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To: Copernicus Gesellschaft mbH Bahnhofsallee 1e 37081 Göttingen Germany **Prof. Dr. Hartmut Herrmann** Head of TROPOS Atmospheric

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Manuscript for Atmospheric Chemistry and Physics Discussions (MS No.: acp-2019-982) "Multiphase MCM/CAPRAM modeling of formation and processing of secondary aerosol constituents observed at the Mt. Tai summer campaign 2014" by Zhu et al.

Dear Prof. Dr. Holger Tost,

please find attached our answers to the reviewer comments for the manuscript mentioned above together with its revised version. The authors would like to thank both reviewers for their constructive and good suggestions to improve our manuscript. We have carefully considered all the review comments and revised the manuscript. Below, we provide responses to the comments in blue, with changes made in the manuscript highlighted in red.

Yours sincerely,

Prof. Dr. Hartmut Herrmann Head of Atmospheric Chemistry Department at Leibniz Institute for Tropospheric Research Leipzig, Germany

Prof. Dr. Likun Xue Environment Research Institute Shandong University, Ji'nan, Shandong, China



Responses to the reviewer comments on

"Multiphase MCM/CAPRAM modeling of formation and processing of secondary aerosol constituents observed at the Mt. Tai summer campaign 2014" by Zhu et al.

The authors would like to thank both reviewers for their constructive and good suggestions to improve our manuscript. We have carefully considered all the review comments and revised the manuscript. Below, we provide responses to the comments in blue, with changes made in the manuscript highlighted in red.

Response to Reviewer 1:

The authors report a detailed multiphase chemical modeling study of the formation and processing of secondary aerosol compositions during transport to the Mt. Tai in summer 2014. The model performance of MCM/CAPRAM was evaluated against the field observations, and the day vs. night and with cloud vs. non-cloud processes were examined. The major formation pathways and key precursors of sulfate, nitrate, ammonium, and DCRCs were identified with the model. The impacts of emissions and glyoxal partitioning constants on the modeling results were also estimated by sensitivity studies. Despite an increasing number of field observational studies of secondary aerosols in recent years in China, such kind of detailed multiphase modeling study is still lacking. This study is helpful for better understanding the regional formation and processing of secondary inorganic and organic aerosols in the North China Plain. Therefore I would recommend that this manuscript can be considered for publication at ACP after the following specific comments being addressed.

Response: We thank reviewer#1 for the helpful comments. Below, we address the comments and have revised the manuscript accordingly. For clarity, the reviewer's comments are listed below in *black italics*, whereas our responses and changes in manuscript are shown in blue and red, respectively. Revised Tables and Figures are put in the end.

1. P2 L58 "However, formation pathways based on measured data are still limited": rephrase this sentence.

Response: We have rephrased this sentence as follows:

However, modeling studies that focus on understanding of DCRCs formation pathways based on field measurements are limited.

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(Page 2, Line 58-59)
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2. P3 L67-68: this sentence is not clear. Do the Yangtze River Delta and Bohai Rim have a total of 410 million populations? Additionally, the commonly used three largest economic zones in China don't include the Bohai Rim.

Response: Yes, the intended meaning was to say that both together have a population of around 410 million. In addition, we checked the China Statistical Yearbook in 2019, and found that together, the Yangtze River Delta and Bohai Rim regions had a population of more than 450 million in 2018. Therefore, we changed the sentence in the revised manuscript as follows:

Mt. Tai is located in Shandong province in the NCP, and between the Bohai Rim (BHR) and the Yangtze River Delta (YRD) regions. Together, the BHR and YRD regions had a population of more than 450 million in 2018 (China Statistical Yearbook in 2019). (Page 3, Line 67-71)

3. P3 L83-84: provide the standard deviations for the average temperature and RH values.

Response: We have added standard deviations for the average temperature and RH values as follows:

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17 ± 6.2 °C
(Page 3, Line 87)
87 ± 13 %
(Page 3, Line 88)
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4. P4 L98: predicted and observed concentrations.....

Response: We have changed "concentration" to "concentrations" in the revised manuscript as follows:

These limitations have to be kept in mind when studying deliquesced particles and comparing predicted and observed concentrations at Mt. Tai.

(Page 4, Line 113-114)

5. P4 Section 2.2: I suggest the authors to provide the air mass cluster figures in the SI so that the readers can easily access the plot.

Response: We have added air mass cluster Figure in the supplement as follows:

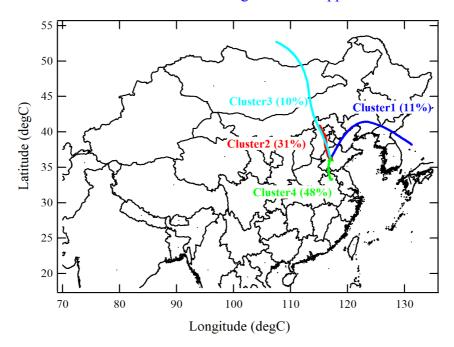


Figure S1. Three-day back-trajectories for Mt. Tai during the sampling period (green triangle: Mt. Tai).

6. P4 L111-112: was only an important source.....

Response: We have added "an" in the revised manuscript as follows:

Additionally, Zhu et al. (2018) have clearly shown that biomass burning was only an important source during the first half of the sampling period (June 4 - 19).

(Page 4, Line 128, Page 5, Line 129)

7. P5 L150-156: this paragraph is a little bit redundant with the last paragraph of the Introduction (P3 L73-79). I suggest the authors may remove this paragraph.

Response: According to the reviewer suggestion, we have removed this paragraph. (Page 6, Line 179-185)

8. P6 L158: replace "oxidations" by "oxidant"

Response: We have replaced "oxidations" by "oxidant" in the revised manuscript as follows:

Due to the key role of radical and non-radical oxidant in the formation processes of secondary aerosol constituents, their concentration variations and corresponding reasons are investigated.

(Page 6, Line 187-188)

9. P6 L 168: non-radical oxidant concentrations...

Response: We have added "oxidant" in the revised manuscript as follows:

To our knowledge, this is the first detailed multiphase chemical modeling study examining radical and non-radical oxidant concentrations along the trajectory to the Mt. Tai under day vs. night and cloud vs. non-cloud cases.

(Page 7, Line 197-198)

10. P6 L 175: I would suggest the authors to delete the citation here as it is only modeling results from this study.

Response: The citations have been deleted.

(Page 7, Line 204-206)

Section 3.2: I suggest to provide the sub-titles for "Sulfate" (e.g., 3.2.1), "Nitrate" and "Ammonium".

Response: We have added sub-titles for "Sulfate", "Nitrate" and "Ammonium" subsections as follows:

3.2.1 Sulfate

(Page 9, Line 282)

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3.2.2 Nitrate
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(Page 10, Line 305)
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3.2.3 Ammonium

(Page 11, Line 331)

11. P9 L253: replace "dominated" by "dominant"

Response: We have replaced "dominated" by "dominant" in the revised manuscript as follows:

In the nighttime cloud, aqueous-phase reaction of HSO₃⁻ with H₂O₂ (42 %), and aqueous reaction of bisulfite with O₃ (28 %) are dominant pathways for sulfate formation.

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(Page 10, Line 295-297)
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12. P10, L286-287 "Potential reasons are discussed below": it is not clear where the potential reasons are discussed. Please clarify.

Response: We have changed the sentence in the revised manuscript as follows:

Potential reasons are discussed in Sect. 3.3.2.

(Page 11, Line 329-330)

13. P11 L 321: replace "shows" by "show"

Response: We have replaced "shows" by "show" in the revised manuscript as follows: In the C2w case, Gly and MGly concentration patterns show a substantial uptake into cloud droplets.

(Page 12, Line 366)

14. P11 L348-349: I suggest to move this sentence to the beginning of this section, i.e., L319.

Response: According to the reviewer suggestion, we have moved the sentence.

(Page 12, Line 362-364)

15. P12 L363: under-estimation

Response: We have changed "underestimation" to "under-estimation" in the revised manuscript as follows:

The over- and under-estimation of the measured concentrations of inorganic and organic aerosol constituents could have the following reasons.

(Page 13, Line 414-415)

16. P12 L365-366: I presume the emission data were obtained from the emission inventory, rather than model calculations.

Response: We have changed "model calculations" to "emission inventory" in the revised manuscript as follows:

The emission data are obtained through a new anthropogenic emission inventory in Asia, which provides monthly emissions in 2010 by sector at $0.25^{\circ} \times 0.25^{\circ}$ resolution. (Page 13, Line 416-418)

17. P13 L401 "reported in in above references": delete one "in".

Response: deleted.

(Page 15, Line 456)

Responses to the reviewer comments on

"Multiphase MCM/CAPRAM modeling of formation and processing of secondary aerosol constituents observed at the Mt. Tai summer campaign 2014" by Zhu et al.

The authors would like to thank both reviewers for their constructive and good suggestions to improve our manuscript. We have carefully considered all the review comments and revised the manuscript. Below, we provide responses to the comments in blue, with changes made in the manuscript highlighted in red.

Response to Reviewer 2:

This paper presents a modeling study with a 0D-sophisticated model of the formation and processing of dicarboxylic acids and related compounds (oxo-carboxylic acids and \alpha-dicarbonyls) (DCRCs) in clouds and deliquesced aerosols at the Mt. Tai. Results are compared to observations of aerosols composition analyzed from night and day filters. This study is very interesting and valuable in doing the effort to compare simulated results to observations. However, I have some main concerns detailed below on the definition of the scenario and on the absence of information on the microphysical structure of simulated cloud. I recommend that the paper should be accepted for publication in Atmospheric Chemistry and Physics after major and specific revisions listed below.

Response: We appreciate the reviewer for the helpful comments and suggestions. Below we address the comments and have revised the manuscript accordingly. For clarity, the reviewer's comments are listed below in *black italics*, while our responses and changes in manuscript are shown in blue and red, respectively. Revised Tables and Figures are in the end.

 $1.\ I-About$ the simulations set-up: I found that the conditions used in simulations (meteorological conditions, emissions, initial chemical concentrations, aerosol parameters) are very confused. I list below some of my main concerns: No value is

given for meteorological conditions contrary to that it states page 5, line 148. I did not understand why the meteorological values along the trajectories are not directly used to drive the air parcel. For instance, in Zhu et al. (2018), WRF simulations and HYSPLIT back-trajectories are mentioned: I guess that typical trajectories to Mt. Tai could be extracted from these runs.

Response: We have used the HYSPLIT back-trajectories results (see Fig. S1 in the supplement) given in Zhu et al. (2018). According to HYSPLIT back-trajectories results in Zhu et al. (2018), we selected clusters 2 and 4 to simulate and investigate the formation processes and the fate of DCRCs. Because the two clusters accounted for 79 % of the total trajectories. Moreover, the sum of DCRC concentrations in clusters 2 and 4 amounted to 73 % of total DCRC concentration during the sampling period (Zhu et al., 2018).

The meteorological data along the cluster trajectories derived by HYSPLIT are mean values of many trajectories. The resolution of HYSPLIT is 1°x1° and the DCRC measurements at Mt. Tai have a 12 h resolution only. Due to the coarse spatial- and time-resolution of both data, we decided to use a representative trajectory for each cluster instead of single trajectories.

Cloud conditions are assumed from satellite pictures, because HYSPLIT data have a very coarse spatial resolution which does not allow the identification of cloud conditions along the trajectory. The relative humidity in the 1°x1° grid cells represents a mean value. Therefore, a grid cell with sub-grid clouds would finally also be characterized by RH values below 100% as only a certain fraction is filled with clouds. In order to illustrate the applied meteorological data in each scenario, we have added the following Figure S2 in the supplement.

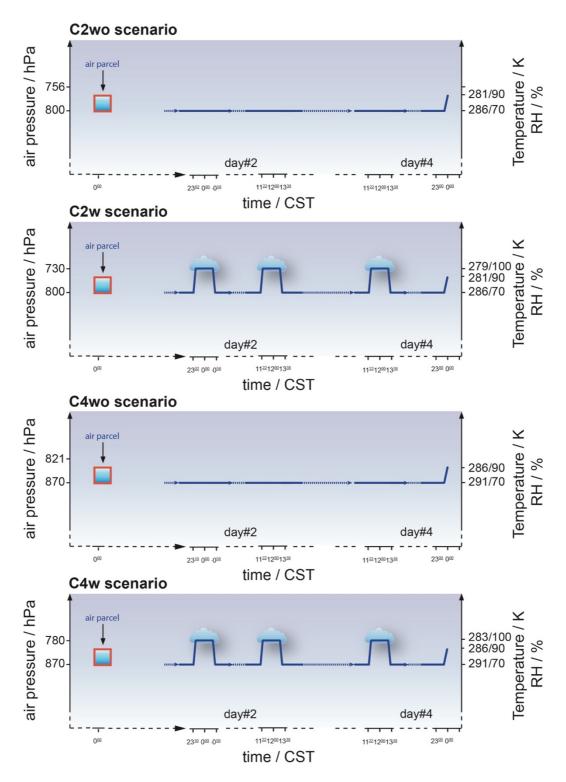


Figure S2. Meteorological data in different scenarios.

We also changed corresponding sentence as follows:

These initial model data, and also aerosol parameters are given in Table S3 and Table S4. The meteorological scenarios are illustrated in Figure S2.

(Page 6, Line 176-177)

Reference:

Zhu, Y., Yang, L., Chen, J., Kawamura, K., Sato, M., Tilgner, A., van Pinxteren, D., Chen, Y., Xue, L., Wang, X., Simpson I. J., Herrmann, H., Blake D. R., and Wang, W. X.: Molecular distributions of dicarboxylic acids, oxocarboxylic acids and α-dicarbonyls in PM2.5 collected at the top of Mt. Tai, North China, during the wheat burning season of 2014, Atmos. Chem. Phys., 18, 10741-10758, 2018.

2. 1 – About the simulations set-up: With regard to scenarios (cloud or cloud-free), in my opinion, a third series of simulation without aqueous chemistry at all should be of interest in assessing the contribution of aqueous pathways to the formation of secondary aerosol constituents. Especially since the ideal solution hypothesis biases the results for aqueous chemistry inside deliquesced aerosols.

Response: The applied gas-phase mechanism MCM is a mechanism that is not able to produce dicarboxylic acids, because of the exclusive treatment of organic acid oxidation via H-abstraction from the OH group destroying the carboxylic acid functionality. As a consequence, only monocarboxylic acids can be formed by gas-phase chemistry within the applied multiphase chemistry system. Therefore, it is evident that most of the compounds in this study are mainly or, in case of the DCRCs, exclusively formed in the aqueous-phase. Thus, in author's opinions, a simulation without aqueous chemistry is not useful. Thus, a comparison with aqueous-phase simulations might result into misleading interpretations.

3. 1 – About the simulations set-up: For biogenic emissions, you should precise, which inventory you used to extract biogenic emissions: is-it MEGAN-MACC, CAMS-GLOB-BIO? Also, why did you consider biogenic emissions only for isoprene and pinenes whereas a lot of other species are available from MEGAN-MACC or CAMS-GLOB-BIO? I suggest that the biogenic and anthropogenic emission values be separated in Table S1.

Response: (1) Biogenic emission data are obtained from MEGAN-MACC, we have

changed the corresponding sentence in the revised manuscript as follows:

Biogenic emission data (isoprene, α - and β -pinenes) are obtained from Emissions of atmospheric Compounds and Compilation of Ancillary Data (ECCAD), MEGAN-MACC dataset (https://eccad.aeris-data.fr/).

(Page 5, Line 158-160)

(2) MEGAN-MACC includes 21 biogenic species, such as ethane, propane, propene, toluene. However, most of the treated compounds are mainly from anthropogenic sources, especially in China. For example, propane is the main components of liquefied petroleum gas/natural gas (McCarthy et al., 2013), ethane, propene and toluene originate from automobile exhaust (Chang et al., 2009), ketones and aldehydes are produced from anthropogenic secondary oxidation (Liu et al., 2015). Therefore, emission data of these species are obtained from a new anthropogenic emission inventory in Asia (Li et al., 2017), not from MEGAN-MACC.

Isoprene and pinenes are mainly from biogenic sources (Brown et al., 2007; Zhang et al., 2013). The new anthropogenic emission inventory in Asia focuses on anthropogenic sources, which inputs to be indefinite in biogenic emission (Li et al., 2017). Therefore, emission of isoprene and pinenes are obtained from MEGAN-MACC.

References:

Brown, S. G., Frankel, A., Hafner, H. R., 2007. Source apportionment of VOCs in the Los Angeles area using positive matrix factorization. Atmospheric Environment 41, 227-237, 2007.

Chang, C. C., Wang, J. L., Lung, S. C. C., Liu, S. C., Shiu, C. J.: Source characterization of ozone precursors by complementary approaches of vehicular indicator and principal component analysis. Atmospheric Environment 43 (10), 1771-1778, 2009.

Li, M., Zhang, Q., Kurokawa, J., Woo, J., He, K. B., Lu, Z. F., et al.: MIX: a mosaic Asian anthropogenic emission inventory under the international collaboration framework of the MICS-Asia and HTAP. Atmospheric Chemistry and Physics, 17, 935–963, 2017.

Liu, Y., Yuan, B., Li, X., Shao, M., Lu, S., Li, Y., et al.: Impact of pollution controls in

Beijing on atmospheric oxygenated volatile organic compounds (OVOCs) during the 2008 Olympic Games: observation and modeling implications. Atmospheric Chemistry and Physics, 15, 3045–3062, 2015.

McCarthy, M. C., Aklilu, Y. A., Brown, S. G., Lyder, D. A.: Source apportionment of volatile organic compounds measured in Edmonton, Alberta. Atmospheric Environment 81 (2), 504-516, 2013.

Zhang, J. K., Sun, Y., Wu, F. K., Sun, J., Wang, Y. S.: The characteristics, seasonal variation and source apportionment of VOCs at Gongga Mountain, China. Atmospheric Environment, 1-9, 2013.

(3) We have separated biogenic and anthropogenic emission values in revised Table S1 as follows:

Table S1. Emission data applied in the SPACCIM (#: anthropogenic emission values, *: biogenic emission values).

Compound	Emission /	Compound	Emission / molec cm ⁻³ s ⁻¹	
	molec cm ⁻³ s ⁻¹			
Acetone#	3.51E+04	Acetaldehyde#	1.44E+04	
Ethane [#]	1.30E+05	Ethylene#	1.76E+05	
Propane [#]	1.87E+05	Glyoxal [#]	1.04E+04	
n-Butane#	6.79E+04	Formaldehyde#	3.15E+04	
Isobutane [#]	2.99E+04	Biacetyl#	1.49E+03	
2,2-Dimethyl Butane#	2.00E+03	Benzaldehyde#	3.37E+02	
Isopentane#	6.75E+04	Methacrolein#	2.08E+03	
n-Pentane#	2.67E+04	Methyl ethyl ketone#	8.68E+03	
2-Methyl Pentane#	1.57E+04	Methanol [#]	2.28E+04	
3-Methylpentane [#]	1.10E+04	Methylglyoxal#	3.93E+03	
n-Hexane#	6.28E+03	Methyl Vinyl Ketone#	2.62E+02	
n-Heptane#	4.71E+03	Propene#	2.96E+04	
2,3-Dimethyl Butane#	4.71E+03	1-Hexene [#]	2.45E+04	
n-Decane#	1.77E+04	1-Butene#	1.22E+04	
3-Methyl Hexane [#]	1.77E+04	1-Pentene [#]	1.12E+04	
n-Nonane#	6.45E+03	3-Methyl-1-Butene#	3.06E+03	
n-Octane#	6.45E+03	cis-2-Pentene#	2.25E+04	
2-Methyl Hexane [#]	4.84E+03	trans-2-Pentene#	2.25E+04	
n-Dodecane#	3.22E+03	1,3-Butadiene#	9.64E+03	
n-Undecane#	1.61E+03	2-Methyl-2-Butene#	8.03E+03	
Toluene#	1.39E+05	Cis-2-Hexene#	8.03E+03	
Ethyl Benzene#	1.86E+04	Trans-2-Hexene#	8.03E+03	

n-Propyl Benzene#	7.43E+03	Propionaldehyde#	6.40E+03
Isopropyl Benzene#	3.72E+03	Limonene#	3.28E+02
m-Xylene#	1.46E+04	Carbon monoxide#	3.04E+07
p-Xylene#	1.46E+04	Carbon dioxide#	1.15E+09
o-Xylene [#]	1.23E+04	Ammonia [#]	3.81E+06
1,2,3-Trimethyl Benzene#	1.01E+04	Nitric Oxide#	2.51E+05
1,3,5-Trimethyl Benzene#	1.01E+04	Nitrogen dioxide#	1.42E+06
m-Ethyl Toluene#	5.61E+03	Sulfur dioxide#	1.91E+06
o-Ethyl Toluene#	5.61E+03	Isoprene*	4.05E+05
p-Ethyl Toluene#	5.61E+03	a-pinene*	2.99E+04
1,2,4-Trimethyl Benzene#	5.61E+03	β-pinene*	1.28E+04

4. I – About the simulations set-up: Where values for deposition velocities come from? Response: Dry deposition velocities were taken from Ganzeveld et al. (1998). We have changed the corresponding sentence as follows:

The deposition velocities used in SPACCIM were taken from Ganzeveld et al. (1998), and presented in Table S2.

(Page 6, Line 164)

Reference:

Ganzeveld, L. N., Roelofs, G. J., Lelieveld, J.: A dry deposition parameterization for sulfur oxides in a chemistry and general circulation model. Journal of Geophysical Research, 103, 5679–5694, 1998.

5. 1 – About the simulations set-up: In Table S3 and S4, the source of each value (literature or urban CAPRAM scenario) should be indicated. Why measured observations of VOC analysed from stainless steel canisters (Zhu et al., 2018) are not used to initialize the model?

