Substantial changes of gaseous pollutants and chemical compositions in fine particles in North China Plain during COVID-19 lockdown period: anthropogenic vs meteorological influences

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Abstract

The rapid response to COVID-19 pandemic led to the unprecedented decreases of economic activities, thereby reducing the pollutant emissions. A random forest (RF) model was applied to determine the respective contributions of meteorology and anthropogenic emissions to the changes of air quality. The result suggested the strict lockdown measures significantly decreased primary components such as Cr (-67%) and Fe (-61%) in $PM_{2.5}$ ($p < 0.01$), whereas the higher relative humidity (RH) and NH_3 level, and the lower air temperature (T) remarkably enhanced the production of secondary aerosol including SO_4^{2-} (29%), NO_3^- (29%), and NH_4^+ (21%) ($p < 0.05$). Positive matrix factorization (PMF) result suggested that the contribution ratios of secondary formation (SF), industrial process (IP), biomass burning (BB), coal combustion (CC), and road dust (RD) changed from 36%, 27%, 21%, 12%, and 4% before COVID-19 outbreak to 44%, 20%, 20%,

9%, and 7%, respectively. The rapid increase of the contribution ratio derived from SF to $PM_{2.5}$ implied the intermittent haze events during COVID-19 period were characterized with secondary aerosol pollution, which was mainly contributed by the unfavorable meteorological conditions and high NH_3 level.

1. Introduction

In December 2019, a cluster of pneumonia cases with unknown etiology were firstly reported in Wuhan and quickly spread around the world (Wu et al., 2020). The continuous global outbreak of coronavirus disease (COVID-19), declared as a public health emergency of international concern by the World Health Organization, resulted in unprecedented public health responses in many countries including lockdown, travel restrictions, and quarantines (Griffiths and Woodyatt, 2020; Horowitz et al., 2020). On January 23, 2020, Chinese government imposed a lockdown in Wuhan and many surrounding cities in Hubei province in order to prevent the spread of epidemic. Afterwards, many similar measures including blocked roads, shutdown of factories, restricted citizen mobility, and checkpoints were soon extended to other cities throughout the entire country. During this period, energy production by coal-fired power plants only remained two thirds levels of the same periods in preceding years (Chang et al., 2020). Besides, the transport volume have been reduced by more than 70% due to the COVID-19 outbreak (Chang et al., 2020). These drastic government-enforced lockdown measures substantially decreased the pollutant emissions, and at least partly improved local air quality. Feng et al. (2020) confirmed that the COVID-19 lockdown have led to more than 70% reduction of NO_x emissions in many large cities over China. Correspondingly, the concentrations of $PM_{2.5}$ and NO_2 decreased by 35% and 60%, respectively (Shi and Brasseur, 2020). The natural experiment provided an unprecedented opportunity to explore

47 the potential for emission reduction and the corresponding response of air quality.

48 A growing body of studies assessed the response of $PM_{2.5}$ and gaseous pollutants to COVID-19
49 lockdown, and found these stringent restrictions resulted in the significant decreases of these
50 pollutant (e.g., $PM_{2.5}$, NO_2, and CO) concentrations (Miyazaki et al., 2020; Marlia et al., 2020).
51 However, some haze events still occurred during this period especially in East China. Huang et al.
52 (2020) employed the chemical transport models (CTMs) to infer that these extraordinary findings
53 might be attributable to enhanced secondary pollution. Understanding the formation mechanism of
54 puzzle haze events depending on CTMs alone might be not very robust, it was highly imperative to
55 perform more field observation to analyze the temporal variations of chemical compositions
56 especially the secondary ions (e.g., SO_4^{2-}, NO_3^-) in $PM_{2.5}$ before and after COVID-19 outbreak and
57 then to validate these inferences.

58 To date, only several field observations analyzed the temporal variations of chemical
59 components in fine particles during COVID-19 lockdown period. Chang et al. (2020) observed a
60 remarkably enhanced nitrate formation in Yangtze River Delta (YRD) counteracted the decreases
61 of primary components in fine particles, which was in good agreement with the modelling result
62 drawn by Huang et al. (2020). In contrast, Xu et al. (2020) found that the marked decreases of fine
63 particle concentrations in Lanzhou during COVID-19 lockdown period was mainly contributed by
64 the lower production rate for secondary aerosols. Under the condition of similar emission control
65 measures, the polarized conclusion might be associated with the local meteorology. He et al. (2017)
66 demonstrated that meteorology might explain more than 70% variances of daily average pollutant
67 levels over China during 2014-2015. Besides, Zhang et al. (2020a) also revealed that the release of
68 primary pollutants and the generation of reactive semi-volatile products partitioned between gas and

aerosol phases were strongly dependent on the temperature and relative humidity (RH). Thus, in order to accurately assess the effects of lockdown measures on air quality and to reveal the key driver of the haze paradox, it was necessary to isolate the contribution of meteorology. Unfortunately, up to date, the respective contributions of emission and meteorology to chemical compositions in $PM_{2.5}$ during COVID-19 period were not quantified yet in most pioneering studies (Chang et al., 2020; Huang et al., 2020; Xu et al., 2020). Moreover, the comparison of source contributions to chemical compositions between pre-lockdown and post-lockdown were scarcely performed. Such knowledge is critical to design effective $PM_{2.5}$ mitigation strategies in the near future.

As a heavily industrialized region, North China Plain (NCP) possesses many energy-intensive industries including coal-fired power plants, non-ferrous smelting industries, textiles, building materials, chemical engineering, and papermaking industries (Ren et al., 2011). Due to these intensive industrial emissions, NCP suffered from poor air quality and frequent aerosol pollution in the past decades (Zhang et al., 2018; Luo et al., 2017). Nevertheless, these strict lockdown measures during COVID-19 period inevitably led to the dramatic decreases of industrial emissions, and thus a study about the response of chemical compositions to emission reduction in the heavy-pollution city might be more sensible.

Here, we selected the typical industrial city (Tangshan) in NCP to determine the concentrations of gaseous pollutants and chemical compositions in $PM_{2.5}$ during January 1-March 31, 2020, and then to analyze their temporal variations before and after COVID-19 outbreak. Besides, a machine-learning approach was applied to separate the contributions of emission reduction and meteorology to the temporal variabilities of chemical compositions and gaseous pollutants. Finally, the source apportionment was performed based on the meteorology-normalized datasets to compare the source

difference for these pollutants before and after COVID-19 lockdown.

2. Materials and methods

2.1 Field observation

Hourly gaseous pollutants and $PM_{2.5}$ chemical compositions including water-soluble ions and trace elements were measured using on-line instruments during January 1-March 31, 2020 at a supersite in Tangshan. The supersite is located in a commercial region without short-distance industrial emissions (Figure 1). SO_2, NO_2, and CO concentrations were determined by the ultraviolet fluorescence analyzer (TEI, Model 43i from Thermo Fisher Scientific Inc., USA), chemiluminescence trace gas analyzer (TEI Model 42i from Thermo Fisher Scientific Inc., USA), and the correlation infrared absorption analyzer (TAPI, model: 300E, USA) (Li et al., 2017; Li et al., 2019). The mass concentration of $PM_{2.5}$ was determined using an oscillating balance analyzer (TH-2000Z, China) (Wang et al., 2014). The NH_3 concentration and water-soluble ions including sulfate (SO_4^{2-}), nitrate (NO_3^-), ammonium (NH_4^+), sodium ion (Na^+), and chloridion (Cl^-) were monitored with a Gas and Aerosol Collector combined with Ion Chromatography (GAC-IC, TH-PKU-303, China) (Wang et al., 2014; Zheng et al., 2019). OC and EC were measured using an OC/EC analyzer (Model RT-4, Sunset Laboratory Inc., Tigard, Oregon, USA). Nine trace elements including Hg, Pb, K, Ca, Cr, Cu, Fe, Ni, and Zn were determined by an online multi-element analyzer (Model Xact 625, Cooper Environment Service, USA). The quality assurance of SO_2, NO_2, CO, and $PM_{2.5}$ were conducted based on HJ 630-2011 specifications. For the quality assurance of NH_3 and water-soluble ions, the concentration gradients of anion and cation standard solutions were set based on the pollution levels of target species, and correlation coefficients of the calibration curve must be higher than 0.99. Besides, a standard sample was collected each day and the relative

standard deviation for the reproducibility test must be less than 5%. The online device agreed well with the result determined by filter sampling coupled with Inductively Coupled Plasma Mass Spectrometry (ICP-MS) and Inductively Coupled Plasma-Atomic Emission Spectroscopy (ICP-AES).

