



1 Simultaneous Measurement of Urban and Rural Single Particles in Beijing, Part I:  
2 Chemical Composition and Mixing State

3 Yang Chen<sup>1</sup>, Jing Cai<sup>2</sup>, Zhichao Wang<sup>1</sup>, Chao Peng<sup>1</sup>, Xiaojiang Yao<sup>1</sup>, Mi Tian<sup>1</sup>, Yiqun  
4 Han<sup>2</sup>, Guangming Shi<sup>1,3</sup>, Zongbo Shi<sup>4,5</sup>, Yue Liu<sup>2</sup>, Xi Yang<sup>2</sup>, Mei Zheng<sup>2\*</sup>, Tong Zhu<sup>2</sup>,  
5 Kebin He<sup>6</sup>, Qiang Zhang<sup>7</sup>, and Fumo Yang<sup>3,1\*</sup>

6 <sup>1</sup>Center for the Atmospheric Environment Research, Chongqing Institute of Green and  
7 Intelligent Technology, Chinese Academy of Sciences, Chongqing 400714, China

8 <sup>2</sup>SKL-ESPC and BIC-ESAT, College of Environmental Sciences and Engineering, Peking  
9 University, Beijing 100871, China

10 <sup>3</sup>College of Architecture and Environment, Sichuan University, Chengdu 610065, China

11 <sup>4</sup>School of Geography Earth and Environmental Sciences, University of Birmingham,  
12 Birmingham B15 2TT, UK

13 <sup>5</sup>Institute of Surface-Earth System Science, Tianjin University, Tianjin 300072, China

14 <sup>6</sup>School of Environment, Tsinghua University, Beijing 100084, China

15 <sup>7</sup>Department of Earth System Science, Tsinghua University, Beijing, China

16 Corresponding to Fumo Yang (fmyang@scu.edu.cn) and Mei Zheng  
17 (mzheng@pku.edu.cn)

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19 **Abstract**

20 Two single particle aerosol mass spectrometers (SPAMS) were deployed simultaneously  
21 at an urban and a rural site in Beijing during an intensive field campaign from 1<sup>st</sup> to 29<sup>th</sup>  
22 Nov 2016 to investigate the source and process of airborne particles in Beijing. In the first  
23 part of this research, we report the single-particle chemical composition, mixing state, and  
24 evolution at both sites. 96% and 98% of collected particles were carbonaceous at the urban  
25 and rural sites, respectively. Five particle categories, including elemental carbon (EC),  
26 organic carbon (OC), internal-mixed EC and OC (ECOC), potassium-rich (K-rich), and  
27 Metals were observed at both sites. The categories were partitioned into particle types  
28 depending on different atmospheric processing stages. Seventeen particle types were  
29 shared at both sites. In the urban area, nitrate-containing particle types, such as EC-Nit and  
30 ECOC-Nit, were enriched, especially at night; sulfate-containing particles were transported  
31 when wind speed was high; ECOC-Nit-Sul were mostly local-aged. In sum, these  
32 processed particles took up to 85.3% in the urban areas. In the rural area, regional particles  
33 were abundant, but freshly emitted ECOC and OC had distinct patterns that were  
34 pronounced at cooking and heating time. Biomass burning, traffic, and coal burning were  
35 major sources of PM<sub>2.5</sub> in both rural and urban areas. Besides, the particles from the steel  
36 industry located in the south were also identified. In summary, the chemical composition  
37 of urban and rural particle types was similar in Beijing; the urban particles were influenced  
38 significantly by rural processing and transport. The work is useful to understand the  
39 evolution of urban and rural particles in Beijing during winter.



## 40 **1. Introduction**

41 China has experienced severe haze events caused by extremely high concentrations of fine  
42 particulate matter (PM<sub>2.5</sub>) since January 2013. In the worst cases, an area of 2.0 million  
43 km<sup>2</sup> and a population of 800 million were affected (Huang et al., 2014). In the Beijing-  
44 Tianjin-Hebei (BTH) area, extreme haze events frequently occur during winter, with PM<sub>2.5</sub>  
45 mass reaching rapidly up to 200 µg m<sup>-3</sup> and sustaining such levels for hours (Guo et al.,  
46 2014).

47 Over the last two decades, comprehensive studies have been conducted on urban PM in  
48 Beijing. He et al. (2001) reported the first characterization of PM<sub>2.5</sub>. Since then, numerous  
49 studies have been published on characterization (Huang et al., 2010), sources (Guo et al.,  
50 2012; Sun et al., 2014), and processing of PM (Sun et al., 2013). The mechanism of rapid-  
51 boosting PM<sub>2.5</sub> in Beijing, including new particle formation and growth (Guo et al., 2014),  
52 regional transport (Li et al., 2015), and both (Du et al., 2017; Sun et al., 2014), have been  
53 proposed. However, discrepancies remain among these studies.

54 Single particle mass spectrometers (SPMS) have been used to investigate the size-resolved  
55 chemical composition and mixing state of atmospheric particles (Gard et al., 1997; Pratt  
56 and Prather, 2012). More recently, single particle aerosol mass spectrometers (SPAMS)  
57 have been used in Chinese megacities such as Beijing (Li et al., 2014), Shanghai (Tao et al.,  
58 2011), Guangzhou (Bi et al., 2011), Xi'an (Chen et al., 2016), Nanjing (Wang et al.,  
59 2015), and Chongqing (Chen et al., 2017).



60 In Beijing, particle types, such as carbonaceous, metal, dust, K-rich, and others during  
61 spring and fall, were reported (Liu et al., 2016b; Li et al., 2014). Besides, lead-containing  
62 particles have also been investigated in recent studies (Ma et al., 2016; Cai et al., 2017).  
63 However, due to the insufficient consideration of mixing state of nitrate, sulfate, and  
64 organics, these studies paid limited attention to the atmospheric particulate processing.

65 This study is a part of the APHH-Beijing (Atmospheric Pollution and Human Health in a  
66 Chinese Megacity of Beijing) intensive field campaign during winter 2016 (Shi et al., 2019).  
67 Two SPAMSs were deployed simultaneously at Peking University (PKU) and Pinggu (PG)  
68 in order to observe both urban and rural particles in the Beijing region. The aims of the  
69 study are 1) to characterize the single-particle chemical composition and mixing state; 2)  
70 to investigate particulate evolution at both sites during haze events. These two objectives  
71 are presented in two parts. In Part I, particle types and their atmospheric processing (e.g.,  
72 origination, source, and diurnal profiles) at both sites are reported; in Part II, the detailed  
73 analysis of haze events, effects of heating activities, and evidence of regional transport are  
74 addressed.

## 75 **2. Methodology**

### 76 **2.1 Sampling sites**

77 The campaigns were performed simultaneously at PKU (116.32°E, 39.99°N) and PG  
78 (117.05°E, 40.17°N) from 11/01/2016 to 11/29/2016. A Description of the PKU site is  
79 available in the literature (Huang et al., 2006). Briefly, the site is located on the rooftop (15  
80 m above the ground) on the PKU campus which is surrounded by residential and



81 commercial blocks. Trace gases (Thermo Inc. series), meteorological parameters (Vaisala  
82 Inc.), and PM<sub>2.5</sub> (TEOM 1430) were recorded during the observation.

83 The PG site (117.053°E, 40.173°N) is 3 km from the PG center. The site is located in the  
84 northeast of the PKU site with a distance of 70 km. The PG site also acts as a host of the  
85 AIRLESS (Effects of AIR pollution on the cardiopulmonary disease in urban and peri-  
86 urban residents in Beijing) Project. The meteorological data is acquired from the local  
87 meteorological office. The PG village is surrounded by orchards and farmland with no  
88 main road nearby on a scale of 3 km. Coal and biomass are used for domestic heating and  
89 cooking in the nearby villages.

