Comprehensive evaluations of diurnal NO₂ measurements during DISCOVER-AQ 2011: Effects of resolution dependent representation of NO_x emissions

Jianfeng Li^{1, a}, Yuhang Wang^{1*}, Ruixiong Zhang¹, Charles Smeltzer¹, Andrew Weinheimer², Jay Herman³, K.
Folkert Boersma^{4, 5}, Edward A. Celarier^{6, 7, b}, Russell W. Long⁸, James J. Szykman⁸, Ruben Delgado³, Anne M.
Thompson⁶, Travis N. Knepp^{9, 10}, Lok N. Lamsal⁶, Scott J. Janz⁶, Matthew G. Kowalewski⁶, Xiong Liu¹¹,

Caroline R. Nowlan¹¹

- 7
- 8 ¹School of Earth and Atmospheric Sciences, Georgia Institute of Technology, Atlanta, Georgia, USA
- 9 ²National Center for Atmospheric Research, Boulder, Colorado, USA
- ³University of Maryland Baltimore County JCET, Baltimore, Maryland, USA
- ⁴Royal Netherlands Meteorological Institute, De Bilt, the Netherlands
- ⁵Wageningen University, Meteorology and Air Quality Group, Wageningen, the Netherlands
- 13 ⁶NASA Goddard Space Flight Center, Greenbelt, Maryland, USA
- ⁷Universities Space Research Association, Columbia, Maryland, USA
- ¹⁵ ⁸National Exposure Research Laboratory, Office of Research and Development, U.S. Environmental Protection
- 16 Agency, Research Triangle Park, NC, USA
- 17 ⁹NASA Langley Research Center, Virginia, USA
- 18 ¹⁰Science Systems and Applications, Inc., Hampton, Virginia, USA
- 19 ¹¹Harvard-Smithsonian Center for Astrophysics, Cambridge, Massachusetts, USA
- 20
- ^anow at: Atmospheric Sciences and Global Change Division, Pacific Northwest National Laboratory, Richland,
- 22 Washington, USA
- 23 ^bnow at: Digital Spec, Tyson's Corner, VA, USA
- ^{*} *Correspondence to* Yuhang Wang (yuhang.wang@eas.gatech.edu)
- 25

26 Abstract

27	Nitrogen oxides ($NO_x = NO + NO_2$) play a crucial role in the formation of ozone and secondary inorganic and
28	organic aerosols, thus affecting human health, global radiation budget, and climate. The diurnal and spatial
29	variations of NO2 are functions of emissions, advection, deposition, vertical mixing, and chemistry. Their
30	observations, therefore, provide useful constraints in our understanding of these factors. We employ a Regional
31	chEmical and trAnsport model (REAM) to analyze the observed temporal (diurnal cycles) and spatial
32	distributions of NO ₂ concentrations and tropospheric vertical column densities (TVCDs) using aircraft in situ
33	measurements, surface EPA Air Quality System (AQS) observations, as well as the measurements of TVCDs by
34	satellite instruments (OMI: the Ozone Monitoring Instrument; and GOME-2A: Global Ozone Monitoring
35	Experiment – 2A), ground-based Pandora, and the Airborne Compact Atmospheric Mapper (ACAM) instrument,
36	in July 2011 during the DISCOVER-AQ campaign over the Baltimore-Washington region. The model
37	simulations at 36- and 4-km resolutions are in reasonably good agreement with the regional mean temporospatial
38	NO ₂ observations in the daytime. However, nighttime mixing in the model needs to be enhanced to reproduce the
39	observed NO ₂ diurnal cycle in the model. Another discrepancy is that Pandora measured NO ₂ TVCDs show much
40	less variation in the late afternoon than simulated in the model. The higher resolution 4-kmRelative to the 36-km
41	model simulations tend to, the 4-km model results show larger biases compared to the observations due largely to
42	the larger spatial variations of $NO_2 NO_x$ emissions in the model when the model spatial resolution is increased
43	from 36 to 4 km, although the biases are often comparable to the ranges of the observations. OMI, GOME-2A,
44	and tThe high-resolution aircraft ACAM observations show a more dispersed distribution of NO2 vertical column
45	densities (VCDs) and lower VCDs in urban regions than corresponding 36- and 4-km model simulations,
46	reflecting likely the spatial distribution bias of NO _x emissions in the National Emissions Inventory (NEI) 2011-at
47	high resolution.

48 **1 Introduction**

49 Nitrogen oxides ($NO_x = NO + NO_2$) are among the most important trace gases in the atmosphere due to their 50 crucial role in the formation of ozone (O_3) , secondary aerosols, and their role in the chemical transformation of 51 other atmospheric species, such as carbon monoxide (CO) and volatile organic compounds (VOCs) (Cheng et al., 52 2017; Cheng et al., 2018; Fisher et al., 2016; Li et al., 2019; Liu et al., 2012; Ng et al., 2017; Peng et al., 2016; 53 Zhang and Wang, 2016). NO_x is emitted by both anthropogenic activities and natural sources. Anthropogenic 54 sources account for about 77% of the global NO_x emissions, and fossil fuel combustion and industrial processes 55 are the primary anthropogenic sources, which contribute to about 75% of the anthropogenic emissions (Seinfeld 56 and Pandis, 2016). Other important anthropogenic sources include agriculture and biomass and biofuel burning. 57 Soils and lightning are two major natural sources. Most NO_x is emitted as NO, which is then oxidized to NO_2 by 58 oxidants, such as O_3 , the hydroperoxyl radical (HO₂), and organic peroxy radicals (RO₂).

59 The diurnal variations of NO₂ controlled by physical and chemical processes reflect the temporal patterns of 60 these underlying controlling factors, such as NO_x emissions, chemistry, deposition, advection, diffusion, and 61 convection. Therefore, the observations of NO_2 diurnal cycles can be used to evaluate our understanding of NO_x 62 related emission, chemistry, and physical processes (Frey et al., 2013; Jones et al., 2000; Judd et al., 2018). For 63 example, Brown et al. (2004) analyzed the diurnal patterns of surface NO, NO₂, NO₃, N₂O₅, HNO₃, OH, and O₃ 64 concentrations along the East Coast of the United States (U.S.) during the New England Air Quality Study 65 (NEAQS) campaign in the summer of 2002 and found that the predominant nighttime sink of NO_x through the 66 hydrolysis of N_2O_5 had an efficiency on par with daytime photochemical loss over the ocean surface off the New 67 England coast. Van Stratum et al. (2012) investigated the contribution of boundary layer dynamics to chemistry 68 evolution during the DOMINO (Diel Oxidant Mechanisms in relation to Nitrogen Oxides) campaign in 2008 in 69 Spain and found that entrainment and boundary layer growth in daytime influenced mixed-layer NO and NO₂

70	diurnal cycles on the same order of chemical transformations. David and Nair (2011) found that the diurnal
71	patterns of surface NO, NO ₂ , and O ₃ concentrations at a tropical coastal station in India from November 2007 to
72	May 2009 were closely associated with sea breeze and land breeze which affected the availability of NO _x through
73	transport. They also thought that monsoon-associated synoptic wind patterns could strongly influence the
74	magnitudes of NO, NO ₂ , and O ₃ diurnal cycles. The monsoon effect on surface NO, NO ₂ , and O ₃ diurnal cycles
75	was also observed in China by Tu et al. (2007) on the basis of continuous measurements of NO, NO ₂ , and O ₃ at
76	an urban site in Nanjing from January 2000 – February 2003.

77 In addition to surface NO₂ diurnal cycles, the daily variations of NO₂ vertical column densities (VCDs) were 78 also investigated in previous studies. For example, Boersma et al. (2008) compared NO₂ tropospheric VCDs 79 (TVCDs) retrieved from OMI (the Ozone Monitoring Instrument) and SCIAMACHY (SCanning Imaging 80 Absorption SpectroMeter for Atmospheric CHartography) in August 2006 around the world. They found that the 81 diurnal patterns of different types of NO_x emissions could strongly affect the NO_2 TVCD variations between 82 OMI and SCIAMACHY and that intense afternoon fire activity resulted in an increase of NO₂ TVCDs from 83 10:00 LT (local time) to 13:30 LT over tropical biomass burning regions. Boersma et al. (2009) further 84 investigated the NO₂ TVCD change from SCIAMACHY to OMI in different seasons of 2006 in Israeli cities and 85 found that there was a slight increase of NO₂ TVCDs from SCIAMACHY to OMI in winter due to increased NO_x 86 emissions from 10:00 LT to 13:30 LT and a sufficiently weak photochemical sink and that the TVCDs from OMI 87 were lower than SCIAMACHY in summer due to a strong photochemical sink of NO_x.

All these above researches, however, exploited only NO₂ surface or satellite VCD measurements. <u>Due to the</u>
 <u>availability of ground-based NO₂ VCD observations, some recent studies tried to investigate the diurnal</u>
 <u>relationships between NO₂ surface concentrations and NO₂ VCDs (Kollonige et al., 2018; Thompson et al.,</u>
 <u>2019</u>). For example, Zhao et al. (2019) converted Pandora direct-sun and zenith-sky NO₂ VCDs to NO₂ surface

92	concentrations using concentration-to-partial-column ratios and found that the derived concentrations well
93	captured the observed NO ₂ surface diurnal and seasonal variations. Knepp et al. (2015) related the daytime
94	variations of NO ₂ TVCD measurements by ground-based Pandora instruments to the variations of coincident NO ₂
95	surface concentrations using a planetary boundary layer height (PBLH) factor over the periods July 2011 -
96	October 2011 at the NASA Langley Research Center in Hampton, Virginia and July 2011 at Padonia and
97	Edgewood sites in Maryland for the DISCOVER-AQ experiment, showing the importance of boundary-layer
98	vertical mixing on NO ₂ vertical distributions and the ability of NO ₂ VCD measurements to infer hourly
99	boundary-layer NO ₂ variations. DISCOVER-AQ, the Deriving Information on Surface conditions from Column
100	and Vertically Resolved Observations Relevant to Air Quality experiment (https://discover-aq.larc.nasa.gov/, last
101	access: April 6, 2019), was designed to better understand the relationship between boundary-layer pollutants and
102	satellite observations (Flynn et al., 2014; Reed et al., 2015). Figure §1 shows the sampling locations of the
103	summer DISCOVER-AQ 2011 campaign in the Baltimore-Washington metropolitan region. In this campaign, the
104	NASA P-3B aircraft flew spirals over six air quality monitoring sites (Aldino - rural/suburban, Edgewood -
105	coastal/urban, Beltsville - suburban, Essex - coastal/urban, Fairhill - rural, and Padonia - suburban) (Table S1)
106	and the Chesapeake Bay (Cheng et al., 2017; Lamsal et al., 2014), and measured 2454 NO ₂ profiles in 14 flight
107	days in July (Zhang et al., 2016). During the same period, the NASA UC-12 aircraft flew across the Baltimore-
108	Washington region at an altitude about 8 km above sea level (ASL), using the Airborne Compact Atmospheric
109	Mapper (ACAM) to map the distributions of NO ₂ VCDs below the aircraft (Lamsal et al., 2017). Furthermore,
110	ground-based instruments were deployed to measure NO2 surface concentrations, NO2 VCDs, and other physical
111	properties of the atmosphere (Anderson et al., 2014; Reed et al., 2015; Sawamura et al., 2014). Satellite OMI and
112	GOME-2A (Global Ozone Monitoring Experiment – 2A) instruments provided NO ₂ TVCD measurements over
113	the campaign region at 13:30 and 9:30 LT, respectively. These concurrent measurements of NO ₂ VCDs, surface
114	NO ₂ , and vertically resolved distributions of NO ₂ during the DISCOVER-AQ 2011 campaign, therefore, provide

a comprehensive dataset to evaluate NO_2 diurnal and spatial variabilities and processes affecting NO_2

116 concentrations.

117	Section 2 describes the measurement datasets in detail. The Regional chEmistry and trAnsport Model
118	(REAM), also described in section 2, is applied to simulate the NO ₂ observations during the DISCOVER-AQ
119	campaign in July 2011. The evaluations of the simulated diurnal cycles of surface NO ₂ concentrations, NO ₂
120	vertical profiles, and NO ₂ TVCDs are discussed in section 3 through comparisons with observations. In section 3,
121	we also investigate the differences between NO ₂ diurnal cycles on weekdays and weekends and their implications
122	for NO _x emission characteristics. To corroborate our evaluation of NO _x emissions based on NO ₂ diurnal cycles,
123	we further compare observed NO _y (reactive nitrogen compounds) concentrations with REAM simulation results
124	in section 3. Moreover, we assess the resolution dependence of REAM simulation results in light of the
125	observations and discuss the potential <u>distribution</u> biases of NO_x emissions at high resolution by comparing the
126	<u>36- and 4-km REAM simulation results with OMI, GOME-2A, and high-resolution ACAM NO₂ VCDs. Finally,</u>
 127	we summarize the study in section 4.

128 **2 Datasets and model description**

129 2.1 REAM

REAM has been widely applied in many studies (Cheng et al., 2017; Choi et al., 2008; Li et al., 2019; Zhang et al., 2018; Zhang et al., 2016; Zhao et al., 2009). The model has a horizontal resolution of 36 km and 30 vertical layers in the troposphere. Meteorology fields are from a Weather Research and Forecasting (WRF, version 3.6) model simulation with a horizontal resolution of 36 km. We summarize the physics parameterization schemes of the WRF simulation in Table S2. The WRF simulation is initialized and constrained by the NCEP coupled forecast system model version 2 (CFSv2) products (http://rda.ucar.edu/datasets/ds094.0/, last access: March 10,

136	2015) (Saha et al., 2011). The chemistry mechanism in REAM is based on GEOS-Chem v11.01 with updated
137	aerosol uptake of isoprene nitrates (Fisher et al., 2016) and revised treatment of wet scavenging processes (Luo et
138	al., 2019). A $2^{\circ} \times 2.5^{\circ}$ GEOS-Chem simulation provides the chemical boundary and initial conditions.

139	Biogenic VOC emissions in REAM are from MEGAN v2.10 (Guenther et al., 2012). Anthropogenic
140	emissions on weekdays are from the National Emission Inventory 2011 (NEI2011) (EPA, 2014) from the Pacific
141	Northwest National Laboratory (PNNL), which has an initial resolution of 4 km and is re-gridded to REAM 36-
142	km grid cells (Figure <u>S</u> 2). Weekday emission diurnal profiles are from NEI2011. The weekday to weekend
 143	emission ratios and weekend emission diurnal profiles are based on previous studies (Beirle et al., 2003; Boersma
144	et al., 2009; Choi et al., 2012; de Foy, 2018; DenBleyker et al., 2012; Herman et al., 2009; Judd et al., 2018;
145	Kaynak et al., 2009; Kim et al., 2016). These studies suggested that weekend NO _x emissions were 20% - 50%
146	lower than weekday emissions, and the weekend NO _x emission diurnal cycles were different from weekdays;
147	therefore, we specify a weekend to weekday NO_x emission ratio of 2/3 in this study. The resulting diurnal
148	variations of weekday and weekend NO _x emissions over the DISCOVER-AQ 2011 region are shown in Figure
149	1 <u>3</u> . The diurnal emission variation is lower on weekends than on weekdays.
 150	To understand the effects of model resolutions on the temporospatial distributions of NO ₂ , we also conduct a

REAM simulation with a horizontal resolution of 4 km during the DISCOVER-AQ campaign. A 36-km REAM simulation (discussed in section $3.\pm2$) provides the chemical initial and hourly boundary conditions. Meteorology fields are from a nested WRF simulation (36 km, 12 km, 4 km) with cumulus parameterization turned off in the 4-km domain (Table S2). Figure \$1 shows a comparison of the 4-km and 36-km REAM grid cells with DISCOVER-AQ observations, and Figure \$2 shows a comparison of NO_x emission distributions between the 4km and 36-km REAM simulations. The comparison of NO_x emission diurnal variations over the DISCOVER-AQ 2011 region between the 4-km and 36-km REAM is shown in Figure ±3 .

158	We evaluate the performances of the 36 km and nested 4 km WRF simulations by comparing temperature
159	and wind from the P-3B spirals (Figure S1) and precipitation from the NCEP (National Centers for
160	Environmental Prediction) Stage IV precipitation dataset with those coincident WRF simulation results in July
161	2011. Generally, P-3B spirals range from ~400 m to ~3.63 km in height above the ground level (AGL). As shown
162	in Figure S3, both the 36-km and nested 4-km WRF simulations predict temperature well with $R^2 = 0.94$ and $R^2 =$
163	0.98, respectively. Both WRF simulations show good agreement with P-3B measurements in U-wind (36-km: R ²
164	= 0.62; 4 km: R ² = 0.71), V wind (36 km: R ² = 0.75; 4 km: R ² = 0.74), wind speed (36 km: R ² = 0.52; 4 km: R ²)
165	= 0.64), and wind direction (36 km: R^2 = 0.40; 4 km: R^2 = 0.50) (Figures S3 and S4). The evaluations above
166	suggest that WRF simulated wind fields are good and comparable at 4 km and 36 km resolutions; they are not the
167	reasons for the differences of the 4-km and 36-km simulations of trace gases by REAM, which is driven by WRF
168	meteorological fields.
169	The NCEP Stage IV precipitation dataset provides hourly precipitation across the contiguous United States
170	(CONUS) with a resolution of ~4 km based on the merging of rain gauge data and radar observations (Lin and
171	Mitchell, 2005; Nelson et al., 2016). The Stage IV dataset is useful for evaluating model simulations, satellite
172	precipitation estimates, and radar precipitation estimates (Davis et al., 2006; Gourley et al., 2011; Kalinga and
173	Gan, 2010; Lopez, 2011; Yuan et al., 2008). We obtain the Stage IV precipitation data in July 2011 from the
174	NCAR/UCAR Research Data Archive (https://rda.ucar.edu/datasets/ds507.5/). As shown in Figures S5 and S6,
175	generally, both the 36-km and nested 4-km WRF simulations predict much less precipitation (in precipitation
176	amount and duration) compared to Stage-IV in July 2011 around the DISCOVER-AQ campaign region,
177	especially for the nested 4-km WRF simulation. We find that large scale precipitation amounts are much less
178	compared to convective precipitation in most regions in the 36-km WRF simulation (Figure S7) during the
178 179	compared to convective precipitation in most regions in the 36-km WRF simulation (Figure S7) during the simulation period, which is contradictory to Li et al. (2020) showing non-convective precipitation accounting for

separated because convection is explicitly resolved. The model low bias is large (Figure S6). The underestimation
 of precipitation in our WRF simulations may lead to high biases of soluble species in REAM, such as HNO₃, due
 to underestimated wet scavenging.

184 2.2 NO₂ TVCD measurements by OMI and GOME-2A

The OMI instrument onboard the sun-synchronous NASA EOS Aura satellite with an equator-crossing time of around 13:30 LT was developed by the Finnish Meteorological Institute and the Netherlands Agency for Aerospace Programs to measure solar backscattering radiation in the visible and ultraviolet bands (Levelt et al., 2006; Russell et al., 2012). The radiance measurements are used to derive trace gas concentrations in the atmosphere, such as O₃, NO₂, HCHO, and SO₂ (Levelt et al., 2006). OMI has a nadir resolution of 13 km × 24 km

- and provides daily global coverage (Levelt et al., 2006).
- 191 Two widely-used archives of OMI NO₂ VCD products are available, NASA OMNO2 (v4.0)
- 192 (https://disc.gsfc.nasa.gov/datasets/OMNO2_003/summary, last access: September 26, 2020) and KNMI
- DOMINO (v2.0) (<u>http://www.temis.nl/airpollution/no2.html, last access: January 14, 2015</u>). Although both use
- 194 Differential Optical Absorption Spectroscopy (DOAS) algorithms to derive NO₂ slant column densities, they
- have differences in spectral fitting, stratospheric and tropospheric NO₂ slant column density (SCD) separation, a
- 196 priori NO₂ vertical profiles, and air mass factor (AMF) calculation, etc. (Boersma et al., 2011; Bucsela et al.,
- 197 2013; Chance, 2002; Krotkov et al., 2017; Lamsal et al., 2020; Marchenko et al., 2015; Oetjen et al., 2013; van
- der A et al., 2010; Van Geffen et al., 2015). Both OMNO2 and DOMINO have been extensively evaluated with
- 199 field measurements and models (Boersma et al., 2009; Boersma et al., 2011; Choi et al., 2020; Hains et al., 2010;
- 200 Huijnen et al., 2010; Ionov et al., 2008; Irie et al., 2008; Lamsal et al., 2014; Lamsal et al., 2020; Oetjen et al.,
- 201 2013). The estimated uncertainty of DOMINO TVCD product <u>includes an absolute component of is</u> 1.0×10^{15}
- molecules cm^{-2} and a relative AMF component of + 25% (Boersma et al., 2011), while the uncertainty of OMNO2

203	TVCD product ranges from ~30% under clear-sky conditions to ~60% under cloudy conditions (Lamsal et al.,
204	2014; Oetjen et al., 2013; Tong et al., 2015). In order to reduce uncertainties in this study, we only use TVCD
205	data with effective cloud fractions < 0.2 , solar zenith angle (SZA) $< 80^{\circ}$, and albedo ≤ 0.3 . Both positive and
206	negative TVCDs are considered in the calculation. The data affected by row anomaly are excluded (Boersma et
l 207	al., 2018; Zhang et al., 2018).

208	For AMF calculation, DOMINO used daily TM4 model results with a resolution of $3^{\circ} \times 2^{\circ}$ as a priori NO ₂
209	vertical profiles (Boersma et al., 2007; Boersma et al., 2011), while OMNO2 v4.0 used monthly mean values
210	from the Global Modeling Initiative (GMI) model with a resolution of $1^{\circ} \times 1.25^{\circ}$. The relatively coarse horizontal
211	resolution of the a priori NO ₂ profiles in the retrievals can introduce uncertainties in the spatial and temporal
212	characteristics of NO ₂ TVCDs at satellite pixel scales. For comparison purposes, we also use 36-km REAM
213	simulation results as the a priori NO ₂ profiles to compute the AMFs and NO ₂ TVCDs with the DOMINO
214	algorithm. The 36-km REAM NO2 data are first regridded to OMI pixels to calculate the corresponding
215	tropospheric AMFs, which are then applied to compute OMI NO ₂ TVCDs by dividing the tropospheric SCDs
216	from the DOMINO product by our updated AMFs.
217	The GOME-2 instrument onboard the polar-orbiting MetOp-A satellite (now referred to as GOME-2A) is an
 217 218	The GOME-2 instrument onboard the polar-orbiting MetOp-A satellite (now referred to as GOME-2A) is an improved version of GOME-1 launched in 1995 and has an overpass time of 9:30 LT and a spatial resolution of
218	improved version of GOME-1 launched in 1995 and has an overpass time of 9:30 LT and a spatial resolution of
218 219	improved version of GOME-1 launched in 1995 and has an overpass time of 9:30 LT and a spatial resolution of $80 \times 40 \text{ km}^2$ (Munro et al., 2006; Peters et al., 2012). GOME-2A measures backscattered solar radiation in the
218 219 220	improved version of GOME-1 launched in 1995 and has an overpass time of 9:30 LT and a spatial resolution of $80 \times 40 \text{ km}^2$ (Munro et al., 2006; Peters et al., 2012). GOME-2A measures backscattered solar radiation in the range from 240 nm to 790 nm, which is used for VCD retrievals of trace gases, such as O ₃ , NO ₂ , BrO, and SO ₂
218 219 220 221	improved version of GOME-1 launched in 1995 and has an overpass time of 9:30 LT and a spatial resolution of 80×40 km ² (Munro et al., 2006; Peters et al., 2012). GOME-2A measures backscattered solar radiation in the range from 240 nm to 790 nm, which is used for VCD retrievals of trace gases, such as O ₃ , NO ₂ , BrO, and SO ₂ (Munro et al., 2006). We use the KNMI TM4NO2A v2.3 GOME-2A NO ₂ VCD product archived on

225 <u>same criteria to filter the NO₂ TVCD dataalso and recalculate the <u>tropospheric</u> AMF values and GOME-2A</u>

226 TVCDs using the daily 36-km REAM NO₂ profiles (9:00 LT - 10:00 LT).

227 2.3 Pandora ground-based NO₂ VCD measurements

228	Pandora is a small direct sun spectrometer, which measures sun and sky radiance from 270 to 530 nm with a
229	0.5 nm resolution and a 1.6° field of view (FOV) for the retrieval of the total VCDs of NO ₂ with a precision of
230	about $\frac{2.75.4}{2.75.4} \times 10^{14}$ molecules/cm ² (2.7 × 10 ¹⁴ molecules/cm ² for NO ₂ SCD) and a nominal accuracy of 2.7×10^{15}
231	molecules cm ⁻² under clear-sky conditions (Herman et al., 2009; Lamsal et al., 2014; Zhao et al., 2020). There
232	were 12 Pandora sites operating in the DISCOVER-AQ campaign (Figure <u>\$</u> 1). Six of them are the same as the P-
l 233	3B aircraft spiral locations (Aldino, Edgewood, Beltsville, Essex, Fairhill, and Padonia) (Table S1 and Figure
234	S1). The other six sites are Naval Academy (Annapolis Maryland) (USNA – ocean), University of Maryland
l 235	College Park (UMCP – urban), University of Maryland Baltimore County (UMBC – urban), Smithsonian
236	Environmental Research Center (SERC - rural/coastal), Oldtown in Baltimore (Oldtown - urban), and Goddard
237	Space Flight Center (GSFC – urban/suburban) (Table S1 and Figure <u>S</u> 1). In this study, we exclude the USNA site
238	as its measurements were conducted on a ship ("Pandora(w)" in Figure S1), and there were no other surface
l 239	observations in the corresponding REAM grid cell. Including the data from the USNA site has a negligible effect
240	on the comparisons of observed and simulated NO2 TVCDs. In our analysis, we ignore Pandora measurements
241	with solar zenith angles (SZA) > 80° (Figure S <u>1</u> 8) and exclude the data when fewer than three valid
l 242	measurements are available within an hour to reduce the uncertainties of the hourly averages due to the
243	significant variations of Pandora observations (Figure S ²⁹).
244	Since Pandora measures total NO ₂ VCDs, we need to subtract stratosphere NO ₂ VCDs from the total VCDs

to compute TVCDs. As shown in Figure S $\underline{310}$, stratosphere NO₂ VCDs show a clear diurnal cycle with an increase during daytime due in part to the photolysis of reactive nitrogen reservoirs such as N₂O₅ and HNO₃ 247 (Brohede et al., 2007; Dirksen et al., 2011; Peters et al., 2012; Sen et al., 1998; Spinei et al., 2014), which is 248 consistent with the significant increase of stratospheric NO₂ VCDs from GOME-2A to OMI. In this study, we use 249 the GMI model simulated stratospheric NO₂ VCDs in Figure S<u>3</u>+0 to calculate the Pandora NO₂ TVCDs. The 250 small discrepancies between the GMI stratospheric NO₂ VCDs and satellite products do not change the pattern of 251 Pandora NO₂ TVCD diurnal variations or affect the conclusions in this study.

252 2.4 ACAM NO₂ VCD measurements

The ACAM instrument onboard the UC-12 aircraft consists of two thermally spectrometers in the ultraviolet/visible/near-infrared range. The spectrometer in the ultraviolet/visible band (304 nm – 520 nm) with a resolution of 0.8 nm and a sampling of 0.105 nm can be used to detect NO₂ in the atmosphere. The native ground resolution of UC-12 ACAM NO₂ measurements is 0.5 km \times 0.75 km at a flight altitude of about 8 km ASL and a nominal ground speed of 100 m s⁻¹ during the DISCOVER-AQ 2011 campaign (Lamsal et al., 2017), thus providing high-resolution NO₂ VCDs below the aircraft.

259 In this study, we mainly use the ACAM NO₂ VCD product described by Lamsal et al. (2017), which applied 260 a pair-average co-adding scheme to produce NO₂ VCDs at a ground resolution of about 1.5 km (cross-track) \times 261 1.1 km (along-track) to reduce noise impacts. In their retrieval of ACAM NO₂ VCDs, they first used the DOAS 262 fitting method to generate differential NO₂ SCDs relative to the SCDs at an unpolluted reference location. Then 263 they computed above/below-aircraft AMFs at both sampling and reference locations based on the vector 264 linearized discrete ordinate radiative transfer code (VLIDORT) (Spurr, 2008). In the computation of AMFs, the a 265 priori NO₂ vertical profiles were from a combination of a high-resolution (4-km) CMAQ (the Community 266 Multiscale Air Quality Modeling System) model outputs in the boundary layer and a GMI simulation $(2^{\circ} \times 2.5^{\circ})$ 267 results elsewhere in the atmosphere. Finally, the below-aircraft NO_2 VCDs at the sampling locations were 268 generated by dividing below-aircraft NO₂ SCDs at the sampling locations by the corresponding below-aircraft

269	AMFs. The below-aircraft NO ₂ SCDs were the differences between the total and above-aircraft NO ₂ SCDs. The
270	total NO ₂ SCDs were the sum of DOAS fitting generated differential NO ₂ SCDs and NO ₂ SCDs at the reference
271	location, and the above-aircraft NO ₂ SCDs were derived based on above-aircraft AMFs, GMI NO ₂ profiles, and
272	OMNO2 stratospheric NO2 VCDs (Lamsal et al., 2017). The ACAM NO2 VCD product had been evaluated via
273	comparisons with other independent observations during the DISCOVER-AQ 2011 campaign, such as P-3B
274	aircraft, Pandora, and OMNO2, and the uncertainty of individual below-aircraft NO2 VCD is about 30% (Lamsal
275	et al., 2017). To keep the consistency of ACAM NO ₂ VCDs, we exclude NO ₂ VCDs measured at altitudes < 8 km
276	ASL, which accounts for about 6.8% of the total available ACAM NO ₂ VCD data. We regrid the 1.5 km \times 1.1
277	km ACAM NO ₂ VCDs to the 4-km REAM grid cells (Figure $\frac{1}{5}$ 1), which are then used to evaluate the distribution
278	of NO ₂ VCDs in the 4-km REAM simulation. As a supplement in section $3.\frac{76}{6}$, we also assess the 4-km REAM
l 279	simulation by using the UC-12 ACAM NO ₂ VCDs produced by the Smithsonian Astrophysical Observatory
280	(SAO) algorithms, archived on https://www-air.larc.nasa.gov/cgi-bin/ArcView/discover-aq.dc-
281	2011?UC12=1#LIU.XIONG/ (last access: December 31, 2019) (Liu et al., 2015a; Liu et al., 2015b). This product
 282	is an early version of the SAO algorithm used to produce the Geostationary Trace gas and Aerosol Sensor
283	Optimization (GeoTASO) and the GEOstationary Coastal and Air Pollution Events (GEO-CAPE) Airborne
284	Simulator (GCAS) airborne observations in later airborne campaigns (Nowlan et al., 2016; Nowlan et al., 2018).

 $285 \qquad 2.5 \ Surface \ NO_2 \ and \ O_3 \ measurements$

286 The measurement of NO_x is based on the chemiluminescence of electronically excited NO_2^* , produced from

the reaction of NO with O_3 , and the strength of the chemiluminescence from the decay of NO_2^* to NO_2 is

- proportional to the number of NO molecules present (Reed et al., 2016). NO₂ concentrations can be measured
- with this method by converting NO_2 to NO first through catalytic reactions (typically on the surface of heated
- 290 molybdenum oxide (MoO_x) substrate) or photolytic processes (Lamsal et al., 2015; Reed et al., 2016). However,
- for the catalytic method, reactive nitrogen compounds other than NO_x (NO_z), such as HNO₃, peroxyacetyl nitrate

292	(PAN), and other organic nitrates, can also be reduced to NO on the heated surface, thus causing an
293	overestimation of NO ₂ . The magnitude of the overestimation depends on the concentrations and the reduction
294	efficiencies of interference species, both of which are uncertain. The photolytic approach, which employs
295	broadband photolysis of ambient NO ₂ , offers more accurate NO ₂ measurements (Lamsal et al., 2015).

