Characterization of organic aerosol across the global remote troposphere: A comparison of ATom measurements and global chemistry models

Alma Hodzic[1], Pedro Campuzano-Jost[2,3], Huisheng Bian[4], Mian Chin[4], Peter R. Colarco[4], Douglas A. Day[2,3], Karl D. Froyd[2,8], Bernd Heinold[6], Duseong S. Jo[2,3], Joseph M. Katich[2,8], John K. Kodros[5], Benjamin A. Nault[2,3], Jeffrey R. Pierce[5], Eric Ray[2,8], Jacob Schacht[6], Gregory P. Schill[2,8], Jason C. Schroder[2,3], Joshua P. Schwarz[2,8], Donna T. Sueper[2,3], Ina Tegen[6], Simone Tilmes[1], Kostas Tsigaridis[7,9], Pengfei Yu[8,10], Jose L. Jimenez[2,3]

[1]*National Center for Atmospheric Research, Boulder, CO, USA*

[2]*Cooperative Institute for Research in Environmental Sciences (CIRES), University of Colorado, Boulder, CO, USA*

[3]*Department of Chemistry, University of Colorado, Boulder, CO, USA*

[4]*NASA Goddard Space Flight Center, Greenbelt, MD, USA*

[5]*Department of Atmospheric Science, Colorado State University, Fort Collins, CO, USA*

[6]*Leibniz Institute for Tropospheric Research, Leipzig, Germany*

[7]*Center for Climate Systems Research, Columbia University, New York, NY, USA*

[8]*NOAA Earth System Research Laboratory (ESRL), Chemical Sciences Division, Boulder, CO, USA*

[9]*NASA Goddard Institute for Space Studies, New York, NY, USA*

[10]*Institute for Environmental and Climate Research, Jinan University, Guangzhou, Guangdong, China*

Key words: organic aerosol, remote atmosphere, ATom field campaign.

Abstract.

The spatial distribution and properties of submicron organic aerosols (OA) are among the key sources of uncertainty in our understanding of aerosol effects on climate. Uncertainties are particularly large over remote regions of the free troposphere and Southern Ocean, where very little data has been available, and where OA predictions from AeroCom Phase II global models span two to three orders-of-magnitude, greatly exceeding the model spread over source regions. The (nearly) pole-to-pole vertical distribution of non-refractory aerosols was measured with an aerosol mass spectrometer onboard the NASA DC8 aircraft as part of the Atmospheric Tomography (ATom) mission during the northern hemisphere summer (August 2016) and winter (February 2017). This study presents the first extensive characterization of OA mass concentrations and their level of oxidation in the remote atmosphere. OA and sulfate are the major contributors by mass to submicron aerosols in the remote troposphere, together with sea salt in the marine boundary layer. Sulfate was dominant in the lower stratosphere. OA concentrations have a strong seasonal and zonal variability, with the highest levels measured in the lower troposphere in the summer and over the regions influenced by the biomass burning from Africa (up to 10 μg sm^{-3}). Lower concentrations (~0.1-0.3 μg sm^{-3}) are observed in the northern mid- and high- latitudes and very low concentrations (< 0.1 μg sm^{-3}) in the southern mid- and high- latitudes. The ATom dataset is used to evaluate predictions of eight current global chemistry models that implement a variety of commonly used representations of OA sources and chemistry, as well as of the AeroCom-II ensemble. The current model ensemble captures the average vertical and spatial distribution of measured OA concentrations, and the spread of the individual models remains within a factor of 5. These results are significantly improved over the AeroCom-II model ensemble, which shows large overestimations over these regions. However, some of the improved agreement with observations occurs for the wrong reasons, as models have the tendency to greatly overestimate the primary OA fraction, and underestimate the secondary fraction. Measured OA in the remote free troposphere are highly oxygenated with organic aerosol to organic carbon (OA/OC) ratios of ~ 2.2-2.8 and are 30-60% more oxygenated than in current models, which can lead to significant errors in OA concentrations. The model/measurement comparisons presented here support the concept of a more dynamic OA system as proposed by Hodzic et al. (2016), with enhanced removal of primary OA, and a stronger production of secondary OA in global models needed to provide a better agreement with observations.

1   Introduction

Organic aerosols (OA) are a complex mixture of directly emitted primary OA (POA) and chemically produced secondary OA (SOA) from anthropogenic and biogenic emission sources. They are associated with adverse health effects (Mauderly and Chow, 2008, Shiraiwa et al., 2017) and contribute radiative forcing in the climate system (Boucher et al., 2013). The currently limited understanding of processes involved in the formation, ageing, and removal of organic compounds results in large uncertainties in (i) the predicted global OA burden, (ii) relative contributions of emissions vs. chemistry to OA formation, (iii) spatial distribution, and (iv) impacts on radiation and clouds (Kanakidou et al., 2005, Hallquist et al., 2009, Heald et al., 2011, Spracklen et al., 2011, Tsigaridis et al., 2014, Hodzic et al., 2016, Shrivastava et al., 2017, Tsigaridis and Kanakidou, 2018, Zhu et al., 2019). The uncertainties are particularly large in the estimated global burden of SOA that range from 12 to 450 Tg y^{-1} (see Fig. 9 of Hodzic et al., 2016), and in their direct and indirect radiative forcing that range from -0.08 to -0.33 W m^{-2}, and -0.60 to −0.77 W m^{-2}, respectively (Spracklen et al., 2011, Myhre et al., 2013, Scott et al., 2014, Hodzic et al., 2016, Tsigaridis and Kanakidou, 2018). Reducing these uncertainties is becoming more important as OA is on a path to becoming the dominant fraction of the submicron anthropogenic aerosol mass globally due to the ongoing efforts to reduce SO_2 emissions and associated sulfate aerosols.

Model performance has been especially poor in the remote regions of the atmosphere where OA measurements available for model evaluation have been sparse (especially aloft). Using data from 17 aircraft campaigns mostly located in the northern hemisphere Heald et al. (2011) showed that the skill of the global GEOS-Chem model in predicting the vertical distribution of OA was significantly decreased in remote regions compared to polluted near-source regions. The study pointed out the limitations of commonly used SOA formation mechanisms that are based on chamber data; these have the tendency to underpredict OA in source regions and overpredict OA in the remote troposphere. For a subset of 9 recent aircraft campaigns, Hodzic et al. (2016) showed that OA is likely a more dynamic system than represented in chemistry-climate models, with both stronger production and stronger removals. These authors suggested that additional removal mechanisms via e.g. photolytic or heterogeneous reactions of OA particles are needed to explain low OA concentrations observed in the upper troposphere where direct cloud scavenging is less efficient. The recent global multi-model comparison study (Tsigaridis

99 et al., 2014) within the AeroCom Phase II project illustrates well the amplitude of model
100 uncertainties simulating OA mass concentrations and the contrast in model performance
101 between near-source and remote regions. The results indicate that model dispersion (the
102 spread between the models with the lowest and highest predicted OA concentrations)
103 increases with altitude from roughly 1 order of magnitude near the surface to 2-3 orders
104 of magnitude in the upper troposphere. Our own analyses of the AeroCom-II models
105 shown in Figure 1a indicate that model dispersion (quantified as the ratio of the average
106 concentration of the highest model to that of the lowest one, in each region) increases not
107 only with altitude but also with distance from the northern mid-latitude source (and data-
108 rich) regions. The model spread is a factor of 10-20 in the free troposphere between the
109 equator and northern mid-latitudes, and increases to a factor of 200-800 over the Southern
110 Ocean and near the tropopause. It is not surprising that model spread is lower closer to
111 source regions where it is mostly driven by uncertainties in emissions and SOA production
112 yields. Spread is expected to be larger in remote regions where models are also impacted
113 by uncertainties in transport, chemical ageing and removal. The lowest model dispersion
114 also coincides with the regions of the northern hemisphere (NH) or the African biomass
115 burning outflow where models have been evaluated the most (Figure 1b), emphasizing
116 the need for further model/observation comparison studies in remote regions (of the
117 southern hemisphere (SH) in particular).

118 Here, we present a unique data set of airborne aerosol mass spectrometer measurements
119 of OA mass concentrations collected onboard the NASA DC-8 as part of the Atmospheric
120 Tomography (ATom) mission. The aircraft sampled the vertical structure of the
121 atmosphere from near-surface (0.2 km) to the lower-stratosphere (LS) regions (12 km
122 altitude) over both the Pacific and Atlantic basins (to limit the influence of source regions)
123 with a quasi-global spatial coverage from 82°N to 67°S. This dataset is used to perform
124 the first systematic global-scale multi-model evaluation of the chemistry-climate models
125 focusing on OA in the remote troposphere over the remote oceans. We focus on the NH
126 summer (August 2016, ATom-1) and NH winter (February 2017, ATom-2) deployments.
127 Overall these ATom missions sampled the marine boundary layer (MBL) for 10% of the
128 flight tracks, 12% of the time the remote lower stratosphere, and the rest the free
129 troposphere. The model-observation comparisons are aimed at identifying discrepancies
130 in terms of OA mass concentrations and vertical distribution, their fractional contribution
131 to submicron aerosols, and their oxidation level in global models.

132　The modeling framework is described in Section 2. Section 3 describes the ATom dataset
133　and the spatial and vertical distributions of OA over the Atlantic and Pacific regions.
134　Section 4 presents the comparisons of ATom-1 and -2 data to multi-model predictions
135　from both the AeroCom-II models, and the ensemble of eight current model simulations of
136　the ATom campaign. Section 5 presents the conclusions of the study and discusses its
137　implications.

138　## 2　Modeling framework

139　### 2.1　ATom model simulations

140　ATom measurements were compared with results of eight global models that simulated
141　the time period of the ATom-1 and 2 campaigns (August 2016 and February 2017), using
142　the emissions and reanalysis meteorology corresponding to this period (and a spin-up
143　time of at least six to twelve months). These are referred hereafter as ATom models and
144　include the NASA global Earth system model GEOS5, the aerosol-climate model
145　ECHAM6-HAM, three versions of the NCAR Community Earth System Model (CESM),
146　and three versions of the global chemistry GEOS-Chem model. Simulations were
147　performed at various horizontal resolutions ranging from relatively high ~50km (GEOS5)
148　and ~100km (CESM2 models) resolutions to somewhat coarser grids of ~200km (CESM1-
149　CARMA, GEOS-Chem) and ~400km for GC10-TOMAS. The advantage of using the same
150　host model (in the cases of variants of CESM2 and GEOS-Chem) is that the dynamics
151　and emissions remain comparable. Models differ greatly in their treatment of emissions,
152　gas-phase chemistry, aerosol chemistry and physical processes, and aerosol coupling
153　with radiation and clouds, among others. Table 1 describes the configuration of various
154　models (e.g. meteorology, emissions), and their treatment of OA. In this section we only
155　summarize the main features and parameters directly impacting the OA simulations. Some
156　models do not include SOA chemistry and instead assume that SOA is directly emitted
157　proportional to the emissions of its precursors (ECHAM6-HAM, CESM2-SMP, GEOS5,
158　GC10-TOMAS), while others have more complex treatments of organic compounds, their
159　chemistry, and partitioning into particles (GC12-REF, GC12-DYN, GC10-TOMAS,
160　CESM1-CARMA, CESM2-DYN). It should be noted that models that directly emit SOA
161　assume that SOA is a non-volatile species that remains irreversibly in the particle phase.
162　In all models POA is treated as a non-volatile directly emitted species. In most models
163　(see below) the primary emitted organic aerosol is artificially aged to transition between
164　hydrophobic to hydrophilic POA. There are some commonalities between simulations for

165 the treatment of biogenic emissions, which are based in all models on the Model of
166 Emissions of Gases and Aerosols from Nature (MEGAN, Guenther et al., 2012) to
167 generate meteorology-dependent emissions of volatile organic compounds. None of the
168 models includes the marine production of OA which is estimated to be ~3 orders of
169 magnitude smaller than the continental production of OA from both isoprene and
170 monoterpene precursors (Kim et al., 2017), but could be important in the MBL. This
171 contribution could however be larger for sea-spray biological material from phytoplankton
172 with predicted contributions of 0.01 to 0.1 μg m^{-3} to surface submicron aerosol over remote
173 oceanic regions (Vergara-Temprado et al., 2017, Middlebrook et al., 1998). Below we only
174 provide a brief description of most important processes that influence OA for each model.

175 GEOS5 was run in a configuration similar to Bian et al. (2019) using the anthropogenic
176 emissions from HTAP v2 (Janssens-Maenhout et al., 2015) and biomass burning
177 emissions from the Quick Fire Emission Dataset (QFED v2.54). Aerosols are simulated
178 within the GOCART bulk aerosol module and include externally mixed particles of black
179 carbon (BC), organic carbon (OC), sulfate, ammonium, nitrate, dust and sea salt (Colarco
180 et al., 2010, Bian et al., 2017). The formation of SOA is based on a prescribed 10%
181 formation yield from the monoterpene emissions. The primary emitted OC and SOA are
182 separated into hydrophobic (50%) and hydrophilic (50%) species, with a 2.5 days e-folding
183 time conversion from hydrophobic to hydrophilic organic particles. All SOAs from biogenic,
184 anthropogenic, and biomass burning sources are treated as hydrophilic particles. Both
185 types of organic particles are dry deposited. The hydrophilic OA is removed by large-scale
186 and convective warm clouds, while hydrophobic OA is removed by ice clouds. The
187 hydrophilic particles undergo hygroscopic growth according to the equilibrium
188 parameterization of Gerber (1985).

189 The ECHAM6.3-HAM2.3 standard version (Tegen et al., 2019) was run using updated
190 anthropogenic emissions (Schacht et al, 2019) combining the ECLIPSE (Klimont et al.,
191 2017) emissions, with the Russian anthropogenic BC emissions from Huang et al. (2015).
192 For biomass burning the Global Fire Assimilation System (GFAS, Kaiser et al., 2012)
193 biomass burning emissions are used, however, without the scaling factor of 3.4 suggested
194 by Kaiser et al. (2012). Aerosol composition and processes are simulated using the
195 Hamburg Aerosol Model (HAM2, Zhang et al., 2012), that considers an aerosol internal
196 mixture of sulfate, BC, OC, sea salt, and mineral dust. The aerosol population and their
197 microphysical interactions are simulated using seven log-normal modes, including the

198 nucleation mode, soluble and insoluble Aitken, accumulation and coarse modes. In the
199 model configuration used in this publication the formation of SOA is based on a prescribed
200 15% mass yield from monoterpene emissions only (Dentener et al., 2006). Aerosol
201 particles are removed by dry and wet deposition. The wet deposition includes the below
202 cloud scavenging by rain and in-cloud cloud scavenging for large-scale and convective
203 systems (Croft et al., 2010).

204 The two simulations with the GEOS-Chem 12.0.1 global chemistry model (Bey et al.,
205 2001) use emissions based on CMIP6 global inventory (CEDS historical emissions up to
206 2014 and future emissions based on climate scenarios, Hoesly et al., 2018; Feng et al.,
207 2019) with regional improvements for anthropogenic sources, and on GFED v.4 for
208 biomass burning emissions (Giglio et al., 2013). Both simulations use the bulk aerosol
209 representation and differ only in the treatment of SOA formation and removal. The first
210 configuration (called hereafter GC12-REF) includes the default
211 (http://wiki.seas.harvard.edu/geos-chem/index.php) representation of SOA formation
212 based on Marais et al. (2016) for isoprene-derived SOA, and on the volatility basis set
213 (VBS) of Pye et al. (2010) for all other precursors. Note that this GEOS-Chem REF
214 simulation is similar to the version 12 default "complex option" which includes non-volatile
215 POA and semi-volatile SOA (semi-volatile POA is an optional switch within this version
216 used in Pai et al. 2020). The second configuration (referred to as GC12-DYN) includes a
217 more dynamic representation of the SOA lifecycle based on Hodzic et al. (2016), with the
218 exception of the treatment of isoprene SOA that is formed in the aqueous aerosols as in
219 Marais et al. (2016). As in Hodzic et al. (2016) the GC12-DYN model version includes
220 updated VBS SOA parameterization, updated dry and wet removal of organic vapors, and
221 photolytic removal of SOA (except for isoprene-SOA). SOA formation is based on wall-
222 corrected chamber yields (Zhang et al., 2014) for the traditional precursors (isoprene,
223 monoterpenes, sesquiterpenes, benzene, toluene, xylene) and on yields derived from an
224 explicit chemical mechanism for higher molecular weight n-alkanes and n-alkenes species
225 (Hodzic et al., 2016). The removal of gas-phase oxidized volatile organics uses updated
226 Henry's law solubility coefficients from Hodzic et al. (2014), and photolytic removal of SOA
227 (Hodzic et al., 2015). In addition to OA, the model includes BC and dust, and simulates
228 the chemistry and gas-particle partitioning of inorganic compounds such as sulfate,
229 ammonium, nitrate and sea salt using the ISORROPIA II thermodynamic model
230 (Fountoukis and Nenes, 2007). In both GEOS-Chem configurations, BC and primary OC
231 are simulated with a hydrophobic and hydrophilic fraction for each. At the time of emission,

232   80% of BC and 50% of primary OC are considered as hydrophobic. Hydrophobic aerosols

233   are converted to hydrophilic aerosols with an e-folding lifetime of 1.15 days. An OA/OC

234   ratio of 2.1 is assumed to convert POC to POA, and SOA is simulated as OA mass (i.e.

235   no OA/OC ratio assumption is needed for SOA, except for comparison with OC

236   measurements). Soluble gases and aerosols are removed by both dry and wet deposition.

237   Wet deposition includes scavenging in convective updrafts, and in-cloud and below-cloud

238   scavenging from large-scale precipitation (Liu et al., 2001). Hydrophobic aerosols (BC and

239   POA) are scavenged in convective updrafts following Wang et al. (2014).

240   GC10-TOMAS is based on the GEOS-Chem version 10.01 coupled with TwO Moment

241   Aerosol Sectional microphysics scheme (TOMAS) and ran in a similar configuration to that

242   described in Kodros et al. (2016). The model computes the evolution of sulfate, sea salt,

243   primary and secondary OA, BC, and dust aerosols described by 15 internally mixed size

244   bins (of which six were analyzed for these comparisons, cf. Table 1). Anthropogenic

245   emissions are based on the EDGAR v4 global inventory with regional improvements, while

246   the biomass burning emissions are from GFED v3. SOA are irreversibly made from the

247   emitted parent precursor, considering a 10% mass yield from monoterpene emissions,

248   and an emission flux of 0.2 Tg of SOA per Tg of CO for the anthropogenic CO emissions.

249   The removal of gases and aerosols are treated similar to the GEOS-Chem 12.0.1 model

250   (GC12-REF, see above).

251   Simulations based on the CESM2.0 Earth system model use the standard version of the

252   Whole Atmosphere Community Climate Model (WACCM6, Gettelman et al., 2019,

253   Emmons et al., 2019). Details on the specific of the model configurations are described in

254   detail in Tilmes et al. (2019) i.e. CESM2-SMP and CESM2-DYN correspond to the

255   specified dynamics WACCM6-SOAG and WACCM6-VBSext simulations described in that

256   work, respectively. Emissions are based on the CMIP6 global inventory for the year 2014

257   for anthropogenic sources, and on the QFED version 2.4 for the wildfires inventory.

258   Aerosols are represented with the modal aerosol scheme (MAM4, Liu et al., 2012) that

259   includes BC, primary and secondary OA, sulfate, dust and sea salt. Four modes are

260   considered including Aitken, accumulation and coarse size modes, and an additional

261   primary carbon mode. Only the accumulation mode was used in this work. The CESM2-

262   SMP and CESM2-DYN simulations differ in their treatment of OA. CESM2-SMP forms OA

263   directly using fixed mass yields from primary emitted precursors (isoprene, monoterpenes,

264   aromatics) without explicitly simulating their oxidation and partitioning. These mass yields

265 are increased by a factor of 1.5 to match the anthropogenic aerosol indirect forcing (Liu et
266 al., 2012). The second configuration (referred to as CESM2-DYN) includes the formation
267 and removal parameterizations of organics of Hodzic et al. (2016), as implemented into
268 CESM2 by Tilmes et al. (2019) for all species based on low-NOx VBS yields only. This is
269 a similar SOA scheme as used in GC12-DYN (with differences in the treatment of
270 isoprene-SOA based on Marais et al. 2016 in GC12-DYN, and the use of both low- and
271 high-NOx VBS yields in GC12-DYN). Organic gases and aerosols undergo dry and wet
272 deposition as described in Liu et al. (2012). It should be noted that CESM2-SMP does not
273 include deposition of intermediate organic vapors. Aerosol wet scavenging considers in-
274 cloud scavenging (the removal of cloud-borne particles that were activated at the cloud
275 base) and below-cloud scavenging for both convective and grid-scale clouds.

276 CESM1-CARMA simulations use the configuration described in Yu et al. (2019) which is
277 based on CESM1 and the sectional Community Aerosol and Radiation Model for
278 Atmospheres (CARMA v3.0). Anthropogenic emissions are those from the Greenhouse
279 gas-Air pollution Interactions and Synergies (GAINS) model, and biomass burning
280 emissions are from the Global Fire Emission Database (GFED v3, van der Werf et al.,
281 2010). In CARMA, 20 size bins are used for both pure sulfate particles (bins from 0.2 nm
282 to 1.3 µm in radius, only used up to 500 nm) and mixed aerosols composed of BC, primary
283 and secondary OC, dust, sea salt, and sea-spray sulfate (bins from 0.05–8.7 µm in radius,
284 again, only analyzed up to 500 nm). SOA formation is based on the VBS approach from
285 Pye et al. (2010). The removal of OA occurs only by dry and wet deposition. Compared to
286 the CESM2-SMP and CESM2-DYN simulations, the convective removal of aerosols uses
287 the modified scheme described in Yu et al. (2019) which accounts for aerosol secondary
288 activation from the entrained air above the cloud base, and the scavenging of activated
289 aerosols in convective updrafts. The default CESM can transport aerosols from the cloud
290 base to the top of the cloud in strong convective updrafts in one time step without
291 scavenging them, while the new scheme allows for a more efficient removal off all aerosols
292 inside convective clouds. A sensitivity simulation is performed for ATom-1 to quantify the
293 effect of this improved removal on OA concentrations (Section 4.5).

