Response to anonymous referee #1

2 General comments:

1

This paper describes the measurement of higher alkanes by NO⁺ PTR-ToF-MS at two sites in China and the authors showed the significant contribution of the higher alkanes to secondary organic aerosol (SOA) formation. I recognize that the measurement of the higher alkanes in ambient air is very challenging, so the data presented here would be precious. But I feel that there are two issues in the present paper.

8 Response: Thank you very much for your valuable comments and suggestions on our
9 manuscript. These comments are all helpful for improving our article. All the authors have
10 seriously discussed about all these comments. According to your comments, we have tried best
11 to modify our manuscript to meet with the requirements for the publication in this journal.
12 Point-by-point responses to the comments are listed below.

Firstly, the authors seemed to fail to suppress the formation of O_2^+ in the ion source (Figure S2). I think that the O_2^+ reaction could substantially interfere with the ion signals of alkanes. In addition, I feel that the ratio of O_2^+ to NO^+ was not stable as far as I looked at Figure S2. The authors should pay attention to the intensity of O_2^+ relative to the NO⁺ intensity.

Response: In the submitted manuscript, we made a mistake on the Fig. S2, where we put a 18 wrong number of Us=40 V (in fact Us=120 V) on the Fig. S2(a). I think that might mislead 19 you into thinking that we faild to suppress the formation of O_2^+ . The corrected data and more 20 21 experiment results were included in the revised Fig. S2. Before field campaigns, we did laboratory tests on the ion source voltages to find suitable volatage settings, where intensities 22 of ion impurites $(O_2^+, H_3O^+ \text{ and } NO_2^+)$ are low. According to the laboratory results, we chose 23 Us=40 V and Uso=100 V as ion source voltage settings for the field measurement of higher 24 alkanes (Fig. S2(a)), under which condition (absolute humidity=~11.6 mmol/mol) the ratio of 25 O_2^+ to NO⁺ is ~10%. 26

We also included the data of NO⁺, O_2^+ , H_3O^+ and NO_2^+ during field campaigns in the revised supporting information. The intensities of primary ion NO⁺ and impurites (O_2^+ , H_3O^+ and NO_2^+) (**Fig. S3**) and the ratio of O_2^+ to NO⁺ (**Fig. S4**) fluctuate during two campaigns. The ratio of O_2^+/NO^+ (**Fig. S4(a**)) is basically stable at 2-4% during the PRD campaign except a little high values of 7-10% for Oct. 26-Nov. 2, 2018. While, for the NCP campaign, the ratio of O_2^+/NO^+ (**Fig. S4(b**)) fluctuates between 10-40% in the early stage of campaign and keeps stable at ~20% in the late stage of campaign. When comparing the O_2^+/NO^+ ratio with the ambient abosolute humidity during campaigns (**Fig. S5**), we find that there was an obvious negative correlation between O_2^+/NO^+ ratio and ambient abosolute humidity (**Fig. S5**), which is also consistent with laboratory humidity experiments. As the result, the fluctuations of O_2^+/NO^+ ratios are attributed to the changes of humidity during the two campaigns.

We agree with the comment that O_2^+ could interfere with the ion signals of alkanes. 38 According to the study from Amador et al. (2016), the reactions of O_2^+ with alkanes can be 39 proceeded by charge transfer and hydride abstraction that might affect the ion signals of alkanes 40 with NO⁺ reactions in our study. As we got the alkanes standard at the very late period of the 41 PRD campaign, we did not have the daily calbirations for this campaign. However, the 42 O_2^+/NO^+ ratios were small with a genaral range of 2-4% (Fig. S4(a)), we think the interfence 43 of O_2^+ in this campaign is negligible. During the NCP camaign, we conducted calibrations 44 every 1-2 days under both dry conditions and ambient humidity conditions. In the revised 45 manuscript, we add a figure (Fig. 2) to show the correlation of sensitivities of *n*-alkanes and 46 O_2^+/NO^+ ratios. As showned in Fig. 2, the sensitivities of *n*-alkanes (C8-C15) varied 47 significantly with the fluctuations of O_2^+/NO^+ ratios in both dry conditions and ambient 48 conditions during the NCP campaign. This indicate that the interference of O_2^+ should be 49 considered. Therefore, we use the daily ambient calibrations results to quantify the 50 51 concentration of higher alkanes for the NCP campaign.

52



54

Figure S2. The variations of NO⁺, H_3O^+ , O_2^+ and NO_2^+ ions on the voltages of ion source (Us and Uso) for NO⁺ PTR-ToF-MS. For each experiment, either Us or Uso is fixed at a voltage and the other was varied to explore the best setting for NO⁺ PTR-ToF-MS. For example, test #1 in (a), we fix Us at 40 V and change Uso from 20 V to 180 V. The dashed line in (a) indicate the setting point in this study (Us=40 V and Uso=100 V).



Figure S3. Time series of NO⁺, H₃O⁺, O₂⁺, NO₂⁺ during the PRD (a) and the NCP (b)
campaigns, respectively.



Figure S4. Time series of O₂⁺ to NO⁺ ratios and absolute humidity during the PRD (a) and
the NCP (b) campaigns, respectively.



Figure S5. Humidity dependence of O_2^+ to NO^+ ratios during the lab experiment and the two

74 field campaigns.



Figure 2. The relationship of sensitivities of *n*-alkanes (C8-C15) versus O_2^+/NO^+ ratios during the NCP campaign. The data points are color-coded using absolute humidity during the calibration.

In the revised manuscript, part of the section of "2.1 NO⁺ PTR-ToF-MS measurement" are
modified as follows:

Line 122-131 on Page 6: "Voltages of ion source and drift chamber were explored 81 (Figure S2) in the laboratory to optimize the generation of NO⁺ ions relative to H_3O^+ , O_2^+ , 82 and NO₂⁺ and minimize alkane fragmentation. The intensities of primary ion NO⁺ and 83 impurites (O₂⁺, H₃O⁺ and NO₂⁺) and the ratio of O₂⁺ to NO⁺ during two campaigns are 84 shown in Figure S3 and Figure S4, respectively. The ratio of O₂⁺/NO⁺ (Figure S4 (a)) is 85 basically stable at 2-4% during the PRD campaign except during Oct. 26-Nov. 2, 2018 (7-86 10%). For the NCP campaign, the ratio of O₂⁺/NO⁺ (Figure S4 (b)) fluctuates between 10-87 40% in the early stage of campaign and keeps stable at ~20% in the later stage of the 88 campaign. Such fluctuations are atrributed to the humidity effect in the ambient air 89 (Figure S5)." 90

Line 147-161 on Page 7: "Sensitivities of *n*-alkanes (C8-C15) standards were 91 obtained during the campaign (Figure S6), which is defined as the normalized signal of 92 hydride abstraction ions for each higher alkane at 1 ppbv with a unit of ncps/ppb. The 93 fluctuations of sensitivities during the NCP campaign may be influenced by the variations 94 of O_2^+ signals (Figure 2), because the reactions of O_2^+ with alkanes can be proceeded by 95 both charge transfer and hydride abstraction (Amador et al., 2016) that may affect the 96 ion signals of alkanes with NO⁺ reactions. Therefore, we use the daily ambient 97 calibrations results to quantify the concentration of higher alkanes during the NCP 98 campaign to reflect the variations of sensitivity from day to day. For the measurements 99 100 without daily calibrations, we used closest calibration results according to corresponding ambient O₂⁺/NO⁺ ratios and ambient humidity. Since we got the alkanes standard at the 101 very late period of the PRD campaign, we did not have the daily calibrations for this 102 campaign. Therefore, we use the sensitivity of each alkane under corresponding O₂⁺/NO⁺ 103 condition obtained from lab experiments after this campaign and also consider the 104 humidity effects (Figure 3(b, c)) to quantify the concentration of higher alkanes during 105 the PRD campaign." 106

107 Secondly, I feel that the diurnal variation of the OH exposure derived from isoprene 108 chemistry method is strange (Figure S9). I cannot understand why the value is low 109 during daytime compared with that in nighttime. The authors indicated two reference 110 (Apel et al., 2002; Carlton et al., 2009), but I could not find the derivation of the OH exposure by "the isoprene chemistry method" in the references. The authors should show how the OH exposure by the isoprene chemistry method was calculated. I think that the low OH exposure during the daytime causes the low SOA formation from isoprenoids during the daytime, as shown in Figure 8(c).

Response: The isoprenoids (i.e. isoprene and monoterpenes in this study) are dominantly 115 emitted from biogenic sources during the PRD campaign and reach the maximam concentration 116 during daytime. These species have less photochemical oxidation time during daytime than 117 that in nighttime. The lower OH exposure of the isoprenoids during daytime than that of 118 nighttime is as a result of much fresh emission during daytime and the aging time is smaller 119 than that of nighttime. We calculated the OH exposure of isoprenoids based on isoprene 120 chemistry for the PRD campaign. The calculation method can be found in Roberts et al. (2006). 121 We will also add a short introduction about the calculation in the supporting information of the 122 revised manuscript. 123

124 Isoprene are mainly photo-oxidized through the reactions with OH radical in the 125 atmosphere and its primary first-generation reaction products are formaldehyde, MVK 126 and MACR (Apel et al., 2002). The reaction processes of isoprene oxidized by OH radical 127 are mainly as follows:

128 Isoprene + $0H \rightarrow 0.63HCHO + 0.32MVK + 0.23MACR$ $k_1=1.0\times10^{-10} cm^{-3} s^{-1}$ (Eq. 129 S1)

130
$$MVK + OH \rightarrow Products$$
 $k_2=1.9\times10^{-11} \text{ cm}^{-3} \text{ s}^{-1}$ (Eq. S2)

131
$$MACR + OH \rightarrow Products \quad k_3 = 3.3 \times 10^{-11} cm^{-3} s^{-1} (Eq. S3)$$

132 where k₁, k₂, k₃ are the rate constants of the reactions. According to above reactions, the 133 relationship between MVK/Isoprene, MACR/Isoprene, (MVK+MACR)/Isoprene and the 134 reaction time Δt can be decribed as follows (Apel et al., 2002):

135
$$\frac{MVK}{Isoprene} = \frac{0.32k_1}{k_2 - k_1} \left(1 - \exp((k_1 - k_2) [OH] \Delta t) \right)$$
 (Eq. S4)

136
$$\frac{MACR}{Isoprene} = \frac{0.23k_1}{k_3 - k_1} \left(1 - \exp((k_1 - k_3) [OH] \Delta t) \right) \quad (Eq. S5)$$

137
$$\frac{MVK+MACR}{Isoprene} = \frac{0.32k_1}{k_2 - k_1} \left(1 - \exp((k_1 - k_2) [OH] \Delta t)) + \frac{0.23k_1}{k_3 - k_1} \left(1 - \exp((k_1 - k_3) [OH] \Delta t)\right) \quad (Eq. S6)$$

138 where $\frac{MVK+MACR}{Isoprene}$ can be derived from the measurements by PTR-ToF-MS. Then the OH 139 exposure ([OH] Δt) of isoprenoids can be obtaind from Eq. S6. SOA production of 140 isoprenoids can be calculated from the following equation.

141
$$[SOA_i]_t = [VOC_i]_t \times (e^{k_{VOC_i} \times ([OH] \times \Delta t)} - 1) \times Yield_i \quad (Eq. S7)$$

142 Where $[SOA_i]_t$ is the calculated SOA production (μ g m⁻³) for a given specific compound VOC_i

143 at time t, $[VOC_i]_t$ is the VOC_i concentration measured at time t (µg m⁻³), Yield_i is the SOA

144 yield data summarized from chamber studies, k_{VOC_i} is the rate constant of VOC_i with the OH

radical (cm³ molecule⁻¹ s⁻¹). $[OH] \times \Delta t$ is the OH exposure (molecules cm⁻³ s). Therefore, the

146 diurnal variations of SOA formation are as result of the combing effects of concentration, the

rate constant with OH radical, OH exposure and SOA yield for each compound.

148 Specific comments:

(1) Page 6, Line 113: In Figure S2, the authors showed that the relative intensity of O_2^+ to that of NO⁺ was ~50 % (Fig. S2(a)) and ~15 % (Fig. S2(b)) under the conditions of Us = 40 V and Uso = 120V. These conditions are not good because I think that the O_2^+ reaction could substantially interfere with the ion signals of alkanes. I am concerned that the ratio of O_2^+ to NO⁺ was not stable. The authors should mention the quality of the data of the alkanes presented in this paper.

