

1 **Flexpart v10.1 simulation of source contributions to Arctic black carbon**

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Abstract

14 **Abstract**
15 The Arctic environment is undergoing rapid changes such as faster warming than the
16 global average and exceptional melting of glaciers in Greenland. Black carbon (BC) particles,
17 which are a short-lived climate pollutant, are one cause of Arctic warming and glacier
18 melting. However, the sources of BC particles are still uncertain. We simulated the potential
19 emission sensitivity of atmospheric BC present over the Arctic (north of 66° N) using the
20 Flexpart Lagrangian transport model (version 10.1). This version includes a new aerosol wet
21 removal scheme, which better represents particle-scavenging processes than older versions
22 did. Arctic BC at the surface (0–500 m) and high altitudes (4750–5250 m) is sensitive to
23 emissions in high latitude (north of 60° N) and mid-latitude (30–60° N) regions, respectively.
24 Geospatial sources of Arctic BC were quantified, with a focus on emissions from
25 anthropogenic activities (including domestic biofuel burning) and open biomass burning
26 (including agricultural burning in the open field) in 2010. We found that anthropogenic
27 sources contributed 82 % and 83 % of annual Arctic BC at the surface and high altitudes,
28 respectively. Arctic surface BC comes predominantly from anthropogenic emissions in
29 Russia (56 %), with gas flaring from the Yamalo-Nenets Autonomous Okrug and Komi
30 Republic being the main source (31 % of Arctic surface BC). These results highlight the need
31 for regulations to control BC emissions from gas flaring to mitigate the rapid changes in the
32 Arctic environment. In summer, combined open biomass burning in Siberia, Alaska, and
33 Canada contributes 56–85 % (75 % on average) and 40–72 % (57 %) of Arctic BC at the
34 surface and high altitudes, respectively. A large fraction (40 %) of BC in the Arctic at high
35 altitudes comes from anthropogenic emissions in East Asia, which suggests that the rapidly
36 growing economies of developing countries could have a non-negligible effect on the Arctic.
37 To our knowledge, this is the first year-round evaluation of Arctic BC sources that has been

38 performed using the new wet deposition scheme in Flexpart. The study provides a scientific
39 basis for actions to mitigate the rapidly changing Arctic environment.
40

41 **1 Introduction**

42 The Arctic region has experienced warming at a rate twice that of the global average in
43 recent decades (Cohen et al., 2014). The Arctic cryosphere has been undergoing
44 unprecedented changes since the mid-1800s (Trusel et al., 2018). Glacier cover in Greenland
45 reached its historically lowest level in summer 2012 (Tilling et al., 2015). Evidence indicates
46 that the emissions and transport of greenhouse gases and aerosols to the Arctic region are
47 contributing to such warming and melting of snow and ice (Keegan et al., 2014; Najafi et al.,
48 2015). Short-lived climate pollutants such as black carbon (BC) particles (e.g., Sand et al.,
49 2016; Yang et al., 2019), sulfate aerosol (Yang et al., 2018), tropospheric ozone, and
50 methane greatly affect the Arctic climate (AMAP, 2015; Quinn et al., 2008).

51 BC particles are emitted during incomplete combustion of fossil fuels, biofuels, and
52 biomass. BC warms the atmosphere by direct absorption of solar radiation. The deposition
53 of BC on snow and ice surfaces accelerates their melting through decreasing albedo, which
54 contributes to the rapid loss of glaciers. In the Arctic region, ground-based observations
55 have indicated that BC shows clear seasonal variations, with elevated mass concentrations
56 in winter and spring (the so-called Arctic haze) and low values in summer (Law and Stohl,
57 2007). Such seasonal variations are explained by increased transport from lower latitudes in
58 the cold season and increased wet scavenging in the warm season (Shaw, 1995; Garrett et
59 al., 2011; Shen et al., 2017).

60 The presence of BC particles in the Arctic is mainly attributed to emissions in high-latitude
61 regions outside the Arctic, such as northern Europe and Russia (Stohl, 2006; Brock et al.,
62 2011). This is partly caused by the polar dome (Stohl, 2006), which is formed because of the
63 presence of constant potential temperature near the surface. The emissions in high-latitude
64 regions are transported to the Arctic region and trapped in the dome, which increases the

65 surface concentration. Recently, Schmale et al. (2018) suggested that local emissions from
66 within the Arctic are another important source, and these are expected to increase in the
67 future.

68 Although numerous studies have been performed, results regarding regional
69 contributions of BC sources in the Arctic are still inconclusive. For example, ground-based
70 observations and Lagrangian transport model results reported by Winiger et al. (2016)
71 showed that BC in Arctic Scandinavia is predominantly linked to emissions in Europe. Over
72 the whole Arctic region (north of 66° N), Russia contributes 62 % to surface BC in terms of
73 the annual mean (Ikeda et al., 2017). Gas flaring in Russia has been identified as a major
74 (42 %) source of BC at the Arctic surface (Stohl et al., 2013). Xu et al. (2017) found that
75 anthropogenic emissions from northern Asia contribute 40–45 % of Arctic surface BC in
76 winter and spring. However, the results of some other studies have suggested that Russia,
77 Europe, and South Asia each contribute 20–25 % of BC to the low-altitude springtime Arctic
78 haze (Koch and Hansen, 2005). Sand et al. (2016) found that the surface temperature in the
79 Arctic is most sensitive to emissions in Arctic countries, and Asian countries contribute
80 greatly to Arctic warming because of the large absolute amount of emissions. With these
81 large disagreements among studies, it is thus necessary to unveil BC sources in the Arctic
82 with high precision simulations.

83 Various models have been used to investigate BC sources in the Arctic. Depending on the
84 simulation method, these models are generally categorized as Lagrangian transport models
85 (Hirdman et al., 2010; Liu et al., 2015; Stohl et al., 2006, 2013), chemical transport models
86 (Ikeda et al., 2017; Koch and Hansen, 2005; Qi et al., 2017; Shindell et al., 2008; Wang et al.,
87 2011; Xu et al., 2017), and global climate models (Ma et al., 2013; Schacht et al., 2019; H.
88 Wang et al., 2014) (Table 1). The treatment of wet-scavenging parameterizations is a key

89 factor affecting the model performance, which determines the uncertainties related to BC
90 particle removal (Kipling et al., 2013; Schacht et al., 2019; Q. Wang et al., 2014). The use of
91 emission inventories is another important factor that affects the simulation results (Dong et
92 al., 2019). The observations of BC that are used for model comparisons may be biased by a
93 factor of 2 depending on the method used (Sinha et al., 2017; Sharma et al., 2017). There
94 are still large uncertainties regarding the sources of BC in the Arctic with respect to emission
95 sectors (anthropogenic sources and open biomass burning) and geospatial contributions
96 (Eckhardt et al., 2015).

97 The FLEXible PARTicle dispersion model (Flexpart) had been used to investigate the
98 transport pathways and source contributions of BC in the Arctic (Stohl et al., 1998, 2006,
99 2013). Of Flexpart model up to version 9, wet removal was treated considering below-cloud
100 and within-cloud scavenging processes (Hertel et al., 1995; McMahon and Denison, 1979),
101 which depends on cloud liquid water content, precipitation rate and the depth of the cloud.
102 However, clouds were parameterized based on relative humidity, clouds frequently
103 extended to the surface and at times no clouds could be found in grid cells, with unrealistic
104 precipitation (Grythe et al., 2017). Recently, version 10 of Flexpart had been developed in
105 which cloud is differentiated into liquid, solid, and mixed phase, the cloud distribution is
106 more consistent with the precipitation data (Grythe et al., 2017). This improvement in the
107 cloud distribution and phase leads to a more realistic distribution of below-cloud and in-
108 cloud scavenging events. In this study, we quantified region-separated sources of BC in the
109 Arctic in 2010 by using Flexpart v10.1. We first evaluated the model performance by
110 comparing the results with those based on observations at surface sites. The source
111 contributions of emission sectors and geospatial contributions were evaluated by
112 incorporating the Arctic BC footprint into the emission inventories.

