Do alternative inventories converge on the spatiotemporal representation of spring ammonia emissions in France?

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Abstract

Agriculture is the main source of ammonia (NH_3) in France, an important gaseous precursor of atmospheric particulate matter (PM). National and even more global emission inventories are known to have difficulty representing the large spatial and temporal variability inherent to atmospheric NH_3. In this study, we compare NH_3 emissions in France during the spring 2011 from **one reference inventory, the TNO inventory, and two alternative inventories that account in different manners for both the spatial and temporal variabilities of the emissions: (i) the NH₃SAT satellite-derived inventory based on IASI NH_3 columns and (ii) the CADASTRE-CIT inventory that combines NH_3 emissions due to nitrogen fertilization calculated with the mechanistic model VOLT'AIR on the database of the CADASTRE_NH₃ framework and other source emissions from the CITEPA.** The total spring budgets, **from March to May 2011**, at the national level are higher when calculated with both alternative inventories than with the reference, the difference being more marked with

CADASTRE-CIT. NH_3SAT and CADASTRE-CIT inventories both yield to large NH_3 **spring** emissions due to fertilization on soils with high pH in the northeastern part of France (65 $ktNH_3$ and 135 $ktNH_3$, respectively, vs 48 $ktNH_3$ for TNO-GEN), while soil properties are not accounted for by the TNO-GEN methodology. For the other parts of France, the differences are smaller. The timing of fertilization and associated ammonia emissions is closely related to the nitrogen requirements and hence the phenological stage of the crops, and therefore to the crop-year's specific weather conditions. Maximum emissions are observed in March for 2011 for some regions for both alternative inventories, while April is the period with maximum emissions for the reference inventory whatever the region or the year. Comparing the inventories at finer temporal resolutions, typically at daily scale, large differences are found. The convergence of alternative, independent and complementary methods on the spatiotemporal representation of the spring NH_3 emissions particularly over areas where the contribution of mineral fertilizer spreading to the spring budget is strong, **encourage** further developments in both prospected complementary directions, as this will help improving national NH_3 emission inventories.

1. Introduction

France is a major crop producer and a major exporter of agricultural and food products. In 2014, it produced 2%, 4%, 5%, 8%, 8% and 14% of the global production of maize, sunflower, wheat, barley, rapeseed and sugar beet, respectively [Food and Agriculture Organization of the United Nations FAO, Schauberger et al., 2018]. Through this food cultivation and also due to animal husbandry, agriculture is the main source of ammonia (NH_3) in the country. As an important gaseous precursor of particulate pollution, harmful to human life [Lelieveld et al., 2015; WHO, 2016], ammonia plays an important role in the regulation of inorganic aerosol concentrations [Erisman and Schaap, 2004, Bauer et al., 2016], and contributes to N deposition and potential exceedance of critical loads of ecosystems [Erisman et al., 2007; EEA European Environment Agency, 2014]. In order to limit air pollution, also responsible for acidification and eutrophication, the new European National Emission Ceilings Directive 2016/2284, replacing the Directive 2001/81/EC, now sets ambitious national reduction commitments for ammonia. Ammonia emissions indeed have to be reduced by 19% in 2030, compared with the 2005 levels [OJEU, 2016].

At the European scale, total NH$_3$ emissions are provided by the European Monitoring and Evaluation Program (EMEP) [Vestreng, 2005] or by the TNO-MACCIII [Kuenen et al., 2014] inventories that rely on national annual declarations and estimates of emission factors. Emissions are accounted for without separating fertilization and livestock. These reference inventories are widely used by the scientific community to study the impact of pollutant emissions on the chemical composition of the troposphere and on air quality. Nevertheless, uncertainties on the quantification of the NH$_3$ emissions are usually estimated to be between 100 and 300% of the annual budgets in the reference inventories [EMEP/EEA, 2016; Kuenen et al., 2014]. In addition, the temporal and spatial variability may not **be** well represented in the reference inventories, as the temporal profiles used do not account for meteorology, soil properties and other local conditions. Moreover, fertilizer spreading is of particular interest, as these are applied during small periods, especially during a few weeks at the end of winter and early spring. However, the exact timing of fertilizer spreading is difficult to predict, as it depends on agricultural practices and meteorological conditions, which is not taken into account in the temporal disaggregation of the reference emission inventories. Both the inaccurate temporal resolutions and the mis-representation of the spreading emissions largely explain the difficulty encountered by models to represent seasonal or daily pattern of NH$_3$ concentrations [Menut et al., 2012], and consequently particulate matter levels [Fortems-Cheiney et al, 2016].

To reduce these uncertainties, a better quantification of agricultural ammonia emissions and its **temporal** and spatial evolution is necessary. In particular, one of the challenges is to capture the right timing of fertilizer spreading at the weekly or even at the daily scale in order to reflect the effect of environmental and agronomical conditions on ammonia emissions. To this end, mechanistic models taking into account meteorological conditions, soil properties and agricultural practices, have been developed (e.g., for Denmark [Skjøth et al., 2004], for the UK [Hellsten et al., 2008], and for mineral fertilization in springtime in France [Hamaoui-Laguel et al., 2014]). Limitations for such approaches come from the fact that detailed agricultural data needed as input to such models are not available for most of the European countries. Moreover, agricultural practices of a specific country cannot be extrapolated to another country [Skjøth et al., 2011].

As an alternative to direct emission modeling, attempts have been made to constrain ammonia emissions through inverse approaches, **based on ammonium wet deposition data** [**Paulot et al., 2014**] **or** on satellite observed atmospheric ammonia distributions (e.g., from the

99 Tropospheric Emission Spectrometer TES [Zhu et al., 2013; **Zhang et al., 2018**], from the

100 Infrared Atmospheric Sounding Interferometer (IASI) [Fortems-Cheiney et al., 2016; Van

101 Damme et al., 2018; Adams et al., 2019] or from the Cross-track Infrared Sounder CrIS [Adams

102 et al., 2019; Dammers et al., 2019]). In principle, such emission estimates can be available

103 shortly after observation. The advantage of satellite-derived estimates is also that these can be

104 derived globally, at a high temporal scale (e.g., daily scale under clear sky). The downside of

105 these however, is that they do not provide information on the underlying sources of the

106 emissions (fertilizers vs husbandry), or e.g. the date of fertilization, the type of fertilizers used,

107 the fertilization rates, etc., that could be important for the regulation of NH_3 emissions.

108

109 In this context, we compare ammonia emissions in France from inventories using the different

110 approaches mentioned above: (i) the reference, **in the sense that this is the method currently**

111 **used,** hereafter called TNO-GEN, is the European inventory based on the annual budgets

112 provided by the TNO-MACCIII inventory [Kuenen et al., 2014] and on seasonal profiles from

113 GENEMIS [Ebel et al., 1997]; (ii) a first alternative inventory, hereafter called NH_3SAT, is

114 based on a top-down approach starting from the IASI derived NH_3 columns; (iii) the other

115 alternative inventory, hereafter called CADASTRE-CIT, is based on a bottom-up approach

116 quantifying NH_3 emissions due to nitrogen fertilization combining spatiotemporal data and

117 calculations performed within the CADASTRE_NH_3 framework with the mechanistic model

118 VOLT'AIR (Ramanantenasoa et al., [2018]; Génermont et al., [2018]) completed with livestock

119 and other source emissions from the French Interprofessional Technical Centre for Studies on

120 Air Pollution (CITEPA). This study aims at assessing the potential contribution of better spatial

121 and temporal representation of fertilization-related ammonia emissions to the quality of

122 ammonia emission inventories. The improvement is assessed in terms of total budget, spatial

123 distribution and timing of the emissions. The study period, spring 2011 **(from March, 1st to**

124 **May 2011, 31st)**, was chosen following three criteria. Firstly, because at the time of the study,

125 the last French agricultural data were available from AGRESTE [AGRESTE, 2014] for the

126 agricultural year 2010-2011, allowing the application of the CADASTRE_NH_3 framework for

127 the quantification of the spatio-temporal distribution of NH_3 emissions due to nitrogen

128 fertilization for this crop year [Ramanantenasoa et al., 2018;Génermont et al., 2018]. Secondly,

129 ammonia emissions are enhanced during spring in accordance with crops requirements [Skjøth

130 et al., 2004; Ramanantenasoa et al., 2018; Génermont et al., 2018]. Finally, unlike autumn and

131 winter months, the NH_3 spring levels are detectable with a better confidence in the IASI satellite

132   observations [Viatte et al., 2020], allowing the extension of the preliminary work of Fortems-
133   Cheiney et al. [2016] to deduce NH_3 emissions from the IASI satellite instrument.