Response: In the originally submitted manuscript, we made a mistake on the Figure S2. The
corrected one is shown as follows. Based on the laboratory results, we chose Us=40 V and
Uso=100 V as ion source voltage settings for the field measurement of higher alkanes (Fig.

158 S2(a)), under which condition (absolute humidity= ~ 11.6 mmol/mol) the ratio of O₂⁺ to NO⁺ is

159 $\sim 10\%$. The more details about the O₂⁺ interference are replied above on pages 1-8.



160

Figure S2. The variations of NO⁺, H_3O^+ , O_2^+ and NO_2^+ ions on the voltages of ion source (Us and Uso) for NO⁺ PTR-ToF-MS. For each experiment, either Us or Uso is fixed at a voltage and the other was varied to explore the best setting for NO⁺ PTR-ToF-MS. For example, test #1 in (a), we fix Us at 40 V and change Uso from 20 V to 180 V. The dashed line in (a) indicate the setting point in this study (Us=40 V and Uso=100 V).

(2) Page 6, Lines 130-131: I cannot agree to the argument by the authors that calibration factors 166 during the campaigns. because the normalized sensitivities of 167 were stable the n-C15 alkane were scattered with a factor of 2. As far as I experienced, the normalized 168 sensitivity is very stable. I feel that this instability of the normalized sensitivity 169 170 may be related to the interference of the O_2^+ reaction.

171 Response: We agree with that the instability of the normalized sensitivity might be related to

the interference of O_2^+ reaction. As showned in **Fig. 2**, the sensitivities of *n*-alkanes (C8-C15)

- varied significantly with the fluctuations of O_2^+/NO^+ ratios in both dry conditions and ambient
- 174 conditions during the NCP campaign. This indicate that the interference of O_2^+ should be
- 175 considered. In the revised manuscript, the Fig. 2 was added to explain the fluctuations of

176 sensitivities of *n*-alkanes. During the NCP camaign, we conducted calibrations every 1-2 days 177 under both dry conditions and ambient humidity conditions. The daily ambient calibrations 178 results were used to quantify the concentration of higher alkanes for the NCP campaign. For 179 the measurements without daily calibrations, we used the closest calibration results according 180 to the corresponding ambient O_2^+/NO^+ ratios and the ambient humidity.

181 In the revised manuscript, part of the section of "2.1 NO⁺ PTR-ToF-MS measurement"
182 are modified as follows:

Line 147-161 on Page 7: "Sensitivities of *n*-alkanes (C8-C15) standards were 183 obtained during the campaign (Figure S6), which is defined as the normalized signal of 184 hydride abstraction ions for each higher alkane at 1 ppbv with a unit of ncps/ppb. The 185 fluctuations of sensitivities during the NCP campaign may be influenced by the variations 186 of O_2^+ signals (Figure 2), because the reactions of O_2^+ with alkanes can be proceeded by 187 both charge transfer and hydride abstraction (Amador et al., 2016) that may affect the 188 ion signals of alkanes with NO⁺ reactions. Therefore, we use the daily ambient 189 calibrations results to quantify the concentration of higher alkanes during the NCP 190 campaign to reflect the variations of sensitivity from day to day. For the measurements 191 without daily calibrations, we used closest calibration results according to corresponding 192 ambient O₂⁺/NO⁺ ratios and ambient humidity. Since we got the alkanes standard at the 193 very late period of the PRD campaign, we did not have the daily calibrations for this 194 campaign. Therefore, we use the sensitivity of each alkane under corresponding O_2^+/NO^+ 195 condition obtained from lab experiments after this campaign and also consider the 196 humidity effects (Figure 3(b, c)) to quantify the concentration of higher alkanes during 197 198 the PRD campaign."



Figure 2. The relationship of sensitivities of *n*-alkanes (C8-C15) versus O_2^+/NO^+ ratios during the NCP campaign. The data points are color-coded using absolute humidity during the calibration.

(3) Page 7, Lines 137-139: When I looked at the mass spectra in Figure S4, I found that the signal intensities of 13C-isotopologue of [M-1]+ are low. For example, the signal at m/z 282 should be 20 % of the intensity at m/z 281 for n-Eicosane. I wonder if the quantitation was guaranteed or not in the present measurements using the NO⁺ PTR-ToF-MS instrument.

208 Response: We obtained the mass spectra (Fig. S4) and the fragmentation patterns (Fig. 3(a)) for each high alkane by introducing commercially acquired pure chemicals with NO⁺ PTR-209 210 ToF-MS measurement. The mass spectra shown in Fig. S4 represent the distributions of product ions of each higher alkane reacted with NO⁺. As shown in the attached figure below, 211 the isotopic signals (green lines) are subtracted when performing the high-resolution peak-212 fitting using Tofware software (Tofwerk AG). That is to say, the signals of masses shown in 213 214 Fig. S4 are the results after subtracting the isotopic signals during the high resolution peak fitting of the mass spectra. 215

The title of Fig. S4 was modified to "Mass spectra of the distributions of product

217 ions from *n*-Dodecane (a), *n*-Pentadecane (b) and *n*-Eicosane (c) with NO⁺ PTR-ToF-

218 MS. The signals of masses shown in the graph are the results after subtracting the

219 isotopic signals during the high resolution peak fitting of the mass spectra."



Figure. High-resolution (HR) peak-fitting to the averaged mass spectra during lab experiments for m/z 169, m/z 170 (a), m/z 211, m/z 212 (b) and m/z 281, m/z 282 (c), at which masses produced by dodecane ($C_{12}H_{25}^+$, $C_{12}H_{26}^+$), pentadecane ($C_{15}H_{31}^+$, $C_{15}H_{32}^+$) and eicosane ($C_{20}H_{41}^+$, $C_{20}H_{42}^+$) produced in NO⁺ PTR-ToF-MS.

(4) Page 9, Lines 187-190: Some are good, but some are not good. The authors should
mention the results correctly and explain the disagreement for some species. Were the
concentrations from PTR-ToF-MS in Figs. S6 and S7 calibrated or calculated?

Response: The concentrations mentioned in the Figs. S6 and S7 were all calibrated with the commercial gas standards. Benzene measurements in H_3O^+ chemistry show large difference with benzene measured from NO⁺ chemistry in the earlier period of PRD campaign (11 Sep.-14 Oct. 2018), but good agreement was obtained for the rest of measurement period.

- Considering good agreement of benzene between NO⁺ PTR-ToF-MS and GC-MS/FID, we
 used benzene data from NO⁺ measurement in this study.
- The sentences on lines 237-244 on page 11 are modified to:

236 "Overall, good agreements between PTR-ToF-MS (both H₃O⁺ and NO⁺ chemistry) and

237 GC-MS/FID are obtained for aromatics and oxygenated VOCs except benzene (Figure

238 S12, S13). Benzene measurements in H_3O^+ chemistry show large difference with benzene

239 measured from NO⁺ chemistry in the earlier period of PRD campaign (11 Sep.-14 Oct.

- 240 2018), but good agreement was obtained for the rest of measurement period. Considering
- 241 good agreement of benzene between NO⁺ PTR-ToF-MS and GC-MS/FID, we used
- 242 benzene data from NO⁺ measurement in this study."
- (5) Page 10, Lines 219-221: The degree of the chemical removal of hydrocarbons by
 the OH reaction was quite different at two sites, when we consider the concentration of
 OH (Figure S19). The authors should mention the results accurately.
- 246 Response: We modified these sentences to "Such diurnal patterns are consistent with other
- 247 primary VOCs species (e.g. aromatics). In PRD, the diurnal variations of higher alkanes

248 are as the result of faster chemical removal in the daytime and shallow boundary layer

249 heights at night. While, since the OH concentrations in NCP during winter are much

250 lower than that in PRD during autumn (Figure S16), the diurnal variations of higher

alkanes in NCP are mainly influenced by the boundary layer effect."

- (6) Page 10, Line 226: The data of naphthalenes were not shown anywhere in thispaper.
- **254** Response: Naphthalenes are measured with H_3O^+ PTR-ToF-MS in these two campaigns. The
- et al., 2020). We added the sentences on line 310-314 of page 14 and a figure of time series of

details about the H₃O⁺ PTR-ToF-MS measurement can be found in our companion paper (Wu

- 257 naphthalenes in the revised supporting information.
- 258 "The concentration data of higher alkanes, isoprenoids and benzene were taken
 259 from measurements of NO⁺ PTR-ToF-MS. The concentration data of naphthalenes
 260 (Figure S18) and monoaromatics except benzene were taken from measurements of H₃O⁺
- 261 PTR-ToF-MS. The detail about the H₃O⁺ PTR-ToF-MS measurements can be found in
- 262 <u>Wu et al. (2020)</u>"



264

Figure S18. Time series of naphthalene, methylnaphthalenes, dimethylnaphthalenes inPRD (a) and NCP (b), respectively.

268 (7) Page 11, Lines 239-242: I could not understand how the authors estimated the OH 269 exposure clearly. Which value was used as the ratio of m+p-xylene to ethylbenzene 270 at t = 0 (the emission ratio)? How was the OH exposure estimated by the oxidation 271 process of isoprene? Why was the OH exposure estimated by the oxidation process of 272 isoprene low during daytime compared with that in nighttime?

273 Response: We included more detail about the calculation of OH exposure in the revised supporting information (Appendix 4, SI). The initial emission ratios of m+p-274 275 xylene/ethylbenzene were estimated according to the correlation of m+p-xylene with ethylbenzene during campaigns. As shown in the attached figure, diurnal variation of m+p-276 277 xylene/ethylbenzene concentration ratios (a) and the correlation of m+p-xylene with ethylbenzene (b) were studied in the PRD campaign. Red dots are measured ratios during the 278 279 campaign. Blue line indicates hourly geometric average, and gray areas are geometric standard deviations. The dashed line in both graphs indicate the estimated initial emission ratio of m+p-280 xylene/ethylbenzene. Details can also be found in Fig. S11 from Wu et al. (2020). 281



Figure S11. Diurnal variations of m+p-xylene/ethylbenzene concentration ratios. Red dots are measured ratios during the PRD. Blue line indicates hourly geometric average, and gray areas are geometric standard deviations. (b) Correlation of m+p-xylene with ethylbenzene. The dashed lines in both graphs indicate the estimated initial emission ratio of m+pxylene/ethylbenzene. (Wu et al., 2020)

The OH exposure estimated by the oxidation process of isoprene are added in the revised supporting information as follows. Isoprene are mainly photo-oxidized through the reactions with OH radical in the atmosphere and its primary first-generation reaction products are formaldehyde, MVK and MACR (Apel et al., 2002). The reaction processes of isoprene oxidized by OH radical are mainly as follows:

293 Isoprene + $OH \rightarrow 0.63HCHO + 0.32MVK + 0.23MACR$ $k_1=1.0\times10^{-10} cm^{-3} s^{-1}$ (Eq. 294 S1)

295
$$MVK + OH \rightarrow Products \quad k_2 = 1.9 \times 10^{-11} \, cm^{-3} \, s^{-1} \, (Eq. \, S2)$$

282

296
$$MACR + OH \rightarrow Products \quad k_3 = 3.3 \times 10^{-1.1} cm^{-3} s^{-1} (Eq. S3)$$

297 where k_1 , k_2 , k_3 are the rate constants of the reactions. According to above reactions, the 298 relationship between MVK/Isoprene, MACR/Isoprene, (MVK+MACR)/Isoprene and the 299 reaction time Δt can be decribed as follows (Apel et al., 2002):

300
$$\frac{MVK}{Isoprene} = \frac{0.32k_1}{k_2 - k_1} \left(1 - \exp((k_1 - k_2) [OH] \Delta t) \right)$$
 (Eq. S4)

301
$$\frac{MACR}{Isoprene} = \frac{0.23k_1}{k_3 - k_1} \left(1 - \exp((k_1 - k_3) [OH] \Delta t) \right) \quad (Eq. S5)$$

302
$$\frac{MVK+MACR}{Isoprene} = \frac{0.32k_1}{k_2 - k_1} \left(1 - \exp((k_1 - k_2) [OH] \Delta t)) + \frac{0.23k_1}{k_3 - k_1} \left(1 - \exp((k_1 - k_3) [OH] \Delta t)\right) \quad (Eq. S6)$$

where $\frac{MVK+MACR}{Isoprene}$ can be derived from the measurements by PTR-ToF-MS. Then

304 the OH exposure ($[OH]\Delta t$) of isoprenoids can be obtaind from Eq. S6.

