

Tracer-based investigation of organic aerosols in marine atmospheres from marginal seas of China to the northwest Pacific Ocean

Tianfeng Guo¹, Zhigang Guo¹, Juntao Wang², Jialiang Feng^{3*}, Huiwang Gao^{2,4}, Xiaohong Yao^{2,4*}

¹ Shanghai Key Laboratory of Atmospheric Particle Pollution and Prevention, Department of Environmental Science and Engineering, Fudan University, Shanghai 200433, China;

² Lab of Marine Environmental Science and Ecology, Ministry of Education, Ocean University of China, Qingdao 266100, China

³ School of Environmental and Chemical Engineering, Shanghai University, Shanghai 200444, China

⁴ Pilot National Laboratory for Marine Science and Technology (Qingdao), Qingdao, China

Correspondence to: Xiaohong Yao (xhyao@ouc.edu.cn); Jialiang Feng (fengjialiang@shu.edu.cn)

Abstract. We investigated the geographic distributions of organic tracers in total suspended particles over marginal seas of China, including the Yellow and Bohai seas (YBS) and the South China Sea (SCS), and the northwest Pacific Ocean (NWPO) in spring, when Asian outflows strongly affect downwind marine atmospheres. The comparison of levoglucosan observed in this study with values from the literature showed that the concentrations of biomass burning aerosols over the NWPO increased largely in 2014. More observations together with our snapshot measurement, however, need to confirm whether the large increase occurred continuously through the last decades. The increase led to a mean value of levoglucosan (8.2 ± 14 ng/m³) observed over the NWPO close to that over the SCS (9.6 ± 8.6 ng/m³) and almost half of that over the YBS (21 ± 11 ng/m³). Small geographic differences in monoterpene-derived and sesquiterpene-derived secondary organic tracer concentrations were obtained among the three atmospheres, although the causes may differ. By contrast, a large difference in isoprene-derived secondary organic tracer concentrations was observed among the three atmospheres, with the sum of tracer concentrations over the SCS (45 ± 54 ng/m³) several times and approximately one order of magnitude greater than that over the YBS (15 ± 16 ng/m³) and the NWPO (2.3 ± 1.6 ng/m³), respectively. The geographic distribution of aromatic-derived secondary organic tracers was similar to that of isoprene-derived secondary organic tracers, with a slightly narrower difference, i.e., 1.8 ± 1.7 ng/m³, 1.1 ± 1.4 ng/m³ and 0.3 ± 0.5 ng/m³ over the SCS, the YBS and the NWPO, respectively. We discuss the causes of the distinctive geographic distributions of these tracers and present the tracer-based estimation of organic carbon.

1 Introduction

Aerosols that emanate from biomass burning (BB) consist primarily of carbonaceous components and inorganic salts, which can affect the climate directly by absorbing solar radiation or indirectly by acting as either cloud condensation nuclei (CCN) or ice nuclei (IN) (Bougiatioti et al., 2016; Chen et al., 2017; Hsiao et al., 2016). High BB aerosol emissions zones include boreal forests (e.g., in Eurasia and North America), tropical forests (e.g., in southeast Asia and the tropical Americas), and agriculture areas where crop residuals are burned (e.g., in developing countries such as China and India, etc.) (van der Werf et al., 2006). BB aerosols can undergo

39 long-range transport in the atmosphere, which can carry them from the continents to the oceans (Ding et al.,
40 2013; Fu et al., 2011; Kanakidou et al., 2005). For example, BB aerosols from boreal forest wildfires in Russia
41 and China reportedly made an appreciable contribution to atmospheric particle loads observed over the Arctic
42 Ocean and northwestern Pacific Ocean (NWPO) based on specific tracers of BB (Ding et al., 2013). Although
43 open wildfires from forests occur sporadically in terms of strength and occurrence frequency, global warming
44 could be conducive to vegetation fires (Running, 2006) and thus increase emissions of BB aerosols. In this
45 century, nine years were among the ten hottest global years on record, with 2014–2018 being ranked as the top
46 five hottest years (<https://www.climatecentral.org/gallery/graphics/the-10-hottest-global-years-on-record>). The
47 question is automatically raised: how do BB aerosols in the marine atmosphere in the hottest global years
48 change against those observations previously reported?

49 In addition to BB aerosols, secondary oxidation of biogenic volatile organic compounds (BVOCs) and
50 anthropogenic VOCs (AVOCs) also contribute to the particulate carbonaceous components of marine
51 atmospheres (Kanakidou et al., 2005). Secondary organic aerosols (SOAs) arising from the oxidation of
52 phytoplankton-derived isoprene have been argued to affect the chemical composition of marine atmospheric
53 aerosols and consequently impact CCN loading and cloud droplet number concentrations (Ekström et al., 2009;
54 Meskhidze and Nenes, 2006), but the importance of the marine isoprene-derived SOA is still debated (Arnold et
55 al., 2009; Claeys et al., 2010; Gantt et al., 2009; Guenther et al., 1995). For example, Gantt et al. (2009)
56 estimated the contribution of marine isoprene-derived SOA to the OC in marine atmospheric particles ranged
57 from <0.2% on a global scale, but to as high as 50% (sub-micron OC) over the vast regions of the oceans during
58 the midday hours when isoprene emissions are highest. Several modeling studies have shown that the NWPO
59 may experience the greatest increases in sea surface temperature and CO₂ input under a future warming climate
60 (John et al., 2015; Lauvset et al., 2017). The Kuroshio Extension current system leads the NWPO to be an active
61 subtropical cyclone basin, promoting biogenic activities (Hu et al., 2018). From the perspective of global change,
62 it is a long-term need to study the dynamic changes in atmospheric aerosols derived from marine sources over
63 the NWPO and adjacent marginal seas of China, as well as their potential effects on climate.

64 More importantly, BVOCs emitted from continental ecosystems and their oxidation products can significantly
65 affect the atmosphere in remote marine areas through long-range transport (Ding et al., 2013; Fu et al., 2011; Hu
66 et al., 2013a; Kang et al., 2018; Kawamura et al., 2017). BVOCs consist primarily of isoprene, monoterpenes,
67 sesquiterpenes, and their oxygenated hydrocarbons such as alcohols, aldehydes, and ketones (Ehn et al., 2014;
68 Guenther et al., 2006) and account for the majority of the global VOC inventory (Heald et al., 2008; Zhu et al.,
69 2016a, b). However, emission fluxes and oxidation processes of BVOCs show great variation, depending on
70 global warming and other factors such as regional landscape, other pollutants in the ambient air, etc. (Ait-Helal
71 et al., 2014; Claeys et al., 2004; Hu and Yu, 2013; Peñuelas and Staudt, 2010). Unlike a potential increase in
72 BVOC-derived organics aerosols in marine atmospheres under global warming, anthropogenic VOCs and
73 carbonaceous particles over the continents have been decreased because of effective mitigation of air pollutants
74 in the last decades (Li et al., 2019; Murphy et al., 2011; Sharma et al., 2004; Zhang et al., 2012). In the northern
75 hemisphere, marine atmospheres are also usually affected by anthropogenic pollutants to some extent, most of
76 which are derived from long-range transport from continents (Bao et al., 2018; Kang et al., 2019; Zhang et al.,
77 2017). The reverse trends in BVOC and anthropogenic VOC would change the composition, sources of
78 carbonaceous particles in marine atmospheres. Updated observations are thereby needed to reveal the change
79 and service the future study of the impacts.

80 In this study, we determined the concentrations of some typical organic tracers in aerosol samples obtained from
81 three cruise campaigns from the marginal seas of China, including in the South China Sea (SCS) in 2017,
82 Yellow Sea and Bohai Sea (YBS), to the NWPO in 2014, both in springtime. We investigated the influences of

83 BB aerosols from continents over three marine atmospheres, quantified the contributions of various precursors
84 to the observed SOA in marine atmospheres using organic tracers established in the literature, and explored the
85 formation pathways of SOA from their precursors during long-range transport in these hottest global years.
86 Particularly, we conducted a comprehensive comparison of this observation with those reported in literature in
87 terms of long-term variations and geographic distributions of these tracers, etc.