## 90 **2.2 Instrumentation and data analysis**

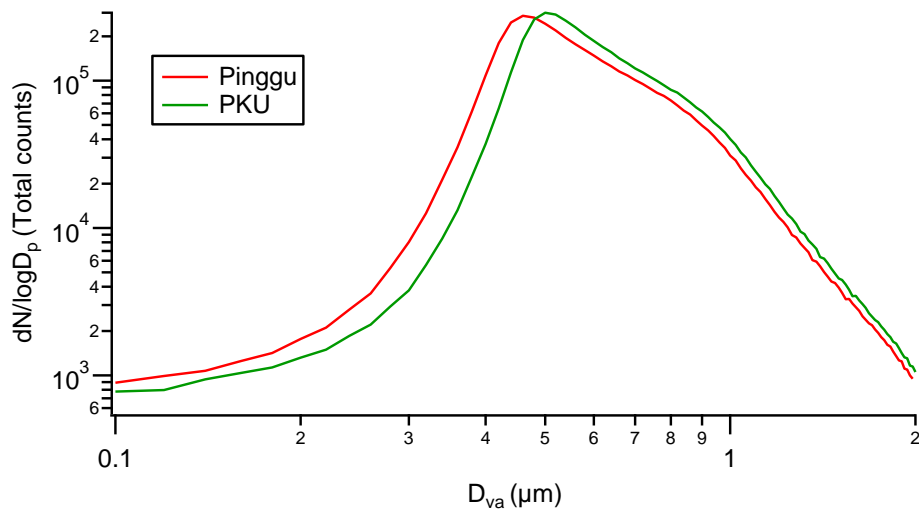
91 Two SPAMSS (Model 0515, Hexin Inc., Guangzhou, China) were deployed at both PKU  
92 and PG. A technical description of SPAMS is available in (Li et al., 2011). Briefly, a  
93 SPAMS has three functional parts: sampling, sizing, and mass spectrometry. In the  
94 sampling part, particles within a 0.1–2.0 μm size range pass efficiently through an  
95 aerodynamic lens. In the sizing unit, the aerodynamic diameter ( $D_{va}$ ) is calculated using  
96 the time-of-flight of particles. The particles are then decomposed and ionized into ions one-  
97 by-one using a 266 nm laser. A bipolar time-of-flight mass spectrometer measures the ions  
98 and generates the positive and negative mass spectra of each particle. The two instruments  
99 were maintained and calibrated following the standard procedures before sampling (Chen  
100 et al., 2017).



101 A neural network algorithm based on adaptive resonance theory (ART-2a) was used to  
102 resolve particle types from both datasets (Song et al., 1999). The parameters used were: a  
103 vigilance factor of 0.70, a learning rate of 0.05, and 20 iterations. This procedure generated  
104 771 and 792 particle groups. Then, the groups were combined into particle types based on  
105 similar mass spectra, temporal trends, and size distributions (Dalosto and Harrison, 2006).  
106 During combining, relative areas of nitrate and sulfate were used to distinguish the stages  
107 of processing, assuming that more sulfate and nitrate can be measured if a particle is more  
108 processed during its lifetime. Thus, particles with relative peak areas of sulfate and nitrate  
109 larger than 0.1 were marked with nitrate (-Nit), sulfate (-Sul), respectively, or both. Finally,  
110 the strategy resulted in 20 and 19 particle types at PKU and PG respectively. Among them,  
111 17 types appeared at both sites, and each type has identical mass spectra ( $R^2 > 0.80$ ) between  
112 each other.

### 113 **3. Results**

114 A total of 4,499,606 and 4,063,522 particles were collected at PKU and PG sites,  
115 respectively. The size distributions peaked at 0.48  $\mu\text{m}$  and 0.52  $\mu\text{m}$  (Figure 1). The smaller  
116 size distribution was due to a more substantial fraction of freshly-emitted particles at PG,  
117 as described in Table 1. Seventeen particle types ( $R^2 > 0.80$ , mass spectra) were observed  
118 both at PKU and PG (Table 1). These particle types were labeled with the suffixes “\_PKU”  
119 or “\_PG” to indicate their locations. The term “particle category” stands for a group of  
120 particle types with variable stages of processing.



121

122 Figure 1. The size distribution of SPAMS particles at PKU and PG sites.

123



124 Table 1. SPAMS particle types identified at PKU and PG sites.

At both sites	Particle type	PKU	PKU	PG	PG
		Number	Percentage	Number	Percentage
EC	EC-Nit	313574	7.0	79082	2.0
	EC-Nit-Sul	473908	10.5	140107	3.5
	EC-Sul	30365	0.7	4096	0.1
	ECOC-Nit-Sul	539533	12.0	755279	18.6
	ECOC-Sul	572548	12.7	397367	9.8
K-rich	K-rich	322731	7.2	259287	6.4
	K-Nit	359281	8.0	334547	8.2
	K-Nit-Sul	717280	16.0	76954	1.9
	K-Sul	26301	0.6	183571	4.5
NaK	NaK	16680	0.4	74943	1.8
	NaK-Nit	289259	6.4	69760	1.7
	NaK-Nit-Sul	114387	2.5	77555	1.9
	NaK-Sul	7509	0.2	16578	0.4
OC	OC-Nit-Sul	334870	7.4	865821	21.3
	OC-Sul	40800	0.9	279322	6.9
	Ca-dust	19869	0.4	3035	0.1
Metal	Fe-rich	137600	3.1	70920	1.8
	ECOC-Nit	137470	3.1%		
	OC-Nit	41159	0.9%		
	K-Amine-Nit-Sul	4482	0.1%		
	ECOC			239953	5.9%
	OC			135345	3.3%

125 Note: Nit stands for nitrate, Sul for sulfate.

### 126 3.1 Meteorological conditions and overview

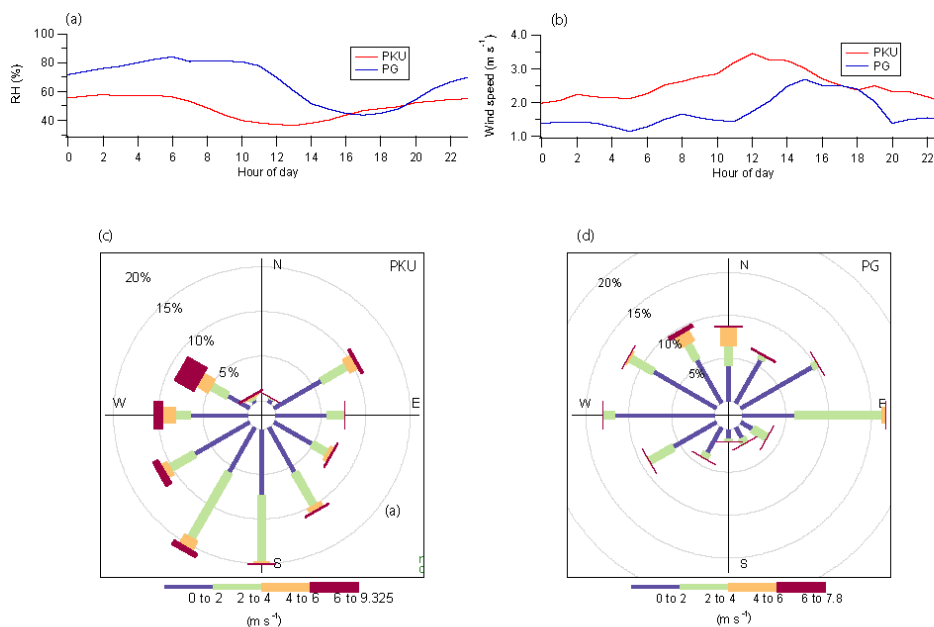
127 Temperature, relative humidity (RH), and wind speed at both sites during the sampling  
 128 period are summarized in Table 2. Their temporal trends are available in Part II. The  
 129 average temperature at PKU (urban,  $5.7 \pm 2.3$  °C) was higher than at PG (rural,  $3.1 \pm 2.2$  °C).  
 130 Correspondingly, relative humidity was higher at PG ( $67 \pm 32\%$ ) than at PKU ( $49 \pm 30\%$ ).  
 131 The wind was stronger at PKU ( $2.5 \pm 1.8$  ms<sup>-1</sup>) than at PG ( $1.7 \pm 0.9$  ms<sup>-1</sup>). As shown in





132 Figure 2, at PKU, wind speed peaked at noon (local time, UTC+8), while at PG, wind speed  
 133 reached its maxima at 15:00. Various wind speeds determined the different dispersion  
 134 patterns of pollutants near the surface. It should be noticed that wind speed up to  $2 \text{ ms}^{-1}$   
 135 representing a scale of 172 km in diurnal transport. Therefore, at PKU, the wind could  
 136 bring the pollutants from Hebei province under stagnant air conditions.

