

1 **Do alternative inventories converge on the spatiotemporal representation of spring**
2 **ammonia emissions in France?**

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19
20 **Abstract**

21 Agriculture is the main source of ammonia (NH₃) in France, an important gaseous precursor of
22 atmospheric particulate matter (PM). National and even more global emission inventories are
23 known to have difficulty representing the large spatial and temporal variability inherent to
24 atmospheric NH₃. In this study, we compare NH₃ emissions in France during the spring 2011
25 from one reference inventory, the TNO inventory, and two alternative inventories that account
26 in different manners for both the spatial and temporal variabilities of the emissions: (i) the
27 NH₃SAT satellite-derived inventory based on IASI NH₃ columns and (ii) the CADASTRE-CIT
28 inventory that combines NH₃ emissions due to nitrogen fertilization calculated with the
29 mechanistic model VOLT' AIR on the database of the CADASTRE_NH₃ framework and other
30 source emissions from the CITEPA. The total spring budgets, from March to May 2011, at the
31 national level are higher when calculated with both alternative inventories than with the
32 reference, the difference being more marked with CADASTRE-CIT. NH₃SAT and

33 CADASTRE-CIT inventories both yield to large NH₃ spring emissions due to fertilization on
34 soils with high pH in the northeastern part of France (65 ktNH₃ and 135 ktNH₃, respectively, vs
35 48 ktNH₃ for TNO-GEN), while soil properties are not accounted for by the TNO-GEN
36 methodology. For the other parts of France, the differences are smaller. The timing of
37 fertilization and associated ammonia emissions is closely related to the nitrogen requirements
38 and hence the phenological stage of the crops, and therefore to the crop-year's specific weather
39 conditions. Maximum emissions are observed in March for 2011 for some regions for both
40 alternative inventories, while April is the period with maximum emissions for the reference
41 inventory whatever the region or the year. Comparing the inventories at finer temporal
42 resolutions, typically at daily scale, large differences are found. The convergence of alternative,
43 independent and complementary methods on the spatiotemporal representation of the spring
44 NH₃ emissions particularly over areas where the contribution of mineral fertilizer spreading to
45 the spring budget is strong, encourage further developments in both prospected complementary
46 directions, as this will help improving national NH₃ emission inventories.

47

48 **1. Introduction**

49 France is a major crop producer and a major exporter of agricultural and food products. In 2014,
50 it produced 2%, 4%, 5%, 8%, 8% and 14% of the global production of maize, sunflower, wheat,
51 barley, rapeseed and sugar beet, respectively [Food and Agriculture Organization of the United
52 Nations FAO, Schauburger et al., 2018]. Through this food cultivation and also due to animal
53 husbandry, agriculture is the main source of ammonia (NH₃) in the country. As an important
54 gaseous precursor of particulate pollution, harmful to human life [Lelieveld et al., 2015; WHO,
55 2016], ammonia plays an important role in the regulation of inorganic aerosol concentrations
56 [Erisman and Schaap, 2004, Bauer et al., 2016], and contributes to N deposition and potential
57 exceedance of critical loads of ecosystems [Erisman et al., 2007; EEA European Environment
58 Agency, 2014]. In order to limit air pollution, also responsible for acidification and
59 eutrophication, the new European National Emission Ceilings Directive 2016/2284, replacing
60 the Directive 2001/81/EC, now sets ambitious national reduction commitments for ammonia.
61 Ammonia emissions indeed have to be reduced by 19% in 2030, compared with the 2005 levels
62 [OJEU, 2016].

63

64 At the European scale, total NH₃ emissions are provided by the European Monitoring and
65 Evaluation Program (EMEP) [Vestreng, 2005] or by the TNO-MACCI3 [Kuenen et al., 2014]

66 inventories that rely on national annual declarations and estimates of emission factors.
67 Emissions are accounted for without separating fertilization and livestock. These reference
68 inventories are widely used by the scientific community to study the impact of pollutant
69 emissions on the chemical composition of the troposphere and on air quality. Nevertheless,
70 uncertainties on the quantification of the NH_3 emissions are usually estimated to be between
71 100 and 300% of the annual budgets in the reference inventories [EMEP/EEA, 2016; Kuenen
72 et al., 2014]. In addition, the temporal and spatial variability may not be well represented in the
73 reference inventories, as the temporal profiles used do not account for meteorology, soil
74 properties and other local conditions. Moreover, fertilizer spreading is of particular interest, as
75 these are applied during small periods, especially during a few weeks at the end of winter and
76 early spring. However, the exact timing of fertilizer spreading is difficult to predict, as it
77 depends on agricultural practices and meteorological conditions, which is not taken into account
78 in the temporal disaggregation of the reference emission inventories. Both the inaccurate
79 temporal resolutions and the mis-representation of the spreading emissions largely explain the
80 difficulty encountered by models to represent seasonal or daily pattern of NH_3 concentrations
81 [Menut et al., 2012], and consequently particulate matter levels [Fortems-Cheiney et al, 2016].

82
83 To reduce these uncertainties, a better quantification of agricultural ammonia emissions and its
84 temporal and spatial evolution is necessary. In particular, one of the challenges is to capture the
85 right timing of fertilizer spreading at the weekly or even at the daily scale in order to reflect the
86 effect of environmental and agronomical conditions on ammonia emissions. To this end,
87 mechanistic models taking into account meteorological conditions, soil properties and
88 agricultural practices, have been developed (e.g., for Denmark [Skjøth et al., 2004], for the UK
89 [Hellsten et al., 2008], and for mineral fertilization in springtime in France [Hamaoui-Laguel
90 et al., 2014]). Limitations for such approaches come from the fact that detailed agricultural data
91 needed as input to such models are not available for most of the European countries. Moreover,
92 agricultural practices of a specific country cannot be extrapolated to another country [Skjøth et
93 al., 2011].

94
95 As an alternative to direct emission modeling, attempts have been made to constrain ammonia
96 emissions through inverse approaches, based on ammonium wet deposition data [Paulot et al.,
97 2014] or on satellite observed atmospheric ammonia distributions (e.g., from the Tropospheric
98 Emission Spectrometer TES [Zhu et al., 2013; Zhang et al., 2018], from the Infrared
99 Atmospheric Sounding Interferometer (IASI) [Fortems-Cheiney et al., 2016; Van Damme et

100 al., 2018; Adams et al., 2019] or from the Cross-track Infrared Sounder CrIS [Adams et al.,
101 2019; Dammers et al., 2019]). In principle, such emission estimates can be available shortly
102 after observation. The advantage of satellite-derived estimates is also that these can be derived
103 globally, at a high temporal scale (e.g., daily scale under clear sky). The downside of these
104 however, is that they do not provide information on the underlying sources of the emissions
105 (fertilizers vs husbandry), or e.g. the date of fertilization, the type of fertilizers used, the
106 fertilization rates, etc., that could be important for the regulation of NH₃ emissions.

107

108 In this context, we compare ammonia emissions in France from inventories using the different
109 approaches mentioned above: (i) the reference, in the sense that this is the method currently
110 used, hereafter called TNO-GEN, is the European inventory based on the annual budgets
111 provided by the TNO-MACCI3 inventory [Kuenen et al., 2014] and on seasonal profiles from
112 GENEMIS [Ebel et al., 1997]; (ii) a first alternative inventory, hereafter called NH₃SAT, is
113 based on a top-down approach starting from the IASI derived NH₃ columns; (iii) the other
114 alternative inventory, hereafter called CADASTRE-CIT, is based on a bottom-up approach
115 quantifying NH₃ emissions due to nitrogen fertilization combining spatiotemporal data and
116 calculations performed within the CADASTRE_NH₃ framework with the mechanistic model
117 VOLT'AIR (Ramanantenasoa et al., [2018]; Générmont et al., [2018]) completed with livestock
118 and other source emissions from the French Interprofessional Technical Centre for Studies on
119 Air Pollution (CITEPA). This study aims at assessing the potential contribution of better spatial
120 and temporal representation of fertilization-related ammonia emissions to the quality of
121 ammonia emission inventories. The improvement is assessed in terms of total budget, spatial
122 distribution and timing of the emissions. The study period, spring 2011 (from March, 1st to May
123 2011, 31st), was chosen following three criteria. Firstly, because at the time of the study, the
124 last French agricultural data were available from AGRESTE [AGRESTE, 2014] for the
125 agricultural year 2010-2011, allowing the application of the CADASTRE_NH₃ framework for
126 the quantification of the spatio-temporal distribution of NH₃ emissions due to nitrogen
127 fertilization for this crop year [Ramanantenasoa et al., 2018; Générmont et al., 2018]. Secondly,
128 ammonia emissions are enhanced during spring in accordance with crops requirements [Skjøth
129 et al., 2004; Ramanantenasoa et al., 2018; Générmont et al., 2018]. Finally, unlike autumn and
130 winter months, the NH₃ spring levels are detectable with a better confidence in the IASI satellite
131 observations [Viatte et al., 2020], allowing the extension of the preliminary work of Fortems-
132 Cheiney et al. [2016] to deduce NH₃ emissions from the IASI satellite instrument.