Response: We have added references for values in Table S3 and S4 as follows:

The initial chemical data include gas-phase concentrations of inorganic gases (NO, NO₂, O₃, SO₂, HNO₃, NH₃, H₂O₂), VOCs (including alkanes, alkenes, aromatics, aldehydes, alcohols and ketones) (Barletta et al., 2005; Duan et al., 2008; An et al., 2009; Liu et al., 2009; He et al., 2010; Ianniello et al., 2011; Meng et al., 2011; Li et al., 2011; Liu

et al., 2012b; Zhao et al., 2013b; Wang et al., 2014b; Li et al., 2015a; Li et al., 2015b; Rao, et al., 2016; Wang et al., 2016b) and particle phase data (Hu et al., 2015; Wang et al., 2015; Sun et al., 2015; Liu et al., 2014; Sun et al., 2013). In case of missing values, values are taken from the CAPRAM urban scenario (http://projects.tropos.de/capram, Herrmann et al., 2005).

(Page 6, Line 170-176)

Mt. Tai (1534 m a.s.l.), the highest mountain in the North China Plain (NCP), is influenced by air pollutants transported from the surrounding or formed along the trajectories. The measured data at Mt. Tai represents the results of transport and prior chemical processing. The aim of the present study was to investigate the processing along trajectories approaching Mt. Tai. Therefore, we have not used the measured data as input, but used VOCs data from the trajectory origin areas instead.

6. 1 – About the simulations set-up: I do not understand aerosol parameters in Table S4. Although, the unit of ng/m^3 or μ m/m^3 is used in the article, aerosol species composition is given in mixing ratio (g/g) in the Table. No initial value is given for aerosol number concentrations.

Response: We have added details about the aerosol number concentration in Table S4 (see below).

Now the Table includes the initial mass fractions of the different aerosol components as well as the initial monodisperse physical aerosol data. Both aerosol mass and single mass fractions are used in the model to calculate the initial concentrations of the aerosol composition. The ratio g/g describes the ratio of component to the overall aerosol weight and is the standard SPACCIM input.

Table S4. Aerosol compositions and parameters applied in the SPACCIM.

Compound		Data / g _{compound} g _{aerosol} ⁻¹	Parameter	Data
Sulfate		0.25	Aerosol radius	2.0E-07 m
Nitrate		0.21	Aerosol Density	1770 kg m ⁻³
Ammonium		0.16	Aerosol number concentration	5.1E+08 m ⁻³
Water-soluble	organic	0.07		

carbon	
HULIS	0.07
Water-insoluble organic	0.05
carbon	0.03
Positive monovalent ions	0.03
Positive divalent ions	0.01
Metals	0.03
Elemental carbon	0.03

7. 2 – About the results: In my opinion, some important information is missing. No results are reported for the microphysical structure of simulated clouds and how they compare to observations.

Response: The microphysical conditions of the clouds are calculated by a microphysical model using temperature, supersaturation and aerosol distribution. Alongside, we have no direct measurements of aerosol/cloud size distribution and liquid water content. Furthermore, information of the cloud measurements at Mt. Tai is not characteristic for the clouds along the trajectory. Hence, a comparison is not possible. As our study focuses on the formation of DCRCs, these are not mandatory and thus, not discussed because it is beyond the scope of the current study.

8. Specific Revisions: Introduction: The recent following review paper should be cited: Shrivastava, M., Cappa, C. D., Fan, J., Goldstein, A. H., Guenther, A. B., Jimenez, J. L., Kuang, C., Laskin, A., Martin, S. T., Ng, N. L., Petaja, T., Pierce, J. R., Rasch, P. J., Roldin, P., Seinfeld, J. H., Shilling, J., Smith, J. N., Thornton, J. A., Volkamer, R., Wang, J., Worsnop, D. R., Zaveri, R. A., Zelenyuk, A. and Zhang, Q.: Recent advances in understanding secondary organic aerosol: Implications for global climate forcing, Rev. Geophys., 2016RG000540, doi:10.1002/2016RG000540, 2017.

Response: We have added the paper in the introduction as follows:

SOA is also a key component of PM_{2.5} and linked to adverse health effects, visibility reduction and climate change (Tabazadeh, 2005; Seagrave et al., 2006; De Gouw and Jimenez, 2009; Shrivastava et al., 2017).

(Page 2, Line 49-50)

9. Specific Revisions: Introduction: The citation of following paper should be added in the paragraph lines 52 to 64: Mouchel-Vallon, C., Deguillaume, L., Monod, A., Perroux, H., Rose, C., Ghigo, G., Long, Y., Leriche, M., Aumont, B., Patryl, L., Armand, P. and Chaumerliac, N.: CLEPS 1.0: A new protocol for cloud aqueous phase oxidation of VOC mechanisms, Geosci. Model Dev., 10(3), 1339–1362, doi:10.5194/gmd-10-1339-2017, 2017.

Response: We have added the paper in the introduction as follows:

Additionally, model studies show growing evidence that substantial amounts of DCRCs are formed by aqueous-phase reactions within aerosol particles, clouds and fog droplets (Sorooshian et al., 2006; Carlton et al., 2007, 2009; Ervens et al., 2008, 2011; Ervens, 2015; Tilgner and Herrmann, 2010; Tilgner et al., 2013; Mouchel-Vallon et al., 2017). (Page 2, Line 60-62)

10. Part 2: The sampling period of observations should be given here. A discussion about the known sources of DCRCs deduced from Zhu et al. (2018) should be done.

Response: We have added sampling period in the revised manuscript as follows:

The sampling period was from June 4 to July 4 2014.

(Page 3, Line 86)

We have added a brief discussion about DCRC sources in the revised manuscript as follows:

Source identification indicated that DCRCs were mainly derived from anthropogenic activities followed by photochemical aging. Secondary sources, fuel combustion, photooxidation of unsaturated fatty acids and waste burning were also significant sources.

(Page 3, Line 89-91)

11. Part 2, 2.1: The limitation of SPACCIM due to the assumptions of ideal solutions concerns deliquesced particles. This should be specified.

Response: We have pointed out when studying deliquesced particles, the limitation of SPACCIM should be kept in mind as follows:

These limitations have to be kept in mind when studying deliquesced particles and comparing predicted and observed concentrations at Mt. Tai. The potential limitations of an ideal solution assumption compared to a non-ideal treatment are discussed in a recent paper by Rusumdar et al. (2020).

(Page 4, Line 113-115)

12. Part 2, 2.1: Microphysical processes include in SPACCIM should be listed. Also it should be mentioned that cloud particles size distribution is spectral.

Response: We have added a decription about the microphysical processes and cloud particles size distribution considered in the present study in Sect. 2.1 as follows:

The microphysical model applied in SPACCIM is based on the work of Simmel and Wurzler, (2006) and Simmel et al. (2005). Droplet formation, evolution and evaporation are realized by a one-dimensional sectional microphysics considering deliquesced particles and cloud droplets. In the present study, the moving bin version of SPACCIM has been applied. In the model, the growth and shrinking of aerosol particles by water vapor diffusion as well as nucleation and growth/evaporation of cloud droplets is considered. The dynamic growth rate in the condensation/evaporation process and the droplet activation is based on the Köhler theory. Due to the emphasis on complex multiphase chemistry, other microphysical processes such as impaction of aerosol particles and collision/coalescence of droplets and thus precipitation were not considered in the present study. Moreover, the air parcel model SPACCIM is not able to reflect the complexity of tropospheric mixing processes. Nevertheless, the complex model enables detailed investigations of the multiphase chemical processing of gases, deliquescent particles and cloud droplets.

(Page 4, Line 101-110)

13. Part 2, 2.1: Does SPACCIM include aerosols? If yes, it should be specified how: which processes are considered for aerosols: microphysics processes (nucleation, aggregation, sedimentation), chemical aging, nucleation and impaction scavenging by cloud particles? Also the method to represent their size distribution should be indicated.

Do you use thermodynamics equilibrium to partition inorganic and organic species between gas phase and particles?

Response: In the multiphase chemistry model, the phase transfer is treated for soluble compounds according to the resistance model of Schwartz (1986). No thermodynamics equilibrium approach is used to calculate the partitioning of soluble compounds. The model uses accommodation coefficients, gas-phase diffusion coefficients, and Henry's law constants of the phase transfer of soluble compounds (Wolke et al., 2005). The phase transfer is treated in the same manner for cloud droplets and aqueous particles following the assumption of deliquesced particles. A detailed description is already given in Wolke et al. (2005) and thus not implemented in our manuscript.

Reference:

Schwartz, S. E.: Mass-transport considerations pertinent to aqueous phase reactions of gases in liquid-water clouds, in: Chemistry of multiphase atmospheric systems. Springer, 415-471, 16, 1986.

Wolke, R., Sehili, A., Simmel, M., Knoth, O., Tilgner, A., Herrmann, H.: SPACCIM: A parcel model with detailed microphysics and complex multiphase chemistry. Atmos. Environ., 39, 4375-4388, 2005.

14. Part 2, 2.1: Could you please indicate if recent findings on isoprene and aromatics gas phase chemistry are included in MCM? As the reference papers for MCM are from 2003, it is probably not the case and could lead to other limitations in the results.

Response: MCMv3.3.1 systematically refined and updated the chemistry of isoprene degradation to reflect recent advances in understanding (Jenkin et al., 2015). However, in SPACCIM, MCMv3.2 was used, which made a general update of isoprene degradation chemistry, including integration of revised chemistry for isoprene-derived hydroperoxides and nitrates. Therefore, the newest findings about isoprene are not included in SPACCIM. However, regarding the review of Wennberg et al. (2018) and other new laboratory studies (e.g., Berndt et al., 2019), the isoprene oxidation scheme in MCMv3.3.1 is also not representative enough. Furthermore, the new scheme is

developed to treat more clean conditions as observed at Mt. Tai. Hence, the main oxidation product MACR and MVK should not be influenced too much by the MCMv3.2 oxidation scheme.

The degradation chemistry of aromatic VOC remains an area of particular uncertainty. Aromatic scheme in MCMv3.2 continue to use these in MCMv3.3.1. The updated aromatic VOC schemes within MCMv3.3.1 were mainly for benzene, toluene, p-xylene and 1, 3, 5-trimethylbenzene and focus on the kinetic rate constants. Therefore, the newest findings about aromatic VOC are also not included, but can be stated as minor bias for the model simulations.

The considered mechanism for isoprene and aromatics indeed impact the modeled results. In the future, we will update MCM mechanism in SPACCIM to better perform more advanced model simulation.

References:

Berndt, A. J., Hwang, J., Islam, M. D., Sihn, A., Urbas, A. M., Ku, Z., Lee, S. J., Czaplewski, D. A., Dong, M., Shao, Q., Wu, S., Guo, Z., Ryu, J. E.: Poly(sulfur-random-(1,3-diisopropenylbenzene)) based mid-wavelength infrared polarizer: Optical property experimental and theoretical analysis. Polymer, 176, 118-126, 2019.

Jenkin, M. E., Young, J. C., Rickard, A. R.: The MCM v3.3.1 degradation scheme for isoprene. Atmospheric Chemistry and Physics, 15, 11433–11459, 2015.

Wennberg, P. O., Bates, K. H., Crounse, J. D., Dodson, L. G., McVay, R. C., Mertens, L. A., Nguyen, T. B., Praske, E., Schwantes, R. H., Smarte, M. D., St Clair, J. M., Teng, A. P., Zhang, X., Seinfeld, J. H.: Gas-Phase Reactions of Isoprene and Its Major Oxidation Products. Chemical Reviews, 118, 3337-3390, 2018.

15. Part 2, 2.2: Could you please indicate the reason to exclude days of the campaign influenced by biomass burning?

Response: We have added reasons for excluding days of the campaign influenced by biomass burning as follows:

The aim of the study was to investigate the secondary formation of aerosol constituents

along the trajectories to Mt. Tai. However, biomass burning can be an important primary source of compounds that are often of secondary origin. Therefore, in this study, we focused on the period that was less impacted by biomass burning.

(Page 5, Line 129-133)

16. Part 2, 2.3: You should use the proper ECCAD address; https://eccad.aeris-data.fr/. Note that you have to acknowledge ECCAD and to reference the dataset citation (see metadata page of the inventory that you used). Why did you use urban CAPRAM scenario for missing values? I suggest to indicate also initial gas-phase conditions in mixing ratio (ppbv or pptv), which is the usual quantity used for representing observations of trace gases.

Response: (1) We have changed the ECCAD address in the revised manuscript as follows:

Biogenic emission data (isoprene, α - and β -pinenes) are obtained from Emissions of atmospheric Compounds and Compilation of Ancillary Data (ECCAD), MEGAN-MACC dataset (https://eccad.aeris-data.fr/).

(Page 5, Line 158-160)

(2) We have acknowledged ECCAD and corresponding dataset in the Acknowledgements as follows:

The authors acknowledge the Emissions of atmospheric Compounds and Compilation of Ancillary Data (ECCAD), MEGAN-MACC dataset.

(Page 21, Line 642-643)

(3) During the sampling period, $PM_{2.5}$ concentration was $98.2 \pm 29.2 \,\mu g \, m^{-3}$ (range from 37.0 to $193 \,\mu g \, m^{-3}$) in daytime, and $98.6 \pm 25.3 \,\mu g \, m^{-3}$ (range from 55.7 to $143 \,\mu g \, m^{-3}$) in nighttime. The $PM_{2.5}$ concentration is equal or higher than that in urban sites in European (Dimitriou and Kassomenos, 2014; Eeftens et al., 2012). Therefore, we choose urban CAPRAM scenario for missing values.

References:

Dimitriou, K., and Kassomenos, P.: A study on the reconstitution of daily PM10 and

PM2.5 levels in Paris with a multivariate linear regression model. Atmospheric Environment, 98, 648-654, 2014.

Eeftens, M., Tsai, M. Y., Ampe, C., Anwander, B., Beelen, R., Bellander, T., et al.: Spatial variation of PM2.5, PM10, PM2.5 absorbance and PM coarse concentrations between and within 20 European study areas and the relationship with NO2 - Results of the ESCAPE project. Atmospheric Environment, 62, 303-317, 2012.

(4) We have changed the unit of initial gas-phase concentrations to mixing ratio as follows:

Table S3. Initial gas-phase concentrations applied in the SPACCIM.

Compound	Concentration	Compound	Concentration	
Nitric oxide	0.32 ppbv	p-Xylene	94.53 pptv	
Nitrogen dioxide	1.72 ppbv	m-Xylene	94.53 pptv	
Ozone	100.33 ppbv	Acetaldehyde	1.00 ppbv	
Nitric acid	0.67 ppbv	Propionaldehyde	70.48 pptv	
Hydrogen peroxide	0.31 ppbv	Butyraldehyde	35.32 pptv	
Formaldehyde	0.70 ppbv	Acetone	1.07 ppbv	
Hydrogen	0.46 ppmv	Methyl ethyl ketone	29.44 pptv	
Carbon monoxide	1.18 ppmv	Methyl isobutyl ketone	13.02 pptv	
Methane	2.06 ppmv	Glyoxal	0.21 ppbv	
Carbon dioxide	332.10 ppmv	Glycolaldehyde	0.21 ppbv	
Sulfur dioxide	2.14 ppbv	Methylglyoxal	18.57 pptv	
Ethane	0.43 ppbv	Peroxyacetyl nitrate	92.87 pptv	
Propane	80.43 pptv	Methyl hydrogen peroxide	0.19 ppbv	
Isoprene	96.19 pptv	Ethyl hydrogen peroxide	18.57 pptv	
n-propanol	1.30 pptv	Peroxyacetic acid	0.19 pptv	
Isopropanol	51.00 pptv	Ammonia	4.39 ppbv	
Butanol	0.75 pptv	Methanol	0.42 ppbv	
Isobutanol	0.56 pptv	Ethanol	0.40 ppbv	
Ethylene glycol	1.17 pptv	Glyoxylic acid	0.11 ppbv	
Ethylene	0.96 ppbv	Glycolic acid	0.11 ppbv	
Toluene	0.31 ppbv			
Cresol	0.19 pptv			
o-Xylene	62.61 pptv			

17. Part 3: 3.1.1: Page 6, lines 178-179, it is stated: "The reduction of OH radical is mainly caused by the reduction of the gas-phase formation pathway of the HO₂+ NO reaction" and page 7, lines 198-200: "Under daytime cloud droplet conditions, OH

aqueous-phase concentrations are increased by a factor of 3, mainly due to the increased direct transfer of OH from the gas phase.". I found these statements contradictory. I guess that a plot showing the total OH concentrations (gaseous + aqueous) would clarify this point.

Response: From the below Figure R1, we can see "The reduction of OH radical in day cloud is mainly caused by the reduction of the gas-phase formation pathway of the HO₂ + NO reaction". However, we forgot to mention that this sentence is related to gas-phase OH concentrations. It is changed now in the revised manuscript as follows:

The reduction of gas-phase OH radical concentrations in daytime cloud is mainly caused by the reduction of the gas-phase formation pathway of the HO₂+ NO reaction. (Page 7, Line 207-208)

Figure R1 also shows that "under daytime cloud droplet conditions, OH aqueousphase concentrations are increased, mainly due to the increased direct transfer of OH from the gas phase."

Figure R2 shows time series of total OH concentrations (gaseous + aqueous phase) in the C2w case, which also suggests the two sentences are not contradicting in their revised form, now.

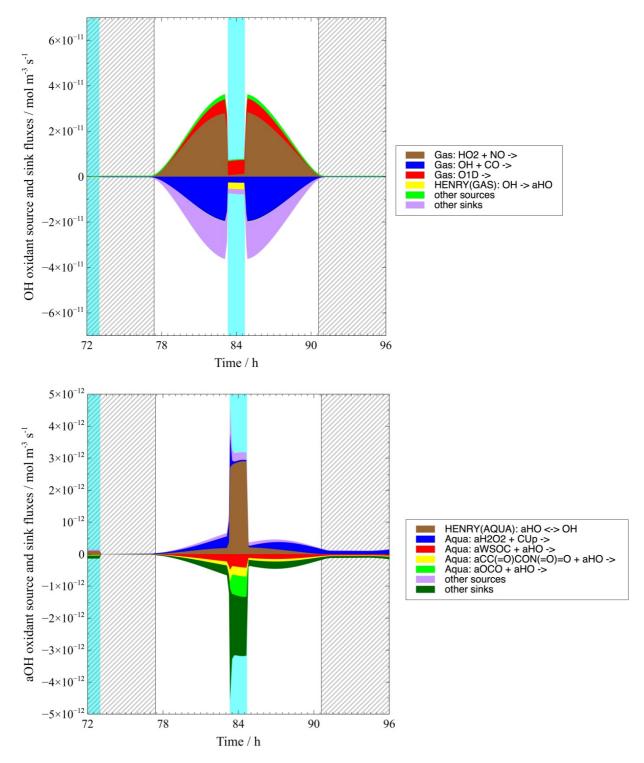


Figure R1. Modeled multiphase source and sink fluxes of gas-phase OH (above) and aqueous-phase aOH (below) oxidant (light blue column: cloud; shadow: night).

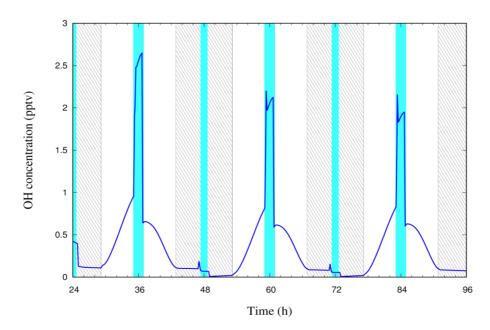


Figure R2. Time series of the modeled gaseous + aqueous phase OH oxidant concentrations in the C2w case (light blue column: cloud; shadow: night).

18. Part 3: 3.1.1: Whereas, the text page 6 line 187 indicates that obtained OH and HO₂ gaseous concentrations are compared to available measurements, it is mainly modelling studies that are used for this comparison. Moreover, no details are given on these studies: which model, which conditions (period of simulation, chemical mechanism used for instance). The only observations cited show discrepancies with results, in particular for OH.

Response: We have added period of simulation and chemical mechanism for available measurements as follows:

However, the simulated maxima of the gas-phase concentrations of OH (C2w: 3.2×10^6 molecules cm⁻³, C2wo: 3.5×10^6 molecules cm⁻³) and HO₂ (C2w: 2.9×10^8 molecules cm⁻³, C2wo: 3.8×10^8 molecules cm⁻³) for Mt. Tai in this study are comparable to the available measurements listed below. Compared with the modeled maximum OH (6.0×10^6 molecules cm⁻³) and HO₂ (7.0×10^8 molecules cm⁻³) concentrations at Mt. Tai in June 2006 using a photochemical box model- that was based on the Regional Atmospheric Chemistry Mechanism (RACM) (Kanaya et al., 2009), the OH and HO₂ concentrations reported here are lower. Moreover, the modeled OH and HO₂ concentrations in this study are lower than those of simulated results over

the Chinese megacity Beijing in August 2007 using a 1-D photochemical model (Regional chEmical and trAnsport Model, REAM-1D), whose chemistry was driven by the standard GEOS-Chem gas-phase chemistry mechanism (OH: 9×10^6 molecules cm⁻³, HO₂: 6.8×10^8 molecules cm⁻³) (Liu et al., 2012) and much lower than the measured data by laser induced fluorescence (LIF) at a rural site downwind of the megacity Guangzhou, China in 3–30 July 2006 (OH: $15\text{-}26 \times 10^6$ molecules cm⁻³, HO₂: $3\text{-}25 \times 10^8$ molecules cm⁻³) (Lu et al., 2012). Additionally, the simulated NO₃ radical maxima (C2w: 1.0×10^8 molecules cm⁻³, C2wo: 1.5×10^8 molecules cm⁻³) are much lower than those observed at the urban site of Shanghai, China by Differential Optical Absorption Spectroscopy (DOAS) from August 15 to October 7, 2011 (2.5×10^9 molecules cm⁻³) (Wang et al., 2013a).

(Page 7, Line 214-Page 8, Line 228)

19. Part 3: 3.1.1: Whereas some estimations exist of OH concentration in cloud droplets (see Arakaki, T., Anastasio, C., Kuroki, Y., Nakajima, H., Okada, K., Kotani, Y., Handa, D., Azechi, S., Kimura, T., Tsuhako, A. and Miyagi, Y.: A General Scavenging Rate Constant for Reaction of Hydroxyl Radical with Organic Carbon in Atmospheric Waters, Environ. Sci. Technol., 47(15), 8196–8203, doi:10.1021/es401927b, 2013), no comparison is discussed. It seems to me that the simulated OH concentrations in droplets are high in comparison to available estimations.