As shown in the attached figure below, isoprene is mainly emitted by vegetation in the 305 PRD campaign with the highest concentration around noon. The lower OH exposure during 306 daytime than that of nighttime is as a result of much fresh emission during daytime and the 307 aging time is smaller than that of nighttime. The photochemical age of the air mass after mixing 308 is the nonlinear addition of the photochemical age before mixing. Generally, the photochemical 309 age of the mixed air mass is closer to that of the fresh air mass (Parrish et al., 1992). The OH 310 exposure values in this study represent the average "photochemical age" that are calculated by 311 some properly chosen VOCs ratios. Parrish et al. (2007) compared the calculated 312 photochemical age of different NMHCs species with the transport time calculated by a back 313 trajectory model, and concluded that the calculated photochemical age could better describe 314 the average transport time of VOCs species by selecting suitable hydrocarbon species pairs. 315



316

Figure. Diurnal variations of isoprene during the 2018 PRD campaign. Blue lines and shaded
areas represent averages and standard deviations, respectively.

(8) Page 13, Lines 288-89: The diurnal variation of the estimated SOA produced from
isoprenoids shown in Figure 8(c) is strange to me. In addition, I cannot understand
the diurnal variations of the calculated SOA produced from higher alkanes, monoaromatics,
naphthalenes, and isoprenoids in Figure 8(d). Why were those concentrations

323 low during the daytime in spite of the fact that the production rates were calculated to 324 be positive during the daytime (Figure S18(d)). Did the authors consider the dynamics 325 (e.g., the movement of the boundary layer) into the calculation? If so, explain in the 326 text.

327 Response: The SOA productions in the Figure 8 are calculated based on the following formula:

328
$$[SOA_i]_t = [VOC_i]_t \times (e^{k_{VOC_i} \times ([OH] \times \Delta t)} - 1) \times Yield_i$$

This means the SOA production for each compound is determined by the concentration, OH exposure, and rate constant with OH and yield the SOA yield. In the Figure 8(c), the SOA productions of isoprenoids present the sum of that from isoprene and monoterpenes. As shown in the attached figure below, the diurnal variations of the SOA production from isoprene exhibit a high level during the daytime and a low level during nighttime. While, the diurnal pattern of isoprenoids which is shown in Figure 8(c) is mainly influenced by that of monoterpenes with a relatively low level during the daytime.



336

Figure. Diurnal variations of SOA productions from isoprene, monoterpenes and isoprenoids
(sum of isoprene and monoterpenes) in PRD.

The low SOA productions during the daytime from higher alkanes, monoaromatics, naphthalenes, and isoprenoids in NCP (Figure 8(d)) are mainly due to the low SOA yields of these compounds during daytime (Figure S27). As shown in Figure 8(c, d), there are still a lot of missing SOA sources in both regions, which implies that some other SOA precursors or formation pathways (e.g. aqueous reactions) (Kuang et al., 2020) are contributing significantly to SOA formation in these regions. In this manuscript, we didn't consider the movement of the

boundary layer into the calculation because we don't have the measurement of the boundary

346 layer.



347

Figure S27. Diurnal variations of SOA yields of *n*-C15 alkane, benzene, naphthalene and α pinene in PRD (a) and NCP (b).

350

(9) Page 32, Figure 8(d): The diurnal variation of the measured SOA in Figure 8(d)seems to be different from that in Figure S11(b). Is it OK?

Response: The duration in the original Figure S11(b) is different from that in Figure 8(d). In
the original submitted manuscript, the durations in the Figure S11(b) and Figure 8(d) are 11.2512.18 and 11.26-12.14, respectively. We modified the Figure S11(b) in the revised supporting
information as follows, in which the duration is consistent with that in Figure 8(d).



- 358 Figure S19. Diurnal variations of concentrations of organic aerosols (OA), secondary organic
- aerosols (SOA) and primary organic aerosols (POA) in PRD (a) and NCP (b). POA and SOA
- 360 were determined by positive matrix factorization (PMF) analysis of OA measured by AMS.
- 361 (10) Page S5, Line 106: Explain "isoprene chemistry method".

Response: Isoprene are mainly photo-oxidized through the reactions with OH radical in
the atmosphere and its primary first-generation reaction products are formaldehyde,
MVK and MACR (Apel et al., 2002). The reaction processes of isoprene oxidized by OH
radical are mainly as follows:

366 Isoprene +
$$OH \rightarrow 0.63HCHO + 0.32MVK + 0.23MACR$$
 $k_1 = 1.0 \times 10^{-10} \text{ cm}^{-3} \text{ s}^{-1}$ (Eq.

367

S1)

368
$$MVK + OH \rightarrow Products \quad k_2 = 1.9 \times 10^{-11} \, cm^{-3} \, s^{-1} \, (Eq. \, S2)$$

369
$$MACR + OH \rightarrow Products$$
 $k_3 = 3.3 \times 10^{-11} cm^{-3} s^{-1}$ (Eq. S3)

370 where k₁, k₂, k₃ are the rate constants of the reactions. According to above reactions, the 371 relationship between MVK/Isoprene, MACR/Isoprene, (MVK+MACR)/Isoprene and the 372 reaction time Δt can be decribed as follows (Apel et al., 2002):

373
$$\frac{MVK}{Isoprene} = \frac{0.32k_1}{k_2 - k_1} \left(1 - \exp((k_1 - k_2) [OH] \Delta t) \right)$$
(Eq. S4)

374
$$\frac{MACR}{Isoprene} = \frac{0.23k_1}{k_3 - k_1} \left(1 - \exp((k_1 - k_3) [OH] \Delta t) \right) \quad (Eq. S5)$$

375
$$\frac{MVK+MACR}{Isoprene} = \frac{0.32k_1}{k_2 - k_1} \left(1 - \exp((k_1 - k_2) [OH] \Delta t)) + \frac{0.23k_1}{k_3 - k_1} \left(1 - \exp((k_1 - k_3) [OH] \Delta t)\right) \quad (Eq. S6)$$

376 where $\frac{MVK+MACR}{Isoprene}$ can be derived from the measurements by PTR-ToF-MS. Then

377 the OH exposure ($[OH]\Delta t$) of isoprenoids can be obtaind from Eq. S6.

378

(11) Page S7, Line 129: Which equation was used in the present paper, eqn. (1) in
Page 11, Line 235 or eqn. (7) in Page S7, Line 129? Don't make readers confused.

381 Response: In this manuscript, we used two methods to evaluate the importance of higher

alkanes to SOA. One is estimated by eqn. (1) in the main body. The other one is estimated by

- eqn. (7) in the supporting information. In the revised supporting information, we added "S"
- before all equation numbers. For example, eqn. (7) was modified to eqn. (S7).

- 385 Technical comments:
- 386 (1) Page 4, Line 72: Anh et al., 2008. ("et al." is missing)
- 387 Response: We added "et al.".
- 388 (2) Page 4, Line 83-84: Erickson et al. (2014) did not use NO+ as the reagent ion.
- **389** Response: We deleted this citation.
- 390 Don't make readers misread. Same comment to Page 5, Line 108.
- **391** Response: We deleted the citation from Erickson et al. (2014).
- (3) Page 6, Line 125: Corbin et al., 2015. (Delete "C.")
- **393** Response: We deleted "C."
- 394 (4) Page 9, Line 208: Table $1 \rightarrow$ Table 2
- 395 Response: Corrected
- (5) Page 14, Line 319-Page 15, Line 347: I feel it strange that some papers are cited in
- 397 "Conclusions". If the authors want to cite the papers, I think that the authors398 change"Conclusions" to "Concluding remarks".
- 556 change conclusions to concluding remarks.
- **399** Response: We changed "Conclusions" to "Concluding remarks".
- 400 (6) Page 17, Line 404: "C. Corbin, J." → "Corbin, J. C."
- 401 Response: Corrected
- 402 (7) Page S6, Line 107: Figure S8 \rightarrow Figure S9
- 403 Response: Corrected
- 404 (8) Page S7, Lines 134-143: These are mentioned in the text (Page 11, Line 249-Page 12, Line
 405 259)
- 406 Response: Here are descriptions for another method to calculate SOA production rate, some of
- 407 the parameters used in this method are same as those in the text.
- 408 (9) Page S17, Figure S9: "expouse" \rightarrow "exposure"
- 409 Response: Corrected
- 410

Response to anonymous referee #2

Wang et al measured concentrations of VOC and IVOC alkanes in two locations in China. They use the measurement data to estimate the importance of alkanes to ambient SOA. Based on their estimates, these alkanes make up ~7% of ambient SOA in both sampling locations in north and south China. Overall the manuscript is topically relevant to ACP. It presents what seems to be a novel application of chemical ionization MS for estimating SOA production. However, I have several comments before I can recommend publication.

418 Response: Thank you very much for your valuable comments and suggestions on our 419 manuscript. These comments are all helpful for improving our article. All the authors have 420 seriously discussed about all these comments. According to your comments, we have tried best 421 to modify our manuscript to meet with the requirements for the publication in this journal. 422 Point-by-point responses to the comments are listed below.

423 Major Comments:

411

(1) I have an issue with calling the measurement technique NO⁺ PTR-MS. The "P" in PTR
stands for proton, which in this case you are switching out for NO⁺ ions. So, the
measurement technique is chemical ionization with NO⁺ or selected ion mass spectrometry,
but not PTR. An exception would perhaps be if you're using PTR-MS to indicate the instrument
itself (e.g., an Ionicon PTR-MS) that has been modified.

429 Response: We understand the reviewer's concern. We use the NO⁺ PTR-TOF-MS as the instrument name to reflect the fact that NO⁺ chemistry has been mainly adopted from PTR-MS 430 431 instruments as a switchable reagent ion technique (Jordan et al. 2009), which is also the case for this work. The first attempt to measure higher alkanes using NO⁺ chemistry (Inomata et al., 432 2013) referred this method as NO⁺ chemical ionization using proton-transfer-reaction mass 433 spectrometry, which seems to reflect the reality of both the ion chemistry and also the 434 instrumentation used for the method. As the result, we changed the title in the revised 435 manuscript: "Measurements of higher alkanes using NO⁺ chemical ionization in PTR-ToF-436

437 MS: important contributions of higher alkanes to secondary organic aerosols in China".

438 In the main text (line 98-99) of the revised manuscript, we also reflect this information:

439 "In this study, we utilize NO⁺ chemical ionization in PTR-ToF-MS (here referred as NO⁺

440 PTR-TOF-MS) to measure higher alkanes at two different sites in China."

441 (2) Line 162-170 introduce that the PTR was run alternately with NO⁺ and H_3O^+ every 442 10-20 minutes. This cycling requires more explanation. How was the cycling achieved? 443 How long did the system take to re-equilibrate when the water flow was turned on and 444 off? Even if the system only took 1-2 minutes to re-equilibrate after switching between 445 ions, a large fraction of the data would be lost. Were data during the transition period 446 considered for analysis? Were voltages and pressures changed or held constant in the 447 NO⁺ and H_3O^+ operating modes?

Response: We included more details about the automatic switches between H₃O⁺ chemistry 448 and NO⁺ chemistry in the revised paper. The built-in software offers the possibility to program 449 sequences where the instrument is switching between different settings. It takes ~ 10 s for H₃O⁺ 450 and ~60 s for NO⁺ to re-stabilize when the measurement automatically switches between these 451 two modes. The ambient measurement data during the transition period ($\sim 1 \text{ min}$) is discarded. 452 The voltages of ion source and drift chamber are changed in H_3O^+ (Us=150 V, Uso=80 V; 453 Udrift=920 V, Udx=46 V) and NO⁺ (Us=40 V, Uso=100 V; Udrift=470 V, Udx=23.5 V) 454 operating modes. The pressures of drift chamber are held constant at 3.8 mbar in both modes 455 during the campaigns. 456

457 The added and modified sentences on line 199-205 of page 9 are listed below:

458 "Switching between H_3O^+ and NO^+ ion mode are provided by the PTR-MS Manager (v3.5) 459 software developed by the Ionicon Anlytik (Table S1). The pressures of drift chamber are 460 held constant at 3.8 mbar in both modes during the campaigns (Figure S10(a)). It usually 461 takes <10 s for H_3O^+ ions and ~60 s for NO^+ ions to re-stabilize after automatically 462 switching between the two measurement modes (Figure S10(b)). The ambient 463 measurement data during the transition period (~1 min) was discarded."

