



Firewood residential heating – local versus regional influence on the aerosol burden

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15 **Abstract.** We report the first-time use of the Lagrangian Particle Dispersion Model (LPDM) FLEXPART to simulate isotope ratios of the biomass burning tracer levoglucosan. Here, we combine the model results with observed levoglucosan concentrations and $\delta^{13}\text{C}$ to assess the contribution of local vs. remote emissions from firewood domestic heating to the particulate matter sampled during the cold season at two measurements stations of the Environmental Agency of North Rhine-Westphalia, Germany.

20 For the investigated samples, the simulations indicate that the largest part of the sampled aerosol is 1 to 2 days old, and thus originates from local to regional sources. Consequently, photo-chemical aging, also limited by low mean OH concentrations in the cold season, has a minor influence the observed levoglucosan concentration and $\delta^{13}\text{C}$. The retro plume ages agree well with those derived from observed $\delta^{13}\text{C}$, (the 'isotopic' ages), demonstrating that the limitation of backwards calculations to seven days for this study doesn't introduce any significant bias. A linear regression analysis applied to the experimental

25 levoglucosan $\delta^{13}\text{C}$ vs. the inverse concentration confirms the young age of aerosol. The high variability in the observed $\delta^{13}\text{C}$ implies that the local levoglucosan emissions are characterized by very different isotopic ratios in the range of -25.3 to -21.4 ‰. These values are in good agreement with previous studies on levoglucosan source specific isotopic composition in biomass burning aerosol. Comparison between measured and estimated levoglucosan concentrations suggest that emissions are underestimated by a factor of two on average. These findings demonstrate that the aerosol burden from home heating in

30 residential areas is not of remote origin and thus it can be mitigated by reducing local emissions. In this work we show that combining Lagrangian modelling with isotope ratios is valuable to obtain additional insight in source apportionment. Error analysis shows that the largest source of uncertainty is limited information on isotope ratios of levoglucosan emissions.



Based on the observed low extent of photochemical processing during cold season, levoglucosan can be used under similar conditions as a conservative tracer without introducing substantial bias.

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1 Introduction

Organic aerosol (OA) has anthropogenic and biogenic sources, being either released as primary OA (POA), or formed as secondary organic aerosol (SOA). Most of the anthropogenic emissions originate from combustion of fossil fuels or biomass. The biogenic particles are predominately SOA formed by the photo oxidation of biogenic VOC. In the atmosphere, OA
40 undergoes various physical and chemical processes, such as aging by photolysis and photo oxidation, or deposition by sedimentation and wash-out. Particles have a direct radiative effect by absorbing and scattering solar radiation. Moreover, they act as cloud condensation nuclei (CCN), leading to cloud formation, which indirectly impacts the radiation budget. Being exposed to OA containing hazardous components, humans experience severe health impairments such as cardiovascular and respiratory diseases (Li et al., 2008) and references therein. Thus, OA affects air quality, health and
45 climate.

Biomass burning is an important source of OA. (Hallquist et al., 2009) estimated that biomass burning releases 42 Tg C a⁻¹ into the atmosphere, which is about a quarter of the global emitted particulate carbon. Such estimates are associated with considerable uncertainties. Parts of the uncertainties result from the lacking knowledge on source distribution and strength, as well as from the incomplete understanding of the loss processes. Since biomass burning substantially contributes to the
50 OA hazards, it is of great scientific and societal interest to accurately apportion its sources and quantify its sinks. For the source apportionment of biomass burning aerosol, factor analyses, chemical mass balance and Lagrangian techniques are employed e.g. (Busby et al., 2016;Zheng et al., 2002). Chemical mass balance modelling, used levoglucosan as the specific non-reactive molecular marker of biomass burning in aerosol e.g. (Fine et al., 2002), because it is only formed by the thermal breakdown of cellulose and it is then emitted in large quantities. The accuracy of such studies is limited by considerable
55 uncertainties in the emission factors, and by the fact that levoglucosan was recently proven to be chemically unstable. Recent laboratory studies have shown that levoglucosan reacts with OH radicals within a lifetime of few days under typical atmospheric conditions (Hennigan et al., 2010;Sang et al., 2016). This new finding opens up new potential applications, especially in the field of isotopic analyses.

Sources of biomass burning aerosol can be significantly better constrained by taking into account the stable carbon isotope ratio of levoglucosan (Gensch et al., 2014). This option is based on the fact that at the emission, levoglucosan has a source specific isotopic composition, the 'isotopic fingerprint'. Furthermore, chemical processing leads to isotopic fractionation due to the kinetic isotope effect (KIE), which is distinct for each reaction (making an 'isotopic footprint'). Consequently, highly innovative source apportionment methods aim to combine trajectory and wind - based models with isotopic analyses, which deliver additional information for validation. Recently, (Gensch et al., 2018) used isotopic measurements together with the
65 Lagrangian particle dispersion model (LPDM) FLEXPART (Stohl et al., 2010) to investigate photo-chemical aging processes in biomass burning aerosol. To this end, the photochemical age of particulate levoglucosan was derived from



observed isotopic ratios, employing the isotopic hydrocarbon clock equation on the one hand, and from back trajectory simulations on the other. For the latter, a post-modelling numerical approach was developed to describe the mixing with freshly emitted levoglucosan and to quantify this impact on the isotopic composition. The results of these two independent methods agreed well on average. Moreover, the agreement between 'isotope age' and 'retro plume age' demonstrates that the modelling results are not significantly influenced by limiting these to a few days for this study. Isotopes are thus a useful tool to evaluate the effect of the finite retro plumes. As a consequence, (Gensch et al., 2018) showed that the degree of photo oxidative aging of particulate levoglucosan can be quantified by combining laboratory KIE studies, observed isotopic composition at sources and in the field, as well as back trajectory analyses. Yet, the scatter in the individual data pairs (model vs. observation) pointed out the need to improve the identification and distinction of contributions from different source types, which is possible by using the full dispersed output of FLEXPART.

As a particular form of biomass burning, home heating with firewood is a major contributor to the fine dust in the cold season in the mid- and high latitudes. According to the German Environmental Agency (UBA), small wood stoves in the residential sector provide only 1.5% of the total energy supply, but contribute 16% to the total PM_{2.5} emissions in Germany. This is comparable to the total road traffic PM_{2.5} exhaust (Amann et al., 2018). For pollution mitigation, an accurate apportionment of local emissions versus remote transport is necessary. The main objective of this study was to implement stable carbon isotopes in the full dispersed output of FLEXPART by explicitly tracking of the levoglucosan fraction containing ¹³C. In order to determine the model performance for given conditions, the sensitivity of the simulation responses to uncertainties of the governing atmospheric processes described in FLEXPART were examined. Finally, the set of selected modelling routines were applied in a case study with the goal to assess the contribution of local vs. remote emissions from firewood domestic heating to the particulate matter (PM) sampled at two measurement stations of the North Rhine-Westphalia Environmental Agency, LANUV. Thereby, the measured levoglucosan concentration and isotopic composition in the sampled aerosol were used to evaluate the model performance.

