

1 Estimating 2010–2015 Anthropogenic and Natural Methane 2 Emissions in Canada using ECCC Surface and GOSAT Satellite 3 Observations

4 Sabour Baray¹, Daniel J. Jacob², Joannes D. Maasackers³, Jian-Xiong Sheng⁴, Melissa P. Sulprizio²,
5 Dylan B.A. Jones⁵, A. Anthony Bloom⁶, and Robert McLaren¹

6 ¹Centre for Atmospheric Chemistry, York University, Toronto, Canada

7 ²Harvard University, Cambridge, MA, USA

8 ³SRON Netherlands Institute for Space Research, Utrecht, The Netherlands

9 ⁴Massachusetts Institute of Technology, Cambridge, MA, USA

10 ⁵University of Toronto, Toronto, Canada

11 ⁶Jet Propulsion Laboratory, California Institute of Technology, Pasadena, CA, USA

12 *Correspondence to:* Sabour Baray (sabour@yorku.ca)

13 **Abstract.** Methane emissions in Canada have both anthropogenic and natural sources. Anthropogenic emissions are estimated to
14 be 4.1 Tg a⁻¹ from 2010–2015 in the National Inventory Report submitted to the United Nation’s Framework Convention on
15 Climate Change (UNFCCC). Natural emissions, which are mostly due to Boreal wetlands, are the largest methane source in Canada
16 and highly uncertain, on the order of ~20 Tg a⁻¹ in biosphere process models. Aircraft studies over the last several years have
17 provided ‘snapshot’ emissions that conflict with inventory estimates. Here we use surface data from the Environment and Climate
18 Change Canada (ECCC) in situ network and space borne data from the Greenhouse Gases Observing Satellite (GOSAT) to
19 determine 2010–2015 anthropogenic and natural methane emissions in Canada in a Bayesian inverse modelling framework. We use
20 GEOS-Chem to simulate anthropogenic emissions comparable to the National Inventory and wetlands emissions using an ensemble
21 of WetCHARTS v1.0 scenarios in addition to other minor natural sources. We conduct a comparative analysis of the monthly
22 natural emissions and yearly anthropogenic emissions optimized by surface and satellite data independently. Mean 2010–2015
23 posterior emissions using ECCC surface data are 6.0 ± 0.4 Tg a⁻¹ for total anthropogenic and 11.6 ± 1.2 Tg a⁻¹ for total natural
24 emissions. These results agree with our posterior using GOSAT data of 6.5 ± 0.7 Tg a⁻¹ for total anthropogenic and 11.7 ± 1.2 Tg a⁻¹
25 for total natural emissions. The seasonal pattern of posterior natural emissions using either dataset shows a slower to start
26 emissions in the spring and a less intense peak in the summer compared to the mean of WetCHARTS scenarios. We combine
27 ECCC and GOSAT data to characterize limitations towards sectoral and provincial level inversions. We estimate Energy +
28 Agriculture emissions to be 5.1 ± 1.0 Tg a⁻¹ which is 59% higher than the National inventory. We attribute 39% higher
29 anthropogenic emissions to Western Canada than the prior. Natural emissions are lower across Canada. Inversion results are
30 verified against independent aircraft data and surface data which show better agreement with posterior emissions. This study shows
31 a readjustment of the Canadian methane budget is necessary to better match atmospheric observations with lower natural emissions
32 partially offset by higher anthropogenic emissions.

33 **1 Introduction**

34 Methane is a significant anthropogenically-influenced greenhouse gas second to carbon dioxide in terms of its direct
35 radiative forcing (Myhre et al., 2013). The mixing ratio of methane has increased from ~720 to ~1800 ppb since pre-
36 industrial times (Hartmann et al., 2013). Present-day global methane emissions are well known to be $550 \pm 60 \text{ Tg a}^{-1}$ (Prather
37 et al., 2012). However recent trends in atmospheric methane since the 1990s are not well understood (Turner et al., 2019).
38 Anthropogenic methane sources include oil and gas activities, livestock, rice cultivation, coal mines, landfills, and
39 wastewater treatment. Natural methane emissions are dominated by wetlands, but also include seeps, termites and biomass
40 burning (Kirschke et al., 2013). The main sink of methane is oxidation by the hydroxyl radical (OH) resulting in a lifetime of
41 9.1 ± 0.9 years (Prather et al., 2012). Improving constraints on national methane emissions is a requirement of mitigation
42 policy (Nisbet et al., 2020). Here we use atmospheric methane observations from the Environment and Climate Change
43 Canada (ECCC) surface network and satellite observations from the Greenhouse Gas Observing Satellite (GOSAT) to
44 estimate Canadian methane emissions and disaggregate anthropogenic and natural sources.

45

46 In the Government of Canada’s submission to the United Nations Framework Convention on Climate Change (UNFCCC),
47 hereafter referred to as the National Inventory, anthropogenic emissions are estimated to be 4.1 Tg a^{-1} in 2015, with 68% of
48 emissions originating from the Western Canadian provinces of Alberta (42%), Saskatchewan (17%) and British Columbia
49 (9%). Sectoral contributions over the entire country are from three categories: Energy (49%), Agriculture (29%) and Waste
50 (22%) (Environment and Climate Change Canada, 2017). Natural emissions, which are mostly due to Boreal wetlands, are
51 highly uncertain, on the order of $\sim 10\text{-}30 \text{ Tg a}^{-1}$ from biosphere process modelling (Miller et al., 2014; Bloom et al., 2017).

52

53 Atmospheric observations provide constraints on methane emissions. Studies constraining anthropogenic and/or natural
54 methane emissions within Canada have included the use of surface in situ measurements (Miller et al., 2016; Atherton et al.,
55 2017; Ishizawa et al., 2019), aircraft campaigns (Johnson et al., 2017; Baray et al., 2018) and satellites (Wecht et al., 2014;
56 Turner et al., 2015; Maasackers et al., 2021). These observations can determine emissions through mass balance methods or
57 be used in conjunction with a chemical transport model (CTM). Bayesian inverse modelling constrains prior knowledge of
58 emissions based on the mismatch between modelled and observed concentrations. This requires reliable mapping of
59 “bottom-up” inventory emissions for the “top-down” observational constraints to be useful (Jacob et al., 2016). Inverse
60 modelling has been more challenging for Canada than the United States due to a) the sparsity of surface stations and satellite
61 data (Sheng et al., 2018a), b) a factor of ~ 10 lower anthropogenic emissions (Maasackers et al., 2019), c) large spatially-
62 overlapping emissions from Boreal wetlands that are highly uncertain (Miller et al., 2014), and d) model biases in the high-
63 latitudes stratosphere (Patra et al., 2011), compromising the interpretation of observed methane columns.

64

65 These observing system challenges have made Canadian methane emissions difficult to quantify. However, studies show a
66 consistent story across different scales and measurement platforms. Miller et al. (2014, 2016) determined that the North
67 American network can successfully constrain Canadian natural emissions and found Boreal wetlands to be lower in 2008
68 when compared to prior fluxes in the WETCHIMP model. Aircraft campaigns over the Alberta oil and gas sector have found
69 higher emissions than inventories in the Red Deer and Lloydminster regions (Johnson et al., 2017) and unconventional oil
70 extraction in the Athabasca Oil Sands region (Baray et al., 2018). Atherton et al. (2017) conducted ground-based mobile
71 measurements of gas production in British Columbia and determined higher emissions than reported, and Zavala-Araiza et
72 al. (2018) conducted similar ground-based measurements in Alberta to show a profile of super-emitters dominating the
73 fugitive methane profile similar to sites in the United States. Ishiziwa et al. (2019) constrained arctic wetland fluxes to be
74 similar in magnitude to the mean of the WetCHARTS inventory but with better identified seasonal and interannual
75 variability. Satellite inversions over North America using the GEOS-Chem CTM and data from SCIAMACHY (Wecht et al.,
76 2014) or GOSAT (Turner et al., 2015; Maasackers et al., 2019) consistently require an increase in anthropogenic emissions
77 in Western Canada and a decrease in natural emissions in Boreal Canada to match observations, even with the use of updated
78 Canadian fluxes in Maasackers et al. (2019) for anthropogenic (Sheng et al., 2017) and wetlands (Bloom et al., 2017)
79 sources. Inverse modelling studies that use both in situ and satellite observations are valuable for intercomparison and for
80 identifying the limits of spatial and temporal discretization that are possible (Lu et al., 2021; Tunnicliffe et al., 2020). The
81 Tropospheric Monitoring Instrument (TROPOMI) launched in 2017 with a data record beginning in 2018 and is expected to
82 provide significant improvements in emissions monitoring through denser observational coverage at a similar precision to
83 GOSAT (Hu et al., 2018). It is necessary to build a reliable historical record of Canadian methane emissions, as
84 anthropogenic emissions are sensitive to changes in policy and economic activity (Rogelj et al., 2018) and natural emissions
85 in Boreal Canada may be sensitive to climate change (Kirschke et al., 2013).

86

87 In this study we use surface observations from the ECCC GHG monitoring network and satellite data from GOSAT to
88 constrain anthropogenic and natural emissions in Canada. We use the GEOS-Chem CTM to simulate 2010–2015 methane
89 concentrations. The model setup includes the use of an improved bottom-up inventory for Canadian oil and gas emissions
90 (Sheng et al., 2017), the WetCHARTS extended ensemble for wetland emissions (Bloom et al., 2017) and EDGAR v4.3.2
91 for other anthropogenic sources. We perform an ensemble forward model analysis which compares six wetlands scenarios to
92 the ECCC surface observation network to assess the influence of process model configurations on Canadian methane. A
93 series of Bayesian inverse analyses are performed that use ECCC and GOSAT data independently and in a joint surface-
94 satellite system. We constrain monthly natural emissions and yearly total anthropogenic emissions from 2010–2015 using
95 ECCC and GOSAT data independently for comparison to produce aggregated-source emissions estimates. We test the
96 limitations of the ECCC and GOSAT joint observation system towards constraining emissions by inventory sector and
97 according to provincial boundaries. We demonstrate where the observation system succeeds in providing strong constraints

98 on major emissions sources and quantify the information content of the system to understand the limitations for resolving all
99 minor Canadian emissions.

100 **2 Data and Methods**

101 We use the GEOS-Chem CTM v12-03 (<http://acmg.seas.harvard.edu/geos/>) to simulate methane fields from 2010–2015 on a
102 $2^\circ \times 2.5^\circ$ global grid and compare to surface observations from the ECCC in situ GHG monitoring network and satellite
103 observations from GOSAT within the Canadian domain. We test for bias in the global model representation of background
104 methane using both surface and aircraft in situ data at Canada’s most westerly site, Estevan Point (ESP), using global
105 GOSAT data, and using global NOAA/HIPPO data. The sensitivity of simulated methane in Canada to the use of different
106 wetlands flux parametrization is evaluated by comparing an ensemble of WetCHARTS v1.0 configurations to ECCC surface
107 observations. The WetCHARTS ensemble mean in addition to other GEOS-Chem prior emissions are used in the Bayesian
108 inverse analysis which optimizes Canadian sources using ECCC surface data and GOSAT satellite data independently for
109 comparative analysis. We show the limitations of the observing system towards subnational level discretization by
110 combining ECCC and GOSAT data in a joint-inversion. Here we describe the observations, the model, and the inverse
111 analysis in further detail.

112 **2.1 Observations**

113 **2.1.1 In situ Surface Observations**

114 We use continuous measurements from eight sites in the ECCC greenhouse gas monitoring network from 2010–2015. Figure
115 1 shows a map of the sites and Table 1 provides a descriptive list. The eight sites are Estevan Point, British Columbia (ESP),
116 Lac La Biche, Alberta (LLB), East Trout Lake, Saskatchewan (ETL), Churchill, Manitoba (CHC), Fraserdale, Ontario
117 (FRA), Egbert, Ontario (EGB), Chibougamau, Quebec (CHM) and Sable Island, Nova Scotia (SBL). All sites use Picarro
118 cavity ring-down spectrometers (G1301, G2301 or G2401) measuring dry-air mole fractions of methane with hourly-average
119 precision better than 1 ppb. For model comparison the measurements are averaged over 4h from 12:00 to 16:00 local time,
120 when the planetary boundary layer is well-mixed. The instruments are calibrated against World Meteorological Organization
121 (WMO) certified standard gases. The westernmost site, ESP, measures methane continuously from a 40 m tower at a
122 lighthouse station on the west coast of Vancouver Island. ESP is surrounded by forests to the north, east, and south and the
123 Pacific Ocean to the west. ESP is used to evaluate boundary conditions and model bias in the methane background as it is the
124 least sensitive to Canadian emissions due to prevailing westerly winds. Sites LLB and ETL are the most sensitive to
125 anthropogenic emissions in Western Canada. LLB measures continuously from a 50 m tower located in a region of peatlands
126 and forest ~200 km NE of Edmonton and ~230 km S of Fort McMurray. ETL measures from a height of 105 m located ~150
127 km north of Prince Albert surrounded by Boreal forest. The sites in the Hudson Bay Lowlands (HBL) region, CHC and
128 FRA, are the most sensitive to natural wetland emissions as this area produces some of the largest methane fluxes from

129 wetlands in North America. CHC measures continuously from a 60 m tower in a small port town on the western edge of
 130 Hudson Bay surrounded by flat tundra. FRA measures from a 40 m tower and is located on the southern perimeter of James
 131 Bay surrounded by extensive wetlands coverage. The site CHM in Quebec is also sensitive to natural wetland emissions and
 132 is excluded in the inverse analysis to be used to verify the posterior results. CHM is substituted by Chapais, Quebec ~50 km
 133 away from 2011 onwards. The remaining Central and Atlantic Canada sites EGB and SBL are sensitive to net outflow from
 134 Canadian sources, both natural and urban, and some emissions from the Eastern United States. EGB is in a small rural
 135 village ~80 km north of Toronto and measures from a 25 m tower. SBL is on a remote uninhabited island 275 km ESE of
 136 Halifax, Nova Scotia and measures from a height of 25 m.

