

**In-situ vertical characteristics of optical properties and heating rates  
of aerosol over Beijing**

Ping Tian<sup>1</sup>, Dantong Liu<sup>2\*</sup>, Delong Zhao<sup>3</sup>, Chenjie Yu<sup>4</sup>, Quan Liu<sup>1</sup>, Mengyu Huang<sup>1</sup>, Zhaoze Deng<sup>5</sup>, Liang Ran<sup>5</sup>, Yunfei Wu<sup>6</sup>, Shuo Ding<sup>2</sup>, Kang Hu<sup>2</sup>, Gang Zhao<sup>7</sup>, Chunsheng Zhao<sup>7</sup>, Deping Ding<sup>1,3</sup>

- 1. Beijing Key Laboratory of Cloud, Precipitation and Atmospheric Water Resources, Beijing, 100089, China.
- 2. Department of Atmospheric Science, School of Earth Sciences, Zhejiang University, Hangzhou, Zhejiang, 310027, China.
- 3. Field Experiment Base of Cloud and Precipitation Research in North China, China Meteorological Administration, Beijing, 100089, China.
- 4. Centre for Atmospheric Sciences, School of Earth and Environmental Sciences, University of Manchester, Manchester, M139PL, UK
- 5. Key Laboratory of Middle Atmosphere and Global Environment Observation, Institute of Atmospheric Physics, Chinese Academy of Sciences, Beijing, 100029, China.
- 6. CAS Key Laboratory of Regional Climate-Environment for Temperate East Asia, Institute of Atmospheric Physics, Chinese Academy of Sciences, Beijing, 100029, China.
- 7. Department of Atmospheric and Oceanic Sciences, Peking University, Beijing, 100871, China

*Corresponds to:* Dantong Liu (dantongliu@zju.edu.cn)

22 **Abstract.** Characterizing vertical profiles of aerosol optical properties is important because only relying  
23 on the surface or column-integrated measurements is unable to unambiguously constrain the radiative  
24 impacts of aerosol. This study presents series of vertical profiles of in-situ measured multi-wavelength  
25 optical properties of aerosols during three pollution events from Nov. to Dec. 2016 over Beijing region.  
26 For all pollution events, the clean periods (CP) before pollution initialization showed a higher scattering  
27 Ångström exponent and a smaller asymmetry parameter ( $g$ ), and relatively uniform vertical structures.  
28 The heavy pollution (HP) periods showed increased particle size, causing these parameters to vary in the  
29 opposite way. During the transition periods (TP), regional transport of aged aerosols at upper level was  
30 found. The AERONET aerosol optical depth (AOD) matched the in-situ measurements within 10 %,  
31 however the AERONET absorption optical depth (AAOD) was 10-20 % higher than in-situ measurements,  
32 and this positive discrepancy increased to 30 % at shorter wavelength. The absorption of brown carbon  
33 (BrC) was identified by increased absorption Angström exponent (AAE), and the heating rate of black  
34 carbon (BC) and BrC was calculated by computing the wavelength-dependent absorption coefficient and  
35 actinic flux by the radiative transfer model. BC and BrC had a heating rate up to 0.18 K/h and 0.05 K/h  
36 in the planetary boundary layer (PBL) respectively during the pollution period. The fraction of BrC  
37 absorption increased from 12 % to 40 % in the PBL from CP to HP period. Notably, higher contribution  
38 of BrC heating was found above the PBL under polluted condition. This study gives a full picture of  
39 shortwave heating impacts of carbonaceous aerosols during different stages of pollution event, and  
40 highlights the increased contribution of BrC absorption especially at higher level during pollution.

## 42 1. Introduction

43 The optical properties of aerosol, which is how aerosol scatters or absorbs solar radiation, have caused  
44 important radiative impacts on earth system (IPCC2013). The optical properties depend on the particle  
45 size (Bergin et al., 2001), refractive index (Ebert et al., 2002; Quinn, 2002) and mixing state of aerosols.  
46 There are still large uncertainties in evaluating the radiative forcing of aerosol especially in east Asia  
47 region due to lack of information on vertical distribution of these parameters (Liao and Seinfeld, 1998;  
48 Ramanathan et al., 2001; Li et al., 2017). Previous studies showed that the surface observation or column-  
49 integrated measurements may not provide sufficient information to derive vertical profiles of aerosol  
50 optical properties (Andrews et al., 2011; Rosati et al., 2016). Modelling studies found the radiative forcing  
51 impact is sensitive to the aerosol vertical distribution (Haywood et al., 1998), and especially for the  
52 absorbing aerosol such as black carbon (BC) will exert different climatic impacts depending on the  
53 location of aerosol layer (Yu et al., 2002; Ban-Weiss et al., 2011; Wilcox et al., 2016). Even though most  
54 aerosol was contained inside the planet boundary layer (PBL), the climatic sensitivity to absorbing aerosol  
55 rapidly increases with altitude (Ramanathan et al., 2001; Hodnebrog et al., 2014; Nazarenko et al., 2017).  
56 Absorbing aerosol above the PBL has the potential to suppress the PBL development and enhance the  
57 inversion cap at top of the PBL (Ding et al., 2016; Wang et al., 2018c), further execrating the pollution.  
58 However, this impact depends on the location of the absorbing layer which may also promote the  
59 convection by heating the layer above (Koch and Del Genio, 2010; Yu et al., 2019). It is therefore  
60 important to characterize the vertical profile of absorbing component in the atmosphere in order to  
61 understand its influence on atmospheric thermodynamics.

62 The North China Plain (NCP) has raised great attention in recent decades because of the severe air  
63 pollution and high frequency of haze days. The causes of pollution have been widely investigated through  
64 surface measurements (Zhang et al., 2013; Zhang et al., 2015; Zhong et al., 2018), however only limited  
65 studies have considered the evolution of pollutants in vertical direction (Tian et al., 2019; Wang et al.,  
66 2018a). It was found the surface aerosol concentration over Beijing not only depended on the emission  
67 but the vertical structure of aerosol distribution, which was largely dependent on local and synoptic  
68 meteorological conditions (Ran et al., 2016a; Zhao et al., 2019a), such as the mountain chimney effect

69 over Beijing region may introduce enhanced aerosol loading to high level (Chen et al., 2009). The light-  
70 absorbing aerosol mainly includes the species of black carbon (Bond et al., 2013), brown carbon (Lack  
71 and Cappa, 2010) and dust (Klingmüller et al., 2019), which have different spectral sensitivities to solar  
72 radiation. Different aerosol components dominate at different environments, and the heating rate caused  
73 by various aerosol sources has been studied over the world, e.g. for the anthropogenic sources over north  
74 America (Gao et al., 2008; Sahu et al., 2012; Liu et al., 2015b), Europe (Ferrero et al., 2014; Ferrero et  
75 al., 2018) and south Asia (Chakrabarty et al., 2012; Shamjad et al., 2015), and biomass burning sources  
76 over north and south America (Saleh et al., 2015; Zhang et al., 2017). However, ~~the data is still sparse~~  
77 regarding the vertical structures of heating rate, in addition to that, the heating was mostly evaluated using  
78 the measurement on the surface (Mallet et al., 2008; Wang et al., 2009) rather than using directly measured  
79 vertical profile. The calculation was performed for single species such as BC or BrC but most did not  
80 consider the co-impacts of all species (Chakrabarty et al., 2012; Chung et al., 2012; Shamjad et al., 2015).  
81 In the lower free troposphere, the ~~heating rate of aerosol in interacting~~ with boundary layer dynamics has  
82 raised much attention recently, as it may play important role in depressing boundary layer development  
83 hereby exacerbating the local pollution (Li et al., 2017). The heating rate caused by light absorbing aerosol  
84 was reported to vary as a function of height and range at 0.3-2.1 K/day for the polluted PBL over Europe  
85 (Kedia et al., 2010; Ferrero et al., 2014; Ferrero et al., 2018), and 0.3-2.5 K/day for south Asia (Tripathi  
86 et al., 2007; Ramana et al., 2007; Ramachandran and Kedia, 2010; Chakrabarty et al., 2012), but limited  
87 reports for the region of polluted east Asia.

88 This study chose three typical pollution events occurring in wintertime over Beijing, and performed  
89 continuous flights on daily basis for each event. The vertical profiles of multi-wavelength aerosol optical  
90 properties were in-situ characterized, accounting for all stages during pollution events from pollution  
91 starts, full development and cease. The directly measured optical parameters were used as inputs for  
92 radiative transfer calculation, hereby the heating rate of light-absorbing aerosols, including black and  
93 brown carbon (BrC) was estimated. The results here provide a full picture of vertical profiles of aerosol  
94 optical properties over Beijing region and investigate the radiative forcing effect of aerosol during the  
95 heavy pollution events.

## 2. Instrumentation and data analysis

A Kingair 350ER turbo aircraft in Beijing weather modification office was employed for the in-situ measurements over Beijing during the 2016 winter in this study. Meteorological parameters including the temperature, relative humidity, pressure, wind direction and wind speed with a time resolution of 1 s were measured by the Aircraft Integrated Meteorological Measurement System (AIMMS-20, Aventech Research Inc, Canada), which was calibrated annually. The aerosol instrumentation inside the cabin was connected to an isokinetic inlet (Model:1200, Brechtel Inc, USA), which can deliver particle with a high transport efficiency (90%) for sub-micrometer particles. The ~~maintained~~ room temperature (25 °C) in the cabin had self-drying effect when the temperature inside was higher than outside the cabin, in addition to which, a silicate drier was utilized ahead of all instruments to maintain the sampling RH lower than 40%. In-situ measurements of aerosol optical properties were performed during three pollution events over Beijing in Nov. 15<sup>th</sup> to Dec. 21<sup>th</sup> 2016, including 14 flights covering the start, development and cease stage for each pollution event. All flights were conducted around midday when the PBL was well developed. Table 1 summarizes the information of each flight. In order to compare the AOD from AERONET and calculate the vertical heating rates, only the cloud-free vertical profiles are used. In this study, three flights (20161117 12:00, 20161117 15:00, 20161118 12:00) were observed with cumulus clouds (Table 1). The in-cloud data in this study was screened out according to the in-situ measured cloud number concentration and liquid water content with a total number concentration of more than 10 cm<sup>-3</sup> and liquid water of more than 0.001 g m<sup>-3</sup> are not included in the following analysis (Deng et al., 2009). A micro pulse lidar (MPL, Sigma Inc, USA) was employed to measure the temporal evolution of aerosol extinction vertical profiles, and the vertical wind profile was measured by a wind profile radar with a vertical resolution of 150 m.

### 2.1 Aerosol optical properties

The aerosol scattering ( $\sigma_{sca}$ ) and hemispheric backscattering ( $\sigma_{bsca}$ ) coefficients at  $\lambda=450$  nm, 525 nm, and 650 nm were measured by an integrating nephelometer (Aurora3000, Ecotech Inc, Australia), and the

122 flowrate of Aurora3000 was maintained at 4 L/min during flight. The baseline of Aurora3000 in real time  
 123 was corrected for Rayleigh scattering of gas molecule at different air pressure (Fig. S1). In addition, the  
 124  $\sigma_{sca}$  and  $\sigma_{bsca}$  at all wavelengths were corrected for truncation affects (Anderson and Ogren, 1998; Müller  
 125 et al., 2009).

126 The scattering Ångström exponent (SAE) measures the wavelength dependence of  $\sigma_{sca}$  assuming a power  
 127 relationship with  $\lambda$ , expressed as:

$$128 \quad SAE = -\frac{\ln(\sigma_{\lambda_1}/\sigma_{\lambda_2})}{\ln(\lambda_1/\lambda_2)}, \quad (1)$$

129 where  $\sigma_{\lambda_l}$  denotes the  $\sigma_{sca}$  at  $\lambda_l$ , the value of SAE could also be used to reflect particles size with larger  
 130 particles showing a smaller SAE (Carrico et al., 1998).

131 The asymmetry parameter ( $g$ ) is obtained from measured backscattering fraction according to the  
 132 empirical function from Andrews et al. (2006).

$$133 \quad g = -7.143889 \cdot \beta^3 + 7.4633439 \cdot \beta^2 - 3.9356 \cdot \beta + 0.9893, \quad (2)$$

134 where  $\beta$  is the hemi-spherical backscatter fraction ( $\sigma_{bsca}/\sigma_{sca}$ ) measured by the Aurora3000.

135 The absorbing coefficient ( $\sigma_{abs}$ ) at different wavelengths (370, 470, 520, 590, 660, 880, and 950nm) was  
 136 measured by an Aethalometer (AE33, Magee Scientific Inc, USA) (Hansen, 2005). The flowrate of AE33  
 137 was maintained at 4 L/min below 3000 m. The shadowing effect of the AE33 was corrected by the two  
 138 spot measurements with different attenuation (Drinovec et al., 2017). The multiple scattering artifact of  
 139 AE33 was corrected by measuring the ambient aerosol in parallel with photoacoustic spectrometer  
 140 (PASS3, DMT Inc, USA), and the latter is independent of the filter artifacts. The PASS3 was calibrated  
 141 using the NO<sub>2</sub> and BC standard. (Arnott et al., 2005). Fig. S2 shows the two weeks' ambient measurements  
 142 between AE33 and PASS3 at three overlapped wavelengths. Multiple scattering correction factor of 2.88  
 143 was consistently found at three  $\lambda$ , which was applied to correct the AE33 measurement.

144 The absorbing Ångström exponent (AAE), which can weight the absorption at different wavelength, is  
 145 calculated using power fitting at seven wavelengths.

$$146 \quad \sigma_{abs}(\lambda) = \sigma_{abs,0} (\lambda/\lambda_0)^{AAE}, \quad (3)$$

We estimated the  $\sigma_{abs}$  of brown carbon (BrC) assuming that BC is the only absorber at  $\lambda=950$  nm, then the absorption of BC at other wavelengths was extrapolated by assuming an AAE of 1 (Kirchstetter et al., 2004; Lack et al., 2013; Massabò et al., 2015), and the contribution of BrC at each wavelength was obtained by subtracting the BC absorption from the total absorption (Schnaiter et al., 2005; Liu et al., 2015a). It should be noted that previous studies point out the  $AAE_{BC}$  may be less than 1, thus assuming  $AAE_{BC}=1$  may lead to underestimation of BrC contribution (Gyawali et al., 2009; Lack and Cappa, 2010; Feng et al., 2013). We therefore consider the results reported here is the lower bound of BrC contribution.

