Secondary organic aerosol formation from photooxidation of furan: effects of NOx and humidity

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Abstract. Atmospheric furan is a primary and secondary pollutant in the atmosphere, and its emission contributes to the formation of ultrafine particles. We investigate the effects of NOx level and humidity on the formation of secondary organic

10 aerosol (SOA) generated from the photooxidation of furan in the presence of NaCl seed particles. SOA mass concentration and yield were determined under different NOx and humidity levels. A significant difference is observed both in the variation of SOA mass concentration and SOA yield with the initial experimental conditions. Varying VOC/NOx ratios over the range 48.1 to 8.2 contributes to the effective formation of SOA in the presence of NaCl seed particles, with the SOA mass concentration and SOA yield ranging from 0.96 μg m^{-3} to 23.46 μg m^{-3} and from 0.04 % to 1.01 %, respectively. We found

15 that there was a favourable relationship between the SOA yields and NOx concentration. Especially, the increase of SOA yield with increasing NOx concentration was continuously observed at high NOx levels owing to a corresponding increase of the amount of low-volatility hydroxyl nitrates and dihydroxyl dinitrates that can partition into particle phase. In addition, varying RH from 5% to 88% increased the SOA yield from 1.01% to 5.03%. The enhanced SOA formation from humid conditions may result from the high OH concentration, rapid furan decay rate, enhanced carbonyl-rich products condensation,

20 and the aqueous-phase reactions. Using hybrid quadrupole-orbitrap mass spectrometer equipped with electrospray ionization (HESI-Q Exactive-Orbitrap MS), three carbonyl-rich products and three kinds of organonitrates were identified in the collected SOA. Based on the MS analysis and FTIR spectra, the reaction mechanism of furan photooxidation was proposed. This study demonstrates the effects of NOx and humidity on SOA formation during the furan-NOx-NaCl photooxidation, and provides new insights into the oxidation regime and SOA composition in furan photooxidation. The results also illustrate

25 the importance of studying SOA formation over a comprehensive range of environmental conditions. Only such evaluations can induce meaningful SOA mechanisms to be implemented in air quality models.

1 Introduction

Atmospheric particulate matter (PM) is primarily composed of organic carbon, elemental carbon, sulfate, nitrate, and other components (Donahue et al., 2009; Zhang et al., 2011), which has adverse effects on human health and global climate

30 forcing (Hallquist et al., 2009; Pope et al., 2013). Secondary organic aerosols (SOA) constitute a substantial portion of the

total ambient aerosol particles, which mainly originate from biomass burning and atmospheric reactions of volatile organic compounds (VOCs) (Kanakidou et al., 2005). Identifying the chemical composition and major precursors of SOA is helpful to better understand their formation mechanism and strategies for the control of PM. Furan is an important five-member heteroatom-containing VOC present in the atmosphere and it can be produced by wood combustion, oil refining, coal

5   mining and gasification, and pyrolysis of biomass, cellulose, and lignin (Shafizadeh, 1982). As shown in previous studies, furan is one of the most common products of the thermal cracking of biomass (Kahan et al., 2013; Gilman et al., 2015). Both furan and furan-related compounds have been detected in the effluent from initial smouldering of fuels, oats and soft pellets (Olsson, 2006; Perzon, 2010). In addition, furan has been identified as one of the major compounds emitted from the combustion of Ponderosa Pine, which constitute a significant fraction (5-37 % by estimated emission factor) of smoke from

10   combustion (Hatch et al., 2015). Field measurements of hydrocarbon emissions from biomass burning in Brazil have also shown that furans consists of 52% and 72% of the oxygenated hydrocarbons emissions in the cerrado (grasslands) and selva (tropical forest) regions, respectively (Greenberg et al., 1984). Moreover, furan has been proven to be the typical trace gas of roasting/burning activities (Gloess et al., 2014; Coggon et al., 2016; Burling et al., 2010; Gilman et al., 2015; Sarkar et al., 2016; Stockwell et al., 2016; Stockwell et al., 2015; Stockwell et al., 2014), and it is also an important contribution to OH

15   reactivity towards biomass burning emissions (Gilman et al., 2015; Sarkar et al., 2016). Furan is also a secondary pollutant produced in the photooxidation of some conjugate alkenes. Previous product studies of 1,3-butadiene oxidation have shown that furan is one of the dominant product during the OH radical-initiated reactions in the presence of NO (Sprengnether et al., 2002; Tuazon et al., 1999). Furthermore, another source of furan is the cyclization of the unsaturated 1,4-hydroxyaldehydes, which will further go through the dehydration reactions (Baker et al., 2005). There is strong evidence that furan can cause

20   malignant tumor formation in a variety of experimental animals (IARC, 1995). Therefore, a better understanding of the atmospheric transformations of furan will be helpful to improve the air quality models for toxic species.

As an aromatic hydrocarbon, furan is very reactive according to its electrophilic substitution reactions (Villanueva et al., 2007). The measured rate constant of furan with OH radicals, O_3 and NO_3 radicals at room temperature is $(4.01\pm0.30)\times10^{-11}$, $(2.42\pm0.28)\times10^{-18}$ and $(1.4\pm0.2)\times10^{-12}$ cm^3 molecule^{-1} s^{-1}, respectively (Atkinson et al., 1983; Atkinson et al., 1985). If the

25   oxidant concentration of OH and NO_3 are assumed to be 1.6×10^6 and 5×10^8 molecule cm^{-3}, lifetimes of furan for reactions with OH and NO_3 are estimated to be 3 h and 18 min, respectively (Cabanas et al., 2004). In the troposphere, atmospheric furan is expected to be mainly oxidized by OH during daytime and NO_3 at night. Although the determination of kinetics and products of furan oxidation has been performed (Cabanas et al., 2004; Liljegren and Stevens, 2013; Tapia et al., 2011), the influence of several factors including NOx level and relative humidity on SOA formation from furan has not been well

30   examined. It is important to assess the SOA formation potential of furan and its role in SOA production in biomass burning plumes. Although the SOA formation potential of furan has been studied in recent years (Cabanas et al., 2004; Liljegren and Stevens, 2013; Tapia et al., 2011), the influence of several factors including NOx levels and relative humidity (RH) on SOA formation should be further investigated. It is generally accepted that NOx level plays a critical role in SOA formation, by governing the reactions of organo-peroxy radicals (RO_2) (Song et al., 2007; Song et al., 2005). The branching ratio of RO_2

reactions with NOx and the hydroperoxyl radical (HO$_2$) is determined by the NOx level during experiments (Kroll and Seinfeld, 2008). A previous study showed that the mechanisms of the two reactions differ sharply under different NOx level conditions. RO$_2$ reacts only with NO under high NOx levels and with HO$_2$ under low NOx conditions (Ng et al., 2007a). The latter case produces lower volatility products than the former one. An increase in SOA yield with increasing NOx was

5   proposed to be due to the formation of low-volatility compounds including multifunctional nitrates and dinitrates, which partitioned to the particle phase and contribute significantly to isoprene SOA under high-NOx conditions (Schwantes et al., 2019). Another study of the SOA formation from the irradiation of propylene also showed that the SOA yield decreased with increasing propylene/NOx ratio (Ge et al., 2017a). Recently, Liu et al. indicated that NO$_2$ could participate in the OH-initiated reaction of guaiacol, consequently resulting in the formation of organic nitrates and promoting SOA formation (Liu

10   et al., 2019a). The effect of NOx level on SOA formation from aromatic precursors has been investigated previously but the results are inconclusive. A study focusing on the photooxidation of toluene and m-xylene has demonstrated that aerosol yields decrease as NOx level increases (Xu et al., 2015), while another study showed that the SOA yield from photooxidation of isoprene under high NO$_2$/NO is 3 times more important than that measured under low NO$_2$/NO level (Chan et al., 2010).

15   Besides the mixed effect of NOx level on SOA formation, the RH can significantly alter the sizes of SOA particles, depending on their hygroscopicity (Varutbangkul et al., 2006), their gas-phase reactions involving water (Jonsson et al., 2006), and the aqueous chemistry occurring at their surfaces (Lim et al., 2010; Grgic et al., 2010). The photooxidation experiments of aromatic compounds, such as toluene (Edney et al., 2000; Cao and Jang, 2010; Faust et al., 2017; Hinks et al., 2018), benzene and ethylbenzene (Jia and Xu, 2014), m-xylene (Zhou et al., 2011), and 1,3,5-trimethylbenzene (Cocker et al.,

20   2001) have been carried out previously to study the effect of RH on SOA formation. The results exhibited large discrepancies under different experimental conditions. The yield of SOA generated under low-NOx conditions was found to be small at high RH (Cao and Jang, 2010). A little RH effect was also observed on SOA particle formation and size distributions (Bonn and Moortgat, 2002; Fry et al., 2009). Nevertheless, positive correlations between RH and SOA yield have also been observed in the presence of hygroscopic seed particles (Zhou et al., 2011), and the role of RH in SOA

25   formation also appears to be mixed. Water vapour does not only participate in the VOC photooxidation reactions and affect the gas-phase oxidation mechanisms, but also alters the partitioning of the reaction products between the gas-phase and the particle-phase, hereby influencing the equilibrium partitioning of generated organic aerosols (Spittler et al., 2006). Moreover, the RH is connected with SOA formation due to its strong influence on seed acidity and particle liquid water concentration (Mahowald et al., 2011). A high RH environment contributes to the increase of aerosol liquid water (ALW), which promotes

30   the hydrolysis of organic compounds and leads to other heterogeneous and aqueous phase reactions (Ervens et al., 2011). Previous studies regarding the atmospheric reactions of furan typically focused on the kinetics and mechanism (Gomez Alvarez et al., 2009; Aschmann et al., 2014).

      In the present study, we used FTIR, in concert with SMPS to elucidate the roles of NOx level and RH in SOA formation from furan-NOx irradiation. All the experiments were conducted in the presence of NaCl seed particles, which acted as the

nuclei and provided sufficient seed surface area at the beginning of the reaction to suppress the effects of vapour wall losses of semi/low-volatility species. Specifically, we evaluate whether the increased ALW affects the SOA mass concentration and the SOA yield. The chemical composition of furan SOA was investigated by HESI-Q Exactive-Orbitrap MS, with a focus on the formation of organic nitrates. Strong evidences that both the RH and different NOx levels have a significant effect on

5   SOA formation from furan photooxidation are presented.

2 Experiments

2.1 SOA sample preparation

All experiments were performed in a 1.3 m^3 Teflon FEP film chamber at the Institute of Atmospheric Physics, Chinese

10   Academy of Sciences, Beijing. To maximize and homogenize the interior light intensity, a mirror surface stainless steel was chosen as the interior wall of the enclosure. A total of 42 black-lamps (F40BLB, GE, Fairfield, CT, USA) with emission band centre at 365 nm were equipped in the facilities to simulate the spectrum of the UV band in solar irradiation. The chamber was equipped with multiple sampling ports, which allowed the introduction of clean air, seed aerosols, gas-phase reagents, and for measurements of both gas-phase and particle-phase compositions. The reaction setup used in the present

15   study and the schematic of the smog chamber facility is shown in Fig. S1 (Liu et al., 2017; Jia and Xu, 2014, 2016; Ge et al., 2017b, a).

