

1 Formation of highly oxygenated organic molecules from chlorine atom
2 initiated oxidation of alpha-pinene

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30 **Abstract**

31

32 Highly oxygenated organic molecules (HOMs) from atmospheric oxidation of alpha-
33 pinene can irreversibly condense to particles and contribute to secondary organic
34 aerosol (SOA) formation. Recently, the formation of nitryl chloride (ClNO₂) from
35 heterogeneous reactions, followed by its subsequent photolysis is suggested to be an
36 important source of chlorine atoms in many parts of the atmosphere. However, the
37 oxidation of monoterpenes such as alpha-pinene by chlorine atoms has received very
38 little attention, and the ability of this reaction to form HOM is completely unstudied.
39 Here, chamber experiments were conducted with alpha-pinene and chlorine under low
40 and high nitrogen oxide (NO_x, NO_x = NO + NO₂) conditions. A nitrate-based CI-API-TOF
41 (Chemical Ionization-Atmospheric Pressure Interface-Time of Flight) was used to
42 measure HOM products. Clear distributions of monomers with 9-10 carbon atoms and
43 dimers with 18-20 carbon atoms were observed under low NO_x conditions. With
44 increased concentration of NO_x within the chamber, the formation of dimers was
45 suppressed due to the reactions of peroxy radicals with NO. We estimated the HOM
46 yields from chlorine-initiated oxidation of alpha-pinene under low-NO_x conditions to
47 be around 1.8 %, though with a substantial uncertainty range (0.8-4 %) due to lack of
48 suitable calibration methods. Corresponding yields at high NO_x could not be
49 determined because of concurrent ozonolysis reactions. Our study demonstrates that
50 chlorine atoms also initiated oxidation of alpha-pinene and yields low volatility organic
51 compounds.

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58 1. Introduction

59

60 Highly oxygenated organic molecules (HOMs) have been identified as key species in
61 the formation of new atmospheric aerosol particles and secondary organic aerosol (SOA)
62 (Ehn et al., 2014, 2017; Kulmala et al., 2013; Bianchi et al., 2019). Recently, the
63 formation of HOMs in the gas phase was described as an autoxidation process of peroxy
64 radicals (RO_2) via multiple intramolecular H atom shifts (Crouse et al., 2013; Jokinen
65 et al., 2014b; Mentel et al., 2015; Rissanen et al., 2014). Oxygen-containing moieties
66 such as carbonyl, carboxylic acid and hydroxyl groups can weaken nearby C-H bonds,
67 making H-abstraction and autoxidation competitive with bimolecular RO_2 reactions,
68 e.g. with NO (Crouse et al., 2013; Praske et al., 2018). Until now, all studies on the
69 formation of HOMs have focused on reactions initiated by oxygen-containing oxidants
70 (O_3 and OH)(Berndt et al., 2016; Ehn et al., 2014; Jokinen et al., 2015).

71

72 Increasing evidence indicates that the chlorine atom (Cl) may also play an important
73 role in transforming atmospheric organics (Tham et al., 2016; Thornton et al., 2010).
74 Chlorine atoms have the greatest reactivity toward volatile organic compounds (VOC),
75 with rate constants that are, with some exceptions, an order of magnitude higher than
76 those of hydroxyl radicals (OH) (Riva et al., 2015). Historically, chlorine atoms were
77 thought to be formed primarily from heterogeneous reaction cycles involving sea salt,
78 and their concentrations estimated to be around 1-10% of that of OH (Thornton et al.,
79 2010). Therefore, the role of chlorine atoms in atmospheric oxidation processes has
80 traditionally been thought to be limited to the marine boundary layer only. In recent
81 years, ClNO_2 , as a significant chlorine atom source, was found in continental regions
82 of America, Canada and Germany, and high concentrations of ClNO_2 were also
83 detected in the urban atmosphere in China (Reyes-Villegas et al., 2018; Tham et al.,
84 2016; Thornton et al., 2010; Wang et al., 2017). The new findings have expanded the
85 potential importance of chlorine atoms from coastal areas to continental urban areas. A
86 recent study also reported that chlorine atoms can be more important than OH radicals

87 for the oxidation of alkanes in the North China Plain (Liu et al., 2017). Therefore, it is
88 desirable to probe the role of chlorine radicals in the degradation of VOCs and related
89 SOA formation.

90
91 Emission of biogenic volatile organic compounds (BVOC) to the atmosphere
92 dominates total hydrocarbon emissions on a global scale, with methane, isoprene and
93 terpenes having the highest source strengths (Guenther et al., 2012). Alpha-pinene is
94 the most abundant monoterpene in the atmosphere and its oxidation products from
95 ozonolysis and photooxidation contribute to a substantial fraction of SOA mass
96 (Riccobono et al., 2014; Zhang et al., 2018). Many of the regions listed earlier, where
97 ClNO₂ was identified as a chlorine atom source, may also have substantial monoterpene
98 emissions, making Cl a relevant oxidant also for BVOC. Chlorine atom initiated
99 reactions of alpha-pinene have also been shown to contribute to the formation of SOA,
100 which implies that low volatile compounds are efficiently produced also in this process
101 (Cai and Griffin, 2006; Ofner et al., 2013).

102
103 Similar to the reaction with OH radicals, the reaction of VOCs with chlorine atoms may
104 proceed either via addition of Cl to unsaturated bonds or via H-abstraction. Wang et al.,
105 (2017) found that the Cl addition to isoprene can lead to the formation of low volatility
106 organic compounds. In principle, Cl-initiated reactions could form HOMs in a similar
107 manner as OH-initiated reactions (Berndt et al., 2016a), as the initial addition or
108 abstraction step is comparable for both oxidants. In view of the increased understanding
109 of the importance of chlorine atoms in atmospheric chemistry, it is desirable to
110 investigate the formation of HOMs from reactions of common atmospheric VOC with
111 Cl.

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113 Here, laboratory chamber experiments were performed to investigate the ability of
114 chlorine atom to form HOMs from the oxidation of alpha-pinene. HOMs were
115 characterized using a nitrate-based chemical ionization mass spectrometer, under both

116 low and high NO_x conditions. The yields of these HOMs were determined under the
117 low NO_x conditions, and the atmospheric implications of this study are discussed.