113 **2 Materials and methods**

114 2.1 Transport model

115 The Flexpart model (version 10.1) was run in backward mode to simulate BC footprints in
116 the Arctic region. The calculation of wet deposition was improved compared with those in
117 previous versions because in-cloud scavenging and below-cloud scavenging of particles were
118 separately calculated (Grythe et al., 2017). In previous versions of Flexpart, in the in-cloud
119 scavenging scheme, the aerosol scavenging coefficient depended on the cloud water
120 content, which was calculated according to an empirical relationship with precipitation rate,
121 in which all aerosols had the same nucleation efficiency (Hertel et al., 1995 ; Stohl et al.,
122 2005). In the new version, the in-cloud scavenging scheme depends on the cloud water
123 phase (liquid, ice, or mixed phase). Aerosols were set as ice nuclei for ice clouds and as
124 cloud condensation nuclei for liquid-water clouds, respectively. For mixed-phase clouds, it
125 was assumed that 10 % of aerosols are ice nuclei and 90 % are cloud condensation nuclei,
126 because BC is much more efficiently removed in liquid water clouds than in ice clouds (Cozic
127 et al., 2007; Grythe et al., 2017). The below-cloud scavenging scheme can parameterize
128 below-cloud removal as a function of aerosol particle size, and precipitation type (snow or
129 rain) and intensity. The biases produced in simulations using the new scheme are therefore
130 smaller than those in the old scheme for wet deposition of aerosols, especially at high
131 latitudes (Grythe et al., 2017).

132 The Arctic region is defined as areas north of 66° N. The potential BC emission
133 sensitivities at two heights in the Arctic region, i.e., the surface (0–500 m) and high altitudes
134 (4750–5250 m), were simulated. The Flexpart outputs were set as gridded retention times.
135 We performed tests at 500, 2000, and 5000 m, and chose 500 m as the upper boundary
136 height of the model output. The model was driven with operational analytical data from the

137 European Centre for Medium-Range Weather Forecasts (ECMWF) at a spatial resolution of
138 $1^\circ \times 1^\circ$ with 61 vertical levels. Temporally, ECMWF has a resolution of 3 h, with 6 h analysis
139 and 3 h forecast time steps. The simulation period was set at 60 days backward starting
140 from each month in 2010. The maximum lifetime of BC was set at 20 days because its
141 suspension time in the upper atmosphere during long-range transport is longer than that at
142 the surface level (Stohl et al., 2013). We implemented the wet deposition scheme in the
143 backward calculations, but it was not represented in the default setting (Flexpart v10.1,
144 <https://www.flexpart.eu/downloads>, obtained 10 April 2017).

145 The chemistry and microphysics could not be resolved by Flexpart. The model therefore
146 ignores hydrophobic to hydrophilic state changes and size changes of BC, and assumes that
147 all BC particles are aged hydrophilic particles. This may lead to an overestimation of BC
148 removal and hence force underestimation of simulated BC concentration, especially of fossil
149 fuel combustion sources where BC could be in the hydrophobic state for a few days. A
150 logarithmic size distribution of BC with a mean diameter of $0.16 \mu\text{m}$ and a standard
151 deviation of 1.96, in accordance with our ship observations in the Arctic, was used (Taketani
152 et al., 2016). The particle density was assumed to be 2000 kg m^{-3} , and 1 million
153 computational particles were randomly generated in the Arctic region for the backward
154 runs.

155 Four ground-based observations made during the period 2007–2011 were used to
156 validate the model performance. The potential BC emission sensitivity at 0–500 m above
157 ground level from a 0.1° grid centered at each site was simulated. Other model
158 parameterizations were consistent with those for the Arctic region, except that 200 000
159 computational particles were released.

160 2.2 Emission inventories

161 We focused on BC sources from anthropogenic emissions and open biomass burning. The
162 Hemispheric Transport of Air Pollution version 2 inventory (HTAP2) for 2010 was used for
163 monthly anthropogenic BC emissions (Janssens-Maenhout et al., 2015), which include
164 sectors from energy, industry, residential and transport. It is worth noting that the
165 residential sector includes not only combustions of fossil fuels, but also biofuels. However,
166 as it has been reported that BC emissions in Russia were underestimated in HTAP2, we used
167 the BC emissions reported by Huang et al. (2015) for Russia, in which the annual BC
168 emissions were 224 Gg yr⁻¹. For open biomass burning, we used the monthly BC emissions
169 from the Global Fire Emissions Database version 3 inventory (GFED3) (van der Werf et al.,
170 2010) for the purposes of intercomparison with other studies, as this version is widely used.
171 The term “open biomass burning” here indicates burning of biomass in the open field as is
172 determined by the remote sensing measurement basis, including forest, agricultural waste,
173 peat fires, grassland and savanna, woodland, deforestation and degradation, where biofuel
174 burning for residential use is not included. Geospatial distributions of emissions from
175 anthropogenic sources and open biomass burning in January and July are shown in Fig. S1.

176 2.3 Calculation of Arctic BC source contributions

177 The source contributions to Arctic BC were derived by incorporating the gridded
178 retention time into the column emission flux, which was derived from the emission
179 inventories in each grid. Calculations for anthropogenic sources and open biomass burning
180 were performed separately and the sum was used. For anthropogenic sources, the regions
181 were separated into North America and Canada (25–80° N, 50–170° W), Europe (30–80° N,
182 0–30° E), Russia (53–80° N, 30–180° E), East Asia (35–53° N, 75–150° E and 20–35° N, 100–
183 150° E), and others (the rest) (Fig. 1a). For open biomass burning sources, the regions were

184 separated into Alaska and Canada (50–75° N, 50–170° W), Siberia (50–75° N, 60–180° E),
185 and others (Fig. 1b).

186 2.3 Observations

187 BC levels simulated by Flexpart were compared with those based on surface observations
188 at four sites: Barrow, USA (156.6° W, 71.3° N, 11 m asl), Alert, Canada (62.3° W, 82.5° N,
189 210 m asl), Zeppelin, Norway (11.9° E, 78.9° N, 478 m asl), and Tiksi, Russia (128.9° E, 71.6°
190 N, 8 m asl). Aerosol light absorption was determined by using particle soot absorption
191 photometers (PSAPs) at Barrow, Alert, and Zeppelin, and an aethalometer at Tiksi. For PSAP
192 measurements, the equivalent BC values were derived using a mass absorption efficiency of
193 $10 \text{ m}^2 \text{ g}^{-1}$. The equivalent BC at Tiksi, which was determined with an aethalometer, was
194 obtained directly. These measurement data were obtained from the European Monitoring
195 and Evaluation Programme and World Data Centre for Aerosols database
196 (<http://ebas.nilu.no>) (Tørseth et al., 2012).

197 It is worth noting that uncertainties could be introduced by using different BC
198 measurement techniques. An evaluation of three methods for measuring BC at Alert,
199 Canada indicated that an average of the refractory BC determined with a single-particle soot
200 photometer (SP2) and elemental carbon (EC) determined from filter samples give the best
201 estimate of BC mass (Sharma et al., 2017). Xu et al. (2017) reported that the equivalent BC
202 determined with a PSAP was close to the average of the values for refractory BC and EC at
203 Alert. In this study, we consider that the equivalent BC values determined with a PSAP at
204 Barrow, Alert, and Zeppelin to be the best estimate. There may be uncertainties in the
205 equivalent BC observations performed with an aethalometer at Tiksi because of co-existing
206 particles such as light-absorptive organic aerosols, scattering particles, and dusts
207 (Kirchstetter et al., 2004; Lack and Langridge, 2013). Interference by the filter and

208 uncertainties in the mass absorption cross section could also contribute to the bias
209 observed in measurements made with an aethalometer at Tiksi.

210 **3 Results and discussion**

211 **3.1. Comparisons of simulations with BC observations at Arctic surface sites**

212 Flexpart generally reproduced the seasonal variations in BC at four Arctic sites well
213 [Pearson correlation coefficient (R) = 0.53–0.80, root-mean-square error (RMSE) = 15.1–56.8
214 ng m^{-3}] (Fig. 2). Winter maxima were observed for the four sites, while a secondary
215 elevation was observed for Alert and Tiksi. At Barrow, the observed high values of BC were
216 unintentionally excluded during data screening in the forest fire season in summer (Stohl et
217 al., 2013); the original observed BC is supposed to be higher as was reflected by the
218 simulation. This seasonality is probably related to relatively stronger transport to the Arctic
219 region in winter, accompanied by lower BC aging and inefficient removal, as simulated by
220 older versions of Flexpart (Eckhardt et al., 2015; Stohl et al., 2013).