134   The three inventories and methods to build them used for this study are presented in Section 2

135   and the results of the comparison are given and discussed in Section 3.

136   **2. Inventories**

137   The three inventories TNO-GEN, NH_3SAT and CADASTRE-CIT compared in this study are

138   described in Table 1 and in the following sections. It is worth noting that only the CADASTRE-

139   CIT inventory provides information on the respective contribution of fertilization and livestock

140   emissions. The spatial **and temporal** resolutions of the inventories are also shown in Table 1.

Name	Spatial Resolution (latitude x longitude)	Temporal Resolution	Fertilization emissions	Livestock emissions
TNO-GEN	0.125°x0.0625°	Monthly	-	
NH_3SAT	0.5°x0.25°	Daily	-	
CADASTRE-CIT	0.015625° x0.03125°	**Hourly**	CADASTRE_NH_3 Ramanantenasoa et al., [2018] and Génermont et al., [2018]	
	0.007825°x0.007825°	Daily		CITEPA national emissions, temporalized according to Skøjth et al., [2011]

141   **Table 1**. *Main characteristics of the different compared inventories **before their***
142   ***aggregation/disaggregation for the inter-comparison**.*

143   The inter-comparison is made at the 0.5° (longitude) x 0.25° (latitude) resolution. The outputs

144   of the TNO-GEN and the CADASTRE-CIT inventories **could** have consequently been

145   aggregated **or disaggregated**.

146   **2.1. TNO-GEN**

147   In this study, the TNO-GEN combines the annual budgets provided by the TNO-MACCIII

148   inventory and the seasonal profiles to deduce the monthly variability of NH_3 emissions. This

149   inventory is based on official annual emission data submitted by countries to EMEP/CEIP

(European Monitoring and Evaluation Programme/Centre on Emission Inventories and Projections) for air pollutants. It is the update of the TNO-MACCII inventory [Kuenen et al., 2014]. It is an inventory at 0.125°x0.0625° resolution providing annual emissions of NH_3 from the agricultural sector, without separating the contributions from fertilization and livestock. Hereafter, we use the TNO-MACIII emissions of year 2011. The seasonal profile of these emissions is prescribed according to the typical national factors provided by GENEMIS. This seasonal temporal profile used for the temporalization of emissions -the same one applied to the entire country- leads to a maximum in NH_3 emissions systematically in April over France [Ebel et al., 1997]. **The emissions remain constant between days in each month and between hours in each day**.

2.2. NH₃SAT

As a first alternative, a mass-balance approach, which is a common method for the quantification of short-lived species surface fluxes [Palmer et al., 2003; Jaeglé et al., 2004; Boersma et al., 2008; Lin et al., 2010] was set-up. We used it to deduce NH_3 emissions from differences between NH_3 total columns observed by the IASI instrument and simulated by the CHIMERE regional chemical transport model (CTM) using the TNO-GEN inventory as inputs data.

2.2.1 The regional CTM CHIMERE

CHIMERE simulates concentrations of gaseous and particulate chemical species [Menut et al., 2013; Mailler et al., 2017]. For this study, we used the CHIMERE version 2013a. The horizontal resolution is given as follows: $0.5° \times 0.25°$ over 17°W/40°E–32°N/70°N, including 115 (longitude) x 153 (latitude) grid-cells. The vertical grid contains 17 layers from the surface to 200 hPa. This model is driven by the European Centre for Medium-Range Weather Forecasts global meteorological fields [Owens and Hewson, 2018]. Climatological values from the LMDz-INCA global model [Szopa et al., 2008] are used to prescribe concentrations at the lateral and top boundaries and the initial atmospheric composition in the domain. For inorganic species, aerosol thermodynamic equilibrium is achieved using the ISORROPIA model [Nenes et al., 1998]. **As most of the models in the world,** the parameterization of NH_3 dry deposition is unidirectional in CHIMERE. **The parameterization of a bidirectional exchange with surfaces in Wichink Kruit et al. [2012] increased their yearly mean modeled LOTOS-**

EUROS European ammonia concentrations almost everywhere, and particularly over agricultural source areas. However, Zhu et al. [2015], with the Goddard Earth Observing System-Chemistry (GEOS-Chem) global CTM, estimated decrease of NH_3 European concentrations in April, when the inclusion of a compensation point for vegetation is included. Further work needs to be done to better investigate the sensitivity of NH_3 concentrations to the deposition. Nevertheless, without such parameterization for bi-directional exchange, Azouz et al. [2019] assessed that regional models such as CHIMERE usually operating with large grid cell sizes simulate quite well the average NH_3 dry deposition flux over a large domain of simulation.

The evaluation of CHIMERE NH_3 and NH_4+ concentrations should be done against NH_3 (as done in Fortems-Cheiney et al., [2016]) and NH_4+ measurements. Nevertheless, to our knowledge, there is no available NH_3 measurement over France for the focused period here. There is interpretable NH_4^+ surface measurements at only one site, making the interpretation of the results difficult. NH_3 and NH_4^+ comparisons during other periods are scarce also. For instance, Tuccella et al., [2019] compared CHIMERE simulated and observed NH_4^+ at the Cabaux supersite and found average concentrations for May 2008 of 1.3 µg/m^3 for both, with a correlation coefficient of 0.52. For the Paris agglomeration between September 2009 and 2010, the modelled regional NH_4^+ burden was 1.8 µg/m^3 while the modelled one was 1.6 µg/m^3 [Petetin et al., 2016]. From June to September 2010, 83% of modelled total NH_x was gaseous, while in the model, it was only 50%, coherent with this NH_3 was underestimated especially during warmer days. Thus, it is concluded for one site and season, that particulate NH_4^+ has a low to medium impact on NH_3.

2.2.2. The IASI observations

We use data from the IASI-A instrument, flying on a low Sun-synchronous polar orbit aboard Metop satellite since October 2006, with equator crossing times of 09:30 (descending mode) and 21:30 (ascending mode) local sidereal time (LST) [Clerbaux et al., 2009, **Clarisse et al., 2009**]. The spatial resolution of its observations is about 12x12 km^2 at nadir. The algorithm used to retrieve NH_3 columns from the radiance spectra is described in Van Damme et al., [2017]. Several improvements have been introduced since the description of Van Damme et al., [2014] and the version v1 used in our previous study [Fortems-Cheiney et al., 2016]. In this study, we use the reanalyzed dataset ANNI-NH3-v2.2R, relying on ERA-Interim meteorological input data from the European Centre for Medium-Range Weather Forecasts

(ECMWF) rather than the operationally provided Eumetsat IASI Level 2 (L2) data used for the standard near-real-time version [Van Damme et al., 2017]. We only consider land measurements from the morning overpass, as IASI is more sensitive at this time to the boundary layer, owing to more favorable thermal conditions [Clarisse et al., 2010, Van Damme et al., 2014].

The IASI total columns are averaged into "super-observations" (average of all IASI data within the 0.5°×0.25° resolution **and for the given CHIMERE physical time-step of about 5/10 minutes**). As suggested by Van Damme et al. [2017], we no longer use weighted averages for this purpose. We performed a sensitivity test by selecting the IASI pixels for which the retrieval error does not exceed 100%: the results **about the temporal and spatial variability of the NH_3 French emissions presented in Section 3** did not significantly change, showing the robustness of the IASI NH_3 product (not shown).