294 *2.2 AeroCom-II model climatology*

295 The ATom measurements are also compared to the global model OA predictions
296 generated within the Phase II Aerosol Comparisons between Observations and Models
297 (AeroCom-II) project (Schulz et al., 2009). We consider the monthly average results of 28

298 global models, which is a subset of those presented in Tsigaridis et al. (2014), based on
299 the availability of model results. It should be noted that the meteorological forcing used in
300 these models is mostly based on the year 2006, while the anthropogenic and biomass
301 burning emissions are mostly representative of the year 2000. For comparison purposes,
302 the monthly mean model outputs for the months of August (ATom-1) and February (ATom-
303 2) are interpolated along the flight path (latitude, longitude, and altitude), and averaged
304 the same way as the measurements (see section 3.2).

3   Description of ATom measurements

3.1   Submicron aerosol data

307 The measurements of non-refractory submicron aerosols were performed onboard the
308 NASA DC8 aircraft as part of the ATom field study (Wofsy et al., 2018) using the University
309 of Colorado Aerodyne High-Resolution Time-of-Flight Aerosol Mass Spectrometer (AMS
310 in the following, Canagaratna et al., 2007, DeCarlo et al., 2006).

311 We use measurements from both the NH summer (August 2016, ATom-1) and winter
312 (February 2017, ATom-2) deployments. Figure 2a shows the flight path and the vertical
313 extent of the ATom-1 dataset colored by OA mass concentrations (see Figure S1 for
314 ATom-2). The aircraft performed systematic vertical sampling with ~140 vertical profiles
315 per campaign throughout the troposphere from the near surface ~0.2 km to the upper
316 troposphere/lower stratosphere region at ~13 km altitude. Details on the operation of the
317 CU AMS on board the DC-8 are reported in Schroder et al. (2018), Nault et al. (2018), and
318 Jimenez et al. (2019b). AMS data was acquired at 1 Hz time resolution and independently
319 processed and reported at both 1 s and 60 s time resolutions (Jimenez et al., 2019a). The
320 later product, with more robust peak fitting at low concentrations was exclusively used as
321 the primary dataset in this work. Detection limits at different time resolutions/geographical
322 bins relevant to this study are discussed in Section 3.3. The overall 2σ accuracies of the
323 AMS measurement (38% for OA, 34% for sulfate and other inorganics) are discussed in
324 Bahreini et al. (2008) and Jimenez et al. (2019b).

325 For ATom, the AMS reported the standard non-refractory aerosol species OA, sulfate,
326 nitrate, ammonium, and chloride, with the response for all the nominally inorganic species
327 characterized by in-field calibrations. In addition, it also reported methanesulfonic acid
328 (MSA, Hodshire et al., 2019a describes the AMS MSA methods and calibrations for ATom)
329 and sea salt for for D_{geo}<450 nm (based on the method of Ovadnevaite et al., 2012). Both

330   of these species were important to achieve closure with the volume calculated from the
331   on-board sizing instruments in the marine boundary layer (Jimenez et al., 2019b). Another
332   important refractory submicron species not captured by the AMS measurements is BC.
333   This was measured on ATom with the NOAA SP2 instrument (Katich et al., 2018). It should
334   be noted that aircraft measurements of aerosol mass concentrations are given in μg sm^{-3}
335   (i.e., under standard conditions of 1 atm and 273.15 K).

336   For ATom the AMS measured particles with geometric diameters (based on the campaign-
337   wide average density of 1640 kg m^{-3}, Jimenez et al., 2019b) of between D_{geo}~60 and 295
338   nm with ~100% efficiency (and between 35 and 460 nm with 50% efficiency). Here we
339   denote the AMS aerosol data as "submicron" mass (based on the more usual definition
340   using aerodynamic diameter, which is larger than the geometric diameter; DeCarlo et al.,
341   2004), with the assumption that non-refractory aerosol are small contributors to mass
342   above the AMS size range. As shown in Brock et al., 2019, the accumulation mode for the
343   ATom sampling environment only extended up to 500 nm, and hence, as expected for a
344   background tropospheric environment, this approximation is appropriate. Very good
345   agreement was observed with the integrated volume calculated from the number size
346   distributions for ATom (Brock et al., 2019). A low bias compared to a typical submicron
347   definition can occur in thick biomass burning plumes and in the lower stratosphere at times
348   (Jimenez et al., 2019b). As detailed in Table 1, the accumulation mode for the bulk models
349   discussed in this study overlaps with the size range of the AMS, and for the sectional
350   models (CESM1-CARMA, GEOS-Chem-TOMAS, ECHAM6-HAM) only the bins that
351   match the AMS size range were used. As expected based on the previous discussion,
352   however, a comparison of the total OA calculated by these sectional models with the
353   modeled OA inside the AMS size-range showed small differences (Slopes for ATom-1
354   linear regressions: CESM1-CARMA:0.91, GC10-TOMAS: 0.94, ECHAM6-HAM 1.00)
355   mostly influenced by the high concentration points in the biomass plumes off Africa that
356   have a large effect on the regression since they are about 10 times larger than the bulk of
357   the dataset).

358   Refractory and non-refractory aerosol composition was also measured using the NOAA
359   Particle Analysis by Laser Mass Spectrometry (PALMS) instrument. PALMS classifies
360   individual aerosol particles into compositional classes including biomass burning (Hudson
361   et al., 2004), sea salt (Murphy et al., 2019), mineral dust (Froyd et al., 2019), and
362   others. Mass concentrations for these particles types are derived by combining PALMS

composition data with aerosol size distribution measurements (Froyd et al., 2019). Good agreement overall was found for OA, sulfate and seasalt between the two particle mass spectrometers during ATom once the AMS and PALMS instrument transmissions were accounted for (Jimenez et al., 2019b). For all PALMS data used in this work (biomass burning fraction and dust) the AMS transmission function was applied to ensure that both instruments were characterizing approximately the same particle range.

For a particular airmass, the mass fraction of biomass burning (BB) aerosol reported by the PALMS instrument f(BB)$_{PALMS}$ (Thompson and Murphy, 2000; Froyd et al., 2019) was then used to evaluate the degree of BB influence. This parameter correlates quite well with other gas-phase BB tracers, and is more useful as a particle tracer since its lifetime follows that of the particles. Importantly, it is not impacted by the long lifetimes of the gas-phase tracers (e.g. 9 months for CH_3CN) and unrelated removal processes (e.g. ocean uptake for CH_3CN and HCN) that result in highly variable backgrounds. Hence f(BB)$_{PALMS}$ has a much higher contrast ratio and linearity for particle BB impacts, compared to the available gas-phase tracers in the ATom dataset. An airmass was classified as non-BB influenced when f(BB)$_{PALMS}$ was lower than 0.30 (Hudson et al, 2004) as shown in Figure 2b. For both ATom-1 and 2, about 74% of measurements were classified as not influenced by biomass burning. f(BB)$_{PALMS}$ was also used to assess the impact of POA on the total OA burden (next section); note that no thresholding was applied in that case.

3.2   Estimation of the POA fraction for the ATom dataset

For model evaluation purposes, it is important to know whether the source of OA is primary or secondary. For ground studies close to sources (e.g. Jimenez et al., 2009) Positive Matrix Factorization of AMS mass spectra (PMF, Ulbrich et al., 2009) can be used to estimate the contribution of primary sources (mostly from transportation, heating, cooking, and biomass burning) to total OA. This approach is not suitable for ATom. To accurately resolve a minor factor such as POA in an AMS dataset, there needs to be a combination of: (a) Sufficient OA mass concentration, so that the signal-to-noise of the spectra is sufficient; (b) Enough fractional mass for the factor to be resolved (>5% in urban areas per Ulbrich et al. (2009), probably a larger fraction at low concentrations such as in ATom); (c) Sufficient spatio-temporal variability ("contrast") in the relative contributions of different factors, since that is part of what PMF uses to extract the factors; (d) Sufficient difference in the spectra of the different factors (for the same reason as (c)), and (e) relatively invariant spectra for each factor across the dataset (as this is a key assumption of the

396   PMF algorithm). As an example of a near ideal case, in Hodshire et al (2019) we extracted

397   MSA by PMF from the ATom-1 data, and were able to match that factor with our

398   independently calibrated MSA species. A very distinct and nearly invariant mass spectrum

399   was measured repeatedly near sources (MBL) (and was mostly absent elsewhere, thus

400   providing strong spatio-temporal contrast) and accounted for about 6% of the fractional

401   mass and 15% of the variance in time. Thus all the conditions were met. For POA, on the

402   other hand, the air sampled in ATom and coming from e.g. Asia has POA and SOA very

403   well mixed, with little change on their relative mass fractions vs. time (as the aircraft flies

404   through that airmass). POA is very low, as documented later in this paper. Atmospheric

405   aging makes the spectra from all OA sources more and more similar as measured by AMS

406   spectra (Jimenez et al., 2009). Thus most of the conditions above are not satisfied for

407   extracting POA by PMF analysis of this dataset.

408   Instead, in this work we have estimated POA based on the fact that it is co-emitted with

409   BC as part of the combustion processes releasing both species in source regions, and

410   that BC is not impacted by chemical aging processes over the lifetime of the airmass. Note

411   that BC can physically age but it is not lost in any significant amount to the gas-phase due

412   to chemical processes in the atmosphere. We assume non-differential removal (and

413   transport) of the BC fraction relative to the rest of the POA (the two are generally internally

414   mixed, Lee et al., 2015). Table S1 summarizes recent POA/BC and POC/EC emission

415   ratio determinations for urban background sites, which best represent real mixes of

416   pollution sources, and for individual sources of POA (from mobile sources – commonly

417   referred as HOA – and cooking aerosol – COA). Based on Table S1 data, we assume

418   POA to be co-emitted with BC for anthropogenic fossil fuel / urban region POA (herein

419   called FF_{ratio} for simplicity, even though much of it is non-fossil, Zotter et al., 2014; Hayes

420   et al., 2015) at a ratio of 1.5 (average of all urban ambient air studies that report POA and

421   BC for best intercomparability to the ATom dataset; including all urban studies results in

422   a very similar number, 1.48). Mobile source measurements in general exhibit lower ratios

423   (POA/OA ratio 0.5-1.5) while COA determination typically ranges from 2 to 3. Hence, the

424   ratio used here is a good estimate for a diverse mix of urban sources as appropriate for

425   ATom. The studies used to derive the emission ratio used ambient data in urban air, where

426   all sources mix together and impact the POA/BC ratio, and thus the ratios include the

427   impact of POA sources that may not emit BC. It should be noted that urban model ratios

428   do not include emissions associated with fugitive dust from road, tire and construction, as

429   those are typically found in larger particles than those studied here (Zhao et al., 2017).

430 For biomass burning sources, we use a value of POA/BC = 11.8 (BB$_{ratio}$), based on the
431 average of the recent review by Andreae (2019), which included over 200 previous
432 determinations for a variety of fuels and burning conditions (since Andreae (2019) used
433 and OA/OC ratio of 1.6 in his work, we have used that value to calculate POA/BC; we note
434 that this is different from the 1.8 OA/OC ratio used for other studies listed in Table S1).
435 We note the measured total OA/BC of ~3.5 (conservatively assuming that all OA is POA)
436 observed on both ATom missions for the large African-sourced BB plumes over the
437 Equatorial Atlantic. We note that using the larger BB$_{ratio}$ from Andreae (2019) leads to a
438 POA fraction >> 100% in the ATom African plumes. We also perform sensitivity studies
439 with values of both FF$_{ratio}$ and BB$_{ratio}$ within the literature range.

440 The PALMS determined mass fraction of biomass impacted aerosol (f(BB)$_{PALMS}$) can then
441 be used to determine a total POA contribution from both types of sources:

442 $$POA = [BC]*(FF_{ratio}+(BB_{ratio}- FF_{ratio})*f(BB)_{PALMS}) \qquad \text{(Eq. 1)}$$

443 Further detail is provided in Table S2, which summarizes the POA/BC ratios used in the
444 emission inventories implemented in current models. Overall, there is reasonable
445 agreement with the measurements in Table S1, with FF$_{ratio}$ ranging from ~0.5 for diesel
446 fuels, to >2 for energy production and ~5 for residential emissions (which include some
447 BB). On the other hand, for biomass burning, the emission inventories ratios range from
448 ~5 for crop, to ~15 for forest, and up to ~50 for peatland. While generally consistent with
449 the values discussed by Andreae (2019), they are on the lower end of the ranges
450 discussed in that work. The averages and ranges of the measurement and model ratios
451 are similar, and thus no significant model bias on the ratios is apparent.

452 PALMS detection efficiency increases with size across the accumulation mode, and
453 therefore the f(BB) number fraction is weighted to the larger size end of the accumulation
454 mode. In very clean regions of the upper troposphere (typically <0.15 µg sm^{-3} submicron
455 mass) particles below the PALMS size range can contribute significantly to aerosol mass
456 (Williamson et al., 2019; Jimenez et al., 2019b). If BB particles are not evenly distributed
457 across the entire accumulation mode (due to preferential removal in convective updrafts
458 of primary aerosol, cf. Yu et al., 2019 and Section 4.5; and preferential condensation of
459 SOA on smaller particles), then the f(BB) reported by PALMS will be an overestimation.
460 For the final analysis these periods where left in the dataset, and therefore for the LS the
461 reported POA is likely overestimated for these regions, although their impact on the mass-
462 weighted campaign average is negligible.

463   The contribution of POA from sea spray is difficult to constrain. As an order-of-magnitude

464   estimate, marine POA is roughly calculated based on preliminary calibrations of OA on

465   mineral dust particles from the PALMS instrument (personal communication K. Froyd).

466   Using this calibration, the average OA by mass on sea salt was <10% for the large majority

467   of MBL sampling (>85%). Since sea salt contributed 4% (11%) of mass in the AMS size

468   range for ATom-1(2) (Figure 2), we estimate that marine POA is on the order of ~1% of

469   aerosol mass in the AMS size range, and possibly much lower. Thus we think that it is

470   reasonable to neglect the contribution of marine POA to this dataset. Future studies will

471   refine this estimate.

472   ### 3.3   Data processing for comparisons

473   For the comparisons between the measurements and the various global models, data

474   were averaged both vertically and zonally to minimize the impact of smaller plumes or

475   vertical gradients in aerosol concentrations that might not be captured by coarse resolution

476   models. For the same reason, all data near airports was removed from the datasets prior

477   to analysis (up to about 3 km on the climb in/out). In order to restrict this analysis to the

478   remote troposphere, the last leg of the ATom-1 mission (over the continental US) was

479   taken out of the dataset as well. Data was binned into 5 large latitude regions as shown in

480   Figure 2a including southern polar (55-80°S, "S.Polar"), southern mid-latitudes (25-55°S,

481   "S.Mid"), equatorial (25°S-25°N, "Equatorial"), northern mid-latitudes (25-55°N, "N.Mid"),

482   northern polar (55-80°N, "N.Polar") and analyzed separately for the Pacific and Atlantic

483   basins. For data in each of these latitude regions, altitude profiles were calculated with a

484   constant 600 m altitude resolution. According to both variability in the cleanest air and

485   statistical analysis of the organic background subtraction (Drewnick et al. 2009), the 1σ

486   precision at low concentrations for one-minute data ranged between 20 and 50 ng sm^{-3},

487   or a 3σ detection limit between 60 and 150 ng sm^{-3} for the one-minute data (confirmed by

488   frequent filter blanks). Per standard statistics, the precision of a measurement decreases

489   (i.e., gets better) with the square root of the number of points (or time interval) sampled.

490   I.e. the precision of an average can be approximated by the standard error of the mean

491   (σ/sqrt(n), where n is the number of measurements averaged), and it is better than the

492   precision of the individual data points (σ). This also applies to the detection limit, since it

493   is just 3 times the precision. Note that a detection limit is not meaningful unless the

494   averaging time is specified. For example let's assume that the detection limit is 20 ng m^{-3}

495   (1-second), and the data points over 60 consecutive seconds are all 10 ng m^{-3}. All 1-

second measurements are below the 1-second DL. However the average (10 ng m^{-3}) is now above the DL for 1-minute averages, which is 20/sqrt(60) = 2.6 ng m^{-3}. On average, each individual point in the profiles represents the average of about 25 min of ATom flight data. At that time resolution, the OA 1σ precision was about 10 ng sm^{-3}. Hence with very few exceptions (10 points for both missions combined), the OA concentrations in the averaged profiles reported are well above the instrumental detection limit in those regions. For model-measurement comparisons along flight tracks, model outputs and measurements were considered at 1-minute time resolution, which corresponds to ~0-700 m vertical resolution and ~0.05-0.15 degrees horizontal resolution. Note that a large fraction of the 1-minute OA values in the remote free troposphere were below the local 3σ detection limit. The data of periods of zero concentration (sampling ambient air through a particle filter) do average to zero. Some negative measurements are present, and this is normal for measurements of very low concentrations in the presence of instrumental noise. Averaging of longer periods, as done for the figures in this paper, reduces the detection limit. We therefore caution future data users that the reported data should be averaged as needed, as replacing below-detection limit (or negative) values by other values introduces biases on averages. For fractional ratio analysis, measurements were averaged to 5-minute time resolution to reduce the noise in the ratios due to noise in the denominator. The results are not very sensitive to the 5-minute averaging (compared to 1-minute) as shown in Figure S12 for OA to sulfate ratios. The same figure also illustrates that excluding ratios affected by negative concentrations (the non-bracketed case, overall these are about 15% of the dataset) does not really affect the fractional distribution, with the variance between the two cases diminishing as the averaging interval increases. To further confirm that there is no inherent bias in the fractional products regardless of the treatment of low concentration values, an additional sensitivity analysis was performed where data was filtered by an independent measurement proxy for aerosol mass, the aerosol volume measured in ATom (Brock et al., 2019). Using a range of value that encompasses the regime where the AMS calculated volume to aerosol measured volume exhibited increased noise (Jimenez et al., 2019b), no systematic bias was found (Figure S13), with variations of about 10% in fractional volume for different filtering conditions.

Some of the performed analysis required separating the dataset into vertical subsets. In this manuscript, we define the marine boundary layer (MBL) as the region below 1.5 times the calculated boundary layer height in the NCEP global model reanalysis. The free troposphere (FT) includes all data points between the top of MBL and the NCEP

tropopause height, and the LS region includes all points above the NCEP tropopause height. The tropopause height varied during ATom between 8 and 16.5 km; given the DC-8 ceiling (42 kft, 12.8 km) the stratosphere was only sampled at latitudes higher than 30 degrees in both hemispheres. The MBL height varied between up to 1.5 km in the mid-latitudes, ~1 km in the tropics, and sometimes <150 m (lowest DC-8 altitude) for some of the sampling in the polar troposphere.

3.4 Submicron aerosol composition

Figure 2b shows that during both NH summer and winter ATom deployments, OA is one of the three dominant components of the measured submicron aerosol in the remote troposphere, together with sulfate and sea salt. During ATom-1, average submicron aerosol concentrations were close to 0.8 μg sm^{-3} in the marine boundary layer and biomass burning outflow regions, and ~2 times lower in the free troposphere and lower stratosphere regions. ATom-2 had overall lower average concentrations below 0.4 μg sm^{-3} (vs. 0.5 μg sm^{-3} for ATom-1). As expected, sulfate (sulfuric acid in the lower stratosphere) is the dominant constituent in the MBL (~50%) and LS (50-70%), while the OA contribution is generally below 10% and 40%, respectively in those regions. A large fraction of sea salt aerosol is found in the MBL especially during the NH winter deployment (~30%, see Murphy et al., 2019).

OA is found to be a major constituent (~50%) of submicron aerosol in the clean (non-BB influenced) free troposphere. The contribution of OA is 1.4 times larger than that of sulfate during the NH summer, and 1.2 times lower than that of sulfate during the NH winter, which is likely due to a large contribution of the NH sources to SOA production in the NH summer. Biomass-burning events increase the OA contribution relative to that of sulfate, and lead to a higher contribution of OA to total during the ATom-1 mission (stronger BB influence).

3.5 Spatial and vertical distribution of OA

Figure 2a (and Fig. S1) shows the spatial and vertical distribution of OA mass concentrations measured during ATom-1 (and ATom-2) campaigns. Most data were taken over remote oceanic regions (and a few remote continental regions, primarily over the Arctic). The measured OA varies between extremely clean conditions (< 0.1 μg sm^{-3}) encountered mostly in the Pacific and Southern Ocean regions and moderately polluted conditions (> 2 μg sm^{-3}) in the biomass burning outflow regions. During ATom-1 (August

562 2016), a strong BB influence is observed in the lower troposphere (below 6 km) over the

563 Atlantic basin off the African coast and over California with OA concentrations exceeding

564 10 µg sm^{-3}. OA associated with biomass burning is also present in the upper troposphere

565 over equatorial regions and over Alaska, associated with the deep convective transport of

566 biomass burning aerosols. The biomass burning contribution to carbonaceous aerosols in

567 those regions during ATom-1 was also apparent in the black carbon measurements

568 (Katich et al., 2019). ATom-2 was generally less polluted than ATom-1, likely due to a

569 more limited global influence of biomass burning emissions during that period, and also to

570 a less active photochemistry during winter months in the NH.