118

119 2. Experiment and method

120 2.1 Experimental setup

121 The experiments were conducted in the “COALA” chamber at the University of
122 Helsinki (Peräkylä et al., 2019; Riva et al., 2019). It is a 2 m³ Teflon chamber, run as a
123 continuously stirred tank reactor, used with a flow of 45 liter per minutes (LPM),
124 resulting in an average residence time of about 45 minutes. The chamber is surrounded
125 by housing to provide dark conditions. No water vapor was added to the chamber,
126 resulting in an RH<1 %, and the temperature was the same as the temperature of the
127 room, around 25°C. The HOM formation targeted in this work is not expected to change
128 markedly as a function of RH, as also indicated in a previous study on both ozone and
129 OH initiated oxidation of monoterpenes (Li et al., 2019). Although SOA mass yields
130 were not studied in this work, these may be affected by RH (Jonsson et al., 2006, 2008).
131 A general schematic of the chamber facility is shown in Figure 1. Our experiment was
132 aimed to probe chlorine atom initiated formation of HOMs under low/high NO_x
133 conditions. We use 400 nm LED lights to photolyze chlorine and NO₂ and produce
134 chlorine atoms and NO as following:



135

136 The concentration of Cl atoms was varied by changing the amount of 400 nm light. In
137 practice this was done by turning on different amounts of the available lights, with the
138 maximum corresponding to seven. We will refer below to the number of lights that were
139 turned on, although each “light” corresponds to a group of LED strips.

140

141 2.2 Instrumentation and data analysis

142

143 A nitrate ion (NO_3^-) based chemical ionization atmospheric pressure interface long
144 time-of-flight (CI-APi-L-TOF) mass spectrometer was used for measuring HOMs. The
145 instrument has been shown to be sensitive towards this group of compounds, detecting
146 them as adducts with the nitrate ion. Due to a lack of suitable calibration methods, the
147 CI-APi-TOF was not calibrated for HOMs during this study. In order to estimate rough
148 HOM concentrations ($[\text{HOM}]$), we directly use the calibration coefficient $C = 1.6 \times 10$
149 molec cm^{-3} utilized by Ehn et al. (2014), to convert the measured HOM ion signals
150 according to the equation below (Jokinen et al., 2014).

$$151 \quad [\text{HOM}] = C * \frac{\text{HOM} \cdot \text{NO}_3^-}{\text{NO}_3^- + \text{HNO}_3 \cdot \text{NO}_3^-}$$

152 This value of C is very close to values utilized in several other studies using a CI-
153 APi-TOF (Jokinen et al., 2014, 2015; Riva et al., 2019). Ehn et al (2014) obtained their
154 calibration factor from a permeation source using a perfluorinated acid, and also
155 showed through calculations that there were theoretical limitations for having a much
156 larger or a much smaller value for C .

157 This approach obviously brings large uncertainties to the HOM concentrations, which
158 we estimate to be at least -50 %/+100 % according to previous calibration results
159 (Jokinen et al., 2014, 2015; Riva et al., 2019). More detailed information about the
160 instrument can be found in (Jokinen et al., 2012), noting that compared with the CI-
161 APi-TOF used before, the long time-of-flight mass spectrometer used here has a
162 doubled mass resolving power enabling a more accurate assignment of molecular
163 formulas. Simultaneously, we also used a high-resolution long time-of-flight aerosol
164 mass spectrometer (HR-L-TOF-AMS) to measure bulk aerosol chemical properties
165 (Decarlo et al., 2006). As no seed aerosol particles were added to the chamber, the VOC
166 oxidation products lead to new particle formation and growth to large enough sizes to
167 be measured by the AMS. We also periodically used a filter in front of the AMS inlet to
168 see the influence of the background signal to measured aerosol mass concentration.
169 There were 10 minutes of filter measurements per hour during our experiments. A PTR-
170 TOF-MS (TOF-8000, Ionicon) was used to measure the concentration of alpha-pinene
171 in the chamber. The instrument background was determined every day for 20 mins by

172 guiding the chamber air through a catalytic converter, which removes the VOCs. Then,
173 the background corrected signals were used to obtain alpha-pinene mixing ratios by
174 using the calibration coefficient determined before the experiments. A description of
175 the used setup employed for the calibration and zero air measurements have been
176 introduced earlier (Schallhart et al., 2018). A custom-built DMPS system was used to
177 measure the particle number size distribution from 10 nm to 400 nm in the chamber.
178 The NO concentration was measured with an ECO-PHYSICS CLD 780 TR instrument
179 with a detection limit of 3 ppt. NO_x (=NO+NO₂) concentrations were determined by
180 using a Thermo-Fisher 42i analyzer. O₃ concentration was measured with a Thermo-
181 Fisher 49i analyzer.

182

183 2.3 Estimation of chlorine atom concentrations

184 During steady state in the chamber, average concentrations of chlorine atom was
185 calculated using the rate coefficients $(4.6 \pm 1.3) \cdot 10^{-10} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$ of Cl atoms
186 with alpha-pinene (Finlayson-Pitts et al., 1999), as following:

$$187 \quad d[\text{AP}]/dt = Q_{\text{in}} - k * [\text{Cl}] * [\text{AP}] - Q_{\text{out}}$$

188 where Q_{in} is the flow rate of alpha-pinene continuously injected into the chamber,
189 and Q_{out} is the flow rate that exited the chamber. The term Q_{in} related
190 concentration was 13.3 ppb, while the term Q_{out} varied depending on the conditions,
191 and is calculated as [AP]/45 min. During steady state, d[AP]/dt is zero, and then [Cl]
192 concentration is calculated accordingly. As shown in Figure 2, the concentration of
193 HOMs decreased and alpha-pinene increased as the number of lights switched on
194 changed from 7 to 4, 2 and 1. We use the variation of alpha-pinene and HOM
195 concentrations during this run to calculate both chlorine atom concentrations and HOM
196 yields. Each change in alpha-pinene concentration was due to the change in Cl atom
197 concentration, and with knowledge of the reaction rate, the concentration of Cl atoms
198 as a function of the number of lights turned on was determined (Figure 3). The
199 calculated [Cl] concentrations are in the range of $(1-5) \times 10^5 \text{ molecules cm}^{-3}$, which is
200 within atmospheric relevant concentration ranges (Tham et al., 2016). Raw data from

201 the CI-API-L-TOF were recorded in 10s resolution in HDF format. We used Toftools
202 for data analysis and detailed protocols of the software have been introduced by
203 Junninen et al. (2010).

204

205 2.4 HOMs molar yield

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207 The change of HOM concentration with time can be described as follows, in analogy
208 with Ehn et al. (2014):

209

$$210 \frac{d[HOMs]}{dt} = k_1 g [\alpha - pinene][Cl] - k_{loss}[HOMs] \quad (1)$$

211

212

$$213 g = \frac{k_{loss}[HOMs]}{k_1[\alpha - pinene][Cl]} \quad (2)$$

214

215 Here, k_1 is the reaction rate coefficient of alpha-pinene with chlorine atoms and γ is the
216 molar yield of HOMs, i.e., the fraction of alpha-pinene + Cl reactions that produced
217 HOMs. k_{loss} is the loss rate of HOMs to the chamber walls and particles, though the
218 latter was negligible in this study due to the low aerosol loadings. We used 300 s as a
219 lifetime of HOMs, i.e. $k_{loss} = 1/300 \text{ s}^{-1}$, in our previous study in the COALA chamber
220 (Riva et al., 2019).