The single scattering albedo (SSA) is the ratio of the scattering coefficient over the extinction coefficient ( $\sigma_{ext}$ ) at a given wavelength.

The parameters  $\sigma_{sca}$ ,  $\sigma_{abs}$ , and  $\sigma_{ext}$  are reported as standard temperature and pressure (STP, 1013.25 hpa, 273.15K) for direct comparison at different altitudes among flights. Note that to compare with the AERONET results and for the radiative transfer calculation (as detailed in the following), these parameters in ambient conditions are used.

Column aerosol optical properties during the aircraft observation period were obtained from Aerosol Robotic Network (AERONET) sun-photometer network (Che et al., 2009; Xia et al., 2008), where the site (AERONET BEIJING\_PKU) was about 10 km away from the location of vertical profiles. The measurement of  $\sigma_{ext}$  was up to 2500m above which the aerosol concentration was low enough to be below the instrument lower detection limit. Given the very low concentration above 2500m, the value on top of 2500m was used to reconstruct the vertical profile up to 5000m. After that the  $\sigma_{ext}$  from 2.5-5 km only accounted for 1-2 % of the integrated columnar extinction.

To evaluate the potential influence of particle hygroscopicity on optical properties, the aerosol hygroscopic growth parameterization ( $f(RH)$ ) was used to calculate the enhancement of  $\sigma_{sca}$  under ambient RH. This function was previously measured by Zhao et al. (2019b) over Beijing region, expressed as:

$$f(RH) = a \cdot (1 - RH/100)^{-\nu(RH/100)} \quad (4)$$

where  $f(RH)$  was obtained by a comparison between a dry and humidified nephelometer in parallel. a

173 /  $\gamma$  was 0.930 / 0.329, 0.971 / 0.372, and 0.988 / 0.356 for clean, moderate, and heavy pollution period,  
174 respectively, according to the study.

175 The RH influence on  $g$  was calculated according to Zhao et al. (2018), expressed as:

176 
$$g(RH)/g(RH < 40\%) = a \cdot (1 - RH/100)^{-\gamma(RH/100)} \quad (5)$$

177 where  $a / \gamma$  was 0.9984 and 0.0849.

178 The resulting  $\sigma_{sca}$ ,  $\sigma_{ext}$ ,  $SSA$ , and  $g$  are all calculated for the hygroscopicity influence.

179

## 180 2.2 Radiative transfer calculation

181 The atmospheric irradiance and actinic flux using the pseudo-spherical version of the Discrete Ordinates  
182 Radiative Transfer Code (DISORT), as implemented in the libRadtran software package (Emde et al.,  
183 2016). The ~~aircraft~~ in-situ measured vertical profiles of AOD, single scattering albedo ( $SSA$ ) and  $g$  are  
184 used as inputs, and the other input parameters for the radiative transfer calculation is summarized in Table  
185 2. The calculation is performed for clear-sky condition only, thus the flights experiencing low-level clouds  
186 are not included in the calculation. The direct, upward diffuse, and downward diffuse irradiance and  
187 actinic flux (AF, in  $\text{mWm}^{-2}$ ) at  $\lambda = 250\text{-}2550$  nm are calculated. The calculation of AF is performed with  
188 and without aerosol input (AOD is set to zero) to evaluate the aerosol net impact. The heating rate is only  
189 calculated with considering the in-situ measured AOD. The spectral instantaneous absorbing power of  
190 BC ( $A_{BC}$ ) or BrC ( $A_{BrC}$ ) can be calculated by multiplying the absorption coefficient of BC (or BrC) and  
191 AF at specified  $\lambda$ , then integrating all  $\lambda$  will obtain the total absorbing power (Gao et al., 2008; Emde et  
192 al., 2016), expressed as:

193 
$$A_{BC \text{ or } BrC} = \int_{250\text{nm}}^{2550\text{nm}} AF(\lambda) \cdot \sigma_{BC \text{ or } BrC}(\lambda) d\lambda, \quad (6)$$

194 By assuming no radiative loss of solar energy and the heat absorbed by aerosol is fully transferred to the  
195 surrounding air, the instantaneous heating rate of BC or BrC to ambient air is hence calculated as:

196 
$$H_{BC, BrC} = A_{BC, BrC} / (\rho \cdot C_p), \quad (7)$$



197 where  $\rho$  and  $C_P$  are the air mass density ( $\text{kg/m}^3$ ) and heat capacity ( $1.007 \text{ J/g}\cdot\text{K}$ ), respectively. The profiles  
198 of aerosol optical properties influenced by hygroscopic growth (as calculated above) is also input in the  
199 calculation to work out its influence on heating rates.

200

### 201 **3. Results and discussions**

#### 202 **3.1 Overview and the pollution events**

203 Three pollution events from Nov. 15<sup>th</sup> to 18<sup>th</sup> (Case 1), Dec. 10<sup>th</sup> to 12<sup>th</sup> (Case2), and Dec. 16<sup>th</sup> to 19<sup>th</sup>  
204 (Case 3) in 2016 were captured. Figure 1 shows the temporal evolution of surface  $\text{PM}_{2.5}$ , AOD (AAOD)  
205 constrained by in-situ aircraft measurements and from AERONET, and vertical profiles of  $\sigma_{ext}$  and wind  
206 information during Case 1 pollution event. The other two events are shown in Fig. S3 and Fig. S4. Aircraft  
207 vertical profiles were performed on daily basis as the flight time indicated by vertical bars (Fig. 1). Each  
208 pollution event was classified as pollution initialization, development and peak pollution periods,  
209 corresponding to the pollution levels as clean period (CP,  $\text{PM}_{2.5, \text{surface}} < 35 \mu\text{g/cm}^3$ ), transition period (TP,  
210  $35 \mu\text{g/cm}^3 < \text{PM}_{2.5, \text{surface}} < 200 \mu\text{g/cm}^3$ ) and heavy pollution (HP,  $\text{PM}_{2.5, \text{surface}} > 200 \mu\text{g/cm}^3$ ). Three flights  
211 (20161117 12:00 , 20161117 15:00, 20161118 12:00) experienced boundary layer cloud (Fig. 1c), as  
212 indicated by the intensive extinction on top of the PBL. There were 3, 4 and 4 profiles in clear sky for LP,  
213 TP and HP period respectively (as detailed in Table 1). As Fig. 1b shows, wind sheer in both wind speed  
214 and direction appeared on top of PBL, consistent with the vertical distribution of  $\sigma_{ext}$  observed by a lidar  
215 (Fig. 1c). During LP, wind profiles (Fig. 1b) showed dominant northwesterly wind with high wind speed  
216 throughout the column, enhancing the pollutant dispersion in more developed PBL (Fig. 1c). During TP,  
217 the southerly air flow dominated and the  $\text{PM}_{2.5}$  mass concentration underwent a rapid increase from 30 to  
218  $100 \mu\text{g m}^{-3}$  in several hours. During HP, the windspeed was relatively low at all altitudes, maintaining the  
219  $\text{PM}_{2.5}$  mass concentration at a high level.

220 Figure 2 summarized the in-situ measured meteorological parameters at different stages of pollution  
221 events. The height of PBL (PBLH) was determined by considering a variety of factors. Firstly, a stable  
222 potential temperature ( $\theta$ ) (Fig. 2d-f) with vertical gradient  $d\theta/dz < 5 \text{ K/km}$  in the PBL indicated an

223 sufficient convective mixing (Su et al., 2017), with an apparent positive gradient above the PBL indicating  
 224 a stable layer (Petra Seibert, 2000). Secondly, there was usually a temperature inversion on top of the PBL  
 225 (Fig. 2a-c). During the CP, the weak temperature inversion ( $\sim 0.15\text{K}/100\text{m}$ ) on top of the PBL allowed  
 226 pollutants to penetrate the PBL and disperse in a higher atmospheric column (Fig. 2b). This inversion was  
 227 significantly enhanced for the TP and HP periods, to  $0.9\text{K}/100\text{m}$  and  $0.7\text{K}/100\text{m}$  respectively. The large  
 228 increase of the inversion during flight 20161211 was caused by regional transport from the south, when  
 229 lower-latitude warmer air mass was imposed onto the measurement point (Tian et al., 2019). Additionally,  
 230 the PBLH decreased gradually as pollution continued during the pollution event, in line with the enhanced  
 231 aerosol concentration in the PBL. The moisture had similar features that a lower moisture content showed  
 232 when lower pollution level and vertically efficiently dispersed, whereas stronger inversion also trapped  
 233 the moisture inside the PBL, leading to a positive vertical gradient with the maximum RH showing on  
 234 top of the PBL. There were some regional transport influences under TP, resulting in enhanced RH when  
 235 airmass was advected from the south (Fig. 2f).

236

### 237 3.2 Vertical profile of $\sigma_{ext}$ , $\sigma_{sca}$ and $\sigma_{abs}$

238 Figure 3 shows the vertical distribution of aerosol optical properties including extinction ( $\sigma_{ext}$ ), scattering  
 239 ( $\sigma_{sca}$ ) and absorbing ( $\sigma_{abs}$ ) coefficients. Different structures of vertical profiles were observed for CP, TP  
 240 and HP periods. During CP, aerosol concentration was low and showed uniform mixing inside the PBL,  
 241 with the  $\sigma_{ext}$ ,  $\sigma_{sca}$  and  $\sigma_{abs}$  ranging from  $220\text{--}270\text{ Mm}^{-1}$ ,  $180\text{--}240\text{ Mm}^{-1}$ , and  $30\text{--}50\text{ Mm}^{-1}$ , respectively. The  
 242 backward trajectories for the CP showed that the air masses were from the northwestern low emission  
 243 region (Fig. S5). TP showed about 4-fold increase of  $\sigma_{ext}$  compared to the CP. During TP, the  $\sigma_{ext}$ ,  $\sigma_{sca}$  and  
 244  $\sigma_{abs}$  had large variation inside the PBL, ranging from  $325\text{--}1435\text{ Mm}^{-1}$ ,  $300\text{--}1275\text{ Mm}^{-1}$ , and  $45\text{--}160\text{ Mm}^{-1}$ ,  
 245 respectively, and the mean PBLH decreased to  $200\text{--}500\text{ m}$ . During these pollution accumulation periods  
 246 (before the pollution reached peak level), two contrast vertical structures was observed. One showed well-  
 247 mixing in the PBL but declined concentration in the free troposphere (FT) (e.g. flight 20161115PM and  
 248 20161210) (Fig. 3a). The other one had the increased aerosol layer on top of the PBL, and showed positive  
 249 vertical gradient for all optical properties at certain level (e.g. flight 20161116 AM, 20161211 and

250 20161216) (Fig. 3b). The former was because of the mostly cleaner northwesterly air mass and higher  
251 wind speed influencing the layer above the PBL, while the latter resulted from the southwesterly regional  
252 transport (Tian et al., 2019).

253 During HP period, most flights showed consistent exponentially-declined vertical profile patterns, and  
254 the PBLH was even lower than that in TP (Fig. 2f). The stronger temperature inversion (Fig. 2c) and lower  
255 wind speed (Fig. 1b) inside the PBL led to high stability of the PBL and promoted the pollutant  
256 accumulation. The aerosol concentration was largely enhanced towards the surface and sharply declined  
257 above the PBL. Interestingly, the absorption showed higher degree of negative vertical gradient than the  
258 scattering at  $\lambda=440\text{nm}$ , which reflected the different sources and mixing ratios of absorbing and non-  
259 absorbing aerosols. The surface emission tends to contain more primary sources of absorbing particles  
260 such as BC and BrC, while enhanced secondary aerosol formation at upper level may add additional  
261 aerosol extinction.

262 The vertical profiles of  $\sigma_{sca}$  and  $\sigma_{abs}$  during HP can be fitted as:

263 
$$\sigma_{sca} = \sigma_{sca,0} \cdot \exp(-a * H); \quad a = 0.0012 \pm 0.0001, \quad (8)$$

264 
$$\sigma_{abs} = \sigma_{abs,0} \cdot \exp(-b * H); \quad b = 0.0015 \pm 0.0001, \quad (9)$$

265 where  $\sigma_0$  represent the surface value of  $\sigma_{sca}$  and  $\sigma_{abs}$ , and  $H$  is the altitude. The  $a$  and  $b$  are the parameters  
266 defining the changing rate with the altitude. This parameterization could be used to represent the vertical  
267 structure of optical properties under heavy pollution condition.

268 The hygroscopic effect on aerosol vertical profiles was mainly controlled by the ambient RH (shown in  
269 blue lines in Fig. 3). For most of the flights, the hygroscopic effect could be neglected due to low RH (<  
270 50%) (Fig. 2). For some of the flights (20161211),  $\sigma_{sca}$  and  $\sigma_{ext}$  especially at top of the PBL could be  
271 enhanced by a factor of 1.3.

272

### 273 3.3 Vertical profile of SSA, SAE, AAE and g

274 Figure 4 shows the vertical profiles of  $SSA$ ,  $SAE$ ,  $AAE$ , and  $g$  for all the flights during different stages of

275 pollution events. Overall, the SSA showed two modes inside the PBL. Under the CP, SSA for most flights  
 276 was populated at 0.85, and had less variation throughout the column in the PBL. Flight 20161115AM  
 277 showed a strong elevation of SSA (0.94) at 2200 m (Fig. 4a), which may be influenced by a dust layer (as  
 278 further discussed below). SSA showed positive vertical gradient for the TP and HP inside the PBL, i.e.  
 279 from the surface to the PBLH, the mean SSA increased from 0.85 to 0.91 and from 0.87 to 0.92 for TP  
 280 and HP period, respectively. This indicates the reduced fraction of absorbing particles, in turn suggesting  
 281 an enhancement of secondary production of non-absorbing particles. There were a few profiles featuring  
 282 with large enhancement of SSA ( $>0.95$ , for flight 20161211) at high altitude (Fig. 4b), and backward  
 283 trajectory analysis (Fig. S5) showed that these resulted from regional transport when more aged pollutants  
 284 were advected to a high altitude. The SSA in the FT was mostly higher than that in the PBL and maintained  
 285 at 0.9-0.95 for TP and HP, suggesting a lower absorbing particle fraction at higher altitude. Comparing  
 286 among different stages during pollution event, it could be concluded that at the initialization stage of  
 287 pollution when the total PM was relatively low, a lower SSA exhibited, while the increase of pollution  
 288 level added more secondary species, hence increasing SSA. This trend was consistent with previous  
 289 ground studies in Beijing (He et al., 2009; Jing et al., 2011).

