The urban canopy meteorological forcing and its impact on ozone and PM2.5: role of the vertical turbulent transport

Huszar, P., Karlický, J., Ďoubalová, J., Šindelářová, K., Nováková, T., Belda, M., Halenka, T., Žák, M., and Pišoft, P.

Dear Referee,

We appreciate the detailed and constructive criticism the reviewer raised in his throughout review. We considered all the comments in our revised manuscript and our point-by-point response follows. Moreover, we prepared the revised manuscript with highlighted changes with respect to the original version to facilitate the reviewers/editors decision how to proceed with the review process.

Referee #1's comments:

COMMENTS TO THE AUTHORS: The manuscript "The urban canopy meteorological forcing and its impact on ozone and PM2.5: role of the vertical turbulent transport" has been reviewed in great detail. The reviewer recognizes the great work carried out by the authors and the potential of the manuscript. However, although the manuscript is well-structured it is not very well-written. Statements are often not link to each other and just reported as "dot sentences" but without dots. The terminology used by the authors, mainly when referring to wind engineering and urban physics aspects, is not adequate and also inexact. Moreover, grammar errors (e.g. singular/plural and/or verb tens) have been found throughout the whole manuscript. Figures are clear and consistent to the body text but the font size of the labels is often too small and unreadable. From a scientific and methodological point of view, the manuscript has a great potential and the reviewer believes that it will be an important contribution for the whole scientific community, but the text should be largely improved, and the text made much more fluent both for an expert (e.g. a reviewer) and a non-expert audience. In conclusion, the reviewer's opinion about this manuscript is that it is not suitable to be published in the present form and an extensive work should be done to make the text understandable, unambiguous and scientifically robust for the audience. In order to help the authors in improving the manuscript the reviewer provided them with some comments/suggestions related to the main sections. However, the reviewer suggests to the authors to write again the manuscript and take care to the aspects raised below.
Authors' response: We appreciate that reviewer's recognition of our work and his potential for publication and we admit that it lacks some exactness and un-ambigiousness to communicate its message in a much clearer way towards the scientific community. We went trough the whole text and reformulated many paragraphs to meet the requirements of the reviewer. This concerns especially the use of some scientific terms which were not exact and better wording had to be used, mainly where the topic touches the field of wind engineering.

Abstract: In general this section is very detailed but also too long to be an abstract. Furthermore, although the scope of the manuscript is well highlighted in the first lines, it is not mentioned at all what the knowledge gap in the literature is. Therefore, the reviewer suggests to the authors to add one line about the "knowledge gap" and try to take out from this section some less important details however already included in other sections.
Authors response:
We added further sentences to clearly formulate the "knowledge gap" in the literature, which our manuscript is trying to fill. Further we shortened the abstract a bit ,however, we still would like to include in the abstract the most important aspects of both the design of the model simulations and the results and given their amount, we were limited in reducing the abstract's length.

Another important comment related to this section and to the whole manuscript as well is about the use of the terminology that can be inappropriate and ambiguous for the wind engineering and urban physics community to which this document is also addressed. Definitions as for example "urban meteorological changes", "model experiments", "urban canopy meteorological forcing" are not commonly used. The reviewer's suggestion is to revise the whole document carefully and check in the literature the terminology usually adopted by publications dealing with similar topics.

Authors response:

We revised the use of the mentioned terms including others regarding on the urban canopy. We clearly defined what we meant under the formulation "the urban canopy meteorological forcing" and removed unambiguous terms as "urban meteorological changes". Further we also removed the term "model experiments", and rather replaced it with "model simulations" although in the atmospheric modeling community, "model experiments" is a rather accepted phrase denoting experiments conducted with model simulations when modeling is considered as experimental method.

At lines 3-5: The sentence "From an air quality . . . an important role too" is inexact and unclear. The wind, from the analytical point of view, can be decomposed into the "mean" and the "fluctuating" components which is also the turbulent part. An increased turbulence belong to a decreased mean wind speed. Furthermore, the reviewer does not understand what the authors mean by "wind stilling". Please rephrase and make the sentence clearer for a reader.

Authors response: We reformulated these lines to be more exact highlighting the role wind and turbulence plays in the transport of pollutants in urbanized environment. The term "wind stilling" is commonly used regarding the influence of urbanized areas on the average windspeeds, e.g. Hou et al. (2013), Huszar et al. (2014) and refers to the decrease of windspeed due to increase drag over urbanized landsurface.

At line 6: what do the authors mean with "urban surfaces"? Do you mean urban city/area/environment? If so, please rephrase here otherwise try to explain exactly what you mean. Apply this comment throughout the whole manuscript whenever needed.

Authors response: Urban surfaces here mean urban landsurface type, i.e. surface changed from natural or rural to urbanized one with typical urban features as buildings, streets, parks etc. However, to be more precise, we replaced these terms with the term "urban canopy", eventually the "urban canopy layer" if referring to the air in the urban canopy.

At line 11: please replace "the sensitivity to model grid resolution" with "the sensitivity of the numerical model to the grid resolution".

Authors response: Changed.

At line 11: what do the authors mean with "model experiments are performed"? Experiments can be performed, experimental models can be used, but model experiments is no sense. It would be better to rephrase here to make the sentence unambiguous for a reader.

Authors response: If we accept the term "model experiment" (see above) as an experiment conducted with numerical model than "model experiments are performed" means the simulations performed. To avoid misunderstanding, we nevertheless changed this to "model simulations" are performed/conducted.

At line 12: the definition "urban canopy meteorological forcing" is not wrong but probably may be defined better and in accordance with the literature. It is well-known that obstacles composing cities (as buildings, bridges, trees, etc.) largely affect the wind flow field inside the urban canopy layer

(UCL) and many papers published in the last 30 years confirmed that the logarithmic profile (representative of neutral stability conditions) does not hold anymore in this layer. It means that the "local-scale forcing effects" on the wind flow inside this layer are mainly caused the "obstacles" than inflow conditions (the so-called "large-scale forcing effects") imposed in the numerical model (for example by the nest of cascading models). Please refer to this comment also to modify the title, eventually.

Author's response: it is well known that urban canopies large influence the air flow and the radiative and thermal conditions in the air which 'fills' the corresponding layer (urban canopy layer) but also above. Accordingly, in the revised manuscript we made it clear that the urban canopy meteorological forcing (UCMF) is the forcing that acts on the meteorological variables over urbanized areas. We have chosen the word 'forcing' as the presence of urbanized land-surface forces the air towards modified values of meteorological variables as temperature, windspeed, humidity etc (compared to the values if natural surface would be present).

Therefore we decided to keep this in the manuscript title as well.

At line 17: "near the surface", to which surface are the author referring to? Please be more specific.
Author's response: we refer to the urban canopy layer and clarified this in the revised manuscript.

At line 20: the definition "urban meteorological changes" is not widely used in the wind engineering and urban physics field. The meteorological changes happening at the scale of cities and/or urban districts can be included, according to the size of cities and districts, into the "microscale" and/or "local scale".
Author's response: we changed this to (at different parts of the manuscript) "urban canopy induced meteorological changes" which tells the reader that we refer to the changes of meteorological variables due to the introduction of urban landsurface (i.e. the UCMF has been imposed).

Introduction:
At lines 4-5, p2: The sentence "Urbanization . . . (Folberth et al., 2015)" is not clear, please rephrase and be more specific. The reference can help the reader in better understanding the meaning but the current manuscript (and all sentences) should stand alone.
Author's response: we extended this sentence to be more clear for the reader without the direct requirement to look at the cited literature.

At lines 5-8, p2: the reviewer probably understood the intention of the authors and what they mean by this sentence but only after reading that few times. It would be better to rephrase here and make the sentence a bit more fluent.
Authors' response: we rephrased the sentence to make it easier to understand and more fluent.

At line 9, p2: if the authors write "First of all" at line 5, the reader supposes that a second point follows soon and they would never expect an indentation when explaining the second aspect ("secondly"). Indentation here is not needed at all.
Author's response: We agree and removed the indentation with a slight rephrasing too.

From line 12 to line 20, p2: this piece of text is very confused for the reviewer and it would also be unclear for a reader. Sentences, are not well-written in English and not very well-linked one to each other. Please rephrase and give to the whole piece of text a logical meaning.
Author's response: here we intended to list the possible modifications of meteorological variables (within the UCMF) to help the reader to understand the the well known urban heat island (UHI) is not

the only change that occurs due to rural-to-urban transition but there are other ones too. We slightly rephrased this part of the text to make it more fluent and sentences to link to each other in a better way.

At line 21, p2: the sentence "Meteorological conditions are, thus, strongly perturbed over urbanized areas" states a concept very well-known since decades and assessed by many publications published in literature. You should cite or at least refer to the most important publications dealing with "urban canopy and boundary layers wind flow modeling".
Author's response: we added the reference to the well known BUBBLE experiment (Rotach et al., 2005) which encompassed a detailed investigation of the boundary layer on both city and street-level scale. This paper revealed most of the urban canopy induced changes we expected.

At line 22, p2: what do the authors mean with "urban induced modifications"?
Author's response: We meant and changed the text to "urban canopy induced modifications of meteorological variables"

At line 25, p2: again, please correct the expression "urban meteorological forcing". Moreover, what do the authors mean with "elements"? Please be specific and unambiguous.
Author's response: We changed the "elements" to "components" referring to the components of the UCMF, i.e. the changes in temperature, windspeed, vertical turbulent diffusion etc.

At line 29, p2: please replace "the main contributor" with "the main contribution is given by".
Author's response: Replaced.

Experimental setup: The reviewer does not understand why the authors have titled here "Experimental setup" but they actually described different numerical models adopted to simulate. If this section, as well the next one, refer to numerical simulations there is no sense to talk about "experimental setup". Please remove also the subtitle "Models".
Author's response: We changed this to "Methodology" which more fits the purpose of the section to introduce the scientific methods used to derive the conclusions of the study. We kept the subsection "Models used" as the used numerical models are key to the study and we think they deserve separate subsections.

RegCM4:
At line 15, p5: please replace "Internation . . . " with "International . . .".
Author's response: Replaced.

At line 20, p5: what do the authors mean with "processes"? Please explain or be more specific and unambiguous. At line 20, p5: please as first time use the full name and the acronym in brackets "University of Washington (UW)".
Author's response: We meant the physical processes characteristic for the PBL, i.e. mechanical/thermal turbulence and interaction with the surface. We rephrased the sentence to be more specific. The acronym UW placed after the full name.

At line 23, p5: if more than two please replace "between" with "among".
Author's response: Replaced.

At line 29, p5: what do the authors mean with "landcover processes"? Please explain or be more specific and unambiguous.

Author's response: Here we mean the processes involved in the interaction of the land-surface and the boundary layer which includes the calculation of fluxes of heat, radiation, water and momentum. We clarified this in the revised manuscript.

At lines 30-31, p5: what do the authors mean with "classical canyon representation of urban geometry"?
Author's response: Here we mean the traditional view of urban geometry in the modeling community as an ensemble of canyons of different orientation, width and depth (height of the buildings on the sides of the canyon). The "urban canyon" concept was introduced by Oke (1987). See also Oke et al. (2017, page 19.).

At lines 3-4, p6: the sentence "Within the urban canyon, momentum fluxes are calculated using a roughness lengths and displacement height typical for the canyon environment" is grammatically wrong and scientifically inexact. How do you calculate the momentum fluxes using the roughness length and the displacement height? Moreover, both the roughness length and the displacement are characteristics of rough terrain or surfaces and not necessarily only of "canyons".
Author's response: Indeed, roughness length and displacement height are characteristic of all landsurface types in atmospheric models however here we described the urban canopy parameterization so we gave a reference how the heat and momentum fluxes are calculated in the urban canyon. We rephrased this sentence to be more correct: "Within the urban canyon, heat and momentum fluxes are calculated using the Monin-Obukhov similarity theory with roughness lengths and displacement heights typical for the canyon environment." The full description of the methods is given by the Technical Note on the urban canopy model in Oleson et al.(2010).

At lines 4-5, p6: it is important here to show some governing equations to make clear for a reader the numerical model used.
Author's response: The equations describing the fluxes of anthropogenic heat and air conditioning waste heat are rather complicated (see Oleson et al., 2010) and our manuscript is already lengthy, so we decided not to include any governing equation in the manuscript but placed appropriate reference for the reader if interested in the details.

Experimental setup and data: The title is wrong since the authors in this section are talking about numerical simulations and not experimental tests. Therefore, it is wrong to use refer to an "experimental setup". Moreover, to which "data" are the authors referring to? This is another error since they are not describing "results" or/and "databases" for example, for which commonly the word "data" are adopted. Similar errors have been repeated systematically throughout the whole section. Just some examples have been stressed by the reviewer below. The section should be probably re-written from scratch.
Author's response: As we already pointed out, numerical model simulations with the purpose of test some hypothesis can be regarded as experimental tool or these experiments as 'model experiments'. Consequently the configuration of such experiments, i.e. the 'experimental design' is in fact the model set-up . However, to avoid misapprehension, we changed the title to "Model set-up and simulations" which clearly states what follows (which includes also the model driving data, which are indeed not experimental data). We rewrote parts of the section to reflect this properly.

At line 15, p7: Please change "model experiments" with "experiments".

Author's response: replaced.

Moreover, please add one extra space between "resolution" and "(size . . .". Moreover, if referring to computational grids, as it seems to be, please use a correct and proper terminology and replace "gridboxes" with "computational grids". Please, apply this comment throughout the whole document whenever necessary.
Author's response: Indeed, we refer to the computational grid, but in particular these numbers refer to the number of gridboxes. There are three Cartesian computational grids with gridboxes as indicated by the numbers (in W-E and N-S direction). We rephrased and made the sentence more clear.

At line 16, p7: what do the authors mean with these numbers in brackets? This is not clear at all. Please add an extra space before each bracket.
Author's response: see above.

At line 16, p7: please rephrase the portion of the sentence "Each domain is centered over Prague"; make this concept clearer for a reader.
Author's response: This mean that the center of the domain matches the center of Prague. We rephrased the sentence appropriately.

At line 17, p7: what do the authors mean with "projection parameters"?
Author's response: we mean the parameters of the geographic projection used to project the Earth's spherical surface into a plane. We clarified this in the revised text.

At line 17, p7: the sentence "Accordingly, the three domain is denoted . . ." is grammatically wrong. Please rephrase here.
Author's response: we rephrased the sentence.

At line 18, p7: what do the authors mean with "The regional climate model simulations were performed over 23 vertical levels"? Please provide the reader with more details about the simulations performed to let him understand what you mean with "23 vertical levels".
Author's response: In the revised manuscript, we explained in more detail the vertical layering of the models used, i.e. the depth of the model atmosphere and the depth of individual layers (which is a function of the number of layers).

At line 19, p7: the sentence "For the higher resolution ones" is wrong if the authors are referring to the grid having the "highest resolution". Please rephrase here and make it clearer for a reader. Similar errors have been found also after line 19 but not reported extensively here.
Author's response: Corrected whenever occurred.

Result: The main title and two sub-titles have been used here without any reason. Please use the title to provide the reader with a "general picture" of the corresponding section. In this specific case the subtitle "Model validation" is too vague. You need to say what really has been done in this section.
Author's response: The model validation here means that the results of the models are validated against available observational data (surface measurements and sounding(included in the revision)). "Model validation" is a very common name when referring to such comparison so we think this title is not wrong, however we added a sentence just after it to make it clear what the section will include.