221 From January to May at Barrow and Alert, the mean BC simulated by Flexpart v10.1 were
222 32.2 ng m^{-3} and 31.2 ng m^{-3} , respectively. Which was 46 % lower than the observations
223 (59.3 ng m^{-3} and 58.2 ng m^{-3} , respectively). This is probably related to the inadequate BC
224 emission in the inventory, although seasonal variations in residential heating are included in
225 HTAP2, which would reduce the simulation bias (Xu et al., 2017). Simulations by GEOS-Chem
226 using the same emission inventories also underestimated BC levels at Barrow and Alert
227 (Ikeda et al., 2017; Xu et al., 2017). The underestimation by Flexpart could also be partly
228 contributed by the assumption that all particles are hydrophilic, where the BC scavenging
229 could be overestimated. The corresponding uncertainties are larger in winter months, when
230 there are more sources from fossil fuel combustion.

231 At Zeppelin, the Flexpart-simulated BC (39.1 ng m^{-3} for annual mean) was 85 % higher
232 than the observed value (21.1 ng m^{-3} for annual mean), especially in winter (112% higher). It
233 has been reported that riming in mixed-phase clouds occurs frequently at Zeppelin (Qi et al.,
234 2017). During the riming process, BC particles act as ice particles and collide with the
235 relatively numerous water drops, which form frozen cloud droplets, and then snow is
236 precipitated. This results in relatively efficient BC scavenging (Hegg et al., 2011). Such a
237 process could not be dealt with by the model. At Tiksi, Flexpart underestimated BC (74.4 ng
238 m^{-3} for annual mean) in comparison with observation (104.2 ng m^{-3} for annual mean). Other
239 than the hydrophilic BC assumption and underestimated BC emission in the simulation as
240 the cases for Barrow and Alert, the observations at Tiksi by an aethalometer could
241 contain light-absorbing particles other than BC, resulting in higher observed
242 concentrations if compared with those obtained by SP2, EC and PSAP.

243 Anthropogenic emissions are the main sources of BC at the four Arctic sites from late
244 autumn to spring, whereas open biomass burning emissions make large contributions in
245 summer. From October to April, anthropogenic emissions accounted for 87–100 % of BC
246 sources at all the observation sites. At Barrow, open biomass burning accounted for 35–
247 78 % of BC in June–September (Fig. 2). There are large interannual variations in both
248 observed and simulated BC (Fig. S2). In June–August 2010, the mean contributions of open
249 biomass burning to BC were 6.3, 2.4, and 8.6 times those from anthropogenic sources at
250 Alert, Zeppelin, and Tiksi, respectively. In this study, we focused on BC in the Arctic region in
251 2010.

252 **3.2 Potential emission sensitivity of Arctic BC**

253 The potential emission sensitivities (footprint) of Arctic BC showed different patterns
254 with respect to altitude. The Arctic surface is sensitive to emissions at high latitudes ($>60^\circ$

255 N). Air masses stayed for over 60 s in each of the 1° grids from the eastern part of northern
256 Eurasia and the Arctic Ocean before being transported to the Arctic surface in the winter,
257 represented by January (Fig. 3a). In comparison, during the summer, represented by July, BC
258 at the Arctic surface was mainly affected by air masses that originated from the Arctic
259 Ocean and the Norwegian Sea (Fig. 3b). These results imply that local BC emissions within
260 the Arctic regions, although relatively weak compared with those from the mid-latitude
261 regions, could strongly affect Arctic air pollution. Local BC emissions are important in the
262 wintertime because the relatively stable boundary layer does not favor pollution dispersion.
263 Recent increases in anthropogenic emissions in the Arctic region, which have been caused
264 by the petroleum industry and development of the Northern Sea Route, are expected to
265 cause deterioration of air quality in the Arctic. Socio-economic developments in the Arctic
266 region would increase local BC emissions, and this will be a non-negligible issue in the future
267 (Roiger et al., 2015; Schmale et al., 2018).

268 BC at high altitudes in the Arctic is more sensitive to mid-latitude (30–60° N) emissions,
269 especially in wintertime. In January, air masses hovered over the Bering Sea and the North
270 Atlantic Ocean before arriving at the Arctic (Fig. 3c). A notable corridor at 30–50° N covering
271 Eurasia and the United States was the sensitive region that affected BC at high altitudes in
272 the Arctic in January. These results indicate that mid-latitude emissions, especially those
273 with relatively large strengths from East Asia, East America, and Europe, could alter the
274 atmospheric constituents at high altitudes in the Arctic. Central to east Siberia was the most
275 sensitive region for BC at high altitudes in the Arctic in July (Fig. 3d). These results suggest
276 that pollutants from frequent and extensive wildfires in Siberia in summer are readily
277 transported to high altitudes in the Arctic. Boreal fires are expected to occur more
278 frequently and over larger burning areas under future warming (Veira et al., 2016),

279 therefore the atmospheric constituents and climate in the Arctic could undergo more rapid
280 changes.

281 **3.3 Seasonal variations and sources of Arctic surface BC**

282 Arctic surface BC showed clear seasonal variations, with a primary peak in winter–spring
283 (December–March, 61.8–82.8 ng m⁻³) and a secondary peak in summer (July, 52.7 ng m⁻³).
284 BC levels were relatively low in May–June (21.8–23.1 ng m⁻³) and September–November
285 (34.1–40.9 ng m⁻³) (Fig. 4a). This seasonality agrees with observations and simulations at
286 Alert, Tiksi, and Barrow if consider the unintentional data exclusion (Stohl et al., 2013), and
287 previous studies targeting the whole Arctic (Ikeda et al., 2017; Xu et al., 2017). Compared
288 with the study reported by Stohl et al. (2013), the current work using the new scheme
289 produced smaller discrepancies between the simulated data and observations. Although the
290 simulation periods (monthly means for 2007–2011 in this study and for 2008–2010 in the
291 old scheme) and the anthropogenic emission inventories (HTAP2 in this study and ECLIPSE4
292 in the previous study) are different, the new scheme shows potential for better representing
293 BC transport and removal processes in the Arctic.

294 The annual mean Arctic BC at the surface was estimated to be 48.2 ng m⁻³. From October
295 to April, anthropogenic sources accounted for 96–100 % of total BC at the Arctic surface.
296 Specifically, anthropogenic emissions from Russia accounted for 61–76 % of total BC in
297 October–May (56 % annually), and was the dominant sources of Arctic BC at the surface.
298 From an isentropic perspective, the meteorological conditions in winter favored the
299 transport of pollutants from northern Eurasia to the lower Arctic, along with diabatic cooling
300 and strong inversions (Klonecki et al., 2003). In comparison, open biomass burning from
301 boreal regions accounted for 56–85 % (75 % on average) of Arctic BC at the surface in
302 summer; open biomass burning emissions from North America and Canada accounted for

303 54 % of total Arctic surface BC in June, and those from Siberia accounted for 59–61 % in
304 July–August. Wildfires in the boreal forests in summer had a major effect on air quality in
305 the Arctic.

306 On an annual basis, anthropogenic sources and open biomass burning emissions
307 accounted for 82 % and 18 %, respectively, of total Arctic surface BC. In which, gas flaring
308 and residential burning (including burning of fossil fuels and biofuels) are accounting for
309 36 % (28–57 % in October–March) and 15 % (13–25 % in October–March), respectively (Fig.
310 5a-b). Our results support Stohl et al. (2013) that residential combustion emissions,
311 especially in winter are important sources of Arctic BC (Table 1). We estimated a
312 contribution of gas flaring to Arctic surface BC of 17.5 ng m^{-3} (36% of total). In comparison,
313 the value was estimated as 11.8 ng m^{-3} using an average Arctic surface BC of 28 ng m^{-3} and
314 a fraction from gas flaring of 42 % evaluated by earlier versions of Flexpart (Stohl et al.,
315 2013; Winiger et al., 2019). The different contribution could be partly attributed to the
316 difference in gas flaring emission inventory. BC emission from gas flaring in Russia by Huang
317 et al. (2015) was used in the current study, where total BC emission from gas flaring in
318 Russia in 2010 was ca. 81.1 kilotonne, which was larger than the estimate of ca. 64.9
319 kilotonne by GAINS inventory (Klimont et al., 2017) used by Stohl et al. (2013). Moreover,
320 Adopting ECLIPSEv5 inventory as was used by Winiger et al. (2019), we estimated that gas
321 flaring was contributing 14.4 ng m^{-3} to Arctic surface BC using Flexpart v10.1, a value 22 %
322 higher than those obtained using Flexpart v9. This difference could be attributed to the
323 improvement of the wet-scavenging scheme by Flexpart v10.1.

