

Response to Anonymous Referees

Response to Anonymous Referee #2

The manuscript reports the spatiotemporal distributions of organic tracers in TSP collected in two marginal seas of China and the NWPO in the spring season and how the East Asian monsoon carries biogenic and anthropogenic aerosols over these oceanic zones. In addition, the authors discussed the origins of SOA_I over the SCS and NWPO. Overall, it is an interesting and inspiring work. However, the follow comments need to be addressed before it can be accepted for publication on ACP.

Response: We thank the reviewer's comments and revise our manuscript accordingly.

Major comments:

1) *Line 152-154: it's better to see if the levo/TSP ratio had been increased. Otherwise it's inconclusive to say the contribution of BB aerosols to particle loading over the NWPO may have increased...*

Response: The origin sentence is misleading and has been revised "Using these previous observations as a reference (Table 1), our observations suggested that the BB aerosols from the long-range transport over the NWPO in 2014 largely increased. Thus, an important question is raised, i.e., does the increase occur continuously and largely over the last decades in marine atmospheres over the NWPO?". The LEVO/TSP was $0.02\% \pm 0.03\%$ (average \pm standard deviation) and $0.02\% \pm 0.01\%$ over the NWPO and over the YBS. It is meaningless to say the contribution of BB aerosols to particle loading over the NWPO.

2) *Line 227: the relative contribution of SOA tracers to TSP in category 2 is much larger than that in category 1. Based on the authors' reasoning, is it realistic to infer that marine sources can contribute around 10% of TSP?*

Response: We carefully check through the whole manuscript. We are sorry for the misleading, but we cannot find where cause this. In revision, we added "The average contribution of SOA tracers to TSP over the SYS was higher in category 1 ($0.4\% \pm 0.6\%$) than in category 2 ($0.06\% \pm 0.07\%$)." And "The average contribution of SOA tracers to TSP over the NWPO was higher in category 1 ($0.008\% \pm 0.005\%$) than that in category 2 ($0.005\% \pm 0.005\%$)."

3) *Line 294: what are the possible major precursors for DHOPA other than BB emission?*

37 **Response:** The sentence has been revised as “leaving emissions other than BB emissions, e.g.,
38 solvent use, oil exploration, marine traffic, etc., as the major precursors for DHOPA in these
39 marine atmospheres”.

40

41 4) *Line 362: it might be attributable to the different stability of 2-MGA and LEVO?*

42

43 **Response:** Thanks for the suggestion. In revision, we added “The decomposition of LEVO
44 reported in literature (Hennigan et al., 2010; Hoffmann et al., 2010; Fraser and Lakshmanan,
45 2000) may lower the correlation between them. However, whether 2-MGA can decompose in
46 ambient air remains poorly understood.”

47

48 **Reference:**

49 Fraser, M. P. and Lakshmanan, K.: Using levoglucosan as a molecular marker for the
50 long-range transport of biomass combustion aerosols, *Environ. Sci. Technol.*, 34, 4560-4564,
51 <https://doi.org/10.1021/es991229l>, 2000.

52 Hennigan, C. J., Sullivan, A. P., Collett Jr., J. L. and Robinson, A. L.: Levoglucosan stability
53 in biomass burning particles exposed to hydroxyl radicals, *Geophys. Res. Lett.*, 37, L09806,
54 <https://doi.org/10.1029/2010GL043088>, 2010.

55 Hoffmann, D., Tilgner, A., Inuma, Y. and Herrmann, H.: Atmospheric stability of
56 levoglucosan: A detailed laboratory and modeling study, *Environ. Sci. Technol.*, 44, 694-699,
57 <https://doi.org/10.1021/es902476f>, 2010.

58

59 **Minor comments:**

60 1) Line 28: change “discuss” to “discussed”

61 2) Line 237: there is a redundant “burning”

62 3) Line 181: change “surprised” to “surprising”

63 4) Line 350: better to change “regarding” to “given that”

64 **Response:** Done. Thanks for your advice.

65

66 **Response to Anonymous Referee #3**

67 *The manuscript of Guo et al. focuses on quantification of primary and secondary organic*
68 *tracers in total suspended particles over the Yellow and Bohai seas (YBS) and the South*
69 *China Sea (SCS) collected both in 2014, and the northwest Pacific Ocean (NWPO) collected*
70 *in 2017. The authors focused on the long-range transport impact to the marine atmosphere in*
71 *spring during East Asian monsoon season. Thus, the influence of continental BB aerosols in*

72 *marine atmosphere was studied and the contributions of different precursors to the observed*
73 *SOA were quantified using specific tracers. Also, SOA formation pathways in marine*
74 *atmosphere have been discussed and related to the literature data. It is high quality work*
75 *focusing on organic aerosols from remote marine areas which are rare compared with*
76 *continental ones.*

77

78 **Response:** We thank the reviewer's comments and revise our manuscript accordingly.

79

80 *Major comment: Major comment is related to the fact that this study presents a snapshot*
81 *situation at particular marine areas and it is hard to distinguish how representative they are*
82 *for general conclusions. With that in mind, the discussion and conclusions related to the SOA*
83 *change in recent decades are questionable. This is especially addressed to the*
84 *continent-derived BB aerosols affected by accidental intensive open-fire events and or*
85 *specific meteorological conditions. Long-term study is needed to enable conclusions on*
86 *increase/decrease in contribution of continental-derived BB aerosols to remote marine areas*
87 *(L17-19; L152-154; L434-443). Despite the authors did great effort to support their*
88 *discussion with the relevant literature, conclusions arising from there should be done more*
89 *carefully since the authors compare a snapshot situation. Authors should comment on that*
90 *and/or all general conclusions related to above should be avoided.*

91

92 **Response:** We agree with the comments and soften the arguments accordingly. In revision,
93 they have been revised as “The comparison of levoglucosan observed in this study with
94 values from the literature showed that the concentrations of biomass burning aerosols over
95 the NWPO increased largely in 2014. More observations together with the snapshot
96 measurement, however, need to confirm whether the large increase occurred continuously
97 through the last decades.”

98

99 “Using these previous observations as a reference (Table 1), our observations suggested that
100 the BB aerosols from the long-range transport over the NWPO in 2014 largely increased.
101 Thus, an important question is raised, i.e., does the increase occur continuously and largely
102 over the last decades in marine atmospheres over the NWPO?”

103

104 “This further implied a large increase in continent-derived BB aerosols in marine
105 atmospheres over the NWPO recently, compared to previous studies. An important question
106 is thereby raised, i.e., does a large increase in continent-derived BB aerosols in marine
107 atmospheres over the NWPO occur continuously and largely in recent decades?”

108

109 *Minor comments*

110 *Line 153 Add . . .as a reference (Table 1),*

111 *L161-L163 Remove –samples- from the sentence or rewrite*

112 *L181 Change –surprised- with to surprise or surprising*

113

114 **Response:** Done. Thanks.

115

116 *L205 I suggest to rewrite the first sentence in figure caption, it is unclearly written*

117 **Response:**

118 The caption was rewritten into “Spatial distribution of LEVO in TSP over the NWPO in
119 spring of 2014 and 72-hrs back trajectory associated with each TSP sample. The red lines
120 represent that air masses can be derived from the continent (a, Category 1); the blue lines
121 represent that air masses may be derived mainly from the oceans (b, Category 2). The red
122 dots represent the locations of fires from Fire Information for Resource Management System
123 (FIRMS, <https://firms.modaps.eosdis.nasa.gov/>). And the base map was from Resource and
124 Environment Data Cloud 210 Platform, DOI: 10.12078/2018110201.”.

125

126 *L212 Replace NWPO with SCS.*

127 **Response:** Done. Thanks.

1 Tracer-based investigation of organic aerosols in marine 2 atmospheres from marginal seas of China to the northwest 3 Pacific Ocean

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13

14 **Abstract.** We investigated the geographic distributions of organic tracers in total suspended particles over
15 marginal seas of China, including the Yellow and Bohai seas (YBS) and the South China Sea (SCS), and the
16 northwest Pacific Ocean (NWPO) in spring, when Asian outflows strongly affect downwind marine
17 atmospheres. ~~The comparison of levoglucosan observed in this study with values from the literature implied that~~
18 ~~the contribution from biomass burning emissions to marine aerosols over the NWPO may have increased largely~~
19 ~~over the last decades. The comparison of levoglucosan observed in this study with values from the literature~~
20 ~~showed that the concentrations of biomass burning aerosols over the NWPO increased largely in 2014. More~~
21 ~~observations together with the snapshot measurement, however, need to confirm whether the large increase~~
22 ~~occurred continuously through the last decades.~~ The increase led to the mean value of levoglucosan (8.2 ± 14 ng
23 m^{-3}) observed over the NWPO closer to that over the SCS (9.6 ± 8.6 ng/ m^3) –and almost half of that over the
24 YBS (21 ± 11 ng/ m^3). Small geographic differences in monoterpene-derived and sesquiterpene-derived secondary
25 organic tracer concentrations were obtained among the three atmospheres, although the causes may differ. By
26 contrast, a large difference in isoprene-derived secondary organic tracer concentrations was observed among the
27 three atmospheres, with the sum of tracer concentrations over the SCS (45 ± 54 ng/ m^3) several times and
28 approximately one order of magnitude greater than that over the YBS (15 ± 16 ng/ m^3) and the NWPO (2.3 ± 1.6
29 ng/ m^3), respectively. The geographic distribution of aromatic-derived secondary organic tracers was similar to
30 that of isoprene-derived secondary organic tracers, with a slightly narrower difference, i.e., 1.8 ± 1.7 ng/ m^3 ,
31 1.1 ± 1.4 ng/ m^3 and 0.3 ± 0.5 ng/ m^3 over the SCS, the YBS and the NWPO, respectively. We discussed the causes
32 of the distinctive geographic distributions of these tracers and present the tracer-based estimation of organic
33 carbon.

34 1 Introduction

35 Aerosols that emanate from biomass burning (BB) consist primarily of carbonaceous components and inorganic
36 salts, which can affect the climate directly by absorbing solar radiation or indirectly by acting as either cloud
37 condensation nuclei (CCN) or ice nuclei (IN) (Bougiatioti et al., 2016; Chen et al., 2017; Hsiao et al., 2016).
38 High BB aerosol emissions zones include boreal forests (e.g., in Eurasia and North America), tropical forests

39 (e.g., in southeast Asia and the tropical Americas), and agriculture areas where crop residuals are burned (e.g., in
40 developing countries such as China and India, etc.) (van der Werf et al., 2006). BB aerosols can undergo
41 long-range transport in the atmosphere, which can carry them from the continents to the oceans (Ding et al.,
42 2013; Fu et al., 2011; Kanakidou et al., 2005). For example, BB aerosols from boreal forest wildfires in Russia
43 and China reportedly made an appreciable contribution to atmospheric particle loads observed over the Arctic
44 Ocean and northwestern Pacific Ocean (NWPO) based on specific tracers of BB (Ding et al., 2013). Although
45 open wildfires from forests occur sporadically in terms of strength and occurrence frequency, global warming
46 could be conducive to vegetation fires (Running, 2006) and thus increase emissions of BB aerosols. In this
47 century, nine years were among the ten hottest global years on record, with 2014–2018 being ranked as the top
48 five hottest years (<https://www.climatecentral.org/gallery/graphics/the-10-hottest-global-years-on-record>). The
49 question is automatically raised: how do BB aerosols in the marine atmosphere in the hottest global years
50 change against those observations previously reported?

51 In addition to BB aerosols, secondary oxidation of biogenic volatile organic compounds (BVOCs) and
52 anthropogenic VOCs (AVOCs) also contribute to the particulate carbonaceous components of marine
53 atmospheres (Kanakidou et al., 2005). Many field and modeling studies have proposed that secondary organic
54 aerosols (SOAs) arising from the oxidation of phytoplankton-derived isoprene may affect the chemical
55 composition of marine atmospheric aerosols and consequently impact CCN loading and cloud droplet number
56 concentrations (Ekström et al., 2009; Meskhidze and Nenes, 2006; Claeys et al., 2004). Several modeling
57 studies have shown that the NWPO may experience the greatest increases in sea surface temperature and CO₂
58 input under a future warming climate in the future (John et al., 2015; Lauvset et al., 2017). The Kuroshio
59 Extension current system leads the NWPO to be an active subtropical cyclone basin, promoting biogenic
60 activities (Hu et al., 2018). From the perspective of global change, it is a long-term need to study the dynamic
61 changes in atmospheric aerosols derived from marine sources over the NWPO and adjacent marginal seas of
62 China, as well as their potential effects on climate.