137



138

139 Figure 2. Diurnal plots of (a) RH and (b) wind speed, and rose plots of wind at (c)PKU and  
 140 (d) PG.

141 Table 2. Meteorological parameters at PKU and PG during the campaign.

	PKU	PG
Temperature ( $^{\circ}\text{C}$ )	$5.7 \pm 2.3$	$3.1 \pm 2.2$
RH (%)	$49 \pm 30$	$67 \pm 32$
Wind speed ( $\text{ms}^{-1}$ )	$2.5 \pm 1.8$	$1.7 \pm 0.9$

142



## 143 **3.2 Common particle categories at both PKU and PG**

### 144 **3.2.1 Elemental carbon (EC)**

145 As shown in Figure 3a, the elemental carbon (EC) particle category was represented by  
146 ions peaking at  $m/z$  12, 24, 36, 48, and 60 in positive mass spectra (Sodeman et al., 2005;  
147 Toner et al., 2008). EC is emitted from solid fuel combustion, traffic (Sodeman et al., 2005;  
148 Toner et al., 2008; Toner et al., 2006), and industrial activities (Healy et al., 2012). Due to  
149 the various ionic intensities of nitrate ( $m/z$  -46 and -62) and sulfate ( $m/z$  -80 and -97), the  
150 EC category has four types including EC-Nitrate (EC-Nit), EC-Sulfate (EC-Sul), and EC-  
151 Nit-Sul. Besides, the EC category was more abundant after the heating began rather than  
152 before (Part II), indicating that coal burning was one of the primary sources.

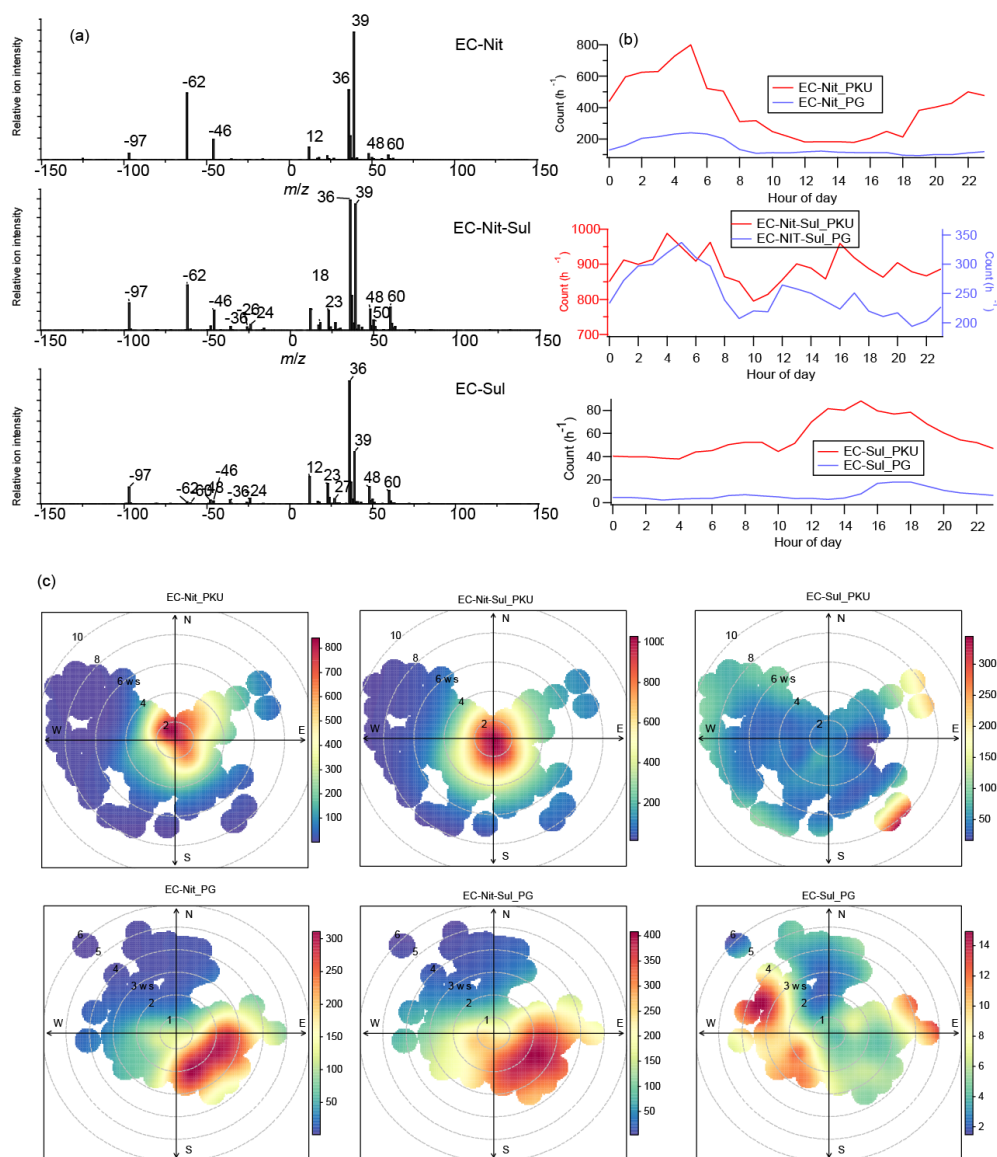
153 EC-Nit\_PKU and EC-Nit\_PG accounted for 7.0% and 2.0% in PKU and PG datasets,  
154 respectively. In the diurnal profiles of EC-Nit\_PKU, there was an apparent early morning  
155 peak at 5:00 (UTC+8, local time), along with an evening peak (22:00). There was also an  
156 early morning NO<sub>x</sub> peak in the urban area of Beijing, providing sufficient precursors for  
157 secondary nitrate (Shi et al., 2019). Wang et al. (2018) validated the role of N<sub>2</sub>O<sub>5</sub> uptake  
158 on the nitrate formation in PM. Therefore, the early morning peak of EC-Nit\_PKU  
159 occurred due to the uptake of nitrate on the freshly emitted EC in the early morning (Sun  
160 et al., 2014). The evening peak could be due to the low temperature after the heating supply  
161 started. Diurnally, EC-Nit\_PG exhibited an early morning peak (5:00) but no evening peak  
162 and mainly came from the southeast.

163 EC-Nit-Sul was more abundant at the rural site (18.6%) than the urban site (11.6%). EC-  
164 Nit-Sul\_PKU (10.5%) had early morning (04:00), morning (7:00), and afternoon peaks



165 (around 16:00), while EC-Nit-Sul\_PG (3.5%) had early morning (04:00), noon, and  
166 afternoon peaks (17:00, Figure 3a). However, they showed relatively small diurnal  
167 variations. For example, EC-Nit-Sul\_PKU varied between  $800 \text{ h}^{-1}$  and  $1,000 \text{ count h}^{-1}$ , and  
168 EC-Nit-Sul\_PG shifted between  $200 \text{ count h}^{-1}$  and  $250 \text{ count h}^{-1}$ . Thus, the EC-Nit-Sul at  
169 both sites was most likely acting as background and regional particles (Dall'Osto et al.,  
170 2016). Additionally, EC-Nit-Sul\_PKU mainly came from the surrounding area in the city  
171 pollutant plume, while EC-Nit-Sul\_PG mainly came from the southeast (Figure 3c).

172 EC-Sul was a minor type at both sites, accounting for 0.7% at PKU and 0.1% at PG. EC-  
173 Sul was pronounced in the afternoon when the wind was strong at both sites. It was unlikely  
174 for either EC-Sul\_PKU or EC-Sul\_PG to be local because their concentrations were  
175 associated with high wind speed, as shown in Figure 3c. More specifically, EC-Sul\_PKU  
176 came from the southeast and northeast of Hebei Province when the wind speed exceeded  
177  $6 \text{ m s}^{-1}$ . EC-Sul\_PG could come from the west when the wind speed exceeded  $2 \text{ m s}^{-1}$  and  
178 the east when the wind speed exceeded  $3 \text{ m s}^{-1}$ , as coal-using industries are located in both  
179 directions. Also, at both sites, the concentrations of  $\text{SO}_2$  were elevated in the afternoon due  
180 to transport, providing sufficient precursors for the formation of sulfate (Shi et al., 2019).