The zero air (generated from Zero Air Supply, Model 111 and Model 1150, Thermo Scientific) was used as the background gas for the experiments, and three big hydrocarbon traps (BHT-4, Agilent, Santa Clara, CA, USA) coupled with three activated-carbon filters were used to get further purified air. The chamber was cleaned by flowing pure air for at least 4

20   h prior to each experiment until residual NO, NO_2, O_3 or any other particles could not be detected in the chamber. The prepared gas-phase furan was introduced into the chamber directly by a syringe. A certain amount of NO_2 was injected into the chamber by a gas-tight syringe to achieve the initial NOx level. For the enhancement of the SOA formation, NaCl seed aerosols were injected into the chamber via atomization of NaCl aqueous solution with a constant-rate atomizer (Model 3076, TSI, USA). An aerosol neutralizer (Model 3087, TSI, USA) was used to bring particles to a steady-state charge distribution

25   before they were introduced into the reactor. The initial seed number and mass concentrations were approximately 5×10^4 cm^{-3} and 6 µg m^{-3} on average, respectively. For different RH conditions, the dry zero air was introduced into a bottle of high-purity water to control the humidity of the background air. The humidity in the chamber was detected with a hygrometer (Model 645, Testo AG, Lenzkirch, Germany). The initial experimental conditions considered in this study are summarized in Table 1. Typically, approximate 750 ppb furan was employed. To study the effect of NOx levels on SOA formation,

30   experiments were conducted with initial furan/NOx ratios ranging from 7.8 to 48.1, whereas to develop and test the role of RH, experiments were performed for RH values varying from 5 to 85% and with an initial average furan to NOx ratio of 7.55. The temperature of the chamber was controlled to be 308-310 K during all the experiments.

Particles and gas-phase species may get lost to the chamber walls on short timescales, hereby influencing the gas-phase chemistry and SOA formation. The wall loss rate constants for O_3, NOx and aerosol particles were 3.3×10^{-7} s^{-1}, 4.1×10^{-7} s^{-1}, and 3.6×10^{-5} s^{-1}, respectively, which were detected from our previous study conducted in the same set-up and similar experimental conditions (Ge et al., 2017a). However, no wall loss of furan was observed within the uncertainty of the detection of the instrument. Furthermore, the light intensity in the reactor was determined to be 0.34 min^{-1} by using the NO_2 photolysis rate constant as the indirect representation. Before the start of each experiment, the static electricity of the Teflon chamber was removed. After all the reactants flowed into the chamber, the reactor was maintained in the dark for at least 40 min without any activities to reach homogeneous mixing.

2.2 Gas and particle monitoring

The concentrations of O_3 and NOx were measured continuously by UV photometric O_3 analyzer (Model 49C, Thermo Environmental Instruments Inc.) and a chemiluminescence analyzer (Model 42C, Thermo Electron Corporation, USA), respectively. The sampling flow rate was 0.75 L min^{-1} for both O_3 and NOx analyzers. The uncertainty in the gas-phase measurements was below ± 1 %. The experiment samples were collected by Tenax absorption tubes (150 mm length × 6 mm O.D., 0.2 g sorbent). Furan concentrations and product samples were detected by thermal desorption-gas chromatography-mass spectrometry (TD-GC-MS): the thermal desorber (Master TD, Dani, Italy) was combined with gas chromatography (Model 7890A, Agilent Tech., USA) interfaced to a mass selective detector (5975C, Agilent Tech., USA). The initial reactants concentrations are given in Table 1. The chromatographic analytical method used for the analysis was as follows: the chromatographic column was kept at 338 K for 4 min, then heated to 598 K at a heating rate of 20 K min^{-1} and held for 5 min.

The particle size distribution and mass concentration were measured with a scanning mobility particle sizer (SMPS, Model 3936, TSI, USA) composed of a TSI 3081 differential mobility analyzer (DMA) and a TSI 3776 condensation particle counter (CPC). The sampling and sheath flow rates of the SMPS were 0.3 L min^{-1} and 3 L min^{-1}, respectively. Given this flow conditions of the SMPS, particle sizes (in counts cm^{-3}) were recorded in 64 size bins for particles in the 13.6 to 710.5 nm size range. The accuracy of the particle number concentration is ±10 % at $<3 \times 10^5$ cm^{-3}. The method of the reduced Dry-Ambient Size spectrometer (DASS) was used to detect the ALW (Engelhart et al., 2011). To get dry particle mass concentrations, a large diameter Nafion dryer (Permapure MD-700-48F-3) and a multi-tube Nafion dryer (Permapure PD-200T-24E-M) were added to the sampling inlet and sheath flow, respectively. This DASS method is generally accepted and widely used in the detection of ALW content, and can remove up to 90% of the water vapour without losing the organic-bound water. It should be noted that the dissloved water-soluble species would evaporate back into the gas phase during the ALW measurement when the aerosol water is removed. In fact, the repartitioning of water-soluble components between gas and particle phases was not taken into consideration. The SOA concentrations for high RH conditions were slightly underestimated, but the underestimation is extremely low and can be negligible. The results were compared with the mass concentrations detected in humid mode, in which the humid air in the chamber was introduced into the SMPS with the

sheath air set at 10 L min^{-1}. The ALW was calculated as the difference between the particle mass concentrations determined at dry and humid modes (Jia and Xu, 2018). On the basis of the recorded particle volume concentration and assumed particle density, a total suspended particulate (TSP in μg m^{-3}) could be translated. As shown in Eq. (1), TSP is the sum of seed aerosol (NaCl in this work), NaNO$_3$, particle ALW, and SOA. By dissolving the SOA collected on ZnSe disks into high

5 purity water, an ion chromatograph (IC, Dionex ICS-900, Thermo Fisher, USA) was then used to analyze the inorganic content (Cl$^-$ and NO$_3^-$) in SOA. With known initial NaCl seed aerosols and their decay rates, the particle mass concentration of NaCl can be calculated, and based on the detected ALW during experiments, the SOA produced in the photooxidation of furan can be estimated from Eq. (1).

$$TSP = NaCl + NaNO_3 + ALW + SOA \qquad (1)$$

10 **2.3 Chemical characterization of products**

  The SOA particles were sampled on ungreased ZnSe disk (25 mm in diameter) using a Dekati low-pressure impactor (DLPI, Dekati Ltd., Kangasala, Finland) after the appearance of the O$_3$ maximum concentration. The sampling flow rate of DLPI was 10 L min^{-1} with the particles sizes from 30 nm to 10 μm, classified into 13 stages. According to the particle distribution explored from the SMPS, the SOA particles size was mainly around 150 nm after 3 h of irradiation. The ZnSe

15 disk was placed on stage 3 to reach the maximum collection of the particles. Afterward, the ZnSe disk was put in the FTIR (Nicolet iS10, Thermos Fisher, USA) sample holder, which had been flushed with N$_2$ to eliminate the impact of ambient H$_2$O and CO$_2$ on the determination of the chemical composition of formed SOA particles. The spectra were recorded at 4 cm^{-1} resolution with 128 scans. The samples collected on the ZnSe disk were then dissolved with high purity water for analysing inorganic species.

20   To obtain chemical characterization information, the ultrahigh performance liquid chromatograph (UPLC, Ultimate 3000, Thermo Scientific, USA) -heated-electrospray ionization-high-resolution orbitrap mass spectrometer (HESI-HRMS, Q Exactive, Thermo Scientific, USA) was used for the analysis of organic compounds produced from the photooxidation of furan. Methanol (Optima™ LC/MSGrade, Fisher Chemical, USA) was used as the eluent in UPLC system. The acquired mass spectrum of SOA was in the range of 30 80-1000 Da. As a "soft ionization", HESI can provide molecular formulae for

25 thousands of OA substances when coupled with a high-resolution mass analyzer. The HESI source was used in both positive and negative ion modes using the optimum method for characterization of organic compounds. A total volume of 300 L was sampled at a flow rate of 0.2 L min^{-1}, and a volume of 20 μL of the extraction was used for the measurement. The addition of acetonitrile allows a lower surface tension of the solution and provides a stable electrospray ionization process (Koch et al., 2005). N$_2$ was used as the sheath and auxiliary gas. The desolvation gas temperature was 320 ℃, and the gas flow was 200

30 μL min^{-1}. The capture voltage was set to 3 kV. The Thermo Scientific Xcalibur software (Thermo Fisher Scientific Inc., USA) was used to analyze the data from HRMS.

3 Results and discussion

3.1 Gas phase and SMPS results

Experiments were divided into two main groups, which were set to study the role of (i) NOx and (ii) RH on SOA formation. To keep the variables of each set of conditions consistent, almost identical initial gas phase conditions were required. The addition of furan and seed aerosol in all the experiments was nearly equal. The first set of experiments (Exp.1-6, Table 1), designed to investigate the effect of NOx level on SOA formation, used similar RH conditions with different NOx concentrations. To further assess the effect of OH produced during the furan-NOx photooxidation, four experiments (Table S1, Exp.13-16) were conducted by injecting H_2O_2 in the chamber before the experiments started. For Exp.6-12, different RH levels coupled with similar furan/NOx ratios were monitored to assess the RH impact on SOA formation. To analyse the SOA composition, five additional experiments (Table S2, Exp. 17-21) were carried out to analyse the HESI-Q Exactive-Orbitrap MS, which intended to compare the role of NOx and RH on the SOA formation. The major gas-phase inorganic chemical reactions occurred during the experiments are presented in Fig. 1. After turning on the light, the photolysis of NO_2 produced NO and O, which further reacted with O_2 to form O_3 in the chamber. As a result, a slight increase in the beginning, followed by the decrease of NOx concentrations are observed (Fig. 2). On one hand, the decrease of NOx is caused by the reaction between NO and alkyl RO_2, which are formed from the OH-initiated oxidation of furan during the experiment. On the other hand, the oxidation of NO by O_3 also contributes to the NOx decreasing trend.