296	There were 11 NO _x monitoring sites operating in the DISCOVER-AQ region during the campaign (Figure
297	S1), including those from the EPA Air Quality System (AQS) monitoring network and those deployed for the
298	DISCOVER-AQ campaign. Nine of them measured NO ₂ concentrations by a catalytic converter. The other two
299	sites (Edgewood and Padonia) had NO2 measurements from both catalytic and photolytic methods. Different
300	stationary catalytic instruments were used during the campaign: Thermo Electron 42C-Y NO _y analyzer, Thermo
301	Model 42C NO _x analyzer, Thermo Model 42I-Y NO _y analyzer, and Ecotech Model 9843/9841 T-NO _y analyzers.
302	In addition, a mobile platform — NATIVE (Nittany Atmospheric Trailer and Integrated Validation Experiment)
303	with a Thermo Electron 42C-Y NO _y analyzer installed, was also deployed in the Edgewood site. The photolytic
304	measurements of NO2 in Edgewood and Padonia were from Teledyne API model 200eup photolytic NOx
305	analyzers. We scale catalytic NO_2 measurements using the diurnal ratios of NO_2 photolytic measurements to NO_2
306	from the corresponding catalytic analyzers (Figure $\frac{24}{2}$). Figure $\frac{24}{2}$ shows the lowest photolytic/catalytic ratios in
307	the afternoon, which reflects the production of nitrates and other reactive nitrogen compounds from NO _x in the
308	daytime. When photolytic measurements are available, we only use the photolytic observations in this study;
309	otherwise, we use the scaled catalytic measurements.

Nineteen surface O₃ monitoring sites were operating in the DISCOVER-AQ region during the campaign
(Figure \$1). They measured O₃ concentrations by using a Federal Equivalent Method (FEM) based on the UV
absorption of O₃ (<u>https://www.arb.ca.gov/aaqm/qa/qa-manual/vol4/chapter6o3.pdf, last access: April 6, 2019</u>)
with an uncertainty of 5 ppb.

314 2.6 Aircraft measurements of NO₂ vertical profiles

In this study, we mainly use the NO₂ concentrations measured by the National Center for Atmospheric Research (NCAR) 4-channel chemiluminescence instrument (P-CL) onboard the P-3B aircraft for the evaluation of REAM simulated NO₂ vertical profiles. The instrument has a NO₂ measurement uncertainty of 10% – 15% and a 1-second, 1-sigma detection limit of 30 pptv.

NO₂ measurements from aircraft spirals provide us with NO₂ vertical profiles. Figure S1 shows the locations of the aircraft spirals during the DISCOVER-AQ campaign, except for the Chesapeake Bay spirals over the ocean. There were only six spirals available over the Chesapeake Bay, which have ignorable impacts on the following analyses. Therefore, we do not use them in this study. The rest 23<u>9</u>8 spirals in the daytime for July 2011 are used to compute the average profiles of NO₂ for the six inland sites (Figure S1).

The aircraft measurements were generally sampled from about a height of <u>3400 m AGL</u> in the boundary layer to 3.63 km AGL in the free troposphere. We bin these measurements to REAM vertical levels. In order to make up the missing observations between the surface and <u>3400 m</u>, we apply quadratic polynomial regressions by using aircraft data below 1 km and coincident NO₂ surface measurements.

In addition to using NO₂ concentrations from the NCAR 4-channel instrument to evaluate REAM simulated NO₂ vertical profiles, we also use P-3B NO, NO₂, and NO_y concentrations measured by the NCAR 4-channel instrument and NO₂, total peroxyacyl nitrates (Σ PNs), total alkyl nitrates (Σ ANs) (include alkyl nitrates and hydroxyalkyl nitrates), and HNO₃ concentrations measured by the thermal dissociation-laser induced fluorescence (TD-LIF) technique (Day et al., 2002; Thornton et al., 2000; Wooldridge et al., 2010) to evaluate the concentrations of NO_y from REAM (Table 1). All these P-3B measurements are vertically binned to REAM grid cells for comparisons with REAM results. In addition, below the P-3B spirals, four NO_y observation sites at

15

- 335 Padonia, Edgewood, Beltsville, and Aldino were operating to provide continuous hourly NO_y surface
- 336 concentrations during the campaign, which we also use to evaluate REAM simulated NO_y surface concentrations
- in this study. We summarize the information of available observations at the 11 inland Pandora sites in Table S1.

338 3 Results and discussion

- 339 <u>3.1 Evaluation of WRF simulated meteorological fields</u>
- We evaluate the performances of the 36-km and nested 4-km WRF simulations using temperature, potential
- temperature, relative humidity (RH), and wind measurements from the P-3B spirals (Figure 1) and precipitation
- data from the NCEP (National Centers for Environmental Prediction) Stage IV precipitation dataset. Generally,
- 843 P-3B spirals range from ~300 m to ~3.63 km in height above the ground level (AGL). As shown in Figure S4,
- both the 36-km and nested 4-km WRF simulations simulate temperature well with $R^2 = 0.98$. Both WRF
- simulations show good agreement with P-3B measurements in U-wind (36-km: $R^2 = 0.77$; 4-km: $R^2 = 0.76$), V-
- 846 wind (36-km: $R^2 = 0.79$; 4-km: $R^2 = 0.78$), wind speed (36-km: $R^2 = 0.67$; 4-km: $R^2 = 0.67$), and wind direction
- $(36-km: R^2 = 0.46; 4-km: R^2 = 0.52)$ (Figures S4 and S5). We further compare the temporal evolutions of vertical
- profiles for temperature, potential temperature, RH, U-wind, and V-wind below 3 km from the P-3B observations
- 349 with those from the 36-km and nested 4-km WRF simulations in Figure S6. Both WRF simulations well capture
- the temporospatial variations of P-3B observed vertical profiles except that RH below 1.5 km is significantly
- <u>underestimated during 9:00 17:00 LT in both WRF simulations. The evaluations above suggest that WRF</u>
- simulated wind fields are good and comparable at 4-km and 36-km resolutions, but potential dry biases exist in
- 353 <u>both WRF simulations.</u>
- The NCEP Stage IV precipitation dataset provides hourly precipitation across the contiguous United States
 (CONUS) with a resolution of ~4 km based on the merging of rain gauge data and radar observations (Lin and

356	Mitchell, 2005; Nelson et al., 2016). The Stage IV dataset is useful for evaluating model simulations, satellite
357	precipitation estimates, and radar precipitation estimates (Davis et al., 2006; Gourley et al., 2011; Kalinga and
358	Gan, 2010; Lopez, 2011; Yuan et al., 2008). We obtain the Stage IV precipitation data for July 2011 from the
359	NCAR/UCAR Research Data Archive (https://rda.ucar.edu/datasets/ds507.5/, last access: December 28, 2019).
360	As shown in Figures S7 and S8, generally, both the 36-km and nested 4-km WRF simulations generally predict
361	much less precipitation (in precipitation amount and duration) compared to the Stage-IV data in July 2011 around
362	the DISCOVER-AQ campaign region, especially for the nested 4-km WRF simulation, consistent with the
363	aforementioned underestimated RH and dry bias in WRF simulations. The precipitation biases in the WRF model
364	will affect REAM simulations of trace gases, leading to high biases of soluble species due to underestimated wet
365	scavenging. Clouds interfere with satellite observations. Therefore, the precipitation bias does not affect model
366	evaluations with satellite measurements of NO2. Aircraft measurements were also taken in non-precipitating days.
367 368	3.21 Effect of boundary layer vertical mixing on the diurnal variations of surface NO ₂ concentrations 3.21.1 36-km model simulation in comparison to the surface observations
369	Figures 53 and 53 b show the observed and 36-km REAM simulated diurnal cycles of surface NO ₂ and O ₃
370	concentrations on weekdays in July 2011 in the DISCOVER-AQ region. REAM with WRF-YSU simulated
 371	vertical diffusion coefficient (k_{zz}) values significantly overestimates NO ₂ concentrations and underestimates O ₃
372	concentrations at night, although it captures the patterns of the diurnal cycles of surface NO2 and O3: an O3 peak
373	and a NO ₂ minimum around noontime. Here, YSU denotes the Yonsei University planetary boundary layer (PBL)
374	scheme (Shin and Hong, 2011) used by our WRF simulations (Table S2). At night, the reaction of $O_3 + NO \rightarrow O_2$
 375	+ NO ₂ produces NO ₂ but removes O ₃ . Since most NO _x emissions are in the form of NO, the model biases of low
376	O ₃ and high NO ₂ occur at the same time. Since there are no significant chemical sources of O ₃ at night, mixing of
377	O ₃ rich air above the surface is the main source of O ₃ supply near the surface. Therefore, the nighttime model

biases with WRF<u>-YSU</u> simulated k_{zz} data in Figure 53 indicate that vertical mixing may be underestimated at night.

380	During the DISCOVER-AQ campaign, WRF simulated vertical wind velocities are very low at night and
381	have little impact on vertical mixing (Figure S ₂₁₁ a). The nighttime vertical mixing is mainly attributed to
382	turbulent mixing. In the YSU scheme the Yonsei University (YSU) planetary boundary layer (PBL) scheme (Shin
383	and Hong, 2011) used by our WRF simulations (Table S2), boundary layer k_{zz} is correlated to PBLH. However,
l 384	Breuer et al. (2014) and Hu et al. (2012) found that the YSU scheme underestimated nighttime PBLHs in WRF,
385	which is consistent with Figure 64 showing that WRF-YSU k_{zz} -determined PBLHs are significantly lower than
386	lidar observations in the late afternoon and at night at the UMBC site during the DISCOVER-AQ campaign.
387	Here, the k_{zz} -determined PBLH refers to the mixing height derived by comparing k_{zz} to its background values
388	(Hong et al., 2006) but not the PBLH outputs from WRF. The lidar mixing depth data were derived from the
l 389	Elastic Lidar Facility (ELF) attenuated backscatter signals by using the covariance wavelet transform method and
390	had been validated against radiosonde measurements, Radar wind profiler observations, and Sigma Space mini-
391	micropulse lidar data (Compton et al., 2013). To improve nighttime PBLHs and vertical mixing in REAM, we
392	increase REAM- k_{zz} below 500 m during 18:00 – 5:00 LT to 5 m s ⁻² if the WRF-YSU computed $k_{zz} < 5$ m s ⁻² ,
393	which significantly increases the k_{zz} -determined PBLHs at night (Figure 64), leading to the decreases of simulated
394	surface NO ₂ and the increases of surface O ₃ concentrations at night (Figure <u>5</u> 3). The assigned value of 5 m s ⁻² is
l 395	arbitrary. Changing this value to 2 or 10 m s ⁻² can also alleviate the biases of model simulated nighttime surface
896	NO_2 and O_3 concentrations (Figure S10). An alternative solution to correct the model nighttime simulation biases
l 397	is to reduce NO_x emissions by 50-67%, but we cannot find good reasons to justify this level of NO_x emission
398	reduction only at night.

899 The updated REAM simulation of surface NO₂ diurnal pattern in Figure 53a is in good agreement with 400 previous studies (Anderson et al., 2014; David and Nair, 2011; Gaur et al., 2014; Reddy et al., 2012; Zhao et al., 401 2019). Daytime surface NO₂ concentrations are much lower compared to nighttime, and NO₂ concentrations 402 reach a minimum around noontime. As shown in Figure S112, under the influence of vertical turbulent mixing, 403 the surface-layer NO_x emission diurnal pattern is similar to the surface NO_y diurnal cycle in Figure 53a. 404 emphasizing the importance of turbulent mixing on modulating surface NO₂ diurnal variations. The highest 405 boundary layer (Figure 64) due to solar radiation leads to the lowest surface-layer NO_x emissions (Figure S112) 406 and, therefore, the smallest surface NO_2 concentrations occur around noontime (Figure 53a). Transport, which is 407 mainly attributed to advection and turbulent mixing, is another critical factor affecting surface NO₂ diurnal 408 variations (Figure S112). The magnitudes of transport fluxes (Figure S112) are proportional to horizontal and 409 vertical gradients of NO_x concentrations and are therefore generally positively correlated to surface NO_2 410 concentrations. However, some exceptions exist, reflecting different strengths of advection (U, V, and W) and 411 turbulent mixing (k_{77}) at different times. For example, in the early morning, NO₂ surface concentrations peak at 412 5:00 - 6:00 LT (Figure 53a), while transport fluxes peak at 7:00 - 8:00 LT (Figure S112). The delay of the peak 413 is mainly due to lower turbulent mixing at 5:00 - 6:00 LT than other daytime hours in the model (Figure 64). 414 Chemistry also contributes to surface NO₂ diurnal variations mainly through photochemical sinks in the daytime 415 and N_2O_5 hydrolysis at nighttime. Chemistry fluxes in Figure S112 are not only correlated to the strength of 416 photochemical reactions and N_2O_5 hydrolysis (chemistry fluxes per unit NO_x) but are also proportional to NO_x 417 surface concentrations. Therefore, chemistry fluxes in Figure S112 cannot directly reflect the impact of solar 418 radiation on photochemical reactions. It can, however, still be identified by comparing afternoon chemistry 419 contributions: from 13:00 to 15:00 LT, surface-layer NO_x emissions and NO₂ concentrations are increasing 420 (Figures S112 and 53a); however, chemistry losses are decreasing as a result of the reduction of photochemical 421 sinks with weakening solar radiation. The contributions of vertical mixing and photochemical sinks to NO_2

422 concentrations can be further corroborated by daytime variations of NO₂ vertical profiles (Figure 6) and TVCDs 423 (Figure S13) discussed in sections 3.32 and 3.43.

Figure 53 c shows the diurnal variation on weekends is also simulated well in the improved 36-km model. The diurnal variation of surface NO₂ concentrations (REAM: 1.5 - 10.24 ppb; observations: 2.1 - 9.8 ppb) is lower than on weekdays (REAM: 2.45 - 12.25 ppb; observations: 3.3 - 14.5 ppb), reflecting lower magnitude and variation of NO_x emissions on weekends (Figure 34). Figure 53 d also shows an improved simulation of surface O₃ concentrations at nighttime due to the improved PBLH simulation (Figure 64).

429 3.<u>2</u>4.2 4-km model simulation in comparison to the surface observations

430 The results of 4-km REAM simulations with original WRF-YSU k_{zz} (not shown) are very similar to Figure 431 53 since WRF simulated nocturnal vertical mixing is insensitive to the model horizontal resolution. Applying the 432 modified nocturnal mixing in the previous section also greatly reduced the nighttime NO₂ overestimate and O₃ 433 underestimate in the 4-km REAM simulations. All the following analyses are based on REAM simulations with 434 improved nocturnal mixing. Figure 75 shows that mean surface NO_2 concentrations simulated in the 4-km model 435 are higher than the 36-km results over Padonia, Oldtown, Essex, Edgewood, Beltsville, and Aldino (Table S1). 436 leading to generally higher biases compared to the observations at night in the daytime. A major cause is that the 437 observation sites are located in regions of high NO_x emissions (Figures S1 and S2). At a higher resolution of 4 438 km, the high emissions around the surface sites are apparent compared to rural regions. At the coarser 36-km 439 resolution, spatial averaging greatly reduces the emissions around the surface sites. On average, NO_x emissions 440 (molecules km⁻² s⁻¹) around the six surface NO₂ observations sites are 67% higher in the 4-km than the 36-km 441 REAM simulations (Table S1). The resolution dependence of model results will be further discussed in the model 442 evaluations using the other in situ and remote sensing measurements.

443 3.32 Diurnal variations of NO₂ vertical profiles

444 Figures 86a and 86c show the temporal variations of P-3B observed and 36-km REAM simulated NO₂ 445 vertical profiles in the daytime on weekdays during the DISCOVER-AO campaign. 36-km REAM reproduces 446 well the observed characteristics of NO₂ vertical profiles in the daytime ($R^2 = 0.8991$), which are strongly 447 affected by vertical mixing and photochemistry (Zhang et al., 2016). When vertical mixing is weak in the early 448 morning (6:00 – 8:00 LT), NO₂, released mainly from surface NO_x sources, is concentrated in the surface layer, 449 and the vertical gradient is large. As vertical mixing becomes stronger after 8:00 LT, NO₂ concentrations below 450 500 m decrease significantly, while those over 500 m increase from 6:00 - 8:00 LT to 12:00 - 14:00 LT. It is 451 noteworthy that PBLHs and NO_x emissions are comparable between 12:00 - 14:00 LT and 15:00 - 17:00 LT 452 (Figures 31 and 64); however, NO₂ concentrations at 15:00 - 17:00 LT are significantly higher than at 12:00 - 10:00 LT are significantly higher than at 12:00 - 10:00 LT are significantly higher than at 12:00 - 10:00 LT are significantly higher than at 12:00 - 10:00 LT are significantly higher than at 12:00 - 10:00 LT are significantly higher than at 12:00 - 10:00 LT are significantly higher than at 12:00 - 10:00 LT are significantly higher than at 12:00 - 10:00 LT are significantly higher than at 12:00 - 10:00 LT are significantly higher than at 12:00 - 10:00 LT are significant than at 12:00 - 10:00 L 453 14:00 LT in the whole boundary layer, reflecting the impact of the decreased photochemical loss of NO_x in the 454 late afternoon. In fact, photochemical losses affect all the daytime NO₂ vertical profiles, which can be easily 455 identified by NO₂ TVCD process diagnostics discussed in section 3.43 (Figure $\frac{8139}{2}$).

Figures <u>86</u>b and <u>86</u>d also show the observed and 36-km REAM simulated vertical profiles on weekends. Similar to Figures <u>53</u> and <u>75</u>, observed and simulated concentrations of NO₂ are lower on weekends than on weekdays. Some of the variations from weekend profiles are due to a lower number of observations (47 spirals) on weekends. The overall agreement between the observed vertical profiles and 36-km model results is good on weekends ($R^2 = 0.874$). At 15:00 – 17:00 LT, the model simulates a larger gradient than what the combination of aircraft and surface measurements indicates. It may be related to the somewhat underestimated PBLHs in the late afternoon in the model (Figure 64).

On weekdays, most simulated vertical profiles at the 4-km resolution (Figure <u>86</u>e) are similar to 36-km
 results in part because the average NO_x emissions over the six P-3B spiral sites are about the same, 4% lower in

465	the 4-km than the 36-km REAM simulations (Table S1). A clear exception is the 4-km REAM simulated vertical
466	profile at $15:00 - 17:00$ LT when the model greatly overestimates boundary layer NO _x mixing and
467	concentrations. The main reason is that WRF simulated vertical velocities (w) in the late afternoon are much
468	larger in the 4-km simulation than the 36-km simulation (Figure S ₂₁₁), which can explain the simulated fully
469	mixed boundary layer at $15:00 - 17:00$ LT. Since it is not designed to run at the 4-km resolution and it is
470	commonly assumed that convection can be resolved explicitly at high resolutions, the Kain-Fritsch (new Eta)
471	convection scheme is not used in the nested 4-km WRF simulation (Table S2); it may be related to the large
472	vertical velocities in the late afternoon when thermal instability is the strongest. Appropriate convection
473	parameterization is likely still necessary for 4-km simulations (Zheng et al., 2016), which may also help alleviate
474	the underestimation of precipitation in the nested 4-km WRF simulation as discussed in section 32.1 .

The same rapid boundary-layer mixing due to vertical transport is present in the 4-km REAM simulated weekend vertical profile (Figure <u>86</u>f), although the mixing height is lower. Fewer spirals (47) and distinct transport effect due to different NO₂ horizontal gradients between the 4-km and 36-km REAM simulations (discussed in detail in Section 3.<u>6</u>5) may cause the overestimation of weekend profiles in the 4-km REAM simulation.

480 3.43 Daytime variation of NO₂ TVCDs

We compare satellite, P-3B aircraft, and model-simulated TVCDs with Pandora measurements, which provide continuous daytime observations. The locations of Pandora sites are shown in Table S1 and Figure \$1. Among the Pandora sites, four sites are located significantly above the ground level: UMCP (~20 m), UMBC (~30 m), SERC (~40 m), and GSFC (~30 m). The other sites are 1.5 m AGL. To properly compare Pandora to other measurements and model simulations, we calculate the missing TVCDs between the Pandora site heights and ground surface by multiplying the Pandora TVCDs with model-simulated TVCD fractions of the corresponding columns. The resulting correction is $2-2\underline{1}0\%$ ($\frac{1}{1-missing TVCD percentage}$) for the four sites significantly above the ground surface, but the effect on the averaged daytime TVCD variation of all Pandora sites is small (Figure S1<u>2</u>4). In the following analysis, we use the updated Pandora TVCD data.

490 The weekday diurnal variations of NO₂ TVCDs from satellites, Pandora, 4- and 36-km REAM, and the P-3B 491 aircraft are shown in Figure <u>107</u>a. We calculate aircraft derived TVCDs by using equation (1):

492
$$TVCD_{aircraft}(t) = \frac{\sum c_{aircraft}(t) \times \rho_{REAM}(t) \times V_{REAM}(t)}{A_{REAM}}$$
 (1),

493 where t is time; $c_{aircraft}(v/v)$ denotes aircraft NO₂ concentrations (mixing ratios) at each level at time t; ρ_{REAM} (molecules / cm^3) is the density of air from 36-km REAM at the corresponding level; V_{REAM} (cm^3) is the volume of 494 495 the corresponding 36-km REAM grid cell; A_{REAM} (cm²) is the surface area (36 × 36 km²). In the calculation, we 496 only use NO₂ concentrations below 3.63 km AGL because few aircraft measurements were available above this 497 height in the campaign. Missing tropospheric NO₂ above 3.63 km AGL in the aircraft TVCD calculation has little 498 impact on our analyses, as 36-km REAM model simulation shows that $85\% \pm 712\%$ of tropospheric NO₂ are 499 located below 3.63 km AGL during 6:00 – 17:00 LT in the DISCOVER-AQ region, which is roughly consistent 500 with the GMI model results with 85% - 90% tropospheric NO₂ concentrated below 5 km (Lamsal et al., 2014). It 501 should be noted that only six P-3B spirals are available during the campaign, less than the samplings of 11 inland 502 Pandora sites.

The 4-km REAM simulated NO₂ TVCDs are mostly higher than the 36-km results and the observations <u>in</u> daytime on weekdays (Figure <u>107</u>a). However, since the standard deviations of the data are much larger than the model difference, the 4- and 36-km model results <u>show generally generally show</u> similar characteristics relative to the observations. REAM simulation results are in reasonable agreement with Pandora, P-3B aircraft, and satellite daytime NO₂ TVCDs, except that NASA-derived OMI (OMNO2) TVCDs are somewhat lower than

508	other datasets, which may be partly due to biased a priori vertical profiles from the GMI model in the NASA
509	retrieval in the campaign (Lamsal et al., 2014; Lamsal et al., 2020). TVCDs derived by using the DOMINO
510	algorithm and 36-km REAM NO2 vertical profiles are in agreement with those from KNMI, which indicates that
511	the TM4 model from KNMI provides reasonable estimates of a priori NO2 vertical profiles on weekdays in the
512	DISCOVER-AQ region in summer.
513	We find evident decreases of NO ₂ TVCDs from GOME-2A to OMI in Figure <u>107</u> a, which is consistent with
514	Pandora, REAM results, and previous studies that showed decreasing NO2 TVCDs from SCIAMACHY to OMI
515	due to photochemical losses in summer (Boersma et al., 2008; Boersma et al., 2009). P-3B aircraft TVCDs also
516	show this decrease feature but have large variations due in part to the limited aircraft sampling data.
517	Pandora NO ₂ TVCD data have different characteristics from REAM simulated and P-3B aircraft measured
518	TVCDs at 5:00 – 7:00 LT and 14:00 – 18:00 LT (Figure <u>10</u> 7a). At 5:00 – 7:00 LT, Pandora data show a
519	significant increase of NO ₂ TVCDs, but REAM and aircraft TVCDs generally decrease except for 4-km REAM
520	TVCDs with a slight increase from 6:00 – 7:00 LT on weekdays. At 14:00 LT – 18:00 LT, Pandora TVCDs have
 521	little variations, but REAM and aircraft TVCDs increase significantly. The relatively flat Pandora TVCDs in the
522	late afternoon compared to REAM and P-3B aircraft measurements are consistent with Lamsal et al. (2017),
523	which found the significant underestimation (26% - 30%) of Pandora VCDs compared to UC-12 ACAM
524	measurements from 16:00 LT to 18:00 LT during the DISCOVER-AQ campaign. We show the simulated effects
525	of emission, chemistry, transport, and dry deposition on NO _x TVCDs in Figure $\frac{\$139}{2}$. The simulated early
526	morning slight decrease of NO ₂ TVCDs is mainly due to the chemical transformation between NO ₂ and NO
l 527	favoring the accumulation of NO under low- O_3 and low- HO_2/RO_2 conditions, thus NO TVCDs increase
528	significantly but NO ₂ TVCDs continue decreasing slowly during the period. The increase in the late afternoon is
 529	primarily due to the decrease of photochemistry-related sinks. The reasons for the discrepancies of NO ₂ TVCDs

530	between Pandora and REAM results during the above two periods are unclear. Large SZAs in the early morning
531	and the late afternoon (Figure S18) lead to the higher uncertainties of Pandora measurements (Herman et al.,
l 532	2009), although we have excluded Pandora measurements with SZA $> 80^{\circ}$. In addition, Pandora is a sun-tracking
533	instrument with a small effective FOV and is sensitive to local conditions within a narrow spatial range which
534	may differ significantly from the average properties of 36- and 4-km grid cells depending upon the time of the
535	day (Figure S139) (Herman et al., 2009; Herman et al., 2018; Herman et al., 2019; Judd et al., 2018; Judd et al.,
536	2019; Judd et al., 2020; Lamsal et al., 2017; Reed et al., 2015). As we mentioned above, ~85% tropospheric NO ₂
537	are located below 3.63 km in the DISCOVER-AQ 2011 region based on the 36-km REAM simulation results.
538	The Pandora FOV of 1.6° is approximately equivalent to a nadir horizontal extension of only 0.1 km
539	$(2 \times 3.63 \text{ km} \times \tan \frac{1.6}{2} = 0.1 \text{ km})$ at 3.63 km AGL and 30 m at 1.0 km AGL. Therefore, Pandora measures
540	different air columns of NO2 at different times of the day, especially in the morning and afternoon when SZA is
541	large, as shown in Figure S13. Considering the potential spatial heterogeneity of boundary-layer NO ₂ , it is
542	possible that the morning (east), noontime (nadir), afternoon (west) NO2 VCDs are significantly different from
543	each other. Unlike Pandora, satellites and aircraft are far from the ground surface and cover large areas; therefore,
544	the impact of SZA on their NO ₂ VCD measurements is insignificant compared to Pandora measurements.
l 545	Another possible reason is that Pandora instruments had few observations in the early morning, and the resulting
546	average may not be representative (Figure $S_{\underline{29}}$).

To further understand the daytime variation of NO₂ TVCDs, we examine P-3B aircraft data derived and REAM simulated NO₂ VCD variations for different height bins (Figure <u>118</u>). NO₂ VCDs below 3.63 km AGL display a "U"-shaped pattern from 5:00 LT to 17:00 LT. In the morning, as vertical mixing becomes stronger after sunrise, high-NO_x air in the lower layer is mixed with low-NO_x air in the upper layer. The increase of NO_x vertical mixing above <u>3400</u> m is sufficient to counter the increase of photochemical loss in the morning. Conversely, the NO₂ VCDs below <u>3400</u> m decrease remarkably from sunrise (about 6:00 LT) to around noontime due to both vertical mixing and the increase of photochemical strength. From 13:00 LT to 16:00 LT, NO₂ VCDs increase slowly, reflecting a relative balance among emissions, transport, chemistry, and dry depositions. The sharp jump of the VCDs from 16:00 LT to 17:00 LT is mainly due to dramatically reduced chemical loss. And 4km REAM simulated NO₂ VCDs at 0.<u>340-3.63 km at 16:00-17:00 LT are much higher than 36-km results partly</u> because of the rapid vertical mixing in the 4-km REAM simulation (Figures <u>86</u> and S<u>911</u>).

558 Similar to NO_2 surface concentrations and vertical profiles in Figures 75 and 86, the NO_2 TVCD variation is 559 also smaller on weekends than on weekdays, but the day-night pattern is similar (Figure 107). Although the 4-km 560 REAM NO₂ TVCDs are generally higher than the 36-km results and observations in the daytime, considering 561 their large standard deviations, NO₂ TVCDs from both simulations are comparable to satellite products, Pandora, 562 and P-3B aircraft observations most of the time on weekends. The exception is that Pandora TVCDs have much 563 less different variation patterns in the early morning and late afternoon than-from REAM simulations, similar to 564 those found on weekdays and aircraft datasets. Another anomaly is that KNMI GOME 2A TVCDs at 9:30 LT are 565 much larger than the other datasets, while the GOME 2A TVCDs retrieved using 36 km REAM profiles shows 566 comparable TVCDs to Pandora, REAM, and aircraft datasets, reflecting possible biased NO₂ a priori profiles 567 from the TM4 model on weekends used in the KNMI GOME-2A retrieval.

568 3.<u>5</u>4 Model comparisons with NO_y measurements

NO_y is longer-lived than NO_x, and NO_y concentrations are not affected by chemistry as much as NO_x. We obtain two types of NO_y concentrations from the P-3B aircraft in the DISCOVER-AQ campaign: one is NO_y concentrations directly measured by the NCAR 4-channel instrument, corresponding to the sum of NO, NO₂, $\sum PNs$, $\sum ANs$, HNO₃, N₂O₅, HNO₄, HONO, and the other reactive nitrogenic species in REAM (all the other species are described in Table 1); the other one, which we name as "derived-NO_y", is the sum of NO from the

574	NCAR 4-channel instrument and NO ₂ (NO ₂ _LIF), Σ PNs, Σ ANs, and HNO ₃ measured by the TD-LIF technique,
575	corresponding to NO, NO ₂ , Σ PNs, Σ ANs, and HNO ₃ in REAM (Table 1). On average, P-3B derived-NO _y
576	concentrations (2.88 \pm 2.24 ppb) are 17% higher than coincident P-3B NO _y concentrations (2.46 \pm 2.06 ppb) with
577	$R^2 = 0.75$, generally reflecting consistency between these two types of measurements. As shown in Table 1, on
578	weekdays, the 36-km REAM NO _y concentrations are 45% larger than P-3B with $R^2 = 0.33$, and the 36-km
579	REAM derived-NO _y concentrations are 8% larger than P-3B with $R^2 = 0.41$. 4-km REAM show similar results,
580	suggesting that REAM simulations generally reproduce the observed NO _y and derived-NO _y concentrations within
581	the uncertainties, although the average values from REAM are somewhat larger than the observations due in part
582	to the underestimate of precipitation in the WRF model simulations resulting in underestimated wet scavenging
583	of HNO ₃ in REAM. The concentrations of weekday NO, NO ₂ , and \sum PNs from REAM simulations are also
l 584	comparable to the observations. However, weekday ∑ANs concentrations are 68% lower in the 36-km REAM
585	than observations, suggesting that the chemistry mechanism in REAM may need further improvement to better
586	represent isoprene nitrates. It is noteworthy that, since $\sum ANs$ only account for a small fraction (~11%) in
587	observed derived-NO _y , the absolute difference between REAM simulated and P-3B observed \sum ANs
588	concentrations is still small compared to HNO ₃ . Weekday HNO ₃ concentrations are significantly higher in
589	REAM simulations (36-km: 57%, 0.65 ppb; 4-km: 740%, 0.862 ppb) than P-3B observations, which is the main
 590	reason for the somewhat larger NO _y and derived-NO _y concentrations in REAM compared to P-3B observations.
591	The higher HNO ₃ concentrations in REAM may be related to the underestimation of precipitation in the
592	corresponding WRF simulations, as discussed in section 32.1 (Figures S 75 and S8), leading to the underestimated
 593	wet scavenging of HNO ₃ , especially for the 4-km REAM simulation.

