

1 **The increasing atmospheric burden of the greenhouse gas sulfur**
2 **hexafluoride (SF₆)**

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38 **Abstract**

39

40 We report a 40-year history of SF₆ atmospheric mole fractions measured at the Advanced
41 Global Atmospheric Gases Experiment (AGAGE) monitoring sites, combined with archived
42 air samples to determine emission estimates from 1978-2018. Previously we reported a global
43 emission rate of 7.3 ± 0.6 Gigagrams (Gg) yr⁻¹ in 2008 and over the past decade emissions
44 have continued to increase by about 24% to 9.04 ± 0.35 Gg yr⁻¹ in 2018. We show that
45 changing patterns in SF₆ consumption from developed (Kyoto Protocol Annex-1) to
46 developing countries (non-Annex-1) and the rapid global expansion of the electric power
47 industry, mainly in Asia, have increased the demand for SF₆-insulated switchgear, circuit
48 breakers and transformers. The large bank of SF₆ sequestered in this electrical equipment
49 provides a substantial source of emissions from maintenance, replacement and continuous
50 leakage. Other emissive sources of SF₆ occur from the magnesium, aluminium, electronics
51 industries and more minor industrial applications. More recently, reported emissions,
52 including those from electrical equipment and metal industries, primarily in the Annex-1
53 countries, have declined steadily through substitution of alternative blanketing gases and
54 technological improvements in less emissive equipment and more efficient industrial
55 practices. **Nevertheless, there are still demands for SF₆ in Annex-1 countries due to economic
56 growth, as well as continuing emissions from older equipment and additional emissions from
57 newly installed SF₆-insulated electrical equipment, although at low emissions rates.** In
58 addition, in the non-Annex-1 countries, SF₆ emissions have increased due to an expansion in
59 the growth of the electrical power, metal and electronics industries to support their continuing
60 development.

61 There is an annual difference of 2.5-5 Gg yr⁻¹ (1990-2018) between our modelled top-
62 down emissions and the UNFCCC reported bottom-up emissions, which we attempt to
63 reconcile through analysis of the potential contribution of emissions from the various
64 industrial applications which use SF₆. We also investigate regional emissions in East Asia
65 (China, S. Korea) and Western Europe and their respective contributions to the global
66 atmospheric SF₆ inventory. On an average annual basis, our estimated emissions from the
67 whole of China are approximately 10 times greater than emissions from Western Europe. In
68 2018, our modelled Chinese and Western European emissions accounted for ~36% and 3.1
69 %, respectively, of our global SF₆ emissions estimate.

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71 **Keywords**

72 *Chemical tracers; atmospheric dispersion; models atmospheric transport*

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79 **1. Introduction**

80 Of all the greenhouse gases regulated under the Kyoto Protocol, SF₆ is the most
81 potent, with a global warming potential (GWP) of 23,500 over a 100-year time horizon
82 (Myhre et al., 2013). In practical terms, this high GWP means that 1 ton of SF₆ released to the
83 atmosphere is equivalent to the release of 23,500 tonnes of carbon dioxide (CO₂). However,
84 the low atmospheric mixing ratio of SF₆ relative to CO₂ limits its current contribution to total
85 anthropogenic radiative forcing to about 0.2 % (Engel, Rigby, 2019). Nevertheless, with a
86 long atmospheric residence time of 3,200 years, almost all the SF₆ released so far will have
87 accumulated in the atmosphere and will continue to do so (Ravishankara et al., 1993).

88 Vertical profiles of SF₆ mixing ratios, collected from balloon flights up to an altitude
89 of about 37 km, indicated that there is very little loss of SF₆ due to photochemistry in the
90 troposphere and lower stratosphere (Harnish et al., 1996; Patra et al., (1997). Using an
91 improved atmospheric-chemical-transport model (Patra et al., 2018) reported significantly
92 older 'age of air' (AoA) in the stratosphere and Krol et al., (2018), based on a comparison of
93 six global transport models showed that upper stratospheric AOA varied from 4-7 years
94 among the models. It has been suggested that SF₆ may have a shorter atmospheric lifetime
95 ranging from 1937 ± 432 years (Patra et al.,1997), 580-1400 years (Ray et al., 2017) and
96 1120-1475 years (Kovács et al., 2017). However, these shorter, but still very long, SF₆
97 lifetimes would not significantly affect SF₆ emissions estimated from atmospheric trends
98 (Engel and Rigby, 2019) .Given the very long lifetime of SF₆, compared to the period of our
99 study, uncertainties in this term had a small influence on the outcome. For example, changing
100 the lifetime from 3000 to 1000 years changed the derived emissions by around 1%, which is
101 smaller than the derived uncertainties.

102 Since the 1970s, SF₆ has been used mainly in high voltage electrical equipment as a
103 dielectric and insulator in gas-insulated switchgear, gas circuit breakers, high voltage lines
104 and transformers. Sales compiled from 1996-2003 by producers in Europe, Japan, USA and
105 South Africa (not including China and Russia) showed that, on an annual average basis, 80%
106 of the SF₆ produced during this period was consumed by electric utilities and equipment
107 manufacturers for electric power systems (EPA, 2018). Percentage sales, averaged from
108 1996-2003, for other end-use applications included the magnesium industry (4%), electronics
109 industry (8%), and uses relating to the adiabatic properties of SF₆ (3%) e.g., incorporating
110 SF₆ into tyres, tennis balls and the soles of trainers as a gas cushioning filler (Palmer, 1996).
111 For example, in 1997 Nike used 277 tonnes (~0.25 Gigagrams, Gg) of SF₆ as a filler in its
112 shoes (Harnish and Schwarz, 2003). Other uses in particle accelerators, optical fibre
113 production, lighting, biotechnology, medical refining, pharmaceutical, laboratory, university
114 research and sound-proof windows accounted for around 5% of sales (Smythe, 2004).

115 Emissions from electrical equipment can occur during production, routine maintenance,
116 refill, leakage, and disposal (Neimeyer and Chu, 1992; Ko et al., 1993). Random failure or
117 deliberate or accidental venting of equipment may also cause unexpected and rapid high
118 levels of emissions. For example, a ruptured seal caused the release of 113 kg of SF₆ in a
119 single event in 2013 (Scottish Hydro Electric, 2013). We assume that such random events are
120 generally not recorded when tabulating bottom-up emission estimates, which would lead to
121 an under-estimate in the reported inventories.

122 Historically, significant emissions of SF₆ occurred in magnesium smelting, where it was
123 used as a blanketing gas to prevent oxidation of molten magnesium, in the aluminium
124 industry, also as a blanketing gas, and in semi-conductor manufacturing (Maiss and
125 Brenninkmeijer, 1998). These industries and the electrical power industry accounted for the
126 majority of SF₆ usage in the USA (Ottinger et al., 2015). A report on limiting SF₆ emissions
127 in the European Union also provided estimates of emissions from sound-proof windows (60
128 Mg) and car tyres (125 Mg) in 1998, although these applications appear to have been largely
129 discontinued due to environmental concern (Schwarz, 2000).

130 Sulphur hexafluoride has also been used as a tracer in atmospheric transport and
131 dispersion studies (Collins et al., 1965; Saltzman et al., 1966; Turk et al., 1968; Simmonds et
132 al., 1972; Drivas et al., 1972; Drivas and Shair, 1974). The combined SF₆ emissions from
133 reported tracer studies (Martin et al., 2011) were approximately 0.002 Gg. Unfortunately, the
134 amounts of tracer released are often not reported and we conservatively assume that these
135 also amounted to ~ 0.002 Gg, providing a total estimate of about 0.004 Gg (4 tonnes)
136 released from historical SF₆ tracer studies. Emissions from natural sources are very small
137 (Busenberg and Plummer, 2000; Vollmer and Weiss, 2002; Deeds et al., 2008).

138 The earliest measurements of SF₆ in the 1970s reported a mole fraction of < 1 pmol mol⁻¹
139 (or ppt, parts per trillion) (Lovelock, 1971; Krey et al., 1977; Singh et al., 1977, 1979).
140 Intermittent campaign-based measurements during the 1970s and 1980s reported an
141 increasing trend. However, it was not until the 1990s that a near-linear increase in the
142 atmospheric burden, throughout the 1980s, was reported (Maiss and Levin, 1994; Maiss et
143 al., 1996, Geller et al., 1997). Fraser et al. (2004), described gas chromatography-electron
144 capture detection (GC-ECD) measurements of SF₆ at Cape Grim, Tasmania and noted a long-
145 term trend of 0.1 pmol mol⁻¹ yr⁻¹ in the late 1970s increasing to 0.24 pmol mol⁻¹ yr⁻¹ in the
146 mid-1990s. However, after 1995 the annual average growth rate from 1996-2000 declined by
147 12.5% to 0.21 pmol mol⁻¹ yr⁻¹, coincident with a ~ 32% decrease in annual sales and prompt
148 releases of SF₆ over this same time period (as noted in Table S2 of the Rand report).

149 Subsequent reports noted a continuing growth in global mole fractions, with an average
150 growth rate of 0.29 ± 0.02 pmol mol⁻¹ yr⁻¹ after 2000 (Rigby et al., 2010), reaching 6.7 pmol mol⁻¹
151 at the end of 2008 (Levin et al., 2009). This increase in the atmospheric burden of SF₆ was also
152 reported by Elkins and Dutton (2009). Measurement of SF₆ in the lower stratosphere and upper
153 troposphere was reported to be 3.2 ± 0.5 pmol mol⁻¹ at 200 mbar in 1992 (Rinsland et al., 1993).
154 These atmospheric observations have been used to infer global emissions rates ('top-down'
155 estimates). Geller et al., (1997) derived a global emission rate of 5.9 ± 0.2 Gg yr⁻¹ in 1996, which
156 by 2008 had increased to 7.2 ± 0.4 Gg yr⁻¹ (Levin et al., 2010) or 7.3 ± 0.6 Gg yr⁻¹ (Rigby et al.,
157 2010), and to 8.7 ± 0.4 Gg yr⁻¹ by 2016 (Engel, Rigby, 2019).

158 Regional inverse modelling studies indicated that emissions have increased substantially
159 from non-Annex-1 parties to the UNFCCC, particularly in eastern Asia, and that these
160 increases have offset the reduction in emissions from Annex-1 countries (Rigby et al., 2011,
161 2014; Fang et al., 2014). Rigby et al., (2010) showed an increasing trend in emissions from
162 Asian countries growing from 2.7 ± 0.3 Gg yr⁻¹ in 2004–2005 to 4.1 ± 0.3 Gg yr⁻¹ in 2008.
163 This rise was large enough to account for all the global emissions growth between these two
164 periods. Similarly, Fang et al. (2014) found that eastern Asian emissions accounted for
165 between 38 ± 5 % and 49 ± 7 % of the global total between 2006 and 2012, with China the

166 major contributor of emissions from this region. Consistent regional estimates, within the
167 uncertainties, were also reported for China Vollmer et al., (2009); 0.8 (0.53-1.1) Gg yr⁻¹ from
168 October 2006-March 2008; Kim et al., (2010); 1.3 (0.23-1.7) Gg yr⁻¹ in 2008 and Li et al.,
169 (2011); 1.2 (0.9 – 1.7) Gg yr⁻¹ from November 2007-December 2008. Emissions from other
170 Asian countries were found to be substantially smaller by Li et al., (2011) with South Korea
171 emitting 0.38 (0.33-0.44) Gg yr⁻¹ in 2008 and Japan 0.4 (0.3-0.5) Gg yr⁻¹. For North America,
172 SF₆ emission estimates of 2.4 ± 0.5 Gg yr⁻¹ were inferred in 1995 (Bakwin et al., 1997),
173 whereas Hurst et al., (2006) reported emissions of 0.6 ± 0.2 Gg yr⁻¹ in 2003, consistent with
174 an expectation of declining Annex-1 emissions during this period. Top-down SF₆ emissions
175 for Western Europe have been reported by Ganesan et al., (2014), indicating larger modelled
176 emission estimates than those reported to the UNFCCC.

177

178 **1. Methods.**

179 Here, we use a 40-year (1978-2018) time series of SF₆ measurements made in situ, and in
180 archived air samples, in combination with a global atmospheric box model and inverse
181 modelling techniques to examine how the growth rate of SF₆ has changed, and we estimate
182 global and regional emissions in a top-down approach.