324 A recent study based on isotope observations at the Arctic sites and Flexpart v9.2
325 simulation suggested that open biomass burning, including open field burning and
326 residential biofuel burning, contributed 39 % of annual BC in 2011–2015 (Winiger et al.,

327 2019) (Table 1). In comparison, we estimated that residential burning and open biomass
328 burning together account for 33 % of total Arctic surface BC. As the residential burning in
329 our study includes burning of both biofuels and fossil fuels, our results indicated that
330 biomass burning has a relatively smaller contribution. Other than the differences in BC
331 removal treatment between different versions of the model, the contribution difference
332 could also be attributed to the different emission inventories and years (2010 versus 2011-
333 2015).

334 The geospatial contributions of anthropogenic sources and open biomass burning
335 emissions can be further illustrated by taking January and July as examples. In January, high
336 levels of anthropogenic emissions from Russia (contributing 64 % of Arctic surface BC),
337 Europe (18 %), and East Asia (9 %) were identified (Fig. 6a). Specifically, Yamalo-Nenets
338 Autonomous Okrug in Russia, which has the largest reserves of Russia's natural gas and oil
339 (Filimonova et al., 2018), was the most notable emission hotspot, which suggests gas-flaring
340 sources. The Komi Republic in Russia was also identified as a strong anthropogenic emitter
341 contributing to Arctic surface BC. These gas-flaring industrial regions in Russia (58–69° N,
342 68–81°E) together contributed 33 % and 31 % of Arctic surface BC for January and the
343 annual mean, respectively. Recently, Dong et al. (2019) evaluated BC emission inventories
344 using GEOS-Chem and proposed that using the inventory compiled by Huang et al. (2015)
345 for Russia, in which gas flaring accounted for 36 % of anthropogenic emissions, had no
346 prominent impact on the simulation performance in Russia and the Arctic. They suggested
347 that use of a new global inventory for BC emissions from natural gas flaring would improve
348 the model performance (Huang and Fu, 2016). These results suggest that inclusion of BC
349 emissions from gas flaring on the global scale is necessary for further BC simulations.

350 In Europe, a relatively high contribution of anthropogenic emissions to Arctic surface BC
351 in January was made by Poland (50–55° N, 15–24° E, contributing 4 % of Arctic surface BC)
352 because of relatively large emission fluxes in the region (Fig. S1a). Anthropogenic emissions
353 from East China, especially those north of ~33° N (33–43° N, 109–126° E), contributed
354 perceptibly (5 %) to Arctic surface BC.

355 In July, contributions from anthropogenic sources shrank to those from Yamalo-Nenets
356 Autonomous Okrug and Komi Republic in Russia, and contributed a lower fraction (3 % of
357 Arctic surface BC) (Fig. 6b). Few open biomass burning sources contributed in January (Fig.
358 6c), but contributions from open biomass burning to Arctic surface BC in July can be clearly
359 seen, mainly from the far east of Russia, Canada, and Alaska (Fig. 6d). Open biomass burning
360 emissions from Kazakhstan, southwest Russia, southern Siberia, and northeast China also
361 contributed to Arctic surface BC, although at relatively low strengths (Fig. 5d and Fig. S1d).
362 However, the contributions from open biomass burning could be higher, as the MODIS
363 burned area, the basis of GFED emission inventories, was underestimated for northern
364 Eurasia by 16 % (Zhu et al., 2017). Evangeliou et al. (2016) estimated a relatively high
365 transport efficiency of BC from open biomass burning emissions to the Arctic, which led to a
366 high contribution, i.e., 60 %, from such sources to BC deposition in the Arctic in 2010. A
367 recent study suggested that open fires burned in western Greenland in summer (31 July to
368 21 August 2017) could potentially alter the Arctic air composition and foster glacier melting
369 (Evangeliou et al., 2019). Although the footprint of Arctic surface BC showed a relatively
370 weak sensitivity to areas such as forests and tundra, in the boreal regions, pollutants from
371 boreal wildfires could have greater effects on the Arctic air composition in summer under
372 future warming scenarios (Veira et al., 2016).

373 **3.4 Sources of Arctic BC at high altitudes**

374 Arctic BC levels at high altitudes showed the highest levels in spring (March–April, 40.5–
375 53.9 ng m⁻³), followed by those in late autumn to early winter (November–January, 36.5–
376 40.0 ng m⁻³), and summer (July–August, 33.0–39.0 ng m⁻³) (Fig. 4c). The annual mean Arctic
377 BC at high altitudes was estimated to be 35.2 ng m⁻³, which is ca. 73 % of those at the
378 surface. Such a vertical profile is in accordance with those based on aircraft measurements
379 over the High Canadian Arctic (Schulz et al., 2019). Similarly to the case for the surface,
380 anthropogenic sources dominated by residential sectors, transport, industry and energy
381 (excluding gas flaring), accounted for 94–100 % of Arctic BC at high altitudes in October–
382 May (Figs. 4c, 5c). East Asia accounted for 34–65 % of the total BC in October–May (40 %
383 annually). In comparison, using the Community Atmosphere Model version 5 driven by the
384 NASA Modern Era Retrospective-Analysis for Research and Applications reanalysis data and
385 the IPCC AR5 year 2000 BC emission inventory, H. Wang et al. (2014) found that East Asia
386 accounted for 23% of BC burden in the Arctic for 1995–2005. In summer, open biomass
387 burning in the boreal regions accounted for 40–72 % (57 % on average) of Arctic BC at high
388 altitudes, similar to the source contributions to Arctic surface BC. Specifically, open biomass
389 burning sources from Siberia accounted for 40–42 % of Arctic BC at high altitudes in July–
390 August. Annually, anthropogenic sources and open biomass burning accounted for 83 % (in
391 which residential sources accounted for 34%) and 17 %, respectively, of total Arctic BC at
392 high altitudes (Figs. 4d, 5d).

393 Further investigations of geospatial contributions to Arctic BC at high altitudes in January
394 and July provided more details regarding BC sources. In January, the main anthropogenic BC
395 source in East Asia covered a wide range in China (Fig. 7a). Not only east and northeast
396 China, but also southwest China (Sichuan and Guizhou provinces) were the major
397 anthropogenic sources of Arctic BC at high altitudes. In July, anthropogenic sources made a

398 relatively weak contribution to Arctic BC at high altitudes. The regions that were sources of
399 open biomass burning contributions to Arctic BC at high altitudes were mainly the far east of
400 Siberia, Kazakhstan, central Canada, and Alaska, i.e., similar to the sources of Arctic surface
401 BC. Unlike Arctic surface BC, for which the dominant source regions are at high latitudes in
402 both winter and summer, Arctic BC at high altitudes mainly originates from mid-latitude
403 regions (Figs. 6 and 7). In terms of transport pathways, air masses could be uplifted at low-
404 to-mid latitudes and transported to the Arctic (Stohl, 2006). Further investigations are
405 needed to obtain more details of the transport processes.

406 **3.5 Comparison of Flexpart and GEOS-Chem simulations of BC sources**

407 Data for BC sources simulated with Flexpart were compared with those obtained with
408 GEOS-Chem (Ikeda et al., 2017), which is an Eulerian atmospheric transport model, using the
409 same emission inventories. The simulated seasonal variations in Arctic BC levels and source
410 contributions obtained with Flexpart agreed well with those obtained with GEOS-Chem (Fig.
411 S3). The annual mean BC levels at the Arctic surface obtained by Flexpart and GEOS-Chem
412 simulations were 48 and 70 ng m^{-3} , respectively; the high-altitude values simulated by
413 Flexpart and GEOS-Chem were 35 and 38 ng m^{-3} , respectively. The magnitude difference
414 between the BC levels at the Arctic surface could be related to meteorology. ECMWF ERA-
415 Interim data were used as the input for the Flexpart simulation, whereas the GEOS-Chem
416 simulation was driven by assimilated meteorological data from the Goddard Earth
417 Observation System (GEOS-5).