Table S1. The settings of the voltages of ion source voltages (Us, Uso), drift tube (Udrift, Udx)
and pressure of drift tube (pDrift) during automatical switching between NO⁺ mode and H₃O⁺
mode, respectively.

Setting	NO ⁺ mode	H_3O^+ mode
Us	40 V	150 V
Uso	100 V	80 V
Udrift	470 V	920 V
Udx	23.5 V	46 V
pDrift	3.8 mbar	3.8 mbar



468

Figure S10. An example of the voltages of ion source voltages (Us, Uso), drift tube (Udrift,
Udx) and pressure of drift tube (a), and the signal changes of primary ions (b) during
automatical switching between NO⁺ mode and H₃O⁺ mode, respectively.

(3) This paper relies on predicted SOA production to compare the importance of alkanes versus 472 compound classes. However, the 473 aromatics and other SOA estimates rely on a number of assumptions (e.g., SOA yields) that are uncertain. There are other 474 ways to compare the importance of alkanes to other compounds. There is very little 475 discussion of absolute concentrations, which is one way to make the comparison. The 476 authors could also compare something like OH reactivity. 477

Response: We included the comparison of average concentrations from higher alkanes (C8C21 alkanes), monoaromatics (benzene, toluene, C8 aromatics, C9 aromatics and styrene),
naphthalenes (naphthalene, methylnaphthalenes, dimethylnaphthalenes) and isoprenoids
(isoprene and monoterpenes) in PRD and NCP in the revised supporting information.
Compared to monoaromatics, higher alkanes have lower average concentrations. But due to
the high SOA yields, higher alkanes also play an important role in SOA formation.

484

We added sentences on line 354-356 of page 15 in the revised manuscript as follows:

485 "Compared to monoaromatics, higher alkanes are associated with lower concentrations
486 (Figure S26). However, higher alkanes play an important role in SOA formation due to

487 their high SOA yields (Figure S27)."



Figure S26. The average concentrations from higher alkanes (C8-C21 alkanes), monoaromatics
(benzene, toluene, C8 aromatics, C9 aromatics and styrene), naphthalenes (naphthalene,
methylnaphthalenes, dimethylnaphthalenes) and isoprenoids (isoprene and monoterpenes) in
PRD (a) and NCP (b), respectively.



494 Figure S27. Diurnal variations of SOA yields of *n*-C15 alkane, benzene, naphthalene and α-495 pinene in PRD (a) and NCP (b).

(4) The comparisons in Figure 8 (and associated discussion) do not seem to place 497 alkanes and aromatics on equal footing. The alkane signal is essentially the sum of 498 all C x alkanes (i.e., C12 is the sum of dodecane, cyclododecane, and all branched 499 isomers). This is analogous to integrating individual peaks in the GC along with the 500 entire C12 "bin" of UCM (unresolved complex mixture) as shown by Zhao et al (2016a, 501 502 b) and others. However, the authors only consider specific aromatics (e.g., BTEX) but not the aromatic UCM of larger substituted benzenes. A more complete comparison of 503 504 alkanes to aromatics would include these species as well.

Response: Many thanks for your nice suggestion. Since we didn't have the measurement of 505 aromatic UCM of larger substituted benzenes during these two campaigns, we only consider 506 the specific aromatics (benzene, toluene, C8 aromatics, C9 aromatics and styrene) in this study. 507 The aromatics used in this study are also the most considered aromatic hydrocarbon species in 508 509 current SOA research. We agree that if the aromatic UCM of larger substituted benzenes are included, the contribution of aromatics to SOA may be even greater. However, in this study we 510 want to emphasize the importance of higher alkanes to SOA formation, which is at least 511 comparable to these common aromatic species, rather than figure out which VOCs class 512 contribute most to SOA. The measurement of aromatic UCM of larger substituted benzenes 513 and their contributions to SOA will also be considered in our future studies. 514

28

Minor comments: 515

(1) Lines 131-136 and Figure 2 - The figure would be easier to interpret if the relative humidity 516 was also shown. 517

518 Response: The relative humidity is relative to the temperature. During the campaigns, especially in 2018 NCP campaign, the relative humidity varies a lot due to the temperature 519 520 changes. While, the sensitivities of higher alkanes are depended on the water vapour content in the sampling air. Previous studies indicate that water vapour in the gas phase can lead to the 521 522 formation of hydrated hydronium ions ($(H_2O)_n H_3O^+$) in the drift chamber (Blake et al., 2009). These hydrated hydronium ions can then themselves act as regent ions. Moreover, these 523 hydrated hydronium ions can result in additional reactions (e.g. ligand switching) other than 524 proton transfer in the drift chamber (Midey et al., 2000). Therefore, we use the absolute 525 526 humidity to present the humidity dependence of higher alkanes in this paper.

(2) Line 192, in the comparison of GC-MS and NO⁺ ionization for alkanes: "Similar temporal 527 trends for these alkanes are observed from the two instruments." I can't really tell this from the 528 figure. The diurnal trends suggest that the PTR-derived measurements have a deeper afternoon 529 trough than the GC-MS, which seems to show a flatter concentration profile. 530

531 Response: We modified the figures to make the comparison clearer and easier to read. Since we don't have sufficient data and evidence, the deeper diurnal trends of higher alkanes in the 532 afternoon from the PTR-derived measurements compared to GC-MS measurements are not 533 discussed in this study. 534



0.25

0.20

0.15

0.10

0.05

0.00

GC-MS/FID (ppt

- 536 Figure 5. Comparisons of times series and diurnal variations of alkanes measured by NO⁺ PTR-
- 537 ToF-MS and GC-MS/FID in PRD. (a) Time series of C8 alkanes measured by NO⁺ PTR-ToF-
- 538 MS, C8 *n*-alkane and four branched isomers measured by GC-MS/FID. (b) Diurnal variations
- of C8 alkanes. (c-e) Diurnal variations of C9-C11 alkanes with NO⁺ PTR-ToF-MS and C9-
- 540 C11 *n*-alkanes with GC-MS/FID.
- 541 (3) Line 193-194 PTR data are a factor of 3-6 higher than the GC data. It looks like the GC
- 542 data were only analyzed for species that are resolved by the GC (e.g., no UCM). How does the
- closure look if the entire chromatogram (or the entire m/z 57 signal) is binned and analyzed,
- similar to Zhao et al's (2016a, b) work?
- 545 Response: During the PRD campaign, the online GC-MS/FID system was operated in the
- selected ion monitoring (SIM) mode without scanning m/z 57 all the time. Therefore, we are
- 546 selected ion monitoring (SIM) mode without scanning m/
- 547 not able to do the work as you suggested here.
- 548 (4) Why does Figure 5 only show PRD data for carbon numbers 8-11? The instrument549 measured up to C21.
- 550 Response: The online GC-MS/FID system we used in this campaign can only measure alkanes
- up to C11. So, we compared only C8-C11 alkanes, which were measured by both GC-MS/FID
- 552 and NO^+ ToF-MS.

553 Grammar:

- (1) Line 80 states that PTR "is response to large alkanes" please edit. I think the authors meanthat PTR responds to large alkanes.
- 556 Response: We modified "is response to large alkanes" to "is responsive to large alkanes".
- 557 (2) Line 160 replace expect with except
- 558 Response: We replaced expect with except.

Response to anonymous referee #3

In their manuscript, "Measurements of higher alkanes using NO⁺ PTR-ToF-MS: significant 561 contributions of higher alkanes to secondary organic aerosols in China," the authors tackle an 562 important but often overlooked issue: the contribution of high-molecularweight gas-phase 563 alkanes to SOA formation. A relatively large body of work over the last decade has provide a 564 fairly thorough examination of the composition of vehicle emissions and the importance of 565 branched (and cyclic) alkanes; this manuscript uses this work as a starting place to explore their 566 contribution to SOA in China. The work is made possible in large part by the recent 567 development of the NO⁺ PTR. Overall, I think this work addresses interesting questions, and 568 makes real efforts to tackle the complexity of calibrating a new instrument for compound 569 classes that may contain hundreds of isomers. However, there are a few scientific weakspots 570 that need to be addressed before this work is ready for publication. These larger comments 571 include some needs for clarification, and some more fundamental issues with how OH exposure 572 and SOA potential are being calculated. I commend the authors for taking a step into some 573 unknown territory, but some additional discussion (and possibly work) needs to be included. 574

575 Response: Thank you very much for your valuable comments and suggestions on our manuscript. These comments are all helpful for improving our article. All the authors have 576 577 seriously discussed about all these comments. According to your comments, we have tried best to modify our manuscript to meet with the requirements for the publication in this journal. The 578 details about the calculation of OH exposure was added in the supporting information of the 579 revised manuscript. Two methods to estimate SOA formation from different VOCs classes 580 were all presented in the manuscript and the supporting information. Point-by-point responses 581 to the comments are listed below. 582

583 General Comments:

(1) There are some typo and english issues throughout. It is not sufficient to seriously detract
from the science, but it is to the degree that the authors should have a native english speaker
review and edit this manuscript.

587 Response: Thanks for your nice suggestion.

(2) Considering the focus of this work is on large alkanes, and the semi-volatile nature of these components, particularly at the lower temperatures of some of these measurements, this manuscript really does not provide enough detail on the sample line and inlet. How long is this 591 sample line? Is it heated all the way to the sample inlet, or just in the room? If it is not heated 592 all the way to the inlet, I would have some misgivings about alkanes about C14 or so, there 593 could be substantial losses or time lags for larger alkanes. Is there any evidence (observations 594 or models) that suggest line losses and delays can be ignored? If the authors want to focus on 595 gas-phase measurements of S/IVOC alkanes, more detail needs to be provided for the sampling.

Response: We included more information about the sampling in the revised manuscript. The 596 schematic drawing of the inlet system for the measurement during campaigns is shown in the 597 attached figure (Fig S2, Wu et al., 2020). The sampling line is ~8 m and ~3 m during PRD 598 campaign and NCP campaign respectively. Ambient air was continuously introduced into ToF-599 MS through a Teflon tubing (1/4") with an external pump at 5.0 L/min. The calculated 600 residence time for the inlet was ~ 3 s for PRD campaign and ~ 1 s for NCP campaign, 601 respectively. The sampling line was heated all the way to the sampling inlet by an insulating 602 tube with a self-controlled heater wire (40 $^{\circ}$ C) wrapping outside. We conducted tubing loss 603 experiments in the laboratory by introducing standards (2 ppb/1 ppb/0.5 ppb) of higher alkanes 604 (n-C8-C15), monoaromatics (benzene, toluene, o-xylene, 1,2,4-trimethylbenzene), isoprenoids 605 (isoprene, α -pinene) and naphthalene into PTR-ToF-MS through a 8 m Teflon tubing (1/4") at 606 room temperature with an external pump at 5.0 L/min (Figure S11). The tubing loss of these 607 standard compounds is found to be <5% except *n*-C15 of ~8% and naphthalene of ~10%. Given 608 the fact that the sampling line was heated at 40 $^{\circ}$ C all the way to the inlet during the campaigns, 609 610 we think the tubing loss would be ignored.



611

Figure S2. Schematic drawing of the inlet system for PTR-ToF-MS during the campaign. (Wuet al., 2020)





615 Figure S11. The tubing loss experiments of higher alkanes (*n*-C8-C15), 1,2,4-616 trimethylbenzene,α-pinene and naphthalene at room temperature using PTR-ToF-MS with an 617 external pump at 5.0 L/min.



The sentences on line 205-216 of page 9-10 are modified as follows:

619 "Ambient air was continuously introduced into PTR-ToF-MS through a Teflon tubing 620 (1/4") with an external pump at 5.0 L/min, with tubing length of ~8 m and ~3 m during 621 the PRD and the NCP campaign, respectively. The inlet tubing was heated all the way to 622 the sampling inlet to avoid water vapour condensation by an insulating tube with a self-623 controlled heater wire (40 °C) wrapping outside. The calculated residence time for the 624 inlet was ~3 s for PRD campaign and ~1 s for NCP campaign, respectively. The tubing 625 loss experiments were conducted in the laboratory by introducing standards of higher alkanes (*n*-C8-C15), monoaromatics (benzene, toluene, o-xylene, 1,2,4-trimethylbenzene), isoprenoids (isoprene, α -pinene) and naphthalene into PTR-ToF-MS through a 8 m Teflon tubing (1/4") at room temperature with an external pump at 5.0 L/min (Figure S11). The tubing loss of these compounds is found to be <5% except *n*-C15 (~8%) and naphthalene (~10%)."

