

1 Trends and source apportionment of aerosols in Europe during  
2 1980–2018

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## 22 **Abstract**

23       Aerosols have significantly affected health, environment and climate in Europe.  
24       Aerosol concentrations have been declining since 1980s in Europe, mainly owing to  
25       the reduction of local aerosol and precursor emissions. Emissions from other source  
26       regions of the world, which have been changing rapidly as well, may also perturb the  
27       historical and future trends of aerosols and change their radiative impact in Europe.  
28       This study examines trends of aerosols in Europe during 1980–2018 and quantify  
29       contributions from sixteen source regions using the Community Atmosphere Model  
30       version 5 with an Explicit Aerosol Source Tagging technique (CAM5-EAST). The  
31       simulated near-surface total mass concentration of sulfate, black carbon and primary  
32       organic carbon had a 62% decrease during 1980–2018, of which the majority was  
33       contributed by reductions of local emissions in Europe and 8%-9% was induced by  
34       the decrease in emissions from Russia-Belarus-Ukraine. With the decreases in the  
35       fractional contribution of local emissions, aerosols transported from other source  
36       regions are increasingly important to air quality in Europe. During 1980–2018, the  
37       decrease in sulfate loading leads to a warming effect of  $2.0 \text{ W m}^{-2}$  in Europe, with  
38       12% coming from changes in non-European sources, especially from North America  
39       and Russia-Belarus-Ukraine. According to the Shared Socioeconomic Pathways  
40       (SSP) scenarios, contributions to the sulfate radiative forcing over Europe from both  
41       European local emissions and non-European emissions would decrease at a  
42       comparable rate in the next three decades, suggesting that future changes in non-

43 European emissions are as important as European emissions in causing possible  
44 regional climate change associated with aerosols in Europe.

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## 48 **1. Introduction**

49 Aerosols are main air pollutants that contribute to excess morbidity and  
50 premature mortality through damaging cardiovascular and respiratory systems  
51 (Lelieveld et al., 2019). They also have adverse effects on atmospheric visibility for  
52 road and air traffic (Vautard et al., 2009). During the 1952 London Fog, high fatality  
53 associated with extremely high level of aerosols caused thousands of premature  
54 deaths (Bell and Davis, 2001), which resulted in a number of air quality legislations  
55 for reducing air pollution in Europe (Brimblecombe et al., 2006).

56 Besides the health and environment effects, aerosols can significantly impact  
57 regional and global climate through perturbing the Earth's radiation fluxes and  
58 influencing cloud microphysics (Boucher et al., 2013). Globally, anthropogenic  
59 aerosols exert a net cooling effect in the Earth system, which have dampened the  
60 warming driven by greenhouse gases since the pre-industrial era. Due to a strong  
61 surface albedo feedback over polar regions, per unit aerosol emission from western  
62 Europe was reported to have the greatest cooling effect than other major source  
63 regions of the globe (Persad and Caldeira, 2018), revealing the importance of  
64 understanding aerosol variations in Europe.

65 Significant reductions in near-surface aerosol concentrations and aerosol optical  
66 depth (AOD) have been observed in Europe during the last few decades from long-  
67 term station measurements and satellite retrievals (de Meij et al., 2012; Tørseth et  
68 al., 2012). The decrease in aerosols has been considered as a cause of the increase  
69 in surface solar radiation over Europe since the 1980s (Wild, 2009), as well as the

70 contributor of the eastern European warming (Vautard et al., 2009), Arctic  
71 amplification (Acosta Navarro et al., 2016), and the increased atmospheric visibility  
72 over Europe (Stjern et al., 2011) during the past three decades.

73       The decrease in aerosols over Europe was mainly attributed to the continuous  
74 reductions in European local anthropogenic emissions of aerosols and precursor  
75 gases since the 1980s (Smith et al., 2011), as a result of legislations for improving  
76 air quality. In addition to local emissions, aerosol levels can also be affected by  
77 aerosol transport at continental scales (Zhang et al., 2017; Yang et al., 2018a).  
78 Aerosol emissions in major economic regions of the world have been changing  
79 rapidly during the past few decades owing to economic development and  
80 environmental measures. North America has started reducing emissions since the  
81 1980s, and emissions in Russia also showed decreasing trends after the dissolution  
82 of the Soviet Union (Smith et al., 2011). In the meantime, aerosol emissions from  
83 East Asia and South Asia have largely increased due to economic growth, although  
84 emissions in China have been undergoing a remarkable reduction in the most recent  
85 years, as a result of strict air quality regulations (Streets et al., 2000; Li et al., 2017).  
86 It is important to understand the relative roles of local emissions and regional  
87 transport in affecting long-term variation of aerosols in Europe from both air quality  
88 and climate perspectives.

89       Source apportionment is useful for quantifying contributions to aerosols from  
90 specific source regions and/or sectors, which is beneficial to the emission control  
91 strategies. The traditional method of examining the source-receptor relationship in

92 aerosol models is to zero out or perturb a certain percent of emissions from a given  
93 source region or sector in parallel sensitivity simulations along with a baseline  
94 simulation, which has been used in many studies to examine source contributions of  
95 particulate matter (PM) in Europe from different sectors (e.g., Sartelet et al., 2012;  
96 Tagaris et al., 2015; Aksoyoglu et al., 2016). Recently, source region contributions to  
97 European CO and O<sub>3</sub> levels, as well as global and regional aerosol radiative forcing,  
98 were examined under the Hemispheric Transport of Air Pollution model experiment  
99 phase 2 (HTAP2) protocol, in which sensitivity simulations were conducted with  
100 anthropogenic emissions from different source regions reduced by 20% (Jonson et  
101 al., 2018). This method suffers a large computational cost for the excessive model  
102 simulations when estimating contributions from a large number of sources, and  
103 contributions from all sources do not sum up to 100% of the total concentration in the  
104 default simulation (Koo et al., 2009; Wang et al., 2014).

105       The explicit aerosol tagging method, which simultaneously tracks contributions  
106 from many different sources, is a useful tool for assessing source-receptor  
107 relationship of aerosols. This method has previous been adopted in regional air  
108 quality models such as CAMx (the Comprehensive Air quality Model with  
109 Extensions) and CMAQ (the Community Multi-scale Air Quality model). Using  
110 regional air quality models with aerosol tagging, contributions from different source  
111 sectors and local/regional sources to European PM and its health impact were  
112 studied (Brandt et al., 2013; Skyllakou et al., 2014; Karamchandani et al., 2017).  
113 However, due to the limitation in domain size of regional air quality models,

114 contributions of intercontinental transport from sources outside the domain are  
115 difficult to be accounted.

116 Anthropogenic emissions of aerosols and their precursor gases from different  
117 economic regions of the world have changed substantially during the past few  
118 decades. Very few studies have examined the source apportionment of aerosols in  
119 Europe from sources all over the changing world. In this study, source attributions of  
120 concentrations, column burden, optical depth of aerosols in four major areas of  
121 Europe from sixteen source regions of the globe over 1980–2018 are quantified,  
122 which is facilitated by the explicit aerosol source tagging technique that were recently  
123 implemented in a global aerosol-climate model (CAM5-EAST). This technique has  
124 lately been used to examine source attribution of aerosol trends in China and U.S.  
125 during 1980–2014 (Yang et al., 2018a,b). The source apportionment analysis is  
126 extended to year 2018 using the Shared Socioeconomic Pathways (SSPs) scenario,  
127 with a focus on Europe here.

128 The CAM5-EAST model, along with the aerosol source tagging technique, and  
129 aerosol emissions are described in Sect. 2. Section 3 evaluates the model  
130 performance in simulating aerosols in Europe. Section 4 show the analysis of  
131 source-receptor relationships of aerosols in Europe in climatological mean. Source  
132 contributions to long-term variations of European aerosols and their direct radiative  
133 forcing (DRF) during 1980–2018, as well as future forcing prediction, are provided in  
134 Sect. 5. Section 6 summarizes these results and conclusions.