63 Not limited by phytoplankton-derived isoprene, BVOCs emitted from continental ecosystems and their
64 oxidation products can also affect the atmosphere in remote marine areas through long-range transport (Hu et al.,
65 2013a; Ding et al., 2013; Kang et al., 2018; Fu et al., 2011; Kawamura et al., 2017). BVOCs consist primarily of
66 isoprene, monoterpenes, sesquiterpenes, and their oxygenated hydrocarbons such as alcohols, aldehydes, and
67 ketones (Guenther et al., 2006; Ehn et al., 2014) and account for the majority of the global VOC inventory (Zhu
68 et al., 2016a, b; Heald et al., 2008). However, emissions fluxes and oxidation processes of BVOCs show great
69 variation, depending on global warming and other factors such as regional landscape, other pollutants in the
70 ambient air etc. (Ait-Helal et al., 2014; Hu and Yu, 2013; Peñuelas and Staudt, 2010). Unlike a potential
71 increase in BVOC-derived organics aerosols in marine atmospheres under global warming, anthropogenic
72 VOCs and carbonaceous particles over the continents have been decreased because of effective mitigation of air
73 pollutants in the last decades ([Li et al., 2019](#), Sharma, 2004; Murphy et al., 2011; Zhang et al., 2012). In the
74 northern hemisphere, marine atmospheres are also usually affected by anthropogenic pollutants to some extent,
75 most of which are derived from long-range transport from continents (Kang et al., 2019; Bao et al., 2018; Zhang
76 et al., 2017). The reverse trends in BVOC and anthropogenic VOC would change the composition, sources of
77 carbonaceous particles in marine atmospheres. Update observations are thereby needed to reveal the change and
78 service the future study of the impacts.

79 In this study, we analyzed the concentrations of some typical organic tracers in aerosol samples obtained from
80 three cruise campaigns from the marginal seas of China, including in the South China Sea (SCS) in 2017,
81 Yellow Sea and Bohai Sea (YBS), to the NWPO in 2014, both in springtime. We investigated the influences of
82 BB aerosols from continents over three marine atmospheres, quantified the contributions of various precursors

83 to the observed SOA in marine atmospheres using organic tracers established in the literature, and explored the
84 formation pathways of SOA from their precursors during long-range transport in these hottest global years.
85 Particularly, we conducted a comprehensive comparison of this observation with those reported in literature in
86 terms of long-term variations and geographic distributions of these tracers, etc.

87 **2 Materials and Methods**

88 Total suspended particulate (TSP) samples were collected over the NWPO from 19 March to 21 April 2014,
89 over the YBS from 30 April to 17 May 2014, and over the SCS from 29 March to 4 May 2017. All samples were
90 collected on the upper deck of the R/V Dong Fang Hong II, which sits ~8 m above the sea surface. To avoid
91 contamination from the ship's exhaust, samples were collected only when the ship was sailing, and the wind
92 direction ranged from -90° to 90° relative to the bow. TSP samples were collected on quartz fiber filters
93 (Whatman QM-A) that had been pre-baked for 4 h at 500°C prior to sampling using a high-volume sampler
94 (KC-1000, Qingdao Laoshan Electric Inc., China). The sampling duration was 15–20 h at a flow rate of $\sim 1000\text{ L}$
95 /min. After sampling, the sample filters were wrapped in baked aluminum foil and sealed in polyethylene bags,
96 then stored at -20°C and transported to the laboratory. Field blanks were collected during each sampling period.
97 However, one sampler was out of service during the cruise on the SCS. As a compromise, cellulose filters
98 (Whatman 41) previously intended for elemental analyses were used for analyses of the organic tracers in TSP.
99 The method for determining the concentrations of tracers was adapted from Kleindienst et al. (2007) and Feng et
100 al. (2013). Briefly, 20 mL dichloromethane/methanol (1:1, v/v) was used for ultrasonic extraction of 40 cm^2 of
101 each filter at room temperature three times. The combined extracts were filtered, dried under a gentle stream of
102 ultrapure nitrogen, and then derivatized with 100 μL N,O-bis-(trimethylsilyl)-trifluoroacetamide (BSTFA,
103 containing 1% trimethylchlorosilane as a catalyst) and 20 μL pyridine at 75°C for 45 min. Gas chromatography
104 mass spectrometry (GC-MS) analyses were conducted with an Agilent 6890 GC/5975 MSD. Prior to solvent
105 extraction, methyl- β -D-xylanopyranoside (MXP) was spiked into the samples as an internal/recovery standard.
106 Hexamethylbenzene was added prior to injection as an internal standard to check the recovery of the surrogates.
107 Like those reported by Feng et al. (2013), the primary organic tracers analyzed in this study included
108 levoglucosan (LEVO), mannosan, and galactosan. Four types of secondary organic tracers were used:
109 isoprene-derived secondary organic tracers (SOA_I) including 2-methylglyceric acid (2-MGA), C5-alkene triols
110 (cis-2-methyl-1,3,4-trihydroxy-1-butene, 3-methyl-2,3,4-trihydroxy-1-butene and
111 trans-2-methyl-1,3,4-trihydroxy-1-butene), and MTLs (2-methylthreitol and 2-methylerythritol);
112 monoterpene-derived secondary organic tracers (SOA_M) including 3-hydroxyglutaric acid (HGA),
113 3-hydroxy-4,4-dimethylglutaric acid (HDMGA), and 3-methyl-1,2,3-butanetricarboxylic acid (MTBCA); the
114 sesquiterpene-derived secondary organic tracer (SOA_S) β -caryophyllinic acid; and the aromatic
115 (toluene)-derived secondary organic tracer (SOA_A) 2,3-dihydroxy-4-oxopentanoic acid (DHOPA). LEVO was
116 quantified based on authentic standards in this study. While the SOA tracers without available commercial
117 standards were quantified using methyl- β -D-xylanopyranoside (MXP) as a surrogate. To reduce the uncertainty
118 of quantification, relative response factors of the target tracers to MXP were estimated by comparing the area
119 ratio of typical target ions to MXP to that of total ions in selected samples that showed high concentrations of
120 the target tracers (Feng et al., 2013).

121 Field blanks and laboratory blanks (run every 10 samples) were extracted and analyzed in the same manner as
122 the ambient samples. Target compounds were nearly always below the detection limit in field and laboratory
123 blanks. Recoveries of the surrogate (MXP) were in the range of 70–110%. The reported results were corrected
124 for recovery, assuming that the target compounds had the same recovery as the surrogate. Duplicate analyses
125 indicated that the deviation was less than 15%.

126 However, the substitution of cellulose filters (Whatman 41) during the cruise on the SCS led to increased field
127 blank values for some tracers. The tracer concentrations in those samples were, however, over three times higher
128 than the field blank values, except for those of mannosan and galactosan. Data for mannosan and galactosan
129 were thus not available, nor were the total organic carbon concentrations, for samples collected during the cruise
130 on the SCS.

131 The concentrations of organic carbon (OC) and element carbon (EC) in each sample were analyzed with a DRI
132 2001A thermal/optical carbon analyzer (Atmoslytic Inc., Calabasas, CA, USA) using the IMPROVE
133 temperature program (Wang et al., 2015). All filters before and after sampling were weighted at a glovebox
134 under the controlled ambient temperature and relative humidity. Mass concentrations of TSP, however, should
135 be treated as semi-quantitative results by considering analytic errors of quartz fiber filters (Yao et al., 2009).

136 3. Results and Discussion

137 3.1 Spatiotemporal distributions of LEVO

138 Levoglucosan, mannosan, and galactosan produced by the pyrolysis of cellulose and hemicellulose have been
139 widely used as organic tracers of BB aerosols in ambient air (Ding et al., 2013; Fu et al., 2011; Feng et al.,
140 2013). The mean levels of LEVO in TSP collected during the cruises on the NWPO and the SCS were
141 comparable, at 8.2 ng/m³ and 9.6 ng/m³, respectively (Figure S1, Table 1). They were almost half of the mean
142 value of 21 ng/m³ during the cruise on the YBS, where high concentrations of BB aerosols have been observed
143 in continental atmospheres upwind of the YBS mainly from wildfires and the burning of ~~burning~~ crop residue,
144 wildfire, etc. (Yang et al., 2014; Feng et al., 2012; Feng et al., 2013). Unlike the smaller difference among the
145 means values, the concentration of LEVO fluctuated greatly among TSP samples in each oceanic zone, ranging
146 from 0.5 to 65 ng/m³ over the NWPO, from 1.0 to 30 ng/m³ over the SCS and from 2.5 to 42 ng/m³ over the
147 YBS (Fig. S1). High spatiotemporal variation in LEVO in TSP has also been observed in literature, with
148 concentrations of LEVO fluctuating around 0.2–41 ng/m³ during Arctic to Antarctic cruises from July to
149 September 2008 and from November 2009 to April 2010 (Hu et al., 2013b). Hu et al. (2013b) also reported the
150 highest LEVO concentrations occurring at mid-latitudes (30°–60° N and S) and the lowest at Antarctic and
151 equatorial latitudes over the several months of sampling. This distinctive geographical distribution was not
152 observed in the present study, as there were no significant differences in LEVO in TSP between the SCS and
153 NWPO ($P > 0.05$).

154 Narrow spatiotemporal variation in LEVO in TSP has been reported during summer sampling over the North
155 Pacific Ocean and the Arctic in 2003, with maximum and mean values as low as 2.1 ng/m³ and 0.5 ng/m³,
156 respectively (Ding et al., 2013). A lower mean value of LEVO of 1.0 ng/m³ has also been reported in the spring
157 over the island of Chichi-jima from 2001 to 2004 (Mochida et al., 2010), while the levels increased to 3.1 ± 3.7
158 ng/m³ in TSP collected on the island of Okinawa in 2009–2012 (Zhu et al., 2015). ~~Using these previous~~
159 ~~observations as a reference, our observations suggested that the contribution of BB aerosols to particle loading~~
160 ~~over the NWPO may have increased continuously and largely over the last decades. Using these previous~~
161 ~~observations as a reference (Table 1), our observations suggested that the BB aerosols from the long-range~~
162 ~~transport over the NWPO in 2014 largely increased. Thus, an important question is raised, i.e., does the increase~~
163 ~~occur continuously and largely over the last decades in marine atmospheres over the NWPO?~~

164 Due to the lack of BB sources in oceans, large spatiotemporal variation in the concentrations of LEVO in the
165 marine atmosphere may be related to the long-range transport of atmospheric particles from continents. Thus, 72
166 h back trajectories of air masses at a height of 1000 m during our sampling periods (Figs. 1, 2) were calculated

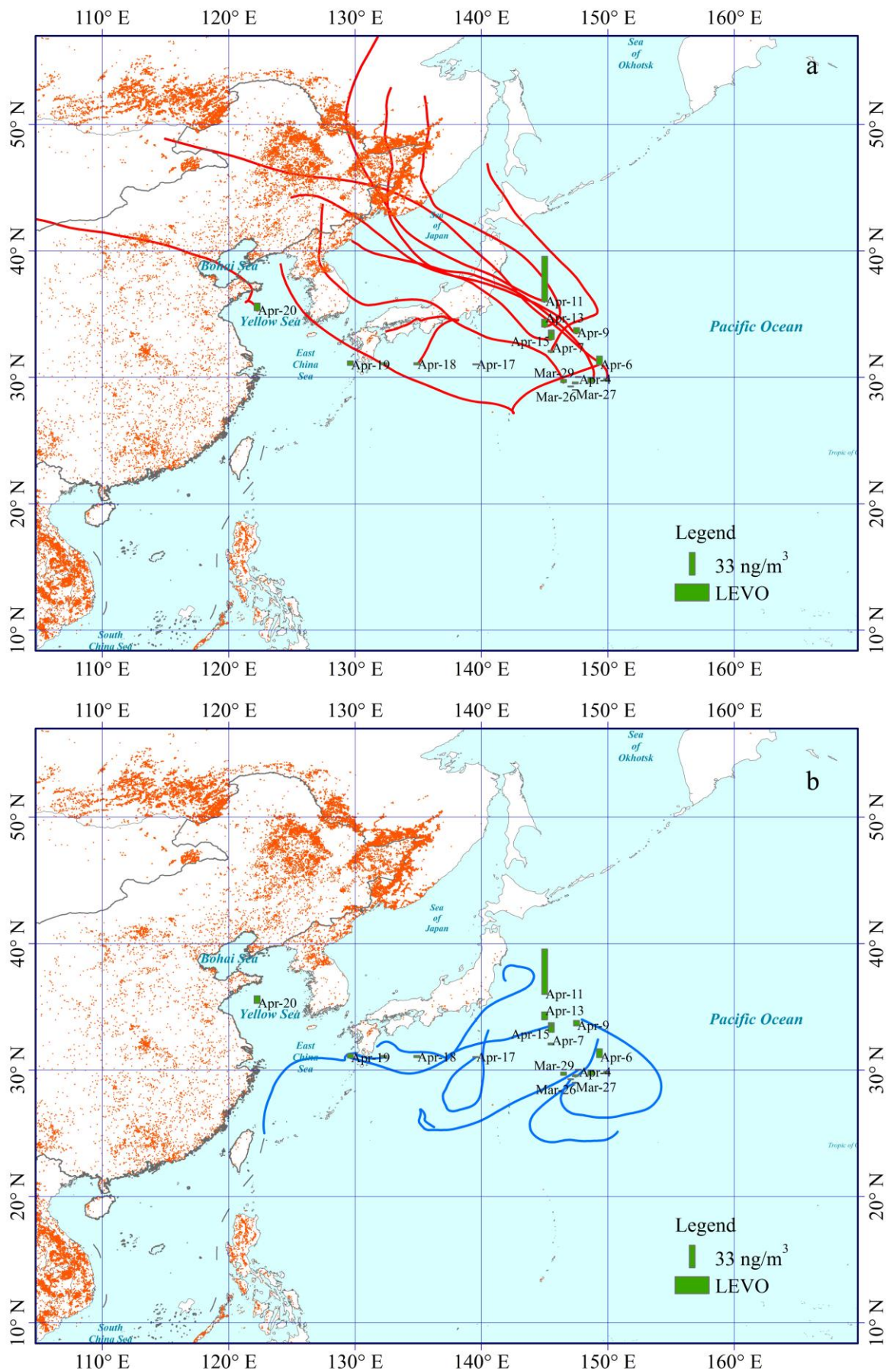
167 using the HYSPLIT model (<https://ready.arl.noaa.gov/HYSPLIT>). Based on the calculated back trajectories, TSP
168 samples could be classified into two categories with Category 1 representing continent-derived aerosol samples
169 and Category 2 being ocean-derived aerosol samples. All 12 samples collected over the YBS fell into Category 1
170 (Fig. 2). Half (11/19) of the samples collected over the NWPO were classified into Category 1 (Fig. 1). A
171 significant difference ($p < 0.05$) was obtained between the concentrations of LEVO in Category 1 ($13 \pm 18 \text{ ng/m}^3$)
172 and Category 2 ($2.0 \pm 1.8 \text{ ng/m}^3$) ~~samples~~ over the NWPO. The values in Category 2 were closer to the
173 springtime observations reported by Mochida et al. (2010) and Zhu et al. (2015) as well as the summer
174 observations reported by Ding et al. (2013), reflecting the marine background value less affected by continental
175 air masses. On the other hand, the much higher values in Category 1 than Category 2 further indicated a large
176 increase in contribution of BB aerosols being transported from the continents to the remote marine atmosphere
177 in 2014.