183

184 **1.1 In situ AGAGE measurements**

185 In situ high frequency (every 30 mins) measurements were recorded at Cape Grim,
186 Tasmania beginning in 2001 using a modified Shimadzu gas chromatograph GC fitted with a
187 Ni⁶³ electron capture detector ECD, (Fraser et al., 2004). Beginning in 2003, newly developed
188 GC-mass spectrometers (GC-MS) equipped with an automated sample processing system,
189 known as the ‘Medusa’ were progressively deployed at the AGAGE stations, thereby providing
190 calibrated SF₆ measurements every 2 hours (Miller et al., 2008; Arnold et al., 2012). Here we
191 use “Medusa” measurements through 2018, acquired at the five core AGAGE stations; Mace
192 Head, Ireland (beginning in 2003); Trinidad Head, California (beginning in 2005); Ragged
193 Point, Barbados (beginning in 2005); Cape Matatula, American Samoa (beginning in 2006);
194 and Cape Grim, Tasmania (beginning in 2005). At Monte Cimone, Italy (an affiliated AGAGE
195 station), SF₆ measurements were measured every 15 minutes using a GC-ECD (Maione et al.,
196 2013). Each real air sample is bracketed with a calibrated (NOAA-2014 scale) air sample
197 analysis resulting in 2 measurements per hour, with a precision of 0.6%.

198 A complete description of the equipment used in the AGAGE station network is given in
199 Prinn et al., (2000, 2018). We combine these measurements with the Medusa-GC-MS analysis
200 of samples from the Cape Grim Air Archive (GCAA) and a collection of Northern Hemisphere
201 (NH) archived air samples to extend the time series back to 1978 (Rigby et al., 2010). Estimated
202 uncertainties during propagation of calibration standards from Scripps Institution of
203 Oceanography (SIO) to the AGAGE measurement sites was ~0.6% with a calibration scale
204 uncertainty of ~2.0% (Prinn et al., 2018). All archived air and in situ measurements are reported
205 on the SIO-05 calibration scale. The difference between the SIO and National Oceanic and
206 Atmospheric Administration (NOAA) calibration scales is <0.5% (0.03 pmol mol⁻¹) (Rigby et
207 al., 2010).

208 Measurements of SF₆ from the UK Deriving Emissions linked to Climate Change network
209 (UK, DECC. <https://www.metoffice.gov.uk/research/approach/monitoring/atmospheric-trends/index>)
210 started in 2012 at Tacolneston (52.5° N, 1.1° E) and Ridge Hill (52.0° N, 2.5° W), and later in
211 2013 at Bilsdale (54.4° N, 1.2° W) and Heathfield (51.0° N, 0.2° E), using Agilent GC-ECDs
212 (Stanley et al. 2018; Stavert et al. 2019). At these four sites, SF₆ measurements were acquired
213 every 10 minutes and air samples are bracketed with calibrated air samples. In addition to the
214 GC-ECD at Tacolneston, a Medusa GC-MS was installed at the site and has been measuring
215 SF₆ since 2012. The GC-ECD at Tacolneston was decommissioned in spring 2018. All
216 calibration gases are on the same scale as the AGAGE stations. Stanley et al., (2018) and
217 Stavert et al., (2019) provide a complete description of the measurement capabilities at the UK
218 sites.

219

220 **2. Bottom-up emission estimates**

221 We compare our model-derived top-down emissions with bottom-up estimates, using
222 reports from the 43 Annex-1 countries that submit annual emissions to the UNFCCC
223 ([unfccc.int/process-and-meetings/transparency-and-reporting/reporting-and-review-under-the-](http://unfccc.int/process-and-meetings/transparency-and-reporting/reporting-and-review-under-the-convention/greenhouse-gas-inventories-annex-i-parties/national-inventory-submissions-2019)
224 [convention/greenhouse-gas-inventories-annex-i-parties/national-inventory-submissions-2019](http://unfccc.int/process-and-meetings/transparency-and-reporting/reporting-and-review-under-the-convention/greenhouse-gas-inventories-annex-i-parties/national-inventory-submissions-2019), last
225 access: May 1, 2019). This contrasts with the non-Annex-1 countries that are not required to
226 report to the UNFCCC (2010); however, some non-Annex-1 countries do voluntarily submit
227 annual emissions, whereas others report infrequently. For infrequent reporting countries we
228 have linearly interpolated emissions for missing years to provide revised non-Annex-1
229 emissions. Acknowledging that these bottom-up estimates will have large uncertainties, we
230 see a substantial increase in total emissions from non-Annex-1 countries after 2005, with 50-
231 80% from China. We also compare our estimates with those estimated in EDGAR v4.2 from
232 1970-2010 (EDGAR, 2010).

233 In the next section we compile bottom-up emissions estimates based on the usage and
234 release of SF₆ in the electrical power, metal and electronics industries. Here we follow the
235 approach of previous publications where SF₆ emissions are scaled to electrical production
236 (Fang et al., 2013; Victor and MacDonald, 1999) and attempt to calculate potential emissions
237 from the electrical power industry in China and the rest of the World (ROW) using reported
238 emissions factors for each region.

239

240 **3.1 Calculation of SF₆ emissions from the electrical power, metal and electronics** 241 **industries in China and the Rest of the World (ROW).**

242

243 3.1.1. Electrical Power

244 Chinese SF₆ emissions, mainly from electrical equipment, account for 60-72% of total
245 emissions from the East Asian region (Fang et al., 2014). Following the method of Zhou et al.
246 (2018), we first determine SF₆ consumption (Table 1) from the Chinese electric power
247 industry, using an initial filling factor (FF) of 52 t/GW (range 40-66 t/GW) and then calculate
248 emissions using the highest suggested emission factors, EFs (8.6% manufacture and
249 installation, 4.7% operation and maintenance). For the ROW we also use a median FF of
250 52t/GW and a 12% loss during manufacture and installation of new equipment and assume

251 3% loss from banked SF₆ in electrical equipment in 1980 and then decreasing linearly to 1%
252 in 2018, reflecting the change from older to newer equipment, with the reduced leakage of
253 SF₆ (Olivier and Bakker, 1999).

254

255 3.1.2 Magnesium Industry

256 In the magnesium industry (dye casting, sand casting and recycling), where SF₆ is used as
257 a cover or blanketing gas to prevent oxidation, it is assumed that emissions are equal to
258 consumption and all the SF₆ historically used in the magnesium industry has been emitted
259 (<https://www.ipcc-nggip.iges>). The consumption of SF₆ in magnesium production in China
260 was apparently halted after 2010 and largely replaced with SO₂. (National Bureau of
261 Statistics, 2017). Average annual sales of SF₆ to the magnesium industry were estimated to be
262 ~0.25 Gg yr⁻¹ from 1996-2003 (Smythe, 2004). Given current regulations and the availability
263 of substitute blanketing gases and the assumption that China and Russian producers use SO₂
264 as the preferred blanketing gas, we assume that current emissions from the magnesium
265 industry are equal to or less than the 1996-2003 average of about 0.25 Gg yr⁻¹.

266 3.1.3 Aluminium Industry

267 For the aluminium industry, historical emissions of SF₆ are poorly understood, as it is
268 generally assumed to be largely destroyed during the production process by reaction with the
269 aluminium (Victor and MacDonald, 1994); nevertheless, any surviving SF₆ will clearly be
270 emitted (IPCC, 1997). Maiss and Brenninkmeijer (1998) roughly quantified SF₆ consumption
271 from aluminium degassing (USA and Canada) and SF₆-insulated windows (Europe,
272 predominately Germany) and numerous small specialized applications to about 450 t yr⁻¹ in
273 1995. Since the use of SF₆ in these applications have been substantially reduced or eliminated
274 in the Annex-1 countries, we assume that current global emissions primarily from aluminium
275 degassing are unlikely to be greater than 0.2 Gg yr⁻¹.

276 3.1.4 Electronics Industry

277 Sulphur hexafluoride is used as a general etching agent in the electronics and
278 semiconductor industry including the production of thin film transistor liquid crystal displays
279 (TFT-LCDs) and in the cleaning of Chemical Vapor Deposition Chambers (CVD). Fang et.
280 (2013) reported emissions of 0.15 Gg in 2005 and 0.4 Gg in 2010 from the semiconductor
281 industry, which has rapidly expanded in China, and emissions from this industry were
282 reported to be 0.2 – 0.25 Gg yr⁻¹ during 2004-2011 (Cheng et al., 2013). Also annual average
283 consumption by the semiconductor industry from 2012-2018 was reported to be 0.51 Gg yr⁻¹
284 (range 0.41-0.55 Gg), (World semiconductor council, 2020). Due to commercial
285 confidentiality, there is very little information on the consumption of SF₆ in electronics
286 manufacturing. However, Asian electronics industries, which dominate TFT-LCD
287 production, have adopted substitute gases, mainly nitrogen trifluoride (NF₃), carbon
288 tetrafluoride (CF₄) and HFC-134a (CH₂FCF₃) in preference to SF₆ in recent years. We
289 therefore assume that global emissions of SF₆ from these industries are in the range 0.15 –
290 0.55 Gg yr⁻¹ in the absence of any new information.

291 3.1.5 Production

292 We also need to consider losses of SF₆ that occur during production. Fugitive emissions
293 during SF₆ production were estimated to be 0.5% for developed countries (IPCC, 2006).
294 Chinese SF₆ production accounts for ~50% of global production and Fang et al. (2013)
295 suggested an EF of 2.2% (1.7-3.3%) for China. We use these EFs for China and an EF of
296 0.5% for the rest of the world to estimate an average annual SF₆ loss from production of ~0.1
297 Gg yr⁻¹ (1990-2018).

298 The combined emissions from SF₆ production, magnesium, aluminium and electronics
299 industries estimated above are approximately 1.1 Gg yr⁻¹, which are prone to large
300 uncertainties. This assumes electronics emissions of 0.55 Gg yr⁻¹, the highest reported by the
301 World semiconductor council.

302

303 **4. Top-down global emissions estimates**

304 Global emissions were derived using a two-dimensional box model of the atmosphere and
305 a Bayesian inverse method. The AGAGE 12-box model has been used extensively for global
306 emissions estimation and is described in Cunnold et al., (1978, 1996) and Rigby et al.,
307 (2013). The model solves for advective and diffusive fluxes between four zonal average
308 ‘bands’ separated at 30° north and south and the equator, and between three vertical levels
309 separated at 500 hPa and 200 hPa. A Bayesian inverse modelling approach was adopted that
310 constrained emissions growth rate *a priori*, as described in Rigby et al., (2011; 2014) and
311 used most recently to derive SF₆ emissions in Engel, Rigby et al., (2019). Briefly, the
312 approach assumed *a priori* that emissions did not change from one year to the next, with a
313 Gaussian 1-sigma uncertainty in the emissions growth rate set to 20% of the maximum
314 EDGAR v4.2 emissions. The inversion then uses an analytical Bayesian method to find a
315 solution that best fits the observations and this prior constraint. This approach was chosen so
316 that independent constraints on absolute emissions magnitudes (e.g. as in Rigby et al., 2010),
317 which were not available for the entire time period, were not required. Following Rigby et al.
318 (2014), uncertainties applied to the in situ data were assumed to be equal to the variability in
319 the monthly baseline data points, representing the sum of measurement repeatability and a
320 model-data ‘mismatch’ term parameterising the inability of the model to resolve sub-monthly
321 timescales. For the archive air data, this mismatch uncertainty was taken to have the same
322 relative magnitude as the average mismatch error found during the in situ data period. This
323 term was added to the estimated measurement repeatability of the archive air samples. The
324 influence of these uncertainties, and those of the prior constraint, was propagated through to
325 the *a posteriori* emissions estimate, the uncertainty in which was augmented by an additional
326 term representing the uncertainty in the calibration scale (2%, applied as described in Rigby
327 et al. 2014).