Conversely, the second subtitle "Model Climate" is completely useless. Once again, the terminology is not correctly and properly used and it may cause a large number of misunderstandings. The reviewer

does not understand why the authors switch continuously terminology to describe the same things. What do the authors mean with "observational data"? Do they mean "model climate" or what? And what do they mean with "simulated model results"?

Author's response: We changed the "Model climate" to "Meteorology" as we validate here the meteorological variables (results from the simulations with the regional climate model RegCM). Similarly to the next subsection "Air-quality", where we validated the chemical results (results from the simulations with the chemistry transport model CAMx). Observational data here means data directly observed, i.e. surface measurements or sounding data. The "simulated model results" are the outputs of model simulations which we compare to the measured/observational data. We rewrote this part to make it unambiguous and clear.

Another problem of this section concerns the "description" of results. The authors have provided the reader with a lot of observations without any reasonable explanation for these. The usage of an improper terminology makes this section unclear for a reader.

Author's response: In the description of the results we focused strictly on the description only, without detailed explanation and interpretation of the results. This is done in the Discussion section further. We further reformulated many paragraphs of this section to reflect the reviewers concerns on the improper or unambiguous terminology.

Literature:

Hou, A., Ni, G., Yang, H., and Lei, Z.: Numerical Analysis on the Contribution of Urbanization to Wind Stilling: an Example over the Greater Beijing Metropolitan Area,~J. Appl. Meteorol. Clim., 52, 1105–1115, 2013.

Huszar, P., Halenka, T., Belda, M., Zak, M., Sindelarova, K., and Miksovsky, J.: Regional climate model assessment of the urban land-surface forcing over central Europe, Atmos. Chem. Phys., 14, 12393-12413, doi:10.5194/acp-14-12393-2014, 2014.

Oke, T.: Boundary Layer Climates (2nd edition). Routledge, London and New York., 1987.

Oke, T., Mills, G., Christen, A., and Voogt, J.: Urban Climates, Cambridge University Press, https://doi.org/10.1017/9781139016476, 2017.

Oleson, K.W., Bonan, G.B., Feddema, J., Vertenstein, M., and Kluzek, E.: Technical description of an urban parameterization for the Community Land Model (CLMU). NCAR Technical Note NCAR/TN-480+STR, National Center for Atmospheric Research, Boulder, CO, 169 pp, 2010.

Rotach, M. W., Vogt, R., Bernhofer, C., Batchvarova, E., Christen, A., Clappier, A., Feddersen, B., Gryning, S.-E., Martucci, G., Mayer, H., Mitev, V., Oke, T. R., Parlow, E., Richner, H., Roth, M., Roulet, Y.-A., Ruffieux, D., Salmond, J. A., Schatzmann, M., and Voogt, J. A.: BUBBLE–an urban boundary layer meteorology project, Theor. Appl. Climatol., 81, 231--261, 2005.

The urban canopy meteorological forcing and its impact on ozone and PM2.5: role of the vertical turbulent transport

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Abstract. ~~Urban surfaces due to specific geometry and physical properties bring modified transport of momentum, moisture and heat between them and the air above and perturb the radiative, thermal and mechanical balance resulting in changed meteorological condition (e.g. the UHI - urban heat island phenomenon). From an air quality perspective, many studies argue that~~ It is well known that the urban canopy (UC) layer, i.e. the layer of air corresponding to the assemblage of the buildings, roads, park, trees and other objects typical to cities is characterized by specific meteorological conditions at city scales generally differing from those over rural surroundings. We refer to the forcing that acts on the meteorological variables over urbanized areas as the urban canopy meteorological forcing (UCMF). UCMF has multiple aspects while one of the most ~~important changes is the increased turbulence enhancing vertical~~ studied is the generation of the urban heat island (UHI) as an excess of heat due to increased absorption and trapping of radiation in street canyons. However, enhanced drag plays role too reducing mean wind speeds and increasing vertical eddy mixing of pollutants~~above cities, although increased temperatures and wind stilling play an important role too. Using~~ . As air quality is strongly tied to meteorological conditions, the UCMF leads to modifications of air chemistry and transport of pollutants. Although it has been recognized in the last decade that the enhanced vertical mixing has a dominant role in the impact of the UCMF on air quality, very little is known about the uncertainty of vertical eddy diffusion arising from different representation in numerical models and how this uncertainty propagates to the final species concentrations as well as to the changes due to the UCMF.

To bridge this knowledge gap, we set-up the regional climate model RegCM4 coupled to chemistry transport model CAMx over central Europe ~~we study how urban surfaces affect~~ and designed a series of simulations to study how UC affects the vertical turbulent transport of selected pollutants through modifications of the vertical eddy diffusion coefficient (Kv) ~~. For the period of 2007-2011 and over central Europe numerous model experiments are carried out in order to evaluate the impact of~~ using six different methods for Kv calculation~~on the surface concentrations as well as vertical profiles~~. The mean concentrations of ozone and PM2.5 ~~over selected cities (Prague and Berlin). Three cascading domains are set up at 27km, 9km and 3km resolutions~~in selected city canopies are analyzed. Calculations are performed over cascading domains at 27-9-3 km horizontal resolution, which further enables to analyze the sensitivity ~~to model~~ of the numerical model to grid resolution. A number of model ~~experiments are performed where urban surfaces~~ simulations are carried out where either urban canopies are considered

or replaced by rural ones in order to isolate the ~~urban canopy~~ UC meteorological forcing. Apart from the well pronounced and expected impact on temperature (increases up to $2°$ C) and wind (decreases up to -2 ms^{-1}) there is strong impact on vertical eddy diffusion in all of the six Kv methods. The Kv enhancement ranges from a few 0.5 up to 30 m^{2}s^{-1} at the surface and from 1 to 100 m^{2}s^{-1} at higher levels depending on the methods, while the turbulent kinetic energy (TKE) based methods produce the largest impact.

The range of impact on the vertical eddy diffusion coefficient propagates to a range of ozone (O$_3$) increase of 0.4 to 4 ppbv ~~near the surface~~ in both summer and winter~~, while at higher levels, decreases occur from a few -0.4 ppbv to as much as -2 ppbv~~. In case of PM2.5 ~~, enhanced vertical eddy diffusion leads to decrease of near surface concentrations ranging from almost zero~~ we obtained decreases up to -1 µgm^{-3} in summer and ~~to decreases from -0.5~~ up to -2 µgm^{-3} in winter. Comparing these results to the "total-impact", i.e. to the impact of all ~~considered urban meteorological changes~~meteorological modifications due to UCMF, we can conclude that much of ~~the overall urban meteorological forcing~~ UCMF is explained by ~~acting of~~ the enhanced vertical eddy diffusion, which counterweights the opposing effects of other components of this forcing (temperature, humidity and wind~~impact~~). The results further show that this conclusion holds regardless of the resolution chosen and in both the warm and cold part of the year. ~~Our study demonstrates the dominant role of turbulent transport of pollutants above urban areas and stresses the need for further investigation how variation of urban land-use influence the pollutant transport from the urban canopy.~~

1 Introduction

~~Urbanization has~~ Cities have numerous effects on the environment while the impact on the atmospheric environment is ~~considered as the most~~ probably the most 'far-reaching' as it acts not only locally but also on regional and global scales (Folberth et al., 2015). This impact ~~acts via multiple~~ has many pathways. First of all, cities ~~are~~ represent intense emission hot-spots affecting the air-quality and ~~atmospheric chemistry in general (Im and Kanakidou, 2012; Huszar et al., 2016a) while these modifications occur not only over city scales but propagate to regional and even global scale (??).~~

the atmospheric chemistry over multiple scales (Im and Kanakidou, 2012; Huszar et al., 2016a). Secondly, cities represent distinct surfaces compared to their rural counterparts due to high percentage of artificial coverage with specific geometric layout. These surfaces, comprising the urban canopy (UC), modify the thermal and radiative balance of the overlaying air ~~resulting~~ which results in the well-known and documented urban heat island effect (UHI; Oke, 1982; Oke et al., 2017), when urban temperatures are higher than those over rural surroundings depending on the synoptic conditions (Žák et al., 2019). ~~Urban surfaces further represent~~ UC has however impact on other meteorological variables. It represents enhanced drag on winds which results in decrease of average wind speed (Huszar et al., 2014; Jacobson et al., 2015; Zha et al., 2019). On the other hand, this drag trigger mechanical turbulence ~~contributing~~ contributes to the increase of the planetary boundary layer (PBL) height (Roth, 2000; Flagg and Taylor, 2011) and enhanced vertical mixing (Barnes et al., 2014; Huszar et al., 2018b; Ren et al., 2019). It has been also recognized that increased run-off and suppressed evaporation from urban ~~surfaces~~ land-surfaces reduce the humidity in cities (e.g. Richards, 2004) and the so called urban dry island (UDI) effect can ~~form~~occur, as recently

defined by Hao et al. (2018). Huszar et al. (2014) argued that urbanization contributes to warming of whole regions and largely determines their climate (Květoň and Žák, 2007). Not surprisingly, mitigation strategies of adverse urban climate conditions is an up-to-date research area (e.g. Zhao et al., 2017).

~~Meteorological conditions are , thus,~~ As seen above, meteorological conditions are strongly perturbed over urbanized ar-
5 eas~~. As~~, especially in the boundary layer Rotach et al. (2005) thus the urban canopy represents a significant forcing on the meteorological variables (the urban canopy meteorological forcing, UCMF). Given that air chemistry is closely associated with ~~meteorological variables, the urban induced~~ them, modifications in meteorological conditions due to the rural-to-urban transition (i.e. the UCMF) will result in perturbation of species concentrations, as already shown by many authors ~~regarding~~ dealing with the meteorological changes due to climate-change and their impact on air pollution (Huszar et al., 2011; Juda-
10 Rezler et al., 2012).

The urban meteorological forcing encompasses many ~~elements~~ components and each has a specific impact on air quality, often counterbalancing ~~each other~~ other components. Higher urban temperatures in connection with UHI modify chemical reaction rates and aerosol nucleation as well as they modify dry deposition and wet scavenging rates (Seinfeld and Pandis, 1998). Although, higher temperatures favor ozone formation (Im et al., 2011), in urban areas the situation can be different.
15 Huszar et al. (2018a) showed that due to higher temperatures alone, surface ozone in urban areas is reduced, while the main ~~contributor to this reduction is the~~ contribution is given by increased dry-deposition velocities and increased flux of nitrogen oxides (NO_x) towards nitric acid (HNO_3). Sarrat et al. (2006) concluded too that, especially during nighttime, UHI influences the $NO + O_3 \rightarrow NO_2 + O_2$ reaction and ozone dry deposition reducing its concentrations. Tao et al. (2013) calculated increase of dry-deposition due to rural-to-urban transition as well, probably due to higher temperatures. Regarding aerosols, Huszar et
20 al. (2018b) showed that due to elevated urban temperatures, gas-to-particle partitioning is limited leading to decrease of the secondary inorganic component of PM2.5 (particles of diameter $< 2\ \mu m$).

Another component of the urban meteorological forcing is the changed wind pattern and average speed. Wang et al. (2009), Hidalgo et al. (2010), Ryu et al. (2013a) and Ryu et al. (2013b) modeled the UHI-induced urban-breeze circulation resulting in pollutant transport from and to cities. This depends on the daytime but also on the surrounding terrain and coast (Ganbat
25 et al., 2015; Li et al., 2017) and usually leads to increases of urban ozone concentrations. Urban surfaces have however an opposite role too: higher drag due to the urban architecture induces wind stilling which consequently reduces the dispersion of urban emissions and secondary pollutants into larger scales. According to Jacobson et al. (2015), the total column pollution over a megacity is enhanced due to air-stagnation. Due to wind stilling only, Huszar et al. (2018a, b) modeled large increases of primary pollutants (NO_x, SO_2) and primary components of PM2.5, however O_3 is reduced due to increased titration. de la
30 Paz et al. (2016) found that lower wind over urban areas is the main driver of urban surface induced air-quality changes.

Due to surface heterogeneities typical to urban areas, mechanical turbulence is increased and eddy-transport helps pollutant removal from near the surface towards upper layers of urban PBL (Stutz et al., 2004). Indeed, a very strong link is identified between air pollution and the structure of the urban PBL (Masson et al., 2008). It has to be noted here, that emissions occur on street levels that of course cannot be resolved by regional scale models. However, these models can resolve the turbulent
35 layer over the building level where turbulent mixing in the vertical detrains scalars from streets canyons into the turbulent

boundary layer above the buildings and the conventional atmospheric turbulence becomes dominant (Belcher et al., 2015). Accordingly, many studies adopted regional scale modeling techniques to describe the urbanization impact on species concentrations. Martilli et al. (2003) and Sarrat et al. (2006) focused on Paris and Athens and found significant pollutant decrease, mainly due to enhanced turbulence when urban surfaces are considered. Reduction of primary pollutants (NO_x and CO) due to enhanced vertical mixing due to urbanization is modeled by Struzewska and Kaminski (2012) too. If mitigation measures are implemented to reduce UHI, the consequent reduction of PBL height and vertical turbulent transport causes increase of primary pollutants but decrease of ozone over the surface (Fallmann et al., 2016). Large Chinese agglomerations of Pearl River Delta and Yangtze River Delta (PRD and YRD) have been subject of numerous studies and all argued that urbanization induced increase of vertical turbulent transport favors the reduction of primary pollutants (e.g. NO_x) but this leads to enhancement of e.g. ozone (Wang et al., 2007, 2009). Zhu et al. (2015) arrives to similar conclusions but finds larger ozone changes over higher model levels. ~~Xie et al. (2016a) and Xie et al. (2016b)~~ Xie et al. (2016a, b) argued that the urban induced enhancement of vertical eddy-transport is important especially during summer and arrives to expected conclusions, i.e. decrease of primary and increase of secondary pollutants (ozone). Both Zhong et al. (2017) and Zhong et al. (2018) predict stronger vertical transport too and emphasize its major role, but they also look at the simultaneous effect of urban emissions and their radiative effects and conclude that the decrease of surface concentrations is out-weighed partly by the PBL stabilization due to aerosol radiative cooling. Enhanced vertical eddy-transport due to the introduction of urban surfaces is the main driver of PM10 (particle matter of diameter $< 10\mu m$) as found by Zhu et al. (2017). Liao et al. (2014) applied a range of urban canopy models (UCM) within a mesoscale model and found that those UCMs that produce deeper PBL predict stronger reduction of PM10, underlining the dominant role of urban turbulence. Urban enhanced vertical eddy-diffusion was found to be the primary factor that led to SO_2 decreases over Chinese cities in Chen et al. (2014). Kim et al. (2015) showed for Paris (France) that using urban canopy models that support stronger vertical mixing in the model over urban areas, $PM_{2.5}$ concentrations become lower and fit better to observations. In Ren et al. (2019), the turbulent effects caused by urban expansion reduces the urban canopy pollution under otherwise same weather and emission intensities. Li et al. (2019a) recently analyzed the benefits to use a detailed large-eddy simulation (LES) of urban boundary layer compared to mesoscale model representation of turbulence and found that vertical turbulence is a dominant process that determines the pollutant's removal from urban areas, however LES provides more homogeneous ozone and NO_x profiles that have a better agreement with observational data, as they argued. Li et al. (2019b) using the WRF model analyzed the urbanization induced air-quality changes over California (U.S.) and found that the main driver of ozone and PM2.5 changes are the changes in ventilation, i.e. both wind speed and vertical eddy diffusion seems to play a very important role. Janssen et al. (2017) analyzed the modifications of primary and secondary organic aerosol (POA/SOA) and found that the PBL increase and enhanced turbulence over cities have opposing effects on these two aerosol components: while POA decreases due to increased removal, increase are encountered for SOA due to stronger downward transport from the residual layer (RL). Finally, the role of the intermittent turbulence in removing PM2.5 from near the surface over urbanized areas was recently examined by Wei et al. (2018).