290 The SAE reflects the particle size with larger size having a smaller SAE. A decreasing SAE was shown  
 291 for increasing pollution levels inside the PBL (Fig. 4), i.e., from CP to HP, the SAE in the PBL showed  
 292 an average value of 1.74, 1.45, and 1.21, respectively. For most of the profiles, SAE showed enhancement  
 293 at higher altitude. This means smaller particle sizes at high altitude, which may result from a higher  
 294 scavenging efficiency for larger particles where smaller particles remained un-scavenged in the upper  
 295 level (Liu et al., 2009). These was an exception of flight 20161211, when regional advection transported  
 296 larger and aged particles to the higher altitude. The particle size also corresponded with asymmetry  
 297 parameter ( $g$ , Fig. 4j-i), with larger particle presents more fraction of forward scattering (larger  $g$ ). Note  
 298 that there only one flight (flight 20161211) under  $RH > 80\%$ , where the particle hygroscopicity had  
 299 appreciable influence on *SSA* (increased by 0.05), *SAE* (decreased by 0.2) and  $g$  (increased by 0.1).

300 AAE reflects the degree of absorption towards shorter wavelength, such as the presence of BrC will  
 301 enhance the absorption in the UV. A lower AAE  $1.2 \pm 0.2$  was shown for the CP (Fig. 4g), but increased  
 302 to  $1.56 \pm 0.3$  for TP in the PBL (Fig. 4h), and additional higher mode of AAE showed at 1.8-2.0 for the

HP period (Fig. 4i). There was weak variation of AAE for CP throughout the column, but became largely spreading for TP, i.e., with either positive or negative vertical gradient at different levels. Notably, the AAE showed consistent positive vertical gradient for most of the HP profiles (Fig. 4i). This implied the enhancement of BrC contribution at higher altitude for polluted troposphere. Flight 20161115AM showed a notably increased AAE up to 2 at altitude 2 km (Fig. 4g), which may reflect the influence of dust (Cazorla et al., 2013). The ground AAE had strong seasonal variation with winter normally showing a higher AAE due to higher emissions of solid fuel burning (Sun et al., 2017; Wang et al., 2018b). However, there is still lack of results on the vertical characteristics of AAE due to limited measurements, and the results here highlight the enhancement of BrC at high level, mainly for polluted environment.

312

### 313 3.4 Comparison of column integrated and in-situ constrained AOD/AAOD

314 To compare the AOD and AAOD between AERONET and that constrained by in-situ aircraft  
315 measurements, the AERONET data was chosen to match with the aircraft profiles in time ( $\pm 3h$ ) and  
316 location (within 10 km) (the PEK site). The comparison was performed at overlapped wavelengths  
317 (440nm, 675nm, and 870nm) between AERONET and aircraft instruments. As Fig. 5a-c shows, high  
318 correlation ( $R^2 > 0.95$ ) was found between columnar and in-situ measurement. In particular, the  
319 correlation was ~~most unit under~~ dry condition ( $RH < 40\%$ ), while the AERONET was about 10-20%  
320 higher than in-situ measurement when  $RH > 60\%$ . Improved agreement was achieved by 8-15% if  
321 considering aerosol hygroscopic growth (open circle in Fig. 5a-c), despite that in-situ constrained AOD  
322 was still 2-5% lower than AERONET after the hygroscopic correction.

323 Figure 5g-i shows at three wavelengths the AAOD had lower correlations between both methods compare  
324 to AOD, with  $R^2 = 0.75, 0.58, \text{ and } 0.49$  at 440 nm, 675 nm, and 870nm, respectively. The columnar AAOD  
325 was overall about 10-25% higher than in-situ measurement, and this AERONET AAOD overestimation  
326 was higher under higher AOD condition. This is consistent with previous findings conducted over ~~US~~ that  
327 the ~~retrieved~~ AAOD from AERONET was biased higher when compared to in-situ measurement  
328 (Andrews et al., 2017).

Note that there was better agreement during CP, when lower pollution level and lower RH (shown in blue dots). This suggests a lower moisture and less AOD interface may improve the agreement of AAOD. As previous studies pointed out that the retrieval of the AERONET was sensitive to the variation of aerosol vertical distribution (Torres et al., 2014). We speculate that the better agreement for CP was due to the vertically homogeneous distribution of aerosol optical properties, and larger bias for CP and HP periods might be caused by the significant variations of the vertical profiles. Other factors like the aerosol hygroscopic growth under higher RH may introduce factors in enhancing the absorption, e.g. more lensing effect on BC absorption via thicker and moisture coating (Wu et al., 2017). Though this study is not able to rule out the exact influencing factor in causing this discrepancy, an overestimation of 25% in the AERONET AAOD under polluted condition is shown for the dataset here.

### 3.5 Heating impacts of BC and BrC

Figure 6 shows vertical profiles of irradiances from radiative transfer calculation using in-situ measurements as model inputs (Table 1). The results show that the presence of aerosols reduced the direct irradiance reaching the surface (Fig. 6a-c) but increased the upward diffuse (Fig. 6d-f) and downward diffuse irradiances, especially above the PBL (Fig. 6g-i). The direct irradiance on the surface ranged from  $1 \times 10^9$  to  $3.5 \times 10^9$  mW m<sup>-2</sup>, with an average of  $2.2 \times 10^9$  mW m<sup>-2</sup> during CP (Fig. 6a), which was about two-fold and three-fold larger than that during TP (Fig. 6b) and HP period (Fig. 6c), respectively. The combined direct, diffuse upward and downward irradiance which forms the actinic flux (AF), showed an enhancement above the PBL and a reduction within the PBL (Fig. 7a-c), but to what extent the enhancement or reduction occurred depends on the aerosol vertical profile. The vertical gradient of AF was slightly modified by aerosol loadings during CP, whereas for the TP and HP, aerosol effects caused AF about two times smaller within the PBL and 20 % larger above the PBL, leading to an increased vertical gradient of AF. The AF received at lower level was reduced by up to 10 % by incorporating the aerosol hygroscopicity influence (Fig. 7) due to enhanced AOD, and AF was further redistributed to give larger vertical gradient (Fig. 7a-c).

The vertical profiles of absorbing power and heating rate of BC are shown in Fig. 7d-f. Vertically homogeneous BC heating rates of 0.05 K/h was found inside the PBL during CP (Fig. 7d). During the

regional transport cases (flight20161211 and flight20161216) for TP, positive vertical gradient (increase with increasing altitude) of BC heating rates was observed, and as high as 0.1 K/h heating rate could occur at top of the PBL height (Fig. 7e). During the HP period, negative heating rate (decrease with increasing altitude) of BC was found except for one flight on 20161212 in Case 2, and the BC heating rate at the surface could reach as high as 0.15 K/h (Fig. 7f). The reason causing negative vertical gradient of BC heating rate was the higher degree of negative gradient of  $\sigma_{abs}$  (Fig. 3i) than the positive gradient of AF (Fig. 7). The results here show that the atmospheric heating by aerosol was mainly inside the PBL and for polluted period the BC-induced heating was 0.05-0.17 K/h, generally consistent with previous studies over the polluted Asia region, with 0.02-0.17 K/h (Ramana et al., 2007; Ramana et al., 2010; Kedia et al., 2010).

The contribution of BrC to absorbing power and heating rates was computed as the integrated portion of absorption over visible wavelength (370–950nm in this study) by subtracting the BC absorption. Figure 7g-i shows the vertical profile of BrC heating rate. Continuously increase of BrC heating rate in the PBL was observed from CP to HP, with mean heating rate of 0.02 K/h, 0.03 K/h, and 0.05 K/h during CP, TP and HP respectively. Though the BC was the main contributor to the heating in the PBL, the heating of BrC was more evenly distributed and could be comparable with the BC heating rate at high altitude especially during HP period (Fig. 7i). The contribution of BrC to the total absorption was reported to be 10-27 % over polluted region of Europe (Ferrero et al., 2018) and south Asia (Chung et al., 2012; Shamjad et al., 2015), in general consistent with results during polluted periods here.

Corresponding with the aerosol hygroscopicity influence on the actinic flux, the heating rate showed lowered intensity but enhanced vertical gradient for the flights with high ambient RH (Fig. 7b). The vertical gradient of the overall heating rate from absorbing components, i.e. increase or decrease heating rate with altitude, will importantly determine the influence on atmospheric stability. If the heating occurred near surface (Case 3), the lower layer will be heated leading to enhanced convective mixing (Sühring et al., 2014; Petaja et al., 2016); whereas if heating was above the PBL (Case 2), an increase of temperature inversion will occur hence inhibit the PBL development trapping the pollutants in the PBL (Chakrabarty et al., 2012; Tripathi et al., 2007). This study showed positive vertical gradient for 30 % of the flights especially under high pollution, and in particular during regional transport when pollutants

were advected from outside of Beijing and showed elevation of absorption at higher altitude (Fig. 8). The rest of the flights showed highly accumulated aerosol concentration near surface, also found by a previous study (Ferrero et al., 2014), when BC ~~would~~ potentially promoted the dispersion in the PBL and decreased its stability.

### 3.6 The importance of BrC heating effects

Figure 8 shows the measured absorption coefficient of BrC and BC inside and above the PBL at different  $\lambda$  for CP, TP and HP period, respectively. The results suggested that both  $\sigma_{abs}$  of BC and BrC increased with the pollution level, e.g. the  $\sigma_{abs}$  at  $\lambda = 440\text{nm}$  was  $42.8 \text{ Mm}^{-1}$  and  $7.2 \text{ Mm}^{-1}$  on average in the PBL and above the PBL respectively under HP period, and was  $4.7 \text{ Mm}^{-1}$  and  $1.3 \text{ Mm}^{-1}$  for LP. The contribution of BrC to total  $\sigma_{abs}$  was found to increase from CP to HP period (Fig. 8c, f). This is in line with previous studies in urban Beijing that more BrC contribution to total absorption was found under higher pollution level (Xie et al., 2019; Ran et al., 2016b), suggesting the important role of BrC on absorption under polluted condition.

The contribution of BrC to total heating rate showed notably different vertical structures. During CP, all profiles showed consistently low BrC contribution throughout the column, with about 7 % at the surface and 9 % in the FT (Fig. 8g). This means the low primary emission or the emission after being diluted by clean air mass did not contained large fraction of absorbing organics. During TP, BrC contribution inside the PBL increased to 22 % and showed considerable variation at higher level (Fig. 8h). During HP period, the surface contribution was comparable with that in TP, but showed remarkably enhanced BrC heating contribution at higher altitude, with a vertically increasing rate of 1.5 %/m in the PBL and reached as high as 45 % in the LFT. The higher heating contribution of BrC at higher altitude means the BrC absorption played an important role in heating at upper level, which may enhance the temperature inversion at that level hereby inhibit the convective mixing under the heated layer.

By comparing the BrC heating contribution at the surface, there was an increase from CP to TP, however, not from TP to HP. This suggests the primary emission will increase the BrC fraction from CP to TP, but for even more pollution environment from TP to HP, the primary emission may provide limited further



increased fraction of BrC. The primary BrC may result from a range of combustion sources, with the polluted region at the south of Beijing may contain higher fraction of residential coal burning sources (Sun et al., 2017; Xie et al., 2019) which may influence the Beijing region under polluted period. The relatively consistent BrC contribution at ~20 % from TP to HP suggested the relatively uniform BrC profiles for the primary sources. During TP, the BrC contribution above the PBL had rarely been above 30 % (Fig. 8h), however during HP, there was further enhancement of BrC contribution up to 45 % above the PBL (Fig. 8i). Note that there was no direct injection of biomass burning plume to the high altitude during the study period, the higher portion of BrC absorption above the PBL during HP may be formed through secondary production in addition to the primary source contribution. As Fig. 7a-c shows, there was more intensive actinic flux received at higher altitude and this will promote the photochemical reactions of gas-phase species, allowing more secondary formation of aerosol which may contain a fraction of BrC (Feng et al., 2013; Nakayama et al., 2013). Previous studies also found promoted BrC formation with light source under certain RH (Nguyen et al., 2012; Updyke et al., 2012; Laskin et al., 2015; Zhao et al., 2015). The positive gradient of BrC heating contribution was more likely resulted from enhanced RH from the surface to the top of PBL (Fig. 2i), because increased moisture will promote the aqueous reaction and gas-to-aerosol conversion which may also form part of the BrC observed here (Ervens et al., 2011; Nakayama et al., 2013). The secondary formation of BrC also requires the inorganic or VOC precursors being transported to the high level, therefore the enhancement of BrC mostly occurred under higher pollution level when sufficient gas precursors was transport to the level. The BrC may be also subject to bleaching process and lose the absorbance (Sareen et al., 2013; Lee et al., 2014; Wong et al., 2019), because the profiles in this study were conducted over an urban megacity where the sampled pollutants were fairly young and may have not experienced sufficient ageing for BrC to be degraded.

#### 4. Conclusions

This study provides detailed characterization of vertical profiles of aerosol optical properties over the Beijing region by continuous aircraft in-situ measurements at different stages during the pollution events. The results combining the direct measurements of scattering and absorption at multiple wavelength, give a full picture of how the optical properties had evolved at different layers during typical pollution event.

437 During clean period for pollution initialization (CP), the aerosols showed relatively uniform  
438 characteristics throughout the planetary boundary layer (PBL) and lower free troposphere (FT), such as  
439 lowing scattering or absorption coefficient, larger SAE (due to smaller particle size) and lower fraction  
440 of brown carbon (BrC) reflected by smaller AAE. The transition period (TP) when pollution was  
441 developing had large variation of all optical properties, and enhanced aerosol loadings at higher altitude  
442 were encountered when being influenced by regional advection. The fully developed heavy pollution  
443 period (HP) featured with the shallow PBL accumulated over 80 % of the scattering and absorption within  
444 the PBL, and deceased SAE due to enlarged particles size. Notably the absorption towards shorter  
445 wavelength became larger under more polluted environment, especially for the higher altitude.