We also examine the weekday diurnal variations of derived-NO_y vertical profiles from P-3B and REAM simulations in Figure S1<u>45</u>. Generally, both 36- and 4-km REAM simulations capture the variation characteristics of observed vertical profiles, which are similar to those for NO₂ in Figure <u>86</u>. REAM derived-NO_y concentrations 597 are comparable to P-3B observations at most vertical levels on weekdays. Some larger derived-NO_y

598 concentrations in the model results can be partially explained by larger HNO₃ concentrations in REAM, such as 599 those below 1 km at 9:00 - 11:00 LT for the 36-km REAM and those below 2.0 km at 12:00 - 17:00 LT for the 500 4-km REAM (Figure S156).

Figure <u>129</u> shows the comparison of the diurnal cycles of surface NO_y concentrations observed at Padonia, Edgewood, Beltsville, and Aldino during the DISCOVER-AQ campaign with those from the REAM simulations. Generally, the REAM simulations reproduce the observed surface NO_y diurnal cycles except for the spikes around 17:00 – 20:00 LT due to still underestimated PBLHs (Figure <u>64</u>). 4-km simulation results have a higher bias than 36-km results relative to the observations <u>in the daytime</u>, similar to the comparisons of NO₂ surface concentrations and TVCDs in Figures <u>75</u> and <u>107</u> due to higher emissions around the observation sites in 4- than 36-km simulations (Table S1 and Figure <u>\$2</u>).

3.65 Resolution dependence of NO_x emission distribution

609 We show previously that the 4-km REAM simulated NO₂ and NO_y surface concentrations and NO₂ TVCDs 610 are higher than observations at in the daytime in comparison to the corresponding 36-km REAM results (Figures 611 7, 10, and 12). An examination of monthly mean NO₂ surface concentrations and TVCDs for July 2011 also 612 shows that 4-km simulation results are significantly higher than the 36-km results over the 11 inland Pandora 613 sites in the daytime (Figure $13\frac{817}{1}$). The process-level diagnostics in Figure $9\frac{813}{1}$ indicate that the mean contribution of NO_x emissions to NO_x Δ TVCDs in the 4-km simulation is 1.32×10^{15} molecules cm⁻² h⁻¹ larger 614 615 than that in the 36-km simulation between 9:00 LT and 16:00 LT, while the absolute mean contributions of 616 chemistry and transport (they are negative in Figure <u>\$139</u>, so we use absolute values here) in the 4-km simulation are 0.262×10^{15} and 0.8799×10^{15} molecules cm⁻² h⁻¹ larger than the 36-km simulation, respectively. The 617 618 contributions of dry deposition to NO_x Δ TVCDs are negligible compared to other factors in both simulations

619	(Figure $\frac{S139}{2}$). Therefore, the 34% higher NO _x emissions over the 11 inland Pandora sites (Table S1 and Figure
620	3^{-1}) is the main reason for the larger daytime NO ₂ surface concentrations and TVCDs in the 4-km than the 36-km
621	REAM simulations (Figure $\frac{\$1713}{13}$). The significantly different contribution changes between NO _x emissions
622	$(1.32 \times 10^{15} \text{ molecules cm}^{-2} \text{ h}^{-1} \text{ or about one third})$ and chemistry $(0.2\underline{62} \times 10^{15} \text{ molecules cm}^{-2} \text{ h}^{-1} \text{ or about } \underline{87\%})$
623	reflect potential chemical nonlinearity (Li et al., 2019; Silvern et al., 2019; Valin et al., 2011) and transport effect.
624	Different transport contributions between the 4-km and the 36-km REAM are mainly caused by their different
625	NO _x horizontal gradients (Figures S2-and, 14, and 150), while the impact of wind fields is small since we do not
626	find significant differences in horizontal wind components between the two simulations except for some lower
627	wind speeds below 1000 m for the 36-km WRF simulation compared to the nested 4-km WRF simulation (Figure
628	S168). Our sensitivity tests with the WRF Single-Moment 3-class (WSM3) simple ice scheme (not shown) can
629	improve the wind speed comparison below 1000 m between the 36-km and nested 4-km WRF simulations but
630	still produce similar NO _x simulation results as WSM6 shown here. Therefore, the somewhat lower wind speeds
631	below 1000 m in the 36-km WRF simulation are not the reason for the difference between the 4-km and 36-km
632	<u>REAM simulations.</u> The impact of transport on the two <u>REAM</u> simulations can be further verified by the
633	comparison of NO ₂ TVCDs over the six P-3B spiral sites between the two simulations (Figure S1 $\frac{79}{2}$). Mean NO _x
634	emissions over the six P-3B spiral sites are close (relative difference $< 4\%$) between the two simulations (Table
635	S1 and Figure S179). From 9:00 to 12:00 LT, the contributions of NO _x emissions to NO _x Δ TVCDs are 2.50 ×
636	10^{15} and 2.49×10^{15} molecules cm ⁻² h ⁻¹ for the 36-km and 4-km REAM simulations, respectively, and the
637	contributions of chemistry are also close between the two simulations (36-km: $-2.64-62 \times 10^{15}$ molecules cm ⁻² h ⁻¹ ;
638	4-km: -2.698×10^{15} molecules cm ⁻² h ⁻¹). However, the contributions of transport are $-0.32-39 \times 10^{15}$ and $0.04-03$
l 639	\times 10 ¹⁵ molecules cm ⁻² h ⁻¹ for the 36-km and 4-km REAM simulations, respectively, leading to larger NO ₂ TVCDs
640	in the 4-km REAM simulation than the 36-km REAM from $9:00 - 12:00$ LT (Figure S179c). Since horizontal
641	wind fields over the six P-3B spiral sites are comparable between two simulations (Figures S43, S45, S6, and

 $S1\underline{68}$) and larger NO_x horizontal gradients are found near the P-3B spiral sites for the 4-km REAM (Figure S2), we attribute the different transport contributions between the two simulations to a much larger NO_x emission gradient around the measurement locations in 4-km than 36-km emission distributions.

645 We re-grid the 4-km REAM results into the grid cells of the 36-km REAM, which can significantly reduce 646 the impact of different NO_x emission distributions and associated transport on the two simulations. Compared to 647 the original 4-km REAM results, the re-gridded surface NO₂ concentrations and TVCDs over the 11 inland 648 Pandora sites are much closer to the 36-km REAM results (Figure <u>\$1713</u>). After re-gridding the 4-km REAM 649 results into 36-km REAM grid cells, we also find more comparable NO_v surface concentrations between the re-650 gridded 4-km results and the 36-km REAM results (Figure S1820). The remaining discrepancies between the re-651 gridded results and the 36-km REAM results may be due to chemical nonlinearity and other meteorological 652 effects, such as larger vertical wind in the 4-km REAM (Figure S⁹⁴⁴) and their different k_{zz} values in the PBL. 653 Although other factors, such as chemical nonlinearity and vertical diffusion, may affect the 36-km and 4-km 654 REAM simulations differently, the difference between 4- and 36-km simulations of reactive nitrogen is largely 655 due to that of NO_x emissions.

656 The 4- and 36-km simulation difference depends on the location of the observations. In some regions, the 657 NO_x emission difference between 4- and 36-km simulations is small. The comparison of NO_y measurements from 658 P-3B spirals with coincident REAM results in Table 1 suggests that the 4-km and 36-km REAM simulations 659 produce similar NO_v (relative difference $\sim \frac{24}{9}$) and derived-NO_v (relative difference $\sim \frac{64}{9}$) concentrations on 660 weekdays, and both simulation results are comparable to the observations. The NO_y similarity over the P-3B 661 spiral sites between the 36-km and 4-km REAM simulations is consistent with the comparable NO_x emissions 662 over (relative difference < 4%) the six P-3B spiral sites between the two simulations (Table S1). The differences 663 between the 4-km model simulation results and P3-B observations are larger on weekends than on weekdays

(Table 1) due to the limited weekend sampling since model simulated monthly mean values show similardifferences between the 4-km and 36-km REAM simulations on weekends as on weekdays (not shown).

666	3. <u>7</u> 6 Evaluation of <u>36- and </u> 4-km NO _x distribution with <u>OMI, GOME-2A, and</u> ACAM measurements
667	The evaluation of model simulations of surface, aircraft, and satellite observations tends to point out a high
668	bias in 4- than 36-km model simulations. We note that this comparison is based on the averages of multiple sites.
669	NO _x emissions at individual sites are not always higher in the 4-km than 36-km REAM, such as SERC, Fairhill,
670	and Essex, with much higher 36-km NO _x emissions than 4-km NO _x emissions (Table S1). We conduct
671	individual-site comparisons of surface NO ₂ concentrations, surface NO _y concentrations, NO ₂ vertical profiles,
672	derived-NO _y vertical profiles, and NO ₂ TVCDs of the 36-km REAM and the 4-km REAM results relative to the
673	corresponding observations in Figures S19 – S23. The 36-km simulation results can be larger, smaller, or
674	comparable to the 4-km simulation results, and both simulations can produce higher, lower, or similar results as
675	the observations for different variables at different sites. The varying model biases depending on the observation
676	site reflect the different spatial distributions of NO _x emissions between the 36- and 4-km REAM simulations
677	(Figure 2) and suggest potential distribution biases of NO_x emissions in both simulations. However, we note that
678	the uncertainties of the observations and model data are often comparable or larger than the model differences.
679	Here we examine the 4-km model simulated NO ₂ VCDs with high-resolution ACAM measurements onboard
680	the UC-12 aircraft in Figures 140 and S241, respectively. The spatial distributions of ACAM and 4-km REAM
681	NO ₂ VCDs are generally consistent with $R^2 = 0.357$ on weekdays and $R^2 = 0.50$ on weekends. The domain
682	averages of ACAM and 4-km REAM NO ₂ VCDs are 4.7 ± 2.0 and $4.65 \pm 3.2 \times 10^{15}$ molecules cm ⁻² on weekdays
683	and 3.0 ± 1.7 and $3.3 \pm 2.28 \times 10^{15}$ molecules cm ⁻² on weekends, respectively. The spatial distributions of ACAM
l 684	and 4-km REAM NO ₂ VCDs are highly correlated with the spatial distribution of 4-km NEI2011 NO _x emissions.
685	All three distributions capture two strong peaks around Baltimore and Washington, D.C. urban regions and

686	another weak peak in the northeast corner of the domain (Wilmington city in Delaware) (Figures 140 and S241).
687	However, Figures 140 and S241 clearly show that NO ₂ VCDs from the 4-km REAM simulation are more
688	concentrated in Baltimore and Washington, D.C. urban regions than ACAM, which are also reflected by the
689	higher NO ₂ VCD standard deviations of the 4-km REAM results than ACAM. Several Pandora sites are in the
690	highest NO ₂ VCD regions where the 4-km REAM generally produces larger NO ₂ VCDs than ACAM, which
691	explains why the NO ₂ TVCDs over the 11 Pandora sites from the 4-km REAM simulation are higher than the
692	observations (Figure 107) and the 36-km REAM results (Figure 13817) around noontime. Horizontal transport
693	cannot explain the NO ₂ VCD distribution biases in the 4-km REAM simulation due to the following reasons.
694	Firstly, horizontal wind fields are simulated as well by the nested 4-km WRF simulation as the 36-km WRF
695	compared to P-3B measurements, as discussed in section 32.1 . Secondly, the prevailing northwest wind in the
696	daytime (Figure S ⁵⁴) should move NO _x eastward, but we find no significant eastward shift of NO ₂ VCDs
697	compared to NO _x emissions in both ACAM and 4-km REAM distributions (Figure 1 <u>4</u> θ). Lastly, we find a local
698	minimum of NO ₂ VCDs in the middle of the Baltimore urban region (the purple circle in Figure $\frac{10b14b}{14b}$) in the
699	ACAM distribution, which cannot be explained by horizontal transport or chemical nonlinearity due to the
700	surrounding high NO _x emissions in the 4-km REAM simulation. Therefore, we attribute the distribution
701	inconsistency between ACAM and the 4-km REAM to the distribution biases of NEI2011 NO_x emissions at the
702	4-km resolution since the average below-aircraft NO ₂ VCDs between ACAM and the 4-km REAM are about the
703	same.

704It is noteworthy that about 91% ACAM NO2 VCD data are measured from 8:00 - 16:00 LT, and only using705ACAM NO2 VCDs between 8:00 and 16:00 LT for the above comparison does not affect our results shown here.706Moreover, to minimize the effect of overestimated afternoon vertical mixing (Figure <u>86</u>) on the 4-km REAM707simulation results, we also examine the comparison between ACAM NO2 VCDs from 9:00 - 14:00 LT with708coincident 4-km REAM results, which produces similar results as shown here.

709	We also evaluate the NC	² VCD distributions	from the 4-km REAM	simulation on weekdays	s and weekends
-----	-------------------------	--------------------------------	--------------------	------------------------	----------------

vith ACAM NO₂ VCDs below the U-12 aircraft obtained from <u>https://www-air.larc.nasa.gov/cgi-</u>

711 <u>bin/ArcView/discover-aq.dc-2011?UC12=1#LIU.XIONG/</u> in Figures S2<u>5</u>² and S2<u>6</u>³. Although the domain mean

ACAM NO₂ VCDs in Figures S252 and S263 are higher than coincident 4-km REAM results due to the different

retrieval method from Lamsal et al. (2017), such as different above-aircraft NO₂ VCDs and different a priori NO₂

vertical profiles, we can still find clear distribution inconsistencies between the 4-km REAM and ACAM NO₂

- VCDs. The 4-km REAM NO₂ VCDs are more concentrated in the Baltimore and Washington, D.C. urban regions
 than this set of ACAM data, which is consistent with the conclusions derived from the ACAM dataset retrieved
- 717 by Lamsal et al. (2017).

718 The potential distribution bias of the NEI2011 NO_x emissions at 36-km resolution is analyzed by comparing 719 the 36-km REAM simulated NO₂ TVCDs with those retrieved by OMI and GOME-2A, as shown in Figures 15 720 (OMI, 13:00 LT) and S27 (GOME-2A, 9:30 LT). Both KNMI and our retrievals based on the 36-km REAM NO2 721 vertical profiles show that OMI and GOME-2A NO₂ TVCDs have lower spatial variations than the corresponding 722 36-km REAM simulation results. OMI and GOME-2A retrievals have lower NO₂ TVCDs around the Baltimore 723 and Washington, D.C. urban regions and higher values in relatively rural regions than the 36-km REAM. The 724 distribution bias of the 36-km REAM NO₂ TVCDs is also identified on weekends through their comparison with 725 OMI and GOME-2A retrievals (not shown). The good agreement between simulated and observed wind suggests 726 that the model horizontal transport error cannot explain such an urban-rural contrast between satellite 727 observations and 36-km REAM simulation results. However, two caveats deserve attention. Firstly, the 36-km 728 REAM cannot resolve urban areas as detailed as the 4-km REAM (Figure 14), and urban and rural regions may

- coexist in one 36-km grid cell. Secondly, the OMI and GOME-2A pixels can be much larger than 36-km REAM
- grid cells, possibly leading to more spatially homogenous distributions of satellite NO₂ TVCD data.

731 $3.\underline{87}$ Implications for NO_x emissions

732 The analysis of section 3.76 indicates that the NEI2011 NO_x emission distributions at-the 36- and 4-km 733 resolutions is are likely biased for the Baltimore-Washington region. The distribution bias of the high resolution 734 NO_x emission inventories is corroborated by the comparison of the NO_x emission inventory derived from the 735 CONsolidated Community Emissions Processor Tool, Motor Vehicle (CONCEPT MV) v2.1 with that estimated 736 by the Sparse Matrix Operator Kernel Emissions (SMOKE) v3.0 model with the Motor Vehicle Emissions 737 Simulator (MOVES) v2010a (DenBleyker et al., 2012). CONCEPT with finer vehicle activity information as 738 input produced a wider-spread but less-concentrated running exhaust NO_x emissions compared to MOVES in the 739 Denver urban area for July 2008 (DenBleyker et al., 2012). In addition, Canty et al. (2015) found that CMAQ 740 4.7.1, with on-road emissions from MOVES and off-road emissions from the National Mobile Inventory Model 741 (NMIM), overestimated NO₂ TVCD over urban regions and underestimated NO₂ TVCDs over rural areas in the 742 northeastern U.S. for July and August 2011 compared to the OMNO2 product. The urban-rural contrast was also 743 found in Texas during the 2013 DISCOVER-AQ campaign in the studies of Souri et al. (2016) and Souri et al. 744 (2018), implying distribution uncertainties in NO_x emissions, although these studies and Canty et al. (2015) 745 focused more on polluted regions with overestimated NO_x emissions in their conclusions. The emission 746 distribution bias may also explain why Anderson et al. (2014) have different results from our simulated 747 concentrations in Table 1. In their study, they compared in-situ observations with a nested CMAQ simulation 748 with a resolution of 1.33 km. It is difficult to build up a reliable emission inventory for the whole U.S. at very 749 high resolutions with currently available datasets due to the significant inhomogeneity of NO_x emissions (Marr et 750 al., (2013), but we can still expect significant improvements of the temporal-spatial distributions of NO_x 751 emissions in the near future as GPS-based information start to be used in the NEI estimates (DenBleyker et al., 752 2017).

753	Although the NEI2011 NO _x emission distribution is likely biased at 4-km resolution, the similar average
754	below-aircraft NO ₂ -VCDs between ACAM and the 4-km REAM (Figure 10), as well as the good performance of
755	the 36-km REAM compared to NO ₂ and NO _y observations (Figures 3 and 5-9 and Table 1), suggest that NEI2011
756	may provide reliable estimates of NO _* emissions over the Baltimore-Washington region at coarser resolutions. It
757	should be noted that although the good performance of the 36-km REAM in reproducing P-3B, Pandora, and
758	surface observations may be limited by representative errors due to the relatively coarse resolution of 36-km
759	REAM, ACAM and satellite NO ₂ -VCDs are all observations covering large areas, alleviating the problem caused
760	by observations limited to small areas. Our conclusion on the potential reliability of NEI2011 is consistent with
761	Salmon et al. (2018), who found NEI2011 and NEI2014 were in agreement with aircraft observation-derived NO _*
762	emissions during the Wintertime INvestigation of Transport, Emissions, and Reactivity (WINTER) campaign in
763	February March 2015 around the Washington, D.C. Baltimore area. The agreement was further confirmed
764	through the investigation of observed and NEI NO _* /CO ₂ , CO/NO _* , and CO/CO ₂ ratios (Salmon et al., 2018).
765	However, our evaluation of NEI NO _x emissions is different from Travis et al. (2016) and Anderson et al. (2014).
766	Travis et al. (2016) compared the GEOS-Chem simulation results with the observations of NO _* and its oxidation
767	products from the SEAC ⁴ RS campaign in the Southeast US, nitrate wet deposition fluxes from the National Acid
768	Deposition Program (NADP) network, and NO2 TVCDs from OMI, and suggested that NEI2011 overestimated
769	mobile and industrial NO _x -emissions by 30% 60%. The GEOS-Chem chemical mechanism from Travis et al.
770	(2016) is almost the same as what we use in REAM, and their model simulation has a horizontal resolution of
771	$0.25^{\circ} \times 0.3125^{\circ}$, which is also close to REAM (36 km \times 36 km). We attribute the discrepancies between Travis et
772	al. (2016) and our study to the regional discrepancies. Travis et al. (2016) derived their conclusions based on the
773	averages of large domains (the whole CONUS or the Southeast U.S.), while our study focuses on the much
774	smaller Baltimore-Washington region. If limited to the Baltimore-Washington region, Figures 3 and 5 in Travis
775	et al. (2016) shown that nitrate wet deposition fluxes from NADP and TVCDs from the Berkeley High-
I	

776	Resolution (BEHR) retrieval and NASA (OMNO2) were significantly higher than their simulations with NEI
777	non-power-plant NO _* emissions reduced by 60%, implying their conclusions about the overestimation of NEI
778	NO _x -emissions at least not applicable to the Baltimore-Washington region. Anderson et al. (2014) evaluated
779	NEI2011 emissions with the observed concentration ratios of CO to NO _y and CO to NO _x from the same
780	DISCOVER-AQ campaign. They concluded that NEI overestimated NO _x emissions by 51% – 70% in Maryland
781	in the summer of 2011. However, the uncertainties of scaling the emission ratios of CO/NO _* to the concentrations
782	ratios of CO/NO _y or CO/NO _* can be large due in part to the large contribution of biogenic isoprene oxidation to
783	the variation of CO (Cheng et al., 2017). Furthermore, base CMAQ simulated Σ PNs (1.4 ppbv) and Σ ANs (0.96)
784	ppbv) concentrations by Anderson et al. (2014) are 0.79 and 0.64 ppbv higher than the corresponding P-3B
785	observations, respectively, leading to the overestimation of NO _y in their model; these biases are much larger than
786	those (~0.2 ppbv) from our 36 km REAM simulation as shown in Table 1, possibly due to different chemical
787	mechanisms used by CMAQ and REAM. Another possible reason, as mentioned above, is that the base CMAQ
788	used by Anderson et al. (2014) has a horizontal resolution of 1.33 km, much higher than our REAM simulations;
789	NO _x emission distribution may be a potential issue causing the overestimation of the base CMAQ results.
790	Here, we emphasize that our study is not necessarily contradictory to recent studies concerning the
791	overestimation of NEI NOx emissions (Anderson et al., 2014; Canty et al., 2015; McDonald et al., 2018; Souri et
792	al., 2016; Souri et al., 2018; Travis et al., 2016). Different types of observations in different periods and locations
793	are analyzed for various purposes. This study focuses more on the spatial distribution of NO_x emissions in
794	NEI2011 at the 4-km resolution, while previous studies are concerned more about the NO _x emission magnitudes
l 795	in highly polluted sites, although the spatial distribution issue was also mentioned in some of the studies. If we
796	limit our analyses to those observations in Figures $\frac{75}{107}$, and $\frac{129}{107}$ and the 4-km REAM, we would also
797	conclude an overestimation of NEI NO _x emissions. In this study, through comprehensive evaluations of NO ₂ and
798	NOy measurements at 36- and 4-km resolutions, we provide another possible explanation for the overestimation
I	

799	of high resolution model results at polluted sites. Although our assessment of NEI2011 NO _* emissions at 36 km
800	resolution is contingent upon potential representative errors, the further comparisons of below-aircraft NO2 VCDs
801	between the 4-km REAM and ACAM from Lamsal et al. (2017) suggest that total NO _* emissions seem accurate
802	but the emission distribution seems biased in the 4-km NEI2011 in our defined domain (Figure 10). Considering
803	the significant heterogeneity of NO _x emissions, different conclusions on NO _x emission biases may be made in
804	other regions, but the spatial distribution of NO _x emissions is a critical factor in evaluating NO _x emissions and
805	improving emission estimation and air quality models, which deserves more attention in future studies, especially
806	when chemical and transport models are moving to higher and higher resolutions.

4 Conclusions

We investigate the diurnal cycles of surface NO₂ concentrations, NO₂ vertical profiles, and NO₂ TVCDs using REAM model simulations on the basis of the observations from air quality monitoring sites, aircraft, Pandora, OMI, and GOME-2A during the DISCOVER-AQ 2011 campaign. We find that WRF simulated nighttime <u> k_{zz} -determined</u> PBLHs are significantly lower than ELF lidar measurements. Increasing nighttime mixing from 18:00 – 5:00 LT in the REAM simulations, we significantly improve REAM simulations of nighttime surface NO₂ and O₃ concentrations.

The REAM simulation reproduces well the observed <u>regional mean</u> diurnal cycles of surface NO₂ and NO_y concentrations, NO₂ vertical profiles, and NO₂ TVCDs on weekdays. Observed NO₂ concentrations in the boundary layer and TVCDs on weekends are significantly lower than on weekdays. By specifying a weekend to weekday NO_x emission ratio of 2:3 and applying a less variable NO_x emission diurnal profile on weekends than weekdays, REAM can simulate well the weekend observations. <u>A fewTwo</u> issues are also noted. First, Pandora TVCDs show different variations from aircraft-derived and REAM-simulated TVCDs in the early morning and late afternoon, which may be due to the uncertainties of Pandora measurements at large SZAs and the small effective FOV of Pandora. Second, the weekday OMI NO₂ TVCDs derived by NASA are somewhat lower than
the KNMI OMI product, P-3B aircraft-derived TVCDs, Pandora, and REAM results; the difference may be
caused by the a priori vertical profiles used in the NASA retrieval. Lastly, the weekend OMI NO₂ TVCDs
derived by KNMI are larger than those from Pandora, P-3B aircraft, REAM, and the OMI retrieval with REAM
NO₂ vertical profiles, indicating a possible bias of the TM4 simulated a priori NO₂ vertical profiles in the
weekend mornings during DISCOVER AQ 2011.

827 While a higher-resolution simulation is assumed to be superior at a priori, the large observation dataset 828 during DISCOVER-AO 2011 offers the opportunity of a detailed comparison of 4-km and 36-km model 829 simulations. In general, the 4 km simulation results tend to have a high bias relate to the 36 km results in light of 830 the observations. Through the comparison, wWe find two areas that have not been widely recognized for high-831 resolution model simulations. The first is not using convection parameterization in high-resolution WRF 832 simulations since convection can be resolved explicitly and most convection parameterizations are not designed 833 for high-resolution simulations. We find that 4-km WRF tends to overestimate boundary-layer mixing and 834 vertical transport in the late afternoon, leading to a high model bias in simulated NO₂ vertical profiles compared 835 to P-3B aircraft observations. The reasons for this late-afternoon bias in 4-km WRF simulations and model 836 modifications to mitigate this bias need further studies.

A second issue is related to the spatial distribution of NO_x emissions in NEI2011. <u>In general, the 4-km</u>
 <u>simulation results tend to have a high bias relative to the 36-km results on the regional mean observations.</u>
 <u>However, for individual sites, relative to the 36-km model simulations, the 4-km model results can show larger,</u>
 <u>smaller, or similar biases compared to the observations depending upon observation location. Based on process</u>
 <u>diagnostics and analyses, we find that the bias discrepancies between the 36-km and 4-km REAM simulations are</u>
 <u>mainly attributed to their different NO_x emissions and their spatial gradients at different sites. At 4-km, the grid</u>

843	cells over the 11 inland Pandora sites have 34% higher NO _x emissions than the 36-km grid cells. Consequently,
844	4 km REAM overestimates NO2 concentrations and TVCDs than the observations. The 4 km grid cells over four
845	surface NOy measurement sites have about a factor of 2 higher NOx emissions than the corresponding 36 km grid
846	cells, leading to significantly overestimated NOy concentrations in the 4 km REAM simulation compared to the
847	observations. After we re-grid the 4-km NO2 and NOy results to the 36-km grid cells, the results of 4-km
848	simulations are similar to the original 36-km simulations. The comparison of 4-km ACAM NO2 VCD
l 849	measurements from the UC-12 aircraft with coincident 4-km REAM results shows that 4-km REAM NO2 VCDs
850	are more concentrated in urban regions than the ACAM observations. OMI and GOME-2A data also show less
851	spatially varying NO ₂ TVCD distributions with lower NO ₂ TVCDs around the Baltimore-Washington urban
852	regions and higher TVCDs in surrounding rural areas than corresponding 36-km REAM simulation results.
853	Further model analysis indicates that the <u>36- and 4-km VCD</u> discrepancies are due primarily to the distribution
854	bias of 4-km-NEI2011 NO _x emissions at 36- and 4-km resolutions. At high resolutions, potential biases in the
855	emission inventories are accentuated when model results are evaluated with the observations. Our results
856	highlight the research need to improve the methodologies and datasets used to improve the spatial distributions in
857	emission estimates at high resolutions.

858 Data availability

- 859 The DISCOVER-AQ 2011 campaign datasets are archived on https://www-air.larc.nasa.gov/cgi-
- 860 bin/ArcView/discover-aq.dc-2011 (last access: March <u>146</u>, 202<u>1</u>0). EPA air quality monitoring datasets are from
- https://www3.epa.gov/airdata/ (last access: June 23, 2015). The NASA OMI NO₂ product is from
- 862 https://disc.gsfc.nasa.gov/datasets/OMNO2_003/summary (last access: September 26, 2020). The KNMI OMI
- 863 NO₂ product is from http://www.temis.nl/airpollution/no2.html (last access: January 14, 2015). We obtain the
- 864 KNMI GOME-2A NO₂ VCD archives from http://www.temis.nl/airpollution/no2col/no2colgome2_v2.php (last
- access: January 22, 2015). The GMI MERRA-2 simulation results are from

- 866 https://portal.nccs.nasa.gov/datashare/dirac/gmidata2/users/mrdamon/Hindcast-
- 867 Family/HindcastMR2/2011/stations/ (last access: May 14, 2019). We obtain the UC-12 ACAM NO₂ VCD
- 868 product by X. Liu from https://www-air.larc.nasa.gov/cgi-bin/ArcView/discover-aq.dc-
- 2011?UC12=1#LIU.XIONG/ (last access: December 31, 2019). The Stage IV precipitation data is downloaded
- 870 from https://rda.ucar.edu/datasets/ds507.5/ (last access: December 28, 2019). The NCEP CFSv2 6-hourly product
- 871 is available at <u>http://rda.ucar.edu/datasets/ds094.0/</u> (last access: March 10, 2015). REAM simulation results for
- this study and the UC-12 ACAM NO₂ VCD product by Lamsal et al. (2017) are available upon request.

873 Author contribution

- JL and YW designed the study. JL, RZ, and CS updated the REAM model. JL conducted model simulations.
- 875 KFB developed the DOMINO algorithm, CS applied the algorithm to REAM vertical profiles, and JL updated the
- 876 retrieval algorithm and did the retrieval by using REAM NO₂ vertical profiles. AW, JH, EAC, RWL, JJS, RD,
- 877 AMT, TNK, LNL, SJJ, MGK, XL, CRN made various measurements in the DISCOVER-AQ 2011 campaign. JL
- conducted the analyses with discussions with YW, RZ, CS, AW, JH, KFB, EAC, RWL, JJS, RD, AMT, TNK,
- 879 LNL, SJJ, MGK, XL, and CRN. JL and YW led the writing of the manuscript with inputs from all other
- coauthors. All coauthors reviewed the manuscript.

881 Competing interests

882 The authors declare that they have no conflict of interest.

883 Acknowledgments

- 884 This work was supported by the NASA ACMAP Program. We thank Chun Zhao for providing us the PNNL
- 885 NEI2011 emission inventory. We thank Yuzhong Zhang and Jenny Fisher for providing the updated GEOS-
- 886 Chem chemistry mechanism files and thank Yuzhong Zhang, Yongjia Song, Hang Qu, Ye Cheng, Aoxing Zhang,

- 887 Yufei Zou and Ziming Ke for discussion with J. Li. We thank Susan Strahan for providing the GMI outputs
- download link.