The photochemical oxidation of VOCs in the presence of NOx and OH radical is an important source of tropospheric O_3 (Labouze et al., 2004). Although the atmospheric degradation mechanisms of furan is complex, the fundamental features of troposphere O_3 formation is relatively simple. The main formation pathway of O_3 in the atmosphere is the photolysis of NO_2 as presented in Fig. 1 (R1 and R2). The RO_2 and HO_2 radicals generated from the photooxidation of furan can further react with NO to form NO_2 (R5 and R6), whose photolysis will produce the ground-state oxygen atom (R1) and thus contribute to the net O_3 formation through the reaction of O_2 with O (R2). The very similar VOCs level may result in a high O_3 production with higher NOx concentrations conditions (Bowman and Seinfeld, 1994). As presented in a previous work of the ethylene-NOx-NaCl irradiations, both experimental results and model simulations showed an abundant O_3 production (approximately 500 ppb) with an initial NOx level of 142 ppb (Jia and Xu, 2016). Therefore, it is important to evaluate the O_3 production potential of VOCs in the atmosphere. As shown in Fig. 2, it shows a consistent increasing trend between O_3 concentration and SOA mass concentration, indicating that O_3 was generated during the furan photooxidation reactions with the formation of SOA. However, it should be noted that the apparent high O_3 concentration monitored by the O_3 analyzer was due to species other than O_3 and that have absorption around 254 nm. The generated gas-phase reaction products that have carbonyl, carboxylic acid, and ester groups conjugated with C=C bonds can absorb strongly in the ~200-300 nm range and, accordingly, contribute to the apparent intense signal of O_3 detected by the O_3 monitor (Strollo and Ziemann, 2013).

The O_3 maximum concentration, SOA mass concentration and SOA yield dependences on the C_4H_4O/NOx ratio are presented in Fig. 2. The results show that the O_3 maximum concentration produced during the experiments decreases with

increasing C_4H_4O/NOx ratio. When the C_4H_4O/NOx ratio decreases from 16.9 to 7.8, there is a sharp increase in O_3 concentration from 197 to 382 ppb. Interestingly, experiments for which the C_4H_4O/NOx ratio changed from 48.1 to 36.60 exhibit a change of 24 ppb in O_3 maximum concentration. This result indicates that under different C_4H_4O/NOx ratio regimes, the concentration of O_3 produced varies much. There is a great change in O_3 accumulation when the C_4H_4O/NOx ratios are relatively low (<16.9), which is consistent with a previous observation from C_3H_6-NOx-NaCl irradiation experiments by Ge et al. (Ge et al., 2017a) who found that when the C_3H_6/NOx ratio was less than 11, the O_3 concentration decreased considerably with increasing ratio, whereas when the ratio was larger than 11, the O_3 concentration slightly decreased with increasing ratio. Fig. S2 shows the comparisons of the observed concentrations of the apparent O_3 and NOx from furan irradiations at different experimental conditions. With similar NOx level, the amount of O_3 formed at 5% RH was larger than that at 42% RH. The apparent O_3 maximum concentration was reduced by almost 60% as the C_4H_4O/NOx ratio increased from 7.8 to 36.6. However, the effect of RH on O_3 formation is not so obvious compared to the difference in initial C_4H_4O/NOx ratio. The influence of NOx level on O_3 is likely related to its formation mechanism. The formation of O_3 is directly connected with the NOx in two ways: the photolysis of NO_2, producing the O atom that reacts with O_2, and the NOx reacting directly with RO. However, the RH has little to no effect on O_3 formation. This formation is slightly favoured at low RH whereas, at high RH, the ONO_2- are easily transferred into the aerosol phase, thereby suppressing the O_3 formation (Jia and Xu, 2014). Additionally, the slight change in O_3 maximum concentration under different RH conditions may also be caused by the consumption of gas-phase reaction products that contain functional groups conjugated with C=C bonds and respond to the O_3 analyser. This assumption is reasonable because these carbonyl-rich products were favourable to condense on the moist surface of particles and thus lowered the O_3 concentration detected by the O_3 analyser, which has been further confirmed by the MS results with enhanced intensities of corresponding organonitrates. The wall loss of the organonitrates species would also explain in the decreasing trend of O_3 concentration as RH increases. The appearance time of O_3 maximum concentration at high C_4H_4O/NOx ratio is almost half an hour earlier than that at low initial C_4H_4O/NOx ratio. On the contrary, there is no significant difference in the balance time of NOx concentration under different experimental conditions. Similar initial C_4H_4O/NOx ratios result in a similar profile of NOx concentration during the experiment, although the final NOx concentration shows a slight difference.

The calculated SOA yield is defined as the ratio of the mass concentrations of the maximum SOA formed (M_{SOA} in µg m^{-3}, corrected by wall loss) and the reacted furan (ΔC_4H_4O in µg m^{-3}):

$$Y_{SOA} = (M_{SOA}/\Delta C_4H_4O) \times 100\% \tag{2}$$

where M_{SOA} is obtained by subtracting the amount of NaCl and $NaNO_3$ from the total particle mass concentration measured by the SMPS, and ΔC_4H_4O is the consumption of furan during the experiments. Experiments were started with an average of 6 µg m^{-3} of NaCl seed aerosol and were conducted under different initial C_4H_4O/NOx ratios. The dependences of the SOA mass concentration and SOA yield on the C_4H_4O/NOx ratios and RH conditions are shown in Fig. 3 and Fig. 4,

respectively. The numerical values of the aerosol mass concentration and SOA yield are given in Table 1. After the photooxidation reactions, the SOA mass concentrations reached maxima between 1.0 and 23.5 μg m^{-3}, with SOA yield ranging from 0.04% to 1.01%. In other experiments carried out under varying RH conditions, SOA maxima were in the range of 23.5-119.2 μg m^{-3}, with SOA yield ranging from 1.01% to 5.03%.

5   It is generally accepted that experiments with low NOx levels lead to higher SOA yields than those with higher NOx levels at the same VOC concentration (Song et al., 2005). However, as shown in Fig. 3, an increasing SOA mass concentration and SOA yield with increasing NOx was observed. There are two possible explanations to this phenomenon: (i) the concentration of OH radicals produced *in situ* in the present study before additional source of OH was insufficient to produce a considerable amount of SOA under low-NOx conditions. As shown in Fig. S3, the OH concentration exhibits a

10   gradual increase with NOx concentration and there appears to be a correlation between NOx concentration, OH concentration and SOA yield. Therefore, at low-NOx conditions, the increase of SOA yield was attributed to an increase of OH concentration, which was affected by OH recycling following reaction (R6) (see Fig. 1) and contributed to the enhancement of SOA formation. This result is consistent with a previous study concerning the impact of NOx and OH on SOA formation from β-pinene photooxidation, which has proved that the positive correlation between SOA yield and NOx

15   levels ([VOC]$_0$/[NOx]$_0$ > 10 ppbC/ppb) was caused by the NOx–induced increase of OH concentration (Sarrafzadeh et al., 2016b). (ii) Differently, Sarrafzadeh et al. found that after eliminating the effect of OH concentration on SOA mass growth, SOA yield only decreased with increasing NOx levels (Sarrafzadeh et al., 2016). To further investigate the NOx effect on furan generated SOA formation under adequate OH conditions, four more experiments (see Table S1) were carried out with additional injection of H$_2$O$_2$ as the OH radical source before the start of each experiment. The SOA yield trend at different

20   C$_4$H$_4$O/NOx ratios is also shown in Fig. 3. The continuous growth trend of SOA yield with increasing NOx concentration at a relative high NOx level may result from the partitioning of generated semi/low-volatility compounds (multifunctional nitrates and dinitrates) into the particle phase, leading to significant furan SOA formation under high NOx conditions. Similarly, SOA from OH-initiated isoprene oxidation under high-NOx conditions was comprehensively investigated by Schwantes et al., who suggested that low volatility hydroxyl nitrates and dihydroxyl dinitrates generated conspicuously more

25   aerosol than previously thought (Schwantes et al., 2019). Our results showing the increase of SOA mass formation by high-NOx conditions also agree with a previous study, which indicated that a high level of NO$_2$ can participate in the OH-induced reaction of guaiacol, consequently leading to the formation of organic nitrates and the enhancement of guaiacol SOA formation (Liu et al., 2019a).

    It is worth note that under high RH conditions, as shown in Fig. 1, the NO$_2$ hydrolysis (reaction (R11)) can generate

30   nitrous acid (HONO), which has been considered as a major source of OH. As indicated in Fig. 4, the SOA yields obtained in the present work clearly show a gradual increase with RH. Also shown in Fig. S3 is the dependence of OH and furan concentrations on RH during the experiments determined from the decay of furan using a reaction rate coefficient of k(OH+furan) = 4.01× 10^{-11} cm^3 molecules^{-1} s^{-1} (Atkinson et al., 1983). It is therefore probable that the increase of RH results in high levels of HONO formation in the chamber, which leads to an increase in OH concentration, a faster furan decay rate,

3.2 SOA chemical composition

To get detailed information on the functional groups in SOA formed during the photooxidation of furan, the collected particles were measured by FTIR, which has been proven to be an ideal technique for the detection of functional groups and bond information in aerosol samples. The FTIR spectra of particles collected from furan-NOx-NaCl with different values of

10   C_4H_4O/NOx ratio are shown in Fig. 5. The absorptions of organic functional groups were detected, which further confirmed the SOA formation from the photooxidation of furan. The assignment of the FTIR absorption frequencies is summarized in Table 2. The organic nitrate exhibits typical NO symmetric stretching at 868 cm^{-1}, NO_2 symmetric stretching at around 1341 cm^{-1}, and NO_2 asymmetric stretching at 1614 cm^{-1} (Jia and Xu, 2016). The absorption at 1067 cm^{-1} matches the C-O stretching vibration in C-O-C, while the sharp absorption at 1724 cm^{-1} is the C=O stretching vibration in carboxylic acid and

15   ketones (Sakamoto et al., 2013). The carbon skeleton corresponds to the vibrations between 2850 cm^{-1} and 3100 cm^{-1}, where the ν(C-H) stretching vibration can be found. Absorptions in the ranges 2850-3000 cm^{-1} and 3000-3100 cm^{-1} represent the C-H stretching vibration in saturated carbon ring and unsaturated alkenes, respectively. Correspondingly, the C-H bending vibrations are represented by the absorption between 1350 and 1515 cm^{-1}. The strong broad vibrations at 2400-3500 cm^{-1} are interpreted as the O-H stretching vibration in carboxyl groups and hydroxyl groups (Ge et al., 2017a).