2.2 Deweathered model development

The air pollutants were influenced by the combined effects of meteorological conditions and emissions. In order to quantify the contributions of anthropogenic emissions, the impacts of meteorological conditions should be removed. In our study, a random forest (RF) approach was employed to serve as the site-specific modeling platform (Chen et al., 2018). All of gaseous pollutants and chemical compositions in $PM_{2.5}$ were regarded as the dependent variables. The meteorological parameters including wind speed (WS), wind direction (WD), air temperature (T), relative humidity (RH), precipitation (Prec), and air pressure (P), and time predictors (year, day of year (DOY), day of week (DOW), hour) served as the independent variables. The original dataset was randomly classified into a training dataset (90% of input dataset) for developing the RF model and the remained one was treated as the test dataset. After the building of the RF model, the deweathered technique was applied to predict the air pollutant level at a specific time point (e.g., 2020/01/01 12:00). The differences of original pollutant concentrations and deweathered pollutant concentrations were regarded as the concentrations contributed by meteorology. Some statistical indicators including R^2 value, RMSE, and MAE were regarded as the major criteria to evaluate the modelling performance. In our study, the RF model with the R^2 value lower than 0.50 was treated as the unreliable result and cannot reflect the impacts of emission and meteorology on air pollutants accurately because more than 50% variability of the training model cannot be appropriately

explained. After the model evaluation, only the species with the cross-validation R^2 values larger than 0.50 were selected to assess the respective contributions of emission and meteorology to their concentrations.

2.3 Source apportionment

Positive matrix factorization (PMF 5.0) model version was used to perform the $PM_{2.5}$ source apportionment. The deweathered gaseous pollutants and chemical compositions in $PM_{2.5}$ were incorporated into the model. The objective of PMF is to resolve the issues of chemical mass balance between measured concentration of each species and its source contributions by decomposing the input matrix into factor contribution and factor profile. The detailed equation is shown in Eq. (1)-(2). Briefly, the basic principle of PMF is to calculate the least object function Q when the g_{ik} must be a positive-definite matrix based on Eq. (2) (Chen et al., 2014; Sharma et al., 2016).

$$x_{ij} = \sum_{k=1}^{p} g_{ik} f_{kj} + e_{ij} \quad (1)$$

$$Q = \sum_{i=1}^{n} \sum_{j=1}^{m} \left[\frac{x_{ij} - \sum_{k=1}^{p} g_{ik} f_{kj}}{u_{ij}} \right]^2 \quad (2)$$

where x_{ij} and e_{ij} represent the concentration and uncertainty of jth species, respectively. g_{ik} represents the contribution ratio of kth source to ith sample, f_{kj} represents the ratio of jth species in kth source, and e_{ij} indicates the residual of jth species in the i sample. The uncertainties associated with factor profiles were evaluated using three error calculation methods including bootstraps (BS) method, displacement (DISP) analysis, and the combination method of DISP and BS (BS-DISP). For the BS method, 100 runs were performed and the result has been believed to be valid since all of the factors showed a mapping of above 90%. DISP analysis also confirmed that the solution was considered to be stable because the observed drop in the Q value was less than 0.1% and no factor

swap occurred. For the BS-DISP analysis, the solution has been verified to be useful because the observed drop in the Q value was less than 0.5%. Furthermore, both of the results from BS and BS-DISP did not suggest any asymmetry or rotational ambiguity for all of the factors (Manousakas et al., Brown et al., 2015).

3. **Results and discussion**

3.1 The concentration changes of gaseous pollutants and $PM_{2.5}$ chemical compositions

Figure 2, Figure 3, Figure 4, and Figure 5 show the temporal variations of gaseous pollutants and chemical compositions in $PM_{2.5}$ from January 1-March 31, which could be divided into two periods including before and after COVID-19 outbreak. In this study, January 23 was regarded as the breakpoint because China's government imposed a lockdown in Wuhan and surrounding cities. Before COVID-19 outbreak, the average observed concentrations of SO_2, NO_2, CO, 8-h O_3, and NH_3 during January 1-22 were 34 μg/m^3, 64 μg/m^3, 2.0 mg/m^3, 19 μg/m^3, and 14 ppb, respectively. After COVID-19 lockdown, the mean concentrations of these gaseous pollutants changed to 25 μg/m^3, 39 μg/m^3, 1.6 mg/m^3, 49 μg/m^3, and 18 ppb, respectively. Overall, CO, SO_2, and NO_2 concentrations decreased by 18%, 27%, and 39%, respectively ($p < 0.05$). However, the NH_3 and O_3 concentration increased by 35% ($p < 0.05$) and 160% ($p < 0.01$).

As shown in Figure 2, the chemical compositions in $PM_{2.5}$ also showed dramatic changes during January 1-March 31 due to the impact of COVID-19 lockdown. The observed SO_4^{2-}, $PM_{2.5}$, Na^+, and Cl^- concentrations decreased by 6% ($p > 0.05$), 13% ($p > 0.05$), 29% ($p < 0.05$), and 48% ($p < 0.01$), respectively, while observed NO_3^- (2%) and NH_4^+ (7%) levels showed slight increases ($p > 0.05$). In Shanghai, Chen et al. (2020) revealed that SO_4^{2-}, and NH_4^+ concentrations displayed significant decreases after COVID-19 outbreak due to the obvious decreases of precursor

concentrations (e.g., SO_2, NO_x). However, both of observed NO_3^- and NH_4^+ concentrations in Tangshan even showed slight increases though the NO_2 concentration suffered from remarkable decrease. It was assumed that the adverse meteorological conditions might be beneficial to the pollutant accumulation (Zheng et al., 2019; Zhang et al., 2019b). Besides, the concentrations of nine trace elements were also determined. The observed values of Fe (25%), Ca (39%), Pb (41%), Cr (41%), and Zn (48%) suffered from dramatic decreases ($p < 0.05$), while the K (0%), Ni (1%), and Hg (8%) concentrations still displayed slight increases ($p > 0.05$). As a whole, the temporal variability of these elements in Tangshan before and after COVID-19 lockdown was in agreement with the result in Beijing (He et al., 2017). However, the K concentration in Beijing showed rapid decrease after COVID-19 outbreak, which was not in coincident with our study (He et al., 2017). It suggested that the slight increase of K in Tangshan might be linked with the unfavorable meteorological conditions (He et al., 2017). The observed concentrations of OC (-19%) and EC (-39%) also suffered from rapid decreases after COVID-19 lockdown (Figure 4) ($p < 0.05$), which was in good agreement with the sea-salt ions (e.g., Na^+, Cl^-) and most trace elements (e.g., Zn, Pb).