88 **2 Materials and Methods**

89 Total suspended particulate (TSP) samples were collected over the NWPO from 19 March to 21 April 2014,
90 over the YBS from 30 April to 17 May 2014, and over the SCS from 29 March to 4 May 2017. All samples were
91 collected on the upper deck of the R/V Dong Fang Hong II, which sits ~8 m above the sea surface. To avoid
92 contamination from the ship's exhaust, samples were collected only when the ship was sailing, and the wind
93 direction ranged from -90° to 90° relative to the bow. TSP samples were collected on quartz fiber filters
94 (Whatman QM-A) that had been pre-baked for 4 h at 500°C prior to sampling using a high-volume sampler
95 (KC-1000, Qingdao Laoshan Electric Inc., China). The sampling duration was 15–20 h at a flow rate of ~ 1000 L
96 /min. After sampling, the sample filters were wrapped in baked aluminum foil and sealed in polyethylene bags,
97 then stored at -20°C and transported to the laboratory. Field blanks were collected during each sampling period.
98 However, one sampler was out of service during the cruise on the SCS. As a compromise, cellulose filters
99 (Whatman 41) previously intended for elemental analyses were used for analyses of the organic tracers in TSP.
100 The method for determining the concentrations of tracers was adapted from Kleindienst et al. (2007) and Feng et
101 al. (2013). Briefly, 20 mL dichloromethane/methanol (1:1, v/v) was used for ultrasonic extraction of 40 cm^2
102 of each filter at room temperature three times. The combined extracts were filtered, dried under a gentle stream of
103 ultrapure nitrogen, and then derivatized with 100 μL N,O-bis-(trimethylsilyl)-trifluoroacetamide (BSTFA,
104 containing 1% trimethylchlorosilane as a catalyst) and 20 μL pyridine at 75°C for 45 min. Gas chromatography
105 mass spectrometry (GC-MS) analyses were conducted with an Agilent 6890 GC/5975 MSD. Prior to solvent
106 extraction, methyl- β -D-xylanopyranoside (MXP) was spiked into the samples as an internal/recovery standard.
107 Hexamethylbenzene was added prior to injection as an internal standard to check the recovery of the surrogates.
108 Like those reported by Feng et al. (2013), the primary organic tracers analyzed in this study included
109 levoglucosan (LEVO), mannosan, and galactosan. Four types of secondary organic tracers were used:
110 isoprene-derived secondary organic tracers (SOA_I) including 2-methylglyceric acid (2-MGA), C5-alkene triols
111 (cis-2-methyl-1,3,4-trihydroxy-1-butene, 3-methyl-2,3,4-trihydroxy-1-butene and
112 trans-2-methyl-1,3,4-trihydroxy-1-butene), and MTLs (2-methylthreitol and 2-methylerythritol);
113 monoterpene-derived secondary organic tracers (SOA_M) including 3-hydroxyglutaric acid (HGA),
114 3-hydroxy-4,4-dimethylglutaric acid (HDMGA), and 3-methyl-1,2,3-butanetricarboxylic acid (MTBCA); the
115 sesquiterpene-derived secondary organic tracer (SOA_S) β -caryophyllinic acid; and the aromatic
116 (toluene)-derived secondary organic tracer (SOA_A) 2,3-dihydroxy-4-oxopentanoic acid (DHOPA). LEVO was
117 quantified based on authentic standards in this study. While the SOA tracers without available commercial
118 standards were quantified using methyl- β -D-xylanopyranoside (MXP) as a surrogate. To reduce the uncertainty
119 of quantification, relative response factors of the target tracers to MXP were estimated by comparing the area
120 ratio of typical target ions to MXP to that of total ions in selected samples that showed high concentrations of
121 the target tracers (Feng et al., 2013).
122 Field blanks and laboratory blanks (ran every 10 samples) were extracted and analyzed in the same manner as
123 the ambient samples. Target compounds were nearly always below the detection limit in field and laboratory
124 blanks. Recoveries of the surrogate (MXP) were in the range of 70–110%. The reported results were corrected
125 for recovery, assuming that the target compounds had the same recovery as the surrogate. Duplicate analyses

126 indicated that the deviation was less than 15%.
127 However, the substitution of cellulose filters (Whatman 41) during the cruise on the SCS led to increased field
128 blank values for some tracers. The tracer concentrations in those samples were, however, over three times higher
129 than the field blank values, except for those of mannosan and galactosan. Data for mannosan and galactosan
130 were thus not available, nor were the total organic carbon concentrations, for samples collected during the cruise
131 on the SCS.

132 The concentrations of organic carbon (OC) and element carbon (EC) in each sample were analyzed with a DRI
133 2001A thermal/optical carbon analyzer (Atmoslytic Inc., Calabasas, CA, USA) using the IMPROVE
134 temperature program (Wang et al., 2015). All filters before and after sampling were weighted at a glovebox
135 under controlled ambient temperature and relative humidity. Mass concentrations of TSP, however, should be
136 treated as semi-quantitative results by considering analytic errors of quartz fiber filters (Yao et al., 2009).

137 3. Results and Discussion

138 3.1 Spatiotemporal distributions of LEVO

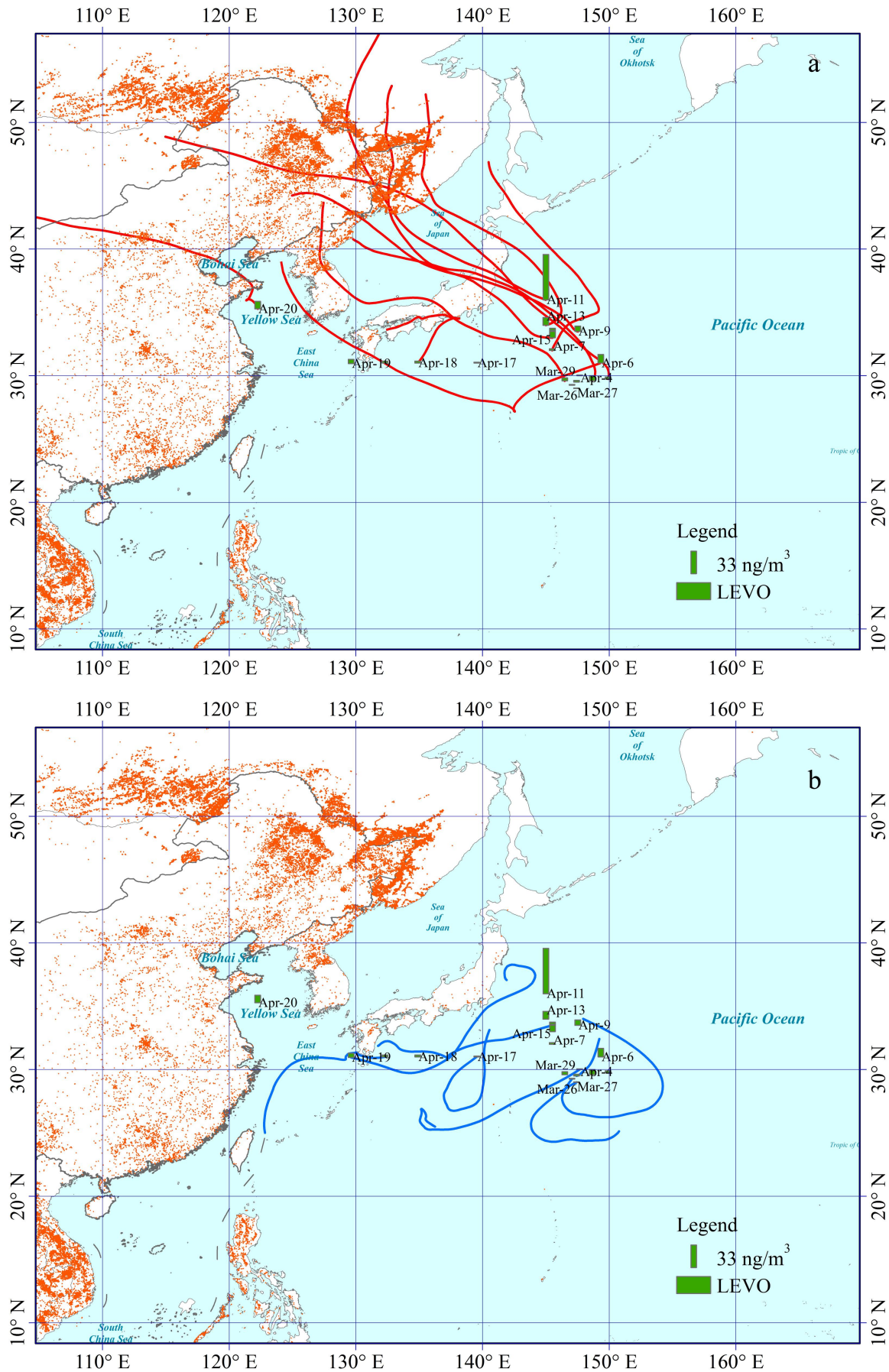
139 Levoglucosan, mannosan, and galactosan produced by the pyrolysis of cellulose and hemicellulose have been
140 widely used as organic tracers of BB aerosols in ambient air (Ding et al., 2013; Fu et al., 2011; Feng et al.,
141 2013). The mean levels of LEVO in TSP collected during the cruises on the NWPO and the SCS were
142 comparable, at 8.2 ng/m³ and 9.6 ng/m³, respectively (Figure S1, Table 1). They were almost half of the mean
143 value of 21 ng/m³ during the cruise on the YBS, where high concentrations of BB aerosols have been observed
144 in continental atmospheres upwind of the YBS mainly from wildfires and the burning of crop residue, wildfire,
145 etc. (Feng et al., 2012; Feng et al., 2013; Yang et al., 2014). Unlike the small difference among the mean values,
146 the concentration of LEVO fluctuated greatly among TSP samples in each oceanic zone, ranging from 0.5 to 65
147 ng/m³ over the NWPO, from 1.0 to 30 ng/m³ over the SCS and from 2.5 to 42 ng/m³ over the YBS (Fig. S1).
148 High spatiotemporal variation in LEVO in TSP has also been observed in literature, with concentrations of
149 LEVO fluctuating around 0.2–41 ng/m³ during Arctic to Antarctic cruises from July to September 2008 and
150 from November 2009 to April 2010 (Hu et al., 2013b). Hu et al. (2013b) also reported the highest LEVO
151 concentrations occurring at mid-latitudes (30°–60° N and S) and the lowest at Antarctic and equatorial latitudes
152 over the several months of sampling. This distinctive geographical distribution was not observed in the present
153 study, as there were no significant differences in LEVO in TSP between the SCS and NWPO ($P > 0.05$).