Response: We have added comparison about OH concentration in cloud droplets in the revised manuscript as follows:

Compared with OH concentrations measured in remote clouds from laboratory studies (average: 7.2×10^{-15} mol l⁻¹; Arakaki et al., 2013), the modeled average aqueous-phase OH concentration in daytime clouds (9.6×10^{-14} mol l⁻¹) is much higher. The difference between measured and modeled OH concentrations is comprehensively discussed in Tilgner and Herrmann, (2018). The chapter outlined that both model results and laboratory investigations of field samples are biased. However, it should be mentioned that more comprehensive aqueous phase mechanism tends to lower OH predictions due

to higher number of possible OH sinks. On the other hand, laboratory investigations of field samples most likely tend to underestimate the OH sources due to the limitation of present offline methods. For instance, during the time period on the way from the measurement site to the laboratory, the OH radical can still be consumed by oxidation processes that cannot be resolved by the laboratory protocol, and OH sources related to the uptake of OH precursors (H₂O₂, ROOHs, etc.) are also excluded. Therefore, an adequate comparison is rather difficult at present.

(Page 8, Line 233- 242)

(Page 9, Line 264-270)

20. Part 3: 3.1.2: I suggest using mixing ratio for gas-phase instead of concentrations when comparing simulated results with in-situ measurements. As mixing ratio is a relative quantity, it is not dependent on altitude (via pressure and temperature).

Response: We have changed the unit of gas-phase O₃ and H₂O₂ to ppbv as follows: In the C2wo case, measured gas-phase O₃ concentrations at Mt. Tai ranged from 78.6 to 108.3 ppbv (Fig. 2), which is typical in a Chinese suburban regime (Wang et al., 2013b). However, these mixing ratios are reached even at the high altitude of Mt. Tai. Additionally, the simulated maxima gas-phase H₂O₂ concentrations (C2w: 1.0 ppbv, C2wo: 2.3 ppbv) are lower than those observed at a rural site downwind of the more polluted area of Hebei, China (11.3 ppbv) (Wang et al., 2016b). The simulated O₃ maxima (C2w: 94.2 ppbv, C2wo: 105.1 ppbv) are lower than those observed at the Nanjing urban area in China (133.9 ppbv) (An et al., 2015).

21. Part 3: 3.2: As no information is given on how the interactions between aerosols and cloud water are considered, it is difficult to interpret the results. Why on Fig. 3 sulphate remains constant when cloud dissipates whereas nitrate decreases at the same time?

Response: Sulfate is a very strong acid and is the main driver for cloud droplet and aerosol acidity. When the cloud dissipates, the liquid water content is reduced by three orders of magnitude. This will decrease the pH down by a factor of three. As nitrate is

a less strong acid compared with sulfate, a higher amount of nitrate is present as nitrous acid and driven out into the gas phase, which is a common known phenomenon.

22. Part 3: 3.3.1: I suggest recalling in legend of Fig.5 what Gly, wC2, C2, Pyr, MGly and C3 means. Does this figure show the aerosol mass concentrations as shown in Figure 3? If yes, I do not understand why the text page 11, line 316 refers to aqueous phase concentrations. Does this mean that aqueous phase concentration is the same than aerosol mass concentration? These results are difficult to interpret without knowing if thermodynamics equilibrium between gas phase and aerosols is considered for organics species. For instance, for malonic acid, it should be mainly in the aerosols and not in the gas phase (Limbeck, A., Kraxner, Y. and Puxbaum, H.: Gas to particle distribution of low molecular weight dicarboxylic acids at two different sites in central Europe (Austria), Journal of Aerosol Science, 36, 991-1005, 2005).

Response: Figure 5 shows the aerosol mass concentrations of Gly, ωC_2 , C_2 , MGly, Pyr and C_3 . We have changed the legend of Figure 5 as follows:

Figure 5. Time series of the modeled aerosol mass concentrations of selected DCRCs (top: Gly, ω C₂, C₂; bottom: MGly, Pyr, C₃) in the C2w and C2wo cases (light blue column: cloud; shadow: night; green triangle: the maximum (above), average (middle) and minimum (below) values of measured concentrations at Mt. Tai). (Page 39)

We have changed "aqueous-phase" in page 11, line 316 to "aerosol mass" in the revised manuscript as follows:

Figure 5 shows the modeled aerosol mass concentrations of Gly, ωC_2 , C_2 , methylglyoxal (MGly), Pyr, and malonic acid (C_3) both in the C2w and C2wo cases as well as the values measured at Mt. Tai.

(Page 12, Line 361-362)

23. Part 3: 3.3.1: Dicarbonyl compounds: I find that the remarkable result is that Gly and MGly are very similar concentrations at the end of the simulation considering or not cloud chemistry. In addition, the small increase of MGly in the cloud case in

comparison to the no cloud one, seems to be related to its small production inside cloud droplets during night. So, I disagree with this statement: "This might have been caused by the fact that the aqueous oxidation fluxes under nighttime cloud conditions are lower than the ones under daytime".

Response: We have added sentence about similar concentrations of Gly or MGly at the end of the simulation with or without cloud as follows:

It's worth noting that Gly or MGly have similar concentrations at the end of the simulation with or without cloud chemistry.

(Page 12, Line 374-375)

However, OH radical concentrations in the nighttime cloud are one order of magnitude lower than those during daytime cloud (see Figure R2). The result leads a much lower oxidation rate of the OH radical under nighttime cloud conditions (see Figure R1). Furthermore, the increased aqueous-phase NO₃ radical concentrations are not comparable with the decreased aqueous-phase OH radical concentrations (see Figure S3 in the supplement). Thus, slightly increased MGly is related to the lower nighttime MGly oxidation rate. We have changed corresponding sentence in the revised manuscript as follows:

This might have been caused by the fact that the aqueous oxidation fluxes under nighttime cloud conditions are lower than the ones under daytime, because of much lower OH radical concentrations under nighttime cloud conditions (Fig. S3).

(Page 12, Line 370-372)

24. Part 3: 3.3.1: C2 carboxylic acids: Could you please rewrite this passage, which is not clear, only describes the curves and does not give hypotheses about the observed trends.

Response: We have added sentence in the revised manuscript as follows:

Compared with the C2wo case, the C2w case shows higher ω C₂ concentrations, which suggests that cloud processes play important roles in ω C₂ formation and oxidation.

(Page 12, Line 380-381)

25. Part 3: 3.3.1: C3 carboxylic acids: I don't see an increase of Pyr in the nighttime cloud.

Response: During the nighttime cloud, although having a sharp increase and decrease at the beginning, Pyr concentrations are indeed increased. For example, during nighttime cloud time interval ranging from 24-24.96 h, Pyr concentrations are increased from 28.3 to 33.6 ng m⁻³; during nighttime cloud time interval ranging from 47.3-48.96 h, Pyr concentrations are increased from 122.2 to 125.6 ng m⁻³; during nighttime cloud time interval ranging from 71.3-72.96 h, Pyr concentrations are increased from 182.5 to 185.0 ng m⁻³.

26. Part 3: 3.3.2: As described in Zhu et al. (2018), DCRCs observations are available for daytime and nighttime. Thus, I suggest indicating in Table 3, ratio for day and night. Could you please specify on which simulated period the averages are done (the three days or only the third day)?

Response: We only use the third day data to calculate the ratio of the modeled and measured DCRC compounds, and Table 3 has been changed as follows:

Table 3. Ratios of the concentrations of the modeled and measured DCRC compounds in the different model trajectories at Mt. Tai.

Compound	Model case			
	C2w-day	C2w-night	C2wo-day	C2wo-night
C_2	0.30	0.27	0.23	0.21
ωC_2	7.07	6.94	3.35	3.43
C_3	1.82	1.86	1.57	1.58
Pyr	8.95	7.12	4.34	3.22
Gly	0.19	0.23	0.13	0.16
MGly	2.30E-3	2.72E-3	1.35E-3	1.67E-3

(Page 33-34)

27. Part 3: 3.3.2: By this sentence "SPACCIM overestimates the measured ω C2 concentrations, but underestimates the measured C2 ones, suggesting the conversion of ω C2 might be implemented less efficiently into CAPRAM.", do you mean that the aqueous phase oxidation of ω C2 producing C2 implemented in CAPRAM seems to be less efficient than in the field?

Response: Yes. We have rephrased the sentence to make it clear as follows:

SPACCIM overestimates the measured ωC_2 concentrations, but underestimates the measured C_2 ones, suggesting that the conversion of ωC_2 is implemented less efficiently into CAPRAM than it is seen in the field.

(Page 13, Line 404-406)

28. Part 3: 3.3.2: Page 12, lines 356-362: the hypothesis of missing processes enhancing the partitioning of MGly implies that the model is capable of simulating realistic total MGly concentrations (gas phase + aerosols): is-it the case?

Response: Missing partitioning processes of MGly is an important reason for the extremely low MGly concentration in model simulation, which is about three orders of magnitude lower than the measured data. Even if complementing the missing MGly partitioning processes, SPACCIM model may not simulate realistic MGly concentrations in the field observation. Because other factors, such as model input data (e.g. emission inventory, initial gas- and aqueous-phase concentrations of key species at originated areas), also impact the modeled MGly concentration. We have changed the corresponding sentence in the revised manuscript as follows:

Phase partitioning between gas and aqueous phase, a key process for modeled MGly concentration, may be not sufficient enough to predict the measured MGly aerosol concentrations in the field because of model simplicity.

(Page 13, Line 406-408)

29. Part 3: 3.3.2: I disagree with this statement "The emission data are obtained through model calculations, not field measurements". Emission data comes from inventory, which is developed in part based upon measurements. I guess that the more probable error coming from emission data is due to the horizontal and temporal resolution of inventories used. This point should be discussed.

Response: We have changed "model calculations" to "emission inventory", and added the impact of horizontal and temporal resolution of inventory in the revised manuscript as follows: The emission data are obtained through a new anthropogenic emission inventory in Asia, which provides monthly emissions in 2010 by sector at $0.25^{\circ} \times 0.25^{\circ}$ resolution. However, the model simulation period in this study is in 2014 and the spatial resolution is less than $5^{\circ} \times 2^{\circ}$. Therefore, errors in conversion of the emission data can occur. (Page 13, Line 416-419)

30. Part 3: 3.3.2: Why do you mean by "The height of Mt. Tai (about 1500 m) also causes its input to be indefinite."?

Response: Anthropogenic sources in emission inventory, such as power, industry, residential and agriculture, emitted near the Earth's surface. However, they can be transported over long distances and a wider area. A key factor on the transport is the boundary layer, which is the lowest atmospheric layer immediately affected by the Earth's surface. Anthropogenic pollution generates a strong lid on the top of the boundary layer, hindering turbulent mixing of pollutants from the surface to higher up. Aerosol particles increase the boundary layer stability and cause any subsequent emissions to remain in a lower boundary layer, reduce the mixing height even further. The boundary layer height varies from a few dozen meters to a few kilometers. Therefore, the height of Mt. Tai (about 1500 m) also causes its input to be indefinite.

31. Part 3: 3.3.2: MCM mechanism: I guess you could cite other limitations, in particular in biogenic VOC oxidation as the Mt. Tai is surrounded by deciduous forest. Response: This is correct, but the height of Mt. Tai is more than 1500 m, and often above the planetary boundary layer. Therefore, it is decoupled from direct ground emissions most of the time (Zhu et al. 2018). Hence, these missing values are low in their limitation and not discussed further.

Reference:

Zhu, Y., Yang, L., Chen, J., Kawamura, K., Sato, M., Tilgner, A., van Pinxteren, D., Chen, Y., Xue, L., Wang, X., Simpson I. J., Herrmann, H., Blake D. R., and Wang, W. X.: Molecular distributions of dicarboxylic acids, oxocarboxylic acids and α-

dicarbonyls in PM2.5 collected at the top of Mt. Tai, North China, during the wheat burning season of 2014, Atmos. Chem. Phys., 18, 10741-10758, 2018.

32. Part 3: 3.3.2: I disagree that ratios are acceptable except for C3. I think that it is an interesting result to show that, even a sophisticated model as the one used in this study, is not able to reproduce DCRCs observations. A discussion trying to assess why the C3 ratio is close to 1 should be interesting.

Response: We have deleted "which can be regarded as an acceptable range due to the model and input data limitations" in the manuscript.

(Page 14, Line 436-437)

A discussion about the reasons of C_3 ratio is close to 1 has been added in the revised manuscript as follows:

Interestingly, the ratio of C₃ is close to 1, which might be related to a good representation of the concentrations of C₃ precursors. The comparison indicates that formation pathways of DCRCs implemented in CAPRAM4.0 is realistic, but highly dependent on the input data of precursors.

(Page 14, Line 437-439)

33. Part 3: 3.4.1: ω C2: I see a low net formation flux during the last night for non-cloud period.

Response: We have checked the net formation flux and net sink flux of ωC_2 , and found really an extremely low net formation flux during the non-cloud period. We have changed the corresponding sentences in the revised manuscript as follows:

The results reveal a net formation flux that mainly occurs during cloud conditions as well as a net degradation mainly during non-cloud periods. About 94 % of the net formation flux of ωC_2 is simulated under cloud condition. However, the non-cloud conditions represent 99 % of the net sink flux of ωC_2 .

(Page 15, Line 464-467)

34. Part 3: 3.4.2: C2: Please moderate the last sentence: one of the reasons and not the reason.

Response: We have changed "the reason" to "one of the reasons" in the revised manuscript as follows:

The possible overestimation of the photolytic decay leads to a significantly low C_2 concentration and might be thus one of the reasons for the underestimated C_2 concentration.

(Page 16, Line 501-503)

35. Part 3: 3.4.3: Pyr: Do you consider the photolysis of Pyr in aqueous phase (Reed Harris, A. E., Ervens, B., Shoemaker, R. K., Kroll, J. A., Rapf, R. J., Griffith, E. C., Monod, A. and Vaida, V.: Photochemical Kinetics of Pyruvic Acid in Aqueous Solution, J. Phys. Chem. A, doi:10.1021/jp502186q, 2014.)?

Response: No, we don't consider the Pyr photolysis in the present aqueous phase mechanism. It will be considered in CAPRAM in the future. However, it needs to be mentioned that a large fraction of Pyr is in the hydrated form in acidic solutions. Therefore, the photochemical active carbonyl group is deactivated and other processes might play more important roles in aqueous solution for the degradation of Pyr, such as H_2O_2 reaction.

36. Part 3: 3.5.1: Green lines on Fig. 7 are difficult to see: could you please use another colour (grey for instance)?

Response: We have changed green lines to tawny lines in Figure 7 as follows:

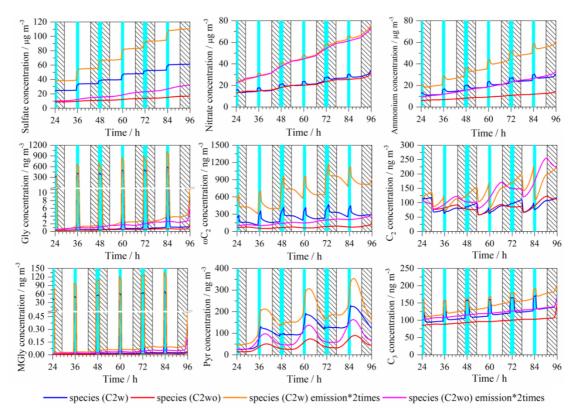


Figure 7. Concentration variations of modeled sulfate, nitrate, ammonium, Gly, ω C₂, C₂, MGly, Pyr and C₃ when doubling emission data (light blue column: cloud; shadow: night).

(Page 41-42)

37. Part 3: 3.5.2: Could you please explain how negative and positive RIR values have to be interpreted? It would help the reader to follow the discussion on DCRCs precursors. Page 17, line 519: please suppress As.

Response: We have explained positive and negative RIR values in the revised manuscript as follows:

The positive or negative RIR value reveals that reducing precursor emissions would weaken or aggravate DCRCs formation, respectively.

(Page 18, Line 554-555)

We have deleted "As" as follows:

For Pyr, in both C2w and C2wo cases, alkenes are the dominant precursor group with the largest RIRs.

(Page 18, Line 575)

Multiphase MCM/CAPRAM modeling of formation and processing of secondary aerosol constituents observed at the Mt. Tai summer campaign 2014

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Abstract. Despite the high abundance of secondary aerosols in the atmosphere, their formation mechanisms remain poorly understood. In this study, MCM/CAPRAM mechanism is used to investigate the multiphase formation and processing of secondary aerosol constituents during the advection of air masses towards the measurement site of Mt. Tai in North China. Trajectories with and without chemical cloud interaction are modeled. Modeled radical and non-radical concentrations demonstrate that the summit of Mt. Tai, with an altitude of ~1.5 km a.m.s.l., is characterized by a sub-urban oxidants budget. The modeled maximum gas-phase concentrations of OH radical are 3.2×10^6 molecules cm⁻³ and 3.5×10^6 molecules cm⁻³ in simulations with and without cloud passages in the air parcel, respectively. Different to previous studies at Mt. Tai, this study has modeled chemical formation processes of secondary aerosol constituents under day vs. night and cloud vs. noncloud cases along the trajectories to Mt. Tai in detail. The model studies show that sulfate is mainly produced in simulations where the air parcel is influenced by cloud chemistry. Under the simulated conditions, the aqueous reaction of HSO3 with H₂O₂ is the major contributor to sulfate formation, contributing 67 % and 60 % in the simulations with cloud and non-cloud passages, respectively. The modeled nitrate formation is higher at nighttime than at daytime. The major pathway is aqueousphase N₂O₅ hydrolysis, with a contribution of 72 % when cloud passages are considered and 70 % when not. Secondary organic aerosol (SOA) compounds, e.g. glyoxylic, oxalic, pyruvic and malonic acid, are found to be mostly produced from the aqueous oxidations of hydrated glyoxal, hydrated glyoxylic acid, nitro 2-oxopropanoate and hydrated 3-oxopropanoic acid, respectively. Sensitivity studies reveal that gaseous VOC emissions have a huge impact on the concentrations of modeled secondary aerosol compounds. Increasing the VOC emissions by a factor of two leads to linearly increased concentrations of the corresponding SOA compounds. Studies using the relative incremental reactivity (RIR) method have identified isoprene, 1,3-butadiene and toluene as the key precursors for glyoxylic and oxalic acid, but only isoprene is found

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to be a key precursor for pyruvic acid. Additionally, the model investigations demonstrate that an increased aerosol partitioning of glyoxal can play an important role in the aqueous-phase formation of glyoxylic and oxalic acid. Overall, the present study is the first that provides more detailed insights in the formation pathways of secondary aerosol constituents at Mt. Tai and clearly emphasizes the importance of aqueous-phase chemical processes on the production of multifunctional carboxylic acids.

1 Introduction

- 40 Secondary aerosols are more abundant than primary aerosols (Volkamer et al., 2006). Their constituents are formed on a regional scale and transported over long distances and thus have a direct impact on the air quality of a wider area (Kim et al., 2007; Matsui et al., 2009; DeCarlo et al., 2010). Secondary aerosols are usually divided into two classes: secondary inorganic aerosol (SIA) and secondary organic aerosol (SOA). A number of studies have been conducted aiming at investigating their formation mechanisms (Yao et al., 2002; Duan et al., 2006; Wang et al., 2006; Guo et al., 2010; Zhao et al., 2013). The SIA components, including sulfate, nitrate and ammonium, are important contributors to fine particulate matter (PM2.5) and play an important role in haze formation (Volkamer et al., 2006; Sun et al., 2014; Wang et al., 2014; Zhang et al., 2014). The SIA formation processes are relatively well understood, but some indefiniteness still remains, such as multiphase formation, particularly under highly polluted conditions such as in China (Wang et al., 2014; Wang et al., 2016a). SOA is also a key component of PM2.5 and linked to adverse health effects, visibility reduction and climate change (Tabazadeh, 2005; Seagrave et al., 2006; De Gouw and Jimenez, 2009; Shrivastava et al., 2017). However, their formation mechanisms are still not well understood (Huang et al., 2014).
 - Dicarboxylic acids and related compounds (oxo-carboxylic acids and α -dicarbonyls) (DCRCs) are ubiquitous water-soluble components of SOA (Kawamura and Sakaguchi, 1999; Kawamura and Yasui, 2005; Pavuluri et al., 2010). They are mainly produced by secondary processes of precursors via gas-phase and subsequent aqueous-phase reactions (Glasius et al., 2000; Legrand et al., 2007; Kundu et al., 2010; Tilgner and Herrmann, 2010). A detailed knowledge of the formation processes of
- DCRCs is helpful to better understand the fate of SOAs in the troposphere. A number of studies have proposed that aromatic hydrocarbons, isoprene and ethene are important precursors for DCRCs (Warneck, 2003; Ervens et al., 2004; Bikkina et al., 2014; Tilgner and Herrmann, 2010). However, modeling studies that focus on understanding of DCRCs formation pathways based on field measurements are limited. Additionally, model studies show growing evidence that substantial amounts of
 - DCRCs are formed by aqueous-phase reactions within aerosol particles, clouds and fog droplets (Sorooshian et al., 2006; Carlton et al., 2007, 2009; Ervens et al., 2008, 2011; Ervens, 2015, Tilgner and Herrmann, 2010, Tilgner et al., 2013, Mouchel-Vallon et al., 2017). Nevertheless, the applied mechanisms are still incomplete and the formation processes are therefore not completely understood. Hence, in this study, a near-explicit multiphase model is applied to investigate the chemical processing of DCRCs in both gas and aqueous phases in order to understand the formation processes and the fate of

65 DCRCs in the atmosphere.

Gelöscht: However, formation pathways based on measured data are still limited.

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The present study focuses on the multiphase formation mechanism of key secondary aerosol constituents measured in June 2014 at Mt. Tai, which is the highest mountain in the North China Plain (NCP). Mt. Tai is located in Shandong province in the NCP, and between the Bohai Rim (BHR) and the Yangtze River Delta (YRD) regions. Together, the BHR and YRD regions had a population of more than 450 million in 2018 (China Statistical Yearbook in 2019). In summer, clouds frequently occur over the summit of Mt. Tai. Despite a little emission from temples and small restaurants at Mt. Tai's peak, the sampling site on top of Air Force Hotel, Houshiwu was typically not much influenced by tourists and temples (Sun et al., 2016). The special altitude and geographical location of Mt. Tai provide a suitable site to measure regional secondary aerosol constituents and to investigate their formations pathways along the advection to the measurement site.