889 References

- Anderson, D. C., Loughner, C. P., Diskin, G., Weinheimer, A., Canty, T. P., Salawitch, R. J., Worden, H. M.,
 Fried, A., Mikoviny, T., and Wisthaler, A.: Measured and modeled CO and NO_y in DISCOVER-AO: An
- evaluation of emissions and chemistry over the eastern US, Atmos. Environ., 96, 78-87,
- 893 https://doi.org/10.1016/j.atmosenv.2014.07.004, 2014.
- Beirle, S., Platt, U., Wenig, M., and Wagner, T.: Weekly cycle of NO₂ by GOME measurements: A signature of
 anthropogenic sources, Atmos. Chem. Phys., 3, 2225-2232, https://doi.org/10.5194/acp-3-2225-2003, 2003.
- Boersma, K. F., Eskes, H. J., Veefkind, J. P., Brinksma, E. J., Van Der A, R. J., Sneep, M., Van Den Oord, G. H.
 J., Levelt, P. F., Stammes, P., and Gleason, J. F.: Near-real time retrieval of tropospheric NO₂ from OMI, Atmos.
 Chem. Phys., 7, 2103-2118, https://doi.org/10.5194/acp-7-2103-2007, 2007.
- 899 Boersma, K. F., Jacob, D. J., Eskes, H. J., Pinder, R. W., Wang, J., and Van Der A, R. J.: Intercomparison of
- SCIAMACHY and OMI tropospheric NO₂ columns: Observing the diurnal evolution of chemistry and emissions
 from space, J. Geophys. Res.-Atmos., 113, https://doi.org/10.1029/2007JD008816, 2008.
- Boersma, K. F., Jacob, D. J., Trainic, M., Rudich, Y., De Smedt, I., Dirksen, R., and Eskes, H. J.: Validation of
 urban NO₂ concentrations and their diurnal and seasonal variations observed from the SCIAMACHY and OMI
 sensors using in situ surface measurements in Israeli cities, Atmos. Chem. Phys., 9, 3867-3879,
 https://doi.org/10.5194/acp-9-3867-2009, 2009.
- Boersma, K. F., Eskes, H. J., Dirksen, R. J., Veefkind, J. P., Stammes, P., Huijnen, V., Kleipool, Q. L., Sneep,
 M., Claas, J., and Leitão, J.: An improved tropospheric NO₂ column retrieval algorithm for the Ozone Monitoring
 Instrument, Atmos. Meas. Tech., 4, 1905-1928, https://doi.org/10.5194/amt-4-1905-2011, 2011.
- 909 Boersma, K. F., Eskes, H. J., Richter, A., De Smedt, I., Lorente, A., Beirle, S., van Geffen, J. H., Zara, M., Peters,
- E., and Roozendael, M. V.: Improving algorithms and uncertainty estimates for satellite NO₂ retrievals: results
- from the quality assurance for the essential climate variables (QA4ECV) project, Atmos. Meas. Tech., 11, 66516678, https://doi.org/10.5194/amt-11-6651-2018, 2018.
- Breuer, H., Ács, F., Horváth, Á., Németh, P., and Rajkai, K.: Diurnal course analysis of the WRF-simulated and
 observation-based planetary boundary layer height, Advances in Science and Research, 11, 83-88,
 https://doi.org/10.5194/asr-11-83-2014, 2014.
- 916 Brohede, S., McLinden, C. A., Berthet, G., Haley, C. S., Murtagh, D., and Sioris, C. E.: A stratospheric NO₂
- 917 climatology from Odin/OSIRIS limb-scatter measurements, Can. J. Phys., 85, 1253-1274,
- 918 https://doi.org/10.1139/p07-141, 2007.

- 919 Brown, S. S., Dibb, J. E., Stark, H., Aldener, M., Vozella, M., Whitlow, S., Williams, E. J., Lerner, B. M.,
- Jakoubek, R., and Middlebrook, A. M.: Nighttime removal of NO_x in the summer marine boundary layer,
 Geophys. Res. Lett., 31, https://doi.org/10.1029/2004GL019412, 2004.
- Bucsela, E. J., Krotkov, N. A., Celarier, E. A., Lamsal, L. N., Swartz, W. H., Bhartia, P. K., Boersma, K. F.,
 Veefkind, J. P., Gleason, J. F., and Pickering, K. E.: A new stratospheric and tropospheric NO₂ retrieval
 algorithm for nadir-viewing satellite instruments: applications to OMI, Atmos. Meas. Tech., 6, 2607-2626,
 https://doi.org/10.5194/amt-6-2607-2013, 2013.
- Canty, T., Hembeck, L., Vinciguerra, T., Goldberg, D., Carpenter, S., Allen, D., Loughner, C., Salawitch, R., and
 Dickerson, R.: Ozone and NO_x chemistry in the eastern US: evaluation of CMAQ/CB05 with satellite (OMI)
 data, Atmos. Chem. Phys., 15, 10965, https://doi.org/10.5194/acp-15-10965-2015, 2015.
- 929 Chance, K.: OMI Algorithm Theoretical Basis Document: OMI Trace Gas Algorithms, available at
 930 https://ozoneaq.gsfc.nasa.gov/media/docs/ATBD-OMI-04.pdf, Smithsonian Astrophysical Observatory,
 931 Cambridge, MA, USA2.0, 78, 2002.
- Cheng, Y., Wang, Y., Zhang, Y., Chen, G., Crawford, J. H., Kleb, M. M., Diskin, G. S., and Weinheimer, A. J.:
 Large biogenic contribution to boundary layer O₃-CO regression slope in summer, Geophys. Res. Lett., 44, 70617068, https://doi.org/10.1002/2017GL074405, 2017.
- 935 Cheng, Y., Wang, Y., Zhang, Y., Crawford, J. H., Diskin, G. S., Weinheimer, A. J., and Fried, A.: Estimator of 936 surface ozone using formaldehyde and carbon monoxide concentrations over the eastern United States in
- surface ozone using formaldehyde and carbon monoxide concentrations over the eastern United States in
 summer, J. Geophys. Res.-Atmos., 123, 7642-7655, https://doi.org/10.1029/2018JD028452, 2018.
- Choi, S., Lamsal, L. N., Follette-Cook, M., Joiner, J., Krotkov, N. A., Swartz, W. H., Pickering, K. E., Loughner,
 C. P., Appel, W., Pfister, G., Saide, P. E., Cohen, R. C., Weinheimer, A. J., and Herman, J. R.: Assessment of
 NO₂ observations during DISCOVER-AQ and KORUS-AQ field campaigns, Atmos. Meas. Tech., 13, 25232546, https://doi.org/10.5194/amt-13-2523-2020, 2020.
- 942 Choi, Y., Wang, Y., Zeng, T., Cunnold, D., Yang, E. S., Martin, R., Chance, K., Thouret, V., and Edgerton, E.:
 943 Springtime transitions of NO₂, CO, and O₃ over North America: Model evaluation and analysis, J. Geophys.
 944 Res.-Atmos., 113, https://doi.org/10.1029/2007JD009632, 2008.
- Choi, Y., Kim, H., Tong, D., and Lee, P.: Summertime weekly cycles of observed and modeled NO_x and O₃
 concentrations as a function of satellite-derived ozone production sensitivity and land use types over the
 Continental United States, Atmos. Chem. Phys., 12, 6291-6307, https://doi.org/10.5194/acp-12-6291-2012, 2012.
- 948 Compton, J. C., Delgado, R., Berkoff, T. A., and Hoff, R. M.: Determination of planetary boundary layer height
 949 on short spatial and temporal scales: A demonstration of the covariance wavelet transform in ground-based wind
 950 profiler and lidar measurements, Journal of Atmospheric and Oceanic Technology, 30, 1566-1575,
 954 https://doi.org/10.1175/JTECULD.12.00116.1.2012
- 951 https://doi.org/10.1175/JTECH-D-12-00116.1, 2013.

David, L. M., and Nair, P. R.: Diurnal and seasonal variability of surface ozone and NO_x at a tropical coastal site:
 Association with mesoscale and synoptic meteorological conditions, J. Geophys. Res.-Atmos., 116,
 https://doi.org/10.1020/2010JD015076_2011

954 https://doi.org/10.1029/2010JD015076, 2011.

Davis, C., Brown, B., and Bullock, R.: Object-based verification of precipitation forecasts. Part I: Methodology
and application to mesoscale rain areas, Monthly Weather Review, 134, 1772-1784,
https://doi.org/10.1175/MWR3145.1, 2006.

Day, D. A., Wooldridge, P. J., Dillon, M. B., Thornton, J. A., and Cohen, R. C.: A thermal dissociation laser induced fluorescence instrument for in situ detection of NO₂, peroxy nitrates, alkyl nitrates, and HNO₃, J.
Geophys. Res.-Atmos., 107, ACH 4-1-ACH 4-14, https://doi.org/10.1029/2001JD000779, 2002.

de Foy, B.: City-level variations in NOx emissions derived from hourly monitoring data in Chicago, Atmos.
Environ., 176, 128-139, https://doi.org/10.1016/j.atmosenv.2017.12.028, 2018.

DenBleyker, A., Morris, R. E., Lindhjem, C. E., Parker, L. K., Shah, T., Koo, B., Loomis, C., and Dilly, J.:
Temporal and Spatial Detail in Mobile Source Emission Inventories for Regional Air Quality Modeling, 2012
International Emission Inventory Conference, Florida, U.S., August 13 - 16, 2012, 2012.

966 DenBleyker, A., Koupal, J., DeFries, T., and Palacios, C.: Improvement of Default Inputs for MOVES and

967 SMOKE-MOVES: CRC Project A-100, available at https://crcao.org/reports/recentstudies2017/A-

968 100/ERG_FinalReport_CRCA100_28Feb2017.pdf, Eastern Research Group, Inc., Austin, TX, 86, 2017.

Dirksen, R. J., Boersma, K. F., Eskes, H. J., Ionov, D. V., Bucsela, E. J., Levelt, P. F., and Kelder, H. M.:

Evaluation of stratospheric NO₂ retrieved from the Ozone Monitoring Instrument: Intercomparison, diurnal cycle,
and trending, J. Geophys. Res.-Atmos., 116, https://doi.org/10.1029/2010JD014943, 2011.

972 EPA: Profile of the 2011 National Air Emissions Inventory, available at

https://www.epa.gov/sites/production/files/2015-08/documents/lite_finalversion_ver10.pdf, U.S. Environmental
 Protection Agency, 2014.

Fisher, J. A., Jacob, D. J., Travis, K. R., Kim, P. S., Marais, E. A., Chan Miller, C., Yu, K., Zhu, L., Yantosca, R.
M., and Sulprizio, M. P.: Organic nitrate chemistry and its implications for nitrogen budgets in an isoprene-and
monoterpene-rich atmosphere: constraints from aircraft (SEAC⁴RS) and ground-based (SOAS) observations in
the Southeast US, Atmos. Chem. Phys., 16, 5969-5991, https://doi.org/10.5194/acp-16-5969-2016, 2016.

Flynn, C. M., Pickering, K. E., Crawford, J. H., Lamsal, L., Krotkov, N., Herman, J., Weinheimer, A., Chen, G.,
Liu, X., and Szykman, J.: Relationship between column-density and surface mixing ratio: Statistical analysis of
O₃ and NO₂ data from the July 2011 Maryland DISCOVER-AQ mission, Atmos. Environ., 92, 429-441,
https://doi.org/10.1016/j.atmosenv.2014.041, 2014.

Frey, M. M., Brough, N., France, J. L., Anderson, P. S., Traulle, O., King, M. D., Jones, A. E., Wolff, E. W., and
Savarino, J.: The diurnal variability of atmospheric nitrogen oxides (NO and NO₂) above the Antarctic Plateau

driven by atmospheric stability and snow emissions, Atmos. Chem. Phys., 13, 3045-3062,

986 https://doi.org/10.5194/acp-13-3045-2013, 2013.

- Gaur, A., Tripathi, S. N., Kanawade, V. P., Tare, V., and Shukla, S. P.: Four-year measurements of trace gases
 (SO₂, NO_x, CO, and O₃) at an urban location, Kanpur, in Northern India, Journal of Atmospheric Chemistry, 71,
 283-301, https://doi.org/10.1007/s10874-014-9295-8, 2014.
- Gourley, J. J., Hong, Y., Flamig, Z. L., Wang, J., Vergara, H., and Anagnostou, E. N.: Hydrologic evaluation of
 rainfall estimates from radar, satellite, gauge, and combinations on Ft. Cobb basin, Oklahoma, Journal of
 Hydrometeorology, 12, 973-988, https://doi.org/10.1175/2011JHM1287.1, 2011.
- Guenther, A. B., Jiang, X., Heald, C. L., Sakulyanontvittaya, T., Duhl, T., Emmons, L. K., and Wang, X.: The
 Model of Emissions of Gases and Aerosols from Nature version 2.1 (MEGAN2.1): an extended and updated
 framework for modeling biogenic emissions, Geosci. Model Dev., 5, 1471-1492, https://doi.org/10.5194/gmd-51471-2012, 2012.

Hains, J. C., Boersma, K. F., Kroon, M., Dirksen, R. J., Cohen, R. C., Perring, A. E., Bucsela, E., Volten, H.,
Swart, D. P. J., and Richter, A.: Testing and improving OMI DOMINO tropospheric NO₂ using observations
from the DANDELIONS and INTEX - B validation campaigns, J. Geophys. Res.-Atmos., 115,
https://doi.org/10.1029/2009JD012399, 2010.

Herman, J., Cede, A., Spinei, E., Mount, G., Tzortziou, M., and Abuhassan, N.: NO₂ column amounts from
ground-based Pandora and MFDOAS spectrometers using the direct-Sun DOAS technique: Intercomparisons and
application to OMI validation, J. Geophys. Res.-Atmos., 114, https://doi.org/10.1029/2009JD011848, 2009.

Herman, J., Spinei, E., Fried, A., Kim, J., Kim, J., Kim, W., Cede, A., Abuhassan, N., and Segal-Rozenhaimer,
M.: NO₂ and HCHO measurements in Korea from 2012 to 2016 from Pandora spectrometer instruments
compared with OMI retrievals and with aircraft measurements during the KORUS-AQ campaign, Atmos. Meas.
Tech., 11, 4583-4603, https://doi.org/10.5194/amt-11-4583-2018, 2018.

Herman, J., Abuhassan, N., Kim, J., Kim, J., Dubey, M., Raponi, M., and Tzortziou, M.: Underestimation of
column NO₂ amounts from the OMI satellite compared to diurnally varying ground-based retrievals from
multiple PANDORA spectrometer instruments, Atmos. Meas. Tech., 12, 5593-5612, https://doi.org/10.5194/amt12-5593-2019, 2019.

Hong, S.-Y., Noh, Y., and Dudhia, J.: A new vertical diffusion package with an explicit treatment of entrainment processes, Monthly weather review, 134, 2318-2341, https://doi.org/10.1175/MWR3199.1, 2006.

Hu, X., Doughty, D. C., Sanchez, K. J., Joseph, E., and Fuentes, J. D.: Ozone variability in the atmospheric
boundary layer in Maryland and its implications for vertical transport model, Atmos. Environ., 46, 354-364,
https://doi.org/10.1016/j.atmosenv.2011.09.054, 2012.

Huijnen, V., Eskes, H. J., Poupkou, A., Elbern, H., Boersma, K. F., Foret, G., Sofiev, M., Valdebenito, A.,
Flemming, J., and Stein, O.: Comparison of OMI NO₂ tropospheric columns with an ensemble of global and
European regional air quality models, Atmos. Chem. Phys., 10, 3273-3296, https://doi.org/10.5194/acp-10-32732010, 2010.

Ionov, D. V., Timofeyev, Y. M., Sinyakov, V. P., Semenov, V. K., Goutail, F., Pommereau, J. P., Bucsela, E. J.,
Celarier, E. A., and Kroon, M.: Ground - based validation of EOS - Aura OMI NO₂ vertical column data in the
midlatitude mountain ranges of Tien Shan (Kyrgyzstan) and Alps (France), J. Geophys. Res.-Atmos., 113,
https://doi.org/10.1029/2007JD008659, 2008.

Irie, H., Kanaya, Y., Akimoto, H., Tanimoto, H., Wang, Z., Gleason, J. F., and Bucsela, E. J.: Validation of OMI
tropospheric NO₂ column data using MAX-DOAS measurements deep inside the North China Plain in June
2006: Mount Tai Experiment 2006, Atmos. Chem. Phys., 8, 6577-6586, https://doi.org/10.5194/acp-8-6577-2008,
2008.

Irie, H., Boersma, K. F., Kanaya, Y., Takashima, H., Pan, X., and Wang, Z.: Quantitative bias estimates for
 tropospheric NO₂ columns retrieved from SCIAMACHY, OMI, and GOME-2 using a common standard for East
 Asia, Atmos. Meas. Tech., 5, 2403-2411, https://doi.org/10.5194/amt-5-2403-2012, 2012.

Jones, A. E., Weller, R., Wolff, E. W., and Jacobi, H. W.: Speciation and rate of photochemical NO and NO₂
 production in Antarctic snow, Geophys. Res. Lett., 27, 345-348, https://doi.org/10.1029/1999GL010885, 2000.

Judd, L. M., Al-Saadi, J. A., Valin, L. C., Pierce, R. B., Yang, K., Janz, S. J., Kowalewski, M. G., Szykman, J. J.,
Tiefengraber, M., and Mueller, M.: The Dawn of Geostationary Air Quality Monitoring: Case Studies from Seoul
and Los Angeles, Front. Environ. Sci., 6, 85, https://doi.org/10.3389/fenvs.2018.00085, 2018.

Judd, L. M., Al-Saadi, J. A., Janz, S. J., Kowalewski, M. G., Pierce, R. B., Szykman, J. J., Valin, L. C., Swap, R.,
Cede, A., Mueller, M., Tiefengraber, M., Abuhassan, N., and Williams, D.: Evaluating the impact of spatial
resolution on tropospheric NO₂ column comparisons within urban areas using high-resolution airborne data,
Atmos. Meas. Tech., 12, 6091-6111, https://doi.org/10.5194/amt-12-6091-2019, 2019.

Judd, L. M., Al-Saadi, J. A., Szykman, J. J., Valin, L. C., Janz, S. J., Kowalewski, M. G., Eskes, H. J., Veefkind,
J. P., Cede, A., Mueller, M., Gebetsberger, M., Swap, R., Pierce, R. B., Nowlan, C. R., Abad, G. G., Nehrir, A.,
and Williams, D.: Evaluating Sentinel-5P TROPOMI tropospheric NO₂ column densities with airborne and
Pandora spectrometers near New York City and Long Island Sound, Atmos. Meas. Tech. Discuss., 2020, 1-52,
https://doi.org/10.5194/amt-2020-151, 2020.

Kalinga, O. A., and Gan, T. Y.: Estimation of rainfall from infrared - microwave satellite data for basin - scale
hydrologic modelling, Hydrological processes, 24, 2068-2086, https://doi.org/10.1002/hyp.7626, 2010.

Kaynak, B., Hu, Y., Martin, R. V., Sioris, C. E., and Russell, A. G.: Comparison of weekly cycle of NO₂ satellite
 retrievals and NO_x emission inventories for the continental United States, J. Geophys. Res.-Atmos., 114,
 https://doi.org/10.1029/2008JD010714, 2009.

Kim, S. W., McDonald, B., Baidar, S., Brown, S., Dube, B., Ferrare, R., Frost, G., Harley, R., Holloway, J., and
Lee, H. J.: Modeling the weekly cycle of NO_x and CO emissions and their impacts on O₃ in the Los AngelesSouth Coast Air Basin during the CalNex 2010 field campaign, J. Geophys. Res.-Atmos., 121, 1340-1360,
https://doi.org/10.1002/2015JD024292, 2016.

Knepp, T., Pippin, M., Crawford, J., Chen, G., Szykman, J., Long, R., Cowen, L., Cede, A., Abuhassan, N., and
Herman, J.: Estimating surface NO₂ and SO₂ mixing ratios from fast-response total column observations and
potential application to geostationary missions, Journal of atmospheric chemistry, 72, 261-286,
https://doi.org/10.1007/s10874-013-9257-6, 2015.

Kollonige, D. E., Thompson, A. M., Josipovic, M., Tzortziou, M., Beukes, J. P., Burger, R., Martins, D. K., van
 Zyl, P. G., Vakkari, V., and Laakso, L.: OMI Satellite and Ground - Based Pandora Observations and Their
 Application to Surface NO₂ Estimations at Terrestrial and Marine Sites, J. Geophys. Res.-Atmos., 123, 1441 1459, https://doi.org/10.1002/2017JD026518, 2018.

Krotkov, N. A., Lamsal, L. N., Celarier, E. A., Swartz, W. H., Marchenko, S. V., Bucsela, E. J., Chan, K. L.,
Wenig, M., and Zara, M.: The version 3 OMI NO₂ standard product, Atmos. Meas. Tech., 10, 3133-3149,
https://doi.org/10.5194/amt-10-3133-2017, 2017.

Lamsal, L. N., Krotkov, N. A., Celarier, E. A., Swartz, W. H., Pickering, K. E., Bucsela, E. J., Gleason, J. F.,
Martin, R. V., Philip, S., and Irie, H.: Evaluation of OMI operational standard NO₂ column retrievals using in situ
and surface-based NO₂ observations, Atmos. Chem. Phys., 14, 11587-11609, https://doi.org/10.5194/acp-1411587-2014, 2014.

Lamsal, L. N., Duncan, B. N., Yoshida, Y., Krotkov, N. A., Pickering, K. E., Streets, D. G., and Lu, Z.: US NO₂
trends (2005–2013): EPA Air Quality System (AQS) data versus improved observations from the Ozone
Monitoring Instrument (OMI), Atmos. Environ., 110, 130-143, https://doi.org/10.1016/j.atmosenv.2015.03.055,
2015.

Lamsal, L. N., Janz, S. J., Krotkov, N. A., Pickering, K. E., Spurr, R. J. D., Kowalewski, M. G., Loughner, C. P.,
Crawford, J. H., Swartz, W. H., and Herman, J.: High - resolution NO₂ observations from the Airborne Compact
Atmospheric Mapper: Retrieval and validation, J. Geophys. Res.-Atmos., 122, 1953-1970,
https://doi.org/10.1002/2016JD025483, 2017.

Lamsal, L. N., Krotkov, N. A., Vasilkov, A., Marchenko, S., Qin, W., Yang, E. S., Fasnacht, Z., Joiner, J., Choi,
S., Haffner, D., Swartz, W. H., Fisher, B., and Bucsela, E.: OMI/Aura Nitrogen Dioxide Standard Product with
Improved Surface and Cloud Treatments, Atmos. Meas. Tech. Discuss., 2020, 1-56, https://doi.org/10.5194/amt2020-200, 2020.

Levelt, P. F., Hilsenrath, E., Leppelmeier, G. W., van den Oord, G. H. J., Bhartia, P. K., Tamminen, J., de Haan,
J. F., and Veefkind, J. P.: Science objectives of the ozone monitoring instrument, IEEE Transactions on
Geoscience and Remote Sensing, 44, 1199-1208, https://doi.org/10.1109/TGRS.2006.872336, 2006.

Li, J., Wang, Y., and Qu, H.: Dependence of summertime surface ozone on NO_x and VOC emissions over the
United States: Peak time and value, Geophys. Res. Lett., 46, 3540-3550, https://doi.org/10.1029/2018GL081823,
2019.

Li, J., Feng, Z., Qian, Y., and Leung, L. R.: A high-resolution unified observational data product of mesoscale
 convective systems and isolated deep convection in the United States for 2004–2017, Earth Syst. Sci. Data
 Discuss., 2020, 1-48, https://doi.org/10.5194/essd-2020-151, 2020.

Lin, Y., and Mitchell, K. E.: the NCEP stage II/IV hourly precipitation analyses: Development and applications,
19th Conf. Hydrology, American Meteorological Society, San Diego, CA, USA, 2005,

Liu, C., Liu, X., Kowalewski, M., Janz, S., González Abad, G., Pickering, K., Chance, K., and Lamsal, L.:
 Analysis of ACAM data for trace gas retrievals during the 2011 DISCOVER-AQ campaign, Journal of
 Spectroscopy, 2015, https://doi.org/10.1155/2015/827160, 2015a.

Liu, C., Liu, X., Kowalewski, M. G., Janz, S. J., González Abad, G., Pickering, K. E., Chance, K., and Lamsal, L.
N.: Characterization and verification of ACAM slit functions for trace-gas retrievals during the 2011
DISCOVER-AQ flight campaign, Atmos. Meas. Tech., 8, 751-759, https://doi.org/10.5194/amt-8-751-2015,
2015b.

Liu, Z., Wang, Y., Gu, D., Zhao, C., Huey, L. G., Stickel, R., Liao, J., Shao, M., Zhu, T., and Zeng, L.:
Summertime photochemistry during CAREBeijing-2007: RO_x budgets and O₃ formation, Atmos. Chem. Phys.,
12, 7737-7752, https://doi.org/10.5194/acp-12-7737-2012, 2012.

Lopez, P.: Direct 4D-Var assimilation of NCEP stage IV radar and gauge precipitation data at ECMWF, Monthly
Weather Review, 139, 2098-2116, https://doi.org/10.1175/2010MWR3565.1, 2011.

1105 Luo, G., Yu, F., and Schwab, J.: Revised treatment of wet scavenging processes dramatically improves GEOS-

Chem 12.0.0 simulations of surface nitric acid, nitrate, and ammonium over the United States, Geosci. Model
 Dev., 12, 3439-3447, https://doi.org/10.5194/gmd-12-3439-2019, 2019.

Marchenko, S., Krotkov, N., Lamsal, L., Celarier, E., Swartz, W., and Bucsela, E.: Revising the slant column
density retrieval of nitrogen dioxide observed by the Ozone Monitoring Instrument, J. Geophys. Res.-Atmos.,
120, 5670-5692, https://doi.org/10.1002/2014JD022913, 2015.

Marr, L. C., Moore, T. O., Klapmeyer, M. E., and Killar, M. B.: Comparison of NO_x Fluxes Measured by Eddy
Covariance to Emission Inventories and Land Use, Environ. Sci. Technol., 47, 1800-1808,
https://doi.org/10.1021/es303150y, 2013.

McDonald, B., McKeen, S., Cui, Y. Y., Ahmadov, R., Kim, S.-W., Frost, G. J., Pollack, I., Peischl, J., Ryerson,
T. B., and Holloway, J.: Modeling Ozone in the Eastern US using a Fuel-Based Mobile Source Emissions
Inventory, Environ. Sci. Technol., https://doi.org/10.1021/acs.est.8b00778, 2018.

Munro, R., Eisinger, M., Anderson, C., Callies, J., Corpaccioli, E., Lang, R., Lefebvre, A., Livschitz, Y., and
Albinana, A. P.: GOME-2 on MetOp, Proc. of The 2006 EUMETSAT Meteorological Satellite Conference,
Helsinki, Finland, 2006, 48,

1120 Nelson, B. R., Prat, O. P., Seo, D.-J., and Habib, E.: Assessment and implications of NCEP Stage IV quantitative

1121 precipitation estimates for product intercomparisons, Weather and Forecasting, 31, 371-394,

1122 https://doi.org/10.1175/WAF-D-14-00112.1, 2016.

Ng, N. L., Brown, S. S., Archibald, A. T., Atlas, E., Cohen, R. C., Crowley, J. N., Day, D. A., Donahue, N. M.,
Fry, J. L., and Fuchs, H.: Nitrate radicals and biogenic volatile organic compounds: oxidation, mechanisms, and
organic aerosol, Atmos. Chem. Phys., 17, 2103-2162, https://doi.org/10.5194/acp-17-2103-2017, 2017.

Nowlan, C. R., Liu, X., Leitch, J. W., Chance, K., González Abad, G., Liu, C., Zoogman, P., Cole, J., Delker, T.,
Good, W., Murcray, F., Ruppert, L., Soo, D., Follette-Cook, M. B., Janz, S. J., Kowalewski, M. G., Loughner, C.
P., Pickering, K. E., Herman, J. R., Beaver, M. R., Long, R. W., Szykman, J. J., Judd, L. M., Kelley, P., Luke, W.
T., Ren, X., and Al-Saadi, J. A.: Nitrogen dioxide observations from the Geostationary Trace gas and Aerosol
Sensor Optimization (GeoTASO) airborne instrument: Retrieval algorithm and measurements during
DISCOVER-AQ Texas 2013, Atmos. Meas. Tech., 9, 2647-2668, https://doi.org/10.5194/amt-9-2647-2016,

1132 2016.

Nowlan, C. R., Liu, X., Janz, S. J., Kowalewski, M. G., Chance, K., Follette-Cook, M. B., Fried, A., González
Abad, G., Herman, J. R., Judd, L. M., Kwon, H. A., Loughner, C. P., Pickering, K. E., Richter, D., Spinei, E.,
Walega, J., Weibring, P., and Weinheimer, A. J.: Nitrogen dioxide and formaldehyde measurements from the
GEOstationary Coastal and Air Pollution Events (GEO-CAPE) Airborne Simulator over Houston, Texas, Atmos.
Meas. Tech., 11, 5941-5964, https://doi.org/10.5194/amt-11-5941-2018, 2018.

Oetjen, H., Baidar, S., Krotkov, N. A., Lamsal, L. N., Lechner, M., and Volkamer, R.: Airborne MAX-DOAS
measurements over California: Testing the NASA OMI tropospheric NO₂ product, J. Geophys. Res.-Atmos., 118,
7400-7413, https://doi.org/10.1002/jgrd.50550, 2013.

Peng, J., Hu, M., Guo, S., Du, Z., Zheng, J., Shang, D., Zamora, M. L., Zeng, L., Shao, M., and Wu, Y.-S.:
Markedly enhanced absorption and direct radiative forcing of black carbon under polluted urban environments,

1143 Proc. Natl. Acad. Sci. U.S.A., 201602310, https://doi.org/10.1073/pnas.1602310113, 2016.

Peters, E., Wittrock, F., Großmann, K., Frieß, U., Richter, A., and Burrows, J. P.: Formaldehyde and nitrogen
dioxide over the remote western Pacific Ocean: SCIAMACHY and GOME-2 validation using ship-based MAXDOAS observations, Atmos. Chem. Phys., 12, 11179-11197, https://doi.org/10.5194/acp-12-11179-2012, 2012.

Reddy, B. S. K., Kumar, K. R., Balakrishnaiah, G., Gopal, K. R., Reddy, R. R., Sivakumar, V., Lingaswamy, A.
P., Arafath, S. M., Umadevi, K., and Kumari, S. P.: Analysis of diurnal and seasonal behavior of surface ozone and its precursors (NO_x) at a semi-arid rural site in Southern India, Aerosol Air Oual Res, 12, 1081-1094,

1150 https://doi.org/10.4209/aaqr.2012.03.0055 2012.

Reed, A. J., Thompson, A. M., Kollonige, D. E., Martins, D. K., Tzortziou, M. A., Herman, J. R., Berkoff, T. A.,
Abuhassan, N. K., and Cede, A.: Effects of local meteorology and aerosols on ozone and nitrogen dioxide
retrievals from OMI and pandora spectrometers in Maryland, USA during DISCOVER-AQ 2011, Journal of
atmospheric chemistry, 72, 455-482, https://doi.org/10.1007/s10874-013-9254-9, 2015.

1155 Reed, C., Evans, M. J., Carlo, P. D., Lee, J. D., and Carpenter, L. J.: Interferences in photolytic NO₂

1156 measurements: explanation for an apparent missing oxidant?, Atmos. Chem. Phys., 16, 4707-4724, 1157 https://doi.org/10.5104/gep.16.4707.2016.2016

1157 https://doi.org/10.5194/acp-16-4707-2016, 2016.