137
 138 **Table 1:** Descriptive list of ECCC in situ observation sites used in the analysis.

Site Code	Full Name, Province	Latitude	Longitude	Elevation (asl) / Sampling Height (agl) (m)
ESP	Estevan Point, British Columbia	49.4° N	126.5° W	7 / 40
LLB	Lac La Biche, Alberta	55.0° N	112.5° W	548 / 50
ETL	East Trout Lake, Saskatchewan	54.4° N	105.0° W	500 / 105
CHC	Churchill, Manitoba	58.7° N	93.8° W	16 / 60
FRA	Fraserdale, Ontario	49.8° N	81.5° W	210 / 40
EGB	Egbert, Ontario	44.2° N	79.8° W	225 / 25
SBL	Sable Island, Nova Scotia	43.9° N	60.0° W	2 / 25
CHM ^{*†}	Chibougamau, Quebec	49.7° N	74.3° W	383 / 30
CHA ^{*†}	Chapais, Quebec	49.8° N	75.0° W	381 / 30

140 ^{*}Chibougamau, Quebec is replaced by Chapais, Quebec ~50 km away from 2011 onwards, overlapping in Fig.1

141 [†] Site is used to evaluate the posterior inversion results, and is not used in the inversion itself

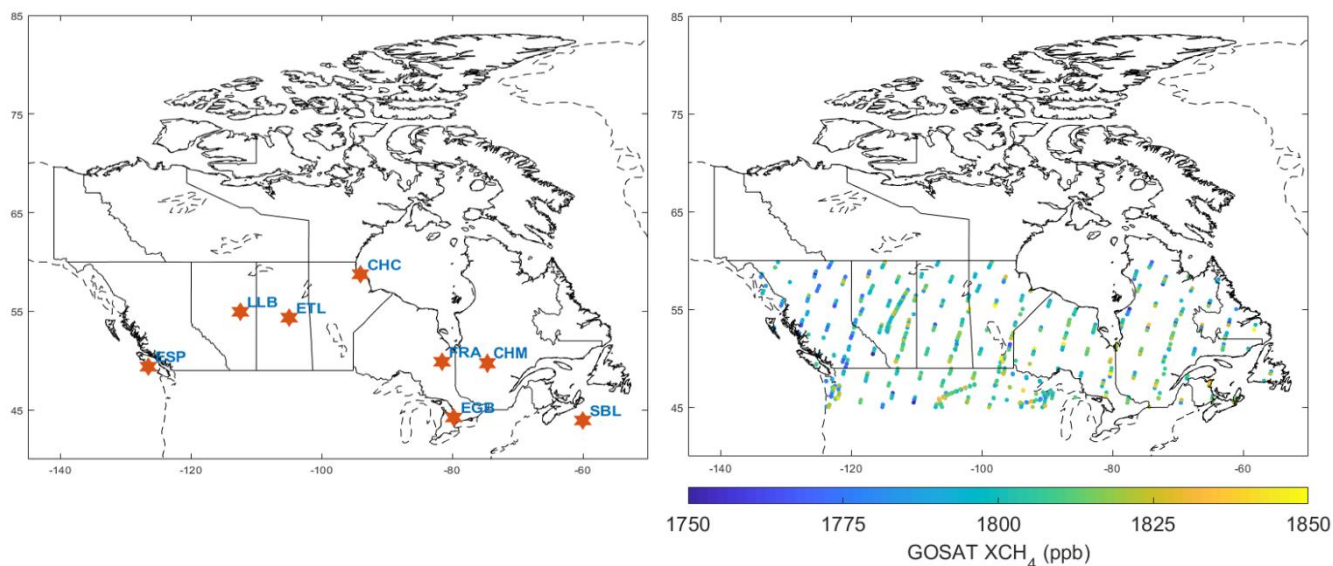
142

143 2.1.2 GOSAT Satellite Observations

144 The Greenhouse Gas Observing Satellite (GOSAT) was launched in January 2009 by the Japan Aerospace Exploration
 145 Agency (JAXA). GOSAT is in a low-Earth polar sun-synchronous orbit with an equator overpass around 13:00 local time.
 146 The TANSO-FTS instrument on-board GOSAT retrieves column-averaged dry air mol fractions of methane using short-
 147 wave infrared (SWIR) solar backscatter in the 1.65 μm absorption band (Butz et al., 2011). Observation pixels in the default
 148 mode are 10 km in diameter separated by 260 km along the orbit track with repeated observations every 3 days. Target mode
 149 observations provide denser spatial coverage over areas of interest. There has been no observed degradation of GOSAT data
 150 quality since the beginning of data collection (Kuze et al., 2016). Here we use version 7 of the University of Leicester proxy

151 methane retrieval over land from January 2010 to December 2015 (Parker et al., 2011, 2015; ESA CCI GHG project team,
152 2018). The single-observation precision of GOSAT XCH₄ data is 13 ppb, and the relative bias is 2 ppb when validated
153 against the Total Column Carbon Observing Network (TCCON; Buchwitz et al., 2015). Figure 1 shows the GOSAT
154 observations over Canada used in our analysis within the domain of 45° N–60° N latitude and 50° W–150° W longitude. The
155 observations used have passed all quality assurance flags for a total of 45,936 observations from 2010–2015, or
156 approximately ~7600 observations per year. Our analysis excludes glint data over oceans, and cloudy conditions are
157 accounted for by the quality assurance flags. We avoid using data above 60° N latitude due to higher uncertainty in the
158 satellite retrieval and the model comparison (Maasakkers et al., 2019; Turner et al., 2015).

159



160
161 **Figure 1:** ECCC surface (left) and GOSAT satellite (right) observations used in the inverse analysis. A descriptive list of the
162 ECCC sites is shown in Table 1. GOSAT data shown is from a single year in 2013 and is filtered to the Canadian domain
163 within 45°N–60°N latitude and 50°W–150°W longitude. There are ~600 GOSAT observations per month in this domain
164 with a minimum Nov–Jan (112–248) and maximum Jul–Sep (872–1098), individual months are shown in the Supplement
165 (Fig. S1).

166

167 2.2 Forward Model

168 We use the GEOS-Chem CTM v12-03 at 2° × 2.5° grid resolution driven by 2009–2015 MERRA-2 meteorological fields
169 from the NASA Global Modeling and Assimilation Office (GMAO). Initial conditions from January 2009 are from a
170 previous GOSAT inversion by Turner et al. (2015) which was shown to be unbiased globally when compared to surface and
171 aircraft data. Bottom-up anthropogenic emissions in GEOS-Chem are from the 2013 ICF Canadian oil and gas inventory

172 (Sheng et al., 2017) and the 2012 EDGAR v4.3.2 global inventory for other Canadian and global sources, and the gridded
 173 US 2012 EPA Inventory for the United States (Maasackers et al., 2016). For wetlands, six configurations from the 2010–
 174 2015 extended ensemble of WetCHARTS (Bloom et al., 2017) are used in the ensemble forward model analysis (Section
 175 3.1) and the ensemble mean is used as the prior for the inverse analysis (Sections 3.2–3.4). Figure 2 shows the spatial
 176 distribution of the prior methane emissions in Canada from the major anthropogenic and natural sources. The two largest
 177 sources are from the ICF oil and gas inventory, (Sheng et al., 2017) and wetland emissions from the ensemble mean of the
 178 WetCHARTS inventory (Bloom et al., 2017), with significant emissions from livestock and waste emissions from EDGAR.
 179 Oil and gas are 54% of the anthropogenic total and wetlands are 94% of the natural total. The prior emissions estimates in
 180 this simulation are summarized in Table 2, which organizes emissions by Canadian source categories and are compared to
 181 sector attribution in the National Inventory (Environment and Climate Change Canada, 2017). Our totals for Energy,
 182 Agriculture and Waste are 2.4, 1.0, and 0.9 Tg a⁻¹ respectively compared to 2.0, 1.2 and 0.9 Tg a⁻¹ in the National Inventory.
 183 In the absence of a spatially disaggregated Canadian inventory for methane, we consider these prior estimates reasonably
 184 similar for the purpose of comparing our posterior emissions to the National Inventory, however we cannot compare the
 185 spatial pattern of emissions which will likely show more discrepancies. Natural emissions are divided into wetlands, which
 186 are 14.0 Tg a⁻¹ in the ensemble mean, and other natural sources, which are 0.8 Tg a⁻¹ from biomass burning, seeps, and
 187 termites. Each component of other natural emissions has a separate spatially disaggregated inventory as described in
 188 Maasackers et al. (2019). Emissions from the United States and the rest of the world are included in the model but not
 189 optimized in the inversions. Loss of methane from oxidation due to OH is computed using archived 3-D monthly fields of
 190 OH from a previous GEOS-Chem full-chemistry simulation (Wecht et al., 2014).

191

192

193 **Table 2:** Mean 2010–2015 prior estimates of Canadian methane emissions used in GEOS-Chem arranged according to
 194 categories in the National Inventory (Environment and Climate Change Canada, 2017).

195

Category	Source Type ^a	Emissions (Tg a ⁻¹) ^a	Total (Tg a ⁻¹) ^a	Inventory (Tg a ⁻¹) ^b	
Anthropogenic	Oil	0.52	2.42	2.00	
	Energy	Gas			1.81
	Coal	0.09			
	Agriculture	Livestock	1.00	1.00	1.20
	Waste	Landfills	0.66	0.94	0.92
		Wastewater	0.19		
		Other Anthropogenic	0.09		

	Wetlands	-	14.0	14.0	-
Natural	Other	Biomass Burning	0.28		
	Natural	Seeps	0.28	0.84	-
		Termites	0.28		

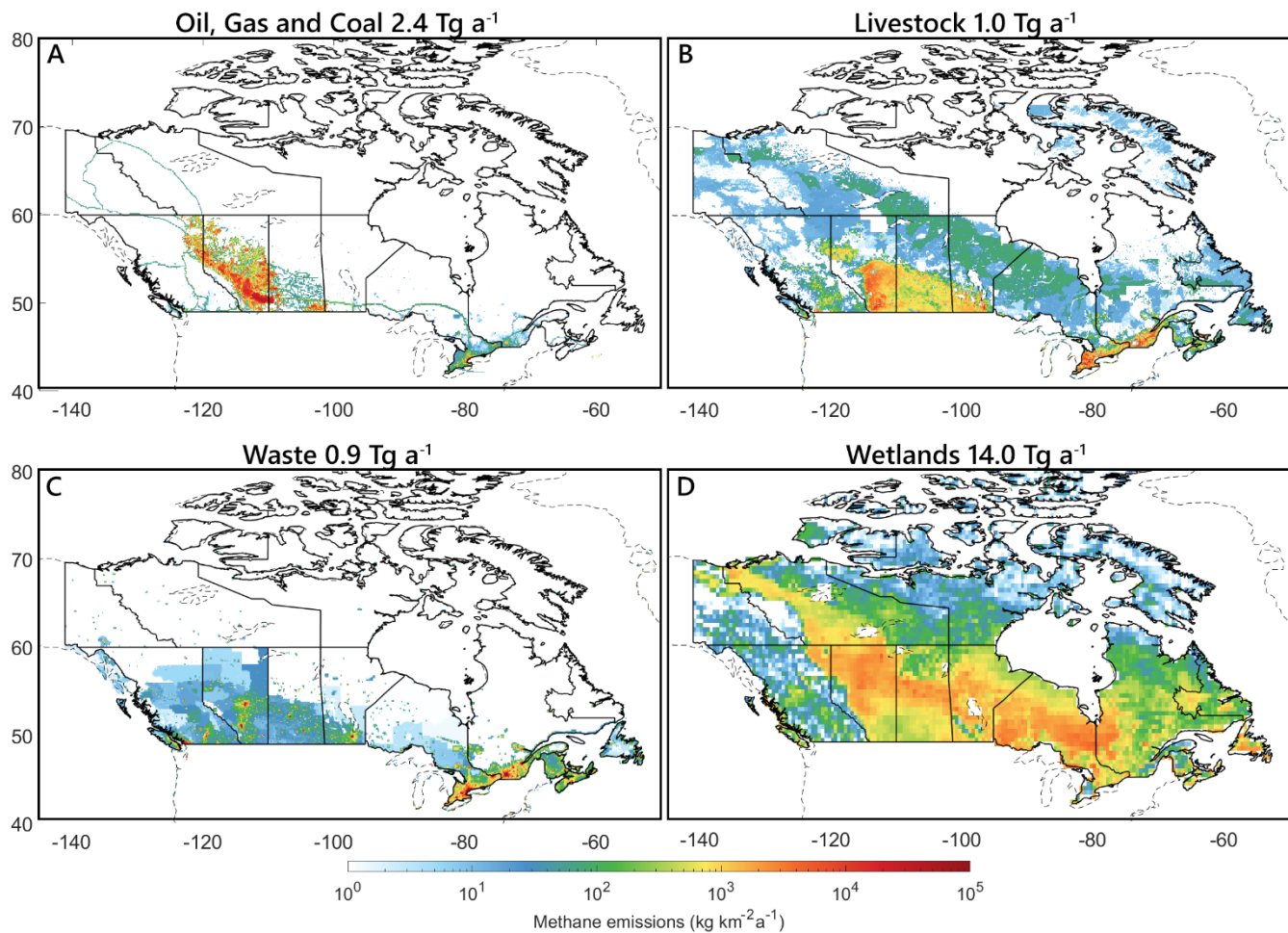
196 ^aEmissions inputs for GEOS-Chem. These are shown for the individual source types and summed over the categories
197 Energy, Agriculture and Waste. In Canada, oil and gas are from Sheng et al. (2017), coal, livestock, landfills, wastewater and
198 other anthropogenic are from EDGAR v4.3.2, wetlands are from Bloom et al. (2017). Biomass burning is from QFED
199 (Darmenov and da Silva, 2013) and termite emissions are from Fung et al. (1991). Seeps and other global sources are
200 described in Maasakkers et al. (2019).