The detailed objectives of the present study are as follows: (i) characterization of modeled radical and non-radical oxidant concentrations; (ii) assessment of modeled concentrations and formation processes of key secondary inorganic compounds; (iii) study of modeled concentrations of DCRCs and a comparison with field observations to assess the model predictions; (iv) investigation of source and sink pathways of selected DCRCs; (v) examination of the impact of emission data on modeled secondary aerosol concentrations; (vi) identification of the key precursors of selected DCRCs and (vii) the impact of higher glyoxal (Gly) partitioning constants on the modeled concentrations of Gly, glyoxylic acid (ωC2) and oxalic acid (C2).

Gelöscht: Mt. Tai is located in the Shandong province in the NCP, between the Yangtze River Delta and the Bohai Rim, two of China's three largest economic zones with a population of more than 410 million

5 2 Multiphase modeling and model setup

Detailed descriptions about the sampling site, the sampling instruments and the analysis methods can be found in a previous publication (Zhu et al., 2018). Campaign observation data, meteorological conditions and corresponding findings are also given there. The sampling period was from June 4 to July 4 2014. The meteorological data during the campaign were as follows: temperatures ranged from 10 to 25 °C with an average of 17 ± 6.2 °C; relative humidity (RH) ranged from 58 to 100 % with an average of 87 ± 13 %; the prevailing wind direction was northwest; wind speeds ranged from 1 to 7 m/s. The weather conditions were mostly cloudy and occasionally foggy. Source identification indicated that DCRCs were mainly derived from anthropogenic activities followed by photochemical aging. Secondary sources, fuel combustion, photooxidation of unsaturated fatty acids and waste burning were also significant sources. DCRCs exhibited mostly similar concentrations at daytime and at nighttime, e.g. C2, pyruvic acid (Pyr) and ωC2 (Zhu et al. 2018). These results differed from previous studies at urban (New Delhi, India) (Miyazaki et al., 2009), suburban (the northern vicinity of Beijing) (He et al., 2013) and mountain sites (Central Himalayan) (Hegde and Kawamura, 2012), where diurnal variations were observed. Therefore, this study focused on the formation processes of these compounds to better understand the similar concentrations during daytime and nighttime summer conditions at Mt. Tai.

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105 2.1 Model and mechanism description

In this study, we applied the air parcel model SPACCIM (SPectral Aerosol Cloud Chemistry Interaction Model, Wolke et al., 2005) to simulate multiphase chemistry along main trajectories during a simulated campaign period. SPACCIM combines a multiphase chemical model with a microphysical model, simulating aqueous-phase chemistry in deliquesced particles and cloud droplets. The microphysical model applied in SPACCIM is based on the work of Simmel and Wurzler, (2006) and Simmel et al. (2005). Droplet formation, evolution and evaporation are realized by a one-dimensional sectional microphysics considering deliquesced particles and cloud droplets. In the present study, the moving bin version of SPACCIM has been applied. In the model, the growth and shrinking of aerosol particles by water vapor diffusion as well as nucleation and growth/evaporation of cloud droplets is considered. The dynamic growth rate in the condensation/evaporation process and the droplet activation is based on the Köhler theory. Due to the emphasis on complex multiphase chemistry, other 115 microphysical processes such as impaction of aerosol particles and collision/coalescence of droplets and thus precipitation were not considered in the present study. Moreover, the air parcel model SPACCIM is not able to reflect the complexity of tropospheric mixing processes. Nevertheless, the complex model enables detailed investigations of the multiphase chemical processing of gases, deliquescent particles and cloud droplets. More detailed descriptions of SPACCIM can be found in Wolke et al. (2005), Sehili et al. (2005) and Tilgner et al. (2013). However, SPACCIM cannot assess the complexity of the tropospheric mixing processes along the transport and the effects of non-ideal solutions on the occurring multiphase chemistry. These limitations have to be kept in mind when studying deliquesced particles and comparing predicted and observed concentrations at Mt. Tai. The potential limitations of an ideal solution assumption compared to a non-ideal treatment are discussed in a recent paper by Rusumdar et al. (2020).

The applied multiphase chemistry mechanism is comprised of the Master Chemical Mechanism (MCM3.2 scheme with 13927 reactions, http://mcm.leeds.ac.uk/MCM/, Jenkin et al., 2003; Saunders et al., 2003) and the chemical aqueous-phase radical mechanism (CAPRAM4.0 scheme with 7129 reactions, Bräuer et al., 2019). The MCM3.2 is a near-explicit gas-phase chemistry mechanism, while CAPRAM4.0 explicitly describes the aqueous-phase chemical and phase transfer processes of inorganic compounds and organic compounds of up to 4 carbon atoms. Phase exchange processes (in total 275) are implemented based on the resistance model of Schwartz (1986), in which the mass accommodation coefficient, the gas phase diffusion coefficient and the Henry's law constant are considered.

2.2 Trajectories and performed model simulations

Zhu et al. (2018) have shown that during the sampling period (4 June – 4 July 2014), air masses arriving at Mt. Tai mainly came from the north (named cluster 2) and the south (named cluster 4) (Fig. S1). The two clusters accounted for 79 % of the total trajectories. Moreover, the sum of DCRC concentrations in clusters 2 and 4 amounted to 73 % of total DCRC concentration during the sampling period. Therefore, in this study, we selected clusters 2 and 4 to simulate and investigate the formation processes and the fate of DCRCs. Additionally, Zhu et al. (2018) have clearly shown that biomass burning was

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only an important source during the first half of the sampling period (4 – 19 June). The aim of the study was to investigate the secondary formation of aerosol constituents along the trajectories to Mt. Tai. However, biomass burning can be an important primary source of compounds that are often of secondary origin. Therefore, in this study, we focused on the period that was less impacted by biomass burning. In addition, both clusters 2 and 4 exhibited a rather stable transport above the mixing layer to Mt. Tai site.

A total simulation time of 96 h is chosen (4 days), representing a typical aerosol lifespan (Willams et al., 2002). The first 24 h are considered as a model initialization day. Thus, only the model results from 24 h to 96 h are presented in this study. With the help of measured RH at Mt. Tai and meteorological values of clusters 2 and 4 that are obtained by HYSPLIT4.9 (Draxler and Rolph, 2003) and MODIS satellite pictures (Li et al., 2005), we have obtained that clouds most likely occurred at the Mt. Tai top and advected to Mt. Tai at the altitude of the trajectories (Zhu et al., 2018). Radiosonde data (http://weather.uwyo.edu/upperair/sounding.html) also shows that clouds could occur in the trajectories of clusters 2 and 4. Cloud interactions are considered on the basis of the measured RH values at Mt. Tai and RH values in radiosonde data at about 1500 m. The fractions of RH values above 90 % are calculated and used as a representative for the time possibly spent by a trajectory inside the clouds. For cluster 2, three 1.28 h long daytime and three 1.92 h long nighttime duration cloud cycles are repeated every 24 h. For cluster 4, daytime and nighttime cloud durations are 1.28 h and 1.68 h long, respectively, and are also repeated every 24 h. In order to better identify the impact of cloud droplet chemistry, we have also considered a model scenario without clouds. During the non-cloud period, RH is 70 % from 24 h to 96 h. Acronyms for the different

We have also carried out sensitivity runs in this study, investigating the following three aspects: (i) the impact of considered emission data on modeled secondary aerosol concentrations; (ii) the identification of key precursors of C_2 , Pyr and ωC_2 by relative incremental reactivity (RIR) (Xue et al., 2014), and (iii) the assessment of impact of increased Gly aerosol partitioning on concentrations of Gly, ωC_2 and C_2 . RIR is defined as the ratio of the decrease in the maximum concentrations of the DCRCs and the decrease in the emission data of the corresponding precursor (a 20 % reduction is adopted in this study).

2.3 Chemical and physical model initialization

model simulations performed in this study can be found in Table 1.

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Zhu et al. (2018) have reported that the pollutant concentrations during the campaign at Mt. Tai were largely controlled by long-range transport. The formation processes of secondary aerosols during long-range transport strongly depend on the emission of precursors. Therefore, emission data in clusters 2 and 4 passed over areas are implemented in the model. Biogenic emission data (isoprene, α- and β-pinenes) are obtained from Emissions of atmospheric Compounds and Compilation of Ancillary Data (ECCAD), MEGAN-MACC dataset (https://eccad.aeris-data.fr/), while other input emission data (volatile organic compounds (VOCs including alkanes, alkenes, aromatics, aldehydes, alcohols), CO, CO₂, NH₃, NO, NO₂, SO₂) are obtained from a new anthropogenic emission inventory in Asia (http://www.meicmodel.org/dataset-mix.html) (Li et al., 2017b). Emission data applied in the model can be found in Table S1. The deposition velocities used in SPACCIM

Gelöscht: In order to minimize the impact of biomass burning including primary emissions, the second half of the sampling period (20 June – 4 July) has been chosen as the model simulation period.

Gelöscht: the emissions of the atmospheric compound compilation of ancillary data

(http://eccad.sedoo.fr/eccad_extract_interface/JSF/page_login.jsf),

were taken from Ganzeveld et al. (1998), and presented in Table S2. Additionally, the multiphase chemistry mechanism is also constrained by the initial concentrations of key species at the corresponding originated areas. The initial chemical data include gas-phase concentrations of inorganic gases (NO, NO₂, O₃, SO₂, HNO₃, NH₃, H₂O₂), VOCs (including alkanes, 180 alkenes, aromatics, aldehydes, alcohols and ketones) (Barletta et al., 2005; Duan et al., 2008; An et al., 2009; Liu et al., 2009; He et al., 2010; Ianniello et al., 2011; Meng et al., 2011; Li et al., 2011; Liu et al., 2012b; Zhao et al., 2013b; Wang et al., 2014b; Li et al., 2015a; Li et al., 2015b; Rao, et al., 2016; Wang et al., 2016b) and particle phase data (Hu et al., 2015; Wang et al., 2015; Sun et al., 2015; Liu et al., 2014; Sun et al., 2013). In case of missing values, values are taken from the CAPRAM urban scenario (http://projects.tropos.de/capram, Herrmann et al., 2005). These initial model data, and also aerosol parameters are given in Table S3 and Table S4. The meteorological scenarios are illustrated in Figure S2.

3 Model results and discussion

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3.1 Modeled concentrations of important oxidants

Due to the key role of radical and non-radical exidation in the formation processes of secondary aerosol constituents, their concentration variations and corresponding reasons are investigated. Several publications have already focused on the oxidant budget in China. Kanaya et al. (2009) modeled gas-phase concentrations of OH, HO2 and O3 oxidants at Mt. Tai by Regional Atmospheric Chemistry Mechanism (RACM), but they didn't consider the effect of aqueous-phase conditions, such as cloud cases. Kanaya et al. (2013) just observed O₃ gas-phase concentration at Mt. Tai, and modeled photochemical O₃ production processes by RACM. Suhail et al. (2019) only observed gas-phase NO₃ concentration at Wangdu site in the NCP. Ren et al. (2009) and Wang et al. (2016b) reported measurement of gas-phase peroxides at Mt. Tai and Wangdu site in the NCP, respectively, but no other radical or non-radical oxidants. Liu et al. (2012) modeled gas-phase concentrations of OH, HO₂ and RO₂, and investigated RO_x budgets and O₃ formation in Beijing in the NCP using a 1-D photochemical model. These studies partly investigated the radical or non-radical oxidants, and were lack of aqueous-phase reactions. To our knowledge, this is the first detailed multiphase chemical modeling study examining radical and non-radical oxidant concentrations along the trajectory to the Mt. Tai under day vs. night and cloud vs. non-cloud cases.

200 3.1.1 Key radical oxidants

Figure 1 shows the modeled gas- and aqueous-phase concentrations of important radical oxidants in the C2w and C2wo cases. The gas- and aqueous-phase OH, HO2 and NO3 radical concentrations and the corresponding time profiles in the C4w and C4wo cases are quite similar to those in the C2w and C2wo cases. Therefore, the corresponding plots of the C4w and C4wo cases are presented in Fig. S3.

Due to photochemistry, the gas-phase OH and HO₂ oxidants showed a distinct diurnal profile, but they are significantly influenced by cloud occurrences. Under daytime cloud droplet conditions, OH and HO2 concentrations in the gas phase decreased by about 90 % and 98 %, respectively. The reduction of gas-phase OH radical concentrations in daytime cloud is

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Gelöscht: which are obtained from related literature (Table S3 and Table S4). In case of missing values, they are obtained from the CAPRAM urban scenario (http://projects.tropos.de/capram, Ervens et al., 2003). The initial chemical data include gas-phase concentrations of inorganic gases (NO, NO2, O3, SO2, HNO3, NH3, H2O2), VOCs (including alkanes, alkenes, aromatics, aldehydes, alcohols and ketones) and particle phase data

Gelöscht: These initial model data, and also meteorological and aerosol parameters can be found in Table S3 and Table S4

Gelöscht: In this section, first, the modeled concentrations of key radical and non-radical oxidants are discussed. Second, modeled concentrations and formation processes of key secondary inorganic compounds are outlined. Third, DCRC concentrations are discussed and compared with the measured data at Mt. Tai. Then, formation and degradation pathways of selected DCRCs are investigated through detailed reaction flux analyses. Finally, the impact of precursor emissions on the modeled concentrations of secondary aerosol is investigated, and the most important precursors of the selected DCRCs are identified through sensitivity runs. Moreover, the impact of increased aerosol partitioning of Glv on the concentrations of Gly and its oxidation products (ωC₂ and C₂) is also examined.

Gelöscht: oxidations

Gelöscht: (Crosley, 1995)

Gelöscht: , which are consistent with former model studies (Tilgner et al., 2013; Harris et al., 2014; Whalley et al., 2015)

mainly caused by the reduction of the gas-phase formation pathway of the HO2 + NO reaction. Due to high water solubility. 235 a direct phase transfer of HO2 into cloud droplets contributed significantly to its gas-phase concentration reduction. The NO₃ radical exhibited higher gas-phase concentrations during the night than during the day. The levels are also substantially lower during cloud conditions. Under nighttime cloud droplet conditions, gas-phase NO₃ concentrations decreased by about 72 %. The decreased NO₃ radical concentration mainly resulted from the increased N₂O₅ uptake during cloud conditions. 240 Unfortunately, we did not perform measurements of key radicals during the campaign. However, the simulated maxima of the gas-phase concentrations of OH (C2w; 3.2×10^6 molecules cm⁻³, C2wo; 3.5×10^6 molecules cm⁻³) and HO₂ (C2w; 2.9×10^6 molecules cm⁻³) and HO₂ (C2w; 2.9×10^6 molecules cm⁻³). 108 molecules cm⁻³, C2wo: 3.8 × 108 molecules cm⁻³) for Mt. Tai in this study are comparable to the available measurements listed below. Compared with the modeled maximum OH $(6.0 \times 10^6 \text{ molecules cm}^3)$ and HO₂ $(7.0 \times 10^8 \text{ molecules cm}^3)$ concentrations at Mt. Tai in 2006 in June 2006 using a photochemical box model-that was based on the Regional 245 Atmospheric Chemistry Mechanism (RACM) (Kanaya et al., 2009), the OH and HO₂ concentrations reported here are only slightly lower. Moreover, the modeled OH and HO2 concentrations in this study are lower than those of simulated results over the Chinese megacity Beijing in August 2007 using a 1-D photochemical model (Regional chEmical and trAnsport Model, REAM-1D), whose chemistry was driven by the standard GEOS-Chem gas-phase chemistry mechanism (OH: 9 × 106 molecules cm⁻³, HO₂: 6.8 × 108 molecules cm⁻³) (Liu et al., 2012) and much lower than the measured data by laser induced fluorescence (LIF) at a rural site downwind of the megacity Guangzhou, China in 3-30 July 2006 (OH: 15-26 × 10⁶ molecules cm⁻³, HO₂: 3-25 × 10⁸ molecules cm⁻³) (Lu et al., 2012). Additionally, the simulated NO₃ radical maxima (C2w: 1.0 × 108 molecules cm⁻³, C2wo: 1.5 × 108 molecules cm⁻³) are much lower than those observed at the urban site of Shanghai, China by Differential Optical Absorption Spectroscopy (DOAS) from August 15 to October 7, 2011 (2.5 × 10⁹ molecules cm⁻³) (Wang et al., 2013a). These comparisons suggest that the predicted model data represented a sub-urban 255 oxidant budget along the trajectory above the boundary layer to Mt. Tai. Similar to the gas-phase, aqueous-phase concentrations of OH and HO2 also display a distinct diurnal profile with peak concentration around noon, and are largely impacted by cloud interactions. Under daytime cloud droplet conditions, OH aqueous-phase concentrations are increased by a factor of 3, mainly due to the increased direct transfer of OH from the gas phase. Compared with OH concentrations measured in remote clouds from laboratory studies (average: 7.2 × 10⁻¹⁵ mol 1⁻¹; Arakaki et al., 2013), the modeled average aqueous-phase OH concentration in daytime clouds (9.6 × 10⁻¹⁴ mol l⁻¹) is much higher. The difference between measured and modeled OH concentrations is comprehensively discussed in Tilgner and Herrmann, (2018). The chapter outlined that both model results and laboratory investigations of field samples are biased. However, it should be mentioned that more comprehensive aqueous phase mechanism tends to lower OH predictions due to higher number of possible OH sinks. On the other hand, laboratory investigations of field samples most likely tend to

underestimate the OH sources due to the limitation of present offline methods. For instance, during the time period on the

way from the measurement site to the laboratory, the OH radical can still be consumed by oxidation processes that cannot be

excluded. Therefore, an adequate comparison is rather difficult at present. On the other hand, HO₂ aqueous-phase concentrations are decreased by a factor of 0.9 due to aqueous-phase reactions of HO₂ with Cu²⁺ or Cu⁺. However, higher HO₂ aqueous-phase concentrations appear at the beginning of the daytime cloud. This is caused by the increased H₂O₂ aqueous concentrations due to uptake at the beginning of the daytime cloud period. In the aqueous-phase, H₂O₂ reactions with Mn³⁺ or FeO²⁺ led to a formation of HO₂.

The NO₃ radical presents higher aqueous-phase concentrations during the night, with a minor cloud impact. Under nighttime cloud conditions, the NO₃ radical concentration decreases by about 13 %. In Table 2, average aqueous-phase concentrations of OH, HO₂ and NO₃ under different conditions are presented. Aqueous-phase NO₃ concentrations are much higher than those of aqueous-phase OH. Considering the normally lower reaction rate constant of aqueous NO₃ radical compared to aqueous OH (Herrmann et al. 2010, Herrmann et al. 2015, Ng et al. 2017), both OH and NO₃ radicals are able to initiate aqueous-phase oxidations under the simulated conditions, contributing to the chemical processing of SOA constituents.

3.1.2 Key non-radical oxidants

Figure 2 depicts the modeled gas- and aqueous-phase concentrations of H₂O₂ and O₃ for the C2w and C2wo cases. The corresponding plots for the C4w and C4wo cases can be found in the supplement information because of their similar concentration levels and patterns to C2w and C2wo, respectively (Fig. S4).

Figure 2 shows that, due to active photochemistry, gas-phase concentrations of H₂O₂ and O₃ mainly are increased during the day and decreased during the night. During cloud periods, H₂O₂ gas-phase concentrations are largely decreased by 96 % due to direct phase transfer and corresponding aqueous-phase chemistry. The predicted cloud impact is minor for gas-phase O₃ concentrations, which is consistent with former studies (Tilgner et al., 2013). During daytime cloud periods, O₃ gas-phase concentrations are decreased by about 4 %. In the C2wo case, H₂O₂ concentrations are substantially higher than those in C2w because of the missing cloud phase sinks (e.g. hydrogen sulfide oxidation). However, O₃ concentrations in C2wo and C2w are similar because of the minor cloud impact on O₃ in C2w.

In the C2wo case, measured gas-phase O₃ concentrations at Mt. Tai ranged from 78.6 to 108.3 ppbv (Fig. 2), which is typical in a Chinese suburban regime (Wang et al., 2013b). However, these concentrations are reached even at the high altitude of Mt. Tai. Additionally, the simulated maxima gas-phase H₂O₂ concentrations (C2w: 1.0 ppbv C2wo: 2.3 ppbv) are lower than those observed at a rural site downwind of the more polluted area of Hebei, China (11.3 ppbv) (Wang et al., 2016b). The simulated O₃ maxima (C2w: 94.2 ppbv C2wo: 105.1 ppbv) are slightly lower than those observed at the Nanjing urban area in China (133.9 ppbv) (An et al., 2015).

The aqueous-phase H_2O_2 presents a similar concentration pattern as the gas-phase H_2O_2 . Higher H_2O_2 aqueous-phase concentrations at the beginning of the daytime cloud are caused by the fast phase-transfer. The subsequent sharp decrease during the first minute is connected to the efficient hydrogen sulfide oxidation. In the C2wo case, aqueous-phase O_3 concentrations increase during the day, and decrease throughout the night. In both daytime and nighttime clouds, O_3

Gelöscht: 1.67 to 2.24 × 10¹² molecules cm⁻³

Gelöscht: 2.0 × 10¹⁰ molecules cm⁻³

Gelöscht: 4.7 × 10¹⁰ molecules cm⁻³

Gelöscht: 3.0 × 10¹¹ molecules cm⁻³

Gelöscht: 1.8 × 10¹² molecules cm⁻³

Gelöscht: 2.1 × 10¹² molecules cm⁻³

Gelöscht: 3.6 × 10¹² molecules cm⁻³

aqueous-phase concentrations are increased by about 18 %. The average aqueous-phase concentrations of H_2O_2 and O_3 in the different cases can be found in Table 2.