328 **4.1. Regional emission estimates using the UK Met Office (InTEM), Empa (EBRIS) and** 329 **Urbino (FLITS) inverse modelling frameworks.**

330

331 Three different inverse methods, (1) Inverse Technique for Emission Modelling (InTEM),
332 (2) Swiss Federal Laboratories for Materials Science and Technology (Empa) Bayesian
333 Regional Inversion System (EBRIS), (3) FLexpart Inversion iTalian System (FLITS),
334 Urbino, Italy and two different chemical transport models and were used to estimate regional
335 SF₆ emissions. A brief description of the three inverse methods is given below and a more

336 detailed description of the InTEM and EBRIS models is provided in the supplementary
337 information.

338 InTEM. (Arnold et al., 2018) uses the NAME (Numerical Atmospheric dispersion
339 Modelling Environment, V7.2) [Jones et al., 2007] atmospheric Lagrangian transport model.
340 NAME is driven by re-analysis 3-D meteorology from the UK Met Office Unified Model
341 (Cullen, 1993). We provide estimated emissions for Western Europe (United Kingdom,
342 Ireland, Benelux countries (Belgium, the Netherlands, and Luxembourg), Germany, France,
343 Denmark, Switzerland, Austria, Spain, Italy, and Portugal) and, in a separate analysis,
344 emission estimates for China, using observations recorded at the Gosan station on Jeju Island,
345 South Korea (33°N, 126°E). Gosan receives air masses mainly from eastern mainland China
346 during the winter months, with winds from the north-northwest (Rigby et al., 2019; Fang et
347 al., 2013). We subsequently scale SF₆ emissions to a China total by population.

348

349 EBRIS. (Henne et al., 2016) employs source sensitivities as derived from the Lagrangian
350 particle dispersion model FLEXPART (Version 9.1; Stohl et al., 2005) and observed
351 atmospheric concentrations to optimally estimate spatially resolved surface emissions to the
352 atmosphere. Here, EBRIS was applied to Western Europe and provided country/region
353 estimates of *a posteriori* SF₆ emissions.

354 FLITS. (FLexpart Inversion iTalian System. U, Urbino), another modelling approach has
355 been used for a regional inversion. The model is based on an inversion approach developed by
356 Stohl et al. (2009). The modelling cascade is composed of the Lagrangian particle dispersion
357 model (LPDM) FLEXPARTv9.1 (<http://www.flexpart.eu>, downloaded 13 May 2019), in
358 conjunction with in situ high-frequency observations from four atmospheric monitoring sites
359 and a Bayesian inversion technique. Here, FLEXPART was driven by operational 3-hourly
360 meteorological data from the European Centre for Medium-Range Weather Forecasts
361 (ECMWF) at 1°×1° latitude and longitude resolution, from 2013 to 2018. We run the model in
362 backward mode, releasing from each measurement sites and every three hours, 40,000 particles
363 followed backward in time for 20 days. Due to the long atmospheric lifetime of SF₆, the model
364 simulation does not account for atmospheric removal process. For the West European *a priori*
365 emission field we disaggregated 2 Kt yr⁻¹ of SF₆ emissions within each country borders
366 according to a gridded population density data set (CIESIN, Center for International Earth
367 Science Information Network, www.ciesin.org), and we set 200% of uncertainty of the
368 emissions for every grid cells. Parametrisation details used here are described in Graziosi et
369 al., 2015.

370 In Table 2 we provide details of the East Asian setup of the inversion system (InTEM) and in
371 Table 3 we provide details of the European setup of the inversion systems (InTEM, EBRIS,
372 FLITS).

373

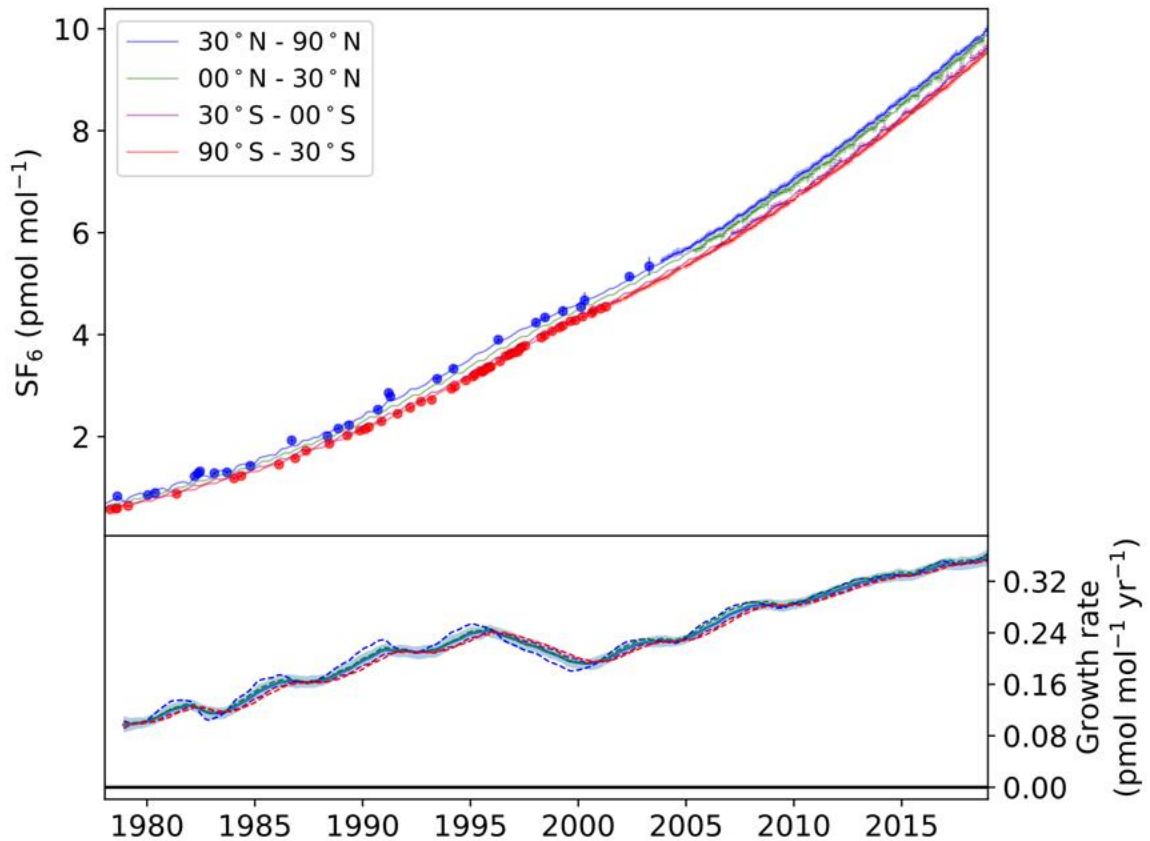
374 5. Results

375

376 Figure 1 and Table 4 shows the AGAGE SF₆ mole fractions from 1978-2018, averaged
377 into semi-hemispheres. In the lower panel of Fig.1, we report the annual SF₆ growth rate
378 increasing from 0.097 ± 0.013 pmol mol⁻¹ yr⁻¹ in 1978 to reach an early maximum average
379 growth rate in 1995 of 0.24 ± 0.01 pmol mol⁻¹ yr⁻¹ (a Kolmogorov-Zurbenko filter was used

380 to estimate annual mean growth rates, as described in Rigby et al. (2014). The growth rate
 381 then gradually drops to 0.19 ± 0.01 pmol mol⁻¹ yr⁻¹ in 2000, before increasing to reach $0.36 \pm$
 382 0.01 pmol mol⁻¹ yr⁻¹ in 2018. Between 1978 and 2018, the SF₆ loading of the atmosphere has
 383 increased by a factor of about 15. Assuming a radiative efficiency of 0.57 W m⁻² nmol mol⁻¹
 384 (WMO, 2018), SF₆ contributed around 5.5 ± 0.1 mW m⁻² in 2018 to global radiative forcing.
 385 In the supplementary material (Fig. S1), we show the model/measurement comparison for the
 386 AGAGE 12-box model.

387



388

389 Figure 1. Observed and model-derived SF₆ mole fractions and annual growth rates from the
 390 AGAGE 12-box model. Upper panel shows measured atmospheric SF₆ mole fractions in each
 391 semi-hemisphere (points with 1-sigma error bars) and archived air samples collected from
 392 1978 in the NH (blue filled circles) and archived air samples collected at Cape Grim
 393 Tasmania in the SH (red filled circles). Solid lines indicate modelled mole fractions using the
 394 mean emissions derived in the global inversion. Semi-hemispheric averages for both the
 395 model and data are shown for 30°–90° N (blue), 0° N–30° N (green), 30° S–0° S (purple) and
 396 90° S–30° S (red). The lower panel shows the model-derived growth rate, smoothed with an
 397 approximately 1-year filter, for each semi-hemisphere (dotted lines), and the global mean and
 398 its 1σ uncertainty (solid line, and shading, respectively).

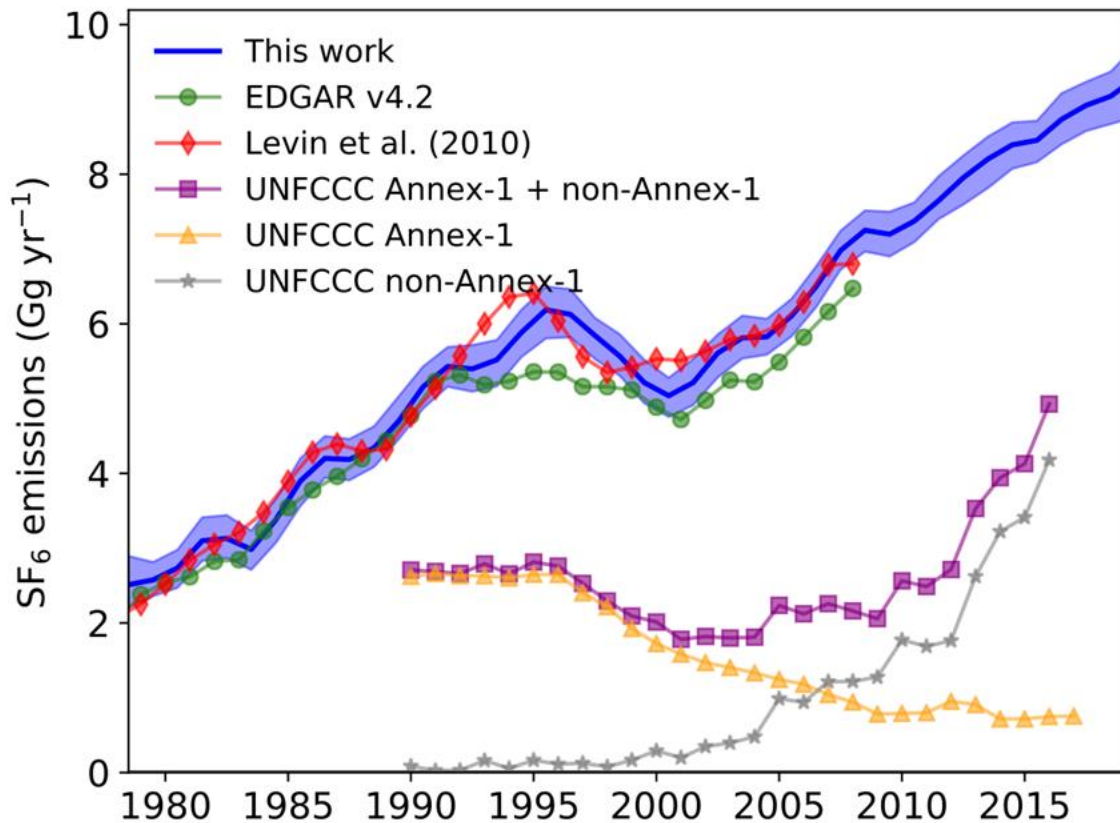
399 Our model estimated annual global emissions are shown in Fig. 2 and listed in Table 5.
 400 Here we extend and update the emission estimates prior to 2008, previously described in
 401 Rigby et al. (2010), that reported a global SF₆ emission rate of 7.3 ± 0.6 Gg yr⁻¹ (1-σ
 402 uncertainty unless specified otherwise) in 2008 and Engel, Rigby et al., (2019) estimated

403 emissions of 8.7 ± 0.4 Gg yr⁻¹ in 2016. We show that, during the last decade (2008-2018),
404 emissions have increased by approximately 24%, to 9.04 ± 0.35 Gg yr⁻¹. At 2018 levels, SF₆
405 emissions are equivalent to 212 ± 8 Tg CO₂ (assuming a 23,500 100-year global warming
406 potential, Myhre et al., 2013). Our results demonstrate that, relative to 1978, global SF₆
407 emissions have increased by around 260%, with cumulative global emissions through 2018 of
408 234 ± 7 Gg (5500 ± 170 Tg CO₂-equivalent). **Our estimates are in close agreement through**
409 **2008 with the independent top-down estimates of Levin et al. (2010). Our estimates show**
410 **similar trends to EDGAR v4.2, although our global total is on average 8.9% higher. It should**
411 **be noted that the EDGAR estimate includes some information from atmospheric observations**
412 **(Rigby et al., 2010). On the other hand, it is likely that Annex-I countries are underreporting**
413 **to the UNFCCC (Weiss and Prinn, 2011) and non-Annex-I countries are not required to**
414 **report to UNFCCC which explains the much lower UNFCCC totals. There is also close**
415 **agreement, within the uncertainties, of our modelled global SF₆ emission estimates and those**
416 **reported by Krol et al., (2018). The annual average difference was 0.2 Gg yr⁻¹ (range 0.01-**
417 **0.49 Gg yr⁻¹).**

418 Figure 2 and Table 3 record the individual Annex-1 and our revised non-Annex-1
419 emissions and their combined emissions. UNFCCC emissions reported after 2008 for the
420 non-Annex-1 countries exceed emissions from Annex-1 countries, as SF₆ consumption
421 moved from Annex-1 countries to non-Annex-1 countries, particularly in Asia. We note that
422 the significant downward trend in our top-down emission estimate between 1996-2000
423 matches the UNFCCC reported emissions, furthermore this decline is also consistent with the
424 drop in sales and prompt emissions listed in the Rand Report (Table S2).