3.2 The impact of emission reduction on air quality

Although the observed concentrations of air pollutants can be applied to analyze the impact of COVID-19 lockdown, the role of emission reduction on air quality might be not clearly revealed because the meteorological factors were also important variables influencing the air pollutant concentrations. In order to accurately reflect the response of air quality to emission reduction during COVID-19 lockdown period, the meteorological conditions were isolated by machine-learning model. In our study, we developed a random forest model to remove the impacts of meteorological conditions on air pollutants. Based on the result in Figure S1, Figure S2, Figure S3, and Figure S4,

200 the RF models for all of the species showed the better performance because all of the R^2 values were
201 higher than 0.50 and the slope of all of the fitting curve were also close to the R^2 values. Thus, we
202 believed that the developed model was robust to remove the impact of meteorological conditions.
203 The deweathered concentrations of gaseous pollutants and chemical compositions in $PM_{2.5}$ are
204 depicted in Figure 2, Figure 3, Figure 4, and Figure 5. Compared with the period before COVID-
205 19, the deweathered NH_3, SO_2, CO, and NO_2 concentrations decreased by 27%, 31%, 32%, and 42%
206 after COVID-19 lockdown period outbreak, respectively ($p < 0.05$), while the deweathered 8-h O_3
207 concentration increased by 80% ($p < 0.01$). Meanwhile, the normalized-meteorology NH_4^+, NO_3^-,
208 SO_4^{2-}, Cl^-, $PM_{2.5}$, and Na^+ and concentrations decreased by 14%, 27%, 35%, 35%, 38%, and 47%,
209 respectively. For trace elements, deweathered Cu, K, Ni, Ca, Pb, Fe, Cr, and Zn levels reduced by
210 15%, 23%, 27%, 54%, 59%, 61%, 67%, and 69%, respectively ($p < 0.05$). Nevertheless, the
211 deweathered Hg concentration still kept stable increase by the rate of 6% compared with the period
212 before COVID-19 outbreak ($p > 0.05$).
213 The deweathered concentrations for most of the pollutants showed significant decreases after
214 COVID-19 outbreak compared with the period before COVID-19 (Figure 2, Figure 3, Figure 4,
215 Figure 5). It was assumed that many cities proposed the lockdown measures, which significantly
216 minimized industrial, transportation, and commercial activities. Among all of the pollutants, the
217 deweathered Zn, Cr, Fe, Pb, and Ca experienced more than 50% decrease rates due to the lockdown
218 measures. It was well known that Zn, Cr, and Fe originated mainly from metallurgical industry (Sun
219 et al., 2018; Zhu et al., 2018), while Pb might be derived from coal-fired power plants (Cui et al.,
220 2019; Meng et al., 2020). During the COVID-19 outbreak, most of the industries have been shut
221 down and energy production by coal-fired power plants was reduced by one third (Chang et al.,

2020). Based on the adjustment factor estimated by Doumbia et al. (2020), the contributions of industrial activity and power sector have decreased by 40% after COVID-19 outbreak, which was close to the decrease ratios of Zn, Cr, Fe, and Pb concentrations. It should be noted that the deweathered Ca concentration also decreased by more than 50%. It was well documented that the Ca was often associated with the dust resuspension (Chang et al., 2018). In fact, the Ca was known as one of the most abundant elements in the upper continental crust, which likely originated from the fugitive dust (Chang et al., 2018; Shen et al., 2016). More than 70% reduction of vehicle transportation and domestic flights facilitated the rapid decrease of Ca concentration (Chang et al., 2020). Although the observed K concentration did not show marked decrease after the COVID-19 lockdown, the deweathered K level suffered from rapid decrease (-22%) ($p < 0.05$). It was widely acknowledged that K was considered to be a key fingerprint of biomass burning (Zheng et al., 2020a), and thus the result suggested that the open biomass burning was also restricted during the period. Both of the deweathered concentrations of OC (-22%) and EC (-45%) also experienced remarkable decreases. In our study, both of OC and EC concentrations showed significant correlation with K level ($p < 0.05$), indicating that the restriction of biomass burning also led to the decreases of OC and EC. Besides, $PM_{2.5}$ and some water-soluble ions including deweathered SO_4^{2-} and NO_3^- concentrations experienced marked decreases after COVID-19 lockdown, which was in good agreement with their gaseous precursors. It might be attributable to the rapid decreases of precursor emissions. Zheng et al. (2020b) verified that the SO_2 emission in the industrial sector and NO_x emission in the transportation sector in Hebei province have decreased by 19% and 13%, respectively. The deweathered Na^+ concentration showed the rapid decrease after COVID-19 lockdown, which suggested that the Na^+ in the $PM_{2.5}$ of Tangshan was probably derived from waste

incineration rather than sea-salt aerosol (Deshmukh et al., 2016).

Although most of pollutant concentrations suffered from remarkable decreases, the decrease ratios of deweathered NH_3 and NH_4^+ concentrations after COVID-19 outbreak were far lower than those of many other gaseous pollutants and water-soluble ions. It was attributable to the fact that ambient NH_3 was mainly sourced from the fertilizer application and livestock, which did not show significant decrease during the COVID-19 period (Kang et al., 2016; Zheng et al., 2020b; Doumbia et al., 2020). Although the transportation volume suffered from dramatic decrease, the contribution of transportation to NH_3 was generally less than 5% (Kang et al., 2016). Furthermore, the contribution of urban waste source slightly increased after COVID-19 outbreak, offsetting the effect of traffic outage (Zhang et al., 2020b). Besides, it should be noted that the normalized-meteorology 8-h O_3 and Hg concentrations still remained the stable increase. Liu et al. (2020) have confirmed that uncoordinated decreases of NO_x and VOCs emissions (decrease ratio: NO_x > VOCs) dominated the 8-h O_3 increase in urban areas because most of urban areas belonged to VOC-limited region. Besides, the excessive decrease of $PM_{2.5}$ from primary emission significantly increased the HO_2 radical concentration on the surface of aerosol, thereby promoting the O_3 formation (Shi and Brasseur, 2020). The minor increase of deweathered Hg level was attributable to that the coal combustion for domestic heating was not restricted during the COVID-19 lockdown period (Zhou et al., 2018). Based on the updated global anthropogenic emission adjustment factor during COVID-19, the contribution of residential sector to air pollutants did not decrease after COVID-19 lockdown (Doumbia et al., 2020).

3.3 The role of meteorology and potential chemical reactions on air quality

Compared with the observed values, the deweathered concentrations of most pollutants were

significantly reduced. Meanwhile, the deweathered decrease ratios of pollutants were significantly higher than those of observed values (Figure 6). The result suggested the meteorology conditions during the COVID-19 lockdown period were not favorable to the pollutant dispersion, as evidenced by some recent studies (Chang et al., 2020; Huang et al., 2020). In our study, six meteorological parameters including WS, WD, T, RH, Prec, and P have been integrated into the random forest model to assess the response of each species to different meteorological variables. The variable importance of each meteorological to all of the species are shown in Figure 7, Figure 8, Figure 9, and Figure 10.

Among all of the gaseous pollutants, the meteorological conditions played the significantly positive roles on NH_3 (62%) and 8-h O_3 concentrations (80%) (Figure 6). As shown in Figure 7, T was the most important factor for the rapid elevation of NH_3 concentration after COVID-19 lockdown. It was assumed that the higher T enhanced the emissions of NH_3 from soil and urban wastes and promoted the volatilization of NH_3 from aerosol NH_4^+ pools (Zhang et al., 2020). In our study, the hourly mean air temperature have increased from 0℃ before COVID-19 outbreak to 5℃ after COVID-19 lockdown, which strongly supported the inference. For 8-h O_3 concentration (Figure 7), T was also treated as the most important variable. On the one hand, the higher T generally enhanced biogenic isoprene emissions, which was the most abundant biogenic VOC and showed the highest ozone formation potential (Liu and Wang, 2020). On the other hand, high T often increased chemical reaction rates and accelerated the O_3 formation (Shi et al., 2020). Besides, WS also played an important role on the 8-h O_3 concentration. Shi et al. (2020) have demonstrated that weaker winds often slowed down the advection and convection of NO_x and VOCs, which was beneficial to O_3 formation.