20   Fig. 5 illustrates the influence of NOx level on the SOA components during the photooxidation of furan. For each of the functional groups, experiments with the most pronounced SOA formation are those with the lowest initial C_4H_4O/NOx ratio. This is further confirmed by comparing the SOA yield for different NOx level conditions under approximately similar gas-phase and initial seed conditions. The calculated variations of the relative abundance of FTIR functional groups at different NOx levels are presented in Fig. S4, in which the absolute abundances are normalized with respect to the corresponding

25   functional abundance detected at $C_4H_4O/NOx=7.8$. The carbonyl compound formed with 7.8 initial C_4H_4O/NOx ratio is approximately thrice more abundant than that formed with 48.1 initial C_4H_4O/NOx ratio. The absorbance of NO_2 functional groups exhibits a much stronger enhancement under initial C_4H_4O/NOx ratio of 7.8 compared to 48.1. Although at different C_4H_4O/NOx ratio conditions the intensities of C-O-C and O-H functional groups show similar trends, their variations are substantially different from the variations of other functional groups. In sum, the increased absorbance of functional groups

30   with decreasing initial C_4H_4O/NOx ratios demonstrates that relatively high NOx level contributes to the formation of SOA.

The FTIR results in the investigation of furan photooxidation under different RH conditions are shown in Fig. 6. The absorbance of FTIR characteristic peak increases with the RH raising from 5% to 85%. Note that obvious intensities of the functional groups were observed when the RH exceeded the efflorescence RH of NaCl. In contrast, the absorptions of

corresponding functional groups enhance gently when the RH is lower than 42%. This phenomenon is consistent with the results of the SOA yield discussed above and, can also be interpreted as the increasing ALW components contributing to SOA formation. The ratio of the absorbance intensities detected at low RH to that at 85% RH is used as the relative abundance to show more intuitive FTIR results. As shown in Fig. S5, for RH between 5% and 37%, the intensities of all functional groups vary weakly and are approximately one-third of the intensity at 85% RH. However, when the RH rises from 42% to 85%, the absorption intensities of O-H and NO_2 functional groups increase by factors of 2.0 and 1.9, respectively. In this RH range, the variations of relative intensities are even stronger for C=O, C-H and C-O-C, being increased by factors of 2.3, 2.3 and 2.5, respectively. A previous study has observed that in an urban environment containing aromatic hydrocarbons and NOx, the SOA yield increased by a factor of 2 to 3 under high RH compared to lower RH (Zhou et al., 2011). This trend is similar to that found in experiments with toluene in an urban aromatic hydrocarbon-NOx mixture system (Kamens et al., 2011). It should be pointed out that observed FTIR results show a high degree of consistency with the SOA mass concentration and SOA yield.

To further identify and confirm the structure of generated SOA components, more techniques, such as GC/MS and HESI-Q Exactive-Orbitrap MS were used to analyze the chemical composition. The chromatography before ESI-MS analyzes is helpful to remove the disturbances of the inorganic salts and determine the appropriate molecular compound (Surratt et al., 2007; Gao et al., 2006). However, the collected PM products were not sufficient for the chromatography owing to the limitation of the chamber volume. Furthermore, the low SOA yield of the furan photooxidation make it harder to remain enough SOA components for the better response of ESI-MS signal. Consequently, direct infusion analyzes were carried out for HESI-Q Exactive-Orbitrap MS in the present study. Five experiments of furan-NOx-NaCl photooxidation that were conducted under different initial C_4H_4O/NOx ratios and RH conditions were analyzed by HESI-Q Exactive-Orbitrap MS. The mass spectra recorded in different ion modes represent the detected compounds ionisable in either positive or negative modes (Walser et al., 2008). The MS spectra of generated species from different NOx level and RH conditions, which show evidence for the OH-furan reaction, are presented in Fig. 7 and Fig. 8, respectively. The major peaks are m/z^+ 85.0018, 101.0894, and 185.0504 in the positive ion mode, and m/z^- 146.0161, 225.0125 and 263.0132 in the negative ion mode. The prominent peaks in the HR-MS spectra detected on negative ion mode are comprised of various functionalized hydroxyl nitrates and dihydroxyl dinitrates. However, in the positive ion mode analysis, most carbonyl components were detected. The assignments of these ion peaks, the molecular weights of the products observed, and proposed structures are summarized in Table 3. These detected compounds provide additional evidence for the proposed radical reaction mechanism.

According to the identified products in this work and based on previous kinetic (Atkinson et al., 1983; Lee and Tang, 1982) and products (Villanueva et al., 2009; Aschmann et al., 2014; Tapia et al., 2011; Villanueva et al., 2007; Strollo and Ziemann, 2013) studies, a proposed chemical mechanism for SOA formation from furan in the presence of NOx is shown in Scheme 1. Additionally, on the basis of well-established mechanisms for atmospheric volatile organic compounds (Atkinson and Arey, 2003), it can be concluded that the reaction is initiated by OH addition to a C=C bond at C2 or C3 positions. Addition at the C2 position forms two cyclic alkyl radicals (a, c), one of which (a) can isomerize to form a ring-opened alkyl

radical (d), whereas addition at the C3 position forms a single alkyl radical (b). The OH radical addition leading to an hydrogen abstraction generates the alkyl radicals (R·) followed by reaction with O_2 to form alkylperoxy radicals (RO_2·) (Pan and Wang, 2014). Moreover, the formed alkylperoxy radicals (RO_2·) can either react with RO_2/HO_2 or NO to yield the corresponding alkoxy radical (RO·), which can (i) decompose and then react with O_2 to yield a 1,4-aldoester (A), (ii) react with $NO_2/NO/O_2$ to form hydroxyl nitrate compound isomers with m/z^- 146, or (iii) react with O_2 to form unsaturated products and hydroxyfuranone (B) and 1,4-aldoacid (C). The ring-opened alkylperoxy radical generated from (d) can decompose to generate an unsaturated 1,4-dialdehyde (D). The formation of 1,4-dialdehyde with m/z^+ 85 detected in the positive ion mode suggests that these unsaturated 1,4-dicarbonyls are formed after initial OH addition at 2- or 5-positions. We note that OH radical addition at 2, 3-positions would lead to carbonyl product isomers with same m/z^+ 101. In addition, some dihydroxyl dinitrates with m/z^- 225 were also detected in negative ion mode. However, the pathways favouring the generation of these dihydroxyl dinitrates could only take place under high NOx levels. Scheme 1 shows the formation of second-generation products hemiacetals (E) via the reactions of the hydroxyfuranone (B) with 1,4-dialdehyde (D). After uptake from the gas-phase, the combination of hydroxyfuranone with 4-dialdehyde appeared to occur by H-abstraction, followed by dehydration, thus forming m/z^+ 185 compounds. This reaction pathway has also been identified in previous studies of OH-initiated reactions of furans (Strollo and Ziemann, 2013; Aschmann et al., 2014). According to the results of HR-MS, this aqueous phase reaction is more favoured in aqueous particles.

3.3 Effects of NOx on SOA formation

To study the NOx level effect on SOA formation from the photooxidation of furan, the experiments were conducted with varying initial C_4H_4O/NOx ratios ranging from 7.8 to 48.1. The SOA formation is found to have much lower yield under a high C_4H_4O/NOx ratio. The SOA mass concentration and SOA yield increased from 1.0 to 23.5 μg m^{-3} and 0.04% to 1.01%, respectively, as the initial concentration ratio of C_4H_4O/NOx decreased from 48.1 to 7.8 (ppbC/ppb). This trend is consistent with previous studies on propylene photooxidation, which found that the SOA yield was enhanced under a low VOC/NOx ratio (Ge et al., 2017a). Another laboratory study of aerosol assessment from isoprene photooxidation concluded that under low NOx level (< 129 ppb), the SOA mass increased with increasing initial NOx level (Kroll et al., 2006). By comparing the studies of Ge et al. and Kroll et al., there is a positive correlation between NOx effect and SOA formation when the initial NOx level is relatively low (nearly below 100 ppb). These results further support our findings in the experiments of furan photooxidation. As shown in Fig. S6, SOA formed under a low C_4H_4O/NOx ratio also have a faster increase than those under large C_4H_4O/NOx ratio conditions. However, the particle number distribution under C_4H_4O/NOx=7.8 and C_4H_4O/NOx=36.6 conditions shows similar profiles.

Generally, the NOx level has two different effects in the reaction process. Firstly, increasing NOx concentration will promote the O_3 and HONO formation, leading to more OH radical formed, which in turns is favorable to SOA formation (Sarrafzadeh et al., 2016). In addition, sufficient NOx can facilitate the competition between NO and HO_2 to react with RO_2. Products with high volatility will be generated more by the NO+RO_2 reaction than by HO_2+RO_2 reaction (Kroll and Seinfeld,

2008). However, the formation of lower volatility products favors the increase of SOA yields (Chen et al., 2018). In this regard, the increasing NOx level is not conducive to SOA formation. It was shown that the yields of SOA generated from the photooxidation of *m*-xylene increased firstly and then decreased with the increase of the NOx level (Chen et al., 2018). In the present study, an increasing trend of SOA formation was observed with the increase of NOx concentration. As shown in the HESI-Q Exactive Orbitrap-MS results, all the detected primary products are carbonyl-rich, and even the organonitrates have at least two carbonyl functional groups. These carbonyl-containing products have lower volatility and contribute to the SOA formation. The peak intensities in the MS of the products (m/z^+ 85, 101) generated by the pathways involving HO$_2$ (as indicated by the scheme 1) decreased with the increase of NOx concentration. Additionally, more products of dihydroxyl dinitrates (m/z^- 225) with multifunctional groups were detected under high NOx conditions. As shown in Scheme 1, the multifunctional organonitrates detected in negative ion mode are produced mostly from later-generation chemistry. Hydroxyl nitrates with m/z^- 149 can be formed through pathways 1-a, 2-a, and 3-a by the reaction of RO$_2$ with NOx. We note that the m/z^- 149 compound was detected both at low NOx levels and high NOx levels. However, the peak intensity of this product was decreased with increasing NOx concentration. This phenomenon might be caused by the later-reaction of the unsaturated hydroxyl nitrates going through a second OH-initiated reaction and leading to the formation of the dihydroxyl dinitrate with m/z^- 225. In addition, the peak intensities changes of SOA products detected in the positive mode, such as, the peaks at m/z^+ 85 and 101 were reduced under high NOx conditions, which resulted from the fact that the RO$_2$ radical fate was dominated by the pathway of RO$_2$+NO or RO$_2$+NO$_2$. This result supports that the fate of RO$_2$ is not a single channel reaction. There exists a competition between RO$_2$ reacting with NOx and with HO$_2$ under high NOx conditions but the former pathway is more favourable. There are two pathways for hydroxyl nitrates formation from RO$_2$ radicals in the presence of NOx according to which RO$_2$ radicals may react with NO and NO$_2$ to form RONO$_2$ and ROONO$_2$, respectively (Kroll and Seinfeld, 2008). However, the formed peroxynitrates could easily thermally dissociate and convert to RONO$_2$.