2 Experimental

Aerosol PM_{2.5} fraction was sampled on quartz filters at two of the numerous LANUV monitoring network stations (Pfeffer et al., 2013). The sampling time was 24h and filters were daily changed at 00:00 UTC+1. For this study two sampling sites with contrasting characteristics were chosen. The 'urban background' station, hereinafter referred to as STYR, is situated in Mülheim-Styrum (51.453459°N, 6.865050°E, with site information available at <https://www.lanuv.nrw.de/luqs/messorte/pdf/STYR.pdf>), while the 'remote/rural' one, hereinafter referred to as EIFE, is located in the hilly Eifel region (50.653234°N, 6.281008°E, with site information available at <https://www.lanuv.nrw.de/luqs/messorte/pdf/EIFE.pdf>). For laboratory and model analyses, 25 pairs of aerosol filters collected on the same day in the cold seasons 2015-2017 at each of the two sites were decided on. The main criterion in selecting those was to provide a broad geographical coverage for the wind directions for the sampled air masses around the measurement sites. A list of the selected samples can be found in the



100 Supporting Information (Table S1 in section S1, containing the sampling dates, levoglucosan loading of the filters and the main origin of the sampled air masses estimated by HYSPLIT trajectory analyses).
Levoglucosan concentration was measured by ion chromatography (Kuepper et al., 2018) at the LANUV. Isotopic analyses were carried out at IEK-8, Forschungszentrum Jülich, by liquid extraction-thermal desorption-two dimensional gas chromatography, coupled with isotope ratio mass spectroscopy (LE-TD-2D-GC-IRMS) (Gensch et al., 2018). Details on the
105 experimental approaches are given in the Supporting Information (section S2). Basic statistical analysis of the measurement results can be found in Supporting Information (Table S3 in section S3).

3 Modelling stable carbon isotopes with FLEXPART

Stable carbon isotope ratios of a VOC can be calculated in numerical atmospheric models by considering ^{12}C and ^{13}C
110 isotopologues as separate species. These are treated individually during the simulations, taking into account the slightly different physical and chemical behaviour. Mostly, molecules containing ^{13}C react a little slower than the light isotopologue due to the kinetic isotope effect (KIE). Denoting the rate constants for these two reactions by ^{13}k and ^{12}k , respectively, KIE is defined as $^{12}k/^{13}k$ and is typically expressed using the epsilon notation in parts per thousand (‰):

$$\varepsilon = (1 - KIE) \times 1000 \text{‰} = \left(1 - \frac{^{12}k}{^{13}k}\right) \times 1000 \text{‰} \quad (1)$$

115 KIE can be experimentally determined in laboratory studies (Anderson et al., 2003; Sang et al., 2016). Carbon isotope effects are generally so small, that the relation between the change in the isotopic ratios and the extent of chemical processing can be linearized without introducing significant bias (Rudolph and Czuba, 2000). Based on the proposed ‘isotopic hydrocarbon clock’ equation, the photochemical age of a VOC can be determined from its isotopic composition when the source signature and the KIE of the atmospheric degradation reaction are known.

$$120 \delta^{13}\text{C}_t = \delta^{13}\text{C}_0 + t_{av}[\text{OH}]_{av} k_{OH} {}^{OH}\varepsilon \quad (2)$$

where $\delta^{13}\text{C}_0$ and $\delta^{13}\text{C}_t$ represent the isotopic composition at the source and observation site, respectively, $t_{av}[\text{OH}]_{av}$ is the average photochemical age, $[\text{OH}]_{av}$ is the average OH concentration during the photochemical processing, k_{OH} is the rate coefficient of the species of interest with OH, and ${}^{OH}\varepsilon$ is the KIE of the latter oxidation reaction.

Not only are the fractionation effects small, but also the ratio of the rare isotope to the abundant one is very low. Therefore
125 carbon isotope ratios $^{13}R = \frac{^{13}\text{C}}{^{12}\text{C}}$ are given using the delta value:

$$\delta^{13}\text{C} = \frac{(^{13}\text{C}/^{12}\text{C})_{\text{sample}} - (^{13}\text{C}/^{12}\text{C})_{\text{VPDB}}}{(^{13}\text{C}/^{12}\text{C})_{\text{VPDB}}} \times 1000 \text{‰} \quad (3)$$

where $(^{13}\text{C}/^{12}\text{C})_{\text{VPDB}}$ is the internationally accepted Vienna Peedee Belemnite (VPDB) value of 0.0111828 (Brand et al., 2010; Craig, 1957). The source specific carbon isotope ratios of atmospheric trace organic components are introduced in the simulations considering the emission data for the investigated VOC (details in the following sections).



130 In this study, detailed information on origin and pathway of the collected aerosol particles was obtained by calculating retro
plumes backwards from the sampling sites with the LPDM FLEXPART, Version 10.2beta (Seibert and Frank, 2004;Stohl et
al., 2010), source code available at git@git.nilu.no:flexpart/flexpart.git). ECMWF 3 hourly data with a resolution of $1^\circ \times 1^\circ$
on 91 vertical levels was used as driving meteorology (Owens and Hewson, 2018). For every investigated day, ca. 200 000
model particles were hourly released for 24h at the measurement stations. Levoglucosan (LG) was implemented as an
135 aerosol biomass burning tracer which is subjected to photochemical degradation by OH, as well as to wet and dry deposition.
The aerosol particle population defined in the input has a lognormal size distribution with $\bar{d} = 0.25 \mu\text{m}$, $\sigma_{\bar{d}} = 1.5$ and a
density of 1.4 kg m^{-3} (Fiebig et al., 2003). To simulate the wet aerosol particle removal, the new deposition module from
(Grythe et al., 2017) was activated, using three dimensional cloud water fields from the ECMWF data. For below-cloud
scavenging, a coalescence probability of one was set for both rain and snow (Grythe et al., 2017). Since anthropogenic
140 biomass burning aerosol is emitted into the lower mixing layer, in-cloud scavenging is not likely. For the chemical loss, the
OH-decay rate constant was set to $2.67 \times 10^{-12} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$ (Sang et al., 2016). The model was run backwards over 7 days.
This is at the higher end of the expected levoglucosan life time for a mean OH concentration of $0.5 \times 10^6 \text{ molec cm}^{-3}$ in the
cold season in Europe (Gerasopoulos et al., 2012;Rohrer and Berresheim, 2006). The output resolution was set to $0.25^\circ \times$
 0.25° on 5 vertical levels. The resulting retro plumes entail the source-receptor-relationship reflecting the deposition- and
145 decay-corrected receptor sensitivity to potential upwind sources. See Supporting Information (section S4) for details.
To investigate the contribution of relevant domestic heating sources to the biomass burning aerosol sampled at the receptors,
the retro plumes were folded with monthly-mean-gridded levoglucosan emissions during the cold season in Europe (Seibert
and Frank, 2004). The emission inventories were derived from firewood consumption in the targeted European countries,
population density and levoglucosan emission factors of firewood burned in common wood stoves (Akagi et al., 2011;Fine et
150 al., 2004;Jimenez et al., 2017;Schauer et al., 2001), Firewood consumption by UN data: data.un.org, access March 10th
2017, population density data by NASA: neo.sci.gsfc.nasa.gov, access February 6th 2017. Details are given in the
Supporting Information (section S5). Domestic heating is the main source of the sampled levoglucosan in this study, since
FIRMS (Fire Information for Resource Management System a NASA product: firms.modaps.eosdis.nasa.gov, access
January 10th 2019) fire inventories show no larger scale open fires affecting the sampling during the considered periods. To
155 determine the receptor sensitivity to home heating emissions, a dynamic footprint layer from 100 to 300m height was
considered (Hüser et al., 2017). The resulting folded retro plumes quantify the contribution of each individual source in kg
 m^{-3} to the levoglucosan sampled at the measurement site. The corresponding concentration is then derived by summing up all
contributions.
Levoglucosan $\delta^{13}\text{C}$ at the sampling sites were determined by introducing ^{13}C -LG as an additional model tracer (Gensch et al.,
160 2011;Stein and Rudolph, 2007). It has the same physical and chemical properties as ^{12}C -LG, except for a reduced OH
reactivity due to the kinetic isotope effect. In the simulations, the rate of the ^{13}C -LG chemical loss was derived by using KIE
 $= 1.00229 \pm 0.00018$ (Sang et al., 2016). ^{13}C -LG at the source was calculated using ^{12}C -LG emissions together with the



specified source isotopic ratios. The isotopic composition at the sampling point was derived as the ratio between the slightly different $^{13}\text{C-LG}/^{12}\text{C-LG}$ retro plumes folded with the corresponding $^{13}\text{C-LG}/^{12}\text{C-LG}$ emission inventories (for details see
165 Supporting Information, Sections S4, S5).