571 The measured OA is characterized by a strong latitudinal gradient. Figure 2c shows the

572 average vertical profiles of measured OA over the selected latitudinal bands during August

573 2016. The cleanest airmasses are observed over the remote oceanic regions of the

574 Southern Hemisphere (SH, 25-80°S) with OA mass concentrations below 0.06 µg sm^{-3}.

575 These extremely low OA concentrations can be explained by the very low influence from

576 continental emission sources, and presumably low marine POA and SOA precursor

577 emissions. This is consistent with low concentrations of gas-phase pollutants (e.g. CO,

578 ethane, propane). An enhancement can be noticed above 10 km in the lower stratosphere.

579 In some cases, this could be related to the long-range transport of biomass burning

580 aerosols from the tropics. By comparison, the Arctic region is more polluted with an order

581 of magnitude higher OA levels compared to its analog of the SH (i.e. OA loadings ranging

582 from 0.1 to 0.5 µg sm^{-3}). These concentrations are comparable to FT levels measured in

583 the extratropical regions (25-55°N) of the NH. The equatorial marine regions (25°S-25°N)

584 display the highest OA concentrations with a strong gradient between lower and upper

585 troposphere. In the lower troposphere OA, concentrations are close to 1 µg sm^{-3}, and

586 decrease down to 0.1 µg sm^{-3} at altitudes above 4km. The highest OA levels are

587 associated with the African outflow over the southeastern Atlantic Ocean, which results

588 from the transport of the biomass burning smoke from the sub-Saharan regions and

589 increasing urban and industrial air pollution in southern West Africa (Flamant et al., 2018).

590 Figure 2d shows that the Atlantic basin is often more polluted than the Pacific basin, not

591 only because of the African biomass burning influence but also due to the contribution of

592 anthropogenic pollution in the lower troposphere of the NH. It should be noted that Asian

593 pollution was likely an important contributor to the North Pacific Basin, especially between

594 2 and 6 km, in both ATom deployments (see figures 2a and S1). Several-fold higher OA

concentrations are found near the surface (below 1km) over the southern Pacific compared to that same location in the southern Atlantic, which could be indicative of the stronger emission of marine OA in the Pacific basin.

In addition to spatial gradients, a strong summer-to-winter contrast is observed in OA concentrations. Figure 2e shows the ratio between OA vertical profiles measured in the NH summer ATom-1 vs. in the NH winter ATom-2. The NH is more polluted during the NH summer due to the photochemical production of SOA, as well as biomass burning emissions, leading to the tripling of OA concentrations in the extratropical regions (25-80°N) on average regardless of altitude. The doubling of OA loading in the lower troposphere at the equator (25°S-25°N) in the NH summer (August, ATom-1) is strongly influenced by the biomass burning activity in the sub-Saharan African region as already mentioned above. Likewise, OA concentrations are found to be generally higher in the SH during the SH summer. These zonal trends are broadly similar to the ones described in Katich et al (2018) for BC.

4   Model-measurement comparisons

4.1   Evaluation of predicted OA concentrations

Prior to evaluating model performance in simulating OA, we have assessed the ATom models' ability to simulate sulfate aerosols. According to the model evaluation shown in Table S3, the predicted sulfate concentrations are generally within 40% of the measured values, which is comparable to the AMS measurement uncertainties. The only exception is found for the ECHAM6-HAM model, which overestimates sulfate aerosols by a factor of two. These results imply that most ATom models capture relatively well the overall sulfate burden. However, large root mean square error (RMSE > 0.4 μg sm^{-3} for ATom-1 and > 0.2 μg sm^{-3} for ATom-2) is indicative of their limited skill in reproducing the observed variability in sulfate concentrations.

For OA, model evaluation metrics for the entire ATom-1 and ATom-2 campaigns are given in Table 2 for the eight ATom models and their ensemble, as well as the AeroCom-II ensemble. The results show that the normalized mean bias is substantially lower for the ATom model ensemble compared to AeroCom-II decreasing from 74% to 4% for ATom-1 and from 137% to 23% to ATom-2, which is within the measurement uncertainty range. The mean temporal correlations are substantially improved from 0.31 (0.38) for AeroCom-II to 0.66 (0.48) for ATom model ensemble during ATom-1 (ATom-2). However, results

627 vary strongly among ATom models. Models using prescribed emissions of non-volatile

628 SOA have the tendency to overestimate the OA concentrations during both NH summer

629 and winter deployments (with ~35-60% overestimation for CESM2-SMP, ~70-100% for

630 ECHAM6-HAM, and up to 150% for GC10-TOMAS during ATom-2), with the exception of

631 the GEOS5 model that on the contrary underestimates OA concentrations by 5-25%.

632 During the NH summer (ATom-1), models using the VBS parameterization from Pye et al.

633 (2010) tend to underpredict the OA concentrations by 43% for GC12-REF and 33% for

634 CESM1-CARMA for ATom-1, most likely due to the excessive evaporation of the formed

635 SOA in remote regions and low yields for anthropogenic SOA (Schroder et al., 2018; Shah

636 et al., 2019). Models using the VBS parameterization from Hodzic et al. (2016) (CESM2-

637 DYN and GC12-DYN) where OA is less volatile and also OA yields are corrected for wall

638 losses show an improved agreement with observations especially for CESM2-DYN (with

639 NMB of ~5%), and to a lesser extent for GC12-DYN (NMB of ~33%). During the NH winter

640 (ATom-2) characterized by a lower production of SOA, both VBS approaches lead to an

641 overestimation of the predicted OA. This is likely caused by excessively high levels of

642 primary emitted OA as discussed in section 4.4.

643 Figure 3 compares the average median ratios between modeled and observed OA

644 concentrations for the ATom and AeroCom-II model ensembles for different regions (BB,

645 MBL, FT, LS). The results show that the median ratio for the ATom model ensemble is

646 close to unity in all regions. This is at least a factor of two improvement compared to

647 AeroCom-II models, which were almost always biased high for the remote regions

648 sampled in ATom. The model spread has also been reduced by a factor of 2-3 in all

649 regions. This reduction in the ensemble spread may partially be explained by a smaller

650 size of the ATom model ensemble (see Fig. S2), which also includes models with a more

651 up-to-date OA representation. In order to explore this point further, results for a subset of

652 AeroCom-II models (using earlier versions of models in the ATom ensemble) show only a

653 slight reduction (~10%) in the model spread, with however some regional differences i.e.

654 an improved agreement with observations in the MBL, but an increase in the model bias

655 and spread in the LS (Figure S2). Thus, model improvement for the more recent models

656 appears to be the main reason for the reduced spread.

657 *4.2   Evaluation of predicted OA vertical distribution*

658 Figure 4 compares the mean vertical profiles of OA measured during ATom-1 and -2 with

659 the predictions of the model ensemble average based on the eight ATom models (Table

660 1) and 28 AeroCom-II models for the different latitudinal regions of the Pacific and Atlantic
661 basins. Note that the use of a wide logarithmic scale (to be able to span all the
662 observations) may make the observed differences appear small, although they often reach
663 factors of 2-10 and larger (Figure S5 shows the results on a linear scale). For AeroCom-
664 II, large latitudinal differences exist in the results with a better performance closer to source
665 regions and large disagreement in the lower stratosphere and remote regions, as already
666 suggested by the mission medians shown in Figure 3. The best AeroCom-II model
667 performance is found over the equator in both basins, where the model ensemble captures
668 within a factor of 2 the observed OA concentrations throughout the troposphere in the
669 Pacific basin, and matches remarkably well the observations in the lower troposphere of
670 the Atlantic basin that is heavily influenced by biomass burning emissions. Reasonable
671 agreement is found for the OA vertical distribution over the NH Atlantic and Pacific oceans,
672 especially in the lower troposphere (< 4 km). The largest model discrepancies (1-2 orders
673 of magnitude) are found in the remote regions of the Southern Ocean and SH mid-latitudes
674 during both seasons and basins. The model overestimation is also large over the NH mid-
675 latitude Pacific basin in the upper troposphere. A spread of 2-3 orders of magnitude is
676 observed around the ensemble average indicating a very large variability in individual
677 model predictions. This evaluation of AeroCom-II models in remote regions is an extension
678 of that performed at the surface for urban and remote stations by Tsigaridis et al. (2014)
679 (as in that previous study, the data and model simulations compared are not synchronous
680 in time). The tendency of the model ensemble to overpredict OA concentrations by a factor
681 of 2 on average in the remote regions is consistent with the transition from the large
682 underprediction in OA near the source region to a slight overprediction of OA in remote
683 continental sites that was reported for most AeroCom-II models (Tsigaridis et al., 2014),
684 and also observed for default parameterizations in other studies (Heald et al., 2011;
685 Hodzic et al., 2016).

686 By comparison, the results of the ATom model ensemble show a much better agreement
687 with observations. The model spread is still substantial, but mostly below a factor of 5.
688 Figures S6 and S7 show OA vertical profiles for individual ATom models and the spread
689 in their results. In most regions, the ATom model ensemble captures reasonably well both
690 the absolute concentrations as well as the shape of the vertical profiles. In the biomass
691 burning outflow and NH mid-latitude regions, the ATom ensemble average better captures
692 the higher OA concentrations in the boundary layer and lower OA concentrations in the
693 lower stratosphere than the AeroCom-II ensemble. We note that using the ensemble

694   median OA profiles instead of ensemble mean OA profiles (as shown in Figure 5 and S7)
695   results in a slightly lower values of OA but does not change the conclusions of the model-
696   measurement comparisons (Figure S18).

697   ### *4.3   Oxidation level of organic aerosols (OA/OC ratios)*

698   In addition to OA mass concentrations, we also evaluate the model's ability to simulate
699   their degree of oxygenation, an indicator of their oxidation and aging (Aiken et al., 2008;
700   Kroll et al., 2011). Ambient measurements of the oxidation level of organic particles are
701   limited (Aiken et al., 2008, Canagaratna et al., 2015), and the ATom dataset provides the
702   first global distribution of O/C and OA/OC ratios for the remote aerosol. The OA/OC ratio
703   is an estimate of the average molecular weight of organic matter per carbon weight, and
704   it mostly depends on the oxygen content (i.e. the O/C ratio), in the absence of significant
705   concentrations of organonitrates and -sulfates. It is needed to compare measurements of
706   organic aerosol mass (from e.g. AMS) with organic carbon measurements (from e.g.
707   thermooptical methods). It is also needed to compare the various types of measurements
708   to model concentrations, which are sometimes carried internally as OA and sometimes as
709   OC. A low OA/OC ratio is indicative of freshly emitted OA from fossil fuel combustion
710   (typically ~1.4), and its value increases with increased processing of organics in the
711   atmosphere. Figure 5 shows that in the remote regions the bulk of measured OA/OC
712   ratios during ATom-1 and -2 range between 2.2 and 2.5, and are larger than values of 2.1
713   $\pm$ 0.2 found in the polluted US continental outflow regions that were sampled during
714   SEAC4RS, WINTER and DC3 field campaigns (Schroder et al., 2018). These values
715   indicate that remote OA is highly oxidized and chemically processed.

716   Note that for organosulfates (R-O-SO$_2$H and organonitrates (R-O-NO$_2$, pRONO$_2$ in the
717   following) only one oxygen is included in the reported OA/OC, as the fragments of these
718   species are typically the same as for inorganic species in the AMS (Farmer et al., 2010).
719   However in ATom organosulfates are estimated to account for ~1% of the total sulfate
720   (based on PALMS data, see Liao et al., 2015 for the methodology). Since sulfate and OA
721   concentrations are comparable, organosulfates would only increase the OA/OC by ~1%
722   on average. Organonitrates are reported from the AMS for ATom. Their impact on OA/OC
723   is not propagated for the default values, to maintain consistency with a large set of OA/OC
724   measurements by AMS in the literature, and since they would increase OA/OC on average
725   by only 4.5% (ATom-1) and 2.2% (ATom-2), which is smaller than the uncertainty of this

726   measurement. However, we show the results with both methods in Fig. 5 to fully document
727   this topic.

728   Importantly, this ratio is also used to calculate the total OA mass concentration for models
729   that provided their outputs in terms of organic carbon concentrations ($[OA]_i = [OC]_i$ x
730   OA/OC_{ratio}). Most Models use a constant OA/OC ratio, but the value used varies
731   substantially. OA/OC of 1.4 is used in ECHAM6-HAM, whereas 1.8 is used in GEOS5 and
732   GC10-TOMAS simulations for both POA and SOA. Other models calculated directly SOA
733   concentrations without applying this conversion (CESM1-CARMA, CESM2-SMP, CESM2-
734   DYN, GC12-REF and GC12-DYN), but for POA used the ratio of 1.8 (CESM1-CARMA,
735   CESM2-SMP, CESM2-DYN) and 2.1 (GC12-REF and GC12-DYN). Most of the AeroCom-
736   II models used the ratio of 1.4 for all primary and secondary OA (Tsigaridis et al., 2014).
737   The comparison with measurements shows that the measured values are ~40% larger
738   than those assumed in some of the ATom models, and 60-80% larger than used in
739   AeroCom-II models. The comparison between the observed and predicted OA/OC vertical
740   profiles (Fig. S3) shows that AeroCom-II models tend to generally underpredict this ratio,
741   and do not capture its increase in remote regions. As a result, this underestimation of
742   OA/OC ratios and the use of a constant value could substantially impact the comparisons
743   of OA mass concentrations for several models considered in this study (ECHAM6-HAM,
744   GEOS5, CESM1-CARMA and GC10-TOMAS). If we correct for the underestimated
745   OA/OC ratio using the ATom measured values of 2.2 (to be conservative) and compare
746   to previously discussed biases in Table 2, the overprediction of the ECHAM6-HAM model
747   is increased to ~110-160%, and that of GC10-TOMAS to 180% during ATom-2 while
748   having ~15% bias in ATom-1, whereas GEOS5 results now overestimate up to 30% during
749   ATom1, and perform much better during ATom-2.

750   These results demonstrate that current global chemistry-climate models use unrealistically
751   low OA/OC ratios, which results in a large underestimate of the degree of oxidation of OA
752   in remote regions. Inaccurate prediction of OA oxidation as it ages could impact not only
753   the calculations of OA burden, but also its optical properties as the absorption of OA
754   changes with its degree of oxidation (through the formation and destruction of brown
755   carbon, Laskin et al., 2015, Forrister et al., 2015). However, models used in this study did
756   not include these effects.

4.4 Contribution of primary vs. secondary OA

We further assess whether global models can adequately predict the relative contributions of primary and secondary OA. We strive to quantify these fractions with the most straightforward methods (with the fewest assumptions) for both models and measurements. POA concentrations were estimated from the BC measurements by using an emission ratio appropriate to the airmass origin (biomass burning vs. anthropogenic), and using the f(BB) mass fraction from the PALMS single particle instrument (see Section 3.2). By using the POA/BC ratio at the source regions after most evaporation, but before POA chemical degradation evaporation has taken place, we implicitly assume POA to be chemically inert, while in reality it can slowly be lost to the gas-phase by heterogeneous chemistry (e.g. George and Abbatt, 2010; Palm et al., 2018). Thus, the observation-based method provides an upper limit to the fraction of POA. The model/measurement comparison is only shown for the CESM and GEOS-Chem model variants, as other participating models do not separate or did not report their POA and SOA fractions. In all simulations, POA was treated as a chemically inert directly emitted primary aerosol species that only undergoes transport, transformation from hydrophobic to hydrophilic state with ageing (1-2 days typically), coagulation, and dry and wet deposition. Importantly, the treatment of POA as non-volatile (rather than semi-volatile) in models is fully consistent with the assumptions for POA estimation from the measurements.

Figure 6 compares the vertical profiles of measurement-derived POA during ATom-1 and predicted by the CESM2-DYN model over clean remote regions of the Pacific basin and northern polar Atlantic that are not influenced by biomass burning. Comparisons for other models are similar (not shown). Observations show that POA is extremely small in remote regions, whereas the model predicts that about half of the OA is made of POA in those areas. Although the model reproduces quite well the measured total OA, it tends to severely overpredict the amount of POA and underpredict that of SOA over clean remote regions (with the two errors canceling each other when it comes to total OA). Over the biomass burning regions (not shown here) it can be difficult to directly quantify POA and SOA with this method, as total OA remains about constant, while POA decreases with aging and SOA increases (Cubison et al., 2011; Jolleys et al., 2015; Hodshire et al., 2019b). However, given this evolution the method used here would lead to an overestimate of POA for this reason.

789 A more general comparison is made in Figure 7, using the frequency distributions of the
790 measured and simulated fraction of POA/OA, for the free troposphere only (Figure S8
791 shows the corresponding cumulative distributions). Observations indicate that most
792 remote FT airmasses contain less than 10% POA, except for biomass burning plumes that
793 are considered mostly primary. A slightly higher proportion of POA is seen in ATom-2,
794 which is consistent with a slower photochemical production of SOA during NH winter.
795 These results indicate that the remote OA is consistently dominated by SOA regardless
796 of the season and location. The comparison with models reveals a very large discrepancy
797 in the predicted vs. measured POA vs. SOA contributions. Models have a general
798 tendency to severely overpredict the fraction of POA and underpredict that of SOA,
799 displaying a much wider frequency distribution than the measurements (as also shown for
800 POA and SOA vertical profiles for individual models on Figures S6 and S7). In GC12-REF,
801 CESM2-DYN and CESM1-CARMA (without improved in-cloud removal) predictions for
802 ATom-1, more than a half of the remote OA is POA, while that is very rarely observed in
803 the free troposphere (possibly only during strong biomass burning events). Most models
804 fail to reproduce the overwhelming dominance of SOA that is inferred from the
805 measurements during ATom-1, while the discrepancies are less severe during NH winter
806 (ATom-2). These seasonal differences suggest that model errors could be partially due to
807 inefficient production of SOA, although removal errors also probably play a major role (see
808 next section).

809 The differences are so large that they are pretty insensitive to details of the POA estimation
810 method from the measurements, mostly because for the vast majority of the ATom track
811 BC/OA ratios were extremely low and hence the exact magnitude of the multiplicative
812 factor is secondary to the estimation of POA (Figure S11). As Figure S9 illustrates, the
813 choice of FF_{ratio} has very little impact on the overall distribution of POA. On the other hand,
814 while the BB_{ratio} does impact the overall distribution of POA, it mostly affects the points in
815 the vicinity of the large Atlantic plumes. Since the POA/BC ratio in those plumes is fairly
816 low, (see Section 3.2), using a very large BB_{ratio} mostly leads to an increase of the fraction
817 of the points where POA > 100%. While the large range of published BB_{ratio} for different
818 sources precludes a more accurate estimation by our method, for the purposes of the
819 comparison with the model results we emphasize that even using the largest BB_{ratio},
820 fraction of SOA is still significantly larger in the ATom dataset that in any of the models.

821 Additional sensitivity tests were performed to investigate the impact of noisy data and
822 uncertainties of f(BB) on the estimation of POA. Figure S11 clearly shows that the impact
823 of a misattribution of the aerosol type by the stated PALMS uncertainty (Froyd et al., 2019)
824 is completely negligible. Figure S10 details how the choice of averaging interval (with
825 longer averaging times reducing both the fraction of OA measurements under the DL and
826 below zero) impact the distribution of POA. Overall, no large changes are observed for
827 averaging times >5 min, and hence a 5 min averaging interval was used for the analysis
828 in Figure 7. Figure S10 also illustrates how capping the histogram impacts the POA
829 distribution. To capture the most realistic $f_{(POA)}$ distribution, the data in Fig 7 was capped
830 at the extremes (so $f_{(POA)}<0$ is taken as $f_{(POA)}=0$, and $f_{(POA)}>1$ is taken as $f_{(POA)}=1$). As Figure
831 S10 shows, data with $f_{(POA)}<0$ is almost exclusively due to very small (and always positive,
832 since BC cannot go negative) POA values being divided by small, negative noise in total
833 OA, and hence treating that fraction of the histogram as essentially fPOA~0 is justified.
834 On the other end of the distribution, data where POA is larger than OA is mostly due to
835 our average BB_{ratio} being larger than the one encountered in most of the BB plumes in
836 ATom. Choosing a lower BB_{ratio}, as Figures S9b and S9d illustrate, leads to $f_{(POA)}>1$
837 basically trending to zero, confirming our interpretation. This is a limitation of the dataset,
838 and it does not seem appropriate to remove these points, since some fraction are likely
839 dominated by POA. However, it shows that the POA estimation, especially for this part of
840 the distribution likely overstates the importance of POA.

841 A comparison between simulations that have the same treatment of POA, and only differ
842 in their chemistry and removal of SOA (e.g. CESM2-SMP vs. CESM2-DYN; GC12-REF
843 vs. GC12-DYN) indicate that a more complex SOA treatment does not always result in a
844 more accurate simulation of the primary/secondary character of OA, a result that was also
845 found in the AeroCom-II multi-model intercomparison (Tsigaridis et al., 2014).

846 Finally, we have examined whether the non-volatile treatment of POA in models could
847 lead to these unrealistically high POA fractions in the remote regions. Figure S16 shows
848 a comparison of POA vertical profiles as predicted by the GC12-REF simulations that use
849 non-volatile POA and a sensitivity simulation GC12-REF-SVPOA that uses semi-volatile
850 POA similar to the standard treatment in GEOS-Chem as described in Pai et al. (2020).
851 Note, however, that Pai et al. (2020) included marine POA emissions, used different
852 reanalysis meteorology, and a different model version (12.1.1 rather than 12.0.1 here), so
853 their resulting comparisons to ATom measurements are somewhat different than found

854 here for GC12-REF-SVPOA. The comparison indicates that the POA concentrations

855 increase substantially in most regions when the semi-volatile POA parameterization is

856 used. These results suggest that non-volatile treatment of POA is not responsible of the

857 model bias.