425 The average annual difference between our global top-down estimates and UNFCCC
426 reports (Annex-1 plus revised non-Annex-1) listed in Table 5 was 4.5 Gg yr⁻¹, reaching a
427 maximum difference of 5.2 Gg in 2012. This difference subsequently decreased to an annual
428 average of ~ 4 Gg yr⁻¹ between 2013-2018, implying improved or more comprehensive
429 reporting from non-Annex-1 countries, although we recognise that these differences are prone
430 to large uncertainties, given the limited emissions data submitted to UNFCCC from the non-
431 Annex-1 countries.

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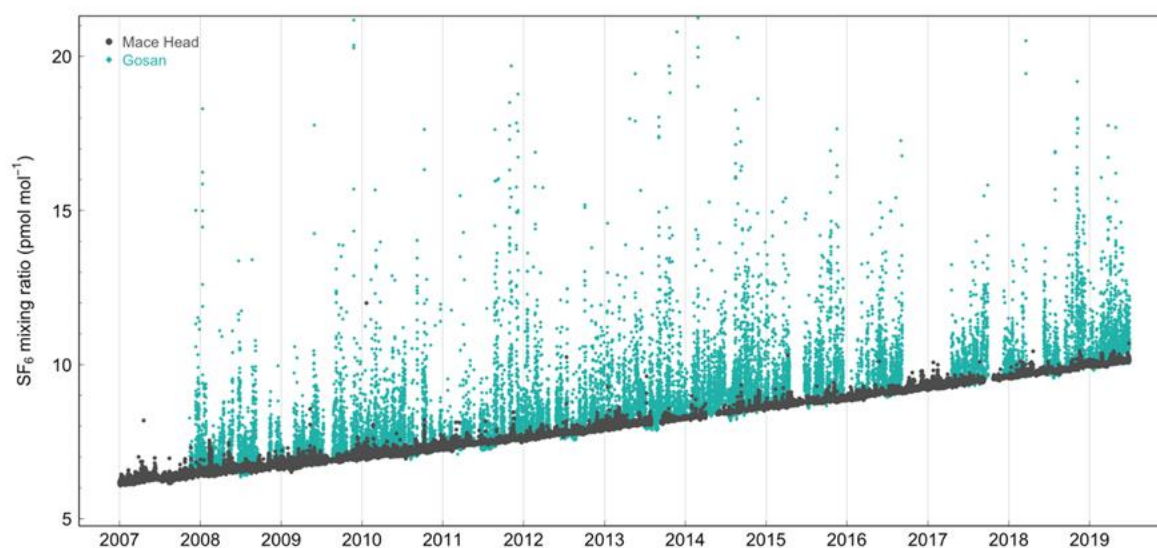


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Figure 2. Optimised global SF₆ emissions using AGAGE measurements (solid blue line) and shaded line shows the 1 σ uncertainties; emissions from Levin et al. (2010, red diamonds); EDGAR v4.2 emissions (green circles); UNFCCC Annex-1 reported emissions (orange triangles); UNFCCC non-Annex-1 reported emissions (grey stars); combined non-Annex-1 and Annex-1 UNFCCC emissions (purple squares).

441 5.1 Regional Emission Estimates

442 Top-down regional emission estimates have been calculated for two major emission
 443 regions of the world, East Asia (China, S. Korea) and North West Europe. As described
 444 below, observations from Gosan (Jeju Island, South Korea 33.3° N, 126.2° E) were used to
 445 estimate the Chinese emissions, and, for Europe, observations from the UK DECC network
 446 and three European AGAGE stations were used (Mace Head, Ireland, MHD; Jungfraujoch,
 447 JFJ, Switzerland; and Monte Cimone CMN, Italy). Figure 3 records the high frequency mole
 448 fractions of SF₆ measured at two AGAGE sites, MHD (53° N, 10° W) and Gosan, GSN, Jeju
 449 Island, South Korea (33° N, 126° E). Compared to Mace Head, the Gosan data show very
 450 large enhancements (10-30 ppt, compared to 1–2 ppt) above the background mixing ratio of
 451 ~5-10 ppt, reflecting significant regional emissions. The Gosan enhancements are associated
 452 with the transport of polluted air masses from the north-east part of China, the Korean
 453 Peninsula and Japan (Kim et al., 2010; Fang et al., 2014).



454 Figure 3. Atmospheric mixing ratios (ppt) recorded at Mace Head, Ireland (black) are
 455 shown on top of the measurements at Gosan, Jeju Island, South Korea (green). Elevated
 456 mixing ratios represent pollution events associated with regional emissions.

457 Note: **GSN occasionally shows lower mixing ratios than MHD during the summer months**
 458 **when the monsoon transports oceanic background air from the Southern regions to the Gosan**
 459 **site on Jeju Island, South Korea, which accounts for the cases in which GSN mole fractions**
 460 **are lower than those at MHD.**

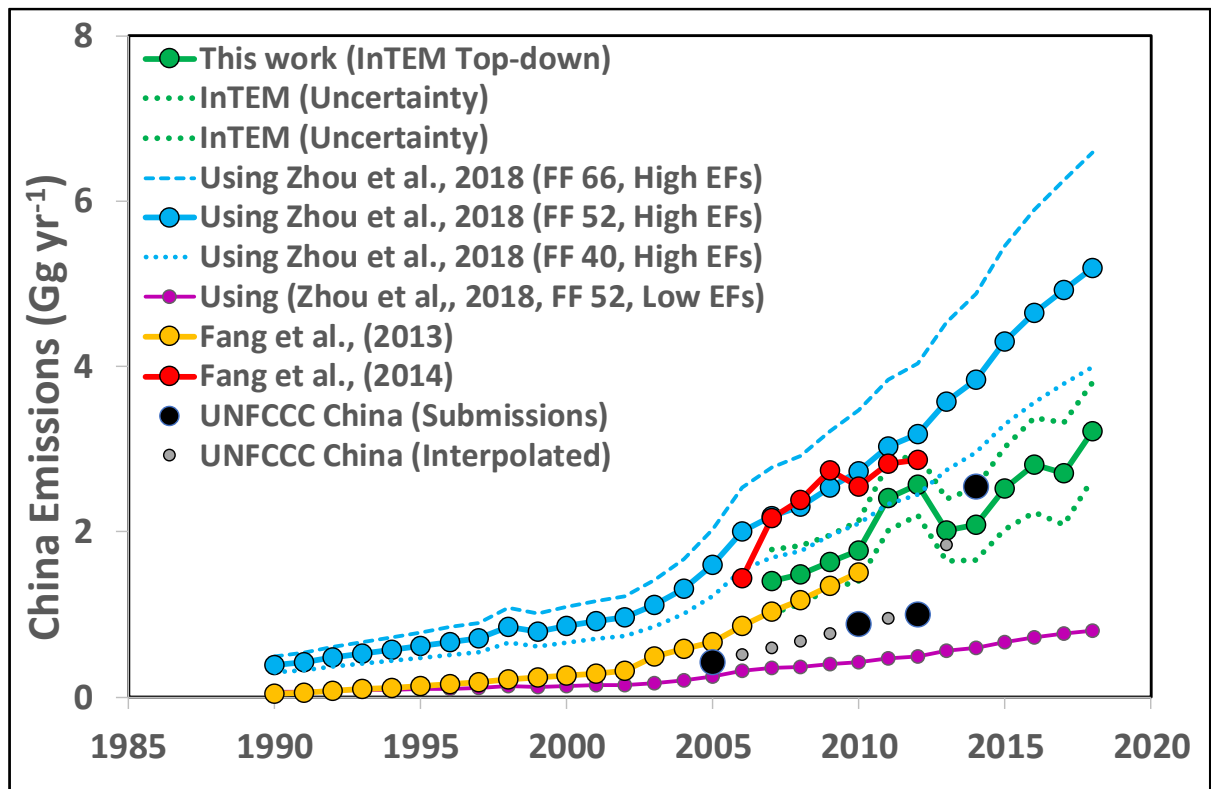
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462 5.2 East Asian estimated emissions

463 Regional top-down estimated emissions for Eastern mainland China, inferred using
 464 InTEM and Gosan measurements, are shown in Fig. 4 and listed in Table 5. Derived Chinese
 465 emissions (from an area representing 34% of China's population) were subsequently scaled to
 466 the whole country by population. China emissions increased from 1.4 (1.0-1.8) Gg yr⁻¹ in
 467 2007 to 3.2 (2.6-3.8) Gg yr⁻¹ in 2018, an increase of 130 %. Based on the InTEM regional
 468 emission estimates, China accounted for 36% (29-42%) of our model estimated global 2018
 469 emissions. **The InTEM results show an increasing emission from China, the temporary rise in**
 470 **the mean value in 2011-2012 needs to be understood within the context of the uncertainty**
 471 **estimates, it is plausible, within 1 sigma, that there was no enhanced emissions during this**
 472 **period.** Higher (average 38%, 2006-2012) emissions for China (Fang et al., 2014) have been
 473 derived using observations from three stations (Gosan, South Korea; Hateruma, Japan; Cape
 474 Ochi-ishi, Japan) rather than one station (Gosan), coarser and different meteorology, a
 475 detailed spatial prior and solved for the whole of China. Also shown in Fig. 4 are our bottom-
 476 up estimated emissions calculated from the usage of SF₆ in the electrical power industry
 477 (Section 3.1), following the methodology published in Zhou et al. (2018) for different filling
 478 factors (FF of 40, 52, and 66 t/GW) and high and low emission factors EFs. The assumed
 479 high EFs were 8.6% (manufacture and installation) and 4.7% (operation and maintenance)
 480 and low EFs of 1.7% (manufacture and installation) and 0.7% (operation and maintenance),

481 Our bottom-up estimated emissions, using the high EFs, are generally larger than the bottom-
 482 up estimated China emissions determined by Fang et al. (2013), while China estimates based
 483 on the lower EFs suggested by Zhou et al. (2018), are much lower than the other Chinese
 484 emission estimates. Notably from 2007-2012 the bottom-up estimates, after Zhou et al.,
 485 (2018), with a FF of (52 t/GW) and high EF are in close agreement with the top-down
 486 estimates of Fang et al., (2014), They also agree with our results within uncertainties.
 487 However, after about 2014 the increase in these bottom-up estimates especially with the
 488 highest FF (66 t/GW), appear to represent an unrealistically large percentage of global
 489 emissions.