Besides, the contributions of meteorological conditions to some secondary ions (e.g., SO_4^{2-} (29%), NO_3^- (29%), and NH_4^+ (21%)) were remarkably higher than those to other ions and some trace elements, suggesting that the chemical reactions and formation pathways of these species were more sensitive to meteorological variations. Deshmukh et al. (2016) confirmed that the high RH promoted the aqueous-phase oxidation of SO_2 and the production of sulfate. Tian et al. (2019) also demonstrated that RH-dependent heterogeneous reactions significantly contributed to the sulfate generation and the high RH enhanced gas- to aqueous-phase dissolution of NH_3 and HNO_3.These pioneering experiments suggested that secondary aerosols were often formed under the condition of high RH. Very recently, Chang et al. (2020) observed that the nitrate concentration in YRD experienced unusual increase during COVID-19 period, while Xu et al. (2020) obtained the opposite result in Lanzhou. It was assumed that the persistent increase of T and decrease of RH in Lanzhou during this period was not beneficial to the generation of secondary aerosol, while the high RH in YRD significantly elevated local nitrate level. Although air temperature in Tangshan suffered from increase after COVID-19 lockdown, RH displayed rapid increase from 47% to 57% during this period. Moreover, the increased O_3 could promote the secondary aerosol formation and partially offset the decreased $PM_{2.5}$ compositions triggered by the primary emission reduction (Liu et al., 2020). Similar to secondary ions, both of OC and EC were also sensitive to RH. It was supposed that high RH could increase the secondary organic aerosol (SOA) levels, which accounted for the major fraction of OC (Zheng et al., 2020).

In addition, some trace elements such as Fe, Ni, and Cr were also significantly affected by the meteorological conditions. As shown in Figure 9, these element concentrations were mainly sensitive to WD. It was assumed that the neighboring industrial points including cement plants and

coal-fired power plants could influence the concentrations of trace elements via long/short-range transport, which was strongly dependent on WD. Following WD, RH was also an important factor for the variation of these trace elements. Under the condition of high RH, Fe and Cr could catalyze the heterogeneous generation of sulfate and nitrate on the mineral/soot surface (Hu et al., 2015).

Unlike the trace elements, water-soluble ions and OC were less sensitive to WD. Major water-soluble ions in $PM_{2.5}$ including SO_4^{2-}, NO_3^-, and NH_4^+ were mainly derived from secondary formation rather than the direct emission (Feng et al., 2020a; Zhang et al., 2020a), and thus they were not very sensitive to WD.

3.4 The enhanced secondary aerosol formation during COVID-19 lockdown period

The deweathered chemical compositions suggested that the sulfate and nitrate chemistry changed slightly after COVID-19 outbreak. The oxidation ratio of sulfate (SOR, the ratio of sulfate concentration and the sum of sulfate and SO_2 concentrations) decreased from 0.26 to 0.22, while the oxidation ratio of nitrate (NOR, the ratio of nitrate concentration and the sum of nitrate and NO_2 concentrations) increased from 0.22 to 0.25 (Table 1). The decreased SOR after COVID-19 outbreak indicated that the decrease rate of sulfate is higher than that of SO_2. In contrast, the increased NOR during COVID-19 lockdown period revealed that the decrease rate of nitrate is lower than that of NO_2. The increased NOR after COVID-19 outbreak suggested the consecutive nitrate production, though the NO_2 emission experienced tremendous reduction, which was in good agreement with the result observed by Chang et al. (2020). It was assumed that the persistently higher observed NH_3 concentration during this period promoted the ammonium nitrate formation though the lower NO_x emission (Zhang et al., 2020b), which also partially explained the abnormal increases of observed concentrations of secondary ions after COVID-19 outbreak. In general, NH_3 firstly tends to react

with H_2SO_4 to form ammonium sulfate, and then the excess NH_3 participated in the reaction with HNO_3 (Chen et al., 2019; Zhang et al., 2019a). However, sulfate concentration suffered from more dramatic decrease compared with SO_2, which might be associated with the aerosol acidity during COVID-19 lockdown period. The ratio of NH_4^+ and the sum of SO_4^{2-}, NO_3^-, and Cl^- named C/A was regarded as an indicator to reflect the aerosol acidity. In our study, the C/A value decreased from 0.33 to 0.28 after COVID-19 outbreak, implicating that the aerosol acidity even showed slight increase during the COVID-19 lockdown period. It was well known that the higher aerosol acidity might prohibit the conversion from SO_2 to sulfate (Liu et al., 2020; Shao et al., 2019), which yielded the lower SOR.

3.5 The impact of COVID-19 lockdown on source apportionment

The emission control measures inevitably triggered the variation of source apportionment (Liu et al., 2017; Meng et al., 2020). In the present study, Positive matrix factorization (PMF 5.0) was employed to identify the major sources of $PM_{2.5}$ in Tangshan before and after COVID-19 outbreak. About 3-9 factor solutions were examined, and a five-factor solution obtained the lowest Q (robust) and Q (true) values. Additionally, the PMF analysis and error diagnostics also suggested the result was robust (Table S2, Table S3, and Table S4).

The source apportionment profiles in pre-COVID and post-COVID resolved by PMF are depicted in Figure 11. For pre-COVID, the first factor contributed 36% to the total species. The factor was characterized with high levels of NH_4^+ (41%), SO_4^{2-} (35%), and NO_3^- (33%). SO_4^{2-} and NO_3^- were generally produced by oxidation of SO_2 and NO_x, respectively. The NH_4^+ was often formed through the heterogeneous reaction of NH_3 and sulfate or HNO_3. Thus, the factor was regarded as the secondary formation (SF). The second factor was characterized with high loadings

of Zn (47%), Cr (42%), Fe (42%), and Pb (31%). Cr and Fe were mainly originated from fuel combustion and metallurgical industry such as chrome plating and steel production(Liu et al., 2018a), while Pb and Zn was derived from the roasting, sintering and smelting process for the extraction of Pb/Zn ores (Wu et al., 2012). Therefore, the factor 2 was treated as the industrial process (IP) source. The predominant species in factor 3 included Na^+ (42%), K (40%), OC (35%), and EC (33%). K was often regarded as the fingerprint of biomass burning (BB) (Chen et al., 2017; Zheng et al., 2019b), whereas the Na^+ was generally regarded as the tracer of waste incineration (Alam et al., 2019; Durlak et al., 1997). Hence, the factor 3 was treated as the BB source. Tangshan suffered from remarkable increasing usage of biomass fuels for domestic heating in winter, which promoted the emissions of K and Na^+ (Chen et al., 2017). The most abundant species in factor 4 were Hg (75%), Pb (68%), K (36%), Cu (35%), Cl^- (33%), and SO_4^{2-} (27%). Pb, Hg, and Cu were typical marker elements for coal combustion, and around 56% of Pb and 47% of Hg were released from coal combustion (Cheng et al., 2015; Zhu et al., 2020). In northern China, the coal-based domestic heating was one of the most important sector of coal consumption (Liu et al., 2018b). Dai et al. (2019) also verified that the residential coal combustion was major source of primary sulfate. Thus, the factor 4 was regarded as the coal combustion (CC) source. The last factor was distinguished by high loadings of Fe (46%), Ni (45%), and Ca (38%). Fe and Ca were main elements enriched in upper crust, and Ni was enriched in the brake wear and tyre wear dusts (Dehghani et al., 2017; Urrutia-Goyes et al., 2018). Thus, these elements in this factor were mainly sourced from traffic-related road dust (RD).