## 135 **2. Methods**

## 136 **2.1 Model Description and Experimental Setup**

137       The global aerosol-climate model CAM5 (Community Atmosphere Model version  
138 5), which was developed as the atmospheric component of CESM (the Community  
139 Earth System Model, Hurrell et al., 2013), is applied to simulate aerosols at a spatial  
140 resolution of 1.9° latitude × 2.5° longitude and 30 vertical layers from the surface to  
141 3.6 hPa. Aerosol species, including sulfate, black carbon (BC), primary organic  
142 aerosol (POA), second organic aerosol (SOA), mineral dust and sea salt, can be  
143 simulated in a modal aerosol module of CAM5. The three-mode aerosol module  
144 (MAM3) configuration is chosen with the consideration of the computational  
145 efficiency of long-term simulation. Details of the MAM3 aerosol representation in  
146 CAM5 are described in Liu et al. (2012). On top of the default CAM5, some aerosol-  
147 related scheme modifications are utilized to improve the model performance in the  
148 aerosol convective transport and wet deposition (Wang et al., 2013).

149       A 40-year (1979–2018) historical AMIP-type (Atmospheric Model  
150 Intercomparison Project) simulation has been performed, following CMIP6 (the  
151 Coupled Model Intercomparison Project Phase 6) configurations and forcing  
152 conditions. Time-varying sea surface temperatures, sea ice concentrations, solar  
153 insolation, greenhouse gas concentrations and aerosol emissions are prescribed in  
154 the simulation. To better reproduce large-scale circulation patterns for aerosol  
155 transport in the model, wind fields are nudged to the MERRA-2 (Modern Era  
156 Retrospective-Analysis for Research and Applications Version 2) reanalysis (Ronald  
157 Gelaro et al., 2017).



158 Aerosol DRF is defined in this study as the difference in clear-sky radiative fluxes  
159 at the top of the atmosphere between two diagnostic calculations in the radiative  
160 transfer scheme with and without specific aerosol species accounted, respectively.  
161 Historical variation of sulfate DRF due to anthropogenic emissions from Europe and  
162 outside Europe are quantified in this study. Rather than sulfate, DRF of other aerosol  
163 species is not calculated in this study due to the computational limitation considering  
164 multiple source regions. However, because sulfate dominates the decrease in total  
165 combustion AOD in Europe shown below, the sulfate DRF is calculated to roughly  
166 represent the DRF caused by the total combustion AOD change. Future DRF of  
167 sulfate aerosol over Europe is also estimated through scaling historical mean (1980–  
168 2018) sulfate DRF by the ratio of SSPs future SO<sub>2</sub> emissions to historical emissions  
169 assuming a linear response of DRF to AOD and regional emissions. This DRF  
170 prediction method has been used to estimate the East Asian contribution to sulfate  
171 DRF in U.S. in 2030s (Yang et al., 2018a).

## 172 **2.2 Aerosol Source Tagging Technique**

173 The Explicit Aerosol Source Tagging (EAST) technique, which was recently  
174 implemented in CAM5 (Wang et al., 2014; Yang et al., 2017a, b), is used to examine  
175 the long-term source apportionment of aerosols in Europe. Unlike the traditional  
176 back-trajectory and emission perturbation methods, EAST has the identical physical,  
177 chemical and dynamical processes considered independently for aerosol species  
178 (defined as new tracers) emitted from each of the tagged source region and/or sector  
179 in the simulation. Sulfate, BC, POA and SOA from pre-defined sources can be

180 explicitly tracked, from emission to deposition, in one CAM5-EAST simulation. Due  
181 to the computational constraint and potentially large model bias from the simplified  
182 SOA treatment (Yang et al., 2018a; Lou et al., 2019), we focus on sulfate, BC and  
183 POA in this study but quantify the potential impact of SOA on the aerosol variation.

184 The global aerosol and precursor emissions are decomposed into sixteen source  
185 regions defined in the HTAP2 protocol, including Europe (EUR), North America  
186 (NAM), Central America (CAM), South America (SAM), North Africa (NAF), South  
187 Africa (SAF), the Middle East (MDE), Southeast Asia (SEA), Central Asia (CAS),  
188 South Asia (SAS), East Asia (EAS), Russia-Belarus-Ukraine (RBU), Pacific-  
189 Australia-New Zealand (PAN), the Arctic (ARC), Antarctic (ANT), and Non-  
190 Arctic/Antarctic Ocean (OCN) (Figure 1). Note that sources from marine and volcanic  
191 eruptions are included in OCN. The focused receptor region in this study is Europe,  
192 which is further divided into Northwestern Europe (NWE or NW Europe),  
193 Southwestern Europe (SWE or SW Europe), Eastern Europe (EAE or E. Europe)  
194 and Greece-Turkey-Cyprus (GTC) based on the finer source region selection in  
195 HTAP2.

### 196 **2.3 Aerosol and Precursor Emissions**

197 Following the CMIP6-AMIP protocol, historical anthropogenic (Hoesly et al.,  
198 2018) and biomass burning (van Marle et al., 2017) emissions of aerosol and  
199 precursor gases are used over 1979–2014. For the remaining four years (2015–  
200 2018), emissions are interpolated from the SSP2-4.5 forcing scenario, in which  
201 aerosol pathways are not as extreme as other SSPs and have been used in many

202 model intercomparison projects for CMIP6 (O'Neill et al., 2016). Figure S1 shows the  
203 spatial distribution and time series of anthropogenic emissions of SO<sub>2</sub> (precursor gas  
204 of sulfate aerosol), BC and POA from Europe over 1980–2018. High emissions are  
205 located over E. Europe and NW Europe, from which the emissions of SO<sub>2</sub>, BC and  
206 POA were reduced by 84–93%, 43–62% and 28–36%, respectively, in 2014–2018  
207 relative to 1980–1984. Although SW Europe had a relatively low total amount of  
208 emissions compared to E. Europe and NW Europe, it had significant reductions in  
209 SO<sub>2</sub> and BC emissions, 91% and 55%, respectively. Over GTC region, SO<sub>2</sub> and BC  
210 emissions were increased by 49% and 48%, respectively. Considering the sub-  
211 regions as a whole, SO<sub>2</sub>, BC and POA emissions from Europe have decreased by  
212 12.57 Tg yr<sup>-1</sup> (83%), 0.22 Tg yr<sup>-1</sup> (46%) and 0.30 Tg yr<sup>-1</sup> (24%) in 2014–2018  
213 compared to 1980–1984 (Table 1). Historical changes in emissions from other  
214 source regions can be found in Hoesly et al. (2018) and Yang et al. (2018b).

### 215 **3 Model Evaluation**

216 EMEP (European Monitoring and Evaluation Programme, <http://www.emep.int>)  
217 networks provide daily near-surface aerosol concentrations in Europe. The annual  
218 mean of daily observations is used to evaluate the model performance in this study.  
219 Compared to the observational data from EMEP networks during 2010–2014, CAM5-  
220 EAST can well reproduce the spatial distribution and magnitude of aerosol  
221 components with normalized mean biases (NMB) of -14%~-23% and correlation  
222 coefficients (R) in a range of 0.43~0.62 for sulfate, BC and organic carbon (OC,  
223 derived from POA and SOA from the model results) (Figures 2).

224 Figure 3 shows the time series of annual mean near-surface sulfate, BC, and OC  
225 concentrations averaged over EMEP sites in Europe and the corresponding model  
226 values during 1993–2018. Variations in near-surface sulfate concentrations are  
227 consistent between the model and observations, with R values higher than 0.9. The  
228 observed variations of BC and OC concentrations in Europe are represented in the  
229 simulation, with R values of 0.52 and 0.65, respectively. However, the observed high  
230 values of BC and OC concentrations are not captured by the model, probably  
231 because very few data were available before 2010 and, therefore, any difference  
232 between model and observation cannot be smoothed out through the spatial  
233 average. This is also indicated by the large spatial variation of BC and OC  
234 concentrations before 2010. Nevertheless, the modeled concentrations are still  
235 within the range of observations. Note that the number of sites used for the spatial  
236 average in Figure 3 is different from year to year and thus the variation or trend here  
237 does not represent that over a sub-region or the entire Europe.