133 The three inventories and methods to build them used for this study are presented in Section 2
 134 and the results of the comparison are given and discussed in Section 3.

135 2. Inventories

136 The three inventories TNO-GEN, NH₃SAT and CADASTRE-CIT compared in this study are
 137 described in Table 1 and in the following sections. It is worth noting that only the CADASTRE-
 138 CIT inventory provides information on the respective contribution of fertilization and livestock
 139 emissions. The spatial and temporal resolutions of the inventories are also shown in Table 1.

Name	Spatial Resolution (latitude x longitude)	Temporal Resolution	Fertilization emissions	Livestock emissions
TNO-GEN	0.125°x0.0625°	Monthly	-	
NH ₃ SAT	0.5°x0.25°	Daily	-	
CADASTRE- CIT	0.015625° x0.03125°	Hourly	CADASTRE_NH ₃ Ramanantenasoa et al., [2018] and Génermont et al., [2018]	
	0.007825°x0.007825°	Daily		CITEPA national emissions, temporalized according to Skøjth et al., [2011]

140 **Table 1.** Main characteristics of the different compared inventories before their
 141 aggregation/disaggregation for the inter-comparison.

142 The inter-comparison is made at the 0.5° (longitude) x 0.25° (latitude) resolution. The outputs
 143 of the TNO-GEN and the CADASTRE-CIT inventories could have consequently been
 144 aggregated or disaggregated.

145 2.1. TNO-GEN

146 In this study, the TNO-GEN combines the annual budgets provided by the TNO-MACCCIII
 147 inventory and the seasonal profiles to deduce the monthly variability of NH₃ emissions. This
 148 inventory is based on official annual emission data submitted by countries to EMEP/CEIP
 149 (European Monitoring and Evaluation Programme/Centre on Emission Inventories and
 150 Projections) for air pollutants. It is the update of the TNO-MACCII inventory [Kuenen et al.,

151 2014]. It is an inventory at $0.125^\circ \times 0.0625^\circ$ resolution providing annual emissions of NH_3 from
152 the agricultural sector, without separating the contributions from fertilization and livestock.
153 Hereafter, we use the TNO-MACIII emissions of year 2011. The seasonal profile of these
154 emissions is prescribed according to the typical national factors provided by GENEMIS. This
155 seasonal temporal profile used for the temporalization of emissions -the same one applied to
156 the entire country- leads to a maximum in NH_3 emissions systematically in April over France
157 [Ebel et al., 1997]. The emissions remain constant between days in each month and between
158 hours in each day.

159

160 **2.2. NH_3 SAT**

161 As a first alternative, a mass-balance approach, which is a common method for the
162 quantification of short-lived species surface fluxes [Palmer et al., 2003; Jaeglé et al., 2004;
163 Boersma et al., 2008; Lin et al., 2010] was set-up. We used it to deduce NH_3 emissions from
164 differences between NH_3 total columns observed by the IASI instrument and simulated by the
165 CHIMERE regional chemical transport model (CTM) using the TNO-GEN inventory as inputs
166 data.

167 2.2.1 The regional CTM CHIMERE

168 CHIMERE simulates concentrations of gaseous and particulate chemical species [Menut et al.,
169 2013; Mailler et al., 2017]. For this study, we used the CHIMERE version 2013a. The horizontal
170 resolution is given as follows: $0.5^\circ \times 0.25^\circ$ over $17^\circ\text{W}/40^\circ\text{E}$ – $32^\circ\text{N}/70^\circ\text{N}$, including 115
171 (longitude) \times 153 (latitude) grid-cells. The vertical grid contains 17 layers from the surface to
172 200 hPa. This model is driven by the European Centre for Medium-Range Weather Forecasts
173 global meteorological fields [Owens and Hewson, 2018]. Climatological values from the
174 LMDz-INCA global model [Szopa et al., 2008] are used to prescribe concentrations at the
175 lateral and top boundaries and the initial atmospheric composition in the domain. For inorganic
176 species, aerosol thermodynamic equilibrium is achieved using the ISORROPIA model [Nenes
177 et al., 1998].

178 As most of the models in the world, the parameterization of NH_3 dry deposition is unidirectional
179 in CHIMERE. The parameterization of a bidirectional exchange with surfaces in Wichink Kruit
180 et al. [2012] increased their yearly mean modeled LOTOS-EUROS European ammonia
181 concentrations almost everywhere, and particularly over agricultural source areas. However,
182 Zhu et al. [2015], with the Goddard Earth Observing System-Chemistry (GEOS-Chem) global

183 CTM, estimated decrease of NH₃ European concentrations in April, when the inclusion of a
184 compensation point for vegetation is included. Further work needs to be done to better
185 investigate the sensitivity of NH₃ concentrations to the deposition. Nevertheless, without such
186 parameterization for bi-directional exchange, Azouz et al. [2019] assessed that regional models
187 such as CHIMERE usually operating with large grid cell sizes simulate quite well the average
188 NH₃ dry deposition flux over a large domain of simulation.

189 The evaluation of CHIMERE NH₃ and NH₄⁺ concentrations should be done against NH₃ (as
190 done in Fortems-Cheiney et al., [2016]) and NH₄⁺ measurements. Nevertheless, to our
191 knowledge, there is no available NH₃ measurement over France for the focused period here.
192 There is interpretable NH₄⁺ surface measurements at only one site, making the interpretation of
193 the results difficult. NH₃ and NH₄⁺ comparisons during other periods are scarce also. For
194 instance, Tuccella et al., [2019] compared CHIMERE simulated and observed NH₄⁺ at the
195 Cabaux supersite and found average concentrations for May 2008 of 1.3 µg/m³ for both, with
196 a correlation coefficient of 0.52. For the Paris agglomeration between September 2009 and
197 2010, the modelled regional NH₄⁺ burden was 1.8 µg/m³ while the modelled one was 1.6 µg/m³
198 [Petetin et al., 2016]. From June to September 2010, 83% of modelled total NH_x was gaseous,
199 while in the model, it was only 50%, coherent with this NH₃ was underestimated especially
200 during warmer days. Thus, it is concluded for one site and season, that particulate NH₄⁺ has a
201 low to medium impact on NH₃.