181

182 Figure 3. (a) average mass spectra of EC-Nit, EC-Nit-Sul, and EC-Sul at both sites; (b)

183 diurnal patterns of EC-Nit, EC-Nit-Sul, and EC-Sul at both sites; (c) polar plots of EC-Nit,

184 EC-Nit-Sul, and EC-Sul; the grey circles indicate wind speed (m s<sup>-1</sup>).



### 185 3.2.2 Organic carbon (OC) category

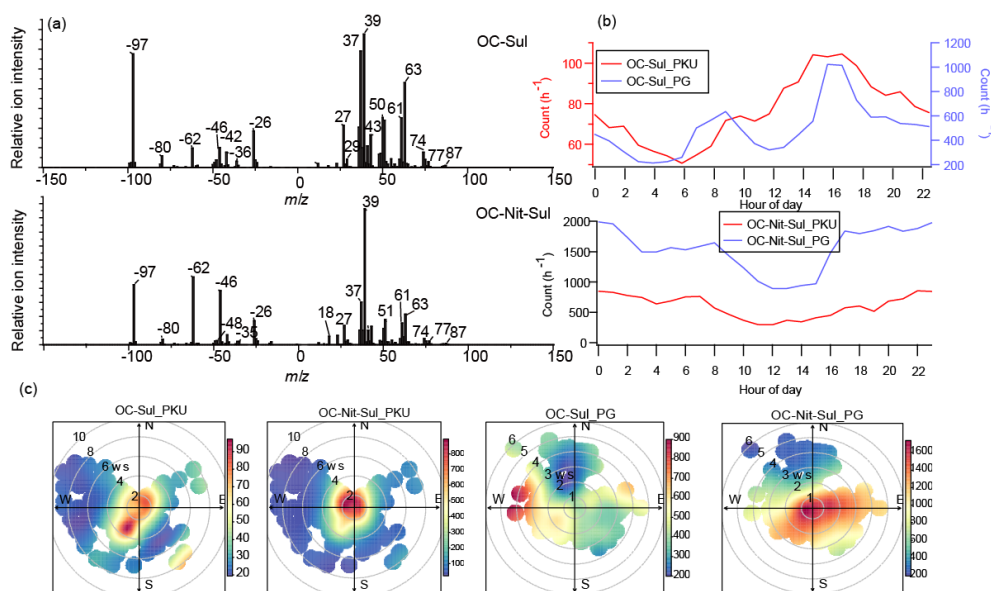
186 The positive mass spectra of both OC-Nit and OC-Nit-Sul contained complicated organic  
187 ions such as  $C_2H_3^+$  ( $m/z$  27),  $C_3H^+$  ( $m/z$  37),  $C_3H_7^+/C_2H_3O^+/CHNO^+$  ( $m/z$  43),  $C_4H_2^+$  ( $m/z$   
188 50), aromatic hydrocarbons ( $C_4H_3^+$ ,  $C_5H_3^+$ , and  $C_6H_5^+$ ), and diethylamine ( $(C_2H_5)_2NH_2^+$ ,  
189  $m/z$  74),  $(C_2H_5)_2NCH_2^+$  ( $m/z$  86)). The negative mass spectra contained  $CN^-$  ( $m/z$  -26),  $Cl^-$   
190 ( $m/z$  -35 and 37),  $CNO^-$  ( $m/z$  -42), nitrate ( $m/z$  -46 and -62), and sulfate ( $m/z$  -97). The  
191 presence of  $CN^-$  and  $CNO^-$  suggests the existence of organonitrogen species (Day et al.,  
192 2010). Peak intensities of organic fragments are relatively high in the OC-Sul particles,  
193 indicating that it was relatively fresh, while OC-Nit-Sul was more processed (Zhai et al.,  
194 2015). The positive mass spectrum had similar ions of Coal Combustion OA (CCOA) with  
195 significant signals of PAHs in AMS studies (Sun et al., 2013). OC-Sul showed different  
196 spatial distributions with 0.9% at PKU and 6.9% at PG.

197 OC-Sul\_PG had morning (8:00) and afternoon (16:00) peaks, while the diurnal profile of  
198 OC-Sul\_PKU showed a trend with an early morning (3:00), morning (10:00), and  
199 afternoon peaks (16:00). The diurnal trends OC-Sul at both PKU and PG were consistent  
200 with the heating pattern depending on the variation of local temperature. Moreover, OC-  
201 Sul\_PG increased after the heating supply began. Polar plots suggest that OC-Sul\_PKU  
202 came from surrounding southwest areas via transport, while OC-Sul\_PG came from  
203 villages to the east and west (Figure 4). These results suggest that OC-Sul\_PG was emitted  
204 from coal burning for residential heating in nearby areas.

205 OC-Nit-Sul accounted for 7.4 % and 21.3 % of all detected particles at PKU and PG,  
206 respectively. OC-Nit-Sul\_PKU had a diurnal peak at 7:00 in rush hours, suggesting that



207 OC-Nit-Sul could be formed due to the uptake of nitrate on OC-Sul. While OC-Nit-Sul\_PG  
208 had a diurnal peak at 8:00 due to traffic in nearby towns. As an aged particle type, OC-Nit-  
209 Sul\_PKU and OC-Nit-Sul\_PG, also acting as a similar type of background types with  
210 hourly counts remained low but elevated to high levels at night. Polar plots suggest that  
211 OC-Nit-Sul\_PKU mainly came from the surrounding areas, while OC-Nit-Sul\_PG mainly  
212 came from the south and east, where populous villages are located (Figure 4).



213

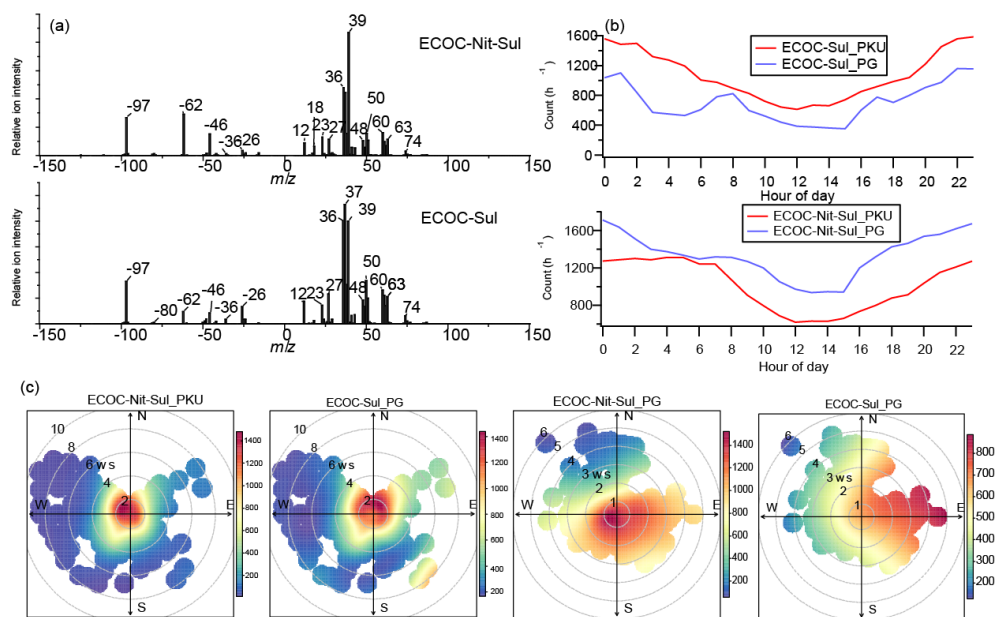
214 Figure 4. (a): average mass spectra of OC-Nit and OC-Nit-Sul observed at both sites; (b):  
215 diurnal patterns of the hourly count of OC-Nit and OC-Nit-Sul at both sites; (c): polar plots  
216 of OC-Sul and OC-Nit-Sul; the grey circles indicate wind speed ( $\text{m s}^{-1}$ ).

### 217 3.2.3 ECOC category

218 As shown in Figure 5a, the ECOC category contained two major particle types: ECOC-  
219 Nit-Sul and ECOC-Sul. The positive mass spectrum of ECOC-Nit-Sul contained  $\text{C}_n^+$  ( $m/z$



220 12, 24, 36...),  $\text{NH}_4^+$  ( $m/z$  18),  $\text{C}_2\text{H}_3^+$  ( $m/z$  27),  $\text{K}^+$  ( $m/z$  39 and 41),  $\text{C}_3\text{H}_7^+/\text{C}_2\text{H}_3\text{O}^+/\text{CHNO}^+$   
221 ( $m/z$  43),  $\text{C}_4\text{H}_2^+$  ( $m/z$  50), and  $[(\text{C}_2\text{H}_5)_2\text{NH}_2]^+$  ( $m/z$  74); in the negative mass spectrum, ions  
222 such as sulfate ( $m/z$  -80 and -97), nitrate ( $m/z$  -46 and -62),  $\text{C}_n^-$ , and  $\text{CN}^-$  ( $m/z$  -26) were  
223 abundant. This mixture of EC and OC particle types was common in single particle studies.  
224 ECOC could be local, and from incomplete combustion processes (Chen et al., 2017), or  
225 regional transport, e.g., after aging (McGuire et al., 2011; Huang et al., 2013). The diurnal  
226 profile of ECOC-Sul\_PG showed early morning (1:00), morning (8:00), and afternoon  
227 (17:00) peaks, which is consistent with local cooking and heating patterns. Also, heating  
228 activities enhanced the fraction of ECOC-Sul\_PG. ECOC-Sul\_PKU did not show a clear  
229 diurnal profile, suggesting that ECOC-Sul\_PKU was mainly a background type. Similarly,  
230 ECOC-Nit-Sul\_PKU and ECOC-Nit-Sul\_PG were also background types with less  
231 obvious diurnal variations (Dall'Osto et al., 2016). Polar plots (Figure 5c) suggested that  
232 both ECOC-Nit-Sul\_PKU and ECOC-Sul\_PKU had both local and regional sources. Wind  
233 speed up to  $4 \text{ m s}^{-1}$  could cause a transport with a distance of 346 km diurnally, indicating  
234 that it was possible for the particles from Hebei province to arrive at the sampling place.