As seen above, the vast majority of the authors argue that the turbulence is a dominant if not the most important factor that determines the overall impact of urban induced meteorological forcing on air-quality. On the other hand, a very few of

them looked at the individual impact of each perturbed meteorological parameters (temperature, wind, turbulence). Recently, in our previous works, Huszar et al. (2018a, b), we modeled the impact of urbanization on meteorological conditions and, consequently, on air-quality using a regional scale, offline coupled climate-chemistry model system. Besides the total impact, we looked at the isolated impact of changed temperature, humidity, wind and vertical eddy diffusion coefficient (Kv) and our

5   results underlined the findings of previous authors. The impact of changed Kv values indeed dominates the overall impact however different part of the day other impacts can counterbalance and become dominant. We found that during night time, the temperature impacts PM2.5 concentrations more (resulting in increase) than the increased turbulence (resulting in decrease). Furthermore, it turned out clear that the total impact of the combined effect of different meteorological parameters is, in fact, a result of counteracting effects of opposite signs but comparable magnitudes. In vast majority of papers listed above, and

10   confirmed by our previous findings, the impact of enhanced vertical eddy-diffusion turned out to be the strongest one. The vertical eddy diffusion coefficient that enters the chemistry transport models coupled online or offline to the driving mesoscale models are usually parameterized or diagnosed from large scale fields as wind, temperature, PBL height or the prognostic turbulent kinetic energy (depending on the PBL scheme used). Question arises here, how the uncertainty that comes from calculating Kv values propagates to the urban impact on species concentration and whether the dominant role of the turbulence

15   impact will hold if using other options for Kv calculation.

Our paper is motivated by the question above and its primary objective is to evaluate, how the regional scale model representation of vertical diffusion (Kv) of scalar variable (e.g. pollutant concentration) affects the impact of urban areas on air-quality via the urban meteorological forcing. In other words, we are interested how a range of methods for Kv calculations translates to a range of Kv values, how this propagates to the impact of urbanization on vertical diffusion and finally, what range of

20   impact on species concentrations this will consequently lead to. Our focus will be ozone and PM2.5 concentrations. Further, the study aims to answer if the dominance of the Kv-impact holds if using Kv calculation methods alternative to the default one (CMAQ; used e.g. in Huszar et al., 2018b). This work is a follow-up study to our previous works, Huszar et al. (2018a) and Huszar et al. (2018b), and extends them by focusing on the vertical eddy-diffusion which appears to be a major factor in the urban-air quality coupling. Moreover, the effect of horizontal resolution, a key factor in regional chemistry-climate modeling,

25   is evaluated here too and our focus is extended to winter months (DJF) as well. A tailored chain of model experiments is implemented to achieve this goal detailed in the next section. The results are then presented in Section 3 which encompasses the ~~model validation~~ validation of the model results, the impact on meteorological parameters as well as the impact on air-quality. Finally, the results are discussed and conclusions are drawn.

2 ~~Experimental setup~~Methodology

2.1 Models used

2.1.1 RegCM4

The regional climate simulations were performed by the model RegCM (version 4.6) (Giorgi et al., 2012) servings as a mete-
5 orological driver for the chemistry transport model simulations (see further). RegCM4 is a non-hydrostatic mesoscale climate
model developed at ~~Internation~~ International Centre for Theoretical Physics (ICTP) based on the MM4 model. It offers multiple
methods for calculating convection; here the Tiedtke scheme was chosen (Tiedtke et al., 1989). Cloud and rain microphysics
is computed using the explicit moisture scheme of Nogherotto et al. (2016) which offers a more comprehensive treatment of
moisture and its transformations in air compared to the older SUBBEX scheme (Pal et al., 2000). For radiative transfer, the
10 NCAR Community Climate Model Version 3 (CCM3; Kiehl et al., 1996) was used.

The planetary boundary layer ~~processes were treated using the UW (~~turbulence was parameterized using the scheme developed
at the University of Washington ~~) method developed~~ by Grenier and Bretheron (2001) and Bretherton et al. (2004) ~~. This scheme
denoted UW. The UW is a local prognostic 1.5-order scheme and~~ provides an alternative to the default non-local diagnostic
Holtslag PBL scheme (HOL; Holtslag et al., 1990) originally included in RegCM. ~~The UW is a local prognostic 1.5-order
scheme.~~ Giorgi et al. (2012) made tests to identify the differences between these two PBL parameterizations and they found
15 excessive vertical turbulent transfer in the HOL scheme for heat and water vapor leading to large biases in winter temperatures
and problems capturing low level stratus. The UW scheme seemed to overcome this shortcoming. Another reason for choosing
this scheme in our simulations was that it provides prognostic turbulent kinetic energy (TKE) values on model output which
enables to use TKE-based estimation of vertical eddy-diffusion coefficient. Moreover, UW itself contains such a method (see
further) and directly supplies Kv values upon model output readily usable in chemistry-transport model calculations.

~~Landcover processes~~ Fluxes of heat, radiation, water and momentum between the land-surface and the atmosphere are
calculated with the Community Land Model version 4.5 (CLM4.5; Lawrence et al., 2011; Oleson et al., 2013) ~~.~~ implemented
inside the driving regional climate model. To account for the specifics of urbanized surfaces, CLM4.5 contains the CLMU
urban canopy module ~~(Oleson et al., 2008)~~ (Oleson et al., 2008, 2010) which considers the classical canyon representation of
25 urban geometry where cities are composed of street canyons. The canyon is bounded by roofs, walls and canyon floor, and
trapping of solar and long-wave radiation within is considered.

Within the urban canyon, heat and momentum fluxes are calculated using ~~a~~ the Monin-Obukhov similarity theory with
roughness lengths and displacement ~~height~~ heights typical for the canyon environment (Oleson et al., 2010). Anthropogenic
heat flux from air conditioning and heating is computed within the urban canopy model from the heat conduction equation based
30 on an interior boundary conditions corresponding to interior temperature of the building. To this heat flux, waste heat from air
heating/conditioning is further added. It is parameterized directly from the amount of energy required to keep the internal
building temperature between a prescribed maximum and minimum values, assuming 50% efficiency of the heating/cooling
systems ~~(?)~~(Oleson et al., 2008).

2.1.2 CAMxv6

The chemical simulations were performed with the chemistry transport model (CTM) CAMx version 6.50 ENVIRON (2018). CAMx is an Eulerian photochemical CTM that implements multiple gas phase chemistry schemes (CB5, CB6, SAPRC07TC). The CB5 scheme (Yarwood et al., 2005) was invoked in this study having optimal complexity for long term climate scale
5 simulations. Particles sizes are considered using static two mode approach. Dry deposition is solved using the Zhang et al. (2003) approach while for wet deposition the Seinfeld and Pandis (1998) method is applied. The ISORROPIA thermodynamic equilibrium model (Nenes and Pandis, 1998) is also activated in our set-up to calculate the composition and phase state of the ammonia-sulfate-nitrate-chloride-sodium-water inorganic aerosol system in equilibrium with gas phase precursors. Secondary organic aerosol (SOA) is computed with the semi-volatile equilibrium scheme SOAP (Strader et al., 1999).

10   CAMx is offline coupled to RegCM using the meteorological preprocessor RegCM2CAMx originally developed by Huszar et al. (2012). For the diagnostic calculations of the vertical eddy diffusion coefficients (Kv), originally, the OB70 (O'Brien, 1970) method was implemented in RegCM2CAMx that uses a simple prescription of the Kv profile. Huszar et al. (2016a) extended RegCM2CAMx by the CMAQ scheme (Byun, 1999) implemented in CMAQ which applies the similarity theory for different stability regimes of the boundary layer. The stability regime in CMAQ method is defined using the dimensionless
15 ratio of the height above the ground and the Monin-Obukhov length. Here, we further extended the range of possible turbulent diffusion packages with a nonlocal turbulent mixing scheme YSU (Hong et al., 2006) which contains an explicit treatment of entrainment processes at the PBL top. We also added the ACM2 method (Pleim, 2007) which is a new version of the original asymmetric convective model (ACM) and includes the nonlocal scheme of ACM combined with an eddy diffusion scheme. ACM2 is thus capable to represent both the supergrid and subgrid components of turbulent transport. The fifth Kv calculation
20 method (MYJ – Mellow-Yamada-Jancic) is based on the TKE approach of Mellor and Yamada (1982) as implemented by Janjic (1994). MYJ implements 1.5-order (level 2.5) turbulence closure mode and determines the eddy diffusion coefficients from prognostic TKE. Finally, the last method of calculating Kv for CAMx is to read them directly from RegCM output (denoted as the DIRECT method). In fact DIRECT and MYJ differ only in the implementation but are based on the same physical principles. In summary, a range of six Kv calculation methods (CMAQ, DIRECT, ACM2, OB70, MYJ and YSU)
25 are available to translate RegCM outputs to CAMx ready Kv fields representing a wide range of values to drive vertical eddy diffusion (see further). It is clear that by this approach the calculation of Kv values are based sometimes on different concept than the calculation of PBL characteristics in the driving meteorological model (e.g, TKE based PBL scheme in RegCM and similarity theory in CMAQ Kv-scheme). However, all Kv methods use only large scale characteristic from the meteorological model as input (like wind, temperature, humidity, TKE profile, PBL height etc.) without any a priory expectation on how these
30 physical quantities have been obtained. In this regard we assume this "non-consistency" a minor issue. Furthermore, Lee et al. (2011) showed too that using "non-consistent" method in calculating Kv for CTMs does not implicate less accurate results than directly coupling the PBL parameters.

    It has to be noted, that the dry deposition scheme used in CAMx does not depend directly on the Kv values provided on CAMx inputs. Instead, for the aerodynamic resistance used for calculation of diffusion trough the first model layer to the

ground is computed using the scheme of Louis (1979) based on the on solar insolation, wind speed, surface roughness, and nearsurface temperature lapse rate. Consequently, different Kv computation methods does not directly impact dry deposition velocities.

Further developments of RegCM2CAMx here include that it takes cloud/rain/snow water directly from RegCM output, which in the version used already enables to output these variables. No feedbacks of the modeled species concentrations on RegCM radiation/microphysical processes were considered. Huszar et al. (2016b), using a similar setup than here showed that urbanization induced chemical changes have a very small radiative feedback in long-term average.

2.2 ~~Experimental~~ Model setup and ~~data~~simulations

Model ~~experiments~~ simulations were performed over three telescoping domains of the following horizontal resolution (and size – as gridboxes): 27 km (189 x 141), 9km (189 x 165) and 3 km (93 x 69). Each computational domain is centered over Prague, Czech republic (50.075° N, 14.44° E) and uses the same ~~projection parameters~~ map projection (Lambert Conic Conformal). ~~Accordingly, the three domain is~~ The three domains are denoted PHA27, PHA09 and PHA03~~. The regional climate model simulations were performed over~~, accordingly. In vertical, the model grid is made of 23 ~~vertical levels~~ layers for the 27 km domain. For the higher resolution ~~ones~~domains (PHA09 and PHA03), this is increased to 41 levels. The lowermost level is about 60-70 m thick while the model top is at 50 hPa for each domain. Within the RegCM runs, the outer 27 km domain was forced by the ERA-interim reanalysis (Simmons et al., 2010). The nested 9 and 3 km domains are forced by the corresponding parent domain using one-way nesting. The 27 km simulations were calculated in hydrostatic mode while the rest, due to higher resolution, required non-hydrostatic approach.

The chemistry-transport model experiments were performed over same domains and use 18 vertical layers. These are identical to the first 18 layers of the PHA27 domain. For the 9 and 3 km runs, the lowermost CAMx layers are identical to RegCM layers too, for higher ones, layer collapsing was applied. The chemical simulation for the 27 km domain were forced by the MOZART-4 global CTM runs forced by NCEP reanalysis (Emmons et al., 2010). The inner domains were one-way nested similarly to the regional climate model runs.

Landuse information was derived from the CORINE CLC 2012 landcover data (https://land.copernicus.eu/pan-european/corine-land-cover) and the USGS database where CORINE was not available. The urban geometry parameters are taken from the $0.05° \times 0.05°$ resolution LandScan dataset which provides average building heights (H), urban canyon height-to-width ratios (H:W), and fraction of pervious surface (e.g., vegetation), roof area, and impervious surfaces (e.g., roads and sidewalks) are provided. Within CLM4.5, urban landuse type is represented as a fraction in percentages of urban intensity categories (HD, MD, LD and TBD). This gives a reasonable description of urban coverage at all resolutions (even at low resolution, small cities are accounted for as low percentage value).

The TNO MACC-III ~~(an update of the previous version II Kuenen et al., 2014, ;)~~ (an update of the previous version II; Kuenen et al., 20 were used as emissions for Europe except Czech republic, where a high resolution national Register of Emissions and Air Pollution Sources (REZZO) dataset issued by the Czech Hydrometeorological Institute (www.chmi.cz) and the ATEM Traffic Emissions dataset provided by ATEM (Studio of ecological models; www.atem.cz) was used. The listed emissions sources

contain annual emissions of the main pollutants, namely NO_x, SO_2, ~CO, volatile organic compounds (VOC), PM2.5 and PM10. MACC-III data are gridded data, while the Czech REZZO and ATEM datasets are defined as area, point and line (for road transportation) shapefiles of irregular shapes corresponding to counties, major sources and roads.

The raw emission data is preprocessed using the Flexible Universal Processor for Modeling Emissions (FUME) emission model (Benešová et al., 2018, ; http://fume-ep.org/). FUME is intended primarily for the preparation of CTM ready emissions files. As such, FUME is responsible for preprocessing the raw input files and the spatial distribution, chemical speciation, and time disaggregation of input emissions. Emission are provided in 11 activity sectors (SNAP – Selected Nomenclature for sources of Air Pollution) and sector specific time-dissaggregation (van der Gon et al., 2011) and speciation factors (Passant, 2002) are applied to spatially interpolated emissions to derive hourly speciated emissions for CAMx. Biogenic emissions of hydrocarbons (BVOC) are calculated using the MEGANv2.1 emissions model (Guenther et al., 2012).

Being an offline couple, first the RegCM model experiments were carried out. After, the meteorology-dependent BVOC emissions were computed by MEGAN while for each gridbox, the fractional cover of different plant functional types and their emission factors determine the actual BVOC emission flux (i.e. for urban gridboxes it can be even zero). In next, the meteorological inputs for CAMx are generated. Finally, BVOC emissions are combined with the anthropogenic emissions calculated by FUME. Having or inputs prepared, a series of CAMx simulations for the 2007-2011 period were conducted for each model domain and these are summarized in Tab. 1. For RegCM, a pair of model experiments were performed denoted "URBAN" and "NOURBAN" where urban landsurface was considered (and modeled with the CLMU model within RegCM) or replaced by rural surface most typical for the surroundings of the particular city. This meant crops in most of the times. The urban effect were calculated thus using the "annihilation method" (Baklanov et al., 2016), which has been often employed for urban studies but also for transport related impact assessment (Huszar et al., 2013, e.g.).