202

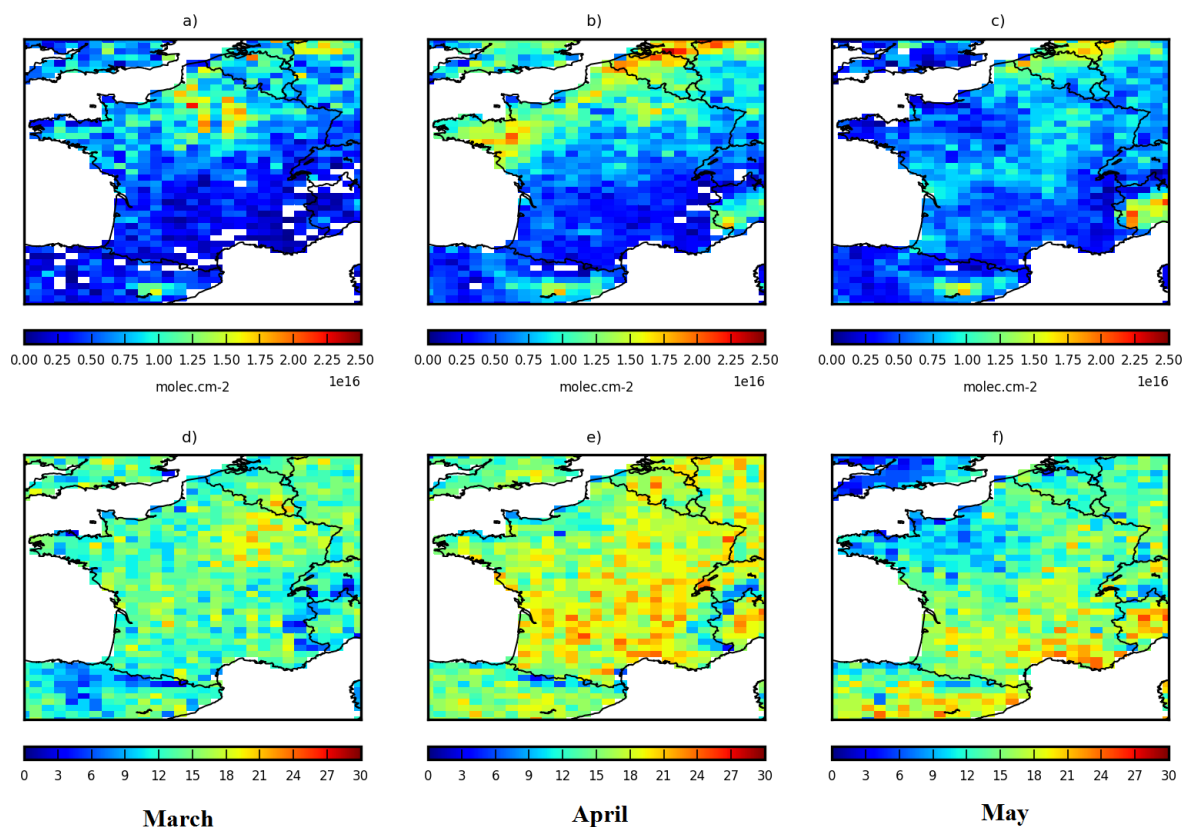
203 2.2.2. The IASI observations

204 We use data from the IASI-A instrument, flying on a low Sun-synchronous polar orbit aboard
205 Metop satellite since October 2006, with equator crossing times of 09:30 (descending mode)
206 and 21:30 (ascending mode) local sidereal time (LST) [Clerbaux et al., 2009, Clarisse et al.,
207 2009]. The spatial resolution of its observations is about 12x12 km² at nadir. The algorithm
208 used to retrieve NH₃ columns from the radiance spectra is described in Van Damme et al.,
209 [2017]. Several improvements have been introduced since the description of Van Damme et al.,
210 [2014] and the version v1 used in our previous study [Fortems-Cheiney et al., 2016]. In this
211 study, we use the reanalyzed dataset ANNI-NH3-v2.2R, relying on ERA-Interim
212 meteorological input data from the European Centre for Medium-Range Weather Forecasts
213 (ECMWF) rather than the operationally provided Eumetsat IASI Level 2 (L2) data used for the
214 standard near-real-time version [Van Damme et al., 2017]. We only consider land
215 measurements from the morning overpass, as IASI is more sensitive at this time to the boundary

216 layer, owing to more favorable thermal conditions [Clarisse et al., 2010, Van Damme et al.,
217 2014].

218 The IASI total columns are averaged into “super-observations” (average of all IASI data within
219 the $0.5^\circ \times 0.25^\circ$ resolution and for the given CHIMERE physical time-step of about 5/10
220 minutes). As suggested by Van Damme et al. [2017], we no longer use weighted averages for
221 this purpose. We performed a sensitivity test by selecting the IASI pixels for which the retrieval
222 error does not exceed 100%: the results about the temporal and spatial variability of the NH_3
223 French emissions presented in Section 3 did not significantly change, showing the robustness
224 of the IASI NH_3 product (not shown).

225 The resulting monthly means of IASI NH_3 columns from March to May 2011 are shown in
226 Figure 1 (a to c). The spatio-temporal variability -with the highest values over northeastern
227 France in March, and over northwestern France in April- is confirmed by the IASI 10-year and
228 by the CrIS 5-year monthly means shown in Viatte et al. [2020]. Note that the potential of IASI
229 to provide information at high temporal resolution, up to daily scale, can be hampered by the
230 cloud coverage as only observations with a cloud coverage lower than 10% are delivered [Van
231 Damme et al., 2017]. To evaluate the impact of this limitation, the number of IASI super-
232 observations used to calculate these monthly means, which represents the number of days over
233 a month covered by IASI, is shown in Figure 1 (d to f). On average, more than half of the month
234 is sampled by IASI during spring, except in May in the northwestern part of France. The regions
235 showing large IASI NH_3 values are consequently well sampled.



236

237 **Figure 1.** (top) Monthly means of IASI "super-observations" for (a) March 2011, (b) April
 238 2011 and (c) May 2011. Units are molec.cm^{-2} . (bottom) Total number of IASI super-
 239 observations per month in (d) March 2011, (e) April 2011 and (f) May 2011.

240

2.2.3. Deducing NH_3 SAT emissions

241 Relative differences between simulated columns by the CHIMERE regional CTM (described
 242 in Section 2.2.1, using the TNO-GEN emissions for the year 2011, described in Section 2.1)
 243 and observed IASI total columns (described in Section 2.2.2) are applied as a corrective factor
 244 to the reference emissions at daily and at grid-cell resolutions over France, from February to
 245 May 2011. As IASI "super-observations" provide one piece of information per day, the diurnal
 246 time profile of reference emissions cannot be improved: we apply the same daily correction
 247 factor to all hourly NH_3 emissions. When IASI is not available (i.e., observations with a cloud
 248 coverage higher than 10%), the correction is not applied and the emissions remain equal to the
 249 TNO-GEN ones. To compare the emissions with the CADASTRE-CIT inventory, and their
 250 respective simulations with CHIMERE, the correction is only applied over France here.

251 With the mass-balance approach, the transport to neighboring cells is assumed negligible
 252 following Palmer et al., [2003]. This approach has been debated by Turner et al. [2012], who

253 found that non-local sources contribute substantially to columns of short-lived species. Li et al.
254 [2019] evaluated the ability of both a mass-balance approach and a variational assimilation to
255 recover known NH₃ emissions at different spatial resolutions. At a 2°x2.5° resolution, they
256 found that both methods yielded similar values. At a 0.25°x0.3125°, the mass-balance approach
257 led to values about 20% higher compared to the variational ones. With our 0.5°x0.25°
258 resolution, the use of a mass-balance approach would lead to additional errors of less than about
259 20% for the quantification of NH₃ emissions. This uncertainty is acceptable and much lower
260 than the uncertainty existing in the annual and national budgets provided by emission
261 inventories [EMEP/EEA, 2016]. In this context, we choose to perform such mass-balance
262 approach to deduce NH₃ emissions from IASI ANNI-NH₃-v2.2R super-observations.
263 Additional uncertainty comes from the IASI observations. The IASI minimum detection limit
264 is of about 2-3 ppbv ($\sim 4\text{-}6 \cdot 10^{15}$ molecules.cm⁻²) [Clarisse et al., 2010]. The signal-to-noise ratio
265 therefore presents better performance for regions with high local concentrations (e.g., northern
266 part of France, Figure 1) than over low local concentration areas (e.g southern parts of France,
267 Figure 1). There is no evaluation available for the IASI ANNI-NH₃-v2.2R product used here
268 yet.

269 **2.3. CADASTRE-CIT**

270 As a second alternative, a bottom-up approach was set-up based on the finest national
271 inventories available for anthropogenic sources of ammonia. The CADASTRE_NH₃
272 framework provides such an inventory for organic and mineral fertilization practices. This is
273 however not the case for the other anthropogenic sources. For livestock emissions, with the
274 exception of the stage of effluent spreading in the field, the less detailed inventory of the French
275 Interprofessional Technical Centre for Studies on Air Pollution CITEPA is used. To meet the
276 objectives of better specialization and temporalization, specific procedures are applied. These
277 inventories are completed by the TNO-GEN inventory for the emissions of the other sectors.

278 2.3.1. Fertilization emissions from CADASTRE NH₃

279 The CADASTRE_NH₃ was implemented in order to represent in a realistic way spatio-
280 temporal variability of French NH₃ emissions due to mineral and organic N fertilization, and is
281 fully described in Ramanantenasoa et al. [2018] and in Générumont et al. [2018]. It has been
282 constructed through the combined use of two types of resources: the process-based VOLT' AIR
283 model and geo-referenced and temporally explicit databases for soil properties, meteorological
284 conditions and N fertilization.