- Richter, A., Begoin, M., Hilboll, A., and Burrows, J. P.: An improved NO₂ retrieval for the GOME-2 satellite
 instrument, Atmos. Meas. Tech., 4, 1147-1159, https://doi.org/10.5194/amt-4-1147-2011, 2011.
- Russell, A. R., Valin, L. C., and Cohen, R. C.: Trends in OMI NO₂ observations over the United States: effects of
 emission control technology and the economic recession, Atmos. Chem. Phys., 12, 12197-12209,
- 1162 https://doi.org/10.5194/acp-12-12197-2012, 2012.
- 1163 Saha, S., Moorthi, S., Wu, X., Wang, J., Nadiga, S., Tripp, P., Behringer, D., Hou, Y. T., Chuang, H.-y., and
- 1164 Iredell, M.: NCEP climate forecast system version 2 (CFSv2) 6-hourly products, available at
- 1165 https://rda.ucar.edu/datasets/ds094.0/, https://doi.org/10.5065/D61C1TXF, 2011 (last access: Mar 10, 2015).
- 1166 Salmon, O., Shepson, P., Ren, X., He, H., Hall, D., Dickerson, R., Stirm, B., Brown, S., Fibiger, D., and
- 1167 McDuffie, E.: Top down Estimates of NO_{*} and CO Emissions from Washington, DC Baltimore During the
- 1168 WINTER Campaign, J. Geophys. Res. Atmos., https://doi.org/10.1029/2018JD028539, 2018.
- 1169 Sawamura, P., Müller, D., Hoff, R. M., Hostetler, C. A., Ferrare, R. A., Hair, J. W., Rogers, R. R., Anderson, B.
- 1170 E., Ziemba, L. D., and Beyersdorf, A. J.: Aerosol optical and microphysical retrievals from a hybrid
- 1171 multiwavelength lidar data set–DISCOVER-AQ 2011, Atmos. Meas. Tech., 7, 3095-3112,
- 1172 https://doi.org/10.5194/amt-7-3095-2014, 2014.
- Seinfeld, J. H., and Pandis, S. N.: Atmospheric chemistry and physics: from air pollution to climate change, John
 Wiley & Sons, Inc, Hoboken, New Jersey, 2016.
- 1175 Sen, B., Toon, G. C., Osterman, G. B., Blavier, J.-F., Margitan, J. J., Salawitch, R. J., and Yue, G. K.:
- Measurements of reactive nitrogen in the stratosphere, J. Geophys. Res.-Atmos., 103, 3571-3585,
 https://doi.org/10.1029/97JD02468, 1998.
- Shin, H. H., and Hong, S.-Y.: Intercomparison of planetary boundary-layer parametrizations in the WRF model
 for a single day from CASES-99, Boundary-Layer Meteorology, 139, 261-281, https://doi.org/10.1007/s10546010-9583-z, 2011.
- Silvern, R. F., Jacob, D. J., Mickley, L. J., Sulprizio, M. P., Travis, K. R., Marais, E. A., Cohen, R. C., Laughner,
 J. L., Choi, S., Joiner, J., and Lamsal, L. N.: Using satellite observations of tropospheric NO₂ columns to infer
- 1183 long-term trends in US NO_x emissions: the importance of accounting for the free tropospheric NO₂ background,
- 1184 Atmos. Chem. Phys., 19, 8863-8878, https://doi.org/10.5194/acp-19-8863-2019, 2019.
- Souri, A. H., Choi, Y., Jeon, W., Li, X., Pan, S., Diao, L., and Westenbarger, D. A.: Constraining NO_x emissions
 using satellite NO₂ measurements during 2013 DISCOVER-AQ Texas campaign, Atmos. Environ., 131, 371381, https://doi.org/10.1016/j.atmosenv.2016.02.020, 2016.
- 1188 Souri, A. H., Choi, Y., Pan, S., Curci, G., Nowlan, C. R., Janz, S. J., Kowalewski, M. G., Liu, J., Herman, J. R.,
- and Weinheimer, A. J.: First top down estimates of anthropogenic NO_x emissions using high resolution
- airborne remote sensing observations, J. Geophys. Res.-Atmos., 123, 3269-3284,
- 1191 https://doi.org/10.1002/2017JD028009, 2018.

Spinei, E., Cede, A., Swartz, W. H., Herman, J., and Mount, G. H.: The use of NO₂ absorption cross section
temperature sensitivity to derive NO₂ profile temperature and stratospheric–tropospheric column partitioning
from visible direct-sun DOAS measurements, Atmos. Meas. Tech., 7, 4299-4316, https://doi.org/10.5194/amt-74299-2014, 2014.

Spurr, R.: LIDORT and VLIDORT: Linearized pseudo-spherical scalar and vector discrete ordinate radiative
transfer models for use in remote sensing retrieval problems, in: Light Scattering Reviews 3, Springer, 229-275,
2008.

Thompson, A. M., Stauffer, R. M., Boyle, T. P., Kollonige, D. E., Miyazaki, K., Tzortziou, M., Herman, J. R.,
 Abuhassan, N., Jordan, C. E., and Lamb, B. T.: Comparison of Near - Surface NO₂ Pollution With Pandora Total
 Column NO₂ During the Korea - United States Ocean Color (KORUS OC) Campaign, J. Geophys. Res.-Atmos.,

1202 <u>124, 13560-13575, https://doi.org/10.1029/2019JD030765, 2019.</u>

Thornton, J. A., Wooldridge, P. J., and Cohen, R. C.: Atmospheric NO₂: In situ laser-induced fluorescence
detection at parts per trillion mixing ratios, Anal. Chem., 72, 528-539, https://doi.org/10.1021/ac9908905, 2000.

1205 Tong, D., Lamsal, L., Pan, L., Ding, C., Kim, H., Lee, P., Chai, T., Pickering, K. E., and Stajner, I.: Long-term

1206 NO_x trends over large cities in the United States during the great recession: Comparison of satellite retrievals,

1207 ground observations, and emission inventories, Atmos. Environ., 107, 70-84,

1208 https://doi.org/10.1016/j.atmosenv.2015.01.035, 2015.

1209 Travis, K. R., Jacob, D. J., Fisher, J. A., Kim, P. S., Marais, E. A., Zhu, L., Yu, K., Miller, C. C., Yantosca, R.

M., and Sulprizio, M. P.: Why do models overestimate surface ozone in the Southeast United States?, Atmos.
Chem. Phys., 16, 13561-13577, https://doi.org/10.5194/acp-16-13561-2016, 2016.

1212 Tu, J., Xia, Z.-G., Wang, H., and Li, W.: Temporal variations in surface ozone and its precursors and

1213 meteorological effects at an urban site in China, Atmospheric Research, 85, 310-337,

1214 https://doi.org/10.1016/j.atmosres.2007.02.003, 2007.

Valin, L. C., Russell, A. R., Hudman, R. C., and Cohen, R. C.: Effects of model resolution on the interpretation
of satellite NO₂ observations, Atmos. Chem. Phys., 11, 11647-11655, https://doi.org/10.5194/acp-11-116472011, 2011.

van der A, R. J., Eskes, H. J., Roozendael, M. V., De Smedt, I., Blond, N., Boersma, F., Weiss, A., and van Peet,
J. C. A.: Algorithm Document Tropospheric NO₂, available at http://www.temis.nl/docs/AD_NO2.pdf1.0, 23,
2010.

Van Geffen, J., Boersma, K., Van Roozendael, M., Hendrick, F., Mahieu, E., De Smedt, I., Sneep, M., and
Veefkind, J.: Improved spectral fitting of nitrogen dioxide from OMI in the 405–465 nm window, Atmos. Meas.
Tech., 8, 1685-1699, https://doi.org/10.5194/amt-8-1685-2015, 2015.

Van Stratum, B. J. H., Vilà-Guerau de Arellano, J., Ouwersloot, H. G., Dries, K. d., Van Laar, T. W., Martinez,
M., Lelieveld, J., Diesch, J.-M., Drewnick, F., and Fischer, H.: Case study of the diurnal variability of chemically

active species with respect to boundary layer dynamics during DOMINO, Atmos. Chem. Phys., 12, 5329-5341,
 https://doi.org/10.5194/acp-12-5329-2012, 2012.

Wooldridge, P. J., Perring, A. E., Bertram, T. H., Flocke, F. M., Roberts, J. M., Singh, H. B., Huey, L. G.,
Thornton, J. A., Wolfe, G. M., and Murphy, J. G.: Total Peroxy Nitrates ([Sigma] PNs) in the atmosphere: the
Thermal Dissociation-Laser Induced Fluorescence (TD-LIF) technique and comparisons to speciated PAN
measurements, Atmos. Meas. Tech., 3, 593, https://doi.org/10.5194/amt-3-593-2010, 2010.

Yuan, H., McGinley, J. A., Schultz, P. J., Anderson, C. J., and Lu, C.: Short-range precipitation forecasts from
time-lagged multimodel ensembles during the HMT-West-2006 campaign, Journal of Hydrometeorology, 9, 477491, https://doi.org/10.1175/2007JHM879.1, 2008.

Zhang, R., Wang, Y., Smeltzer, C., Qu, H., Koshak, W., and Boersma, K. F.: Comparing OMI-based and EPA
 AQS in situ NO₂ trends: towards understanding surface NO_x emission changes, Atmos. Meas. Tech., 11, 3955 3967, https://doi.org/10.5194/amt-11-3955-2018, 2018.

Zhang, Y., and Wang, Y.: Climate-driven ground-level ozone extreme in the fall over the Southeast United
States, Proc. Natl. Acad. Sci. U.S.A., 113, 10025-10030, https://doi.org/10.1073/pnas.1602563113, 2016.

1240 Zhang, Y., Wang, Y., Chen, G., Smeltzer, C., Crawford, J., Olson, J., Szykman, J., Weinheimer, A. J., Knapp, D.

1241 J., and Montzka, D. D.: Large vertical gradient of reactive nitrogen oxides in the boundary layer: Modeling

analysis of DISCOVER - AQ 2011 observations, J. Geophys. Res.-Atmos., 121, 1922-1934,

1243 https://doi.org/10.1002/2015JD024203, 2016.

Zhao, C., Wang, Y., Choi, Y., and Zeng, T.: Summertime impact of convective transport and lightning NO_x
 production over North America: modeling dependence on meteorological simulations, Atmos. Chem. Phys., 9,
 4315-4327, https://doi.org/10.5194/acp-9-4315-2009, 2009.

Zhao, X., Griffin, D., Fioletov, V., McLinden, C., Davies, J., Ogyu, A., Lee, S. C., Lupu, A., Moran, M. D.,
 Cede, A., Tiefengraber, M., and Müller, M.: Retrieval of total column and surface NO₂ from Pandora zenith-sky
 measurements, Atmos. Chem. Phys., 19, 10619-10642, https://doi.org/10.5194/acp-19-10619-2019, 2019.

1250 Zhao, X., Griffin, D., Fioletov, V., McLinden, C., Cede, A., Tiefengraber, M., Müller, M., Bognar, K., Strong,

1251 K., Boersma, F., Eskes, H., Davies, J., Ogyu, A., and Lee, S. C.: Assessment of the quality of TROPOMI high-

spatial-resolution NO₂ data products in the Greater Toronto Area, Atmos. Meas. Tech., 13, 2131-2159,
 https://doi.org/10.5194/amt-13-2131-2020, 2020.

Zheng, Y., Alapaty, K., Herwehe, J. A., Del Genio, A. D., and Niyogi, D.: Improving high-resolution weather
 forecasts using the Weather Research and Forecasting (WRF) Model with an updated Kain–Fritsch scheme,

1256 Monthly Weather Review, 144, 833-860, https://doi.org/10.1175/MWR-D-15-0005.1, 2016.

Table 1. Comparison of the concentrations of NO_v and its components between REAM and P-3B aircraft measurements during the 1257 1258 **DISCOVER-AO** campaign

			NO_y / ppb^1	NO / ppb	NO2_NCAR / ppb	NO2_LIF / ppb ²	$\sum PNs / ppb$	∑ANs / ppb	HNO ₃ / ppb	Derived- NO _y / ppb ²
	Weekday ⁵	P-3B	2.51 ± 2.09	0.18 ± 0.29	0.85 ± 1.13	0.68 ± 0.95	0.70 ± 0.58	0.31 ± 0.23	1.15 ± 0.73	2.86 ± 2.26
		REAM	3.64 ±	$0.18 \pm$	0.74 ± 1.0 <u>4</u> 6	$0.68 \pm$	0.5 <u>4</u> 3 ±	0.10 ± 0.09	$1.80 \pm$	3.10 ± 2.7
			3.1 <u>3</u> 5	0.3 <mark>0</mark> 1		0. <u>89</u> 90	0.45		1.6 <u>1</u> 2	
		\mathbb{R}^2	0.33	0.3 <u>5</u> 3	0.3 <u>8</u> 7	0.3 <u>4</u> 3	0.37	0.38	0.24	0.41
36-km ⁴	Weekend	D 2D	3.0 <mark>0</mark> 1 ±	0.15 ± 0.20	0.71 ± 0.80	0.63 ± 0.72	0.91 ± 0.53	0.36 ± 0.21	1.15 ± 0.79	2.96 ± 2.1
		P-3B	2.1 <u>8</u> 9							
		nd REAM	3.7 <u>8</u> 6 ±	0.15 ± 0.17	0.5 <u>4</u> 3 ±	$0.53 \pm$	0.5 <u>3</u> 2 ±	0.09 ± 0.06	2.3 <u>1</u> 0 ±	3.4 <u>3</u> 1 ±
		KEAM	2.2 <mark>0</mark> 4		0. <u>59</u> 61	0. <u>58</u> 60	0.29		1. <u>38</u> 40	2. <u>26</u> 30
		\mathbb{R}^2	0.29	0.28	0.41	0.45	0.27	0.3 <mark>98</mark>	0. <u>50</u> 4 9	0.51
	Weekday	P-3B	2.51 ± 2.15	0.19 ± 0.30	0.8 <u>6</u> 5 ± 1.2 <u>7</u> 9	0.6 <u>8</u> 7 ±	0.70 ± 0.59	0.31 ± 0.22	1.17 ± 0.74	2.90 ± 2.2
						0.9 <mark>86</mark>				
		REAM	3. <u>81</u> 67 ±	0.1 <u>9</u> 7 ±	0.7 <mark>9</mark> 4 ±	0.7 <u>6</u> 2 ±	0.4 <u>6</u> 5 ±	0.08 ± 0.10	<u>2.03</u> 1.99 ±	3. <u>31</u> 23 ±
		KEAM	3. <u>81</u> 63	0. <u>35</u> 29	1. <u>31</u> 12	1. <mark>2</mark> 40	0.51		1.9 <u>1</u> 2	3.2 <mark>8</mark> 4
4.1.00		\mathbb{R}^2	0.2 <mark>85</mark>	0. <u>22</u> 4 6	0. <u>26</u> 48	0. <u>32</u> 55	0.37	0.2 <mark>97</mark>	0. <u>38</u> 60	0.4 <u>7</u> 4
4-km	Weekend	P-3B	2.9 <mark>6</mark> 7 ±	0.1 <u>4</u> 5 ±	0.69 ± 0.7	0.6 <u>3</u> 5 ±	0.9 <u>1</u> 0 ±	0.35 ± 0.21	1.15 ±	2.9 <mark>43</mark> ±
		г-эв	2.13	0.18		0. <u>71</u> 85	0.51		0. <u>80</u> 79	2.0 <u>9</u> 8
		nd DEAM	4. <u>3622-</u> ±	0.2 <u>5</u> 4 ±	0.8 <u>5</u> 3 ±	0. <u>81</u> 78 ±	0. <u>41</u> 38 ±	0.0 <u>8</u> 7 ±	2. <u>54</u> 48 ±	3. <u>72</u> 59 ±
		REAM	3. <u>66</u> 79	0.4 <mark>02</mark>	1. <u>28</u> 40	1. <u>23</u> 34	0.2 <mark>9</mark> 7	0.08	<u>1.99</u> 2.04	3.5 <u>2</u> 9
		\mathbb{R}^2	0.210	0.15 33	0.19 37	0.18 30	0.1 <mark>65</mark>	0.2 <mark>30</mark>	0.38 59	0.37 5

1259 ¹ For P-3B, the concentrations of NO_v, NO, and NO₂ NCAR were measured by using the NCAR 4-channel chemiluminescence instrument. The measurement

1260 uncertainties are 10%, 10 - 15%, and 10% for NO, NO₂, and NO_y, respectively. The 1-second, 1-sigma detection limits are 20 ppty, 30 ppty, and 20 ppty for NO, 1261 NO2, and NOy, respectively (https://discover-aq.larc.nasa.gov/pdf/2010STM/Weinheimer20101005 DISCOVERAO AJW.pdf). For REAM, NOy is the sum of

1262 NO, NO₂, total peroxyacyl nitrates (Σ PNs), total alkyl nitrates (Σ ANs) (include alkyl nitrates and hydroxyalkyl nitrates), HNO₃, HONO, 2 × N₂O₅, HNO₄, first

1263 generation C5 carbonyl nitrate (nighttime isoprene nitrate ISN1: $C_5H_8NO_4$), 2 × C5 dihydroxydinitrate (DHDN: $C_5H_{10}O_8N_2$), methyl peroxy nitrate (MPN:

1264 CH₃O₂NO₂), propanone nitrate (PROPNN: CH₃C(=O)CH₂ONO₂), nitrate from methyl vinyl ketone (MVKN: HOCH₂CH(ONO₂)C(=O)CH₃), nitrate from

1265 methacrolein (MARCN: HOCH₂C(ONO₂)(CH₃)CHO), and ethanol nitrate (ETHLN: CHOCH₂ONO₂).

1266 ² For P-3B, the concentrations of NO₂ LIF, Σ PNs, Σ ANs, and HNO₃ were measured by applying the thermal dissociation-laser induced fluorescence (TD-LIF) 1267 technique. The accuracy of TD-LIF measurements of NO₂, Σ PNs, Σ ANs, and HNO₃ is better than 15%, and the detection limit for the sum of NO₂, Σ PNs, Σ ANs, 1268 and HNO₃ is ~ 10 ppt 10 s⁻¹ (Day et al., 2002).

1269 ³ To compare NO_V concentrations from TD-LIF measurements with those from REAM, we calculate derived-NO_V as the sum of NO, NO₂ LIF, Σ PNs, Σ ANs, and

1270 HNO₃. Only when the concentrations of all the five species are available at the same hour in the same grid cell, we can calculate derived-NO_y at the given hour in

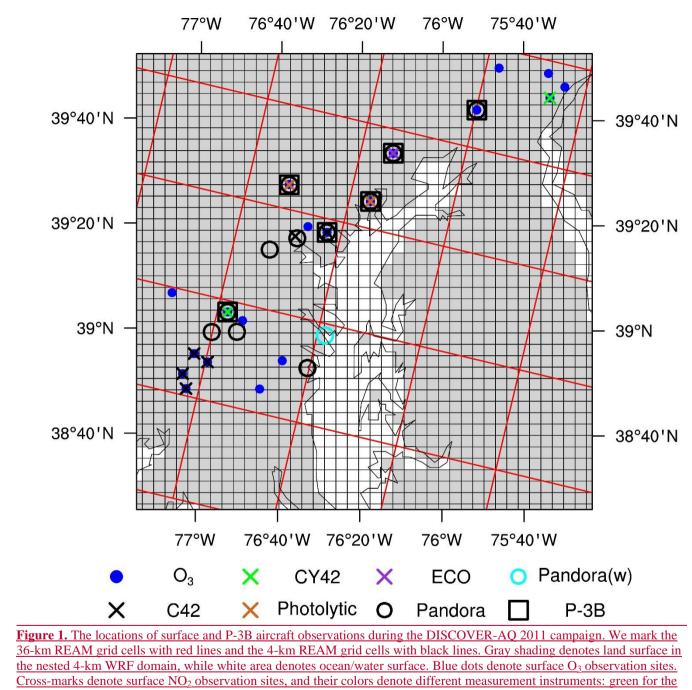
1271 the given grid cell. Therefore, in Table 1, the averaged derived-NO_v values are not exactly equal to the sum of averaged NO, NO₂ LIF, Σ PNs, Σ ANs, and HNO₃

1272 concentrations that only depend on the availability of a single species. In addition, the measurement times and frequencies between NO_y and derived-NO_y are not

1273 the same. A comparison between these two types of data needs coincident sampling, as described in the main text.

1274 ⁴ Mean NO_x emissions over the six P-3B spiral sites are close (relative difference < 4%) between the 36-km and 4-km REAM (Table S1).

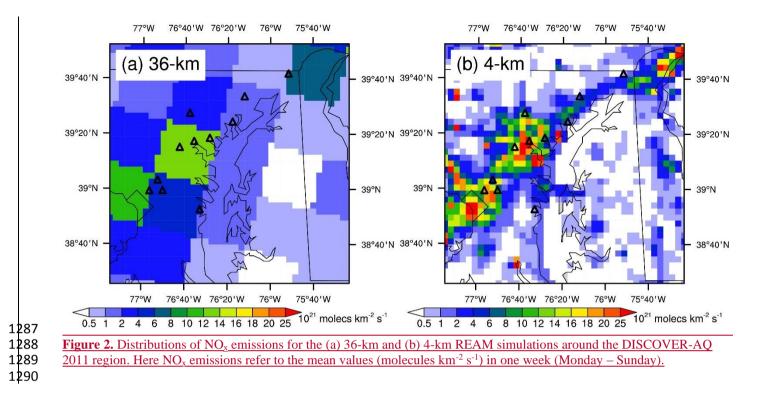
1275 ⁵ Due to different sampling times and locations between weekdays and weekends, we do not recommend a direct comparison between weekday and weekend 1276 values here.



1282 Thermo Electron 42C-Y NO_y analyzer, dark orchid for the Ecotech Model 9841/9843 T-NO_y analyzers, black for the
 1283 Thermo Model 42C NO_x analyzer, and chocolate for the Teledyne API model 200eup photolytic NO_x analyzer. Circles

1284 denote Pandora sites, and the cyan circle denotes a Pandora site (USNA) on a ship. Black squares denote the inland P-3B

1285 <u>aircraft spiral locations.</u>



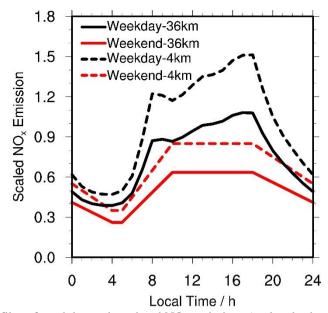
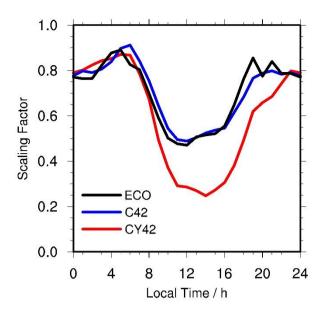
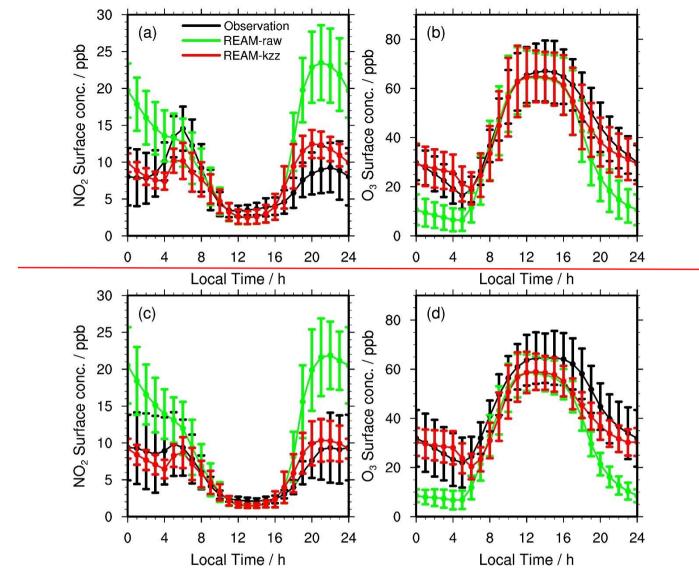


Figure 31. Relative diurnal profiles of weekday and weekend NO_x emissions (molecules km⁻² s⁻¹) in the DISCOVER-AQ 2011 region (the 36/4 km grid cells over the 11 inland Pandora sites shown in Figure \$1) for the 36-km and 4-km REAM.

All the profiles are scaled by the 4-km weekday emission average value (molecules $\text{km}^{-2} \text{ s}^{-1}$).



1296 1297 Figure 42. Hourly ratios of NO₂ measurements from the Teledyne API model 200 eup photolytic NO_x analyzer to NO₂ from 1298 coincident catalytic instruments for 2011 July. "CY42" denotes the ratios of photolytic NO₂ to NO₂ from the Thermo 1299 Electron 42C-Y NO_v analyzer in Edgewood, "C42" denotes the ratios of photolytic NO₂ to NO₂ from the Thermo Model 1300 42C NO_x analyzer in Padonia, and "ECO" denotes the ratios of photolytic NO₂ to NO₂ from the Ecotech Model 9841 T-NO_y 1301 analyzer in Padonia. "ECO" ratios are also used to scale NO₂ measurements from the Ecotech Model 9843 T-NO_y analyzer. 1302 Thermo Model 42I Y NO_x analyzer was used only in Padonia, where photolytic measurements were available, so we do not 1303 use the Thermo Model 42I Y NO₊ analyzer measurements in this study.





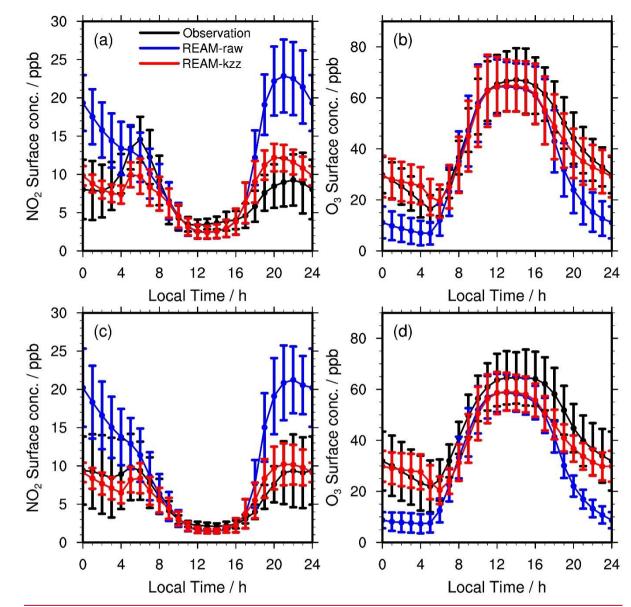
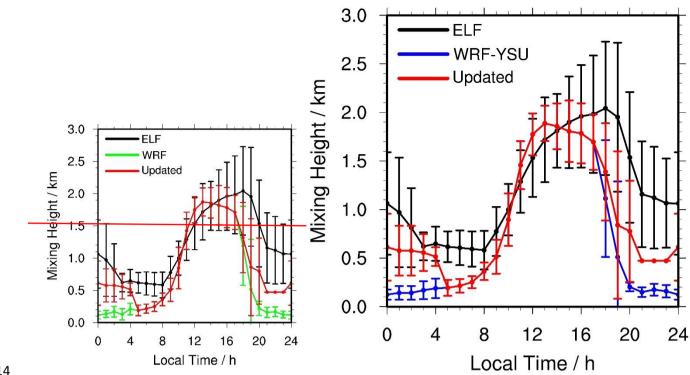
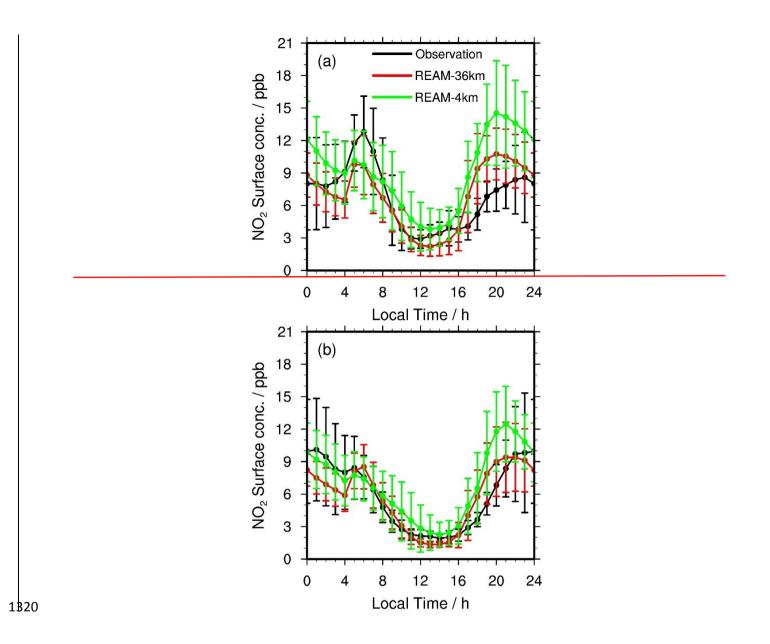
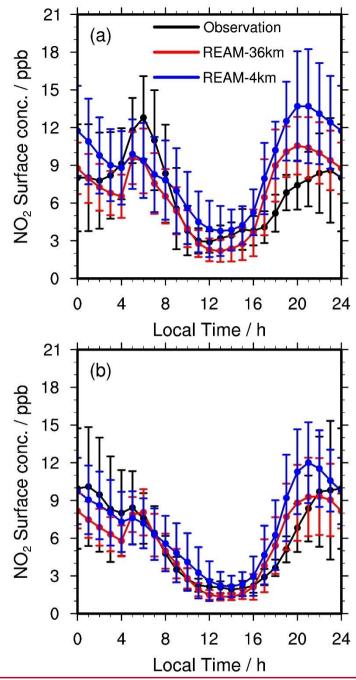


Figure 53. Diurnal cycles of surface (a, c) NO₂ and (b, d) O₃ concentrations on (a, b) weekdays and (c, d) weekends during the DISCOVER-AQ campaign in the DISCOVER-AQ region (the 36-km grid cells over the 11 inland Pandora sites shown in Figure \$1). Black lines denote the mean observations from all the 11 NO₂ surface monitoring sites and 19 O₃ surface sites during the campaign (Figure \$1), as mentioned in Section 2.5. "REAM-raw" (green-blue lines) denotes the coincident 36-km REAM simulation results with WRF-YSU simulated k_{zz} data, and "REAM-kzz" (red lines) is the coincident 36-km REAM simulation results with updated k_{zz} data. See the main text for details. Vertical bars denote corresponding standard deviations.



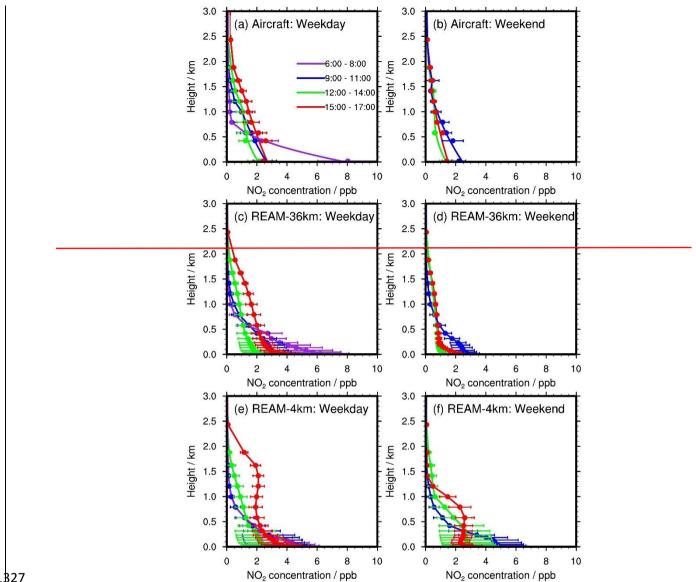
1315Figure <u>64.</u> ELF observed and model simulated diurnal variations of PBLH at the UMBC site during the Discover-AQ1316campaign. "ELF" denotes ELF derived PBLHs by using the covariance wavelet transform method. "WRF-<u>YSU</u>" denotes the131736-km WRF-<u>YSU</u> k_{zz} -determined PBLHs, and "Updated" denotes updated k_{zz} -determined PBLHs. See the main text for1318details. Vertical bars denote standard deviations.