221

222 3. Results and discussion

223 3.1 Formation of HOMs under low NO condition

224

225 Figure 4 (a, b, c and d) shows mass spectra measured by the NO₃-CI-API-TOF during
226 steady state alpha-pinene oxidation with different amounts of lights switched on. The

227 x-axis represents mass to charge ratio, in units of Thomson (Th). The y-axis represents
228 signals in units of counts per second. As we can see, both monomers (280-400 Th) and
229 dimers (440-580 Th) showed increased signals with increased number of lights, and
230 consequently increased [Cl]. The most abundant peaks are labeled in Figure 4d, with
231 some of the largest signals in the monomer range attributed to $C_9H_{12}O_{7,8}$ and
232 $C_{10}H_{14}O_{8,9,10}$. The formation of both groups could correspond to the oxidation being initiated by H-
233 atom abstraction by Cl, and final termination (from uni- or bimolecular reactions) leading to loss of
234 OH or HO₂. For the C₉ compounds, an additional loss of formaldehyde (CH₂O) during the oxidation
235 process would explain the amount of observed C and H atoms. During the oxidation of C₁₀H₁₆,
236 in the absence of NO, the fate of RO₂ radicals depends on the concentrations of HO₂
237 and RO₂. Autoxidation competes with bimolecular reactions, becoming more likely at
238 lower RO₂ and HO₂ concentrations.

239

240 As we show in the Figure 4(d), $C_{10}H_{14}O_{8-12}$ compounds are large peaks in the monomer
241 range observed with the NO₃-Cl-API-TOF. These compounds with 14 hydrogens may
242 come from decomposition of $C_{10}H_{15}O_n$ peroxy radicals via loss of OH or HO₂, or
243 following reactions with other RO₂, as depicted schematically in Figure 5. Another
244 abundant group is $C_{10}H_{16}O_{6-12}$, which may result from RO₂ terminated by HO₂. In the
245 dimer range, the most abundant compounds are $C_{19}H_{28}O_{8-14}$ and $C_{20}H_{30}O_{11-14}$. These
246 compounds come from RO₂ cross reactions, as has been shown in multiple earlier
247 studies (Ehn et al., 2014b; Jokinen et al., 2015; Mentel et al., 2015). The $C_{20}H_{30}O_n$
248 dimers are most likely formed from reactions of two $C_{10}H_{15}O_x$ radicals, as were many
249 abundant monomers. As noted earlier, Cl oxidation of alkenes may occur via a Cl
250 addition (forming an initial radical containing 16 H-atoms and one Cl atom) or via an
251 H-abstraction reaction (forming a radical with 15 H-atoms and no Cl) (Figure 5). The
252 abstraction pathway leads to HOM formation, or the Cl atom is lost during the
253 subsequent reaction in the oxidation processes. With our data, we cannot rule out either
254 of these explanations for this result. Loss of HCl from alpha-pinene products from Cl
255 oxidation have, to our knowledge, only been reported to take place in the aerosol phase

256 (Ofner et al., 2013).

257

258 Figure 6 shows the variation of several close-shelled HOM products and the peroxy
259 radical $C_{10}H_{15}O_{10}$ measured by NO_3 -CI-API-TOF when we changed the lights from
260 dark conditions to 1, 2, 4 and 7 lights switched on. Given the low Cl atom concentration,
261 it is expected that no multi-generation oxidation by Cl can take place, and the behavior
262 of all closed shell oxidation products should follow similar patterns. As seen in Figure
263 6, this was the case both for monomers and dimers. The less steep increase of the radical
264 is also according to expectations, as the formation of RO_2 is linear with the alpha-pinene
265 oxidation rate, but the loss rate (when dominated by RO_2 cross reactions) is proportional
266 to $(RO_2)^2$. For closed shell species, the wall loss-driven loss rate stays constant
267 throughout the experiment, and therefore they increase linearly with the alpha-pinene
268 oxidation rate while the RO_2 radicals increase as the square root of the oxidation rate.
269 For more detailed discussion on RO_2 dynamics in a steady state chamber, see Ehn et al.
270 (2014).

271

272 In Figure7, we plotted time series of the particle number size distribution and the total
273 number concentration, together with mass concentrations of particulate chloride and
274 organics as we changed the number of lights. Particle formation was detected even at
275 the lowest Cl atom concentration, as indicated by the increases in aerosol number
276 concentration. An increase in aerosol mass concentration as detected by the AMS only
277 took place at the two highest Cl atom concentrations, when the particles were able to
278 grow into a size range measurable by the AMS. Particulate chloride mass
279 concentrations also increased relatively linearly with the concentration of organics as
280 we increased the number of lights. The Chl/Org ratio was only around 3 %, suggesting
281 that the majority of condensed OVOC did not contain Cl atoms. However, the exact Chl
282 quantification from organochlorides using the AMS may contain uncertainties (Wang and Ruiz,
283 2017), and we avoid drawing too far-reaching conclusions from this value, keeping also in mind
284 that some fraction of the size distribution was below the lowest detectable size of our AMS. In

285 addition, some part of the chloride signal may also result from adsorption of HCl to
286 particles.