285 VOLT' AIR is a 1D process-based model predicting NH₃ emissions from N fertilizers on bare
286 soils, from physical, chemical and biological processes [Le Cadre, 2004; Garcia et al., 2012].
287 It incorporates current knowledge on NH₃ volatilization after application of the main types of
288 organic manure and mineral N fertilizers in the field. It takes into account the major factors
289 known to influence NH₃ volatilization in the field, i.e., soil properties, weather conditions,
290 cultural practices and properties of mineral fertilizers and organic products. It runs at an hourly
291 time step at the field scale for a period of several weeks covering thus the entire volatilization
292 duration of fertilization events.

293 Local features are attributed to each simulation units, the Small Agricultural Regions (SAR):
294 local weather conditions (SAFRAN, Météo-France), the dominant soil type of the SAR from
295 the European Soil Data Center (ESDC), with soil properties provided by the Harmonized World
296 Soil Database (HWSD) of the Food and Agriculture Organization (FAO); areas cultivated in
297 crop year 2010-2011 per crop per region derived from the European Land Parcel Identification
298 System (LPIS, Common Agricultural Policy (CAP) regulations); Nitrogen fertilization
299 management practices are derived from data of the national AGRESTE survey of cultural
300 practices for arable crops and grassland (Department of Statistics and Forecasting of the French
301 Ministry of Agriculture) [AGRESTE, 2014]. All input data required by VOLT' AIR are
302 geographically overlaid and intersected with a Geographical Information System to generate
303 input combinations in each SAR. Each input combination is used as the input data for a virtual
304 300m x 300m field for a simulation using VOLT' AIR. Exact times and dates of fertilizations
305 are required to run VOLT' AIR, but for the sake of robustness, the statistical analysis of the
306 survey data has been performed on the basis of two-week intervals for the date of fertilization.
307 Fertilizations are thus randomly distributed within these two-week intervals in proportion to
308 their respective representation following Ramanantenasoa et al. [2018]. Each simulation of NH₃
309 emissions is run at an hourly time step for a period of two months, starting one month before
310 the fertilization in order to calculate soil water content at the time of application, and ending
311 one month after fertilization, in order to cover the whole volatilization event.

312 About 160 000 runs with the VOLT' AIR model have been performed over the crop-year 2010-
313 11 to produce ammonia emissions per hour, per ha, per crop type, per SAR. Emissions can be
314 aggregated at different spatial and temporal scales. At the spatial scale, they are weighted with
315 the contribution of (i) each N fertilization management applied to each crop in each SAR and
316 (ii) the area of the crop cultivated in the SAR. A procedure allows producing ammonia
317 emissions at the required grid scale for the inventory comparison: it is based on cultivated areas

318 for each crop as the key of desegregation-reaggregation from the SAR to the
319 $0.015625^\circ \times 0.03125^\circ$ grid. At the temporal scale, emissions are aggregated over daily, weekly
320 or monthly bases for the sake of comparison with TNO-GEN and NH₃SAT inventories.
321 Volatilization taking place over several days, from few days to several weeks, one fertilization
322 in one field contributes to ammonia emissions over several days or weeks. Weather condition
323 effects on overall ammonia emissions is thus the result of both their effects on fertilization
324 timing and their effects on volatilization intensity and dynamics over 30 days from fertilizer
325 application.

326 As they are not available in the agricultural practice survey, N fertilizations of vegetables, fruits,
327 and vines are not accounted for in CADASTRE_NH₃: their contribution is minor for France
328 overall, only accounting for 5% of the total agricultural area [AGRESTE, 2010] but are
329 important in particular regions. As the agricultural practice survey does not provide information
330 over Corsica, this inventory is completed by the TNO-GEN inventory over this region (Figure
331 2).

332 2.3.2 Livestock emissions

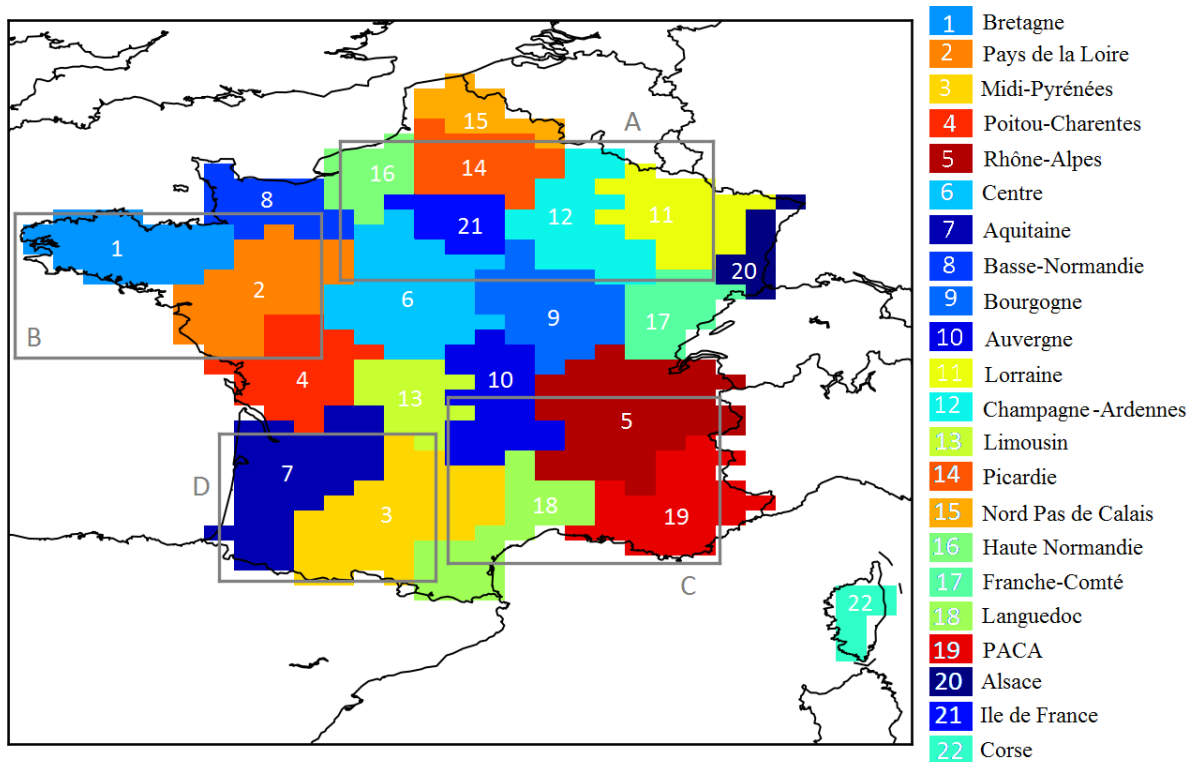
333 As for the TNO-GEN inventory, French NH₃ emissions from livestock for the CADASTRE-
334 CIT inventory are generated by using annual national emissions provided by CITEPA for 2011
335 (Figure 3a). Nevertheless, here these emissions have been spatially distributed differently than
336 for the TNO-GEN inventory. This has been done by using the FAO Gridded Livestock of the
337 World database with a resolution of 30 seconds of arc. The temporalization of the emissions
338 has been performed as a function of temperature and wind speed with the parameterizations of
339 Skøjth et al. [2011] for the different subsectors.

340 **3. Results and discussion**

341 First, we analyze the different contributions of livestock and of fertilization to the spring budget
342 in the CADASTRE-CIT inventory. Then, the comparison of the two alternative inventories
343 NH₃SAT and CADASTRE-CIT versus the reference inventory TNO-GEN, and their inter-
344 comparison are made at different temporal and spatial resolutions. We evaluate the inventories
345 at the national scale and at the scale of the different French administrative regions (the
346 administrative division in France on level 2 of the unified NUTS territory classification,
347 NUTS2, shown in Figure 2). We also analyze their spatial variability at the 0.5° (longitude)
348 $\times 0.25^\circ$ (latitude) resolution in order to draw a first picture of the consistency of the inventories
349 in terms of the spring NH₃ total budget and to identify regions of interest. Finally, we focus on

350 the temporal variability of the identified regions and discuss the agricultural practices that can
 351 influence the variability but also down to which temporal resolution the comparison of the
 352 inventories is relevant.

353



354

355 **Figure 2.** Localization and names of the different French regions as taken into account in this
 356 study. The regions are listed according to the TNO-GEN annual budget, in descending order.
 357 The grey boxes A, B, C, D describe the domains we respectively call the northeastern,
 358 northwestern, southeastern and southwestern parts of France in the following.