1321 1322 Figure 75. Diurnal cycles of observed and simulated average surface NO₂ concentrations over Padonia, Oldtown, Essex, 1323 Edgewood, Beltsville, and Aldino (Table S1) on (a) weekdays and (b) weekends. Black lines denote mean observations from 1324 the six sites. Red lines denote coincident 36-km REAM simulation results, and green-blue lines are for coincident 4-km 1325 REAM simulation results. Error bars denote standard deviations.



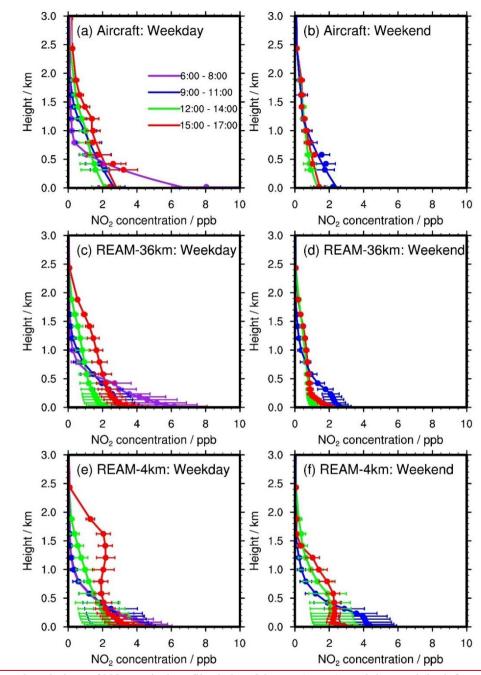


Figure <u>86</u>. Temporal evolutions of NO₂ vertical profiles below 3 km on (a, c, e) weekdays and (b, d, f) weekends from the (a, b) P-3B aircraft and (c, d) 36-km and (e, f) 4-km REAM during the DISCOVER-AQ campaign. Horizontal bars denote the corresponding standard deviations. In (a) and (b), dots denote aircraft measurements, while lines below 1 km are based on quadratic polynomial fitting, as described in section 2.6. The fitting values are mostly in reasonable agreement with the aircraft and surface measurements in the boundary layer. On weekends, no aircraft observations were made at 6:00 – 8:00 LT, and therefore no corresponding model profiles are shown.

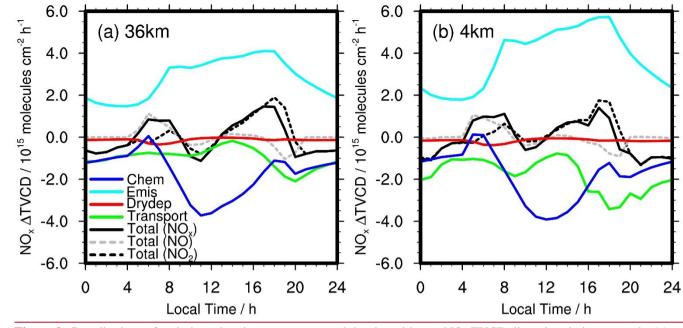
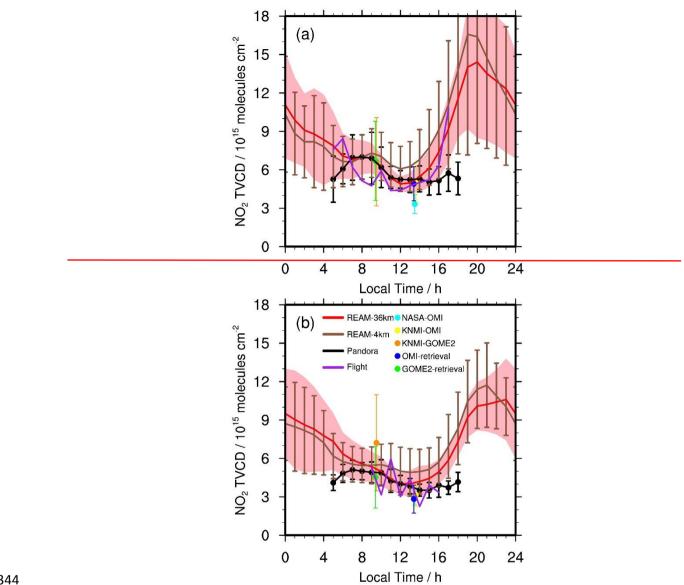
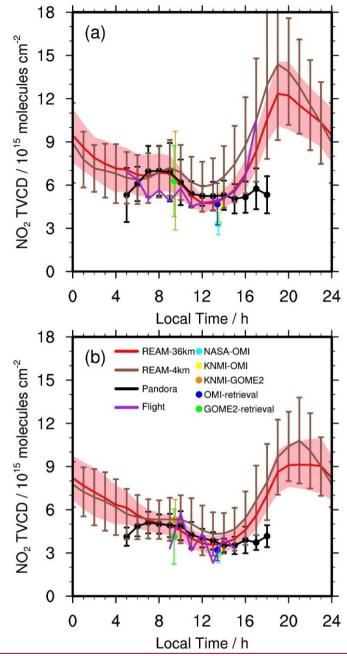
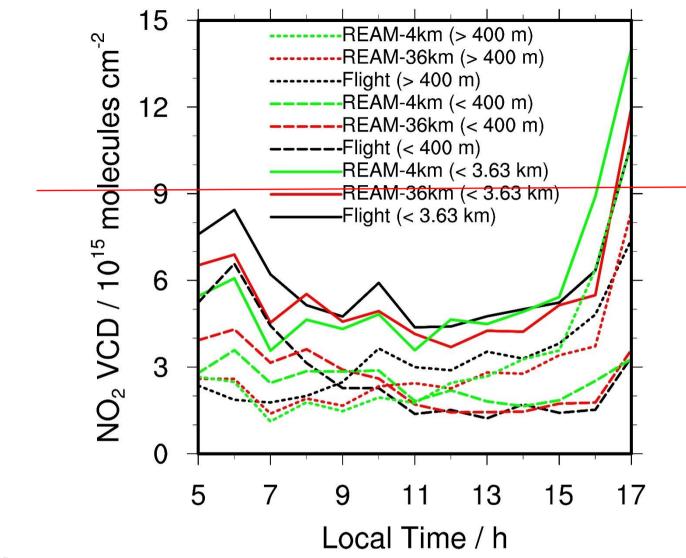


Figure 9. Contributions of emission, chemistry, transport, and dry deposition to NOx TVCD diurnal variations over the 11**1337Figure 9.** Contributions of emission, chemistry, transport, and dry deposition to NOx TVCD diurnal variations over the 11**1338**inland Pandora sites (Table S1 and Figure 1) on weekdays in July 2011 for the (a) 36-km and (b) 4-km REAM simulations.**1339**"Chem" refers to net NOx chemistry production; "Emis" refers to NOx emissions; "Drydep" denotes NOx dry depositions;**1340**"Transport" includes advection, turbulent mixing, lightning NOx production, and wet deposition. "Total (NOx)" is the hourly**1341**change of NOx TVCDs ($\triangle(TVCD) = TVCD_{t+1} - TVCD_t$). "Total (NO2)" is the hourly change of NO2 TVCDs, and "Total**1342**(NO)" is the hourly change of NO TVCDs.





1346 Figure 107. Daily variations of NO₂ TVCDs on (a) weekdays and (b) weekends during the DISCOVER-AQ campaign. 1347 "REAM-36km" refers to the 36-km REAM simulation results over the 11 inland Pandora sites. "REAM-4km" refers to the 1348 4-km REAM simulation results over the 11 inland Pandora sites. "Pandora" refers to updated Pandora TVCD data. "Flight" 1349 denotes P-3B aircraft-derived NO₂ VCDs below 3.63 km. "NASA-OMI" denotes the OMI NO₂ TVCDs retrieved by NASA 1350 over the Pandora sites; "KNMI-OMI" denotes the OMI NO2 TVCDs from KNMI; "KNMI-GOME2" is the GOME-2A NO2 1351 TVCDs from KNMI. "OMI-retrieval" and "GOME2-retrieval" denote OMI and GOME-2A TVCDs retrieved by using the 1\$52 KNMI DOMINO algorithm with corresponding 36-km REAM vertical profiles, respectively. We list NO2 TVCD values at 1353 9:30 and 13:30 LT in Table S3.



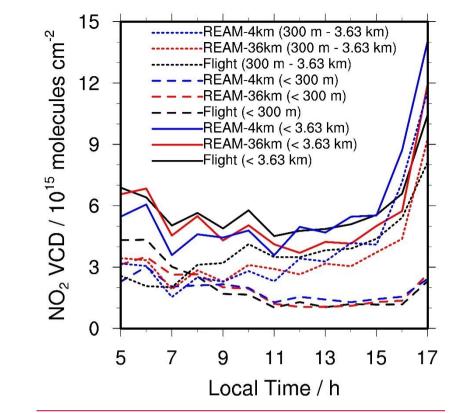
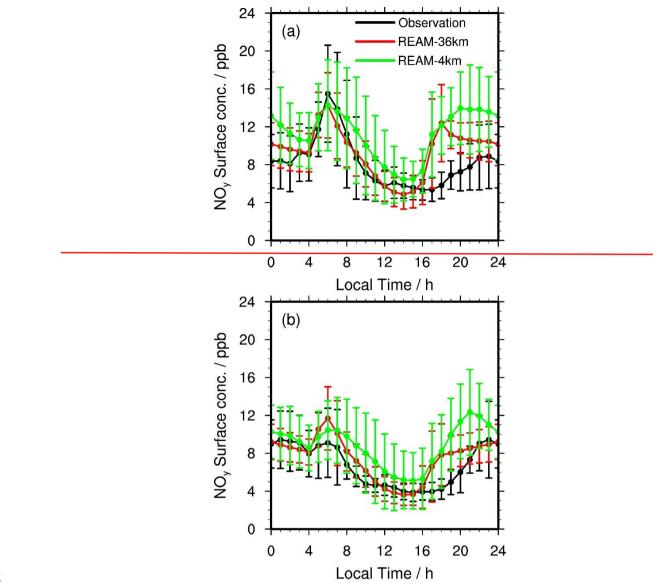


Figure 118. Weekday hourly variations of NO₂ VCDs at different height (AGL) bins (< 3.63 km AGL, < 3400 m AGL, and 3400 m ~ 3.63 km AGL) based on P-3B aircraft-derived datasets and the 36-km and 4-km REAM results. "Flight" denotes P-3B aircraft-derived NO₂ VCDs, "REAM-36km" denotes coincident 36-km REAM simulated VCDs, and "REAM-4km" denotes coincident 4-km REAM simulated VCDs.



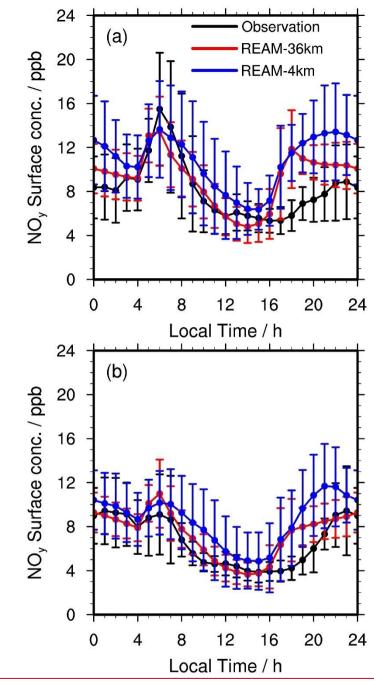
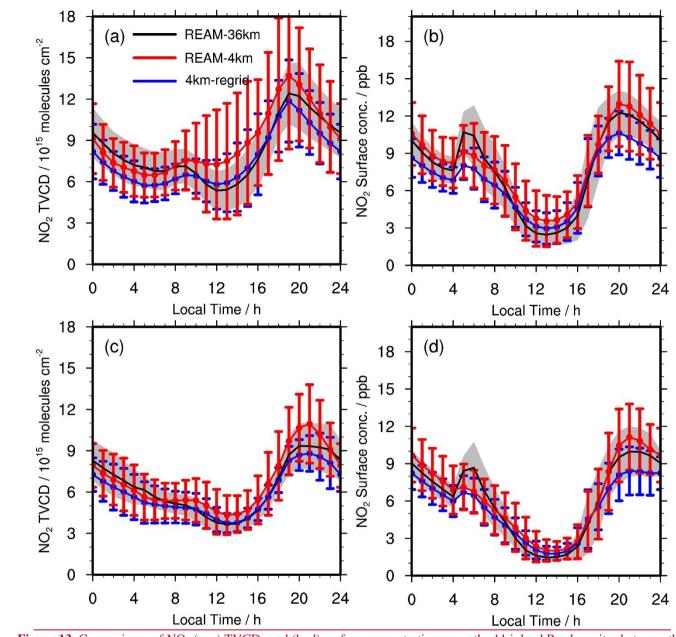


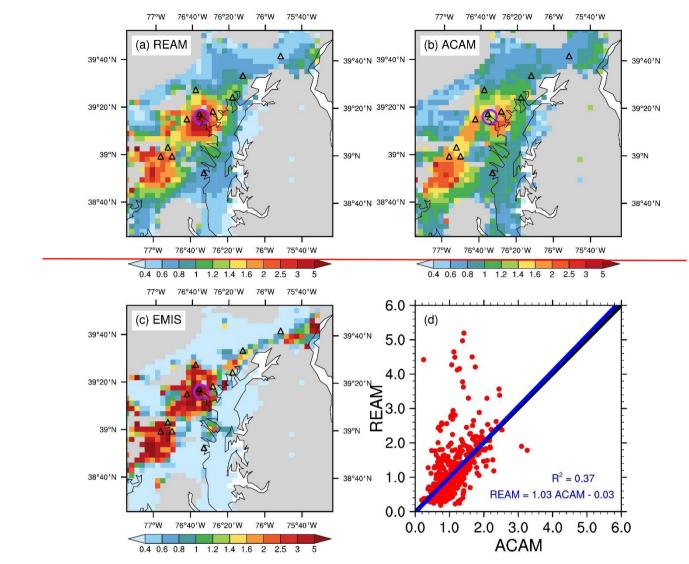
Figure 129. Diurnal cycles of observed and simulated average surface NO_y concentrations at Padonia, Edgewood, Beltsville, and Aldino on (a) weekdays and (b) weekends. Vertical bars denote the corresponding standard deviations. It is noteworthy that the mean NO_x emissions over Padonia, Edgewood, Beltsville, and Aldino are 99% higher in the 4-km than the 36-km REAM simulations (Table S1-and Figure S1).

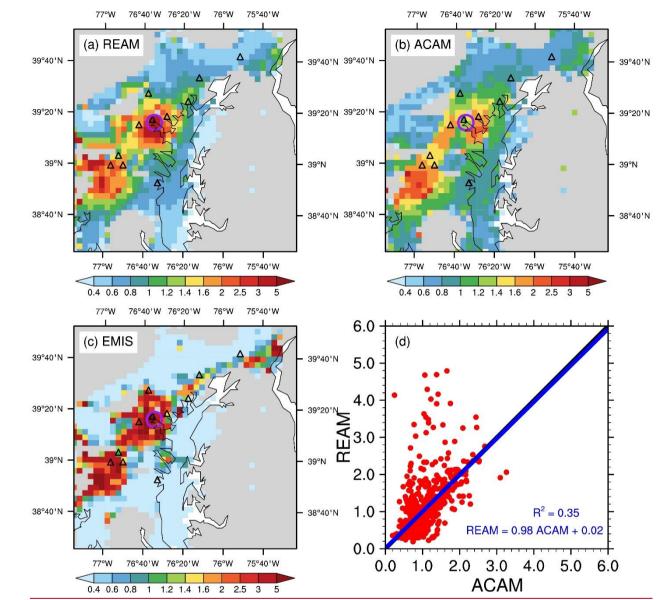


1369

1370 1\$71 1372 1373

Figure 13. Comparisons of NO₂ (a, c) TVCDs and (b, d) surface concentrations over the 11 inland Pandora sites between the 4-km and 36-km REAM simulations on (a, b) weekdays and (c, d) weekends for July 2011. "REAM-36km" (black lines) denotes the 36-km REAM simulation results; "REAM-4km" (red lines) denotes the 4-km REAM simulation results; "4kmregrid" (blue lines) refers to the 36-km values by re-gridding the 4-km REAM simulation results into 36-km REAM grid 1374 cells. Error bars denote standard deviations.





1\$77

Figure 140. Distributions of the scaled mean (a) ACAM NO₂ VCDs below the UC-12 aircraft and (b) coincident 4-km
REAM simulation results on weekdays in July 2011. (c), the distribution of the scaled NEI2011 NO_x emissions on
weekdays. The purple circles denote a small region surrounded by high-NO_x emission pixels and with high NO₂ VCDs in the
4-km REAM but low NO₂ VCDs in ACAM. (d) is the scatter plot of the scaled ACAM and 4-km REAM NO₂ VCDs from
(a) and (b). Here, we scale all values (VCDs and NO_x emissions) based on their corresponding domain averages. The domain averages of ACAM and coincident 4-km REAM NO₂ VCDs are 4.7 ± 2.0 and 4.65 ± 3.2 × 10¹⁵ molecules cm⁻², respectively.

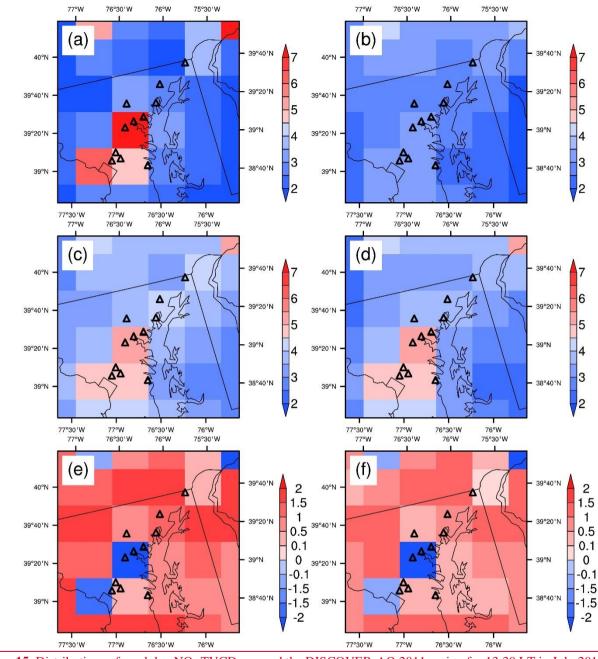
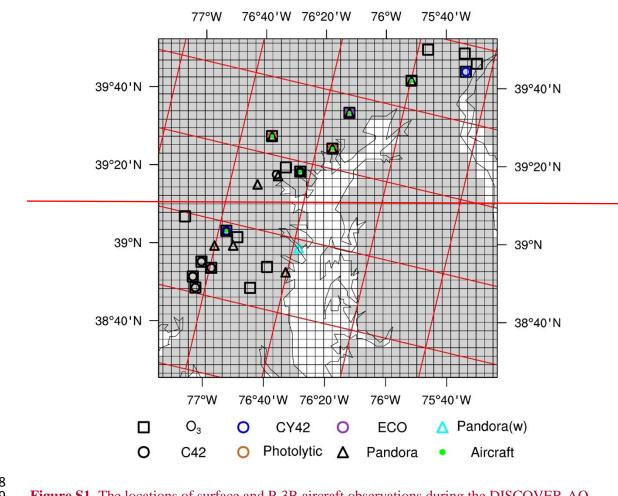
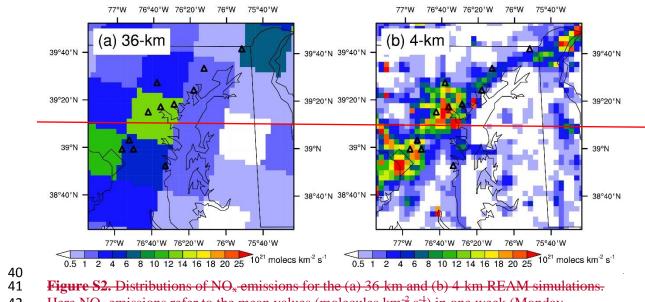


Figure 15. Distributions of weekday NO₂ TVCDs around the DISCOVER-AQ 2011 region for 13:30 LT in July 2011: (a)
 the 36-km REAM simulation results, (b) the NASA OMI product (OMNO2), (c) the KNMI OMI product, (d) the retrieved
 OMI NO₂ TVCDs by using the KNMI DOMINO algorithm with corresponding 36-km REAM vertical profiles, (e) the
 distribution of the NO₂ TVCD differences (c minus a) between KNMI OMI and 36-km REAM, and (f) the difference (d
 minus a) between retrieved OMI NO₂ TVCDs and the 36-km REAM results. The NO₂ TVCD unit is 10¹⁵ molecules cm⁻².

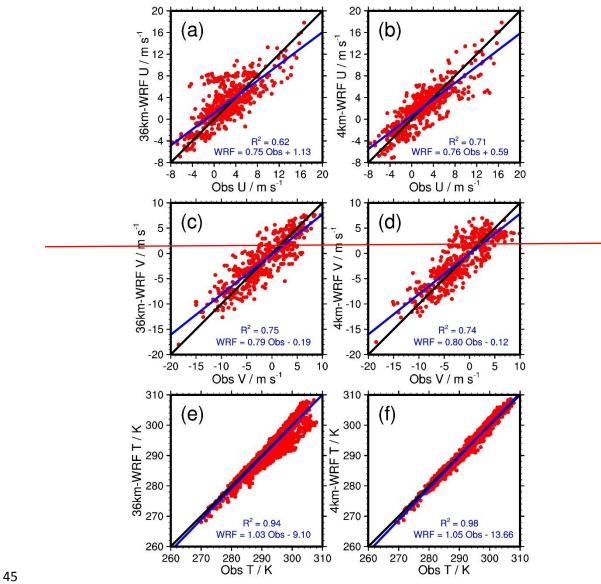
1	Supporting figures for
2	Comprehensive evaluations of diurnal NO2 measurements during
3	DISCOVER-AQ 2011: Effects of resolution dependent representation of NO _x
4	emissions
5 6 7 8	Jianfeng Li ^{1, a} , Yuhang Wang ^{1*} , Ruixiong Zhang ¹ , Charles Smeltzer ¹ , Andrew Weinheimer ² , Jay Herman ³ , K. Folkert Boersma ^{4, 5} , Edward A. Celarier ^{6, 7, b} , Russell W. Long ⁸ , James J. Szykman ⁸ , Ruben Delgado ³ , Anne M. Thompson ⁶ , Travis N. Knepp ^{9, 10} , Lok N Lamsal ⁶ , Scott J Janz ⁶ , Matthew G Kowalewski ⁶ , Xiong Liu ¹¹ , Caroline R. Nowlan ¹¹
9 10	¹ School of Earth and Atmospheric Sciences, Georgia Institute of Technology, Atlanta, Georgia, USA
11	² National Center for Atmospheric Research, Boulder, Colorado, USA
12	³ University of Maryland Baltimore County JCET, Baltimore, Maryland, USA
13	⁴ Royal Netherlands Meteorological Institute, De Bilt, the Netherlands
14	⁵ Wageningen University, Meteorology and Air Quality Group, Wageningen, the Netherlands
15	⁶ NASA Goddard Space Flight Center, Greenbelt, Maryland, USA
16	⁷ Universities Space Research Association, Columbia, Maryland, USA
17	⁸ National Exposure Research Laboratory, Office of Research and Development, U.S.
18	Environmental Protection Agency, Research Triangle Park, NC, USA
19	⁹ NASA Langley Research Center, Virginia, USA
20	¹⁰ Science Systems and Applications, Inc., Hampton, Virginia, USA
21	¹¹ Harvard-Smithsonian Center for Astrophysics, Cambridge, Massachusetts, USA
22 23	anow at Atmospheric Sciences and Clobal Change Division, Desific Northwest National
23 24	^a now at: Atmospheric Sciences and Global Change Division, Pacific Northwest National Laboratory, Richland, Washington, USA
24 25	^b now at: Digital Spec, Tyson's Corner, VA, USA
25	
26	* Correspondence to Yuhang Wang (yuhang.wang@eas.gatech.edu)



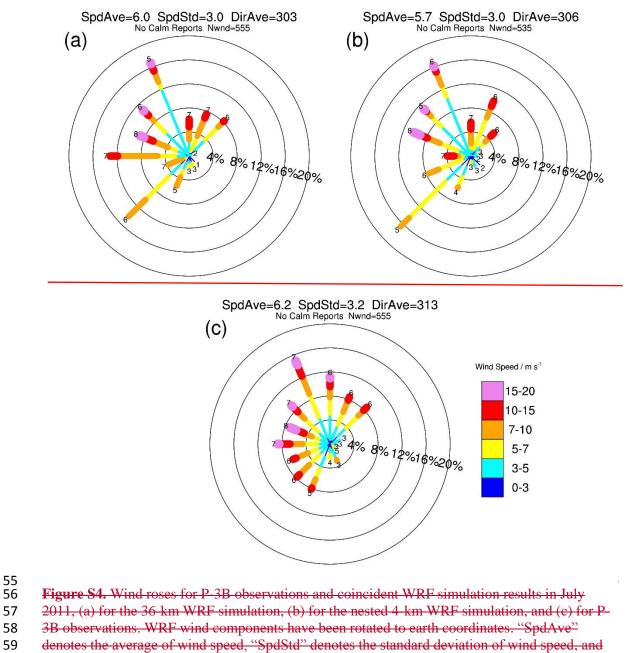
29 Figure S1. The locations of surface and P 3B aircraft observations during the DISCOVER AQ 30 2011 campaign. We mark the 36 km REAM grid cells with red lines and the 4 km REAM grid 31 cells with black lines. Gray shading denotes land surface in the nested 4 km WRF domain, while 32 white denotes ocean/water surface. Black squares denote surface O3 observation sites. Circles 33 denote surface NO₂-observation sites, and their colors denote different measurement instruments: 34 blue for the Thermo Electron 42C-Y NO_x analyzer, dark orchid for the Ecotech Model 9841/9843 35 T-NO_{*} analyzers, black for the Thermo Model 42C NO_{*} analyzer, and chocolate for the Teledyne 36 API model 200eup photolytic NO_{*} analyzer. Triangles denote Pandora sites, and the cyan triangle 37 denotes a Pandora site (USNA) on a ship. Green dots denote the inland P 3B aircraft spiral 38 locations. 39



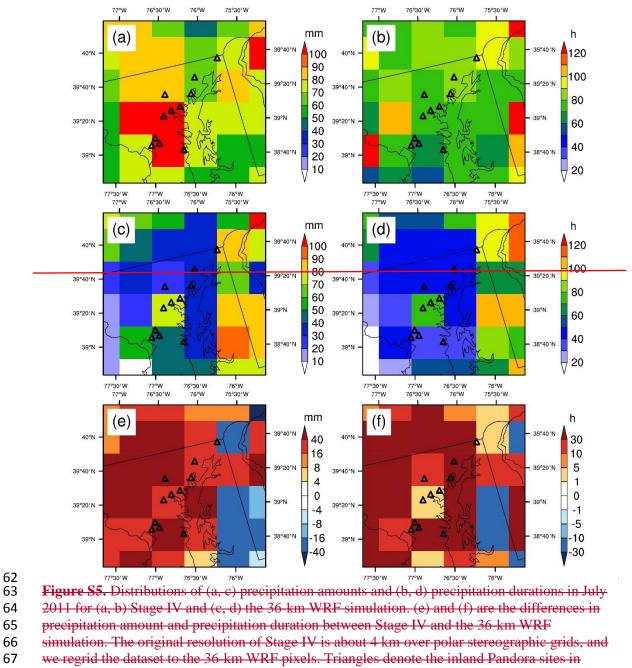
- 42 Here NO_{*} emissions refer to the mean values (molecules km⁻² s⁻⁴) in one week (Monday—
- 43 Sunday).
- 44



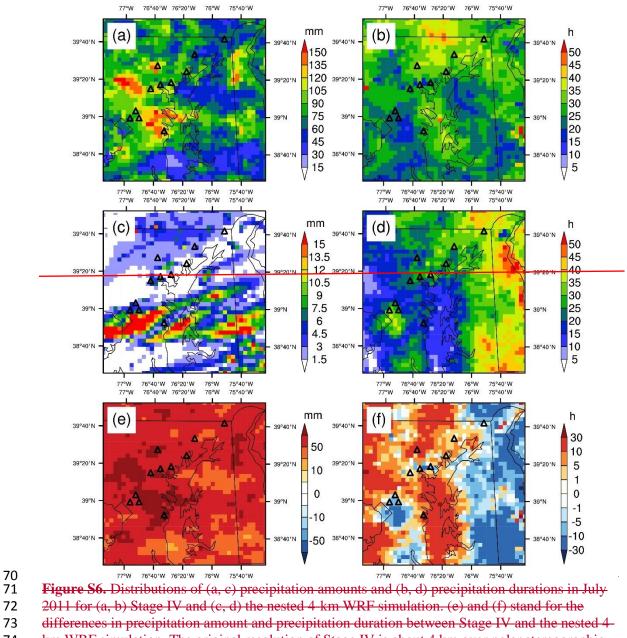
46 Figure S3. Comparisons of (a, b) U wind, (c, d) V wind, and (e, f) temperature (T) between P-3B 47 spirals and coincident WRF simulation results in July 2011. For P-3B observations, we derive U 48 wind and V wind from wind speed and wind direction (Figure S4), which were measured via a 49 Honeywell INS sensor. The accuracies of P-3B wind speed and wind direction are 1 m s⁺ and 50 ±5°, respectively. Temperature from P-3B was measured by using a Rosemount model 102 sensor 51 with an accuracy of ±0.2 °C. The left panel is for comparisons between P-3B and the 36 km WRF 52 simulation, and the right panel is for comparisons between P 3B and the nested 4 km WRF 53 simulation. WRF wind components have been rotated to earth coordinates.