After COVID-19 outbreak, the chemical compositions in $PM_{2.5}$ were also classified into five sources including SF, IP, BB, CC, and RD. However, the contribution ratios of these sources varied

greatly after the implementation of serious lockdown measures. The contribution ratio of IP experienced the largest decrease from 27% to 20%, whereas the apportionment of SF showed the marked increase from 36% to 44%. The contributions of other three sources only suffered from slight variations. The rapid decrease of IP contribution might be associated with the shutdown of many industries during COVID-19 period (Zheng et al., 2020), while the obvious increase of SF contribution was attributable to more heterogeneous or aqueous reactions of precursors (Chang et al., 2020). For nearly all of the species, the contribution ratios of IP suffered from remarkable decreases after COVID-19 outbreak. The contribution ratios of SF for SO_4^{2-}, NO_3^-, and NH_4^+ increased from 35%, 33%, and 41% to 48%, 44%, and 52% after COVID-19 outbreak, respectively. However, the contribution ratios of SF for other species remained relatively stable. It was assumed that SO_4^{2-}, NO_3^-, and NH_4^+ were mainly produced from secondary formation of precursors (Jiang et al., 2019; Yao et al., 2020), while other species especially the trace elements were mainly derived from the primary emission (Wu et al., 2020b). Although the COVID-19 pandemic led to the shutdown of many coal-fired power plants and industries and decreased the CC emissions from these sectors (Kraemer et al., 2020), the government-enforced home order might increase the electricity consumption (Venter et al., 2020), which offset the decreases of CC contributions to industrial activities. Therefore, the contribution ratios of CC did not experience dramatic variation after COVID-19 outbreak.

4. Conclusions and implications

The lockdown measures led to the shutdown of many industries, in turn resulting in the significant decreases of primary components in $PM_{2.5}$. We employed RF model to determine the respective contributions of meteorology and emission reduction on the variations of gaseous

pollutants and $PM_{2.5}$ chemical compositions during COVID-19 lockdown period. The deweathered levels of some trace elements (e.g., Pb (-59%), Zn (-69%)) derived from industrial emissions experienced more than 50% decrease rates due to the stringent lockdown measures. However, the higher relative humidity (RH) and lower air temperature (T) significantly prohibited the decreases of water-soluble ion concentrations because they were beneficial to the heterogeneous or aqueous reaction of sulfate and nitrate. Trace elements were very sensitive to wind direction (WD) due to the long-range transport of anthropogenic emissions. Besides, the contributions of secondary formation to $PM_{2.5}$ increased from 36% to 44% after COVID-19 outbreak. The finding also explained that the opposite change trends of the secondary aerosols in East and West China found by previous studies was not only attributable to the large difference in meteorological conditions, but also the discrepancy of NH_3 concentration.

In the future work, it is necessary to seek multi-pollutants (e.g., VOC, NO_x) emission control measures to reduce the concentrations of primary and secondary components simultaneously since adverse meteorological conditions coupled with slightly higher oxidation capability especially in winter still caused the haze formation. Our results also highlight that more NH_3 emission control measures are urgently needed because the excess NH_3 could exacerbate the generation of secondary aerosols. Besides, the generation of primary pollutants was very sensitive to RH and WD. Thus, the primary pollutant emissions from the industries in the upwind direction should be strictly restricted.

In addition, the present study still suffered from some uncertainties. At first, only six meteorological factors were incorporated into the RF model to quantify the contributions of emission and meteorology of air pollutants. Especially, the missing of solar radiation could affect the accuracy of 8-h O_3 estimation. Besides, solar radiation could change the concentrations of

hydroxyl radicals, thereby affecting the NO_3^- formation. In the future work, the solar radiation should be integrated into the model. In addition, some temporal indicators such as hour and DOY were applied to reflect the COVID-19 lockdown intensity because hourly emission inventory during this period was not available, which should be integrated into the RF model after the development of real-time emission inventory

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Author contributions

Hongbo Fu designed the study. Rui Li wrote the manuscript. Yilong Zhao analyzed the data.

Competing interests

The authors declare that they have no conflict of interest.

Data availability

The meteorological data are available in http://data.cma.cn/.

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Figure 1 The topographic map of China indicating the location of Tangshan (a), sampling site (b), and some key industrial points (b). The population density of Tangshan is also depicted in (b). The red circle in Fig. (b) represents the industrial points, and the pink pentagram denotes the sampling site.

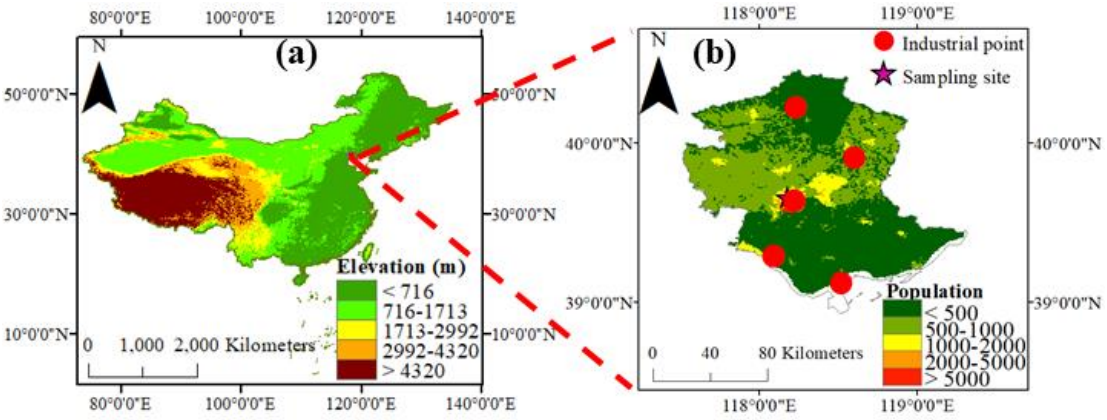

Figure 2 Observed and deweathered weekly concentrations and changes of gaseous pollutants during January 1st-March 31th. The black solid line and dotted line represent the decrease ratio of observed concentration and simulated concentration from Pre-COVID to Post-COVID, respectively. The white background denotes the changes of gaseous pollutants before COVID-19, while the faint yellow one represents the chemical components after COVID-19 outbreak.

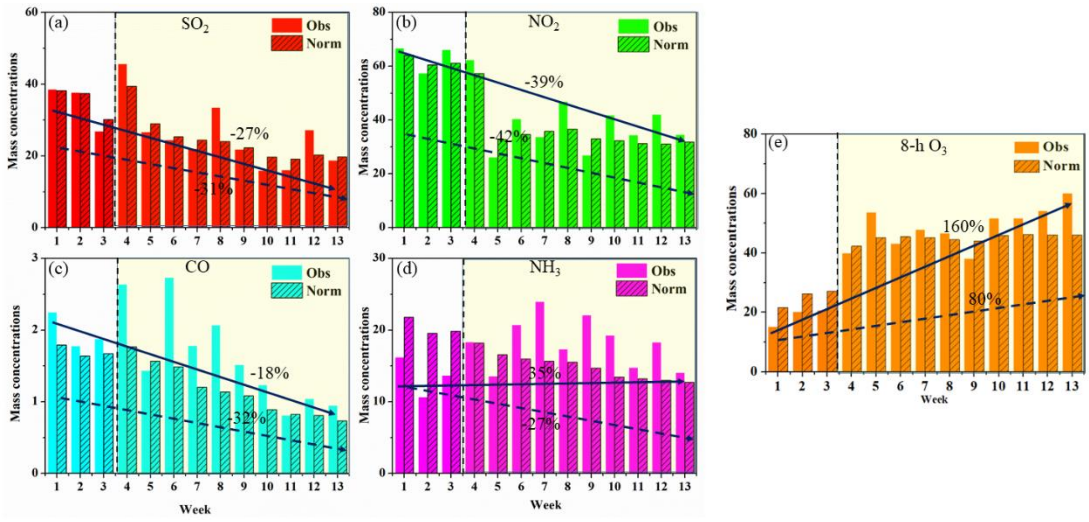

Figure 3 Observed and deweathered weekly concentrations and changes of $PM_{2.5}$ and water-soluble ions during January 1st-March 31th. The black solid line and dotted line represent the decrease ratio of observed concentration and simulated concentration from Pre-COVID to Post-COVID, respectively. The white background denotes the changes of gaseous pollutants before COVID-19, while the faint yellow one represents the chemical components after COVID-19 outbreak.