4.5 Sensitivity to OA formation and removal

859 In this section, we further investigate some of the possible reasons for the incorrect model

860 predictions of the relative contributions of POA and SOA in remote regions. Given the

861 tendency of models to underestimate OA close to anthropogenic source regions and

862 overestimate OA downwind in past studies (e.g. Heald et al., 2011; Tsigaridis et al., 2014;

863 Hodzic et al., 2016), in this section we investigate the sensitivity of OA to increasing

864 sources and increasing removals. We have performed two additional model simulations

865 to test the sensitivity of the POA/SOA fractions to uncertainties in the representation of (i)

866 wet scavenging, based on the CESM1-CARMA simulations in which we have removed

867 the improvements in the aerosol removal by the convective updrafts (Yu et al., 2019); and

868 of (ii) SOA formation based on the GC12-REF simulations in which we have replaced the

869 SOA formation VBS mechanism (Pye et al., 2010) by an updated VBS mechanism that

870 uses chamber wall-loss corrected SOA yields (Hodzic et al., 2016, same formation

871 scheme that is used in GC12-DYN and CESM2-DYN runs, but with removals kept identical

872 to GC12-REF). The results of these two sensitivity simulations are displayed on Figure 8,

873 which shows measured and predicted mass concentrations of OA, POA, SOA and sulfate

874 for ATom-1 as a function of the number of days since the air mass was processed through

875 convection. One should keep in mind that this is an averaged plot that included airmasses

876 from various regions and altitudes, and not a Lagrangian plot following the same airmass.

877 **Sensitivity to in-cloud scavenging in convective clouds.** Inefficient wet removal of

878 primary OA could contribute to the POA overprediction in global models, especially in the

879 tropics. Previous global model studies have reported two to three orders of magnitude

880 overestimation of primary carbonaceous species such as BC in the free troposphere when

881 the removal in the convective updrafts was not included (e.g. Schwarz et al., 2013, Yu et

882 al., 2019). A strong reduction due to convective removal is also expected for POA

883 concentrations, as POA is a primary species co-emitted with BC at the surface and

884 internally mixed with it (Lee et al., 2015), and that is typically coated by secondary

885 inorganics and organics over short timescales (Petters et al., 2006; Jiang et al., 2010;

886 Wang et al., 2010). Figures 7a and 8 compare the simulations of CESM1-CARMA with

887 and without improved convective in-cloud scavenging during ATom-1. The improved in-
888 cloud scavenging scheme considers aerosol activation into cloud droplets from entrained
889 air above the cloud base, which is more realistic and results in a more efficient removal of
890 aerosols in the upper troposphere by convection. E.g. a two order of magnitude reduction
891 in BC in the upper FT was reported by Yu et al. (2019), resulting in much improved
892 agreement with observations. Similar results were observed for sea salt aerosols in
893 Murphy et al. (2019). Figure 8 shows that all submicron aerosol species simulated in
894 CESM1-CARMA are strongly impacted by the in-cloud removal above the cloud base.
895 POA concentrations are reduced by an order of magnitude while sulfate is reduced by
896 30% leading in both cases to a much-improved agreement with observations. SOA is
897 reduced by ~30% as well, which leads to an underprediction of measured SOA
898 concentrations. The overall impact on OA concentrations is a significant reduction, which
899 leads to ~20% underestimation of OA in the aged remote air during ATom-1.

900 For the CESM2-DYN model that does not have improved in-cloud removal, the reasonable
901 agreement (within 20%) with the observed OA concentrations thus results from
902 coincidental error compensation between the overpredicted POA and underpredicted
903 SOA. The prescribed SOA formation and the artificial 50% adjustment of SOA emissions
904 based on Liu et al. (2012) in CESM2-SMP leads to an overestimation of observed SOA in
905 aged remote airmasses.

906 ***Sensitivity to SOA formation.*** In addition, we have also tested the sensitivity of the OA
907 composition to the choice of the SOA formation mechanism. Figure 8 compares the results
908 of the GC12-REF model that uses SOA formation yields derived from traditional chamber
909 experiments (Pye et al., 2010) and those corrected for loses of organic vapors onto
910 chamber walls as proposed in Hodzic et al. (2016). Previous studies have reported that
911 chamber wall losses could lead up to a factor of 4 underprediction of formed SOA (Zhang
912 et al., 2014; Krechmer et al., 2016). It should be noted that in both cases, isoprene-SOA
913 is formed in aqueous aerosols following Marais et al. (2016). The comparison shows a
914 factor of 3 increase in SOA concentrations when the updated SOA formation is considered
915 leading to a much better agreement with the observed SOA as well as the observed total
916 OA. GC12-REF predicts well the amount of POA and overpredicts somewhat the amount
917 of sulfate aerosols, which is expected as it already includes the improved aerosol removal
918 in convective updrafts (Wang et al., 2014). Figure S6 also shows that POA vertical
919 distribution is well captured in GEOS-Chem in most regions, except over the polar north

920   Pacific. It should be noted that these results are consistent with the POA/OA frequency

921   distribution shown in Figure 7 (the POA/OA ratio predicted by GC12-REF is larger than

922   the measured ratio, which is consistent with the fact that POA is about the right amount,

923   and OA is underpredicted in Figure 8).

924   These sensitivity simulations suggest that a stronger convective removal of POA and a

925   stronger production of SOA might be needed to correctly represent not only the total OA

926   concentrations but also its primary and secondary nature in remote free troposphere and

927   remote ocean regions. Accurate predictions of the OA concentration, composition, and

928   source contributions for the right reasons are key for accurately predicting their lifecycle

929   and radiative impacts. Only when there is confidence that the sources are accurately

930   predicted, we can have confidence in OA predictions for pre-industrial and future

931   conditions, as well as to evaluate PM mitigation strategies.

932   ***4.6   OA and sulfate relative contributions in FT***

933   Finally, we assess the model ability to predict relative amounts of OA and sulfate in the

934   free troposphere where they are the two major constituents of the submicron aerosol

935   (Figure 2b). Accurate predictions of their relative contributions are crucial to determine the

936   hygroscopicity of the submicron aerosol, and its ability to serve as a cloud condensation

937   nuclei (CCN) in the remote free troposphere (Carslaw et al., 2013; Brock et al., 2016).

938   Figure 9a compares the average measured relative fractions of sulfate (36%) and

939   carbonaceous aerosols (OA=59% and BC=5%) in the FT with those predicted by

940   individual models during ATom-1. The CESM2 models best reproduce the observed

941   relative contributions, with a slight underestimation of OA (57% instead of 59%) for

942   CESM2-DYN, and a slight overestimation of OA (63% instead of 59%) for CESM2-SMP.

943   GEOS5 has 15% more OA relative to sulfate than observed. All other models

944   underestimate both OA and BC relative fractions. For instance, in GC12-REF and -DYN,

945   both the BC and OA fractions are ~40% (relative) lower than observed.

946   Figure 9b shows the frequency distribution of observed and predicted fractions of OA

947   relative to sulfate during ATom-1 and -2 in the free troposphere. Most models fail to

948   reproduce the relatively uniform nature of the observed distributions during ATom-1, with

949   typically narrower model shapes around a preferred ratio. The NH summer measurements

950   indicate that OA > sulfate in ~55% of the samples (consistent with Fig. 2b), while models

951   generally tend to underestimate the relative OA contribution. In particular, GEOS-Chem

952 and ECHAM6-HAM tend to overestimate the relative contribution of sulfate. A better
953 agreement is found for GEOS5, CESM1-CARMA and CESM2-DYN, which follow more
954 closely the shape of the observed distribution. The comparisons also suggest that the
955 more complex SOA treatment of SOA formation and removal proposed by Hodzic et al.
956 (2016) in the same host model leads to an improved agreement with observations (e.g.
957 CESM2-DYN vs. CESM2-SMP; GC12-DYN vs. GC12-REF). It should be noted that
958 CESM2-SMP uses fixed SOA yields that were increased by 50% as suggested by Liu et
959 al. (2012), leading to an overestimation of the relative contribution of OA compared to that
960 of sulfate in the free troposphere. During the NH winter (ATom-2), measurements show a
961 somewhat higher proportion of sulfate aerosols (vs. ATom-1), which is consistent with a
962 slower production of SOA in the NH during winter and a reduced influence of biomass
963 burning. Similar conclusions are found for the evaluation of different models. It is worth
964 mentioning that the comparison performed for the whole ATom-1 and 2 dataset (not
965 shown) leads to similar results with even slightly stronger overestimation of the sulfate
966 relative contribution compared to OA.

967 The discrepancies between the observed and predicted composition of submicron aerosol
968 over remote regions can be quite large for other constituents as well. Figure 10 shows the
969 comparison of measured and predicted composition of the submicron aerosol over the
970 Southern Ocean (during the NH winter) where the disagreement in simulated sea salt,
971 nitrates, ammonium, and MSA often exceeds the contribution of OA. While the
972 observations show a more uniform distribution of non-marine aerosol with higher values
973 in the mid and upper troposphere, respectively, most models tend to simulate highest
974 fractions of OA (and sulfate) towards the tropopause. This may also be explained by the
975 uncertainties in modeled wet removal of aerosol that has been discussed above. Specific
976 studies have discussed and continue to investigate the ATom measurements and
977 simulations of different components in more detail, including particle number (Williamson
978 et al., 2019), black carbon (Katich et al., 2018; Ditas et al., 2019), MSA (Hodshire et al.,
979 2019), sulfate-nitrate-ammonium (Nault et al., 2019), and sea salt (Yu et al, 2019; Bian et
980 al., 2019; Murphy et al., 2019).

981

5   Conclusions and implications

Our understanding and representation in global models of the lifecycle of the OA remain highly uncertain, especially in remote regions where constraints from measurements have been very sparse. We have performed a systematic evaluation of the performance of eight global chemistry climate models and of 28 AeroCom-II models in simulating the latitudinal and vertical distribution of OA and its composition in the remote regions of the Atlantic and Pacific marine boundary layer, free troposphere and lower stratosphere, using the unique measurements from the ATom campaign. Our simulations are conducted for both ATom-1 and ATom-2 deployments that took place in August 2016 and February 2017, respectively. The main conclusions of the comparison are as follows:

- The AeroCom-II ensemble average tends to be biased high by a factor of 2-5 in comparison to measured vertical OA profiles in the remote atmosphere during both NH summer and NH winter. The ensemble spread increases from a factor of 40 in the NH source regions to a factor of 1000 in remote regions of the Southern Ocean. The evaluation of AeroCom-II models in the remote regions provides an extension of the previous evaluation with continental ground data by Tsigaridis et al. (2014). We note that the data from the AeroCom-II models were based on monthly mean values from a different simulated year than the ATom campaigns; however, the consistent model biases are strong enough that we would not expect our conclusions to change for a different modeled year.

- The results of the ATom model ensemble used in this work show a much better agreement with the OA observations in all regions and reduced model variability. However, some of the agreement is for the wrong reasons, as most models severely overestimate the contribution of POA and underestimate the contribution of SOA to total OA. Sensitivity simulations indicate that the POA overestimate in CESM could be due to an inadequate representation of primary aerosol removal by convective clouds, (additional convective removal per Yu et al. (2019) in CESM1-CARMA led to a better agreement with observations). Most models have insufficient production of SOA, and sensitivity studies indicate that a stronger production of SOA is needed to capture the measured concentrations. The photochemical ageing of POA which was not considered here (unlike for SOA) could also contribute to the model overestimation. The non-volatile POA treatment in models is consistent with the assumption of inert POA particles used to estimate POA from measurements, and cannot explain the

1015   model bias. Indeed, sensitivity simulations with semi-volatile POA lead to a much
1016   larger model bias for OA in the upper troposphere and remote regions. The
1017   compensation between errors in POA and SOA in remote regions is however a
1018   recurring issue in OA modeling (de Gouw and Jimenez, 2009). For instance, it was
1019   found in the urban outflow regions such as Mexico City during MILAGRO 2006 field
1020   campaign (Fast et al., 2009; Hodzic et al., 2009); Paris during MEGAPOLI 2009
1021   (Zhang et al., 2013); the Los Angeles area during CalNex-2010 (Baker et al., 2015;
1022   Woody et al., 2016); the NE US outflow during WINTER 2015 (Schroder et al., 2018;
1023   Shah et al., 2019).

1024 • Additional errors in simulated OA concentrations can arise from the use of too low
1025   OA/OC ratios when model results (often calculated as OC) are converted to OA for
1026   comparison with measurements. We note that OA is the most atmospherically-relevant
1027   quantity, while OC is an operational quantity, partially a relic from a period in which
1028   only OC could be separately quantified (although also of some use for carbon budget
1029   studies). It should also be noted that most emission inventories still use OC as the
1030   primary variable, which is why the use of accurate OA/OC ratios is still key for all
1031   models. We show that the OA/OC ratio used in most models is too low compared to
1032   measured values that range mostly from 2.2 to 2.5, resulting in errors in OA mass of
1033   ~70% for AeroCom-II models and ~30% for current models that use organic carbon to
1034   track OA mass. Remote OA is thus highly oxidized and chemically processed. These
1035   results demonstrate that current global chemistry-climate models underestimate the
1036   degree of oxidation of OA in remote regions and need to consider further chemical
1037   ageing of OA, which could impact the calculations of its burden, and optical and
1038   hygroscopic properties.

1039 • The results also show that in most models (except CESM2) the predicted OA
1040   contribution to the total submicron aerosol is underestimated relative to sulfate in the
1041   remote FT where OA and sulfate are the dominant submicron aerosols (important for
1042   climate). Accurate predictions of composition of submicron particles remains
1043   challenging in remote regions and should be the topic of future studies.

1044 Key implications of our results are: (i) Model errors on the relative contribution of POA and
1045 SOA to OA reduce our confidence on the ability to simulate radiative forcing over time or
1046 OA health impacts; (ii) Model errors for the relative contributions of sulfate and organics
1047 to the submicron aerosol in the free troposphere could lead to errors in the predicted CCN
1048 or radiative forcing of aerosols as inorganics are more hygroscopic than OA; (iii) the OA

system seems to be more dynamic with a need for an enhanced removal of primary OA, and a stronger production of secondary OA in global models to provide a better agreement with observations.

Code/Data availability: Data can be obtained from the ATom website: https://doi.org/10.3334/ORNLDAAC/1581.

L2 Measurements from CU High-Resolution Aerosol Mass Spectrometer (HR-AMS) can be obtained from the ORNL DAAC, Oak Ridge, Tennessee, USA. https://doi.org/10.3334/ORNLDAAC/1716.

Author contribution: A. Hodzic, P. Campuzano-Jost and J.L. Jimenez performed the measurement / model comparisons, wrote and revised the manuscript. P. Campuzano-Jost, D.A. Day, B.N. Nault, J.C. Schroder, D.T. Sueper, and J. L. Jimenez performed and analyzed the AMS measurements. K.D. Froyd and G.P. Schill performed and analyzed the PALMS measurements. J.P. Schwarz and J.M. Katich performed the BC measurements. H. Bian, M. Chin, P.R. Colarco, B. Heinold, A. Hodzic, D.S. Jo, J.K.

References

Aiken, A.C., P.F. DeCarlo, J.H. Kroll, D.R. Worsnop, J.A. Huffman, K. Docherty, I.M. Ulbrich, C. Mohr, J.R. Kimmel, D. Sueper, Q. Zhang, Y. Sun, A. Trimborn, M. Northway, P.J. Ziemann, M.R. Canagaratna, T.B. Onasch, R. Alfarra, A.S.H. Prevot, J. Dommen, J. Duplissy, A. Metzger, U. Baltensperger, and J.L. Jimenez. O/C and OM/OC Ratios of Primary, Secondary, and Ambient Organic Aerosols with High Resolution Time-of-Flight Aerosol Mass Spectrometry Environmental Science & Technology, 42, 4478–4485, doi: 10.1021/es703009q, 2008.

Andreae, M. O.: Emission of trace gases and aerosols from biomass burning – An updated assessment, Atmos. Chem. Phys. Discuss., 1–27, doi:10.5194/acp-2019-303, 2019.

Bahreini, R., Ervens, B., Middlebrook, A. M., Warneke, C., de Gouw, J. A., DeCarlo, P. F., Jimenez, J. L., Brock, C. A., Neuman, J. A., Ryerson, T. B., Stark, H., Atlas, E., Brioude, J., Fried, A., Holloway, J. S., Peischl, J., Richter, D., Walega, J., Weibring, P., Wollny, A. G. and Fehsenfeld, F. C.: Organic aerosol formation in urban and industrial plumes near Houston and Dallas, Texas, J. Geophys. Res., 114, D00F16, doi:10.1029/2008JD011493, 2009.

Baker, K. R., Carlton, A. G., Kleindienst, T. E., Offenberg, J. H., Beaver, M. R., Gentner, D. R., Goldstein, A. H., Hayes, P. L., Jimenez, J. L., Gilman, J. B., de Gouw, J. A., Woody, M. C., Pye, H. O. T., Kelly, J. T., Lewandowski, M., Jaoui, M., Stevens, P. S., Brune, W. H., Lin, Y.-H., Rubitschun, C. L., and Surratt, J. D.: Gas and aerosol carbon in California: comparison of measurements and model predictions in Pasadena and Bakersfield, Atmos. Chem. Phys., 15, 5243-5258, https://doi.org/10.5194/acp-15-5243-2015, 2015.

Bey, I., Jacob, D. J., Yantosca, R. M., and Logan, J. A.: Global modeling of tropospheric chemistry with assimilated meteorology: model description and evaluation, J Geophys. Res., 106, 23073–23095, 2011.

Bian, H., Froyd, K., Murphy, D. M., Dibb, J., Chin, M., Colarco, P. R., Darmenov, A., da Silva, A., Kucsera, T. L., Schill, G., Yu, H., Bui, P., Dollner, M., Weinzierl, B., and Smirnov, A.: Observationally constrained analysis of sea salt aerosol in the marine atmosphere, Atmos. Chem. Phys. Discuss., https://doi.org/10.5194/acp-2019-18, 2019.

1112   Bian, H., Chin, M., Hauglustaine, D. A., Schulz, M., Myhre, G., Bauer, S. E., Lund, M. T.,

1113   Karydis, V. A., Kucsera, T. L., Pan, X., Pozzer, A., Skeie, R. B., Steenrod, S. D., Sudo, K.,

1114   Tsigaridis, K., Tsimpidi, A. P., and Tsyro, S. G.: Investigation of global nitrate from the

1115   AeroCom Phase III experiment, Atmos. Chem. Phys., 17, 12911-12940,

1116   https://doi.org/10.5194/acp-17-12911-2017, 2017.

1117   Boucher, O., D. Randall, P. Artaxo, C. Bretherton, G. Feingold, P. Forster, V.-M.

1118   Kerminen, Y. Kondo, H. Liao, U. Lohmann, P. Rasch, S.K. Satheesh, S. Sherwood, B.

1119   Stevens and X.Y. Zhang: Clouds and Aerosols. In: Climate Change 2013: The Physical

1120   Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the

1121   Intergovernmental Panel on Climate Change [Stocker, T.F., D. Qin, G.-K. Plattner, M.

1122   Tignor, S.K. Allen, J. Boschung, A. Nauels, Y. Xia, V. Bex and P.M. Midgley (eds.)].

1123   Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 2013.

1124   Bowman, K. P.: Large-scale isentropic mixing properties of the Antarctic polar vortex from

1125   analyzed winds. J. Geophys. Res., 98 (D12), 23 013–23 027, 1993.

1126   Brock, C. A., Wagner, N. L., Anderson, B. E., Attwood, A. R., Beyersdorf, A., Campuzano-

1127   Jost, P., Carlton, A. G., Day, D. A., Diskin, G. S., Gordon, T. D., Jimenez, J. L., Lack, D.

1128   A., Liao, J., Markovic, M. Z., Middlebrook, A. M., Ng, N. L., Perring, A. E., Richardson, M.

1129   S., Schwarz, J. P., Washenfelder, R. A., Welti, A., Xu, L., Ziemba, L. D., and Murphy, D.

1130   M.: Aerosol optical properties in the southeastern United States in summer – Part 1:

1131   Hygroscopic growth, Atmos. Chem. Phys., 16, 4987-5007, https://doi.org/10.5194/acp-16-

1132   4987-2016, 2016.

1133   Brock, C. A., Williamson, C., Kupc, A., Froyd, K. D., Erdesz, F., Wagner, N., Richardson,

1134   M., Schwarz, J. P., Gao, R.-S., Katich, J. M., Campuzano-Jost, P., Nault, B. A., Schroder,

1135   J. C., Jimenez, J. L., Weinzierl, B., Dollner, M., Bui, T. and Murphy, D. M.: Aerosol size

1136   distributions during the Atmospheric Tomography Mission (ATom): methods,

1137   uncertainties, and data products, Atmos. Meas. Tech., 12(6), 3081–3099,

1138   doi:10.5194/amt-12-3081-2019, 2019.

1139   Canagaratna, M. R., Jayne, J. T., Jimenez, J. L., Allan, J. D., Alfarra, M. R., Zhang, Q.,

1140   Onasch, T. B., Drewnick, F., Coe, H., Middlebrook, A. M., Delia, A., Williams, L. R.,

1141   Trimborn, A. M., Northway, M. J., DeCarlo, P. F., Kolb, C. E., Davidovits, P., and Worsnop,

1142   D. R.: Chemical and microphysical characterization of ambient aerosols with the Aerodyne

Aerosol Mass Spectrometer, Mass Spectrom. Rev., 26, 185–222, https://doi.org/10.1002/mas.20115, 2007.