Using RegCM meteorology, numerous of CAMx runs were carried out depending on which urban meteorological effects are considered and which Kv calculation method is employed. The "NOURBAN" reference CAMx run is driven by RegCM meteorology that does not consider any urban meteorological forcing: i.e. no temperature, humidity, wind and turbulence effects, which means it is driven by the NOURBAN RegCM runs. These CAMx experiments implement the CMAQ method for Kv calculation which is the default option and was used also in Huszar et al. (2018a, b). Within the impact of the simulated meteorological changes on chemistry, the following effects were taken into account 1) modified temperature (t-impact); 2) modified absolute humidity (q-impact); 3) modified wind speed (uv-impact) and 4) modified vertical eddy diffusion coefficient; kv-impact). In a further set of simulations, denoted "URB_t+q+uv+kv", all the listed effects were considered and finally, in the "URB_t+q+uv" simulations, the kv–impact was removed. In addition, the "URB_t+q+uv+kv" and the "URB_t+q+uv" simulations were repeated with all listed Kv calculation methods. In this way, the total urban impact can be evaluated as the difference between the corresponding URB_t+q+uv+kv and NOURBAN model experiments. However, the main focus of the paper is the kv-impact (as URB_t+q+uv+kv minus URB_t+q+uv), which is now possible to evaluate by six different representations of vertical eddy diffusion. Each listed simulation is repeated for all model domain which allows to assess the sensitivity of results on horizontal model resolution too. In case of the representation of meteorological conditions in AQ

modeling, this can be relatively large (Tie et al., 2010). Finally, it has to be noted that for chemical simulations, the landuse was kept the same for all model experiments in order to isolate the effect of meteorological changes on air quality.

3 Results

3.1 Model validation

3.1.1 ~~Model climate~~

~~The simulated model results were validated against observational data . For meteorology~~Here we provide a basic comparison of the most important modeled quantities to measured data (both the meteorology and air-quality).

3.1.1 Meteorology

For meteorological variables, the gridded E-OBS van der Besselaar et al. (2011) data~~were chosen which enables spatial comparison~~, that enable spatial comparison, were chosen to reflect the model values. The modeled average summer (JJA) and winter (DJF) near surface temperatures and precipitation are evaluated for all model ~~resolution~~resolutions. Note that from the 27 km model domain result are shown only over a subdomain corresponding roughly to the 9 km domain ~~for easier~~ area to facilitate comparison. In Fig. 1, the difference between the ~~model~~ modeled and measured near surface temperature ~~and observation~~ is presented. During summer, the 27 and 9 km resolution ~~runs~~ simulations tend to underestimate surface temperatures by 2° C, while the 9 km ~~run~~ one have smaller biases and even some overestimation occurs over the eastern part of the domain up to 1° C. Largest differences are encountered over mountainous areas where the main reason lies probably in the poor model representation of complex orography. The 3 km ~~run~~ simulation shows a warm bias almost everywhere up to 1-2° C (except a few areas near the Czech border)~~up to 1-2 C~~. For the winter months, ~~all three resolutions shows~~ a warm model bias ~~, which is~~ is seen at each resolutions, being lowest in the 27 km ~~run~~ one (up to 2° C) and reaches 3° C ~~in the~~ at 3 km ~~model experiment~~resolution. Notably, the warm bias is largest over Prague (indicated by its borders) which ~~suggest that the model~~ suggests that the regional climate model used overestimates the urban temperature effects.

The difference between the model total precipitation and observation is shown in Fig. 2 in mmday^{-1}. During winter, precipitation is overestimated in RegCM at each resolution reaching up to 3-4 mmday^{-1} above mountains. The medium resolution model ~~experiment~~ simulation shows somewhat smaller bias while at 3 km and around Prague, the model overestimated rain by 2-3 mmday^{-1}. A different picture is seen during JJA: the low resolution ~~run~~ simulation show a strong overestimation of the precipitation by up to 2-3 mmday^{-1} all over the domain. A much smaller positive model bias is encountered for ~~the~~ 9 km with values usually lower than 1 mmday^{-1} (with even some negative bias over the domain edges). ~~Regarding the 3 km domain~~At the highest resolution, both positive and negative biases are present in the range of -3 to 2 mmday^{-1}. For the area of Prague, the bias is however small, lying between -0.5 to 0.5 mmday^{-1}.

3.1.2 Air quality

The modeled surface concentrations were confronted with the European Environment Agency Airbase (https://www.eea.europa.eu/data-and-maps/data/aqereporting-8) ~~observational~~ station data. Rural and urban background stations were selected which are not affected by local sources unresolved by the models (i.e. traffic stations were not considered). The validation focuses on ~~the~~ two key pollutants, O_3 and PM2.5. Note that under the term 'modeled surface values' we mean the uniform concentration of the lowermost model layer at the corresponding grid-box, which corresponds roughly to the urban canopy layer.

In Fig. 3 the scatter plots of ~~the comparison of~~ measured and modeled ozone values are shown. It is seen that the vast majority of the values lies within a factor of 2 (FAC2) except in winter ~~season~~ when observations exhibit very low values that are not resolved by the model, regardless of the relatively high resolution (3km). In summer, model values are usually overestimated and the overestimation is slightly higher ~~in~~ at 27 and 9 km ~~domain~~resolutions. It is also evident that the model provides a narrower range of values than ~~the~~ measurements, especially during summer ~~months~~ when it is unable to capture low ozone situations (large number of observation near zero, while model values are around 50 to 150 μgm^{-3}).

The deviations seen in the scatter plots are better understood looking ~~and~~ at the comparison of ~~average seasonal diurnal cycles. This is shown~~ seasonal average diurnal cycles in Fig. 4. For summer~~months~~, the average daily maximum ozone is reasonably captured with a little positive bias around 3-5 μgm^{-3} ~~over the~~ at 27 and 9 km ~~domains~~ resolutions while urban stations have this bias slightly larger. The daily maximum ozone values are almost perfectly captured for ~~the~~ 3 km resolution with a small overestimation for ~~the~~ urban stations. The timing ~~the of maximum~~ of the of the maxima is reasonably captured too. However, the model tends to strongly overestimate nighttime values ~~of ozone for all resolution~~ for all resolutions and especially for urban stations. This explains the overall overestimation of average daily ~~ozone~~ values seen in the scatter ~~plot~~plots. For winter~~months~~, model biases are smaller. For rural stations, model underestimates measured values for the 27 km ~~run~~ resolution and slightly overestimates for the higher ~~resolution domains~~ones. Here, however, the daily maximum ozone is well modeled. The observed urban values are overestimated by the model, especially during morning hours by up to 10 μgm^{-3}, however, again, maximum daily values are reasonably captured. The comparison of monthly means in Fig. 5 confirms the overestimation of ozone values during JJA (by 10-20 μgm^{-3}), where the main contributor are the too high ozone values during night as seen from previous figure. On the other hand, winter (and spring) averages are modeled with higher accuracy and even some underestimation by model occurs for the 27 km ~~domain~~ resolution (seen also in the diurnal cycles). Again, the positive model bias is somewhat larger for urban stations than background ones.

The daily PM2.5 scatter plots in Fig. 6 and the corresponding normalized mean bias (NMB) values suggest that PM2.5 is underestimated in the model, except for summer over rural stations ~~in the~~ at 27 km ~~model experiment~~resolution. On the other hand, a very good agreement in terms of this metrics is achieved for the 3 km ~~run~~ resolution during JJA. It is seen that the underestimation is caused mostly by the high-end of the observed values which CAMx is not able to ~~resolve. The ratio of modeled value within the ratio of 2 (~~reproduce. The FAC2 ~~)~~is about 0.6-0.7 at each resolution and season and~~surprisingly ,~~ surprisingly, it is usually higher for urban stations. The gain in using higher resolution is not clear and the model improvement

depends which metrics is analyzed. For rural stations however, the 3 km simulations seem to be more accurate than the lower resolution ones.

The monthly values in Fig. 7 bring some light to the root of model biases seen in scatter plots. Winter concentrations are underestimated by the model in each case by about 5-10 μgm^{-3} over rural stations and 10-15 μgm^{-3} over urban ones. During JJA and for rural stations, model concentrations of PM2.5 tend to be higher than the measured ones by about 3-4 μgm^{-3} for the 27 km resolution, somewhat smaller for the 9 km one (about 2 μgm^{-3}) and are in a very good agreement for the fine resolution simulations. Over urban stations, model is consistently lower in PM2.5 concentrations – here, however, the highest underestimation for JJA occurs at 3 km resolution (around 3-5 μgm^{-3}).

It has to be noted, that many urban background stations selected in the analysis are taken from small urban areas their emission is not resolved by model well. Therefore the average urban model concentrations are only slightly different from rural ones (higher values for PM2.5 in urban areas and lower ones for O_3 due to titration effects).

Finally, to validate the model's ability to resolve the vertical transport and the sensitivity to the choice of Kv method, we contrasted the modeled pollutant profiles with available ozone sounding data for Prague, Czech republic measured for January to April at 12:00 UTC (see http://portal.chmi.cz/files/portal/docs/meteo/oa/sondaz_ozon.html). Fig. 8 shows a systematic underestimation of observed O_3 values occurring for elevations (above-sea-level) higher than about 1000 m reaching 10 ppbv or even exceeding. it is also clear that the coarse resolution experiment tends to agree with the sounding data best, at least for the two winter months. On the other hand, the lowest ozone values are systematically modeled at the highest model resolution (3 km). Near the surface, ozone is usually overstimated by different simulations while the 3 km resolution shows the best match. From the shape of the modeled profiles, it is also clear that the TKE based MYJ method results in the most straight curve (highest values near the surface, lowest at high elevations) meaning that it produces the strongest mixing for ozone (see further in Sec. 3.2.3). On the other hand the OB70 and YSU methods result in the skewest ozone profiles at lower elevations (up to 1000 m) with best agreement with observed values. In summary, it is difficult to conclude with resolution or Kv methods result in the best model-observation agreement. It seems that for higher elevations, coarse resolutions are more accurate while at lower ones the high resolution simulations match observations better. Further, within the PBL Kv methods producing lower Kv values (YSU and OB70) lead to ozone vertical pattern which most resemble the observations.

3.2 Impact on meteorology

In our recent papers, Huszar et al. (2018a, b), we showed that urban canopies largely influence the local and regional summer values of temperature, humidity, wind speed and the vertical eddy diffusion coefficient (which determines the vertical turbulent transport). Here, we extend our analysis to winter months as well as to the sensitivity on the chosen horizontal grid resolution. It is widely known, that during winter months at stable stratification, buoyant turbulence is suppressed and the mechanical one dominates. Our analysis extended for winter thus brings new insight on how the urban canopy meteorological forcing acts on air-quality under substantially different weather conditions compared to the hot season. Further, an important parameter to meso-scale modeling of urban meteorological effects is the model resolution which determines how well the urban landuse heterogeneity is represented along with

the (meso-)synoptic weather features that strongly influence the urban canopy meteorological ~~effects~~ phenomenon (e.g. UHI; Žák et al., 2019). The urbanization induced meteorological effects (i.e. the UCMF) will be evaluated as the difference between RegCM experiments URBAN and NOURBAN~~(applying the annihilation method)~~. In spatial figures, ~~we will show results~~ results will be shown for central European region that covers two large cities we focus on, Berlin and Prague. These two cities

5 will be in focus in the diurnal cycle figures too (expect for results on the 3 km ~~domain which cover~~ resolution which covers only Prague).

3.2.1 Temperature

In Fig. 9 the spatial impact of urban ~~surfaces~~ canopy on near surface temperature is shown. In general, the DJF impact on temperature is ~~, in average,~~ higher for both cities and exceeds 2 °C. In summer, the impact lies between 1.5 and 2 °C. Over

10 Berlin, the 9 km simulation results in a more pronounced impact in both seasons. Over Prague, the impact is highest for the 27 km simulation in both seasons. It is also seen that if spatially averaged over coarse resolution, the ~~3 km~~ high resolution impact over Prague will be lower than over the 9 km and, especially, over the 27 km simulation. Further the figure shows that most of the area analyzed exhibits statistically significant temperature impact suggesting that even minor urbanized areas (small cities, villages) play role in modulating ~~temperature~~near surface temperatures.

15 The diurnal cycle of the ~~absolute urban~~ urban canopy absolute temperatures and the ~~urban impact is show~~ difference compared to the non-urban case is shown in Fig. 10. It is seen, that higher resolutions exhibit warmer urban canopy temperatures in correspondence with the conclusions made in the validation. According to the ~~expectation~~expectations, maximum urban warming occurs during evening around 8-10 pm UTC (10-12 pm local time) for JJA. For DJF, the maximum difference occurs at different times over each city: for Berlin, it is around midnight UTC (1 am local time) however, over Prague it occurs

20 earlier around 4-8 pm UTC (5-9 pm local time). Over Berlin, the 9 km simulation results in almost 2 times higher impact (1 ° C vs. 2 ° C). Over Prague, the situation differs: the highest impact, in contrary to the absolute values, occurs for the 27 km simulation for ~~winter~~DJF, however, for ~~summer~~ JJA the results for the three resolutions are very close to each other and show maximum warming around 2.4 ° C.

3.2.2 Wind speed

25 In Fig.11 the ~~summer and winter~~ JJA and DJF average impact on 10 m wind speed is shown for the three resolutions. It is seen that the wind is significantly decreased over ~~urban~~ urbanized areas while the decrease is higher in DJF (around -1.5 to -2 ms^{-1} ~~) than during JJA (~~compared to -1 ms^{-1}). The ~~smalled~~ smallest decrease is modeled for the 27 km simulation for both cities and seasons. Statistically significant changes on the 98% level are modeled over large part of the analyzed ~~regions~~region suggesting that even small urban areas contribute to the wind ~~stilling~~reduction significantly.

30 Regarding the diurnal cycle of the ~~wind impact,~~ wind speed and its urban canopy induced changes in Fig. 12, it is seen ~~in terms of absolute values,~~ that the two analyzed cities behave somewhat differently: while over Berlin, the 27 km ~~resolution run produce~~simulation produces lower winds than the 9 km ~~ones~~one, over Prague the lowest wind speeds are modeled for the highest resolution~~(and the opposite for the 27 km run)~~. A more unique picture ~~,~~ (in accordance with the spatial figures) is seen

for the impact itself. The higher the model resolution the stronger is the impact on winds. For Berlin, it reaches -1.5 ms^{-1} for both seasons during noon. For Prague, the wind impact reaches -1.4 ms^{-1} for ~~summer~~ JJA and can be as strong as -1.8 ms^{-1} ~~during winter~~for DJF.

3.2.3   Vertical eddy diffusivities

The main focus of this paper is the urban impact on the vertical eddy diffusion coefficient as a key factor determining urban pollution transport. Here we present this impact for each of the Kv ~~methods~~ method that are listed in Section. 2.1.2. In Fig. 13 the urbanization induced changes of the JJA eddy diffusion coefficient at the first model level (approximately 65m, i.e. ~~above the at~~ urban canopy layer height) are shown for all three model resolutions (rows) and six Kv-methods (columns). It is clear from each ~~methods~~method/resolution that Kv values are affected the most over the two selected large cities (Berlin and Prague), however statistically significant changes occur over rural areas too. The most striking feature is the wide range of Kv changes (predominantly increases). The smallest ~~urban increase isin generally~~ urbanization induced increase is, in general, obtained for the CMAQ and YSU schemes: in both cases the rural-to-urban transition results in about 1-2 m^{2}s^{-1} increase of Kv over both ~~Prague and Berlin~~cities. The strongest increase is ~~achieved~~ modeled with the TKE based DIRECT and MYJ methods where it reaches 15 m^{2}s^{-1} ~~for DIRECT~~ and 30 m^{2}s^{-1}~~for the MYJ methods~~, respectively. The ACM2 and OB70 ~~method~~methods lie in the middle range with increases up to 6-10 m^{2}s^{-1}. Regarding the sensitivity of the resolution, ~~the~~its effect seems to be small and the three ~~gridsizes~~resolutions result in comparable change of Kv values while often the 27 km resolution produces the strongest impact, especially when spatially averaging of the higher resolution results to 27 km. In summary, the urbanization induced Kv changes above the urban canopy layer during JJA encompass a relatively wide range from 1 to 30 m^{2}s^{-1}.