287

288 3.2 Formation of HOMs at high NO_x

289

290 Anthropogenic emissions have a significant influence on the formation of SOA, to a
291 large part due to the influence of NO_x on the atmospheric oxidation chemistry (Lee et
292 al., 2016). In general, the fate of peroxy radicals in chamber experiments can be
293 dominated by reactions with other RO₂, HO₂ or NO, depending on the exact conditions.
294 In our experiments without NO_x addition, RO₂ was expected to be the main terminator,
295 as also supported by the high number of detected ROOR dimers. In the atmosphere, all
296 of the three mentioned reaction partners may be relevant at the same time. However,
297 with increased anthropogenic influence, the reaction of RO₂ with NO will often become
298 dominant. Therefore, we added NO_x to the chamber as it allowed for the isolation of
299 the formation pathways leading to HOMs in cases where NO was the main terminator
300 for RO₂ radicals. Figure 8 depicts a HOM mass spectrum at steady state during alpha-
301 pinene oxidation by chlorine radicals in the presence of ~10 ppb NO_x, with the
302 maximum 7 lights turned on. As anticipated, the dimers above 440 Th were greatly
303 reduced compared to the runs without NO_x. As more lights were turned on, both the Cl
304 atoms and NO formation increased, as the 400 nm lights photolyze both Cl₂ and NO₂.
305 This coupling, together with the fact that the NO₂ photolysis leads to ozone formation,
306 which subsequently can react with alpha-pinene to form HOMs, limits our quantitative
307 analysis of these experiments. However, we conclude that efficient HOM formation
308 took place also under these high-NO_x conditions, and thus the autoxidation occurs
309 rapidly enough to still compete with RO₂ termination reactions. The NO_x addition also
310 formed an abundance of organonitrate compounds like C₁₀H₁₅NO_{8,9,10,11,12}, as shown in
311 Figure 8. This family of compounds may also form from H-abstraction by the chlorine
312 radical, followed by autoxidation and finally radical termination by NO. We calculated
313 reacted alpha-pinene from ozone and Chlorine atom as following: $[AP_{\text{ozone}}]/[AP_{\text{Cl}}]$

314 $=[\text{AP}][\text{Ozone}]2.5\text{e}10k_{\text{ap+ozone}}/[\text{AP}][\text{Cl}]k_{\text{ap+Cl}}$, the result is around 0.2. With the results, we
315 conclude that alpha-pinene reaction with chlorine atom is the main reaction in the system.

316 The concurrent formation of ozone means that also some alpha-pinene ozonolysis
317 reaction will take place, though oxidation by Cl atoms was still the main loss for alpha-
318 pinene also under these conditions.

319

320 Figure 9 shows variation of some nitrogen-containing HOMs and variation of alpha-
321 pinene, ozone, NO and NO_x, as we changed the lights from dark conditions to 1, 2, 4
322 or 7 lights switched on. The concentrations of alpha-pinene and NO₂ decreased because
323 of the consumption by chlorine radicals and photolysis of NO₂ into NO. Importantly,
324 we did not observe any SOA when we had NO in the chamber. NO may have suppressed
325 the particle formation by suppressing the dimer formation, as these have been shown to
326 be important for initial particle formation (Tröstl et al., 2016).

327

328 3.3 Estimated HOMs production yields

329

330 Quantifying the molar yields of HOMs is essential to know their potential importance
331 from a specific system. We attempt to estimate the molar yield in the case of Cl
332 oxidation of alpha-pinene in the absence of NO_x. The initial C₁₀H₁₆ concentration is
333 around 13.3 ppb without any UV lights switched on in the chamber. As we changed the
334 lights, alpha-pinene and HOM concentrations varied as we showed in Figure 3. In
335 addition, we calculated the concentration of Cl radicals as introduced in the Methods
336 section. With this information, we can calculate the formation rate of HOM, which in
337 steady state equals the HOM loss rate $[\text{HOM}]k_{\text{loss}}$. We can also calculate the oxidation
338 rate of alpha-pinene as $[\text{alpha-pinene}][\text{Cl}]k_{\text{AP+Cl}}$. The ratio of these two numbers
339 corresponds to the HOM molar yield. We selected the same runs as in Fig. 3, used also
340 for calculating the chlorine radical concentration, and calculated the ratio as a linear fit
341 to these four conditions (Figure 10). We get a slope of 0.018, meaning a HOM yield of
342 1.8%. Considering the uncertainty in estimating absolute HOM concentrations, we

343 conservatively estimate that the molar HOM yield from alpha-pinene + Cl is within the
344 range of 0.8-4 %. These values are similar to HOM yields reported for alpha-pinene
345 oxidation by ozone and OH (Berndt et al., 2016; Ehn et al., 2014).

346

347 4.Conclusion

348

349 We have systematically explored the reactions of alpha-pinene with chlorine atoms in
350 a simulation smog chamber under atmospherically relevant conditions. We measured
351 substantial amounts of highly oxidized organic molecules (HOM) with a NO₃-Cl-API-
352 TOF. With increasing UV lights, and consequently higher chlorine radical
353 concentrations, the concentrations of both HOM and secondary organic aerosol
354 increased. With addition of NO_x, HOM monomer formation was still efficient, but the
355 particle formation decreased greatly. We estimated HOM molar yields of around 1.8 %
356 (0.8-4 %) from the reaction of alpha-pinene with Cl atoms. Our study thus indicates
357 that in regions where chlorine atom oxidation is of importance, its possible reactions
358 with monoterpenes can be an important source of HOM, and consequently, SOA.

359

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364 providing the software.

365

366 **Competing financial interests**

367 The authors declare no competing financial interests.

368

369 **Author contributions**

370 Y. H. W, H. B. X and M. E had the original idea of the study. Y. H. W, M. R and H. B.

371 X conducted the chamber experiments. Y. H. W, M. R, H. B. X, L.H and M. E

372 interpreted the data. Y.H.W plotted the figures, wrote the manuscript with comments

373 and suggestions from all co-authors.

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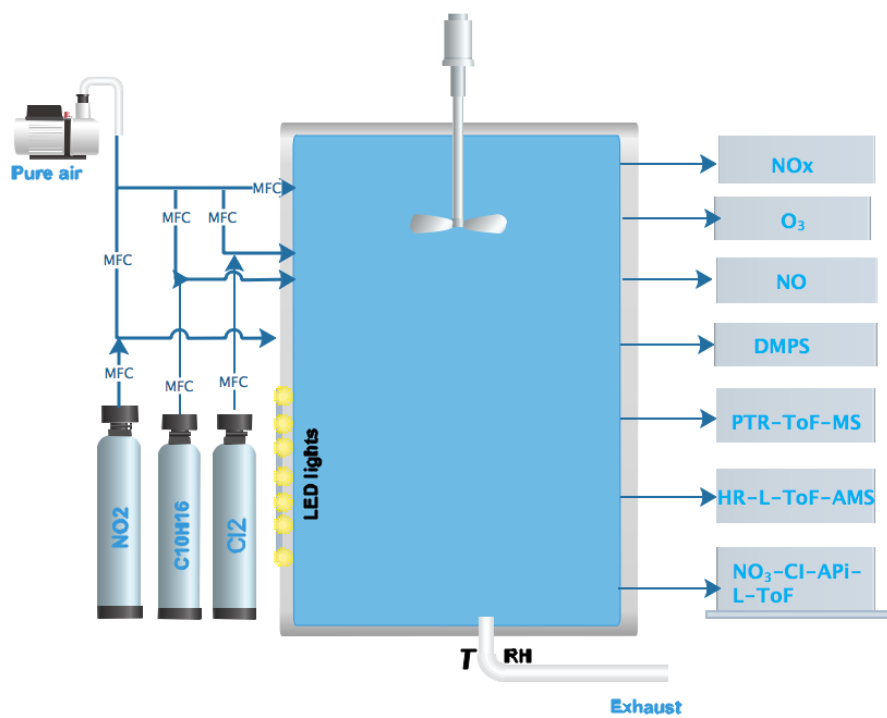
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Figure and table captions

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620 Figure 1 A schematic of the chamber setup and instruments used in the experiment.