238 The modeled AOD is evaluated against the AERONET (Aerosol Robotic  
239 Network, <https://aeronet.gsfc.nasa.gov>) data in Figure 8. Both the modeled and  
240 observed AOD show decreasing trends during 2001–2018. The model  
241 underestimates AOD in all four sub-regions of Europe probably due to the lack of  
242 nitrate aerosol. The variations of AOD in Western Europe (combined NW and SW  
243 Europe) are well predicted with R values of about 0.75, but the model barely  
244 reproduces the AOD variations in E. Europe and the GTC region ( $R < 0.5$ ). The  
245 difference of the interannual variation in AOD between the model simulation and

246 observation can be caused by many factors such as aerosol emissions, aerosol  
247 parameterizations in model, aerosol mixing state, inaccurate meteorological fields  
248 from reanalysis data, and biases in measurements. However, identifying the  
249 contribution of each factor to the difference is beyond the scope of this paper.

#### 250 **4. Source Apportionment of Aerosols in Europe**

251 Based on the tagging technique in CAM5-EAST, near-surface concentrations of  
252 total sulfate-BC-POA can be attributed to emissions within and outside Europe, as  
253 shown in Figures 4a and 4b, and the relative contributions in percentage are given in  
254 Figures 4c and 4d. Averaged over 2010–2018, due to the relatively high local  
255 emissions, annual mean sulfate-BC-POA concentrations contributed by European  
256 emissions show peak values of  $4 \mu\text{g m}^{-3}$  in E. Europe. The slight increase in  $\text{SO}_2$   
257 emission from the GTC region (Figure S1), which is opposite to the decreases in the  
258 other three sub-regions of Europe, also leads to high concentrations in GTC, with 2–  
259  $4 \mu\text{g m}^{-3}$  contributed by European emissions. Due to the atmospheric transport from  
260 surrounding regions including North Africa, the Middle East and Russia-Belarus-  
261 Ukraine, non-European emissions account for  $0.5\text{--}1 \mu\text{g m}^{-3}$  over SW Europe, E.  
262 Europe and GTC area. Overall, European local emissions are the dominant sources  
263 of sulfate-BC-POA near-surface concentrations in Europe with contributions larger  
264 than 80% (60%) in central areas (most of Europe). Non-European emissions are  
265 responsible for 30–50% of the near-surface concentrations near the coastal areas  
266 and boundaries of the Europe that are easily influenced by aerosol regional  
267 transport.

268 Figure 5 illustrates the source contributions in percentage of emissions from  
269 different regions of the globe to near-surface aerosol concentrations and column  
270 burdens over the four sub-regions of Europe averaged over 2010–2018. Different  
271 aerosols have fairly different local/remote source attributions depending on the local  
272 to remote emission ratio and transport efficiency. European emissions explain 54%–  
273 68% of near-surface sulfate concentrations over the four sub-regions of Europe, with  
274 the largest local contribution in E. Europe due to the relatively high emission rate.  
275 The emissions from Europe dominate BC and POA concentrations in Europe with  
276 contributions in the range of 78%–95% and 58%–78%, respectively. The impact of  
277 local emissions on near-surface sulfate concentration is smaller than BC and POA.  
278 This is partially due to the less efficient gas scavenging than particles and the  
279 additional SO<sub>2</sub>-to-sulfate conversion process that increases the atmospheric  
280 residence time of sulfur. On the other hand, the higher initial injection height of SO<sub>2</sub>  
281 emissions from some sources (e.g., industrial sector and power plants) facilitates the  
282 lifting of SO<sub>2</sub> and sulfate aerosol into the free atmosphere and, therefore, favors the  
283 long-range transport (Yang et al., 2019). The efficient reduction in local SO<sub>2</sub>  
284 emissions in Europe also caused the lower influences of local emissions on sulfate  
285 concentrations in recent years.

286 Anthropogenic emissions over oceans (e.g., international shipping) and natural  
287 emissions of oceanic dimethyl sulfide (DMS) and volcanic activities together account  
288 for 16%–28% of near-surface sulfate concentrations in the four sub-regions of  
289 Europe. About 10% of sulfate and 5%–10% of BC and POA in E. Europe and GTC

290 come from Russia-Belarus-Ukraine emissions. North Africa contributes to 7% of  
291 sulfate, 17% of BC and 24% of POA over SW Europe. The contributions of  
292 emissions, from the Middle East, to aerosol concentrations in GTC are between 5%  
293 and 10%.

294 The transboundary and intercontinental transport of aerosols occur most  
295 frequently in the free troposphere rather than near the surface, as horizontal transport  
296 pathways at the surface and 500 hPa are indicated on the spatial distribution map of  
297 the relative contributions shown in Figures S2 and S3. This also leads to larger  
298 relative contributions from non-European sources to aerosol column burdens than to  
299 the near-surface concentrations (Figure 5). The European emissions only contribute  
300 32%–47% of column burden of sulfate, 57%–75% of BC and 51%–71% of POA over  
301 the four sub-regions of Europe. Over NW Europe and SW Europe, about 10%–15%  
302 of the sulfate burden is from East Asia and Russia-Belarus-Ukraine, respectively.  
303 Sources in North Africa are responsible for 27% and 14% of BC and 19% and 11%  
304 of POA burden over SW Europe and GTC, respectively. Emissions from North  
305 America account for 15% and 10% POA burden over NW Europe and SW Europe,  
306 respectively. Emissions from Russia-Belarus-Ukraine explain 12% and 19% of POA  
307 burden over E. Europe and GTC, respectively. Since near-surface aerosol  
308 concentrations directly affect air quality and column burden is more relevant to  
309 radiative impact, the differences of relative contributions between near-surface  
310 concentrations and column burden highlight the possible roles of non-local emissions  
311 in either air quality or energy balance over Europe.

312 Source contributions to aerosols in Europe vary with season due to the  
313 seasonality of emissions and meteorology. In general, local sources have the largest  
314 contributions to both near-surface concentration and column burden of European  
315 aerosols in winter and smallest contributions in summer averaged over 2010–2018  
316 (outer rings in Figure 6). With the contributions normalized by the ratio of seasonal  
317 anthropogenic emission to annual mean for each source, the impact of emission  
318 seasonal variation on the source contributions can be removed (inner rings in Figure  
319 6) (Yang et al., 2019). Without the influence of emission seasonality, local source  
320 contributions decrease in winter and increase in summer, indicating that it was the  
321 higher local anthropogenic emissions that result in the larger local source  
322 contributions to wintertime aerosols in Europe relative to other seasons. Sulfur  
323 sources over oceans account for one fourth to one third of European sulfate  
324 concentration and burden in spring likely due to the strong westerlies in this season  
325 that transport aerosols from the North Atlantic Ocean to the Europe. Source  
326 contributions from Russia-Belarus-Ukraine and North America to BC and POA in  
327 Europe show strong seasonal variabilities, which can be explained by the changes in  
328 biomass burning emissions considering its large seasonal variability.

## 329 **5. Source Apportionment of Long-term Trends**

330 Total sulfate-BC-POA concentrations decreased during 1980–2018 over all of  
331 the four sub-regions of Europe (Figure 7), since that near-surface aerosol  
332 concentrations in Europe are dominated by its local emissions and the European  
333 anthropogenic emissions have significantly decreased during this time period.



334 Averaged over the entire Europe, near-surface concentrations of sulfate, BC and  
335 POA decreased by 70%, 43% and 23%, respectively, between 1980–1984 and  
336 2014–2018, which is consistent with the decreases in local emissions (Table 1). The  
337 total sulfate-BC-POA concentrations decreased by 62%. With SOA included, this  
338 value does not have a substantial change (from 62% to 59%) and the decreasing  
339 trends in the four sub-regions of the Europe are not largely affected either. The  
340 column burden of sulfate, BC, POA and the sum of these three decreased by 60%,  
341 28%, 4% and 55%, respectively, which is less than the decrease in corresponding  
342 near-surface concentration. It is because non-local emissions have larger influences  
343 at high altitudes than at the surface, which partly dampened the contribution of near-  
344 surface aerosol decrease (induced by reductions in local emissions) to the  
345 column integration.