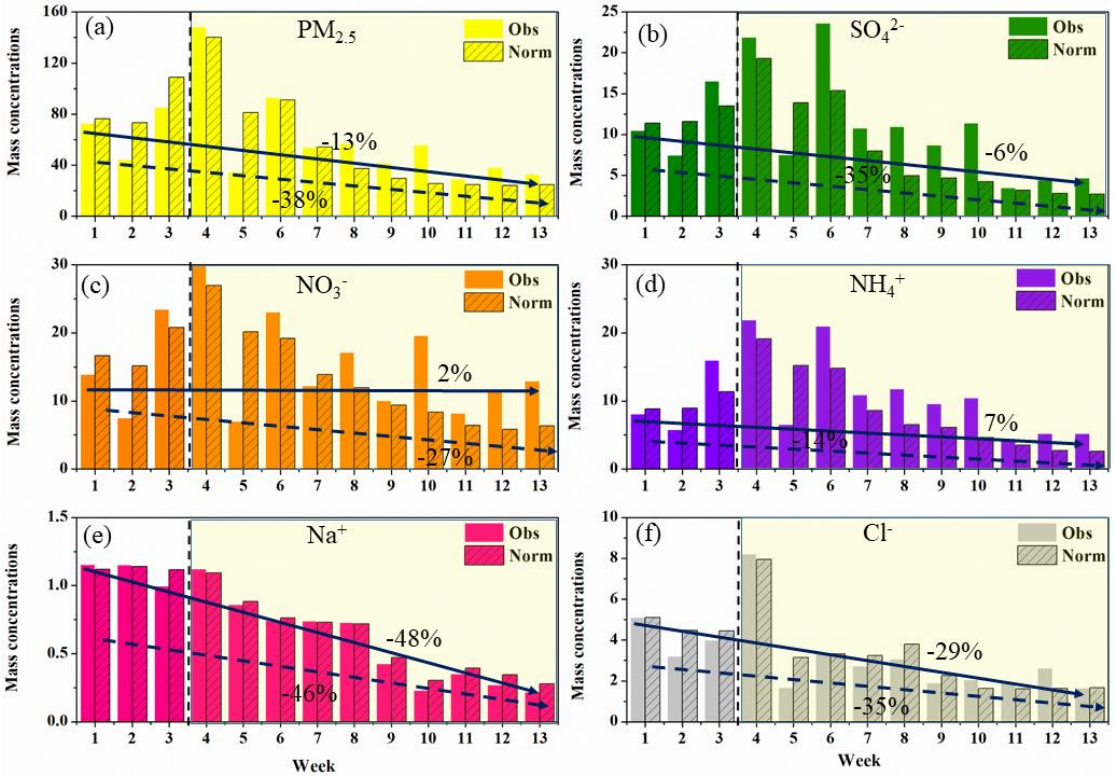

Figure 4 Observed and deweathered weekly concentrations and changes of trace elements during January 1st-March 31th. The black solid line and dotted line represent the decrease ratio of observed concentration and simulated concentration from from Pre-COVID to Post-COVID, respectively. The white background denotes the changes of gaseous pollutants before COVID-19, while the faint yellow one represents the chemical components after COVID-19 outbreak.

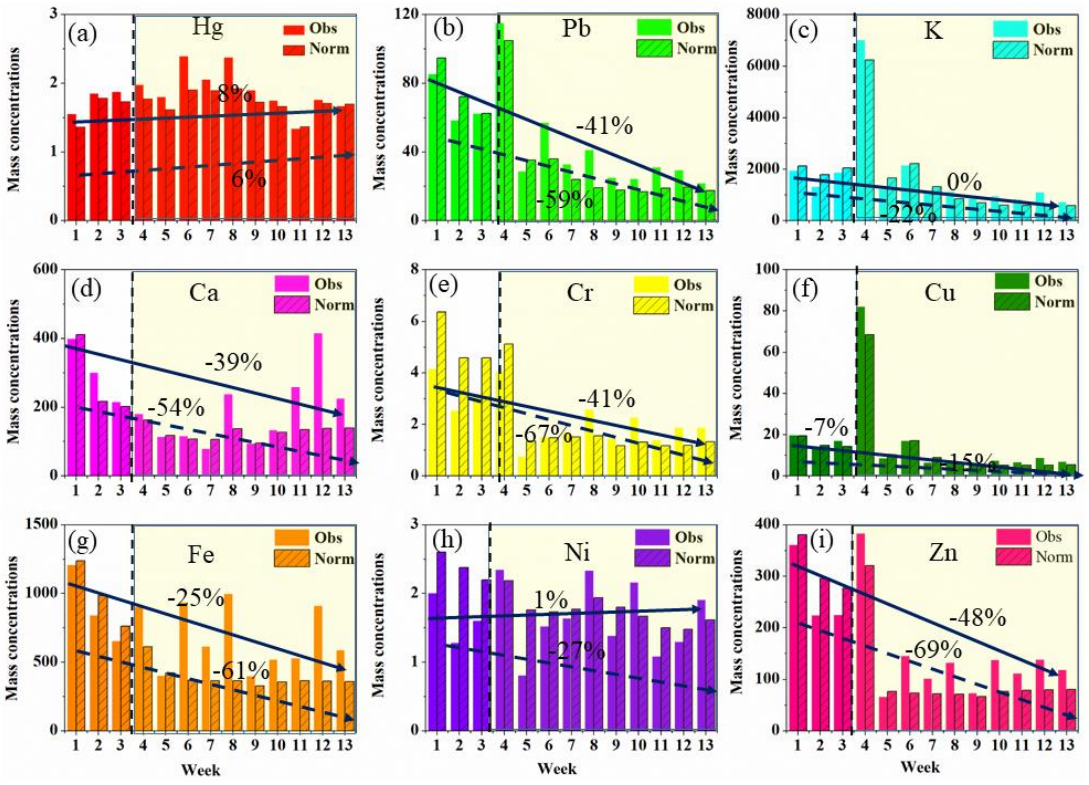

Figure 5 Observed and deweathered weekly concentrations and changes of organic carbon (OC) and elemental carbon (EC) during January 1st-March 31th. The black solid line and dotted line represent the decrease ratio of observed concentration and simulated concentration from from Pre-COVID to Post-COVID, respectively. The white background denotes the changes of gaseous pollutants before COVID-19, while the faint yellow one represents the chemical components after COVID-19 outbreak.

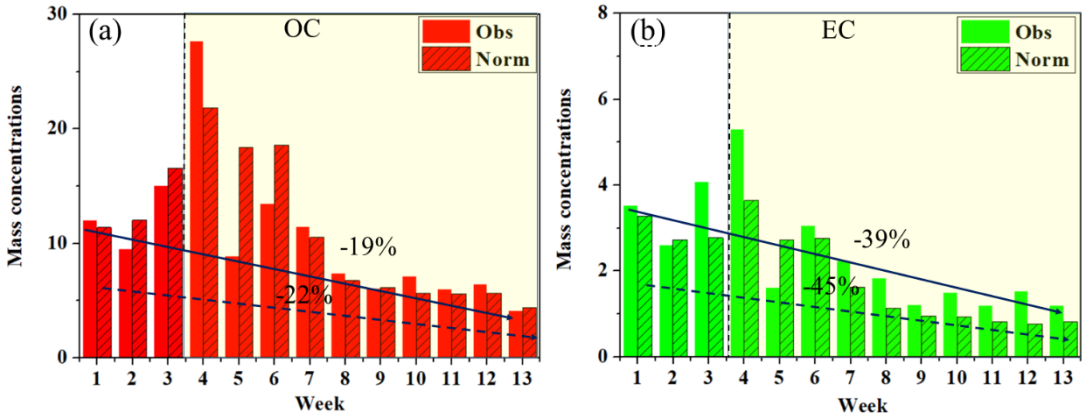

Figure 6 The changes of observed concentrations of multiple components between pre-lockdown (week 1-3) and post-lockdown (week 4-13) against the changes derived from the emission and meteorological changes. The gaseous pollutants, water-soluble ions and carbonaceous aerosols, and trace metals are shown in (a), (b), (c), respectively.