201

202 ^bEmissions from the National Inventory (Environment and Climate Change Canada, 2017) that correspond to the Energy,
203 Agriculture and Waste categories. These are used in the discussion of results but are not included in the inverse model.

204

205



206
 207 **Figure 2:** Prior estimates of anthropogenic and natural methane emissions. Colour bars are in log scale in units of kg CH_4
 208 $\text{km}^{-2} \text{a}^{-1}$. Most anthropogenic emissions fall under the energy category (A) which are oil and gas in the ICF inventory (Sheng
 209 et al., 2017) plus minor emissions from coal in EDGAR 4.3.2. Livestock (B) and waste (C) are from EDGAR. Natural
 210 emissions are primarily wetlands from the WetCHARTS inventory (D; Bloom et al., 2017).

211 2.3 Inverse Model Methodology

212 We optimize emissions in the inverse analysis by minimizing the Bayesian cost function $J(\mathbf{x})$ (Rodgers, 2000).

213

$$214 J(\mathbf{x}) = \frac{1}{2} (\mathbf{x} - \mathbf{x}_a)^T \mathbf{S}_a^{-1} (\mathbf{x} - \mathbf{x}_a) + \frac{1}{2} (\mathbf{y} - \mathbf{F}(\mathbf{x}))^T \mathbf{S}_o^{-1} (\mathbf{y} - \mathbf{F}(\mathbf{x})) \quad (1)$$

215

216 Where \mathbf{x} is the vector of emissions being optimized, \mathbf{x}_a is the vector of prior emissions (Table 2), $\mathbf{F}(\mathbf{x})$ is the simulation of
 217 methane concentrations corresponding to the observation vector \mathbf{y} of ECCC surface and/or GOSAT data. \mathbf{S}_a is the prior error
 218 covariance matrix and \mathbf{S}_o is the observational error covariance matrix. The observational error matrix includes both

219 instrument and model transport error. The GEOS-Chem model relating methane concentrations to emissions $F(\mathbf{x})$ is
 220 essentially linear and can be represented by the Jacobian matrix \mathbf{K} such that $F(\mathbf{x}) = \mathbf{K}\mathbf{x} + \mathbf{b}$, where \mathbf{b} is the model
 221 background. The background includes initial conditions from Turner et al. (2015) and methane from global emissions that
 222 are held constant in the inversion. Possible bias in the background is evaluated in detail in the Supplement Section 1.3 and
 223 shown to be minimal. The \mathbf{K} matrix is of m by n size where n is the number of state vector elements being optimized and m
 224 is the number of ECCC surface and/or GOSAT observations being used. The \mathbf{K} matrix is constructed using the forward
 225 mode of GEOS-Chem and the tagged tracer output for Canadian sources which describes the sensitivity of concentrations to
 226 emissions dy/dx in ppb Tg^{-1} .

227

228 GEOS-Chem continuously simulates global emissions with a global source-sink imbalance of $+13 \text{ Tg a}^{-1}$ in the budget as
 229 described in Maasakkers et al. (2019). We show in Section 1.3 of the Supplement that this configuration of the model
 230 reliably reproduces the global growth rate in atmospheric methane with adjustments only needed for 2014 and 2015
 231 primarily due to differences in tropical wetland emissions (Maasakkers et al., 2019), with reduced transport errors at the $2^\circ \times$
 232 2.5° resolution (Stanevich et al., 2020). This gives a well-represented background for methane which is tested using global
 233 GOSAT and NOAA data, as well as in situ data at Canadian background sites. We improve the model representation of
 234 methane using bias corrections which are discussed in Section 1.3 of the Supplement, and we show the consistency of the
 235 inversion results without adjustments to the model. A high resolution inversion over North America over the 2010–2015
 236 time-period using the same prior has shown adjustments to US emissions near the Canadian border are also relatively
 237 minimal, (Maasakkers et al., 2021), so we treat US emissions as constant. The assumption of constant US emissions is tested
 238 in Section 1.3.2 of the Supplement by removing ECCC stations near the US border from the inversion, which show
 239 consistent results. Hence, we can attribute the model-observation mismatch ($\mathbf{y} - F(\mathbf{x})$) using observations limited to Canada
 240 to Canadian emissions which are optimized in the inversion. In the main text we show three inversions with a different
 241 number of state vector elements: a) the monthly inversion ($n = 78$) optimizes monthly natural emissions in Canada and
 242 yearly anthropogenic emissions from 2010–2015, b) the sectoral inversion ($n = 5$) optimizes emissions according to the
 243 major inventory categories in Table 2 individually for each year, and c) the provincial inversion ($n = 16$) optimizes emissions
 244 according to subnational boundaries which is also repeated for each year. The monthly inversion provides higher temporal
 245 resolution relative to the other approaches in this study to constrain the seasonality of natural emissions, assuming the spatial
 246 distribution is correct. The sectoral inversion provides direct constraints on inventory categories, and the provincial inversion
 247 provides relatively higher spatial resolution for subnational attribution. Substituting $F(\mathbf{x}) = \mathbf{K}\mathbf{x}$ in eq. 1 and subtracting the
 248 background \mathbf{b} , the analytical solution of the cost function $dJ(\mathbf{x})/d\mathbf{x} = 0$ yields the optimal posterior solution $\hat{\mathbf{x}}$ (Rodgers,
 249 2000):

250

$$251 \hat{\mathbf{x}} = \mathbf{x}_a + \mathbf{S}_a \mathbf{K}^T (\mathbf{K} \mathbf{S}_a \mathbf{K}^T + \mathbf{S}_0)^{-1} (\mathbf{y} - \mathbf{K} \mathbf{x}_a) \quad (2)$$

252

253 The analytical solution provides closed-form error characterization, such that the the posterior error covariance $\hat{\mathbf{S}}$ of the
254 posterior solution $\hat{\mathbf{x}}$ is given by:

255

$$256 \quad \hat{\mathbf{S}} = (\mathbf{K}^T \mathbf{S}_o^{-1} \mathbf{K} + \mathbf{S}_a^{-1})^{-1} \quad (3)$$

257

258 The averaging kernel matrix \mathbf{A} is used to evaluate the surface and satellite observing systems and is given by:

259

$$260 \quad \mathbf{A} = \mathbf{I}_n - \hat{\mathbf{S}} \mathbf{S}_a^{-1} \quad (4)$$

261

262 where \mathbf{I}_n is the identity matrix of length n corresponding to the number of state vector elements. The averaging kernel matrix
263 \mathbf{A} describes the sensitivity of the posterior solution $\hat{\mathbf{x}}$ to the true state \mathbf{x} ($\mathbf{A} = d\hat{\mathbf{x}}/d\mathbf{x}$). The trace of \mathbf{A} provides the degrees of
264 freedom for signal (DOFS), which is the number of pieces of information of the state vector that is gained from the inversion
265 (DOFS $\leq n$). The diagonal values of \mathbf{A} provide information on which Canadian state vector elements can be constrained by
266 ECCO surface and GOSAT satellite observations above the noise, and higher DOFS closer to n correspond to better
267 constrained sources in total. As a further diagnostic of the inversion we conduct a singular value decomposition of the
268 prewhitened Jacobian $\check{\mathbf{K}} = \mathbf{S}_o^{-1/2} \mathbf{K} \mathbf{S}_a^{1/2}$ (Rodgers, 2000). The number of singular values greater than one is the effective rank
269 of $\check{\mathbf{K}}$, which shows the independence of the state vector elements and the number of pieces of information above the noise
270 that are resolved in the inversion (Heald et al., 2004). The comparison between this eigenanalysis and the DOFS are
271 discussed in the Supplement Section 1.4 and is used to inform the limitations of the observation system.

272

273 We construct the prior error covariance matrix \mathbf{S}_a based on aggregated error estimates for source categories and regions. We
274 use 50% error standard deviation for the aggregated anthropogenic emissions which includes the Sheng et al. (2017) oil and
275 gas inventory and other EDGAR sources, 60% for wetland emissions from the Bloom et al. (2017) WetCHARTS inventory
276 and 100% for non-wetlands natural sources. We assume no correlation between state vector elements so that \mathbf{S}_a is diagonal.
277 Anthropogenic emissions have been shown to be spatially uncorrelated (Maasackers et al., 2016) however wetlands show
278 spatial correlation (Bloom et al., 2017). Here we optimize broadly aggregated categories, so our method assumes the spatial
279 pattern of each state vector element is correct, however correlations between state vector elements in the eigenanalysis are
280 used to assess the limitations of source discretization in the observing systems.

281

282 We construct the diagonal observation error matrix \mathbf{S}_o which captures instrument and model error using the relative residual
283 error method (Heald et al., 2004). In this approach the vector of observed-modelled differences $\Delta = y_{\text{GEOS-Chem}} - y_{\text{observations}}$ is
284 calculated and the mean observed-modelled difference $\overline{\Delta} = \overline{y_{\text{GEOS-Chem}} - y_{\text{observations}}}$ is attributed to the emissions that will
285 be optimized. Hence, the standard deviation in the residual error $\Delta' = \Delta - \overline{\Delta}$ represents the observational error and is used

286 as the diagonal elements of S_o . For our Canadian inversion we find positive model-observation biases in the warmer months
287 (April to September) and negative biases in the colder months (October to March). We calculate the relative residual error
288 for growing and non-growing seasons separately, such that Δ' is partitioned into Δ'_g (October to March) and Δ'_{ng} (April to
289 September) which is then used to calculate the diagonal elements of S_o . For surface observations the mean observational
290 error is 65 ppb. Since the instrument error is <1 ppb for afternoon mean methane measurements, the observational error is
291 entirely attributed to transport and representation error of surface methane in the model grid pixels. For satellite observations
292 the mean observational error is 16 ppb where the instrument error is 11 ppb, showing most of the observational error is from
293 the instrument rather than the forward model representation of the total column. Column-averaged methane concentrations
294 are less sensitive to surface emissions resulting in the lower model error (Lu et al., 2021).

295

296 In summary, the inverse model is designed to suit the objectives of this study, which are to: (1) optimize anthropogenic and
297 natural emissions in Canada at the national-scale and (2) compare the results of inversions using surface and satellite
298 observations, and (3) characterize the limitations of the observing system towards subnational-scale emissions discretization.
299 The spatial and temporal resolution of the inversion is limited by the precision of GOSAT data, the precision of the model
300 representation of surface methane for ECCC data, and the sparse coverage of both systems relative to the smaller magnitude
301 of Canadian emissions. This simplified approach, where Canadian emissions are optimized using only observations in
302 Canada, may be sensitive to errors in the global model that are projected onto the Canadian domain. This is minimized if
303 errors in the regional representation of methane, which are corrected in the inversion, are much larger than errors in the
304 background from the global model, or if the background methane is corrected using global observations outside of the
305 Canadian domain. We show an analysis of the global model alongside sensitivity tests of the inversions in Section 1.3 of the
306 Supplement which produce consistent results. Future studies may deploy a more sophisticated, high resolution inverse model
307 that will match more sophisticated observations, which include an expanded ECCC surface network, as well as satellites
308 with higher density (TROPOMI; Hu et al., 2018) or higher precision (GOSAT-2; Nakajima et al., 2017) observations outside
309 of the years of this analysis.