154 Narrow spatiotemporal variation in LEVO in TSP has been reported during summer sampling over the North
155 Pacific Ocean and the Arctic in 2003, with maximum and mean values as low as 2.1 ng/m³ and 0.5 ng/m³,
156 respectively (Ding et al., 2013). A lower mean value of LEVO of 1.0 ng/m³ has also been reported in the spring
157 over the island of Chichi-jima from 2001 to 2004 (Mochida et al., 2010), while the levels increased to 3.1 ± 3.7
158 ng/m³ in TSP collected on the island of Okinawa in 2009–2012 (Zhu et al., 2015). Using these previous
159 observations as a reference (Table 1), our observations suggest that the BB aerosols from the long-range
160 transport over the NWPO in 2014 largely increased. Thus, an important question is raised, i.e., does the increase
161 occur continuously and largely over the last decades in marine atmospheres over the NWPO? Due to the lack of
162 BB sources in oceans, large spatiotemporal variation in the concentrations of LEVO in the marine atmosphere
163 may be related to the long-range transport of atmospheric particles from continents. Thus, 72 h back trajectories
164 of air masses at a height of 1000 m during our sampling periods (Figs. 1, 2) were calculated using the HYSPLIT
165 model (<https://ready.arl.noaa.gov/HYSPLIT>). Based on the calculated back trajectories, TSP samples could be
166 classified into two categories with Category 1 representing continent-derived aerosol samples and Category 2

167 being ocean-derived aerosol samples. All 12 samples collected over the YBS fell into Category 1 (Fig. 2). Half
168 (11/19) of the samples collected over the NWPO were classified into Category 1 (Fig. 1). A significant
169 difference ($p < 0.05$) was obtained between the concentrations of LEVO in Category 1 ($13 \pm 18 \text{ ng/m}^3$) and
170 Category 2 ($2.0 \pm 1.8 \text{ ng/m}^3$) over the NWPO. The values in Category 2 were closer to the springtime
171 observations reported by Mochida et al. (2010) and Zhu et al. (2015) as well as the summer observations
172 reported by Ding et al. (2013), reflecting the marine background value less affected by continental air masses.
173 On the other hand, the much higher values in Category 1 than Category 2 further indicate a large increase in
174 contribution of BB aerosols being transported from the continents to the remote marine atmosphere in 2014.
175 On 11 April 2014 over the NWPO, an episode of high LEVO concentration of 65 ng/m^3 occurred (Fig. 1). Like
176 LEVO, the concentrations of galactosan and mannosan in the sample were also the highest among all samples
177 collected over the NWPO. This sample was collected in the oceanic zone, approximately 500 km from the
178 continent of Japan. A combination of air mass back trajectories and NASA's FIRMS Fire Map indicated strong
179 BB aerosol emissions from intense fire events in Siberia, followed by long-range transport with the westerly
180 wind as the major contributors to this anomaly (Fig. 1). A similar episodic concentration of LEVO of 27 ng/m^3
181 in TSP was observed once previously over the NWPO during a circumnavigation cruise (Fu et al., 2011). By
182 combining satellite data with other observations, many studies in literature have found that BB aerosols from
183 major forest fires and smoke events in Siberia could be transported downwind to remote marine regions not only
184 in spring, but also in summer (Ding et al., 2013; Generoso et al., 2007; Huang et al., 2009). In a few cases, BB
185 aerosols have been reported to have reached as far as the adjacent Arctic region (Generoso et al., 2007; Warneke
186 et al., 2010). Van der Werf et al. (2006) estimated the emissions of BB aerosols from Eurasia to be much larger
187 than those from North America. Thus, it is not surprising that the concentrations of LEVO over the NWPO were
188 much higher than those over the eastern North Pacific and western North Atlantic at similar latitudes (Hu et al.,
189 2013b).

190 In addition, both galactosan and mannosan showed strong linear correlations with LEVO ($R^2 = 0.98$, $p < 0.05$)
191 in TSP collected over the NWPO and YBS in this study. These strong correlations indicate that the three tracers
192 were probably derived from the same BB sources. Previous studies have reported LEVO/mannosan (L/M) ratios
193 of 3–10, 15–25, and 25–40 from softwood, hardwood, and crop-residue burning, respectively (Kang et al., 2018;
194 Zhu et al., 2015). The calculated L/M ratios in TSP collected over the NWPO were 19 ± 4 in this study, which
195 implies dominant contributions from herbaceous plants and hardwood. The calculated L/M ratios in TSP
196 collected over the YBS were 14 ± 11 , indicating mixed sources.