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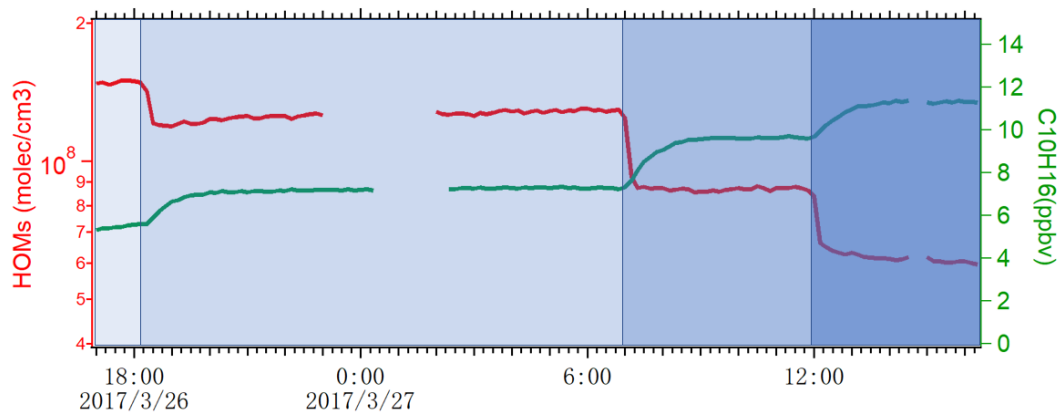
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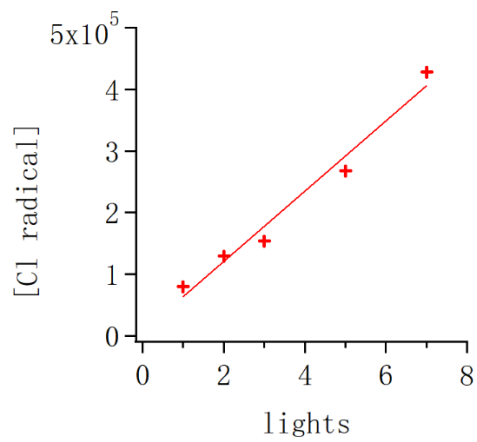
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Figure 2. The variation of total HOM concentration and alpha-pinene during four experiments where the 400 nm lights were decreased stepwise from 7 lights to 4, 2 and 1 light, respectively.



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Figure 3. The variation of chlorine radical concentration as a function of lights. The

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input alpha-pinene concentration was kept constant throughout the experiments.