490 China inventory compiled from biennial submissions to the UNFCCC National
 491 Communications and Biennial Update Report (2018) are also included in Fig. 4 (black filled
 492 circles and interpolated values grey filled circles for missing years). China reported emissions
 493 to the UNFCCC, that were consistently lower than the observation based InTEM modelled
 494 emission estimates through 2012, substantially increase in 2014 to within the uncertainties of
 495 the modelled emissions.



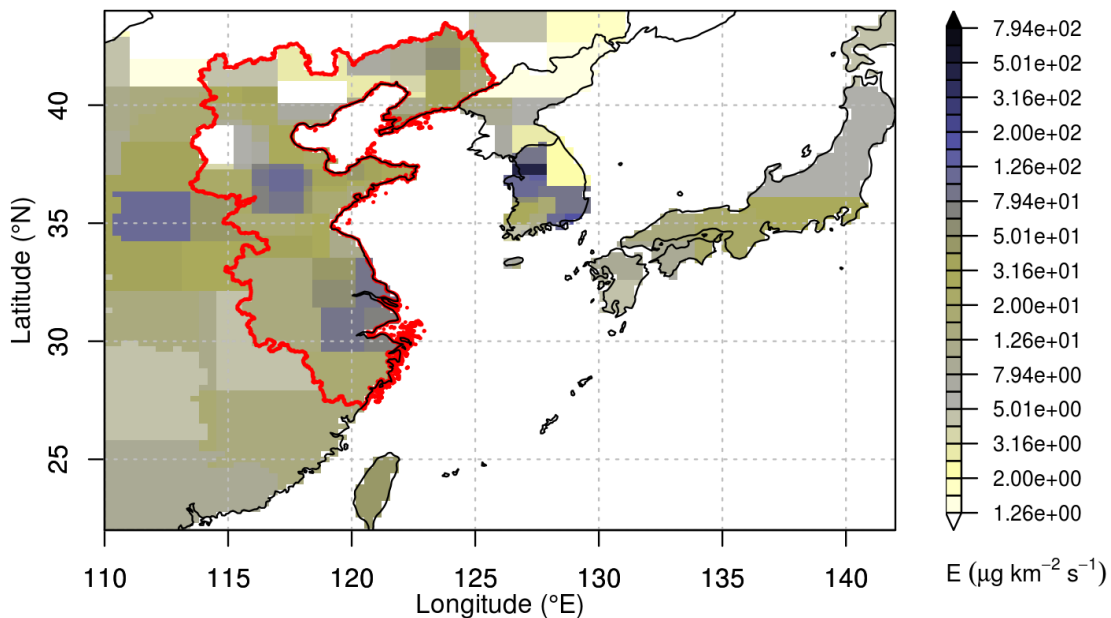
496
 497 Figure 4. Bottom-up and Top-down emission estimates scaled for the whole of China.
 498

499 Top-down Chinese emission estimates, Fang et al. (2014), agree within the uncertainties,
 500 with our bottom-up emission estimates. Conversely our top-down InTEM Chinese emission
 501 estimates fall between our bottom-up and the Fang et al. (2013) bottom-up emission
 502 estimates. Regardless of the EF used, our bottom-up emission estimates would require lower
 503 Chinese EFs to obtain closer agreement with our top-down emission estimate.
 504

505 Figure 5 shows the footprint of the mapped China emission magnitudes determined from
 506 InTEM, based on measurements recorded at the Gosan station, South Korea. Although our
 507 main focus has been on emissions from China, it is clear from Fig. 5 that there are also
 508 emissions from South Korea. The 2007-2018 average annual SF₆ InTEM emission estimate
 509 for South Korea (population ~52 M) is $0.26 \pm 0.05 \text{ Gg yr}^{-1}$ with a slight upward trend ($+0.007$
 510 Gg yr^{-2}). This compares well with the reported average value of 0.36 Gg yr^{-1} over the period
 511 2007-2014 with an upward trend of $+0.006 \text{ Gg yr}^{-2}$ (South Korea, 2017, second biennial
 512 report). The emissions for South Korea are higher per head of population ($\sim 0.005 \text{ Gg/M}$) than
 513 those estimated for China (population $\sim 1400 \text{ M}$, $\sim 0.002 \text{ Gg/M}$ in 2018). For Western Europe,
 514 discussed in the next section, the equivalent value is $\sim 0.001 \text{ Gg/M}$.

515 In Supplementary Table S3 we list the InTEM SF₆ emission estimates for South Korea.
 516 The average annual emissions of South Korea (0.26 Gg yr^{-1}) are similar to those of Western
 517 Europe (0.22 Gg yr^{-1}) and in both cases are approximately 1/10 of Chinese average annual
 518 emissions (2.2 Gg yr^{-1}).

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 522 Figure 5. Map of the top-down emission estimate from China and East Asia. The red line
 523 indicates the boundary of the region we denote ‘eastern mainland China’, to which the
 524 measurements at Gosan and the inversion method are most sensitive.

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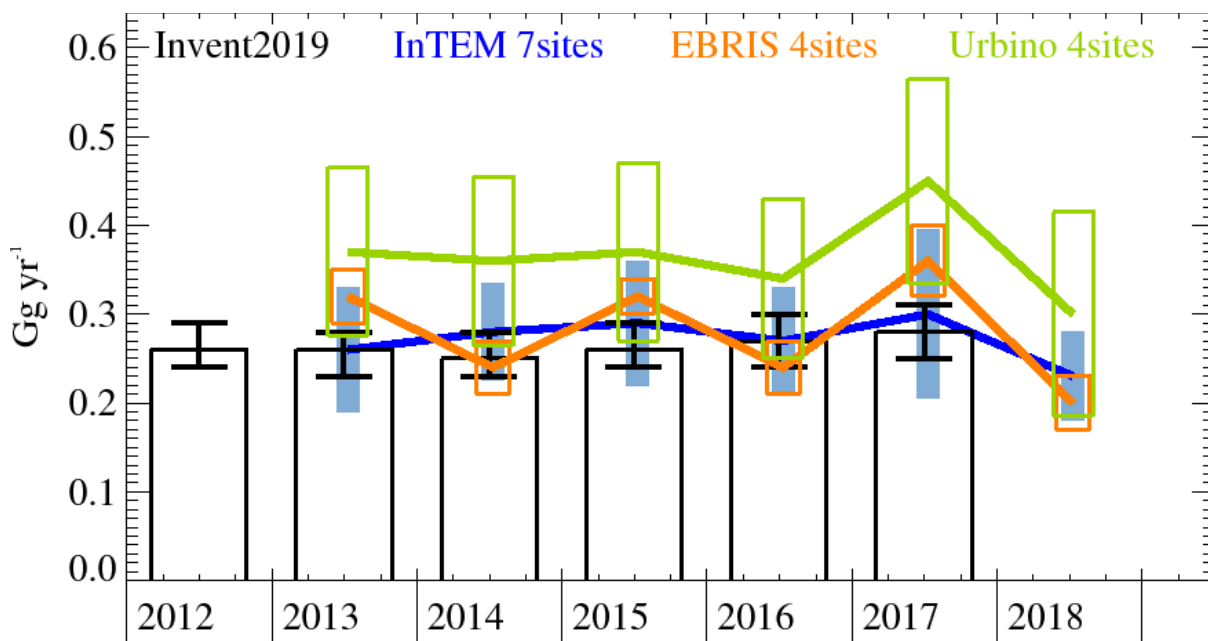
526 5.3 Western Europe emission estimates

527 InTEM estimated top-down emissions (2013-2018) for Western Europe (United
 528 Kingdom, Ireland, Benelux, Germany, France, Denmark, Switzerland, Austria, Spain, Italy,
 529 and Portugal) from measurements at 7 sites (Mace Head (MHD), Ireland, Bilsdale, UK
 530 (BSD), Heathfield, UK (HFD), Ridge Hill, UK (RGL) and Tacolneston, UK (TAC),
 531 Jungfrauoch (JFJ), Switzerland, and Monte Cimone (CMN), Italy, are presented in Fig. 6 and
 532 listed in Table 6. EBRIS used observations from 4 sites (MHD, TAC, JFJ, CMN) to estimate
 533 top-down emissions for the period 2013-2018. Emissions from the InTEM and ERBRIS

534 inversion models are in close agreement with inventory emissions (UNFCCC 2019). FLITS
 535 also used observations (2013-2018) from 4 sites (MHD, TAC, JFJ, CMN) and an inverse
 536 model to estimate top-down emissions, which are higher than the other two results but follow
 537 a similar trend. The emission flux uncertainty decreases from 200% for the *a priori* to ~25 %
 538 for the *a posteriori* emission field (average over the study period), supporting the reliability
 539 of the results. Top-down emissions for Western Europe from 4 inversion systems for the year
 540 2011 were reported to be 47% higher than UNFCCC, with Germany identified as the
 541 principal emitter (Brunner et al., 2017).

542 The contribution of Western European SF₆ emissions to the global total in 2018 was 3.1%
 543 (2.4-3.9 %, Table 6, average of all inversions). Comparing the model estimated SF₆ emissions
 544 from Western Europe and China, it is apparent that China is a much larger contributor to the
 545 global SF₆ inventory. On an annually averaged basis, top-down Chinese emissions exceed
 546 those emitted from Western Europe by a factor of ~10. For Western Europe, EFs are
 547 generally expected to be lower, representing better maintenance practices and more efficient
 548 SF₆ capture during re-filling (EU Commission, 2015). The faster uptake of SF₆ substitutes
 549 and vacuum-insulated units would also explain the much lower emission estimates in
 550 Western European countries.

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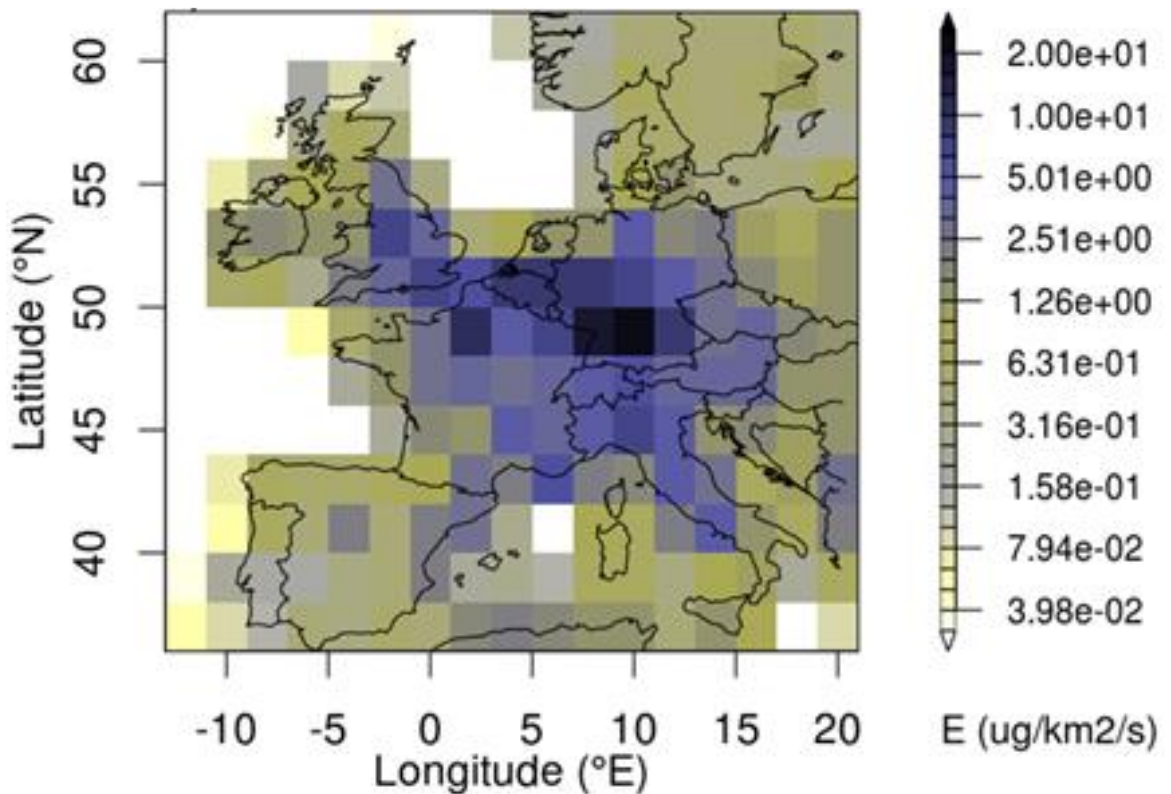
552

553 Figure 6. SF₆: Western Europe emission estimates (Gg yr⁻¹) from the UNFCCC Inventory
 554 black; InTEM inversion (2013-2018, blue, 7 sites: MHD, JFJ, CMN, TAC, RGL, HFD,
 555 BSD); EBRIS (2013-2018, orange, 4 sites: MHD, TAC, JFJ, CMN). FLITS (2013-2018,
 556 green, 4 sites: MHD, TAC, JFJ, CMN).