4 Results and discussion

FLEXPART sensitivity studies

Different FLEXPART modules, describing chemical decay, dry and wet deposition were successively activated. Modelling
170 parameters, such as meteorology, levoglucosan lifetime and emission data were varied to reveal the governing simulated processes and to assess the modelling performance. Changes in output depending on the input parameter modifications were evaluated for consistency.

Driving meteorological input data

Unpredictability of the driving meteorology is one of the major error sources in the Lagrangian modelling (Angevine et al.,
175 2014; Davis and Dacre, 2009; Lin, 2013), leading to concentration uncertainties of up to 40%. In this study, two global numerical weather prediction models, ECMWF and Global Forecast System (GFS, with the same horizontal resolution, but with 39 vertical levels) (Global Climate & Weather Modeling Branch, 2003), delivered the input meteorological fields for two otherwise identical simulations with FLEXPART. The derived concentration and $\delta^{13}\text{C}$ values were compared.

Tables S6.1 and S6.2 present model results obtained with the two meteorology sets (Supporting Information, section S6).
180 Differences in the isotopic composition are well below the experimental uncertainty of 0.6‰. Levoglucosan concentration calculations based on ECMWF fairly agree with the measurements, showing mean deviations of 9.1 and 4.2% for the EIFE and STYR datasets, respectively. In general, the concentrations obtained when using the GFS meteorology are higher than for ECMWF initiated calculations, with a difference mean of 6.9 and 4.6%, respectively (Supporting Information, Figures 6.1 and 6.2).

185 Two extreme cases are presented in Figure 1. For February 18th, 2016, the simulations show a 1.5% deviation between the concentrations yielded using the ECMWF and GFS meteorology. This difference is largest (86.2%) for March 5th, 2016. In the former example, similar mixing heights (H_{mix}) and mixing events (when the centroid height $H_{\text{traj}} < H_{\text{mix}}$) are predicted. For the latter case, higher GFS H_{mix} cause more frequent mixing events and thus, fresh emission entrainment in the sampled air masses. Additionally, mixing heights up to 2500 m imply strong convection, which might result in higher residence time of
190 model particles in the footprint layer and thus in higher derived levoglucosan concentrations.

Overall, the simulated concentration and $\delta^{13}\text{C}$ based on ECMWF meteorological fields show good agreement with the values obtained when using GFS data, the uncertainties being in the same range as those of the experimental observations. Small differences between the two models due to their vertical mixing parameterizations are expected during the more stable cold season over Europe. Due to the higher vertical resolution and ability to more accurately account for topography, ECMWF
195 meteorology was chosen to initialize the model for the further runs.



Photo-chemical aging

Based on the FLEXPART full dispersed output of the 'inert tracer' scenario, the retro plumes were divided into age classes, separating the data by the time the particles in the considered portion need until they reach the sampling point. Percentages of the total levoglucosan contribution to the sampled air were calculated. The results show that emissions during the last 24 h before sampling contribute on average 49 % of the sampled aerosol, being considered 'one day old'. Detailed model results are given in Table S7.1 (Supporting Information, section S7). 30 % of the contributed emissions occurred between 48 and 24h, 10 % between 72 and 48 h before sampling, being considered 'two and three days old', respectively. Only 11 % of the total emitted particles were older than three days. The simulations thus show that the major part of the sampled aerosol originates from local sources being emitted during the sampling day and the day before. Exemplarily, Figure 2 shows emission contributions of different ages for April 1, 2017. Due to the young age of the sampled aerosol and the typically low OH concentration in the cold season photo-chemical aging is not expected to be the governing loss process in this study. To investigate the influence of the photo-chemical aging of the sampling aerosol, simulations were carried out implementing the chemical degradation of levoglucosan as described by (Sang et al., 2016). Indeed, differences between the results of the 'inert' and 'reactive tracer' scenarios, are on average 10% and 7% for the EIFE and STYR stations, respectively, both being within the experimental error range. Changes in the isotopic composition are at both stations 0.2 ‰ on average, being smaller than the measurement precision.

Overall, under the investigated conditions, both concentration and isotopic composition at the sampling site are rather determined by mixing with fresh emissions than by chemical loss processes. This agrees well with the study by (Busby et al., 2016), which pointed out that levoglucosan is relatively stable during winter due to the low OH concentration.

215 Deposition

Simulations including dry as well as dry and wet deposition were carried out. The calculated concentration and isotope ratios were compared with the 'reactive tracer' – without deposition – scenario to quantify the contribution of dry and wet removal, respectively.

The initialized aerosol particle population has a mean diameter of 0.25 μm with a corresponding average settling velocity is 220 5.6×10^{-6} m/s. Since the aerosol in this study is relatively young, most of it stays in the accumulation mode. Thus, gravitational settling concentration losses are negligible, amounting to 1.5 and 0.7% at the EIFE and STYR stations, respectively. Consequently, changes in the isotopic composition are minor as well (details in Supporting Information, section S8).

Further, the existent simulations show that wet deposition removed minimal amounts of the emissions. This might be 225 explained by a short exposure of aerosol to weak precipitation of less than 5mm in 6h in the investigated periods. Moreover, due to the low injection height of the levoglucosan emissions from domestic heating, there are no occurrences of in-cloud scavenging. Wet deposition had no significant influence on the isotopic composition of the sampled aerosol either.

Domestic biomass burning emission estimates



Given the lack of levoglucosan emission data from residential heating an approach was developed here to estimate these
230 from available information. Therefore, population density, country-specific firewood consumption and levoglucosan
emission factors for typical fuel used in the residential heating were considered (details in the Supporting Information,
section S5). Accordingly, uncertainties arising from potential spatial and temporal variability of the emission intensities are
brought into the calculations, such as errors due to seasonal and regional differences in the wood acquisition and
consumption. By considering the monthly mean consumption of firewood, which is provided by a personal survey, weekly to
235 diurnal variances in the emissions are neglected. Additionally, weighting with the population density doesn't reflect the real
spatial distribution of the wood consumption. As an example, fireplace heating is rather unusual in cities with high
population densities. The conversion factor from wood weight to levoglucosan emission depends on the wood type and the
combustion process (Akagi et al., 2011; Fine et al., 2004; Jimenez et al., 2017; Schauer et al., 2001).

The injection height of a fire emission is usually parametrized based on exhaust magnitude and temperature. These are quite
240 similar for domestic woodstoves, unlike in the case of large open fires. According to (Zhang et al., 2014), the footprint layer
for domestic heating emissions stretches from 100m - 300 m. The model results are not sensitive to the footprint layer height
as long as it is inside an effectively mixed layer (Hüser et al., 2017). This is valid for the investigated period, since the
FLEXPART simulated mixing heights drop under 300m in less than 10% of cases. Furthermore, increasing the thickness of
the footprint layer has no major influence on the model outcome because of two counteracting effects. A greater dilution
245 reduces the impact of a source, while a wider spread of the emission increases the residence time of model particles in the
footprint layer.

Case study

Based on the sensitivity study findings, modelling routines and parameters were selected and applied in a case study aiming
250 to assess the contribution of local vs. remote emissions from firewood domestic heating to the aerosol sampled at EIFE and
STYR sites.