346 The decrease in European local emissions explains 93% of the reduced  
347 concentration and 91% of the reduced burden in Europe between the first and last  
348 five-year period of 1980–2018, while 8%–9% is contributed by the reduction in  
349 emissions from Russia-Belarus-Ukraine (Table 2). The decrease in emissions from  
350 North America also explains 10% of the reduced column burden of sulfate-BC-POA  
351 in Europe from 1980–1984 to 2014–2018. Along with the decreases in local emission  
352 contributions to near-surface sulfate-BC-POA concentrations in Europe, the fraction  
353 of non-European emission contributions increased from 10%–30% to 30%–50%  
354 during 1980–2018 (Figure 7), indicating that aerosols from foreign emissions through  
355 long-range transport have become increasingly important to air quality in Europe.

356 Regulations for further improvement of air quality in Europe in the near future need  
357 to take changes in non-European emissions into account.

358 Similar to the declining trend in column burden, simulated total AOD also  
359 decreased from 0.12–0.16 to 0.06–0.08 in NW Europe and SW Europe and from  
360 0.19–0.21 to 0.09–0.13 in E. Europe and GTC region during the past four decades  
361 (Figure 8). Sulfate AOD accounts for the largest portion of total combustion AOD  
362 (sum of sulfate, BC, POA and SOA) over the four sub-regions of Europe. The  
363 combustion AOD has decreased by 0.065 from 1980–1984 to 2014–2018 (Table 1),  
364 with 0.059 (91%) contributed by the decrease in sulfate AOD. Therefore, we focus  
365 on sulfate aerosol when examining the decadal changes in AOD and DRF in Europe  
366 below.

367 The decreased sulfate AOD can also be decomposed into different contributions  
368 from individual source regions in CAM5-EAST. European local emissions contribute  
369 to 89% of the decrease, followed by 9% and 7% attributed to changes in emissions  
370 from Russia-Belarus-Ukraine and North America, respectively, with the residual  
371 offset by other source regions (Table 2). Over the last four decades, model simulated  
372 sulfate AOD decreased at a rate of 0.017, 0.017, 0.026 and 0.012 decade<sup>-1</sup>,  
373 respectively, over NW Europe, SW Europe, E. Europe and GTC. Decreases in  
374 European local SO<sub>2</sub> emissions result in 78% of the sulfate AOD decreases over GTC  
375 and about 90% over the other three sub-regions (Figure 9). For the remote sources,  
376 emission changes in North America explain 5%–10% of the European sulfate AOD  
377 decrease, while Russia-Belarus-Ukraine sources contribute 29% of the sulfate AOD

378 decrease over GTC and 6%–8% over NW Europe and E. Europe, indicating a  
379 possible warming enhancement effect of changes in emissions from North America  
380 and Russia-Belarus-Ukraine.

381 Averaged over 1980–2018, sulfate imposed a cooling effect over Europe with the  
382 maximum negative DRF at the top of the atmosphere (TOA) exceeding  $-3 \text{ W m}^{-2}$  in  
383 E. Europe (Figure 10). Compared to 1980–1984, the magnitude of sulfate DRF  
384 decreased in 2014–2018, leading to a  $1\text{--}3 \text{ W m}^{-2}$  warming mainly in E. Europe. The  
385 warming effect mostly came from local  $\text{SO}_2$  emission reduction, while non-European  
386 emission changes only contributed less than  $0.4 \text{ W m}^{-2}$  over most regions of the  
387 Europe. Considering Europe as a whole, the decrease in sulfate DRF caused a  
388 warming effect of  $2.0 \text{ W m}^{-2}$ , with 88% and 12% coming from reductions European  
389 local emissions and changes in non-European emissions, respectively (Tables 1 and  
390 2).

391 Future changes in sulfate DRF associated with European and non-European  
392 emissions based on eight SSP scenarios are also estimated and shown in Figure 11  
393 and Figure S4 gives the estimate for each SSP scenario. Sulfate DRF contributed by  
394 both European and non-European emissions would decrease in the near future but  
395 has large variabilities between different SSPs. The sulfate DRF (cooling) over  
396 Europe contributed from European local emissions shows a decrease from  $-0.48 \text{ W}$   
397  $\text{m}^{-2}$  in year 2015 to  $-0.18$  ( $-0.08 \sim -0.33$ )  $\text{W m}^{-2}$  in year 2030 and  $-0.14$  ( $-0.05 \sim -0.29$ )  
398  $\text{W m}^{-2}$  in year 2050. Unlike their contributions to the historical (1980–2018) change,  
399 non-European emissions have an increasingly significant impact on the future sulfate

400 DRF changes in Europe. The contributions of non-European emissions decrease  
401 from  $-0.68 \text{ W m}^{-2}$  in year 2015 to  $-0.39$  ( $-0.13 \sim -0.64$ )  $\text{W m}^{-2}$  in year 2030 and  $-0.26$   
402 ( $-0.08 \sim -0.63$ )  $\text{W m}^{-2}$  in year 2050, with the changes in a magnitude similar to that of  
403 European local emissions. It suggests that future changes in non-European  
404 emissions are as important as European emissions to radiative balance and  
405 associated regional climate change in Europe.

## 406 **6. Conclusions**

407 Using a global aerosol-climate model with an explicit aerosol source tagging  
408 technique (CAM5-EAST), we examine the long-term trends and source  
409 apportionment of aerosols in Europe over 1980–2018 from sixteen source regions  
410 covering the globe in this study. CAM5-EAST can well capture the spatial distribution  
411 and temporal variation of aerosol species in Europe during this time period.

412 Averaged over 2010–2018, European emissions account for 54%–68%, 78%–  
413 95% and 58%–78% of near-surface sulfate, BC, and POA concentrations over  
414 Europe, respectively. Russia-Belarus-Ukraine emissions explain 10% of sulfate in E.  
415 Europe and GTC. North Africa contributes to 17% of BC and 24% of POA over SW  
416 Europe. Anthropogenic emissions over oceans (e.g., from international shipping) and  
417 natural emissions from marine and volcanic activities together account for 16%–28%  
418 of sulfate near-surface concentrations in Europe. European emissions only account  
419 for 32%–47%, 57%–75% and 51%–71% of column burden of sulfate, BC and POA,  
420 respectively, in Europe, with the rest contributed by emissions from East Asia,  
421 Russia-Belarus-Ukraine, North Africa and North America. Source contributions of

422 aerosols in Europe vary with seasons driven by the seasonality of emissions and  
423 meteorology.

424 Compared to 1980–1984, simulated total sulfate-BC-POA near-surface  
425 concentration and column burden over 2014–2018 had a decrease of 62% and 55%,  
426 respectively, the majority of which was contributed by reductions in European local  
427 emissions. The decrease in emissions from Russia-Belarus-Ukraine contributed 8%–  
428 9% of the near-surface concentration decrease, while the decrease in emissions  
429 from North America accounted for 10% of the reduced column burden. With the large  
430 decrease in local emission contribution, aerosols from foreign sources became  
431 increasingly important to air quality in Europe. The decrease in sulfate led to a 2.0 W  
432 m<sup>-2</sup> warming in Europe, with 12% coming from changes in non-European emissions,  
433 especially in North America and Russia-Belarus-Ukraine. Based on the SSP  
434 scenarios and the assumed relationship between DRF and emissions, we estimated  
435 that sulfate DRF over Europe contributed from European emissions and non-  
436 European emissions would decrease at a comparable rate in the near future. This  
437 suggests that future changes in non-European emissions are as important as  
438 European emissions in affecting regional climate change associated with aerosols in  
439 Europe. It should also be noted that the model currently does not have the ability to  
440 simulate nitrate and ammonium aerosols and, therefore, the conclusions may not  
441 hold with all aerosols.

442

443

444 ***Data availability.***

445 The default CAM5 model is publicly available at  
446 <http://www.cesm.ucar.edu/models/cesm1.2/> (last access: 16 August 2019). Our  
447 CAM5-EAST model code and results can be made available through the National  
448 Energy Research Scientific Computing Center (NERSC) servers upon request.

449

450 ***Competing interests.***

451 The authors declare that they have no conflict of interest.

452

453 ***Author contribution.***

454 YY, SL, and HW designed the research; YY performed the model simulations; YY,  
455 and SL analyzed the data. All the authors discussed the results and wrote the paper.