3) In the discussion around "sensitivity", is that the response per mass of the m-1 ion, or the
sum of all ions? If the former, it should be discussed in some of the relationships shown, as
discussed in comments below. If the latter, how are those ions summed and attributed in the
complex real-world data? Understanding of this terminology is critical for interpreting some of
the figures.

- Response: The "sensitivity" is that the response per mass of the m-1 ion with a unit of ncps/ppb 636 637 in this study. We measured the fragmentation patterns of selected higher alkanes with NO⁺ PTR-ToF-MS by introducing commercially acquired pure chemicals (Figure S8). We found 638 that higher alkanes are mainly reacted through hydride abstraction by NO⁺ forming mass (m-639 1) ions (m is the molecular mass). The fractions of m-1 ions account for more than 60% of total 640 ion signals for >C12 *n*-alkanes (Figure 4 (a)). Thus, we use the response per mass of the m-1 641 ion for each higher alkane when performing sensitivity experiments using a gas standard with 642 a series of *n*-alkanes. We added a sentence on line 147-150 of page 7 in the revised manuscript 643 to explain the "sensitivity" where this word was first mentioned. 644
- 645 "Sensitivities of *n*-alkanes (C8-C15) standards were obtained during the campaign
 646 (Figure S6), which is defined as the normalized signal of hydride abstraction ions for each
 647 higher alkane at 1 ppby with a unit of ncps/ppb."



648

Figure S8. Mass spectra of the distributions of product ions from *n*-Dodecane (a), *n*-Pentadecane (b) and *n*-Eicosane (c) with NO⁺ PTR-ToF-MS. The signals of masses shown in the graph are the results after subtracting the isotopic signals during the high resolution peak fitting of the mass spectra.



Figure 4. (a) The fractions of product ions (m-1) from hydride abstraction of C8-C20 *n*-alkanes
in NO⁺ PTR-ToF-MS. (b) Scatterplot of sensitivities under dry conditions versus the fractions
of hydride abstraction ions for C8-C15 *n*-alkanes.

4) More detail is necessary on how they calculate OH exposure. I'm not familiar with the
isoprene chemistry method - please include a description. And for the ratio of aromatics method,
don't you need to start by assuming some ratio at the time of emissions? What is assumed here?
Why does the OH exposure diurnal (Figure S9) look very different for the biogenics and the
anthropogenics? Notably, in the NCP campaign this is also a large missing daytime source,
could this be related to a "mistiming" of SOA caused by a bias in the OH exposure calculation?
Or what might be the missing source?

665 Response: We included more detail about the calculations of OH exposure in the revised 666 supporting information. The observed ratios between m+p-xylene and ethylbenzene were 667 used to estimate the OH exposure ($[OH] \times \Delta t$) by Roberts et al. (1984):

668
$$[OH] \times \Delta t = \frac{1}{k_{m+p-xylene} - k_{ethylbenzene}} \times [ln(\frac{m+p-xylene}{ethylbenzene})_{t=0} - ln(\frac{m+p-xylene}{ethylbenzene})_t]$$

669 Where the initial emission ratios of m+p-xylene/ethylbenzene were estimated according to 670 the correlation of m+p-xylene with ethylbenzene during campaigns. The ratio of 4 and

671 1.5 were used in the PRD campaign and the NCP campaign, respectively (Figure S29).



Figure S29. Correlation of m+p-xylene with ethylbenzene in PRD (a) and NCP (b). The dashed
lines in both graphs indicate the estimated initial mission ratio of m+p-xylene/ethylbenzene.

Isoprene are mainly photo-oxidized through the reactions with OH radical in the
atmosphere and its primary first-generation reaction products are formaldehyde, MVK
and MACR (Apel et al., 2002). The reaction processes of isoprene oxidized by OH radical
are mainly as follows:

679 Isoprene + $OH \rightarrow 0.63HCHO + 0.32MVK + 0.23MACR$ $k_1 = 1.0 \times 10^{-10} \text{ cm}^{-3} \text{ s}^{-1}$ (Eq.

680 **S1**)

672

681
$$MVK + OH \rightarrow Products \quad k_2 = 1.9 \times 10^{-11} \, cm^{-3} \, s^{-1} \, (Eq. \, S2)$$

$$MACR + OH \rightarrow Products \quad k_3 = 3.3 \times 10^{-1.1} cm^{-3} s^{-1} (Eq. S3)$$

683 where k_1 , k_2 , k_3 are the rate constants of the reactions. According to above reactions, the 684 relationship between MVK/Isoprene, MACR/Isoprene, (MVK+MACR)/Isoprene and the 685 reaction time Δt can be decribed as follows (Apel et al., 2002):

686
$$\frac{MVK}{Isoprene} = \frac{0.32k_1}{k_2 - k_1} \left(1 - \exp((k_1 - k_2) [OH] \Delta t) \right)$$
(Eq. S4)

687
$$\frac{MACR}{Isoprene} = \frac{0.23k_1}{k_3 - k_1} \left(1 - \exp((k_1 - k_3) [OH] \Delta t) \right) \quad (Eq. S5)$$

688
$$\frac{MVK+MACR}{Isoprene} = \frac{0.32k_1}{k_2 - k_1} \left(1 - \exp((k_1 - k_2) [OH] \Delta t)) + \frac{0.23k_1}{k_3 - k_1} \left(1 - \exp((k_1 - k_3) [OH] \Delta t)\right) \quad (Eq. S6)$$

689 where $\frac{MVK+MACR}{Isoprene}$ can be derived from the measurements by PTR-ToF-MS. Then the OH 690 exposure ([*OH*] Δt) of isoprenoids can be obtaind from Eq. S6.

The different diurnal patterns for the biogenics and the anthropogenics are mainly due 691 to their different emission patterns. Much more fresh emissions for biogenic compounds during 692 the daytime than nighttime, leading to the low OH exposure during the daytime for these 693 compounds. The photochemical age of the air mass after mixing is the nonlinear addition of 694 the photochemical age before mixing. Generally, the photochemical age of the mixed air mass 695 is closer to that of the fresh air mass (Parrish et al., 1992). The large missing daytime source in 696 the NCP might be some other SOA precursors or formation pathways (e.g. aqueous reactions) 697 (Kuang et al., 2020) 698

The math on Eq. 1 makes sense, but I'm not sure the assumptions do. This assumes that the air behaves like a plug flow reactor from some emissions source, and then just oxidizes until the measurement site. Wouldn't local emissions (that perhaps have a different oxidation history) change things dramatically? Given that benzene and TMB aren't present in diesel or motor oil (which is the source of higher alkanes), this might be an issue. Though they are both from vehicle emissions, so maybe not. This might all be more compelling with more detail on how OH exposure is calculated.

Response: When calculating OH exposure for anthropogenic compounds, we use the ratio between m+p-xylene and ethylbenzene. Currently, given the fact that we don't have sufficient data to evaluate the specific contributions from different emission sources for all these anthropogenic compounds, we assume that m+p-xylene, ethylbenzene and higher alkanes have the same dominant emission sources during campaigns. In the near future, we will conduct VOCs measurements from typical emission sources in China.

712 Specific comments:

Line 135. The relationship between humidity and sensitivity seems fairly robust, so correctable, but I'm not sure a would classify a 50% drop in sensitivity as a "slight" decrease. If the "sensitivity" is to m-1, is the drop due to a change in fragmentation, or a drop in all ions? As it also due to a shift of NO⁺ to H_3O^+ as water content increases, or does increasing humidity not increase H_3O^+ ?

718 Response: We modified the description about the relationship between humidity and sensitivity.

719 We did several times humidity experiments in the laboratory by diluting higher alkanes

- standard into humidified air (relative humidity of 0-95% at 25%) to reach approximately 1 ppb
- 721 mixing ratio. The Figure 3 (b, c) summarize all the experiments data at different times to derive

the relationships for C12 and C15 alkanes. The "sensitivity" is to m-1 ion. As shown in the attached Figure (a), the normalized signals of all product ions and the fragment ions are decreasing with the increase of humidity. These decreasing patterns are probably due to the decreasing reagent ions (NO⁺ and O₂⁺) as the humidity increases. Thus, the humidity correction should be applied for the quantitation of higher alkanes using NO⁺ PTR-ToF-MS.

We modified and added sentences on line 163-168 of page 7-8 in the revised manuscriptas follows:

"As shown in Figure 3(b, c) and Figure S7 (a), the normalized signal of all product ions
(m-1) and the fragment ions of *n*-alkanes (C8-C15) standards are decreasing with the
increase of humidity. These decreasing patterns are probably due to the decreasing
primary reagent ions (NO⁺ and O₂⁺) as the humidity increases Figure S7(b). Thus, the
humidity correction should be applied for the quantitation of higher alkanes using NO⁺
PTR-ToF-MS."

735



Figure 3. (a) Calibrations of *n*-Dodecane and *n*-Pentadecane under dry conditions; (b) Humidity
dependence of *n*-Dodecane. (c) Humidity dependence of *n*-Pentadecane.



740

Figure S7. Humidity dependence of all product ions and the fragment ions for *n*-alkanes (C8-C15) (a), and primary ions (NO⁺, O_2^+ , H_3O^+) (b).

Line 144. Is this slope between fraction of m-1 and sensitivity just due to the decrease m-1? In other words, is the total ion count produced the same, and just the fragmentation different, or is the total ion count lower for lower alkanes?

Response: We used a mixed standards of *n*-alkanes (C8-C15) to do the sensitivity experiments with NO⁺ PTR-ToF-MS. Since almost all the higher alkanes produce the fragmentation ions of m/z 57, m/z 71, m/z 85, and m/z 99, we are not able to calculate the total ion count for each higher alkane with our experiments in this study. But the reason for this slope might be explored in the future when we have sufficient experimental conditions.

- 751 Line 144-145. Are calibrations of C16-C21 not available directly just because it is hard to get 752 them into the gas phase, or was there some other reason?
- 753 Response: The commercial mixed standards of *n*-alkanes (C8-C15) we used in this study are
- all at 100 ppb except *n*-C15 at 30 ppb, because the lower vapour pressure of *n*-C15 make it

difficult to obtain higher concentration of 100 ppb in the same cylinder. The vapour pressures

of n-C16-n-C21 are even lower than that of n-C15. It is not possible to prepare n-C16-n-

- 757 C21standards with concentrations of > 30 ppb into the mixed standards of *n*-alkanes (C8-C15)
- in the same cylinder. Moreover, the very low vapour pressure compounds are not stable in the
- 759 cylinder.

Line 160. Should be "except" instead of "expect Line 162. How was this switching achieved?
Was it just a change in voltages? Are there residual effects observed, and/or do some transitional data need to be removed?

- Response: We replaced the "expect" with "except" in the revised manuscript. More details about the automatic switches between H_3O^+ chemistry and NO^+ chemistry are also included on
- 765 line 199-205 of page 9 in the revised paper.

"Switching between H₃O⁺ and NO⁺ ion mode are provided by the PTR-MS Manager (v3.5)

767software developed by the Ionicon Anlytik (Table S1). The pressures of drift chamber are768held constant at 3.8 mbar in both modes during the campaigns (Figure S10(a)). It usually769takes <10 s for H_3O^+ ions and ~60 s for NO⁺ ions to re-stabilize after automatically770switching between the two measurement modes (Figure S10(b)). The ambient771measurement data during the transition period (~1 min) was discarded."

Table S1. The settings of the voltages of ion source voltages (Us, Uso), drift tube (Udrift, Udx) and pressure of drift tube (pDrift) during automatical switching between NO⁺ mode and H_3O^+ mode, respectively.

NO^+ mode	H_3O^+ mode
40 V	150 V
100 V	80 V
470 V	920 V
23.5 V	46 V
3.8 mbar	3.8 mbar
	NO ⁺ mode 40 V 100 V 470 V 23.5 V 3.8 mbar



Figure S10. An example of the voltages of ion source voltages (Us, Uso), drift tube (Udrift,
Udx) and pressure of drift tube (a), and the signal changes of primary ions (b) during
automatical switching between NO⁺ mode and H₃O⁺ mode, respectively.