446 The AOD and AAOD measured by passive remote sensing was for the first time compared with in-situ  
447 measurements over this polluted region. AOD showed high correlation between AERONET and in-situ  
448 measurement within 10 %, and the most discrepancy between both could be possibly resolved by  
449 considering the hygroscopic growth of aerosols under high RH condition. The AAOD however showed  
450 10-25 % higher for remote sensing especially at shorter wavelength, consistent with other studies(Müller  
451 et al., 2012;Andrews et al., 2017). The possibilities of causing this could be the non-homogeneously  
452 vertically structure of optical properties, mixing state of light-absorbing aerosol, and also the particle  
453 hygroscopic growth, which are unable to be ruled out only using the results here.

454 BC was the main heating species, having 0.05 K/h, 0.1 K/h and 0.15 K/h heating rate at local time 12:00  
455 to 15:00 in the PBL during pollution initialization, transition and full development respectively, and  
456 showed positive vertical gradient of heating during regional transport period when pollution was advected  
457 at high level from the polluted south region outside of Beijing (Tian et al., 2019). The contribution of BrC  
458 to heating rate was found to increase by 20 % throughout the column from CP to HP period, in particular  
459 the increased BrC contribution was pronounced at the layer above the PBL during HP, which was  
460 proposed to result from the intensive photochemical reactions above the PBL. The BrC present at this  
461 layer will have the potential to contribute to the heating at this layer, hence enhancing the temperature  
462 inversion on top of the PBL hereby the capping effect to the pollutants trapped in the PBL. Particles at  
463 higher altitude may be transported to wider region spatially in both vertical and horizontal directions  
464 through convection, which may lead BrC present at this layer to have wilder and longer radiative impacts.

465 Different mechanism of BrC formation at different levels such as above the PBL (where more solar flux  
466 received) or within the PBL (where more moisture was constrained) warrants future studies.

467

468

469 **Data availability.** All data in this paper are available from the authors upon request  
470 (tianping@bj.cma.gov.cn).

471 **Competing interests.** The authors declare no conflicts of interest.

472 **Author contribution.** D. D., and M. H., led and designed the study; P. T., and D. L., designed the study,  
473 set up the experiment, analyzed the data, and wrote the paper. P. T., D. Z., and Q. L., conducted the aircraft  
474 observation. C. Y., performed the radiative transfer model calculation. P. T., D. L., Z. D., L. R., and Y. W.,  
475 contributed to the aircraft data analysis. S. D., and K. H., contribute to the surface data analysis. G. Z.,  
476 and C. Z., conducted the aerosol absorption comparison experiment.

477 **Acknowledgment.** This research was supported by the National key Research and Development Program  
478 of China (2016YFA0602001), and the National Natural Science Foundation of China (41875044,  
479 41675038, 41875167). Part of this work is supported by the National Center of Meteorology, Abu Dhabi,  
480 UAE under the UAE Research Program for Rain Enhancement Science.

481

482 **Reference**

- 483 Anderson, T. L., and Ogren, J. A.: Determining Aerosol Radiative Properties Using the TSI 3563  
 484 Integrating Nephelometer, *Aerosol Sci. Technol.*, 29, 57-69, 10.1080/02786829808965551, 1998.
- 485 Andrews, E., Sheridan, P. J., Fiebig, M., McComiskey, A., Ogren, J. A., Arnott, P., Covert, D., Elleman,  
 486 R., Gasparini, R., Collins, D., Jonsson, H., Schmid, B., and Wang, J.: Comparison of methods for deriving  
 487 aerosol asymmetry parameter, *J. Geophys. Res.*, 111, 10.1029/2004jd005734, 2006.
- 488 Andrews, E., Sheridan, P. J., and Ogren, J. A.: Seasonal differences in the vertical profiles of aerosol  
 489 optical properties over rural Oklahoma, *Atmos. Chem. Phys.*, 11, 10661-10676, 10.5194/acp-11-10661-  
 490 2011, 2011.
- 491 Andrews, E., Ogren, J. A., Kinne, S., and Samset, B.: Comparison of AOD, AAOD and column single  
 492 scattering albedo from AERONET retrievals and in situ profiling measurements, *Atmos. Chem. Phys.*, 17,  
 493 6041-6072, 10.5194/acp-17-6041-2017, 2017.
- 494 Arnott, W. P., Hamasha, K., Moosmüller, H., Sheridan, P. J., and Ogren, J. A.: Towards Aerosol Light-  
 495 Absorption Measurements with a 7-Wavelength Aethalometer: Evaluation with a Photoacoustic  
 496 Instrument and 3-Wavelength Nephelometer, *Aerosol Sci. Technol.*, 39, 17-29,  
 497 10.1080/027868290901972, 2005.
- 498 Ban-Weiss, G. A., Cao, L., Bala, G., and Caldeira, K.: Dependence of climate forcing and response on  
 499 the altitude of black carbon aerosols, *Clim. Dynam.*, 38, 897-911, 10.1007/s00382-011-1052-y, 2011.
- 500 Bergin, M. H., Cass, G. R., Xu, J., Fang, C., Zeng, L. M., Yu, T., Salmon, L. G., Kiang, C. S., Tang, X.  
 501 Y., Zhang, Y. H., and Chameides, W. L.: Aerosol radiative, physical, and chemical properties in Beijing  
 502 during June 1999, *J. Geophys. Res.*, 106, 17969-17980, Doi 10.1029/2001jd900073, 2001.
- 503 Bond, T. C., Doherty, S. J., Fahey, D. W., Forster, P. M., Berntsen, T., DeAngelo, B. J., Flanner, M. G.,  
 504 Ghan, S., Kärcher, B., Koch, D., Kinne, S., Kondo, Y., Quinn, P. K., Sarofim, M. C., Schultz, M. G.,  
 505 Schulz, M., Venkataraman, C., Zhang, H., Zhang, S., Bellouin, N., Guttikunda, S. K., Hopke, P. K.,  
 506 Jacobson, M. Z., Kaiser, J. W., Klimont, Z., Lohmann, U., Schwarz, J. P., Shindell, D., Storelvmo, T.,  
 507 Warren, S. G., and Zender, C. S.: Bounding the role of black carbon in the climate system: A scientific  
 508 assessment, *J. Geophys. Res. Atmos.*, 118, 5380-5552, 10.1002/jgrd.50171, 2013.
- 509 Carrico, C. M., Rood, M. J., and Ogren, J. A.: Aerosol light scattering properties at Cape Grim, Tasmania,

510 during the First Aerosol Characterization Experiment (ACE 1), *J. Geophys. Res.*, 103, 16565,  
 511 10.1029/98jd00685, 1998.

512 Cazorla, A., Bahadur, R., Suski, K. J., Cahill, J. F., Chand, D., Schmid, B., Ramanathan, V., and Prather,  
 513 K. A.: Relating aerosol absorption due to soot, organic carbon, and dust to emission sources determined  
 514 from in-situ chemical measurements, *Atmos. Chem. Phys.*, 13, 9337-9350, 10.5194/acp-13-9337-2013,  
 515 2013.

516 Chakrabarty, R. K., Garro, M. A., Wilcox, E. M., and Moosmüller, H.: Strong radiative heating due to  
 517 wintertime black carbon aerosols in the Brahmaputra River Valley, *Geophys. Res. Lett.*, 39, n/a-n/a,  
 518 10.1029/2012gl051148, 2012.

519 Che, H., Zhang, X., Chen, H., Damiri, B., Goloub, P., Li, Z., Zhang, X., Wei, Y., Zhou, H., Dong, F., Li,  
 520 D., and Zhou, T.: Instrument calibration and aerosol optical depth validation of the China Aerosol Remote  
 521 Sensing Network, *J. Geophys. Res.*, 114, 10.1029/2008jd011030, 2009.

522 Chen, Y., Zhao, C., Zhang, Q., Deng, Z., Huang, M., and Ma, X.: Aircraft study of Mountain Chimney  
 523 Effect of Beijing, China, *J. Geophys. Res.*, 114, 10.1029/2008jd010610, 2009.

524 Chung, C. E., Ramanathan, V., and Decremier, D.: Observationally constrained estimates of carbonaceous  
 525 aerosol radiative forcing, *PNAS*, 109, 5, 10.1073/pnas.1203707109/-/DCSupplemental, 2012.

526 Ding, A. J., Huang, X., Nie, W., Sun, J. N., Kerminen, V. M., Petäjä, T., Su, H., Cheng, Y. F., Yang, X. Q.,  
 527 Wang, M. H., Chi, X. G., Wang, J. P., Virkkula, A., Guo, W. D., Yuan, J., Wang, S. Y., Zhang, R. J., Wu,  
 528 Y. F., Song, Y., Zhu, T., Zilitinkevich, S., Kulmala, M., and Fu, C. B.: Enhanced haze pollution by black  
 529 carbon in megacities in China, *Geophys. Res. Lett.*, 43, 2873-2879, 10.1002/2016gl067745, 2016.

530 Drinovec, L., Gregorič, A., Zotter, P., Wolf, R., Bruns, E. A., Prévôt, A. S. H., Petit, J.-E., Favez, O.,  
 531 Sciare, J., Arnold, I. J., Chakrabarty, R. K., Moosmüller, H., Filep, A., and Močnik, G.: The filter-loading  
 532 effect by ambient aerosols in filter absorption photometers depends on the coating of the sampled particles,  
 533 *Atmos. Meas. Tech.*, 10, 1043-1059, 10.5194/amt-10-1043-2017, 2017.

534 Ebert, M., Weinbruch, S., Rausch, A., Gorzawski, G., Helas, G., Hoffmann, P., and Wex, H.: Complex  
 535 refractive index of aerosols during LACE 98 as derived from the analysis of individual particles, *J.*  
 536 *Geophys. Res. Atmos.*, 107, LAC 3-1-LAC 3-15, 10.1029/2000jd000195, 2002.

537 Emde, C., Buras-Schnell, R., Kylling, A., Mayer, B., Gasteiger, J., Hamann, U., Kylling, J., Richter, B.,

538 Pause, C., Dowling, T., and Bugliaro, L.: The libRadtran software package for radiative transfer  
539 calculations (version 2.0.1), *Geosci. Model Dev.*, 9, 1647-1672, 10.5194/gmd-9-1647-2016, 2016.

540 Ervens, B., Turpin, B. J., and Weber, R. J.: Secondary organic aerosol formation in cloud droplets and  
541 aqueous particles (aqSOA): a review of laboratory, field and model studies, *Atmos. Chem. Phys.*, 11,  
542 11069-11102, 10.5194/acp-11-11069-2011, 2011.

543 Feng, Y., Ramanathan, V., and Kotamarthi, V. R.: Brown carbon: a significant atmospheric absorber of  
544 solar radiation?, *Atmos. Chem. Phys.*, 13, 8607-8621, 10.5194/acp-13-8607-2013, 2013.

545 Ferrero, L., Castelli, M., Ferrini, B. S., Moscatelli, M., Perrone, M. G., Sangiorgi, G., amp, apos, Angelo,  
546 L., Rovelli, G., Moroni, B., Scardazza, F., Močnik, G., Bolzacchini, E., Petitta, M., and Cappelletti, D.:  
547 Impact of black carbon aerosol over Italian basin valleys: high-resolution measurements along vertical  
548 profiles, radiative forcing and heating rate, *Atmos. Chem. Phys.*, 14, 9641-9664, 10.5194/acp-14-9641-  
549 2014, 2014.

550 Ferrero, L., Mocnik, G., Cogliati, S., Gregoric, A., Colombo, R., and Bolzacchini, E.: Heating Rate of  
551 Light Absorbing Aerosols: Time-Resolved Measurements, the Role of Clouds, and Source Identification,  
552 *Environ Sci Technol*, 52, 3546-3555, 10.1021/acs.est.7b04320, 2018.

553 Gao, R. S., Hall, S. R., Swartz, W. H., Schwarz, J. P., Spackman, J. R., Watts, L. A., Fahey, D. W., Aikin,  
554 K. C., Shetter, R. E., and Bui, T. P.: Calculations of solar shortwave heating rates due to black carbon and  
555 ozone absorption using in situ measurements, *J. Geophys. Res.*, 113, 10.1029/2007jd009358, 2008.

556 Gyawali, M., Arnott, W. P., and Moosmuller, H.: In situ aerosol optics in Reno, NV, USA, during and  
557 after the summer 2008 California wildfires and the influence of absorbing and non-absorbing organic  
558 coatings on spectral light absorption, *Atmos. Chem. Phys.*, 9, 8009-8015, 2009.

559 Hansen, A. D. A.: *The Aethalometer Manual*, Berkeley, California, USA, pp., 2005.

560 Haywood, J. M., Schwarzkopf, M. D., and Ramaswamy, V.: Estimates of radiative forcing due to modeled  
561 increases in tropospheric ozone, *J. Geophys. Res. Atmos.*, 103, 16999-17007, 10.1029/98jd01348, 1998.

562 He, X., Li, C. C., Lau, A. K. H., Deng, Z. Z., Mao, J. T., Wang, M. H., and Liu, X. Y.: An intensive study  
563 of aerosol optical properties in Beijing urban area, *Atmos. Chem. Phys.*, 9, 8903-8915, 2009.

564 Hodnebrog, O., Myhre, G., and Samset, B. H.: How shorter black carbon lifetime alters its climate effect,  
565 *Nat. Commun.*, 5, 1-7, 10.1038/ncomms6065, 2014.

566 Jing, J. S., Zhang, R. J., Tao, J., and Zhang, L. M.: Observations of Aerosol Optical Properties in the  
567 Beijing Urban Area in Summer, *Atmos. Sci. Lett.*, 4, 338-343, 2011.

568 Kedia, S., Ramachandran, S., Kumar, A., and Sarin, M. M.: Spatiotemporal gradients in aerosol radiative  
569 forcing and heating rate over Bay of Bengal and Arabian Sea derived on the basis of optical, physical,  
570 and chemical properties, *J. Geophys. Res.*, 115, 10.1029/2009jd013136, 2010.

571 Kirchstetter, T. W., Novakov, T., and Hobbs, P. V.: Evidence that the spectral dependence of light  
572 absorption by aerosols is affected by organic carbon, *J. Geophys. Res. Atmos.*, 109, n/a-n/a,  
573 10.1029/2004jd004999, 2004.