178 On 11 April 2014 over the NWPO, an episode of high LEVO concentration of 65 ng/m^3 occurred (Fig. 1). Like
179 LEVO, the concentrations of galactosan and mannosan in the sample were also the highest among all samples
180 collected over the NWPO. This sample was collected in the oceanic zone, approximately 500 km from the
181 continent of Japan. A combination of air mass back tracers and NASA's FIRMS Fire Map indicated strong BB
182 aerosol emissions from intense fire events in Siberia, followed by long-range transport with the westerly wind as
183 the major contributors to this anomaly (Fig. 1). A similar episodic concentration of LEVO of 27 ng/m^3 in TSP
184 was observed once previously over the NWPO during a circumnavigation cruise (Fu et al., 2011). By combining
185 satellite data with other observations, many studies in literature have found that BB aerosols from major forest
186 fires and smoke events in Siberia could be transported downwind to remote marine regions not only in spring,
187 but also in summer (Generoso et al., 2007; Ding et al., 2013; Huang et al., 2009). In a few cases, BB aerosols
188 have been reported to have reached as far as the adjacent Arctic region (Warneke et al., 2010; Generoso et al.,
189 2007). Van der Werf et al. (2006) estimated the emissions of BB aerosols from Eurasia to be much larger than
190 those from North America. Thus, it is not ~~surprised-surprising~~ that the concentrations of LEVO over the NWPO
191 were much higher than those over the eastern North Pacific and western North Atlantic at similar latitudes (Hu
192 et al., 2013b).

193 In addition, both galactosan and mannosan showed strong linear correlations with LEVO ($R^2 = 0.98$, $p < 0.05$)
194 in TSP collected over the NWPO and YBS in this study. These strong correlations indicate that the three tracers
195 were probably derived from the same BB sources. Previous studies have reported LEVO/mannosan (L/M) ratios
196 of 3–10, 15–25, and 25–40 from softwood, hardwood, and crop-residue burning, respectively (Kang et al., 2018;
197 Zhu et al., 2015). The calculated L/M ratios in TSP collected over the NWPO were 19 ± 4 in this study, which
198 implies dominant contributions from herbaceous plants and hardwood. The calculated L/M ratios in TSP
199 collected over the YBS were 14 ± 11 , indicating mixed sources.

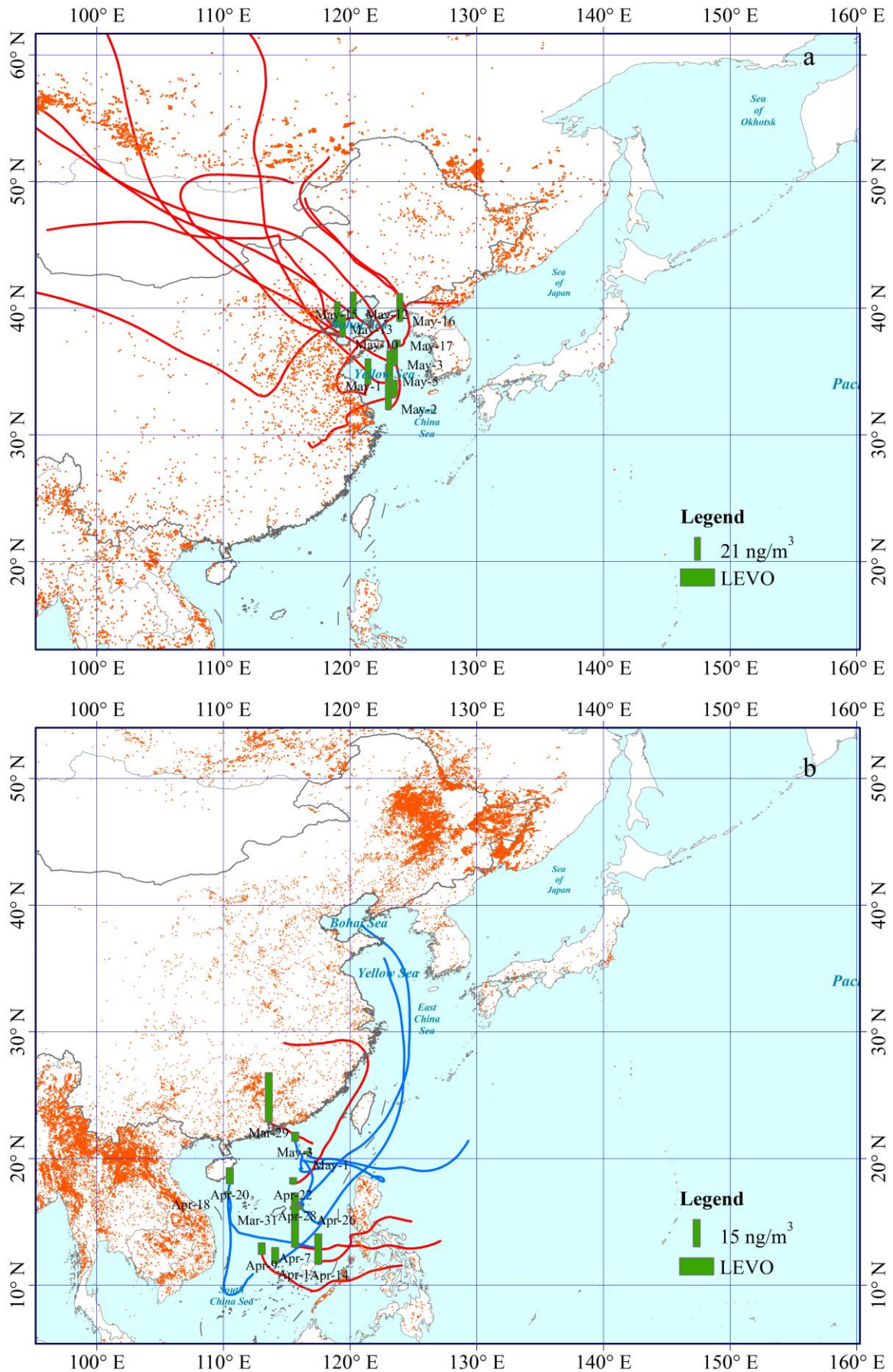
200 In all, 5 of 13 samples collected over the SCS were classified into Category 1, with air masses identified as
201 originating from either the continental areas of South China or the Philippines (Fig. 2). The concentration of
202 LEVO fluctuated around $17 \pm 12 \text{ ng/m}^3$ in Category 1 but decreased to $3.6 \pm 3.4 \text{ ng/m}^3$ in Category 2. However, no
203 significant differences were found between categories due to the large variation in LEVO concentration among
204 the limited samples in Category 1 ($p > 0.05$). Forest fires occur accidentally, leading to the large variation in
205 LEVO in Category 1. Southern Asia has been reported to be one of the greatest emissions sources of BB
206 aerosols worldwide (van der Werf et al., 2006), which likely led to the higher mean value of LEVO in Category
207 1. However, the LEVO level observed over the SCS in Category 2 was closer to that reported from low-latitude
208 regions ($2.7 \pm 1.1 \text{ ng/m}^3$, Table 1) collected during a global circumnavigation cruise (Hu et al., 2013b). Hu et al.
209 (2013b) argued that their low observed concentrations may have been associated with intense wet deposition,
210 degradation as well as intensive moist convection that occurred in the tropical region during their summer cruise.

211 Unfortunately, no previous observations of LEVO in spring can allow us analyzing the long-term variation in
212 contribution of BB aerosols therein. However, this observation can be used for future comparison.



213
214 **Figure 1. Spatial distribution of LEVO over the NWPO (2014) and BTs corresponding to the samples; the**
215 **samples with the backward trajectories (red lines) indicate land-influenced aerosols (a, Category 1) and**

216 ~~the blue line denotes ocean-influenced aerosols collected (b, Category 2). The red dots represent the~~
217 ~~locations of fires from Fire Information for Resource Management System (FIRMS,~~
218 ~~<https://firms.modaps.eosdis.nasa.gov/>). The base map was from Resource and Environment Data Cloud~~
219 ~~Platform, DOI: 10.12078/2018110201. Spatial distribution of LEVO in TSP over the NWPO in spring of~~
220 ~~2014 and 72-hrs back trajectory associated with each TSP sample. The red lines represent that air masses~~
221 ~~can be derived from the continent (a, Category 1); the blue lines represent that air masses may be derived~~
222 ~~mainly from the oceans (b, Category 2). The red dots represent the locations of fires from Fire~~
223 ~~Information for Resource Management System (FIRMS, <https://firms.modaps.eosdis.nasa.gov/>). And the~~
224 ~~base map was from Resource and Environment Data Cloud 210 Platform, DOI: 10.12078/2018110201.~~



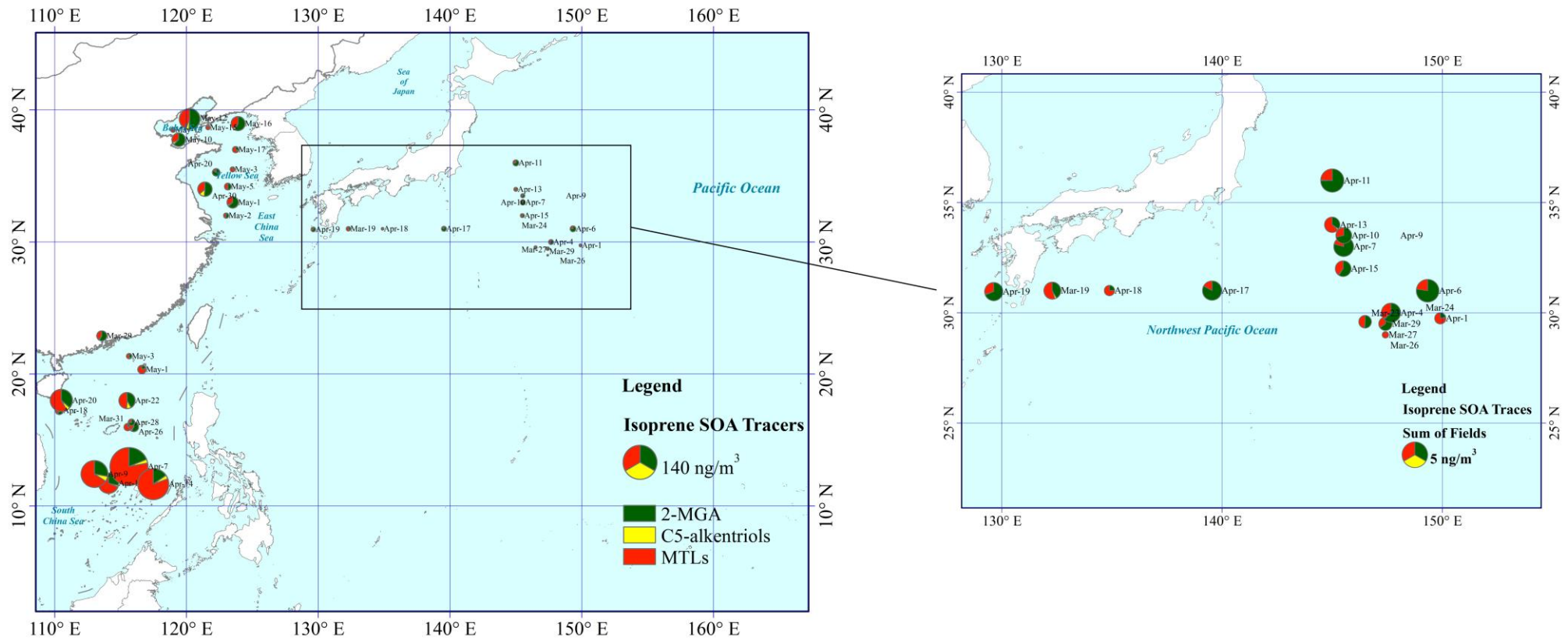
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Figure 2. Spatial distribution of LEVO over the YBS (a, 2014), and **NWPO-SCS** (b, 2017), detailed information described in Figure 1. [And the base map was from Resource and Environment Data Cloud 210 Platform, DOI: 10.12078/2018110201.](https://www.researchgate.net/publication/318110201)

230 3.2 Spatiotemporal distributions of SOA_I tracers

231 SOA_I tracers were detected during all three cruises. The sum of SOA_I tracers showed a decreasing trend of up to
232 approximately one order of magnitude from marginal seas to the open ocean, i.e., 45 ± 54 ng/m³ in TSP over the
233 SCS, 15 ± 16 ng/m³ over the YBS and 2.3 ± 1.6 ng/m³ over the NWPO (Fig. S1). The highest sum value of SOA_I
234 tracers over the SCS was 176 ng/m³, indicating strong photochemical formation of SOA from biogenic volatile
235 organics (Fig. 3). The geographical distribution of SOA_I tracers in this study was generally consistent with those
236 reported by Hu et al. (2013a), with higher concentrations of these tracers in atmospheric particles collected from
237 low-latitude oceanic zones (30° S–30° N) due to large emissions from tropical forests and strong photochemical
238 reactions. Their reported average contents of SOA_I tracers in low-latitude oceanic zones fluctuated around
239 9.2 ± 6.7 ng/m³, much lower than those measured in this study.