310 **3 Results and Discussion**

311 **3.1 Evaluation of WetCHARTS Extended Ensemble for Wetland Emissions in Canada**

312 Wetlands are the largest methane source in Canada with uncertainties in the magnitude, seasonality, and spatial distribution
313 of emissions. Our inverse analysis constrains the magnitude and seasonality of emissions with observations. Ideally, the prior
314 emissions in the model should be the best possible representation of emissions to reduce error in the optimization problem
315 (Jacob et al., 2016). Table 2 shows 2010–2015 mean wetland emissions in Canada to be 14.0 Tg a^{-1} from the mean of the
316 WetCHARTS v1.0 inventory (Bloom et al., 2017). These emissions are more than three times the total of anthropogenic
317 emissions 4.4 Tg a^{-1} . The much larger signal from wetland emissions poses a difficulty for constraining anthropogenic

318 emissions (Miller et al., 2014). In this section, we evaluate our use of the mean of the WetCHARTS v1.0 extended ensemble
319 by running a series of forward model runs using alternate ensemble members in GEOS-Chem and comparing model output
320 to ECCC in situ observations.

321

322 The WetCHARTS extended ensemble for 2010–2015 contains an uncertainty dataset of 18 possible global wetlands
323 configurations as described in Bloom et al. (2017). These depend on three processing parameters which are: three CH₄:C
324 temperature-dependent respiration fractions ($q_{10} = 1, 2, \text{ and } 3$; where 1 is the highest temperature dependency), two
325 inundation extent models (GLWD vs. GLOBCOVER; where GLWD corresponds to higher inundation in Canada) and three
326 global scaling factors for global emissions to amount to 124.5, 166 or 207.5 Tg CH₄ yr⁻¹ ($3 \times 2 \times 3 = 18$). We find using the
327 scaling factors corresponding to 124.5 and 207.5 Tg CH₄ yr⁻¹ within GEOS-Chem results in an imbalance in the global
328 budget beyond what is observed in our measurements and degrades the representation of background methane, so we limit
329 the extended ensemble to six members which depend on three temperature parameterizations and two inundation scenarios
330 ($3 \times 2 = 6$). Figure 3 shows the magnitude and spatial distribution of wetland emissions in the six scenarios. The total wetland
331 emissions within Canada show nearly an order of magnitude difference between ensemble members from 3.9 Tg a⁻¹ to 32.4
332 Tg a⁻¹. Compared to the rest of North America, Boreal Canada shows the largest variability between ensemble members,
333 with the Southeast United States as the second most uncertain (Sheng et al., 2018b).

334

335 We use ECCC in situ observations to better constrain the range of wetlands methane emissions in the ensemble members.
336 All six configurations are used in GEOS-Chem to produce a series of forward model runs for a subrange of years between
337 2013–2015. Figure 4 shows GEOS-Chem simulated methane concentrations using the six WetCHARTS configurations and
338 compares them to four ECCC in situ measurement sites in Canada (LLB, ETL, FRA, EGB). This subset of available data is
339 representative of sites sensitive to both anthropogenic and natural emissions. Most of Canadian anthropogenic emissions are
340 from Western Canada (Fig. 2), which we use sites LLB and ETL to evaluate (Fig. 1), and a significant amount of Canadian
341 natural emissions are from regions surrounding the Hudson’s Bay Lowlands, which we use sites FRA and EGB to evaluate.
342 Methane concentrations from GEOS-Chem show large differences when compared to ECCC observations, ranging from
343 +1050 to –150 ppb. The boundary-condition site ESP (Fig. S3) showed a mean bias of 5.3 ppb for all of 2010–2015. Since
344 there is no similar mismatch in the global representation of methane, these biases up to 1050 ppb can therefore be attributed
345 to misrepresented local Canadian emissions plus associated transport and representation error. Two types of biases with
346 opposite signs appear from this comparison. The first type is a positive summertime bias where the modelled methane
347 concentrations significantly exceed the observations; this bias is more pronounced in sites FRA (Fig. 4-C) and EGB (Fig. 4-
348 D), which are in Ontario and sensitive to the Hudson Bay Lowlands. The bias is also visible in the western sites LLB (Fig. 4-
349 A) and ETL (Fig. 4-B) to a lesser extent. As we use a smaller magnitude of wetlands methane emissions corresponding to
350 the ensemble members in Figure 3 (from 32.4 Tg a⁻¹ to 3.9 Tg a⁻¹), this summertime bias decreases proportionately.
351 Therefore, we can attribute these large positive summertime biases to growing season wetland emissions that are

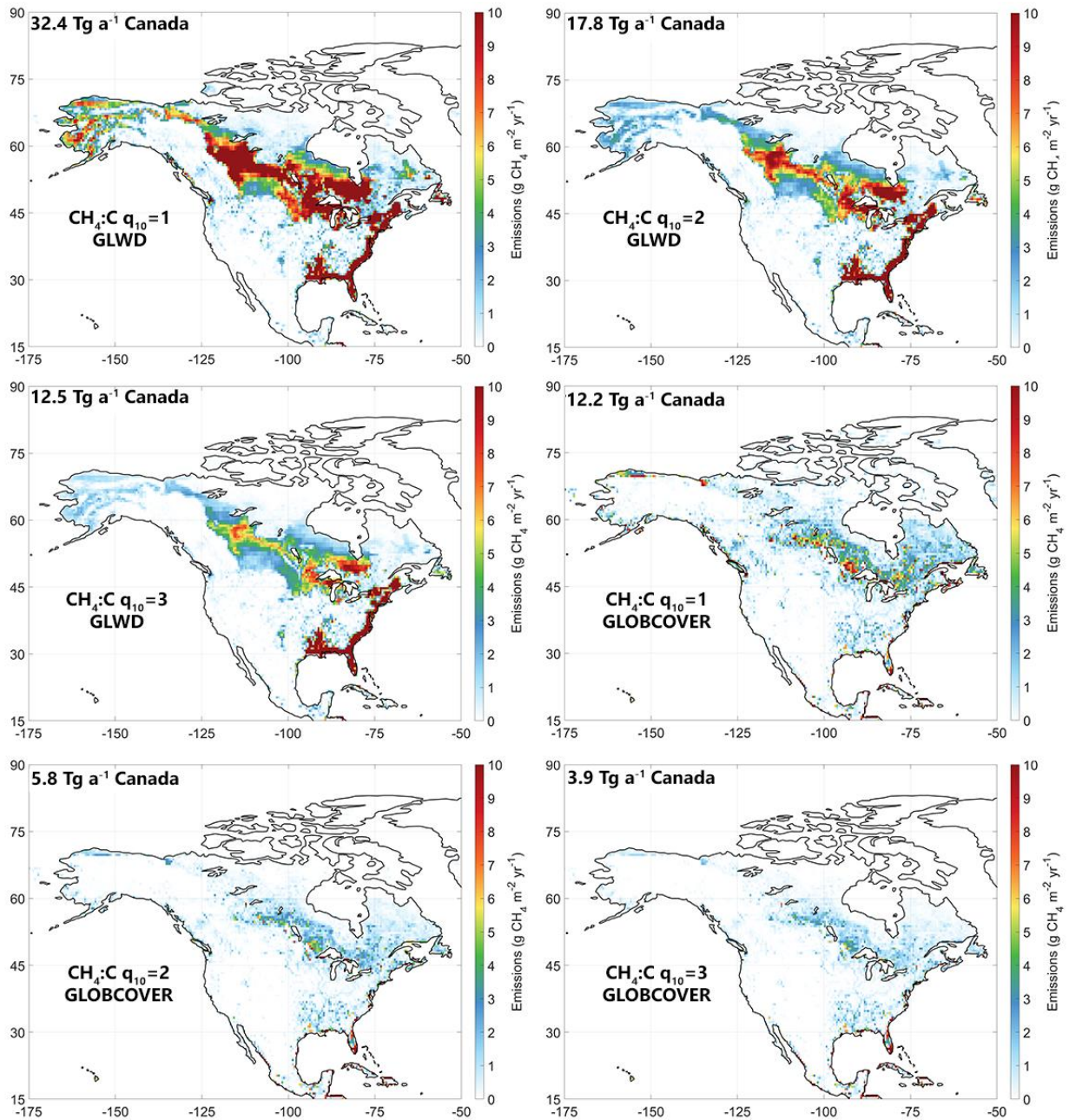
352 overestimated in the process model configurations. The second type of bias is a year-long negative bias that appears most in
353 site LLB (Fig. 4-A) and is magnified with the use of lower-magnitude wetland emissions. This suggests the presence of year-
354 round anthropogenic emissions in Western Canada that are underestimated in the prior, or that winter-time wetland
355 emissions could also be underestimated in WetCHARTS due to the lack of explicit soil water and temperature dependencies.
356 The inverse modelling results in the next section attribute this bias to anthropogenic emissions.

357

358 Miller et al. (2016) conducted a study constraining North American Boreal wetland emissions from the WETCHIMP
359 inventory modelled in WRF-STILT by comparing to observations in 2008. Their study included the use of three of the
360 ECCC stations described here (CHM, FRA and ETL). The model comparison to observations in that study showed a similar
361 pattern of modelled methane exceeding observations in the summer and a low bias at ETL. They suggested wetland
362 emissions were overestimated in most model configurations and that the wetlands bias may be masking underestimated
363 anthropogenic emissions. These conclusions are corroborated by the 2013–2015 comparison shown here, we show high
364 wetland emissions configurations in WetCHARTS produce a high bias that exceed measured summertime methane
365 concentrations, and the use of lower wetlands configurations reveal a year-long low bias apparent in Western Canada. Our
366 results suggest the combined use of higher inundation extent and lower temperature dependencies (GLWD and $q_{10} = 3$), or
367 the use of lower inundation extent and higher temperature dependencies (GLOBCOVER and $q_{10} = 1$) best reproduce
368 observations near the mean of the range of emissions, although the ensemble forward model analysis is unable to specify
369 more detailed process model constraints.

370

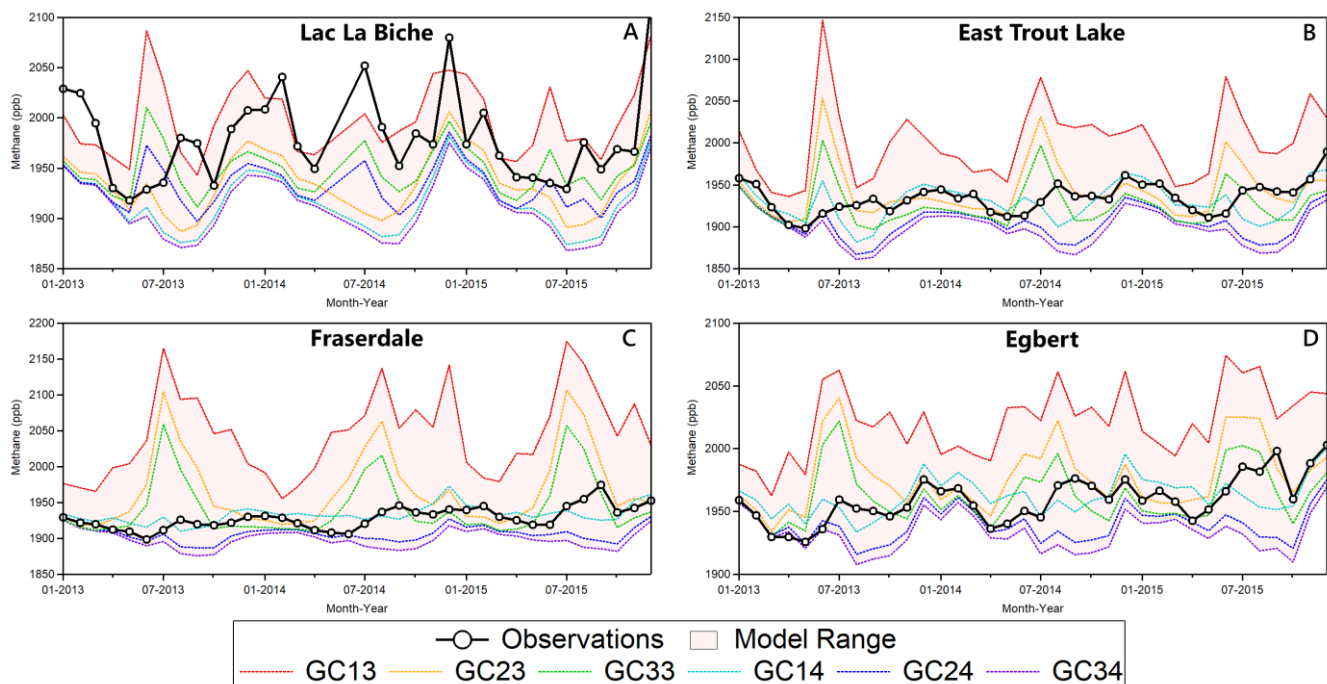
371 The forward model analysis in this section is a direct evaluation of wetlands configurations. This approach allows us
372 *manually* tune wetlands scenarios and diagnose the sensitivity of the modelled-observed differences to the process modelling
373 parameters. The inverse analysis shown subsequently is a statistical optimization that applies scaling factors to emissions
374 based on the same model-observation differences. The inverse analysis can be viewed analogously as an *automatic* approach.
375 These results show the challenge with optimizing Canadian methane emissions when wetland emissions are largely
376 uncertain. Our approach of optimizing anthropogenic and natural emissions simultaneously in an inversion is useful because
377 attempting to constrain either emissions category, anthropogenic or natural, obfuscates the analysis on the other. We exploit
378 the different pattern of anthropogenic and natural emissions in time and space (Fig. 4). Natural emissions peak in the
379 summertime and are concentrated in Boreal Canada, while anthropogenic emissions are persistent year-round and are
380 concentrated in Western Canada (Fig. 2). Hence when optimizing the model-observation mismatch in a Bayesian inverse
381 framework, some elements of the observation vector will correspond to high biases from summertime observations in Boreal
382 Canada and some elements will correspond to low biases in Western Canada. As the choice of prior for the inversion we use
383 the mean of the WetCHARTS configurations (14.0 Tg a^{-1}) which corresponds to the middle of the range shown shaded in red
384 in Figure 4. The 60% range of uncertainty in the prior error covariance matrix \mathbf{S}_a appropriately excludes the extreme
385 scenarios in Fig. 3 and 4.