359

360 3.1. Respective contributions of different sources

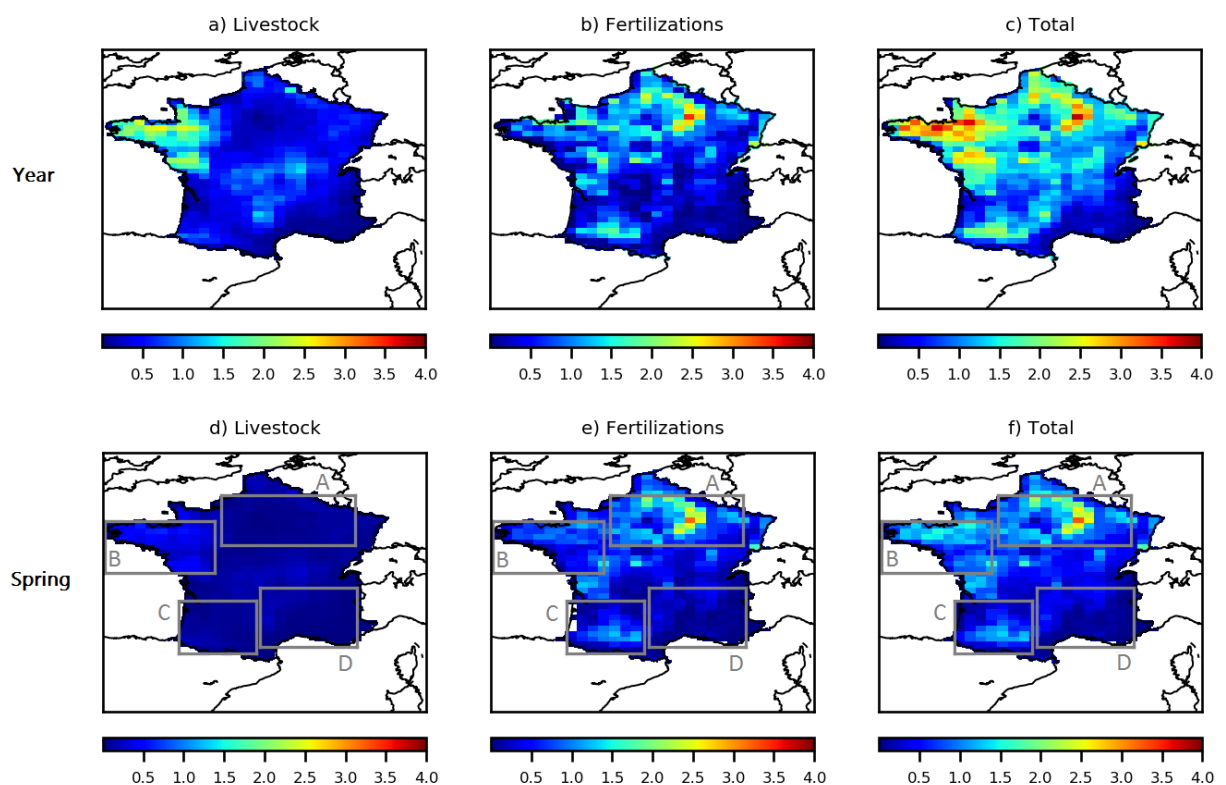
361 The different contributions of livestock and of fertilization to the annual and to the spring
 362 budget in the CADASTRE-CIT inventory are shown in Figure 3. This figure shows that the
 363 contribution of fertilization to the annual French budget can be strong, with emissions occurring
 364 mainly during spring.

365 The different contributions of livestock and of fertilization to the spring budget in the
 366 CADASTRE-CIT inventory highlight four different domains of interest. We can see that the
 367 contribution of fertilization on the high emissions of the northeastern part of France (box A) is
 368 strong. For example, the contribution of fertilization is about 99% in the region Champagne-
 369 Ardennes, and about 85% in the region Picardie (Table 2). These emissions are mainly due to

370 the use of mineral fertilizer over barley, sugar beet, and potato [Ramanantenasoa et al., 2018;
 371 Générumont et al., 2018]. In particular, the use of urea or nitrogen solution and the high soil pH
 372 [Hamaoui-Laguel et al., 2014; Ramanantenasoa et al., 2018; Générumont et al., 2018] -
 373 parameters not taken into account by the TNO-GEN inventory- seem to be the factors
 374 responsible for the high emissions in this domain.

375 The second domain of interest is the northwestern part of France (box B in Figure 3). Over this
 376 domain with high emissions, the NH_3 emissions are due in roughly equal parts to livestock
 377 (including animal housing, manure storage, and grazing) and to fertilizations, with a high use
 378 of organic manure [Ramanantenasoa et al., 2018]. Livestock farming indeed produces large
 379 amounts of livestock manure available for application on grassland and on arable crop.

380 The third domain of interest is the southeastern part of France (box C), showing the smallest
 381 spring NH_3 emissions. Finally, the contribution of fertilization on the emissions of the
 382 southwestern part of France (box D) is strong.



383
 384 **Figure 3.** top) Yearly NH_3 emissions due to a) livestock husbandry and manure storage, b) N
 385 fertilization (organic and mineral) and c) all sources in the CADASTRE-CIT inventory, in
 386 ktNH_3 . bottom) the same for spring NH_3 emissions, from March to May 2011.

387

388

3.2. French spring NH₃ total budget and its main spatial features

389 The spring NH₃ total budget is shown in Table 2 at the national scale and at the French regional
 390 scale. The French spring ammonia budgets, calculated for the period from March to May 2011,
 391 estimated by the NH₃SAT (264 ktNH₃) and the CADASTRE-CIT (354 ktNH₃) inventories are
 392 both higher than the TNO-GEN reference one (234 ktNH₃). The CADASTRE-CIT inventory
 393 estimates higher NH₃ spring emissions, by about 30%, than NH₃SAT.

394

395 The relative agreement on national budget between TNO-GEN and NH₃SAT must be nuanced
 396 as total budget values from NH₃SAT and TNO-GEN are close but large differences in the
 397 spatial distribution of the French NH₃ emissions between TNO-GEN and both NH₃SAT and
 398 CADASTRE-CIT can be observed (Figure 4).