574 Klingmüller, K., Lelieveld, J., Karydis, V. A., and Stenchikov, G. L.: Direct radiative effect of dust–  
575 pollution interactions, *Atmos. Chem. Phys.*, 19, 7397-7408, 10.5194/acp-19-7397-2019, 2019.

576 Koch, D., and Del Genio, A. D.: Black carbon semi-direct effects on cloud cover: review and synthesis,  
577 *Atmos. Chem. Phys.*, 10, 7685-7696, 10.5194/acp-10-7685-2010, 2010.

578 Lack, D. A., and Cappa, C. D.: Impact of brown and clear carbon on light absorption enhancement, single  
579 scatter albedo and absorption wavelength dependence of black carbon, *Atmos. Chem. Phys.*, 10, 4207-  
580 4220, 10.5194/acp-10-4207-2010, 2010.

581 Lack, D. A., Bahreni, R., Langridge, J. M., Gilman, J. B., and Middlebrook, A. M.: Brown carbon  
582 absorption linked to organic mass tracers in biomass burning particles, *Atmos. Chem. Phys.*, 13, 2415-  
583 2422, 10.5194/acp-13-2415-2013, 2013.

584 Laskin, A., Laskin, J., and Nizkorodov, S. A.: Chemistry of atmospheric brown carbon, *Chemical reviews*,  
585 115, 4335-4382, 10.1021/cr5006167, 2015.

586 Lee, H. J., Aiona, P. K., Laskin, A., Laskin, J., and Nizkorodov, S. A.: Effect of solar radiation on the  
587 optical properties and molecular composition of laboratory proxies of atmospheric brown carbon, *Environ.*  
588 *Sci. Technol.*, 48, 10217-10226, 10.1021/es502515r, 2014.

589 Li, Z., Guo, J., Ding, A., Liao, H., Liu, J., Sun, Y., Wang, T., Xue, H., Zhang, H., and Zhu, B.: Aerosol  
590 and boundary-layer interactions and impact on air quality, *National Science Review*, 4, 810-833,  
591 10.1093/nsr/nwx117, 2017.

592 Liao, H., and Seinfeld, J. H.: Radiative forcing by mineral dust aerosols: Sensitivity to key variables, *J.*  
593 *Geophys. Res. Atmos.*, 103, 31637-31645, 10.1029/1998jd200036, 1998.

594 Liu, D., Taylor, J. W., Young, D. E., Flynn, M. J., Coe, H., and Allan, J. D.: The effect of complex black  
 595 carbon microphysics on the determination of the optical properties of brown carbon, *Geophys. Res. Lett.*,  
 596 42, 613-619, 10.1002/2014gl062443, 2015a.

597 Liu, J., Scheuer, E., Dibb, J., Diskin, G. S., Ziemba, L. D., Thornhill, K. L., Anderson, B. E., Wisthaler,  
 598 A., Mikoviny, T., Devi, J. J., Bergin, M., Perring, A. E., Markovic, M. Z., Schwarz, J. P., Campuzano-Jost,  
 599 P., Day, D. A., Jimenez, J. L., and Weber, R. J.: Brown carbon aerosol in the North American continental  
 600 troposphere: sources, abundance, and radiative forcing, *Atmos. Chem. Phys.*, 15, 7841-7858,  
 601 10.5194/acp-15-7841-2015, 2015b.

602 Liu, P., Zhao, C., Liu, P., Deng, Z., Huang, M., Ma, X., and Tie, X.: Aircraft study of aerosol vertical  
 603 distributions over Beijing and their optical properties, *Tellus B: Chemical and Physical Meteorology*, 61,  
 604 756-767, 10.1111/j.1600-0889.2009.00440.x, 2009.

605 Müller, D., Lee, K. H., Gasteiger, J., Tesche, M., Weinzierl, B., Kandler, K., Müller, T., Toledano, C., Otto,  
 606 S., Althausen, D., and Ansmann, A.: Comparison of optical and microphysical properties of pure Saharan  
 607 mineral dust observed with AERONET Sun photometer, Raman lidar, and in situ instruments during  
 608 SAMUM 2006, *J. Geophys. Res. Atmos.*, 117, n/a-n/a, 10.1029/2011jd016825, 2012.

609 Müller, T., Nowak, A., Wiedensohler, A., Sheridan, P., Laborde, M., Covert, D. S., Marinoni, A., Imre, K.,  
 610 Henzing, B., Roger, J.-C., dos Santos, S. M., Wilhelm, R., Wang, Y.-Q., and de Leeuw, G.: Angular  
 611 Illumination and Truncation of Three Different Integrating Nephelometers: Implications for Empirical,  
 612 Size-Based Corrections, *Aerosol Sci. Technol.*, 43, 581-586, 10.1080/02786820902798484, 2009.

613 Mallet, M., Pont, V., Lioussé, C., Gomes, L., Pelon, J., Osborne, S., Haywood, J., Roger, J. C., Dubuisson,  
 614 P., Mariscal, A., Thouret, V., and Goloub, P.: Aerosol direct radiative forcing over Djougou (northern  
 615 Benin) during the African Monsoon Multidisciplinary Analysis dry season experiment (Special  
 616 Observation Period-0), *J. Geophys. Res.*, 113, 10.1029/2007jd009419, 2008.

617 Massabò, D., Caponi, L., Bernardoni, V., Bove, M. C., Brotto, P., Calzolari, G., Cassola, F., Chiari, M.,  
 618 Fedi, M. E., Fermo, P., Giannoni, M., Lucarelli, F., Nava, S., Piazzalunga, A., Valli, G., Vecchi, R., and  
 619 Prati, P.: Multi-wavelength optical determination of black and brown carbon in atmospheric aerosols,  
 620 *Atmos. Environ.*, 108, 1-12, 10.1016/j.atmosenv.2015.02.058, 2015.

621 Nakayama, T., Sato, K., Matsumi, Y., Imamura, T., Yamazaki, A., and Uchiyama, A.: Wavelength and



622 NO<sub>x</sub> dependent complex refractive index of SOAs generated from the photooxidation of toluene, *Atmos.*  
623 *Chem. Phys.*, 13, 531-545, 10.5194/acp-13-531-2013, 2013.

624 Nazarenko, L., Rind, D., Tsigaridis, K., Del Genio, A. D., Kelley, M., and Tausnev, N.: Interactive nature  
625 of climate change and aerosol forcing, *J. Geophys. Res. Atmos.*, 122, 3457-3480, 10.1002/2016jd025809,  
626 2017.

627 Nguyen, T. B., Lee, P. B., Updyke, K. M., Bones, D. L., Laskin, J., Laskin, A., and Nizkorodov, S. A.:  
628 Formation of nitrogen- and sulfur-containing light-absorbing compounds accelerated by evaporation of  
629 water from secondary organic aerosols, *J. Geophys. Res. Atmos.*, 117, n/a-n/a, 10.1029/2011jd016944,  
630 2012.

631 Petaja, T., Jarvi, L., Kerminen, V. M., Ding, A. J., Sun, J. N., Nie, W., Kujansuu, J., Virkkula, A., Yang,  
632 X. Q., Fu, C. B., Zilitinkevich, S., and Kulmala, M.: Enhanced air pollution via aerosol-boundary layer  
633 feedback in China, *Sci. Repts.*, 6, 18998, 10.1038/srep18998, 2016.

634 Petra Seibert, F. B., Sven-Erik Gryning, Sylvain Joffre, Alix Rasmussen, Philippe Tercier: Review and  
635 intercomparison of operational methods for the determination of the mixing height, *Atmos. Environ.*, 34,  
636 27, 2000.

637 Quinn, P. K.: Aerosol optical properties during INDOEX 1999: Means, variability, and controlling factors,  
638 *J. Geophys. Res.*, 107, 10.1029/2000jd000037, 2002.

639 Ramachandran, S., and Kedia, S.: Black carbon aerosols over an urban region: Radiative forcing and  
640 climate impact, *J. Geophys. Res.*, 115, 10.1029/2009jd013560, 2010.

641 Ramana, M. V., Ramanathan, V., Kim, D., Roberts, G. C., and Corrigan, C. E.: Albedo, atmospheric solar  
642 absorption and heating rate measurements with stacked UAVs, *Quarterly Journal of the Royal*  
643 *Meteorological Society*, 133, 1913-1931, 10.1002/qj.172, 2007.

644 Ramana, M. V., Ramanathan, V., Feng, Y., Yoon, S. C., Kim, S. W., Carmichael, G. R., and Schauer, J. J.:  
645 Warming influenced by the ratio of black carbon to sulphate and the black-carbon source, *Nature*  
646 *Geoscience*, 3, 542-545, 10.1038/ngeo918, 2010.

647 Ramanathan, V., Crutzen, P. J., Kiehl, J. T., and Rosenfeld, D.: Aerosols, climate, and the hydrological  
648 cycle, *Science*, 294, 2119-2124, 10.1126/science.1064034, 2001.

649 Ran, L., Deng, Z., Xu, X., Yan, P., Lin, W., Wang, Y., Tian, P., Wang, P., Pan, W., and Lu, D.: Vertical

650 profiles of black carbon measured by a micro-aethalometer in summer in the North China Plain, *Atmos.*  
651 *Chem. Phys.*, 16, 10441-10454, 10.5194/acp-16-10441-2016, 2016a.

652 Ran, L., Deng, Z. Z., Wang, P. C., and Xia, X. A.: Black carbon and wavelength-dependent aerosol  
653 absorption in the North China Plain based on two-year aethalometer measurements, *Atmos. Environ.*, 142,  
654 132-144, 10.1016/j.atmosenv.2016.07.014, 2016b.

655 Rosati, B., Herrmann, E., Bucci, S., Fierli, F., Cairo, F., Gysel, M., Tillmann, R., Größ, J., Gobbi, G. P.,  
656 Di Liberto, L., Di Donfrancesco, G., Wiedensohler, A., Weingartner, E., Virtanen, A., Mentel, T. F., and  
657 Baltensperger, U.: Studying the vertical aerosol extinction coefficient by comparing in situ airborne data  
658 and elastic backscatter lidar, *Atmos. Chem. Phys.*, 16, 4539-4554, 10.5194/acp-16-4539-2016, 2016.

659 Sührling, M., Maronga, B., Herbolt, F., and Raasch, S.: On the Effect of Surface Heat-Flux Heterogeneities  
660 on the Mixed-Layer-Top Entrainment, *Boundary-Layer Meteorol.*, 151, 531-556, 10.1007/s10546-014-  
661 9913-7, 2014.

662 Sahu, L. K., Kondo, Y., Moteki, N., Takegawa, N., Zhao, Y., Cubison, M. J., Jimenez, J. L., Vay, S., Diskin,  
663 G. S., Wisthaler, A., Mikoviny, T., Huey, L. G., Weinheimer, A. J., and Knapp, D. J.: Emission  
664 characteristics of black carbon in anthropogenic and biomass burning plumes over California during  
665 ARCTAS-CARB 2008, *J. Geophys. Res. Atmos.*, 117, n/a-n/a, 10.1029/2011jd017401, 2012.

666 Saleh, R., Marks, M., Heo, J., Adams, P. J., Donahue, N. M., and Robinson, A. L.: Contribution of brown  
667 carbon and lensing to the direct radiative effect of carbonaceous aerosols from biomass and biofuel  
668 burning emissions, *J. Geophys. Res.*, 120, 10285-10296, 10.1002/, 2015.

669 Sareen, N., Moussa, S. G., and McNeill, V. F.: Photochemical aging of light-absorbing secondary organic  
670 aerosol material, *The journal of physical chemistry. A*, 117, 2987-2996, 10.1021/jp309413j, 2013.

671 Schnaiter, M., Schmid, O., Petzold, A., Fritzsche, L., Klein, K. F., Andreae, M. O., Helas, G., Thielmann,  
672 A., Gimmler, M., Möhler, O., Linke, C., and Schurath, U.: Measurement of Wavelength-Resolved Light  
673 Absorption by Aerosols Utilizing a UV-VIS Extinction Cell, *Aerosol Sci. Technol.*, 39, 249-260,  
674 10.1080/027868290925958, 2005.

675 Shamjad, P. M., Tripathi, S. N., Pathak, R., Hallquist, M., Arola, A., and Bergin, M. H.: Contribution of  
676 Brown Carbon to Direct Radiative Forcing over the Indo-Gangetic Plain, *Environ Sci Technol*, 49, 10474-  
677 10481, 10.1021/acs.est.5b03368, 2015.