The resulting monthly means of IASI NH_3 columns from March to May 2011 are shown in Figure 1 (a to c). The spatio-temporal variability -with the highest values over northeastern France in March, and over northwestern France in April- is confirmed by the IASI 10-year and by the CrIS 5-year monthly means shown in Viatte et al. [2020]. Note that the potential of IASI to provide information at high temporal resolution, up to daily scale, can be hampered by the cloud coverage as only observations with a cloud coverage lower than 10% are delivered [Van Damme et al., 2017]. To evaluate the impact of this limitation, the number of IASI super-observations used to calculate these monthly means, which represents the number of days over a month covered by IASI, is shown in Figure 1 (d to f). On average, more than half of the month is sampled by IASI during spring, except in May in the northwestern part of France. The regions showing large IASI NH_3 values are consequently well sampled.
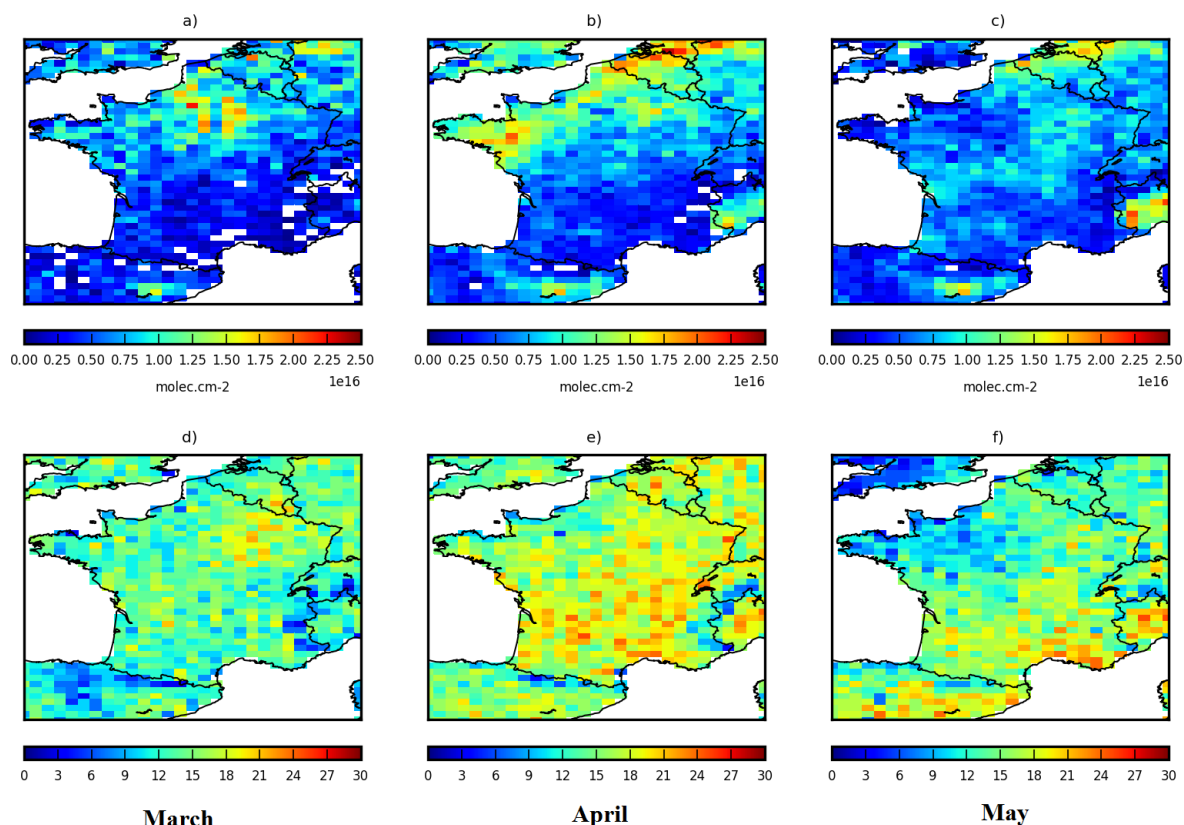
Figure 1. (*top) Monthly means of IASI "super-observations" for (a) March 2011, (b) April 2011 and (c) May 2011. Units are molec.cm^{-2}. (bottom) Total number of IASI super-observations per month in (d) March 2011, (e) April 2011 and (f) May 2011.*

2.2.3. Deducing NH$_3$SAT emissions

Relative differences between simulated columns by the CHIMERE regional CTM (described in Section 2.2.1, using the TNO-GEN emissions for the year 2011, described in Section 2.1) and observed IASI total columns (described in Section 2.2.2) are applied as a corrective factor to the reference emissions at daily and at grid-cell resolutions over France, from February to May 2011. As IASI "super-observations" provide one piece of information per day, the diurnal time profile of reference emissions cannot be improved: we apply the same daily correction factor to all hourly NH$_3$ emissions. When IASI is not **available** (i.e., observations with a cloud coverage higher than 10%), the correction is not applied and the emissions remain equal to the TNO-GEN ones. To compare the emissions with the CADASTRE-CIT inventory, and their respective simulations with CHIMERE, the correction is only applied over France here.

With the mass-balance approach, the transport to neighboring cells is assumed negligible following Palmer et al., [2003]. This approach has been debated by Turner et al. [2012], who

found that non-local sources contribute substantially to columns of short-lived species. Li et al. [2019] evaluated the ability of both a mass-balance approach and a variational assimilation to recover known NH_3 emissions at different spatial resolutions. At a 2°x2.5° resolution, they found that both methods yielded similar values. At a 0.25°x0.3125°, the mass-balance approach led to values about 20% higher compared to the variational ones. With our 0.5°x0.25° resolution, the use of a mass-balance approach would lead to additional errors of less than about 20% for the quantification of NH_3 emissions. This uncertainty is acceptable and much lower than the uncertainty existing in the annual and national budgets provided by emission inventories [EMEP/EEA, 2016]. In this context, we choose to perform such mass-balance approach to deduce NH_3 emissions from IASI ANNI-NH_3-v2.2R super-observations. Additional uncertainty comes from the IASI observations. The IASI minimum detection limit is of about 2-3 ppbv ($\sim$4-6.10^{15}molecules.cm^{-2}) [Clarisse et al., 2010]. The signal-to-noise ratio therefore presents better performance for regions with high local concentrations (e.g., northern part of France, Figure 1) than over low local concentration areas (e.g southern parts of France, Figure 1). There is no evaluation available for the IASI ANNI-NH_3-v2.2R product used here yet.

2.3. CADASTRE-CIT

As a second alternative, a bottom-up approach was set-up based on the finest national inventories available for anthropogenic sources of ammonia. The CADASTRE_NH_3 framework provides such an inventory for organic and mineral fertilization practices. This is however not the case for the other anthropogenic sources. For livestock emissions, **with the exception of the stage of effluent spreading in the field**, the less detailed inventory of the French Interprofessional Technical Centre for Studies on Air Pollution CITEPA is used. To meet the objectives of better specialization and temporalization, specific procedures are applied. These inventories are completed by the TNO-GEN inventory for the emissions of the other sectors.

2.3.1. Fertilization emissions from CADASTRE_NH_3

The CADASTRE_NH_3 was implemented in order to represent in a realistic way spatio-temporal variability of French NH_3 emissions due to mineral and organic N fertilization, and is fully described in Ramanantenasoa et al. [2018] and in Génermont et al. [2018]. It has been constructed through the combined use of two types of resources: the process-based VOLT'AIR

286 model and geo-referenced and temporally explicit databases for soil properties, meteorological

287 conditions and N fertilization.

288 VOLT'AIR is a 1D process-based model predicting NH_3 emissions from N fertilizers on bare

289 soils, from physical, chemical and biological processes [Le Cadre, 2004; Garcia et al., 2012].

290 It incorporates current knowledge on NH_3 volatilization after application of the main types of

291 organic manure and mineral N fertilizers in the field. It takes into account the major factors

292 known to influence NH_3 volatilization in the field, i.e., soil properties, weather conditions,

293 cultural practices and properties of mineral fertilizers and organic products. It runs at an hourly

294 time step at the field scale for a period of several weeks covering thus the entire volatilization

295 duration of fertilization events.