557 The uncertainty bars are ± 1 std.

558

559 Figure 7 shows the footprint of the average emission estimates for Western Europe calculated
 560 using three inverse models (InTEM, EBRIS and FLITS), illustrating that significant
 561 emissions are located in southern Germany, a region with a substantial number of semi-
 562 conductor producers (<https://prtr.eea.europa.eu/#home>).



564

565 Figure 7. Top-down inversion emission estimate for Western Europe (2013-2018).
 566 Average of InTEM (7 observation sites), EBRIS (4 observation sites) and FLITS (4
 567 observation sites).

568

569 **6. Increasing global SF₆ emissions and the deficit between bottom-up and top-down** 570 **emissions estimates**

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572 Weiss and Prinn (2011) noted that SF₆ bottom-up estimates derived from industrial
 573 accounting and reported to the UNFCCC by Annex-1 countries are likely under-reported,
 574 actually representing 80% of the total in the mid-1990s and 60% of the total in 2006, leading
 575 to poor agreement (under-reported by a factor of 2) with top-down emissions estimates
 576 determined from atmospheric observations. However, for Western Europe (Sect.5.3) our
 577 estimated emissions (2013-2018) from the model inversions are in close agreement with the
 578 UNFCCC reported inventory. Limitations imposed by commercial secrecy and the lack of
 579 consistent reporting of SF₆ emissions, both from Annex-1 and non-Annex-1 countries,
 580 continue to contribute to the discrepancies between bottom-up and top-down methods.

581 We next explore if the increasing global emissions of SF₆ may be related to changing
 582 patterns of source location and usage in electrical equipment, magnesium smelting,
 583 aluminium production and electronics manufacturing, and attempt to reconcile the large
 584 average annual discrepancy of $\sim 4.5 \text{ Gg yr}^{-1}$ between bottom-up and top-down emissions
 585 estimates.

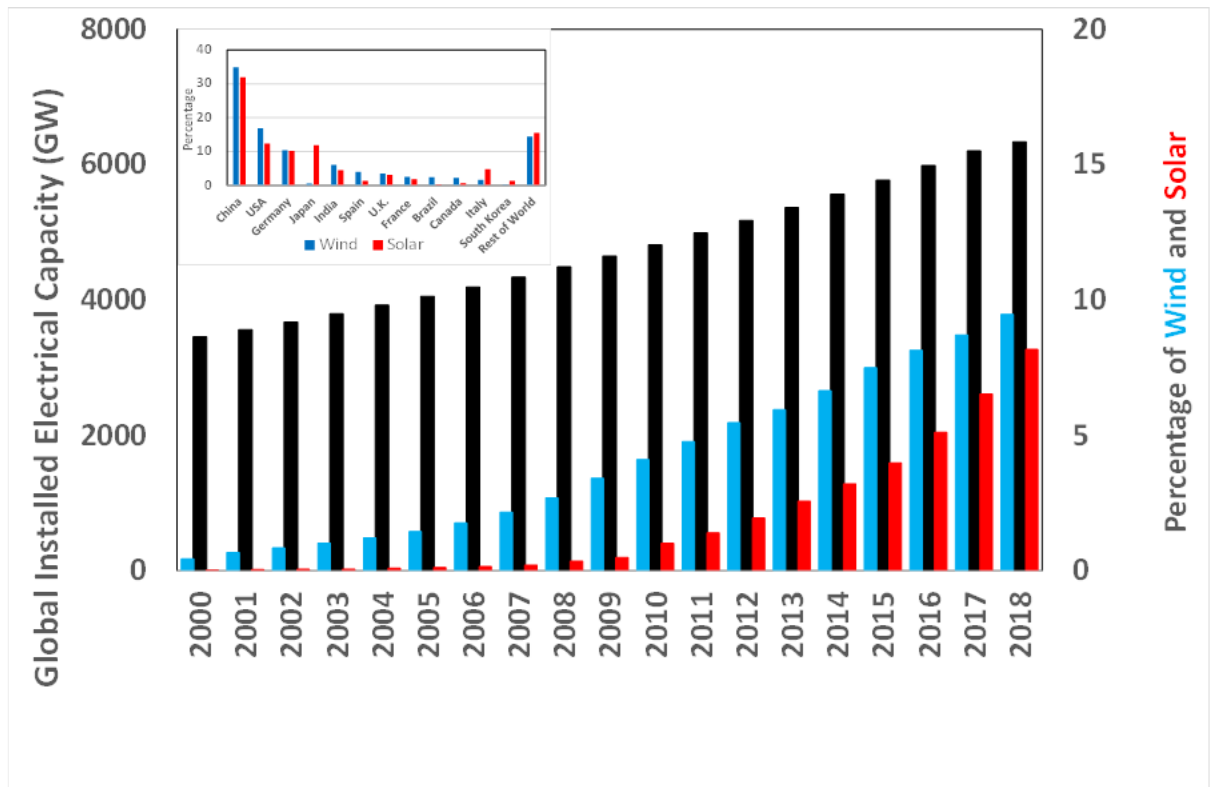
586 Previous reports on SF₆ emissions from the electrical power industry have noted that
 587 emission factors (EFs) may vary widely depending on the type of equipment and different

588 maintenance and servicing practices (Capiel/Unipede, 1999). About 12% of SF₆ consumed in
589 the manufacture and commissioning of electrical equipment is estimated to be directly
590 emitted. Industry assessments of the maximum leakage during operation from older
591 equipment (manufactured before 1980) was 3% yr⁻¹, although higher leakage rates for some
592 countries continued into the 1990s. For example, in 1995 USA annual refill and leakage of
593 circuit breakers was estimated to be 20% of total installed stock; however, with the
594 installation of improved self-contained equipment leakages steadily reduced to 0.5-1% yr⁻¹
595 (Olivier and Bakker, 1999). A recent study of the SF₆ losses from gas-insulated electrical
596 equipment in the UK calculated an average annual leakage rate of 0.46% yr⁻¹ from the
597 inventory of SF₆ held in the installed equipment and 1.29% yr⁻¹ from the transmission
598 network, with an overall average (2010-2016) leakage rate of 1% yr⁻¹ from the UK electrical
599 power industry (Widger and Haddad, 2018).

600 Figure 8 shows the global installed electrical capacity and the percentage contribution of
601 wind and solar power capacity from 2000-2018. Installed electrical capacity grew by 62%
602 (2412 GW) during this period. Of this rise, ~45% was due to solar and wind, illustrating the
603 very rapid growth rate of the renewable sector, as utility companies invested in renewable
604 energy (GWEC, 2018; CWEA, 2018; IRENA, 2019). The inset panel records the percentage
605 of solar and wind power by country during 2017 – led by China, USA, and Germany. The
606 global adoption of renewable technologies, especially hydroelectric, wind and solar power
607 has been particularly strong in the non-Annex-1 countries to support their continuing
608 development (Fang et al., 2013). For example, Chinese installed electrical capacity, relative
609 to the ROW, increased from about 3% in 1980 to ~43% in 2018, as noted in Table 1.

610 We assume that with the wider geographical distribution of renewables, compared with
611 the localised gas or oil-fired power stations, this has resulted in many more connections to the
612 electricity grid and a consequent rise in the number of gas-insulated electrical switches,
613 circuit breakers, and transformers. **With the adoption of more technologically advanced GIS
614 with lower emissions we might expect there to be a reduction in overall SF₆ emissions over
615 time. There is clearly a balance between the very substantial increase in the global number of
616 newly installed GIS equipment and the major advances in reducing the leakage of SF₆ from
617 GIS equipment and the recovery and substitution of SF₆. At present the larger number of
618 global GIS installations appear to be overpowering the success in reducing SF₆ emissions.**

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622 Figure 8. Global installed electrical capacity (GW) and the percentage contribution from
623 wind power (blue bars) and from solar power (red bars) from 2000-2018. Insert: Percentage
624 of wind and solar power by country in 2017 (IRENA, 2019).

625 Sulphur hexafluoride in the electrical power industry is primarily used in high voltage
626 gas-insulated switchgear (GIS) which consumes > 80% of the SF₆ used, with medium voltage
627 GIS consuming only about 10% (Niemeyer and Chu, 1992; Dervos and Vassiliou, 2000; Zu
628 et al., 2011; Xiao et al., 2018). Since this electrical equipment can be operational for 30-40
629 years, there is a large bank of SF₆ in older equipment that will be a continuing source of
630 global SF₆ emissions through routine maintenance, decommissioning, catastrophic failure of
631 components (as noted previously) and long-term leakage. Sulphur hexafluoride is also used
632 by the utility companies in gas-insulated transmission lines (Ecofys, 2018).

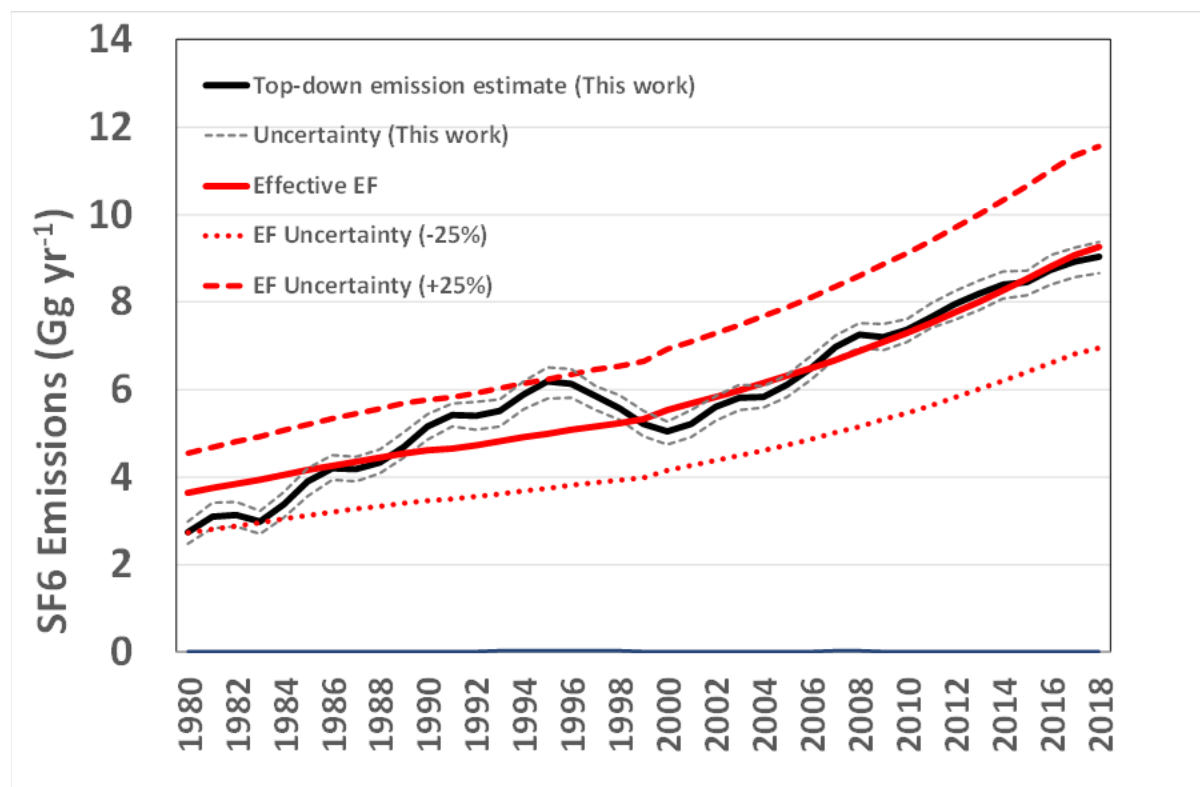
633 Recently, regulations have been introduced to mitigate the environmental impact of SF₆
634 emissions. The European Commission reinforced a 2006 F-Gas regulation in 2015 (No.
635 517/2014) with the aim of reducing the EU's F-gas emissions by two-thirds from 2014 levels
636 by 2030 (EU Commission, 2015). It is important to realise that under these current European
637 regulations there are no restrictions on the use of SF₆ in switchgear, but there are
638 requirements to recover SF₆ where possible (Biasse, 2014). Historically, SF₆ has been the
639 preferred insulator and arc-quenching gas, although technological advances and alternative
640 gases to SF₆ have been introduced to reduce overall emissions. Substitute gases include
641 perfluoroketones, perfluoronitriles and trifluoroiodomethane (CF₃I) (Okubo, 2011; Li et al.,
642 2018, Xiao et al., 2018). Some wind turbine manufacturers have recently started to offer SF₆-
643 free equipment or vacuum insulated switch gear. In 1995 the U.S. Environmental Protection
644 Agency (EPA) established an SF₆ Emissions Reduction Partnership for electric power
645 systems to improve equipment reliability and reduce SF₆ emissions by technological

646 innovation. They subsequently reported that by 2016 there had been a 74% reduction of SF₆
647 emissions by the industrial partners (EPA, 2018).