240 When the sum of SOA_I tracers in each sample was examined separately according to the air mass source, a
241 significant difference was found over the SCS between Category 1 (85 ± 66 ng/m³) and Category 2 (19 ± 22
242 ng/m³), with significance at $p < 0.01$. The average contribution of SOA tracers to TSP mass concentration over
243 the SYS was higher in category 1 ($0.4\% \pm 0.6\%$) than in category 2 ($0.06\% \pm 0.07\%$). The tracer values were
244 2.7 ± 1.8 ng/m³ in Category 1 and 1.7 ± 1.0 ng/m³ in Category 2 over the NWPO, where no significant difference
245 between the two categories was found ($p > 0.05$). The average contribution of SOA tracers to TSP mass
246 concentration over the NWPO was higher in category 1 ($0.008\% \pm 0.005\%$) than that in category 2 ($0.005\% \pm$
247 0.005%). Supposed that concentrations of the tracers in Category 2 were completely contributed by marine
248 sources, it can be inferred that SOA_I carried by continental air masses increased sharply over the SCS. However,
249 it was not the case over the NWPO. Because all samples over the YBS fell into Category 1, this comparison
250 could not be made for the YBS.



251
 252 **Figure 3. Spatial distribution of SOA_I tracer compounds over three marine regions, ECS and NWPO in 2014, SCS in 2017. The area of the pie indicates the concentration of total**
 253 **SOA_I tracers. The base map was from Resource and Environment Data Cloud Platform, DOI: 10.12078/2018110201.**

254 3.3 Spatiotemporal distributions of SOA_M, SOA_s tracers

255 The sum of SOA_M tracers including HGA, HD-MGA, and MBTCA was greatest over the SCS region (3.5±6.0
256 ng/m³), where the concentration was approximately double that over the YBS (1.6±2.0 ng/m³) and NWPO
257 regions (1.6±2.7 ng/m³) (Fig. S1), but no significant differences were identified between any two campaigns.
258 The concentrations of SOA_M tracers were almost one magnitude lower than those of SOA_I tracers. Due to the
259 unique contribution of terpene-derived SOA to nucleation and growth of newly formed particles in the
260 atmospheres (Gordon et al., 2017; Zhu et al., 2019; Ehn et al., 2014), the SOA_M may primarily cause indirect
261 climate effects rather than direct effects of aerosols in the marine atmosphere. The difference in mean SOA_M
262 concentration between the SCS and NWPO narrowed to a factor of two, in contrast to the differences of
263 approximately one order of magnitude in mean SOA_I between the two types of atmospheres. The precursors of
264 SOA_M tracers derive mainly from coniferous forests (Duhl et al., 2008) and the decreasing proportion of
265 coniferous forests in subtropical and tropical regions may partially explain the smaller spatial difference in
266 SOA_M tracers over the SCS compared to the YBS and NWPO. However, the comparable SOA_M levels over the
267 YBS and NWPO have not yet been explained.

268 Only three SOA_M tracers were measured in this study, but other SOA_M tracers have been measured and reported
269 in marine atmospheres (Kang et al., 2018; Fu et al., 2011). In order to compare our results with the total amount
270 of SOA_M tracers in the literature, the total amounts measured in this study were multiplied by a factor of 3.1
271 (described in supporting information Sect. S1, Fig. S4) according to the chamber results obtained by Kleindienst
272 et al. (2007). The adjusted values over the SCS were closer to the mean of 11.6 ng/m³ observed over the ECS
273 (Kang et al., 2018) and the lower values of 9.80–49.0 ng/m³ observed among 12 continental sites in China (Ding
274 et al., 2016). The adjusted total amounts of SOA_M over the NWPO and YBS were comparable to previous
275 observations of 3.0±5.0 ng/m³ collected from the Arctic to Antarctic in 2008-2010 (Hu et al., 2013a), but much
276 higher than observations of 63±49 pg/m³ over the North Pacific and Arctic in 2003 (Ding et al., 2013). This may
277 also imply a substantial increase in SOA_M in the last decades, although more investigations are needed to
278 confirm.

279 β-Caryophyllene is a major sesquiterpene emitted from plants such as Scots pine and European birch (Duhl et al.,
280 2008; Tarvainen et al., 2005). β-Caryophyllinic acid is formed through the ozonolysis or photo-oxidation of
281 β-caryophyllene. The highest levels of β-caryophyllinic acid were observed over the YBS (0.13±0.03 ng/m³),
282 followed by the SCS (0.08±0.11 ng/m³) and NWPO (0.05±0.09 ng/m³) (Fig. S1). The spatial distribution of
283 β-caryophyllinic acid clearly did not follow the general trend of biogenic SOA, with the highest values over the
284 SCS followed by the YBS. Compared to values from the literature, our results are much higher than those over
285 the North Pacific and Arctic Oceans (2.4±5.4 pg/m³) (Ding et al., 2013) but much lower than observations over
286 the East China Sea reported by Kang et al. (2018), where β-caryophyllinic acid was reported to be in the range
287 of 0.16–17.2 ng/m³ with a mean of 2.9 ng/m³. The large differences in β-caryophyllinic acid content observed in
288 various campaigns remains unexplained.

289 3.4 Spatiotemporal distributions of SOA_A tracers

290 When the concentrations of DHOPA in TSP were examined, the highest concentrations occurred over the SCS
291 (1.8±1.7 ng/m³), followed by the YBS (1.1±1.4 ng/m³), and the lowest values were recorded in the NWPO
292 region (0.3±0.5 ng/m³) (Fig. S1). The decreasing extent of the DHOPA ~~decrease~~ from the SCS to the NWPO
293 was approximately three times less than that of SOA_I tracers but approximately three times larger than that of

294 SOA_M tracers. Ding et al. (2017) reported annual averages of DHOPA among various sites in China, which
295 ranged from 1.2 to 8.8 ng/m³. The concentrations of DHOPA observed over the SCS and the YBS were similar
296 to the lower values observed in upwind continental atmospheres.

297 Formation of DHOPA depends on the molecular structures of aromatics, as well as concentrations of free
298 radicals and oxidants, etc. (Li et al., 2016; Henze et al., 2008). The mean value of DHOPA in Category 1
299 (0.43±0.65 ng/m³) was nearly twice that in Category 2 (0.20±0.31 ng/m³) over the NWPO ($p > 0.05$). With two
300 samples with high DHOPA (1.2, 2.1 ng/m³) in Category 1 to be excluded, the recalculated average DHOPA
301 decrease down to 0.17±0.21 ng/m³. The continent-derived DHOPA seemingly yielded a minor contribution to
302 the observed values over the NWPO, except during strong long-range transport episodes. Similarly, the mean
303 values of DHOPA were same in Category 1 (1.8±2.1 ng/m³) and Category 2 (1.8±1.5 ng/m³) samples collected
304 over the SCS and no significant differences were observed between two categories. Much stronger UV radiation
305 occurs over the SCS than the YBS, which may contribute to the elevated DHOPA level over the SCS. Aside
306 from continent-derived precursors, oil exploration and heavy marine traffic over the SCS are also potential
307 contributors to the higher DHOPA levels therein, and this link requires further investigation. Previous field
308 observations in China have demonstrated that biofuel or biomass combustion emissions act as important sources
309 of aromatics in the atmosphere (Zhang et al., 2016), as evidenced by the association between the nationwide
310 increase in DHOPA during the cold period and the enhancement of BB emissions (Ding et al., 2017). In this
311 study, no linear correlation was obtained between DHOPA and LEVO in samples collected over the SCS and
312 other two campaigns, ~~leaving emissions other than BB emissions, e.g., solvent use, oil exploration, marine~~
313 ~~traffic, etc., as the major precursors for DHOPA in these marine atmospheres~~ ~~leaving emissions other than BB~~
314 ~~emissions as the major precursors for DHOPA in these marine atmospheres~~ (Li et al., 2013).

315 3.5 Causes for high photochemical yields of SOA_I over the SCS

316 Because higher concentrations of SOA_I were observed in TSP samples collected over the SCS, the composition
317 of SOA_I tracers was further investigated in terms of their formation pathways and sources. Based on the results
318 of chamber experiments, Surratt et al. (2010) proposed different formation mechanisms for 2-MGA and MTLs.
319 2-MGA is a C₄-dihydroxycarboxylic acid, which forms through a high-NO_x pathway. MTLs and C₅-alkene
320 triols are mainly products of the photooxidation of epoxydiols of isoprene under low-NO_x conditions.

321 MTLs acted as the dominant compounds among SOA_I tracers in most TSP samples collected over the SCS, with
322 concentrations of 31±42 ng/m³ (Fig. 3). The ratio of 2-MGA/MTLs ranged from 0.2 to 3.1, with a median value
323 of 0.6. The ratio exceeded the unity in only 4 of 13 samples. This result allowed us to infer that the observed
324 SOA_I tracers were generated mainly under low-NO_x conditions. Although the concentration of
325 2-methylerythritol was nearly double that of 2-methylthreitol, they were highly correlated ($R^2 = 0.99$, $p < 0.05$)
326 because of their shared formation pathway. Satellite data showed that the NO₂ levels in South China and the
327 Philippines were low, except in a few hotspots (Fig. S2). Such low-NO_x conditions favor the formation of
328 MTLs rather than 2-MGA over the tropical SCS. The isoprene emitted from plants growing on oceanic islands
329 may also undergo chemical conversion to SOA under low-NO_x conditions, and low-NO_x conditions are always
330 expected in remote marine atmospheres (Davis et al., 2001).

331 In general, zonally and monthly averaged OH concentrations around 15°N are ~50% were greater than those
332 around 35°N (Bahm and Khalil, 2004). Thus, enhanced formation of MTLs is theoretically expected under the
333 strong UV radiation of tropical regions. However, no significant correlation between the concentrations of
334 MTLs and UV radiation was obtained over the SCS (data not shown) possibly due to the influences of various
335 air masses. A field study showed that MTL yields were positively correlated with ambient temperature in

336 continental atmospheres (Ding et al., 2011). 2-MGA yields, in contrast, showed no significant correlation with
337 ambient temperature in this study. Moreover, lower relative humidity may enhance the formation of 2-MGA in
338 the particulate phase but not for MTLs (Zhang et al., 2011). Variation in ambient temperature and relative
339 humidity may complicate the relationship between the concentrations of SOA_I tracers and UV radiation over the
340 SCS.

341 In addition, the MTLs concentration in Category 1 (62 ± 55 ng/m³) was larger than that ~~for~~in Category 2 (11 ± 14
342 ng/m³). The more abundant MTLs associated with Category 1 was most likely related to long-range transport of
343 these chemicals from upwind continental areas, the oxidation of continental precursors in the marine atmosphere,
344 or both. Large emissions of isoprene were expected from tropical forests upwind of the SCS due to the high
345 vegetation coverage and high ambient temperature of such areas (Ding et al., 2011; Rinne et al., 2002). Global
346 estimates show tropical trees to be responsible for ~80% of terpenoid emissions and ~50% of other VOC
347 emissions (Guenther et al., 2012).

348 In a clean marine atmosphere, phytoplankton is the sole source of isoprene emissions over the oceans (Bonsang
349 et al., 1992; Broadgate et al., 1997). Chlorophyll-a has been widely employed as a measure of phytoplankton
350 abundance and a proxy for predicting isoprene concentrations in water (Hackenberg et al., 2017). The
351 satellite-derived chlorophyll-a level during the study period over the SCS was below 0.45 mg/m³, excluding
352 coastal areas (Fig. S3). The MTLs observations of 11 ± 14 ng/m³ in Category 2 should be considered as the upper
353 limitation value derived from marine phytoplankton in the SCS. Although air masses differed between
354 Categories 1 and 2, a good correlation was obtained between MTLs and 2-MGA when the data in the two
355 categories was pooled for analyses ($R^2 = 0.77$, $P < 0.01$). This strong correlation indicated these tracers are
356 primarily formed through shared pathways. However, this correlation was poor over the NWPO, as discussed
357 below.