Furthermore, by analysing the OH concentration and products components, we conclude that there are two possible explanations for the increasing trend of SOA yield as the NOx level increases: (i) the SOA production is closely related to the oxidation capacity in the photooxidation experiments. Experiments conducted under different NOx levels indicate that the OH concentration is controlled by the NOx level if there is no additional OH precursor added before the start of the experiment. As shown in Fig. S3, an increase in NOx level results in more OH generation and a faster furan decay rate. This justifies the observed higher SOA mass concentration and higher SOA yield. (ii) HRMS fragments associated with multifunctional organonitrates are enhanced under high NOx conditions (Fig. 7). As presented in Scheme 1, the furan dihydroxyl dinitrates are generated from the first-generation hydroxyl nitrate reacting with OH to form a peroxy radical, which reacts thereafter with NO. Together with multifunctional hydroxyl nitrates, these low-volatility species can easily partition into the particle phase and increase the SOA mass concentration. More importantly, the seed particle added initially plays an important role in the processes of gas-particle partitioning as indicated by a recent study, which showed that sufficient seed surface area at the start of the reaction largely suppressed the effects of vapour wall losses of low-volatility

compounds (Schwantes et al., 2019). Therefore, the NaCl seed particle added in the present work promoted the partitioning of the formed low-volatility functional organonitrates.

3.4 Effect of RH on SOA formation

Experiments 6-12 were conducted under seven different RH conditions ranging from 5% to 85%. In this RH range, the SOA yield increases from 1.01% to 5.03%. With almost identical initial conditions except RH, the yield of furan-derived SOA formed at high RH can be a factor of two higher than that formed at low RH. A similar trend was also observed by Yu et al., who found that the SOA mass concentrations increased by a factor of six when RH increased from 18% to 82% (Yu et al., 2011). As shown in Fig. S3, an increase in RH leads to higher OH concentration resulting from higher HONO levels generated by the reaction of NO_2 with H_2O. Previously, Anglada et al. confirmed, using quantum mechanical calculations, that the water component could increase the OH production (Anglada et al., 2011). The positive correlation between initial water vapour concentration and OH concentration has also been previously observed experimentally (Healy et al., 2009; Tillmann et al., 2010). Additionally, Healy et al. have also reported that increasing OH concentration promoted the decay of VOC and enhanced SOA formation (Healy et al., 2009). Similarly, in the present work, a faster decay rate of furan was also observed as RH increased, as shown in Fig. S3. It is possible that the faster rate of gas phase oxidation under higher OH concentrations will lead to the generation of less volatile compounds as presented previously (Chan et al., 2007). A higher OH concentration promotes oxidation reactions, influences the distribution of organic products, and facilitates the SOA formation (Sarrafzadeh et al., 2016).

An obvious increase of SOA yield was observed when the RH increased from 37% to 54%. This phenomenon was mainly caused by the efflorescence transition when the seed particles were coated with SOA. It has been previously shown that SOA formation decreases both the efflorescence RH and deliquescence RH of the seed particles and results in the uptake of water by the particles (Liu et al., 2018; Takahama et al., 2007; Smith et al., 2012). It is highly possible that the NaCl seeds effloresce and deliquesce early after being coated by the new formed SOA. The effect of efflorescence contributes to the water uptake by the particles, leading to the obvious trend-changing of SOA yield. It is noted that with the NaCl seed aerosols serving as nuclei, the ALW was high at high RH. Products with water solubility produced from the photooxidation of furan can dissolve into the ALW of aerosol particles. As a result, ALW in the formed aerosols plays an important role in gas/particle partitioning. As shown in Fig. 4, the ALW was detected when the RH was higher than 54%, which was based on the deliquescence of NaCl under high RH conditions. The increase of ALW could partially explain the increase of SOA mass concentration and SOA yield. It is highly probable that the particle surface area increases with increasing amount of ALW as shown in Fig. S7, which likely promotes the dissolution of semi-volatile matters produced during the experiments. According to the HESI-Q Exactive-Orbitrap MS results shown in Fig. 8, the intensities of multifunctional hydroxyl nitrates and dihydroxyl dinitrate (m/z^- 146 and 225, respectively) exhibited positive correlations with RH. Slight peak intensities increases of m/z^+ 85 and 101 products were also observed under high RH conditions. This phenomenon indicates that the gas-particle phase partitioning of low-volatility compounds was enhanced under high RH conditions. Furthermore, the

increased surface area under high RH conditions may also be attributed to the condensation of the produced multifunctional compounds.

Another possibility for the increasing trend of SOA yield with the increase of RH might result from the SOA formation through aqueous chemistry in wet aerosols (Grgic et al., 2010; Lim et al., 2010). In these atmospheric processes, alcohols, aldehydes, and ketones formed from the photooxidation of furan in the gas phase can be absorbed into the humid surface of the hygroscopic SOA at high RH. This process further contributes to the formation of low-volatility products on the SOA surface. In addition, the aqueous photochemistry of highly soluble small compounds that portioned in ALW could produce additional organic compounds and result in larger SOA yield under high RH conditions (Faust et al., 2017; Jia and Xu, 2014). The appearance of m/z^+ 185 and m/z^- 262 detected by the HESI-Q Exactive-Orbitrap MS further demonstrated that aqueous phase reactions indeed took place under high RH conditions. As shown in Scheme 1, the peak of m/z^+ 185 could form by aqueous phase reaction of the hydroxyfuranone (B) and 1,4-dialdehyde (D). The formation of hemiacetal (E) has also been detected by a previous study of the OH-initiated reaction of 3-methylfuran in the presence of NOx (Strollo and Ziemann, 2013). The proposed hemiacetal compound (E) plays a substantial role in the obvious increase of m/z^+ 185 product formation under high RH conditions. A pathway of organonitrate (m/z^- 262) formation in the aqueous particles with the presence of NO$_3^-$ was suggested in the present study based on a previous work, which indicated a radical-radical reaction pathway for organosulfate formation from aqueous OH oxidation of glycolaldehyde in the presence of sulfuric acid (Perri et al., 2010). Increasing the RH also resulted in an overall addition of peak intensities in the negative ion mode, due to the fact that the sample obtained at high RH during the SOA generation had a larger particle surface. Specifically, a relatively stronger intense band of $C_8H_8O_9N^-$ compound (m/z^-=262) was found under a high RH. Consequently, the heterogeneous products in wet seed particles will further contribute to the formation of SOA, because higher aerosol liquid water content enables more aqueous phase reactions.

In conclusion, the reasons for the increasing trend of SOA formation under high RH conditions may be summarized as following: firstly, higher concentration of OH radical will certainly promote the SOA formation as RH increases. A faster decay rate of furan will also contribute to the formation of products that can partition into the particle phase. In addition, it is possible that the aqueous surface of seed particles provides a new substrate for the photooxidation of furan. Previously, N$_2$O$_5$ and HNO$_3$ have been proven to be the key products in the VOC-NOx irradiation experiments (Wang et al., 2016). The moist surface under high RH conditions is more favorable for the condensation of products with low vapor pressure, leading to the increasing production of SOA formation. The high RH environment favors the formation of the hemiacetal compound. Moreover, the effect of RH on SOA formation in furan photooxidation can also be determined by the aqueous photochemistry under high RH conditions as discussed above. The aqueous phase reactions at the surface of particles promotes the formation of hemiacetal-like products, which likely plays an important role in the process of SOA formation. Previously, unsaturated first-generation reaction product of 3-methyl furan has also been suggested to undergo acid-catalyzed condensed-phase reactions, with SOA yields up to 15% (Strollo and Ziemann, 2013). In addition, the reinforced

effect of RH on SOA yield was also ascribed from the photooxidation of other aromatic compounds, such as, benzene (Ng et al., 2007b), toluene (Hildebrandt et al., 2009; Kamens et al., 2011), and xylene (Zhou et al., 2011).

4 Conclusion

The effects of NOx and RH on SOA formation from the photooxidation of furan in the presence of NaCl seed particles have been investigated in this study. The results demonstrated that the formation of SOA were promoted when the initial VOC/NOx ratio decreases. The increase of SOA yield at high VOC/NOx ratio was caused by the NOx-induced increase of OH concentration. The reason for the promotion of SOA mass at a low furan/NOx ratio is that more low-volatility products were generated from the furan photooxidation, which contributes to the formation of SOA. Additionally, the increase of RH results in the increase in the mass concentration of the produced SOA. The mechanisms controlling SOA formation may include the gas-phase photooxidation of furan, physical water uptake as RH increases, and the gas-phase reaction of water with the first-generation products and the aqueous chemistry of low-volatility products reacting at the wet surface of NaCl seed particles. Seed aerosols are important for the growth of atmospheric particles and, therefore, affect aerosol-cloud-climate interactions. Organic nitrates, detected with FTIR and ESI-Exactive-Orbitrap MS, were found as significant composition of newly formed particles. A significant amount of carbonyl-rich products were also detected in the SOA products from the photooxidation of furan. A recent study showed that the reactive nitrogen chemistry in aerosol water can be a source of atmospheric sulfate during haze events (Cheng et al., 2016). In addition, the ALW is deeply linked with air quality (Malm et al., 1994) and aqueous SOA formation (Sareen et al., 2017). The influence of the ALW component on SOA yield was examined and it was found that increased ALW amounts lead to higher SOA mass concentration and yield, therefore highlighting the importance of the ALW in photooxidation reactions. ALW plays a crucial role in atmospheric physicochemical processes. The current results could also be used to interpret ambient gas-phase measurements and reaction mechanisms inference.

Data availability. Data are available by contacting the corresponding author.

The Supplement related to this article is available online.

Competing interests

The authors declare that they have no conflict of interest.

Author contributions

LD and XJ conceived and led the studies. XJ, NT, LJ, SL and HZ carried out the experiments and analyzed the data. XJ, LD and SL interpreted the results. NT, LJ and YX discussed the results and commented on the manuscript. XJ prepared the manuscript with contributions from all co-authors.