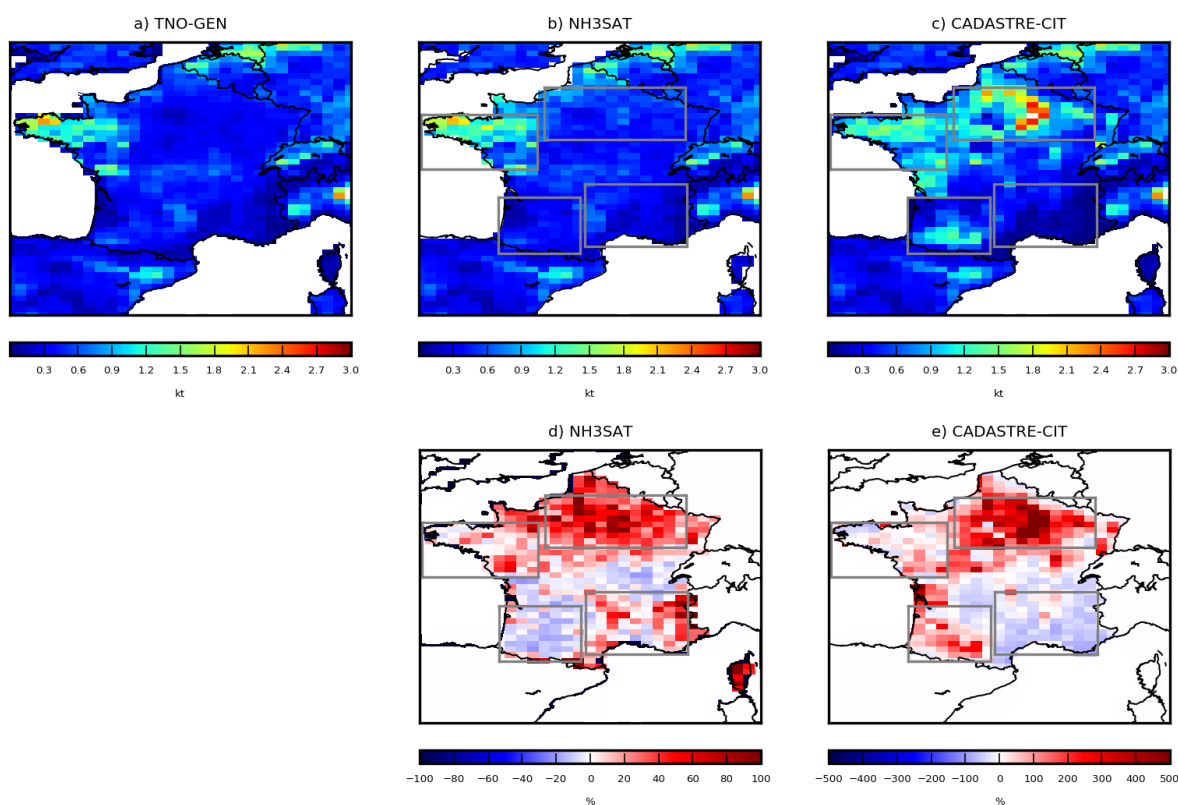
399

	TNO-GEN (in ktNH ₃)	NH ₃ SAT (in ktNH ₃)	CADASTRE-CIT (in ktNH ₃)	Contribution of the fertilization to the spring budget in CADASTRE-CIT (in %)
Regions in the northeastern part of France (box A in Figure 2)				
Champagne-Ardennes	8	13 (+63%)	35 (+337%)	99
Centre	13	15 (+15%)	32 (+146%)	81
Lorraine	9	12 (+33%)	17 (+88%)	80
Picardie	8	12 (+50%)	30 (+275%)	85
Haute-Normandie	7	9 (+28%)	12 (+71%)	75
Ile de France	3	4 (+33%)	9 (+200%)	77
Regions in the northwestern part of France (box B in Figure 2)				
Bretagne	34	34 (=)	30 (-12%)	61
Pays de la Loire	25	28 (+12%)	29 (+16%)	45
Regions in the southeastern part of France (box C in Figure 2)				
Rhône-Alpes	13	14 (+8%)	12 (-8%)	37
Auvergne	12	12 (=)	11 (-8%)	41
Languedoc	5	6 (+20%)	3 (-40%)	53
PACA	5	5 (=)	2 (-60%)	47
Regions in the southwestern part of France (box D in Figure 2)				
Midi-Pyrénées	18	18 (=)	26 (+44%)	60
Aquitaine	12	11 (-8%)	15 (+25%)	50
Other regions				
Alsace	4	5 (+25%)	7 (+175%)	87
Basse-Normandie	11	15 (+36%)	15 (+36%)	68

Bourgogne	11	13 (+18%)	20 (+81%)	55
Franche-Comté	6	6 (=)	8 (+33%)	43
Limousin	8	8 (=)	6 (-25%)	21
Nord-Pas-de-Calais	8	11 (+38%)	10 (+25%)	97
Poitou-Charentes	14	14 (=)	25 (+79%)	63
France				
	234	264 (+13%)	354 (+51%)	67

400 **Table 2.** French national and regional budgets of NH₃ spring emissions, from March to May
401 2011, in ktNH₃. The relative differences compared to the TNO-GEN are presented between
402 brackets, in %. Regions for which the inventories NH₃SAT and CADASTRE-CIT propose the
403 same sign of relative differences are marked in bold. The contributions of the fertilization
404 emissions to the NH₃ regional spring budget in the CADASTRE-CIT inventory are shown in %.

405 Indeed, over France and for the spring budget, the spatial correlation compared to the
406 CADASTRE-CIT inventory, which should better represent the agricultural practices and their
407 spatial distribution, is improved when using the NH₃SAT inventory instead of using TNO-GEN
408 (Pearson correlation coefficient $r=0.78$ with NH₃SAT against $r=0.72$ with TNO-GEN).



409 **Figure 4.** French NH₃ spring emissions estimated by a) the TNO-GEN inventory, b) the
410 NH₃SAT inventory and c) the CADASTRE-CIT inventory, in ktNH₃. Relative differences of d)
411 the NH₃SAT, and e) the CADASTRE-CIT inventories compared to the TNO-GEN inventory, in
412

413 %, at the pixel resolution, for the period from March to May 2011. Note that the scale is
414 different between 4b) and 4c).

415 The northeastern part of France presents the largest difference with the TNO-GEN inventory
416 (48 ktNH₃) for both NH₃SAT and CADASTRE-CIT inventories (65 and 135 ktNH₃,
417 respectively). The high emissions in the northeastern part of France are in agreement with the
418 MASAGE_NH₃ inventory [Paulot et al., 2014], the magnitude of their annual NH₃ emissions
419 from mineral fertilizer being calculated by combining an inventory of crop acreages, crop- and
420 country-specific fertilizer application rates and fertilizer-, crop-, and application-specific
421 emission factors. Emissions are higher for both NH₃SAT and CADASTRE-CIT inventories
422 compared to TNO-GEN over the Champagne-Ardennes (+337% and +63%, respectively for
423 CADASTRE-CIT and NH₃SAT, Table 3), Picardie (+275%, and +50%, respectively), Centre
424 (+146% and +15%, respectively), Haute-Normandie (+71% and +28%, respectively), Lorraine
425 (+88%, and +33%, respectively) and Ile de France regions (+200%, and +33%, respectively).

426 The northwestern part of France presents the largest NH₃ emissions according to the TNO-GEN
427 inventory (Figure 4a). The TNO-GEN, NH₃SAT and CADASTRE-CIT inventories lead to
428 similar spring budget (68, 73 and 71 ktNH₃, respectively), over this domain.

429 Over the southeastern part of France, CADASTRE-CIT is about 23% lower than NH₃SAT (28
430 and 37 ktNH₃, respectively, Table 2). One hypothesis to explain the lower NH₃ emissions in
431 CADASTRE-CIT is that market gardening is important in this area and not taken into account
432 in the CADASTRE-CIT inventory [Ramanantenasoa et al., 2018; Générmon et al., 2018].
433 Nevertheless, market gardening is not included, to our knowledge, in the TNO-GEN inventory.
434 TNO-GEN and NH₃SAT inventories being in quite good agreement in terms of budget (35 and
435 37 ktNH₃, respectively, Table 2), further work is required to understand these discrepancies.

436

437 Finally, over the southwestern part of France (box D in Figure 4), IASI observations only trigger
438 slight corrections to the TNO-GEN inventory over this area (29 and 30 ktNH₃, respectively)
439 and CADASTRE-CIT is 36% higher than TNO-GEN and NH₃SAT (Table 2).

440 3.3. Temporal variability of the NH₃ emissions at the sub-seasonal scale

441 Monthly regional budgets have been calculated for the three inventories. Figure 5 presents the
442 monthly variability of the NH₃ emissions from February to May 2011 for the four domains of
443 interest presented above. February is only displayed here as a baseline to show the sharp peak

444 of NH₃ emissions in March over some domains. The contribution of the emissions due to
445 livestock in the CADASTRE-CIT spring budget is also given.

446
447 The TNO-GEN inventory shows rather similar NH₃ emissions from March to May for all
448 regions, with a slight maximum in April (Figure 5a), imposed by the used GENEMIS monthly
449 profiles for the temporalization of emissions [Ebel et al., 1997]. On the contrary, over the
450 northeastern part of France, both NH₃SAT and CADASTRE-CIT inventories show a maximum
451 in March, and a decrease until May by about a factor of 1.5 to 2 (Figure 5a). This maximum in
452 March is also noticed by Tournadre et al. [2020], providing nine years of total column
453 observations from ground-based infrared remote sensing over the Paris megacity. We
454 calculated the monthly contribution of livestock to the NH₃ emissions based on CADASTRE-
455 CIT, which allows one to separate this contribution from the fertilization one. As in Figure 3,
456 Figure 5a confirms that NH₃ emissions are mainly due to fertilization in the northeastern part
457 of France, and shows that the seasonal variation is mainly driven by this contribution.
458 CADASTRE-CIT shows larger values than NH₃SAT which might be partly due to a possible
459 low bias in the IASI observations [Dammers et al., 2017; Tournadre et al., 2020] combined to
460 the fact that the TNO-GEN inventory (negatively biased compared to CADASTRE-CIT) is
461 used as a priori for the mass balance approach with no correction applied to the a priori when
462 IASI observations are not available.