3.2 Modeled concentrations and formation of key secondary inorganic aerosol constituents

310 In Fig. 3, modeled concentrations of the most important SIA constituents are plotted, including (i) sulfate (sum of all sulfur compounds with oxidation state +6, mainly SO₄²/HSO₄⁻), (ii) nitrate (sum of all nitrogen compounds with oxidation state +5, mainly NO₃⁻), and (iii) ammonium (sum of all nitrogen compounds with oxidation state -3, mainly NH₄⁺). The corresponding plots for the C4w and C4wo cases are presented in Fig. S5.

√3.2.1 Sulfate

315 Conducted field observations together with estimated sulfur oxidation rates using a tracer method in previous studies at Mt. Tai have suggested that sulfate formation is highly related to cloud chemistry (Zhou et al., 2009; Shen et al., 2012; Guo et al., 2012). However, these studies are not able to comprehensively quantify the impact of cloud chemistry on sulfate concentration, and have not performed detailed investigations on chemical formation pathways of sulfate during the transport to Mt. Tai. In this study, we primarily present modeled concentration of sulfate, and discuss the differences between the 320 different day vs. night and cloud vs. non-cloud cases using a multiphase chemistry model. Moreover, findings of sulfate source and sink chemical reactions are presented for the different model cases.

Figure 3 shows that sulfate concentrations mainly increase under cloud condition throughout the whole simulation due to active in-cloud chemical sulfur oxidation pathways. Although in-cloud residence time is slightly higher during the night, sulfate concentrations increase more in the daytime clouds (35 %) than the nighttime clouds (15 %) because of the increased aqueous reaction of HSO₃⁻ with H₂O₂ in daytime cloud droplets (Fig. 4). As shown in Fig. 4, about 97 % of sulfate formation fluxes are related to cloud conditions, and mostly occurred in daytime clouds. The aqueous-phase reaction of HSO₃⁻ with H₂O₂ represents the most important sulfate source in daytime cloud with a contribution of 78 %. In the nighttime cloud, aqueous-phase reaction of HSO₃⁻ with H₂O₂ (42 %), and aqueous reaction of bisulfite with O₃ (28 %) are volument pathways for sulfate formation.

330 In the C2wo case, sulfate concentrations gradually increase over time (Fig. 3). The highest increase occurs during the day as a consequence of the gas-phase SO₂ oxidation by OH (Fig. S6). However, the formed sulfate mass in C2wo case is substantially lower than in the C2w case. At the end of the simulation, the modeled sulfate concentrations are 76.7 and 24.7 μg m⁻³ in the C2w and C2wo cases, respectively. Compared to the maximum (40.0 μg m⁻³), average (32.0 μg m⁻³) and minimum (18.8 μg m⁻³) values of the measured sulfate concentrations at Mt. Tai (Fig. 3), SPACCIM model overestimates measured concentrations of sulfate in the C2w case, but slightly underestimated the mean sulfate concentration in the C2wo case.

3.2.2 Nitrate

Studies at Mt. Tai focused on nitrate suggested that photochemical formation of HNO₃ has important contribution to nitrate formation (Zhou et al., 2009; Guo et al., 2012). Wen et al. (2018) found that partitioning of HNO₃ from gas to particulate

Gelöscht: Sulfate

Gelöscht: dominated

Gelöscht: Nitrate

phase and hydrolysis of N₂O₅ is the predominant daytime and nighttime formation pathway of nitrate at Mt. Tai, respectively. However, these studies did not consider nitrate formation processes related to cloud conditions. In this study, we do not only focus on the concentration time profiles of nitrate under day vs. night, but also compared cloud vs. non-cloud cases. Furthermore, we have analyzed nitrate source and sink reactions rates and their contributions in different cases.

As can be seen in Fig. 3, nitrate concentrations are increased throughout the simulation. Under cloud condition, nitrate concentrations are increased by about 10 % and 24 % during the day and the night, respectively. The concentration time profiles in C2w and C2wo cases show only small differences, indicating that most of the nitrate formation occurs during non-cloud periods. Therefore, the end concentrations of C2w and C2wo do not differ significantly. An analysis of chemical sink and source in the C2w case (Fig. 4) has revealed that nitrate is mainly produced by aqueous-phase N_2O_5 hydrolysis during the night. During the day, nitrate is mainly produced by the aqueous-phase reaction of HNO4 and HSO3 $^{\circ}$, gas-phase reaction of OH and NO_2 , and aqueous-phase N_2O_5 hydrolysis.

A comparison of daytime and nighttime fluxes in the C2wo case has revealed that 31 % and 69 % of nitrate formation fluxes occur at day and at night, respectively. In the C2wo case, nighttime nitrate is mainly produced by aqueous-phase N₂O₅ hydrolysis with a contribution of 92% (Fig. S6), which agrees with the result in Wen et al. (2018). However, gas-phase reaction of OH + NO₂ to gaseous HNO₃ and further partitioning into the aerosols is the most important source for daytime nitrate with a contribution of 56% (Fig. S6). Aqueous-phase N₂O₅ hydrolysis is also significant for daytime nitrate formation with a contribution of 30%. In comparison, Wen et al. (2018) have suggested the partitioning of HNO₃ from gas to the particulate phase to be the most important formation pathway for daytime nitrate with a contribution of 94%. The N₂O₅ hydrolysis has been identified as the second most important pathway with a contribution of 4-6 %.

The modeled nitrate concentrations are 69.5 and 65.3 μ g m³ at 96 h in the C2w and C2wo cases, respectively. Compared to the maximum (25.0 μ g m³), average (14.0 μ g m³) and minimum (6.8 μ g m³) values of the measured nitrate concentrations at Mt. Tai (Fig. 3), the model simulations overestimate the measured concentrations in both cases. <u>Potential reasons are</u>

discussed in Sect. 3.3.2.

▼3.2.3 Ammonium

Measured ammonium concentrations at Mt. Tai can be strongly impacted by acidification and cloud chemistry (Guo et al., 2012; Li et al., 2017a). Still, a detailed analysis of the occurring processes is missing. Therefore, we provide a detailed insight into the ammonium concentration variation trends, and the impact of acidification and cloud processing along the simulated trajectories to Mt. Tai.

Similar to sulfate and nitrate, ammonium concentrations also gradually increased throughout the simulation due to the included emissions rates and the followed uptake of gaseous NH₃ into the aqueous phase. Ammonium concentrations raised in cloud droplets both during the day and night, because a more efficient uptake into the larger cloud water volume. After cloud evaporation, ammonium aerosol concentrations are increased due to stronger partitioning into more acidified deliquesced particles. In the C2wo case, ammonium concentrations are also increased both during the day and night. However, modeled aerosol mass of ammonium in the C2wo case is lower than that in the simulation case with cloud

Gelöscht: Potential reasons are discussed below.

Gelöscht: Ammonium

interaction (C2w case). In the C2wo case, less sulfate is formed. Consequently, deliquesced aerosol particles are less acidified and a larger fraction of ammonium stays in the gas phase as gaseous NH₃. The simulated ammonium concentrations after 96 h are 42.7 and 25.9 μ g m⁻³ in the C2w and C2wo cases, respectively, which are higher than the measured concentrations (maximum: 18.0 μ g m⁻³, average: 15.6 μ g m⁻³, minimum: 7.9 μ g m⁻³).

The differences between the modeled and measured concentrations of sulfate, nitrate and ammonium can be attributed to several issues such as (i) the indefiniteness of the input emission data, (ii) the initial concentrations, (iii) the missing entrainment/detrainment processes, and (iv) the performed heating of the inlet during the sampling of wet aerosol (see Sect. 3.3.2 for further details).

3.3 Modeled concentrations of DCRCs and comparison with field data

In recent years, a number of field observations on DCRCs have been conducted in the NCP. For example, He et al. (2013), Ho et al. (2015), Zhao et al. (2018) and Yu et al. (2019) have observed DCRCs in Beijing; Wang et al. (2009), Kawamura et al. (2013), Meng et al. (2018) and Zhao et al. (2019) have measured DCRCs at Mt. Tai. Our field observation about DCRCs at Mt. Tai has been reported in Zhu et al. (2018). However, these studies are focused on DCRC concentrations, molecular compositions, temporal variations, size distributions, source implications and stable carbon isotopic composition. They have not investigated the chemical formation of DCRC concentrations along the trajectory as well as the impact of cloud and non-cloud conditions on DCRC concentrations. To our knowledge, a multiphase chemical model study investigating the DCRCs concentration variations and their chemical processing along the trajectory to Mt. Tai considering day/night and cloud/non-cloud cases has not been yet reported.

3.3.1 Modeled concentrations of DCRCs

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Figure 5 shows the modeled aerosol mass concentrations of Gly, ωC₂, C₂, methylglyoxal (MGly), Pyr, and malonic acid (C₃) both in the C2w and C2wo cases as well as the values measured at Mt. Tai. In the C4w and C4wo cases, the modeled Gly, ωC₂, C₂, MGly, Pyr, and C₃ concentrations show similar patterns to those in the C2w and C2wo cases, respectively (Fig. S8). Dicarbonyl compounds

In the C2w case, Gly and MGly concentration patterns show a substantial uptake into cloud droplets. Gly concentrations decreased during the daytime and nighttime cloud droplet periods due to in-cloud oxidation processes. On the other hand, MGly concentrations display a decrease in the daytime cloud droplets, but an increase under nighttime cloud conditions. This might have been caused by the fact that the agreeous oxidation fluxes under nighttime cloud conditions are lower than

This might have been caused by the fact that the aqueous oxidation fluxes under nighttime cloud conditions are lower than the ones under daytime. This might have been caused by the fact that the aqueous oxidation fluxes under nighttime cloud conditions are lower than the ones under daytime, because of much lower OH radical concentrations under nighttime cloud conditions (Fig. S3). In the C2wo case, Gly and MGly concentrations are very low due to the low partitioning towards aqueous particles that has been predicted by the model. The effect of a potentially higher partitioning constant of Gly

Gelöscht: aqueous-phase

[1] verschoben (Einfügung)

Gelöscht: shows

(Volkamer et al., 2009; Ip et al., 2009) is investigated in Sect. 3.5.3. It's worth noting that Gly or MGly have similar concentrations at the end of the simulation with or without cloud chemistry.

C2 carboxylic acids

- 415 In the C2w case, aqueous-phase concentrations of ωC2 are increased under both daytime and nighttime cloud conditions and early in the night of non-cloud periods. On the other hand, ωC2 concentrations are lowered during the day and later in the night under non-cloud conditions. In the C2wo case, ωC2 is decreased during the morning periods but is increased in the late afternoon and at night. Compared with the C2wo case, the C2w case shows higher ωC2 concentrations, which suggests that cloud processes play important roles in ωC2 formation and oxidation.
- 420 In the C2w case, modeled aqueous-phase concentrations of C2 are increased under daytime cloud conditions and daytime aqueous particle conditions, but are lowered during nighttime cloud periods and under nighttime aqueous particle conditions. In the C2wo case, C2 is increased during the day but is decreased during the night.

C3 carboxylic acids

- In the C2w case, Pyr concentrations are raised during the daytime and nighttime cloud conditions as well as in the late mornings of the non-cloud periods. Pyr concentrations are decreased in the early morning, afternoon and nighttime non-cloud conditions (caused by the efficient degradation from the reaction with aqueous-phase H₂O₂). In the C2wo case, aqueous-phase concentration profile of Pyr is characterized by an increase during the morning and early afternoon period and by a strong decrease during the late afternoon and night. Pyr shows a high correlation with aqueous-phase H₂O₂ in the C2wo case due to its efficient H₂O₂ decay (Fig. S7).
- 430 The aqueous-phase C₃ concentrations are increased during cloud formation due to the uptake of gaseous C₃ into cloud droplets. Moreover, C₃ concentrations are increased during non-cloud periods. In the C2wo case, C₃ concentrations are increased slightly during the night but even less during the day. In comparison to the C2wo case, the C2w case shows higher concentrations of C₃, indicating that cloud oxidation processes are very important for C₃ aqueous-phase formation under the simulated conditions. The production of C₃ is 27 % higher in the C2w case than in the C2wo case.

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3.3.2 Comparison to field observations

The ratios of the average concentration of modeled and measured DCRCs can be found in Table 3. The results show that model predictions are higher than the measured concentrations of C_3 , Pyr and ωC_2 in both C_2 and C_3 and C_4 and C_5 are the concentration ratios of Pyr and C_5 in the C_5 are much higher than in the C_5 are C_5 on the other hand, model results underpredict the C_5 Gly and MGly concentrations in both cases.

SPACCIM overestimates the measured ω C₂ concentrations, but underestimates the measured C₂ ones, suggesting the conversion of ω C₂ is implemented less efficiently into CAPRAM than it is seen in the field. Phase partitioning between gas and aqueous phase, a key process for modeled MGly concentration, may be not sufficient enough to predict the measured MGly aerosol concentrations in the field because of model simplicity. Kroll et al. (2017) have found that a possible

[1] nach oben verschoben: In the C4w and C4wo cases, the modeled Gly, oC2, C2, MGly, Pyr, and C3 concentrations show similar patterns to those in the C2w and C2wo cases, respectively (Fig. S8).

Gelöscht: the conversion of ω C2 might be implemented less efficiently into CAPRAM.

Gelöscht: The partitioning treatment of MGly

hydration of MGly in the gas-phase might lead to an enhanced uptake into aqueous particles. Thus, maybe the MGly uptake is underestimated. Additionally, other important processes that can effectively enhance partitioning of MGly are not yet considered, such as reversible oligomerizations (Ervens and Volkamer, 2010). As a result, the modeled aqueous-phase MGly concentration is rather low and about three orders of magnitude lower than the measured data. Based on this finding, a sensitivity study has been performed (see Sect. 3.5 for details).

The over- and <u>under-estimation</u> of the measured concentrations of inorganic and organic aerosol constituents could have the following reasons:

(1) Input data: Indefiniteness of emission data and initial concentrations. The emission data are obtained through a new anthropogenic emission inventory in Asia, which provides monthly emissions in 2010 by sector at 0.25° × 0.25° resolution. However, the model simulation period in this study is in 2014 and the spatial resolution is less than 5° × 2°. Therefore, errors in conversion of the emission data can occur. The height of Mt. Tai (about 1500 m) also causes its input to be indefinite. Furthermore, initial concentrations in originated areas are obtained through related references rather than field measurements, which also lead to indefiniteness.

- 465 (2) Field measurement: PM_{2.5} samples are the only ones available, so a possible contribution of larger particles might have been missed. Moreover, heating the inlet during the sampling of wet aerosol has definitely lowered the measured concentrations of more volatile compounds such as ammonium nitrate.
- (3) MCM mechanism: Some species' crucial gaseous precursors are efficiently destroyed by MCM. For example, it is recommended that gas-phase oxidation of acetic acid by OH proceed via both an H-abstraction from the OH-group and the CH₃-group (Sun et al., 2009). Nonetheless, only an H-abstraction for OH-group is implemented in the MCM. This oxidation scheme is implemented for all carboxylic acids. The disadvantage of the MCM probably leads to an underestimation of the functionalized acids and need to be resolved in more detail. However, the development of an improved gas-phase acid oxidation scheme for the MCM goes beyond the scope of this study.
 - (4) CAPRAM mechanism: Missing sources of selected DCRCs from higher organic chemistry, such as the aqueous-phase chemistry of aromatic compounds (Hoffmann et al. 2018).
 - (5) SPACCIM model: The model neglects non-ideal solution effects and do not consider organic salt formation. These factors possibly result in overestimated or underestimated results. Recent studies by Rusumdar et al. (2020) have demonstrated that non-ideal chemistry treatment led to much higher concentrations of C_2 and ωC_2 .

Apart from MGly, the concentration ratios of the modeled and measured species ranged from 0.1 to 8.3. Interestingly, the ratio of C₃ is close to 1, which might be related to a good representation of the concentrations of C₃ precursors. The comparison indicates that formation pathways of DCRCs implemented in CAPRAM4.0 is realistic, but highly dependent on the input data of precursors. The SPACCIM model with an implemented MCM3.2/CAPRAM4.0 is a numerical tool that can help us understand the complexity of the multiphase processing of DCRCs better. However, the present study also demonstrates that there are still huge gaps in knowledge about the formation and degradation of secondary aerosol constituents. Hence, further laboratory investigations and modeling work are necessary.

Gelöscht: underestimation

Gelöscht: model calculations

Gelöscht: not field measurements.

Gelöscht: , which can be regarded as an acceptable range due to the model and input data limitations

3.4 Model investigations of chemical sources and sinks of selected DCRCs

Although field observations have speculated about several potential formation pathways of some DCRCs species by correlations or ratios analyses (Hegde and Kawamura, 2012; Kawamura et al., 2013; Zhao et al., 2019), the detailed pathways of DCRCs need to be studied.

Multiphase model simulations are a suitable tool to investigate DCRCs formation processes. In recent years, DCRCs formation processes have been examined by several model studies. For example, Tilgner and Herrmann (2010) have modeled gas and aqueous phases processing of C₂-C₄ carbonyl compounds and carboxylic acids by SPACCIM; Ervens et al. (2004) have discussed formation pathways of Pyr and C₂-C₆ dicarboxylic acids in gas and aqueous phases; Lim et al. (2005) have reported the formation pathways of Gly, MGly, Pyr and C₂ by isoprene oxidation in gas and aqueous phases using a photochemical box model; Huisman et al. (2011) have investigated the driving processes of Gly chemistry by Master Chemical Mechanism (MCM, v3.1). These studies have suggested that DCRC formations are related to the oxidations of anthropogenic (e.g. toluene and ethylene) and biogenic (e.g. isoprene) gas-phase VOCs precursors. The emissions of these anthropogenic and biogenic VOC in China are much higher than those reported in above references (Sindelarova et al., 2014; Bauwens et al., 2016). However, multiphase model simulations are sparsely used to study DCRCs formations in China.

Therefore, the present study aimed at a detailed investigation of the formation pathways of selected DCRCs under day vs. night and cloud vs. non-cloud cases along the trajectories approaching to Mt. Tai.

Due to the similar concentration levels and corresponding variation trends of ωC_2 , C_2 , Pyr and C_3 in clusters 2 and 4, the source and sink flux analyses are only presented and discussed for the C2 case. Additionally, the corresponding plots of the four compounds in the C2wo case are presented in the supplement information (Fig. S9).

510 3.4.1 Glyoxylic acid (ωC₂)

In Fig. 6, the multiphase source and sink fluxes of ωC_2 (C2w case) on the third model day are plotted. The results reveal a net formation flux that mainly occurs during cloud conditions as well as a net degradation <u>mainly</u> during non-cloud periods. About 94 % of the net formation flux of ωC_2 is simulated under cloud condition. However, the non-cloud conditions represent 99 % of the net sink flux of ωC_2 .

Under daytime and nighttime cloud conditions, the major formation pathways of ωC₂ are aqueous-phase reactions of

hydrated Gly with the OH and SO_x radical (contribution: 60 % at day, 86 % at night), which is similar with results in Tilgner and Herrmann (2010) and Tilgner et al. (2013). The aqueous-phase oxidation of glycolate by OH is also significant under daytime cloud conditions with a contribution of 18 %. Under daytime and nighttime non-cloud conditions, aqueous reactions of hydrated Gly (contribution: day: 20 %, night: 20 %) and gas-phase decay of 3-methyl-6-nitrocatechol (contribution: day: 14 %, night: 24 %) are significant for ωC₂ formation. Other reactions contributed less than 5 % to the overall source flux. Under daytime clouds, ωC₂ sink is dominated by aqueous-phase reaction of hydrated glyoxylate with OH (contribution: 88 %), which is consistent with Ervens et al. (2004). During nighttime clouds, however, aqueous-phase reactions of hydrated

Gelöscht: in

Gelöscht: 100 %

glyoxylate with NO₃ (45 %) and OH (28 %) are the most important sinks. In contrast to those under cloud conditions, gasphase ω C₂ photolysis (57 %) and gas-phase reaction of ω C₂ with OH (18 %) acts as key sinks of ω C₂ under daytime noncloud conditions. Under nighttime non-cloud conditions, the sink fluxes of ω C₂ are low and therefore unimportant.

3.4.2 Oxalic acid (C2)

Figure 6 also depicts the source and sink fluxes of C₂ in the C2w case. The model has simulated a net formation flux during the non-cloud periods and a net degradation in the early morning hours when non-clouds are present but the photolysis of iron-oxalate complexes is efficient. A net formation of about 94 % C₂ is simulated under non-cloud conditions. About 74 % of the net C₂ sink fluxes are predicted during the early morning non-cloud period, and 26 % are related to the cloud oxidation fluxes.

The most important source of C₂ in the aqueous phase under cloud condition is aqueous oxidations of hydrated glyoxylate with the OH radical, which agrees with other model results (Tilgner and Herrmann, 2010; Ervens et al., 2004). This formation pathway contributes to C₂ formation with about 72 % during the day and 87 % during the night. In contrast to that under cloud condition, the most important C₂ formation pathway during daytime and nighttime non-cloud conditions is the aqueous-phase oxidation of hydrated ωC₂ with the OH radical (contribution: day: 39 %, night: 52 %). The field observations also suggested that aqueous-phase oxidation of ωC₂ is the most important formation pathway of C₂ (Kundu et al., 2010; Kawamura et al., 2013; He and Kawamura, 2010), but they are not able to quantify contribution and the responsible specific

oxidation pathways. Other reactions contribute less than 5 % to the overall source flux.

The most important sink of C₂ under daytime cloud conditions is the photolysis of iron-oxalate complexes, with a contribution of about 93 %. The result is similar to reported laboratory experiment findings (Pavuluri and Kawamura, 2012) and aqueous model simulation (Tilgner and Herrmann, 2010). On the other hand, aqueous reaction of oxalate with NO₃ (80 %) is the most important sink in the nighttime cloud case. Under nighttime non-cloud conditions, C₂ sink is dominated by the reaction of the oxalate monoanion with NO₃ (81 %). Under daytime non-cloud conditions, only significant C₂ sink is the photolysis of iron-oxalate complexes. However, photolysis of iron-oxalate complexes under aqueous particle conditions is most likely overestimated in the present SPACCIM model studies. Recent studies by Rusumdar et al. (2020) – using an advanced SPACCIM model with a non-ideality treatment, but with a more reduced chemical CAPRAM scheme – reveal that the formation and consequently the photolysis of iron-oxalate complexes is inhibited under aqueous particle conditions. The possible overestimation of the photolytic decay leads to a significantly low C₂ concentration and might be thus one of the

3.4.3 Pyruvic acid (Pyr)

reasons for the underestimated C2 concentration.