Canagaratna, M. R., Jimenez, J. L., Kroll, J. H., Chen, Q., Kessler, S. H., Massoli, P., Hildebrandt Ruiz, L., Fortner, E., Williams, L. R., Wilson, K. R., Surratt, J. D., Donahue, N. M., Jayne, J. T., and Worsnop, D. R.: Elemental ratio measurements of organic compounds using aerosol mass spectrometry: characterization, improved calibration, and implications, Atmos. Chem. Phys., 15, 253-272, https://doi.org/10.5194/acp-15-253-2015, 2015.

Carslaw, K. S., Lee, L. a, Reddington, C. L., Pringle, K. J., Rap, A., Forster, P. M., Mann, G. W., Spracklen, D. V, Woodhouse, M. T., Regayre, L. a and Pierce, J. R.: Large contribution of natural aerosols to uncertainty in indirect forcing., Nature, 503(7474), 67–71, doi:10.1038/nature12674, 2013.

Colarco, P., da Silva, A., Chin, M., and Diehl, T.: Online simulations of global aerosol distributions in the NASA GEOS-4 model and comparisons to satellite and ground based aerosol optical depth, J. Geophys. Res., 115, D14207, doi:10.1029/2009JD012820, 2010.

Croft, B., Lohmann, U., Martin, R. V., Stier, P., Wurzler, S., Feichter, J., Hoose, C., Heikkilä, U., van Donkelaar, A., and Ferrachat, S.: Influences of in-cloud aerosol scavenging parameterizations on aerosol concentrations and wet deposition in ECHAM5-HAM, Atmos. Chem. Phys., 10, 1511–1543, https://doi.org/10.5194/acp-10-1511-2010, 2010.

Cubison, M. J., Ortega, A. M., Hayes, P. L., Farmer, D. K., Day, D., Lechner, M. J., Brune, W. H., Apel, E., Diskin, G. S., Fisher, J. A., Fuelberg, H. E., Hecobian, A., Knapp, D. J., Mikoviny, T., Riemer, D., Sachse, G. W., Sessions, W., Weber, R. J., Weinheimer, A. J., Wisthaler, A., and Jimenez, J. L.: Effects of aging on organic aerosol from open biomass burning smoke in aircraft and laboratory studies, Atmos. Chem. Phys., 11, 12049-12064, https://doi.org/10.5194/acp-11-12049-2011, 2011.

DeCarlo, P., Slowik, J., Worsnop, D., Davidovits, P., Jimenez, J., Stainken, K., Williams, L., Jayne, J., Kolb, C. and Rudich, Y.: Particle Morphology and Density Characterization by Combined Mobility and Aerodynamic Diameter Measurements. Part 1: Theory, Aerosol Sci. Technol., 38(12), 1185–1205, doi:10.1080/027868290903907, 2004.

DeCarlo, P. F., Kimmel, J. R., Trimborn, A., Northway, M. J., Jayne, J. T., Aiken, A. C., Gonin, M., Fuhrer, K., Horvath, T., Docherty, K. S., Worsnop, D. R., and Jimenez, J. L.:

Field-Deployable, High-Resolution, Time-of-Flight Aerosol Mass Spectrometer, Anal. Chem., 78, 8281–8289, https://doi.org/10.1021/ac061249n, 2006.

de Gouw, J., and J.L. Jimenez. Organic Aerosols in the Earth's Atmosphere. Environmental Science & Technology, 43, 7614–7618, 2009. DOI: 10.1021/es9006004

Dentener, F., Kinne, S., Bond, T., Boucher, O., Cofala, J., Generoso, S., Ginoux, P., Gong, S., Hoelzemann, J. J., Ito, A., Marelli, L., Penner, J. E., Putaud, J.-P., Textor, C., Schulz, M., van der Werf, G. R., and Wilson, J.: Emissions of primary aerosol and precursor gases in the years 2000 and 1750 prescribed data-sets for AeroCom, Atmospheric Chemistry and Physics, 6, 4321–4344, https://doi.org/10.5194/acp- 6-4321-2006, https://www.atmos-chem-phys.net/6/4321/2006/, 2006.

Ditas, J., Ma, N., Zhang, Y., Assmann, D., et al.: Strong impact of wildfires on the abundance and aging of black carbon in the lowermost stratosphere, Proc. Natl. Acad. Sci., 811595-11603, doi:10.1073/pnas.1806868115, 2018.

Drewnick, F., Hings, S. S., Alfarra, M. R., Prevot, A. S. H. and Borrmann, S.: Aerosol quantification with the Aerodyne Aerosol Mass Spectrometer: detection limits and ionizer background effects, Atmos. Meas. Tech., 2(1), 33–46, 2009.

Emmons, L. K., Orlando, J. J., Tyndall, G., Schwantes, R. H., Kinnison, D. E., Marsh, D. R., Mills, M. J., Tilmes, S., and Lamarque, J.-F.: The MOZART Chemistry Mechanism in the Community Earth System Model version 2 (CESM2), to be Submitted to J. Adv. Modeling Earth Systems, 2019.

Farmer, D.K., A. Matsunaga, K.S. Docherty, J.D. Surratt, J.H. Seinfeld, P.J. Ziemann, and J.L. Jimenez. Response of an Aerosol Mass Spectrometer to Organonitrates and Organosulfates and implications for Atmospheric Chemistry. Proceedings of the National Academy of Sciences of the USA, 107, 6670-6675, doi: 10.1073/pnas.0912340107, 2010.

Fast, J., Aiken, A. C., Allan, J., Alexander, L., Campos, T., Canagaratna, M. R., Chapman, E., DeCarlo, P. F., de Foy, B., Gaffney, J., de Gouw, J., Doran, J. C., Emmons, L., Hodzic, A., Herndon, S. C., Huey, G., Jayne, J. T., Jimenez, J. L., Kleinman, L., Kuster, W., Marley, N., Russell, L., Ochoa, C., Onasch, T. B., Pekour, M., Song, C., Ulbrich, I. M., Warneke, C., Welsh-Bon, D., Wiedinmyer, C., Worsnop, D. R., Yu, X.-Y., and Zaveri, R.: Evaluating simulated primary anthropogenic and biomass burning organic aerosols during MILAGRO: implications for assessing treatments of secondary organic aerosols, Atmos. Chem. Phys., 9, 6191-6215, https://doi.org/10.5194/acp-9-6191-2009, 2009.

Feng, L., Smith, S. J., Braun, C., Crippa, M., Gidden, M. J., Hoesly, R., Klimont, Z., van Marle, M., van den Berg, M., and van der Werf, G. R.: Gridded Emissions for CMIP6, Geosci. Model Dev. Discuss., https://doi.org/10.5194/gmd-2019-195, in review, 2019.

Flamant, C., Knippertz, P., Fink, A. H., Akpo, A., Brooks, B., Chiu, C. J., Coe, H., Danuor, S., Evans, M., Jegede, O., Kalthoff, N., Konaré, A., Liousse, C., Lohou, F., Mari, C., Schlager, H., Schwarzenboeck, A., Adler, B., Amekudzi, L., Aryee, J., Ayoola, M., Batenburg, A. M., Bessardon, G., Borrmann, S., Brito, J., Bower, K., Burnet, F., Catoire, V., Colomb, A., Den- jean, C., Fosu-Amankwah, K., Hill, P. G., Lee, J., Lothon, M., Maranan, M., Marsham, J., Meynadier, R., Ngamini, J.-B., Rosenberg, P., Sauer, D., Smith, V., Stratmann, G., Taylor, J. W., Voigt, C., and Yoboué, V.: The Dynamics-Aerosol-Chemistry- Cloud Interactions in West Africa field campaign: Overview and research highlights, B. Am. Meteorol. Soc., 99, 83–104, https://doi.org/10.1175/BAMS-D-16-0256.1, 2018.

Forrister, H., Liu, J., Scheuer, E., Dibb, J., Ziemba, L., Thornhill, K. L., Anderson, B., Diskin, G., Perring, A. E., Schwarz, J. P., Campuzano-Jost, P., Day, D. A., Palm, B. B., Jimenez, J. L., Nenes, A. and Weber, R. J.: Evolution of brown carbon in wildfire plumes, Geophys. Res. Lett., 42(11), doi:10.1002/2015GL063897, 2015.

Fountoukis, C. and Nenes, A.: ISORROPIA II : a computationally efficient thermodynamic equilibrium model for K+-Ca2+-Mg2+-Na+-SO42--NO3--Cl--H2O aerosols, Atmos. Chem. Phys., 7, 4639–4659, 2007.

Froyd, K. D., Murphy, D. M., Brock, C. A., Campuzano-Jost, P., Dibb, J. E., Jimenez, J.-L., Kupc, A., Middlebrook, A. M., Schill, G. P., Thornhill, K. L., Williamson, C. J., Wilson, J. C., and Ziemba, L. D.: A new method to quantify mineral dust and other aerosol species from aircraft platforms using single particle mass spectrometry, Atmos. Meas. Tech. Discuss., https://doi.org/10.5194/amt-2019-165, in review, 2019.

George, I. J. and Abbatt, J. P. D.: Heterogeneous oxidation of atmospheric aerosol particles by gas-phase radicals, Nature Chemistry, 2, 713–722, 2010.

Gettelman, A., Mills, M. J., Kinnison, D. e., Garcia, R. R., Smith, A. K., Marsh, D. R., Tilmes, S., Vitt, F., Bardeen, C. G., McInerny, J., Liu, H.-L., Solomon, S. C., Polvani, L. M., Emmons, L. K., Lamarque, J.-F., Richter, J. H., Glanville, A. S., Bacmeister, J. T., Phillips, A. S., Neale, R. B., Simpson, I. R., DuVivier, A. K., Hodzic, A., Randel, W. J.: The

1238 Whole Atmosphere Community Climate Model Version 6 (WACCM6), J. of Geophys. Res.,
1239 in review, 2019.

1240 Gerber, H. E.: Relative-humidity parameterization of the Navy Aerosol Model (NAM),
1241 Tech. Rep. NRL Report 8956, Naval Research Laboratory, 1985.

1242 Giglio, L., Randerson, J. T. and Werf, G. R.: Analysis of daily, monthly, and annual burned
1243 area using the fourth-generation global fire emissions database (GFED4), J. Geophys.
1244 Res. Biogeosciences, 118, 317–328, 2013.

1245 Guenther, A. B., Jiang, X., Heald, C. L., Sakulyanontvittaya, T., Duhl, T., Emmons, L. K.,
1246 and Wang, X.: The Model of Emissions of Gases and Aerosols from Nature version 2.1
1247 (MEGAN2.1): an extended and updated framework for modeling biogenic emissions,
1248 Geosci. Model Dev., 5, 1471–1492, https://doi.org/10.5194/gmd-5-1471-2012,
1249 https://www.geosci-model-dev.net/5/1471/2012/, 2012.

1250 Hayes, P. L., Carlton, A. G., Baker, K. R., Ahmadov, R., Washenfelder, R. A., Alvarez, S.,
1251 Rappenglück, B., Gilman, J. B., Kuster, W. C., de Gouw, J. A., Zotter, P., Prévôt, A. S. H.,
1252 Szidat, S., Kleindienst, T. E., Offenberg, J. H., Ma, P. K., and Jimenez, J. L.: Modeling the
1253 formation and aging of secondary organic aerosols in Los Angeles during CalNex 2010,
1254 Atmos. Chem. Phys., 15, 5773-5801, https://doi.org/10.5194/acp-15-5773-2015, 2015.

1255 Hallquist, M., Wenger, J. C., Baltensperger, U., Rudich, Y., Simpson, D., Claeys, M.,
1256 Dommen, J., Donahue, N. M., George, C., Goldstein, A. H., Hamilton, J. F., Herrmann, H.,
1257 Hoffmann, T., Iinuma, Y., Jang, M., Jenkin, M. E., Jimenez, J. L., Kiendler-Scharr, A.,
1258 Maenhaut, W., McFiggans, G., Mentel, Th. F., Monod, A., Prévôt, A. S. H., Seinfeld, J. H.,
1259 Surratt, J. D., Szmigielski, R., and Wildt, J.: The formation, properties and impact of
1260 secondary organic aerosol: current and emerging issues, Atmos. Chem. Phys., 9, 5155-
1261 5236, https://doi.org/10.5194/acp-9-5155-2009, 2009.

1262 Heald, C. L., Coe, H., Jimenez, J. L., Weber, R. J., Bahreini, R., Middlebrook, A. M.,
1263 Russell, L. M., Jolleys, M., Fu, T.-M., Al- lan, J. D., Bower, K. N., Capes, G., Crosier, J.,
1264 Morgan, W. T., Robinson, N. H., Williams, P. I., Cubison, M. J., DeCarlo, P. F., and Dunlea,
1265 E. J.: Exploring the vertical profile of atmospheric organic aerosol: comparing 17 aircraft
1266 field campaigns with a global model, Atmos. Chem. Phys., 11, 12673–12696,
1267 doi:10.5194/acp-11-12673-2011, 2011.

1268 Hodshire, A. L., Campuzano-Jost, P., Kodros, J. K., Croft, B., Nault, B. A., Schroder, J.
1269 C., Jimenez, J. L. and Pierce, J. R.: The potential role of methanesulfonic acid (MSA) in

1270 aerosol formation and growth and the associated radiative forcings, Atmos. Chem. Phys.,
1271 19(5), 3137–3160, doi:10.5194/acp-19-3137-2019, 2019a.

1272 Hodshire, A., A. Akherati, M. Alvarado, B. Brown-Steiner, S. Jathar, J.L. Jimenez, S.
1273 Kreidenweis, C. Lonsdale, T. Onasch, A. Ortega, J. Pierce. Aging Effects on Biomass
1274 Burning Aerosol Mass and Composition: A Critical Review of Field and Laboratory
1275 Studies. Environ. Sci. Technol., submitted, 2019b.

1276 Hodzic, A., Jimenez, J. L., Madronich, S., Aiken, A. C., Bessagnet, B., Curci, G., Fast, J.,
1277 Lamarque, J.-F., Onasch, T. B., Roux, G., Schauer, J. J., Stone, E. A., and Ulbrich, I. M.:
1278 Modeling organic aerosols during MILAGRO: importance of biogenic secondary organic
1279 aerosols, Atmos. Chem. Phys., 9, 6949-6981, https://doi.org/10.5194/acp-9-6949-2009,
1280 2009.

1281 Hodzic, A., Aumont, B., Knote, C., Lee-Taylor, J., Madronich, S., and Tyndall, G.: Volatility
1282 dependence of Henry's law constants of condensable organics: Application to estimate
1283 depositional loss of secondary organic aerosols, Geophys. Res. Lett., 41, 4795–4804,
1284 doi:10.1002/2014GL060649, 2014.

1285 Hodzic, A., Madronich, S., Kasibhatla, P. S., Tyndall, G., Aumont, B., Jimenez, J. L., Lee-
1286 Taylor, J., and Orlando, J.: Organic photolysis reactions in tropospheric aerosols: effect
1287 on secondary organic aerosol formation and lifetime, Atmos. Chem. Phys., 15, 9253-9269,
1288 https://doi.org/10.5194/acp-15-9253-2015, 2015.

1289 Hodzic, A., Kasibhatla, P. S., Jo, D. S., Cappa, C. D., Jimenez, J. L., Madronich, S., and
1290 Park, R. J.: Rethinking the global secondary organic aerosol (SOA) budget: stronger
1291 production, faster removal, shorter lifetime, Atmos. Chem. Phys., 16, 7917-7941,
1292 https://doi.org/10.5194/acp-16-7917-2016, 2016.

1293 Hoesly, R. M., Smith, S. J., Feng, L., Klimont, Z., Janssens-Maenhout, G., Pitkanen, T.,
1294 Seibert, J. J., Vu, L., Andres, R. J., Bolt, R. M., Bond, T. C., Dawidowski, L., Kholod, N.,
1295 Kurokawa, J.-I., Li, M., Liu, L., Lu, Z., Moura, M. C. P., O'Rourke, P. R., and Zhang, Q.:
1296 Historical (1750–2014) anthropogenic emissions of reactive gases and aerosols from the
1297 Community Emissions Data System (CEDS), Geosci. Model Dev., 11, 369–408,
1298 https://doi.org/10.5194/gmd-11-369-2018, 2018.

1299 Huang, K., Fu, J. S., Prikhodko, V. Y., Storey, J. M., Romanov, A., Hodson, E. L., Cresko,
1300 J., Morozova, I., Ignatieva, Y., and Cabaniss, J.: Russian anthropogenic black carbon:

Emission reconstruction and Arctic black carbon simulation, J. Geophys. Res. Atmos., 120, 11,306– 11,333, doi:10.1002/2015JD023358, 2015.

Hudson, P. K., Murphy, D. M., Cziczo, D. J., Thomson, D. S., de Gouw, J. A., Warneke, C., Holloway, J., Jost, H. J. and Hübler, G.: Biomass-burning particle measurements: Characteristics composition and chemical processing, J. Geophys. Res. D Atmos., 109(23), 1–11, doi:10.1029/2003JD004398, 2004.

Janssens-Maenhout, G., Crippa, M., Guizzardi, D., Dentener, F., Muntean, M., Pouliot, G., Keating, T., Zhang, Q., Kurokawa, J., Wankmüller, R., Denier van der Gon, H., Kuenen, J. J. P., Klimont, Z., Frost, G., Darras, S., Koffi, B., and Li, M.: HTAP_v2.2: a mosaic of regional and global emission grid maps for 2008 and 2010 to study hemispheric transport of air pollution, Atmos. Chem. Phys., 15, 11411–11432, https://doi.org/10.5194/acp-15-11411-2015, 2015.

Jimenez, J. L., Canagaratna, M. R., Donahue, N. M., Prevot, a. S. H., Zhang, Q., Kroll, J. H., DeCarlo, P. F., Allan, J. D., Coe, H., Ng, N. L., Aiken, a. C., Docherty, K. S., Ulbrich, I. M., Grieshop, A. P., Robinson, a. L., Duplissy, J., Smith, J. D., Wilson, K. R., Lanz, V. a., Hueglin, C., Sun, Y. L., Tian, J., Laaksonen, A., Raatikainen, T., Rautiainen, J., Vaattovaara, P., Ehn, M., Kulmala, M., Tomlinson, J. M., Collins, D. R., Cubison, M. J., Dunlea, E. J., Huffman, J. A., Onasch, T. B., Alfarra, M. R., Williams, P. I., Bower, K., Kondo, Y., Schneider, J., Drewnick, F., Borrmann, S., Weimer, S., Demerjian, K., Salcedo, D., Cottrell, L., Griffin, R., Takami, A., Miyoshi, T., Hatakeyama, S., Shimono, A., Sun, J. Y., Zhang, Y. M., Dzepina, K., Kimmel, J. R., Sueper, D., Jayne, J. T., Herndon, S. C., Trimborn, a. M., Williams, L. R., Wood, E. C., Middlebrook, A. M., Kolb, C. E., Baltensperger, U., Worsnop, D. R., Worsnop, D. R., Dunlea, J., Huffman, J. A., Onasch, T. B., Alfarra, M. R., Williams, P. I., Bower, K., Kondo, Y., Schneider, J., Drewnick, F., Borrmann, S., Weimer, S., Demerjian, K., Salcedo, D., Cottrell, L., Griffin, R., Takami, A., Miyoshi, T., Hatakeyama, S., Shimono, A., Sun, J. Y., Zhang, Y. M., Dzepina, K., Kimmel, J. R., Sueper, D., Jayne, J. T., Herndon, S. C., Trimborn, a. M., Williams, L. R., Wood, E. C., Middlebrook, A. M., Kolb, C. E., Baltensperger, U., Worsnop, D. R., Dunlea, E. J., et al.: Evolution of Organic Aerosols in the Atmosphere, Science 80., 326(5959), 1525–1529, doi:10.1126/science.1180353, 2009.

Jimenez, J.L., P. Campuzano-Jost, D.A. Day, B.A. Nault, D.J. Price, and J.C. Schroder: ATom: L2 Measurements from CU High-Resolution Aerosol Mass Spectrometer (HR-

AMS). ORNL DAAC, Oak Ridge, Tennessee, USA. https://doi.org/10.3334/ORNLDAAC/1716, 2019a.

Jimenez, J.L., et al.: Evaluating the Consistency of All Submicron Aerosol Mass Measurements (Total and Speciated) in the Atmospheric Tomography Mission (ATom), Abstract A31A-08, presented at 2019 Fall Meeting, AGU, San Francisco, CA, 9-13 Dec., 2019b.

Jolleys, M. D., Coe, H., McFiggans, G., Taylor, J. W., O'Shea, S. J., Le Breton, M., Bauguitte, S. J.-B., Moller, S., Di Carlo, P., Aruffo, E., Palmer, P. I., Lee, J. D., Percival, C. J. and Gallagher, M. W.: Properties and evolution of biomass burning organic aerosol from Canadian boreal forest fires, Atmos. Chem. Phys., 15(6), 3077–3095, doi:10.5194/acp-15-3077-2015, 2015.

Kaiser, J. W., Heil, A., Andreae, M. O., Benedetti, A., Chubarova, N., Jones, L., Morcrette, J.-J., Razinger, M., Schultz, M. G., Suttie, M., and van der Werf, G. R.: Biomass burning emissions estimated with a global fire assimilation system based on observed fire radiative power, Biogeosciences, 9, 527–554, https://doi.org/10.5194/bg-9-527-2012, https://www.biogeosciences.net/9/527/2012/, 2012.