387

388

389 **Figure 3:** Ensemble members from the WetCHARTS v1.0 inventory (Bloom et al., 2017) with totals for wetland methane
 390 emissions within Canada for each configuration shown in Tg CH₄ a⁻¹. Ensemble members vary according to the use of three
 391 CH₄:C q₁₀ temperature dependencies and two inundation extent scenarios (GLWD vs. GLOBCOVER) for 3×2=6 scenarios.



392

393 **Figure 4:** Time series of 2013–2015 modelled and observed methane concentrations. Monthly-mean methane from ECCC in
 394 in situ observations (black) are shown and compared to six GEOS-Chem simulations differing in the use of WetCHARTS
 395 ensemble members for wetland emissions, with other emissions corresponding to Table 2. The six configurations are labelled
 396 GCXY where first digit (X=1,2,3) corresponds to the CH₄:C q₁₀ temperature dependency, which decreases the sensitivity of
 397 emissions to temperature with increasing value. The second digit (Y=3,4) corresponds to the model used for inundation
 398 extent (3 = GLWD, 4 = GLOBCOVER) where GLOBCOVER produces lower emissions in Canada. Emissions
 399 configurations are those shown in Fig. 3 in order of magnitude from red to purple lines, with the shaded red showing the
 400 range of concentrations. Sites are LLB, Alberta (A), ETL, Saskatchewan (B), FRA, Northern Ontario (C) and EGB, Southern
 401 Ontario (D).

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412 3.2 Comparative Analysis of Inversions using ECCC in situ and GOSAT Satellite Data

413 We optimize 2010–2015 emissions in Canada using an $n = 78$ state vector element inversion setup with GOSAT and ECCC
414 data independently. Elements 1–72 of the inversion are monthly total natural emissions (wetlands + other natural) from
415 2010–2015 and elements 73–78 are yearly total anthropogenic emissions (energy + agriculture + waste) for the same years.
416 These categories correspond to the emissions shown in Table 2. We do not optimize emissions according to clustered grid
417 boxes like other satellite inversions using GEOS-Chem (Wecht et al., 2014; Turner et al., 2015; Maasakkers et al., 2019) and
418 instead scale the amplitudes of these two aggregated categories. This approach is a trade-off of time for space, due to the
419 limitations of the observations, giving up finer spatial resolution for finer temporal resolution. This is useful for optimizing
420 Canadian methane emissions since a) anthropogenic emissions are largely concentrated in Western Canada and require less
421 spatial discretization over the entire country and b) natural emissions are the largest source and have an uncertain seasonality
422 – as shown in the previous section – and require finer temporal discretization. The limitations of this method are that natural
423 emissions are very unlikely to be spatially homogenous and vary due to hydrological differences even at the
424 microtopographic level (Bubier et al., 1993). Perfectly resolving Canadian emissions sources in time and space is challenged
425 by the sparsity and precision of the observing system and the model representation of the observations. We show the
426 limitations of the combined ECCC and GOSAT observing system towards resolving subnational emissions in more detail in
427 the subsequent section.

428

429 Figure 5 (top) shows 2010–2015 posterior emissions using this 78 state vector approach with ECCC in situ data (blue) and
430 GOSAT satellite data (green). Error bars are from the diagonal elements of the posterior error covariance matrix $\hat{\mathbf{S}}$. Posterior
431 anthropogenic emissions averaged over the 6 year period are $6.0 \pm 0.4 \text{ Tg a}^{-1}$ (1σ year-to-year variability) using ECCC data
432 and $6.5 \pm 0.7 \text{ Tg a}^{-1}$ using GOSAT data. Posterior estimates are 36% and 48% higher than the prior of 4.4 Tg a^{-1} for ECCC
433 and GOSAT results, respectively. There does not appear to be a significant year-to-year trend above the noise for the
434 anthropogenic emissions optimized by either dataset. The posterior anthropogenic emissions using ECCC and GOSAT data
435 show agreement with each other in each year but 2011, where the GOSAT derived emissions are statistically higher. The
436 error from the diagonal of the posterior error covariance matrix $\hat{\mathbf{S}}$ may be overly optimistic, particularly for GOSAT data.
437 This is due to the observational error covariance matrix \mathbf{S}_o being treated as diagonal when realistically there are correlations
438 between GOSAT observations that are difficult to quantify (Heald et al., 2004). Our results for anthropogenic emissions
439 show agreement with top-down aircraft estimates of methane emissions in Alberta that are higher than bottom-up inventories
440 (Johnson et al., 2017; Baray et al., 2018) and previous satellite inverse-modelling studies over North America that upscale
441 emissions in Western Canada (Turner et al., 2015; Maasakkers et al., 2019; Maasakkers et al., 2021; Lu et al., 2021). We
442 show source attribution through a sectoral and subnational scale analysis of anthropogenic emissions in the subsequent
443 section.

444

445 Inversion results for monthly natural emissions from 2010–2015 are also shown in Figure 5 (bottom). The total of posterior
446 natural emissions averaged over the 6 year period is $11.6 \pm 1.2 \text{ Tg a}^{-1}$ using ECCC data and $11.7 \pm 1.2 \text{ Tg a}^{-1}$ using GOSAT
447 data. The prior for natural emissions is 14.8 Tg a^{-1} from the mean of the WetCHARTS extended ensemble (14.0 Tg a^{-1}) plus
448 other natural (biomass burning + termites + seeps = 0.8 Tg a^{-1}). There is some interannual variability in the prior due to
449 higher emissions in 2010 and 2015. Posterior results averaged over the six years are 22% lower than the prior using ECCC
450 data and 21% lower using GOSAT data, with both posterior results showing agreement with each other. These results are
451 within the uncertainty range of the WetCHARTS extended ensemble, and we show the magnitude of emissions from the
452 larger uncertainty dataset (3.9 to 32.4 Tg a^{-1}) can be better constrained with both ECCC and GOSAT observations.

453

454 While our results show lower natural emissions in all years, a linear fit to the posterior annual emissions using ECCC data
455 shows a trend of increasing natural emissions at a rate of $\sim 0.56 \text{ Tg a}^{-1}$ per year from 2010–2015. The posterior with GOSAT
456 data does not corroborate this result, the overall emissions trend using GOSAT data is not robust and shows a decreasing
457 trend of $\sim 0.2 \text{ Tg a}^{-1}$ per year. The lack of corroboration of trends between ECCC and GOSAT data may be reflective of the
458 lower overall sensitivity of total column methane to these surface fluxes (Sheng et al., 2017; Lu et al., 2021) or the inability
459 of this inverse system to constrain trends sufficiently. The combined ECCC+GOSAT inversion using this setup is consistent
460 with the results of the individual inversions, it is shown in the Supplement (Fig S11) while the intercomparison is
461 emphasized here, although we note the combined inversion also does not corroborate this trend. We evaluate the possible
462 influence of errors in the global model on the projection of a trend onto the ECCC inversion in Section 1.3.2 of the
463 Supplement. While the mean natural emissions over 2010–2015 show consistent results in the sensitivity tests, the
464 limitations of the observation system, the inversion procedure and the timescale of the analysis limit the interpretation of
465 trends. Poulter et al. (2017) estimated global wetland emissions using biogeochemical process models constrained by
466 inundation and wetlands extend data. They estimated mean annual emissions over all of Boreal North America to be $25.1 \pm$
467 11.3 Tg a^{-1} in 2000–2006, $26.1 \pm 11.8 \text{ Tg a}^{-1}$ in 2007–2012 and $27.1 \pm 12.5 \text{ Tg a}^{-1}$ which suggests a small increasing trend.
468 Observational constraints over longer timescales are necessary to investigate the possibility of trends in Canadian natural
469 methane emissions. Improvements to the observation network and a better understanding of climate sensitivity in
470 WetCHARTS are necessary to understand how wetlands methane emissions will evolve in future climates.

471

472 Figure 6 shows the 2010–2015 average seasonal pattern of natural emissions in the prior and posterior results. The
473 seasonality of natural methane emissions in the prior shows a sharp peak in July with a narrow methanogenic growing
474 season. The posterior with ECCC data shows a peak 1-month later in August in most years instead of July, with lower than
475 prior emissions in the spring months before the peak (March to May) and similar emissions to the prior in the autumn
476 months after the peak (September to November). Posterior emissions with GOSAT show a peak in July and corroborates the
477 pattern of slower-to-begin spring emissions and the lower intensity summer peak seen from the ECCC inversion. The

478 posterior results show the seasonality of emissions is not symmetrical around the temperature peak in July. August emissions
479 are higher than June, September emissions are higher than May, and October emissions are higher than April. This pattern
480 around July is present in the prior emissions from WetCHARTS, however the inversion results constrained by ECCC or
481 GOSAT observations intensify the relative difference between emissions before and after July. Miller et al. (2016) found a
482 similar seasonal pattern of emissions in the Hudson Bay Lowlands using an inverse model constrained by 2007–2008 in situ
483 data. They found a less narrow and less intense peak of summertime emissions with higher autumn over spring emissions.
484 Warwick et al. (2016) used a forward model and isotopic measurements of $\delta^{13}\text{C}\text{-CH}_4$ and $\delta\text{D}\text{-CH}_4$ from 2005–2009 to show
485 northern wetland emissions should peak in August-September with a later spring kick-off and later autumn decline. This is
486 further corroborated by Arctic methane measurements (Thonat et al., 2017) and high latitude eddy covariance measurements
487 (Peltola et al., 2019; Treat et al., 2018; Zona et al., 2016) that show a larger contribution from the nongrowing season. Our
488 inverse model results using ECCC and GOSAT data both show agreement with slower to start emissions in the spring and a
489 less intense summertime peak for Canadian wetland emissions.