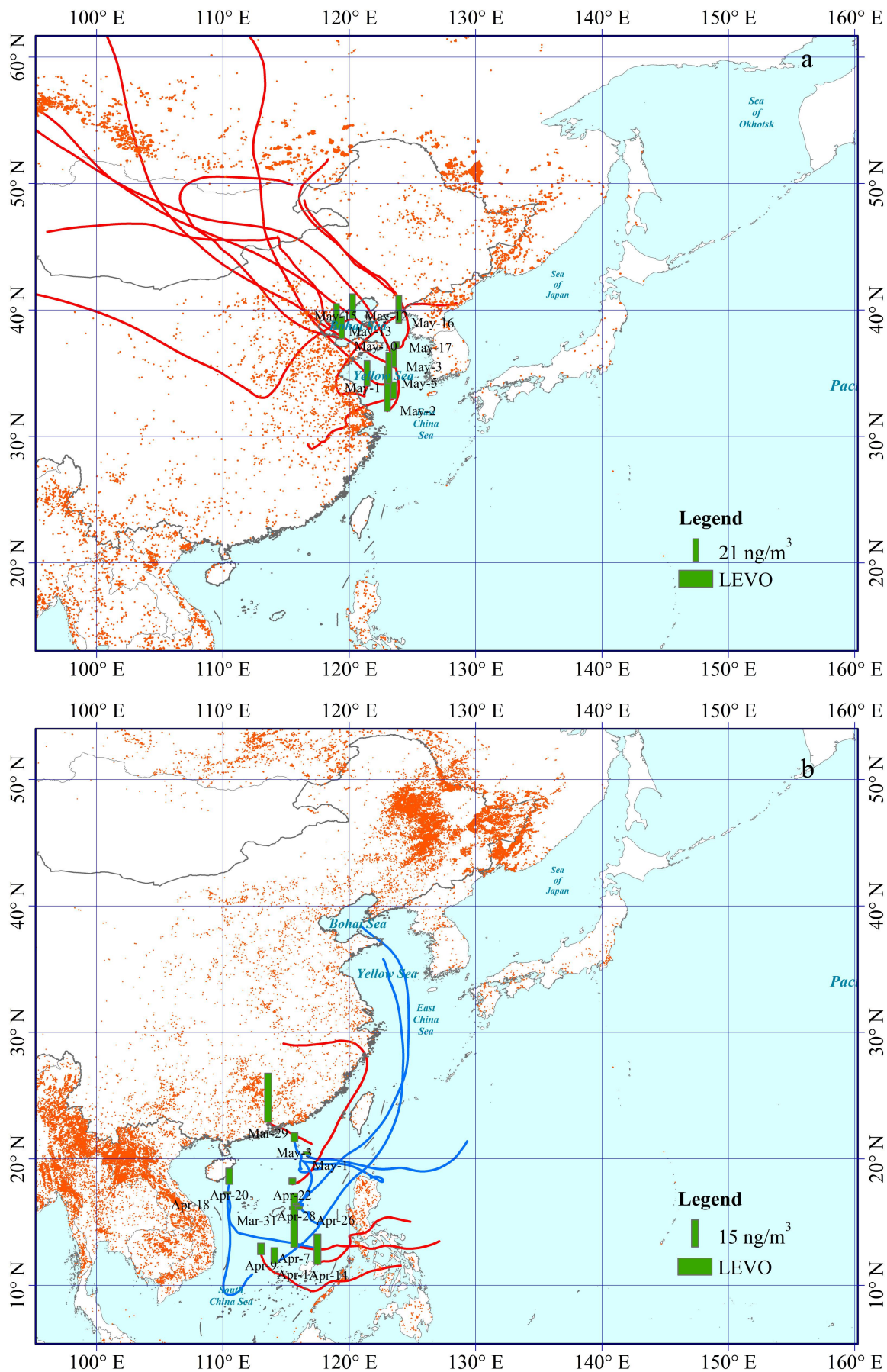
197 In all, 5 of 13 samples collected over the SCS were classified into Category 1, with air masses identified as
198 originating from either the continental areas of South China or the Philippines (Fig. 2). The concentration of
199 LEVO fluctuated around $17 \pm 12 \text{ ng/m}^3$ in Category 1 but decreased to $3.6 \pm 3.4 \text{ ng/m}^3$ in Category 2. However, no
200 significant difference was found between categories due to the large variation in LEVO concentration among the
201 limited number of samples in Category 1 ($p > 0.05$). Forest fires occur accidentally, leading to the large variation
202 in LEVO in Category 1. Southern Asia has been reported to be one of the greatest emission sources of BB
203 aerosols worldwide (van der Werf et al., 2006), which likely led to the higher mean value of LEVO in Category
204 1. However, the LEVO level observed over the SCS in Category 2 was closer to that reported from low-latitude
205 regions ($2.7 \pm 1.1 \text{ ng/m}^3$, Table 1) collected during a global circumnavigation cruise (Hu et al., 2013b). Hu et al.
206 (2013b) argued that their low observed concentrations may have been associated with intense wet deposition,
207 degradation as well as intensive moist convection that occurred in the tropical region during their summer cruise.
208 Unfortunately, no previous observations of LEVO in spring can allow us analyzing the long-term variation in
209 contribution of BB aerosols therein. However, this observation can be used for future comparison.



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Figure 1. Spatial distribution of LEVO in TSP over the NWPO in spring of 2014 and 72-hrs back trajectory associated with each TSP sample. The red lines represent that air masses can be derived from the continent (a, Category 1); the blue lines represent that air masses may be derived mainly from the oceans (b, Category 2). The red dots represent the locations of fires from Fire Information for Resource

215 Management System (FIRMS, <https://firms.modaps.eosdis.nasa.gov/>). And the base map was from
216 Resource and Environment Data Cloud 210 Platform, DOI: 10.12078/2018110201.

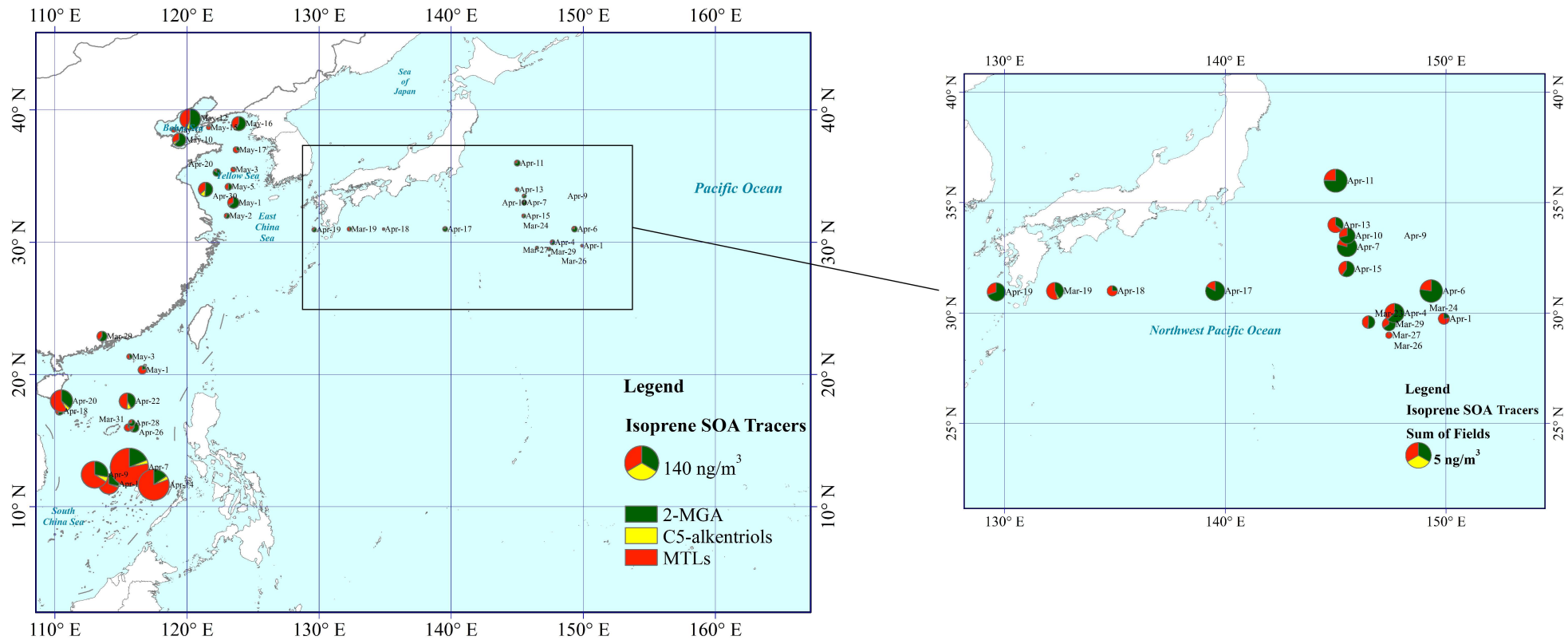


217
218 **Figure 2. Spatial distribution of LEVO over the YBS (a, 2014), and SCS (b, 2017), detailed information**
219 **described in Figure 1. And the base map was from Resource and Environment Data Cloud 210 Platform,**

221 3.2 Spatiotemporal distributions of SOA_I tracers

222 SOA_I tracers were detected during all three cruises. The sum of SOA_I tracers showed a decreasing trend of up to
223 approximately one order of magnitude from marginal seas to the open ocean, i.e., 45 ± 54 ng/m³ in TSP over the
224 SCS, 15 ± 16 ng/m³ over the YBS and 2.3 ± 1.6 ng/m³ over the NWPO (Fig. S1). The highest sum value of SOA_I
225 tracers over the SCS was 176 ng/m³, indicating strong photochemical formation of SOA from biogenic volatile
226 organics (Fig. 3). The geographical distribution of SOA_I tracers in this study was generally consistent with those
227 reported by Hu et al. (2013a), with higher concentrations of these tracers in atmospheric particles collected from
228 low-latitude oceanic zones (30° S–30° N) due to large emissions from tropical forests and strong photochemical
229 reactions. Their reported average contents of SOA_I tracers in low-latitude oceanic zones fluctuated around
230 9.2 ± 6.7 ng/m³, much lower than those measured in this study.

231 When the sum of SOA_I tracers in each sample was examined separately according to the air mass source, a
232 significant difference was found over the SCS between Category 1 (85 ± 66 ng/m³) and Category 2 (19 ± 22
233 ng/m³), with significance at $p < 0.01$. The average contribution of SOA_I tracers to TSP mass concentration over
234 the SYS was higher in category 1 ($0.4\% \pm 0.6\%$) than in category 2 ($0.06\% \pm 0.07\%$). The tracer values were
235 2.7 ± 1.8 ng/m³ in Category 1 and 1.7 ± 1.0 ng/m³ in Category 2 over the NWPO, where no significant difference
236 between the two categories was found ($p > 0.05$). The average contribution of SOA_I tracers to TSP mass
237 concentration over the NWPO was higher in category 1 ($0.008\% \pm 0.005\%$) than that in category 2 ($0.005\% \pm$
238 0.005%). Supposed that concentrations of the tracers in Category 2 were completely contributed by marine
239 sources, it can be inferred that SOA_I carried by continental air masses increased sharply over the SCS. However,
240 it was not the case over the NWPO. Because all samples over the YBS fell into Category 1, this comparison
241 could not be made for the YBS.