296 Local features are attributed to each simulation units, the Small Agricultural Regions (SAR):

297 local weather conditions (SAFRAN, Météo-France), the dominant soil type of the SAR from

298 the European Soil Data Center (ESDC), with soil properties provided by the Harmonized World

299 Soil Database (HWSD) of the Food and Agriculture Organization (FAO); areas cultivated in

300 crop year 2010-2011 per crop per region derived from the European Land Parcel Identification

301 System (LPIS, Common Agricultural Policy (CAP) regulations); Nitrogen fertilization

302 management practices are derived from data of the national AGRESTE survey of cultural

303 practices for arable crops and grassland (Department of Statistics and Forecasting of the French

304 Ministry of Agriculture) [AGRESTE, 2014]. All input data required by VOLT'AIR are

305 geographically overlaid and intersected with a Geographical Information System to generate

306 input combinations in each SAR. Each input combination is used as the input data for a virtual

307 300m x 300m field for a simulation using VOLT'AIR. Exact times and dates of fertilizations

308 are required to run VOLT'AIR, but for the sake of robustness, the statistical analysis of the

309 survey data has been performed on the basis of two-week intervals for the date of fertilization.

310 Fertilizations are thus randomly distributed within these two-week intervals in proportion to

311 their respective representation following Ramanantenasoa et al. [2018]. Each simulation of NH_3

312 emissions is run at an hourly time step for a period of two months, starting one month before

313 the fertilization in order to calculate soil water content at the time of application, and ending

314 one month after fertilization, in order to cover the whole volatilization event.

315 About 160 000 runs with the VOLT'AIR model have been performed over the crop-year 2010-

316 11to produce ammonia emissions per hour, per ha, per crop type, per SAR. Emissions can be

317 aggregated at different spatial and temporal scales. At the spatial scale, they are weighted with

the contribution of (i) each N fertilization management applied to each crop in each SAR and (ii) the area of the crop cultivated in the SAR. A procedure allows producing ammonia emissions at the required grid scale for the inventory comparison: it is based on cultivated areas for each crop as the key of desegregation-reagregation from the SAR to the 0.015625° x 0.03125°grid. At the temporal scale, emissions are aggregated over daily, weekly or monthly bases for the sake of comparison with TNO-GEN and NH$_3$SAT inventories. Volatilization taking place over several days, from few days to several weeks, one fertilization in one field contributes to ammonia emissions over several days or weeks. Weather condition effects on overall ammonia emissions is thus the result of both their effects on fertilization timing and their effects on volatilization intensity and dynamics over 30 days from fertilizer application.

As they are not available in the agricultural practice survey, N fertilizations of vegetables, fruits, and vines are not accounted for in CADASTRE_NH$_3$: their contribution is minor for France overall, only accounting for 5% of the total agricultural area [AGRESTE, 2010] but are important in particular regions. As the agricultural practice survey does not provide information over Corsica, this inventory is completed by the TNO-GEN inventory over this region (Figure 2).

2.3.2 Livestock emissions

As for the TNO-GEN inventory, French NH$_3$ emissions from livestock for the CADASTRE-CIT inventory are generated by using annual national emissions provided by CITEPA for 2011 (Figure 3a). Nevertheless, here these emissions have been spatially distributed differently than for the TNO-GEN inventory. This has been done by using the FAO Gridded Livestock of the World database with a resolution of 30 seconds of arc. The temporalization of the emissions has been performed as a function of temperature and wind speed with the parameterizations of Skøjth et al. [2011] for the different subsectors.

3. Results and discussion

First, we analyze the different contributions of livestock and of fertilization to the spring budget in the CADASTRE-CIT inventory. Then, the comparison of the two alternative inventories NH$_3$SAT and CADASTRE-CIT versus the reference inventory TNO-GEN, and their inter-comparison are made at different temporal and spatial resolutions. We evaluate the inventories at the national scale and at the scale of the different French administrative regions (the administrative division in France on level 2 of the unified NUTS territory classification,

NUTS2, shown in Figure 2). We also analyze their spatial variability at the 0.5° (longitude) x 0.25° (latitude) resolution in order to draw a first picture of the consistency of the inventories in terms of the spring NH₃ total budget and to identify regions of interest. Finally, we focus on the temporal variability of the identified regions and discuss the agricultural practices **that** can influence the variability but also down to which temporal resolution the comparison of the inventories is relevant.
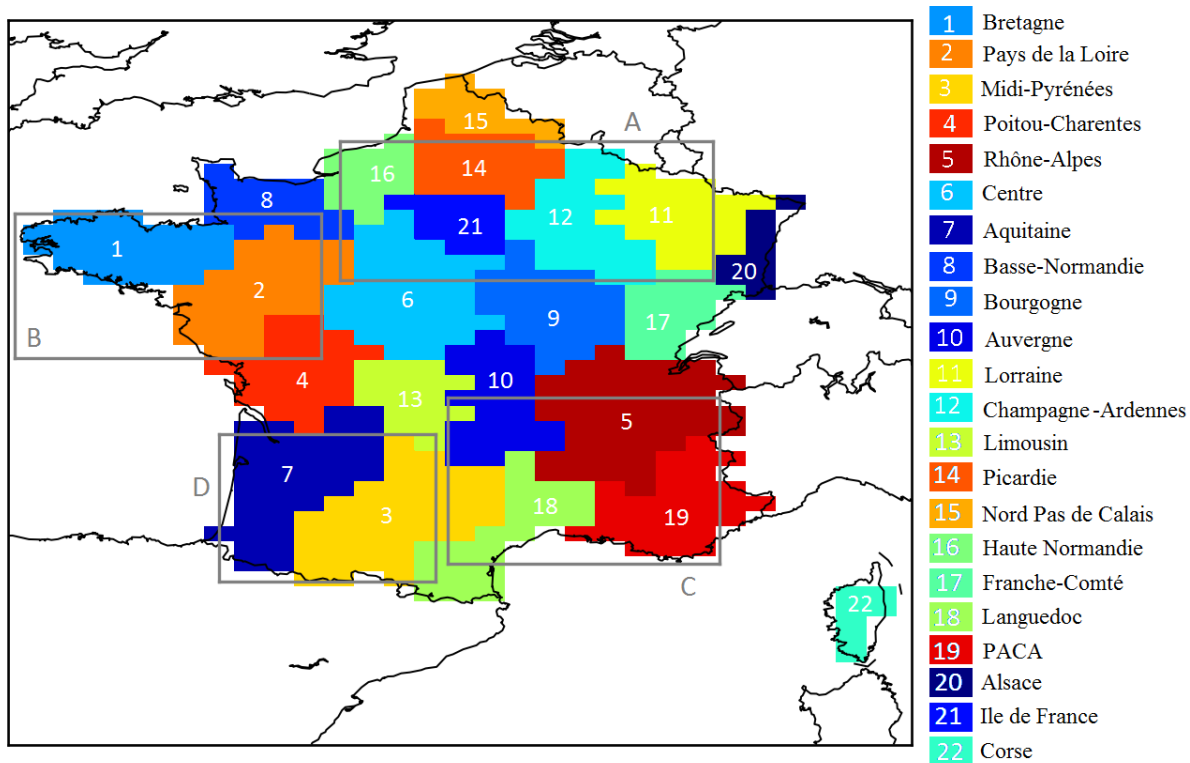
Figure 2. *Localization and names of the different French regions as taken into account in this study. The regions are listed according to the TNO-GEN annual budget, in descending order. The grey boxes A, B, C, D describe the domains we respectively call the northeastern, northwestern, southeastern and southwestern parts of France in the following.*

3.1. Respective contributions of different sources

The different contributions of livestock and of fertilization to the annual and to the spring budget in the CADASTRE-CIT inventory are shown in Figure 3. This figure shows that the contribution of fertilization to the annual French budget can be strong, with emissions occurring mainly during spring.

The different contributions of livestock and of fertilization to the spring budget in the CADASTRE-CIT inventory highlight four different domains of interest. We can see that the

370 contribution of fertilization on the high emissions of the northeastern part of France (box A) is
371 strong. For example, the contribution of fertilization is about 99% in the region Champagne-
372 Ardennes, and about 85% in the region Picardie (Table 2). These emissions are mainly due to
373 the use of mineral fertilizer over barley, sugar beet, and potato [Ramanantenasoa et al., 2018;
374 Génermont et al., 2018]. In particular, the use of urea or nitrogen solution and the high soil pH
375 [Hamaoui-Laguel et al., 2014; Ramanantenasoa et al., 2018; Génermont et al., 2018] -
376 parameters not taken into account by the TNO-GEN inventory- seem to be the factors
377 responsible for the high emissions in this domain.