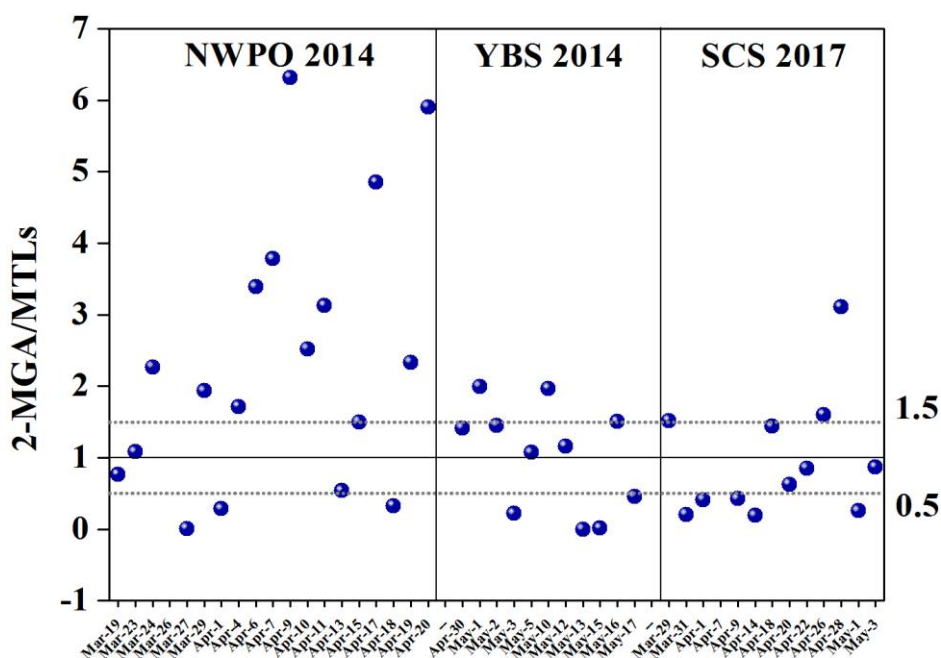
358 3.6 Origin and formation of SOA_I over the NWPO

359 Over the NWPO, the concentration of 2-MGA was 1.6 ± 1.5 ng/m³, which was generally dominant among SOA_I
360 tracers, followed by MTLs (0.7 ± 0.3 ng/m³) and C5-alkene triols (0.03 ± 0.02 ng/m³). When the ratio of
361 2-MGA/MTLs was further examined, it varied greatly from <0.1 to 6.3, with a median value of 2.1. Most ratios
362 observed over the NWPO in this study were far greater than the values of 0.18–0.59 reported by Hu et al. (2013a)
363 from a global circumnavigation cruise, and also greater than 0.87–1.8 reported in urban areas of California
364 (Lewandowski et al., 2013) and the maximum value of 2.0 obtained over the YBS. Ding et al. (2013) also
365 reported ratios that fluctuated greatly from 0.5 to 10 with a median value of 3.3 during a summer cruise in the
366 NWPO and Arctic Ocean in 2003. The large 2-MGA/MTL ratios over the NWPO appeared to be highly
367 consistent over two independent sampling campaigns.

368 The compound profile of SOA_I tracers over the NWPO implied high-NO_x conditions allowing oxidation of
369 isoprene to generate the SOA_I present in most samples. Such high-NO_x conditions are impossible in a remote
370 marine atmosphere, as indicted in Figure S2. Regarding Given that the lifespan of isoprene in the atmosphere is
371 only several hours (Bonsang et al., 1992), the long-range transport of oxidation products formed under high NO_x
372 levels over the continents likely led to the 2-MGA-dominated composition of SOA_I. Based on air mass back
373 trajectories, this long-range transport may involve 2-MGA originating from Siberia, northeastern China, or
374 Japan.

375 Organic aerosols over the NWPO were strongly influenced by forest fires that take place in Siberia during
376 spring and summer almost every year (Ding et al., 2013; Huang et al., 2009). Previous emissions inventory
377 studies have reported high isoprene and NO_x emissions from various BB types (Akagi et al., 2011; Andreae and

378 Merlet, 2001). Ding et al. (2013) thus argued that an increase in emissions of isoprene in the presence of BB,
 379 followed by its chemical conversion under high-NO_x conditions, may lead to transport over thousands of
 380 kilometers and hold at the detectable concentrations in the remote marine atmosphere over the NWPO. The
 381 same argument may hold true for the elevated ratios of 2-MGA/MTLs observed over the NWPO in this study
 382 (Fig. 4). However, we did not find a significant correlation between 2-MGA and LEVO over the NWPO. The
 383 decomposition of LEVO reported in literature (Hennigan et al., 2010; Hoffmann et al., 2010; Fraser and
 384 Lakshmanan, 2000) may lower the correlation between them. However, whether 2-MGA can decompose in
 385 ambient air remains poorly understood.
 386 On the other hand, the ratios of 2-MGA/MTLs in 3 of 19 samples collected over the NWPO were below 0.5
 387 (Figure 4). In these cases, the oxidation of isoprene under low-NO_x conditions likely dominated the generation
 388 of SOA_I. The ratios of 2-MGA/MTLs were 0.5–1.5 in 4 of 19 samples, suggesting mixed contributions to SOA_I
 389 from the oxidation of isoprene under low-NO_x conditions and high-NO_x conditions. As the major formation
 390 pathways of 2-MGA and MTLs varied greatly among samples, no significant correlation ($R^2 = 0.12$, $p > 0.05$)
 391 was obtained between 2-MGA and MTLs over the NWPO. Recall that the tracer values of SOA_I were 2.7 ± 1.8
 392 ng/m^3 in Category 1 and $1.7 \pm 1.0 \text{ ng/m}^3$ in Category 2. This implied that SOA_I derived from marine sources was
 393 comparable to that derived from the continent outflows.



394
 395 **Figure 4. Spatial ratio of 2-MGA/MTLs among SOA_I tracers over three marine regions.**

396 3.7 Source apportionment of secondary organic carbon (SOC)

397 The tracer-based approach developed by Kleindienst et al. (2007) was applied to estimate the concentrations of
 398 SOC and WSOC_{BB}, as follows:

$$[SOC] = \frac{\sum_i [tri]}{f_{soc}} \quad (1)$$

399

$$[WSOC_{BB}] = \frac{C_{tracer}}{f_{tracer/WSOC_{BB}}} \quad (2)$$

400
 401 where $\Sigma_i(\text{tri})$ is the sum of concentrations of the selected suite of tracers for a precursor, and f_{SOC} is the mass
 402 fraction of tracer compounds in SOC generated from the precursor in chamber experiments. Assuming that the
 403 f_{SOC} values in ambient air match those in the chamber, the f_{SOC} values for precursors such as isoprene,
 404 monoterpenes, β -caryophyllene, and aromatics were $0.155 \pm 0.039 \mu\text{g}/\mu\text{gC}$, $0.231 \pm 0.111 \mu\text{g}/\mu\text{gC}$, 0.023 ± 0.0046
 405 $\mu\text{g}/\mu\text{gC}$, and $0.00797 \pm 0.0026 \mu\text{g}/\mu\text{gC}$, respectively (Kleindienst et al., 2007), with uncertainty described in
 406 Sect. S2. The fraction of LEVO in WSOC ($0.0994 \mu\text{g}/\mu\text{gC}$) from the BB plume was used for $WSOC_{BB}$ (Ding et
 407 al., 2008). The f_{SOC} value for monoterpenes was scaled up by a factor of 3.1 based on experimental observations,
 408 as these two tracers (HGA+HD-MGA) accounted for 2/9 of the total tracers of monoterpenes, as described in
 409 the supporting information (Kleindienst et al., 2007).

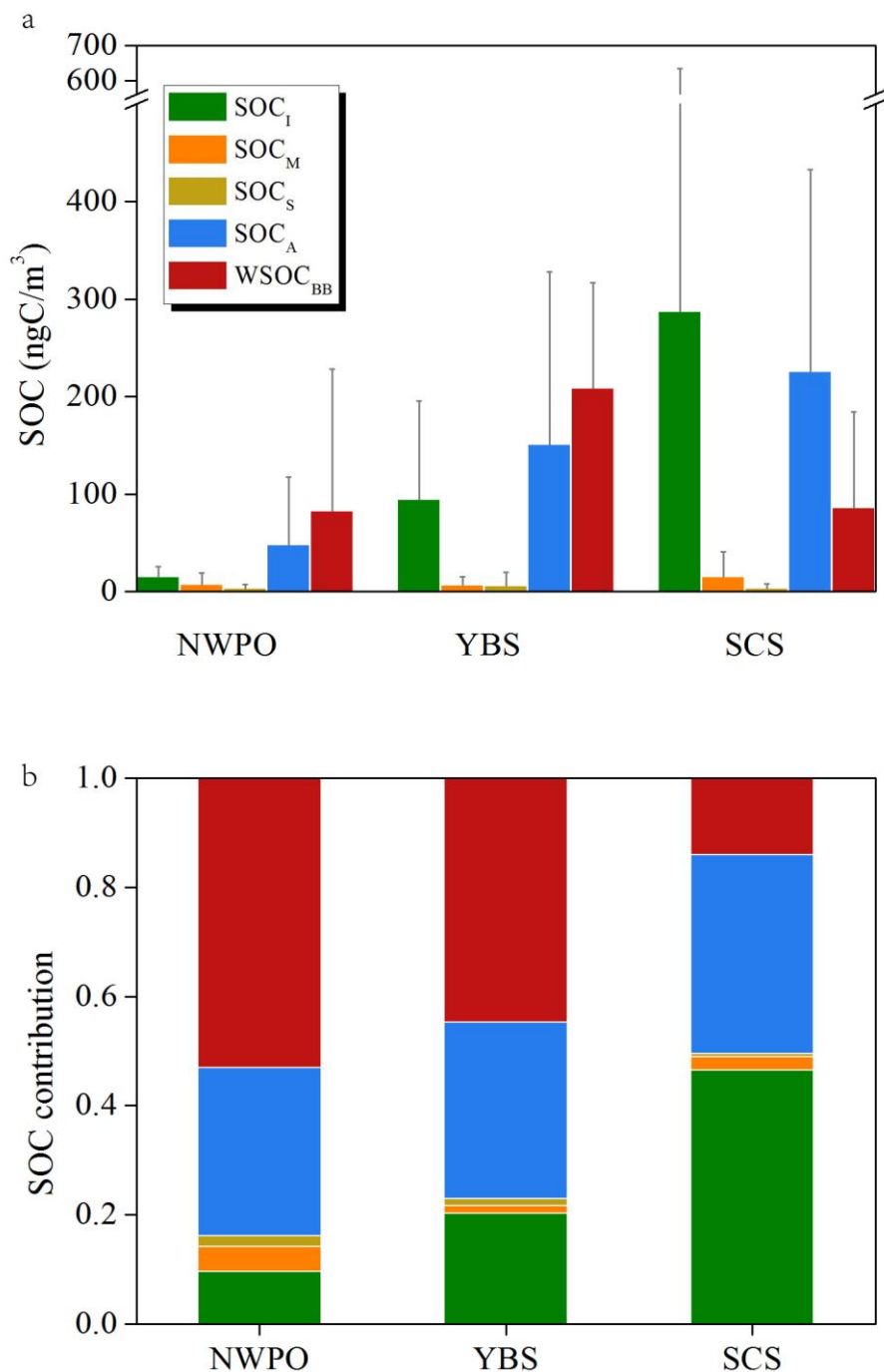
410 Over the SCS, nearly half of the sum of SOC and $WSOC_{BB}$ was in the form of SOC_I (47%), followed by SOC_A
 411 (36%), $WSOC_{BB}$ (14%) and a minor contribution of 2.5% from SOC_M (Fig. 5). This composition pattern over
 412 the SCS could be attributed to abundant biogenic SOA formation in low-latitude tropical marine atmospheres.
 413 Over tropical marine regions, atmospheric oxidation products can account for 47–59% of the total organic
 414 content estimated, with biomass burning emissions making up only 2–7% based on source apportionment using
 415 organic tracers (Fu et al., 2011). A model study by Fu et al. (2012) showed that secondary formation accounts
 416 for as much as 62% of OC estimated using tracers in eastern China in summer. A reverse pattern was observed
 417 over the YBS, with $WSOC_{BB}$ as the dominant contributor (45%) to the sum of SOC and $WSOC_{BB}$, followed by
 418 SOC_A (32%) and SOC_I (20%). The contribution of SOC_M was also minor, at 1.5%. Notably, the chemical
 419 composition observed over the NWPO was similar to that over the YBS, with $WSOC_{BB}$ contributing up to 53%.
 420 In addition, Kang et al. (2018) used the PMF method to identify various sources of OC in marine aerosols over
 421 the ECS such as secondary nitrate, BSOA, BB, and fungal spores.