456

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674 **Table 1.** Annual emissions (Tg yr<sup>-1</sup>), concentrations (µg m<sup>-3</sup>), column burden (mg m<sup>-2</sup>)  
675 <sup>2</sup>), AOD (scaled up by a factor of 100) and DRF (W m<sup>-2</sup>) of Sulfate, BC, POA, SBP  
676 (sulfate-BC-POA) and SBP-SOA (sulfate-BC-POA-SOA) in Europe averaged over  
677 1980–1984 and 2014–2018, as well as the differences between 1980–1984 and  
678 2014–2018. Differences in percentage relative to mean values in 1980–1984 are  
679 presented in parentheses.  
680

		Emis.	Conc.	Burden	AOD*100	DRF
Sulfate	1980–1984	15.10	6.00	14.35	9.13	-3.27
	2014–2018	2.53	1.80	5.79	3.24	-1.24
	Δ	-12.57 (-83.2)	-4.20 (-70.0)	-8.55 (-59.6)	-5.89 (-64.6)	2.04 (-62.2)
BC	1980–1984	0.47	0.4	0.38	0.7	--
	2014–2018	0.25	0.23	0.28	0.5	--
	Δ	-0.22 (-45.8)	-0.17 (-43.0)	-0.11 (-27.6)	-0.21 (-29.2)	--
POA	1980–1984	1.24	1.12	1.12	0.63	--
	2014–2018	0.94	0.86	1.08	0.58	--
	Δ	-0.30 (-24.4)	-0.26 (-23.2)	-0.04 (-3.8)	-0.05 (-7.5)	--
Sulfate-BC-POA	1980–1984	--	7.52	15.85	10.46	--
	2014–2018	--	2.89	7.15	4.32	--
	Δ	--	-4.63 (-61.6)	-8.70 (-54.9)	-6.15 (-58.7)	--
SBP-SOA	1980–1984	--	10.48	19.58	11.92	--
	2014–2018	--	4.34	8.55	5.44	--
	Δ	--	-6.14 (-58.6)	-11.03 (-56.3)	-6.48 (-54.37)	--

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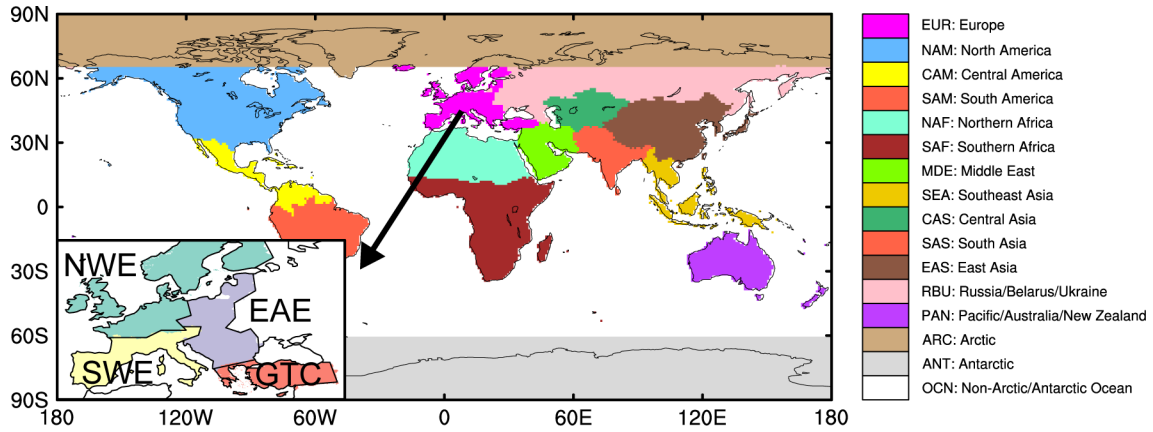
683 **Table 2.** Relative contributions (%) of emissions from major source regions to the  
 684 changes in near-surface concentrations, column burden, AOD and DRF in Europe  
 685 between 1980–1984 and 2014–2018.  
 686

	Sulfate-BC- POA		
	$\Delta$ Conc.	$\Delta$ Burden	$\Delta$ AOD
EUR	92.8	91.2	91.2
NAM	1.8	10.0	6.5
NAF	-1.0	-1.5	-1.6
MDE	-0.9	-1.9	-1.5
EAS	-0.3	-3.1	-1.7
RBU	8.0	9.2	8.5
OTH	-0.1	-4.2	-2.0
OCN	-0.3	0.2	0.6

	Sulfate			$\Delta$ DRF
	$\Delta$ Conc.	$\Delta$ Burden	$\Delta$ AOD	
EUR	91.3	89.2	88.9	88.2
NAM	2.1	10.5	6.9	
NAF	-0.6	-0.9	-0.8	
MDE	-0.8	-1.7	-1.3	
EAS	-0.3	-2.8	-1.4	11.8
RBU	8.6	9.5	8.7	
OTH	-0.1	-4.0	-1.8	
OCN	-0.3	0.3	0.7	

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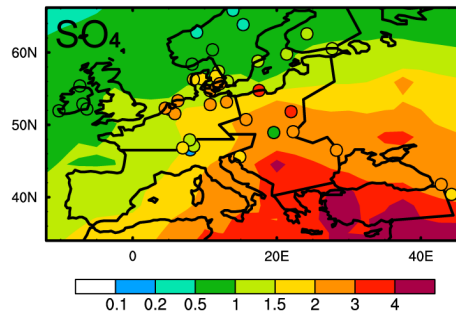
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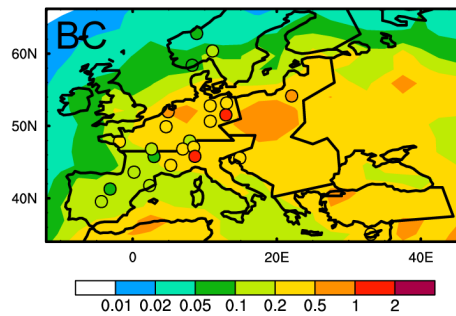
692 **Figure 1.** Source regions that are selected for the Explicit Aerosol Source Tagging  
 693 (EAST) in this study, including Europe (EUR), North America (NAM), Central  
 694 America (CAM), South America (SAM), North Africa (NAF), South Africa (SAF), the  
 695 Middle East (MDE), Southeast Asia (SEA), Central Asia (CAS), South Asia (SAS),  
 696 East Asia (EAS), Russia-Belarus-Ukraine (RBU), Pacific-Australia-New Zealand  
 697 (PAN), the Arctic (ARC), Antarctic (ANT), and Non-Arctic/Antarctic Ocean (OCN).  
 698 The embedded panel (at bottom left) is Europe, as the receptor region, which is  
 699 further divided to Northwestern Europe (NWE), Southwestern Europe (SWE),  
 700 Eastern Europe (EAE) and Greece-Turkey-Cyprus (GTC).

701

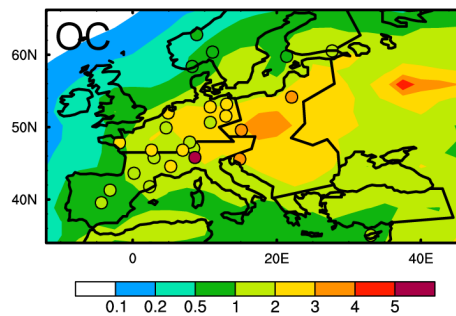
a NMB = -14% R = 0.62



b NMB = -13% R = 0.43



c NMB = -23% R = 0.57



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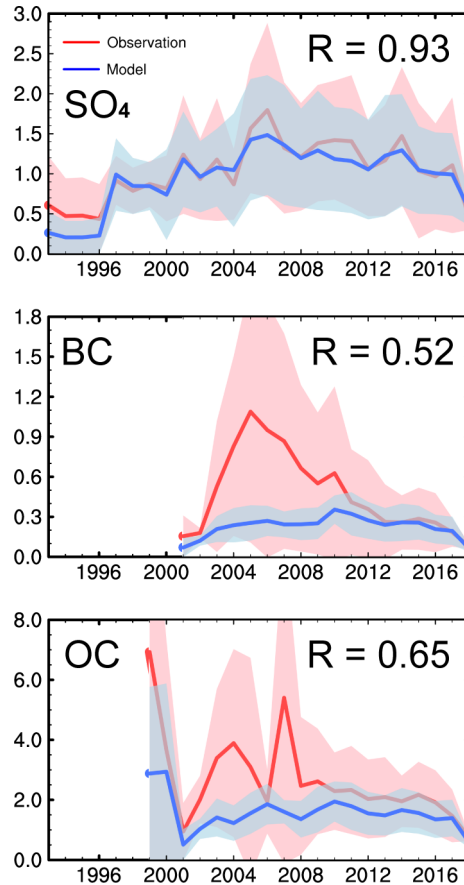
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**Figure 2.** Spatial distribution of simulated (contour) and observed (color-filled circles) annual mean near-surface (a) sulfate, (b) BC, and (c) OC (derived as (POA+SOA)/1.4 in model) concentrations ( $\mu\text{g m}^{-3}$ ) over 2010–2014. Observations are from EMEP (European Monitoring and Evaluation Programme) networks. Normalized mean bias ( $\text{NMB} = 100\% \times \sum(\text{Model}_{\text{site}} - \text{Observation}_{\text{site}}) / \sum \text{Observation}_{\text{site}}$ ) and correlation coefficient (R) between observed and simulated concentrations are noted at the top of each panel.