463 To go further in the comparison, we analyzed the daily variability of the NH₃SAT and
464 CADASTRE-CIT inventories -TNO-GEN representing no daily variability (Figure 6). To
465 interpret the results, some limitations of the inventories have to be considered. For NH₃SAT,
466 the corrections applied to the TNO-GEN emissions are only applied for clear-sky conditions
467 when IASI observations are available. In the CADASTRE-CIT, fertilization days are randomly
468 selected within two-week intervals of application extracted from the farm survey analysis
469 [Ramanantenasoa et al., 2018], thus the actual day of fertilization is unknown. However, the
470 NH₃ volatilization is continuous over several days after spreading, reducing the uncertainty
471 introduced by this random selection. Moreover, the random selection is made at the field scale
472 (see section 2.3), then spatially averaging at the CHIMERE resolution should also smoothed
473 the random selection effect. Both effects of weather conditions on fertilization timing, on the
474 one hand, and on volatilization intensity and dynamics at the time of application and after
475 application, on the other hand, are realistically produced. Hence, the NH₃ volatilization is
476 continuous over several days after spreading introducing additional smoothing.

477 The CADASTRE-CIT inventory presents a high day-to-day variability from March to May
478 2011 (Figure 6a) with several strong maxima of emissions - characteristic of emissions due to
479 fertilizer application and a significant effect of the varying meteorological conditions. Also for
480 NH₃SAT, day-to-day variability is large. However, NH₃SAT and CADASTRE-CIT maxima
481 are not very well correlated. Over the first 15-days in March 2011, the high emissions occurring
482 from the 9th to 16th in CADASTRE-CIT are not reproduced by the NH₃SAT inventory,
483 potentially because of a lack of IASI coverage for this period (only about 40% of the domain –
484 Figure 6a). NH₃SAT shows an emission maximum one week earlier from the 1st to 7th. This
485 time gap in emission maxima could be explained by the random selection of fertilization days
486 in CADASTRE-CIT. Over the last 15 days in March 2011 and over the first 15 days in April
487 2011, the maxima of emissions estimated by NH₃SAT and CADASTRE-CIT are more
488 correlated, (e.g., from 7th to 11th April in CADASTRE-CIT versus from 9th to 11th in NH₃SAT).
489 Finally, CADASTRE-CIT still shows high emissions in the last 15 days of April and in May
490 2011, particularly related to the use of fertilizer over corn. Despite a good coverage with IASI
491 observations, no specific high emissions are derived from IASI for the same periods.

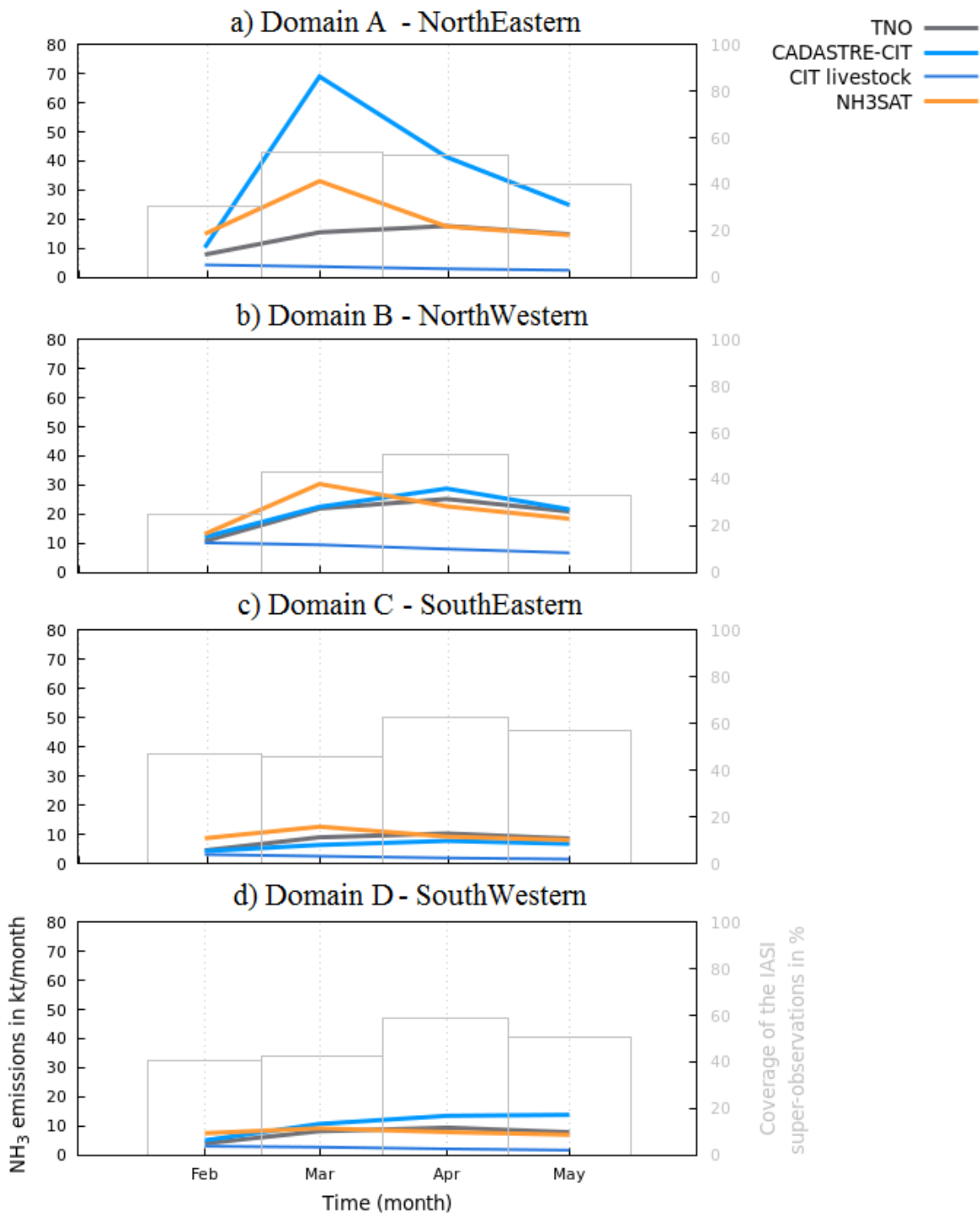
492 Over the northwestern part of France, the CADASTRE-CIT inventory is in agreement with
493 TNO-GEN, with a slight maximum in April (Figure 5b). The NH₃SAT inventory shows a
494 maximum in March, related to the high emissions seen during the first week of March 2011
495 (Figure 6b). However, the monthly differences are much smaller than for the northeastern part
496 of France. Also for this domain, the day-to-day variabilities provided by CADASTRE-CIT and
497 NH₃SAT are mostly uncorrelated: a strong sharp maximum of emissions during the first week
498 of March 2011 is seen in NH₃SAT -but not reproduced in CADASTRE-CIT (Figure 6b). The
499 highest daily emissions in CADASTRE-CIT occur during the two last weeks of April 2011 and
500 are not reproduced by NH₃SAT.

501

502 Over the southeastern and southwestern parts of France, month-to-month variations and
503 emission amounts are much smaller than for the two previous domains (Figures 5-6c and
504 Figures 5-6d, respectively). In the southeastern part of France, NH₃SAT emissions are slightly
505 larger than TNO-GEN and CADASTRE-CIT at the beginning of the period (February and
506 March, Figure 5c). The fertilization contribution to CADASTRE-CIT emissions slightly
507 decreases at the end of the period. In the southwestern part of France, NH₃SAT and TNO-GEN
508 are very similar (Figure 5d) and CADASTRE-CIT is slightly larger at the end of spring. This
509 increase is mainly related to fertilization emissions, the livestock contribution being stable.