Kanakidou, M., Seinfeld, J. H., Pandis, S. N., Barnes, I., Dentener, F. J., Facchini, M. C., Van Dingenen, R., Ervens, B., Nenes, A., Nielsen, C. J., Swietlicki, E., Putaud, J. P., Balkanski, Y., Fuzzi, S., Horth, J., Moortgat, G. K., Winterhalter, R., Myhre, C. E. L., Tsigaridis, K., Vignati, E., Stephanou, E. G., and Wilson, J.: Organic aerosol and global climate modelling: a review, Atmos. Chem. Phys., 5, 1053-1123, https://doi.org/10.5194/acp-5-1053-2005, 2005.

Katich, J. M., Samset, B. H., Paul Bui, T., Dollner, M., Froyd, K.D., Campuzano-Jost, P., Nault, B. A., Schroder, J. C., Weinzierl, B., Schwarz, J. P.: Strong Contrast in Remote Black Carbon Aerosol Loadings Between the Atlantic and Pacific Basins, J. Geophys. Res. Atmos., 123 (23), pages 13,386-13,395, https://doi.org/10.1029/2018JD029206, 2018.

Kim, M. J., G. A. Novak, M. C. Zoerb,M. Yang, B. W. Blomquist, B. J. Huebert, C. D. Cappa, and T. H. Bertram: Air-Sea exchange of biogenic volatile organic compounds and the impact on aerosol particle size distributions, Geophys. Res. Lett.,44, 3887–3896, doi:10.1002/2017GL072975, 2017.

1363   Klimont, Z., K. Kupiainen, C. Heyes, P. Purohit, J. Cofala, P. Rafaj, J. Borken-Kleefeld,
1364   and W. Schöpp,: Global anthropogenic emissions of particulate matter including black
1365   carbon, Atmos. Chem. Phys., 17(14), 8681-8723, doi: 10.5194/acp-17-8681-2017, 2017.

1366   Kodros, J. K., Cucinotta, R., Ridley, D. A., Wiedinmyer, C., and Pierce, J. R.: The aerosol
1367   radiative effects of uncontrolled combustion of domestic waste, Atmos. Chem. Phys., 16,
1368   6771-6784, https://doi.org/10.5194/acp-16-6771-2016, 2016.

1369   Krechmer, J.E., D. Pagonis, P.J. Ziemann, and J.L. Jimenez. Quantification of gas-wall
1370   partitioning in Teflon environmental chambers using rapid bursts of low-volatility oxidized
1371   species generated in-situ. Environmental Science and Technology, 50, 5757–5765,
1372   doi:10.1021/acs.est.6b00606, 2016.

1373   Kroll, J. H., Donahue, N. M., Jimenez, J. L., Kessler, S. H., Canagaratna, M. R., Wilson,
1374   K. R., Altieri, K. E., Mazzoleni, L. R., Wozniak, A. S., Bluhm, H., Mysak, E. R., Smith, J.
1375   D., Kolb, C. E. and Worsnop, D. R.: Carbon oxidation state as a metric for describing the
1376   chemistry of atmospheric organic aerosol., Nat. Chem., 3(2), 133–9,
1377   doi:10.1038/nchem.948, 2011.

1378   Laskin, A., Laskin, J., and Nizkorodov, S.: Chemistry of Atmospheric Brown Carbon,
1379   Chem. Rev. 2015, 115, 10, 4335-4382, 2015.

1380   Lee, A. K. Y., Willis, M. D., Healy, R. M., Onasch, T. B., and Abbatt, J. P. D.: Mixing state
1381   of carbonaceous aerosol in an urban environment: single particle characterization using
1382   the soot particle aerosol mass spectrometer (SP-AMS), Atmos. Chem. Phys., 15, 1823-
1383   1841, https://doi.org/10.5194/acp-15-1823-2015, 2015.

1384   Liao, J., K.D. Froyd, D.M. Murphy, F.N. Keutsch, G. Yu, P.O.Wennberg, J.St. Clair, J.D.
1385   Crounse, A. Wisthaler, T. Mikoviny, T.B. Ryerson, I.B. Pollack, J. Peischl, J.L. Jimenez,
1386   P. Campuzano Jost, D.A. Day, B.E. Anderson, L.D. Ziemba, D.R. Blake, S. Meinardi, G.
1387   Diskin. Airborne organosulfates measurements over the continental US. Journal of
1388   Geophysical Research-Atmospheres, 120, 2990–3005, doi:10.1002/2014JD022378,
1389   2015.

1390   Liu, H. Y., D. J. Jacob, I. Bey, and R. M. Yantosca: Constraints from Pb-210 and Be-7 on
1391   wet deposition and transport in a global three-dimensional chemical tracer model driven
1392   by assimilated meteorological fields, J. Geophys. Res., 106(D11), 12,109–12,128, 2001.

Liu, X., Easter, R. C., Ghan, S. J., Zaveri, R., Rasch, P., Shi, X., Lamarque, J.-F., Gettelman, A., Morrison, H., Vitt, F., Conley, A., Park, S., Neale, R., Hannay, C., Ekman, A. M. L., Hess, P., Mahowald, N., Collins, W., Iacono, M. J., Bretherton, C. S., Flanner, M. G., and Mitchell, D.: Toward a minimal representation of aerosols in climate models: description and evaluation in the Community Atmosphere Model CAM5, Geosci. Model Dev., 5, 709-739, https://doi.org/10.5194/gmd-5-709-2012, 2012.

Marais, E. A., Jacob, D. J., Jimenez, J. L., Campuzano-Jost, P., Day, D. A., Hu, W., Krechmer, J., Zhu, L., Kim, P. S., Miller, C. C., Fisher, J. A., Travis, K., Yu, K., Hanisco, T. F., Wolfe, G. M., Arkinson, H. L., Pye, H. O. T., Froyd, K. D., Liao, J., and McNeill, V. F.: Aqueous-phase mechanism for secondary organic aerosol formation from isoprene: application to the southeast United States and co-benefit of SO_2 emission controls, Atmos. Chem. Phys., 16, 1603-1618, https://doi.org/10.5194/acp-16-1603-2016, 2016.

Myhre, G., Samset, B. H., Schulz, M., Balkanski, Y., Bauer, S., Berntsen, T. K., Bian, H., Bellouin, N., Chin, M., Diehl, T., Easter, R. C., Feichter, J., Ghan, S. J., Hauglustaine, D., Iversen, T., Kinne, S., Kirkevåg, A., Lamarque, J.-F., Lin, G., Liu, X., Lund, M. T., Luo, G., Ma, X., van Noije, T., Penner, J. E., Rasch, P. J., Ruiz, A., Seland, Ø., Skeie, R. B., Stier, P., Takemura, T., Tsigaridis, K., Wang, P., Wang, Z., Xu, L., Yu, H., Yu, F., Yoon, J.-H., Zhang, K., Zhang, H., and Zhou, C.: Radiative forcing of the direct aerosol effect from AeroCom Phase II simulations, Atmos. Chem. Phys., 13, 1853-1877, https://doi.org/10.5194/acp-13-1853-2013, 2013.

Mauderly, J. L., & Chow, J. C: Health Effects of Organic Aerosols, Inhalation Toxicology, 20:3, 257-288, DOI: 10.1080/08958370701866008, 2008.

Middlebrook, A. M., Murphy, D. M. and Thomson, D. S.: Observations of organic material in individual marine particles at Cape Grim during the First Aerosol Characterization Experiment (ACE 1), J. Geophys. Res. Atmos., 103(D13), 16475–16483, doi:10.1029/97JD03719, 1998.

Murphy, D. M., Froyd, K. D., Bian, H., Brock, C. A., Dibb, J. E., DiGangi, J. P., Diskin, G., Dollner, M., Kupc, A., Scheuer, E. M., Schill, G. P., Weinzierl, B., Williamson, C. J., and Yu, P.: The distribution of sea-salt aerosol in the global troposphere, Atmos. Chem. Phys., 19, 4093-4104, https://doi.org/10.5194/acp-19-4093-2019, 2019.

Nault, B. A., Campuzano-Jost, P., Day, D. A., Schroder, J. C., Anderson, B., Beyersdorf, A. J., Blake, D. R., Brune, W. H., Choi, Y., Corr, C. A., de Gouw, J. A., Dibb, J., DiGangi,

1425 J. P., Diskin, G. S., Fried, A., Huey, L. G., Kim, M. J., Knote, C. J., Lamb, K. D., Lee, T.,
1426 Park, T., Pusede, S. E., Scheuer, E., Thornhill, K. L., Woo, J.-H., and Jimenez, J. L.:
1427 Secondary organic aerosol production from local emissions dominates the organic aerosol
1428 budget over Seoul, South Korea, during KORUS-AQ, Atmos. Chem. Phys., 18, 17769-
1429 17800, https://doi.org/10.5194/acp-18-17769-2018, 2018.

1430 Ovadnevaite, J., Ceburnis, D., Canagaratna, M., Berresheim, H., Bialek, J., Martucci, G.,
1431 Worsnop, D. R. and O'Dowd, C.: On the effect of wind speed on submicron sea salt mass
1432 concentrations and source fluxes, J. Geophys. Res. Atmos., 117(D16), 1–11,
1433 doi:10.1029/2011JD017379, 2012.

1434 Pai, S. J., Heald, C. L., Pierce, J. R., Farina, S. C., Marais, E. A., Jimenez, J. L.,
1435 Campuzano-Jost, P., Nault, B. A., Middlebrook, A. M., Coe, H., Shilling, J. E., Bahreini,
1436 R., Dingle, J. H., and Vu, K.: An evaluation of global organic aerosol schemes using
1437 airborne observations, Atmos. Chem. Phys. Discuss., https://doi.org/10.5194/acp-2019-
1438 331, in press, 2020.

1439 Palm, B. B., de Sá, S. S., Day, D. A., Campuzano-Jost, P., Hu, W., Seco, R., Sjostedt, S.
1440 J., Park, J.-H., Guenther, A. B., Kim, S., Brito, J., Wurm, F., Artaxo, P., Thalman, R., Wang,
1441 J., Yee, L. D., Wernis, R., Isaacman-VanWertz, G., Goldstein, A. H., Liu, Y., Springston,
1442 S. R., Souza, R., Newburn, M. K., Alexander, M. L., Martin, S. T., and Jimenez, J. L.:
1443 Secondary organic aerosol formation from ambient air in an oxidation flow reactor in
1444 central Amazonia, Atmos. Chem. Phys., 18, 467–493, https://doi.org/10.5194/acp-18-467-
1445 2018, 2018.

1446 Petters, M. D., Prenni, A. J., Kreidenweis, S. M., DeMott, P. J., Matsunaga, A., Lim, Y. B.,
1447 and Ziemann, P. J. Chemical aging and the hydrophobic-to-hydrophilic conversion of
1448 carbonaceous aerosol, Geophys. Res. Lett., 33, L24806, doi:10.1029/2006GL027249,
1449 2006.

1450 Pye, H. O. T., Chan, A. W. H., Barkley, M. P., and Seinfeld, J. H.: Global modeling of
1451 organic aerosol: The importance of reactive nitrogen (NOx and NO_3), Atmos. Chem. Phys.,
1452 10, 11,261–11,276, doi:10.5194/acp-10-11261-2010, 2010.

1453 Robinson, A. L., Donahue, N. M., Shrivastava, M. K., Weitkamp, E. A., Sage, A. M.,
1454 Grieshop, A. P., Lane, T. E., Pandis, S. N., and Pierce, J. R.: Rethinking organic aerosols:
1455 Semivolatile emissions and photochemical aging, Science, 315, 1259–1262, 2007.

Schroder, J. C., Campuzano-Jost, P., Day, D. A., Shah, V., Larson, K., Sommers, J. M., Sullivan, A. P., Campos, T., Reeves, J. M., Hills, A., Hornbrook, R. S., Blake, N. J., Scheuer, E., Guo, H., Fibiger, D. L., McDuffie, E. E., Hayes, P. L., Weber, R. J., Dibb, J. E., Apel, E. C., Jaeglé, L., Brown, S. S., Thornton, J. A. and Jimenez, J. L.: Sources and Secondary Production of Organic Aerosols in the Northeastern United States during WINTER, J. Geophys. Res. Atmos., 123(14), 7771–7796, doi:10.1029/2018JD028475, 2018.

Schacht, J., Heinold, B., Quaas, J., Backman, J., Cherian, R., Ehrlich, A., Herber, A., Huang, W. T. K., Kondo, Y. Massling, A, Sinha, P.R., Weinzierl, B. Zanatta, M., and Tegen, I.: The importance of the representation of air pollution emissions for the modeled distribution and radiative effects of black carbon in the Arctic, Atmos. Chem. Phys. Discuss., 2019, 1-39, doi: 10.5194/acp-2019-71, 2019.

Schulz, M., Chin, M., and Kinne, S.: The aerosol model comparison project, AeroCom, phase II: clearing up diversity, IGAC newsletter, No. 41, 2–11, 2009.

Scott, C. E., Rap, A., Spracklen, D. V., Forster, P. M., Carslaw, K. S., Mann, G. W., Pringle, K. J., Kivekäs, N., Kulmala, M., Lihavainen, H., and Tunved, P.: The direct and indirect radiative effects of biogenic secondary organic aerosol, Atmos. Chem. Phys., 14, 447-470, https://doi.org/10.5194/acp-14-447-2014, 2014.

Shah, V., Jaeglé, L., Jimenez, J. L., Schroder, J. C., Campuzano-Jost, P., Campos, T. L., et al.: Widespread pollution from secondary sources of organic aerosols during winter in the Northeastern United States. Geophysical Research Letters, 46, 2974– 2983. https://doi.org/10.1029/2018GL081530, 2019.

Shiraiwa, M., Ueda, K., Pozzer, A., Lammel, G., Kampf, C. J.: Aerosol Health Effects from Molecular to Global Scales, Environ. Sci. Technol. 51 (23), pp 13545–13567, 2017.

Shrivastava, M., Cappa, C.D., Fan, J., et al: Recent advances in understanding secondary organic aerosol: Implications for global climate forcing, Rev. Geophys., 55, 509–559, doi:10.1002/2016RG000540, 2017.

Spracklen, D. V., Jimenez, J. L., Carslaw, K. S., Worsnop, D. R., Evans, M. J., Mann, G. W., Zhang, Q., Canagaratna, M. R., Allan, J., Coe, H., McFiggans, G., Rap, A., and Forster, P.: Aerosol mass spectrometer constraint on the global secondary organic aerosol budget, Atmos. Chem. Phys., 11, 12109–12136, doi:10.5194/acp-11-12109-2011, 2011.

Tegen, I., Neubauer, D., Ferrachat, S., Siegenthaler-Le Drian, C., Bey, I., Schutgens, N., Stier, P., Watson-Parris, D., Stanelle, T., Schmidt, H., Rast, S., Kokkola, H., Schultz, M., Schroeder, S., Daskalakis, N., Barthel, S., Heinold, B., and Lohmann, U.: The global aerosol–climate model ECHAM6.3–HAM2.3 – Part 1: Aerosol evaluation, Geosci. Model Dev., 12, 1643-1677, https://doi.org/10.5194/gmd-12-1643-2019, 2019.

Thomson, D. S., Schein, M. E. and Murphy, D. M.: Particle Analysis by Laser Mass Spectrometry WB-57F Instrument Overview, Aerosol Sci. Technol., 33(1–2), 153–169, doi:10.1080/027868200410903, 2000.

Tilmes, S., Hodzic, A., Emmons, L.K., Mills, M.J., Gettelman, A., Kinnison, D.E., Park, M., Lamarque J.-F., Vitt, F., et al.: Climate forcing and trends of organic aerosols in the Community Earth System Model (CESM2), to be submitted to JAMES, 2019.

Tsigaridis, K., Daskalakis, N., Kanakidou, M., Adams, P. J., Artaxo, P., Bahadur, R., Balkanski, Y., Bauer, S. E., Bellouin, N., Benedetti, A., Bergman, T., Berntsen, T. K., Beukes, J. P., Bian, H., Carslaw, K. S., Chin, M., Curci, G., Diehl, T., Easter, R. C., Ghan, S. J., Gong, S. L., Hodzic, A., Hoyle, C. R., Iversen, T., Jathar, S., Jimenez, J. L., Kaiser, J. W., Kirkevåg, A., Koch, D., Kokkola, H., Lee, Y. H., Lin, G., Liu, X., Luo, G., Ma, X., Mann, G. W., Mihalopoulos, N., Morcrette, J.-J., Müller, J.-F., Myhre, G., Myriokefalitakis, S., Ng, N. L., O'Donnell, D., Penner, J. E., Pozzoli, L., Pringle, K. J., Russell, L. M., Schulz, M., Sciare, J., Seland, Ø., Shindell, D. T., Sillman, S., Skeie, R. B., Spracklen, D., Stavrakou, T., Steenrod, S. D., Takemura, T., Tiitta, P., Tilmes, S., Tost, H., van Noije, T., van Zyl, P. G., von Salzen, K., Yu, F., Wang, Z., Wang, Z., Zaveri, R. A., Zhang, H., Zhang, K., Zhang, Q., and Zhang, X.: The AeroCom evaluation and intercomparison of organic aerosol in global models, Atmos. Chem. Phys., 14, 10845–10895, doi:10.5194/acp-14-10845-2014, 2014.

Tsigaridis, K., and Kanakidou, M.: The Present and Future of Secondary Organic Aerosol Direct Forcing on Climate, Current Climate Change Reports, 2018, Volume 4, Issue 2, pp 84–98, 2018.

Turpin, B. J. and Lim, H. J.: Species contributions to PM2.5 mass concentrations: Revisiting common assumptions for estimating organic mass, Aerosol Sci. Tech., 35, 602–610, doi:10.1080/02786820152051454, 2001.

Ulbrich, I. M., Canagaratna, M. R., Zhang, Q., Worsnop, D. R. and Jimenez, J. L.: Interpretation of organic components from Positive Matrix Factorization of aerosol mass

1519 spectrometric data, Atmos. Chem. Phys., 9(9), 2891–2918, doi:10.5194/acp-9-2891-2009,
1520 2009.

1521 van der Werf, G. R., Randerson, J. T., Giglio, L., Collatz, G. J., Mu, M., Kasibhatla, P. S.,
1522 Morton, D. C., DeFries, R. S., Jin, Y., and van Leeuwen, T. T.: Global fire emissions and
1523 the contribution of deforestation, savanna, forest, agricultural, and peat fires (1997–2009),
1524 Atmos. Chem. Phys., 10, 11707-11735, https://doi.org/10.5194/acp-10-11707-2010,
1525 2010.

1526 Vergara-Temprado, J., Murray, B. J., Wilson, T. W., O'Sullivan, D., Browse, J., Pringle, K.
1527 J., Ardon-Dryer, K., Bertram, A. K., Burrows, S. M., Ceburnis, D., DeMott, P. J., Mason,
1528 R. H., O'Dowd, C. D., Rinaldi, M., and Carslaw, K. S.: Contribution of feldspar and marine
1529 organic aerosols to global ice nucleating particle concentrations, Atmos. Chem. Phys., 17,
1530 3637-3658, https://doi.org/10.5194/acp-17-3637-2017, 2017.

1531 Wang, J., Cubison, M. J., Aiken, A. C., Jimenez, J. L., and Collins, D. R.: The importance
1532 of aerosol mixing state and size-resolved composition on CCN concentration and the
1533 variation of the importance with atmospheric aging of aerosols, Atmos. Chem. Phys., 10,
1534 7267-7283, https://doi.org/10.5194/acp-10-7267-2010, 2010.

1535 Williamson, C. J., Kupc, A., Axisa, D., Bilsback, K.R., Bui, T., Campuzano-Jost, P.,
1536 Dollner, M., Froyd, K., Hodshire, A. L., Jimenez, J. L., Kodros, J. K., Luo, G., Murphy, D.
1537 M., Nault, B. A., Ray, E. A., Weinzierl, B. B., Wilson, J. C., Yu, F., Yu, P., Pierce, J.F.,
1538 Brock C. A.: A Large Source of Cloud Condensation Nuclei from New Particle Formation
1539 in the Tropics, Nature, 574, https://doi.org/10.1038/s41586-019-1638-9, 2019.

1540 Wofsy, S.C., S. Afshar, H.M. Allen, E. Apel, E.C. Asher, B. Barletta, J. Bent, H. Bian, B.C.
1541 Biggs, D.R. Blake, N. Blake, I. Bourgeois, C.A. Brock, W.H. Brune, J.W. Budney, T.P. Bui,
1542 A. Butler, P. Campuzano-Jost, C.S. Chang, M. Chin, R. Commane, G. Correa, J.D.
1543 Crounse, P. D. Cullis, B.C. Daube, D.A. Day, J.M. Dean-Day, J.E. Dibb, J.P. DiGangi,
1544 G.S. Diskin, M. Dollner, J.W. Elkins, F. Erdesz, A.M. Fiore, C.M. Flynn, K. Froyd, D.W.
1545 Gesler, S.R. Hall, T.F. Hanisco, R.A. Hannun, A.J. Hills, E.J. Hintsa, A. Hoffman, R.S.
1546 Hornbrook, L.G. Huey, S. Hughes, J.L. Jimenez, B.J. Johnson, J.M. Katich, R.F. Keeling,
1547 M.J. Kim, A. Kupc, L.R. Lait, J.-F. Lamarque, J. Liu, K. McKain, R.J. Mclaughlin, S.
1548 Meinardi, D.O. Miller, S.A. Montzka, F.L. Moore, E.J. Morgan, D.M. Murphy, L.T. Murray,
1549 B.A. Nault, J.A. Neuman, P.A. Newman, J.M. Nicely, X. Pan, W. Paplawsky, J. Peischl,
1550 M.J. Prather, D.J. Price, E. Ray, J.M. Reeves, M. Richardson, A.W. Rollins, K.H. Rosenlof,

T.B. Ryerson, E. Scheuer, G.P. Schill, J.C. Schroder, J.P. Schwarz, J.M. St.Clair, S.D. Steenrod, B.B. Stephens, S.A. Strode, C. Sweeney, D. Tanner, A.P. Teng, A.B. Thames, C.R. Thompson, K. Ullmann, P.R. Veres, N. Vieznor, N.L. Wagner, A. Watt, R. Weber, B. Weinzierl, P. Wennberg, C.J. Williamson, J.C. Wilson, G.M. Wolfe, C.T. Woods, and L.H. Zeng. 2018. ATom: Merged Atmospheric Chemistry, Trace Gases, and Aerosols. ORNL DAAC, Oak Ridge, Tennessee, USA. https://doi.org/10.3334/ORNLDAAC/1581, 2018.