418 The treatments of the BC removal processes could also lead to different simulation
419 results, depending on the model. In terms of BC loss processes, dry and wet depositions
420 were the removal pathways, depending on the particle size and density, in Flexpart. The
421 treatment of meteorology, especially cloud water and precipitation, would therefore affect

422 the uncertainties of the simulations. In Flexpart version 10.1, BC particles are separately
423 parameterized as ice nuclei for ice clouds, cloud condensation nuclei for liquid-water clouds,
424 and 90 % as cloud condensation nuclei for mixed-phase clouds. The separation of mixed-
425 phase clouds is realistic, as 77 % of in-cloud scavenging processes occurred in the mixed
426 phase over a 90 day period starting from December 2006 (Grythe et al., 2017).

427 In GEOS-Chem simulations, the BC aging was parameterized based on the number
428 concentration of OH radicals (Liu et al., 2011). The BC was assumed to be hydrophilic in
429 liquid clouds ($T \geq 258$ K) and hydrophobic when serving as ice nuclei in ice clouds ($T < 258$ K)
430 (Wang et al., 2011), with modifications because the scavenging rate of hydrophobic BC was
431 reduced to 5 % of water-soluble aerosols for liquid clouds (Bourgeois and Bey, 2011). Such a
432 treatment is expected to improve the simulation accuracy (Ikeda et al., 2017).

433 In Lagrangian models, the trajectories of particles are computed by following the
434 movement of air masses with no numerical diffusion, although some artificial numerical
435 errors could be generated from stochastic differential equations (Ramli and Esler, 2016). As
436 a result, long-range transport processes can be well simulated (Stohl, 2006; Stohl et al.,
437 2013). In comparison, Eulerian chemical transport models such as GEOS-Chem have the
438 advantage of simulating non-linear processes on the global scale, which enables treatment
439 of the BC aging processes (coating with soluble components) (Bey et al, 2001; Eastham et
440 al., 2018). However, with GEOS-Chem, the capture of intercontinental pollution plumes is
441 difficult because of numerical plume dissipation (Rastigejev et al., 2010). Nevertheless, the
442 agreement between the Flexpart and GEO-Chem simulations of BC source contributions
443 indicates improved reliability of evaluated source contributions to Arctic BC.

444 **4 Conclusions**

445 The source contributions to Arctic BC were investigated by using a Flexpart (version 10.1)
446 transport model that incorporated emission inventories. Flexpart-simulated BC data agreed
447 well with observations at Arctic sites, i.e., Barrow, Alert, Zeppelin, and Tiksi. The source
448 regions and source sectors of BC at the surface and high altitudes over a wide region in the
449 Arctic north of 66° N were simulated. BC at the Arctic surface was sensitive to local
450 emissions and those from nearby Nordic countries (>60° N). These results emphasize the
451 role of anthropogenic emissions such as gas flaring and development of the Northern Sea
452 Route in affecting air quality and climate change in the Arctic. Anthropogenic emissions in
453 the northern regions of Russia were the main source (56 %) of Arctic surface BC annually. In
454 contrast, BC in the Arctic at high altitudes was sensitive to mid-latitude emissions (30–60°
455 N). Although they are geospatially far from the Arctic, anthropogenic emissions in East Asia
456 made a notable (40 %) contribution to BC in the Arctic at high altitudes annually. Open
457 biomass burning emissions, which were mainly from Siberia, Alaska, and Canada, were
458 important in summer, contributing 56–85 % of BC at the Arctic surface, and 40–72 % at
459 Arctic high altitudes. Future increases in wildfires as a result of global warming could
460 therefore increase the air pollution level during the Arctic summer. This study clarifies the
461 source regions and sectors of BC in the Arctic. This information is fundamental for
462 understanding and tackling air pollution and climate change in the region.

463

464 *Data Availability.* The data set for simulated footprint and BC source contributions is
465 available on request to the corresponding author.

466

467 *Author contributions.* CZ and YK designed the study. CZ, MT, and IP optimized the Flexpart
468 model. CZ performed Flexpart model simulations, conducted analyses, and wrote the

469 manuscript. KI and HT provided data for GEOS-Chem simulations and site observations. All
470 authors made comments that improved the paper.

471

472 *Competing interests.* The authors declare that they have no conflict of interest.

473

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476

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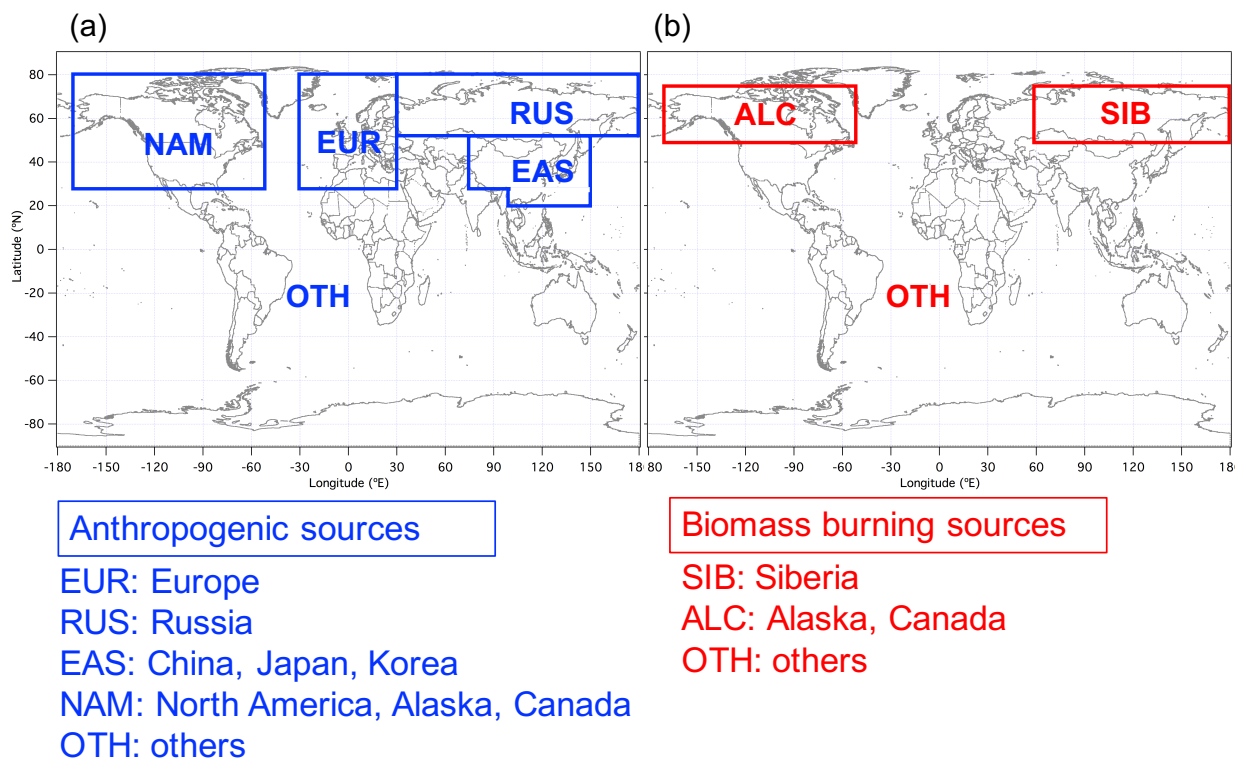
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Table 1. Comparison of BC source contributions in the Arctic surface