422 Geographically, the estimated SOC values from BVOCs ranked at the highest level of $306 \pm 343 \text{ ngC}/\text{m}^3$ over the
 423 SCS, decreasing to $107 \pm 99 \text{ ngC}/\text{m}^3$ over the YBS and $24 \pm 22 \text{ ngC}/\text{m}^3$ over the NWPO. The estimates of
 424 aromatic SOC exhibited the same geographic trend, with values of $225 \pm 208 \text{ ngC}/\text{m}^3$ over the SCS, 151 ± 177
 425 ngC/m^3 over the YBS and $48 \pm 69 \text{ ngC}/\text{m}^3$ over the NWPO. Recent modeling results have also shown that
 426 aromatic emissions are the predominant precursors of SOA during springtime in China in comparison with
 427 BVOCs and other AVOCs (Han et al., 2016). Among estimates of $WSOC_{BB}$, the highest values of 209 ± 108
 428 ngC/m^3 were recorded over the YBS, followed by comparable levels of $86 \pm 98 \text{ ngC}/\text{m}^3$ (SCS) and 83 ± 145
 429 ngC/m^3 (NWPO).

430 In our study, the calculated $WSOC_{BB}$ estimate accounted for $4.1 \pm 5.0\%$ and $3.3 \pm 1.7\%$ of measured OC over the
 431 NWPO and YBS, respectively, and these values are higher than that obtained over the ECS during summer
 432 (1.4%) (Kang et al., 2018). Estimated SOC from BVOCs accounted for only $1.5 \pm 1.4\%$ and $1.8 \pm 1.7\%$ to the
 433 measured OC over the NWPO and YBS, respectively, which is lower than that over ECS (4.21%) (Kang et al.,
 434 2018). However, the mean values obtained in this study were similar to the total SOC level estimated using
 435 tracers as a proportion of measured WSOC (4%) during a cruise on the North Pacific and Arctic Oceans,
 436 supposed that WSOC accounted for half of the total OC in atmospheric particles (Ding et al., 2013).

437 The calculated SOC level derived from organic tracers accounted for less than 68% of total measured OC in
 438 these study areas. However, this SOC compounds are expected to derive mainly from photochemical reactions
 439 in the gas phase, followed by gas-aerosol partitioning. These compounds likely play an important role in the
 440 growth of newly formed particles alongside pre-existing nucleation mode or Aitken mode particles. However,
 441 most organic matter detected in bulk samples may originate from primary sources, heterogenous reactions and

442 in-cloud processing (Ervens et al., 2011; Kanakidou et al., 2005; Nichols, 2016), and these compounds may be
443 major drivers of the direct climate effects of aerosols, rather than indirect climate effects. In the future, a
444 comprehensive combination measurement of organic tracers and organic matter with an aerosol mass
445 spectrometer should be used to elucidate the formation and growth processes of atmospheric nanoparticles.
446



447
448 **Figure 5. Average SOC levels calculated using the tracer-SOC/WSOC method over three marine regions**
449 **(ECS and NWPO in 2014, SCS in 2017) and their contributions based on five organic tracers measured in**
450 **this study.**

451 4. Conclusions

452 This study investigated the geographical distributions of tracer-based organic matter observations in TSP
453 collected over two marginal seas of China and the NWPO in the spring season, when the East Asian monsoon
454 carries biogenic and anthropogenic aerosols over these oceanic zones. We found that a significantly large
455 difference in LEVO over the NWPO between two categories of air masses originating from upwind continents
456 or oceanic regions, as Category 1 (continental) contained $13 \pm 18 \text{ ng/m}^3$ and Category 2 (oceanic) had 2.0 ± 1.8
457 ng/m^3 ; the concentrations of LEVO in Category 2 were closer to the low values reported in the literature. This
458 further implied a large increase in ~~contribution of~~ continent-derived BB aerosols ~~to~~ in marine atmospheres over
459 the NWPO in recent decades, compared to previous studies. An important question is thereby raised, i.e., does a
460 large increase in continent-derived BB aerosols in marine atmospheres over the NWPO occur continuously and
461 largely in recent decades? Combining the L/M ratios of 19 ± 4 over the NWPO with the calculated air mass back
462 trajectories indicates that the increase was very likely associated with enhanced emissions of BB aerosols from
463 wildfires in Siberia and northeastern China. Moreover, the mean level of BB aerosols over the SCS nearly
464 matched that over the NWPO. The contents of LEVO in Category 2 air masses, derived from oceanic zones over
465 the SCS, were comparable with those reported in the literature, but the mean value was only about a quarter of
466 that in Category 1, representing air masses from upwind continents. However, the limited data available over the
467 SCS in the literature cannot support inferences about whether BB aerosols emitted from upwind tropical forests
468 have increased in recent decades.

469 The concentrations of SOA_I over the SCS were approximately one order of magnitude greater than those
470 observed over the NWPO and several times larger than those over the YBS. The larger values observed over the
471 SCS in Category 1 than in Category 2 were likely driven by high emissions of isoprene from upwind tropical
472 forests and strong solar radiation. The MTLs dominance of SOA_I over the SCS strongly suggested that SOC
473 from BVOCs was generated primarily under low- NO_x conditions. On the other hand, 2-MGA dominance over
474 the YBS implied that most SOC was generated under high- NO_x conditions. Elevated ratios of 2-MGA/MTLs
475 of >1.5 were obtained for 11 of 19 total samples collected over the NWPO, consistent with those reported in the
476 literature. Larger ratios may be attributed to possible emissions of BVOCs in the presence of BB. However, the
477 comparable concentrations of SOA_I in Category 1 and Category 2 samples collected over the NWPO implied a
478 large contribution of SOA_I from marine sources. The aromatic SOA tracers' levels were highest over the SCS,
479 followed by values obtained over the YBS and NWPO. The high values observed over the SCS may be related
480 to strong solar radiation, but the sources of precursors remain unexplained. Based on the concentrations in
481 Category 1 and 2 air samples collected over the SCS and NWPO, mixed sources of aromatic VOCs should exist,
482 including continent-derived precursors, oil exploration and heavy marine traffic.

483 Over the NWPO and the YBS, the estimated WSOC_{BB} levels were nearly equal to the sum of SOC estimated
484 from the oxidation of aromatics and BVOCs. Over the SCS, SOC estimated from the oxidation of BVOCs was
485 significantly larger than the estimated WSOC_{BB} . The geographical difference may be related to emissions of
486 primary particulate organics and gaseous precursors as well as formation processing of secondary organics in
487 various atmospheres.

488 The atmospheric composition of SOA in different geographical locations is, however, highly complex and is
489 regulated by many factors including local meteorological conditions, anthropogenic emissions, plant species,
490 vegetation cover and regional chemistry, and therefore warrants further quantification and analyses. Particularly,
491 whether BB aerosols and other biogenic organic aerosols in marine atmospheres will continuously increase
492 under warming conditions.

493 **Table 1. Sum of organic tracer contents (ng/m³) at different locations worldwide.**

Site	Date	Sampler	LEVO	SOA _I	SOA _M	SOA _S	SOA _A	Reference
Wakayama, Japan (Forest)	August 20–30, 2010, Day	TSP	2.5±2.1	281±274	54.6±50.2	1.2±1.2		(Zhu et al., 2016a)
	Night		1.1±0.9	199±207	36.3±33.6	0.9±0.8		
Across China	summer 2012	Anderson sampler		123±79	10.5±6.6	5.0±4.0	2.9±1.5	(Ding et al., 2014)
Beijing (PKU) (urban site)	summer 2007	PM2.5	37-148	59±32	30±14	2.7±1.0		(Yang et al., 2016)
Beijing (YUFA) (suburban site)			34-149	75±43	32±14	3.9±1.5		
Shanghai (BS) (Suburban site)	Apr-May 2010	PM2.5	88.8±57.2	3.8±3.9	6.1±3.7	1.0±0.7	1.1±0.7	(Feng et al., 2013)
Shanghai (XJH) (Urban site)			58.3±27.5	2.5±1.7	2.7±1.3	0.4±0.3	0.6±0.4	
Mt. Tai	summer 2014	PM2.5		56.4±45.6	34.4±28.4			(Zhu et al., 2017)
Central Pearl River Delta	fall-winter 2007	PM2.5		30.8±15.9	6.6±4.4	0.5±0.6		(Ding et al., 2011)
Central Tibetan Plateau	2012-2013	Anderson sampler		26.6±44.2	1.0±0.6	0.09±0.1	0.3±0.2	(Shen et al., 2015)
Mumbai, India	winter 2007	PM10		4.1±2.4	29±22		0.6±0.6	(Fu et al., 2016)
	summer 2007			1.1±0.7	9.4±4.7		0.05±0.1	
Alaska	Spring 2009	TSP		2.4	3.6	0.9		(Haque et al., 2016)
	2008-2009	TSP		4.1	2.0	1.5		
SYS	Spring 2017	TSP	9.6±8.6	45±54	3.5±6.0	0.07±0.1	1.8±1.7	This study
YBS	Spring 2014	TSP	21±11	15±16	1.6±2.0	0.1±0.3	1.1±1.4	This study
NWPO	Spring 2014	TSP	8.2±14	2.3±1.6	1.6±2.7	0.05±0.09	0.3±0.5	This study
East China Sea	18 May to 12 June 2014	TSP	0.09–64.3 (7.3)	0.15–64.0 (8.4)	0.26–87.2 (11.6)	0.16–17.2 (2.9)		(Kang et al., 2018)
Arctic to Antarctic	July to September 2008; November 2009 to April 2010	TSP	5.4±6.2	8.5±11	3.0±5.0			(Hu et al., 2013a; Hu et al., 2013b)

North Pacific Ocean and the Arctic	2003	TSP	0.5±0.4	0.6±0.4	0.06±0.05	0.002±0.005	(Ding et al., 2013)	494
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495 **Data availability.** Most of the data are shown in supplement. Other data are available by contacting the
496 corresponding author.

497 **Supplement.** The supplement related to this article is available.

498 **Author contributions.** XY, TG and JF conceived and led the studies. TG, JW and JF carried out the
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506 **References:**

- 507 Ait-Helal, W., Borbon, A., Sauvage, S., de Gouw, J. A., Colomb, A., Gros, V., Freutel, F., Crippa, M., Afif, C.,
508 Baltensperger, U., Beekmann, M., Doussin, J.-F., Durand-Jolibois, R., Fronval, I., Grand, N., Leonardis, T.,
509 Lopez, M., Michoud, V., Miet, K., Perrier, S., Prévôt, A. S. H., Schneider, J., Siour, G., Zapf, P., and Locoge, N.:
510 Volatile and intermediate volatility organic compounds in suburban Paris: variability, origin and importance for
511 SOA formation, *Atmos. Chem. Phys.*, 14, 10439-10464, <https://doi.org/10.5194/acp-14-10439-2014>, 2014.
- 512 Akagi, S. K., Yokelson, R. J., Wiedinmyer, C., Alvarado, M. J., Reid, J. S., Karl, T., Crouse, J. D., and
513 Wennberg, P. O.: Emission factors for open and domestic biomass burning for use in atmospheric models,
514 *Atmos. Chem. Phys.*, 11, 4039-4072, <https://doi.org/10.5194/acp-11-4039-2011>, 2011.
- 515 Andreae, M. O. and Merlet, P.: Emission of trace gases and aerosols from biomass burning, *Global Biogeochem.*
516 *Cy.*, 15, 955-966, 2001.
- 517 Bahm, K. and Khalil, M. A. K.: A new model of tropospheric hydroxyl radical concentrations, *Chemosphere*, 54,
518 143-166, <https://doi.org/143-166>, 10.1016/j.chemosphere.2003.08.006, 2004.
- 519 Bao, H., Niggemann, J., Luo, L., Dittmar, T., and Kao, S.: Molecular composition and origin of water-soluble
520 organic matter in marine aerosols in the Pacific off China, *Atmos. Environ.*, 191, 27-35,
521 <https://doi.org/10.1016/j.atmosenv.2018.07.059>, 2018.
- 522 Bonsang, B., Polle, C., and Lambert, G.: Evidence for marine production of isoprene, *Geophys. Res. Lett.*, 19,
523 1129-1132, 1992.
- 524 Bougiatioti, A., Bezantakos, S., Stavroulas, I., Kalivitis, N., Kokkalis, P., Biskos, G., Mihalopoulos, N.,
525 Papayannis, A., and Nenes, A.: Biomass-burning impact on CCN number, hygroscopicity and cloud formation
526 during summertime in the eastern Mediterranean, *Atmos. Chem. Phys.*, 16, 7389-7409,
527 <https://doi.org/10.5194/acp-16-7389-2016>, 2016.
- 528 Broadgate, W. J., Liss, P. S., and Penkett, S. A.: Seasonal emissions of isoprene and other reactive hydrocarbon
529 gases from the ocean, *Geophys. Res. Lett.*, 24, 2675-2678, 1997.
- 530 Chen, J., Li, C., Ristovski, Z., Milic, A., Gu, Y., Islam, M. S., Wang, S., Hao, J., Zhang, H., He, C., Guo, H., Fu,
531 H., Miljevic, B., Morawska, L., Thai, P., LAM, Y. F., Pereira, G., Ding, A., Huang, X., and Dumka, U. C.: A
532 review of biomass burning: Emissions and impacts on air quality, health and climate in China, *Sci. Total*

533 Environ., 579, 1000-1034, <https://doi.org/10.1016/j.scitotenv.2016.11.025>, 2017.

534 Claeys, M., Graham, B., Vas, G., Wang, W., Vermeylen, R., Pashynska, V., Cafmeyer, J., Guyon, P., Andreae,
535 M. O., Artaxo, P., and Maenhaut, W.: Formation of secondary organic aerosols through photooxidation of
536 isoprene, *Science*, 303, 1173-1176, 2004.