The measured levoglucosan concentration for the investigated samples varies over more than one order of magnitude (from ~
10 to 500 ng m⁻³), being overall lower at EIFE (54.2 ng m⁻³ on average) than at STYR (152.1 ng m⁻³ on average). The
histograms depicted in Figure 3 show further differences between the sites with a unimodal vs. a multimodal distribution of
255 the concentrations observed at EIFE and STYR, respectively. According to this, there are two major types of sources
contributing to the sampled aerosol. The urban site region is affected not only by regional upwind emissions, like the remote
one, but also by sources very near to the receptor.

Measured versus simulated concentrations are depicted in Figure . Generally, the model results are at the lower end of the
observation range, showing mean absolute percentage deviations of 42.0 and 53.8% for the EIFE and STYR sites,
260 respectively. For the rural station, the model overestimates in 20% and underestimates in 44% of cases the concentration of
levoglucosan outside the experimental error ranges. By contrast, at the urban stations, there is only one case of model



overestimation. Here, the model predominantly underestimates levoglucosan concentrations, far outside the experimental error ranges.

Figure shows the measured δ -values of the sampled levoglucosan for both measurement stations. The isotopic composition ranges between -26.3‰ and -21.3‰ . In theory, such high variance can be explained either by different source specific isotopic ratios of the contributing emissions, or by different extent of the chemical processing of levoglucosan in the sampled aerosol.

We previously showed that most contributing emissions are only one day old. Chemical decay is insignificant for this study also due to the low OH concentrations in the cold season. Thus, mixing of sources characterized by manifold isotopic composition likely explain the observations. The $\delta^{13}\text{C}$ distribution shown in Figure 3 is narrower for the EIFE site than for STYR, where mixing of individual different sources seems likely.

A model-observation-comparison analysis gives the possibility to assign individual isotopic signatures to different source regions. Within this study, retro plume analyses providing the main wind direction were used to confine geographically source regions for the sampled aerosol (details in the Supporting Information, section S9). Weighted mean measured δ -values of these assigned source regions are given in Figure 6. Within measurement error ranges there is no significant difference in $\delta^{13}\text{C}$ for the different source regions, indicating that there is on average no significant spatial dependence of the isotopic signature of levoglucosan emissions.

Keeling plot analysis

Due to absence of chemical degradation of levoglucosan in aerosol particles originating from domestic heating in the cold season, a Keeling plot approach was at first employed to investigate the mixing of two reservoirs, in this study of fresh - 'isotopically-light', high-concentrated emissions with aged - 'heavier', low-concentrated background (Lin, 2013). To this end, the measured isotopic ratio was plotted vs. the inverse concentration (Figure 7). A linear regression analysis was carried out. Remarkably, according to the 95% confidence interval analysis, the yielded y-intercept range of -25.3 to -21.4‰ agrees well within error ranges with the published isotopic composition measured in aerosol from the combustion of various C3 plants (Sang et al., 2012). This analysis shows that the variability in the observed δ -values is likely due to the contribution of local sources that possess very different isotopic ratios in the above mentioned range. A y-intercept of $-23.2\pm 0.1\text{‰}$ was derived, being defined as the isotope ratio of fresh emissions. Accordingly, the model runs were initialized using a $\delta^{13}\text{C}_0$ of -23.2‰ . The slope of the fitted line to the experimental data was found to be negative. Since background levoglucosan concentration data were not available, the lowest measured concentration (12.4 ng m^{-3} at the EIFE station on November 10th, 2015) was considered as constant background value. A corresponding $\delta^{13}\text{C}$ of $-24.0\pm 0.3\text{‰}$ was calculated (Figure 7).

This isotopic ratio, lower than that of sources, cannot describe the 'expected' photo-chemically-aged background aerosol. This result calls into question whether the initially postulated reservoirs explain the observed concentrations and delta values. The assumed background levoglucosan can originate either from a diffuse source that is not related to the population



295 density, or from air masses that are not taken into account by the 7 day retro plumes. The former hypothesis is not plausible
in winter. The latter is likely in terms of the concentrations, but not compatible with the delta values, since the isotopic ratio
of 7-days older levoglucosan increases by 2‰. The Keeling plot is therefore a strong indication that a significant background
cannot be reconciled with the observed isotope values. Thus, a systematic underestimation of the source strength might be a
better explanation for the observed data. In the first model runs, the background levoglucosan concentration and isotopic
ratio were set to 12.4 ng m⁻³ and -24.0 ‰, respectively. Subsequently, model runs were repeated by gradually reducing the
300 background. The best fit between the observed and calculated concentrations was reached by setting the background to zero
(Supporting Information, Table 8.3). The linear regression analysis additionally shows a significant difference in the
emission factors between EIFE and STYR. Nevertheless, emissions are underestimated by a factor of two on average.

The modelled $\delta^{13}\text{C}$ results shown in Figure 7 are located next to the line fitted to the observations. Based on the isotopic
305 hydrocarbon clock equation (Eq. 3), the 'isotopic age' was calculated from modelled and observed levoglucosan isotopic
ratios, using a $\delta^{13}\text{C}_0$ of -23.2‰ and kinetic data from (Sang et al., 2016). The mean OH concentration was considered
 0.5×10^6 molec cm⁻³. This yields a 'negative' age for the observations (Supporting Information, Table 7.3), probably a
consequence of the inaccuracy of the used emission isotope ratio. The trajectory model has an age of around 1.5 days, with a
small but very significant difference between EIFE and STYR (EIFE is older, as expected). The observed delta values show
310 a similar difference under the premise used here that the emissions contributing to EIFE and STYR have the isotopic
composition. This is to be expected for the trajectory analyses in this study, since the mean age is greater than one day and
therefore is regionally averaged. Moreover, due to the number and duration of the sample collection, the averaging includes
also a variety of prevailing conditions. In reverse, when assuming that the trajectory analyses correctly reflect the mean age,
one can determine the emission isotopic ratios from the observations. Noticeably here, the values for EIFE and STYR are
315 almost identical supporting the good averaging conclusion.

Overall, this model-observation-comparison study show good agreement between the model and observations. This
demonstrates that FLEXPART well describes the atmospheric processes investigated in this study. Despite unknowns
expected to introduce biases in the analyses, FLEXPART simulations provide a good description of the sources and
background.

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5 Conclusions

In this study we have combined Lagrangian particle dispersion modeling with concentration- and isotopic measurements of
levoglucosan in PM sampled during the cold season at two LANUV stations as innovative tool to investigate home heating
aerosol sources and its fate. To this end, we have successfully implemented ¹²Levoglucosan and ¹³Levoglucosan as separate
325 chemical species into the LPDM FLEXPART, to calculate the isotopic ratio distribution of the specific biomass burning
tracer from the source to the sampling site. The analysis of the full dispersed model output, in combination with emission
inventories using the ,folded retro plume technique 'yielded very detailed information on the source–receptor relationships.
Thus, aerosol source contribution to the receptor sites and its loss processes during the atmospheric transport are quantified.



Sensitivity studies show that for this special case, varying model variables, such as meteorology, rate constant of the photo-oxidation reaction or deposition processes, does not yield significant changes in the simulation results. This is justified by similar vertical mixing parameterization in the wind models during the more stable cold season over Europe. Lower OH concentrations cause less photo-chemical degradation of levoglucosan. This can be also associated with the young age of the sampled aerosol derived from the simulations. Furthermore, sedimentation is an insignificant loss process for the fresh biomass burning aerosol. The activation of the wet deposition module leads to minor levoglucosan concentration reductions due to missing strong precipitation events in winter. Young levoglucosan can be explained either by deposition, or by local / regional sources and dispersion. Since dry and wet deposition are insignificant, the former hypothesis is the most likely.