The DJF impact on Kv at the canopy layer height is shown in Fig. 14 and very similar patterns are seen when compared ~~summer season~~to JJA, both qualitatively and quantitatively. The strongest impact is obtained for the two TKE based ~~methods:~~ method (DIRECT and MYJ) reaching 15 to 30 m^{2}s^{-1}. On the other hand, a one order smaller impact is calculated for the CMAQ and YSU ~~methods~~method (up to 2 m^{2}s^{-1}). Similarly to the JJA impact, the ACM2 and OB70 methods lie in the middle range of the Kv-methods with the former one somewhat stronger. During both seasons, a few areas encounter statistically significant Kv decrease in the DIRECT and MYJ methods. This might be connected to the general wind ~~stilling~~reduction over large areas and the corresponding reduced TKE generation resulting in lower Kv values, but this would require more analysis.

In order to understand the Kv evolution during the day we plotted its diurnal cycle in Fig. 15 for Berlin. We are interested here not only on the urban canopy values, but on the impact on the whole Kv profile (within the PBL) too and the absolute values from the URBAN model experiments are plotted ~~too as contour layer~~as well as contour lines. Regarding the absolute values, it is seen that largest Kv are generated during early afternoon hours, in line with the expectations. The level of maximum Kv is higher in JJA (about 150 - 600 m) than during DJF (100-500) and also higher for the 9 km simulation compared to 27 km one. The TKE based methods ~~,~~(DIRECT and MYJ) produce higher values, reaching 200 m^{2}s^{-1} for the DIRECT method and 120 m^{2}s^{-1} for MYJ in JJA. The lowest Kv values are calculated ~~by~~using the OB70 and YSU, reaching about 20 m^{2}s^{-1} in ~~summer~~JJA. Winter Kv values are much lower, as expected. It is also seen that Kv values are usually larger for the 9 km simulation in both seasons. Here, the MYJ and DIRECT methods are exception with sightly higher values for the 27 km

resolution. Somewhat distinct diurnal distribution is obtained with the ACM2 method. Kv remain relatively large throughout the whole day and the maximum value is reached at much higher levels than for other methods, around 500–800 m. Turning our attention to the urban induced Kv changes (shaded colors), it is seen, again, that the highest impact is obtained using the DIRECT and MYJ methods, reaching 100 $\mathrm{m^2 s^{-1}}$. It is also clear that the impact is higher during ~~summer~~ JJA and stronger

5 for the 9 km ~~run~~ resolution for all methods. The highest impact occurs at comparable levels than the absolute values and consistently around late afternoon to early evening hours for each method. This means that the maximum of the impact is shifted by 3 to 6 hours later than the occurrence of the maximum absolute values. The smallest impact is simulated for the YSU and OB70 methods, reaching 10-20 $\mathrm{m^2 s^{-1}}$ at ~~its maximum during~~ their maximum occuring at afternoon/early evening hours.

10 In Fig. 16, the absolute Kv values and the urban canopy impact is shown for Prague for JJA in the same manner as for Berlin, only extended by the 3 km simulation. In general, both the absolute Kv values as well as the impact is very similar to the impact over Berlin. The absolute eddy diffusion coefficient increases with increasing resolution and is, again, highest for the DIRECT a MYJ methods reaching 350 and 150 $\mathrm{m^2 s^{-1}}$, respectively. The lowest Kv values are obtained when calculated by the YSU and OB70 methods (up to 20 $\mathrm{m^2 s^{-1}}$)). It is also clear, that at higher ~~resolution~~resolutions, the maximum Kv

15 occurs at higher model levels. Regarding the impact, there is a very clear increase when going into higher resolutions and the change is especially large between the 27km and 9 km resolutions.

During ~~winter~~DJF, Fig. 17, absolute Kv values are, of course, smaller compared to summer ones, however one cannot conclude clearly an increase when turning to higher resolutions. E.g. for the DIRECT and MYJ methods, the 27 km results are higher than those obtained for the 9km and 3 km simulations. However, the supremacy of these two methods still holds

20 producing vertical diffusivities up to 40–50 $\mathrm{m^2 s^{-1}}$ (MYJ being higher). Regarding the urban canopy impact, the DIRECT and MYJ methods result in strongest change, however, in contrary to Berlin (or to the ~~summer Prague~~ Prague JJA results), the strongest impact is modeled ~~over the~~ at 27 km ~~domain~~ resolution while for other Kv methods, the difference between individual resolution is not significant.

3.3 Impact on the air-quality

25 The chemical changes, in particular the changes in the concentrations of O_3 and PM2.5 ~~due to~~ as a result of the simulated urban ~~meteorological effects~~ canopy meteorological forcing (seen abova) are presented here. This includes the effects of modified temperature, wind and vertical diffusivity~~changes~~, which where analyzed in the previous section. The effect of the urban induced modifications of humidity is included too, however, we showed in Huszar et al. (2018b) that the impact is negligible~~and the main contributors to the overall change are the modifications of the three meteorological parameters listed above~~. In accor-

30 dance with Huszar et al. (2018b), we will distinguish different impacts based upon which urban meteorological ~~perturbations are considered,~~ perturbation is considered: e.g. the "t+q+uv-impact" means the combined impact of temperature-, humidity- and wind changes; the "kv-impact" stands for the chemical changes triggered by ~~perturbed~~ modified vertical eddy diffusion values only and the "t+q+uv+kv-impact" means the impact of all considered urban ~~meteorological changed~~ canopy meteorological

changes which, in this paper, will be equivalent with the "total-impact". We will start with the impact of enhanced turbulent transport which is the main focus of this paper.

3.3.1 The effect of perturbed diffusivities

In Figures 13–17 we presented the range of Kv ~~values perturbed~~ perturbation caused by the introduction of urban ~~surfaces~~ land-surfaces and it was seen that ~~this change~~ it covers two orders of magnitude (increases from a few $\mathrm{m^2 s^{-1}}$ to tens of $\mathrm{m^2 s^{-1}}$). Here, our attention moves to what range of perturbations of O_3 and PM2.5 urban canopy concentrations this leads (i.e. the kv-impact is evaluated for individual Kv methods).

Ozone

Fig. 18 presents the ~~summer perturbation of surface concentrations of ozone~~ JJA changes of O_3 due to the ~~urban~~ urbanization induced Kv enhancement for the three resolutions and six Kv methods. Ozone is increased in all cases ranging from 0.2 ~~ppbv~~ to 3 ppbv, as expected ~~according to~~ following Huszar et al. (2018a). They showed that the main contributor to this increase is the reduced destruction due to the turbulence enhanced vertical removal of ~~NOx from the surface~~ NOx from the canopy layer and the increased turbulent flux from the RL during night. The smallest increase is modeled by the CMAQ and YSU methods. At 27 km resolution, the largest effect is obtained using the DIRECT method, at 9 km the four remaining methods ~~gives~~ give rather comparable impact for both cities. At 3 km resolution, the largest impact is seen for the ACM2 and OB70 methods. For Berlin, higher resolution leads to stronger impact for each method. For Prague, it is most often the 27 km resolution where the highest impact is modeled (expect YSU). In general, the impact over Berlin is stronger than over Prague.

The ~~winter~~ DJF impact in Fig.19 is stronger than the ~~summer~~ JJA one often reaching 4–5 ppbv. Here again, the smallest effect is obtained using the CMAQ and YSU methods. The strongest one is seen for the DIRECT, MYJ and ACM2 ~~for~~ methods at each resolution. It is also clear that higher resolution usually leads to smaller modeled impact.

To gain a more detailed insight into the range of kv-impacts, we also plotted the diurnal cycle of the vertical profile above both analyzed cities along with the absolute values. In Fig. 20, we present the results for Berlin for both seasons. Regarding the absolute values in JJA, ~~values~~ at higher levels they are higher and this elevated maximum (usually around 200-500 m) "reaches" the surface as the usual early afternoon summer ozone maxima occur (about 40-50 ppbv). These are somewhat higher ~~in the~~ at 9 km resolution. Turning our attention to the impact, a very clear maximum is seen near the surface during early evening reaching 6 ppbv while the effect is stronger for the 9 km resolution. This maximum is not visible only for the OB70 method at ~~the~~ 27 km resolution. Another striking feature is the ozone decrease up to -2 ppbv at higher levels (200-400 m) with maximum intensity coinciding with the maximum surface increase and slightly shifted compared to maxima of the absolute values. In general the impact (similarly to the absolute values) is more pronounced for the 9 km resolution and reaches higher levels. A secondary ozone maximum in the daily cycle is detectable during noon to evening at altitudes around 500-1500 m (higher altitudes at 9 km resolution) reaching 0.3-0.4 ppbv. During ~~winter~~ DJF, the pattern of absolute values is much simpler with gradually increasing concentrations with increasing altitude ~~with~~ including the expected weak afternoon maximum. The 9 km resolution profiles have usually lower values compared to 27 km ones, except near the surface when the ACM2 method

provides higher values. The kv-impact on ~~ozone~~ O$_3$ is characterized with a clear increase at the surface model layer with a weak maximum during late afternoon up to 6 ppbv while the strongest effect is modeled for the DIRECT and ACM2 methods. The diurnal amplitude of the impact is much smaller than during JJA and ~~is~~ it is clearly stronger for the 9 km resolution than for the 27 km one. The ~~higher level ozone decrease~~ O$_3$ decrease at higher levels is well seen (around 300-400 m; up to -2 ppbv

5   reduction) and it is much stronger for the 9 km resolution.

The pattern of the kv-impact on O$_3$ as well as that of the absolute values is very similar for Prague compared to Berlin, as seen in Fig. 21. The ~~ozone~~ O$_3$ increase is limited to the lowermost layers and there is again a clear early evening maximum reaching 5-6 ppbv. ~~Also the~~ The impact reaches almost zero values during midday near the surface as seen for Berlin too. The decrease at higher levels (200-400 m) with maximum during late afternoon/early evening is ~~visible~~ evident from model reults

10 too and reaches -2 to -4 ppbv. Finally, the secondary maximum of ~~ozone~~ O$_3$ increase during noon hours at around 500-1500 m ~~is visible too reaching~~ occurs as well and reaches 1 ppbv. The impact for the 3 km resolution seems to be the smallest, probably in connection with the fact that the absolute values are the smallest at this resolution~~too~~.

During ~~winter~~ DJF over Prague, Fig. 22, both the absolute values and the kv-impact resembles the pattern seen for Berlin. The diurnal amplitude is much smaller for the impact with maxima usually around early evening and occasionally a weaker

15 maximum is seen during morning hours for the DIRECT, ACM2, OB70 and MYJ methods. Again, the smallest impact is modeled for the CMAQ and YSU methods and the values for the 3 km resolution are below those for lower resolutions. The ozone decrease at higher altitudes (around 200-500 m) is detectable too, reaching -2 ppbv.

In summary, different Kv-methods lead to not only different average kv-impact on ozone values, but substantially different vertical profiles and different shape of the daily cycle including the timing of the maximum value and occurrence of secondary

20 maxima at higher altitudes.

PM2.5

Our attention now turns to the UCMF induced PM2.5 changes and we first look at the changes in near surface concentrations due to the kv-impact. The results for JJA and DJF are presented in Fig. 23 and Fig. 24, respectively. For ~~summer~~ JJA

25 and for Prague, there is a clear decrease of ~~PM2.5~~ PM2.5 in line with the expectation of higher vertical turbulent removal (dispersion) of aerosol as well as their precursors from the urban canopy layer (Huszar et al., 2018b). In general, the 27 km resolution ~~produces~~ leads to stronger decrease up to -2 µgm^{-3}, especially for the DIRECT, ACM2 and OB70 methods. For higher resolutions, these methods ~~lead to~~ result in the largest impacts too (-1 to -2 µgm^{-3}) while very weak impact is modeled using the CMAQ and YSU methods from -0.2 to -0.4 µgm^{-3}. A slightly different picture is obtained for the kv-impact for

30 Berlin. While at 9 km resolution, the conclusions are very similar to the ones seen for Prague (decreases of PM2.5 of similar magnitude, ACM2 and OB70 providing the strongest impact), at 27 km resolution the impact is either a very small decrease (DIRECT and ACM2, up to -0.2 µgm^{-3}) or a slight increase up to 0.3 µgm^{-3} for the ~~rest of the~~ remaining methods. It is seen that there is a general increase of PM2.5 over large part of the analyzed region and this increase dominates the impact over Berlin. This ~~increase~~ is probably caused by the PM2.5 removed from the urban atmosphere (due to higher Kv values) and

35 transported ~~to other regions where it causes~~ and deposited at other regions causing an opposite effect.

During ~~winter~~DJF, the kv-impact leads to clear increase of PM2.5 concentrations for all resolutions, cities and methods. The strongest impact is seen for the DIRECT, ACM2 and MYJ methods peaking at -2 μgm^{-3} while, again, the smallest impact is ~~seen~~ obtained for the CMAQ and YSU methods (below -1 μgm^{-3}). It is also clear, that the impact calculated for lower resolutions is usually slightly stronger, especially over Prague.

5   In order to see, how the kv-impact on PM2.5 evolves during the day, we plotted ~~on~~ in Fig. 25 the diurnal cycle of the impact of ~~urban~~ urbanization induced Kv enhancement on the PM2.5 vertical profiles along with the absolute values. At the surface, the absolute values are higher during ~~winter~~ DJF (18 μgm^{-3}) ~~than during summer~~ compared to JJA (up to 15 μgm^{-3}) as expected due to the more stagnant meteorological conditions. However, at higher elevations ~~summer~~ JJA concentrations are higher due to enhanced vertical transport from the canopy layer. It is also clear that for the 9 km resolution, PM2.5 decreases

10 with height slower than in the 27 km run, which is in line with the slightly stronger vertical mixing in the 9 km resolution compared to 27 km one (see Fig. 15). Further it is seen that the Kv methods giving stronger vertical turbulent diffusivities (e.g. DIRECT) result in lower near surface concentrations and vice versa (e.g. OB70), especially for ~~summer~~JJA. The diurnal cycle of the absolute values are characterized with clear maximum during morning hours and a minimum (especially in JJA) during afternoon. Looking at the kv-impact on PM2.5 concentrations, it is seen that the near surface values are characterized

15 with decreases except the 27 km ~~resolution~~ simulation values during JJA (in line with Fig. 23). At 9 km resolution during JJA, this decrease encompasses two peaks: a primary peak during afternoon reaching -1 μgm^{-3} and a secondary ~~peak~~ one during morning (up to -0.6 μgm^{-3}). The strongest impact is provided by DIRECT, ACM2 and OB70 methods as seen in the spatial figure above. At higher altitudes, the PM2.5 removed from the ~~surface levels~~ urban canopy layer causes a positive impact "region" (at about 100–300 m) up to 0.3–0.4 μgm^{-3} ~~and occurring~~ occuring during afternoon. At 27 km resolution however,

20 this elevated maximum reaches the ground and leads to almost complete disappearance of surface decreases. During DJF, the two resolutions ~~qualitatively behave~~ behave qualitatively in a very similar way. Both lead ~~two~~ to a well pronounced surface PM2.5 decrease while it is stronger in the 9 km resolution (up to -2 to -3 μgm^{-3}). The DIRECT and ACM2 methods provide the largest impact. The double peak shape of the near surface impact is less clear or missing, instead, the kv-impact remain high from morning to evening hours~~(i.e. during daytime)~~. The elevated increase of PM2.5 is more pronounced during the cold

25 season and is, similarly to JJA, stronger over ~~the~~ at 9 km ~~domain~~resolution, reaching 0.5–0.6 μgm^{-3} at about 200–500 m.