780

781 Line 167. I'm not quite sure what the "insulated tube" is - do they mean heater rope or 782 something?

783 Response: We revised the sentence. "The inlet tubing was heated all the way to the sampling

inlet to avoid water vapour condensation by an insulating tube with a self-controlled

- 785 heater wire (40 ℃) wrapping outside."
- Line 185. Do the authors mean there are limitations to its application? Or just that it hasn'tbeen applied very often?
- 788 Response: We revised the sentence. "Although NO⁺ chemistry has been shown to be

valuable in measuring many organic species, the applications in real atmosphere of

790 different environments are still rare".

- Line 187-190. The comparison between H3O+ and NO+ and GC-MS/FID is very encouraging.
- There are some significantly non-unity slopes though (0.47 for benzene) -do the authors have
- an explanation for this?

- **794** Response: Yes, the benzene of H_3O^+ measurement is not well consistent with that of NO^+
- measurement in the formal period of PRD campaigns. We don't know the exact reason for this.
- 796 Considering the good consistency of benzene between NO⁺ measurement and GC-MS
- 797 measurement, we used the benzene from NO^+ measurement in this study.
- Line 276. A Table of SOA yields used would be helpful for the SI
- 799 Response: We included a table of SOA yields in the revised supporting information (Table S3).
 - Formula **Average SOA** Average SOA Compounds Octane C₈H₁₈ 0.003 ± 0.002 0.006 ± 0.004 Nonane $C_{9}H_{20}$ 0.010 ± 0.005 0.017 ± 0.010 Decane 0.026 ± 0.012 0.040 ± 0.021 $C_{10}H_{22}$ Undecane $C_{11}H_{24} \\$ 0.058 ± 0.020 0.080 ± 0.036 Dodecane $C_{12}H_{26}$ 0.106 ± 0.032 0.142 ± 0.059 Tridecane C₁₃H₂₈ 0.249 ± 0.061 0.305 ± 0.103 Tetradecane 0.329 ± 0.070 0.388 ± 0.118 $C_{14}H_{30} \\$ Pentadecane 0.386 ± 0.081 0.450 ± 0.135 $C_{15}H_{32}$ 0.492 ± 0.141 Hexadecane 0.428 ± 0.086 $C_{16}H_{34}$ Heptadecane $C_{17}H_{36}$ 0.488 ± 0.096 0.556±0.156 Octadecane $C_{18}H_{38}$ 0.664 ± 0.079 0.704 ± 0.139 Nonadecane 0.792 ± 0.105 $C_{19}H_{40}$ 0.773 ± 0.056 Eicosane 0.860 ± 0.025 0.863 ± 0.054 $C_{20}H_{42}$ 0.877 ± 0.025 0.870 ± 0.046 Heneicosane $C_{21}H_{44}$
- Table S3. The calculated average SOA yields of higher alkanes in PRD and NCP.

Line 301-304. Are there emissions sources for only alkanes? I would guess not, in which case these alkanes must be getting emitted alongside cyclic saturated hydrocarbons (e.g., cycloalkanes), which actually contribute similar or more to petroleum fuels (e.g., Gentner et al., 2012, Table S5). These compounds are expected to have broadly similar (or slightly higher) SOA yields. This would have you substantially overestimating SOA in NCP at night - the authors should comment on possible explanations or biases.

Response: In this study, the higher alkanes represent only the total concentration of linear and
branched isomers, without considering the cyclic alkanes. Yes, some studies have indicated
that cyclic alkanes also contribute importantly to SOA formation (Tkacik et al., 2012; Gentner

812 Including these cyclic alkanes would enhance the estimation of SOA production shown in this

et al., 2012). In the near future, we will also try the possibility to measure these cyclic alkanes.

813 study.

811

814 Figure 5. I'm not quite sure what data is being shown here. Gentner provides diesel fuel data in table S5 of that publication (this is used to approximate exhaust in that manuscript). That 815 data doesn't agree with what is shown here though - for C8 it is 37%, and all the values are 816 near or above 21%. The Isaacman paper is actually just a single fuel sample that is included in 817 the Gentner data set, so could probable be excluded. That paper also provides gasoline data, 818 which is not shown. In Figure 9 of the Chan paper, the branched-to-normal ratios are given for 819 vehicle exhaust, which convert to between 13 and 41% normal for C24 through C21, which 820 are fairly different than those shown. Similarly, in Figure 5, the branched-to-normal ratios for 821 LA are shown at multiple oxidation times, and reach levels of 20 to 50% for those same carbon 822 numbers - this is simlar to the values shown, but I think not quite right. I think the numbers for 823 Bakersfield (based on Figure 4 from that paper) should be much lower, 5-10%, for all carbon 824 numbers. Values should also be included for direct vehicle emissions from Worton et al., 2014, 825 826 dx.doi.org/10.1021/es405375j. Basically, some explanation for how these numbers were all compiled is necessary (even if it's just in the SI), because they don't look quite right to me. 827

828 Response: Linear alkanes and branched alkanes follow the same molecular formula: C_nH_{2n+2} . 829 When using NO⁺ ToF-MS, we measure the total concentrations of linear alkanes and branched 830 alkanes with the same formulas since this technique doesn't differentiate isomers. We collected 831 data from literatures which are derived from GC-based techniques to calculate the mass 832 fractions of *n*-alkanes in higher alkanes with the same formula from various ambient and 833 emission studies, with the purpose of emphasizing the importance of contribution of branched isomers to higher alkanes concentrations at each carbon number. The fraction of *n*-alkane ateach carbon number is calculated as follows:

836

$$Fraction_{linear} = \frac{Concentration_{linear}}{Concentration_{linear} + Concentration_{branched}} \times 100\%$$

As shown in the figure, *n*-alkanes contribute ~5-60% to higher alkanes concentrations from C8 to C25, indicating that branched isomers contribute up to ~40-95%. This indicates the measurement of branched isomers is also important. We checked the data collected from references and modified the figure (Figure 6) as suggested. A table of data used in this figure was also included in the revised supporting information (Table S2). The data from Worton et al., 2014 was not included because those high alkanes data are particle phase. In this study, we only focus on the gas phase alkanes and their importance to SOA.



844

Figure 6. Fractions of *n*-alkanes in higher alkanes with same formulas derived from this study,

ambient air in Los Angeles, Bakersfield and in vehicle exhausts.

848 Table S2. Fractions of *n*-alkanes in higher alkanes with same formulas derived from this study,

Carbon Number	Fraction of <i>n</i> -alkanes in higher alkanes with same formulas			same formulas	
	PRD ^a	Los Angeles ^b	Bakersfield ^b	Diese exhaust ^c	Liquid gasoline exhaust ^c
8	10.82	/	/	37.04	5.39
9	21.48	/	/	51.22	7.71
10	17.56	/	/	23.81	8.81
11	17.81	/	/	20.91	10.88
12	/	/	/	22.54	29.82
13	/	/	/	21.98	/
14	/	/	/	19.84	5.41
15	/	/	/	22.86	40
16	/	/	/	25.44	/
17	/	/	/	32.16	/
18	/	/	/	28.57	/
19	/	/	/	20.83	/
20	/	34.78	24.85	20.87	/
21	/	53.16	12.47	24.82	/
22	/	42.85	9.11	25.51	/
23	/	40.24	8.14	21.05	/
24	/	21.85	6.92	24.44	/
25	/	27.17	27.87	60	/

849 ambient air in Los Angeles, Bakersfield and in vehicle exhausts.

850 ^a:This work; ^b: Chan et al., 2013; ^c: Gentner et al., 2012

Response to anonymous referee #4

853 General comments:

This work is a nice, extensive manuscript using an NO⁺ CIMS (see below) to measure ambient alkanes in the PRD of China and evaluate their importance to SOA chemistry. The experimental methods are high quality and very well-documented. The importance of this manuscript is high, and I find it to be novel and useful.

It is, however, difficult to understand in places and suffers from some lack of precision in language. I recommend publication after these relatively minor, but common issues are fixed.

Response: Thank you very much for your valuable comments and suggestions on our
manuscript. These comments are all helpful for improving our article. All the authors have
seriously discussed about all these comments. According to your comments, we have tried best
to modify our manuscript to meet with the requirements for the publication in this journal.
Point-by-point responses to the comments are listed below.

865 Specific comments:

(1) Instrument name: The title of the paper and description of the instrument is not correct.
PTR is a widely used term that refers to the transfer of a proton from H₃O⁺ under controlled
conditions. NO⁺ ionization uses different ionization mechanisms, particularly charge transfer.
I understand the instrument is a commercial PTR-MS, but the authors are now using a different
reagent ion and the authors should change the name to something like NO⁺ ToF-MS or NO⁺
CIMS. Analytical chemistry acronyms are confusing enough as it is. Removing all meaning
from them will make them unintelligible to everyone.

Response: We understand the reviewer's concern. We use the NO⁺ PTR-TOF-MS as the 873 instrument name to reflect the fact that NO⁺ chemistry has been mainly adopted from PTR-MS 874 instruments as a switchable reagent ion technique (Jordan et al. 2009), which is also the case 875 for this work. The first attempt to measure higher alkanes using NO⁺ chemistry (Inomata et al., 876 877 2013) referred this method as NO⁺ chemical ionization using proton-transfer-reaction mass spectrometry, which seems to reflect the reality of both the ion chemistry and also the 878 879 instrumentation used for the method. As the result, we changed the title in the revised manuscript: "Measurements of higher alkanes using NO⁺ chemical ionization in PTR-ToF-880

881 MS: important contributions of higher alkanes to secondary organic aerosols in China".

882 In the main text (line 98-99) of the revised manuscript, we also reflect this information:

883 "In this study, we utilize NO⁺ chemical ionization in PTR-ToF-MS (here referred as NO⁺ 884 PTR-TOF-MS) to measure higher alkanes at two different sites in China."

(2) SOA Yields: There would appear to be a large amount of possible error in these
measurements and calculations and it needs to be reflected when yields (SOA production) are
displayed across the paper. Some yield calculations are presented with errors and some are not.
Figure 7 and 8, in particular have issues with this.

- 889 Response: We included the errors when presenting the concentrations, SOA yields and SOA
- production. The errors displayed in this study refer to the standard deviations (1 δ) over the averaging period during the campaigns.
- (3) P2 L29: Is it novel? This method has previously been published (with GC comparisons) bysome of the co-authors of this paper in Koss et al.
- 894 Response: We deleted the word "novel".
- (4) P12 L280: Is this really a surprise? Why? Do those references use different yields or inputsthan this work?
- Response: We deleted the word "interesting and the sentence was modified to "The
 distribution of contributions from alkanes with different carbon number to SOA
 formation shown here is in good agreement with the previous results referred from
 volatility calculation for precursors (de Gouw et al., 2011;Liggio et al., 2016)."
- 901 (5) Figure 1: The mass-to-charge labels on the top of the graph are entirely illegible and the
 902 labels on the axes are too small to be conveniently legible. Please make larger. I understand
 903 that the authors used the default Tofware labels, but "IndivHRfitCurves" will not mean a lot to
 904 most readers. Please change the labels to something clearer in each graph.
- 905 Response: We modified this figure in the revised manuscript as follows.



Figure 1. High-resolution (HR) peak-fitting to the averaged mass spectra on a typical day (12 October 2018) for m/z 169 (a), m/z 211 (b) and m/z 281 (c), at which masses produced by dodecane ($C_{12}H_{25}^+$), pentadecane ($C_{15}H_{31}^+$) and eicosane ($C_{20}H_{41}^+$) produced in NO⁺ PTR-ToF-MS.

911 (6) Figure 5 gives the impression that the authors suggest the dominant source of alkanes in the
912 region is diesel vehicles. Is that correct? If not, I'm not sure I understand the point of putting
913 those traces on the graph with ambient measurements.

914 Response: This figure was used to emphasize the importance of contribution of branched isomers to higher alkanes concentrations at each carbon number. Linear alkanes and branched 915 alkanes follow the same molecular formula: C_nH_{2n+2}. When using NO⁺ PTR-ToF-MS, we 916 measure the total concentrations of linear alkanes and branched alkanes with the same formulas 917 since this technique doesn't differentiate isomers. We collected data from literatures which are 918 derived from GC-based techniques to calculate the mass fractions of *n*-alkanes in higher 919 alkanes with the same formula from various ambient and emission studies. Because of quite 920 limited studies published about the emission sources of higher alkanes, we only collected few 921 922 data about the vehicle exhaust. The other emission sources of higher alkanes are still unknown and are needed to study in the future. 923



924

925 Figure 6. Fractions of *n*-alkanes in higher alkanes with same formulas derived from this study,926 ambient air in Los Angeles, Bakersfield and in vehicle exhausts.