378 The second domain of interest is the northwestern part of France (box B in Figure 3). Over this
379 domain with high emissions, the NH$_3$ emissions are due in roughly equal parts to livestock
380 (including animal housing, manure storage, and grazing) and to fertilizations, with a high use
381 of organic manure [Ramanantenasoa et al., 2018]. Livestock farming indeed produces large
382 amounts of livestock manure available for application on grassland and on arable crop.

383 The third domain of interest is the southeastern part of France (box C), showing the smallest
384 spring NH$_3$ emissions. Finally, the contribution of fertilization on the emissions of the
385 southwestern part of France (box D) is strong.

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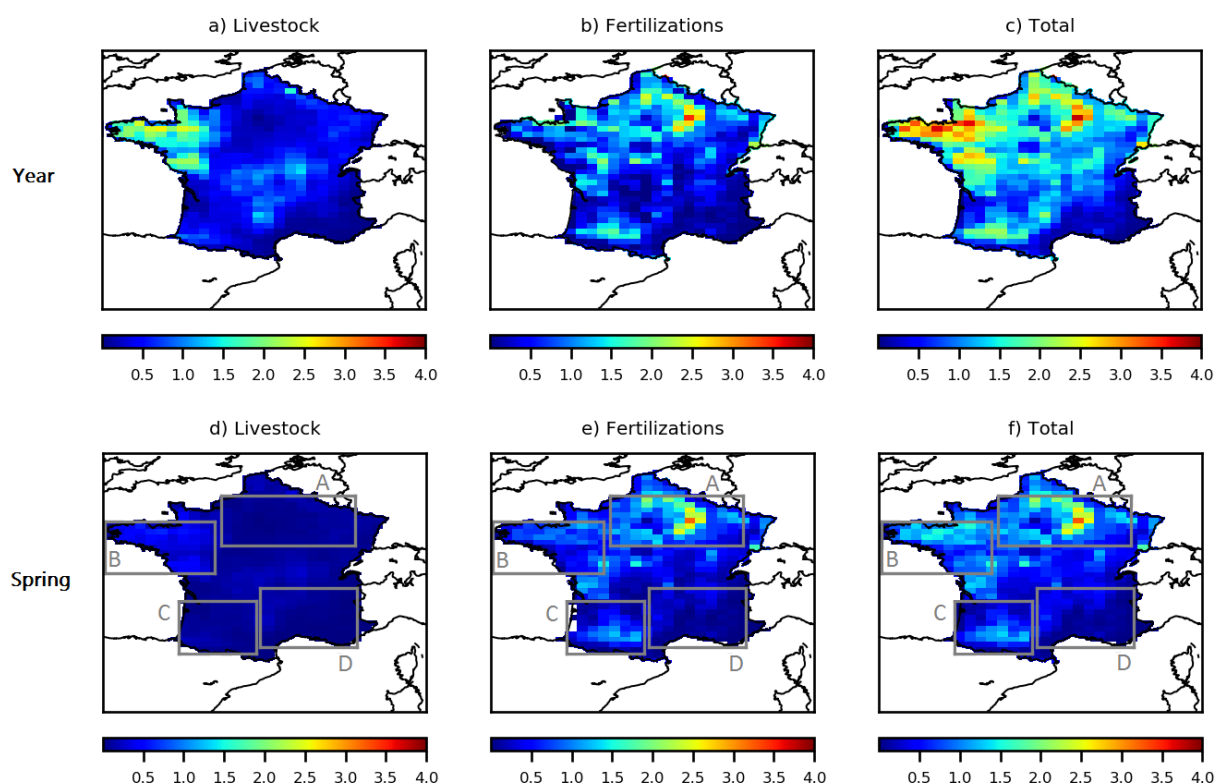
 Figure 3. *top) Yearly NH₃ emissions due to a) livestock husbandry and manure storage, b) N*

388 *fertilization (organic and mineral) and c) all sources in the CADASTRE-CIT inventory, in*

389 *ktNH₃. bottom) the same for spring NH₃ emissions, from March to May 2011.*

390

391 ⁠ ⁠ ⁠ ⁠ ⁠ ⁠ 3.2. French spring NH₃ total budget and its main spatial features

392 The spring NH₃ total budget is shown in Table 2 at the national scale and at the French regional

393 scale. The French spring ammonia budgets, calculated for the period from March to May 2011,

394 estimated by the NH₃SAT (264 ktNH₃) and the CADASTRE-CIT (354 ktNH₃) inventories are

395 both higher than the TNO-GEN reference one (234 ktNH₃).The CADASTRE-CIT inventory

396 estimates higher NH₃ spring emissions, by about 30%, than NH₃SAT.

397

398 The relative agreement on national budget between TNO-GEN and NH₃SAT must be nuanced

399 as total budget values from NH₃SAT and TNO-GEN are close but large differences in the

400 spatial distribution of the French NH₃ emissions between TNO-GEN and both NH₃SAT and

401 CADASTRE-CIT can be observed (Figure 4).

402

	TNO-GEN (in ktNH₃)	NH₃SAT (in ktNH₃)	CADASTRE-CIT (in ktNH₃)	Contribution of the fertilization to the spring budget in CADASTRE-CIT (in %)
Regions in the northeastern part of France (box A in Figure 2)				
Champagne-Ardennes	**8**	**13 (+63%)**	**35 (+337%)**	99
Centre	**13**	**15 (+15%)**	**32 (+146%)**	81
Lorraine	**9**	**12 (+33%)**	**17 (+88%)**	80
Picardie	**8**	**12 (+50%)**	**30 (+275%)**	85
Haute-Normandie	**7**	**9 (+28%)**	**12 (+71%)**	75
Ile de France	**3**	**4 (+33%)**	**9 (+200%)**	77
Regions in the northwestern part of France (box B in Figure 2)				
Bretagne	34	34 (=)	30 (-12%)	61
Pays de la Loire	**25**	**28 (+12%)**	**29 (+16%)**	45
Regions in the southeastern part of France (box C in Figure 2)				
Rhône-Alpes	13	14 (+8%)	12 (-8%)	37
Auvergne	12	12 (=)	11 (-8%)	41
Languedoc	5	6 (+20%)	3 (-40%)	53
PACA	5	5 (=)	2 (-60%)	47

Regions in the southwestern part of France (box D in Figure 2)				
Midi-Pyrénées	18	18 (=)	26 (+44%)	60
Aquitaine	12	11 (-8%)	15 (+25%)	50
Other regions				
Alsace	**4**	**5 (+25%)**	**7 (+175%)**	87
Basse-Normandie	**11**	**15 (+36%)**	**15 (+36%)**	68
Bourgogne	**11**	**13 (+18%)**	**20 (+81%)**	55
Franche-Comté	6	6 (=)	8 (+33%)	43
Limousin	8	8 (=)	6 (-25%)	21
Nord-Pas-de-Calais	**8**	**11 (+38%)**	**10 (+25%)**	97
Poitou-Charentes	14	14 (=)	25 (+79%)	63
France				
	234	264 (+13%)	354 (+51%)	67

403 **Table 2**. *French national and regional budgets of NH₃spring emissions, from March to May*
404 *2011, in ktNH₃. The relative differences compared to the TNO-GEN are presented between*
405 *brackets, in %.* **Regions for which the inventories NH₃SAT and CADASTRE-CIT propose**
406 **the same sign of relative differences are marked in bold**. *The contributions of the fertilization*
407 *emissions to the NH₃regional spring budget in the CADASTRE-CIT inventory are shown in %.*

408 Indeed, over France and for the spring budget, the spatial correlation compared to the
409 CADASTRE-CIT inventory, which should better represent the agricultural practices and their
410 spatial distribution, is improved when using the NH₃SAT inventory instead of using TNO-GEN
411 (Pearson correlation coefficient r=0.78 with NH₃SAT against r=0.72 with TNO-GEN).