648 The combined bottom-up emissions estimated in Sect. 3.1 from SF₆ production and the
649 various industrial applications which use SF₆ are ~1.1 Gg yr⁻¹. Sales of SF₆ to these industries
650 are listed in the Rand report (Smythe, 2004) from 1996-2003 (Supplement: Table S2) which
651 does not include recent data and only covers an unspecified part of the globe, implying a
652 potential underestimation of actual emissions. Assuming SF₆ consumption from these
653 industries are emitted promptly (i.e. not banked), we calculate an average annual emission
654 from 1996-2003 of 1.1 Gg yr⁻¹ (0.83-1.42 Gg yr⁻¹), that includes the magnesium, electronics,
655 adiabatic and the fraction of ‘other uses’ that are emitted promptly. The agreement between
656 our bottom-up estimate of industrial emissions and the estimate derived from sales are
657 consistently lower than modelled top-down emission estimates. Even accepting these many
658 assumptions and uncertainties the dominant emissions are attributable to the electrical
659 industry and its use of SF₆ insulated equipment.

660 We can also obtain an ‘effective’ EF for the electrical industry by first subtracting prompt
661 emissions from our top-down emission estimate and then calculating the amount of SF₆
662 required to match the remaining annual top-down emission estimate, given global installed
663 electrical capacity and an assumed FF (52t/GW; Zhou et al., 2018). Based on this simplified
664 method we estimate an ‘effective’ average EF of 2.5% for the entire time period. In Figure 9
665 we compare our top-down emissions estimate and a bottom-up estimate with ± 25%
666 uncertainty (prompt emissions + electrical industry emissions using a median FF of 52 t/GW
667 and the inferred effective EF). Notwithstanding some disagreement between the top-down
668 estimate and the simple bottom-up model during certain periods (e.g. an overestimate in the
669 early 1980s and underestimate during the 1990s, it is notable that decadal trends in SF₆ can be
670 broadly explained by the rise in installed electrical capacity and a single effective EF. **This
671 suggests that, when considered on ~10 year timescales, reductions in EF achieved in certain
672 countries through new technologies or improved GIS management, have been offset by the
673 growth in higher-EF GIS from other parts of the world, such that the effective EF has not
674 changed substantially on a global scale.**

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Figure 9. Top-down emission estimate (solid black line) with 1-sigma uncertainties and a bottom-up estimate, based on SF₆ prompt emissions plus an emissions estimate from the electrical power industry (FF=52 t/GW) and an inferred effective EF (2.5%) with ±25% uncertainty (dashed lines).

684 4. Conclusions

685

686 New atmospheric SF₆ mole fractions are presented which extend and update our
687 previously reported time series from the 1970s to 2008 by a further 10 years to 2018 in both
688 hemispheres. We estimate global emissions of SF₆ using data from the 5 core AGAGE
689 observing sites and archived air samples with a 12-box global chemical transport model and
690 an inverse method. SF₆ emissions exhibited an almost linear increase from 2008-2018
691 reaching $9.0 \pm 0.4 \text{ Gg yr}^{-1}$ in 2018, a decadal increase of ~24%. Chinese emissions in 2018
692 based on InTEM regional emission estimates, with large uncertainties, account for 36% (29-
693 42%) of total global emissions relative to our model estimated global 2018 emissions.

694 We find that on an annually averaged basis emissions from China are about 10 times
695 larger than emissions from South Korea and Western Europe. Relative to 1978, global SF₆
696 emissions have increased by ~260% with cumulative global emissions through December
697 2018 of $234 \pm 6 \text{ Gg}$ or CO₂ equivalents $5.5 \pm 0.2 \text{ Pg}$. To further mitigate the large
698 uncertainties will require an increase in the number of monitoring sites, improved transport
699 models and a substantial improvement in the accuracy and transparency of emissions
700 reporting.

701 We note that the rapid expansion of global power demand and the faster adoption of
702 renewable technologies, such as wind and solar capacity over the past decade, particularly in
703 the Asia region, has provided a large bank of SF₆ which currently contributes to the

704 atmospheric burden of SF₆ and will continue throughout the lifetime (30-40 years) of the
705 installed equipment. The resultant increase in SF₆ emissions from the non-Annex-1 countries
706 has overwhelmed the substantial reductions in overall emissions in the Annex-1 countries,
707 where less emissive industrial practices are used in the handling of SF₆ (EPA, 2018; EU
708 Commission, 2015). This also suggests that any decrease in emission factor from Annex-1
709 countries has been offset by an increase in non-Annex-1 emission factors. The non-Annex-1
710 countries are progressively using improved and less emissive GIS electrical equipment
711 however, it is only in the last few years that alternative gases to SF₆ or SF₆-free equipment
712 have been commercially available for switchgear and other electrical systems. We conclude
713 that the observed increase in global installed electrical capacity, in both developed and
714 developing countries, is consistent with the temporal rise in SF₆ global emissions.

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Table 1. Estimated SF₆ emissions (Gg) calculated from China and the Rest of the World (ROW) installed electrical capacity.

Year	¹ China installed electrical capacity (GW)	² China estimated emissions (Gg)	³ ROW installed electrical capacity (GW)	⁴ ROW estimated emissions (Gg)	Global emissions (ROW + China) (Gg)
1980	66	0.17 (0.13-0.21)	1910.8	3.48 (2.79-4.28)	3.65 (2.93-4.50)
1981	69	0.18 (0.14-0.23)	1996.2	3.58 (2.75- 4.54)	3.76 (2.89-4.77)
1982	72	0.19 (0.15-0.24)	2068.0	3.53 (2.72-4.48)	3.72 (2.87-4.72)
1983	76	0.20 (0.16-0.26)	2133.3	3.52 (2.71-4.46)	3.72 (2.87-4.72)
1984	80	0.21 (0.16-0.27)	2225.3	3.74 (2.88-4.75)	3.95 (3.04-5.02)
1985	87	0.24 (0.19-0.31)	2303.5	3.69 (2.84-4.68)	3.93 (3.02-4.99)
1986	94	0.26 (0.20-0.33)	2368.6	3.62 (2.78-4.59)	3.88 (2.98-4.92)
1987	103	0.29 (0.23-0.37)	2430.1	3.59 (2.76-4.56)	3.88 (2.99-4.93)
1988	115	0.33 (0.26-0.43)	2486.7	3.55 (2.73-4.51)	3.89 (2.99-4.93)
1989	127	0.36 (0.28-0.46)	2555.7	3.63 (2.79-4.61)	3.99 (3.07-5.07)
1990	138	0.39 (0.30-0.49)	2594.4	3.40 (2.62-4.32)	3.79 (2.91-4.81)
1991	151	0.43 (0.33-0.54)	2617.9	3.25 (2.50-4.12)	3.68 (2.93-4.66)
1992	167	0.48 (0.37-0.61)	2661.5	3.33 (2.56-4.23)	3.81 (2.93-4.84)
1993	184	0.53 (0.40-0.67)	2712.4	3.34 (2.57-4.24)	3.87 (2.98-4.91)
1994	201	0.57 (0.44-0.72)	2766.0	3.32 (2.56-4.22)	3.89 (2.99-4.94)
1995	219	0.62 (0.47-0.78)	2800.9	3.15 (2.42-4.00)	3.77 (2.90-4.78)
1996	238	0.67 (0.51-0.85)	2858.5	3.25 (2.50-4.13)	3.92 (3.02-4.98)
1997	256	0.71 (0.54-0.90)	2905.0	3.13 (2.41-3.98)	3.84 (2.95-4.87)
1998	289	0.85 (0.66-1.08)	2922.8	2.87 (2.21-3.64)	3.72 (2.87-4.73)
1999	302	0.80 (0.61-1.01)	2984.6	3.10 (2.393.94)	3.90 (3.00-4.95)
2000	319.3	0.86 (0.66-1.09)	3135.7	3.69 (2.84-4.68)	4.55 (3.50-5.77)
2001	338.5	0.91 (0.70-1.16)	3224.8	3.27 (2.52-4.15)	4.18 (3.22-5.31)

2002	357.6	0.96 (0.74-1.22)	3319.7	3.27 (2.52-4.15)	4.23 (3.26-5.37)
2003	392.4	1.11 (0.86-1.42)	3404.8	3.16 (2.43-4.02)	4.27 (3.29-5.43)
2004	443.5	1.31 (1.01-1.67)	3479.4	3.04 (2.34-3.85)	4.35 (3.35-5.52)
2005	517.8	1.60 (1.23-2.03)	3536.9	2.85 (2.19-3.62)	4.45 (3.42-5.65)
2006	624.1	2.00 (1.54-2.54)	3568.8	2.59 (1.99-3.29)	4.59 (3.53-5.83)
2007	720.6	2.19 (1.69-2.78)	3617.5	2.61 (2.00-3.31)	4.80 (3.69-6.09)
2008	798.5	2.30 (1.77-2.92)	3692.2	2.69 (2.07-3.41)	4.99 (3.84-6.33)
2009	883.1	2.54 (1.95-3.22)	3767.2	2.61 (2.01-3.31)	5.15 (3.96-6.53)
2010	966.4	2.73 (2.10-3.47)	3850.9	2.58 (1.98-3.27)	5.31 (4.09-6.74)
2011	1062.5	3.03 (2.33-3.84)	3929.5	2.45 (1.89-3.11)	5.48 (4.22-6.96)
2012	1146.8	3.18 (2.45-4.04)	4027.9	2.49 (1.91-3.16)	5.67 (4.36-7.19)
2013	1257.7	3.57 (2.75-4.53)	4107.9	2.27 (1.75-2.88)	5.84 (4.49-7.41)
2014	1369.2	3.84 (2.96-4.88)	4195.7	2.21 (1.70-2.81)	6.05 (4.66-7.69)
2015	1506.7	4.30 (3.31-5.46)	4265.8	1.99 (1.53-2.52)	6.29 (4.83-7.97)
2016	1645.8	4.64 (3.57-5.89)	4343.1	1.91 (1.47-2.42)	6.55 (5.04-8.32)
2017	1777.0	4.93 (3.79-6.26)	4424.6	1.81 (1.39-2.30)	6.74 (5.19-8.56)
2018	1899.7	5.19 (3.99-6.59)	4435.7	1.65 (1.39-1.96)	6.84 (5.38-8.55)

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721 ¹ China installed electrical power capacity compiled from
722 www.statistica.com/statistics/302269 and www.iea.com (IEA 2017).

723 ² estimated China emissions derived from method of Zhou et al. (2018), using an initial
724 filling of 52 t/GW (in parenthesis, range 40-66 t/GW) and emission factors of 8.6%
725 (manufacture and installation) and 4.7% (operation and maintenance).

726 ³ Rest of the World (ROW) installed electrical power capacity, compiled from
727 www.data.UN.org (www.iea.com) and mecometer.com.