     For Prague, the JJA absolute values in Fig. 26 look quantitatively and qualitatively very similar to those over Berlin with peak values around 10-15 μgm^{-3} during morning hours while higher values are modeled with Kv ~~methods~~methods producing lower Kv values (e.g. OB70). The kv-impact manifests again as two maxima of the near surface PM2.5 decrease reaching -2 to -3 μgm^{-3}. The strongest impact is seen for the DIRECT, ACM2 and OB70 methods as seen for Berlin too. The elevated

30 positive impact seen for Berlin is present here too and reaches 0.5–0.6 μgm^{-3} with two maxima, one during morning hours and one during evening hours. In general, the impact over the 9 and 3 km ~~resolution~~resolutions (up to -3 μgm^{-3}) is stronger than over ~~the coarse~~ at 27 km ~~domain~~ (up to -1.5 μgm^{-3}).

     Fig. 27 presents the absolute kv-values and the kv-impact for DJF for Prague~~, it~~. It is clear that the 9 and 3 km resolutions result in higher near surface concentrations and the vertical spread of the PM2.5 is stronger than for ~~the~~ at 27 km~~domain~~.

35 Regarding the impact, it reaches higher values in the 9 km and 3 km resolution and the maximum kv-impact is reached for

DIRECT, ACM2 and OB70 methods (up to -3 μgm^{-3} decrease) during daytime. For the 27 km resolution and for the YSU and CMAQ methods, the PM2.5 decrease is smaller (up to about -1 μgm^{-3}). Similar to Berlin, the PM2.5 increase and higher levels is evident too and is stronger for the 9 and 3 km resolutions. It occurs between 150–400 m and reaches 0.8-1 μgm^{-3}, especially for the DIRECT, ACM2 and OB70 methods.

3.3.2 The total urban impact

One of the most important questions to answer in this paper concerns the dominance of enhanced turbulence within the urban canopy impact on air-quality via the urban meteorological effects. To answer this question, we evaluated also the total- or the t+q+uv+kv-impact for both analyzed ~~pollutant~~ pollutants as the difference between the URB_t+q+uv+kv and NOURBAN ~~model experiments~~ simulations for each resolution. As the NOURBAN model experiment is calculated using only the CMAQ Kv-method, the total-impact is given also only for this method. Here, based on short 1 month test simulations, we ~~checked~~ verified that the t-, q- and uv-impacts depend on the choice of the Kv calculation only weakly (in contrast with the kv-impact itself which is strongly dependent on the choice of the Kv scheme).

Fig. 28 presents the total-impact of the urban meteorological changes on the mean ~~surface~~ canopy concentrations of O_3 for each resolution and both seasons. Huszar et al. (2018a) predicted that near surface ~~ozone~~ O_3 should increase due to the increased removal (~~trubulent~~ turbulent dispersion) of ~~NOx~~ NOx from urban areas and thus reduced titration as well as due to enhanced turbulent transport from higher levels, although they pointed out that the overall effect is always a result of competitive impact of multiple influences. Indeed, in our results ozone usually increases too but the magnitude of the increase changes substantially across resolutions and it is different for the two cities. For Berlin, it reaches around 0.2–0.3 ppbv and 0.4–0.6 ppbv for the 27 and 9 km simulation, respectively. For Prague, increases are encountered for the 9 and 3 km resolution, reaching 0.3 and 0.8 ppbv, respectively. However, for the 27 km simulation, the overall urban impact turns to be negative (about -0.2 ppbv). During DJF, ozone increases over both cities, mostly for the 27 km resolution (reaching 0.8 ppbv). However, for Prague ~~,~~ in the 3 km resolution ~~,~~ O_3 decreases (up to -0.6 ppbv), in contrast with ~~JJA~~ the JJA results. In order to see, whether the simulated total-impact is uniform during the day or it behaves qualitatively and quantitatively differently during different hours, we plotted further the diurnal cycle of this impact along with the absolute concentrations for both cities and seasons, as shown in Fig. 29. For JJA the absolute values (solid lines) are in line with the expectation ~~with~~ of maximum during afternoon while the difference between the different resolutions is not greater than 6-8 ppbv (greatest differences in the daily maxima and nighttime values). The impact (dashed lines) is characterized with a clear main maximum during afternoon hours reaching ~~up to~~ 3 ppbv for Berlin, while it is less pronounced in the coarse resolution simulation. A secondary maximum is visible too during morning hours, especially for Prague. The minimum of the impact is encountered during early afternoon and here, the differences between the resolutions are very large for Prague ranging from -2 ~~to 0 ppbv~~ ppbv to almost zero, explaining the overall negative impact of the urban canopy meteorological forcing on ozone in the previous figure. During ~~winter~~DJF, absolute ozone concentrations encounter two maxima, one during early morning and one during early afternoon. In ~~winter~~DJF, ozone titration is the dominant process in cities while ozone is transported here by turbulence from upper levels. The titration rate is highest when ~~NOx~~ NOx emissions are peaking and this occurs in morning and ~~later~~ late afternoon, putting the two ozone

maxima in-between. In general, the 9 and 3 km ~~resolution results~~ resolutions result in higher absolute ozone for this season, especially during nighttime. The total-impact is, again, characterized with a clear late-afternoon maximum reaching 2-3 ppbv and being strongest in the 27 km simulation. It is present also in the 3 km simulation for Prague, however, a negative peak forms here during morning (reaching -1.5 ppbv), which results in the overall negative impact on ozone seen in Fig.28.

The total-impact on PM2.5 surface concentrations is presented in Fig. 30. The impact is negative in all cases confirming the expectations that the effect of turbulent removal dominates (Huszar et al., 2018b)~~, however~~. However the magnitude of the change varies greatly between the cities and resolutions. It reaches -1.5 to -2 μgm^{-3} for Prague in both ~~summer and winter~~analyzed seasons, and is strongest over ~~the~~ at 27 km ~~domain~~resolution, while at 3 km, the decrease is only about -0.6 μgm^{-3}. Over Berlin, the decrease is in the range of -0.4 to -1 μgm^{-3}; in ~~summer~~JJA, it is slightly stronger ~~over~~ in the 9 km ~~run and in winter~~simulation while in DJF, the opposite holds.

The diurnal cycle of the absolute near surface PM2.5 concentrations as well as the total-impact over Berlin and Prague is presented in Fig. 31. A clear maximum occurs in the absolute values in ~~summer~~ JJA during morning hours over both cities in line with the emission's temporal evolution, while the difference between individual resolutions is within 5 μgm^{-3} and is highest during nighttime when the 27 km ~~run~~ simulation produces the largest concentrations. Regarding the total-impact, in ~~summer~~ JJA there is a clear negative peak during evening hours for both cities reaching -2 μgm^{-3}. However, ~~in the~~at 3 km ~~run~~ resolution this peak is less pronounced and reaches only about -0.6 μgm^{-3}. Here, another peak is present during morning hours and this is seen also ~~over the~~ at 27 and 9 km ~~domains~~resolutions, especially for Berlin. During ~~winter~~DJF, absolute concentrations exhibit a main maximum during morning hours while ~~especially for Prague,~~ a secondary peak is present during evening hours, probably in connections with the diurnal cycle of the emissions. The diurnal cycle of the impact is very similar between individual resolutions and cities with one main peak during morning hours when the impact is the smallest (almost disappears over Berlin ~~in the~~ at 9 km ~~run~~resolution). Later, the impact increases, ~~depending on the resolution,~~ during noon to afternoon hours, reaching -2 μgm^{-3}, ~~similar to the summer~~ similarly to the JJA impact. Again, the smallest impact is modeled ~~over the~~ at 9 km resolution (about -1 μgm^{-3}).

In summary, when comparing the individual resolutions, it is clear that the total-impact for both ~~ozone~~ O$_3$ and PM2.5 can vary largely, in both quantitative and qualitative sense. Further it is evident, that the kv-impact (evaluated in the previous section) will add further uncertainty to the results as its spread is even larger than the spread seen in the total-impact due to different resolutions (bear in mind that the kv-impact represents a major component of the total impact). ~~Remember, that~~ Further, the total-impact was evaluated using the CMAQ method. This resulted in one of the smallest kv-impacts. Even as such, it clearly dominates the total-impact. Consequently, using other Kv methods would lead to even stronger total-impact (increase of ozone and decrease of PM2.5) and thus confirms the dominance of the ~~changes~~ modifications in vertical turbulent transport among other components of the urban canopy meteorological forcing.

4   Discussion and conclusions

~~Results suggest~~ The validation of model results showed a clear overestimation of winter temperatures in RegCM along with a precipitation overestimation. Previously, Giorgi et al. (2012) reported this positive bias too using the same PBL scheme (UW scheme). Along with the positively biased winter temperatures this suggest that the heat removal from the surface towards higher levels is probably underestimated (in contrast with the older HOL scheme) and this leads to higher surface temperatures.

5 Large positive rain bias was showed also in Huszar et al. (2016b) and they attributed it to overestimation of cloudiness which in winter, can lead to positive temperature bias ~~(due to~~ reduced radiative cooling ~~)~~of the surface. In summer, the results are more mixed but the model stays within a reasonable range compared to observation.

The comparison of modeled pollutant concentrations with observations showed several model deficiencies. For ozone, it is especially the strong overestimation of the nighttime values (while daytime peaks are reasonably captured), ~~this~~ which in turn
10 leads to underestimation of the diurnal amplitude~~significantly~~. This was commonly encountered in regional climate chemistry studies (Zanis et al., 2011; Huszar et al., 2016a; Karlický et al., 2017). E.g. Zanis et al. (2011) found that night-time ozone values are captured in the model (they used CAMx too) with less accuracy than daytime ones and argue that this is caused by more stable conditions when ozone values are more sensitive to the vertical profiles of meteorological variables, emissions and the precursor concentrations. It is clear that in future, more emphasis should be given on processes determining nighttime
15 surface ozone, as improving the nighttime model accuracy will yield more reasonable starting point for the development of daytime ozone (Wong and Stutz, 2010) and an overall better model performance. ~~Consequently, the monthly ozone averages are overestimated .~~ The positive nighttime ozone bias results in overestimated monthly ozone.

Higher model ozone may suggest that some of the urban impacts on ~~ozone~~ this gas may be slightly overestimated, assuming that the impact increases with increasing absolute values. It is also evident from the results that higher resolutions did not bring
20 substantial improvement to model accuracy (except that the daily ozone maxima are captured with a higher accuracy in the 3 km model runs). The same conclusion as recently stated by Falasca and Curci (2018), who applied WRF over Italy at similar (cascading) domain resolutions. It also evident from the absolute values of concentrations in the ozone diurnal cycle figures, that Kv methods produce higher diffusivities (compared to the default one) result in higher ozone values (e.g. the TKE based methods), which is due to higher NOx turbulent removal in cities and therefore lower ozone titration. This means that the
25 positive ozone bias will be even higher. This however does not mean that these methods are erroneous as the model bias has in general multiple reasons (inaccurate diurnal profile of emissions, biased nighttime chemistry etc.) rather than caused by one particular process (vertical eddy diffusion).

Regarding the PM2.5, the most important model shortcoming is the strong underestimation of observed values during winter, especially over urban stations. On the other hand, overestimation occurs in summer, especially for the 27 km and 9 km domains.
30 Many stations from western Europe were included in the validation and these showed systematic model overestimation also in Huszar et al. (2018b) and overweight the underestimation seen for other central European countries, especially Czech republic, which encompasses the whole 3 km domain. Note that the emission data used in this study is basically a new version of that used in the former study and differs only slightly. The REZZO/ATEM emissions used for Czech republic are also quantitatively very similar to the TNO data over the this country too. Huszar et al. (2016a), using similar model resolution and emission data,
35 showed similar large DJF underestimations of PM2.5 and attributed it to underestimated nitrate aerosol and black/organic

carbon. Myhre et al. (2006) and Schaap et al. (2004) encountered similar negative bias for these aerosol components. Further we see that, that using other kv calculation method which produce stronger diffusivities will results in higher turbulent removal of PM2.5 from near the surface resulting in more pronounced negative bias. However, similar to the case of ozone, this does not imply that these "stronger" Kv schemes are wrong as, again, the model bias has multiple components (and here is caused

5   most probably bias in the secondary aerosol formation). In such cases, improving model physics can often lead reduced model accuracy.

In summary, a weak sensitivity of modeled concentrations to the resolution of the driving meteorological as well as CTM is seen as concluded earlier by Markakis et al. (2015) who performed similar climate driven air-quality simulations over Paris (France). Indeed, the largest uncertainity of modeled concentrations is associated with emissions, especially over urban

10   areas (Aleksankina et al., 2019), however, recall that in our case, emission where kept constant and the uncertainty to the representation of urban boundary layer was analyzed only.

The average urban canopy impact on temperature is very similar to the values presented in Huszar et al. (2018a) who, for the two analyzed cities (Berlin and Prague), encountered increases up to 1.5–2° C. The diurnal variation of the impact shows also large similarities in both quantitative and qualitative sense, when the maximum impact occurs around late evening. The

15   urban impact on temperature is consistent with previous ~~observation~~ observation (Gaffin et al., 2008) and model based studies (Pichierri et al., 2012; Giannaros and Melas, 2012; Struzewska and Kaminski, 2012) while Sarrat et al. (2006) simulated somewhat later timing of the maximum urban impact. Our simulations, extended to ~~DJF~~winter, suggest that the magnitude of the impact ~~remain~~ remains similarly high during winter and here probably the effects of anthropogenic heat dominate over the radiative and thermal effects associated with radiation trapping ~~(Karlický et al., 2018; Varentsov et al., 2018)~~(Karlický et al., 2018; Varentsov et

20   Although increased resolution results in spatially more detailed impact with maximum temperature ~~impact~~ increases concentrated to the city center (seen especially for Prague in the 3 km ~~run~~simulation), there is no systematic effect of the choice of resolution. ~~It~~The impact however strongly depends on how the model grid covers a particular city and if the city center matches a grid point (Prague) or it lies in-between (Berlin). This suggest that efforts should be made to adapt the grid to the geographic location of city centers which is best achieved by choosing multiple disjunct nested domains under the parent grids (e.g. as in

25   Wang et al., 2012).

The impact on average wind speed with values up to -2 ms^{-1} ~~decreases match~~ matches previous studies too. ~~E.g.~~ For example, Struzewska and Kaminski (2012) simulated similar ~~decreases~~ reductions for central European cities, but Chinese cities in Zhu et al. (2017) ~~encounter comparable decreases as well~~encountered comparable decreases too. Compared to the decreases in Huszar et al. (2018a), up to -1 ms^{-1}, our results suggest that in case of wind, higher resolutions bring stronger

30   impact. Indeed, the 27 km and the 9 km impacts are always less than the 3 km result for Prague. This indicates that higher resolution yield stronger peak wind modifications in city centers. A sudden decrease (in absolute sense) of the wind-impact during evening hours is visible in our simulations too and is probably connected to the evening PBL transition (Lapworth, 2003). The wind impact during winter turned to be somewhat larger than summer, which is connected to more stable stratification when less turbulence is present (see further) and less momentum is transfered to the surface from above the PBL. The timing of both

the minimum and maximum urban canopy wind speeds (as well as those of temperature) is in very good agreement with Zhang and Zheng (2004).