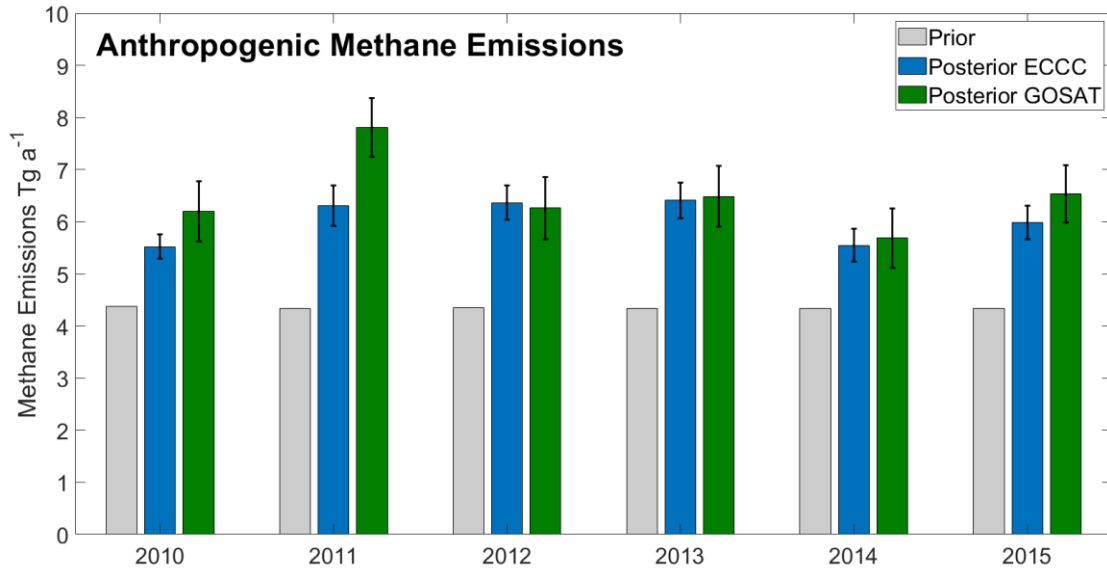
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491 Several mechanisms have been proposed to describe a larger relative contribution from cold season methane emissions.
492 Pickett-Heaps et al. (2011) attributed a delayed spring onset in the HBL to the suppression of emissions by snow cover. The
493 temperature dependency in WetCHARTS is based on surface skin temperature (Bloom et al., 2017), however subsurface soil
494 temperatures may continue to sustain methane emissions while the surface is below freezing. When subsurface soil
495 temperatures are near 0°C , this “zero curtain” period can further continue to release methane for an extended period (Zona et
496 al., 2016). Subsurface soils may remain unfrozen at a depth of 40 cm even until December (Miller et al., 2016).
497 Alternatively, field studies in the 1990’s suggested the seasonality of emissions may be more influenced by hydrology than
498 temperature, with large differences between peatlands sites (Moore et al., 1994). The WetCHARTS extended ensemble
499 inundation extent variable is constrained seasonally by precipitation. While this does not directly constrain water table depth
500 and wetland extent it provides an aggregate constraint on hydrological variability (Bloom et al., 2017). We show the mean
501 seasonal pattern of both air temperature and precipitation from climatological measurements in subarctic Canada are
502 similarly asymmetrical about the July peak (Fig. S2 in the Supplement). August is warmer and wetter than June, September
503 is warmer and wetter than May, and October is wetter and warmer than April – with wetness being more persistent into the
504 autumn than air temperature. Our inversion results showing a delayed spring start in the seasonal pattern of natural methane
505 emissions in Canada may suggest a lag in the response of methane emissions to temperature and precipitation. This may be
506 due to lingering subsurface soil temperatures and/or more complex parametrization necessary for hydrology.

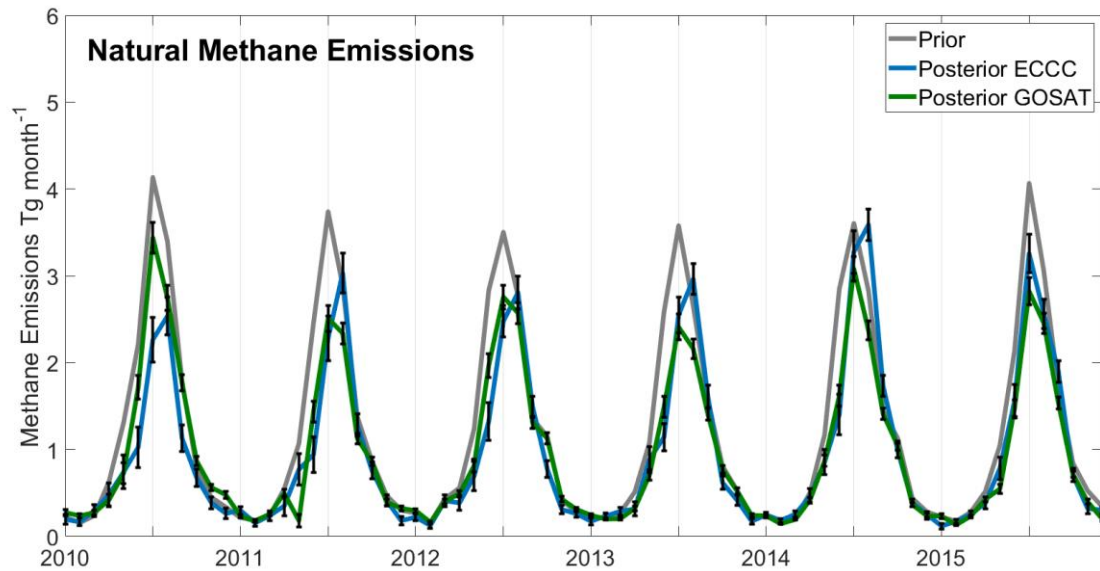
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508 The overall agreement between ECCC and GOSAT inversions shows robustness in the results. While the same model, prior
509 emissions and inversion procedure are used for assimilating ECCC and GOSAT data, the two datasets are produced with
510 very different measurement methodologies (in situ vs. remote sensing) and sample different parts of the atmosphere (surface
511 concentrations or the total vertical column). The posterior error intervals shown from $\hat{\mathbf{S}}$ reflect assumptions about the

512 treatment of observations and may insufficiently account for correlations, however the comparative analysis provides a
513 useful sensitivity test of the posterior emissions since the datasets reflect different treatment of these assumptions.
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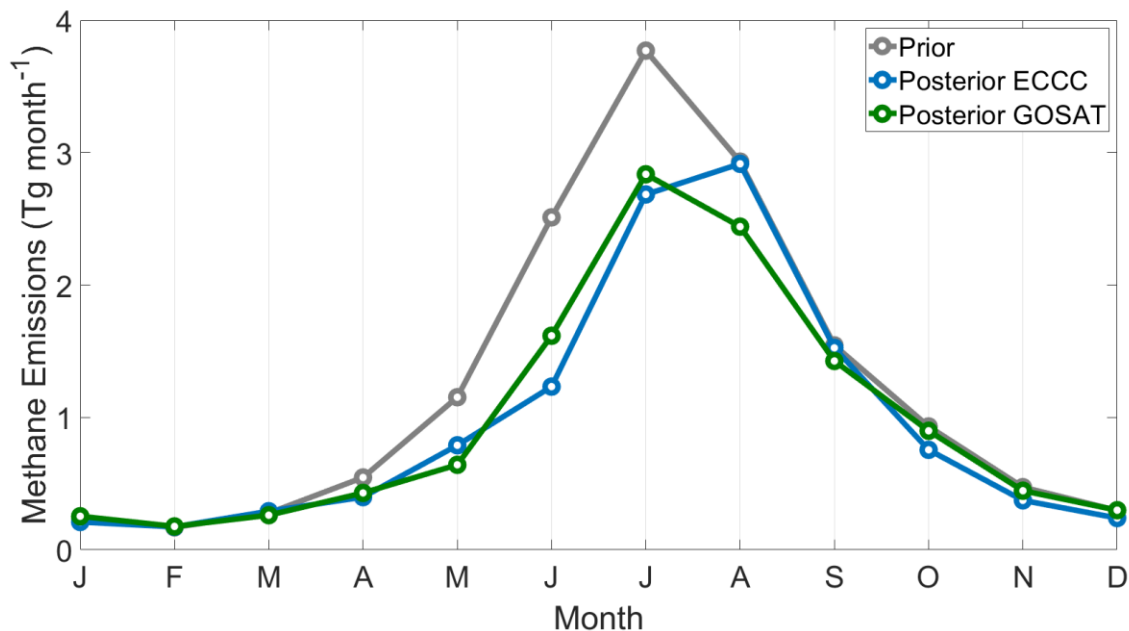
517 **Figure 5:** Comparative analysis of inversion results optimizing annual total Canadian anthropogenic emissions (top) and
518 monthly total natural emissions (bottom) in an $n = 78$ state-vector element setup. The posterior emissions determined using

519 ECCC in situ (blue) and GOSAT satellite (green) data are compared to the prior (gray). Error bars are from the diagonal
520 elements of the posterior error covariance matrix.

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524

525 **Figure 6:** Mean 2010–2015 seasonal pattern of natural methane emissions in Tg month⁻¹. The annual total emissions are
526 14.8 Tg a⁻¹ (prior, gray), 11.6 ± 1.2 Tg a⁻¹ (posterior ECCC, blue) and 11.7 ± 1.2 Tg a⁻¹ (posterior GOSAT, green). The
527 posterior results are within the uncertainty range provided by the WetCHARTS extended ensemble (3.9–32.4 Tg a⁻¹ for
528 Canada).

529

530 3.3 Joint-inversions Combining ECCC In situ and GOSAT Satellite Data

531 We combine the ECCC and GOSAT datasets in two policy-themed inversions: (1) optimizing emissions according to the
532 sectors in the national inventory ($n = 5$ state vector elements; corresponding to the categories in Table 2) and (2) optimizing
533 emissions by provinces split into anthropogenic and natural totals ($n = 16$) and show the results in Figure 7. These inversions
534 are under-determined and show the limitations of the ECCC+GOSAT observing system towards constraining emissions in
535 Canada with very small magnitudes. We conduct the inversions for each year from 2010–2015 individually and present the
536 average from these six samples. Since these two policy inversions use a low number of state vector elements, they are
537 vulnerable to both aggregation error and overfitting of the well-constrained state vector elements and do not necessarily

538 benefit from using a larger data vector from all six years. We discuss the diagnostics and information content for these
539 inversions in detail in Section 1.4 of the Supplement. The error bars are the 1σ standard deviation of the six yearly results
540 and therefore represent both noise in the inversion procedure and year-to-year differences in the state (emissions and/or
541 transport). Here we do not apply a weighting factor to either dataset, the observations are treated equivalently for the cost
542 function in eq. (1). While there are about 5 times more GOSAT observations than ECCO observations for use in the analysis
543 and the in-situ observations have larger observational error in \mathbf{S}_a (due to model error), the surface measurements are much
544 more sensitive to surface fluxes, which offsets the weight of the larger amount of GOSAT data. As further diagnostics we
545 show the inversions using GOSAT and ECCO individually (Table S4 and S5) which show general agreement between the
546 datasets. We also use a singular value decomposition eigenanalysis (Heald et al., 2004) to evaluate the independence of the
547 state vector elements and to demonstrate which sectoral categories and provinces can be reliably constrained above the noise
548 in the system (Fig. S9 and S10 in the Supplement).

549

550 Figure 7 (top) shows the sectoral inversion corresponding to categories in the National Inventory (Table 2). The prior
551 emissions with 50% error estimates (60% for wetlands) are 2.4 Tg a⁻¹ (Energy), 1.0 Tg a⁻¹ (Agriculture), 0.9 Tg a⁻¹ (Waste),
552 14.0 Tg a⁻¹ (Wetlands) and 0.8 Tg a⁻¹ (Other Natural). The posterior emissions are 3.6 ± 0.9 Tg a⁻¹ (Energy), 1.5 ± 0.4 Tg a⁻¹
553 (Agriculture), 0.8 ± 0.2 Tg a⁻¹ (Waste), 9.6 ± 1.1 Tg a⁻¹ (Wetlands), and 1.7 ± 0.9 Tg a⁻¹ (Other Natural). The degrees of
554 freedom for signal and singular value decomposition (Fig. S9) show 3–4 independent pieces of information can be retrieved,
555 which are differentiated in the figure by solid and hatched bars. The singular value decomposition shows strong source
556 signals corresponding to wetlands and energy with signal-to-noise ratios of ~ 37 and ~ 5 , respectively. These are the two
557 largest emissions sources in Canada and show the inverse system can successfully disentangle the major anthropogenic and
558 natural contributors. Emissions from waste have a signal-to-noise ratio of ~ 2 and can be constrained despite the low
559 magnitude of emissions. This is likely due to waste emissions being more concentrated in Central Canada and away from the
560 influence of large energy and agriculture emissions in Western Canada. Emissions from other natural sources are at the noise
561 limit and show a moderate correlation with wetlands, which shows that these two sources are not completely independent.
562 Agriculture emissions are below the noise in the system and highly correlated with energy emissions. This is likely due to
563 the high spatial overlap of energy and agriculture emissions in Western Canada. As a result of these limitations, we present
564 the total of energy and agriculture as 5.1 ± 1.0 Tg a⁻¹ and the total of wetlands and other natural as 11.3 ± 1.4 Tg a⁻¹. Our
565 results for total natural and total anthropogenic emissions are consistent with the results from the previous monthly
566 inversion, with the added benefit of identifying which sectors are responsible for the higher anthropogenic emissions at the
567 cost of lower temporal resolution. Waste emissions are 15% lower than the prior and 14% lower than the National Inventory.
568 The total for energy and agriculture is 49% higher than the prior and 59% higher than the total in the inventory. These results
569 show that energy and/or agriculture are the sectors that are responsible for the higher anthropogenic emissions.