678 Su, T., Li, J., Li, C., Xiang, P., Lau, A. K.-H., Guo, J., Yang, D., and Miao, Y.: An intercomparison of  
 679 long-term planetary boundary layer heights retrieved from CALIPSO, ground-based lidar, and radiosonde  
 680 measurements over Hong Kong, *J. Geophys. Res. Atmos.*, 122, 3929-3943, 10.1002/2016jd025937, 2017.  
 681 Sun, J., Zhi, G., Hitzenberger, R., Chen, Y., Tian, C., Zhang, Y., Feng, Y., Cheng, M., Zhang, Y., Cai, J.,  
 682 Chen, F., Qiu, Y., Jiang, Z., Li, J., Zhang, G., and Mo, Y.: Emission factors and light absorption properties  
 683 of brown carbon from household coal combustion in China, *Atmos. Chem. Phys.*, 17, 4769-4780,  
 684 10.5194/acp-17-4769-2017, 2017.  
 685 Tian, P., Liu, D., Huang, M., Liu, Q., Zhao, D., Ran, L., Deng, Z., Wu, Y., Fu, S., Bi, K., Gao, Q., He, H.,  
 686 Xue, H., and Ding, D.: The evolution of an aerosol event observed from aircraft in Beijing: An insight  
 687 into regional pollution transport, *Atmos. Environ.*, 206, 11-20, 10.1016/j.atmosenv.2019.02.005, 2019.  
 688 Torres, B., Dubovik, O., Toledano, C., Berjon, A., Cachorro, V. E., Lapyonok, T., Litvinov, P., and Goloub,  
 689 P.: Sensitivity of aerosol retrieval to geometrical configuration of ground-based sun/sky radiometer  
 690 observations, *Atmos. Chem. Phys.*, 14, 847-875, 10.5194/acp-14-847-2014, 2014.  
 691 Tripathi, S. N., Srivastava, A. K., Dey, S., Satheesh, S. K., and Krishnamoorthy, K.: The vertical profile  
 692 of atmospheric heating rate of black carbon aerosols at Kanpur in northern India, *Atmos. Environ.*, 41,  
 693 6909-6915, 10.1016/j.atmosenv.2007.06.032, 2007.  
 694 Updyke, K. M., Nguyen, T. B., and Nizkorodov, S. A.: Formation of brown carbon via reactions of  
 695 ammonia with secondary organic aerosols from biogenic and anthropogenic precursors, *Atmos. Environ.*,  
 696 63, 22-31, 10.1016/j.atmosenv.2012.09.012, 2012.  
 697 Wang, F., Li, Z., Ren, X., Jiang, Q., He, H., Dickerson, R. R., Dong, X., and Lv, F.: Vertical distributions  
 698 of aerosol optical properties during the spring 2016 ARIAs airborne campaign in the North China Plain,  
 699 *Atmos. Chem. Phys.*, 18, 8995-9010, 10.5194/acp-18-8995-2018, 2018a.  
 700 Wang, J., Nie, W., Cheng, Y., Shen, Y., Chi, X., Wang, J., Huang, X., Xie, Y., Sun, P., Xu, Z., Qi, X., Su,  
 701 H., and Ding, A.: Light absorption of brown carbon in eastern China based on 3-year multi-wavelength  
 702 aerosol optical property observations and an improved absorption Ångström exponent segregation  
 703 method, *Atmos. Chem. Phys.*, 18, 9061-9074, 10.5194/acp-18-9061-2018, 2018b.  
 704 Wang, Y., Che, H., Ma, J., Wang, Q., Shi, G., Chen, H., Goloub, P., and Hao, X.: Aerosol radiative forcing  
 705 under clear, hazy, foggy, and dusty weather conditions over Beijing, China, *Geophys. Res. Lett.*, 36,

10.1029/2009gl037181, 2009.

Wang, Z., Huang, X., and Ding, A.: Dome effect of black carbon and its key influencing factors: a one-dimensional modelling study, *Atmos. Chem. Phys.*, 18, 2821-2834, 10.5194/acp-18-2821-2018, 2018c.

Wilcox, E. M., Thomas, R. M., Praveen, P. S., Pistone, K., Bender, F. A., and Ramanathan, V.: Black carbon solar absorption suppresses turbulence in the atmospheric boundary layer, *Proc. Natl. Acad. Sci. USA.*, 113, 11794-11799, 10.1073/pnas.1525746113, 2016.

Wong, J. P. S., Tsagkaraki, M., Tsiodra, I., Mihalopoulos, N., Violaki, K., Kanakidou, M., Sciare, J., Nenes, A., and Weber, R. J.: Atmospheric evolution of molecular-weight-separated brown carbon from biomass burning, *Atmos. Chem. Phys.*, 19, 7319-7334, 10.5194/acp-19-7319-2019, 2019.

Wu, Y., Wang, X., Tao, J., Huang, R., Tian, P., Cao, J., Zhang, L., Ho, K.-F., Han, Z., and Zhang, R.: Size distribution and source of black carbon aerosol in urban Beijing during winter haze episodes, *Atmos. Chem. Phys.*, 17, 7965-7975, 10.5194/acp-17-7965-2017, 2017.

Xia, X., Eck, T. F., Holben, B. N., Phillippe, G., and Chen, H.: Analysis of the weekly cycle of aerosol optical depth using AERONET and MODIS data, *J. Geophys. Res.*, 113, 10.1029/2007jd009604, 2008.

Xie, C., Xu, W., Wang, J., Wang, Q., Liu, D., Tang, G., Chen, P., Du, W., Zhao, J., Zhang, Y., Zhou, W., Han, T., Bian, Q., Li, J., Fu, P., Wang, Z., Ge, X., Allan, J., Coe, H., and Sun, Y.: Vertical characterization of aerosol optical properties and brown carbon in winter in urban Beijing, China, *Atmos. Chem. Phys.*, 19, 165-179, 10.5194/acp-19-165-2019, 2019.

Yu, H., Liu, S. C., and Dickinson, R. E.: Radiative effects of aerosols on the evolution of the atmospheric boundary layer, *J. Geophys. Res.*, 107, 10.1029/2001jd000754, 2002.

Yu, P. F., Toon, O. B., Bardeen, C. G., Zhu, Y. Q., Rosenlof, K. H., Portmann, R. W., Thornberry, T. D., Gao, R. S., SDavis, S. M., Wolf, E. T., Gouw, J. D., Peterson, D. A., Fromm, M. D., and Robock, A.: Black carbon lofts wildfire smoke high into the stratosphere to form a persistent plume, *Science*, 365, 14, 2019.

Zhang, R., Jing, J., Tao, J., Hsu, S. C., Wang, G., Cao, J., Lee, C. S. L., Zhu, L., Chen, Z., Zhao, Y., and Shen, Z.: Chemical characterization and source apportionment of PM<sub>2.5</sub> in Beijing: seasonal perspective, *Atmos. Chem. Phys.*, 13, 7053-7074, 10.5194/acp-13-7053-2013, 2013.

Zhang, X. Y., Wang, J. Z., Wang, Y. Q., Liu, H. L., Sun, J. Y., and Zhang, Y. M.: Changes in chemical

734 components of aerosol particles in different haze regions in China from 2006 to 2013 and contribution of  
735 meteorological factors, *Atmos. Chem. Phys.*, 15, 12935-12952, 10.5194/acp-15-12935-2015, 2015.

736 Zhang, Y., Forrister, H., Liu, J., Dibb, J., Anderson, B., Schwarz, J. P., Perring, A. E., Jimenez, J. L.,  
737 Campuzano-Jost, P., Wang, Y., Nenes, A., and Weber, R. J.: Top-of-atmosphere radiative forcing affected  
738 by brown carbon in the upper troposphere, *Nature Geoscience*, 10.1038/ngeo2960, 2017.

739 Zhao, D. L., Huang, M. Y., Tian, P., He, H., Lowe, D., Zhou, W., Sheng, J. J., Wang, F., Bi, K., Kong, S.  
740 F., Yang, Y., Liu, Q., Liu, D. T., and Ding, D. P.: Vertical characteristics of black carbon physical properties  
741 over Beijing region in warm and cold seasons, *Atmos. Environ.*, 213, 296-310, 2019a.

742 Zhao, G., Zhao, C. S., Kung, Y., Bian, Y. X., Tao, J. C., Shen, C. Y., and Yu, Y. L.: Calculating the aerosol  
743 asymmetry factor based on measurements from the humidified nephelometer system, *Atmos. Chem. Phys.*,  
744 18, 9049-9060, 10.5194/acp-18-9049-2018, 2018.

745 Zhao, P. S., Ding, J., Du, X., and Su, J.: High time-resolution measurement of light scattering hygroscopic  
746 growth factor in Beijing A novel method for high relative humidity conditions, *Atmos. Environ.*, 215,  
747 10.1016/j.atmosenv.2019.116912, 2019b.

748 Zhao, R., Lee, A. K. Y., Huang, L., Li, X., Yang, F., and Abbatt, J. P. D.: Photochemical processing of  
749 aqueous atmospheric brown carbon, *Atmos. Chem. Phys.*, 15, 6087-6100, 10.5194/acp-15-6087-2015,  
750 2015.

751 Zhong, J., Zhang, X., Dong, Y., Wang, Y., Liu, C., Wang, J., Zhang, Y., and Che, H.: Feedback effects of  
752 boundary-layer meteorological factors on cumulative explosive growth of  $PM_{2.5}$   
753 during winter heavy pollution episodes in Beijing from 2013 to 2016, *Atmos. Chem. Phys.*, 18, 247-258,  
754 10.5194/acp-18-247-2018, 2018.

755  
756

757 Table 1. Flight summary in this study.

<b>Flight number</b>	<b>Time range Local time</b>	<b>Case</b>	<b>Pollution period</b>	<b>Mixing layer height</b>
RF1	20161115 12:00	Case_1	LP	1450 m
RF2	20161115 14:00	Case_1	LP	1450 m
RF3	20161116 12:00	Case_1	TP	850 m
RF4	20161116 14:00	Case_1	TP	750 m
RF5	20161117 12:00	Case_1	TP(Cloud)	1250 m
RF6	20161117 14:00	Case_1	TP(Cloud)	1150 m
RF7	20161118 12:00	Case_1	HP(Cloud)	1050 m
RF8	20161210 14:00	Case_2	LP	950 m
RF9	20161211 14:00	Case_2	MP	950 m
RF10	20161212 14:00	Case_2	HP	450 m
RF11	20161216 14:00	Case_3	TP	350 m
RF12	20161217 14:00	Case_3	HP	350 m
RF13	20161218 14:00	Case_3	HP	350 m
RF14	20161219 14:00	Case_3	HP	250 m

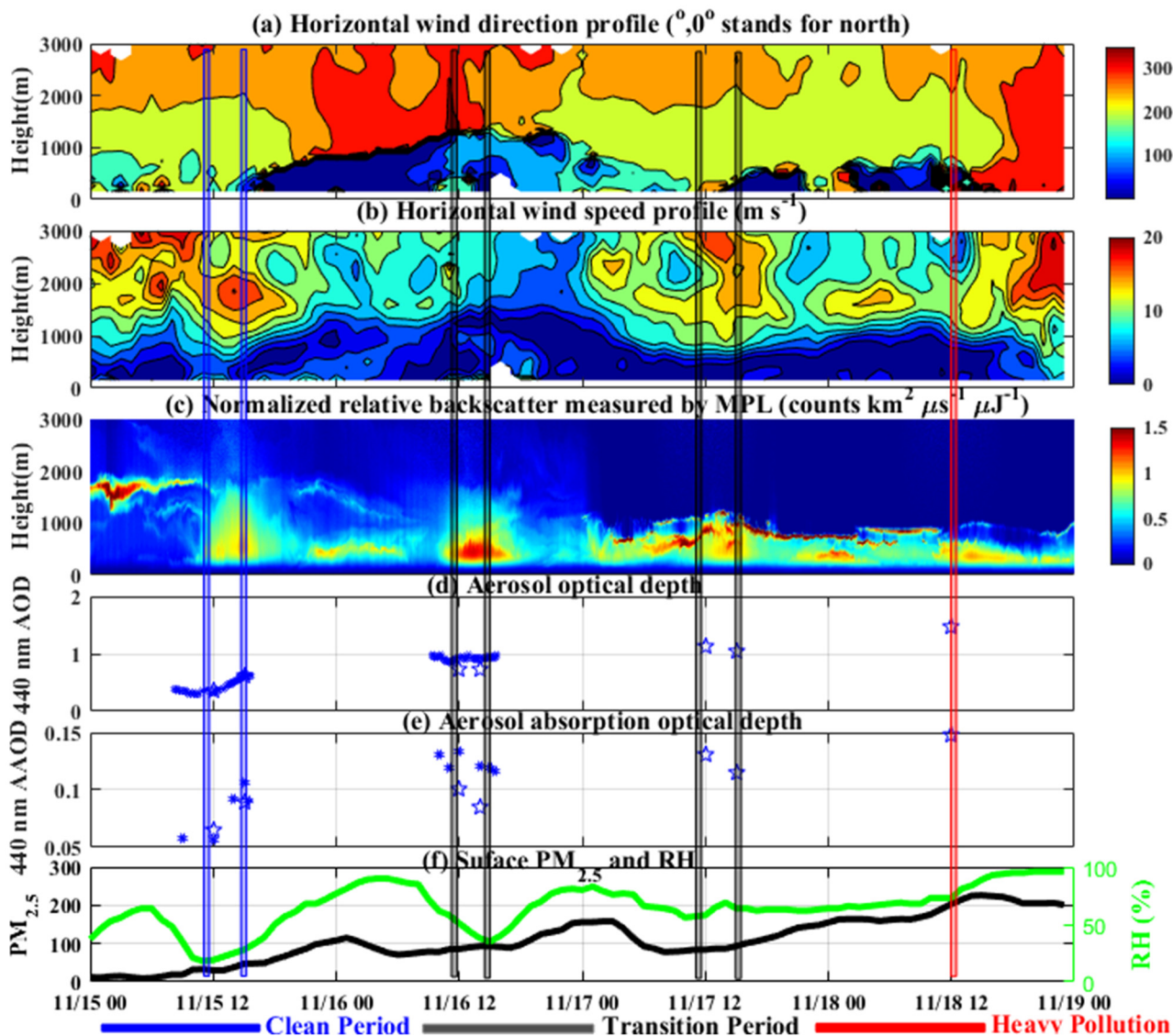
758 <sup>1</sup> LP, TP, and HP represents the low, transition and high pollution period during a pollution event.