537 Davis, D. D., Grodzinsky, G., Kasibhatla, P., Crawford, J., Chen, G., Liu, S., Bandy, A., Thornton, D., Guan, H.,
538 and Sandholm, S.: Impact of ship emissions on marine boundary layer NO_x and SO₂ distributions over the
539 Pacific Basin, *Geophys. Res. Lett.*, 28, 235-238, 2001.

540 Ding, X., Zheng, M., Yu, L., Zhang, X., Weber, R. J., Yan, B., Russell, A. G., Edgerton, E. S., and Wang, X.:
541 Spatial and seasonal trends in biogenic secondary organic aerosol tracers and water-soluble organic carbon in
542 the southeastern United States, *Environ. Sci. Technol.*, 42, 5171-5176, <https://doi.org/10.1021/es7032636>, 2008.

543 Ding, X., Wang, X., and Zheng, M.: The influence of temperature and aerosol acidity on biogenic secondary
544 organic aerosol tracers: Observations at a rural site in the central Pearl River Delta region, South China, *Atmos.*
545 *Environ.*, 45, 1303-1311, <https://doi.org/10.1016/j.atmosenv.2010.11.057>, 2011.

546 Ding, X., Wang, X., Xie, Z., Zhang, Z., and Sun, L.: Impacts of Siberian biomass burning on organic aerosols
547 over the North Pacific Ocean and the Arctic: Primary and secondary organic tracers, *Environ. Sci. Technol.*, 47,
548 3149-3157, <https://doi.org/10.1021/es3037093>, 2013.

549 Ding, X., He, Q., Shen, R., Yu, Q., and Wang, X.: Spatial distributions of secondary organic aerosols from
550 isoprene, monoterpenes, β -caryophyllene, and aromatics over China during summer, *J. Geophys. Res.-Atmos.*,
551 119, 877-891, <https://doi.org/10.1002/2014JD021748>, 2014.

552 Ding, X., Zhang, Y., He, Q., Yu, Q., Shen, R., Zhang, Y., Zhang, Z., Lyu, S., Hu, Q., Wang, Y., Li, L., Song,
553 W., and Wang, X.: Spatial and seasonal variations of secondary organic aerosol from terpenoids over China, *J.*
554 *Geophys. Res.-Atmos.*, 121, 661-678, <https://doi.org/10.1002/2016JD025467>, 2016.

555 Ding, X., Zhang, Y., He, Q., Yu, Q., Wang, J., Shen, R., Song, W., Wang, Y., and Wang, X.: Significant
556 increase of aromatics-derived secondary organic aerosol during fall to winter in China, *Environ. Sci. Technol.*,
557 51, 7432-7441, <https://doi.org/10.1021/acs.est.6b06408>, 2017.

558 Duhl, T. R., Helmig, D., and Guenther, A.: Sesquiterpene emissions from vegetation: a review, *Biogeosciences*,
559 5, 761-777, <https://doi.org/10.5194/bg-5-761-2008>, 2008,

560 Ehn, M., Thornton, J. A., Kleist, E., Sipilä, M., Junninen, H., Pullinen, I., Springer, M., Rubach, F., Tillmann, R.,
561 Lee, B., Lopez-Hilfiker, F., Andres, S., Acir, I.-H., Rissanen, M., Jokinen, T., Schobesberger, S., Kangasluoma,
562 J., Kontkanen, J., Nieminen, T., Kurtén, T., Nielsen, L. B., Jørgensen, S., Kjaergaard, H. G., Canagaratna, M.,
563 Dal Maso, M., Berndt, T., Petäjä, T., Wahner, A., Kerminen, V.-M., Kulmala, M., Worsnop, D. R., Wildt, J.,
564 and Mentel, T. F.: A large source of low volatility secondary organic aerosol, *Nature*, 506, 476-479,
565 <https://doi.org/10.1038/nature13032>, 2014.

566 Ekström, S., Nozière, B., and Hansson, H.: The Cloud Condensation Nuclei (CCN) properties of 2-methyltetrols
567 and C3-C6 polyols from osmolality and surface tension measurements, *Atmos. Chem. Phys.*, 9, 973-980, 2009.

568 Ervens, B., Turpin, B. J., and Weber, R. J.: Secondary organic aerosol formation in cloud droplets and aqueous
569 particles (aqSOA): a review of laboratory, field and model studies, *Atmos. Chem. Phys.*, 11, 11069-11102,
570 <https://doi.org/10.5194/acp-11-11069-2011>, 2011.

571 Feng, J. L., Guo, Z. G., Zhang, T. R., Yao, X. H., Chan, C. K., and Fang, M.: Source and formation of
572 secondary particulate matter in PM_{2.5} in Asian continental outflow, *J. Geophys. Res.-Atmos.*, 117, D03302,
573 <https://doi.org/10.1029/2011JD016400>, 2012.

574 Feng, J., Li, M., Zhang, P., Gong, S., Zhong, M., Wu, M., Zheng, M., Chen, C., Wang, H., and Lou, S.:
575 Investigation of the sources and seasonal variations of secondary organic aerosols in PM_{2.5} in Shanghai with
576 organic tracers, *Atmos. Environ.*, 79, 614-622, <https://doi.org/10.1016/j.atmosenv.2013.07.022>, 2013.

577 [Fraser, M. P. and Lakshmanan, K.: Using levoglucosan as a molecular marker for the long-range transport of](#)
578 [biomass combustion aerosols, *Environ. Sci. Technol.*, 34, 4560-4564, <https://doi.org/10.1021/es9912291>, 2000.](#)

579 Fu, P., Kawamura, K., and Miura, K.: Molecular characterization of marine organic aerosols collected during a
580 round-the-world cruise, *J. Geophys. Res.-Atmos.*, 116, D13302, <https://doi.org/10.1029/2011JD015604>, 2011.

581 Fu, P., Aggarwal, S. G., Chen, J., Li, J., Sun, Y., Wang, Z., Chen, H., Liao, H., Ding, A., Umarji, G. S., Patil, R.
582 S., Chen, Q., and Kawamura, K.: Molecular markers of secondary organic aerosol in Mumbai, India, *Environ.*
583 *Sci. Technol.*, 50, 4659-4667, <https://doi.org/10.1021/acs.est.6b00372>, 2016.

584 Fu, T. M., Cao, J. J., Zhang, X. Y., Lee, S. C., Zhang, Q., Han, Y. M., Qu, W. J., Han, Z., Zhang, R., Wang, Y.
585 X., Chen, D., and Henze, D. K.: Carbonaceous aerosols in China: top-down constraints on primary sources and
586 estimation of secondary contribution, *Atmos. Chem. Phys.*, 12, 2725-2746,
587 <https://doi.org/10.5194/acp-12-2725-2012>, 2012.

588 Generoso, S., Bey, I., Attié, J., and Bréon, F.: A satellite- and model-based assessment of the 2003 Russian fires:
589 Impact on the Arctic region, *J. Geophys. Res.-Atmos.*, 112, D15302, <https://doi.org/10.1029/2006JD008344>,
590 2007.

591 Gordon, H., Kirkby, J., Baltensperger, U., Bianchi, F., Breitenlechner, M., Curtius, J., Dias, A., Dommen, J.,
592 Donahue, N. M., Dunne, E. M., Duplissy, J., Ehrhart, S., Flagan, R. C., Frege, C., Fuchs, C., Hansel, A., Hoyle,
593 C. R., Kulmala, M., Kürten, A., Lehtipalo, K., Makhmutov, V., Molteni, U., Rissanen, M. P., Stozkhov, Y.,
594 Tröstl, J., Tsagkogeorgas, G., Wagner, R., Williamson, C., Wimmer, D., Winkler, P. M., Yan, C., and Carslaw,
595 K. S.: Causes and importance of new particle formation in the present-day and preindustrial atmospheres, *J.*
596 *Geophys. Res.-Atmos.*, 122, 8739–8760, <https://doi.org/10.1002/2017JD026844>, 2017.

597 Guenther, A., Karl, T., Harley, P., Wiedinmyer, C., Palmer, P. I., and Geron, C.: Estimates of global terrestrial
598 isoprene emissions using MEGAN (Model of Emissions of Gases and Aerosols from Nature), *Atmos. Chem.*
599 *Phys.*, 6, 3181-3210, 2006.

600 Guenther, A. B., Jiang, X., Heald, C. L., Sakulyanontvittaya, T., Duhl, T., Emmons, L. K., and Wang, X.: The
601 Model of Emissions of Gases and Aerosols from Nature version 2.1 (MEGAN2.1): an extended and updated
602 framework for modeling biogenic emissions, *Geosci. Model Dev.*, 5, 1471-1492,
603 <https://doi.org/10.5194/gmd-5-1471-2012>, 2012.

604 Hackenberg, S. C., Andrews, S. J., Airs, R., Arnold, S. R., Bouman, H. A., Brewin, R. J. W., Chance, R. J.,
605 Cummings, D., Dall'Olmo, G., Lewis, A. C., Minaeian, J. K., Reifel, K. M., Small, A., Tarran, G. A., Tilstone,
606 G. H., and Carpenter, L. J.: Potential controls of isoprene in the surface ocean, *Global Biogeochem. Cy.*, 31,
607 644-662, <https://doi.org/10.1002/2016GB005531>, 2017.

608 Han, Z., Xie, Z., Wang, G., Zhang, R., and Tao, J.: Modeling organic aerosols over east China using a volatility
609 basis-set approach with aging mechanism in a regional air quality model, *Atmos. Environ.*, 124, 186-198,
610 <https://doi.org/10.1016/j.atmosenv.2015.05.045>, 2016.

611 Haque, M. M., Kawamura, K., and Kim, Y.: Seasonal variations of biogenic secondary organic aerosol tracers in
612 ambient aerosols from Alaska, *Atmos. Environ.*, 130, 95-104, 2016.

613 Heald, C. L., Henze, D. K., Horowitz, L. W., Feddema, J., Lamarque, J.-F., Guenther, A., Hess, P. G., Vitt, F.,
614 Seinfeld, J. H., Goldstein, A. H., and Fung, I.: Predicted change in global secondary organic aerosol
615 concentrations in response to future climate, emissions, and land use change, *J. Geophys. Res.-Atmos.*, 113,
616 D05211, <https://doi.org/10.1029/2007JD009092>, 2008.

617 [Hennigan, C. J., Sullivan, A. P., Collett Jr., J. L. and Robinson, A. L.: Levoglucosan stability in biomass](https://doi.org/10.1029/2010GL043088)
618 [burning particles exposed to hydroxyl radicals, *Geophys. Res. Lett.*, 37, L09806,](https://doi.org/10.1029/2010GL043088)
619 <https://doi.org/10.1029/2010GL043088>, 2010.

620 Henze, D. K., Seinfeld, J. H., Ng, N. L., Kroll, J. H., Fu, T.-M., Jacob, D. J., and Heald, C. L.: Global modeling
621 of secondary organic aerosol formation from aromatic hydrocarbons: high-vs. low-yield pathways, *Atmos.*
622 *Chem. Phys.*, 8, 2405-2420, 2008.

623 [Hoffmann, D., Tilgner, A., Iinuma, Y. and Herrmann, H.: Atmospheric stability of levoglucosan: A detailed](https://doi.org/10.1021/es902476f)
624 [laboratory and modeling study, *Environ. Sci. Technol.*, 44, 694-699, https://doi.org/10.1021/es902476f](https://doi.org/10.1021/es902476f), 2010.

625 Hsiao, T., Ye, W., Wang, S., Tsay, S., Chen, W., Lin, N., Lee, C., Hung, H., Chuang, M., and Chantara, S.:
626 Investigation of the CCN activity, BC and UVBC mass concentrations of biomass burning aerosols during the
627 2013 BASELInE campaign, *Aerosol Air Qual. Res.*, 16, 2742-2756, <https://doi.org/10.4209/aaqr.2015.07.0447>,
628 2016.

629 Hu, D. and Yu, J. Z.: Secondary organic aerosol tracers and malic acid in Hong Kong: seasonal trends and
630 origins, *Environ. Chem.*, 10, 381–394, <https://doi.org/10.1071/EN13104>, 2013.

631 Hu, Q., Xie, Z., Wang, X., Kang, H., He, Q., and Zhang, P.: Secondary organic aerosols over oceans via
632 oxidation of isoprene and monoterpenes from Arctic to Antarctic, *Sci. Rep.-UK*, 3, 2280,
633 <https://doi.org/10.1038/srep02280>, 2013a.

634 Hu, Q., Xie, Z., Wang, X., Kang, H., and Zhang, P.: Levoglucosan indicates high levels of biomass burning
635 aerosols over oceans from the Arctic to Antarctic, *Sci. Rep.-UK*, 3, 3119, <https://doi.org/10.1038/srep03119>,
636 2013b.