Model and versions	Model type	Wet-deposition	Grid resolution	Meteorology	Emissions	Domain/Sites	Year/season	Major source regions/sectors	Reference
Flexpart-WRF 6.2	Lagrangian	Stohl et al. (2005)	unspecified	WRF forecast	ECLIPSE, FINN	continental Norway and Svalbard	spring 2013	Asian anthropogenic	Liu et al. (2015)
Flexpart 6.2	Lagrangian	Stohl et al. (2005)	1° × 1°	ECMWF operational	Unspecified (BC sensitivities were calculated)	Alert, Barrow, Zeppelin	1989-2009	Northern Eurasia	Hirdman et al. (2010)
Flexpart 6.2	Lagrangian	Stohl et al. (2005)	1° × 1°	ECMWF operational	ECLIPSE4(GAINS), GFED3	Arctic (north of 66°N)	2008-2010	Flaring (42%), residential (>20%)	Stohl et al. (2013)
Flexpart 9.2	Lagrangian	Stohl et al. (2005)	1° × 1°	ECMWF operational	ECLIPSE5(GAINS), GFED4.1	Arctic (north of 66.7°N)	2011-2015	Residential and open burning (39%) Flaring (36%), open burning (18%), residential (15%), others (31%)	Winiger et al. (2019)
Flexpart 10.1	Lagrangian	Grythe et al. (2017)	1° × 1°	ECMWF operational	HTAP2, GFED3, Huang et al. (2015) for Russia flaring	Arctic (north of 66°N)	2010	(18%), residential (15%), others (31%)	Current study
GEOS-Chem 9.02	CTM	Wang et al. (2011)	2° × 2.5°	GEOS-5	HTAP2, GFED3, Huang et al. (2015) for Russia flaring	Arctic (north of 66°N)	2007-2011	Russia (62%)	Ikeda et al. (2017)
GEOS-Chem	CTM	Wang et al. (2011)	2° × 2.5°	GEOS-5	Bond et al. (2004), Zhang et al. (2009), GFED3	Alert, Barrow, Zeppelin	April 2008	Asian anthropogenic (35–45%), Siberian biomass burning (46–64%)	Qi et al. (2017)
GEOS-Chem	CTM	Wang et al. (2011)	2° × 2.5°	GEOS-5	Bond et al. (2007), FLAMBE	North America Arctic	April 2008	Open fire (50%)	Wang et al. (2011)

GEOS-Chem10.01	CTM	Wang et al. (2011)	$2^\circ \times 2.5^\circ$	GEOS-5	HTAP2, ECLIPSE5, GFED4	Alert, Barrow, Zeppelin, Arctic (north of 66.5°N)	2009-2011	N-Asian anthropogenic (40–45%) in winter-spring	Xu et al. (2017)
GISS ModelE	GCM	Koch et al. (2006)	$4^\circ \times 5^\circ$	Internal	Bond et al. (2004), Cooke and Wilson (1996)	Arctic (north of $\sim 60^\circ\text{N}$)	Annual general	South Asia	Koch and Hansen (2005)

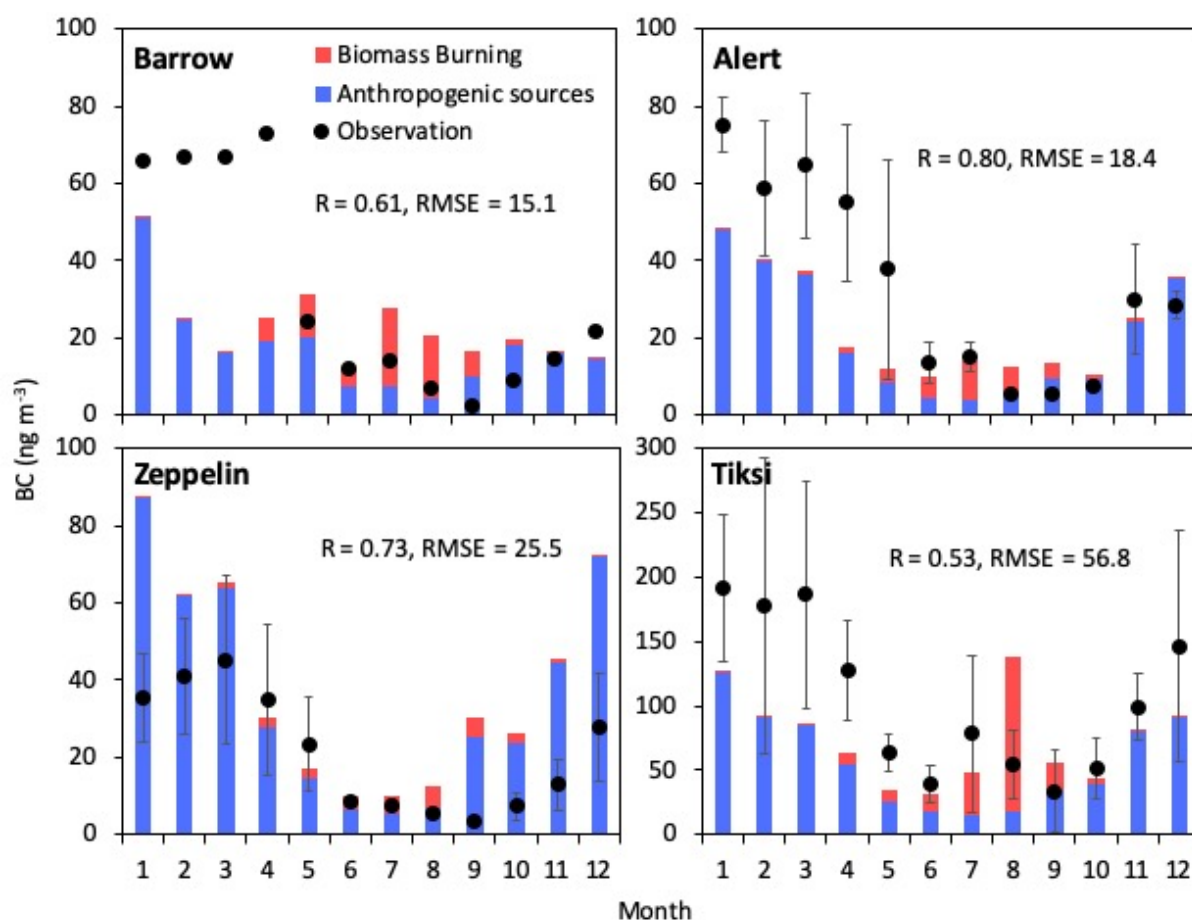


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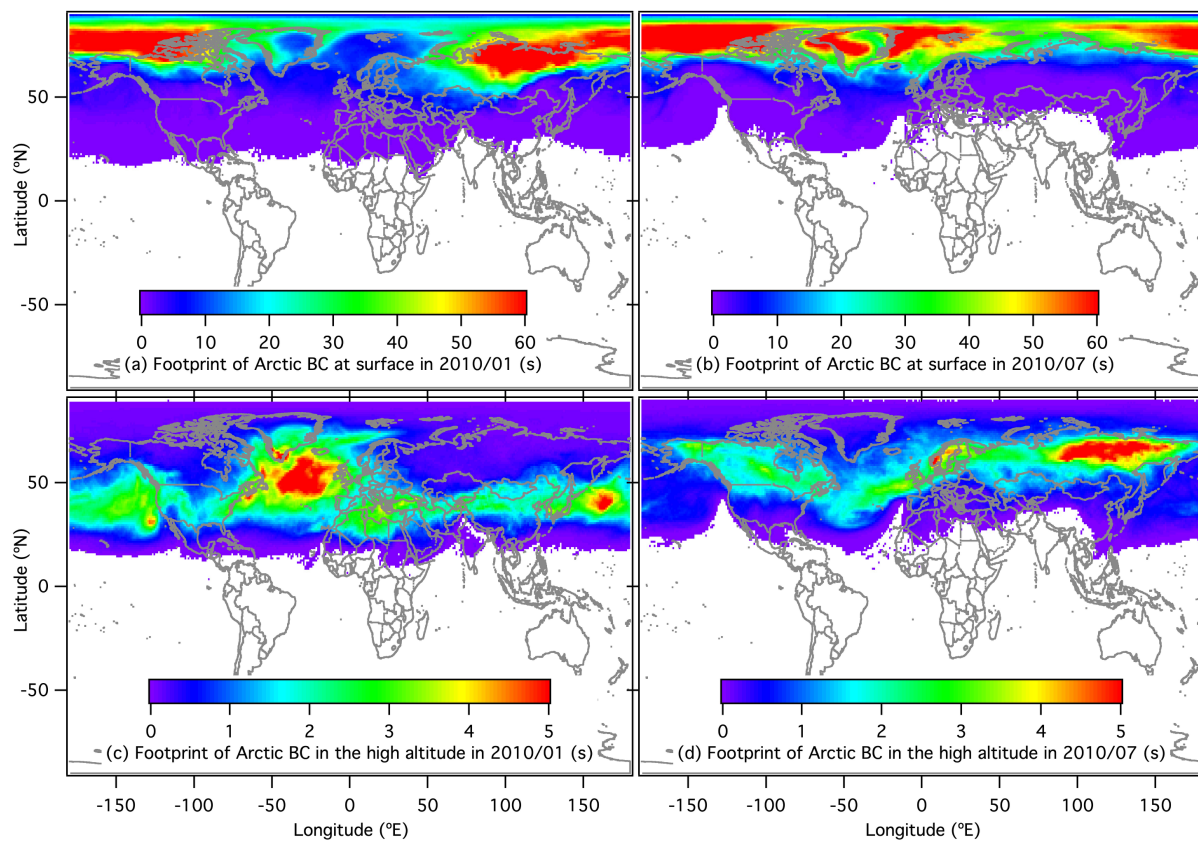
754 Figure 1. Regional separation for quantification of BC in the Arctic from (a) anthropogenic and

755 (b) open biomass burning sources.