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$$NO_2 + h\nu \rightarrow NO + O\,(^3P) \qquad (R1)$$
$$O(^3P) + O_2 + Air \rightarrow O_3 + Air \qquad (R2)$$
$$O_3 + h\nu \rightarrow O(^1D) + O_2 \qquad (R3)$$
$$O(1D) + H_2O \rightarrow 2OH \qquad (R4)$$
$$RO_2 + NO \rightarrow RO + NO_2 \qquad (R5)$$
$$HO_2 + NO \rightarrow OH + NO_2 \qquad (R6)$$
$$NO + O_3 \rightarrow NO_2 + O_2 \qquad (R7)$$
$$NO + RO_2 \rightarrow NO_2 + RO \qquad (R8)$$
$$RO + O_2 \rightarrow HO_2 + R'O \qquad (R9)$$
$$RO_2 + NO \rightarrow RONO_2 + RNO_2 \qquad (R10)$$
$$2\,NO_2 + H_2O \rightarrow HONO + HNO_3 \qquad (R11)$$

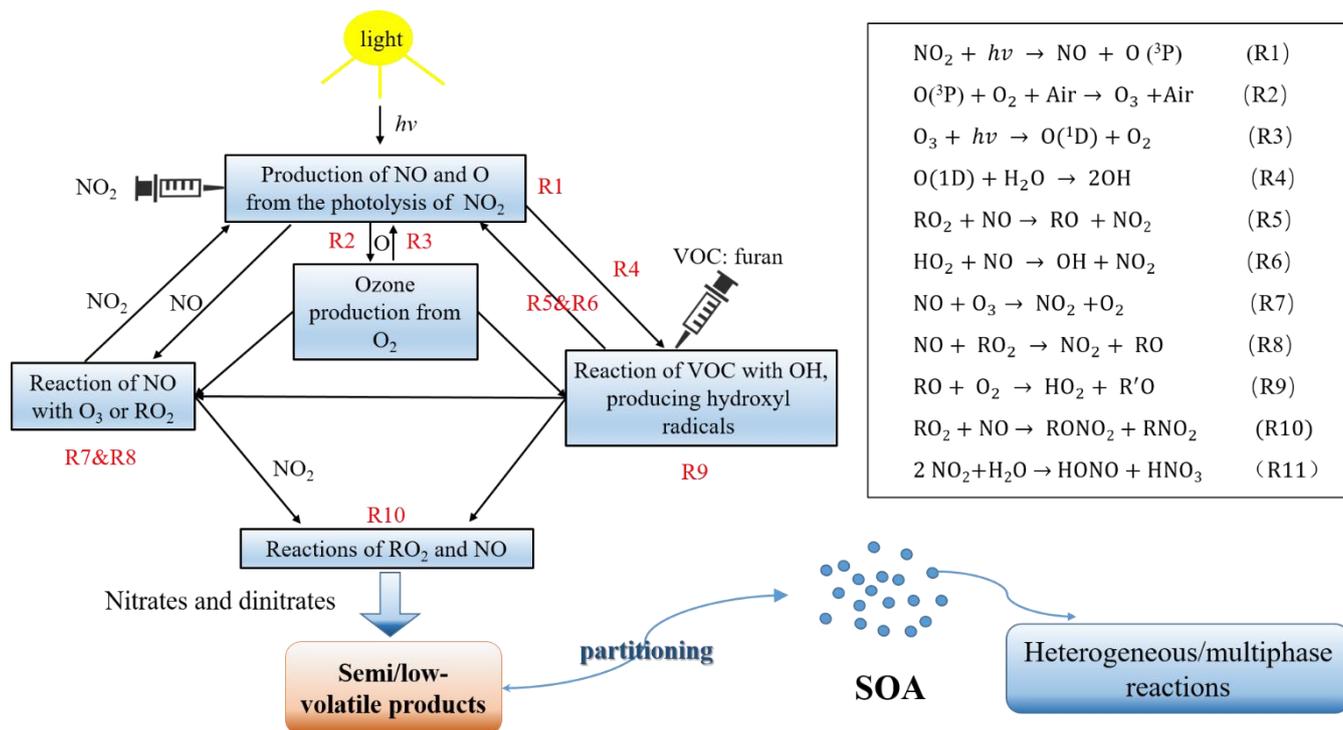



Figure 1: Major chemical reactions during the experiments.

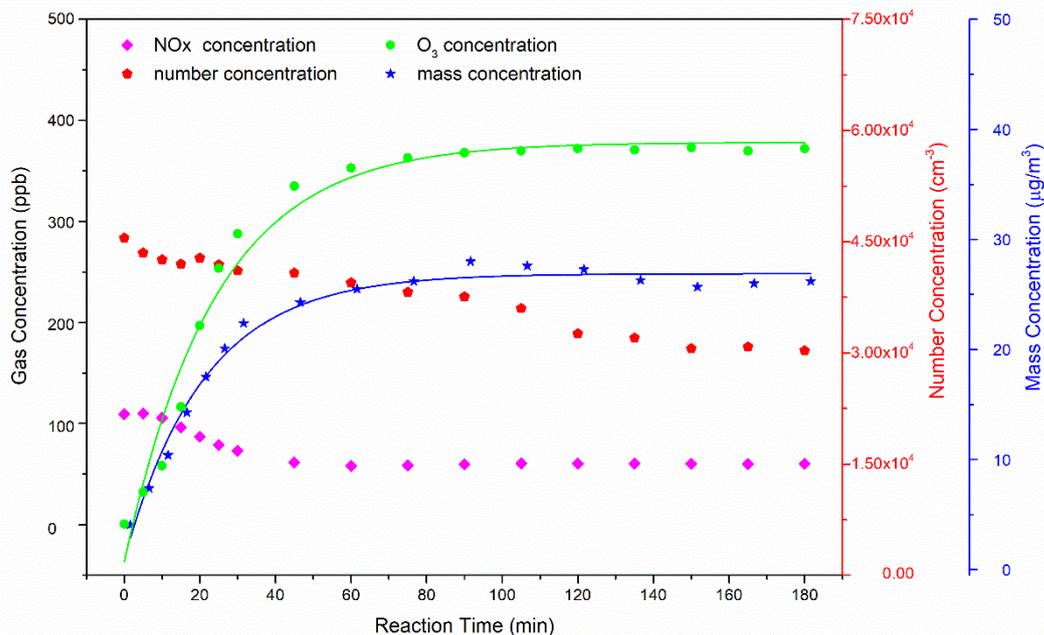



Figure 2: Profile of the gas-phase concentrations of reactants (NO, NO$_2$, NOx and O$_3$) and particle number/mass concentrations (corrected with wall loss) over time. The C$_4$H$_4$O/NOx ratio is 7.9 and RH = 23%. Since the particle wall loss has a weak RH dependence in our chamber, a mean value of 4.7 $\times$ 10^{-5} s^{-1} was used for wall loss correction. A density of 1.4 g cm^{-3} was used in the SMPS (Jia and Xu, 2018; Kostenidou et al., 2007).

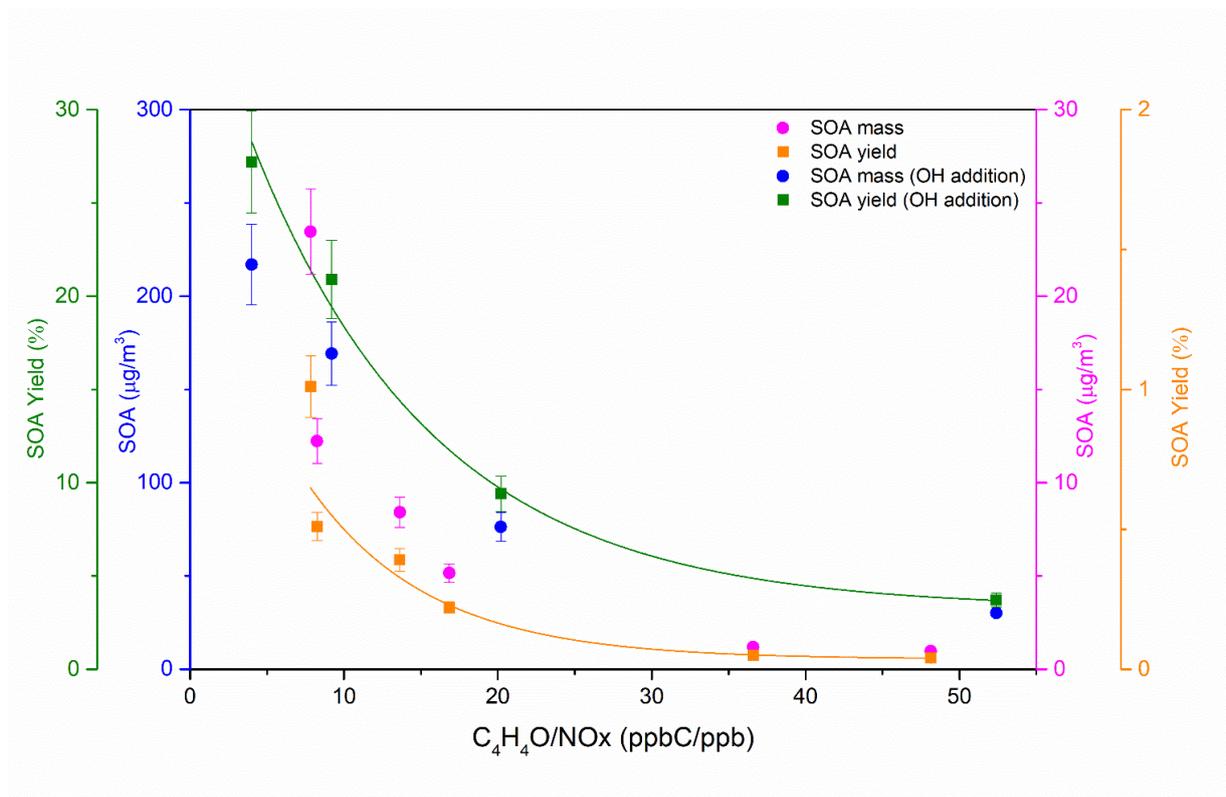



Figure 3: Dependence of the SOA mass concentration and SOA yield on the C₄H₄O/NOx ratio. Since the particle wall loss has a weak RH dependence in our chamber, a mean value of 3.6 × 10⁻⁵ s⁻¹ was used for wall loss correction. A density of 1.4 g cm⁻³ was used in SMPS (Jia and Xu, 2018; Kostenidou et al., 2007).