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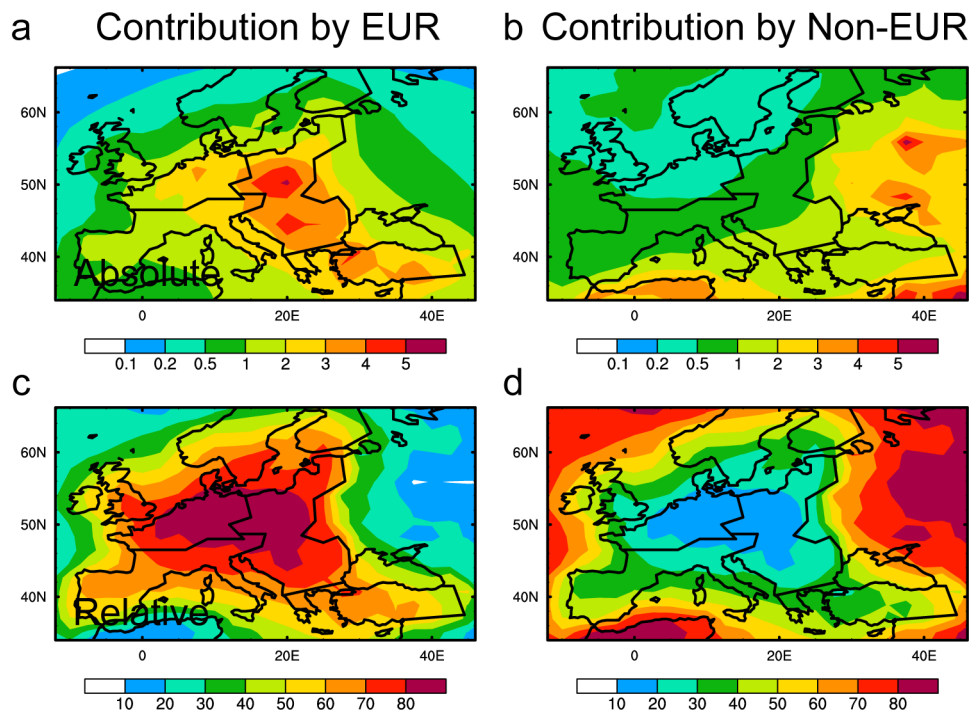
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714 **Figure 3.** Time series (1993–2018) of spatial and annual mean near-surface (a)  
 715 sulfate, (b) BC, and (c) OC concentrations ( $\mu\text{g m}^{-3}$ ) in Europe from model simulation  
 716 (blue lines) and observations (red lines). Model results are plotted only when EMEP  
 717 observational data are available. Shaded areas represent 1- $\sigma$  spatial standard  
 718 deviation of annual mean concentrations for each year. Temporal correlation  
 719 coefficients (R) between observed and simulated spatially averaged concentrations  
 720 are noted on the top-right corner of each panel.

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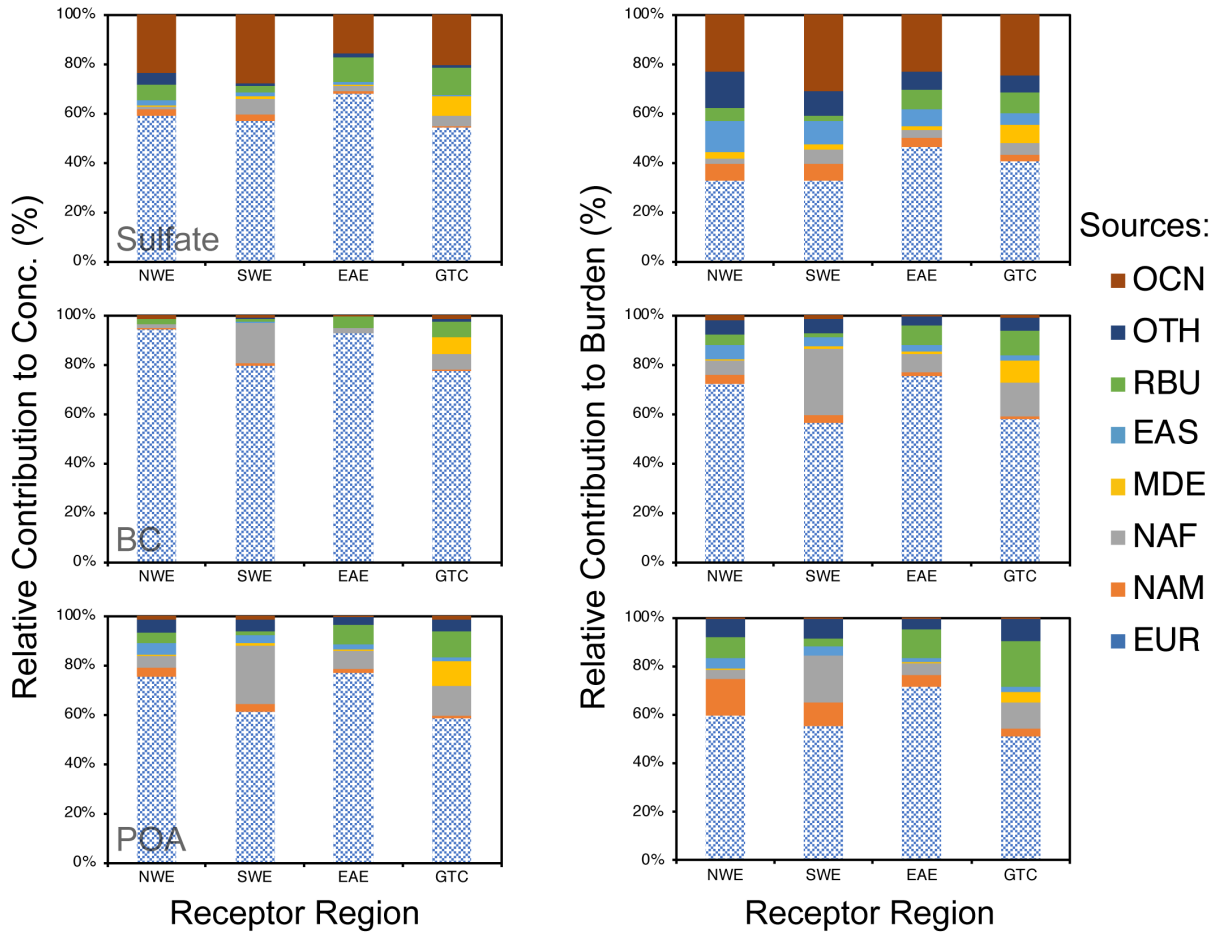
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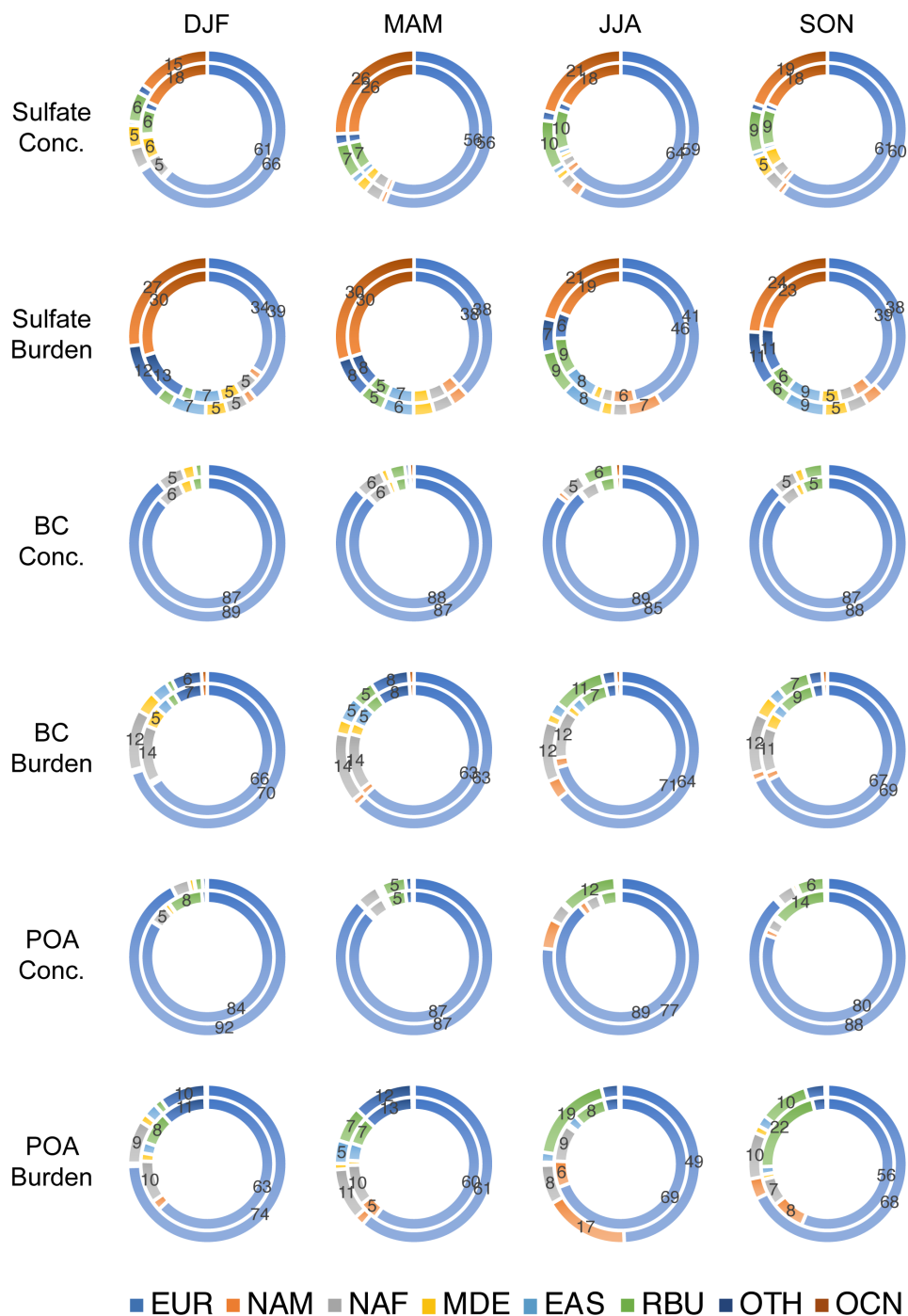
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**Figure 4.** (a,b) Absolute ( $\mu\text{g m}^{-3}$ ) and (c,d) relative contributions (%) to annual mean near-surface concentrations of sulfate-BC-POA from European local emissions (EUR) and emissions outside the Europe (Non-EUR), respectively, averaged over 2010–2018.