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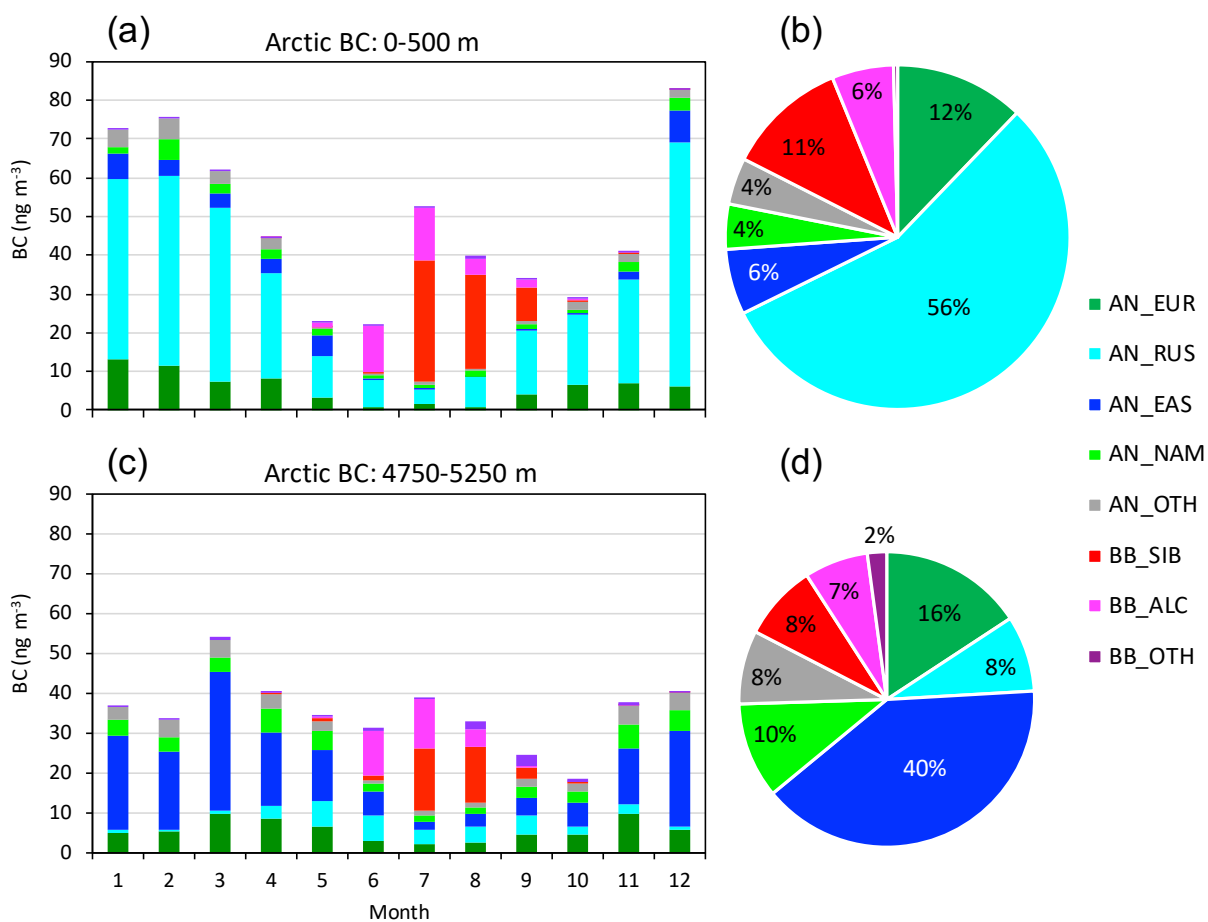
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 758 Figure 2. Observed (filled circles) and modeled (bars) seasonal variations in BC mass
 759 concentrations at Arctic sites. Contributions from anthropogenic sources (blue) and open
 760 biomass burning (red) in each month are shown. Monthly averages of observed (filled circles)
 761 and simulated (bars) BC were conducted for 2007–2011 at Alert, Canada (62.3° W , 82.5° N),
 762 and Zeppelin, Norway (11.9° E , 78.9° N), for 2009 at Barrow, USA (156.6° W , 71.3° N), and for
 763 2010–2014 at Tiksi, Russia (128.9° E , 71.6° N). R and RMSE indicate correlation coefficient and
 764 root-mean-square error (ng m^{-3}), respectively.
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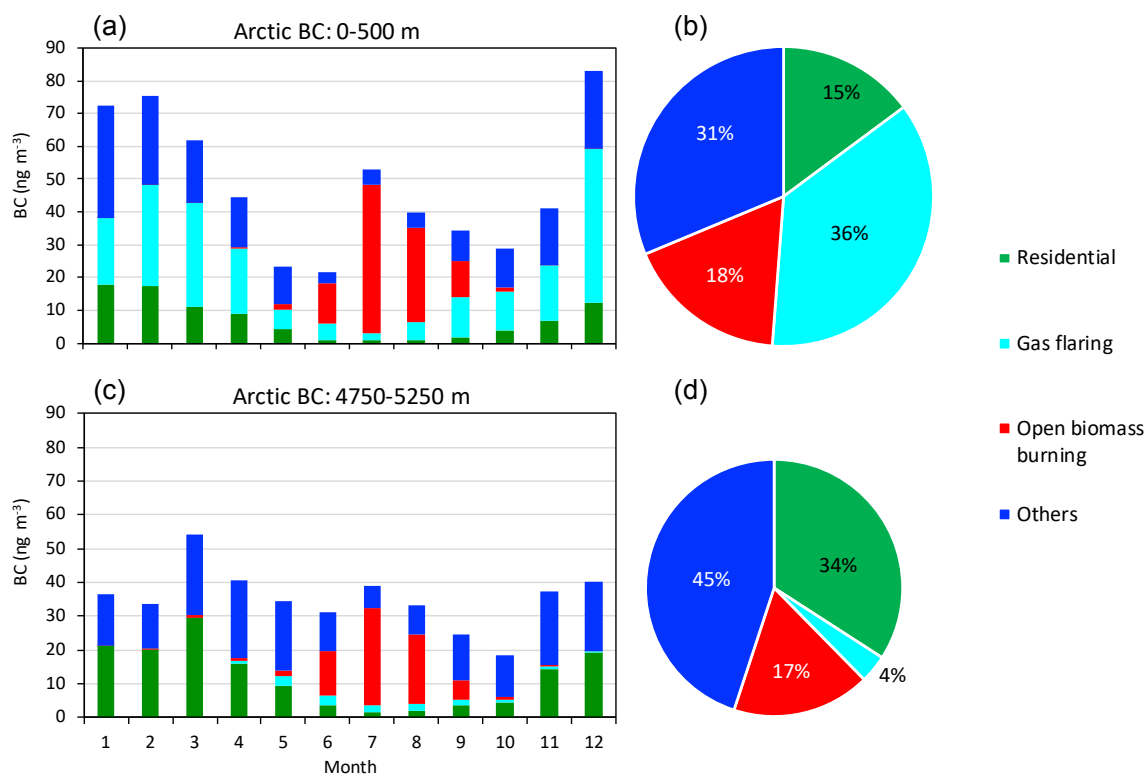
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767 Figure 3. Footprints of Arctic BC shown as retention time(s) of (a) BC at surface (0–500 m) in
 768 January 2010, (b) BC at surface in July 2010, (c) BC at high altitudes (4750–5250 m) in January
 769 2010, and (d) BC at high altitudes in July 2010.