412
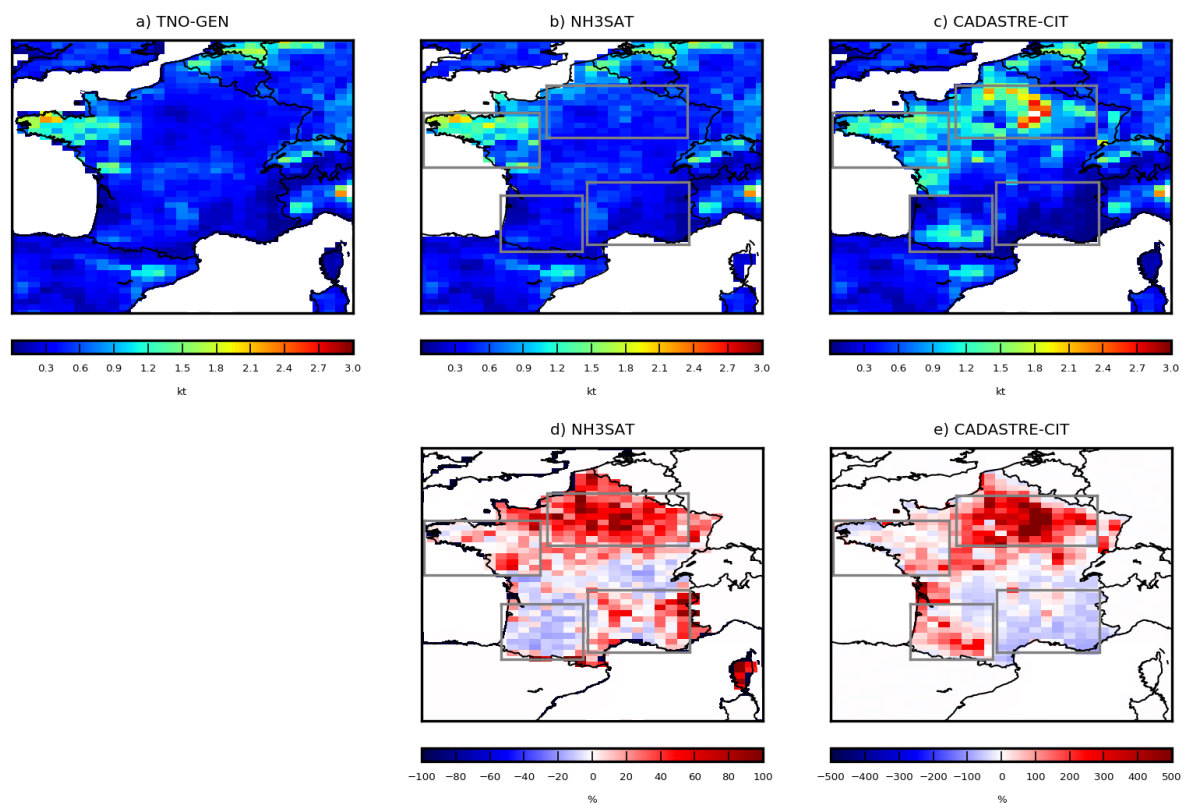
Figure 4. *French NH₃ spring emissions estimated by a) the TNO-GEN inventory, b) the NH3SAT inventory and c) the CADASTRE-CIT inventory, in ktNH₃. Relative differences of d) the NH₃SAT, and e) the CADASTRE-CIT inventories compared to the TNO-GEN inventory, in %, at the pixel resolution, for the period from March to May 2011. Note that the scale is different between 4b) and 4c).*

The northeastern part of France presents the largest difference with the TNO-GEN inventory (48 ktNH₃) for both NH₃SAT and CADASTRE-CIT inventories (65 and 135 ktNH₃, respectively). **The high emissions in the northeastern part of France are in agreement with the MASAGE_NH₃ inventory [Paulot et al., 2014], the magnitude of their annual NH₃ emissions from mineral fertilizer being calculated by combining an inventory of crop acreages, crop- and country-specific fertilizer application rates and fertilizer-, crop-, and application-specific emission factors.** Emissions are higher for both NH₃SAT and CADASTRE-CIT inventories compared to TNO-GEN over the Champagne-Ardennes (+337% and +63%, respectively for CADASTRE-CIT and NH₃SAT, Table 3), Picardie (+275%, and +50%, respectively), Centre (+146% and +15%, respectively), Haute-Normandie (+71% and +28%, respectively), Lorraine (+88%, and +33%, respectively) and Ile de France regions (+200%, and +33%, respectively).

The northwestern part of France presents the largest NH₃ emissions according to the TNO-GEN inventory (Figure 4a). The TNO-GEN, NH₃SAT and CADASTRE-CIT inventories lead to similar spring budget (68, 73 and 71 ktNH₃, respectively), over this domain.

Over the southeastern part of France, CADASTRE-CIT is about 23% lower than NH₃SAT (28 and 37 ktNH₃, respectively, Table 2). One hypothesis to explain the lower NH₃ emissions in CADASTRE-CIT is that market gardening is important in this area and not taken into account in the CADASTRE-CIT inventory [Ramanantenasoa et al., 2018; Génermont et al., 2018]. Nevertheless, market gardening is not included, to our knowledge, in the TNO-GEN inventory. TNO-GEN and NH₃SAT inventories being in quite good agreement in terms of budget (35 and 37 ktNH₃, respectively, Table 2), further work is required to understand these discrepancies.

Finally, over the southwestern part of France (box D in Figure 4), IASI observations only trigger slight corrections to the TNO-GEN inventory over this area (29 and 30 ktNH₃, respectively) and CADASTRE-CIT is 36% higher than TNO-GEN and NH₃SAT (Table 2).

445   3.3. Temporal variability of the NH_3 emissions at the sub-seasonal scale

446 Monthly regional budgets have been calculated for the three inventories. Figure 5 presents the

447 monthly variability of the NH_3 emissions from February to May 2011 for the four domains of

448 interest presented above. February is only displayed here as a baseline to show the sharp peak

449 of NH_3 emissions in March over some domains. The contribution of the emissions due to

450 livestock in the CADASTRE-CIT spring budget is also given.

451

452 The TNO-GEN inventory shows rather similar NH_3 emissions from March to May for all

453 regions, with a slight maximum in April (Figure 5a), imposed by the used GENEMIS monthly

454 profiles for the temporalization of emissions [Ebel et al., 1997]. On the contrary, over the

455 northeastern part of France, both NH_3SAT and CADASTRE-CIT inventories show a maximum

456 in March, and a decrease until May by about a factor of 1.5 to 2 (Figure 5a). **This maximum**

457 **in March is also noticed by Tournadre et al. [2020], providing nine years of total column**

458 **observations from ground-based infrared remote sensing over the Paris megacity**. We

459 calculated the monthly contribution of livestock to the NH_3 emissions based on CADASTRE-

460 CIT, which allows one to separate this contribution from the fertilization one. As in Figure 3,

461 Figure 5a confirms that NH_3 emissions are mainly due to fertilization in the northeastern part

462 of France, and shows that the seasonal variation is mainly driven by this contribution.

463 CADASTRE-CIT shows larger values than NH_3SAT which might be partly due to a possible

464 low bias in the IASI observations [Dammers et al., 2017; **Tournadre et al., 2020**] combined to

465 the fact that the TNO-GEN inventory (negatively biased compared to CADASTRE-CIT) is

466 used as a priori for the mass balance approach with no correction applied to the a priori when

467 IASI observations are not available.

468 To go further in the comparison, we analyzed the daily variability of the NH_3SAT and

469 CADASTRE-CIT inventories -TNO-GEN representing no daily variability (Figure 6). To

470 interpret the results, some limitations of the inventories have to be considered. For NH_3SAT,

471 the corrections applied to the TNO-GEN emissions are only applied for clear-sky conditions

472 when IASI observations are available. In the CADASTRE-CIT, fertilization days are randomly

473 selected within two-week intervals of application extracted from the farm survey analysis

474 [Ramanantenasoa et al., 2018], thus the actual day of fertilization is unknown. However, the

475 NH_3 volatilization is continuous over several days after spreading, reducing the uncertainty

476 introduced by this random selection. Moreover, the random selection is made at the field scale

477 (see section 2.3), then spatially averaging at the CHIMERE resolution should also smoothed

478   the random selection effect. Both effects of weather conditions on fertilization timing, on the

479   one hand, and on volatilization intensity and dynamics at the time of application and after

480   application, on the other hand, are realistically produced. Hence, the NH_3 volatilization is

481   continuous over several days after spreading introducing additional smoothing.