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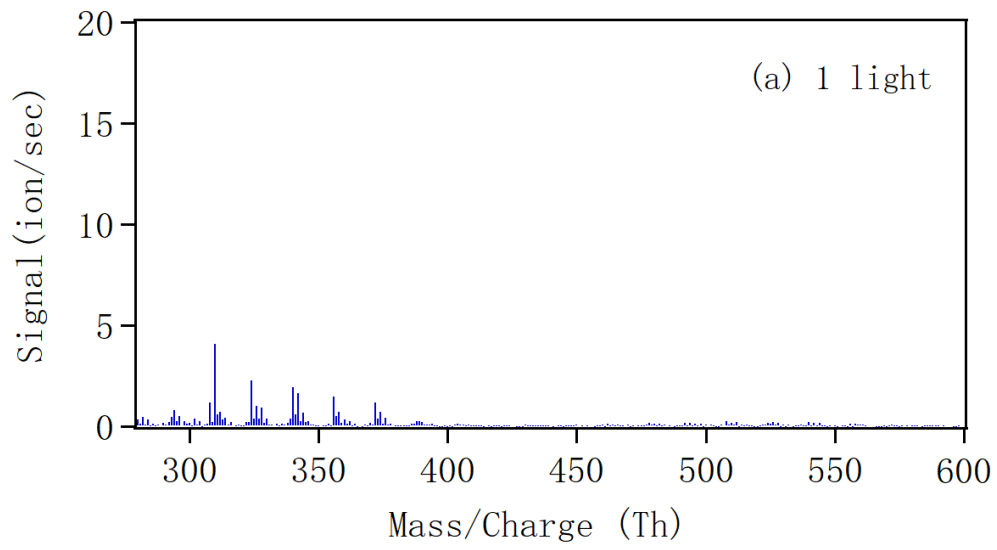
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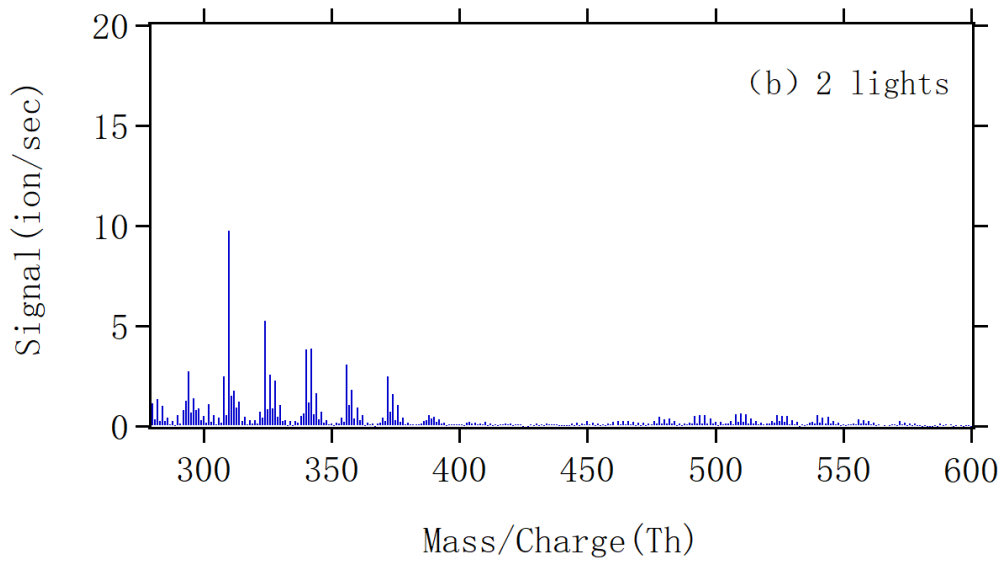
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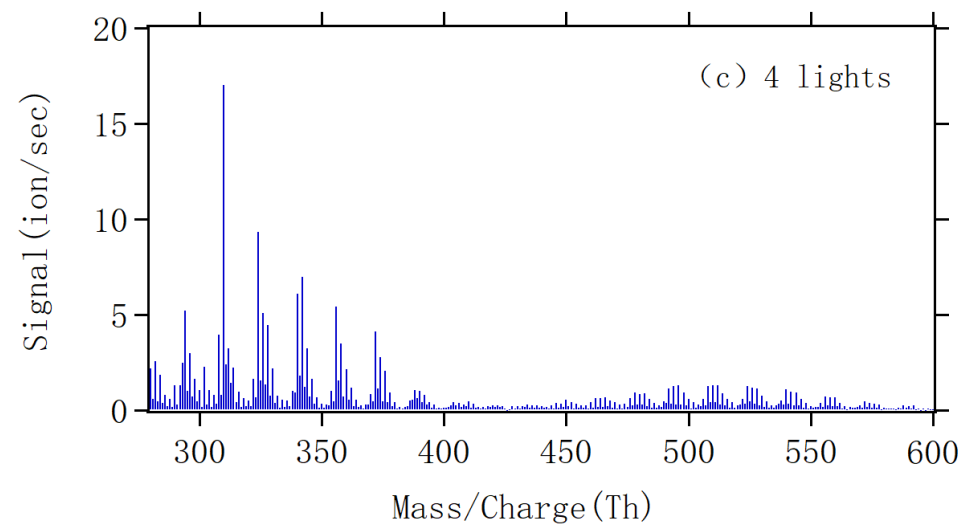
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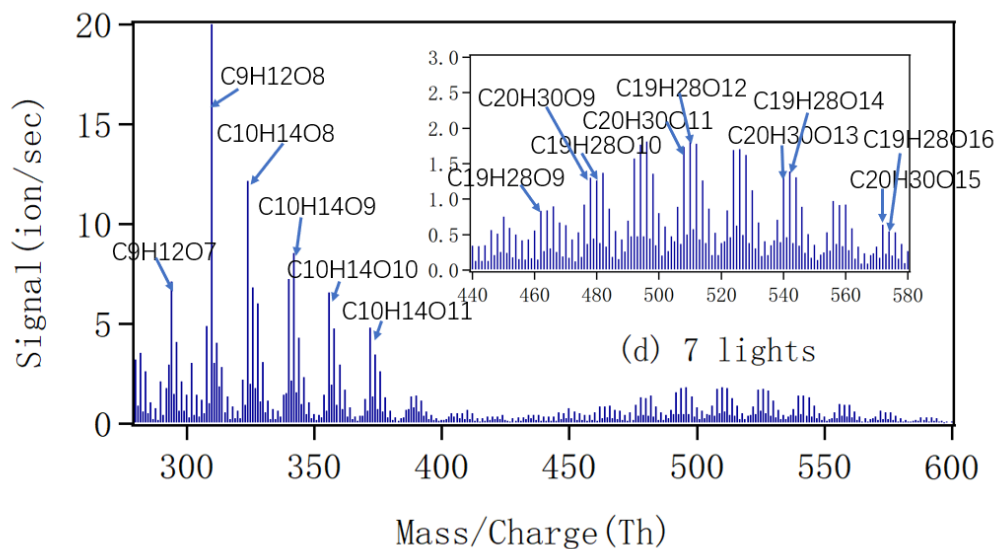
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665 Figure 4. The mass spectra obtained by NO₃-CI-API-TOF during steady state with 1(a),
 666 2(b), 4(c) and 7(d) lights. All peaks are detected as clusters with NO₃⁻. The spectra are
 667 plotted as unit mass resolution, with background signals removed, but the peak
 668 identifications (labeled in panel d) are based on high resolution analyses. The spectra
 669 correspond to the same four steady state conditions depicted in Fig. 2.

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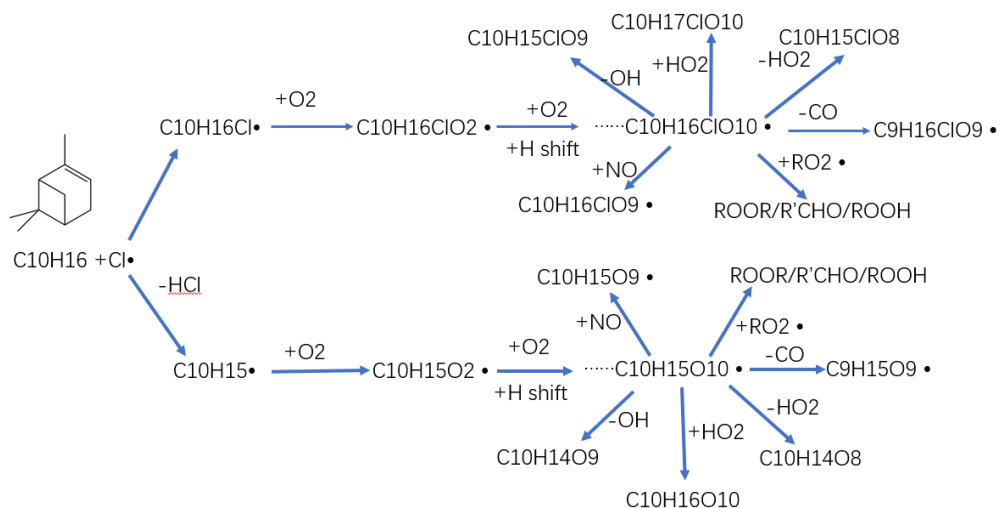
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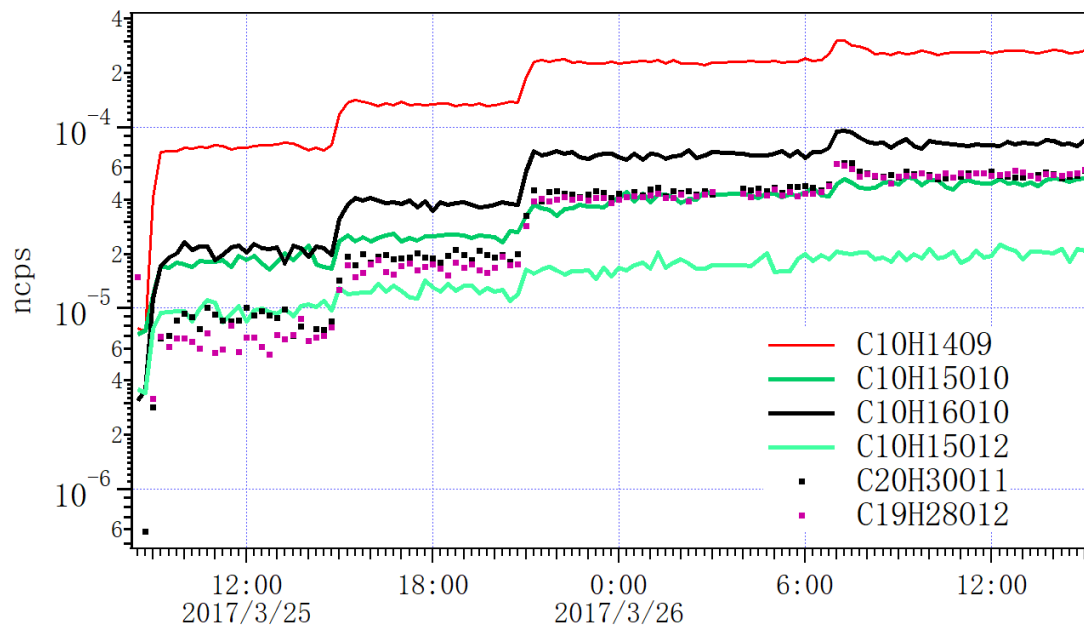
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679 Figure 5. Proposed chemical pathways for chlorine radical oxidation of alpha-pinene,
 680 and subsequent autoxidation and HOM formation. The upper path shows the chlorine
 681 radical addition pathway, while the lower chain shows hydrogen atom abstraction
 682 pathway. In both cases, initially a C-centered radical forms ($C_{10}H_{16}Cl$ or $C_{10}H_{15}$) to
 683 which O_2 adds to form an initial peroxy radical. This peroxy radical may then undergo
 684 multi-step autoxidation to reach the example molecules $C_{10}H_{16}ClO_{10}$ or $C_{10}H_{15}O_{10}$
 685 before termination.