927 (7) Figure 7a needs error bars on the calculated SOA yields. There would appear to be a large
928 amount of possible error in these measurements and calculations and it needs to be reflected
929 when yields are displayed in this work

930 Response: We included the error bars on the calculated SOA yields in Figure 8 in the revised manuscript as follows. In this study, SOA yield data for higher alkanes (Lim and Ziemann, 931 2009;Presto et al., 2010;Loza et al., 2014;Lamkaddam et al., 2017b) were summarized from 932 reported values in the literature, with the consideration of the influence of organic aerosol 933 concentration to SOA yields (Figure S20). The error bars in Figure 8(a) refer to standard 934 deviations (1 δ) over the averaging period of calculated SOA yields. The error bars in Figure 935 936 8(b) refer to standard deviations (1 δ) over the averaging period of calculated SOA production during the campaigns. 937



Figure S20. The reported SOA yields as a function of OA concentrations for higher alkanes
(C8-C21 alkanes) (a-k) under high-NO_x condition from chamber studies.



Figure 8. (a) Measured concentrations by NO⁺ PTR-ToF-MS, calculated consumed
concentrations and average SOA yields for C8-C21 alkanes in PRD and NCP. The error bars
represent standard deviations (1δ) over the averaging period of calculated SOA yields (b)
Calculated average SOA productions for C8-C21 alkanes in PRD and NCP. The error bars
represent standard deviations (1δ) over the averaging period of calculated SOA
generation (1δ) over the averaging period of calculated SOA
generation (1δ) over the averaging period of calculated SOA
generation (1δ) over the averaging period of calculated SOA
generation (1δ) over the averaging period of calculated SOA

- 948 **Technical corrections:**
- 949 P4 L80: should be "is responsive" P4 L82: Not PTR-MS. See above. P4 L88: "in more
- 950 detail." P5 L102: See above. Not PTR-MS P5 L110: "mass resolving power" instead of
- 951 resolution P5 L111: "source" P15 L343: "While: ::" this sentence needs to be re-written

- 952 for grammatical clarity
- 953 Many small language clarity issues throughout the manuscript.
- 954 Response: Thanks. We corrected all these technical issues and checked the language clarity
- 955 issues throughout the manuscript.

Response to anonymous referee #5:

Wang et al. report on measurements of higher alkanes made with a PTR-ToF-MS in an urban 958 and rural location in China. They find relatively abundant quantities of higher alkanes and, 959 based on estimates of OH concentrations and SOA mass yields, argue that they also contribute 960 meaningfully to SOA formation in both regions. Measuring the chemical composition and 961 complexity of VOCs in the atmosphere is an important step in studying the potential of those 962 VOCs to form SOA and identify the sources of fine particulate matter. The study is well 963 964 motivated and mostly well written (some detailed comment about technical communication can be found below). I am not overtly familiar with the online mass spectrometry measurements 965 so the editor should rely on the other reviews to make a judgement about that. The data analysis 966 and modeling sections require some additional detail to better communicate the inputs and 967 assumptions. I generally favor publication of this work in ACP but after the authors have had 968 a chance to respond to my comments. 969

970 Response: Thank you very much for your valuable comments and suggestions on our
971 manuscript. These comments are all helpful for improving our article. All the authors have
972 seriously discussed about all these comments. According to your comments, we have tried best
973 to modify our manuscript to meet with the requirements for the publication in this journal.
974 Point-by-point responses to the comments are listed below.

975 **Comments:**

976 1. Line 2: After reviewing the manuscript, I did not find that the word 'significant' was 977 appropriate in the title. The authors find that alkanes are probably as important as aromatics 978 and biogenic VOCs but the model currently still underestimates the total SOA production in 979 both studied regions. The word 'significant' could be misconstrued to mean that alkanes 980 explain the majority of the SOA in urban and rural china.

981 Response: We replaced the word "significant" with "important".

2. Lines 47-53: This introduction to SOA modeling is not well described and does accurately
represent the historical approaches used to model SOA formation. For instance, precursor
lumping has been used prior to the volatility basis set.

Response: We modified this part of introduction as follows. "A volatility basis set (VBS)
model was developed to advance SOA modeling by improving the modeling of further
multigenerational oxidation processes and incorporating numerous, yet unidentified,

- 988 low-volatility precursors of SOA, which substantially improved the agreement between
 989 SOA simulations and observations."
- 3. Lines 58-61: It might be worthwhile to talk about the differences in the various SOAprecursor classes between gasoline and diesel exhaust.
- 992 Response: We included the introduction about the differences in the various SOA precursor993 classes between gasoline and diesel exhaust on line 69-71 of page 4 in the revised manuscript.
- "Based on vehicle exhaust tests, higher alkanes were found to contribute ~37% to diesel
 exhaust-derived SOA and ~0.8% to gasoline exhaust-derived SOA, respectively (Gentner
 et al., 2012)."
- 4. Lines 61-63, lines 314-316: A recent study by Akherati et al. (ACP, 2019) also modeled
 SOA formation from higher alkanes but did not find them to be as important when compared
 to aromatics, especially after accounting for the effects of vapor wall losses (see #12).
- 1000 Response: The estimation of SOA formation from higher alkanes in this study is calculated based on the field measurements in the urban and rual regions of China, which is different from 1001 1002 the model study in Akherati et al. (ACP, 2019). According to the calculated SOA yield biases for a variety of VOCs when considering the potential influence of vapour wall losses (Table 1003 S4), the literature reported SOA yields are low by factors of $\sim 1.1-2.2$ for the high NO_x 1004 conditions (Zhang et al., 2014). This suggests the SOA calculations in our study might be 1005 correspondingly underestimated. However, we still can find the importance of higher alkanes 1006 to SOA formation in our study. 1007
- 1008 Table S4. Average biases in SOA yields due to vapour wall losses for various VOCs under
- 1009

high-NO_x conditions from Zhang et al. 2014.

VOC	R _{wall}
Benzene	1.25±0.1
Toluene	1.13±0.06
m-xylene	1.2±0.1
Naphthalene	1.2±0.1
Isoprene	2.2±0.5
α-pinene	1.3±0.1
n-dodecane	1.16±0.08
2-methylundecane	1.4±0.2

- 1010 5. Line 80: 'is response' should be 'responds'.
- 1011 Response: We corrected it to "responsive".
- 1012 6. Line105: grammatical mistake in 'makes possible of quantitative of alkenes'.
- 1013 Response: We corrected this sentence.
- 1014 7. Lines 147-148: grammatical mistake in 'group affect little on the degrees of fragmentation1015 for product ions'.
- 1016 Response: We corrected this sentence.
- 1017 8. Lines 193-195: I don't see why the PTR measurement should be higher than the GC.

1018 Shouldn't the GC measurement be close to the PTR if the unresolved mixture is accounted for?

1019 Response: In this study, the higher alkanes standards used by online GC-MS are just linear

1020 alkanes (*n*-alkanes) without branched ones. Therefore, the quantification of branched alkanes

1021 by GC-MS is not available in this study. While, the higher alkanes measured by PTR-ToF-MS

- 1022 include linear and branched isomers for each carbon number.
- 1023 9. Line 212: Should be 'This decreasing pattern'.
- 1024 Response: We corrected this sentence.

1025 10. Lines 239-242: It isn't clear to me why different pairs of species are used to estimate OH 1026 concentrations for the anthropogenic and biogenic species separately. Can the equation for the 1027 OH estimation be provided? Also, is there confounding in the OH estimation associated with 1028 emissions being added to the air parcel while the selected pairs are oxidized?

Response: We included the details about the calculations of OH exposure in the revised 1029 1030 supporting information. During the PRD campaign, since biogenic species (i.e. isoprene and monoterpenes) have different emission sources and patterns from the anthropogenic species, 1031 1032 the photochemical oxidation processes are quite different in these species. Therefore, we use different methods to calculate the OH exposures for anthropogenic species and biogenic 1033 1034 species during PRD campaign. For the anthropogenic species, the observed ratio between m+pxylene and ethylbenzene was used to estimate the OH exposure ($[OH] \times \Delta t$) by Roberts et al. 1035 1036 (1984):

1037
$$[OH] \times \Delta t = \frac{1}{k_{m+p-xylene} - k_{ethylbenzene}} \times [ln(\frac{m+p-xylene}{ethylbenzene})_{t=0} - ln(\frac{m+p-xylene}{ethylbenzene})_t]$$

1038 Where the initial emission ratios of m+p-xylene/ethylbenzene were estimated according 1039 to the correlation of m+p-xylene with ethylbenzene during campaigns. The ratio of 4 and 1.5 1040 were used in the PRD campaign and the NCP campaign, respectively (Figure S30).



1041

Figure S30. Correlation of m+p-xylene with ethylbenzene in PRD (a) and NCP (b). The dashed
lines in both graphs indicate the estimated initial mission ratio of m+p-xylene/ethylbenzene.

The OH exposure of biogenic emissions are estimeted based on isoprene chemistry for the PRD campaign. The calculation method can be found in Roberts et al. (2006). Isoprene are mainly photo-oxidized through the reactions with OH radical in the atmosphere and its primary first-generation reaction products are formaldehyde, MVK and MACR (Apel et al., 2002). The reaction processes of isoprene oxidized by OH radical are mainly as follows:

1049 Isoprene +
$$OH \rightarrow 0.63HCHO + 0.32MVK + 0.23MACR$$
 $k_1 = 1.0 \times 10^{-10} \text{ cm}^{-3} \text{ s}^{-1}$ (Eq. S1)

1050
$$MVK + OH \rightarrow Products \quad k_2 = 1.9 \times 10^{-11} \text{ cm}^{-3} \text{ s}^{-1} (Eq. S2)$$

1051
$$MACR + OH \rightarrow Products \quad k_3 = 3.3 \times 10^{-1.1} cm^{-3} s^{-1} (Eq. S3)$$

1052 where k₁, k₂, k₃ are the rate constants of the reactions. According to above reactions, the 1053 relationship between MVK/Isoprene, MACR/Isoprene, (MVK+MACR)/Isoprene and the 1054 reaction time Δt can be decribed as follows (Apel et al., 2002):

1055
$$\frac{MVK}{Isoprene} = \frac{0.32k_1}{k_2 - k_1} \left(1 - \exp((k_1 - k_2) [OH]\Delta t)\right) \quad (Eq.S \ 4)$$

1056
$$\frac{MACR}{Isoprene} = \frac{0.23k_1}{k_3 - k_1} \left(1 - \exp((k_1 - k_3) [OH]\Delta t)\right) \quad (Eq. S5)$$

1057 $\frac{MVK+MACR}{Isoprene} = \frac{0.32k_1}{k_2 - k_1} \left(1 - \exp((k_1 - k_2) [OH]\Delta t)) + \frac{0.23k_1}{k_3 - k_1} \left(1 - \exp((k_1 - k_3) [OH]\Delta t)\right) \quad (Eq.$ 1058 S6)

1059 where $\frac{MVK+MACR}{Isoprene}$ can be derived from the measurements by PTR-ToF-MS. Then the OH 1060 exposure ($[OH]\Delta t$) of isoprenoids can be obtaind from Eq. S6.

1061 The OH exposure values in this study represent the average "photochemical age" that 1062 are calculated by some properly chosen VOCs ratios. Parrish et al. (2007) compared the 1063 calculated photochemical age of different NMHCs species with the transport time calculated 1064 by a back trajectory model, and concluded that the calculated photochemical age could better 1065 describe the average transport time of VOCs species by selecting suitable hydrocarbon species 1066 pairs.

1067 11. Lines 242-246: If the isoprene and monoterpenes are of anthropogenic origin, shouldn't thebiogenic VOCs be oxidized based on the OH determined from the anthropogenic VOCs?

1069 Response: In this study, we think the isoprene and monoterpenes are of anthropogenic origin 1070 in NCP. Thus, we use the OH exposure estimated from the anthropogenic VOCs for 1071 isoprenoids in NCP campaign. While, because the isoprenoids in PRD campaign are of 1072 biogenic origin, we estimate the OH exposure of isoprenoids based on isoprene chemistry 1073 method as mentioned above.

1074 12. Lines 253-onwards: While I commend the authors for relying on a lot of historical data to determine SOA parameterizations for the VOCs measured in this work, the toluene 1075 parameterizations do not use the more recent work of Zhang et al. (PNAS, 2014) that account 1076 for the influence of vapor wall losses in chambers. On a related note, were the SOA mass yields 1077 adjusted for vapor wall losses? Alternatively, some of the parameterizations can also be found 1078 1079 in Eluri et al. (ACP, 2018). Finally, how was the NOx-dependence on SOA modeled? No NOx data at either of those sites was presented, nor any argument was made for the use of high NOx 1080 1081 SOA parameters.