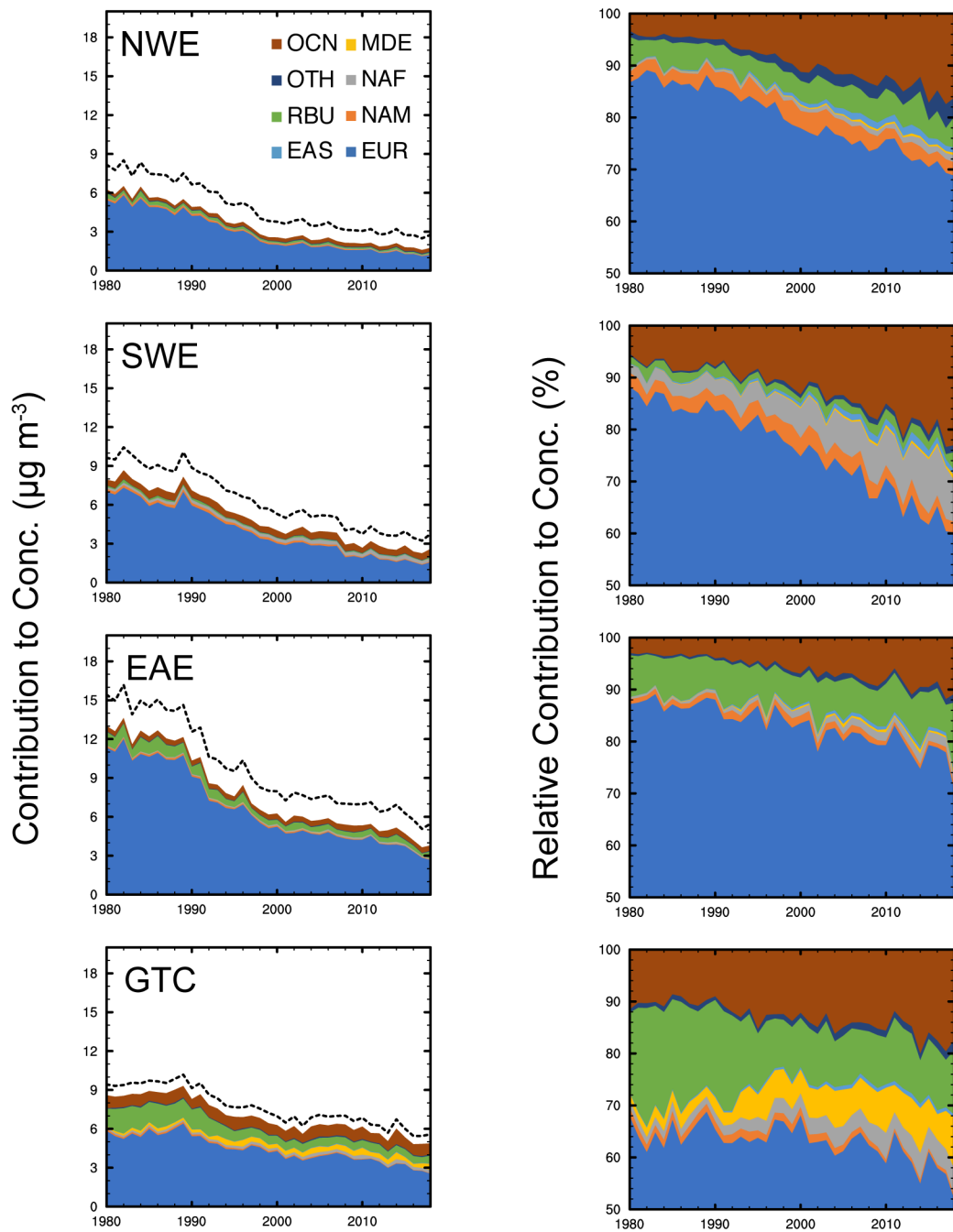
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**Figure 5.** Relative contributions (%) by emissions from major tagged source regions including Europe (EUR), North America (NAM), North Africa (NAF), the Middle East (MDE), East Asia (EAS), Russia-Belarus-Ukraine (RBU), Non-Arctic/Antarctic Ocean (OCN) and other (OTH) regions to near-surface concentrations (left) and column burdens (right) of sulfate, BC and POA (from top to bottom) in the four sub-regions of Europe averaged over 2010–2018. Patterned areas represent EUR local contributions.