242
 243 **Figure 3. Spatial distribution of SOA₁ tracer compounds over three marine regions, YBS and NWPO in 2014, SCS in 2017. The area of the pie indicates the concentration of total**
 244 **SOA₁ tracers. The base map was from Resource and Environment Data Cloud Platform, DOI: 10.12078/2018110201.**

245 3.3 Spatiotemporal distributions of SOA_M, SOA_s tracers

246 The sum of SOA_M tracers including HGA, HD-MGA, and MBTCA was greatest over the SCS region (3.5±6.0
247 ng/m³), where the concentration was approximately double that over the YBS (1.6±2.0 ng/m³) and NWPO
248 regions (1.6±2.7 ng/m³) (Fig. S1), but no significant differences were identified between any two campaigns.
249 The concentrations of SOA_M tracers were almost one magnitude lower than those of SOA_I tracers. Due to the
250 unique contribution of terpene-derived SOA to nucleation and growth of newly formed particles in the
251 atmosphere (Ehn et al., 2014; Gordon et al., 2017; Zhu et al., 2019), the SOA_M may primarily cause indirect
252 climate effects rather than direct effects of aerosols in the marine atmosphere. The difference in mean SOA_M
253 concentration between the SCS and NWPO narrowed to a factor of two, in contrast to the differences of
254 approximately one order of magnitude in mean SOA_I between the two types of atmospheres. The precursors of
255 SOA_M tracers derive mainly from coniferous forests (Duhl et al., 2008) and the decreasing proportion of
256 coniferous forests in subtropical and tropical regions may partially explain the smaller spatial difference in
257 SOA_M tracers over the SCS compared to the YBS and NWPO. However, the comparable SOA_M levels over the
258 YBS and NWPO have not yet been explained.

259 Only three SOA_M tracers were measured in this study, but other SOA_M tracers have been measured and reported
260 in marine atmospheres (Fu et al., 2011; Kang et al., 2018). In order to compare our results with the total amount
261 of SOA_M tracers in the literature, the total amounts measured in this study were multiplied by a factor of 3.1
262 (described in supporting information Sect. S1, Fig. S4) according to the chamber results obtained by Kleindienst
263 et al. (2007). The adjusted values over the SCS were closer to the mean of 11.6 ng/m³ observed over the East
264 China Sea (ECS) (Kang et al., 2018) and the lower values of 9.80–49.0 ng/m³ observed among 12 continental
265 sites in China (Ding et al., 2016). The adjusted total amounts of SOA_M over the NWPO and YBS were
266 comparable to previous observations of 3.0±5.0 ng/m³ collected from the Arctic to Antarctic in 2008-2010 (Hu
267 et al., 2013a), but much higher than observations of 63±49 pg/m³ over the North Pacific and Arctic in 2003
268 (Ding et al., 2013). This may also imply a substantial increase in SOA_M in the last decades, although more
269 investigations are needed to confirm this.

270 β-Caryophyllene is a major sesquiterpene emitted from plants such as Scots pine and European birch (Duhl et al.,
271 2008; Tarvainen et al., 2005). β-Caryophyllinic acid is formed through the ozonolysis or photo-oxidation of
272 β-caryophyllene. The highest levels of β-caryophyllinic acid were observed over the YBS (0.13±0.03 ng/m³),
273 followed by the SCS (0.08±0.11 ng/m³) and NWPO (0.05±0.09 ng/m³) (Fig. S1). The spatial distribution of
274 β-caryophyllinic acid clearly did not follow the general trend of biogenic SOA, with the highest values over the
275 SCS followed by the YBS. Compared to values from the literature, our results are much higher than those over
276 the North Pacific and Arctic Oceans (2.4±5.4 pg/m³) (Ding et al., 2013) but much lower than observations over
277 the East China Sea reported by Kang et al. (2018), where β-caryophyllinic acid was reported to be in the range
278 of 0.16–17.2 ng/m³ with a mean of 2.9 ng/m³. The large differences in β-caryophyllinic acid content observed in
279 various campaigns remain unexplained.

280 3.4 Spatiotemporal distributions of SOA_A tracers

281 When the concentrations of DHOPA in TSP were examined, the highest concentrations occurred over the SCS
282 (1.8±1.7 ng/m³), followed by the YBS (1.1±1.4 ng/m³), and the lowest values were recorded in the NWPO
283 region (0.3±0.5 ng/m³) (Fig. S1). The decreasing extent of the DHOPA from the SCS to the NWPO was
284 approximately three times less than that of SOA_I tracers but approximately three times larger than that of SOA_M

285 tracers. Ding et al. (2017) reported annual averages of DHOPA among various sites in China, which ranged from
286 1.2 to 8.8 ng/m³. The concentrations of DHOPA observed over the SCS and the YBS were similar to the lower
287 values observed in upwind continental atmospheres.

288 Formation of DHOPA depends on the molecular structures of aromatics, as well as concentrations of free
289 radicals and oxidants, etc. (Henze et al., 2008; Li et al., 2016). The mean value of DHOPA in Category 1
290 (0.43±0.65 ng/m³) was nearly twice that in Category 2 (0.20±0.31 ng/m³) over the NWPO ($p > 0.05$). With two
291 samples with high DHOPA (1.2, 2.1 ng/m³) in Category 1 to be excluded, the recalculated average DHOPA
292 decreases down to 0.17±0.21 ng/m³. The continent-derived DHOPA seemingly yielded a minor contribution to
293 the observed values over the NWPO, except during strong long-range transport episodes. Similarly, the mean
294 values of DHOPA were same in Category 1 (1.8±2.1 ng/m³) and Category 2 (1.8±1.5 ng/m³) samples collected
295 over the SCS and no significant difference was observed between two categories. Much stronger UV radiation
296 occurs over the SCS than the YBS, which may contribute to the elevated DHOPA level over the SCS. Aside
297 from continent-derived precursors, oil exploration and heavy marine traffic over the SCS are also potential
298 contributors to the higher DHOPA levels therein, and this link requires further investigation. Previous field
299 observations in China have demonstrated that biofuel or biomass combustion emissions act as important sources
300 of aromatics in the atmosphere (Zhang et al., 2016), as evidenced by the association between the nationwide
301 increase in DHOPA during the cold period and the enhancement of BB emissions (Ding et al., 2017). In this
302 study, no linear correlation was obtained between DHOPA and LEVO in samples collected over the SCS and the
303 other two campaigns, leaving emissions other than BB emissions, e.g., solvent use, oil exploration, marine
304 traffic, etc., as the major precursors for DHOPA in these marine atmospheres (Li et al., 2014).

305 3.5 Causes for high photochemical yields of SOA_I over the SCS

306 Because higher concentrations of SOA_I were observed in TSP samples collected over the SCS, the composition
307 of SOA_I tracers was further investigated in terms of their formation pathways and sources. Based on the results
308 of chamber experiments, Surratt et al. (2010) proposed different formation mechanisms for 2-MGA and MTLs.
309 2-MGA is a C₄-dihydroxycarboxylic acid, which forms through a high-NO_x pathway. MTLs and C₅-alkene
310 triols are mainly products of the photooxidation of epoxydiols of isoprene under low-NO_x conditions.

311 MTLs acted as the dominant compounds among SOA_I tracers in most TSP samples collected over the SCS, with
312 concentrations of 31±42 ng/m³ (Fig. 3). The ratio of 2-MGA/MTLs ranged from 0.2 to 3.1, with a median value
313 of 0.6. The ratio exceeded the unity in only 4 of 13 samples. This result allowed us to infer that the observed
314 SOA_I tracers were generated mainly under low-NO_x conditions. Although the concentration of
315 2-methylerythritol was nearly double that of 2-methylthreitol, they were highly correlated ($R^2 = 0.99$, $p < 0.05$)
316 because of their shared formation pathway. Satellite data showed that the NO₂ levels in South China and the
317 Philippines were low, except in a few hotspots (Fig. S2). Such low-NO_x conditions favor the formation of
318 MTLs rather than 2-MGA over the tropical SCS. The isoprene emitted from plants growing on oceanic islands
319 may also undergo chemical conversion to SOA under low-NO_x conditions, and low-NO_x conditions are always
320 expected in remote marine atmospheres (Davis et al., 2001).