Response: Thank you for the nice suggestion. We agree with that the vapour wall losses might 1082 influence the calculation of SOA formation by using SOA yields from previous chamber 1083 1084 studies where the wall losses were rarely corrected. Cappa and Wilson (2012) developed a Statistical Oxidation Model (SOM) to simulate the multigenerational chemistry and gas-1085 1086 particle partitioning of organic compounds. This SOM has been used to interpret chamber experiments to examine the influence of chamber-based vapour wall losses on yields of SOA 1087 1088 (Zhang et al., 2014). According to the calculated SOA yield biases for a variety of VOCs when considering the potential influence of vapour wall losses (Table 1 in this reference), the 1089

1090 literature reported SOA yields are low by factors of \sim 1.1-2.2 for the high NO_x conditions 1091 (Zhang et al., 2014). This suggests the SOA calculations in our study might be correspondingly 1092 underestimated. In the revised manuscript on line 370-374 of page 16, we added sentences to 1093 mention this influence as follows:

1094 "The influence of chamber-based vapour wall losses on SOA yields was examined in

1095 previous studies (Zhang et al., 2014) and the results show that the literature reported

1096 SOA yields are low by factors of ~1.1-2.2 for the high NO_x conditions (Table S4). This

1097 suggests that the SOA estimations in this study might be correspondingly underestimated."

1098 Table S4. Average biases in SOA yields due to vapour wall losses for various VOCs under

1099 high-NO_x conditions from Zhang et al. 2014.

VOC	R _{wall}		
Benzene	1.25±0.1		
Toluene	1.13±0.06		
m-xylene	1.2±0.1		
Naphthalene	1.2±0.1		
Isoprene	2.2±0.5		
α-pinene	1.3±0.1		
n-dodecane	1.16±0.08		
2-methylundecane	1.4±0.2		

1100

We included the NO_x data in the revised supporting information (Figure S22). The average concentrations of NO_x are 42.6±33.7 ppb in PRD and 81.7±57.0 ppb in NCP, respectively. Under such high NO_x levels, the RO₂ mainly reacts with NO (Bahreini et al., 2009). Thus, we used the high SOA parameters in this study.

1105 We added sentences on line 321-324 of page 14 in the revised manuscript as follows:

"SOA yields under high NO_x conditions are used in this study, as relatively high NO_x
concentrations in PRD (42.6±33.7 ppb) and in NCP (81.7±57.0 ppb) (Figure S22) would
cause RO₂ radicals from organic compounds mainly reacting with NO (Bahreini et al.,
2009)."



1112 Figure S22. Time series of NO_x during the PRD (a) and the NCP (b) campaigns, respectively.

1113 13. Lines 299-300: The sentence reads as if isoprene and monoterpenes are emitted in vehicle1114 exhaust. Please correct.

- 1115 Response: We deleted "including vehicle exhausts and biomass combustions in this region".
- 1116 The sentence is modified to "As mentioned above, we attribute these isoprene and
- 1117 monoterpenes to anthropogenic emissions in this region."
- 1118 14. Line 312: Can you describe how this is 'independent'?
- 1119 Response: We replaced the "independent" with "another".
- 1120 15. Line 332: The previous literature that has examined alkanes in more detail (Gentner et al.,
- 1121 2012; Zhao et al., 2015,2016; Drozd et al., 2019) have found a strong contribution of cyclic
- alkanes, especially in vehicular exhaust. Were these specifically measured/estimated in this
- 1123 work and can the authors comment on their role in contributing to SOA formation?
- 1124 Response: In this study, we didn't estimate the contribution of cyclic alkanes to SOA formation.
- 1125 We are trying to establish the quantification method of these cyclic alkanes.

1126 Reference:

Amador, O., Misztal, P., Weber, R., Worton, D., Zhang, H., Drozd, G., and Goldstein, A.:
Sensitive detection of n -alkanes using a mixed ionization mode proton-transfer-reaction mass
spectrometer, Atmospheric Measurement Techniques, 9, 5315-5329, 10.5194/amt-9-53152016, 2016.

- 1131 Apel, E. C., Riemer, D. D., Hills, A., Baugh, W., Orlando, J., Faloona, I., Tan, D., Brune, W.,
- 1132 Lamb, B., Westberg, H., Carroll, M. A., Thornberry, T., and Geron, C. D.: Measurement and
- 1133 interpretation of isoprene fluxes and isoprene, methacrolein, and methyl vinyl ketone mixing
- 1134 ratios at the PROPHET site during the 1998 Intensive, Journal of Geophysical Research:
- 1135 Atmospheres, 107, ACH 7-1-ACH 7-15, 10.1029/2000JD000225, 2002.

Blake, R. S., Monks, P. S. and Ellis, A. M. Proton-Transfer Reaction Mass Spectrometry.
Chemical Reviews 2009, 109(3): 861-896.

- 1138 Bahreini, R., B. Ervens, A. M. Middlebrook, C. Warneke, J. A. de Gouw, P. F. DeCarlo, J. L.
- 1139 Jimenez, C. A. Brock, J. A. Neuman, T. B. Ryerson, H. Stark, E. Atlas, J. Brioude, A. Fried, J.
- 1140 S. Holloway, J. Peischl, D. Richter, J. Walega, P. Weibring, A. G. Wollny and F. C. Fehsenfeld.
- 1141 Organic aerosol formation in urban and industrial plumes near Houston and Dallas, Texas.
- Journal of Geophysical Research: Atmospheres 114(D7), 10.1029/2008JD011493, 2009.
- Cappa, C. D. and Wilson, K. R.: Multi-generation gas-phase oxidation, equilibrium partitioning,
 and the formation and evolution of secondary organic aerosol, Atmos. Chem. Phys., 12, 95059528,doi:10.5194/acp-12-9505-2012, 2012.
- Chan, A. W. H., Isaacman, G., Wilson, K. R., Worton, D. R., Ruehl, C. R., Nah, T., Gentner,
 D. R., Dallmann, T. R., Kirchstetter, T. W., Harley, R. A., Gilman, J. B., Kuster, W. C.,
 deGouw, J. A., Offenberg, J. H., Kleindienst, T. E., Lin, Y. H., Rubitschun, C. L., Surratt, J.
 D., Hayes, P. L., Jimenez, J. L., and Goldstein, A. H.: Detailed chemical characterization of
 unresolved complex mixtures in atmospheric organics: Insights into emission sources,
 atmospheric processing, and secondary organic aerosol formation, Journal of Geophysical
 Research-Atmospheres, 118, 6783-6796, 10.1002/jgrd.50533, 2013.
- Gentner, D. R., Isaacman, G., Worton, D. R., Chan, A. W. H., Dallmann, T. R., Davis, L., Liu,
 S., Day, D.A., Russell, L. M., Wilson, K. R., Weber, R., Guha, A., Harley, R. A., and Goldstein,
- 1155 A. H.: Elucidating secondary organic aerosol from diesel and gasoline vehicles through
- detailed characterization of organic carbon emissions, Proceedings of the National Academy

- of Sciences of the United States of America, 109, 18318-18323, 10.1073/pnas.1212272109,
 2012.
- Hellén, H., Tykkä, T., and Hakola, H.: Importance of monoterpenes and isoprene in urban air
 in northern Europe, Atmospheric Environment, 59, 59-66, 10.1016/j.atmosenv.2012.04.049,
 2012.
- 1162 Inomata, S., Tanimoto, H., and Yamada, H.: Mass Spectrometric Detection of Alkanes Using
- 1163 NO⁺ Chemical Ionization in Proton-transfer-reaction Plus Switchable Reagent Ion Mass
- 1164 Spectrometry, Chemistry Letters, 43, 538-540, 10.1246/cl.131105, 2013.
- 1165 Jordan, A., Haidacher, S., Hanel, G., Hartungen, E., Herbig, J., Maerk, L., Schottkowsky, R.,
- 1166 Seehauser, H., Sulzer, P., and Maerk, T. D.: An online ultra-high sensitivity Proton-transfer-
- 1167 reaction mass-spectrometer combined with switchable reagent ion capability (PTR+SRI-MS),
- **1168** International Journal of Mass Spectrometry, 286, 32-38, 10.1016/j.ijms.2009.06.006, 2009.
- 1169 Kuang, Y., He, Y., Xu, W., Yuan, B., Zhang, G., Ma, Z., Wu, C., Wang, C., Wang, S., Zhang,
- 1170 S., Tao, J., Ma, N., Su, H., Cheng, Y., Shao, M., and Sun, Y.: Photochemical Aqueous-Phase
- 1171 Reactions Induce Rapid Daytime Formation of Oxygenated Organic Aerosol on the North
- 1172 China Plain, Environmental Science & Technology, 10.1021/acs.est.9b06836, 2020.
- Midey, A. J., Arnold, S. T. and Viggiano, A. A.. Reactions of H₃O⁺(H₂O)_n with Formaldehyde
 and Acetaldehyde. The Journal of Physical Chemistry A, 2000, 104(12): 2706-2709.
- 1175 Roberts, J. M., Fehsenfeld, F. C., Liu, S. C., Bollinger, M. J., Hahn, C., Albritton, D. L., and
- 1176 Sievers, R. E.: Measurements of aromatic hydrocarbon ratios and NOx concentrations in the
- 1177 rural troposphere: Observation of air mass photochemical aging and NOx removal,
- 1178 Atmospheric Environment (1967), 18, 2421-2432, https://doi.org/10.1016/00041179 6981(84)90012-X, 1984.
- 1180 Parrish, D. D., C. J. Hahn, E. J. Williams, R. B. Norton, F. C. Fehsenfeld, H. B. Singh, J. D.
- 1181 Shetter, B. W. Gandrud and B. A. Ridley. Indications of photochemical histories of Pacific air
- 1182 masses from measurements of atmospheric trace species at Point Arena, California. Journal of
- **1183** Geophysical Research: Atmospheres, 1992, 97(D14): 15883-15901.
- 1184 Parrish, D. D., A. Stohl, C. Forster, E. L. Atlas, D. R. Blake, P. D. Goldan, W. C. Kuster and
- 1185 J. A. de Gouw. Effects of mixing on evolution of hydrocarbon ratios in the troposphere. Journal
- 1186 of Geophysical Research: Atmospheres 112(D10), 2007.

- Roberts, J., Marchewka, M., Bertman, S., Goldan, P., Kuster, W., de Gouw, J., warneke, C., 1187 Williams, E., Lerner, B., Murphy, P., Apel, E., and Fehsenfeld, F.: Analysis of the isoprene 1188 chemistry observed during the New England Air Quality Study (NEAQS) 2002 Intensive 1189 Experiment, Journal of Geophysical Research-Atmospheres, 111, D23S12, 1190 10.1029/2006JD007570, 2006. 1191
- 1192 Tkacik, D. S., Presto, A. A., Donahue, N. M., and Robinson, A. L.: Secondary Organic Aerosol
- 1193 Formation from Intermediate-Volatility Organic Compounds: Cyclic, Linear, and Branched
- Alkanes, Environmental Science & Technology, 46, 8773-8781, 10.1021/es301112c, 2012.
- 1195 Wu, C., Wang, C., Wang, S., Wang, W., Yuan, B., Qi, J., Wang, B., Wang, H., Wang, C., Song,
- 1196 W., Wang, X., Hu, W., Lou, S., Ye, C., Peng, Y., Wang, Z., Huangfu, Y., Xie, Y., Zhu, M.,
- 1197 Zheng, J., Wang, X., Jiang, B., Zhang, Z., and Shao, M.: Measurement Report: important
- 1198 contributions of oxygenated compounds to emissions and chemistry of VOCs in urban air,
- 1199 Atmos. Chem. Phys. Discuss., 2020, 1-37, 10.5194/acp-2020-152, 2020.
- 1200 Zhang, X., C. D. Cappa, S. H. Jathar, R. C. McVay, J. J. Ensberg, M. J. Kleeman and J. H.
- 1201 Seinfeld. Influence of vapor wall loss in laboratory chambers on yields of secondary organic
- aerosol. Proceedings of the National Academy of Sciences of the United States of America,
- **1203** 111(16): 5802-5807, 10.1073/pnas.1404727111, 2014.