759

760 Table 2. Summary of input parameters for the radiative transfer calculation using Discrete Ordinates  
761 Radiative Transfer Code (DISORT)  
762

Parameter	Input value
Radiative transfer solver	DISORT, 12-streams, delta-m method
Gas absorption parameterization	LOWTRAN/SBDART parameterization
Wavelength range	250-2550nm
Atmosphere	Standard Mid-latitude atmosphere
Aerosol	The 25 layers from the surface to 5000 m was chosen inside the DISORT, and every 200 m average aerosol optical properties at 550 nm was used; AOD values are derived from in-situ Aurora3000 and AE33 measurements, also applying an exponential $\lambda$ -dependent function SSA values are from in-situ $\sigma_{\text{sca}}$ and $\sigma_{\text{abs}}$ measurement SAE values are from in-situ Aurora3000 measurement AAE values are from in-situ AE33 measurement Asymmetry factor ( $g$ ) is derived from the Aurora3000 measurement and uses Henyey-Greenstein phase function
Location	39.54°N, 116.23°E
Time	Flight time
Solar zenith angle	Effective solar zenith angle Using local time and aircraft location
Surface albedo	IGBP surface type 13 (Urban)

763  
764



766

767

768

769

770

771

772

773

Fig. 1. Temporal variations from Nov. 15<sup>th</sup> to 18<sup>th</sup> of vertical profiles of wind direction (a), and wind speed (b) measured by wind profile radar; (c) particle extinction measured by MPL lidar; (d) aerosol optical depth (d) and aerosol absorption optical depth (e) from AERONET (asterisk) and derived from aircraft in-situ measurements (open star) (f) surface PM<sub>2.5</sub> and RH. The vertical bars denote the periods of flight profiles, with blue, black and red representing the clean period, transition period and heavy pollution during a pollution event respectively. The other two pollution events can refer to supplement



775

776

777

778

779

780

781

782

783

784

785

786

787

788

789

790

791

792

793

794

795

796

797

798

799

800

801

802

803

804

805

806

807

808

809

810

811

812

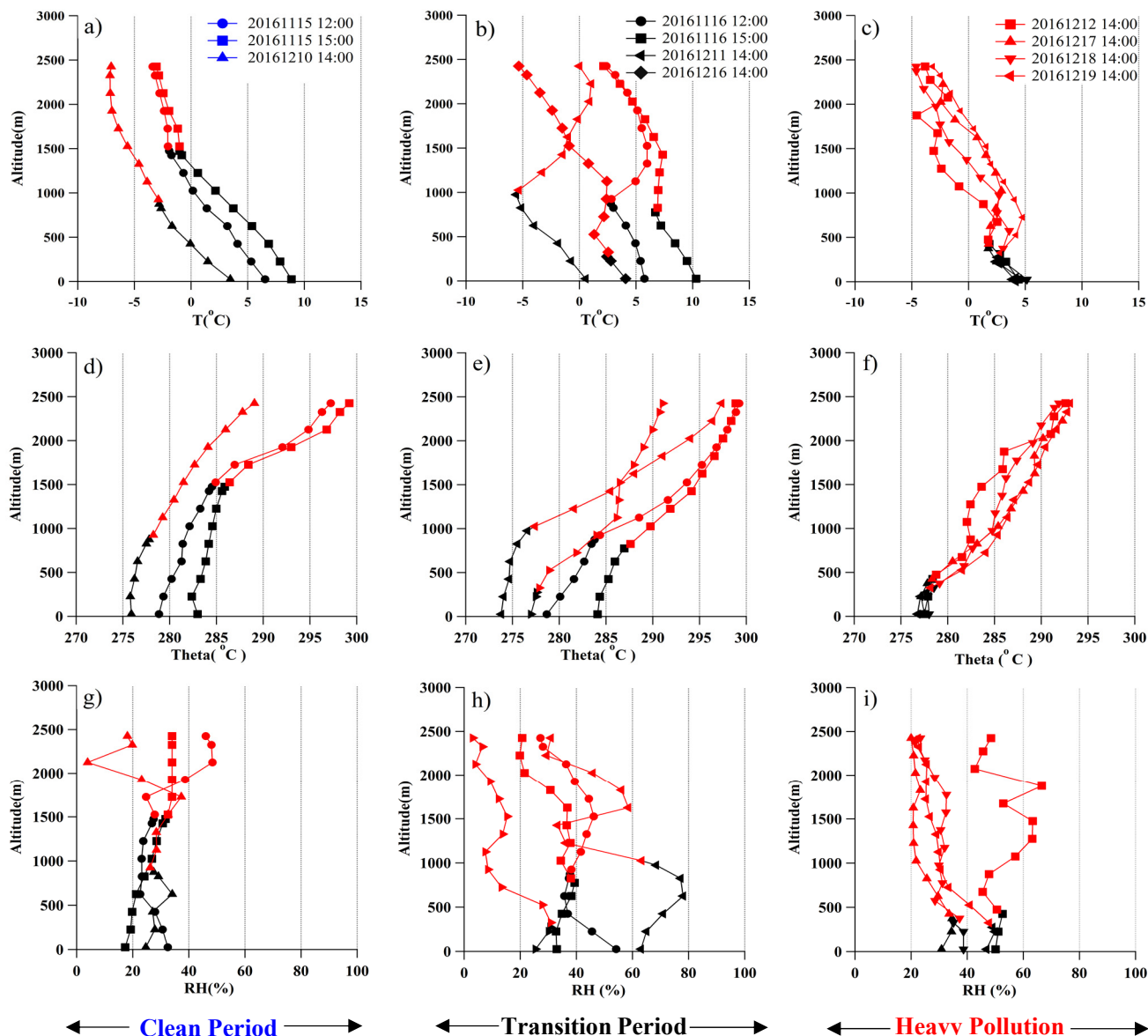
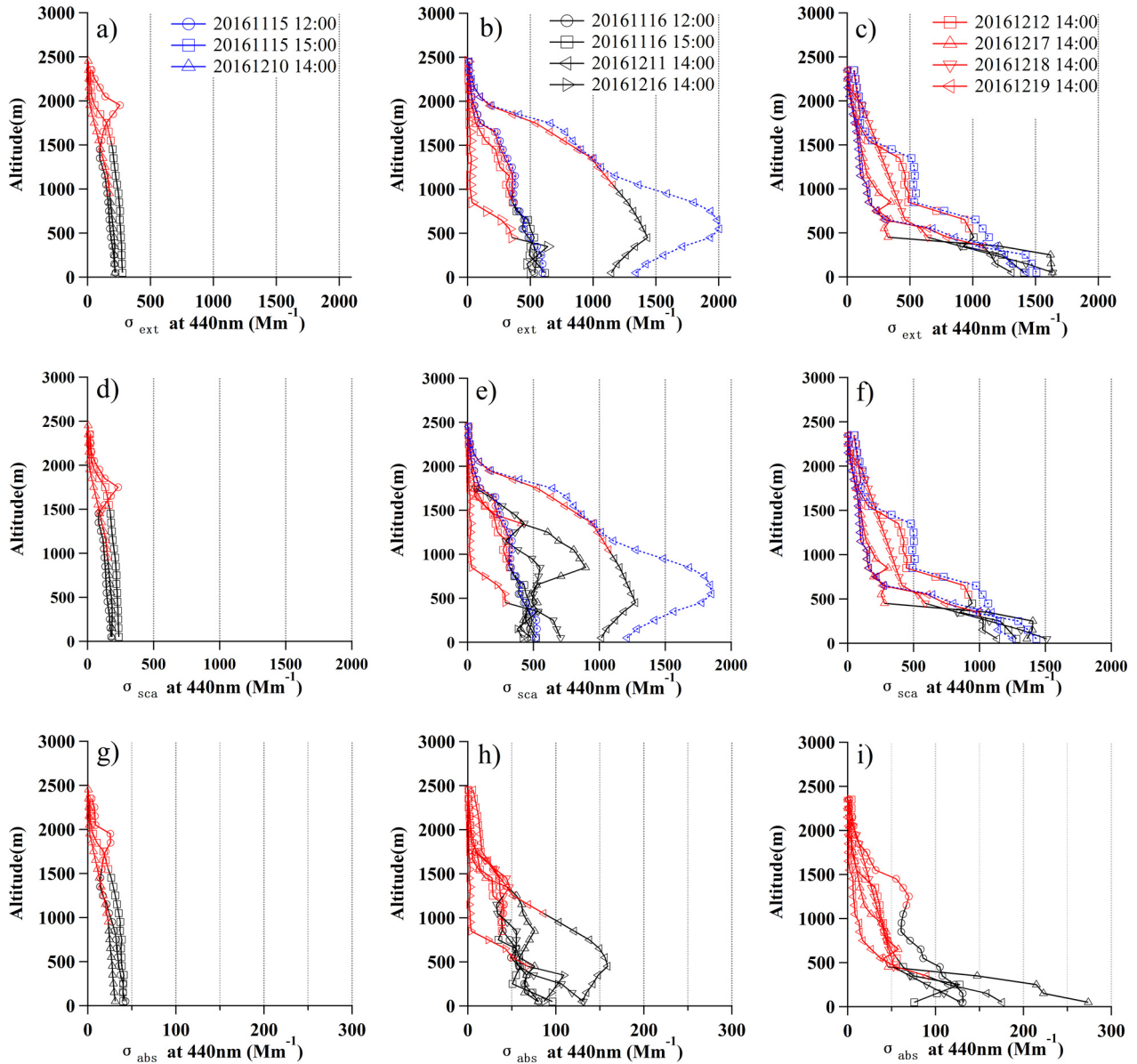


Fig. 2. Vertical profiles of temperature (a, b, c), relative humidity (d, e, f) and potential temperature (g, h, i) for Clean Period, Transition Period and Heavy Pollution period, respectively. The black and red dots represent for inside the PBL and above the PBL.



← Clean Period →      ← Transition Period →      ← Pollution Period →

Fig. 3. Vertical profiles of aerosol extinction, scattering and absorption coefficient at 440 nm for CP (blue), TP (black) and HP period (red), respectively. The black and red lines represent for inside and above the ML, respectively. The hygroscopic-corrected profiles was shown in blue lines.

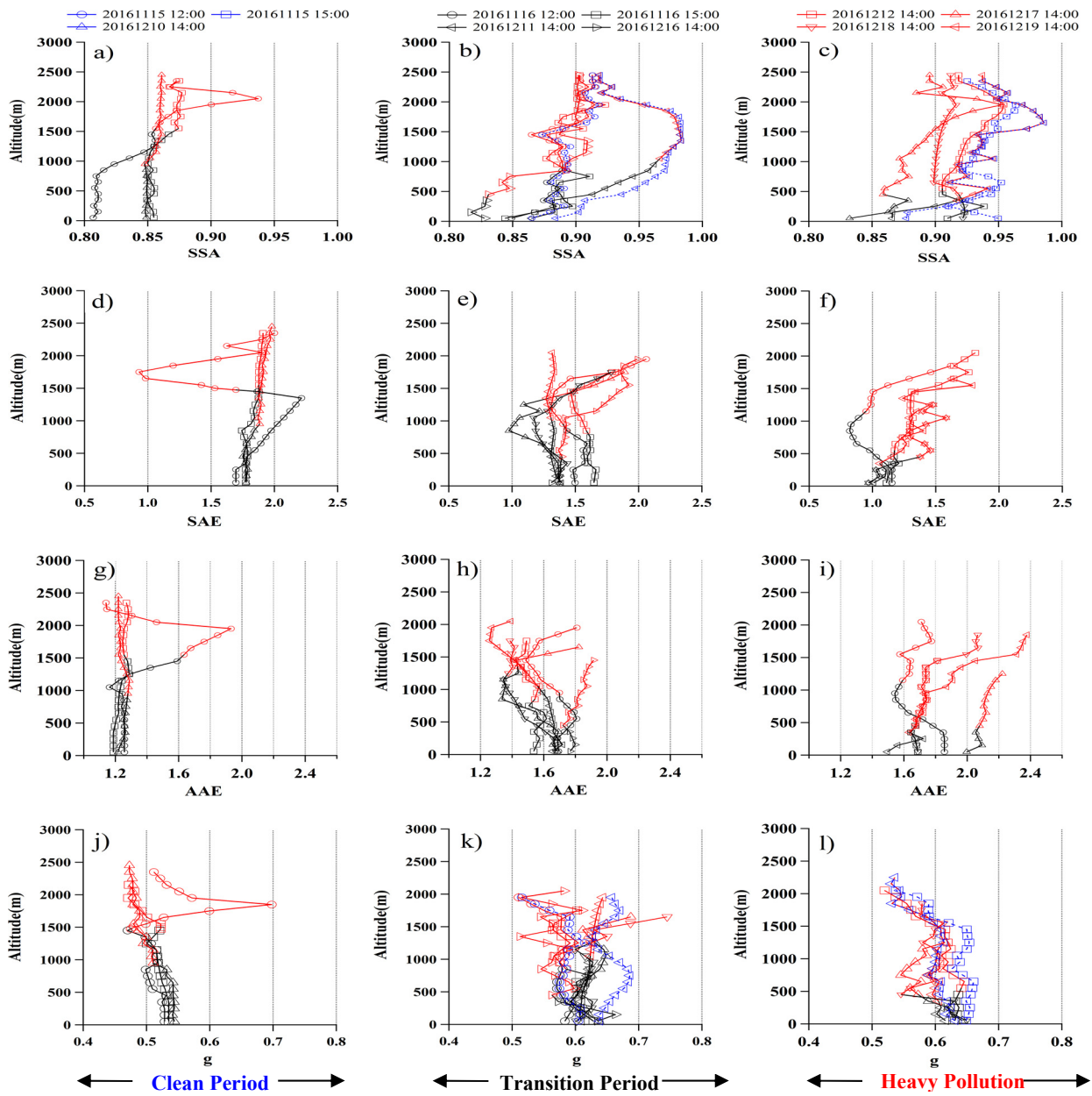


Fig. 4. Vertical profiles of aerosol single scattering albedo at 440 nm (SSA, a - c), scattering Angström exponent (SAE, d - f), absorption Angström exponent (AAE, g - i), and asymmetry parameter (g, j - l) for CP (left panel), TP (middle panel) and HP period (right panel), respectively. The hygroscopicity

corrected profiles was shown in blue lines.