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693 Figure 6. Time series of selected closed-shell HOM monomers, dimers and an RO₂

694 radical (C₁₀H₁₅O₁₀) detected by NO₃-CI-APi-TOF as the lights increased from 0 to 1,

695 2, 4 and 7.

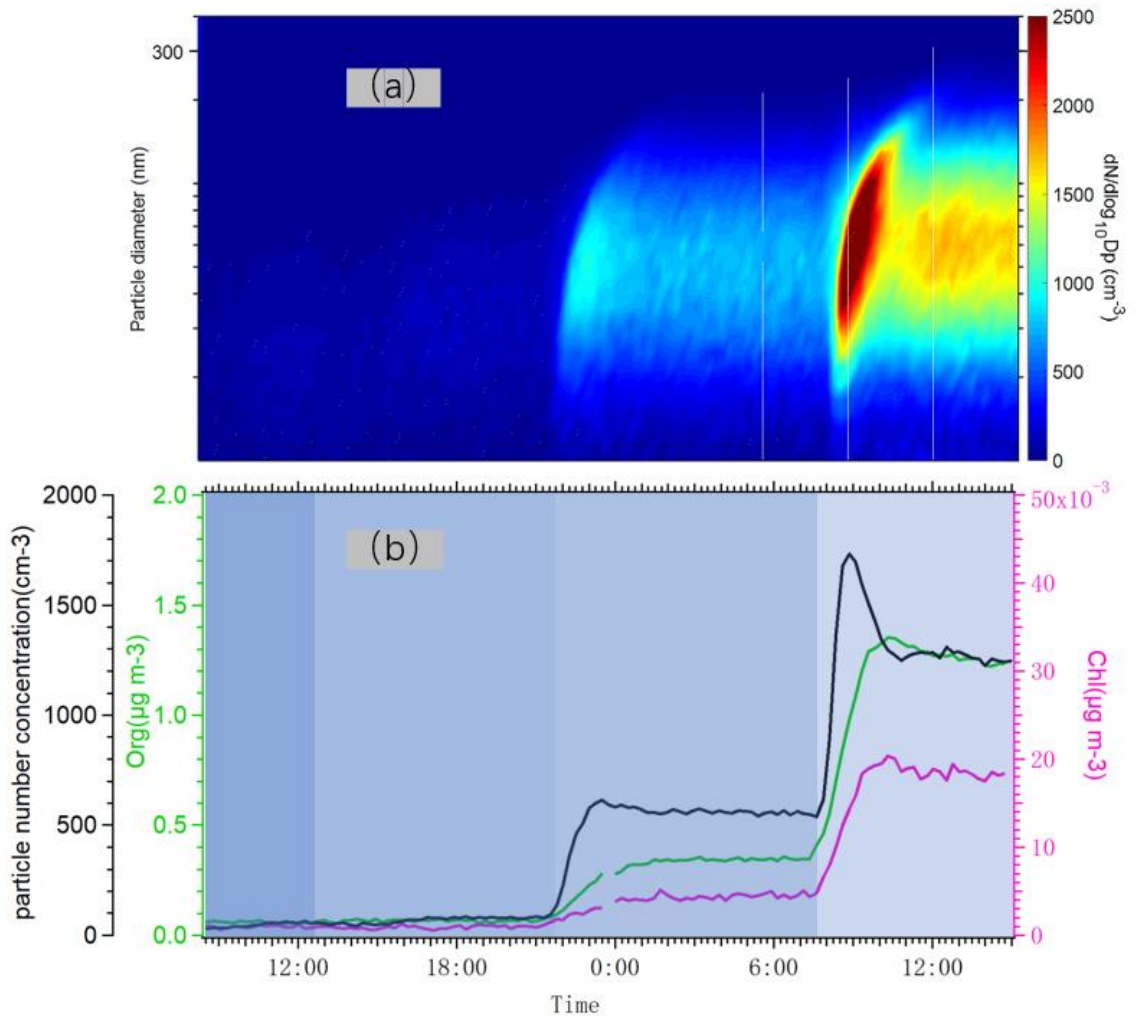
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703 Figure 7. (a) Particle number size distribution measured by DMPS from 10 nm to

704 400nm when the lights varied from 1 to 2, 4 and 7. (b) Time series of total number

705 concentration (black) measured by DMPS, organic aerosol concentration (green) and

706 particulate chloride concentration (pink) measured by AMS.