235

236 Figure 5. (a): average mass spectra of ECOC-Nit and ECOC-Nit-Sul observed at both sites;  
237 (b): diurnal patterns of the hourly count of ECOC-Sul and ECOC-Nit-Sul at both sites; (c):  
238 polar plots of ECOC-Sul and ECOC-Nit-Sul; the grey circles indicate wind speed ( $\text{ms}^{-1}$ ).

### 239 3.2.4 K-rich category

240 Figure 6 shows a series of potassium-rich (K) particle types. K-rich contained  $\text{Na}^+$  ( $m/z$  23),  
241  $\text{C}_2\text{H}_3^+$  ( $m/z$  27),  $\text{C}_n^+$ ,  $\text{C}_3\text{H}^+$  ( $m/z$  37),  $\text{K}^+$ , aromatic hydrocarbons ( $\text{C}_4\text{H}_3^+$ ,  $\text{C}_5\text{H}_3^+$ , and  $\text{C}_6\text{H}_5^+$ ),  
242 levoglucosan ( $m/z$  -45, -59, and -71), sulfate, and nitrate. According to the ionic intensities  
243 of sulfate and nitrate, the K-rich particle category had several branches such as K-rich, K-  
244 Nit, K-Sul, and K-Nit-Sul. K-rich particles are commonly found in biomass burning  
245 emissions (Silva et al., 1999; Pagels et al., 2013; Chen et al., 2017).  $\text{Cl}^-$  was un-abundant in  
246 all K-rich particle types, suggesting that the K-rich particles had undergone aging during

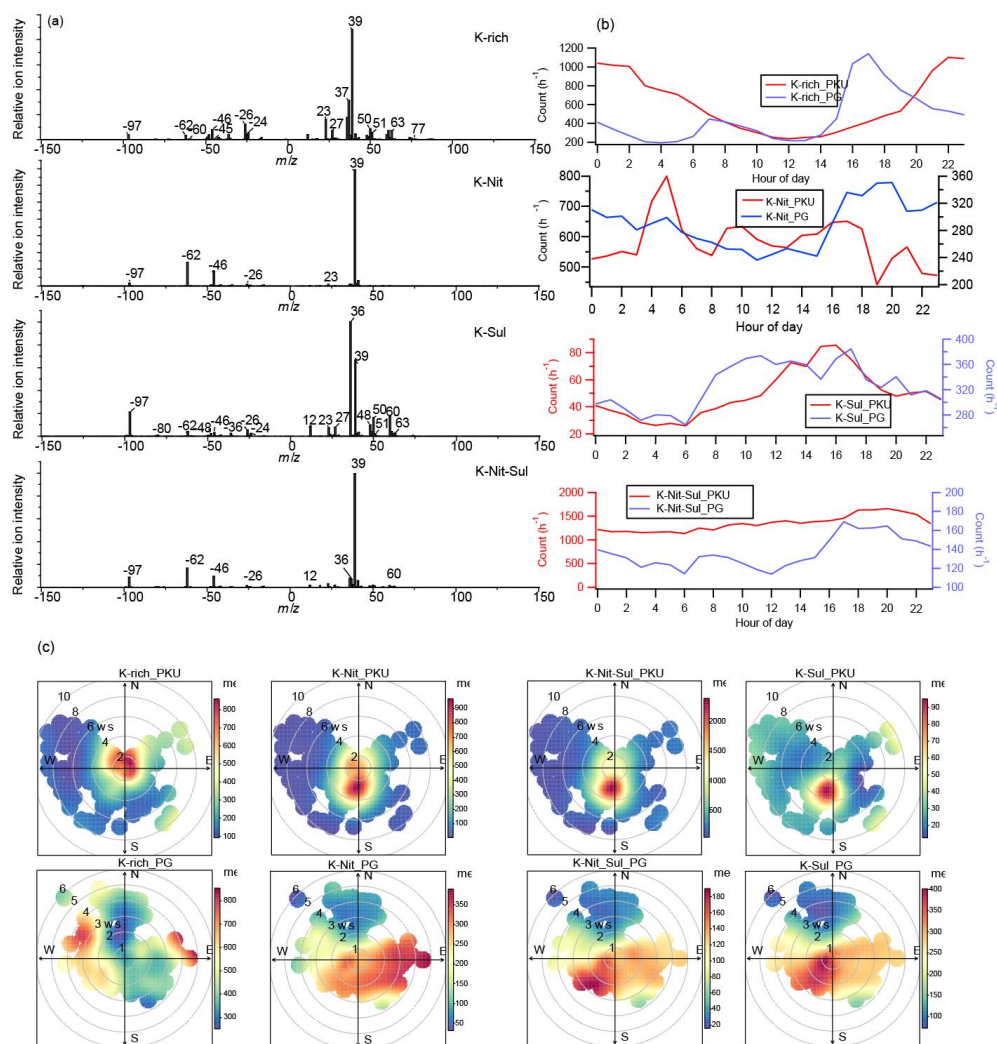




247 atmospheric processing (Sullivan et al., 2007; Chen et al., 2016), but K-Nit, K-Nit-Sul, and  
248 K-Sul were more processed.

249 All K-rich category particles showed different atmospheric evolution process at both PKU  
250 and PG. K-rich\_PKU illustrated a typical pattern that was at low levels in the daytime but  
251 high levels at nighttime (22:00). As shown in Figure 6c, at an average wind speed of 3 m  
252  $s^{-1}$ , it took five hours for particles from a distance of 50 km to arrive at PKU. This is also  
253 the reason why BB-related particles were abundant in urban Beijing where the household  
254 BB is prohibited. The origination of K-rich\_PKU was from nearby and southwest. K-  
255 rich\_PG, however, showed a pattern with cooking and heating activities, peaking at 7:00  
256 and 17:00. The peak at 7:00 was due to the local emissions; the 17:00 could be transported  
257 from a distance of 50 km at a wind speed of 3 m  $s^{-1}$  from the east and west.

258 The secondary process contributed to the early morning peak (5:00) of K-Nit\_PKU due to  
259 the nighttime formation of nitrate via hydrolysis of  $N_2O_5$  in the  $NO_x$ -rich urban areas (Wang  
260 et al., 2017). In the day time, after the rush hours, the number concentration of K-Nit\_PKU  
261 increased again via the uptake of nitrate due to day time photoactivity. K-Nit\_PKU mainly  
262 originated from the local and southerly areas (Figure 6c). Besides the early morning peak,  
263 K-Nit\_PG showed cooking and heating patterns that they were abundant when the  
264 temperature was low in the early morning and afternoon. K-Nit\_PG had wide originated  
265 from both local and region via long-range transport.



266

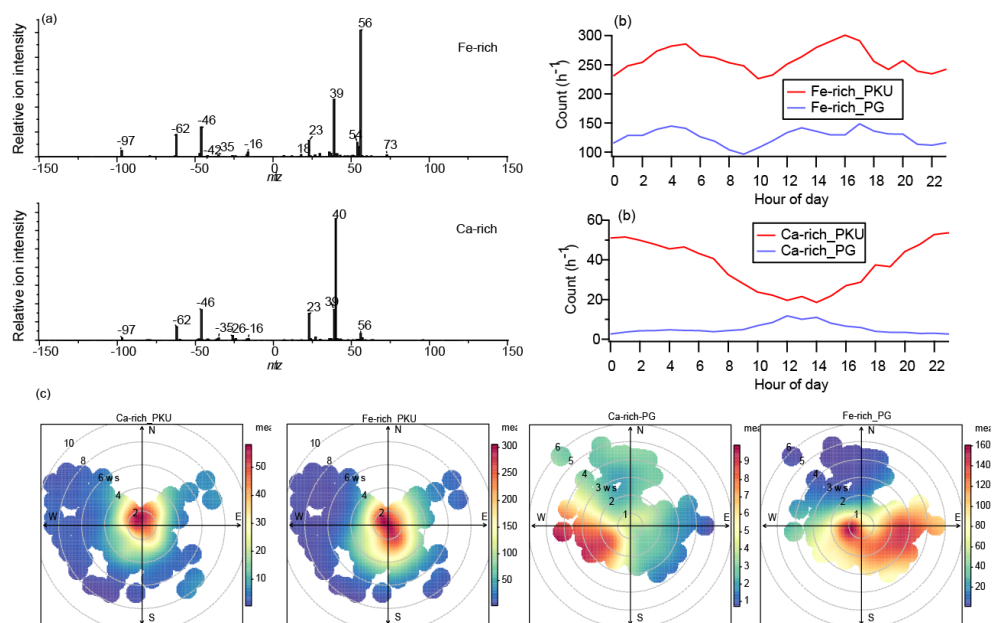
267 Figure 6. (a): average mass spectra of BB, K-Nit, K-Sul, and K-Nit-Sul observed at both  
268 sites; (b): diurnal patterns of the hourly count of K-rich, K-Nit, K-Sul, and K-Nit-Sul at  
269 both sites; (c): polar plots of BB, K-Nit, K-Sul, and K-Nit-Sul; the grey circles indicate  
270 wind speed ( $\text{m s}^{-1}$ ).