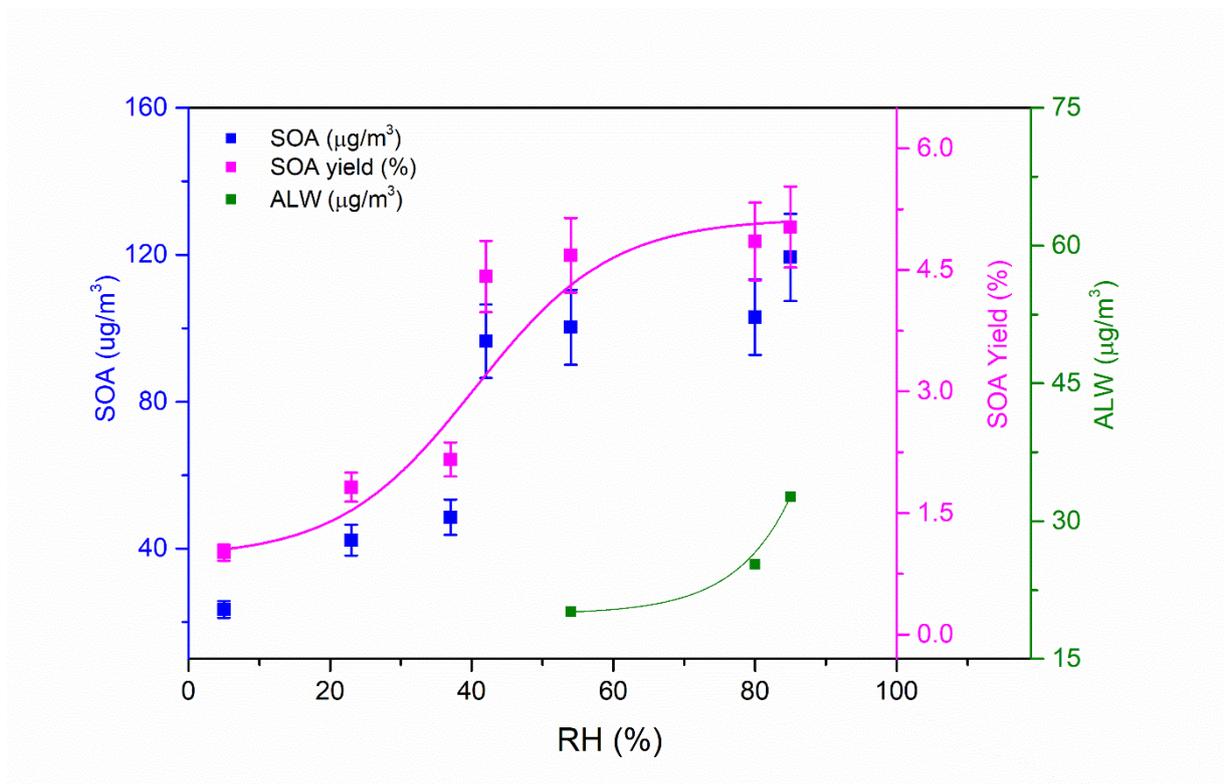



Figure 4: Dependences of the SOA mass concentration, SOA yield and ALW on relative humidity (RH).

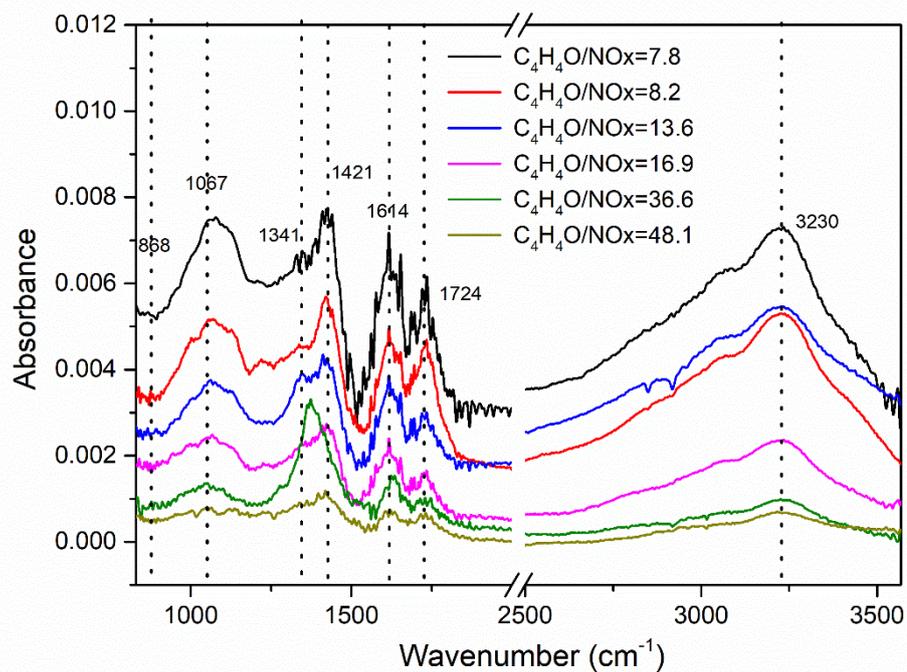



Figure 5: FTIR spectra of particles collected from furan-NOx-NaCl photooxidation experiments with different values of C₄H₄O/NOx ratio ranging from 7.3 to 48.1.

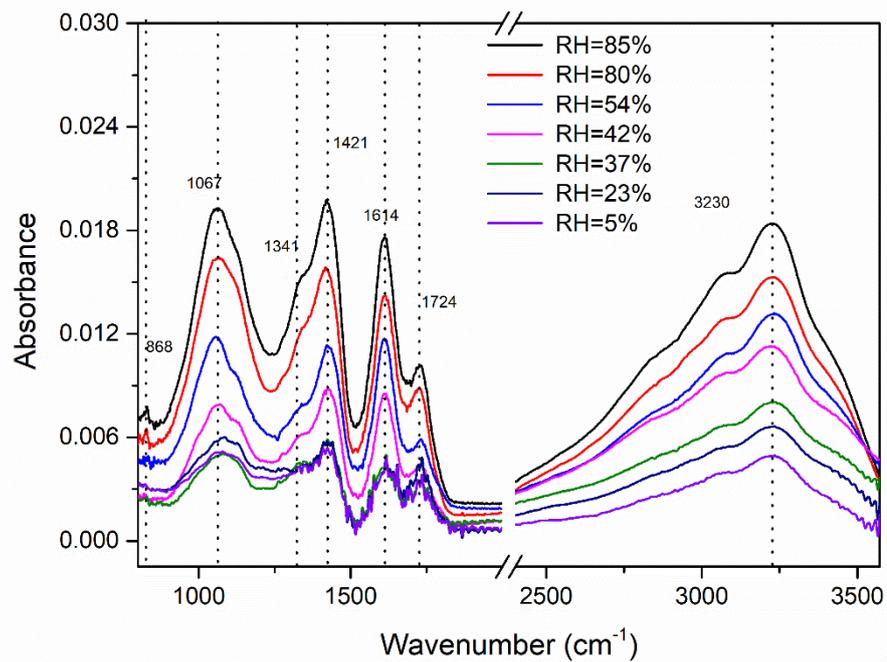



Figure 6: FTIR spectra of particles collected from furan-NOx-NaCl photooxidation experiments with different values of RH ranging from 5 % to 85 %.

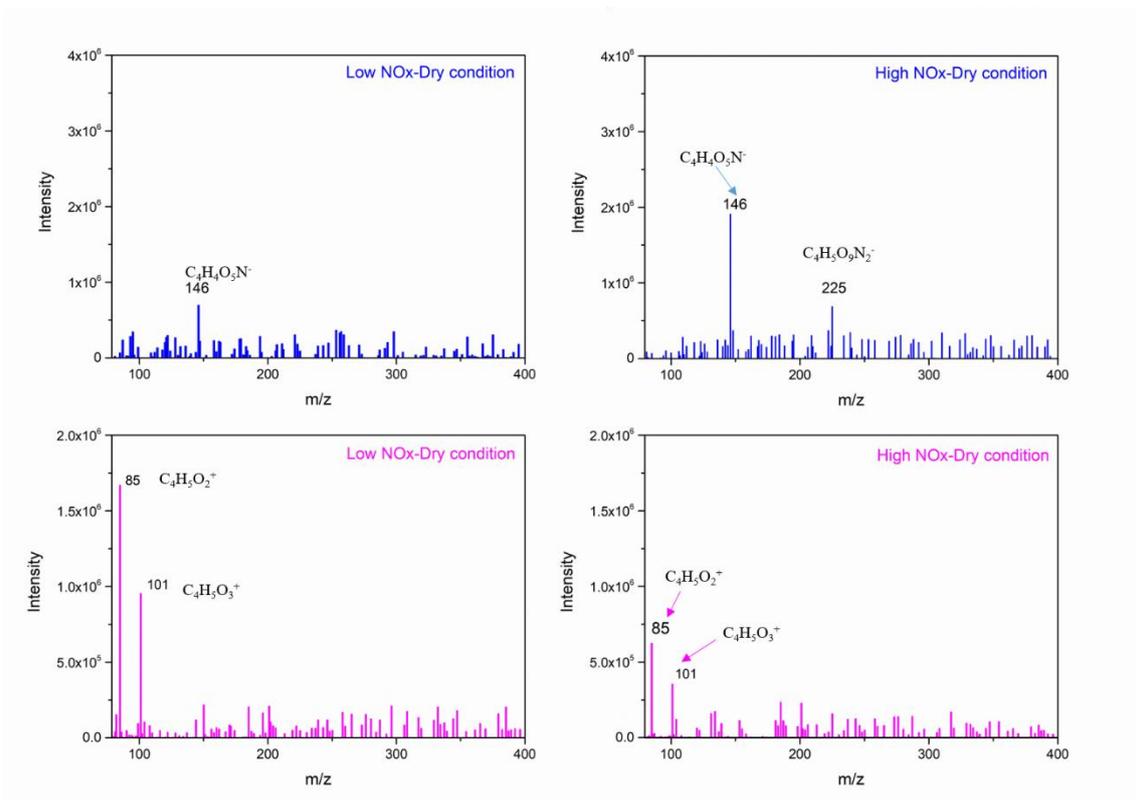



Figure 7: Selected background-subtraction HESI-Q Exactive-Orbitrap MS results of SOA in both negative (blue) and positive (pink) ion modes from the photooxidation of furan under different RH conditions.

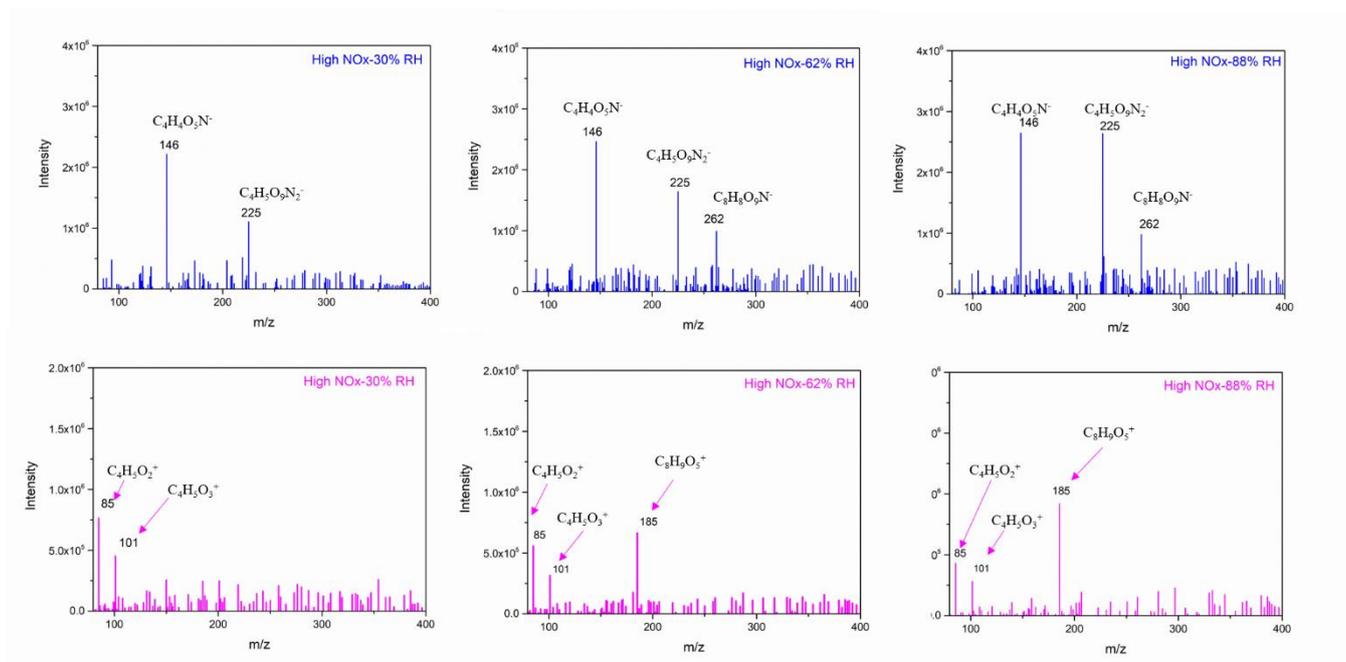



Figure 8: Selected background-subtraction HESI-Q Exactive-Orbitrap MS results of SOA in both negative (blue) and positive (pink) ion modes from the photooxidation of furan under different RH conditions.

