at these two
1017 sites were substantially aged and thus very well internally mixed.

1018 Several studies carried out CCN activity closure analysis. Some studies suggested that the
1019 measured κ could be well quantitatively explained by aerosol composition (Rose et al., 2010;
1020 Gunthe et al., 2011; Wu et al., 2017), while other studies showed that κ estimated using aerosol
1021 composition were either larger (Zhang et al., 2017) or smaller than measured values (Zhang et al.,
1022 2017; Cai et al., 2018). In addition, a few studies investigated aerosol hygroscopic growth and
1023 CCN activities concurrently, and both consistence (Wu et al., 2017; Zhang et al., 2017; Cai et al.,
1024 2018) and discrepancies (Ma et al., 2016; Zhang et al., 2016b) were reported.

1025 **5 Perspectives**

1026 In the last 10-20 years a number of field measurements of hygroscopic properties and CCN
1027 activities of tropospheric aerosols have been carried out in China, and summaries of measured



1028 hygroscopic properties and CCN activities are provided in Sections 3.5 and 4.5. As shown in
1029 Sections 3 and 4, these studies have significantly improved our knowledge of tropospheric aerosol
1030 hygroscopicity in China and provided valuable data to better understand the roles aerosols play in
1031 heterogeneous and multiphase chemistry, as well as direct and indirect radiative forcing. However,
1032 large knowledge gaps still exist for aerosol hygroscopicity in China, as described below, and future
1033 research directions are also discussed.

1034 **Data availability:** In Tables S1-S5 we attempt to compile measurement data reported by
1035 previous studies under a consistent framework in order to enhance their accessibility and usability.
1036 However, important data are not always available from every study published; for example, several
1037 studies presented their main results graphically. It is recommended that in future data in the
1038 numerical form (H-TDMA measurements: including but not limited to diameter, RH, and GF
1039 and/or κ ; CCN activity measurements: including but not limited to supersaturation, activation
1040 diameter and κ) should be provided.

1041 **Geographical coverages:** As shown in Sections 3-4, almost all the measurements of
1042 hygroscopic properties and CCN activities in China were carried out in east regions (e.g., NCP,
1043 YRD and PRD) heavily affected by anthropogenic emissions. Therefore, it will be very desirable
1044 in future to carry out these measurements in other regions; measurements in areas far from by
1045 human activities will be especially important, as they will provide information on aerosol
1046 hygroscopicity in the pristine troposphere.

1047 **Vertical distribution:** Most of previous aerosol hygroscopicity measurements in China were
1048 only carried out at or close to the ground level. However, both aerosol composition and RH, and
1049 as a result aerosol hygroscopic growth and CCN activation, will vary with altitude. For example,
1050 aircraft-based measurements of aerosol size distribution and composition indicated that single



1051 hygroscopicity parameters would increase significantly with altitude (Liu et al., 2020), and it was
1052 revealed from remote sensing that aerosol hygroscopicity at the upper boundary level was different
1053 from that at the ground level (Tan et al., 2020). Therefore, in-situ measurements of vertical profiles
1054 of aerosol composition and hygroscopicity on different platforms (e.g., towers, airships, aircrafts,
1055 and etc.) will be very valuable; in addition, remote sensing may be very useful for retrieving
1056 vertical profiles of aerosol hygroscopicity, as demonstrated by a very recent study (Tan et al., 2020).

1057 **Long-term measurements:** Both aerosol concentration and composition have undergone (and
1058 very likely will undergo) significant changes in China; however, most aerosol hygroscopicity
1059 measurements were carried out for 1-2 months during specific field campaigns. Long-term
1060 measurements of aerosol hygroscopicity will be very important to understand seasonal and annual
1061 variations of aerosol hygroscopicity and the implications for visibility, atmospheric chemistry and
1062 climate change.

1063 **Hygroscopicity of large particles:** Tables S1-S4 reveal that the maximum aerosol diameter
1064 examined in hygroscopic growth studies was 350 nm, which is the upper limit of dry aerosol size
1065 for most of H-TDMA instruments (Tang et al., 2019). As particles larger than 350 nm can
1066 contribute substantially to aerosol surface area and volume (or mass) concentrations,
1067 hygroscopicity of these particles will be very important and should be measured in future, and this
1068 requires technical improvements of H-TDMA. On the other hand, hygroscopicity of >350 nm
1069 particles may not be very important for CCN activation, as these particles can be easily activated
1070 due to their large diameters.

1071 **RH dependence:** Most H-TDMA measurements were carried out at a single RH (usually
1072 ~90%), and a few studies which measured GF as a function of RH suggested that a constant κ
1073 failed to describe hygroscopic growth at different RH. In addition, due to lack of measurement



1074 data at different RH, it is not clear how well widely-used aerosol thermodynamic models can
1075 simulate ALWC at ambient RH. Therefore, measurements of aerosol hygroscopicity at different
1076 RH are certainly warranted.

1077 **The effect of aerosol organics:** As discussed in Section 3, several studies (Liu et al., 2014;
1078 Wu et al., 2016; Cai et al., 2018; Hong et al., 2018; Li et al., 2019b; Jin et al., 2020) suggested that
1079 organics contributed substantially to aerosol water uptake, while some studies also indicated that
1080 the contribution of aerosol organics to ALWC was rather minor. Therefore, aerosol hygroscopicity
1081 closure analysis, with concurrent measurements of aerosol composition and hygroscopicity, is
1082 recommended for future, in order to further understand the effects of aerosol organics on ALWC
1083 and CCN activation; in addition, relevant factors which need consideration include the dependence
1084 of hygroscopicity on composition of aerosol organics (e.g., O/C ratios) and the effects of aerosol
1085 organics on surface tension, phase separation effects, and etc.

1086

1087 **Data Availability.** This is a review paper, and all the data used come from cited literature. In
1088 addition, the data we have compiled can be found in the supplement.

1089 **Author contribution.** Mingjin Tang conceived this work; Chao Peng and Mingjin Tang wrote the
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