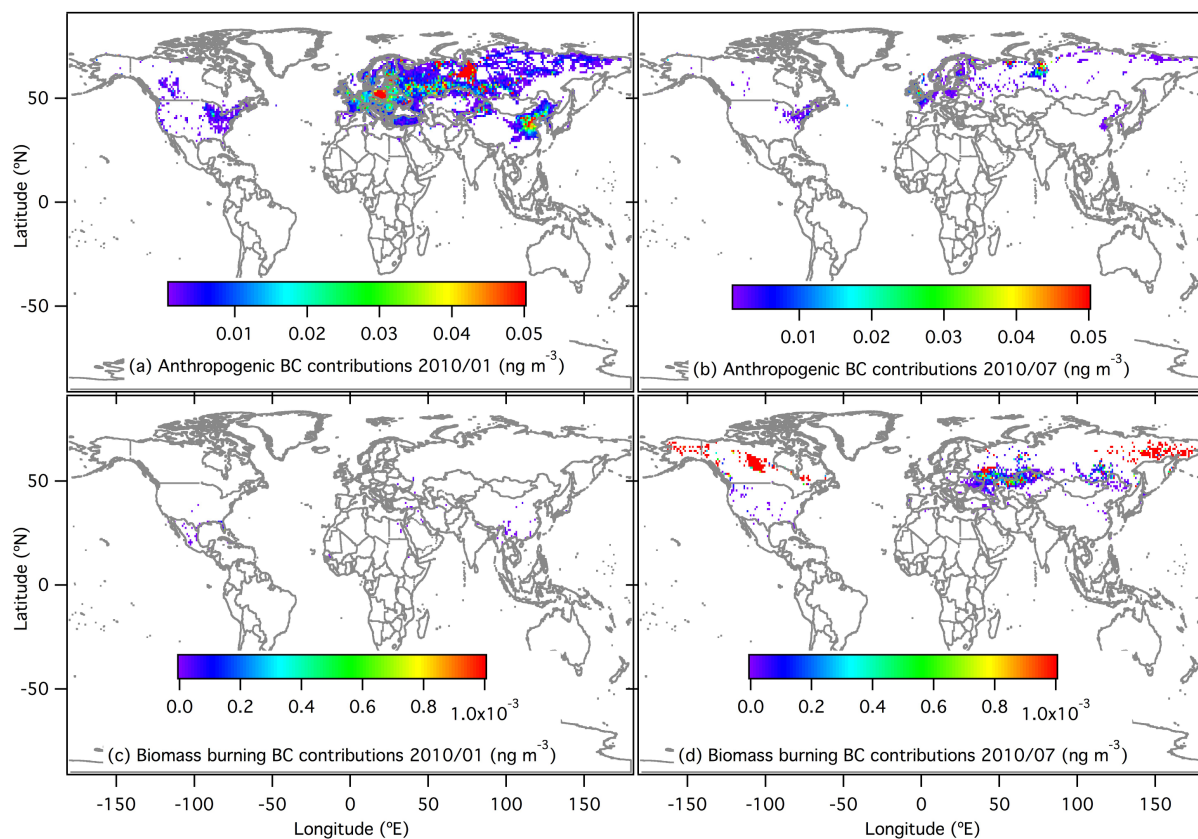
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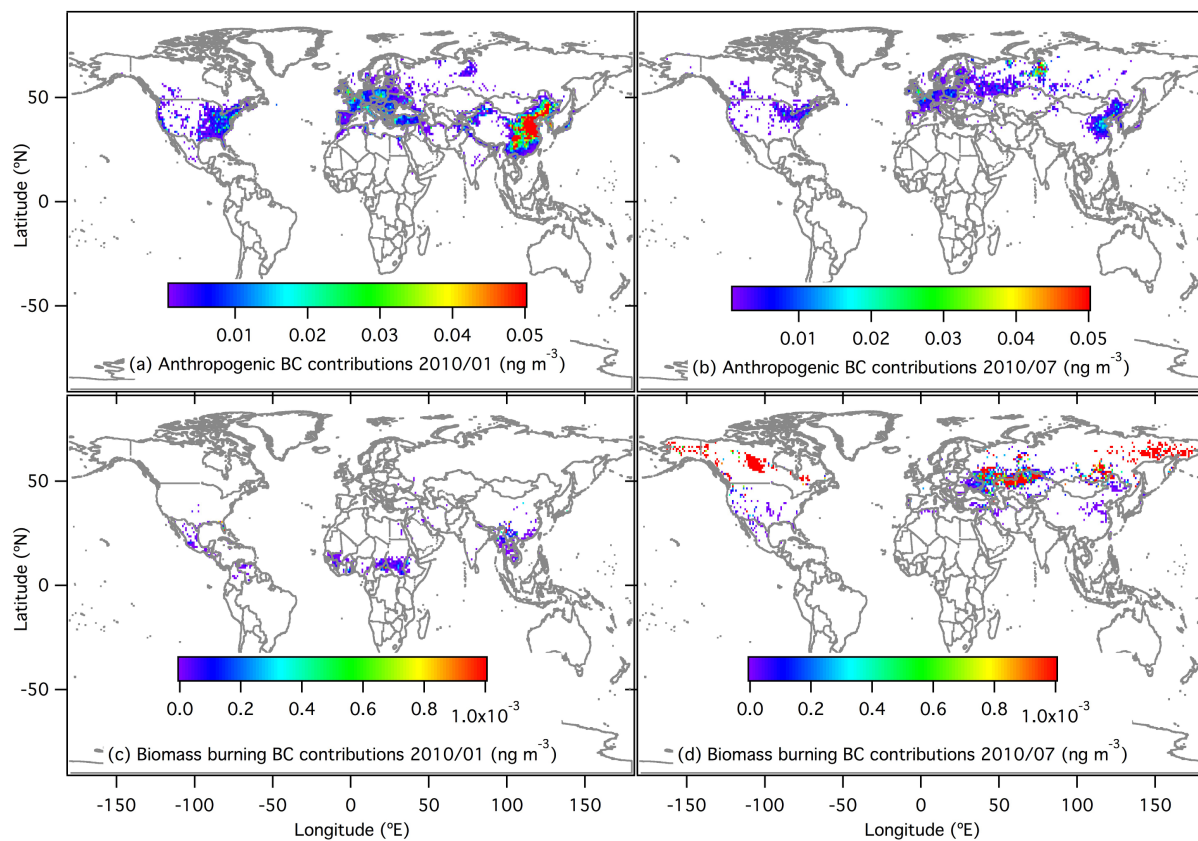
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 772 Figure 4. Contributions of anthropogenic sources (prefixed "AN_" in the legend) and open
 773 biomass burning ("BB_") from each region to (a) seasonal variations in Arctic surface BC, (b)
 774 annual mean Arctic surface BC, (c) seasonal variations in Arctic BC at high altitudes, and (d)
 775 annual mean of Arctic BC at high altitudes.
 776



777
 778 Figure 5. Sectorial contributions from residential combustion (including fossil fuel and biofuel
 779 combustions), gas flaring, open biomass burning and others (energy other than gas flaring,
 780 industry and transport) to (a) seasonal variations in Arctic surface BC, (b) annual mean Arctic
 781 surface BC, (c) seasonal variations in Arctic BC at high altitudes, and (d) annual mean of Arctic
 782 BC at high altitudes.
 783



784
 785 Figure 6. Spatial distributions of contributions to Arctic BC at surface for (a) anthropogenic
 786 contributions in January 2010, (b) anthropogenic contributions in July 2010, (c) open biomass
 787 burning contributions in January 2010, and (d) open biomass burning contributions in July
 788 2010.
 789



790
 791 Figure 7. Spatial distributions of contributions to Arctic BC at high altitudes for (a)
 792 anthropogenic contributions in January 2010, (b) anthropogenic contributions in July 2010, (c)
 793 open biomass burning contributions in January 2010, and (d) open biomass burning
 794 contributions in July 2010.
 795