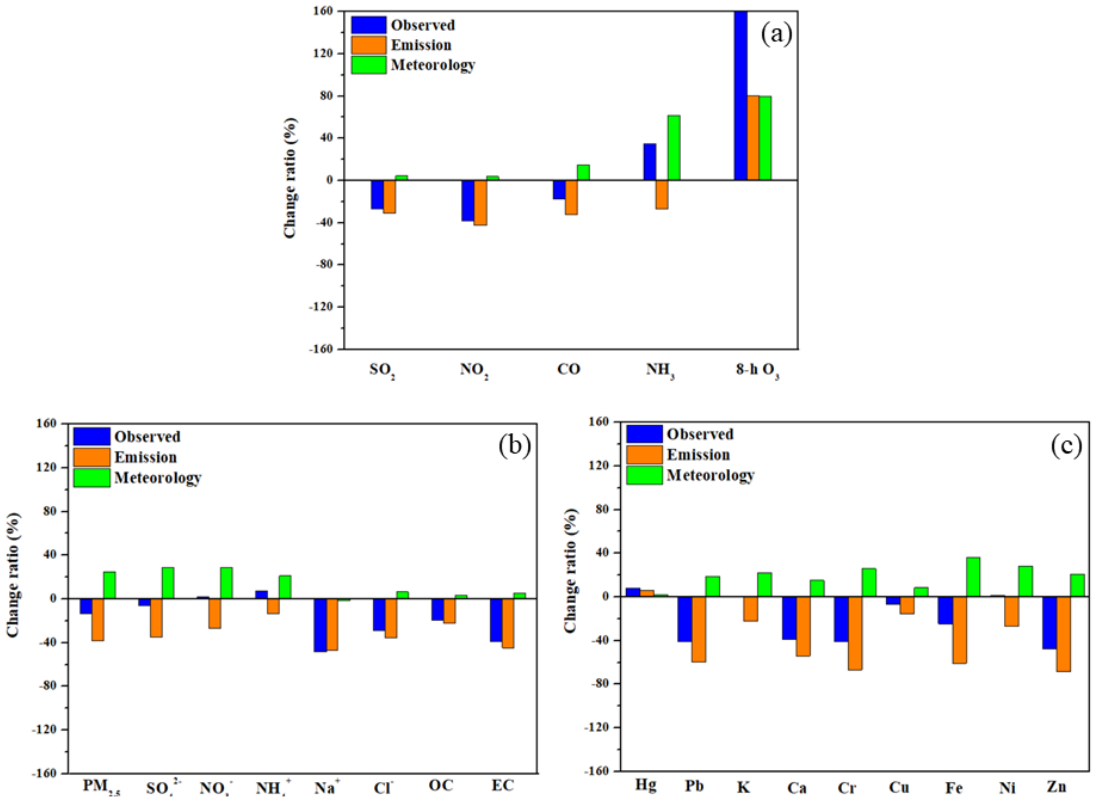

Figure 7 Relative importance of the predictors for the prediction of gaseous pollutants. The match in the figure denotes the variable importance in RF models for various species. DOY, WD, P, RH, Hour, T, DOW, WS, Prec, and Year represent day of year, wind direction, air pressure, relative humidity, hour of the day, air temperature, day of week, wind speed, precipitation, and study year.

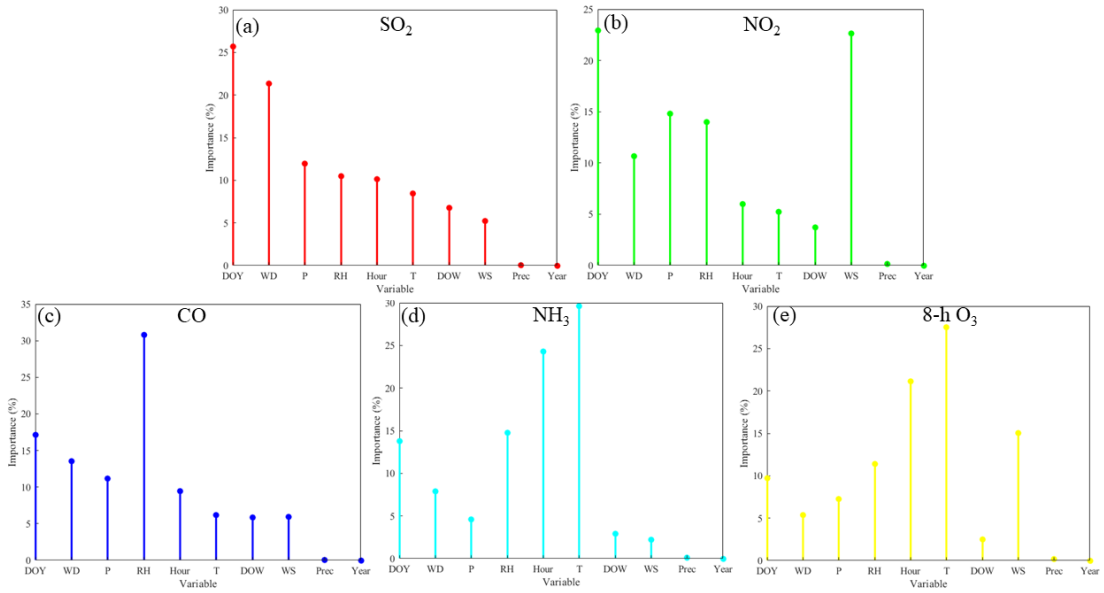

Figure 8 Relative importance of the predictors for the prediction of water-soluble ions in $PM_{2.5}$. The match in the figure denotes the variable importance in RF models for various species. DOY, WD, P, RH, Hour, T, DOW, WS, Prec, and Year represent day of year, wind direction, air pressure, relative humidity, hour of the day, air temperature, day of week, wind speed, precipitation, and study year.

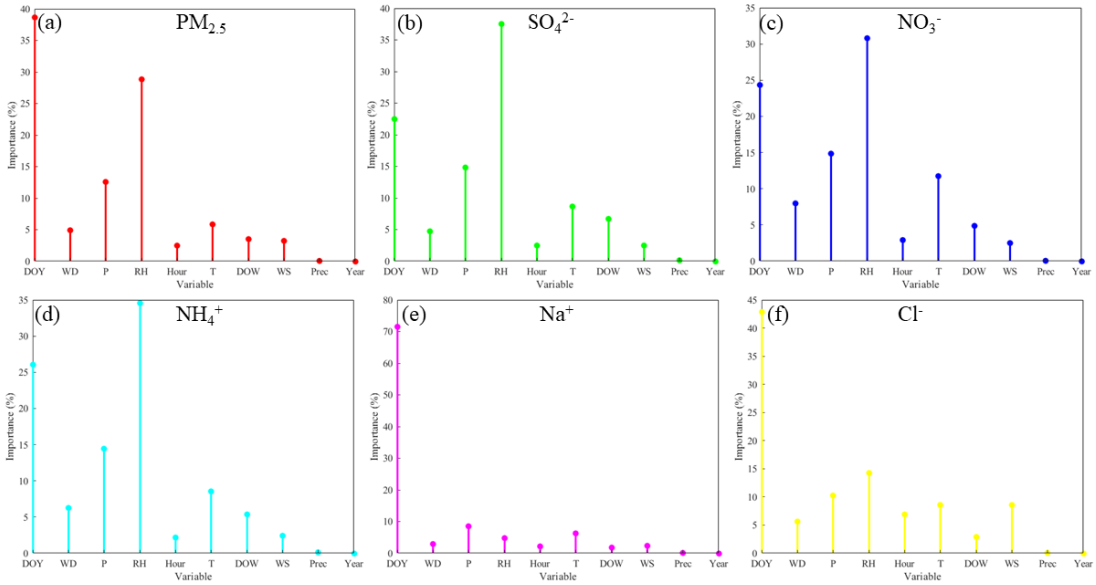

Figure 9 Relative importance of the predictors for the prediction of trace elements in $PM_{2.5}$. The match in the figure denotes the variable importance in RF models for various species. DOY, WD, P, RH, Hour, T, DOW, WS, Prec, and Year represent day of year, wind direction, air pressure, relative humidity, hour of the day, air temperature, day of week, wind speed, precipitation, and study year.