271 **3.2.5 Metal category**

272 Two metal-rich particles types were identified, namely Fe-rich and Ca-rich. Fe-rich  
273 contained iron ( $m/z$  56 and 54),  $K^+$ ,  $Na^+$ ,  $NH_4^+$ ,  $Cl^-$  ( $m/z$  -35 and -37), sulfate, and nitrate.  
274 Ca-rich was composed of  $Ca^+$  ( $m/z$  40),  $CaO$  ( $m/z$  56),  $K^-$ ,  $Na^+$ ,  $Cl^-$ , sulfate, and nitrate. As  
275 shown in Figure 6b, Ca-rich\_PKU (0.4%) and Ca-rich\_PG (0.1%) were likely of regional  
276 origin with no distinct diurnal variations. Since  $SiO_2^-$  or  $SiO_3^-$  ( $m/z$  -60 and -76) were not  
277 abundant in the Ca-rich particles, they are not likely to come from dust (Silva et al., 2000).  
278 According to its weak peaks during the rush hour at PKU, a possible source of the Ca-rich  
279 particles was from road dust re-suspension. Such rush hour peaks were not observed at PG.

280 Fe-rich\_PKU (3.1%) and Fe-rich\_PG (1.8%) had similar diurnal profiles that arose in the  
281 early morning when heavy-duty vehicles were allowed to enter the 5-ring expressway. The  
282 peak occurred earlier at PG (4:00) than (5:00) because these vehicles got close to PG earlier  
283 than to PKU. The daytime peak occurred in the afternoon at both PKU and PG when wind  
284 speed was high. Therefore, there were also multiple sources for Fe-rich particles, including  
285 re-suspended dust particles from traffic and fly ash from the steel industry. In Beijing,  
286 daytime Fe-rich particles were reported and assigned to long-range transport and industrial  
287 sources from Heibei Province (Figure 7c) (Li et al., 2014). The steel industry moved out  
288 of Beijing more than a decade ago (Liu et al., 2016b). Currently, most of these steel  
289 industries were located in the Heibei Province.



290

291 Figure 7. (a): average mass spectra of Fe-rich and Ca-rich observed at both sites; (b):  
292 diurnal patterns of the hourly count of Fe-rich and Ca-rich at both sites; (c): polar plots of  
293 Fe-rich and Ca-rich; the grey circles indicate wind speed ( $\text{ms}^{-1}$ ).

### 294 3.2.6 NaK category

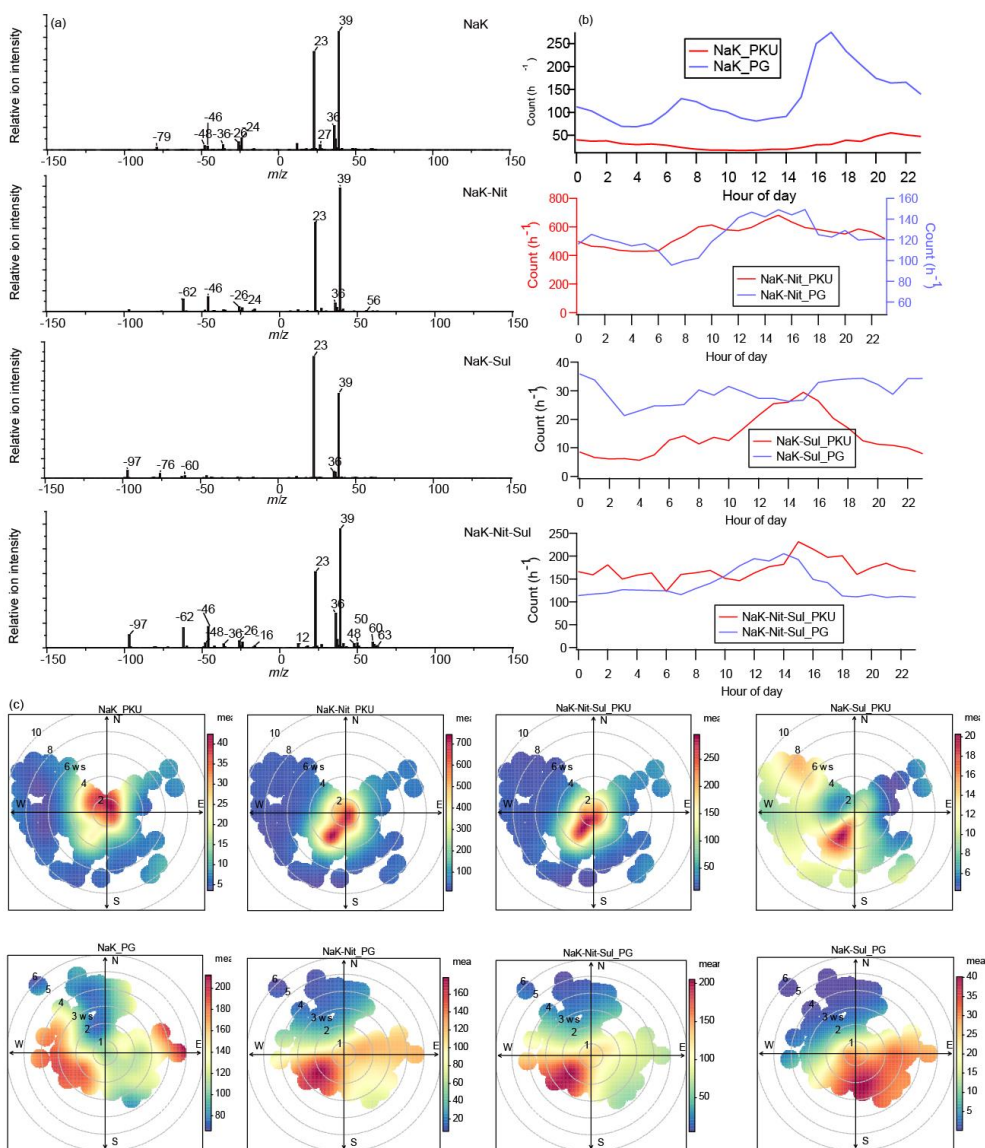
295 As shown in Figure 8, mass spectra of NaK category contained f  $\text{Na}^+$ ,  $\text{K}^+$ ,  $\text{C}_n^+$ ,  $\text{C}_n^-$ , nitrate,  
296 and  $\text{PO}_3^-$  ( $m/z$  -79). The aged NaK particles contained strong signals of nitrate (NaK-Nit),  
297 sulfate (NaK-Sul), or both (NaK-Nit-Sul). In general, the NaK category contained stronger  
298 signals of  $\text{Na}^+$  than the EC and K-rich categories. The NaK category may also come from  
299 incomplete solid fuel combustion processes such as coal, peat, or wood (Chen et al., 2017;  
300 Healy et al., 2010; Xu et al., 2017). NaK category was more abundant at PKU (9.5%) than  
301 PG (5.8%), suggesting a stronger contribution of emission from coal boilers (Xu et al.,



302 2017; Xu et al., 2018). Additionally, after heating began, the fraction of NaK-Nit\_PG and  
303 NaK-Sul-Nit\_PG increased by 1.2 times (see Part II).

304 NaK\_PKU showed no distinct diurnal variations, suggesting that it was a regional particle  
305 type arriving at the PKU site via transport, while NaK\_PG showed an apparent diurnal  
306 variation consistent with cooking and heating pattern. Polar plots also suggest that they are  
307 from the east and the west. NaK-Nit, with a considerable uptake of nitrate, was more  
308 abundant at PKU (6.4%) than PG (1.7%). Both NaK-Nit\_PKU and NaK-Nit\_PG increased  
309 in the afternoon when photochemical activities were most active (Figure 8c). Both of them  
310 may be from regional transport (Figures 8b and 8c).