The presented case study shows good agreement between modelled and observed data within error ranges. The few overestimations of the derived concentration at EIFE might be caused by overestimated emissions rather than underestimated removal (Grythe et al., 2017). The frequent underestimations, especially at STYR might indicate unidentified sources or flaws in the levoglucosan emission strength. This comparison supports the fact that sources which are very close but not captured in the developed emission inventory approach strongly influence the local aerosol burden, particularly for the STYR site. Repeated calculations reducing the background and increasing the emissions indicated that the source strength is underestimated. The measured $\delta^{13}\text{C}$ -values show by far higher variability compared with the simulated isotopic ratios. This can be explained by possible individual source to source variation (e.g. due to differences in the used fuel). The retro plume modelled age of levoglucosan agrees well with the age resulting from the observed isotopic ratios. This agreement demonstrates that the limitation to 7-days backwards calculations does not create any significant bias. Finally, since both observations and the retro plume analyse show that chemical aging does not play a significant role in the cold season in Central Europe, levoglucosan can be used as a 'conservative' tracer under similar conditions. All these findings demonstrate the FLEXPART fitness to simulate aerosol processes occurring between source and receptor. The sensitivity studies revealed individual factors leading to potential biases, while the comparison between simulated and observed concentration assessed the most probable sources and loss processes for the investigated aerosol.

Both sensitivity and case studies unquestionably point out that local domestic heating is the major source contributing to the biomass burning aerosol burden under the investigated conditions. Thus we have demonstrated that the developed modelling strategies are suitable to assess sources of biomass burning aerosol in living areas in winter. Local emission mitigation is possible. Under similar conditions, i.e. cold season in Europe, photo-chemical decay is negligible, therefore levoglucosan can be employed as inert tracer in source-receptor studies, without introducing considerable bias.

For the future, more isotopic measurements of fuels used for domestic heating in Europe are essential to better constrain the isotopic signature of individual sources. Global modelling together with more frequent ambient measurements is necessary to describe more accurately the concentration and isotopic composition of background aerosol. Further studies preferably on summer fires are needed to test processes described by FLEXPART. Isotopic information will likely deliver the additional information to quantify aerosol photo-chemical aging in a OH-radical-rich atmosphere, as well as the wet deposition during strong precipitation leading to heavy levoglucosan removal.



Data availability.

365 Data are available from the corresponding author on request (i.gensch@fz-juelich.de).

Author contribution.

IG, AKS and TK were involved in research planning and experimental design. CB and IG developed the modelling approach, CB performed the simulations. CK performed the levoglucosan isotopic ratio measurements. CB, IG and JR
370 analysed and synthesized the experimental and simulation data. CB wrote the paper together with IG. All other co-authors participated in data collection, experiment operations and participated in paper discussion.

Competing interests.

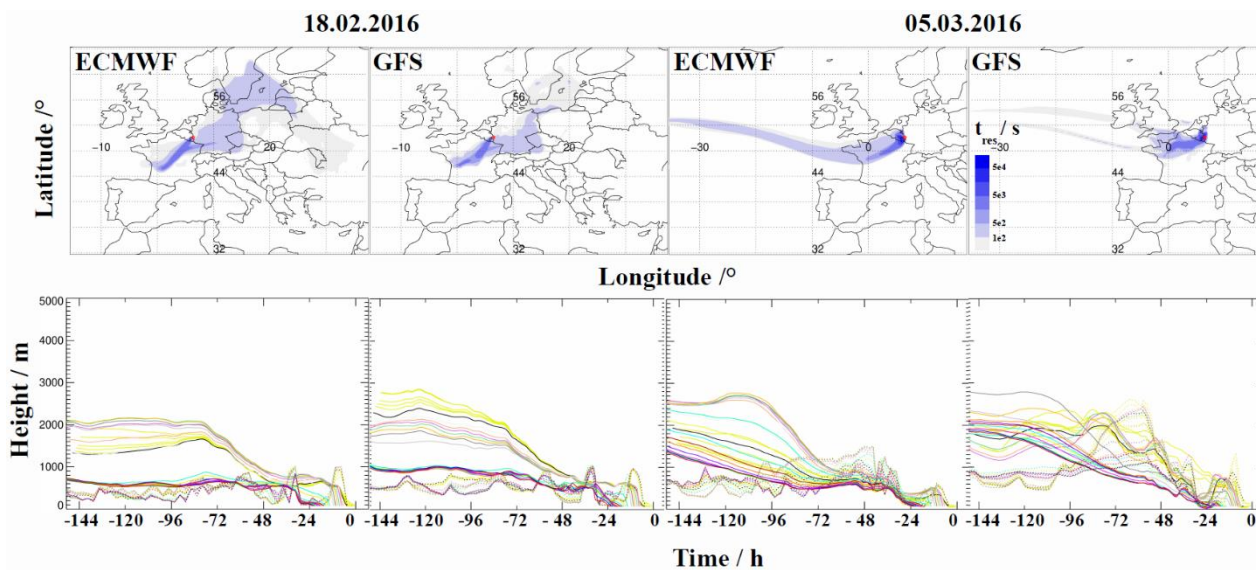
The authors declare that they have no conflict of interest.

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Figure 1: Simulated retrorplumes (top, for the colour code see legend) as well as centroid-back-trajectory (bottom, solid lines) and the corresponding mixing-layer heights (bottom, dashed lines) for each release hour, using the ECMWF and GFS meteorology, exemplarily presented for the February 18th(left), and the March 5th 2016 (right) samples. The FLEXPART analyses are made for EIFE site (red star).

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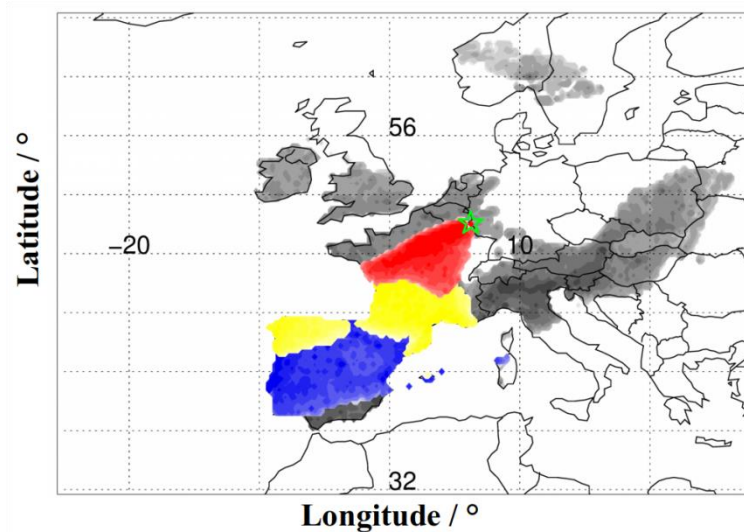


Figure 2: Levoglucosan emission contribution to the EIFE site (green star), divided into age classes, for 1 April 2017. The plot enlightens source regions for one- (35% of the collected levoglucosan, red), two- (35%, yellow), three- (11%, blue) days old particles, as well as older than that (20%, grey).