Our results suggest large sensitivity of the modeled vertical eddy diffusivities on the choice of the ~~methods~~ method for its calculation and the same is true for the urban ~~impact~~canopy induced changes. Near surface Kv modifications ~~range~~ spawn two orders of magnitude from 0.5 to as much as 30 $\mathrm{m^2 s^{-1}}$. The impact over higher ~~levels cover~~ model levels covers even wider range of values from 1 up to 100 $\mathrm{m^2 s^{-1}}$. The TKE based methods systematically generate higher Kv values, this is in line with Kim et al. (2015) who compared a improved version of MYJ used here and the YSU scheme. They concluded that MYJ ~~generate~~ generates stronger vertical mixing at about 400–500 m above the surface than the YSU scheme, although the depth of the PBL was larger in YSU in their simulations. Our Kv values for the CMAQ method are in good agreement with the values of Huszar et al. (2018a) calculated ~~according to~~ using the same method. Kim et al. (2015) predicted Kv increase due to application of urban canopy scheme up to 100 $\mathrm{m^2 s^{-1}}$, same as in our study and the height of the maximum increase well coincides with our results too. In ENVIRON (2011), OB70 and YSU methods generated the smallest Kvs too, in line with our results. They however predicted Kv values for the MYJ methods comparable to CMAQ (and comparable to our CMAQ values), in contrary to our results. ~~These~~ This could suggest that the turbulent kinetic energy (TKE) ~~,~~ from which Kv is derived, is somewhat overestimated in our RegCM simulations. ~~Resolution~~

The resolution seems to have rather minor impact on the result near the surface. However, finer resolution results in stronger impact at higher elevations. This seems to be true for both the absolute diffusivities as well as for the urban canopy impact. Regarding the difference between winter and summer, ~~DJF~~ winter absolute Kv values are lower, as expected from the more frequent stable stratification. The urban canopy impact on Kv at higher levels are stronger in summer, however at the surface the difference between summer and winter is rather small. The maximum Kv impact is simulated in all diffusion schemes during early evening. This is related to the evening PBL transition when urban PBL height ~~decrease~~ decreases more slowly than the rural ~~ones~~ one resulting in relatively large urban-rural difference and hence large impact on Kv (Pal et al., 2012).

The simulated changes in the vertical eddy diffusion coefficient alone led to ozone increases as expected from previous studies of Sarrat et al. (2006); Kim et al. (2015). NOx is removed from the surface layer more efficiently over urbanized areas leading to reduced ozone titration while at the same time the increased transport from the residual layer during nighttime is contributing too, as showed by Huszar et al. (2018a). They used the CMAQ method for Kv calculation, and simulated (for JJA) a higher ozone increase up to 3 and 2 ppbv over Berlin and Prague, respectively, compared to our CMAQ results of about 1.5 and 0.8 ppbv for Berlin and Prague. Higher kv-impact on ozone, up to 4 ppbv is simulated with the TKE based methods, but comparable magnitude is achieved also with the ACM2 and OB70 approaches. It is not clear ~~weather higher resolution~~ whether higher resolutions lead to systematic change of the urban core kv-impact ~~,~~ while often the middle resolution (9 km) ~~exhibit~~ exhibits the largest value. Interestingly, the kv-impact for OB70 is one of the strongest while the urban canopy induced Kv modifications were at the low end in case of this ~~methods~~method. This is also true for the ACM2 method for which the Kv ~~modification where~~ modifications were far not the strongest however this propagated to almost the strongest impact on near surface ozone. The impact during winter is somewhat stronger than in summer, maybe due to more important role of the reduced titration in winter, when NOx levels are higher. Again, the CMAQ and YSU methods generate weakest ozone changes.

In summary, the simulated range of Kv changes of about 0.5 to 30 m^2s^{-1} propagated to a range of the kv-impact near the surface of about 0.6 to 4 ppbv. The vertical profile of ozone changes further showed decreases at higher levels. This is probably due to NOx removed from the surface causing titration at higher levels but the increased transport from RL ~~removing ozone there~~ can contribute too. Zhu et al. (2015) showed that in urban plumes, ozone is increased in upper boundary and decreased in the lower one. Our ~~result~~ results exhibit a small increase of ozone at higher levels and the mentioned decrease often extends to the lower boundary layer, putting our results in line with theirs. The range of elevated ozone decreases due to different Kv methods is relatively large, spanning from -0.4 to -2 ppbv.

The response of ~~the near surface~~ PM2.5 concentrations to urban canopy enhanced vertical turbulent transport shows predominantly decreases ~~as the process of~~ due to the dominant role of the increased eddy diffusion removal ~~dominates~~ (Zhu et al., 2017; Kim et al., 2015). Huszar et al. (2018b) however showed, that individual aerosol components react differently to increased eddy transport and secondary aerosols can even increase slightly ~~. However, in their study, primary aerosol decrease strongly shaping~~ although the decrease of the primary aerosols shaped the total PM2.5 response. Our results suggest PM2.5 decrease from a few -0.1 μgm^{-3} up to -1 μgm^{-3} in summer and even stronger decrease (up to -2 μgm^{-3}) in winter, which is - at least for summer - quantitatively close to Huszar et al. (2018b) who simulated summer decreases for Berlin and Prague up to -2 and -1 μgm^{-3}, respectively. The decreases due to urban induced Kv enhancement in Kim et al. (2015) are somewhat larger, but this can be due to different city chosen (Paris). Our results indicate that in connection with Kv enhancement, even increases can occur. This is probably due to the fact, that the aerosol removed from the surface level to the upper ones is transported to other regions where it deposits back to lower levels overweighting the turbulence enhanced local reduction. This effect is expected to be strong over larger domain where the simulation allows the aerosol to be transported to larger distances. Indeed, our 27 km domain results show some PM2.5 increase, especially over rural areas but also for Berlin. Our results further showed that the winter PM2.5 decreases are larger, which is in line with the fact that the winter urban canopy induced Kv ~~increase~~ increases are higher too. Comparing the individual Kv methods, it is seen that the strongest impact is simulated by the ACM2 scheme, but the OB70 and the TKE based schemes are comparably strong too. On the other hand, the default CMAQ scheme generates much weaker impact along with the YSU scheme. Kim et al. (2015) concluded too that the MYJ scheme resulted in larger impact on PM than the YSU. We also found that during summer the strongest impact is modeled during afternoon/evening hours reaching -3 μgm^{-3}, in line with Huszar et al. (2018b). The choice of the Kv ~~methods~~ method however strongly determines the shape of the diurnal cycle of the impact, especially in summer.

The total urban canopy impact (i.e. the combined impact of temperature-, humidity-, wind- and turbulence) exhibited a relatively large range of values for ozone from -0.6 to 0.8 ppbv, depending on the resolution and season. This underlines the fact, that the total-impact encompasses multiple components that act simultaneously and have opposite sign (Huszar et al., 2018a). This ~~include~~ includes decrease due to increased $\text{NO} + \text{O}_3 \rightarrow \text{NO}_2$ reaction, decrease due to increased dry deposition, decrease due to increased titration caused by decreased wind and finally, increase due to vertical turbulent removal of NOx from the near surface where it causes titration. Over Prague, the wind and the temperature impact in winter was evaluated to be relatively large, while the CMAQ impact on Kv small. As the total-impact was evaluated only for this Kv method, this suggest that decreases due to increased temperature and decreased wind speeds will be strong and indeed, they apparently

overweighted the impact of enhanced turbulence leading to concentration decrease, especially over the fine resolution domain. However, when we compare this results to the much stronger kv-impact gained by other Kv methods, we clearly see that using them would turn ~~to~~ the impact into positive making the turbulence changes again a dominant factor shaping the total-impact. We can conclude here, that the vertical eddy diffusion is a dominant factor that determines the impact of urban canopy

5 forcing on ozone, however there is relatively large uncertainity given the way how Kv values are calculated and under certain conditions this dominance is not clear or can be even slightly overweighted by opposite effects.

In case of the total impact on PM2.5, the dominance of enhanced turbulence is clear for each resolution, city and season, although the magnitude covers a wide range, from a few -0.1 m^2s^{-1} up to -2 m^2s^{-1}. Here, the counteracting effect is the decreased windspeed which reduces the dilution into larger scales and leads to PM2.5 increases (Huszar et al., 2018b). As

10 already said, the CMAQ ~~methods~~ method was used to calculate the total impact and again, the PM2.5 changes due to kv-impact calculated using this method turned to be one of the smallest. We can conclude thus here that using other methods would lead to even larger decreases of PM2.5 due to the total urban impact, confirming the dominance of enhanced vertical eddy transport over urban areas.

In summary, our results confirm that turbulence is a prominent factor that determines the impact of urban canopy meteo-

15 rological forcing on the urban air-quality. In case of ozone, it leads to increased concentrations while PM2.5 responses with decreases. This holds for both summer and winter season. Finally, model resolution seems to play ~~rather a~~ a rather minor role and the effect of urban ~~induced modification~~ canopy induced modifications of vertical eddy diffusion is dominant in both coarse and high resolution model ~~experiments~~simulations. Our study demonstrates the dominant role of turbulent transport of pollutants above urban areas and stresses the need for further investigation how variation of urban land-use (within urban

20 mitigation and adaptation strategies) influence the pollutant transport from the urban canopy into the whole urban PBL.

Code and data availability. The model RegCM4.6 is freely available for public under https://gforge.ictp.it/gf/download/frsrelease/257/1784/RegCM-4.6.1.tar.gz. CAMx version 6.50 is available under http://www.camx.com/download/default.aspx. The RegCM2CAMx meteorological preprocessor used to convert RegCM outputs to CAMx inputs is available upon request from the main author. The complete model configuration and all the simulated data (3D for meteorological variables, 3D for ozone and PM2.5 and 2D for other chemical species) used

25 for the analysis are placed on the Data storage center of the Czech CESNET service (https://www.cesnet.cz/e-infrastruktura-3/data-storage/?lang=en) and counts about 40 TB. These data are available upon request from the main author.

Author contributions. PH provided the scientific idea and the design of the model experiments, and led the writing of the paper; PH, JK and MB assisted in maintaining the models and performing the runs, JD, PP, TN and KS contributed to the evaluation of the results, TH, MZ and JD helped with writing the paper

30 *Competing interests.* The authors declare that they have no conflict of interest.

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Table 1. Model ~~experiments (~~simulations ~~)~~ performed ~~for~~ with RegCM (1st column) and ~~for~~ CAMx (other columns). The 2nd column denotes which urban ~~effects are~~ effect is considered: NOURBAN – none; URB_t+q+uv – effect of temperature, humidity, wind; URB_t+q+uv+kv – effect of temperature, humidity, wind and ~~trubulence~~turbulence. The 3rd column lists the Kv-methods used.

Models			Domains		
RegCM	CAMx	Kv-method	PHA27	PHA09	PHA03
NOURBAN	NOURBAN	CMAQ	*	*	*
URBAN	URB_t+q+uv	CMAQ	*	*	*
		DIRECT	*	*	*
		ACM2	*	*	*
		OB70	*	*	*
		MYJ	*	*	*
		YSU	*	*	*
	URB_t+q+uv+kv	CMAQ	*	*	*
		DIRECT	*	*	*
		ACM2	*	*	*
		OB70	*	*	*
		MYJ	*	*	*
		YSU	*	*	*
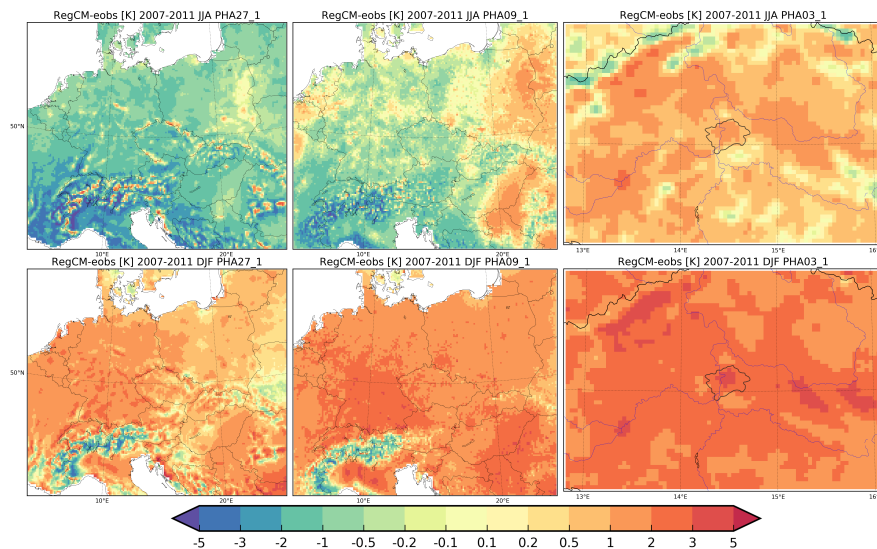
Figure 1. The difference between RegCM model near surface temperature and E-OBS data for 2007-2011 JJA (upper row) and DJF (lower row) for the 27, 9 and 3 km domain (columns) in ° C. Note that the 27 km domain is cropped to show only the part corresponding to the 9 km domain. For the right panel, the administrative boundary of Prague is indicated too.
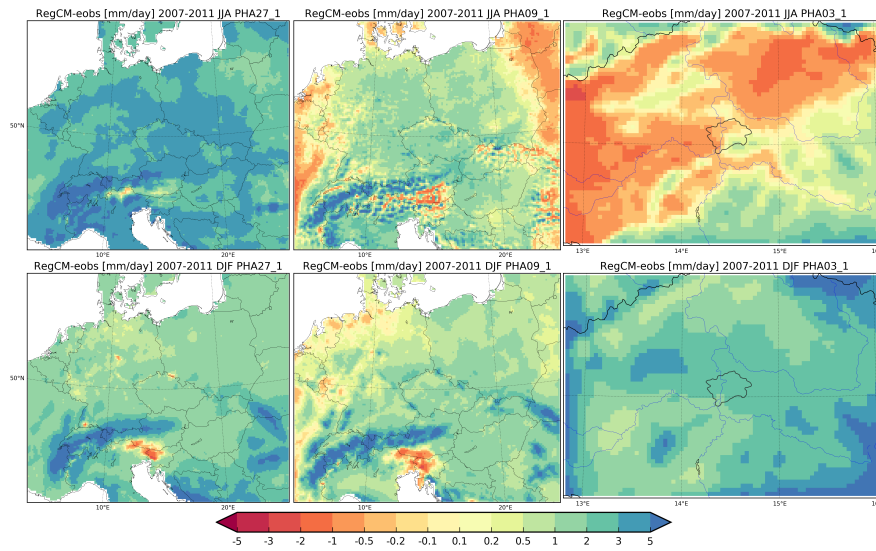
Figure 2. The difference between RegCM model precipitation and E-OBS data for 2007-2011 JJA (upper row) and DJF (lower row) for the 27, 9 and 3 km domain (columns) in mmday^{-1}. Note that the 27 km domain is cropped to show only the part corresponding to the 9 km domain. For the right panel, the administrative boundary of Prague is indicated too.
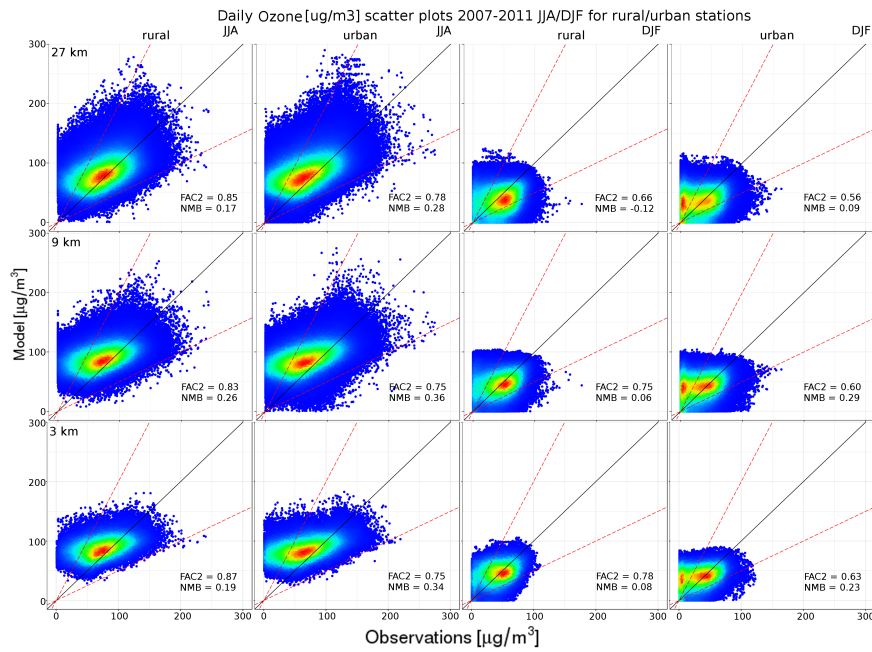
Figure 3. Scatter plots of daily ozone values for JJA (1st and 2nd column) and DJF (3rd and 4th column) for rural and urban stations in μgm^{-3}. The rows represent different model resolutions from 27 km (top) to 3 km (bottom). Dot colors stand for the density of the model-observation pairs. Red dashed lines define the "factor-2" (FAC2) region. Calculated values of FAC2 and the normalized mean bias (NMB) are indicated too.