311 NaK-Sul was a minor particle type at both PG and PKU, accounting for 0.2% and 0.4%,  
312 respectively. The diurnal profile of NaK-Sul\_PG was also following the local cooking and  
313 heating pattern, while NaK-Sul\_PKU showed a typical transport pattern that became  
314 abundant in the afternoon as the southwestern wind speed increased. As a heavily aged  
315 particle type, NaK-Nit-Sul was transported to both PKU and PG from the southwest. In  
316 short, NaK-related particle types mainly came from the solid fuel burning process, e.g.,  
317 coal. Due to its different origins, it showed different levels of processing at PKU and PG,  
318 respectively.



319

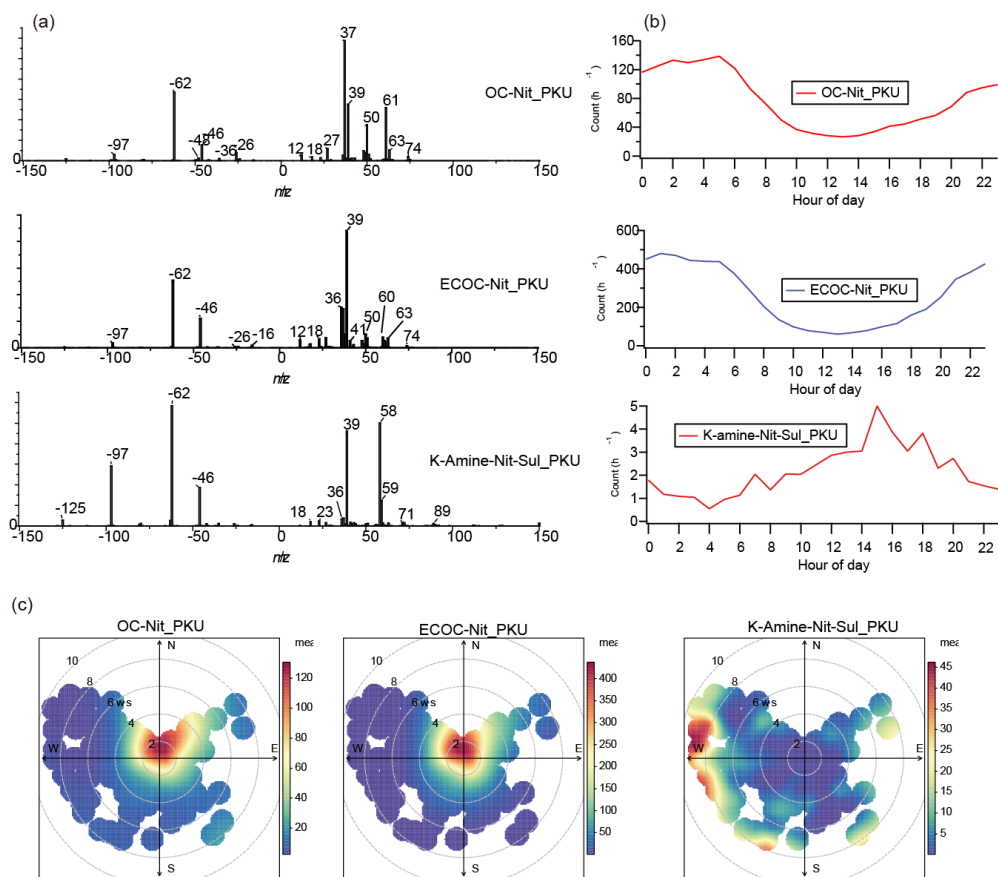
320 Figure 8. (a): average mass spectra of NaK, NaK-Nit, NaK-Nit-Sul, and NaK-Sul observed  
 321 at both sites; (b): diurnal patterns of the hourly count of NaK, NaK-Nit, NaK-Nit-Sul, and  
 322 NaK-Sul at both sites; (c): polar plots of NaK, NaK-Nit, NaK-Nit-Sul, and NaK-Sul; the  
 323 grey circles indicate wind speed ( $\text{m s}^{-1}$ ).



324 **3.3 Unique Particle types at the PKU site**

325 OC-Nit\_PKU (0.9%) and ECOC-Nit\_PKU (3.1%) with strong ion intensities of nitrate  
326 were observed at the PKU site. OC-Nit\_PKU and ECOC-Nit\_PKU showed a peak at night  
327 than at daytime, similar to the diurnal profiles of OC-Nit-Sul\_PKU and ECOC-Nit-  
328 Sul\_PKU. Such nitrate-rich particle types could have come from the uptake of nitrate in  
329 OC and ECOC(Qin et al., 2012; Chen et al., 2016). Polar plots suggest that both types were  
330 formed locally when the wind speed was lower than  $4 \text{ ms}^{-1}$ . The  $\text{NO}_x$ -rich environment in  
331 urban Beijing provides a favorable condition for nitrate formation at night (Wang et al.,  
332 2016; Zou et al., 2015; Shi et al., 2019).

333 A minor amount (0.10%) of amine-containing particles was observed at the PKU site, and  
334 trimethylamine ion fragments ( $m/z$  58 and 59) were influential in the mass spectrum of K-  
335 amine-Nit-Sul\_PKU (Figure 9a). The diurnal profile of K-amine-Nit-Sul\_PKU showed an  
336 afternoon peak, indicating a regional source (Figure 9c). K-amine-Nit-Sul\_PKU was  
337 transported to the site from nearby locations. The amines may come from animal husbandry,  
338 BB, traffic, or vegetation (Chen et al., 2019). Amines were ubiquitous in the atmospheric  
339 environment, playing essential roles in new particle formation and growth, as well as fog  
340 and cloud processing (Ge et al., 2011; Chen et al., 2019).



341

342 Figure 9. (a): average mass spectra of OC-Nit\_PKU, ECOC-Nit\_PKU, and K-amine-Nit-  
343 Sul\_PKU observed at the PKU site; (b): diurnal patterns of the hourly count of OC-  
344 Nit\_PKU, ECOC-Nit\_PKU, and K-amine-Nit-Sul\_PKU at the PKU site; (c): polar plots of  
345 OC-Nit\_PKU, ECOC-Nit\_PKU, and K-amine-Nit-Sul\_PKU, and the grey circles indicate  
346 wind speed ( $\text{m s}^{-1}$ ).

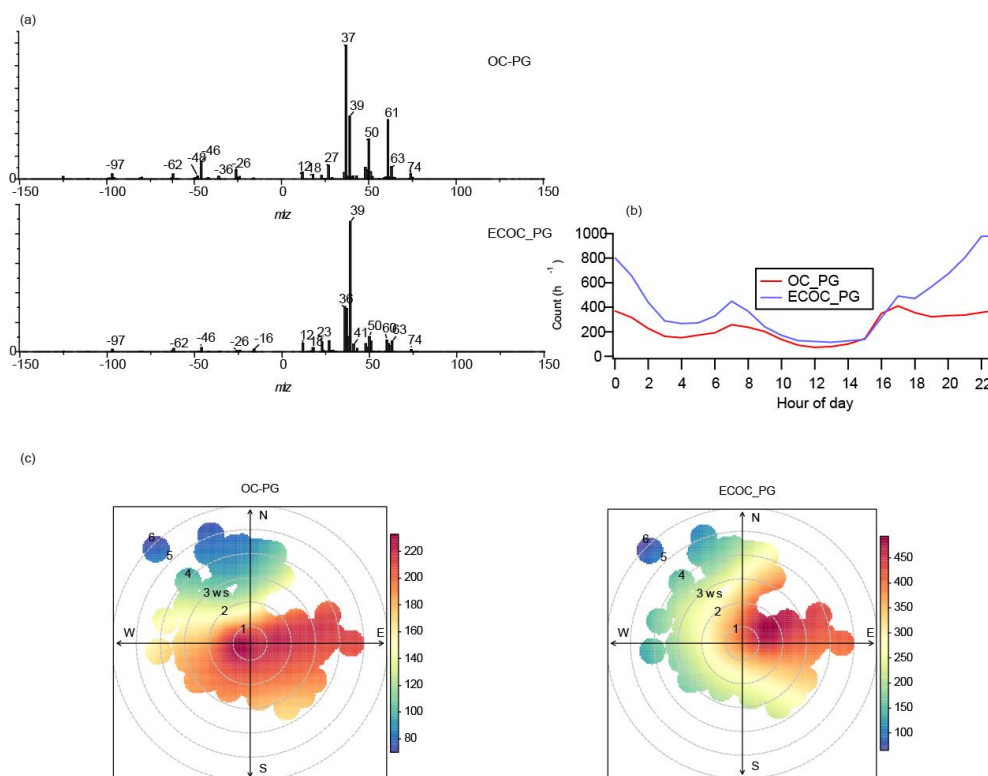
#### 347 3.4 Unique Particle types at the PG site

348 OC\_PG (5.9%) and ECOC\_PG (3.3%) were only observed at the rural site PG (Figure 10).  
349 The major components of these two types were consistent with the OC and ECOC





350 categories, respectively, but with limited uptake of sulfate and nitrate, suggesting that they  
351 were possibly freshly emitted particles. Their diurnal profiles are consistent with cooking  
352 and heating patterns which peaked at 07:00 in the morning and 17:00. Polar plots suggest  
353 that OC\_PG mainly came from nearby and other remote areas in all directions except the  
354 north. ECOC mainly came from the east of the PG site. These results supported the  
355 assumption that the two types were mainly from local emission sources.