The modeled source and sink fluxes of Pyr in the C2w case on the third model day can be found in Fig. 6. A net formation flux is modeled mainly under cloud condition, especially during the day, along with a net degradation during non-cloud

Gelöscht: the reason

periods. About 72% of the net Pyr flux occurs in clouds, whereas 28% is formed under non-cloud conditions. However, 100% of the multiphase Pyr net sink fluxes are related to non-cloud oxidation.

Under cloud condition, the dominant source for Pyr is hydrolysis of the aqueous-phase oxidation product of nitro
2-oxopropanoate, with a contribution of 89 % during the day and 70 % during the night. The result is different from former
model studies, e.g. Ervens et al. (2004), Lim et al. (2005), Tilgner and Herrmann (2010) and Tilgner et al. (2013), which
modeled the aqueous oxidations of MGly as the major formation pathway of Pyr. However, these model studies have also
modeled different environmental conditions with much lower anthropogenic pollution including lower NO_x and NO₃ radical
budgets compared to the Chinese conditions. Furthermore, the former studies have also used rather simple gas-phase
mechanisms and lacks potential production pathways from higher organic compounds. Similarly, the aqueous oxidation of
nitro 2-oxopropanoate is identified as a major source under non-cloud conditions, with a contribution of 87 % during the day
and 74 % during the night.

The key sinks of Pyr under daytime cloud conditions are aqueous-phase reactions of pyruvate with OH (58 %) and H_2O_2 (29 %). This finding is consistent with results in laboratory experiment (Carlton et al., 2006). Under night cloud conditions, the sink fluxes are very low and therefore can be ignored. Under daytime and nighttime non-cloud conditions, dominant sinks are aqueous-phase reactions of pyruvate with H_2O_2 (contribution: day: 57 %, night: 72 %) and free Pyr with H_2O_2 (contribution: day: 13 %, night: 15 %). Additionally, gas-phase Pyr photolysis (15 %) is also important under daytime non-cloud conditions.

3.4.4 Malonic acid (C₃)

575 In Fig. 6, the modeled source and sink fluxes of C₃ in the C2w case are presented for the third model day. A net formation flux can be seen under daytime cloud conditions and both daytime and nighttime non-cloud conditions. A net degradation is only found under nighttime cloud conditions. A C₃ net formation flux is about 82 % under non-cloud conditions and 18 % under daytime cloud conditions.

The major modeled sources of C_3 under daytime and nighttime cloud conditions are aqueous oxidation reactions of hydrated 3-oxopropanoic acid (contribution: day: 48 %, night: 50 %) and hydrated 3-oxopropanoate (contribution: day: 45 %, night: 47 %). However, under non-cloud conditions, aqueous-phase oxidation of hydrated 3-oxopropanoic acid is dominant, with a contribution of 79 % during the day and 88 % during the night.

Differences between the sink fluxes under cloud and non-cloud conditions are modeled. The C₃ sinks under daytime clouds are dominated by aqueous-phase reaction of malonate monoanion with OH. Its contribution to total sink flux under the daytime cloud is 70 %. Contrary to that, aqueous-phase reaction of malonate monoanion with NO₃ is the most important sink under nighttime cloud conditions with a contribution of 75 %. The predominated sink pathway of C₃ in daytime cloud is consistent with Ervens et al. (2004), but that in nighttime cloud is different due to the missing NO₃ radical pathways in their mechanism. The modeled C₃ sinks under non-cloud conditions are much lower than the sinks under cloud conditions, and thus are unimportant.

3.5 Sensitivity studies

Due to the similarity between clusters 2 and 4, as mentioned above, sensitivity tests are only performed under cluster 2 conditions. The present study investigated (i) impact of emissions on modeled compound concentrations, (ii) key precursors of DCRCs, and (iii) impact of increased Gly aerosol partitioning on Gly, ω C₂ and C₂.

3.5.1 Impact of emissions

595 First, sensitivity tests are performed to evaluate the effect of different emission strengths on the concentrations of key secondary inorganic compounds and selected DCRCs during the transport. The emission sensitivities of sulfate, nitrate, ammonium, Gly, ωC₂, C₂, MGly, Pyr and C₃ are investigated by doubling the applied emission fluxes of the base case. The results of the sensitivity tests can be found in Fig. 7. The modeled concentrations of sulfate, nitrate, ammonium, Gly, ωC₂, C₂, MGly, Pyr and C₃ are increased by a factor of about two when the emissions doubled, which suggests an almost linear dependency. The results indicate that the modeled concentrations of secondary aerosol are very sensitive to the emissions in the model

3.5.2 Identification of the key precursors of C2, Pyr and ωC2

Further sensitivity tests are conducted to identify key primary precursors of DCRCs during atmospheric transport. We have adopted the relative incremental reactivity (RIR) method by Carter and Atkinson (1989) for the sensitivity tests. The positive or negative RIR value reveals that reducing precursor emissions would weaken or aggravate DCRCs formation, respectively. The RIR method has been applied already in a former study to investigate the precursors of peroxy acetyl nitrate in urban plume in Beijing (Xue et al., 2014).

As can be seen in Fig. 8, C_2 formation in the C2w case is mostly sensitive to aromatics and alkenes. Among the aromatic compounds, toluene is the most important one for C_2 formation. However, other aromatic species (such as xylene, ethyl benzene, isopropyl benzene) present negative RIRs. Among the alkenes, isoprene and 1,3-butadiene are dominant, but ethene shows negative RIR. The alkane RIRs are all negative. Positive and negative RIRs probably depend on oxidant variations. As shown in Fig. S10, the important sources of C_2 in the C2w case are the oxidation of hydrated ωC_2 by the OH radical and sulfur containing radicals (SO_x*). A reduced concentration ratio of the OH or SO_x* is calculated in case of a 20 % decrease of emission data. After the calculation, a correlation with RIR values has been performed. Fig. 9 shows that OH and SO_x* radicals have high and moderate correlations with C_2 -RIRs in C_2 case, respectively, suggesting that the concentration variations of OH and SO_x* radicals are the reason for the positive and negative C_2 -RIRs in the C_2 case.

In the C2wo case, alkenes account for the highest RIR. The RIR of alkenes is more than two times higher than that of the second highest group (aromatic compounds). Among the alkenes, the dominant compound is isoprene. Contrary to the C2w case, 1,3-butadiene reveals very low RIR under C2wo condition. In the C2wo case, ethene exhibits a positive but low RIR. Among aromatic compounds, toluene shows the highest RIR. Xylene, ethyl benzene and isopropyl benzene also presents

significantly positive RIR values in the C2wo case. Alkanes have negative RIRs again. As shown in Fig. S10, in the C2wo case, the reactions of dissolved ω C2 with OH radicals represent the major pathways. Strong correlations between C2-RIRs and OH radical in the C2wo case (Fig. 9) indicate that the calculated positive and negative RIRs in the C2wo case are due to OH radical concentration variations.

For Pyr, in both C2w and C2wo cases, alkenes are the dominant precursor group with the largest RIRs. The major compound is isoprene. The absolute RIR values for other selected species are less than 0.05. These results indicated that Pyr formation during atmospheric transport is highly sensitive to isoprene.

In the C2w case, aromatic compounds are the most significant precursors of ω C2 with high positive RIR. However, individual aromatic species listed in Fig. 8 are characterized by negative RIRs. Alkanes and alkenes show negative values.

- 630 However, isoprene and 1,3-butadiene have positive RIRs, and their high levels suggests that they are key species controlling ωC₂ formation during the modeled summer conditions. As shown in Fig. S10, oxidations of dissolved Gly by OH and SO_x⁻ radicals are the most important sources for ωC₂ formation in the C2w case. High correlations between OH and SO_x⁻ radicals with ωC₂-RIR values (Fig. 9) suggest that positive and negative ωC₂-RIRs in the C2w case are a result of variations of the two oxidants
- 635 Figure 8 shows that aromatic compounds account for the highest RIR under C2wo condition, and toluene is a major contributor. Ethyl benzene and isopropyl benzene also made significant contributions. The alkene RIRs are the next highest. Isoprene is the most abundant compound during ωC₂ formation. As in cases of C₂ and Pyr, alkanes also have negative ωC₂-RIRs. ωC₂ production mainly depends on the oxidation of dissolved Gly by OH radical in the C2wo case, and variation trend of OH radical are the reason for the positive and negative ωC₂-RIRs (see Fig. S10 and Fig. 9).

640 3.5.3 Identifying the impact of increased Gly aerosol partitioning

Phase partitioning between gas and aqueous phase in a multiphase model can be affected, e.g. by salting-in/salting-out effects and other reversible accretion reactions (Herrmann et al., 2015). For example, Ip et al. (2009) and Kampf et al. (2013) have reported that SO₄²- and ammonium sulfate can have a significant effect on the uptake of Gly into an aqueous solution. Therefore, a sensitivity study considering increased Gly aerosol partitioning has been done to evaluate the changes of Gly, ωC₂ and C₂ concentrations. This has been realised by increasing the CAPRAM Gly Henry's law constant (1.4 mol I⁻¹ atm⁻¹, Betterton and Hoffmann, 1988). An increased Gly Henry's law constant (raised by two orders of magnitude) have been applied, which is close to the value reported in Volkamer et al. (2009). As can be seen in Fig. 10, compared to the base C2wo case performed without an increased Gly Henry's law constant, the modeled ωC₂ and C₂ aerosol concentrations increase by three and two times, respectively. This result suggested that an increased Gly aerosol partitioning might play an important role in ωC₂ and C₂ aqueous-phase formation.

Gelöscht: As for

4 Conclusions

The present study focuses on the formation processes of secondary aerosols constituents along trajectories to Mt. Tai using the multiphase chemistry air parcel model SPACCIM. The modeled radical and non-radical concentrations (such as gasphase OH concentration: 3.2×10^6 molecules cm⁻³ in C2w and 3.5×10^6 molecules cm⁻³ in C2wo) suggest that the atmospheric environment of Mt. Tai (~ 1.5 km a.m.s.l.) is still characterized by a sub-urban oxidants budget at the altitude of about 1.5 km. Compared to previous studies at Mt. Tai, this study is the first that investigates the formation processes of secondary aerosols constituents along different trajectories to Mt. Tai under day vs. night and cloud vs. non-cloud conditions in detail. The agueous reaction of HSO₃ with H₂O₂ has been identified as the major contributor to SO₄² formation (contribution: 67 % in C2w, 60 % in C2wo). NO₃ formation is higher during the night than during the day. The major pathways are aqueous-phase N₂O₅ hydrolysis (contribution: 72 % in C2w, 70 % in C2wo) and gas-phase reaction of OH + NO₂ (contribution: 18 % in C2w, 21 % in C2wo). Aqueous-phase reactions of hydrated Gly, hydrated ωC₂, nitro 2oxopropanoate and hydrated 3-oxopropanoic acid are dominant formation pathways of ωC₂, C₂, Pyr and C₃, respectively. Sensitivity tests indicate isoprene, 1,3-butadiene and toluene as key precursors of ωC₂ and C₂. The model data analyses show that isoprene is the predominant precursor for Pvr. When emissions are doubled, the modeled SOA compound concentrations increase about two times, suggesting that gaseous VOC emissions are a driving factor for the modeled SOA compound concentrations. The results indicate the importance of further emission reduction efforts to gain a better air quality in this part of China.

Additionally, the simulations show that an increased Gly aerosol partitioning plays an important role in ωC₂ and C₂ aqueousformation. Finally, the present study reveals that, in order to better understand the presence, formation, chemical fate
and phase partitioning of DCRCs in the troposphere in the future, comprehensive aerosol and cloud field studies, advanced
mechanistic laboratory studies and more chemical processes model studies are necessary. In case of field investigations,
advanced measurements characterizing chemical gas and aerosol compositions with a high time resolution are needed to
enable better comparison with and evaluations of present multiphase models. On the other hand, for future model
comparisons also more advanced models are required. Those should include a more detailed chemistry description, a detailed
treatment of non-ideal solution effects, and an improved treatment of the phase-partitioning of organic compounds, e.g.
considering salting-in/salting-out effects and other reversible accretion reactions. Thus, the obeserved differences between
modelled data and measurements could be fixed.

Data availability

580 The data used in this study are available from the corresponding author upon request (email: herrmann@tropos.de, xuelikun@sdu.edu.cn).

Author contributions

YHZ, AT and HH designed the SPACCIM modeling work. YHZ, AT and EHH performed the different SPACCIM simulations. YHZ, AT, EHH and HH analysed the SPACCIM simulation results. YHZ and LKX performed and interpreted the RIR analysis. YHZ, AT, KK, LXY and WXW compared the model results with field data. YHZ, AT, EHH, HH and LKX wrote the paper, and prepared the manuscript material with contributions from all the co-authors.

Competing interests

The authors declare that they have no conflict of interest.

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References

- An, J., Zou, J., Wang, J., Lin, X., and Zhu, B.: Differences in ozone photochemical characteristics between the megacity
 Nanjing and its suburban surroundings, Yangtze River Delta, China, Environ. Sci. Pollut. Res., 22, 19607-19617, https://doi.org/10.1007/s11356-015-5177-0, 2015.
 - An, J. L., Zhang, W., and Qu, Y.: Impacts of a strong cold front on concentrations of HONO, HCHO, O3, and NO2 in the heavy traffic urban area of Beijing, Atmos. Environ., 43, 3454–3459, https://doi.org/10.1016/j.atmosenv.2009.04.052, 2009.

 Arakaki, T., Anastasio, C., Kuroki, Y., Nakajima, H., Okada, K., Kotani, Y., Handa, D., Azechi, S., Kimura, T., Tsuhako, A.,
 - and Miyagi, Y.: A General Scavenging Rate Constant for Reaction of Hydroxyl Radical with Organic Carbon in Atmospheric Waters, Environ. Sci. Technol., 47, 8196-8203, https://doi.org/10.1021/es401927b, 2013.
 - Barletta, B., Meinardi, S., Rowland, F. S., Chan, C. Y., Wang, X. M., Zou, S. C., Chan, L. Y., and Blake, D. R.: Volatile organic compounds in 43 Chinese cities, Atmos. Environ., 39, 5979–5990, https://doi.org/10.1016/j.atmosenv.2005.06.029, 2005.

- 710 Bauwens, M., Stavrakou, T., Müller, J., De Smedt., I., Van Roozendael, M., van der Werf, G., Wiedinmyer, C., Kaiser, J., Sindelarova, K., and Guenther, A.: Nine years of global hydrocarbon emissions based on source inversion of OMI formaldehyde observations, Atmos. Chem. Phys., 16, 10133-10158, https://doi.org/10.5194/acp-16-10133-2016, 2016.
 Betterton, E. A., and Hoffmann, M. R.: Henry's law constants of some environmentally important aldehydes, Environ. Sci.
- 715 Bikkina, S., Kawamura, K., Miyazaki, Y., and Fu, P.: High abundances of oxalic, azelaic, and glyoxylic acids and methylglyoxal in the open ocean with high biological activity: Implication for secondary OA formation from isoprene, Geophys. Res. Lett., 41, 3649-3657, https://doi.org/10.1002/2014GL059913, 2014.
 - Bräuer, P., Mouchel-Vallon, C., Tilgner, A., Mutzel, A., Böge, O., Rodigast, M., Poulain, L., van Pinxteren, D., Wolke, R., and Aumont, B.: Development of a protocol for the auto-generation of explicit aqueous-phase oxidation schemes of organic
- 720 compound, Atmos. Chem. Phys. Discuss., https://doi.org/10.5194/acp-2018-1318, 2019.
 Carlton, A., Wiedinmyer, C., and Kroll, J.: A review of Secondary Organic Aerosol (SOA) formation from isoprene, Atmos. Chem. Phys., 9, 4987-5005, https://doi.org/10.5194/acp-9-4987-2009, 2009.
 - Carlton, A. G., Turpin, B. J., Altieri, K. E., Seitzinger, S., Reff, A., Lim, H. J., and Ervens, B.: Atmospheric oxalic acid and SOA production from glyoxal: Results of aqueous photooxidation experiments, Atmos. Environ., 41, 7588-7602,
- 725 https://doi.org/10.1016/j.atmosenv.2007.05.035, 2007.

Technol., 22, 1415-1418, https://doi.org/10.1021/es00177a004, 1988.

- Carlton, A. G., Turpin, B. J., Lim, H. J., Altieri, K. E., and Seitzinger, S.: Link between isoprene and secondary organic aerosol (SOA): Pyruvic acid oxidation yields low volatility organic acids in clouds, Geophys. Res. Lett., 33, L06822, https://doi.org/10.1029/2005GL025374, 2006.
- Carter, W. P., and Atkinson, R.: Computer modeling study of incremental hydrocarbon reactivity, Environ. Sci. Technol., 23,
- 730 864-880, https://doi.org/10.1021/es00065a017, 1989.
 - De Gouw, J., and Jimenez, J. L.: Organic aerosols in the Earth's atmosphere, Environ. Sci. Technol., 43(20), 7614-7618, https://doi.org/10.1021/es9006004, 2009.
 - DeCarlo, P., Ulbrich, I., Crounse, J., Foy, B. d., Dunlea, E., Aiken, A., Knapp, D., Weinheimer, A., Campos, T., and Wennberg, P.: Investigation of the sources and processing of organic aerosol over the Central Mexican Plateau from aircraft
- 735 measurements during MILAGRO, Atmos. Chem. Phys., 10, 5257-5280, https://doi.org/10.5194/acp-10-5257-2010, 2010.
 Draxler, R. and Rolph, G.: HYSPLIT (HYbrid Single-Particle Lagrangian Integrated Trajectory) model access via NOAA
 ARL READY, Silver Spring, Md: NOAA Air Resources Laboratory, available at: http://www.arl.noaa.gov/ready/hysplit4.html, 2003.
 - Duan, F., He, K., Ma, Y., Yang, F., Yu, X., Cadle, S., Chan, T., and Mulawa, P.: Concentration and chemical characteristics
- 740 of PM2.5 in Beijing, China: 2001–2002, Sci. Total Environ., 355, 264-275, https://doi.org/10.1016/j.scitotenv.2005.03.001, 2006.

- Duan, J. C., Tan, J. H., Yang, L., Wu, S., Hao, J. M.: Concentration, sources and ozone formation potential of volatile organic compounds (VOCs) during ozone episode in Beijing, Atmos. Res., 88, 25–35, https://doi.org/10.1016/j.atmosres.2007.09.004, 2008.
- 745 Ervens, B.: Modeling the processing of aerosol and trace gases in clouds and fogs, Chem. Rev., 115, 4157-4198, https://doi.org/10.1021/cr5005887, 2015.
 - Ervens, B., and Volkamer, R.: Glyoxal processing by aerosol multiphase chemistry: towards a kinetic modeling framework of secondary organic aerosol formation in aqueous particles, Atmos. Chem. Phys., 10, 8219-8244, https://doi.org/10.5194/acp-10-8219-2010, 2010.
- 750 Ervens, B., Carlton, A. G., Turpin, B. J., Altieri, K. E., Kreidenweis, S. M., and Feingold, G.: Secondary organic aerosol yields from cloud-processing of isoprene oxidation products, Geophys. Res. Lett., 35, L02816, https://doi.org/10.1029/2007GL031828, 2008.
 - Ervens, B., Feingold, G., Frost, G. J., and Kreidenweis, S. M.: A modeling study of aqueous production of dicarboxylic acids: 1. Chemical pathways and speciated organic mass production, J. Geophys. Res.-Atmos., 109, D15205,
- 755 https://doi.org/10.1029/2003JD004387, 2004.
 - Ervens, B., George, C., Williams, J., Buxton, G., Salmon, G., Bydder, M., Wilkinson, F., Dentener, F., Mirabel, P., and Wolke, R.: CAPRAM2.4 (MODAC mechanism): An extended and condensed tropospheric aqueous phase mechanism and its application, J. Geophys. Res.-Atmos., 108, D14, https://doi.org/10.1029/2002JD002202, 2003.
- Ervens, B., Turpin, B., and Weber, R.: Secondary organic aerosol formation in cloud droplets and aqueous particles 760 (aqSOA): a review of laboratory, field and model studies, Atmos. Chem. Phys., 11, 11069-11102, https://doi.org/10.5194/acp-11-11069-2011, 2011.
 - Ganzeveld, L., Lelieveld, J., and Roelofs, G.: A dry deposition parameterization for sulfur oxides in a chemistry and general circulation model, J. Geophys. Res.-Atmos., 103, 5679-5694, https://doi.org/10.1029/97JD03077, 1998.
 - Glasius, M., Lahaniati, M., Calogirou, A., Di Bella, D., Jensen, N. R., Hjorth, J., Kotzias, D., and Larsen, B. R.: Carboxylic acids in secondary aerosols from oxidation of cyclic monoterpenes by ozone, Environ. Sci. Technol., 34, 1001-1010, https://doi.org/10.1021/es990445r, 2000.
 - Guo, J., Wang, Y., Shen, X. H., Wang, Z., Lee, T., Wang, X. F., Li, P. H., Sun, M. H., Collett Jr., J. L., Wang, W. X., and Wang, T.: Characterization of cloud water chemistry at Mount Tai, China: Seasonal variation, anthropogenic impact, and cloud processing, Atmos. Environ., 60, 467-476, https://doi.org/10.1016/j.atmosenv.2012.07.016, 2012.
- 770 Guo, S., Hu, M., Wang, Z., Slanina, J., and Zhao, Y.: Size-resolved aerosol water-soluble ionic compositions in the summer of Beijing: implication of regional secondary formation, Atmos. Chem. Phys., 10, 947-959, https://doi.org/10.5194/acp-10-947-2010, 2010.
 - Herrmann, H., Tilgner, A., Barzaghi, P., Majdik, Z., Gligorovski, S., Poulain, L., Monod, A.: Towards a more detailed description of tropospheric aqueous phase organic chemistry: CAPRAM 3.0, Atmos. Environ., 39, 4351–4363,
- 775 https://doi.org/10.1016/j.atmosenv.2005.02.016, 2005.