321 In general, zonally and monthly averaged OH concentrations around 15°N are ~50% were greater than those
322 around 35 °N (Bahm and Khalil, 2004). Thus, enhanced formation of MTLs is theoretically expected under the
323 strong UV radiation of tropical regions. However, no significant correlation between the concentrations of
324 MTLs and UV radiation was obtained over the SCS (data not shown) possibly due to the influences of various
325 air masses. A field study showed that MTL yields were positively correlated with ambient temperature in
326 continental atmospheres (Ding et al., 2011). 2-MGA yields, in contrast, showed no significant correlation with

327 ambient temperature in this study. Moreover, lower relative humidity may enhance the formation of 2-MGA in
328 the particulate phase but not for MTLs (Zhang et al., 2011). Variation in ambient temperature and relative
329 humidity may complicate the relationship between the concentrations of SOA_I tracers and UV radiation over the
330 SCS.

331 In addition, the MTLs concentration in Category 1 (62 ± 55 ng/m³) was larger than that in Category 2 (11 ± 14
332 ng/m³). The more abundant MTLs associated with Category 1 was most likely related to long-range transport of
333 these chemicals from upwind continental areas, the oxidation of continental precursors in the marine atmosphere,
334 or both. Large emissions of isoprene were expected from tropical forests upwind of the SCS due to the high
335 vegetation coverage and high ambient temperature of such areas (Ding et al., 2011; Rinne et al., 2002). Global
336 estimates show tropical trees to be responsible for ~80% of terpenoid emissions and ~50% of other VOC
337 emissions (Guenther et al., 2012).

338 In a clean marine atmosphere, phytoplankton is the sole source of isoprene emissions over the oceans (Bonsang
339 et al., 1992; Broadgate et al., 1997). Chlorophyll-a has been widely employed as a measure of phytoplankton
340 abundance and a proxy for predicting isoprene concentrations in water (Hackenberg et al., 2017). The
341 satellite-derived chlorophyll-a level during the study period over the SCS was below 0.45 mg/m³, excluding
342 coastal areas (Fig. S3). The MTLs observation of 11 ± 14 ng/m³ in Category 2 should be considered as the upper
343 limitation value derived from marine phytoplankton in the SCS. Although air masses differed between
344 Categories 1 and 2, a good correlation was obtained between MTLs and 2-MGA when the data in the two
345 categories was pooled for analyses ($R^2 = 0.77$, $P < 0.01$). This strong correlation indicates these tracers are
346 primarily formed through shared pathways. However, this correlation was poor over the NWPO, as discussed
347 below.

348 3.6 Origin and formation of SOA_I over the NWPO

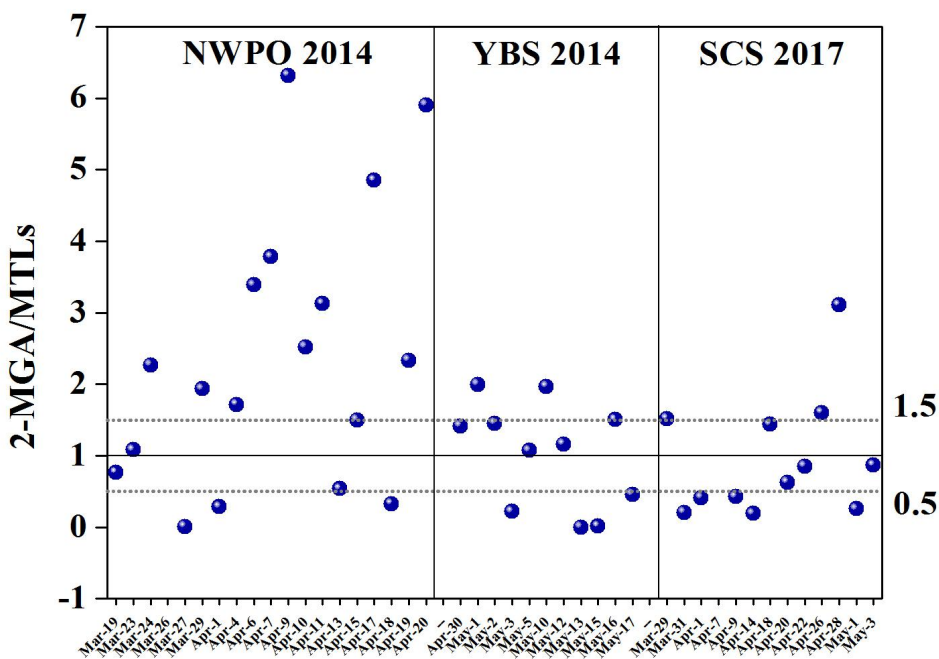
349 Over the NWPO, the concentration of 2-MGA was 1.6 ± 1.5 ng/m³, which was generally dominant among SOA_I
350 tracers, followed by MTLs (0.7 ± 0.3 ng/m³) and C5-alkene triols (0.03 ± 0.02 ng/m³). When the ratio of
351 2-MGA/MTLs was further examined, it varied greatly from <0.1 to 6.3, with a median value of 2.1. Most ratios
352 observed over the NWPO in this study were far greater than the values of 0.18–0.59 reported by Hu et al. (2013a)
353 from a global circumnavigation cruise, and also greater than 0.87–1.8 reported in urban areas of California
354 (Lewandowski et al., 2013) and the maximum value of 2.0 obtained over the YBS. Ding et al. (2013) also
355 reported ratios that fluctuated greatly from 0.5 to 10 with a median value of 3.3 during a summer cruise in the
356 NWPO and Arctic Ocean in 2003. The large 2-MGA/MTL ratios over the NWPO appeared to be highly
357 consistent over two independent sampling campaigns.

358 The compound profile of SOA_I tracers over the NWPO implied high-NO_x conditions allowing oxidation of
359 isoprene to generate the SOA_I present in most samples. Such high-NO_x conditions are impossible in a remote
360 marine atmosphere, as indicted in Figure S2. Given that the lifespan of isoprene in the atmosphere is only
361 several hours (Bonsang et al., 1992), the long-range transport of oxidation products formed under high NO_x
362 levels over the continents likely led to the 2-MGA-dominated composition of SOA_I. Based on air mass back
363 trajectories, this long-range transport may involve 2-MGA originating from Siberia, northeastern China, or
364 Japan.

365 Organic aerosols over the NWPO were strongly influenced by forest fires that take place in Siberia during
366 spring and summer almost every year (Ding et al., 2013; Huang et al., 2009). Previous emissions inventory
367 studies have reported high isoprene and NO_x emissions from various BB types (Akagi et al., 2011; Andreae and
368 Merlet, 2001). Ding et al. (2013) thus argued that an increase in emissions of isoprene in the presence of BB,

369 followed by its chemical conversion under high-NO_x conditions, may lead to transport over thousands of
 370 kilometers and hold at the detectable concentrations in the remote marine atmosphere over the NWPO. The
 371 same argument may hold true for the elevated ratios of 2-MGA/MTLs observed over the NWPO in this study
 372 (Fig. 4). However, we did not find a significant correlation between 2-MGA and LEVO over the NWPO. The
 373 decomposition of LEVO reported in literature (Hennigan et al., 2010; Hoffmann et al., 2010; Fraser and
 374 Lakshmanan, 2000) may lower the correlation between them. However, whether 2-MGA can decompose in
 375 ambient air remains poorly understood.

376 On the other hand, the ratios of 2-MGA/MTLs in 3 of 19 samples collected over the NWPO were below 0.5
 377 (Figure 4). In these cases, the oxidation of isoprene under low-NO_x conditions likely dominated the generation
 378 of SOA_I. The ratios of 2-MGA/MTLs were 0.5–1.5 in 4 of 19 samples, suggesting mixed contributions to SOA_I
 379 from the oxidation of isoprene under low-NO_x conditions and high-NO_x conditions. As the major formation
 380 pathways of 2-MGA and MTLs varied greatly among samples, no significant correlation ($R^2 = 0.12$, $p > 0.05$)
 381 was obtained between 2-MGA and MTLs over the NWPO. Recall that the tracer values of SOA_I were 2.7 ± 1.8
 382 ng/m³ in Category 1 and 1.7 ± 1.0 ng/m³ in Category 2. This implied that SOA_I derived from marine sources was
 383 comparable to that derived from the continent outflows.