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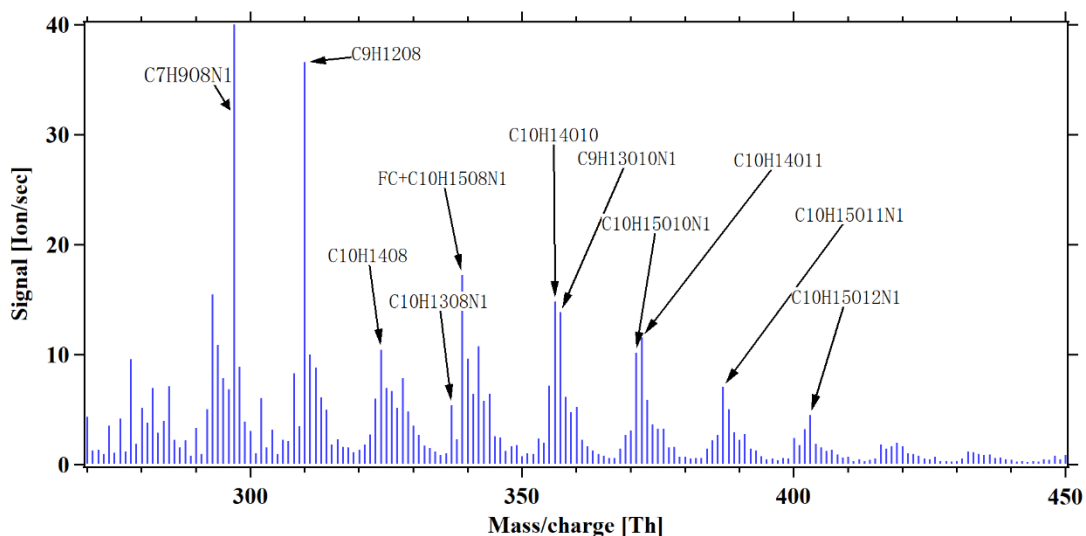
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714 Figure 8. HOM mass spectrum during steady state alpha-pinene oxidation in the
 715 presence of 10 ppb NO_x, with 7 lights switched on. In addition to molecules detected
 716 also in the experiments without NO_x, several abundant organic nitrate peaks are formed.
 717 Note that a fluorinated compounds (FC) overlaps with the organ nitrate C₁₀H₁₅O₈N at
 718 339Th. All peaks are detected as clusters with NO₃⁻.

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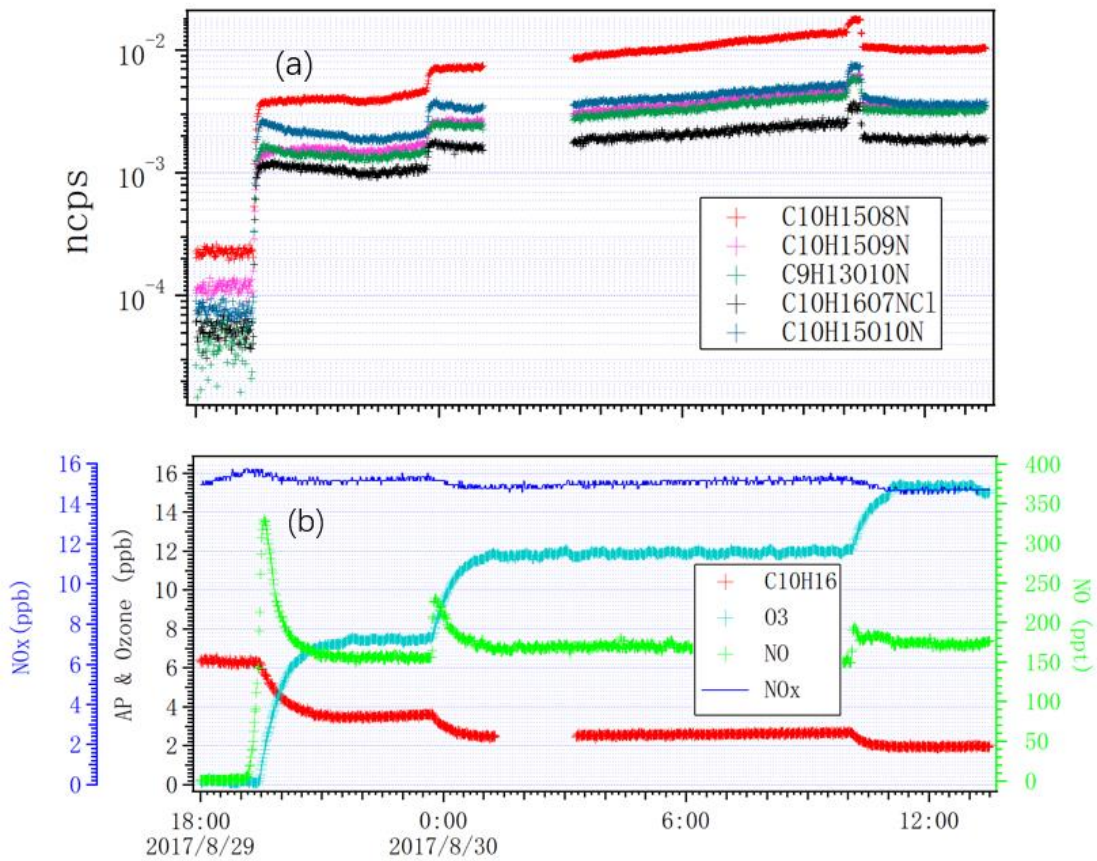
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733 Figure 9. Time series of (a) selected HOMs measured by $\text{NO}_3\text{-Cl-L-API-TOF}$ and (b)

734 NO_x , a-pinene, ozone and NO, as the lights switched on from zero to 2, 4, and 7.

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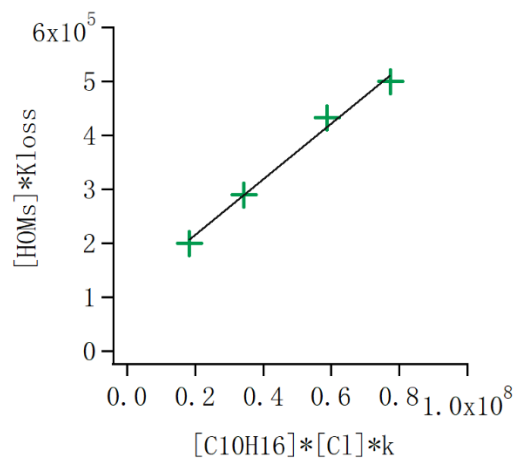
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745 Figure 10. HOM loss rate ($[HOM] \cdot K_{loss}$) as a function of the alpha-pinene
 746 oxidation rate. In steady state the loss rate equals the formation rate, and thus
 747 the slope of the points gives the molar yield of HOM from the alpha-pinene +
 748 Cl reaction. The data corresponds to the conditions with 1, 2, 4 and 7 lights
 749 switched on in the chamber, respectively. The slope indicates an average molar
 750 yield of HOMs of 1.8%.

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