- He, N., and Kawamura, K.: Distributions and diurnal changes of low molecular weight organic acids and α -dicarbonyls in suburban aerosols collected at Mangshan, North China, Geochem. J., 44, 17-22, https://doi.org/10.2343/geochemj.1.0093, 2010.
- He, N., Kawamura, K., Okuzawa, K., Kanaya, Y., and Wang, Z.: Diurnal variations of total carbon, dicarboxylic acids, ketoacids and α-dicarbonyls in aerosols in the northern vicinity of Beijing, Atmos. Chem. Phys. Discuss., 13, 16699-16731, https://doi.org/10.5194/acpd-13-16699-2013, 2013.
 - He, S. Z., Chen, Z. M., Zhao, Y., Huang, D. M., Zhao, J. N., Zhu, T., Hu, M., and Zeng, L. M.: Measurement of atmospheric hydrogen peroxide and organic peroxides in Beijing before and during the 2008 Olympic Games: Chemical and physical factors influencing their concentrations, J. Geophys. Res., 115, https://doi.org/10.1029/2009JD013544, 2010.
- 785 Hegde, P., and Kawamura, K.: Seasonal variations of water-soluble organic carbon, dicarboxylic acids, ketocarboxylic acids, and α-dicarbonyls in Central Himalayan aerosols, Atmos. Chem. Phys., 12, 6645-6665, https://doi.org/10.5194/acp-12-6645-2012, 2012.
 - Herrmann, H., Hoffmann, D., Schaefer, T., Bräuer, P., and Tilgner, A.: Tropospheric aqueous-phase free-radical chemistry: Radical sources, spectra, reaction kinetics and prediction tools, Chem. Phys. Chem., 11, 3796-3822, https://doi.org/10.1002/cphc.201000533, 2010.
 - Herrmann, H., Schaefer, T., Tilgner, A., Styler, S. A., Weller, C., Teich, M., and Otto, T.: Tropospheric aqueous-phase chemistry: kinetics, mechanisms, and its coupling to a changing gas phase, Chem. Rev., 115, 4259-4334, https://doi.org/10.1021/cr500447k, 2015.
- Ho, K. F., Huang, R. J., Kawamura, K., Tachibana, E., Lee, S. C., Ho, S. S. H., Zhu, T., and Tian, L.: Dicarboxylic acids, ketocarboxylic acids, a-dicarbonyls, fatty acids and benzoic acid in PM2.5 aerosol collected during CAREBeijing-2007: an effect of traffic restriction on air quality, Atmos. Chem. Phys., 15, 3111–3123, https://doi.org/10.5194/acp-15-3111-2015, 2015.
- Hoffmann, E. H., Tilgner, A., Wolke, R., Böge, O., Walter, A., and Herrmann, H.: Oxidation of substituted aromatic hydrocarbons in the tropospheric aqueous phase: kinetic mechanism development and modelling, Phys. Chem. Chem. Phys.,
 20, 10960-10977, https://doi.org/10.1039/C7CP08576A, 2018.
 - Hu, G. Y., Sun, J. Y., Zhang, Y. M., Shen, X. J., and Yang, Y.: Chemical Composition of PM2.5 Based on Two-Year Measurements at an Urban Site in Beijing, Aerosol and Air Quality Research, 15: 1748–1759, https://doi.org/10.4209/aaqr.2014.11.0284, 2015.
- Huang, R. J., Zhang, Y., Bozzetti, C., Ho, K. F., Cao, J. J., Han, Y., Daellenbach, K. R., Slowik, J. G., Platt, S. M., and Canonaco, F.: High secondary aerosol contribution to particulate pollution during haze events in China, Nature, 514, 218, https://doi.org/10.1038/nature13774, 2014.
 - Huisman, A. J., Hottle, J. R., Galloway, M. M., DiGangi, J. P., Coens, K. L., Choi, W., Faloona, I. C., Gilman, J. B., Kuster, W. C., de Gouw, J., Bouvier-Brown, N. C., Goldstein, A. H., LaFranchi, B. W., Cohen, R. C., Wolfe, G. M., Thornton, J. A., Docherty, K. S., Farmer, D. K., Cubison, M. J., Jimenez, J. L., Mao, J., Brune, W. H., and Keutsch, F. N.: Photochemical

- 810 modeling of glyoxal at a rural site: observations and analysis from BEARPEX 2007, Atmos. Chem. Phys., 11, 8883–8897, https://doi.org/10.5194/acp-11-8883-2011, 2011.
 - Ianniello, A., Spataro, F., Esposito, G., Allegrini, I., Hu, M., and Zhu, T.: Chemical characteristics of inorganic ammonium salts in PM2.5 in the atmosphere of Beijing (China), Atmos. Chem. Phys., 11, 10803–10822, https://doi.org/10.5194/acp-11-10803-2011, 2011.
- 815 Ip, H., Huang, X., and Yu, J. Z.: Effective Henry's law constants of glyoxal, glyoxylic acid, and glycolic acid, Geophys. Res. Lett., 36, L01802, https://doi.org/10.1029/2008GL036212, 2009.
 Jenkin, M., Saunders, S., Wagner, V., and Pilling, M.: Protocol for the development of the Master Chemical Mechanism, MCMv3 (Part B): tropospheric degradation of aromatic volatile organic compounds, Atmos. Chem. Phys., 3, 181-193, https://doi.org/10.5194/acp-3-181-2003, 2003.
- 820 Kampf, C. J., Waxman, E. M., Slowik, J. G., Dommen, J., Pfaffenberger, L., Praplan, A. P., Prevot, A. S., Baltensperger, U., Hoffmann, T., and Volkamer, R.: Effective Henry's law partitioning and the salting constant of glyoxal in aerosols containing sulfate, Environ. Sci. Technol., 47, 4236-4244, https://doi.org/10.1021/es400083d, 2013.
 - Kanaya, Y., Akimoto, H., Wang, Z. F., Pochanart, P., Kawamura, K., Liu, Y., Li, J., Komazaki, Y., Irie, H., Pan, X. L., Taketani, F., Yamaji, K., Tanimoto, H., Inomata, S., Kato, S., Suthawaree, J., Okuzawa, K., Wang, G., Aggarwal, S. G., Fu,
- P. Q., Wang, T., Gao, J., Wang, Y., and Zhuang, G.: Overview of the Mount Tai Experiment (MTX2006) in central East China in June 2006: studies of significant regional air pollution, Atmos. Chem. Phys., 13, 8265-8283, https://doi.org/10.5194/acp-13-8265-2013, 2013.
 - Kanaya, Y., Pochanart, P., Liu, Y., Li, J., Tanimoto, H., Kato, S., Suthawaree, J., Inomata, S., Taketani, F., Okuzawa, K., Kawamura, K., Akimoto, H., and Wang, Z. F.: Rates and regimes of photochemical ozone production over Central East
- China in June 2006: a box model analysis using comprehensive measurements of ozone precursors, Atmos. Chem. Phys., 9, 7711–7723, https://doi.org/10.5194/acpd-9-12965-2009, 2009.
 - Kawamura, K., and Sakaguchi, F.: Molecular distributions of water soluble dicarboxylic acids in marine aerosols over the Pacific Ocean including tropics, J. Geophys. Res.-Atmos., 104, 3501-3509, https://doi.org/10.1029/1998JD100041, 1999.
- Kawamura, K., and Yasui, O.: Diurnal changes in the distribution of dicarboxylic acids, ketocarboxylic acids and dicarbonyls in the urban Tokyo atmosphere, Atmos. Environ., 39, 1945-1960,
 - https://doi.org/10.1016/j.atmosenv.2004.12.014, 2005.

 Kawamura, K., Tachibana, E., Okuzawa, K., Aggarwal, S. G., Kanaya, Y., and Wang, Z. F.: High abundances of water-soluble dicarboxylic acids, ketocarboxylic acids and a-dicarbonyls in the mountaintop aerosols over the North China Plain
- Kim, H. S., Huh, J. B., Hopke, P. K., Holsen, T. M., and Yi, S. M.: Characteristics of the major chemical constituents of PM2.5 and smog events in Seoul, Korea in 2003 and 2004, Atmos. Environ., 41, 6762-6770, https://doi.org/10.1016/j.atmosenv.2007.04.060, 2007.

during wheat burning season, Atmos. Chem. Phys., 13, 8285-8302, https://doi.org/10.5194/acp-13-8285-2013, 2013.

- Kroll, J. A., Hansen, A. S., Møller, K. H., Axson, J. L., Kjaergaard, H. G., and Vaida, V.: Ultraviolet Spectroscopy of the Gas Phase Hydration of Methylglyoxal, ACS Earth Space Chem., 1, 345-352,
- 845 https://doi.org/10.1021/acsearthspacechem.7b00054, 2017.
 - Kundu, S., Kawamura, K., and Lee, M.: Seasonal variations of diacids, ketoacids, and α -dicarbonyls in aerosols at Gosan, Jeju Island, South Korea: Implications for sources, formation, and degradation during long-range transport, J. Geophys. Res.-Atmos., 115, D19307, https://doi.org/10.1029/2010JD013973, 2010.
- Legrand, M., Preunkert, S., Oliveira, T., Pio, C. A., Hammer, S., Gelencsér, A., Kasper-Giebl, A., and Laj, P.: Origin of C2-
- C5 dicarboxylic acids in the European atmosphere inferred from year-round aerosol study conducted at a west-east transect, J. Geophys. Res.-Atmos., 112, D23S07, https://doi.org/10.1029/2006JD008019, 2007.
 - Li, C. C., Mao, J. T., Alexis, K. H. L., Yuan, Z. B., Wang, M. H., and Liu, X. Y.: Application of MODIS satellite products to the air pollution research in Beijing, Sci. China Ser. D., 48. 209-219, https://doi.org/10.1360/05yd0395, 2005.
- Li, J., Xie, S. D., Zeng, L. M., Li, L. Y., Li, Y. Q., and Wu, R. R.: Characterization of ambient volatile organic compounds and their sources in Beijing, before, during, and after Asia-Pacific Economic Cooperation China 2014, Atmos. Chem. Phys., 15, 7945–7959, https://doi.org/10.5194/acp-15-7945-2015, 2015a.
 - Li, J. R., Wang, X. F., Chen, J. M., Zhu, C., Li, W. J., Li, C. B., Liu, L., Xu, C. H., Wen, L., Xue, L. K., Wang, W. X., Ding, A. J., and Herrmann, H.: Chemical composition and droplet size distribution of cloud at the summit of Mount Tai, China, Atmos. Chem. Phys., 17, 9885–9896, https://doi.org/10.5194/acp-17-9885-2017, 2017a.
- Li, L. Y., Xie, S. D., Zeng, L. M., Wu, R. R., and Li, J.: Characteristics of volatile organic compounds and their role in ground-level ozone formation in the Beijing-Tianjin-Hebei region, China, Atmos. Environ., 113, 247-254, https://doi.org/10.1016/j.atmosenv.2015.05.021, 2015b.
 - Li, M., Zhang, Q., Kurokawa, J. I., Woo, J. H., He, K., Lu, Z., Ohara, T., Song, Y., Streets, D. G., and Carmichael, G. R.: MIX: a mosaic Asian anthropogenic emission inventory under the international collaboration framework of the MICS-Asia
- 865 and HTAP, Atmos. Chem. Phys., 17(2), 935-963, https://doi.org/10.5194/acp-17-935-2017, 2017b.
 - Li, Y., An, J. L., Min, M., Zhang, W., Wang, F., and Xie, P. H.: Impacts of HONO sources on the air quality in Beijing, Tianjin and Hebei Province of China, Atmos. Environ., 45, 4735-4744, https://doi.org/10.1016/j.atmosenv.2011.04.086, 2011.
- Lim, H. J., Carlton, A. G., and Turpin, B. J.: Isoprene forms secondary organic aerosol through cloud processing: model simulations, Environ. Sci. Technol., 39, 4441-4446, https://doi.org/10.1021/es048039h, 2005.
 - Liu, Y., Shao, M., Kuster, W. C., Goldan, P. D., Li, X. H., Lu, S. H., De Gouw, J. A.: Source Identification of Reactive Hydrocarbons and Oxygenated VOCs in the Summertime in Beijing, Environ. Sci. Technol., 43, 75–81, https://doi.org/10.1021/es801716n, 2009.
 - Liu, Y. J., Zhang, T. T., Liu, Q. Y., Zhang, R. J., Sun, Z. Q., and Zhang, M. G.: Seasonal Variation of Physical and Chemical
- Properties in TSP, PM10 and PM2.5 at a Roadside Site in Beijing and Their Influence on Atmospheric Visibility, Aerosol Air Qual. Res., 14, 954–969, https://doi.org/10.4209/aaqr.2013.01.0023, 2014.

- Liu, Z., Wang, Y. H., Vrekoussis, M., Richter, A., Wittrock, F., Burrows, J. P., Shao, M., Chang, C. C., Liu, S. C., Wang, H. L., Chen, C. H.: Exploring the missing source of glyoxal (CHOCHO) over China, Geophys. Res. Lett., 39, L10812, https://doi.org/10.1029/2012GL051645, 2012b.
- Eiu, Z., Wang, Y., Gu, D., Zhao, C., Huey, L., Stickel, R., Liao, J., Shao, M., Zhu, T., and Zeng, L.: Summertime photochemistry during CAREBeijing-2007: ROx budgets and O3 formation, Atmos. Chem. Phys., 12, 7737-7752, https://doi.org/10.5194/acp-12-7737-2012, 2012a.
 - Lu, K., Rohrer, F., Holland, F., Fuchs, H., Bohn, B., Brauers, T., Chang, C., Häseler, R., Hu, M., and Kita, K.: Observation and modelling of OH and HO2 concentrations in the Pearl River Delta 2006: a missing OH source in a VOC rich
- 885 atmosphere, Atmos. Chem. Phys., 12, 1541-1569, https://doi.org/10.5194/acp-12-1541-2012, 2012.
 Matsui, H., Koike, M., Kondo, Y., Takegawa, N., Kita, K., Miyazaki, Y., Hu, M., Chang, S. Y., Blake, D., and Fast, J. D.:
 Spatial and temporal variations of aerosols around Beijing in summer 2006: Model evaluation and source apportionment, J.
 Geophys. Res.-Atmos., 114, D00G13, https://doi.org/10.1029/2008JD010906, 2009.
 - Meng, J. J., Wang, G. H., Hou, Z. F., Liu, X. D., Wei, B. J., Wu, C., Cao, C., Wang, J. Y., Li, J. J., Cao, J. J., Zhang, E. X.,
- Boong, J., Liu, J. Z., Ge, S. S., and Xie, Y. N.: Molecular distribution and stable carbon isotopic compositions of dicarboxylic acids and related SOA from biogenic sources in the summertime atmosphere of Mt. Tai in the North China Plain, Atmos. Chem. Phys., 18, 15069–15086, https://doi.org/10.5194/acp-18-15069-2018, 2018.
 - Meng, Z. Y., Lin, W. L., Jiang, X. M., Yan, P., Wang, Y., Zhang, Y. M., Jia, X. F., and Yu, X. L.: Characteristics of atmospheric ammonia over Beijing, China, Atmos. Chem. Phys., 11, 6139–6151, https://doi.org/10.5194/acp-11-6139-2011, 2011.

- Miyazaki, Y., Aggarwal, S. G., Singh, K., Gupta, P. K., and Kawamura, K.: Dicarboxylic acids and water-soluble organic carbon in aerosols in New Delhi, India, in winter: Characteristics and formation processes, J. Geophys. Res.-Atmos., 114, D19206, https://doi.org/10.1029/2009JD011790, 2009.
- Mouchel-Vallon, C., Deguillaume, L., Monod, A., Perroux, H., Rose, C., Ghigo, G., Long, Y., Leriche, M., Aumont, B., Patryl, L., Armand, P., and Chaumerliac, N.: CLEPS 1.0: A new protocol for cloud aqueous phase oxidation of VOC mechanisms, Geosci. Model Dev., 10(3), 1339–1362, https://doi.org/10.5194/gmd-10-1339-2017, 2017.
 - Ng, N. L., Brown, S. S., Archibald, A. T., Atlas, E., Cohen, R. C., Crowley, J. N., Day, D. A., Donahue, N. M., Fry, J. L., and Fuchs, H.: Nitrate radicals and biogenic volatile organic compounds: oxidation, mechanisms, and organic aerosol, Atmos. Chem. Phys., 17, 2103-2162, https://doi.org/10.5194/acp-17-2103-2017, 2017.
- Pavuluri, C. M., and Kawamura, K.: Evidence for 13-carbon enrichment in oxalic acid via iron catalyzed photolysis in aqueous phase, Geophys. Res. Lett., 39, L03802, https://doi.org/10.1029/2011GL050398, 2012.
 - Pavuluri, C. M., Kawamura, K., and Swaminathan, T.: Water-soluble organic carbon, dicarboxylic acids, ketoacids, and α-dicarbonyls in the tropical Indian aerosols, J. Geophys. Res.-Atmos., 115, D11302, https://doi.org/10.1029/2009JD012661, 2010.

- 910 Rao, Z. H., Chen, Z. M., Liang, H., Huang, L. B., Huang, D.: Carbonyl compounds over urban Beijing: Concentrations on haze and non-haze days and effects on radical chemistry, Atmos. Environ., 124, 207-216, https://doi.org/10.1016/j.atmosenv.2015.06.050, 2016.
- Ren, Y., Ding, A. J., Wang, T., Shen, X. H., Guo, J., Zhang, J. M., Wang, Y., Xu, P. J., Wang, X. F., Gao, J., and Collett, Jr., Jeffrey L.: Measurement of gas-phase total peroxides at the summit of Mount Tai in China, Atmos. Environ., 43, 1702-1711, https://doi.org/10.1016/j.atmosenv.2008.12.020, 2009.
 - Rusumdar, A. J., Tilgner, A., Wolke, R., and Herrmann, H.: Treatment of non-ideality in the multiphase model SPACCIM-Part 2: Impacts on multiphase chemical processing in deliquesced aerosol particles, will be submitted to Atmos. Chem. Phys., 2020.
- Saunders, S. M., Jenkin, M. E., Derwent, R., and Pilling, M.: Protocol for the development of the Master Chemical
- 920 Mechanism, MCMv3 (Part A): tropospheric degradation of non-aromatic volatile organic compounds, Atmos. Chem. Phys., 3, 161-180, https://doi.org/10.5194/acp-3-161-2003, 2003.
 - Schwartz, S. E.: Mass-transport considerations pertinent to aqueous phase reactions of gases in liquid-water clouds, in: Chemistry of multiphase atmospheric systems, Springer, 415-471, https://doi.org/10.1007/978-3-642-70627-1 16, 1986.
 - Seagrave, J., McDonald, J. D., Bedrick, E., Edgerton, E. S., Gigliotti, A. P., Jansen, J. J., Ke, L., Naeher, L. P., Seilkop, S.
- 925 K., and Zheng, M.: Lung toxicity of ambient particulate matter from southeastern US sites with different contributing sources: relationships between composition and effects, Environ. Health Persp., 114(9), 1387-1393, https://doi.org/10.1289/ehp.9234, 2006.

- Sehili, A., Wolke, R., Knoth, O., Simmel, M., Tilgner, A., and Herrmann, H.: Comparison of different model approaches for the simulation of multiphase processes, Atmos. Environ., 39, 4403-4417, https://doi.org/10.1016/j.atmosenv.2005.02.039,
- Shen, X. H., Lee, T., Guo, J., Wang, X. F., Li, P. H., Xu, P. J., Wang, Y., Ren, Y., Wang, W. X., Wang, T., Li, Y., Carn, S. A., and Collett Jr., J. L.: Aqueous phase sulfate production in clouds in eastern China, Atmos. Environ., 62, 502-511, https://doi.org/10.1016/j.atmosenv.2012.07.079, 2012.
- Shrivastava, M., Cappa, C. D., Fan, J. W., Goldstein, A. H., Guenther, A. B., Jimenez, J. L., Kuang, C., Laskin, A., Martin, S. T., Ng, N. L., Petaja, T., Pierce, J. R., Rasch, P. J., Roldin, P., Seinfeld, J. H., Shilling, J., Smith, J. N., Thornton, J. A., Volkamer, R., Wang, J., Worsnop, D. R., Zaveri, R. A., Zelenyuk, A., and Zhang, Q.: Recent advances in understanding secondary organic aerosol: Implications for global climate forcing, Rev. Geophys., 55, 509-559, https://doi.org/10.1002/2016RG000540, 2017.
- Simmel, M., and Wurzler, S.: Condensation and activation in sectional cloud microphysical models, Atmos. Res., 80 (2–3), 940 218–236, https://doi.org/10.1016/j.atmosres.2005.08.002, 2006.
 - Simmel, M., Diehl, K., and Wurzler, S.: Numerical simulation of the microphysics of an orographic cloud: comparison with measurements and sensitivity studies, Atmos. Environ., 39, 4365–4373, https://doi.org/10.1016/j.atmosenv.2005.02.017, 2005.