728 ⁴ ROW emissions estimated using an initial filling of 52 t/GW and a 12% loss during
729 manufacture and installation of new equipment and assuming 3% loss from banked SF₆ in
730 electrical equipment in 1980 and then decreasing linearly to 1.0% in 2018, reflecting the
731 change from older to newer equipment (Olivier and Bakker, 1999).

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743 Table 2. The East Asian setup of the inversion system run 2007-2018 in 2-yr blocks: The
 744 Meteorology, Transport Model (ATM), geographical domains over which the ATM is run,
 745 number of particles released, inversion time-steps, prior information and observations used.

Inversion System	Atmospheric Transport Model	Driving Meteorology	Computational Domain	Inversion Domain	Particles Released	Release Time Step	Prior	Obs
InTEM	NAME	Unified Model 12-40 km horizontal	54.3°E to 192.0°E 5.3°S to 74.3°N	88.1°E to 145.9°E 16.0°N to 57.6°N	20,000 hr ⁻¹	2 hr	Population 2 kt over domain 300% uncertainty per sub-region	GSN

746 Note. GSN=Gosan station, Korea.

747

748 Table 3. The European setup of each inversion system run each year 2013-2018: the
 749 Meteorology, Transport Model (ATM), geographical domains over which the ATM's are run,
 750 number of particles released, inversion time-steps, prior information and observations used.

751

Inversion System	Atmospheric Transport Model	Driving Meteorology	Computational Domain	Inversion Domain	Particles Released	Release Time Step	Prior	Obs
InTEM	NAME	Unified Model 1.5 km nested in 12-17 km horizontal	98.1°W to 39.6°E 10.6°N to 79.2°N	14.3°W to 30.8°E 36.4°N to 66.3°N	20,000 hr ⁻¹	2 hr	Population 2 kt over domain 200% uncertainty per sub-region	MHD JFJ CMN TAC RGL BSD HFD
EBRIS	FLEXPART 9.1_Empa	ECMWF-IFS 0.2° x 0.2° (-4°E - 16°E, 39°N - 51°N) nested in 1° x 1°	Global	12.0°W to 26.4°E 36.0°N to 62.0°N	16,667 hr ⁻¹	3 hr	Population 2kt over domain 100 % uncertainty for whole inversion domain	MHD JFJ CMN TAC
FLITS	FLEXPART 9.0	ECMWF Operational 1° lat x 1° lon horizontal	Global	20.0°W to 50.0°E 0.0°N to 80.0°N	13,333 hr ⁻¹	3 hr	Population 2 kt over domain 200% uncertainty per sub-region	MHD JFJ CMN TAC

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753 Note. Observing stations: MHD (Mace Head, Ireland); JFJ (Jungfraujoch, Switzerland);
 754 CMN (Monte Cimone, Italy); TAC (Tacolneston, UK); RGL (Ridge Hill, UK); BSD
 755 (Bilsdale, UK); HFD (Heathfield, UK).

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757 Table 4. Global SF₆ mole fraction output from the AGAGE 12-box model.

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YEAR	Global mole fraction (ppt)	16%ile	84%ile	YEAR	Global mole fraction (ppt)	16%ile	84%ile
1978	0.66	0.64	0.68	1999	4.34	4.25	4.43
1979	0.76	0.73	0.78	2000	4.53	4.43	4.63
1980	0.86	0.83	0.88	2001	4.73	4.62	4.83
1981	0.98	0.95	1.00	2002	4.94	4.83	5.04
1982	1.10	1.08	1.13	2003	5.17	5.06	5.27
1983	1.22	1.19	1.25	2004	5.39	5.28	5.50
1984	1.34	1.31	1.37	2005	5.62	5.50	5.74
1985	1.49	1.45	1.52	2006	5.87	5.74	5.99
1986	1.65	1.61	1.68	2007	6.14	6.00	6.26
1987	1.81	1.77	1.85	2008	6.42	6.27	6.54
1988	1.98	1.93	2.02	2009	6.70	6.55	6.84
1989	2.15	2.11	2.20	2010	6.99	6.83	7.12
1990	2.35	2.30	2.40	2011	7.28	7.12	7.42
1991	2.56	2.51	2.62	2012	7.59	7.42	7.73
1992	2.77	2.71	2.83	2013	7.90	7.73	8.06
1993	2.98	2.92	3.05	2014	8.23	8.05	8.39
1994	3.21	3.13	3.27	2015	8.56	8.37	8.73
1995	3.45	3.37	3.52	2016	8.89	8.69	9.06
1996	3.69	3.60	3.76	2017	9.24	9.03	9.42
1997	3.92	3.83	4.00	2018	9.59	9.37	9.79
1998	4.14	4.04	4.22				

759 Note: Mole fractions are reported at the mid-point of the year.

760 Table 5. Modelled global and SF₆ emissions from China. (Gg yr⁻¹) and UNFCCC reported
 761 emissions.

YEAR	*This work Global Emissions (Gg yr ⁻¹)	UNFCCC Annex-1	UNFCCC Revised Non-Annex-1	UNFCCC Combined Annex-1 + Revised Non-Annex-1	InTEM: China (Gosan site)	Krol et al., 2018
1978	2.51 (2.11-2.83)					
1979	2.58 (2.29-2.84)					
1980	2.74 (2.47-2.93)					
1981	3.10 (2.83-3.40)					
1982	3.13 (2.86-3.41)					
1983	2.98 (2.65-3.15)					
1984	3.38 (3.08-3.60)					
1985	3.90 (3.61-4.18)					
1986	4.20 (3.92-4.42)					
1987	4.19 (3.89-4.43)					
1988	4.34 (4.05-4.59)					4.30
1989	4.70 (4.41-4.97)					4.33
1990	5.16 (4.85-5.46)	2.62	0.08	2.71		4.77
1991	5.43 (5.12-5.75)	2.66	0.03	2.69		5.14
1992	5.40 (5.07-5.63)	2.63	0.03	2.66		5.59
1993	5.52 (5.20-5.81)	2.63	0.16	2.79		6.00
1994	5.88 (5.49-6.14)	2.60	0.06	2.66		6.36
1995	6.19 (5.86-6.54)	2.65	0.16	2.81		6.41
1996	6.13 (5.75-6.35)	2.65	0.11	2.76		6.06
1997	5.84 (5.55-6.13)	2.40	0.13	2.53		5.56
1998	5.57 (5.31-5.82)	2.22	0.08	2.29		5.35
1999	5.21 (4.94-5.49)	1.93	0.17	2.09		5.42
2000	5.04 (4.68-5.32)	1.72	0.29	2.01		5.55
2001	5.21 (4.89-5.43)	1.58	0.20	1.78		5.51
2002	5.61 (5.31-5.89)	1.47	0.35	1.82		5.63
2003	5.81 (5.53-6.03)	1.40	0.40	1.80		5.79
2004	5.83 (5.57-6.05)	1.33	0.48	1.81		5.86

2005	6.11 (5.87-6.35)	1.25	0.99	2.23		5.98
2006	6.50 (6.21-6.72)	1.18	0.94	2.12		6.29
2007	6.98 (6.64-7.20)	1.04	1.22	2.26	1.40 (1.01-1.79)	6.79
2008	7.25 (6.89-7.45)	0.94	1.22	2.16	1.48 (1.13-1.83)	7.18
2009	7.20 (6.92-7.47)	0.78	1.28	2.06	1.64 (1.31-1.96)	7.26
2010	7.37 (7.05-7.65)	0.79	1.77	2.56	1.77 (1.40-2.13)	7.36
2011	7.65 (7.35-7.98)	0.80	1.69	2.49	2.41 (2.01-2.82)	7.56
2012	7.95 (7.59-8.20)	0.95	1.76	2.71	2.57 (2.20-2.95)	7.78
2013	8.20 (7.86-8.50)	0.91	2.62	3.53	2.02 (1.65-2.40)	7.96
2014	8.39 (8.05-8.65)	0.72	3.22	3.94	2.09 (1.66-2.53)	8.16
2015	8.45 (8.11-8.73)	0.72	3.41	4.13	2.52 (2.03-3.02)	8.36
2016	8.73 (8.37-8.99)	0.75	4.18	4.93	2.81 (2.24-3.38)	4.30
2017	8.92 (8.56-9.24)	NR	NR	NR	2.71 (2.09-3.32)	4.33
2018	9.04 (8.63-9.34)	NR	NR	NR	3.22 (2.64-3.81)	4.77

762 Note: Global emissions are mid-year. NR = Not reported. Revised Non-Annex-1 includes
763 interpolated values for missing years. Uncertainties shown in parenthesis as 16%ile and
764 84%ile. **China SF₆ emissions estimated by InTEM were scaled to total China emissions by**
765 **population.**

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780 Table 6. SF₆ emission estimates for Western Europe; UNFCCC inventory and InTEM,
 781 EMPA and FLITS emissions (Gg yr⁻¹). (Uncertainties in parenthesis).

	Inventory 1yr	InTEM (3sites, 2yr)	InTEM (7sites, 1yr)	EMPA (4sites, 1yr)	FLITS (4sites, 1yr)
1990	0.48 (0.43-0.53)				
1991	0.5 (0.45-0.55)				
1992	0.54 (0.48-0.59)				
1993	0.57 (0.51-0.62)				
1994	0.61 (0.55-0.68)				
1995	0.66 (0.59-0.72)				
1996	0.65 (0.58-0.71)				
1997	0.58 (0.52-0.64)				
1998	0.55 (0.50-0.61)				
1999	0.45 (0.41-0.50)				
2000	0.45 (0.41-0.50)				
2001	0.42 (0.37-0.46)				
2002	0.37 (0.33-0.40)				
2003	0.34 (0.31-0.38)				
2004	0.35 (0.31-0.38)				
2005	0.33 (0.30-0.37)				
2006	0.31 (0.28-0.35)				
2007	0.29 (0.26-0.32)	0.25 (0.16-0.34)			
2008	0.28 (0.26-0.31)	0.23 (0.16-0.31)			
2009	0.27 (0.24-0.29)	0.19 (0.13-0.25)			
2010	0.27 (0.24-0.29)	0.19 (0.13-0.25)			
2011	0.26 (0.23-0.28)	0.21 (0.16-0.27)			
2012	0.26 (0.24-0.29)	0.21 (0.16-0.27)			
2013	0.26 (0.23-0.28)	0.22 (0.16-0.28)	0.26 (0.19-0.33)	0.32 (0.29-0.35)	0.37 (0.27-0.46)
2014	0.25 (0.23-0.28)	0.23 (0.16-0.29)	0.28 (0.22-0.33)	0.24 (0.21-0.27)	0.36 (0.26-0.45)
2015	0.26 (0.24-0.29)	0.22 (0.16-0.29)	0.29 (0.22-0.36)	0.32 (0.30-0.34)	0.37 (0.27-0.47)
2016	0.27 (0.24-0.30)	0.22 (0.15-0.30)	0.27 (0.21-0.33)	0.24 (0.21-0.27)	0.34 (0.25-0.43)
2017	0.28 (0.25-0.31)	0.21 (0.14-0.28)	0.30 (0.20-0.39)	0.36 (0.32-0.40)	0.45 (0.33-0.56)
2018	NR	0.21 (0.15-0.27)	0.23 (0.18-0.28)	0.20 (0.17-0.23)	0.30 (0.19-0.42)

782 NR= Not reported,

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784

785 **Author contributions**

786 S.P., S.O'D., P.B.K., P.J.F., L.P.S., B.M., S.H., S.R., M.M., J.A., J.M., R.F.W., C.M.H.,
 787 M.K.V., M.P., H.P., T.A., D.Y., and C.R contributed observational data. M.R., A.J.M., F.G.,
 788 and R.G.P., carried out atmospheric model simulations and inverse analysis with support
 789 from P.K.S., and R.H.J.W. The authors A.Mc., and K.M.S., made valuable data analysis
 790 contributions to the work.

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793 **Competing interests**

794 We confirm that there are no competing interests.

795 **Data availability**

796 The entire ALE/GAGE/AGAGE data base comprising every calibrated measurement
797 including pollution events is archived on the official AGAGE website
798 <http://agage.mit.edu/data> (note guidelines for use of AGAGE data), and on the ESS-DIVE
799 website <http://cdiac.ess-dive.lbl.gov/ndps/alegagage.html>. UK DECC data is available from
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801 Analysis.

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