570

571 Figure 7 (bottom) shows the provincial inversion corresponding to the six largest emitting provinces (BC British Columbia,
572 AB Alberta, SK, Saskatchewan, MB Manitoba, ON Ontario, QC Quebec) and two aggregated regions (ATL Atlantic
573 Canada, NOR Northern Territories). These regions are further subdivided into total anthropogenic and total methane
574 emissions, with below detection limit anthropogenic emissions from Atlantic Canada and Northern Territories. This
575 inversion especially challenges the limitations of the ECCC+GOSAT observation system, as only about 8 of 16 independent
576 pieces of information are retrieved. This means that half of the posterior provincial emissions are below the noise, and we are
577 unable to constrain province-by-province emissions. The singular value decomposition identifies which regions are well
578 constrained (Fig. S10). For the anthropogenic emissions AB and ON are strongly constrained. For the natural emissions AB,
579 ON, SK and MB are well constrained. BC shows correlation between its own anthropogenic and natural emissions and
580 cannot be completely disaggregated. As a result, we group elements together in Western Canada (BC + AB + SA + MB) and
581 Central Canada (ON + QC) for interpretation. The total for Western Canada anthropogenic emissions is $4.7 \pm 0.6 \text{ Tg a}^{-1}$
582 which is 42% higher than the prior of 3.3 Tg a^{-1} . The total for Central Canada is $0.8 \pm 0.2 \text{ Tg a}^{-1}$ which is 11% lower than the
583 prior of 0.9 Tg a^{-1} .

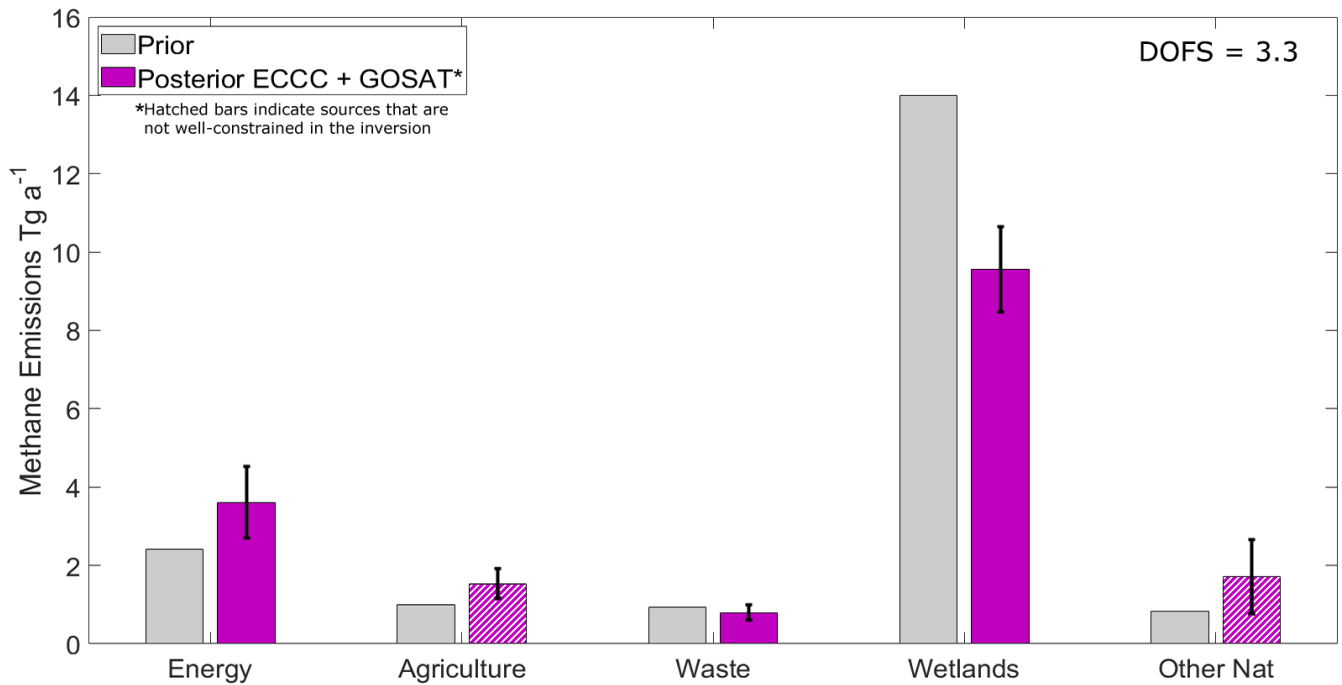
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585 Each of our top-down inversion results show higher total anthropogenic emissions than bottom-up estimates. This is
586 consistent regardless of the observation vector incorporating ECCC data, GOSAT data or ECCC+GOSAT data. The
587 subnational scale emissions are limited in their ability to provide full characterization of minor emissions across Canada but
588 can successfully constrain major emissions for source attribution. The sectoral inversion attributes higher anthropogenic
589 emissions to energy and/or agriculture and applies a small decrease to waste emissions. The provincial inversion attributes
590 higher anthropogenic emissions to Western Canada and a small decrease to Central Canada. These results suggest that
591 anthropogenic emissions in Canada are underestimated primarily because of higher emissions from Western Canada energy
592 and/or agriculture. This interpretation is consistent with previous satellite inverse modelling studies over North America that
593 apply positive scaling factors to grid box clusters in Western Canada to match observations (Maasackers et al., 2019; Turner
594 et al., 2015; Wecht et al., 2014). Aircraft studies in Alberta have also shown higher emissions from oil and gas in Alberta
595 than bottom up estimates (Baray et al., 2018; Johnson et al., 2017). Atherton et al. (2017) estimated higher emissions from
596 natural gas in north-eastern British Columbia using mobile surface in situ measurements (Atherton et al., 2017). Zavala-
597 Araiza et al. (2018) showed a significant amount of methane emissions in Alberta from equipment leaks and venting go
598 unreported due to current reporting requirements and in some regions a small number of sites may be responsible for most
599 methane emissions. Our inverse modelling results from 2010–2015 suggest a consistent presence of under-reported or
600 unreported emissions which require a policy adjustment to reporting practices.

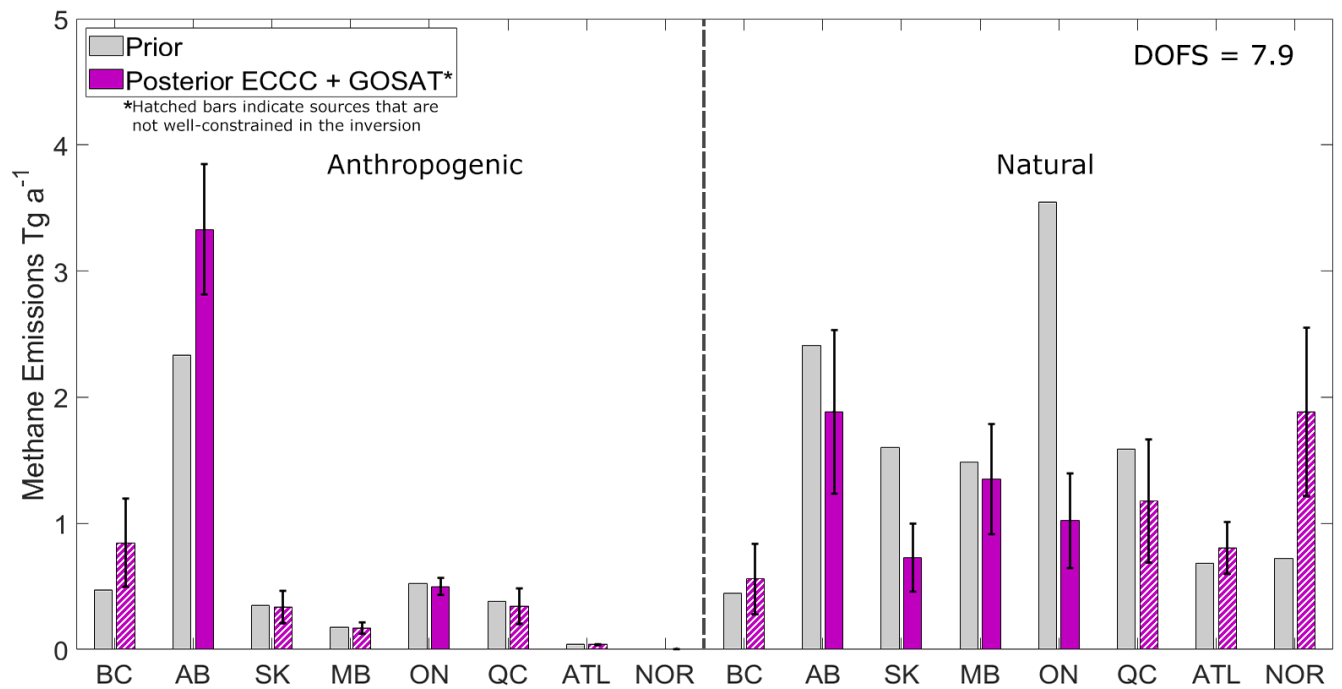
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Figure 7: Joint-inversions combining 2010–2015 ECCC in situ and GOSAT satellite data showing how the combined observing system remains limited towards resolving all Canadian sources. Inversions are done for each year and we present

609 the six-year average with error bars showing the 1σ standard deviation of the yearly results. Hatched bars indicate sources
610 that are not well-constrained, these are defined as state vector elements with averaging kernel sensitivities less than 0.8
611 which are affected by aliasing with other sources (See Supplemental Fig. 9 and 10). The top panel shows the sectoral
612 inversion according to the categories in the National Inventory (Energy, Agriculture, Waste) and two natural categories
613 (Wetlands and Other Natural). As an example, the diagnostics in Figure S9 shows Agriculture emissions are beneath the
614 noise and cannot be distinguished from Energy. The bottom panel shows the subnational regional inversion according to
615 provinces (BC British Columbia, AB Alberta, SK, Saskatchewan, MB Manitoba, ON Ontario, QC Quebec) and aggregated
616 regions (ATL Atlantic Canada, NOR Northern Territories) further subdivided according to total anthropogenic and total
617 natural emissions. The diagnostics in Fig. S10 show more than half of the regions are at or below the noise. For
618 anthropogenic emissions, the best constraints are on provinces AB and ON. For natural emissions, the best constraints are on
619 AB, SK, MB and ON.

620 **3.4 Comparison to Independent Aircraft and In situ Data**

621 We test the robustness of the optimized emissions from each of the three inversions shown (monthly natural, sectoral, and
622 provincial) by comparing to independent measurements not used in the inversions. Prior and posterior simulated methane
623 concentrations are compared to measurements from NOAA ESRL aircraft profiles at East Trout Lake, Saskatchewan (Mund
624 et al., 2017) and ECCC surface measurements in sites Chapais and Chibougamau in Quebec, Canada. The surface data was
625 averaged to daily afternoon means (12:00 to 16:00 local time) in the same manner as the surface measurements used in the
626 inversion. Aircraft data from the NOAA ESRL profiles coincide spatially with the surface measurements at ETL through a
627 joint analysis program with Environment and Climate Change Canada and have occurred on a regular basis approximately
628 once a month from 2005 until present time. Aircraft measurements reach ~ 7000 m above the surface with samples at
629 multiple altitudes accomplished using a programmable multi-flask system that is further discussed in Mund et al. (2017),
630 however we limit the comparison to the lowest 1 km above ground since higher altitude measurements are mostly
631 background. The aircraft data is not averaged however the flights occur around the same time in the early afternoon.