- Sindelarova, K., Granier, C., Bouarar, I., Guenther, A., Tilmes, S., Stavrakou, T., Müller, J., Kuhn, U., Stefani, P., and
- 945 Knorr., W.: Global data set of biogenic VOC emissions calculated by the MEGAN model over the last 30 years, Atmos. Chem. Phys., 14, 9317-9341, https://doi.org/10.5194/acp-14-9317-2014, 2014.
 - Sorooshian, A., Varutbangkul, V., Brechtel, F. J., Ervens, B., Feingold, G., Bahreini, R., Murphy, S. M., Holloway, J. S.,
 - Atlas, E. L., and Buzorius, G.: Oxalic acid in clear and cloudy atmospheres: Analysis of data from International Consortium
- for Atmospheric Research on Transport and Transformation 2004, J. Geophys. Res.-Atmos., 111, D23S45,
- 950 https://doi.org/10.1029/2005JD006880, 2006.
 - Suhail, K., George, M., Chandran, S., Varma, R., Venables, D. S., Wang, M., and Chen, J.: Open path incoherent broadband cavity-enhancedmeasurements of NO3 radical and aerosol extinction in the North China Plain, Spectrochim. Acta. A., 208, 24-31, https://doi.org/10.1016/j.saa.2018.09.023, 2019.
 - Sun, L., Xue, L., Wang, T., Gao, J., Ding, A., Cooper, O. R., Lin, M., Xu, P., Wang, Z., and Wang, X.: Significant increase
- 955 of summertime ozone at Mount Tai in Central Eastern China, Atmos. Chem. Phys., 16, 10637-10650, https://doi.org/10.5194/acp-16-10637-2016, 2016.
 - Sun, W. J., Yang, L. M., Yu, L. Y., and Saeys, M.: Ab Initio Reaction Path Analysis for the Initial Hydrogen Abstraction from Organic Acids by Hydroxyl Radicals, J. Phys. Chem. A., 113, 7852-7860, https://doi.org/10.1021/jp8090792, 2009.
- Sun, Y., Jiang, Q., Wang, Z., Fu, P., Li, J., Yang, T., and Yin, Y.: Investigation of the sources and evolution processes of severe haze pollution in Beijing in January 2013, J. Geophys. Res.-Atmos., 119, 4380-4398, https://doi.org/10.1002/2014JD021641, 2014.
 - Sun, Y. L., Wang, Z. F., Du, W., Zhang, Q., Wang, Q. Q., Fu, P. Q., Pan, X. L., Li, J., Jayne, J., and Worsnop, D. R.: Long-term real-time measurements of aerosol particle composition in Beijing, China: seasonal variations, meteorological effects, and source analysis, Atmos. Chem. Phys., 15, 10149–10165, https://doi.org/10.5194/acp-15-10149-2015, 2015.
- 965 Sun, Y. L., Wang, Z. F., Fu, P. Q., Yang, T., Jiang, Q., Dong, H. B., Li, J., and Jia, J. J.: Aerosol composition, sources and processes during wintertime in Beijing, China, Atmos. Chem. Phys., 13, 4577–4592, https://doi.org/10.5194/acp-13-4577-2013, 2013.
 - Tabazadeh, A.: Organic aggregate formation in aerosols and its impact on the physicochemical properties of atmospheric particles, Atmos. Environ., 39, 5472-5480, https://doi.org/10.1016/j.atmosenv.2005.05.045, 2005.
- 970 Tilgner, A., and Herrmann, H.: Radical-driven carbonyl-to-acid conversion and acid degradation in tropospheric aqueous systems studied by CAPRAM, Atmos. Environ., 44, 5415-5422, https://doi.org/10.1016/j.atmosenv.2010.07.050, 2010.
 - Tilgner, A., and Herrmann, H.: Tropospheric Aqueous-Phase OH Oxidation Chemistry: Current Understanding, Uptake of Highly Oxidized Organics and Its Effects, Multiphase Environ. Chem. in the Atmos., 1299, 49-85, https://doi.org/10.1021/bk-2018-1299.ch004, 2018.
- 975 Tilgner, A., Bräuer, P., Wolke, R., and Herrmann, H.: Modelling multiphase chemistry in deliquescent aerosols and clouds using CAPRAM3.0i, J. Atmos. Chem., 70, 221-256, https://doi.org/10.1007/s10874-013-9267-4, 2013.

- Volkamer, R., Jimenez, J. L., San Martini, F., Dzepina, K., Zhang, Q., Salcedo, D., Molina, L. T., Worsnop, D. R., and Molina, M. J.: Secondary organic aerosol formation from anthropogenic air pollution: Rapid and higher than expected, Geophys. Res. Lett., 33, https://doi.org/10.1029/2006GL026899, 2006.
- 980 Volkamer, R., Ziemann, P., and Molina, M.: Secondary organic aerosol formation from acetylene (C2H2): seed effect on SOA yields due to organic photochemistry in the aerosol aqueous phase, Atmos. Chem. Phys., 9, 1907-1928, https://doi.org/10.5194/acp-9-1907-2009, 2009.
 - Wang, G., Kawamura, K., Umemoto, N., Xie, M. J., Hu, S. Y., and Wang, Z. F.: Water-soluble organic compounds in PM2.5 and size-segregated aerosols over Mount Tai in North China Plain, J. Geophys. Res.-Atmos., 114, D19208, https://doi.org/10.1029/2008JD011390, 2009.
 - Wang, G., Zhang, R., Gomez, M. E., Yang, L., Zamora, M. L., Hu, M., Lin, Y., Peng, J., Guo, S., and Meng, J.: Persistent sulfate formation from London Fog to Chinese haze, PNAS, 113, 13630-13635, https://doi.org/10.1073/pnas.1616540113, 2016a.
 - Wang, H. B., Tian, M., Li, X. H., Chang, Q., Cao, J. J., Yang, F. M., Ma, Y. L., and He, K. B.: Chemical Composition and Light Extinction Contribution of PM2.5 in Urban Beijing for a 1-Year Period, Aerosol Air Qual. Res., 15, 2200–2211, https://doi.org/10.4209/aaqr.2015.04.0257, 2015.
 - Wang, S., Shi, C., Zhou, B., Zhao, H., Wang, Z., Yang, S., and Chen, L.: Observation of NO3 radicals over Shanghai, China, Atmos. Environ., 70, 401-409, https://doi.org/10.1016/j.atmosenv.2013.01.022, 2013a.
- Wang, Y., Chen, Z., Wu, Q., Liang, H., Huang, L., Li, H., Lu, K., Wu, Y., Dong, H., and Zeng, L.: Observation of atmospheric peroxides during Wangdu campaign 2014 at a rural site in the North China plain, Atmos. Chem. Phys., 16, 10985-11000, https://doi.org/10.5194/acp-16-10985-2016, 2016b.
 - Wang, Y., Shen, L., Wu, S., Mickley, L., He, J., and Hao, J.: Sensitivity of surface ozone over China to 2000-2050 global changes of climate and emissions, Atmos. Environ., 75, 374-382, https://doi.org/10.1016/j.atmosenv.2013.04.045, 2013b.
- Wang, Y., Yao, L., Wang, L., Liu, Z., Ji, D., Tang, G., Zhang, J., Sun, Y., Hu, B., and Xin, J.: Mechanism for the formation of the January 2013 heavy haze pollution episode over central and eastern China, Sci. China Earth Sci., 57, 14-25, https://doi.org/10.1007/s11430-013-4773-4, 2014a.
 - Wang, Y., Zhuang, G., Sun, Y., and An, Z.: The variation of characteristics and formation mechanisms of aerosols in dust, haze, and clear days in Beijing, Atmos. Environ., 40, 6579-6591, https://doi.org/10.1016/j.atmosenv.2006.05.066, 2006.
- Warneck, P.: In-cloud chemistry opens pathway to the formation of oxalic acid in the marine atmosphere, Atmos. Environ.,
- 1005 37, 2423-2427, https://doi.org/10.1016/S1352-2310(03)00136-5, 2003.
 - Wang, Y. H., Hu, B., Ji, D. S., Liu, Z. R., Tang, G. Q., Xin, J. Y., Zhang, H. X., Song, T., Wang, L. L., Gao, W. K., Wang, X. K., and Wang, Y. S.: Ozone weekend effects in the Beijing–Tianjin–Hebei metropolitan area, China, Atmos. Chem. Phys., 14, 2419–2429, https://doi.org/10.5194/acp-14-2419-2014, 2014b.

- Wen, L., Xue, L. K., Wang, X. F., Xu, C. H., Chen, T., Yang, L. X., Wang, T., Zhang, Q. Z., and Wang, W. X.: Summertime
- 1010 fine particulate nitrate pollution in the North China Plain: increasing trends, formation mechanisms and implications for control policy, Atmos. Chem. Phys., 18, 11261-11275, https://doi.org/10.5194/acp-18-11261-2018, 2018.
 - Williams, J., Reus, M. d., Krejci, R., Fischer, H., and Ström, J.: Application of the variability-size relationship to atmospheric aerosol studies: estimating aerosol lifetimes and ages, Atmos. Chem. Phys., 2, 133-145, https://doi.org/10.5194/acp-2-133-2002, 2002.
- Wolke, R., Sehili, A., Simmel, M., Knoth, O., Tilgner, A., and Herrmann, H.: SPACCIM: A parcel model with detailed microphysics and complex multiphase chemistry, Atmos. Environ., 39, 4375-4388, https://doi.org/10.1016/j.atmosenv.2005.02.038, 2005.
 - Xue, L., Wang, T., Wang, X., Blake, D. R., Gao, J., Nie, W., Gao, R., Gao, X., Xu, Z., and Ding, A.: On the use of an explicit chemical mechanism to dissect peroxy acetyl nitrate formation, Environ. Pollut., 195, 39-47,
- 1020 https://doi.org/10.1016/j.envpol.2014.08.005, 2014.
 - Yao, X., Chan, C. K., Fang, M., Cadle, S., Chan, T., Mulawa, P., He, K., and Ye, B.: The water-soluble ionic composition of PM2.5 in Shanghai and Beijing, China, Atmos. Environ., 36, 4223-4234, https://doi.org/10.1016/S1352-2310(02)00342-4, 2002
 - Yu, Q., Chen, J., Qin, W. H., Cheng, S. M., Zhang, Y. P., Ahmad, M., and Ouyang, W.: Characteristics and secondary
- formation of water-soluble organic acids in PM1, PM2.5 and PM10 in Beijing during haze episodes, Sci. Total Environ., 669, 175-184, https://doi.org/10.1016/j.scitotenv.2019.03.131, 2019.
 - Zhang, J., Chen, J., Yang, L., Sui, X., Yao, L., Zheng, L., Wen, L., Xu, C., and Wang, W.: Indoor PM2.5 and its chemical composition during a heavy haze–fog episode at Jinan, China, Atmos. Environ., 99, 641-649, https://doi.org/10.1016/j.atmosenv.2014.10.026, 2014.
- O30 Zhao, P. S., Dong, F., He, D., Zhao, X. J., Zhang, X. L., Zhang, W. Z., Yao, Q., and Liu, H. Y.: Characteristics of concentrations and chemical compositions for PM2.5 in the region of Beijing, Tianjin, and Hebei, China, Atmos. Chem. Phys., 13, 4631–4644, https://doi.org/10.5194/acp-13-4631-2013, 2013b.
 - Zhao, W., Fu, P., Yue, S., Li, L., Xie, Q., Zhu, C., Wei, L., Ren, H., Li, P., Li, W., Sun, Y., Wang, Z., Kawamura, K., and Chen, J.: Excitation-emission matrix fluorescence, molecular characterization and compound-specific stable carbon isotopic
- 1035 composition of dissolved organic matter in cloud water over Mt. Tai, Atmos. Environ., 213, 608-619, https://doi.org/10.1016/j.atmosenv.2019.06.034, 2019.
 - Zhao, W. Y., Kawamura, K., Yue, S. Y., Wei, L. F., Ren, H., Yan, Y., Kang, M. J., Li, L. J., Ren, L. J., Lai, S. C., Li, J., Sun, Y. L., Wang, Z. F., and Fu, P. Q.: Molecular distribution and compound-specific stable carbon isotopic composition of dicarboxylic acids, oxocarboxylic acids and a-dicarbonyls in PM2.5 from Beijing, China, Atmos. Chem. Phys., 18, 2749—
- 1040 2767, https://doi.org/10.5194/acp-18-2749-2018, 2018.

Zhao, X., Zhao, P., Xu, J., Meng, W., Pu, W., Dong, F., He, D., and Shi, Q.: Analysis of a winter regional haze event and its formation mechanism in the North China Plain, Atmos. Chem. Phys., 13, 5685-5696, https://doi.org/10.5194/acp-13-5685-2013, 2013a.

Zhou, Y., Wang, T., Gao, X. M., Xue, L. K., Wang, X. F., Wang, Z., Gao, J., Zhang, Q. Z., and Wang, W. X.: Continuous observations of water-soluble ions in PM2.5 at Mount Tai (1534 ma.s.l.) in central-eastern China, J. Atmos. Chem., 64, 107–127, https://doi.org/10.1007/s10874-010-9172-z, 2009.

Zhu, Y., Yang, L., Chen, J., Kawamura, K., Sato, M., Tilgner, A., van Pinxteren, D., Chen, Y., Xue, L., Wang, X., Simpson I. J., Herrmann, H., Blake D. R., and Wang, W. X.: Molecular distributions of dicarboxylic acids, oxocarboxylic acids and α -dicarbonyls in PM2.5 collected at the top of Mt. Tai, North China, during the wheat burning season of 2014, Atmos. Chem.

1050 Phys., 18, 10741-10758, https://doi.org/10.5194/acp-18-10741-2018, 2018.

Tables

Table 1. Acronyms of the performed model simulations.

Acronym	Acronym description		
C2w	Cluster 2 with cloud interaction		
C2wo	Cluster 2 without cloud interaction		
C4w	Cluster 4 with cloud interaction		
C4wo	Cluster 4 without cloud interaction		

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Table 2. Average aqueous-phase concentrations (mol l-1) of modeled radical and non-radical oxidants in different simulations.

	C2w		C2wo			
Oxidant	day	night	day	night	day	night
	cloud	cloud	particle	particle	particle	particle
OH _(aq)	9.6 × 10 ⁻¹⁴	6.1 × 10 ⁻¹⁵	1.8 × 10 ⁻¹⁴	3.0 × 10 ⁻¹⁵	3.0 × 10 ⁻¹⁴	8.9 × 10 ⁻¹⁵
HO _{2(aq)}	2.1×10^{-9}	6.2×10^{-10}	3.5×10^{-9}	1.6×10^{-9}	3.0 × 10 ⁻⁹	1.3×10^{-9}
NO _{3(aq)}	$1.0\times10^{\text{-}13}$	5.0×10^{-13}	$8.8\times10^{\text{-}14}$	$5.6\times10^{\text{-}13}$	9.4 × 10 ⁻¹⁴	$6.9\times10^{\text{-}13}$
$H_2O_{2(aq)}$	7.6×10^{-6}	1.0×10^{-5}	7.7×10^{-5}	6.1×10^{-5}	2.4 × 10 ⁻⁴	2.2×10^{-4}
$O_{3(aq)}$	1.1×10^{-9}	1.0×10^{-9}	8.7×10^{-10}	$8.5\times10^{\text{-}10}$	1.0 × 10 ⁻⁹	1.0×10^{-9}

Gelöscht: Table 3. Ratios of the average concentrations of the modeled and measured DCRC compounds in the different model trajectories at Mt. Tai. ¶

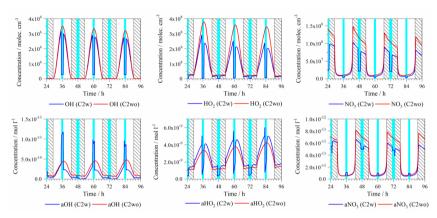
Compound ... [1]

Table 3. Ratios of the concentrations of the modeled and measured DCRC compounds in the different model trajectories at Mt.

Tai.

Compound	Model case					
Compound	C2w-day	C2w-night	C2wo-day	C2wo-night		
<u>C</u> ₂	0.30	0.27	0.23	0.21		
ωC_2	<u>7.07</u>	6.94	3.35	<u>3.43</u>		
<u>C</u> ₃	1.82	1.86	1.57	<u>1.58</u>		
<u>Pyr</u>	<u>8.95</u>	<u>7.12</u>	4.34	3.22		
Gly	0.19	<u>0.23</u>	<u>0.13</u>	<u>0.16</u>		
MGly	2.30E-3	<u>2.72E-3</u>	<u>1.35E-3</u>	<u>1.67E-3</u>		

Figures



1090 Figure 1. Time series of the modeled gas- (above: molec. cm³) and aqueous- (below: mol F¹) phase radical oxidant concentrations of the C2w and C2wo cases, respectively (light blue column: cloud; shadow: night; a: aqueous phase). For acronyms, see Table 1.

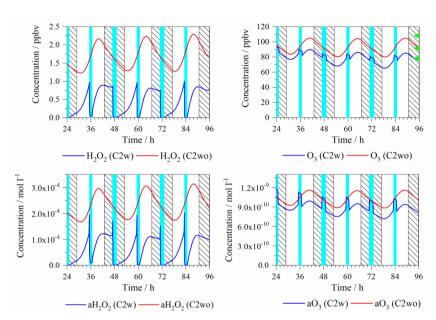


Figure 2. Time series of modeled gas- (above: "ppbv") and aqueous- (below: mol 1⁻¹) phase non-radical oxidant concentrations in the C2w and C2wo cases (light blue column: cloud; shadow: night; a: aqueous phase; green triangle: the maximum (above), average (middle) and minimum (below) value of measured concentration at Mt. Tai).

Gelöscht: molec. cm-3

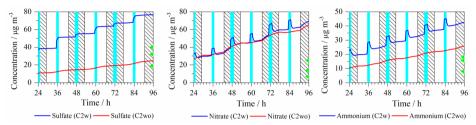


Figure 3. Time series of the modeled aerosol mass concentrations (µg m³) of key secondary inorganic aerosol constituents (sulfate, nitrate and ammonium in the C2w and C2wo cases (light blue column: cloud; shadow: night; green triangle: the maximum (above), average (middle) and minimum (below) value of measured concentration at Mt. Tai)).

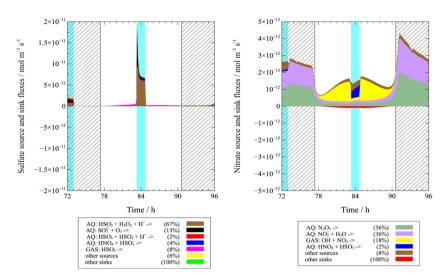


Figure 4. Modeled multiphase (gas + aqueous phase) source and sink fluxes of sulfate and nitrate (light blue column: cloud; shadow: night; percent is for the third model day).

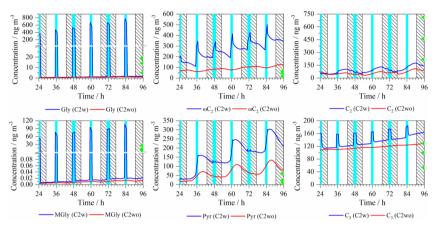


Figure 5. Time series of the modeled aerosol mass concentrations of selected DCRCs (top: Gly, ωC₂, C₂; bottom: MGly, Pyr and C₃) in the C2w and C2wo cases (light blue column: cloud; shadow: night; green triangle: the maximum (above), average (middle) and minimum (below) value of measured concentrations at Mt. Tai).

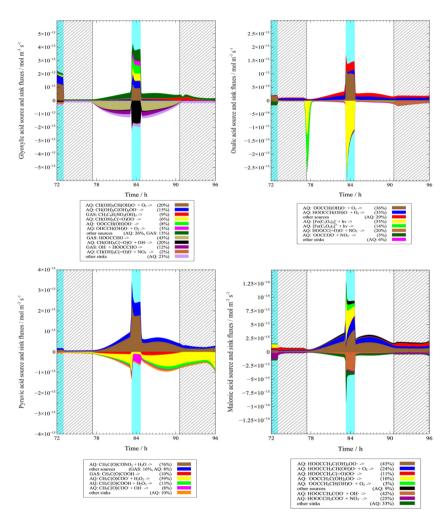


Figure 6. Modeled multiphase (gas + aqueous phase) source and sink fluxes of ωC_2 (above left), C_2 (above right), Pyr (below left) and C_3 (below right) along the trajectory of the third model day (light blue column: cloud; shadow: night; percentage is for the third model day).

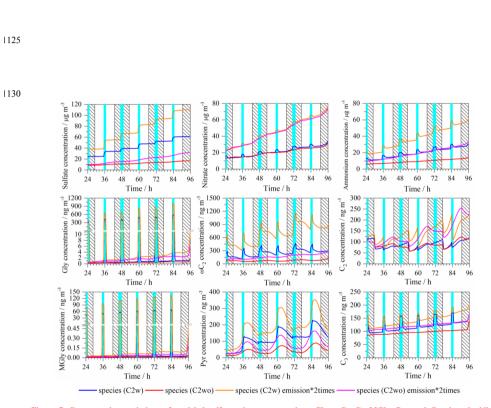
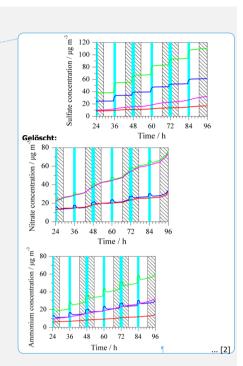


Figure 7. Concentration variations of modeled sulfate, nitrate, ammonium, Gly, ωC_2 , C_2 , MGly, Pyr and C_3 when doubling emission data (light blue column: cloud; shadow: night).



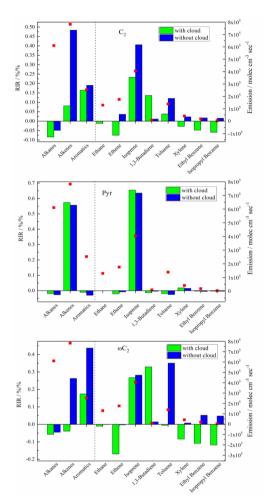


Figure 8. The calculated RIRs for C_2 , Pyr and ωC_2 in both the C2w (green bars) and C2wo (blue bars) cases at Mt. Tai (column: RIR values; red dots: emission data).

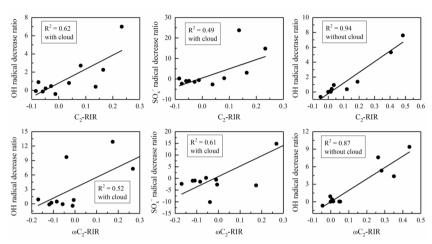


Figure 9. Correlations between the decreasing ratios of radical oxidants and C_2 -RIR (above) and ωC_2 -RIR (below) under C_2 and C_2 -RIR (below) under C_2 and C_3 -RIR (below) under C_3

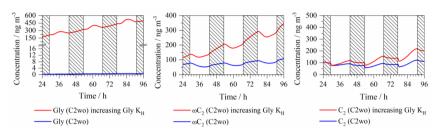


Figure 10. Concentration variations of modeled Gly, ωC_2 and C_2 when increasing Gly' Henry law constant by two orders of magnitude (shadow: night; K_H : Henry law constant).

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