637 Hu, Q., Qu, K., Gao, H., Cui, Z., Gao, Y., and Yao, X.: Large increases in primary trimethylaminium and
638 secondary dimethylaminium in atmospheric particles associated with cyclonic eddies in the Northwest Pacific
639 Ocean, *J. Geophys. Res.-Atmos.*, 123, 133-146, <https://doi.org/10.1029/2018JD028836>, 2018.

640 Huang, S., Siegert, F., Goldammer, J. G., and Sukhinin, A. I.: Satellite-derived 2003 wildfires in southern
641 Siberia and their potential influence on carbon sequestration, *Int. J. Remote Sens.*, 30, 1479-1492,
642 <https://doi.org/10.1080/01431160802541549>, 2009.

643 John, J. G., Stock, C. A., and Dunne, J. P.: A more productive, but different, ocean after mitigation, *Geophys.*
644 *Res. Lett.*, 42, 9836-9845, <https://doi.org/10.1002/2015GL066160>, 2015.

645 Kanakidou, M., Seinfeld, J. H., Pandis, S. N., Barnes, I., Dentener, F. J., Facchini, M. C., Van Dingenen, R.,
646 Ervens, B., Nenes, A., Nielsen, C. J., Swietlicki, E., Putaud, J. P., Balkanski, Y., Fuzzi, S., Horth, J., Moortgat,
647 G. K., Winterhalter, R., Myhre, C. E. L., Tsigaridis, K., Vignati, E., Stephanou, E. G., and Wilson, J.: Organic
648 aerosol and global climate modelling: a review, *Atmos. Chem. Phys.*, 5, 1053–1123,
649 <https://doi.org/10.5194/acp-5-1053-2005>, 2005.

650 Kang, M., Fu, P., Kawamura, K., Yang, F., Zhang, H., Zang, Z., Ren, H., Ren, L., Zhao, Y., Sun, Y., and Wang,
651 Z.: Characterization of biogenic primary and secondary organic aerosols in the marine atmosphere over the East
652 China Sea, *Atmos. Chem. Phys.*, 18, 13947-13967, <https://doi.org/10.5194/acp-18-13947-2018>, 2018.

653 Kang, M., Guo, H., Wang, P., Fu, P., Ying, Q., Liu, H., Zhao, Y., and Zhang, H.: Characterization and source
654 apportionment of marine aerosols over the East China Sea, *Sci. Total Environ.*, 651, 2679-2688,
655 <https://doi.org/10.1016/j.scitotenv.2018.10.174>, 2019.

656 Kawamura, K., Hoque, M. M. M., Bates, T. S., and Quinn, P. K.: Molecular distributions and isotopic
657 compositions of organic aerosols over the western North Atlantic: Dicarboxylic acids, related compounds,
658 sugars, and secondary organic aerosol tracers, *Org. Geochem.*, 113, 229-238,
659 <https://doi.org/10.1016/j.orggeochem.2017.08.007>, 2017.

660 Kleindienst, T. E., Jaoui, M., Lewandowski, M., Offenberg, J. H., Lewis, C. W., Bhave, P. V., and Edney, E. O.:
661 Estimates of the contributions of biogenic and anthropogenic hydrocarbons to secondary organic aerosol at a
662 southeastern US location, *Atmos. Environ.*, 41, 8288-8300, <https://doi.org/10.1016/j.atmosenv.2007.06.045>,
663 2007.

664 Lauvset, S. K., Tjiputra, J., and Muri, H.: Climate engineering and the ocean: effects on biogeochemistry and
665 primary production, *Biogeosciences*, 14, 5675-5691, <https://doi.org/10.5194/bg-14-5675-2017>, 2017.

666 Lewandowski, M., Piletic, I. R., Kleindienst, T. E., Offenberg, J. H., Beaver, M. R., Jaoui, M., Docherty, K. S.,
667 and Edney, E. O.: Secondary organic aerosol characterisation at field sites across the United States during the
668 spring-summer period, *Int. J. Environ. Anal. Chem.*, 93, 1084-1103,
669 <https://doi.org/10.1080/03067319.2013.803545>, 2013.

670 Li, L., Tang, P., Nakao, S., Kacarab, M., and Cocker, D. R.: Novel approach for evaluating secondary organic
671 aerosol from aromatic hydrocarbons: unified method for predicting aerosol composition and formation, *Environ.*
672 *Sci. Technol.*, 50, 6249-6256, <https://doi.org/10.1021/acs.est.5b05778>, 2016.

673 Li, M., Zhang, Q., Streets, D. G., He, K. B., Cheng, Y. F., Emmons, L. K., Huo, H., Kang, S. C., Lu, Z., Shao,
674 M., Su, H., Yu, X., and Zhang, Y.: Mapping Asian anthropogenic emissions of non-methane volatile organic
675 compounds to multiple chemical mechanisms, *Atmos. Chem. Phys.*, 14, 5617-5638, 2014.

676 [Li, R., Wang, Z., Cui, L., Fu, H., Zhang, L., Kong, L., Chen, W., and Chen, J.: Air pollution characteristics in](#)

677 [China during 2015–2016: Spatiotemporal variations and key meteorological factors, *Sci. Total Environ.*, 648,](#)
678 [902-915, 2019.](#)

679 Meskhidze, N. and Nenes, A.: Phytoplankton and cloudiness in the Southern Ocean, *Science*, 314, 1419-1423,
680 <https://doi.org/10.1126/science.1131779>, 2006.

681 Mochida, M., Kawamura, K., Fu, P., and Takemura, T.: Seasonal variation of levoglucosan in aerosols over the
682 western North Pacific and its assessment as a biomass-burning tracer, *Atmos. Environ.*, 44, 3511-3518,
683 <https://doi.org/10.1016/j.atmosenv.2010.06.017>, 2010.

684 Murphy, D. M., Chow, J. C., Leibensperger, E. M., Malm, W. C., Pitchford, M., Schichtel, B. A., Watson, J. G.,
685 and White, W. H.: Decreases in elemental carbon and fine particle mass in the United States, *Atmos. Chem.*
686 *Phys.*, 11, 4679-4686, <https://doi.org/10.5194/acp-11-4679-2011>, 2011.

687 Nichols, M. A.: Spatial and temporal variability of marine primary organic aerosols over the global oceans: a
688 review, University of Maryland College Park, 2016.

689 Peñuelas, J. and Staudt, M.: BVOCs and global change, *Trends Plant Sci.*, 15, 133-144,
690 <https://doi.org/10.1016/j.tplants.2009.12.005>, 2010.

691 Rinne, H. J. I., Guenther, A. B., Greenberg, J. P., and Harley, P. C.: Isoprene and monoterpene fluxes measured
692 above Amazonian rainforest and their dependence on light and temperature, *Atmos. Environ.*, 36, 2421-2426,
693 [https://doi.org/10.1016/S1352-2310\(01\)00523-4](https://doi.org/10.1016/S1352-2310(01)00523-4), 2002.

694 Running, S. W.: Is global warming causing more, larger wildfires? *Science*, 313, 927-928,
695 <https://doi.org/10.1126/science.1130370>, 2006.

696 Sharma, S., Lavoué, D., Cachier, H., Barrie, L. A., and Gong, S. L.: Long-term trends of the black carbon
697 concentrations in the Canadian Arctic, *J. Geophys. Res.-Atmos.*, 109, D15203,
698 <https://doi.org/10.1029/2003JD004331>, 2004.

699 Shen, R., Ding, X., He, Q., Cong, Z., and Wang, X.: Seasonal variation of secondary organic aerosol tracers in
700 Central Tibetan Plateau, *Atmos. Chem. Phys.*, 15, 8781-8793, 2015.

701 Surratt, J. D., Chan, A. W. H., Eddingsaas, N. C., Chan, M., Loza, C. L., Kwan, A. J., Hersey, S. P., Flagan, R.
702 C., Wennberg, P. O., and Seinfeld, J. H.: Reactive intermediates revealed in secondary organic aerosol
703 formation from isoprene, *Proc. Natl. Acad. Sci. U.S.A.*, 107, 6640-6645,
704 <https://doi.org/10.1073/pnas.0911114107>, 2010.

705 Tarvainen, V., Hakola, H., Hellén, H., Bäck, J., Hari, P., and Kulmala, M.: Temperature and light dependence of
706 the VOC emissions of Scots pine, *Atmos. Chem. Phys.*, 5, 989-998, 2005.

707 van der Werf, G. R., Randerson, J. T., Giglio, L., Collatz, G. J., Kasibhatla, P. S., and Arellano Jr, A. F.:
708 Interannual variability in global biomass burning emissions from 1997 to 2004, *Atmos. Chem. Phys.*, 6,
709 3423-3441, 2006.

710 Wang, F., Guo, Z., Lin, T., Hu, L., Chen, Y., and Zhu, Y.: Characterization of carbonaceous aerosols over the
711 East China Sea: The impact of the East Asian continental outflow, *Atmos. Environ.*, 110, 163-173,
712 <https://doi.org/10.1016/j.atmosenv.2015.03.059>, 2015.

713 Warneke, C., Froyd, K. D., Brioude, J., Bahreini, R., Brock, C. A., Cozic, J., de Gouw, J. A., Fahey, D. W.,
714 Ferrare, R., Holloway, J. S., Middlebrook, A. M., Miller, L., Montzka, S., Schwarz, J. P., Sodemann, H.,
715 Spackman, J. R., and Stohl, A.: An important contribution to springtime Arctic aerosol from biomass burning in
716 Russia, *Geophys. Res. Lett.*, 37, L01801, <https://doi.org/10.1029/2009GL041816>, 2010.

717 Yang, F., Gu, Z., Feng, J., Liu, X., and Yao, X.: Biogenic and anthropogenic sources of oxalate in PM_{2.5} in a
718 mega city, Shanghai, *Atmos. Res.*, 138, 356-363, <https://doi.org/10.1016/j.atmosres.2013.12.006>, 2014.

719 Yang, F., Kawamura, K., Chen, J., Ho, K., Lee, S., Gao, Y., Cui, L., Wang, T., and Fu, P.: Anthropogenic and
720 biogenic organic compounds in summertime fine aerosols (PM_{2.5}) in Beijing, China, *Atmos. Environ.*, 124,
721 166-175, 2016.

722 [Yao, X., Xu, X., Sabaliauskas, K., and Fang, M.: Comment on “Atmospheric Particulate Matter Pollution during](#)
723 [the 2008 Beijing Olympics”, *Environ. Sci. Technol.*, 43, 7589, <https://doi.org/10.1021/es902276p>, 2009.](#)

724 Zhang, H., Surratt, J. D., Lin, Y. H., Bapat, J., and Kamens, R. M.: Effect of relative humidity on SOA

725 formation from isoprene/NO photooxidation: enhancement of 2-methylglyceric acid and its corresponding
726 oligoesters under dry conditions, *Atmos. Chem. Phys.*, 11, 6411-6424,
727 <https://doi.org/10.5194/acp-11-6411-2011>, 2011.

728 Zhang, Q., He, K., and Huo, H.: Policy: cleaning China's air, *Nature*, 484, 161-162, 2012.

729 Zhang, Y., Yang, X., Brown, R., Yang, L., Morawska, L., Ristovski, Z., Fu, Q., and Huang, C.: Shipping
730 emissions and their impacts on air quality in China, *Sci. Total Environ.*, 581-582, 186-198,
731 <https://doi.org/10.1016/j.scitotenv.2016.12.098>, 2017.

732 Zhang, Z., Zhang, Y., Wang, X., Lü, S., Huang, Z., Huang, X., Yang, W., Wang, Y., and Zhang, Q.:
733 Spatiotemporal patterns and source implications of aromatic hydrocarbons at six rural sites across China's
734 developed coastal regions, *J. Geophys. Res.-Atmos.*, 121, 6669-6687, <https://doi.org/10.1002/2016JD025115>,
735 2016.

736 Zhu, C., Kawamura, K., and Kunwar, B.: Effect of biomass burning over the western North Pacific Rim:
737 wintertime maxima of anhydrosugars in ambient aerosols from Okinawa, *Atmos. Chem. Phys.*, 15, 1959-1973,
738 <https://doi.org/10.5194/acp-15-1959-2015>, 2015.

739 Zhu, C., Kawamura, K., Fukuda, Y., Mochida, M., and Iwamoto, Y.: Fungal spores overwhelm biogenic organic
740 aerosols in a midlatitudinal forest, *Atmos. Chem. Phys.*, 16, 7497-7506, 2016a.

741 Zhu, C., Kawamura, K., and Fu, P.: Seasonal variations of biogenic secondary organic aerosol tracers in Cape
742 Hedo, Okinawa, *Atmos. Environ.*, 130, 113-119, <https://doi.org/10.1016/j.atmosenv.2015.08.069>, 2016b.

743 Zhu, Y., Yang, L., Kawamura, K., Chen, J., Ono, K., Wang, X., Xue, L., and Wang, W.: Contributions and
744 source identification of biogenic and anthropogenic hydrocarbons to secondary organic aerosols at Mt. Tai in
745 2014, *Environ. Pollut.*, 220, 863-872, 2017.

746 Zhu, Y., Li, K., Shen, Y., Gao, Y., Liu, X., Yu, Y., Gao, H., and Yao, X.: New particle formation in the marine
747 atmosphere during seven cruise campaigns, *Atmos. Chem. Phys.*, 19, 89-113,
748 <https://doi.org/10.5194/acp-19-89-2019>, 2019.