384
 385 **Figure 4. Spatial ratio of 2-MGA/MTLs among SOA_I tracers over three marine regions.**

386 3.7 Source apportionment of secondary organic carbon (SOC)

387 The tracer-based approach developed by Kleindienst et al. (2007) was applied to estimate the concentrations of
 388 SOC and WSOC_{BB}, as follows:

$$[SOC] = \frac{\sum_i [tri]}{f_{SOC}} \quad (1)$$

$$[WSOC_{BB}] = \frac{C_{tracer}}{f_{tracer/WSOC_{BB}}} \quad (2)$$

390

391 where $\Sigma_i(\text{tri})$ is the sum of concentrations of the selected suite of tracers for a precursor, and f_{SOC} is the mass
392 fraction of tracer compounds in SOC generated from the precursor in chamber experiments. Assuming that the
393 f_{SOC} values in ambient air match those in the chamber, the f_{SOC} values for precursors such as isoprene,
394 monoterpenes, β -caryophyllene, and aromatics were $0.155 \pm 0.039 \mu\text{g}/\mu\text{gC}$, $0.231 \pm 0.111 \mu\text{g}/\mu\text{gC}$, 0.023 ± 0.0046
395 $\mu\text{g}/\mu\text{gC}$, and $0.00797 \pm 0.0026 \mu\text{g}/\mu\text{gC}$, respectively (Kleindienst et al., 2007), with uncertainty described in
396 Sect. S2. The fraction of LEVO in WSOC ($0.0994 \mu\text{g}/\mu\text{gC}$) from the BB plume was used for WSOC_{BB} (Ding et
397 al., 2008). The f_{SOC} value for monoterpenes was scaled up by a factor of 3.1 based on experimental observations,
398 as these two tracers (HGA+HD-MGA) accounted for 2/9 of the total tracers of monoterpenes, as described in
399 the supporting information (Kleindienst et al., 2007).

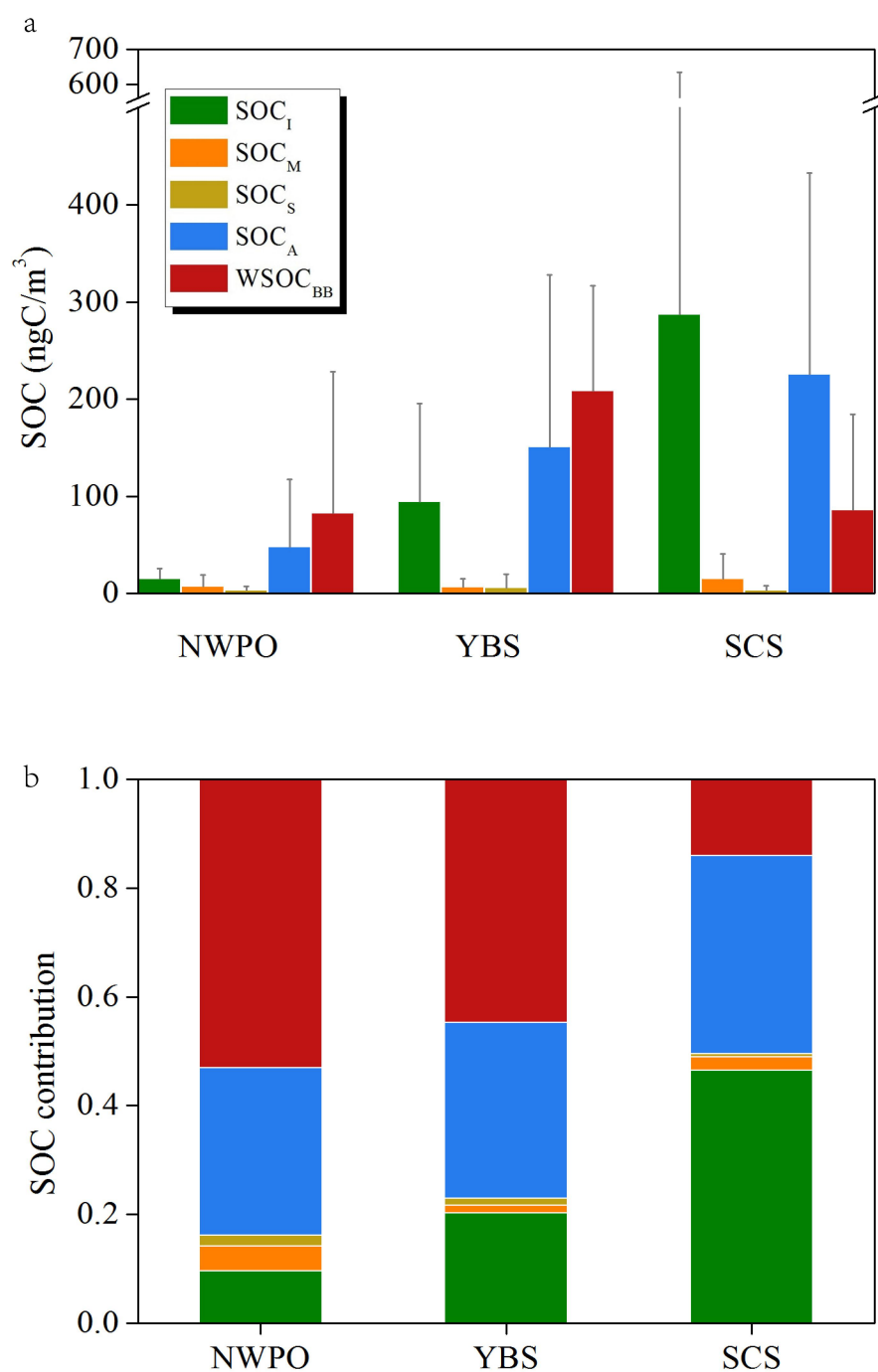
400 Over the SCS, nearly half of the sum of SOC and WSOC_{BB} was in the form of SOC_I (47%), followed by SOC_A
401 (36%), WSOC_{BB} (14%) and a minor contribution of 2.5% from SOC_M (Fig. 5). This composition pattern over
402 the SCS could be attributed to abundant biogenic SOA formation in low-latitude tropical marine atmospheres.
403 Over tropical marine regions, atmospheric oxidation products can account for 47–59% of the total organic
404 content estimated, with biomass burning emissions making up only 2–7% based on source apportionment using
405 organic tracers (Fu et al., 2011). A model study by Fu et al. (2012) showed that secondary formation accounts
406 for as much as 62% of OC estimated using tracers in eastern China in summer. A reverse pattern was observed
407 over the YBS, with WSOC_{BB} as the dominant contributor (45%) to the sum of SOC and WSOC_{BB} , followed by
408 SOC_A (32%) and SOC_I (20%). The contribution of SOC_M was also minor, at 1.5%. Notably, the chemical
409 composition observed over the NWPO was similar to that over the YBS, with WSOC_{BB} contributing up to 53%.
410 In addition, Kang et al. (2018) used the PMF method to identify various sources of OC in marine aerosols over
411 the ECS such as secondary nitrate, BSOA, BB, and fungal spores.

412 Geographically, the estimated SOC values from BVOCs ranked at the highest level of $306 \pm 343 \text{ ngC}/\text{m}^3$ over the
413 SCS, decreasing to $107 \pm 99 \text{ ngC}/\text{m}^3$ over the YBS and $24 \pm 22 \text{ ngC}/\text{m}^3$ over the NWPO. The estimates of
414 aromatic SOC exhibited the same geographic trend, with values of $225 \pm 208 \text{ ngC}/\text{m}^3$ over the SCS, 151 ± 177
415 $\text{ ngC}/\text{m}^3$ over the YBS and $48 \pm 69 \text{ ngC}/\text{m}^3$ over the NWPO. Recent modeling results have also shown that
416 aromatic emissions are the predominant precursors of SOA during springtime in China in comparison with
417 BVOCs and other AVOCs (Han et al., 2016). Among estimates of WSOC_{BB} , the highest values of 209 ± 108
418 $\text{ ngC}/\text{m}^3$ were recorded over the YBS, followed by comparable levels of $86 \pm 98 \text{ ngC}/\text{m}^3$ (SCS) and 83 ± 145
419 $\text{ ngC}/\text{m}^3$ (NWPO).

420 In our study, the calculated WSOC_{BB} estimate accounted for $4.1 \pm 5.0\%$ and $3.3 \pm 1.7\%$ of measured OC over the
421 NWPO and YBS, respectively, and these values are higher than that obtained over the ECS during summer
422 (1.4%) (Kang et al., 2018). Estimated SOC from BVOCs accounted for only $1.5 \pm 1.4\%$ and $1.8 \pm 1.7\%$ to the
423 measured OC over the NWPO and YBS, respectively, which is lower than that over ECS (4.21%) (Kang et al.,
424 2018). However, the mean values obtained in this study were similar to the total SOC level estimated using
425 tracers as a proportion of measured WSOC (4%) during a cruise on the North Pacific and Arctic Oceans,
426 supposed that WSOC accounted for half of the total OC in atmospheric particles (Ding et al., 2013).