Woody, M. C., Baker, K. R., Hayes, P. L., Jimenez, J. L., Koo, B., and Pye, H. O. T.: Understanding sources of organic aerosol during CalNex-2010 using the CMAQ-VBS, Atmos. Chem. Phys., 16, 4081-4100, https://doi.org/10.5194/acp-16-4081-2016, 2016.

Yu, P., O. B. Toon, C. G. Bardeen, M. J. Mills, T. Fan, J. M. English, and R. R. Neely, Evaluations of tropospheric aerosol properties simulated by the community earth system model with a sectional aerosol microphysics scheme, J. Adv. Model. Earth Syst.,7, 865–914, doi:10.1002/2014MS000421, 2015.

Yu, P., Froyd, K. D., Portmann, R. W., Toon, O. B., Freitas, S. R., Bardeen, C. G., et al.: Efficient in-cloud removal of aerosols by deep convection. Geophysical Research Letters, 46. https://doi.org/10.1029/2018GL080544, 2019.

Zhang, K., O'Donnell, D., Kazil, J., Stier, P., Kinne, S., Lohmann, U., Ferrachat, S., Croft, B., Quaas, J., Wan, H., Rast, S., and Feichter, J.: The global aerosol-climate model ECHAM-HAM, version 2: sensitivity to improvements in process representations, Atmos. Chem. Phys., 12, 8911-8949, https://doi.org/10.5194/acp-12-8911-2012, 2012.

Zhang, Q. J., Beekmann, M., Drewnick, F., Freutel, F., Schneider, J., Crippa, M., Prevot, A. S. H., Baltensperger, U., Poulain, L., Wiedensohler, A., Sciare, J., Gros, V., Borbon, A., Colomb, A., Michoud, V., Doussin, J.-F., Denier van der Gon, H. A. C., Haeffelin, M., Dupont, J.-C., Siour, G., Petetin, H., Bessagnet, B., Pandis, S. N., Hodzic, A., Sanchez, O., Honoré, C., and Perrussel, O., 2013. Formation of organic aerosol in the Paris region during the MEGAPOLI summer campaign: evaluation of the volatility-basis-set approach within the CHIMERE model, Atmos. Chem. Phys., 13, 5767-5790, doi:10.5194/acp-13-5767-2013.

Zhang, X., Cappa, C. D., Jathar, S. H., McVay, R. C., Ensberg, J. J., Kleeman, M. J., and Seinfeld, J. H.: Influence of vapor wall loss in laboratory chambers on yields of secondary organic aerosol, P. Natl. Acad. Sci. USA, 111, 5802–5807, 2014.

1582 Zhao, G., Chen, Y., Hopke, P.K., Holsen, T.M., Dhaniyala, S.: Characteristics of traffic-
1583 induced fugitive dust from unpaved roads, Aerosol Science and Technology, 51:11, 1324-
1584 1331, DOI: 10.1080/02786826.2017.1347251, 2017.

1585 Zhu, J., Penner, J. E., Yu, F., Sillman, S., Andreae, M., and Coe, H.: Organic aerosol
1586 nucleation, climate and land use change: Decrease in radiative forcing, Nature
1587 Communications, 10, Article No. 423, https://www.nature.com/articles/s41467-019-
1588 08407-7, 2019.

1589 Zotter, P., I. El-Haddad, Y. Zhang, P.L. Hayes, X. Zhang, Y.H. Lin, L. Wacker, J. Schnelle-
1590 Kreis, G. Abbaszade, R. Zimmermann, J.D. Surratt, R. Weber, J.L. Jimenez, S. Szidat, U.
1591 Baltensperger, A.S.H. Prévôt. Diurnal cycle of fossil and non-fossil carbon using
1592 radiocarbon analyses during CalNex. Journal of Geophysical Research-Atmospheres,
1593 119, 6818–6835, doi:10.1002/2013JD021114, 2014.

1594

1595   Tables

1596   Table 1: ATom global model configurations and their treatment of the most important processes affecting organic aerosols.

Models & horizontal res. & met. fields & config. reference	Aerosol module	Submicron size[6] OA (dust/sea salt)	SOA precursors[1]					SOA production	Emission	POA/POC (SOA/SOC)	Removal		
			ISO	MT	SQ	ANT	C>12				Standard[2]	Improved	Photolytic
CESM1-CARMA (1.9°lon x 2.5°lat) MERRA-2 (Yu et al. 2019)	20 bins	< 500 nm (< 800 nm)	x	x		x		Semi-volatile using VBS (Pye et al. 2010)	GAIS and GFED v3	1.8 (N/A)	x	For convective updrafts (Yu et al. 2019)[3]	
CESM2-DYN (0.9° lon x 1.25° lat) GEOS5 (Tilmes et al. 2019)	4 modes	< 270 nm (< 800 nm)	x	x	x	x	x	Semi-volatile using VBS (Hodzic et al. 2016)	CMIP6 and QFED v2.4	1.8 (N/A)	x	Water solubility of organic gases per Hodzic et al. (2014)	For SOA (Hodzic et al. 2016)
CESM2-SMP GEOS5 (0.9° lon x 1.25° lat) (Tilmes et al. 2019)	4 modes	< 270 nm (< 800 nm)	x	x		x		Non-volatile with prescribed mass yields for all precursors[4]	CMIP6 and QFED v2.4	1.8 (N/A)	x		
ECHAM6-HAM ECHAM6 (1.87°lon x1.87°lat) (Tegen et al. 2019)	7 modes	< 500 nm (< 500 nm)			x			Non-volatile with 15% prescribed mass yields (Dentener et al. 2006)	ECLIPSE[5] and GFAS	1.4 (1.4)	x		
GC12-REF (2° lon x 2.5° lat) GEOS-FP (Bey et al. 2001)	Bulk	Bulk (< 500 nm)	x	x	x	x		Semi-volatile using VBS (Pye et al. 2010); non-volatile isoprene-	CMIP6 and GFED v4	2.1 (N/A)	x	For convective updrafts per Wang et al. 2014	

								SOA (Marais et al. 2016)					
GC12-DYN (2° lon x 2.5° lat) GEOS-FP (Bey et al. 2001)	Bulk	Bulk (< 500 nm)	x	x	x	x	x	Semi-volatile using VBS (Hodzic et al. 2016); non-volatile isoprene-SOA (Marais et al. 2016)	CMIP6 and GFED v4	2.1 (N/A)	x	For convective updrafts (Wang et al. 2014); Water solubility of organic gases (Hodzic et al. 2014)	For SOA (Hodzic et al. 2016)
GC10-TOMAS (5° lon x 4° lat) GEOS-FP (Kodros et al. 2016)	15 bins	< 316 nm (< 316 nm)		x		x		Non-Volatile using 10% mass yields for MT, 0.2 Tg SOA per Tg CO for anthropogenic emissions	EDGAR v4 and GFED v3	1.8 (1.8)	x	For convective updrafts (Wang et al. 2014)	
GEOS5 (0.5°lon x 0.625°lat) MERRA-2 (Bian et al. 2019)	Bulk	bulk (< 1 μm for dust, 500 nm for seasalt)	x	x		x		Non-Volatile, 10% mass yields for all precursors	HTAP and QFED v2.54	1.8 (1.8)	x		

1597 (1) SOA precursors include isoprene (ISO), monoterpenes (MT), sesquiterpenes (SQ), anthropogenics (ANT) including aromatics such as
1598   benzene, toluene and xylene, as well as lumped shorter chain alkanes and alkenes; and higher molecular weight n-alkanes and n-alkenes
1599   (C>12).
1600 (2) Standard removal includes dry deposition and sedimentation, as well as convective and large-scale scavenging of soluble organic gases
1601   and aerosols, and below-cloud scavenging of aerosols.
1602 (3) A sensitivity simulation is performed with CESM1-CARMA without the improved scavenging in convective updrafts.
1603 (4) 5% for lumped C<12 alkanes, 5% for lumped C<12 alkenes, 15% for aromatics, 4% for isoprene, 25% for monoterpenes.
1604 (5) Anthropogenic BC emission are replaced in Russia with the dataset of Huang et al. (2015).
1605 (6) Submicron size range (diameter) used in various models for comparison with the AMS data.

1606 Table 2: Comparison of observed and simulated OA concentrations along ATom-1 and
1607 ATom-2 flights for eight global model simulations and their ensemble. The results of the
1608 model ensemble are also indicated. The statistical indicators are calculated as normalized
1609 mean bias $NMB(\%) = 100 \times \sum_i (M_i - O_i)/\sum_i O_i$; normalized mean error $NME(\%) =$
1610 $100 \times \sum_i |(M_i - O_i)|/\sum_i O_i$; root mean square error $RMSE(\mu g\ m^{-3}) =$
1611 $\sqrt{(1/N)\sum_i (M_i - O_i)^2}$ and correlation coefficient (R^2) between modeled (M_i) and observed
1612 (O_i) data points. The mean of ATom-1 observations is ~0.23 $\mu g\ m^{-3}$ and for ATom-2 is
1613 0.11 $\mu g\ m^{-3}$. Figure S4 shows the normalized mean bias for all individual ATom model
1614 simulations for various latitudinal regions and for both the Atlantic and Pacific basins.

Organic aerosols	Avg.Mod. ($\mu g\ m^{-3}$)	NMB (%)	NME (%)	RMSE ($\mu g\ m^{-3}$)	R^2	Avg.Mod. ($\mu g\ m^{-3}$)	NMB (%)	NME (%)	RMSE ($\mu g\ m^{-3}$)	R^2
Model	ATom-1 scores (August 2016)					ATom-2 scores (February 2017)				
AeroCom-II Ens.	0.400	74.2	127.3	0.560	0.31	0.254	137	175	0.278	0.38
AeroCom-II Sub.[1]	0.335	47.0	111	0.557	0.28	0.242	127	178	0.290	0.27
ATom Ensemble	0.239	-4.5	64.6	0.372	0.66	0.139	23	92.6	0.224	0.48
CESM2-DYN	0.268	4.6	83.7	0.867	0.47	0.140	25.6	111.7	0.317	0.36
CESM2-SMP	0.349	36.3	94.3	0.556	0.51	0.175	57.2	125.4	0.299	0.31
CESM1-CARMA	0.155	-33.2	93.8	0.603	0.12	0.131	22.6	119.6	0.244	0.31
ECHAM6-HAM	0.400	73.6	143.6	0.714	0.24	0.214	100	184.0	0.363	0.23
GC12-DYN	0.142	-32.6	79.4	0.560	0.16	0.174	14.7	96.6	0.312	0.39
GC12-REF	0.122	-43.0	76.5	0.536	0.18	0.147	3.6	96.3	0.292	0.35
GC10-TOMAS	0.218	-14.4	86.5	0.644	0.16	0.313	150.0	223.7	0.537	0.12
GEOS5	0.242	-5.4	86.6	0.975	0.38	0.084	-24.9	86.4	0.268	0.29

1615 (1) This is the subset of AeroCom-II model ensemble that includes only seven
1616 models that are similar to those that are included in the ATom ensemble (either
1617 the same model, or an older model version, or the same aerosol module).
1618 AeroCom-II Sub. incudes CAM5-MAM3, CCSM4-hem, ECHAM5-HAM2,

1619   GEOSChem-APM 8.2, GEOSChem 9, GISS-TOMAS and GMI (see Tsigaridis

1620   et al., 2014 for their description).

1621   Figures:
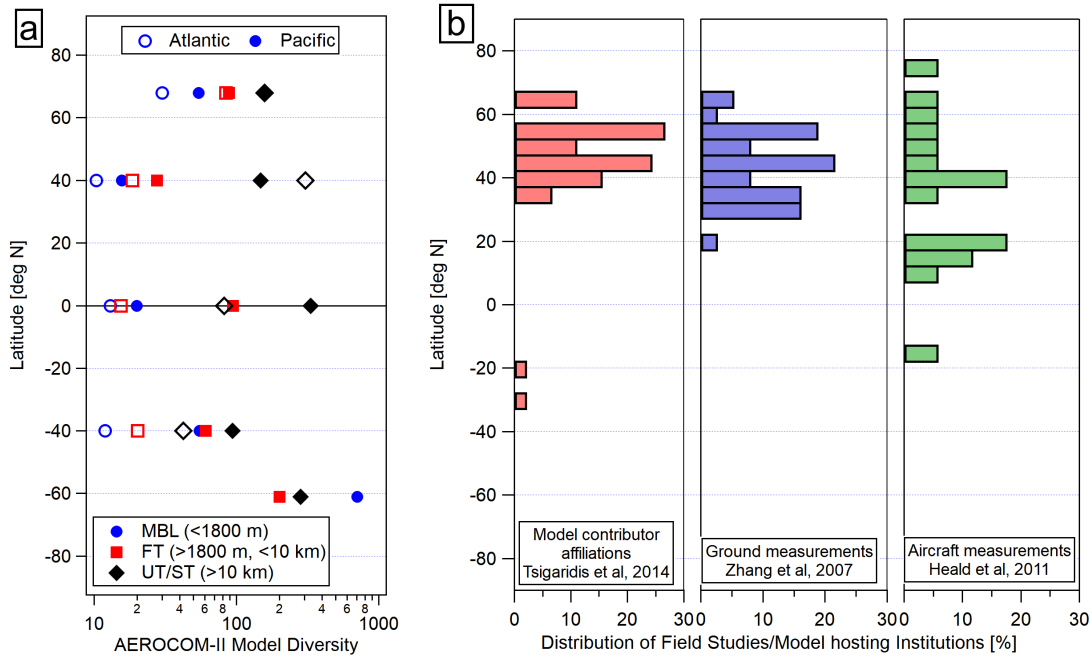
1622

1623   Figure 1: (a, left) The ratio between the average OA concentrations of the highest and the
1624   lowest models (for each region) as predicted among 28 global chemistry transport models
1625   participating in the AeroCom phase II intercomparison study (Tsigaridis et al. 2014); (b,
1626   right) Geographical distribution of institutions at which the AeroCom-II models were
1627   ran/developed (based on author affiliations) and of the field measurements included in two
1628   major literature overview studies (Zhang et al., 2007; Heald et al., 2011) for the OA ground
1629   and aircraft AMS as a function of latitude. For the aircraft campaigns, the average latitude
1630   for the full deployment was taken.

1631

1632

1633

1634

1635
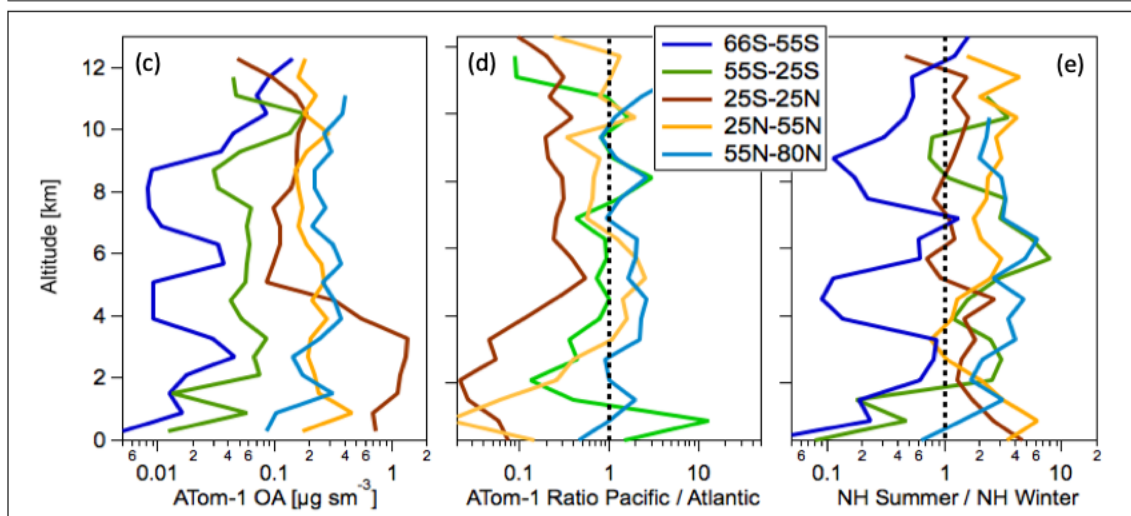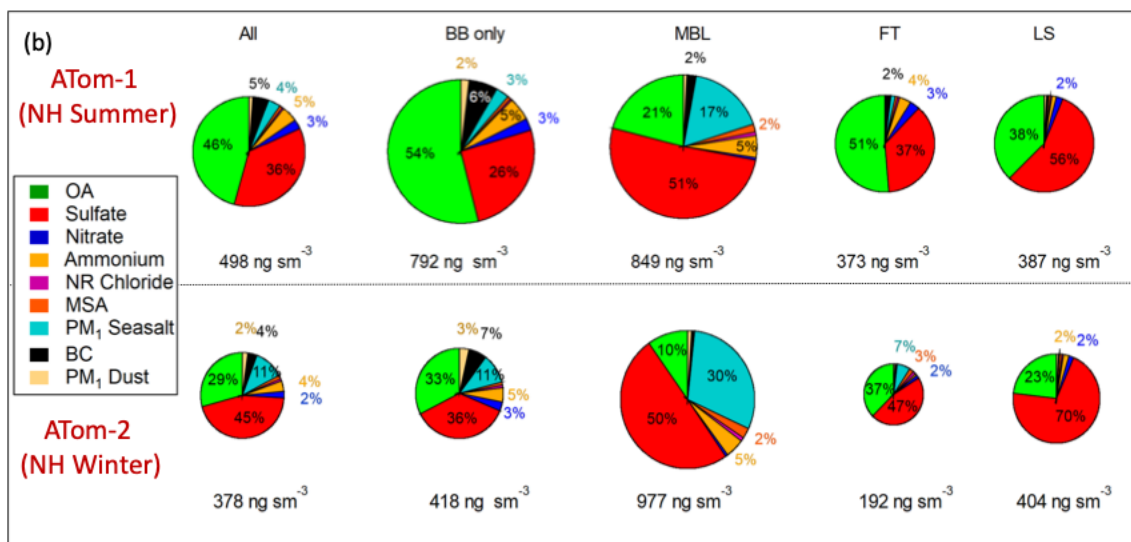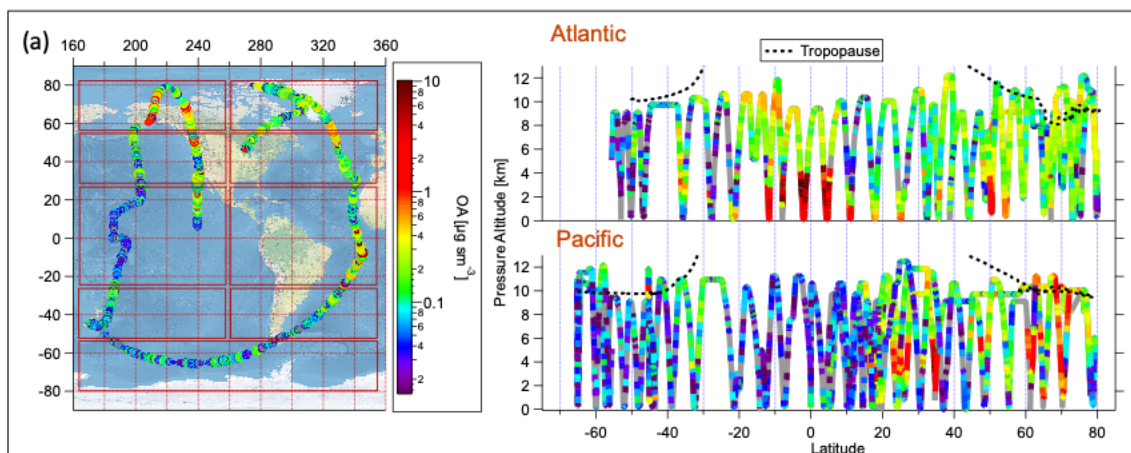
1636

1637   Figure 2: (a, left) ATom-1 DC-8 flights during the August 2016 deployment. Red boxes

1638   indicate regions used for the latitude averaging of the model results. (a, right) Vertical

1639   distribution of OA concentrations (μg sm^{-3}) along ATom-1 flight tracks (b) Average

1640   submicron aerosol composition as measured in the biomass-burning influenced regions

1641   (BB only), and the non-BB influenced regions including the marine boundary layer (MBL),

1642   free troposphere (FT), and lower stratosphere (LS) for ATom-1 (upper plots) and ATom-2

1643   (lower plots). The BB influenced airmasses were filtered using the PALMS data (see

1644   section 3.1). Contributions below 2% are shown but not labeled on the pie chart graph. In

1645   ATom-1, BB-only represents 24% of the data, clean MBL 8%, clean FT 57% and clean

1646   UT 12%, whereas in ATom-2 BB-only represents 3%, clean MBL 8%, clean FT 74%, clean

1647   UT 16%. (c) The average OA vertical profiles are shown for each latitude region as well

1648   as (d) the ratios between the Pacific and Atlantic Oceans in each region. (e) The seasonal