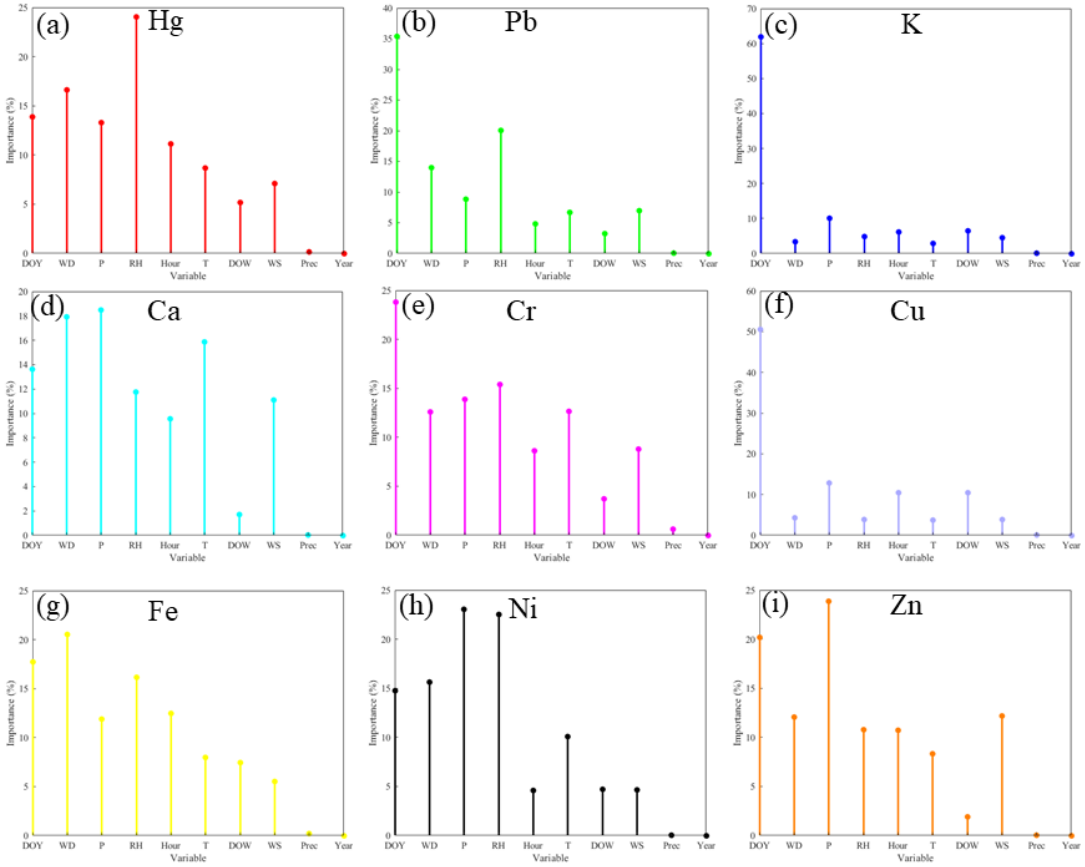

Figure 10 Relative importance of the predictors for the prediction of OC and EC in $PM_{2.5}$. The match in the figure denotes the variable importance in RF models for various species. DOY, WD, P, RH, Hour, T, DOW, WS, Prec, and Year represent day of year, wind direction, air pressure, relative humidity, hour of the day, air temperature, day of week, wind speed, precipitation, and study year.

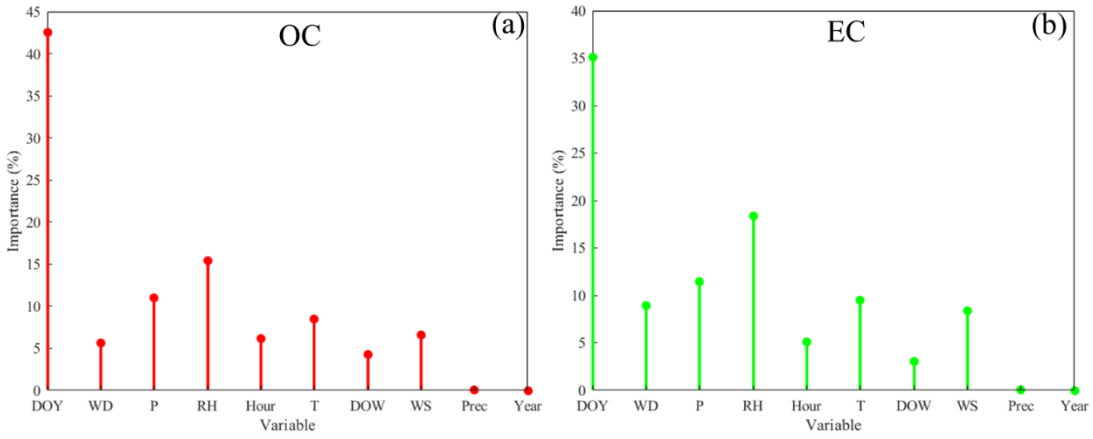

Figure 11 The comparison of source apportionment for $PM_{2.5}$ chemical compositions before (a) and after (b) COVID-19 outbreak. In our study, five major sources were distinguished based on PMF model. The color bar denotes the contributions of these sources to each species. SF, IP, BB, CC, and RD represent secondary formation, industrial process, biomass burning, coal combustion, road dust, respectively.

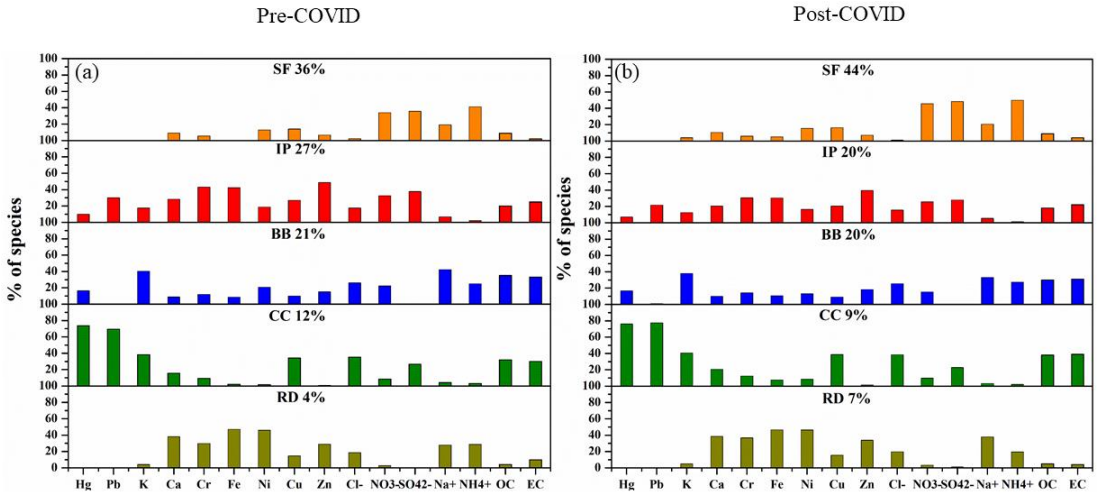

Table 1 SOR, NOR, and C/A values in Pre-COVID and Post-COVID (SOR = $SO_4^{2-}/(SO_4^{2-}+SO_2)$, NOR=$NO_3^-/(NO_3^-+NO_2)$, C/A=$NH_4^+/(SO_4^{2-}+NO_3^-+Cl^-)$).

	SOR	NOR	C/A
Pre-COVID	0.26	0.22	0.33
Post-COVID	0.22	0.25	0.28