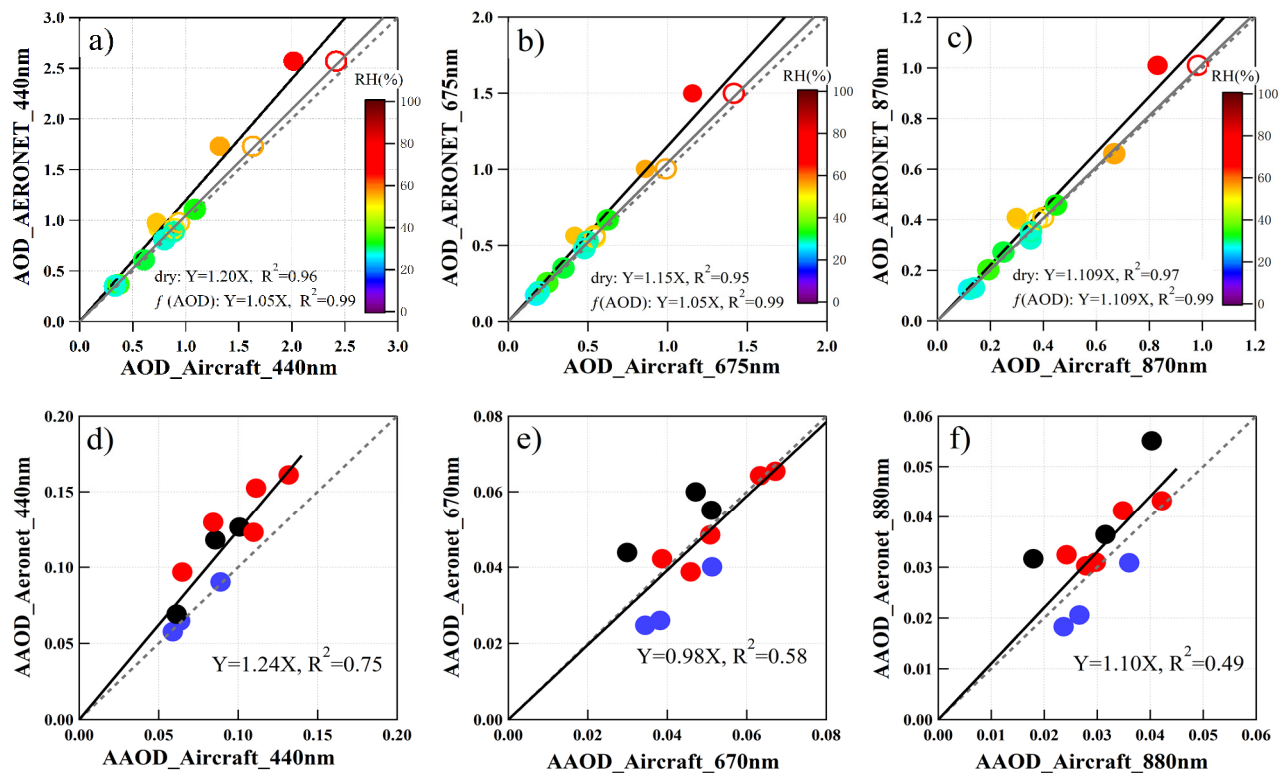
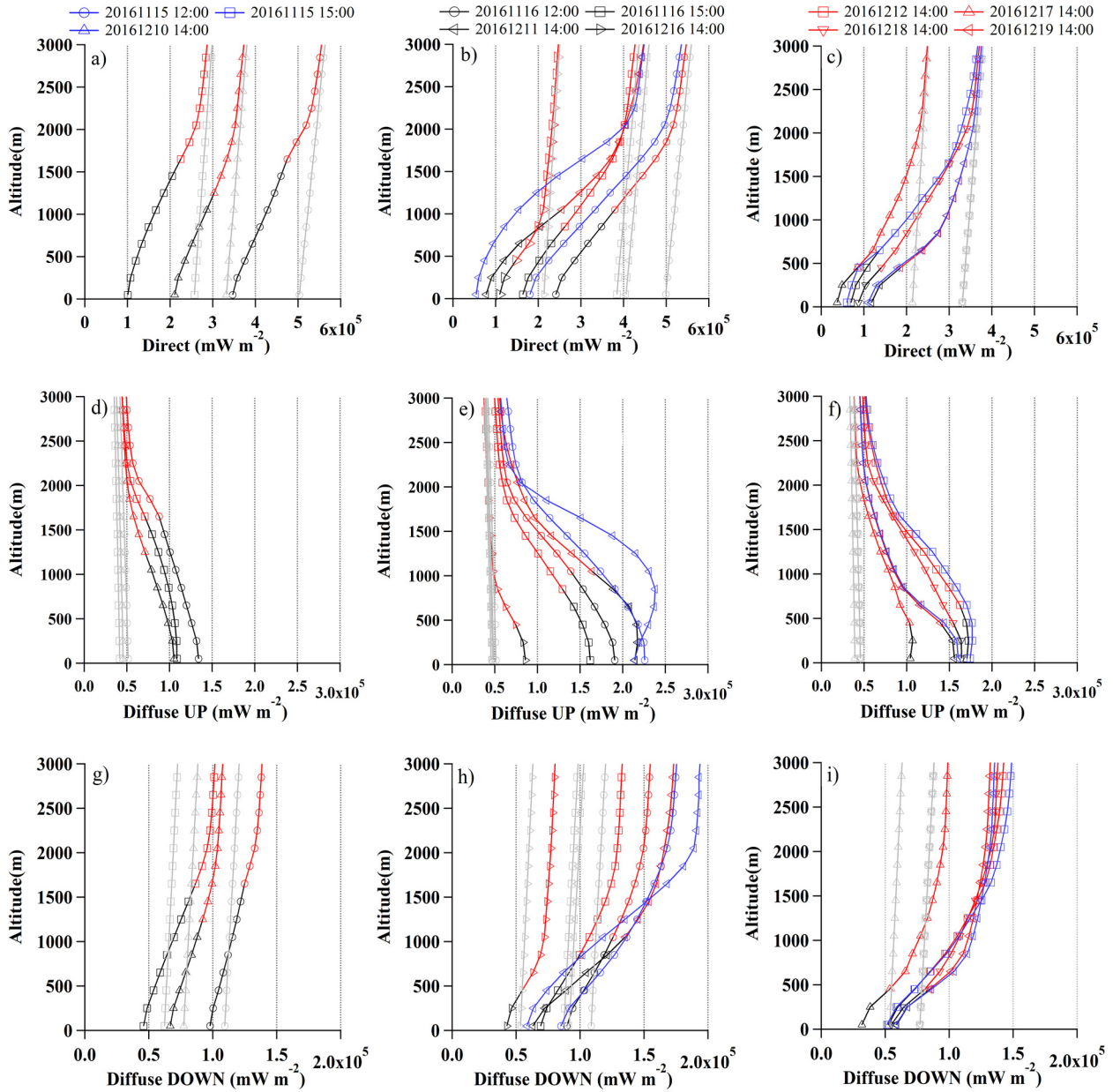


Fig. 5. Comparison between AERONET and aircraft in-situ constrained AOD and AAOD: a) - c) The comparison of AOD at 440nm, 675nm, and 870nm colored by RH; the solid and open markers denote the dry and hygroscopicity-corrected condition; d) – f) Comparison of AAOD at 440nm, 670nm, and 880nm. The blue, black, and red dots represent for CP, TP, and HP period, respectively.

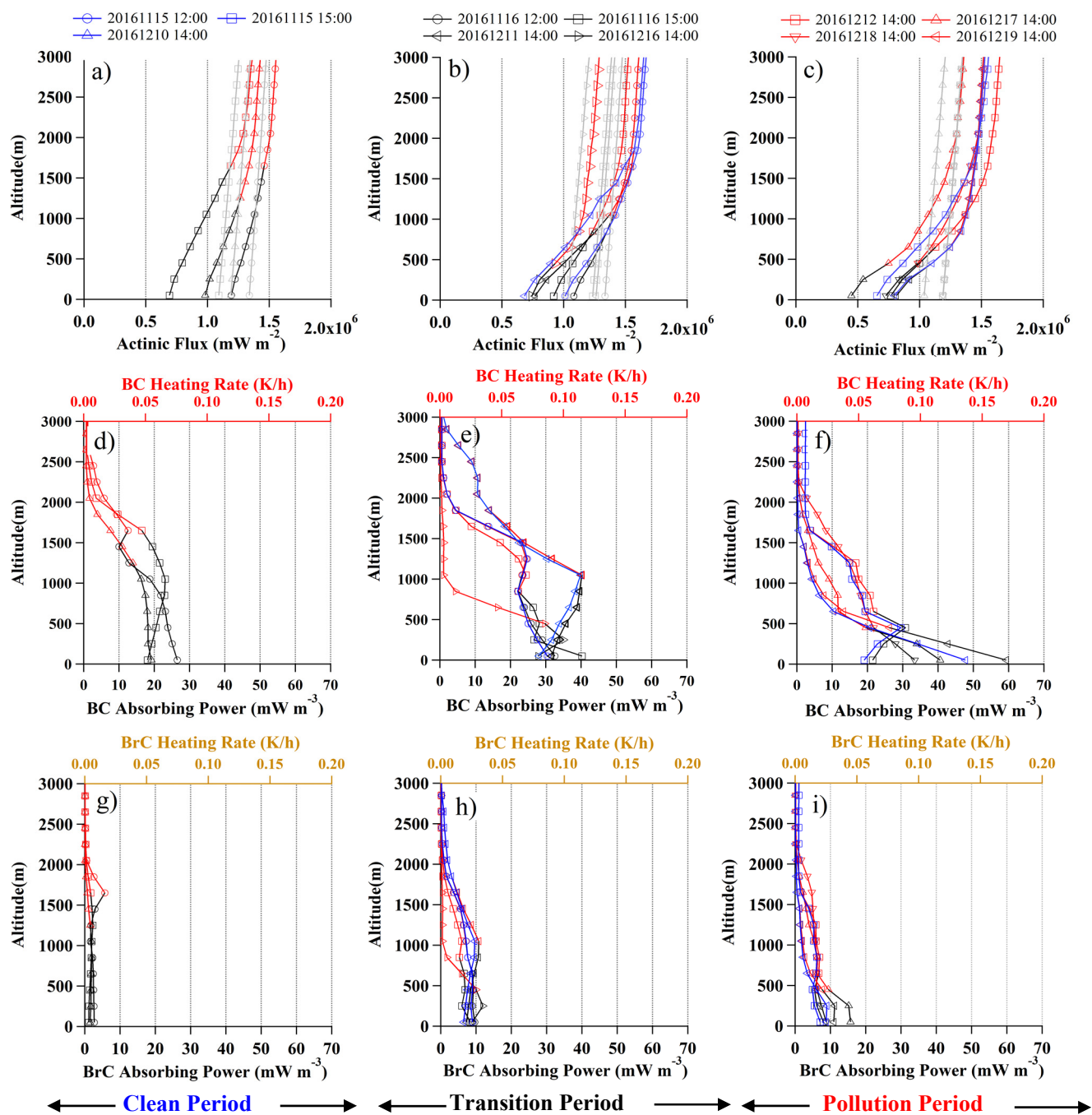




← Clean Period →
← Transition Period →
← Pollution Period →

Fig. 6. Radiative transfer results calculated by DISTORT. a)-c), b)-e), and g)-i) show the direct, diffuse upward and diffuse downward irradiance respectively. The left, middle and right panel represent for CP, TP and HP period respectively, with black and red lines denoting above and within the PBL. The

949 colored and grey lines denote the profiles for with and without aerosol influence, respectively, and the  
 950 blue lines are for corrected hygroscopic effect.



989 Fig. 7. Actinic flux (a-c), BC absorbing power (d-f) and BrC absorbing power (g-i). The left, middle and

right panel was for LP, TP and HP respectively, with the black and red line denoting within and above the PBL. The gray lines in a) to c) show the aerosol free results and the blue line denote the corrected hygroscopic effect. The upper x-axis from d) to i) shows the heating rate.

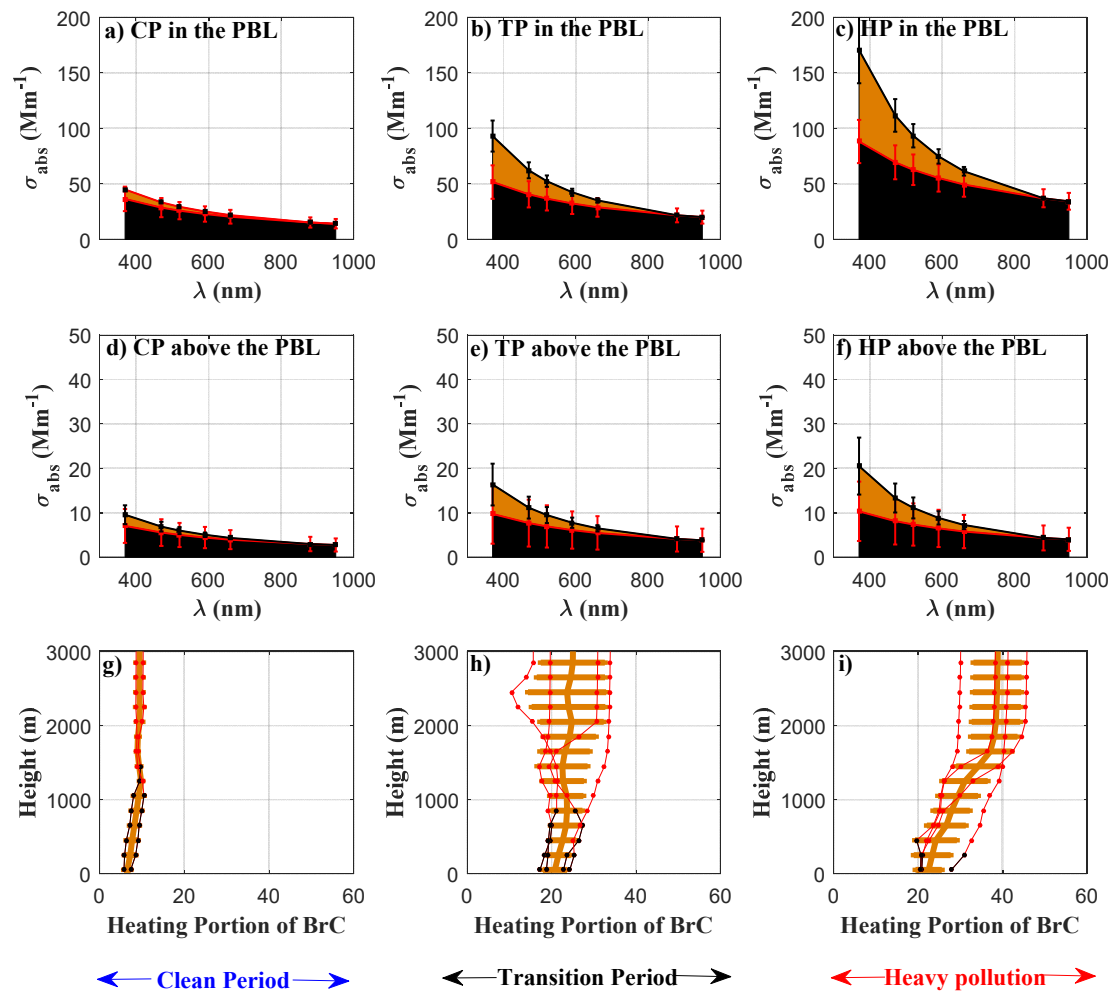


Fig. 8. Spectral absorption coefficient of BC and BrC inside and above the PBL for CP (a, d), TP (b, e) and HP period (c, f), respectively, shown in black and brown carbon color respectively. The vertical profiles of heating portion of BrC for CP, TP and HP period are shown in g) – i).