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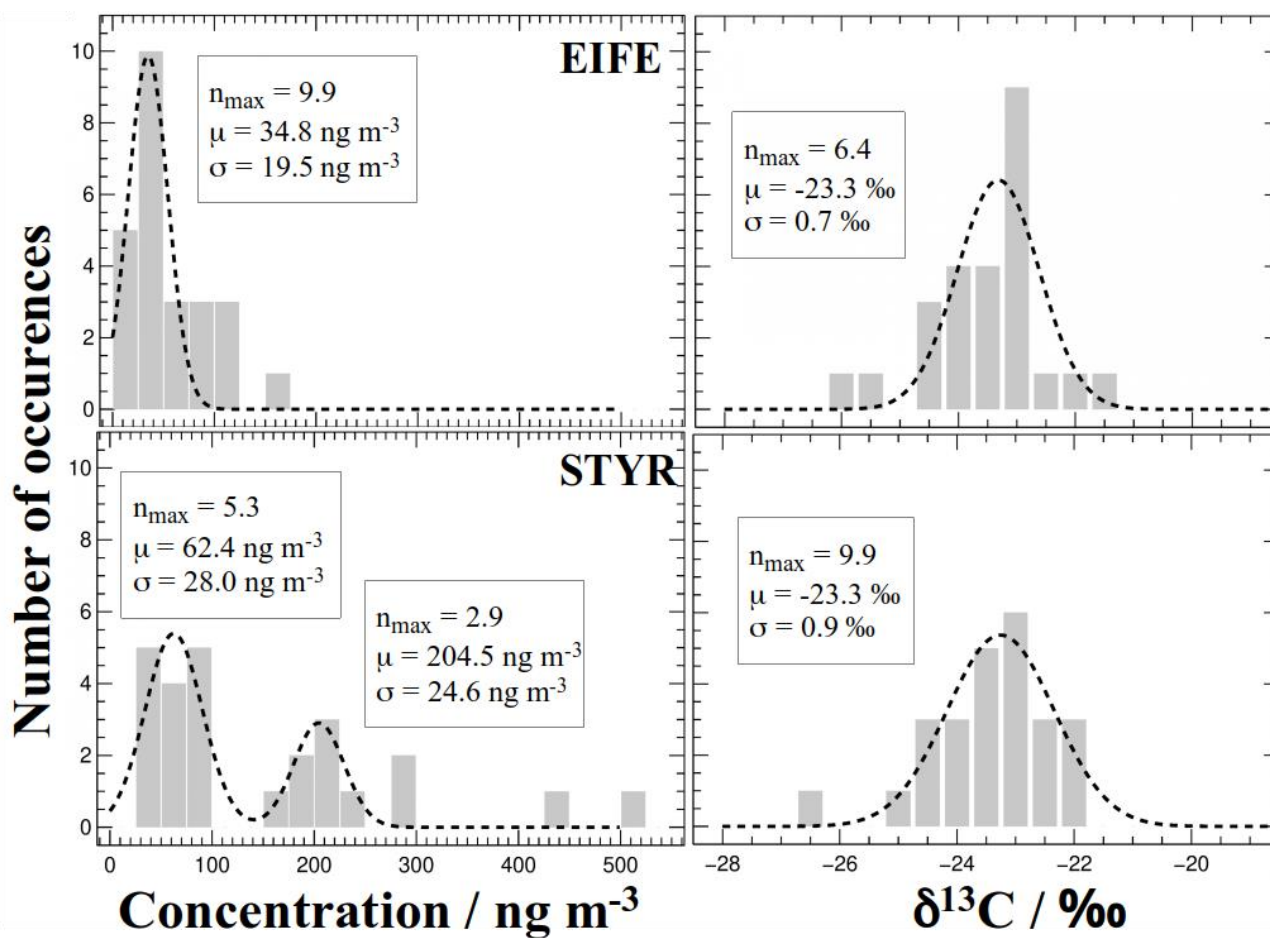
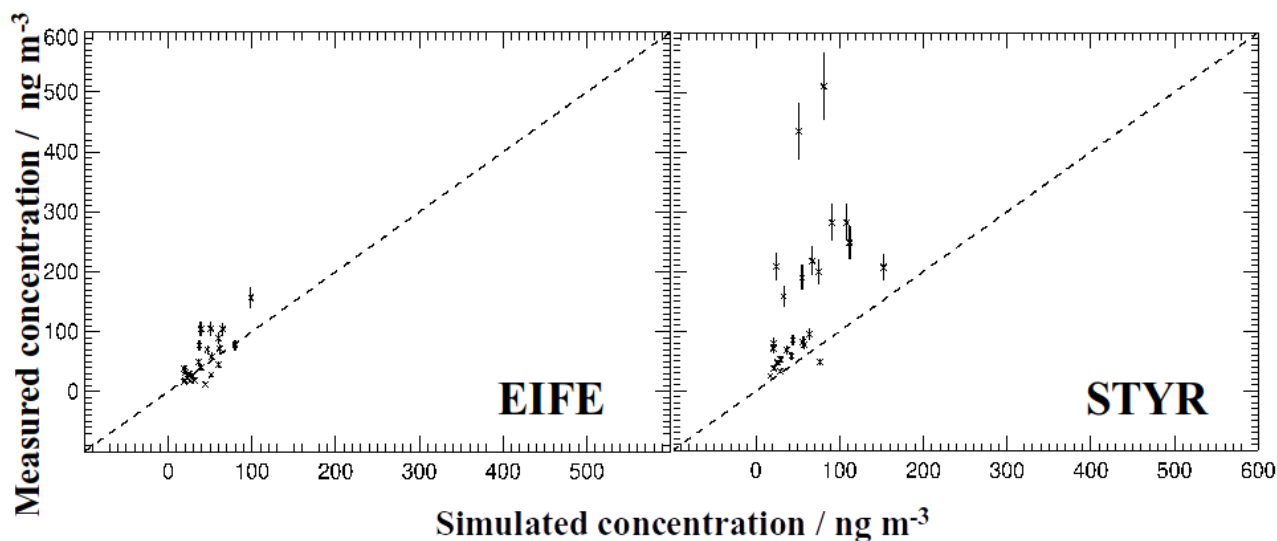
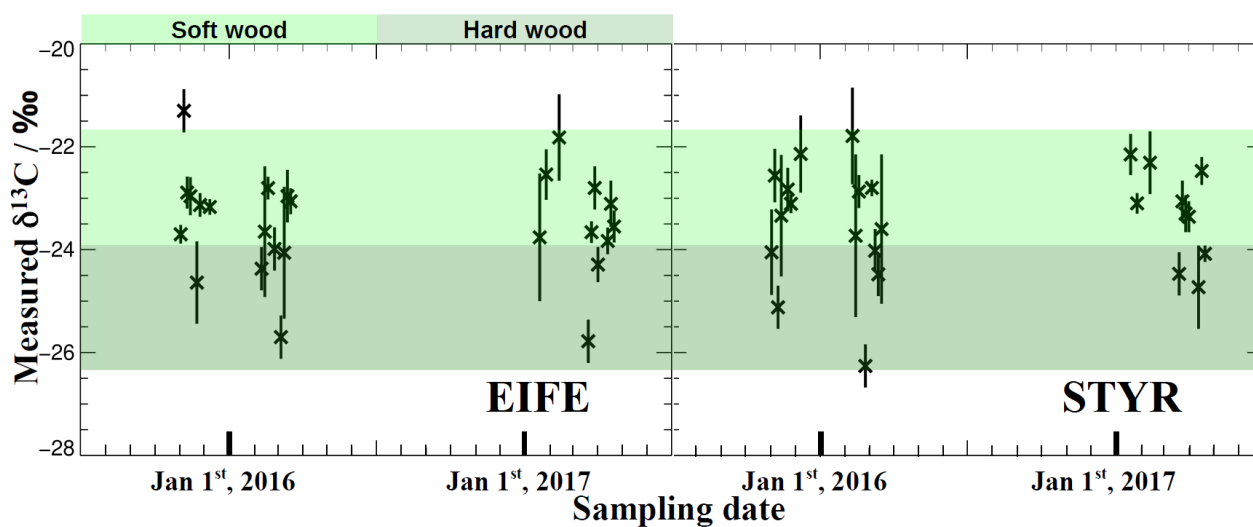