482   The CADASTRE-CIT inventory presents a high day-to-day variability from March to May

483   2011 (Figure 6a) with several strong maxima of emissions - characteristic of emissions due to

484   fertilizer application and a significant effect of the varying meteorological conditions. Also for

485   NH_3SAT, day-to-day variability is large. However, NH_3SAT and CADASTRE-CIT maxima

486   are not very well correlated. Over the first 15-days in March 2011, the high emissions occurring

487   from the 9th to 16th in CADASTRE-CIT are not reproduced by the NH_3SAT inventory,

488   potentially because of a lack of IASI coverage for this period (only about 40% of the domain –

489   Figure 6a). NH_3SAT shows an emission maximum one week earlier from the 1st to 7th. This

490   time gap in emission maxima could be explained by the random selection of fertilization days

491   in CADASTRE-CIT. Over the last 15 days in March 2011 and over the first 15 days in April

492   2011, the maxima of emissions estimated by NH_3SAT and CADASTRE-CIT are more

493   correlated, (e.g., from 7th to 11th April in CADASTRE-CIT versus from 9th to 11th in NH_3SAT).

494   Finally, CADASTRE-CIT still shows high emissions in the last 15 days of April and in May

495   2011, particularly related to the use of fertilizer over corn. Despite a good coverage with IASI

496   observations, no specific high emissions are derived from IASI for the same periods.

497   Over the northwestern part of France, the CADASTRE-CIT inventory is in agreement with

498   TNO-GEN, with a slight maximum in April (Figure 5b). The NH_3SAT inventory shows a

499   maximum in March, related to the high emissions seen during the first week of March 2011

500   (Figure 6b). However, the monthly differences are much smaller than for the northeastern part

501   of France. Also for this domain, the day-to-day variabilities provided by CADASTRE-CIT and

502   NH_3SAT are mostly uncorrelated: a strong sharp maximum of emissions during the first week

503   of March 2011 is seen in NH_3SAT -but not reproduced in CADASTRE-CIT (Figure 6b). The

504   highest daily emissions in CADASTRE-CIT occur during the two last weeks of April 2011 and

505   are not reproduced by NH_3SAT.

506

507   Over the southeastern and southwestern parts of France, month-to-month variations and

508   emission amounts are much smaller than for the two previous domains (Figures5-6c and

509   Figures5-6d, respectively). In the southeastern part of France, NH_3SAT emissions are slightly

510   larger than TNO-GEN and CADASTRE-CIT at the beginning of the period (February and

511   March, Figure 5c). The fertilization contribution to CADASTRE-CIT emissions slightly

512   decreases at the end of the period. In the southwestern part of France, NH_3SAT and TNO-GEN

513   are very similar (Figure 5d) and CADASTRE-CIT is slightly larger at the end of spring. This

514   increase is mainly related to fertilization emissions, the livestock contribution being stable.

515   Again, daily time series between both NH_3SAT and CADASTRE-CIT inventories are

516   uncorrelated (Figures 6c, d).

517

518

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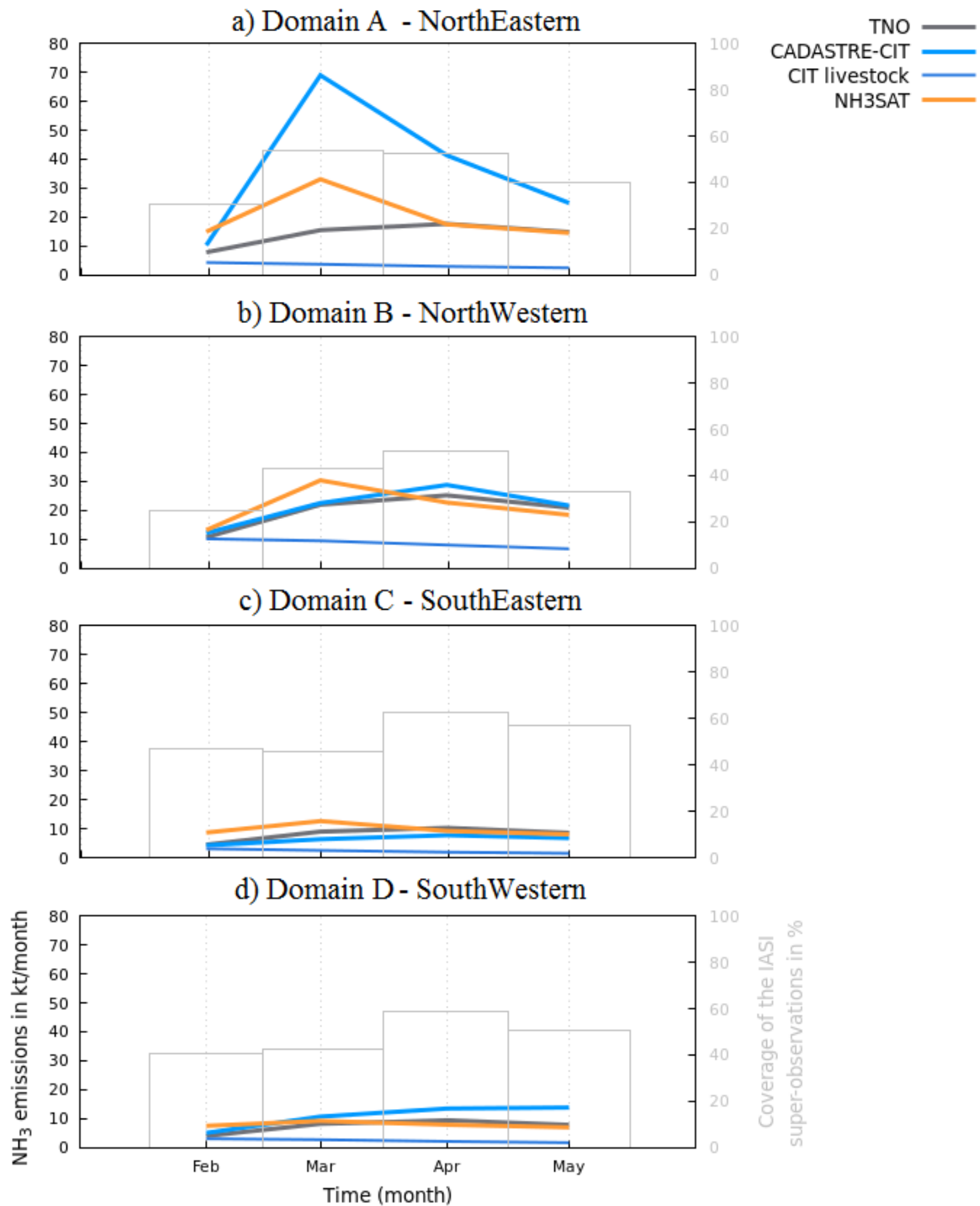
Figure 5. *Time series of monthly NH$_3$ emissions estimated by TNO-GEN (in **black**), by NH$_3$SAT (in orange), and by CADASTRE-CIT (in blue) inventories, from February to May 2011, for (a) domain A (northeastern), (b) domain B (northwestern), (c) domain C (southeastern), and (d) domain D (southwestern), as defined in Figure 2. The contribution of the emissions due to livestock in the CADASTRE-CIT monthly budgets is also given. Units are ktNH$_3$/month. The monthly regional coverage of the IASI super-observations is given in % (in grey).*
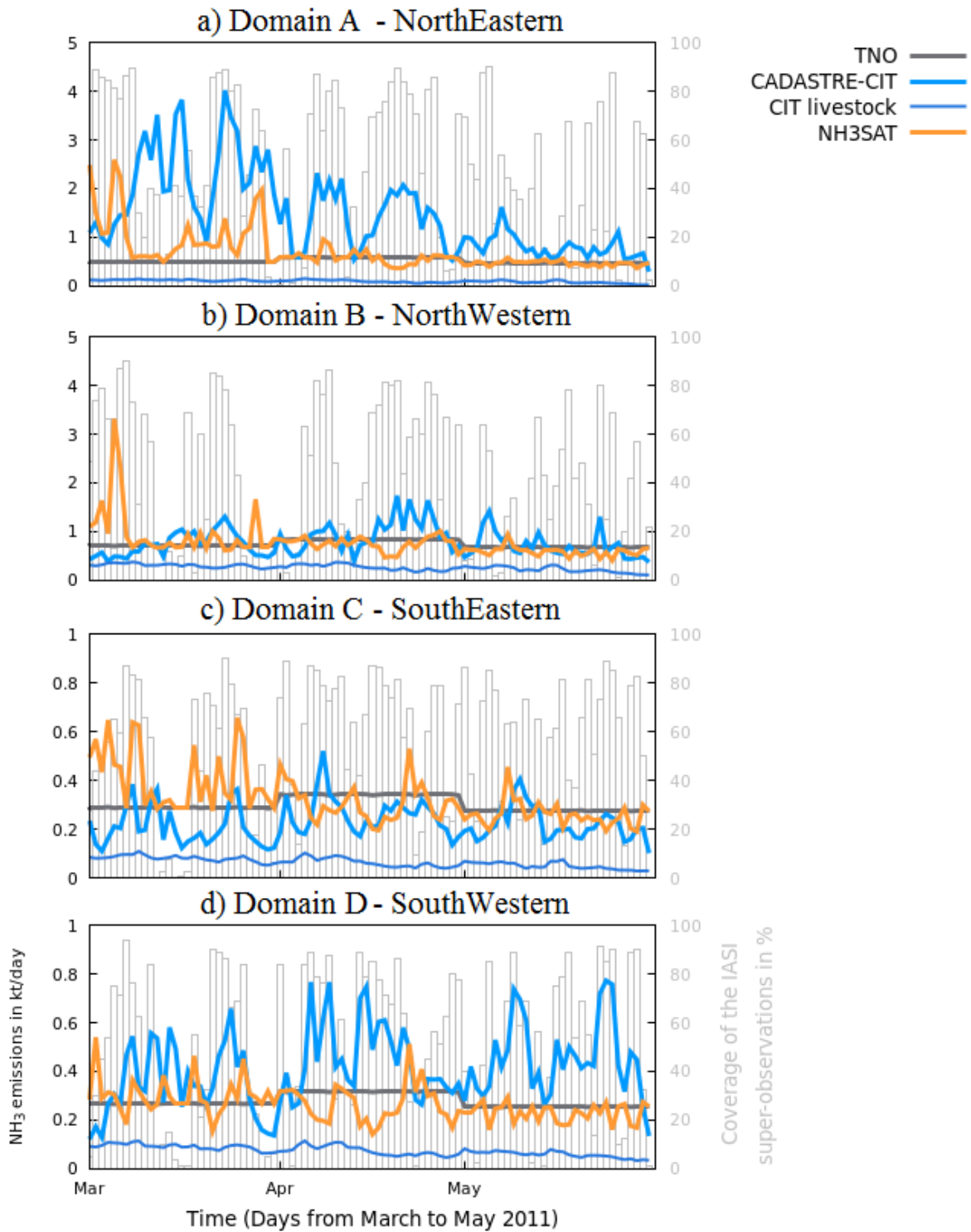
Figure 6. *Time series of daily NH₃ emissions estimated by TNO-GEN (in **black**), by NH₃SAT (in orange), and by CADASTRE-CIT (in blue) inventories, from March to May2011, for (a) domain A (northeastern), (b) domain B (northwestern),(c) domain C (southeastern), and (d) domain D (southwestern), as defined in Figure 2. Note that the scale is different between a),b) and c),d). Units are ktNH₃/day. The contribution of the emissions due to livestock in the*