427 The calculated SOC level derived from organic tracers accounted for less than 8% of total measured OC in these
428 study areas. However, these SOC compounds are expected to derive mainly from photochemical reactions in the
429 gas phase, followed by gas-aerosol partitioning. These compounds likely play an important role in the growth of
430 newly formed particles alongside pre-existing nucleation mode or Aitken mode particles. However, most organic
431 matter detected in bulk samples may originate from primary sources, heterogeneous reactions and in-cloud
432 processing (Ervens et al., 2011; Kanakidou et al., 2005; Nichols, 2016), and these compounds may be major
433 drivers of the direct climate effects of aerosols, rather than indirect climate effects. In the future, a
434 comprehensive combination measurement of organic tracers and organic matter with an aerosol mass

435 spectrometer should be used to elucidate the formation and growth processes of atmospheric nanoparticles.
436



437
438 **Figure 5. Average SOC levels calculated using the tracer-SOC/WSOC method over three marine regions**
439 **(YBS and NWPO in 2014, SCS in 2017) and their contributions based on five organic tracers measured in**
440 **this study.**

441 **4. Conclusions**

442 This study investigated the geographical distributions of tracer-based organic matter observations in TSP
443 collected over two marginal seas of China and the NWPO in the spring season, when the East Asian monsoon
444 carries biogenic and anthropogenic aerosols over these oceanic zones. We found a significantly large difference

445 in LEVO over the NWPO between two categories of air masses originating from upwind continents or oceanic
446 regions, as Category 1 (continental) contained $13 \pm 18 \text{ ng/m}^3$ and Category 2 (oceanic) had $2.0 \pm 1.8 \text{ ng/m}^3$; the
447 concentrations of LEVO in Category 2 were closer to the low values reported in the literature. This further
448 implied a large increase in continent-derived BB aerosols in marine atmospheres over the NWPO in recent
449 decades, compared to previous studies. An important question is thereby raised, i.e., does a large increase in
450 continent-derived BB aerosols in marine atmospheres over the NWPO occur continuously and largely in recent
451 decades? Combining the L/M ratios of 19 ± 4 over the NWPO with the calculated air mass back trajectories
452 indicates that the increase was very likely associated with enhanced emissions of BB aerosols from wildfires in
453 Siberia and northeastern China. Moreover, the mean level of BB aerosols over the SCS nearly matched that over
454 the NWPO. The contents of LEVO in Category 2 air masses, derived from oceanic zones over the SCS, were
455 comparable with those reported in the literature, but the mean value was only about a quarter of that in Category
456 1, representing air masses from upwind continents. However, the limited data available over the SCS in the
457 literature cannot support inferences about whether BB aerosols emitted from upwind tropical forests have
458 increased in recent decades.

459 The concentrations of SOA_I over the SCS were approximately one order of magnitude greater than those
460 observed over the NWPO and several times larger than those over the YBS. The larger values observed over the
461 SCS in Category 1 than in Category 2 were likely driven by high emissions of isoprene from upwind tropical
462 forests and strong solar radiation. The MTLs dominance of SOA_I over the SCS strongly suggested that SOC
463 from BVOCs was generated primarily under low- NO_x conditions. On the other hand, 2-MGA dominance over
464 the YBS implied that most SOC was generated under high- NO_x conditions. Elevated ratios of 2-MGA/MTLs
465 of >1.5 were obtained for 11 of 19 total samples collected over the NWPO, consistent with those reported in the
466 literature. Larger ratios may be attributed to possible emissions of BVOCs in the presence of BB. However, the
467 comparable concentrations of SOA_I in Category 1 and Category 2 samples collected over the NWPO implied a
468 large contribution of SOA_I from marine sources. The aromatic SOA tracers' levels were highest over the SCS,
469 followed by values obtained over the YBS and NWPO. The high values observed over the SCS may be related
470 to strong solar radiation, but the sources of precursors remain unexplained. Based on the concentrations in
471 Category 1 and 2 air samples collected over the SCS and NWPO, mixed sources of aromatic VOCs should exist,
472 including continent-derived precursors, oil exploration and heavy marine traffic.

473 Over the NWPO and the YBS, the estimated WSOC_{BB} levels were nearly equal to the sum of SOC estimated
474 from the oxidation of aromatics and BVOCs. Over the SCS, SOC estimated from the oxidation of BVOCs was
475 significantly larger than the estimated WSOC_{BB} . The geographical difference may be related to emissions of
476 primary particulate organics and gaseous precursors as well as formation processing of secondary organics in
477 various atmospheres.

478 The atmospheric composition of SOA in different geographical locations is, however, highly complex and is
479 regulated by many factors including local meteorological conditions, anthropogenic emissions, plant species,
480 vegetation cover and regional chemistry, and therefore warrants further quantification and analyses. Particularly,
481 whether BB aerosols and other biogenic organic aerosols in marine atmospheres will continuously increase
482 under warming conditions.

483 **Table 1. Sum of organic tracer contents (ng/m³) at different locations worldwide.**

Site	Date	Sampler	LEVO	SOA _I	SOA _M	SOA _S	SOA _A	Reference
Wakayama, Japan (Forest)	August 20–30, 2010, Day	TSP	2.5±2.1	281±274	54.6±50.2	1.2±1.2		(Zhu et al., 2016a)
	Night		1.1±0.9	199±207	36.3±33.6	0.9±0.8		
Across China	summer 2012	Anderson sampler		123±79	10.5±6.6	5.0±4.0	2.9±1.5	(Ding et al., 2014)
Beijing (PKU) (urban site)	summer 2007	PM2.5	37-148	59±32	30±14	2.7±1.0		(Yang et al., 2016)
Beijing (YUFA) (suburban site)			34-149	75±43	32±14	3.9±1.5		
Shanghai (BS) (Suburban site)	Apr-May 2010	PM2.5	88.8±57.2	3.8±3.9	6.1±3.7	1.0±0.7	1.1±0.7	(Feng et al., 2013)
Shanghai (XJH) (Urban site)			58.3±27.5	2.5±1.7	2.7±1.3	0.4±0.3	0.6±0.4	
Mt. Tai	summer 2014	PM2.5		56.4±45.6	34.4±28.4			(Zhu et al., 2017)
Central Pearl River Delta	fall-winter 2007	PM2.5		30.8±15.9	6.6±4.4	0.5±0.6		(Ding et al., 2011)
Central Tibetan Plateau	2012-2013	Anderson sampler		26.6±44.2	1.0±0.6	0.09±0.1	0.3±0.2	(Shen et al., 2015)
Mumbai, India	winter 2007	PM10		4.1±2.4	29±22		0.6±0.6	(Fu et al., 2016)
	summer 2007			1.1±0.7	9.4±4.7		0.05±0.1	
Alaska	Spring 2009	TSP		2.4	3.6	0.9		(Haque et al., 2016)
	2008-2009	TSP		4.1	2.0	1.5		
SYS	Spring 2017	TSP	9.6±8.6	45±54	3.5±6.0	0.07±0.1	1.8±1.7	This study
YBS	Spring 2014	TSP	21±11	15±16	1.6±2.0	0.1±0.3	1.1±1.4	This study
NWPO	Spring 2014	TSP	8.2±14	2.3±1.6	1.6±2.7	0.05±0.09	0.3±0.5	This study
East China Sea	18 May to 12 June 2014	TSP	0.09–64.3 (7.3)	0.15–64.0 (8.4)	0.26–87.2 (11.6)	0.16–17.2 (2.9)		(Kang et al., 2018)
Arctic to Antarctic	July to September 2008; November 2009 to April 2010	TSP	5.4±6.2	8.5±11	3.0±5.0			(Hu et al., 2013a; Hu et al., 2013b)
North Pacific	2003	TSP		0.5±0.4	0.6±0.4	0.06±0.05	0.002±0.005	(Ding et al., 2013)

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485 **Data availability.** Most of the data are shown in supplement. Other data are available by contacting the
486 corresponding author.

487 **Supplement.** The supplement related to this article is available.

488 **Author contributions.** XY, TG and JF conceived and led the studies. TG, JW and JF carried out the
489 experiments and analyzed the data. TG and JF interpreted the results. ZG, JF, HG discussed the results and
490 commented on the manuscript. TG prepared the manuscript with contributions from all the co-authors.

491 **Competing interests.** The authors declare that they have no conflict of interest.

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495

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