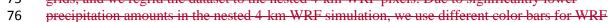
- 60 "DirAve" denotes wind direction derived from averaged U-wind and V-wind.
- 61



- 68 Figure S1.
- 69



74 km WRF simulation. The original resolution of Stage IV is about 4 km over polar stereographic
 75 grids, and we regrid the dataset to the nested 4-km WRF pixels. Due to significantly lower



- and Stage IV data. Triangles denote the inland Pandora sites in Figure S1.
- 78

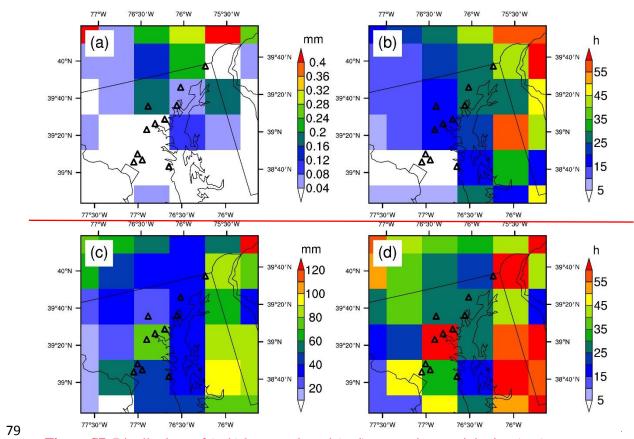
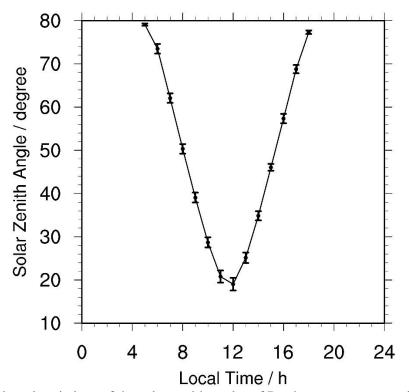


Figure S7. Distributions of (a, b) large scale and (c, d) convective precipitation (a, c) amounts
 and (b, d) durations in July 2011 for the 36 km WRF simulation. Due to significantly lower large-

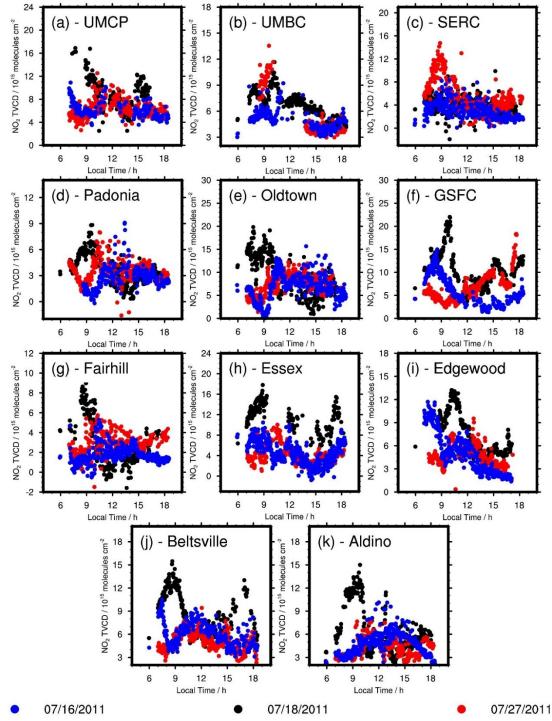
82 scale precipitation amounts compared to convective precipitation, we use different color bars for

- 83 them. Triangles denote the inland Pandora sites in Figure S1.
- 84

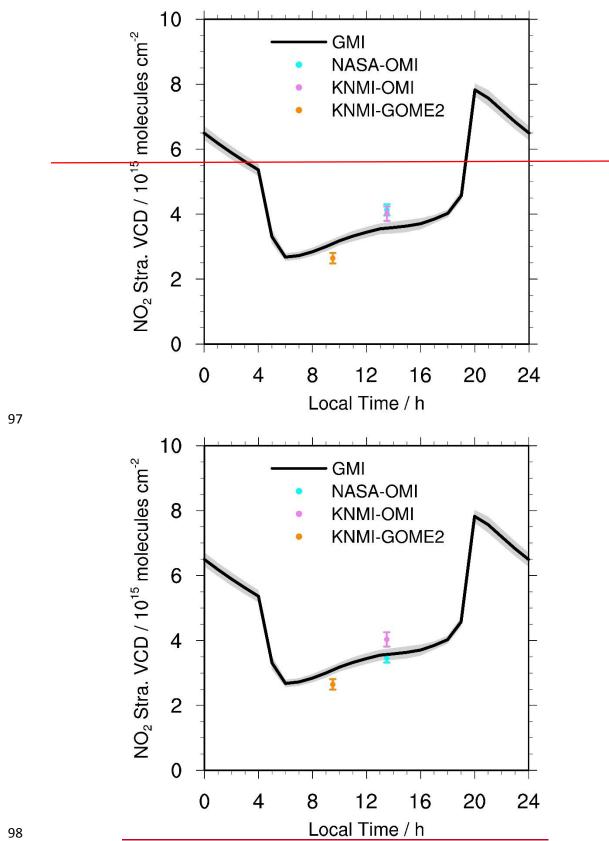


85 86 Figure S18. Diurnal variations of the solar zenith angles of Pandora measurements in July 2011.

- 87 Here we use monthly averages of the 11 inland Pandora sites (Table S1 and Figure S1). Error bars 88 denote standard deviations.
- 89

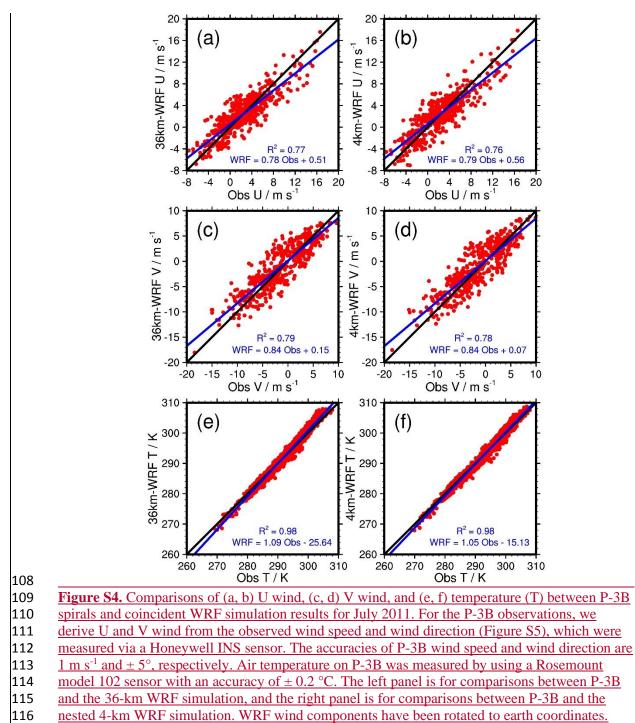


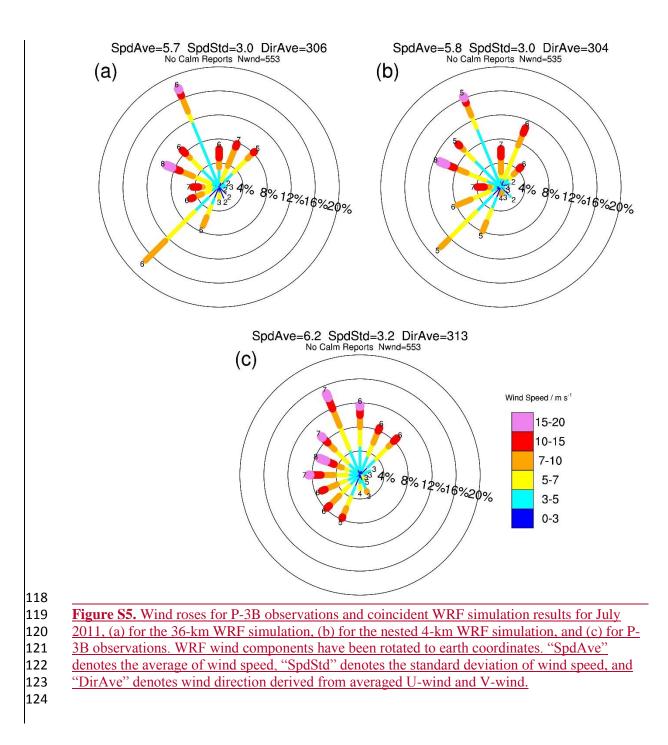
90 07/16/2011 07/18/2011 07/27/2011
91 Figure S29. Daily variations of Pandora NO₂ TVCDs at the 11 inland sites for three randomly
92 selected days in July 2011. Blue dots denote Pandora measurements on July 16, 2011 (Saturday),
93 black dots denote July 18, 2011 (Monday), and red dots denote July 27, 2011 (Wednesday). Here
94 Pandora NO₂ TVCDs are calculated by subtracting stratospheric NO₂ VCDs from Pandora total
95 NO₂ VCDs. See the main text for details.

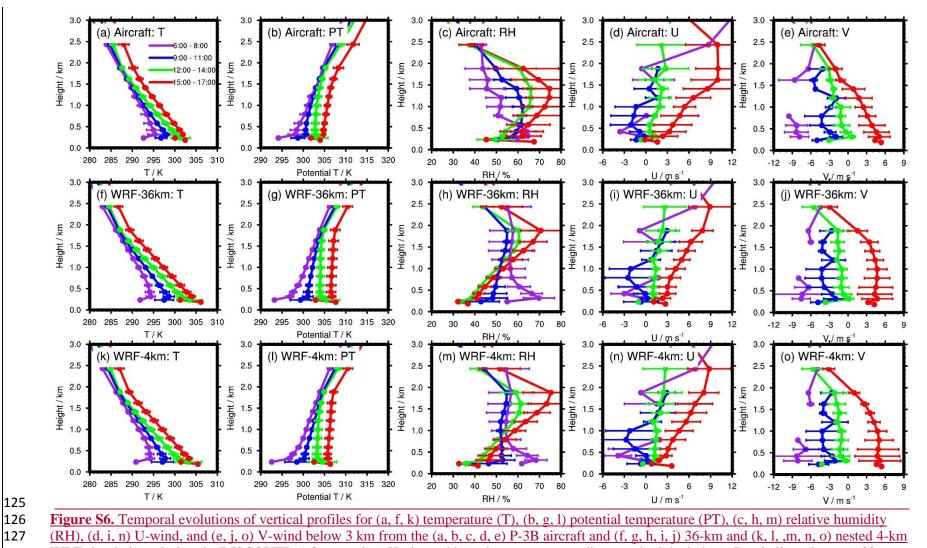


99 **Figure S<u>310</u>**. Stratospheric NO₂ VCD diurnal variations at the Greenbelt station in Maryland (39°

- 100 N, 76.89° W) from the GMI MERRA-2 $1^{\circ} \times 1.25^{\circ}$ simulation
- 101 (<u>https://gmi.gsfc.nasa.gov/merra2hindcast/, last access: May 14, 2019</u>) for July 2011, and the
- 102 corresponding satellite stratospheric NO₂ VCDs in the DISCOVER-AQ region (about 39.2° N,
- 103 76.3° W) (Figure <u>\$1</u>). "NASA-OMI" denotes the OMI NO₂ VCDs from NASA, "KNMI-OMI"
- denotes the OMI NO₂ VCDs from KNMI, and "KNMI-GOME2" denotes the GOME-2A NO₂
- 105 VCDs from KNMI. Gray shading and vertical bars denote the standard deviations of the GMI
- 106 results and satellite stratospheric VCD products, respectively.

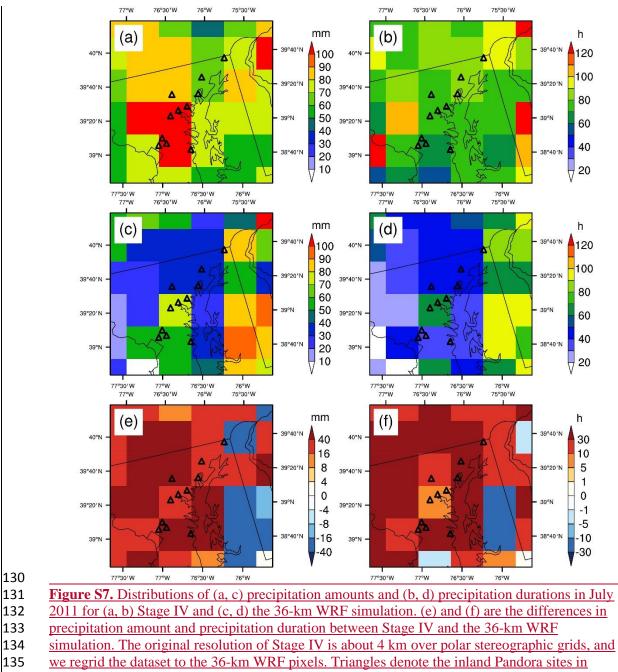




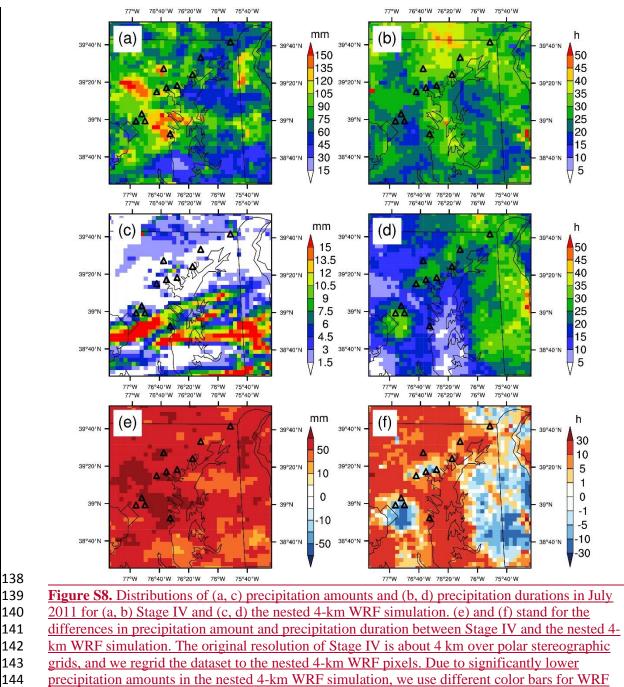


128 WRF simulations during the DISCOVER-AQ campaign. Horizontal bars denote corresponding standard deviations. Purple lines denote 6:00 –

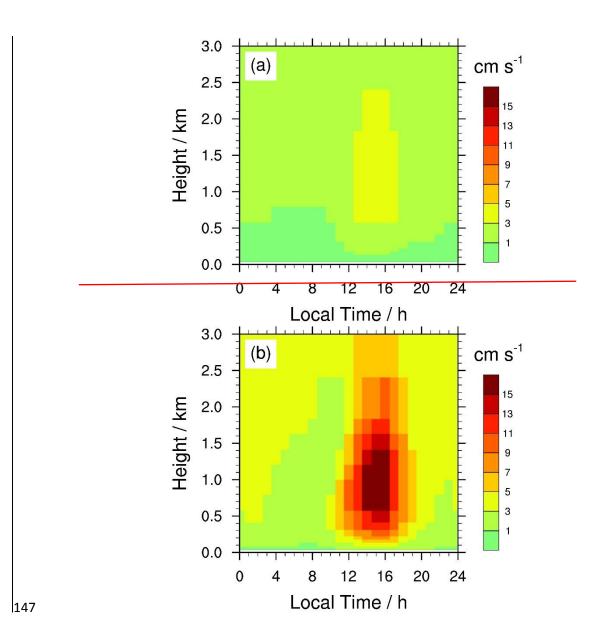
129 <u>8:00 LT, blue lines for 9:00 – 11:00 LT, green lines for 12:00 – 14:00 LT, and red lines for 15:00 – 17:00 LT.</u>

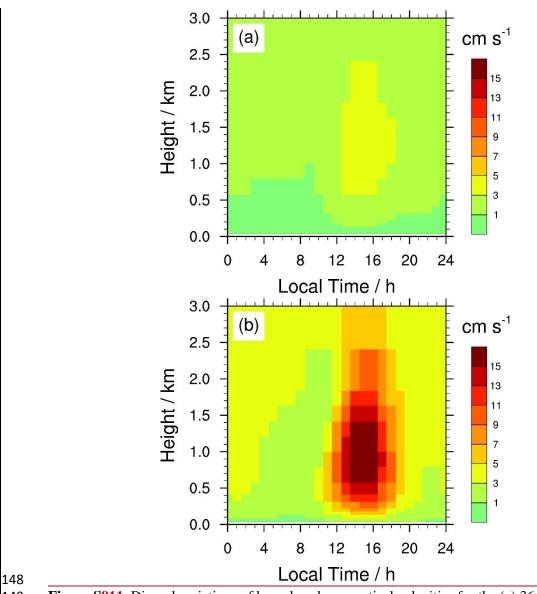


- 136 <u>Figure 1.</u>
- 137

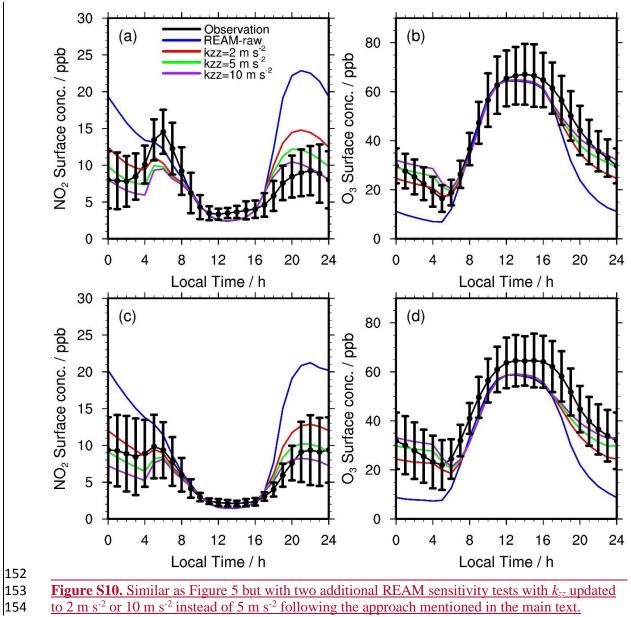


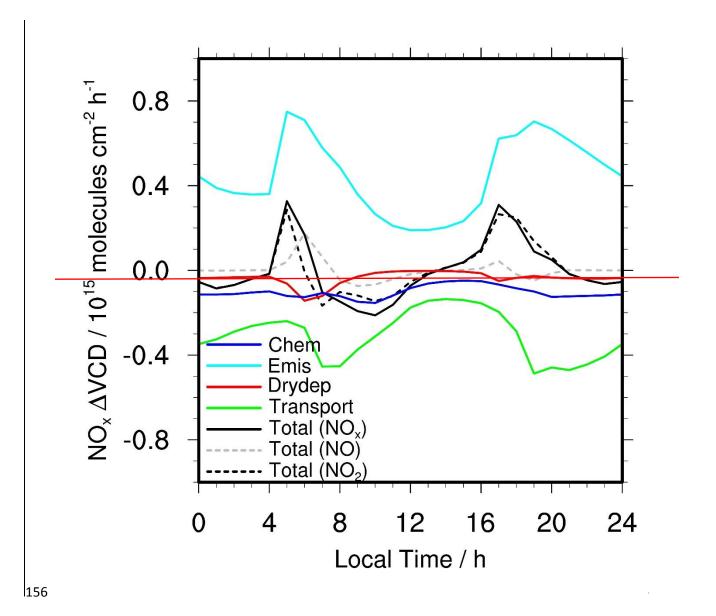
- and Stage IV data. Triangles denote the inland Pandora sites in Figure 1.
- 146

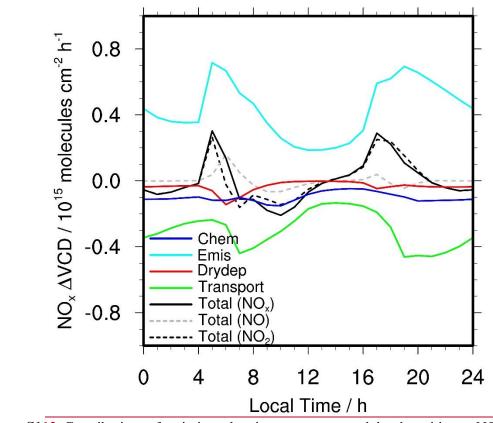




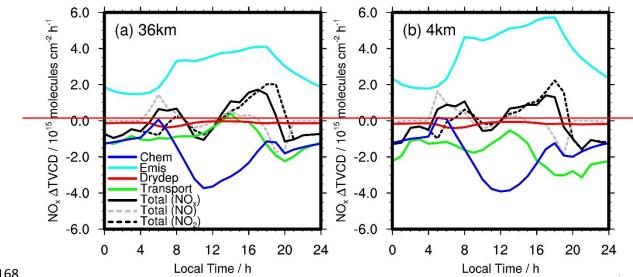








158 Figure S112. Contributions of emission, chemistry, transport, and dry deposition to NO_x VCD diurnal variations in the surface layer of the 36-km REAM simulation in the DISCOVER-AQ 159 region on weekdays in July 2011. "Chem" refers to net NO_x chemistry production in the surface 160 161 layer; "Emis" refers to NO_x emissions in the surface layer with the impact of vertical turbulent mixing; "Drydep" denotes NO_x dry depositions in the surface layer; "Transport" includes 162 advection, turbulent mixing, lightning NO_x production, and wet deposition in the surface layer. 163 "Total (NO_x)" is the hourly change of surface-layer NO_x VCDs (\triangle (*VCD*) = *VCD*_{t+1} - *VCD*_t). 164 "Total (NO₂)" is the hourly change of surface-layer NO₂ VCDs, and "Total (NO)" is the hourly 165 change of surface-layer NO VCDs. 166 167



168Local Time / hLocal Time / h169Figure S13. Contributions of emission, chemistry, transport, and dry deposition to NO_{*} TVCD170diurnal variations over the 11 inland Pandora sites (Table S1 and Figure S1) on weekdays in July1712011 for the (a) 36 km and (b) 4 km REAM simulations. "Chem" refers to net NO_{*} chemistry172production; "Emis" refers to NO_{*} emissions; "Drydep" denotes NO_{*} dry depositions; "Transport"173includes advection, turbulent mixing, lightning NO_{*} production, and wet deposition. "Total174(NO_{*})" is the hourly change of NO_{*} TVCDs ($\triangle(TVCD) = TVCD_{t+T} - TVCD_t)$. "Total (NO₂)" is175the hourly change of NO₂ TVCDs, and "Total (NO)" is the hourly change of NO TVCDs.

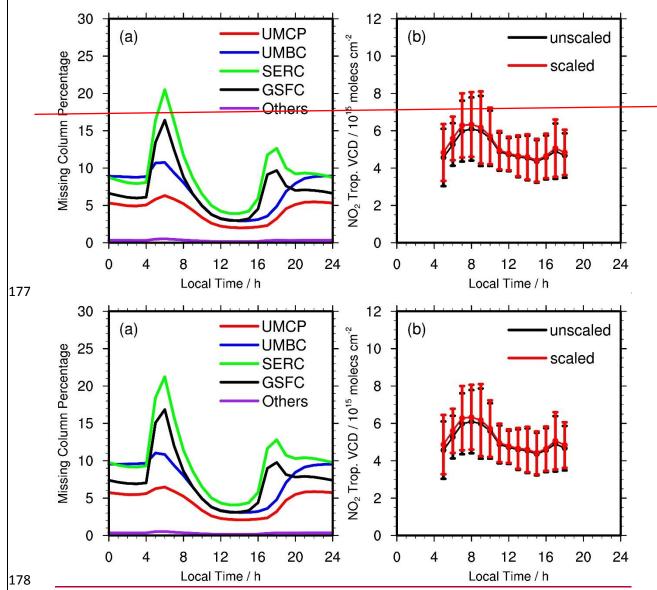
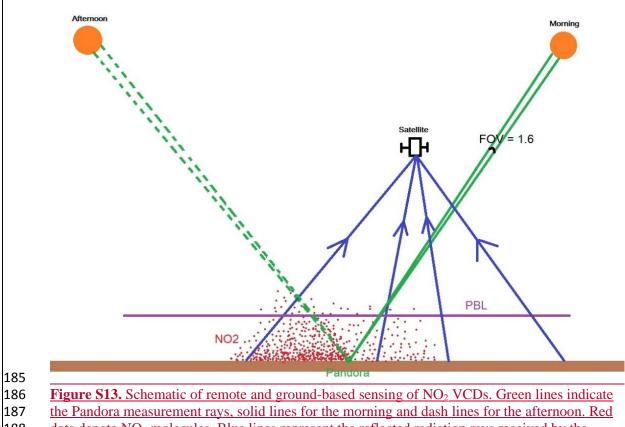
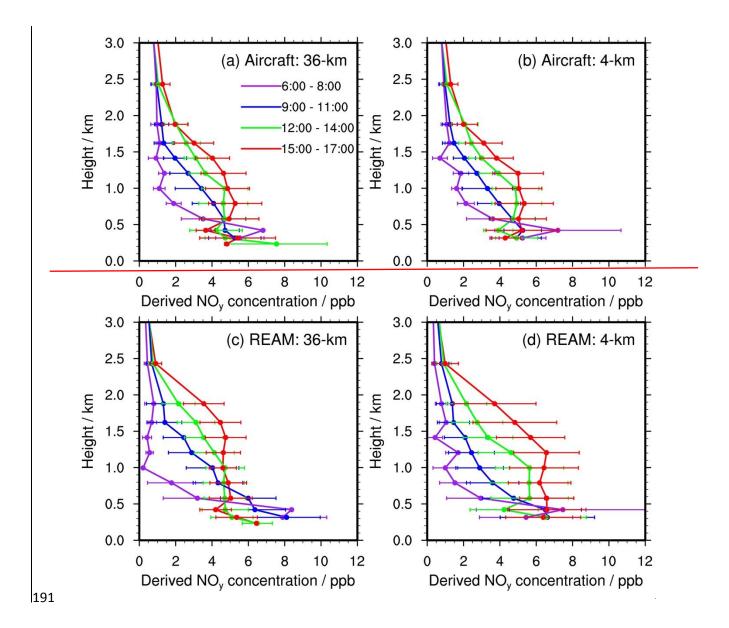


Figure S124. (a) Percentages of NO₂ VCDs below the heights of Pandora instruments in July
2011 based on 36-km REAM results; (b) the comparison between original Pandora TVCDs
("unscaled") and updated Pandora TVCDs ("scaled") with the inclusion of VCDs below the
Pandora instruments. Here we use monthly averages in July 2011. Error bars in (b) denote
standard deviations.



188 dots denote NO₂ molecules. Blue lines represent the reflected radiation rays received by the
 189 satellite. Orange circles denote the sun, and the purple line denote the PBL height.



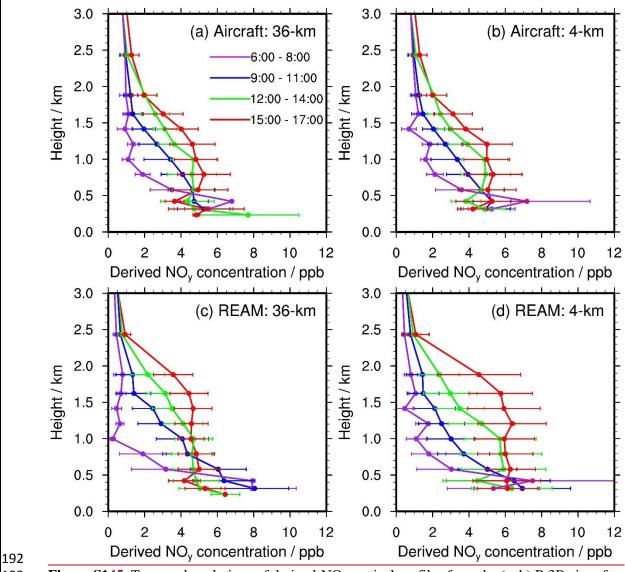
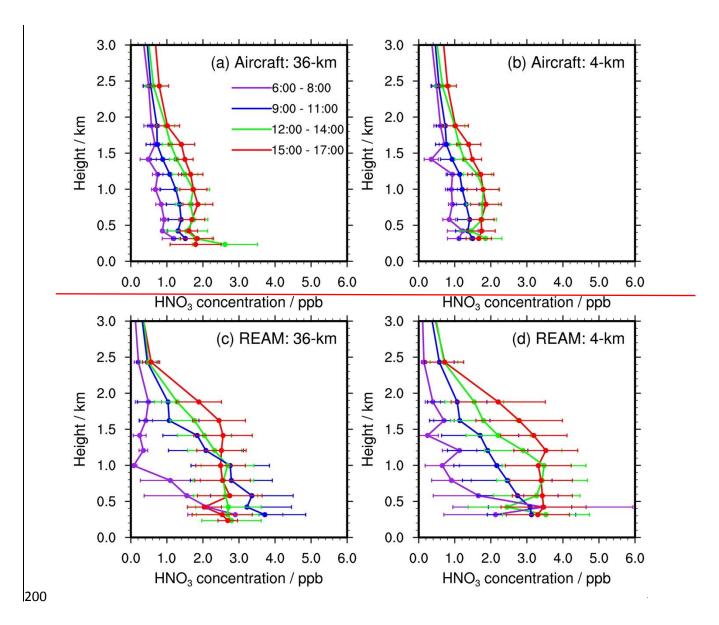




Figure S145. Temporal evolutions of derived-NO_v vertical profiles from the (a, b) P-3B aircraft 194 and (c, d) REAM simulations at (a, c) 36-km and (b, d) 4-km resolutions during the DISCOVER-AQ campaign. Error bars denote the corresponding standard deviations. Due to the limited 195 196 number of P-3B derived-NO_v observations and slightly different heights between 36- and 4-km grid cells from different WRF simulations (Table S2), small differences exist between the 36-km 197 198 and 4-km observations when we bin them vertically to REAM grid cells (see also Table 1). 199



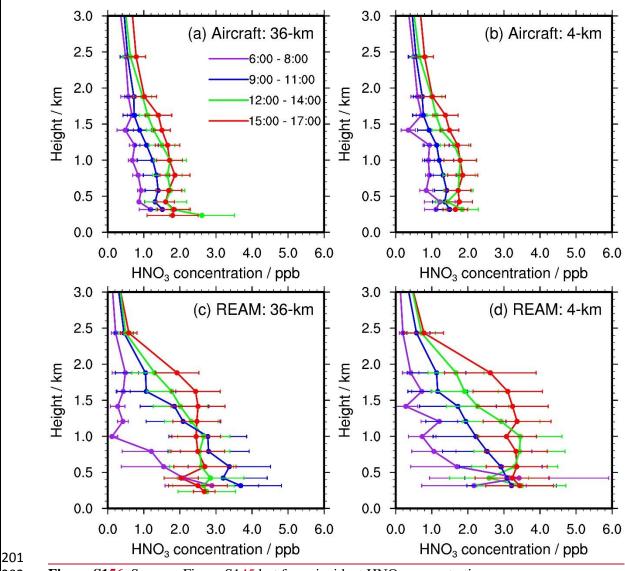
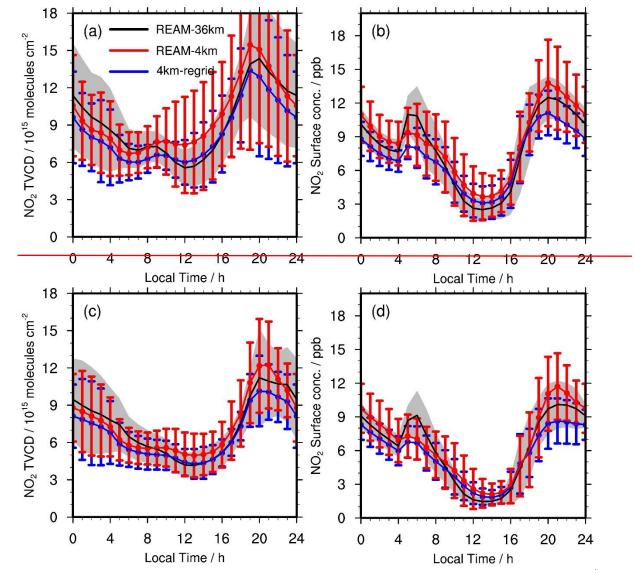
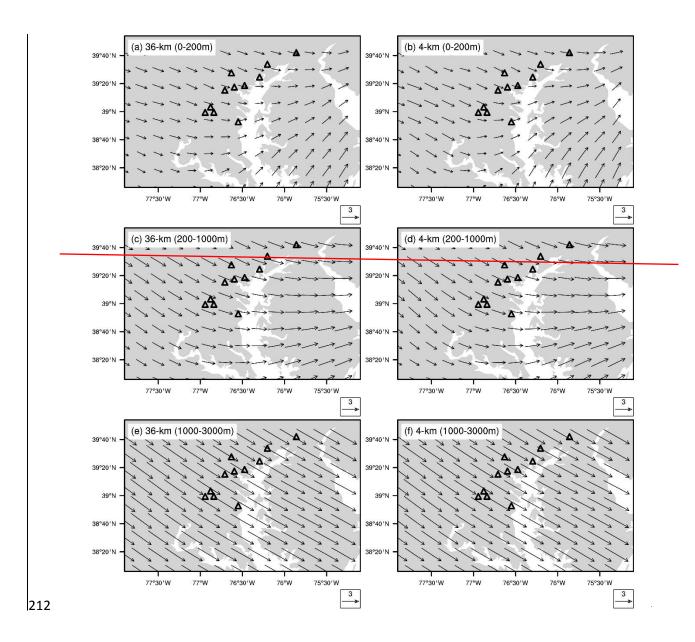


Figure S156. Same as Figure S145 but for coincident HNO₃ concentrations.