532 *CADASTRE-CIT daily variability is also given. The daily regional coverage of the IASI super-*

533 *observations is given in % (in grey).*

534 **Conclusion**

535 In this study, we performed an inter-comparison of two alternative inventories with the TNO-

536 GEN reference inventory that quantify the French NH_3 emissions during spring 2011. One of

537 the main conclusions of this study is that over regions with large mineral fertilizer use, like over

538 the northeastern part of France, induced NH_3 emissions are probably considerably

539 underestimated by the TNO-GEN reference inventory, as both the NH_3SAT (constrained by

540 IASI observations) and CADASTRE-CIT (process level oriented), show much larger values.

541 For instance, over northeastern France, NH_3SAT and CADASTRE-CIT show respectively a

542 factor 1.4 and 2.8 larger spring 2011 emissions than TNO-GEN. Over the whole France, NH_3

543 emissions are still more than 50% larger in CADASTRE-CIT than in TNO-GEN. Average

544 French NH_3SAT emissions are about 10% larger than TNO-GEN ones. Over the southern part

545 of France, with a more diverse agriculture as compared to the crop intensive one in northeastern

546 France, differences between the inventories are on the whole lower, and signs between

547 CADASTRE-CIT / TNO-GEN and NH_3SAT / TNO-GEN corrections are often opposite for

548 different regions.

549

550 Month-to-month variations are again much more pronounced over northeastern France and

551 show a maximum in March for both CADASTRE-CIT and NH_3SAT. Day-to-day variations are

552 large in CADASTRE-CIT and NH_3SAT, roughly a factor of 5 between minimal and maximal

553 values. This shows the interest in evaluating NH_3 emissions at a daily scale because this input

554 is required for chemistry transport modeling of particulate matter formation and thus air quality.

555 However, time-series delivered by CADASTRE-CIT and NH_3SAT are uncorrelated for all

556 considered regions. This result can be partly explained by the lack in IASI NH_3 column

557 observations under partially cloudy conditions, and by the fact that available information on

558 agricultural practices is resolved at a two weeks scale.

559

560 **Yet, current results of our study have important implications for air quality modelling**

561 **over Europe. The important changes in the spatial distribution of NH_3 emissions as a**

562 **function of soil properties are of general concern not only for France, but for whole**

563 **Europe. Soils are alkaline or neutral (pH>6) not only over North-Eastern France, but also**

564 **over large parts of Italy, eastern Spain, or eastern Germany [Reuter, 2008]. Over these**

regions, our study suggests potentially larger NH₃ emissions than with a constant emission factor treatment, with impacts then on fine particle formation. These features should be included in "operational" emission inventories used for air quality modelling.

Thus, as a general conclusion, use of observational constrained or process-oriented emission inventories is clearly of added-value for estimating better monthly averages over French areas with intensive mineral fertilizer use, but capacity for delivering day-to-day variations is not yet proven. This warrants further studies, both refining hypotheses on days chosen by farmers for fertilizer spread out as a function of meteorological conditions, and, acquiring and using continuous surface NH₃ measurements for validating satellite or process derived NH₃ emission variability.

Competing interests

The authors declare that they have no conflict of interest.

Author Contribution

All authors have contributed to the manuscript writing (main authors A.F-C, G.D, S.G and MB). A.F-C has performed the mass-balance approach to deduce NH₃ emissions from NH₃ total columns observed by the IASI satellite instrument. K.D, J-M.G and S.G have performed the bottom-up approach providing the CADASTRE_NH₃ inventory for organic and mineral fertilization practices. F.C has compiled this CADASTRE_NH₃ inventory with livestock emissions from the CITEPA. L.C, P-F.C, M.V.D and C.C are the PIs of the IASI NH₃ product. All authors discussed the results and contributed to the final paper.

Code and Data Availability

The CHIMERE code is available here: www.lmd.polytechnique.fr/chimere/.

The IASI ANNI-NH3-v2.2R data are freely available through the AERIS database: https://iasi.aeris-data.fr/nh3/.

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