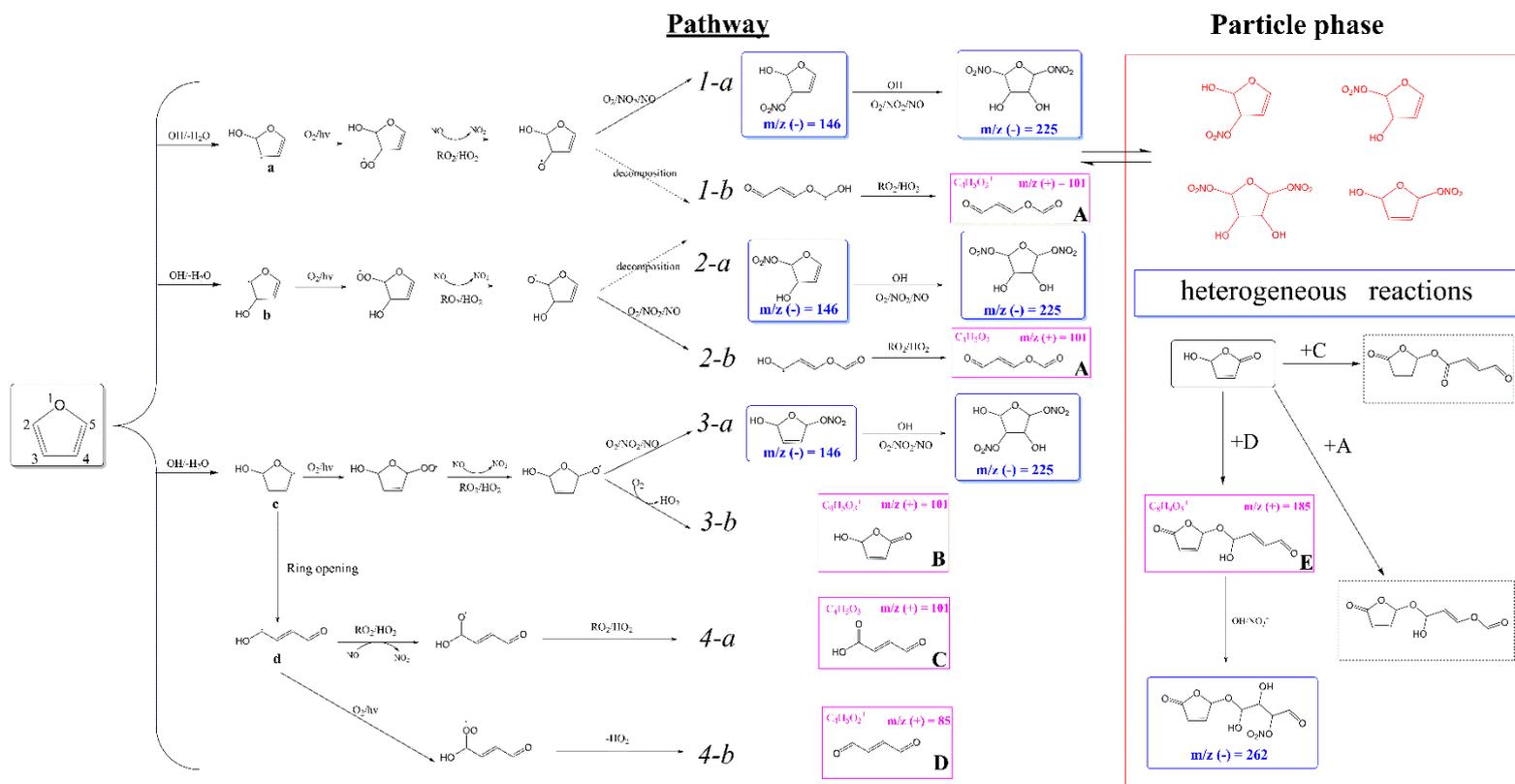
Pathway   **Particle phase**

Scheme 1: Proposed chemical mechanism of furan-NOx photooxidation under different experimental conditions. SOA constituents in blue and pink boxes are proposed SOA constituents detected by HESI-Q Exactive-Orbitrap MS under negative and positive ion mode, respectively. The detected nitrates and dinitrates in red are low-volatility organic species, which can easily partition into the particle phase and enhance the SOA formation.

Table 1. Summary of initial conditions, O_3 concentrations and particle mass concentrations in furan-NOx-NaCl photooxidation experiments.

Exp No	Initial conditions				SOA formation results						
	[furan]$_0$	[NO]$_x$	RH	C_4H_4O/NOx	O_3	PM[a]	NaCl[b]	NaNO$_3$[c]	ALW[d]	SOA[e]	SOA yield
	(ppb)	(ppb)	(%)	(ppbC/ppb)	(ppb)	($\mu g\ m^{-3}$)	($\mu g\ m^{-3}$)	($\mu g\ m^{-3}$)	($\mu g\ m^{-3}$)	($\mu g\ m^{-3}$)	(%)
1	708.4	16.8	<5 %	48.1	91	12.2	11.3	3.5	--	1.0±0.1	0.04±0.01
2	749.0	23.2	<5 %	36.6	115	15.4	10.7	1.6	--	1.2±0.2	0.05±0.01
3	752.5	44.7	<5 %	16.9	197	21.3	14.1	2.1	--	5.1±0.5	0.3±0.02
4	705.8	51.8	<5 %	13.6	250	23.5	12.6	2.5	--	8.4±0.9	0.3±0.03
5	783.4	94.9	<5 %	8.2	372	29.1	13.5	3.9	--	12.2±1.3	0.5±0.05
6	763.4	97.5	<5 %	7.8	382	38.6	11.5	3.5	--	23.5±2.3	1.0±0.1
7	764.8	96.7	23 %	7.9	359	55.5	10.4	3.8	--	42.3±4.2	1.9±0.2
8	740.1	97.3	37 %	7.6	353	64.2	11.2	4.5	--	48.6±4.9	2.2±0.3
9	719.0	100.2	42 %	7.2	329	111.1	9.4	5.2	--	96.4±9.7	4.5±0.5
10	704.8	92.9	54 %	7.6	280	138.7	12.0	6.2	20.1	100.2±10.1	4.7±0.4
11	699.3	102.9	80 %	6.8	253	144.1	8.2	7.6	25.3	103.0±10.3	4.8±0.5
12	780.7	98.1	85 %	8.0	241	173.0	10.0	11.1	32.6	119.2±10.4	5.0±0.5

[a]PM: particle mass concentration in the chamber was determined from the SMPS and was the sum of NaCl, NaNO$_3$, ALW, and SOA at the end of the experiments.22

[b]NaCl: the amount of NaCl at the end of the experiments;

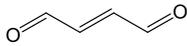
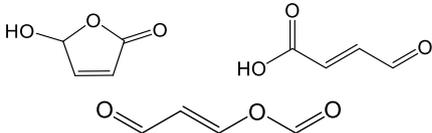
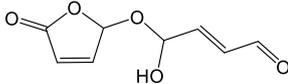
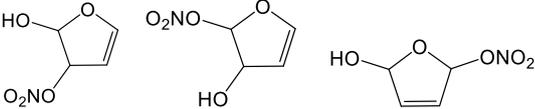
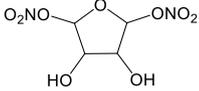
5   [c]NaNO$_3$: the amount of NaNO$_3$ at the end of the experiments;

[d]ALW: the amount of aerosol liquid water content at the end of the experiments;

[e]SOA: the amount of secondary organic aerosol at the end of the experiments.

Table 2. Assignment of the observed FTIR absorption frequencies (cm^{-1}).

Absorption frequencies	Functional group	Assignment
1222-868	C-C and C−O	Stretching in alcohols (Coury and Dillner, 2008)
		C−O of COOH group (Duarte et al., 2005)
1067	C−O−C	C−O stretching (Jang et al., 2002)
1515-1350	−CH	Bending vibration of −C−H group
868	-NO	NO symmetric stretching (Jia and Xu, 2016)
1614, 1341	−ONO$_2$	NO$_2$ symmetric stretching and asymmetric stretching
1724	C=O	Stretching vibration in carboxylic acid and ketones
3000-2850	C−H	C−H stretching vibration in the saturated carbon ring
3100-3000	C−H	C−H stretching vibration in unsaturated alkene
3200-2400	O−H	Stretching vibration in carboxyl groups
3100-3500	O−H	Stretching vibration in hydroxyl groups

Table 3. Ion peaks with the assigned compounds observed in the HESI-Q Exactive-Orbitrap MS. Proposed assignments are based on the formula from HESI-Q Exactive-Orbitrap MS.

Ion mode	No	Mass (m/z)	Ion Mode	Ion Formula	Delta (amu)	Proposed Structure
Positive ion mode	1	85.0018	$[M+H]^+$	$C_4H_5O_2^+$	-0.027	
	2	101.0894	$[M+H]^+$	$C_4H_5O_3^+$	0.066	
	3	185.0504	$[M+H]^+$	$C_8H_9O_5^+$	0.006	
Negative ion mode	4	146.0161	$[M-H]^-$	$C_4H_4O_5N^-$	0.007	
	5	225.0125	$[M-H]^-$	$C_4H_5O_9N_2^-$	0.012	
	6	262.0132	$[M-H]^-$	$C_8H_8O_9N^-$	-0.007	