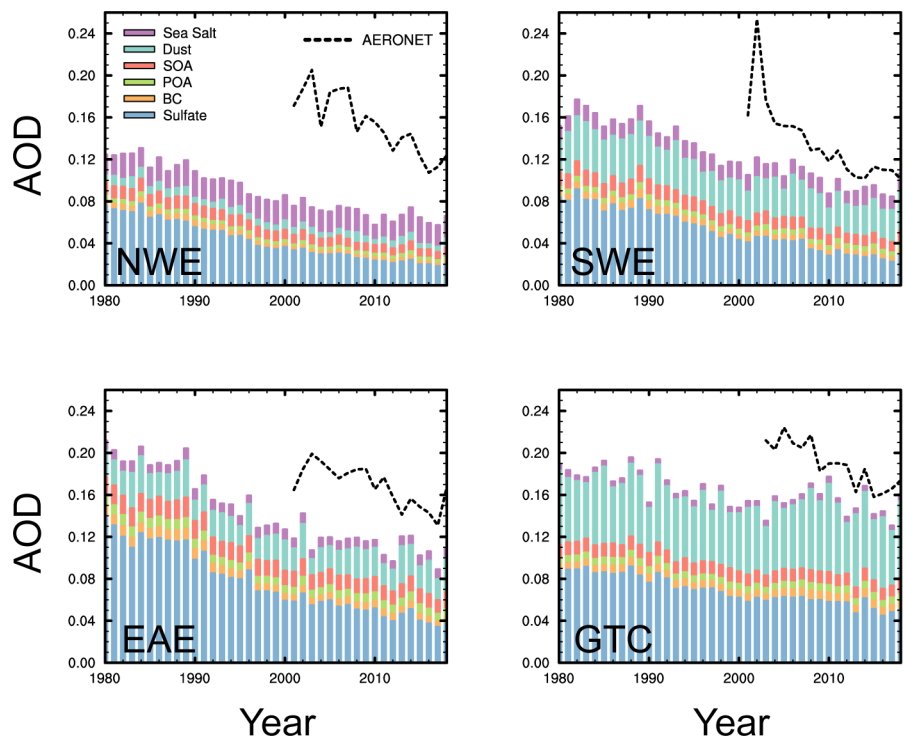
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**Figure 6.** Relative contributions (%) by emissions from major tagged source regions to near-surface concentrations (Conc.) and column burdens of December-January-February (DJF), March-April-May (MAM), June-July-August (JJA) and September-October-November (SON) mean sulfate, BC and POA over the Europe averaged over 2010–2018. Outer rings represent the modeled values and the relative contributions in inner rings is calculated based on absolute values normalized by the ratio of seasonal emission to annual mean. Values larger than 5% are marked.



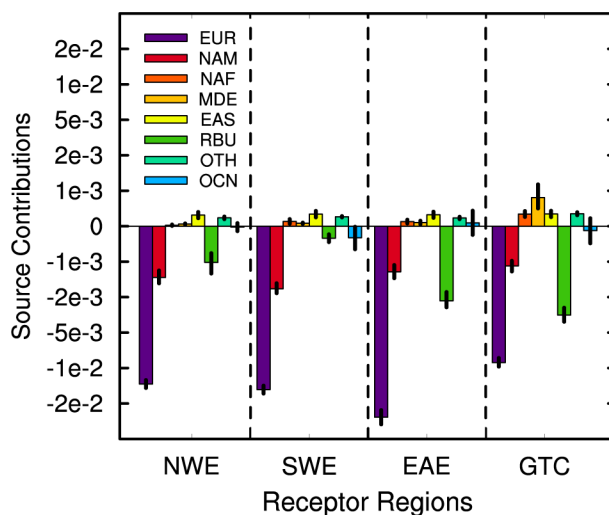
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**Figure 7.** Time series (1980–2018) of absolute (left,  $\mu\text{g m}^{-3}$ ) and relative (right, %) contributions of emissions from major source regions to the simulated annual mean near-surface sulfate-BC-POA concentrations averaged over the four sub-regions of Europe. Dashed lines in left panels represent simulated aerosol concentrations including SOA.



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**Figure 8.** Time series (1980–2018) of simulated annual mean AOD for sulfate, BC, POA, SOA, dust and sea salt averaged over the four sub-regions of Europe. Dashed lines represent AOD from AERONET measurements.



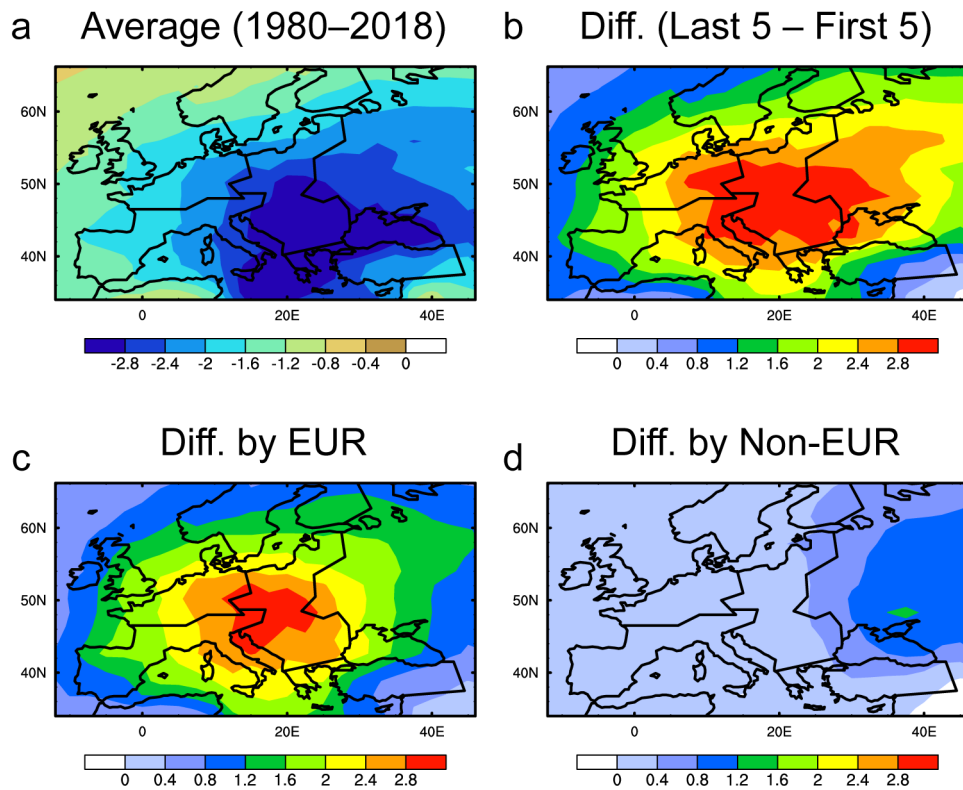
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769 **Figure 9.** Absolute contributions ( $\text{decade}^{-1}$ ) of the emissions from major source  
 770 regions to the trends of sulfate AOD over the four sub-regions of Europe. Error bars  
 771 represent 95% confidence intervals of the linear regression.

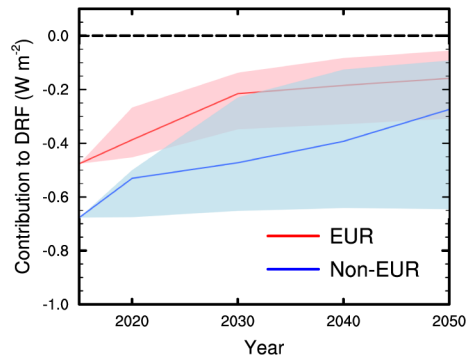
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**Figure 10.** (a) Simulated annual mean DRF ( $W m^{-2}$ ) of sulfate averaged over 1980–2018 and (b) the difference in sulfate DRF between 1980–1984 and 2014–2018. The contributions of European and non-European emissions to the difference are given in (c) and (d), respectively.



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**Figure 11.** Time series (2015–2050) of estimated annual mean sulfate DRF over Europe contributed by European and non-European emissions. Lines and areas represent median values and minimum-to-maximum ranges of the estimated sulfate DRF from eight SSP scenarios, including SSP1-1.9, SSP1-2.6, SSP2-4.5, SSP3-7.0, SSP4-3.4, SSP4-6.0, SSP5-3.4, and SSP5-8.5. Future DRF of sulfate aerosol over Europe is estimated by scaling historical mean (1980–2018) sulfate DRF using the ratio of SSPs future SO<sub>2</sub> emissions to historical emissions assuming a linear response of DRF to regional emissions.