632

633 Figure 8 shows the comparison using reduced-major axis (RMA) regressions with the coefficient of determination (R^2), the
634 slope and the mean-bias shown as metrics to evaluate the agreement. Surface data in CHA, Quebec shows better posterior
635 agreement with observations according to all metrics for each of the three inversions. The R^2 of the prior is 0.36 and
636 improves to a range of 0.44–0.49 for the posterior results, the slope is 1.17 in the prior and improves to a range of 0.92–1.12
637 and the mean bias (model – observations) is +16.4 ppb in the prior and improves to +13.2 and +5.6 ppb. Since this site in
638 Quebec is particularly sensitive to the Hudson Bay Lowlands, the agreement in all metrics suggests our posterior emissions
639 can better represent wetland emissions in this region. This includes the reduced peak seasonality of natural emissions in the
640 monthly inversion, the reduction of wetland emissions in the sectoral inversion and the reduction of natural emissions
641 primarily in Central Canada in the provincial inversion. Aircraft data in Saskatchewan shows improvement in the R^2 and

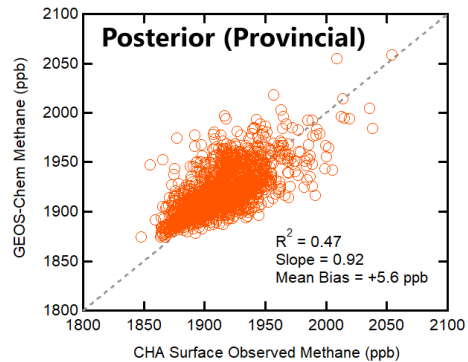
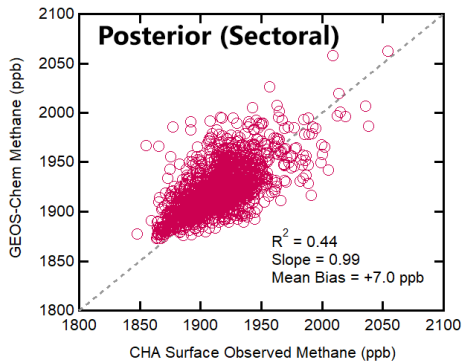
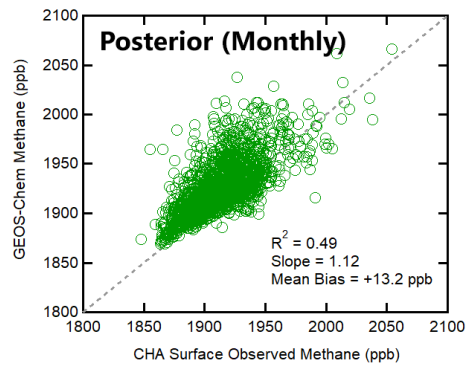
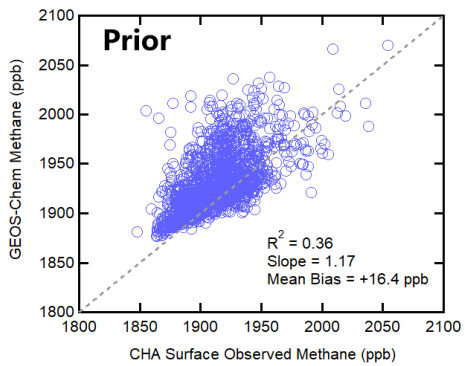
642 mean bias metrics but slightly degrades the slope in one case. The R^2 of the prior is 0.14 and improves to a range of 0.20–
643 0.30, the mean bias of the prior is +6.8 ppb and improves to +1.2 and +3.1 ppb. The slope of the prior is 1.15 which slightly
644 degrades to 0.83 in the monthly inversion and improves to a range of 0.88–0.93 in the provincial and sectoral inversions. The
645 high resolution aircraft measurements are more susceptible to representation error at this $2^\circ \times 2.5^\circ$ grid resolution.
646 Furthermore, the time-series comparison to surface data at East Trout Lake (Fig. 4) shows overall lower sensitivity to
647 summertime wetland emissions than Fraserdale and Egbert, and lower sensitivity to anthropogenic emissions from Alberta
648 than Lac La Biche. Hence the modelled methane concentrations at the aircraft measurement points are adjusted less by the
649 change in posterior emissions. However, improvement in the R^2 and mean bias metrics show there is still a better
650 representation of the variance in the data which suggests the posterior emissions reduce bias due to peak emission episodes.

651

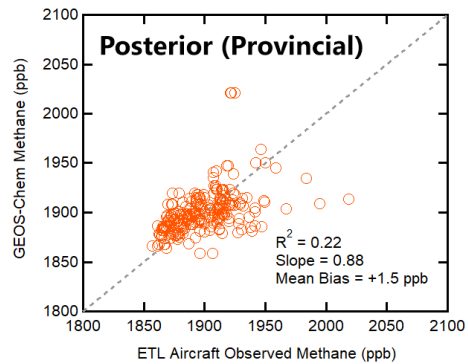
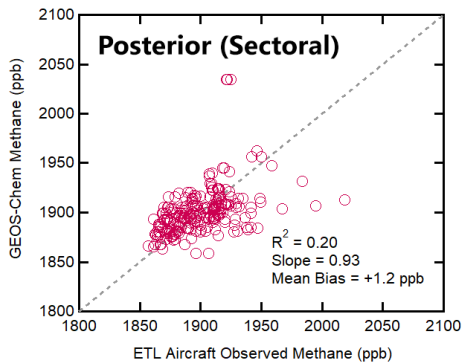
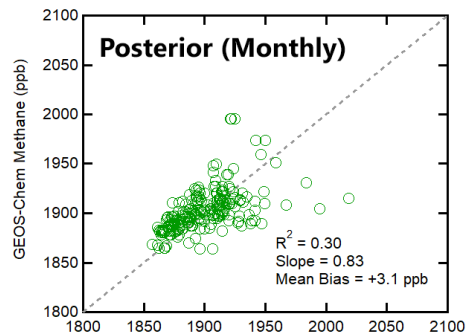
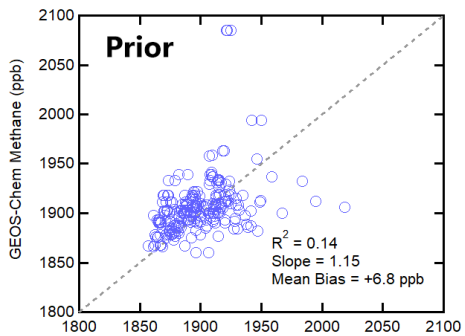
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ECCC Surface at CHA, Quebec



NOAA Aircraft at ETL, Saskatchewan



655 **Figure 8:** Evaluation of inversion results with reduced-major axis (RMA) regressions using independent data. The top four
656 panels show the comparison to ECCC surface observations at Chapais and Chibougamau in Quebec, Canada and the bottom
657 four panels show the comparison to NOAA aircraft profiles at East Trout Lake, Saskatchewan. The agreement of
658 observations with prior simulated methane concentrations (blue) are compared to posterior concentrations using optimized
659 emissions from the monthly inversion (green), the sectoral inversion (magenta), and the provincial inversion (orange). The
660 coefficient of determination (R^2), slope and mean bias are shown as metrics of agreement.

661 **4 Conclusions**

662 We conduct a Bayesian inverse analysis to optimize anthropogenic and natural methane emissions in Canada using 2010–
663 2015 ECCC in situ and GOSAT satellite observations in GEOS-Chem. Methane concentrations are simulated on a $2^\circ \times 2.5^\circ$
664 grid using recently updated prior emissions inventories for energy and wetland emissions in Canada. Modelled background
665 conditions for the Canadian domain are shown to be unbiased in the comparison to surface in situ data at the western most
666 site in Canada, Estevan point, with agreement within 6 ppb. A forward model analysis shows much larger biases between –
667 100 ppb and +1050 ppb at surface sites throughout Canada demonstrating the presence of misrepresented local emissions.
668 We show large positive biases (overestimation of emissions) in the summertime are observed at sites sensitive to wetland
669 emissions, these biases are reduced by using lower magnitude wetland emissions scenarios with lower $\text{CH}_4:\text{C}$ temperature
670 sensitivities or lower inundation extent. We also show the opposite case of negative biases (underestimation of emissions)
671 observed year-round at sites in Western Canada. The forward model analysis is consistent with the results of the inverse
672 analysis that reduce emissions from natural sources and increase emissions from anthropogenic sources to minimize the
673 mismatch between modelled and observed methane.

674

675 We show three approaches for using ECCC and GOSAT data towards inverse modelling of Canadian methane emissions.
676 These approaches differ according to the temporal and spatial resolution of the solution. We show: (1) a relatively higher
677 time-resolution inversion that solves for natural emissions each month from 2010–2015 and anthropogenic emissions as
678 yearly totals, (2) a sectoral inversion that solves for emissions according to categories in the National Inventory, (3) a
679 provincial inversion that solves for total anthropogenic and natural emissions at the subnational level. The monthly inversion
680 provides information on the seasonality of natural emissions (which are ~95% wetlands) but does not provide more depth
681 into anthropogenic emissions beyond yearly scaling. The sectoral inversion provides more information on the categories of
682 anthropogenic emissions that are misrepresented in the prior but without spatial detail. The provincial inversion provides the
683 highest level of spatial discretization but is largely underdetermined due to the limitations of the observing system towards
684 characterizing very low magnitude emissions from smaller contributing provinces.

685

686 Inversion results (1) show mean 2010–2015 posterior emissions for total anthropogenic sources in Canada are $6.0 \pm 0.4 \text{ Tg a}^{-1}$
687 ¹ using ECCC data and $6.5 \pm 0.7 \text{ Tg a}^{-1}$ using GOSAT data. Annual mean natural emissions are $11.6 \pm 1.2 \text{ Tg a}^{-1}$ using
688 ECCC data and $11.7 \pm 1.2 \text{ Tg a}^{-1}$ using GOSAT data. Both inverse modelling estimates are higher than the prior for
689 anthropogenic emissions 4.4 Tg a^{-1} and lower than the prior for natural emissions 14.8 Tg a^{-1} . Inversion results using both
690 datasets show a change in the seasonal profile of natural methane emissions where emissions are slower to begin in the
691 spring and show a less intense peak in the summer. The agreement between two datasets assembled with different
692 measurement methodologies that sample different parts of the atmosphere is a robust result that lends weight to our
693 conclusions. Our results corroborate recent studies showing a less-intense and less-narrow summertime peak in North
694 American Boreal wetland emissions with a higher relative contribution from the cold season (Miller et al., 2016; Zona et al.,
695 2016; Warwick et al., 2016; Thonat et al., 2017; Treat et al., 2018; Peltola et al., 2019). These top-down studies using
696 atmospheric observations show biosphere process models can better account for a more complex response to peak surface
697 soil temperatures.

698

699 We also conduct combined ECCC+GOSAT inversions that aim to resolve finer resolution emissions corresponding to (2) the
700 sectors of the National Inventory and corresponding to (3) provincial boundaries. These policy-themed inversions challenge
701 the capabilities of the ECCC+GOSAT observation system and show the system is not capable of resolving many minor
702 emissions in Canada. The degrees of freedom for signal for these inversions are 3–4 out of 5 state vector elements for the
703 sectoral inversion and 8 out of 16 for the provincial inversion. The limitation of this inverse approach towards constraining
704 sectoral or regional scale emissions in Canada is due to the low magnitude of these emissions, their overlapping nature in
705 concentrated regions, and the sparsity of data available to distinguish them apart. Grouping correlated sectors together, we
706 determine $5.1 \pm 1.0 \text{ Tg a}^{-1}$ for energy and agriculture which is 59% higher than the inventory, $0.8 \pm 0.2 \text{ Tg a}^{-1}$ for waste
707 which is 14% lower than the inventory. For provincial emissions, we show Western Canada is $4.7 \pm 0.6 \text{ Tg a}^{-1}$ which is 42%
708 higher than the prior and Central Canada is 0.8 ± 0.2 which is 11% lower. Both regions show lower natural emissions. These
709 results show that the higher anthropogenic emissions in the posterior results can be attributed to energy and/or agriculture
710 primarily in Western Canada where most of Canadian anthropogenic emissions are concentrated. Our results are consistent
711 with other top-down studies that show higher than reported anthropogenic emissions in Western Canada (Wecht et al., 2014;
712 Turner et al., 2015; Atherton et al., 2017; Johnson et al., 2017; Baray et al., 2018; Maasackers et al., 2019). This may be due
713 to oil and gas emissions that are under-reported or unreported due to current reporting requirements (Zavala-Araiza et al.,
714 2018). These top-down studies show a need for policy readjustment in reporting practices for Canadian anthropogenic
715 methane emissions.

716

717 This study shows the value of using complementary surface and satellite datasets in an inverse analysis. We emphasize the
718 value of comparative analysis using the datasets independently versus as joint inversions, as minor emissions are too low in
719 magnitude for the observational precision to distinguish finer scale discretization above the noise. The comparative analysis

720 has the added benefit of evaluating the datasets against each other and the assumptions that are specific to using either
721 surface or satellite data. The capabilities for combining and intercomparing datasets is expected to improve, with the launch
722 of Copernicus Sentinel-5p satellite (TROPOMI) in 2017 and continued expansions on in situ observation networks. The
723 ability for next generation observations to constrain subnational level emissions in Canada will depend on instrument and
724 model precision, as well as the emissions magnitudes and spatiotemporal overlap of the targets. These technical capabilities
725 should be weighed alongside policy needs for improved methane monitoring.

726

727 **Competing Interests**

728 The authors declare that they have no conflict of interest.

729 **Data Availability**

730 GEOS-Chem is from <http://acmg.seas.harvard.edu/geos/> which includes links to all gridded prior emissions and
731 meteorological fields used in this analysis. GOSAT satellite data is from the University of Leicester v7 proxy retrieval is
732 available through the Copernicus Climate Change Service <https://climate.copernicus.eu/>. ECCC in situ data is available
733 through the World Data Centre for Greenhouse Gases (WDCGG) <https://gaw.kishou.go.jp/>. NOAA/ESRL aircraft data is
734 from the Global Monitoring Laboratory <https://www.esrl.noaa.gov/gmd/ccgg/aircraft/>.

735 **Author Contributions**

736 SB, DJJ and RM designed the study. SB conducted the simulations and analysis with contributions from JDM, JXS, MPS,
737 and DBAJ. AAB provided WetCHARTS emissions and supporting data. SB and RM wrote the paper with contributions
738 from all authors.

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744 Gases (CCGG) cooperative air sampling network measurements.

745

746

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