204 205

Figure S17. Comparisons of NO₂ (a, c) TVCDs and (b, d) surface concentrations over the 11 206 inland Pandora sites between the 4-km and 36-km REAM simulations on (a, b) weekdays and (c, 207 d) weekends for July 2011. "REAM-36km" (black lines) denotes the 36 km REAM simulation 208 results; "REAM 4km" (red lines) denotes the 4-km REAM simulation results; "4km regrid" 209 refers to the 36-km values by re-gridding the 4-km REAM simulation results into 36-km REAM 210 grid cells. Error bars denote standard deviations. 211



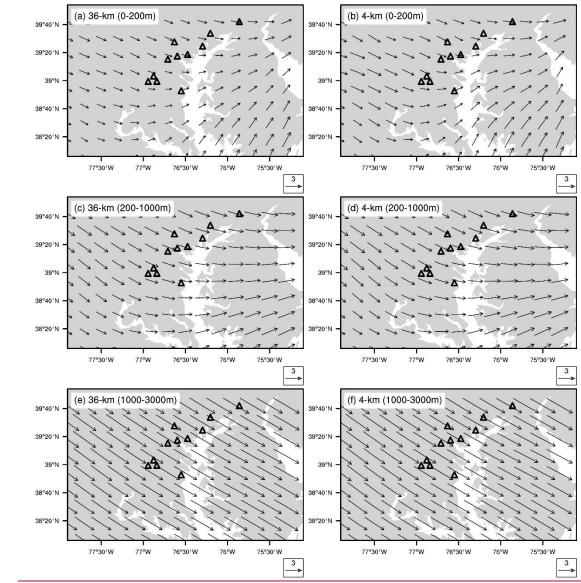
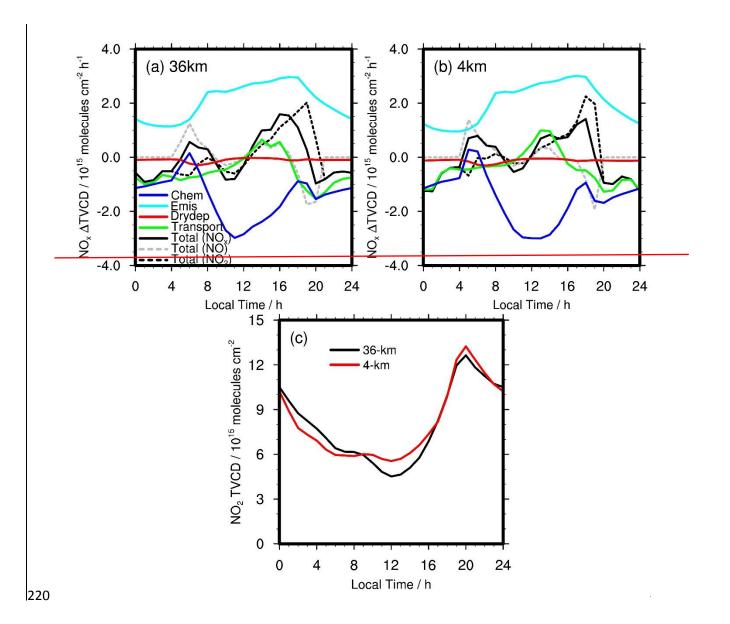
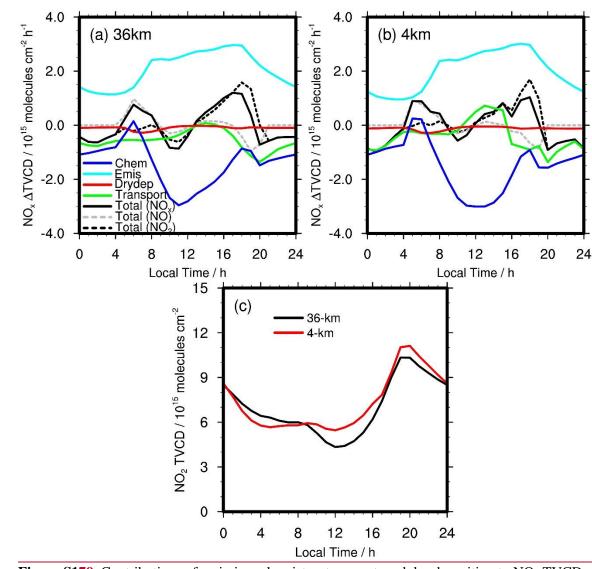
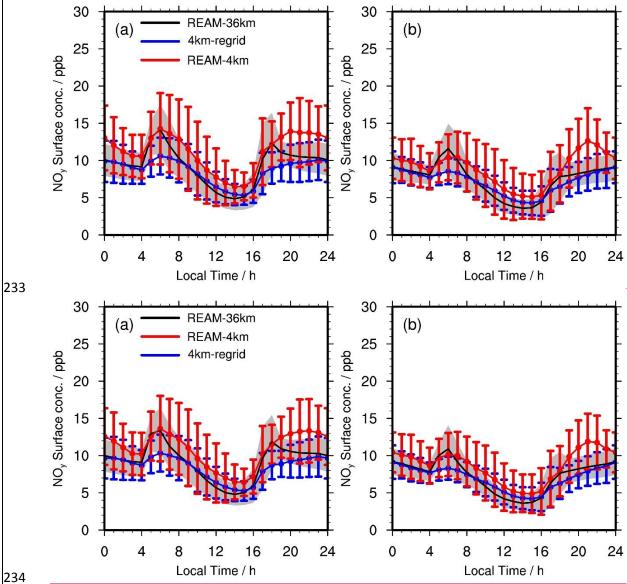


Figure S168. Comparisons of wind fields between the (a, c, e) 36-km and (b, d, f) nested 4-km WRF simulations for different height (AGL) bins in July 2011. (a) and (b) are for mean wind fields below about 200 m, (c) and (d) are for about 200 – 1000 m, and (e) and (f) are for about 1000 - 3000 m. Triangles denote the inland Pandora sites in Figure S1. The unit of wind speed is m s⁻¹.



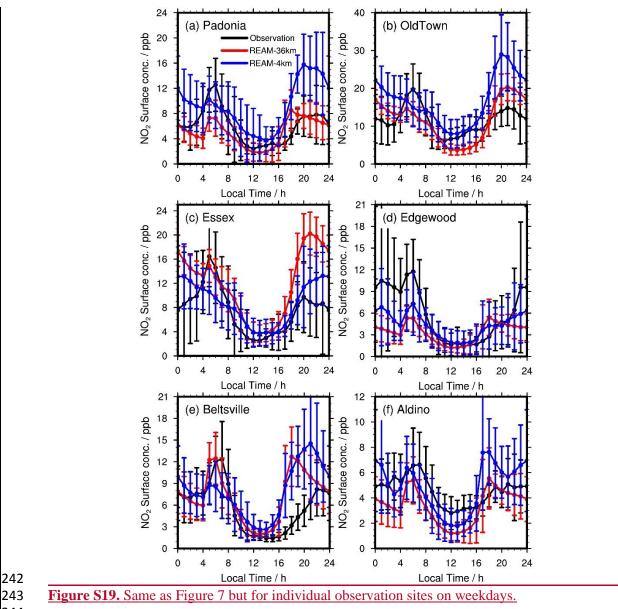


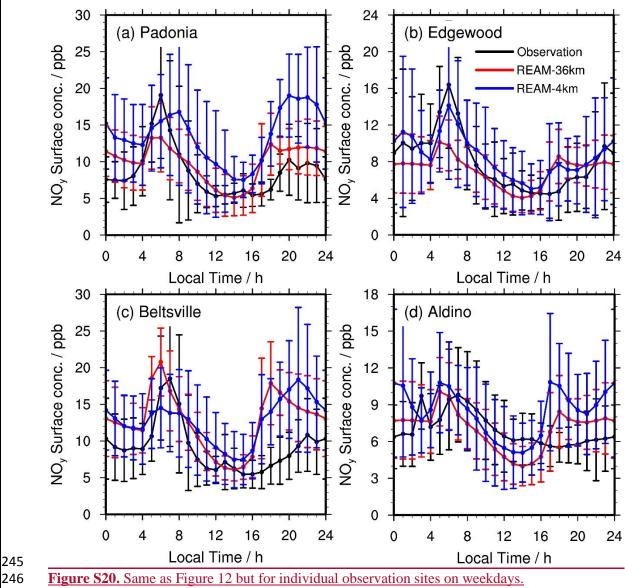
222 Figure S179. Contributions of emission, chemistry, transport, and dry deposition to NO_x TVCD diurnal variations over the six P-3B spiral sites (Figure \$1 and Table \$1) on weekdays in July 223 2011 for the (a) 36-km and (b) 4-km REAM simulations. "Chem" refers to net NO_x chemistry 224 production; "Emis" refers to NO_x emissions; "Drydep" denotes NO_x dry depositions; "Transport" 225 226 includes advection, turbulent mixing, lightning NOx production, and wet deposition. "Total (NO_x)" is the hourly change of NO_x TVCDs (\triangle (*TVCD*) = *TVCD*_{t+1} - *TVCD*_t). "Total (NO₂)" is 227 the hourly change of NO₂ TVCDs, and "Total (NO)" is the hourly change of NO TVCDs. (c), the 228 36-km and 4-km REAM simulated diurnal cycles of NO₂ TVCDs over the P-3B spiral sites on 229 230 weekdays in July 2011. The black line in (c) denotes the 36-km REAM simulation results, and the 231 red line denotes the 4-km REAM simulation results. 232



235 236

Figure S1820. Comparisons of mean surface NO_y concentrations at Padonia, Edgewood, Beltsville, and Aldino between the 4-km and 36-km REAM simulations on (a) weekdays and (b) 237 weekends for July 2011. "REAM-36km" (black lines) denotes the 36-km REAM simulation 238 results; "REAM-4km" (red lines) denotes the 4-km REAM simulation results; "4km-regrid" 239 refers to the 36-km values by re-gridding 4-km REAM simulation results into 36-km REAM grid 240 cells. Error bars denote standard deviations.





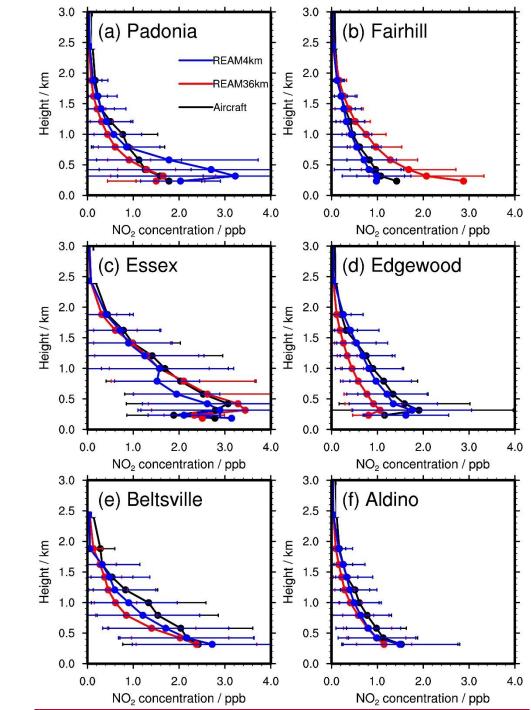


Figure S21. Comparison of NO₂ vertical profiles among P-3B aircraft observations and the 36-249 km and 4-km REAM simulations on weekdays in July 2011 for different spiral sites. Here we 250 calculate the average of all available weekday NO₂ vertical profiles for each spiral site but do not 251 252 consider the temporal evolutions shown in Figure 8.

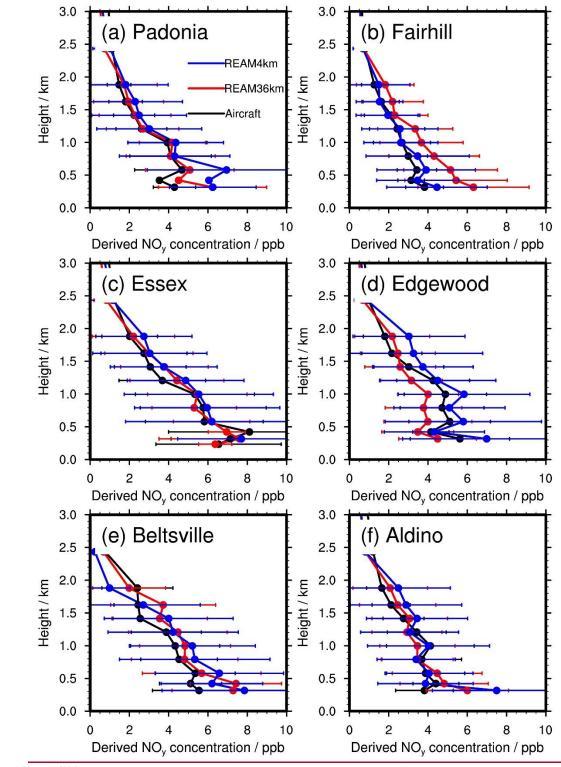
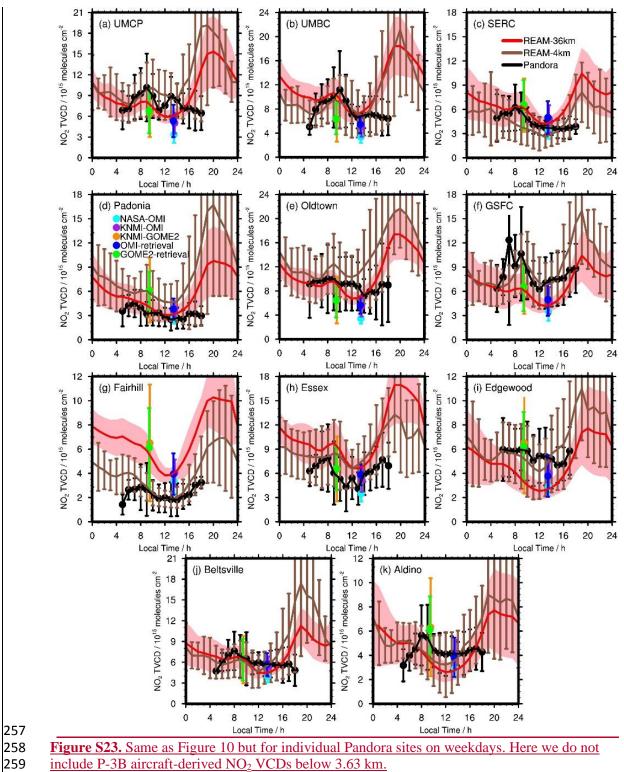
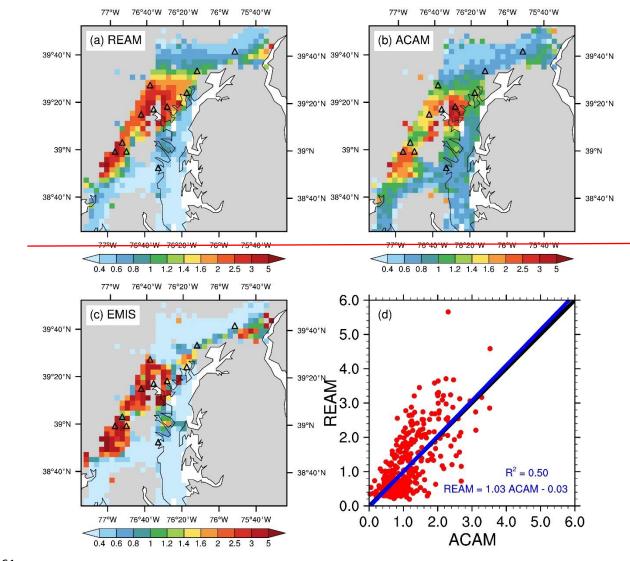




Figure S22. Same as Figure S21 but for derived-NO_y vertical profiles.









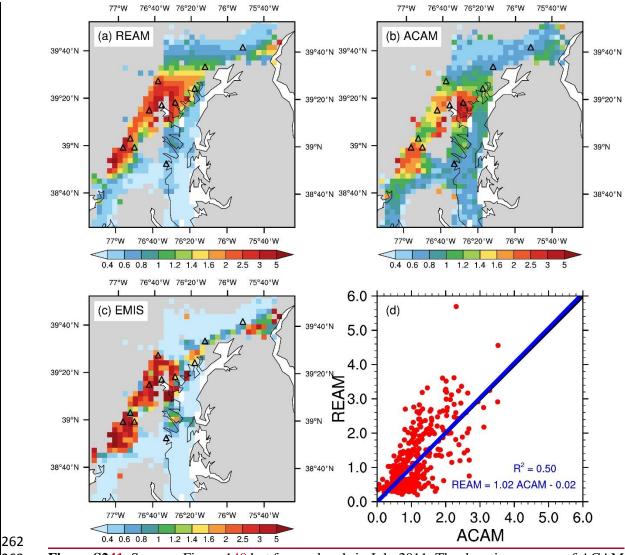
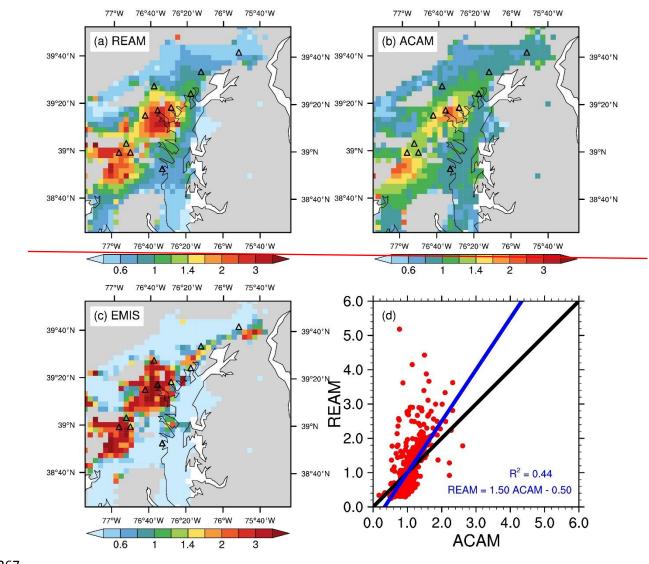
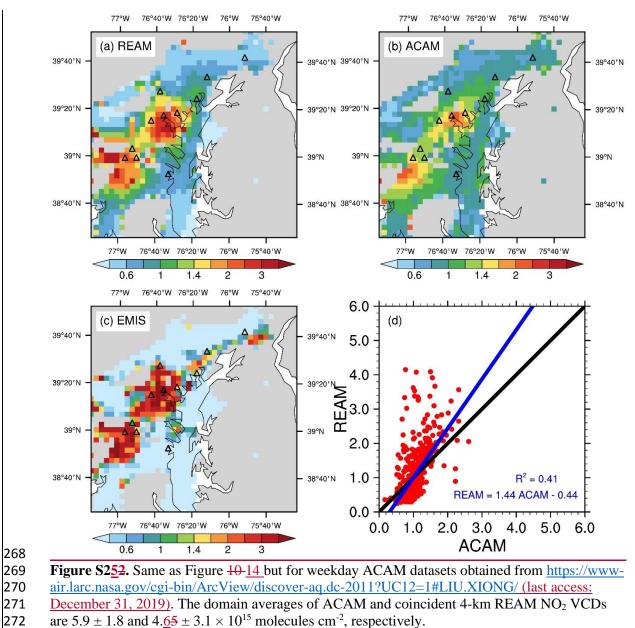
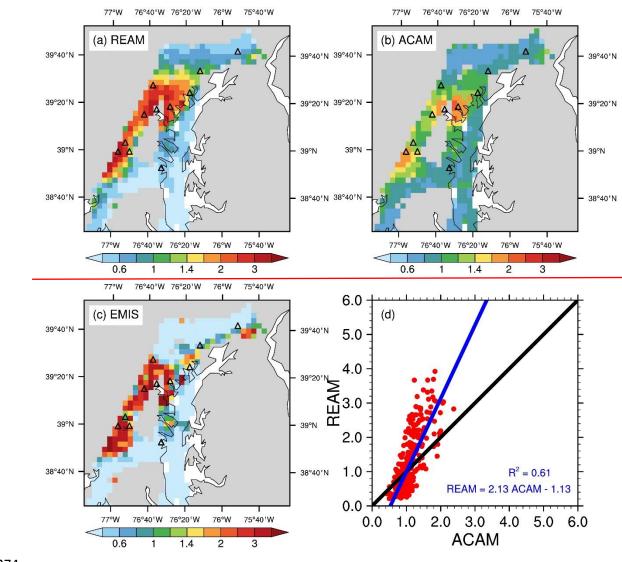


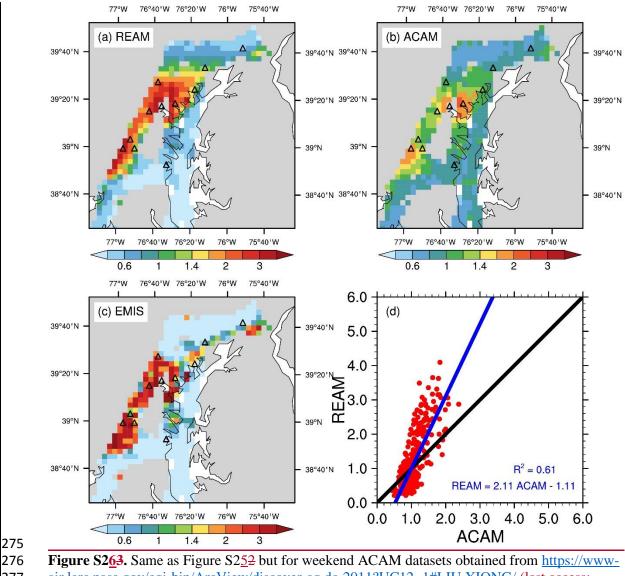
Figure S241. Same as Figure 140 but for weekends in July 2011. The domain averages of ACAM and coincident 4-km REAM NO₂ VCDs are 3.0 ± 1.7 and $3.3 \pm 2.78 \times 10^{15}$ molecules cm⁻², respectively.



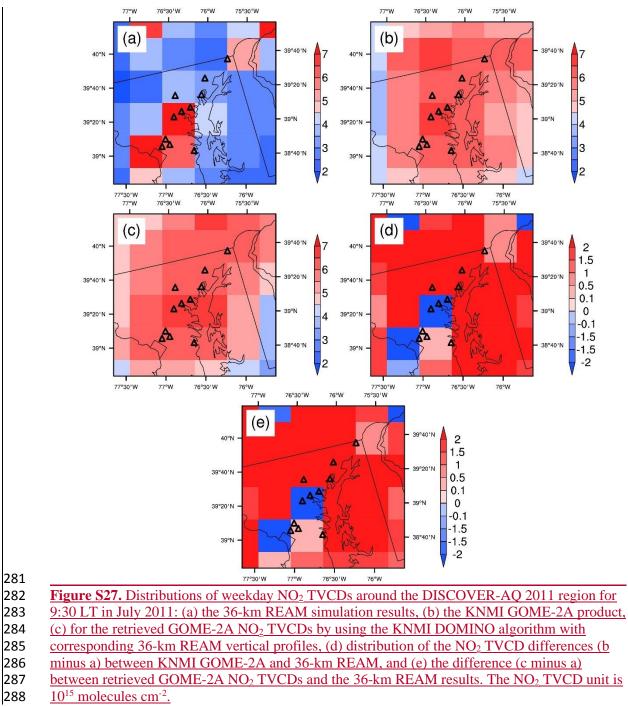








276Figure S263. Same as Figure S252 but for weekend ACAM datasets obtained from https://www277air.larc.nasa.gov/cgi-bin/ArcView/discover-aq.dc-2011?UC12=1#LIU.XIONG/ (last access:278December 31, 2019). The domain averages of ACAM and coincident 4-km REAM NO2 VCDs279are 4.7 \pm 1.4 and 3.4 \pm 2.7 \times 10¹⁵ molecules cm⁻², respectively.280



1	Supporting tables for
2	Comprehensive evaluations of diurnal NO2 measurements during
3	DISCOVER-AQ 2011: Effects of resolution dependent representation of NO _x
4	emissions
5 6 7 8	Jianfeng Li ^{1, a} , Yuhang Wang ^{1*} , Ruixiong Zhang ¹ , Charles Smeltzer ¹ , Andrew Weinheimer ² , Jay Herman ³ , K. Folkert Boersma ^{4, 5} , Edward A. Celarier ^{6, 7, b} , Russell W. Long ⁸ , James J. Szykman ⁸ , Ruben Delgado ³ , Anne M. Thompson ⁶ , Travis N. Knepp ^{9, 10} , Lok N Lamsal ⁶ , Scott J Janz ⁶ , Matthew G Kowalewski ⁶ , Xiong Liu ¹¹ , Caroline R. Nowlan ¹¹
9 10	¹ School of Earth and Atmospheric Sciences, Georgia Institute of Technology, Atlanta, Georgia, USA
11	² National Center for Atmospheric Research, Boulder, Colorado, USA
12	³ University of Maryland Baltimore County JCET, Baltimore, Maryland, USA
13	⁴ Royal Netherlands Meteorological Institute, De Bilt, the Netherlands
14	⁵ Wageningen University, Meteorology and Air Quality Group, Wageningen, the Netherlands
15	⁶ NASA Goddard Space Flight Center, Greenbelt, Maryland, USA
16	⁷ Universities Space Research Association, Columbia, Maryland, USA
17	⁸ National Exposure Research Laboratory, Office of Research and Development, U.S.
18	Environmental Protection Agency, Research Triangle Park, NC, USA
19	⁹ NASA Langley Research Center, Virginia, USA
20	¹⁰ Science Systems and Applications, Inc., Hampton, Virginia, USA
21	¹¹ Harvard-Smithsonian Center for Astrophysics, Cambridge, Massachusetts, USA
22	anow at Atmospheric Sciences and Clobal Change Division Desific Northwest National
23 24	^a now at: Atmospheric Sciences and Global Change Division, Pacific Northwest National Laboratory, Richland, Washington, USA
24 25	^b now at: Digital Spec, Tyson's Corner, VA, USA
25	now at Dignal opec, Tyson's Corner, VA, OSA
26	* Correspondence to Yuhang Wang (<u>yuhang.wang@eas.gatech.edu</u>)

Site #	Site name	Latitude / ° N	Longitude / ° W	Land type	NO _x emission in 36-km REAM ¹ / 10 ²¹ molecules km ⁻² s ⁻¹	NO _x emission in 4-km REAM / 10 ²¹ molecules km ⁻² s ⁻¹	Availability of P-3B aircraft observations ²	Availability of surface NO _y	Availability of surface NO ₂
1	UMCP	38.991	76.943	urban	10.9	19.5			
2	UMBC	39.255	76.709	urban	12.9	14.8			
3	SERC	38.880	76.550	rural/coastal	5.0	0.8			
4	Padonia	39.461	76.631	suburban	2.9	12.8	Y	Y	Y
5	Oldtown	39.291	76.596	urban	12.9	33.0			Y
6	GSFC	38.993	76.840	urban/suburban	5.0	6.2			
7	Fairhill	39.701	75.860	rural	6.6	0.6	Y		
8	Essex	39.311	76.474	coastal/urban	12.9	6.4	Y		Y
9	Edgewood	39.410	76.297	coastal/urban	1.7	1.3	Y	Y	Y
10	Beltsville	39.055	76.878	suburban	5.0	4.4	Y	Y	Y
11	Aldino	39.563	76.204	rural/suburban	1.7	4.2	Y	Y	Y

28 Table S1. Summary of information for the 11 inland Pandora sites in the DISCOVER-AQ campaign

¹ Here, NO_x emissions refer to the mean values in one week (Monday – Sunday). Since we scale weekend emissions based on weekday emissions in this study,

30 the relative differences among different sites and between the 36-km REAM and the 4-km REAM are the same for weekdays and weekends.

² "Y" denotes that P-3B aircraft observations were available at the corresponding site during the DISCOVER-AQ campaign. And blank indicates that no aircraft observations were available. Similar to the "availability of surface NO_y" and the "availability of surface NO₂."

Table S2. Setup of the 36-km and nested 4-km WRF simulations

	36-km WRF	Nested 4-km WRF
Horizontal resolution	36 km	Nested (36 km, 12 km, 4 km)
Domain center	40° N, 97° W	38.94° N, 75.76° W
Microphysics	WRF Single-Moment <u>6</u> 3-class simple ice scheme <u>(WSM6)</u>	Same as 36-km WRF
Surface layer	Revised MM5 Monin-Obukhov scheme	Same as 36-km WRF
Land surface	Unified Noah land-surface model	Same as 36-km WRF
Longwave radiation	RRTM scheme	Same as 36-km WRF
Shortwave radiation	Dudhia scheme	Same as 36-km WRF
Planetary boundary layer	Yonsei University (YSU) scheme	Same as 36-km WRF
Cumulus parameterization	Kain-Fritsch (new Eta) scheme	Kain-Fritsch scheme for outer domains (36-km and 12-km); no cumulus parameterization for the 4-km domain
Urban surface	3-category urban canopy model	Same as 36-km WRF

36 Table S3. Comparison of NO₂ TVCDs among different simulations and datasets during the DISCOVER-AQ campaign for 9:30 and 13:30 LT on

37 weekdays and weekends

	Wee	<u>kday</u>	Weekend		
	<u>9:30 LT¹</u>	<u>13:30 LT</u>	<u>9:30 LT</u>	<u>13:30 LT</u>	
REAM-36km ²	6.5 ± 1.1^3	4.9 ± 0.6	4.7 ± 0.7	3.6 ± 0.6	
REAM-4km	<u>7.0 ± 1.9</u>	<u>6.3 ± 2.0</u>	<u>5.3 ± 1.6</u>	<u>4.3 ± 1.5</u>	
Pandora	<u>6.5±1.8</u>	<u>5.3±1.0</u>	<u>4.9±0.9</u>	<u>3.7±0.5</u>	
<u>Flight</u>	<u>5.3</u>	<u>5.0</u>	<u>4.5</u>	<u>3.2</u>	
KNMI-GOME2	<u>6.3 ± 3.4</u>		<u>5.3 ± 1.4</u>		
GOME2-retrieval	<u>6.3 ± 2.5</u>		<u>4.1 ± 1.9</u>		
NASA-OMI		3.3 ± 0.8		2.9 ± 0.6	
KNMI-OMI		<u>4.6 ± 1.3</u>		3.2 ± 0.7	
OMI-retrieval		4.7 ± 1.4		3.2 ± 0.7	

38 $\frac{1}{1}$ For REAM simulations, we use the average of NO₂ TVCDs at 9:00 and 10:00 LT to represent the value at 9:30 LT, similar to those at 13:30 LT.

39 ² The dataset names have the same meaning as Figure 10, and the NO₂ TVCD values are the same as those shown in Figure 10.

40 ³ The unit is 10^{15} molecules cm⁻².