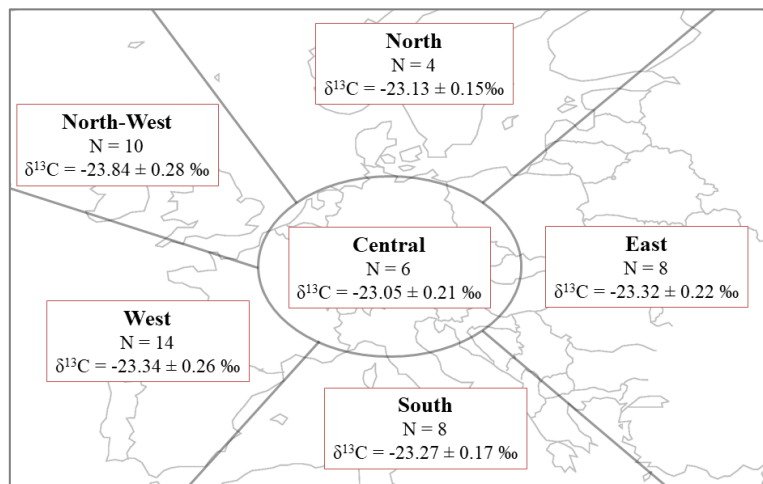
Figure 3: Distribution frequency of the observed levoglucosan concentration and $\delta^{13}\text{C}$.



405 **Figure 4:** Comparison between observed and simulated levoglucosan concentration at the EIFE and STYR sites. Also the 1:1 line is given (dashed line).

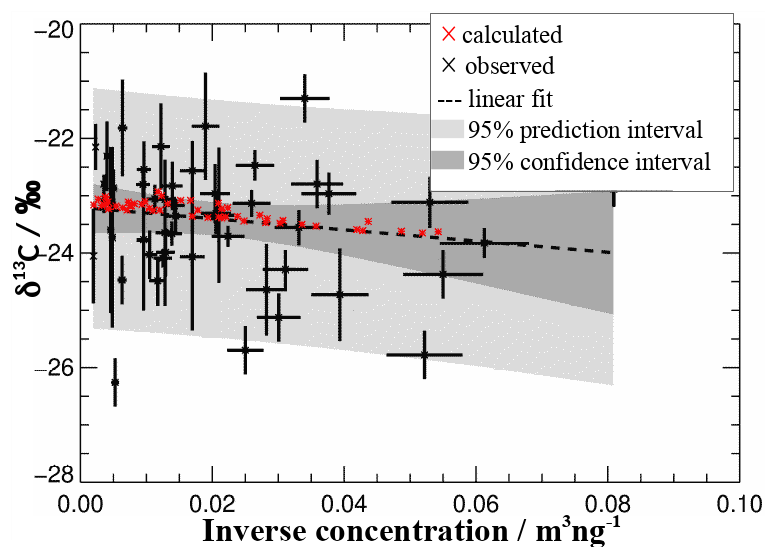


410 **Figure 5:** $\delta^{13}\text{C}$ of the sampled levoglucosan at EIFE and STYR sites. The shaded areas represent ranges of observed levoglucosan source specific isotope ratios in aerosol formed during the combustion of soft (light green) and hard wood (dark green) (Sang et al., 2012).



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Figure 6: Isotopic signatures for different source regions, including corresponding sample number as well as the average $\delta^{13}\text{C}$. The source regions for the sampled aerosol were designated based on the the main wind direction (see text).



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Figure 7 Keeling plot depicting the observed levoglucosan $\delta^{13}\text{C}$ vs the inverse concentration (black symbols). Model results are given (red symbols) as well as the line fitted to the experimental data (dashed). The light and dark grey shaded areas represent the 95% prediction and confidence intervals, respectively.

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- 430 Akagi, S. K., Yokelson, R. J., Wiedinmyer, C., Alvarado, M. J., Reid, J. S., Karl, T., Crouse, J. D., and Wennberg, P. O.: Emission factors for open and domestic biomass burning for use in atmospheric models, *Atmos. Chem. Phys.*, 11, 4039-4072, 10.5194/acp-11-4039-2011, 2011.
- Amann, M., Cofala, J., Klimont, Z., Nagl, C., and Schieder, W.: Measures to address air pollution from small combustion sources, Environment Agency Austria, International Institute for Applied Systems Analysis, 51, 2018.
- Anderson, R. S., Czuba, E., Ernst, D., Huang, L., Thompson, A. E., and Rudolph, J.: Method for Measuring Carbon Kinetic Isotope Effects of Gas-Phase Reactions of Light Hydrocarbons with the Hydroxyl Radical, *J. Phys. Chem. A*, 107, 6191-6199, 10.1021/jp034256d, 2003.
- 435 Angevine, W. M., Brioude, J., McKeen, S., and Holloway, J. S.: Uncertainty in Lagrangian pollutant transport simulations due to meteorological uncertainty from a mesoscale WRF ensemble, *Geoscientific Model Development*, 7, 2817-2829, 10.5194/gmd-7-2817-2014, 2014.
- Brand, W. A., Assonov, S. S., and Coplen, T. B.: Correction for the ^{17}O interference in $\delta(^{13}\text{C})$ measurements when analyzing CO_2 with stable isotope mass spectrometry (IUPAC Technical Report), *Pure Appl. Chem.*, 82, 1719-1733, 10.1351/pac-rep-09-01-05, 2010.
- 440 Busby, B. D., Ward, T. J., Turner, J. R., and Palmer, C. P.: Comparison and Evaluation of Methods to Apportion Ambient $\text{PM}_{2.5}$ to Residential Wood Heating in Fairbanks, AK, *Aerosol and Air Quality Research*, 16, 492-503, 10.4209/aaqr.2015.04.0235, 2016.
- 445 Craig, H.: Isotopic standards for carbon and oxygen and correction factors for mass-spectrometric analysis of carbon dioxide, *Geochimica et Cosmochimica Acta*, 12, 133-149, [http://dx.doi.org/10.1016/0016-7037\(57\)90024-8](http://dx.doi.org/10.1016/0016-7037(57)90024-8), 1957.
- Davis, L. S., and Dacre, H. F.: Can dispersion model predictions be improved by increasing the temporal and spatial resolution of the meteorological input data?, *Weather*, 64, 232-237, 10.1002/wea.421, 2009.
- 450 Fiebig, M., Stohl, A., Wendisch, M., Eckhardt, S., and Petzold, A.: Dependence of solar radiative forcing of forest fire aerosol on ageing and state of mixture, *Atmospheric Chemistry and Physics*, 3, 881-891, 10.5194/acp-3-881-2003, 2003.
- Fine, P. M., Cass, G. R., and Simoneit, B. R. T.: Chemical Characterization of Fine Particle Emissions from the Fireplace Combustion of Woods Grown in the Southern United States, *Environmental Science & Technology*, 36, 1442-1451, 10.1021/es0108988, 2002.
- 455 Fine, P. M., Cass, G. R., and Simoneit, B. R. T.: Chemical Characterization of Fine Particle Emissions from the Wood Stove Combustion of Prevalent United States Tree Species, *Environmental Engineering Science*, 21, 705-721, 10.1089/ees.2004.21.705, 2004.
- Gensch, I., Laumer, W., Stein, O., Kammer, B., Hohaus, T., Saathoff, H., Wegener, R., Wahner, A., and Kiendler-Scharr, A.: Temperature dependence of the kinetic isotope effect in β -pinene ozonolysis, *Journal of Geophysical Research*, 116, 10.1029/2011JD016084, 2011.
- 460 Gensch, I., Kiendler-Scharr, A., and Rudolph, J.: Isotope ratio studies of atmospheric organic compounds: Principles, methods, applications and potential, *Int. J. Mass Spectrom.*, 365-366, 206-221, 10.1016/j.ijms.2014.02.004, 2014.
- Gensch, I., Sang-Arlt, X. F., Laumer, W., Chan, C. Y., Engling, G., Rudolph, J., and Kiendler-Scharr, A.: Using $\delta^{13}\text{C}$ of Levoglucosan As a Chemical Clock, *Environ. Sci. Technol.*, 52, 11094-11101, 10.1021/acs.est.8b03054, 2018.
- 465 Gerasopoulos, E., Kazadzis, S., Vrekoussis, M., Kouvarakis, G., Liakakou, E., Kouremeti, N., Giannadaki, D., Kanakidou, M., Bohn, B., and Mihalopoulos, N.: Factors affecting O_3 and NO_2 photolysis frequencies measured in the eastern Mediterranean during the five-year period 2002-2006, *J. Geophys. Res.: Atmos.*, 117, D22305/22301-D22305/22314, 22314 pp., 10.1029/2012JD017622, 2012.
- Global Climate & Weather Modeling Branch, E.: The GFS atmospheric model, NCEP Office Note 442, 14, 2003.
- 470 Grythe, H., Kristiansen, N. I., Zwaafink, C. D. G., Eckhardt, S., Stroem, J., Tunved, P., Krejci, R., and Stohl, A.: A new aerosol wet removal scheme for the Lagrangian particle model FLEXPART v10, *Geosci. Model Dev.*, 10, 1447-1466, 10.5194/gmd-10-1447-2017, 2017.
- Hallquist, M., Wenger, J. C., Baltensperger, U., Rudich, Y., Simpson, D., Claeys, M., Dommen, J., Donahue, N. M., George, C., Goldstein, A. H., Hamilton, J. F., Herrmann, H., Hoffmann, T., Iinuma, Y., Jang, M., Jenkin, M. E., Jimenez, J. L., Kiendler-Scharr, A., Maenhaut, W., McFiggans, G., Mentel, T. F., Monod, A., Prevot, A. S. H., Seinfeld, J. H., Surratt, J. D., Szmigielski, R., and Wildt, J.: The formation, properties and impact of secondary organic aerosol: current and emerging issues, *Atmos. Chem. Phys.*, 9, 5155-5236, 2009.