1649   contrast in OA concentrations as calculated as the ratio in OA concentrations between the

1650   NH summer (ATom-1) and NH winter (ATom-2) campaigns. The corresponding plots for

1651   ATom-2 can be found in Fig. S1.
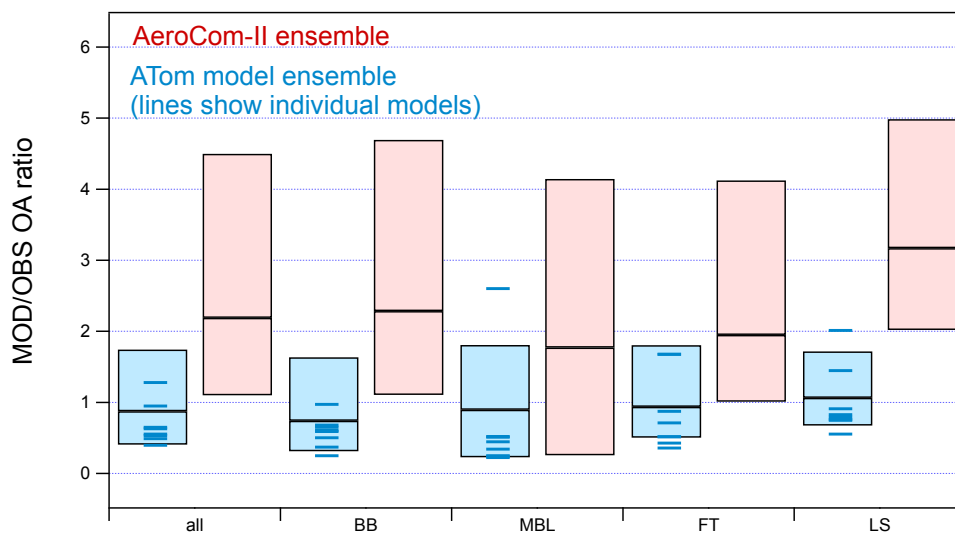
1652

1653   Figure 3: Ratios between predicted and observed OA concentrations for all ATom-1 flights

1654   as calculated for the ATom and AeroCom-II model ensembles in different regions ("BB"

1655   biomass burning influenced regions; "MBL" clean marine boundary layer; "FT" clean free

1656   troposphere' and "LS" lower stratosphere). Median of the ensemble ratio is shown as a

1657   horizontal line, while the boxes indicate 25th and 75th percentiles. Medians for the

1658   individual models included in the current ATom model ensemble are also shown as blue

1659   lines.
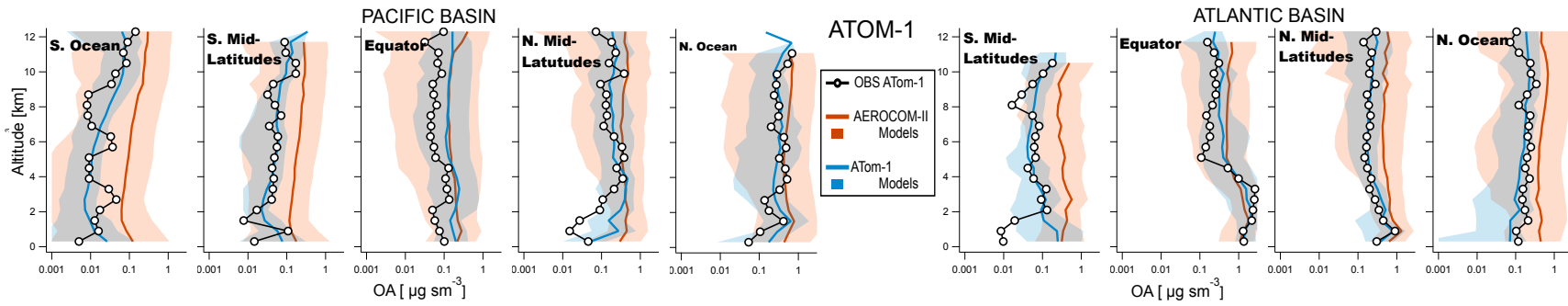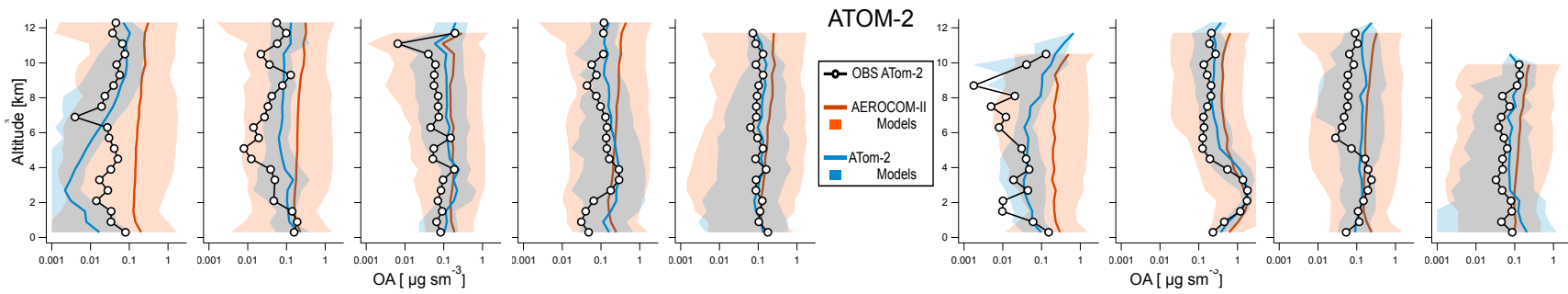
Figure 4: Comparison of latitude-averaged predicted OA vertical profiles with ATom-1 and -2 measurements taken over the Pacific (left side) and Atlantic (right side) basins. Results of the AeroCom-II model ensemble average are shown in red while those of the ATom model ensemble are shown in blue. Shaded areas indicate the variability (two standard deviations) within each model ensemble.
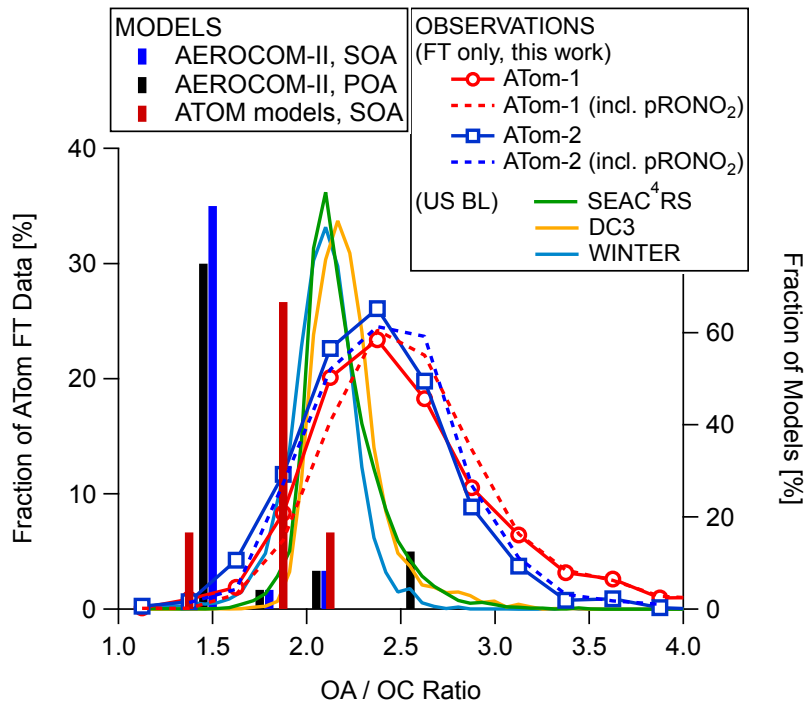
Figure 5: Distribution of the OA / OC ratio as measured during ATom-1 and -2. Values for the recent aircraft campaigns (SEAC4RS, DC3 and WINTER) that took place over continental US regions closer to continental source regions are also shown (Schroder et al., 2018). The bars (right axis) show the OA/OC used for SOA and POA by the models included in the AeroCom and ATom ensemble, with OA/OC=1.4 being the modal value for the former and 1.8 for the latter.
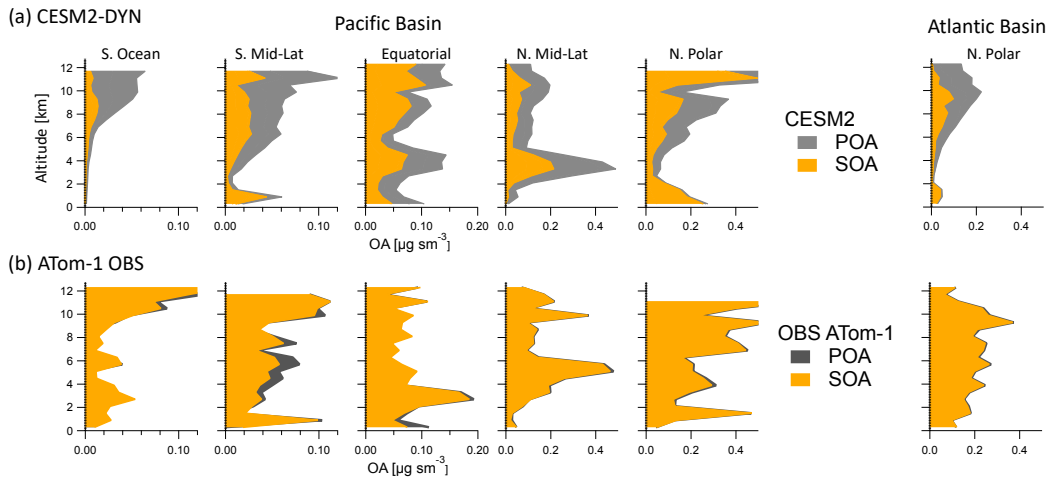
Figure 6: Comparison of averaged POA and SOA vertical profiles as observed during ATom and as predicted by the CESM2-DYN model over the non-BB influenced Pacific and Atlantic basins. The comparison is not shown for the strongly biomass burning influenced regions as all the OA is conservatively allocated to POA in those regions.
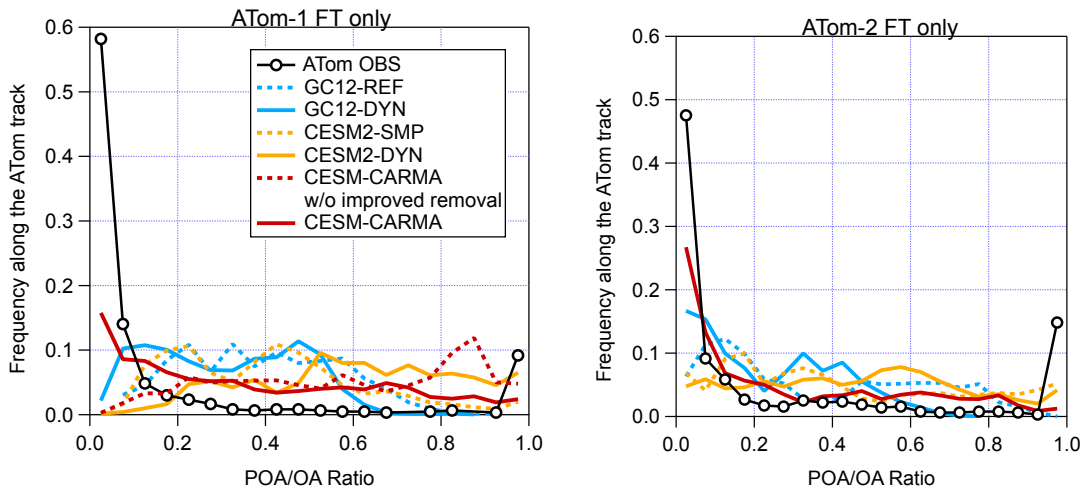
Figure 7: Frequency distribution of observed and simulated ratio of POA to total OA in the free troposphere during ATom-1 and ATom-2 as computed by the GC12-, CESM2-, and CESM1-CARMA models.
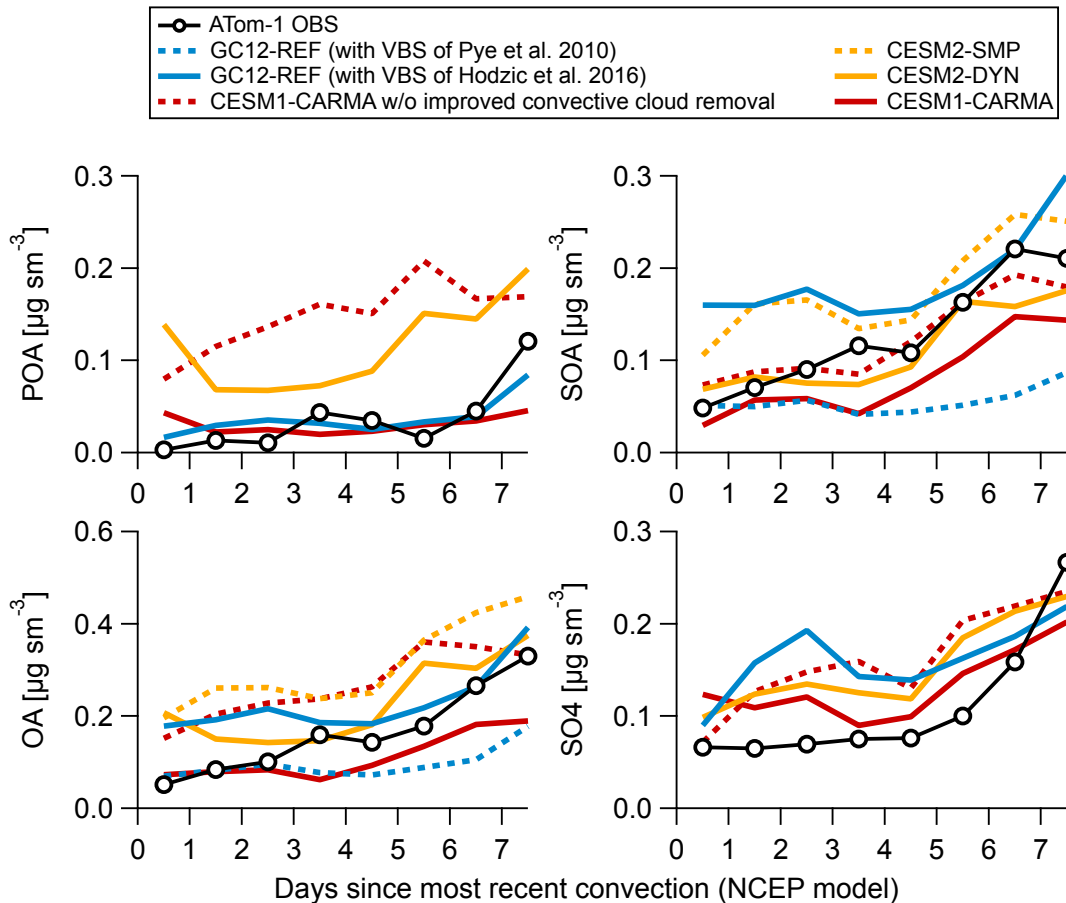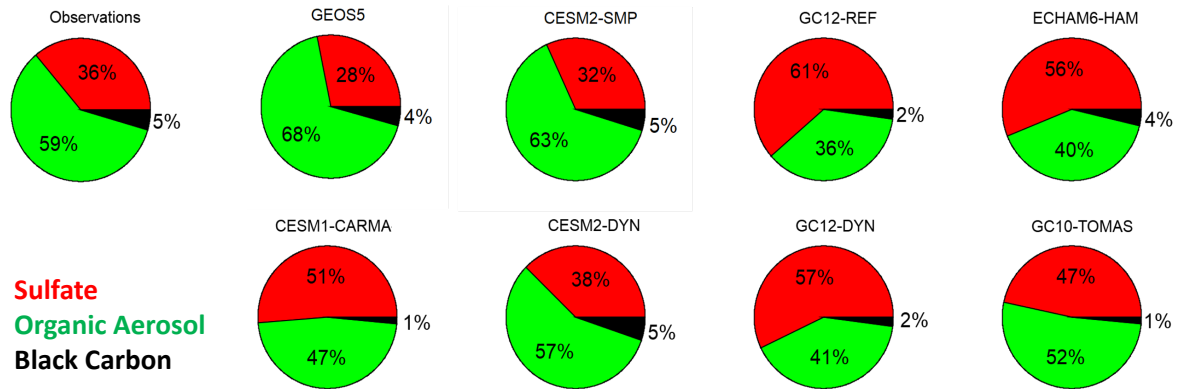
Figure 8: Measured and predicted mass concentrations of POA, SOA, OA and sulfate aerosols during ATom-1 as a function of the number of days since the air mass was processed through convection (based on a trajectory model from Bowman, 1993, and satellite cloud data from NASA Langley, https://clouds.larc.nasa.gov/). CESM2-SMP and CESM2-DYN have the same emissions and processing of POA and sulfate, and thus similar concentrations. The same is true for the two versions of GC12.
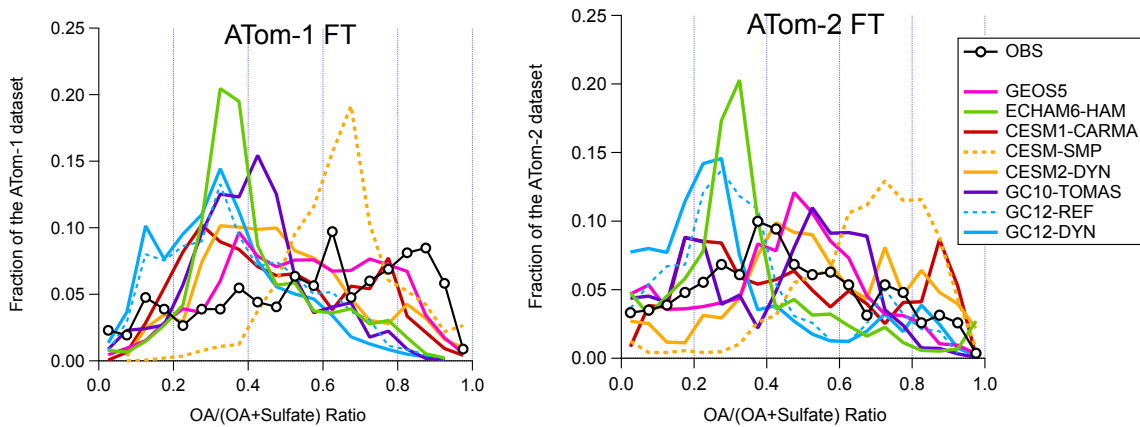
1692   (a)

1693

1694   (b)

1695

1696   Figure 9: (a) Predicted and measured composition of submicron aerosols in the free
1697   troposphere as a function of the submicron aerosol mass concentrations during ATom-1.
1698   (b) Frequency distribution of observed and simulated ratio of organic to organic plus
1699   sulfate aerosols in the free troposphere during ATom-1 and -2.

1700

1701
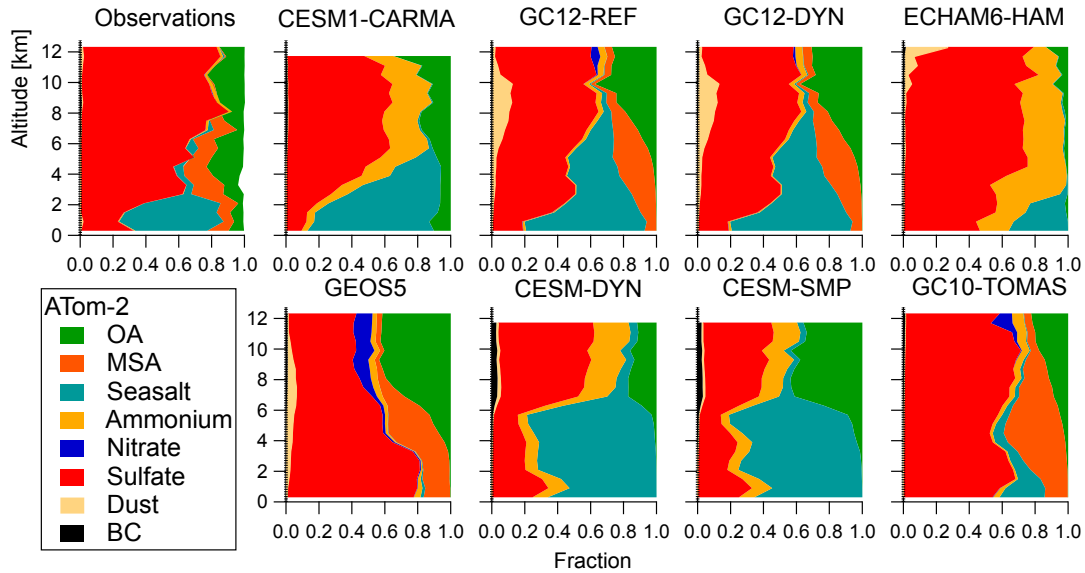
1704   Figure 10: Comparison of measured and predicted composition of submicron aerosols as

1705   a function of altitude over the remote Southern Ocean region during NH Winter (ATom-2).

1706   For models that do not calculate ammonium in the aerosol (such as CESM1-CARMA,

1707   CESM2-SMP, CESM2-DYN and ECHAM6-HAM), ammonium was estimated from the

1708   sulfate mass assuming the formation of ammonium sulfate. Note that while the modeled

1709   and measured submicron sea salt size ranges agree fairly well (Table 1), this is not quite

1710   the case for dust. Given that the accumulation mode dust in the models presented

1711   contains larger sizes than the AMS range (< 500 nm), it is expected for the modeled dust

1712   concentration to be larger than measured.