356

357 Figure 10. Average mass spectra of (a) ECOC-Nit, (b) ECOC-Nit, and (c) OC-Nit. All  
358 these particle types appeared at the PG site.



359 **4. Discussion**

360 Multiple source apportionment models have been used in Beijing to quantify the sources  
361 of particles (Sun et al., 2014; Xu et al., 2015; Zhai et al., 2016). Biomass burning, coal  
362 combustion, traffic, and dust are the key sources of PM (Sun et al., 2014; Liu et al., 2018;  
363 Huang et al., 2014). Multiple studies confirmed that biomass burning is an essential source  
364 of PM in urban Beijing (Gao et al., 2014; Huang et al., 2014; Sun et al., 2014; Zheng et al.,  
365 2017). In this study, biomass burning, and other solid fuel burning were identified as crucial  
366 sources of PM in not only urban but also rural areas of Beijing. We observed that BB-  
367 related particles (K-rich category) were more abundant at PG than at PKU. In particular,  
368 we found fresh-emitted K-containing particles at the Pinggu site, confirmed the importance  
369 of local emissions to PM. Furthermore, K-containing particles in the urban area were more  
370 aged, suggested that they are aged and mostly from the surrounding areas. The result is  
371 consistent with the results from (Liu et al., 2019) based on a combined receptor and  
372 footprint models. Nevertheless, household emissions in the BHT region caused 32% and  
373 15% of primary PM<sub>2.5</sub> and SO<sub>2</sub>. These studies have proved the importance of household  
374 emission from BB in the BHT area (Liu et al., 2016a). Especially at the PG site, the ambient  
375 PM was mainly controlled by long-range transport and household emissions from cooking  
376 and heating.

377 Due to the nature of SPAMS, the chemical composition of PM cannot be precisely  
378 quantified. However, single particle aerosol mass spectrometers have advantages in  
379 studying the chemical composition, mixing state, source, and process of particles (Pratt and  
380 Prather, 2012). Mass-based technologies can not differentiate the origin of the bulk of  
381 nitrate, whether it is transported or formed locally. Indeed, single particle types in urban



382 Beijing have been reported in previous studies (Li et al., 2014; Liu et al., 2016b), and the  
383 major types are consistent with this study. However, in this study, we adopted a cluster  
384 strategy considering the relative ion peak area of sulfate and nitrate as indicators of particle  
385 processing. Therefore, more detailed could be extracted from both two simultaneous  
386 datasets. We confirmed that the source, origination, and processes were different for these  
387 particles in the urban and rural areas. For example, the seriously processed particles, such  
388 as K-Nit-Sul, OC-Nit-Sul, and NaK-Nit-Sul, acted with no distinct diurnal patterns as  
389 background or regional sources. The processed particles, such as OC-Nit, ECOC-Nit, and  
390 NaK-Nit, were affected by emissions and secondary formations.

391 The emission and transport patterns were different in the urban and rural areas, resulting in  
392 different characteristics of PM. For example, EC particles were a key component at PKU  
393 (18.2% in total), but a minor particle type at PG (5.6%). Meanwhile, in the urban area of  
394 Beijing, direct emission of K-rich particles should be small due to strict control measures;  
395 thus, the K-Nit-Sul particles are mainly from long-range transport. More interestingly, the  
396 sulfate-rich particle types, such as EC-Sul and ECOC-Sul, were commonly accompanied  
397 by high wind speed, suggesting the contribution of regional air masses. Sulfate and SO<sub>2</sub>  
398 emissions were controlled strictly in Beijing. However, in the nearby Hebei and Shandong  
399 province, the emission of SO<sub>2</sub> is still significant (Shi et al. 2019). Transported particles  
400 were aged and commonly coated a thick layer of nitrate and sulfate, but the local particles  
401 were affected by both emission and the near-surface aging process. For example, at PKU,  
402 the primary emission sources were traffic and central heating supply, causing a NO<sub>x</sub>-rich  
403 region in which freshly-emitted particle types could undergo processing due to the uptake  
404 of nitrate (Wang et al., 2016). In the nearby villages of PG, domestic heating and cooking



405 were the major contributors of primary particles when the temperature was low in the  
406 morning and afternoon, resulting in the emission of multiple primary particle types such as  
407 OC\_PG and ECOC\_PG. In short, the characteristics of PM in urban and rural areas of  
408 Beijing were affected by local emissions and interacted with each other due to regional  
409 transport.

410 Secondary nitrate formation is still a critical issue in Beijing. The daytime arising of nitrate  
411 has been reported in studies (Sun et al., 2013), and we also found a similar predominant of  
412 nitrogen-containing particles in this study. Recent studies have reported the early morning  
413 peaks of nitrate using a soot particle aerosol mass spectrometer (SP-AMS) (Wang et al.,  
414 2019), which is consistent with our results. Interestingly, the early morning peak was only  
415 observed for several particle types at both sites, including EC-Nit\_PKU, K-Nit\_PKU, EC-  
416 Nit-Sul-PG, and EC-Nit\_PG. This result is not surprising because PG is also a NO<sub>2</sub>-rich  
417 region (Shi et al., 2019). The increasing contribution of nitrate-containing particles  
418 suggests the role of night chemistry in nitrate uptake on particles. Wang et al. (2017)  
419 revealed the importance of night N<sub>2</sub>O<sub>5</sub> chemistry on nocturnal nitrate formation in the  
420 urban area of Beijing. The heterogeneous hydrolysis of N<sub>2</sub>O<sub>5</sub> was most favorable when NO  
421 was at a low level. Moreover, the polar plots suggested a small role of long-range transport  
422 to the nitrate in individual particles. The contribution of local traffic was insignificant at  
423 the PG site as it was far from highways and major roads, the nighttime formation of nitrate  
424 appeared to be important in PG as well.



425 **5. Conclusion**

426 Two SPAMs were simultaneously deployed at urban and rural sites in Beijing in order to  
427 characterize PM during wintertime. The results at both sites indicate that they shared 17  
428 types of common clusters, most of which belonged to particle categories such as EC, OC,  
429 ECOC, BB, and NaK. The origins and sources of these particle types at both sampling sites  
430 are also comprehensively analyzed. Most of the processed PM, including EC-Nit-Sul\_PKU,  
431 ECOC-Nit-Sul\_PKU, and NaK-Nit-Sul\_PKU, were aged locally in a NO<sub>x</sub>-rich  
432 environment, while EC-Nit-Sul\_PG, ECOC-Nit-Sul\_PG, NaK-Nit-Sul\_PG, and OC-Nit-  
433 Sul\_PG were regional. Domestic heating in the rural area was found to be an important  
434 source of PM, and such heating activities typically caused three diurnal peaks in the early  
435 morning, morning, and afternoon (after sunset). Moreover, the early morning peak of  
436 nitrate was observed at both sites, suggesting the contribution of the heterogeneous  
437 hydrolysis of N<sub>2</sub>O<sub>5</sub> in the dark during the winter. The insights gained in this study can  
438 provide useful references for understanding the relationship between regional transport and  
439 local aging in both urban and rural areas in Beijing. In Part II, we focus on haze events  
440 observed at both sites and attempt to determine the effects of heating activities and possible  
441 regional transport between PKU and PG.

442 *Data availability.* All data described in this study are available upon request from the  
443 corresponding authors.

444 *Author contributions.* FY, MZ, TZ, QZ, and KH designed the experiments; YC, JC, ZW,  
445 MT, CP, and HY carried them out; XYang, XYao, YL, GS, and ZS analyzed the  
446 experimental data; YC prepared the manuscript with contributions from all coauthors.



447 *Competing interests.* The authors declare that they have no conflict of interest.

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