- Hennigan, C. J., Sullivan, A. P., Collett, J. L., and Robinson, A. L.: Levoglucosan stability in biomass burning particles exposed to hydroxyl radicals: LEVOGLUCOSAN STABILITY IN AEROSOL, *Geophysical Research Letters*, 37, n/a-n/a, 10.1029/2010GL043088, 2010.
- 480 Hüser, I., Harder, H., Heil, A., and Kaiser, J. W.: Assumptions about footprint layer heights influence the quantification of emission sources: a case study for Cyprus, *Atmospheric Chemistry and Physics*, 17, 10955-10967, 10.5194/acp-17-10955-2017, 2017.
- Jimenez, J., Farias, O., Quiroz, R., and Yañez, J.: Emission factors of particulate matter, polycyclic aromatic hydrocarbons, and levoglucosan from wood combustion in south-central Chile, *Journal of the Air & Waste Management Association*, 67, 485 806-813, 10.1080/10962247.2017.1295114, 2017.
- Kuepper, M., Quass, U., John, A. C., Kaminski, H., Leinert, S., Breuer, L., Gladtko, D., Weber, S., and Kuhlbusch, T. A. J.: Contributions of carbonaceous particles from fossil emissions and biomass burning to PM10 in the Ruhr area, Germany, *Atmos. Environ.*, 189, 174-186, 10.1016/j.atmosenv.2018.06.039, 2018.
- Li, N., Xia, T., and Nel, A. E.: The role of oxidative stress in ambient particulate matter-induced lung diseases and its implications in the toxicity of engineered nanoparticles, *Free Radical Biol. Med.*, 44, 1689-1699, 490 10.1016/j.freeradbiomed.2008.01.028, 2008.
- Lin, J. C.: An Introduction, in: *Lagrangian Modeling of the Atmosphere*, edited by: Lin, J., Brunner, D., Gerbig, C., Stohl, A., Luhar, A., and Webley, P., John Wiley & Sons, Washington, DC, 2013.
- Owens, R. G., and Hewson, T.: *ECMWF Forecast User Guide*, in: ECMWF, Reading, 2018.
- 495 Pfeffer, U., Breuer, L., Gladtko, D., and Schuck, T. J.: Contribution of wood burning to the exceedance of PM10 limit values in north rhine-westphalia, *Gefahrstoffe - Reinhalt. Luft*, 73, 239-245, 2013.
- Rohrer, F., and Berresheim, H.: Strong correlation between levels of tropospheric hydroxyl radicals and solar ultraviolet radiation, *Nature (London, U. K.)*, 442, 184-187, 10.1038/nature04924, 2006.
- Rudolph, J., and Czuba, E.: On the use of isotopic composition measurements of volatile organic compounds to determine the “photochemical age” of an air mass, *Geophysical Research Letters*, 27, 3865-3868, 10.1029/2000gl011385, 2000.
- 500 Sang, X. F., Gensch, I., Laumer, W., Kammer, B., Chan, C. Y., Engling, G., Wahner, A., Wissel, H., and Kiendler-Scharr, A.: Stable Carbon Isotope Ratio Analysis of Anhydrosugars in Biomass Burning Aerosol Particles from Source Samples, *Environmental Science & Technology*, 46, 3312-3318, 10.1021/es204094v, 2012.
- Sang, X. F., Gensch, I., Kammer, B., Khan, A., Kleist, E., Laumer, W., Schlag, P., Schmitt, S. H., Wildt, J., Zhao, R., 505 Mungall, E. L., Abbatt, J. P. D., and Kiendler-Scharr, A.: Chemical stability of levoglucosan: An isotopic perspective, *Geophys. Res. Lett.*, 43, 5419-5424, 10.1002/2016gl069179, 2016.
- Schauer, J. J., Kleeman, M. J., Cass, G. R., and Simoneit, B. R. T.: Measurement of Emissions from Air Pollution Sources. 3. C₁–C₂₉ Organic Compounds from Fireplace Combustion of Wood, *Environmental Science & Technology*, 35, 1716-1728, 10.1021/es001331e, 2001.
- 510 Seibert, P., and Frank, A.: Source-receptor matrix calculation with a Lagrangian particle dispersion model in backward mode, *Atmospheric Chemistry and Physics*, 4, 51-63, 10.5194/acp-4-51-2004, 2004.
- Stein, O., and Rudolph, J.: Modeling and interpretation of stable carbon isotope ratios of ethane in global chemical transport models, *Journal of Geophysical Research*, 112, 10.1029/2006JD008062, 2007.
- Stohl, A., Sodemann, H., Eckhardt, S., Frank, A., Seibert, P., and Wotawa, G.: The Lagrangian particle dispersion model FLEXPART version 8.2, 33, 2010.
- 515 Zhang, B., Owen, R. C., Perlinger, J. A., Kumar, A., Wu, S., Val Martin, M., Kramer, L., Helmig, D., and Honrath, R. E.: A semi-Lagrangian view of ozone production tendency in North American outflow in the summers of 2009 and 2010, *Atmospheric Chemistry and Physics*, 14, 2267-2287, 10.5194/acp-14-2267-2014, 2014.
- Zheng, M., Cass, G. R., Schauer, J. J., and Edgerton, E. S.: Source Apportionment of PM_{2.5} in the Southeastern United States Using Solvent-Extractable Organic Compounds as Tracers, *Environmental Science & Technology*, 36, 2361-2371, 520 10.1021/es011275x, 2002.