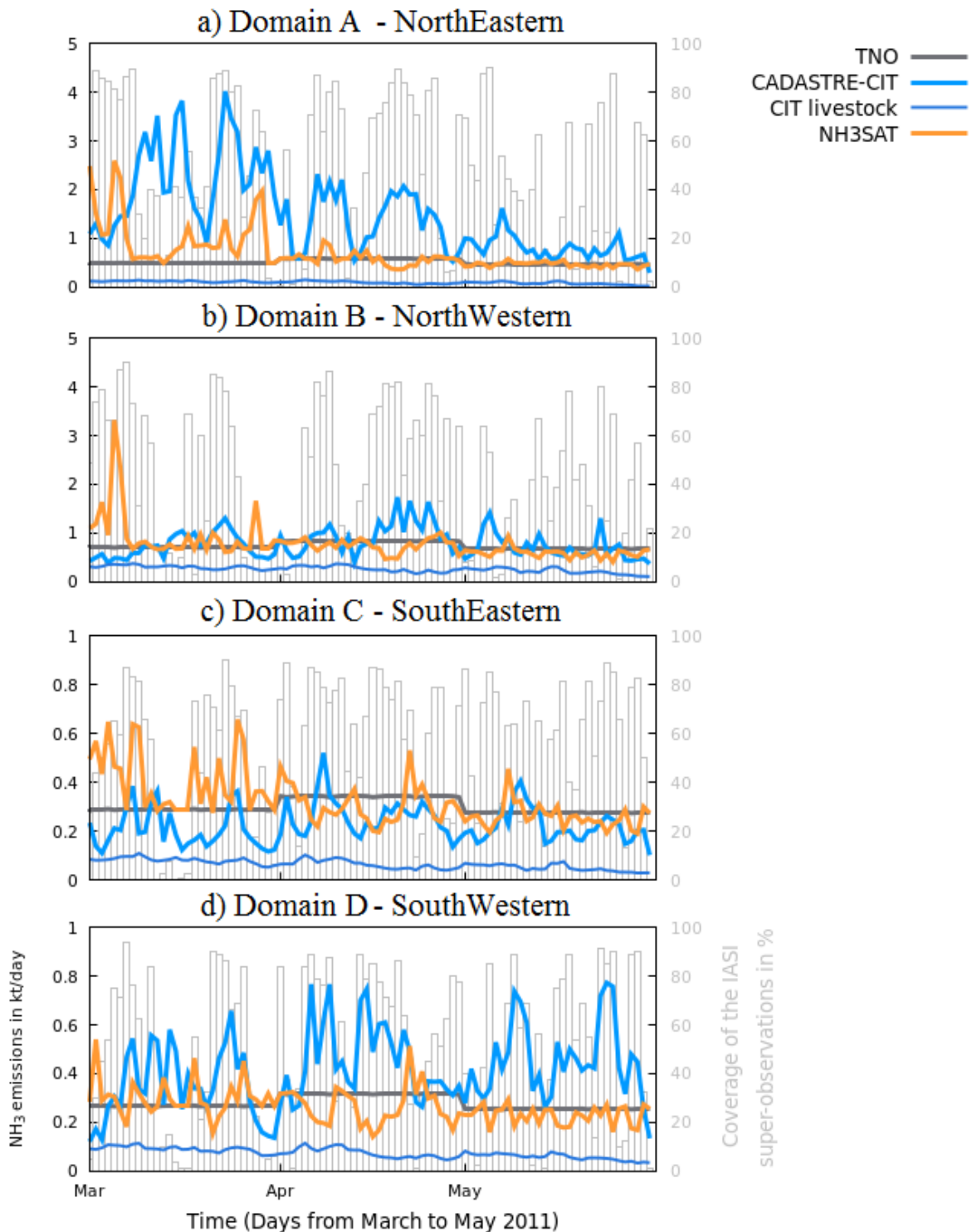
510 Again, daily time series between both NH₃SAT and CADASTRE-CIT inventories are
 511 uncorrelated (Figures 6c, d).

512
 513



514
 515 **Figure 5.** Time series of monthly NH₃ emissions estimated by TNO-GEN (in black), by NH₃SAT
 516 (in orange), and by CADASTRE-CIT (in blue) inventories, from February to May 2011, for (a)
 517 domain A (northeastern), (b) domain B (northwestern), (c) domain C (southeastern), and (d)

518 domain D (southwestern), as defined in Figure 2. The contribution of the emissions due to
 519 livestock in the CADAStRE-CIT monthly budgets is also given. Units are ktNH₃/month. The
 520 monthly regional coverage of the IASI super-observations is given in % (in grey).



521

522 **Figure 6.** Time series of daily NH₃ emissions estimated by TNO-GEN (in black), by NH₃SAT
 523 (in orange), and by CADAStRE-CIT (in blue) inventories, from March to May 2011, for (a)
 524 domain A (northeastern), (b) domain B (northwestern), (c) domain C (southeastern), and (d)

525 *domain D (southwestern), as defined in Figure 2. Note that the scale is different between a),b)*
526 *and c),d). Units are ktNH₃/day. The contribution of the emissions due to livestock in the*
527 *CADASTRE-CIT daily variability is also given. The daily regional coverage of the IASI super-*
528 *observations is given in % (in grey).*

529 **Conclusion**

530 In this study, we performed an inter-comparison of two alternative inventories with the TNO-
531 GEN reference inventory that quantify the French NH₃ emissions during spring 2011. One of
532 the main conclusions of this study is that over regions with large mineral fertilizer use, like over
533 the northeastern part of France, induced NH₃ emissions are probably considerably
534 underestimated by the TNO-GEN reference inventory, as both the NH₃SAT (constrained by
535 IASI observations) and CADASTRE-CIT (process level oriented), show much larger values.
536 For instance, over northeastern France, NH₃SAT and CADASTRE-CIT show respectively a
537 factor 1.4 and 2.8 larger spring 2011 emissions than TNO-GEN. Over the whole France, NH₃
538 emissions are still more than 50% larger in CADASTRE-CIT than in TNO-GEN. Average
539 French NH₃SAT emissions are about 10% larger than TNO-GEN ones. Over the southern part
540 of France, with a more diverse agriculture as compared to the crop intensive one in northeastern
541 France, differences between the inventories are on the whole lower, and signs between
542 CADASTRE-CIT / TNO-GEN and NH₃SAT / TNO-GEN corrections are often opposite for
543 different regions.

544

545 Month-to-month variations are again much more pronounced over northeastern France and
546 show a maximum in March for both CADASTRE-CIT and NH₃SAT. Day-to-day variations are
547 large in CADASTRE-CIT and NH₃SAT, roughly a factor of 5 between minimal and maximal
548 values. This shows the interest in evaluating NH₃ emissions at a daily scale because this input
549 is required for chemistry transport modeling of particulate matter formation and thus air quality.
550 However, time-series delivered by CADASTRE-CIT and NH₃SAT are uncorrelated for all
551 considered regions. This result can be partly explained by the lack in IASI NH₃ column
552 observations under partially cloudy conditions, and by the fact that available information on
553 agricultural practices is resolved at a two weeks scale.

554

555 Yet, current results of our study have important implications for air quality modelling over
556 Europe. The important changes in the spatial distribution of NH₃ emissions as a function of
557 soil properties are of general concern not only for France, but for whole Europe. Soils are

558 alkaline or neutral ($\text{pH}>6$) not only over North-Eastern France, but also over large parts of Italy,
559 eastern Spain, or eastern Germany [Reuter, 2008]. Over these regions, our study suggests
560 potentially larger NH_3 emissions than with a constant emission factor treatment, with impacts
561 then on fine particle formation. These features should be included in “operational” emission
562 inventories used for air quality modelling.

563

564 Thus, as a general conclusion, use of observational constrained or process-oriented emission
565 inventories is clearly of added-value for estimating better monthly averages over French areas
566 with intensive mineral fertilizer use, but capacity for delivering day-to-day variations is not yet
567 proven. This warrants further studies, both refining hypotheses on days chosen by farmers for
568 fertilizer spread out as a function of meteorological conditions, and, acquiring and using
569 continuous surface NH_3 measurements for validating satellite or process derived NH_3 emission
570 variability.

571

572 **Competing interests**

573 The authors declare that they have no conflict of interest.

574 **Author Contribution**

575 All authors have contributed to the manuscript writing (main authors A.F-C, G.D, S.G and MB).
576 A.F-C has performed the mass-balance approach to deduce NH_3 emissions from NH_3 total
577 columns observed by the IASI satellite instrument. K.D, J-M.G and S.G have performed the
578 bottom-up approach providing the CADASTRE_ NH_3 inventory for organic and mineral
579 fertilization practices. F.C has compiled this CADASTRE_ NH_3 inventory with livestock
580 emissions from the CITEPA. L.C, P-F.C, M.V.D and C.C are the PIs of the IASI NH_3 product.
581 All authors discussed the results and contributed to the final paper.

582 **Code and Data Availability**

583 The CHIMERE code is available here: www.lmd.polytechnique.fr/chimere/.

584 The IASI ANNI- NH_3 -v2.2R data are freely available through the AERIS database:
585 <https://iasi.aeris-data.fr/nh3/>.

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