Figure 4. Comparison of average diurnal cycles of ozone with measurements μgm^{-3} for JJA (1st and 2nd columns) and DJF (3rd and 4th columns) for rural and urban background stations for the three domains (27 km, 9 km and 3 km, from top to bottom): model (red), observation (green).
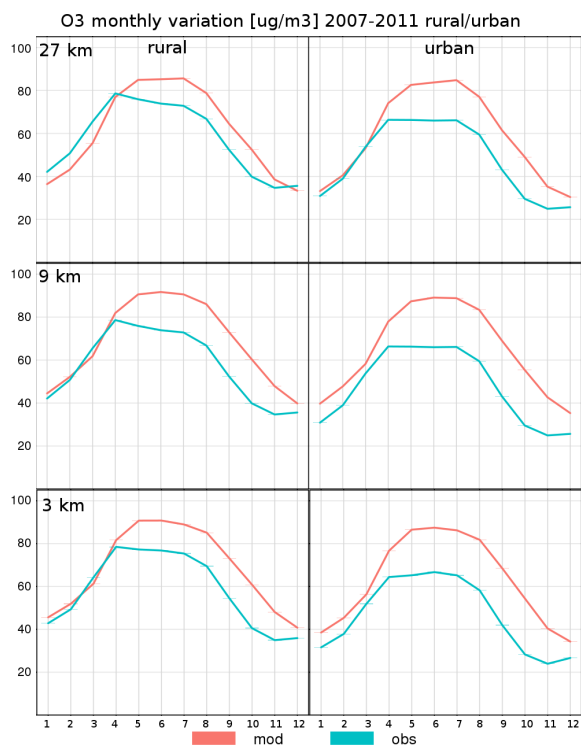
Figure 5. Comparison of monthly ozone values with measurements daily in μgm^{-3} for rural (left) and urban background stations (right) for the three domains (27 km, 9 km and 3 km, from top to bottom): model (red), observation (green).
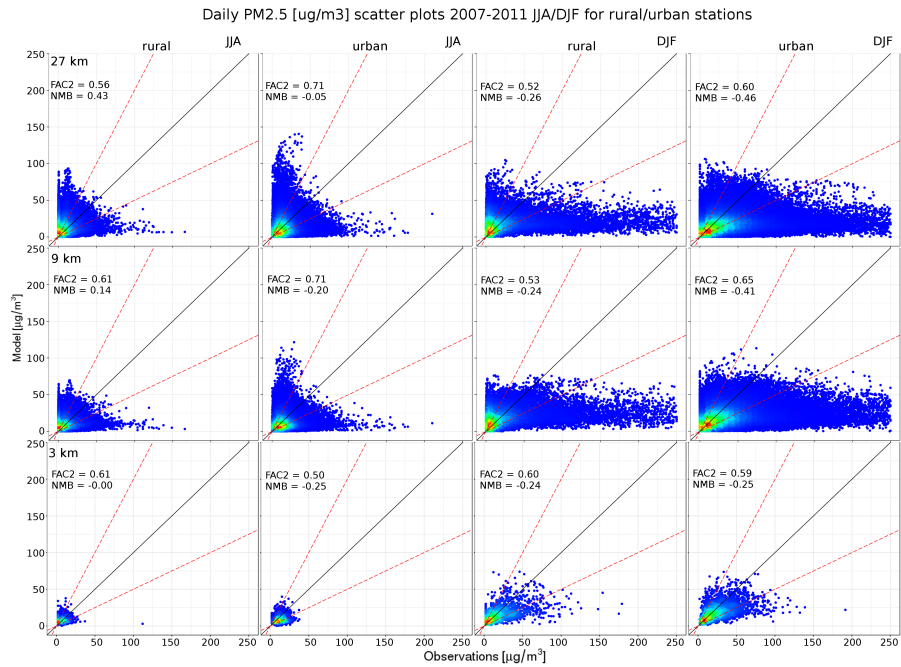
Figure 6. Scatter plots of daily PM2.5 values for JJA (1st and 2nd column) and DJF (3rd and 4th column) for rural and urban stations in $\mu\mathrm{gm}^{-3}$. The rows represent different model resolutions from 27 km (top) to 3 km (bottom). Dot colors stand for the density of the model-observation pairs. Red dashed lines define the "factor-2" (FAC2) region. Calculated values of FAC2 and the normalized mean bias (NMB) are indicated too.
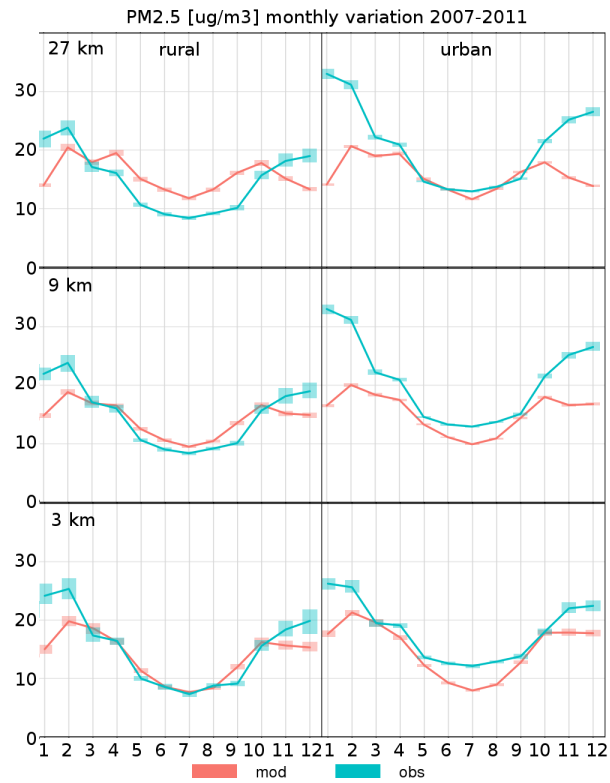
Figure 7. Comparison of monthly PM2.5 concentrations with measurements daily in μgm^{-3} for rural (left) and urban background stations (right) for the three domains (27 km, 9 km and 3 km, from top to bottom): model (red), observation (green).
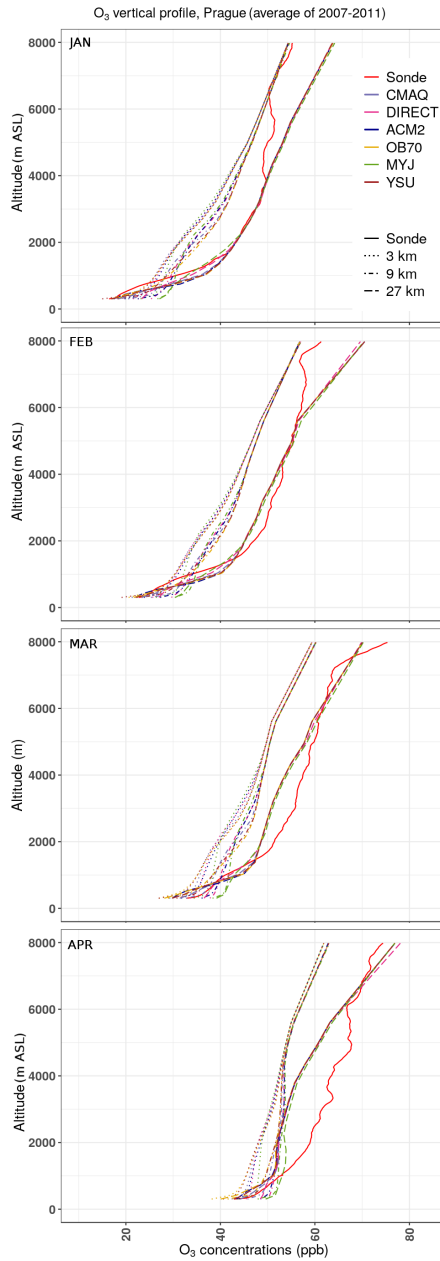
Figure 8. Comparison of monthly O_3 vertical profiles with sounding data for Prague, Czech republic for January-April. Red solid lines denote the measured data. Dashed, dotdashed and dotted lines stand for the 27, 9 and 3 km resolution model data and different colors mean different Kv methods. Measurements are from 12:00 UTC each day. Concentrations are in ppbv. Vertical axis means elevation above sea level.
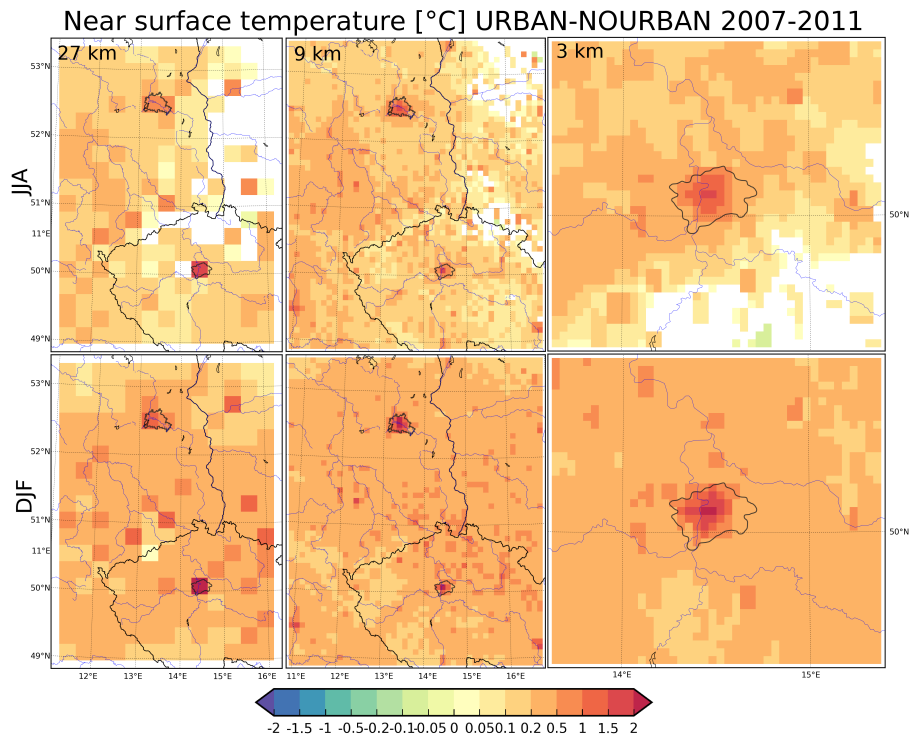
Figure 9. Impact of urban surfaces on near surface temperature in °C for JJA (upper) and DJF (lower row) for the three resolutions (27, 9 and 3 km). Shaded areas represent statistically significant changes over the 98% threshold using two tailed t-test. The geographic location of Berlin and Prague is indicated by their administrative boundary.
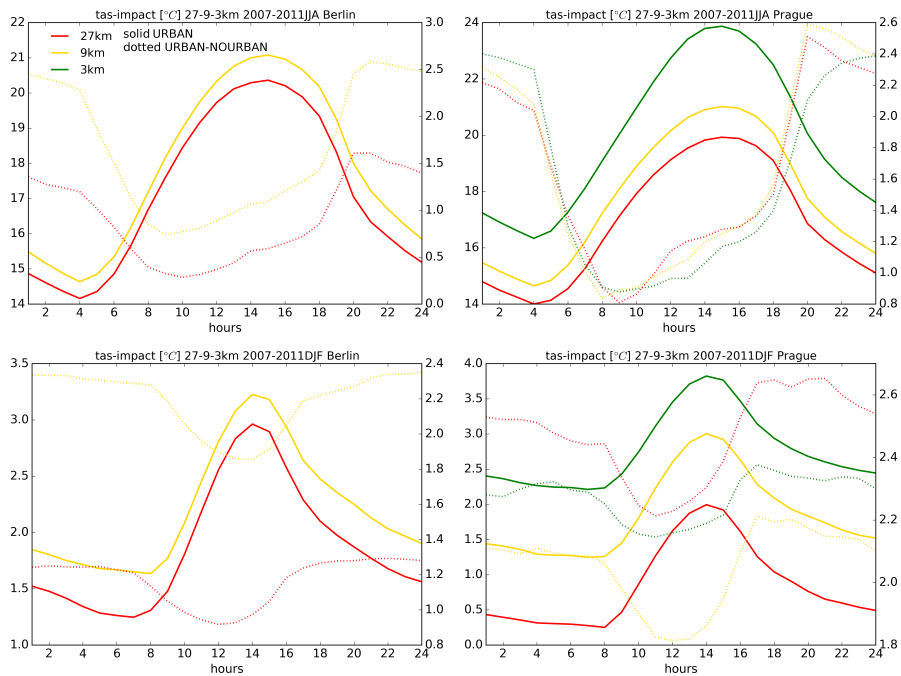
Figure 10. Impact of urban surfaces on near surface temperature diurnal cycle in °C for JJA (upper) and DJF (lower row) for the three resolutions (27 km – red, 9 km – orange and 3 km – dark green). Solid lines are the absolute values from the URBAN model experiment, dashed lines represent the urban impact.
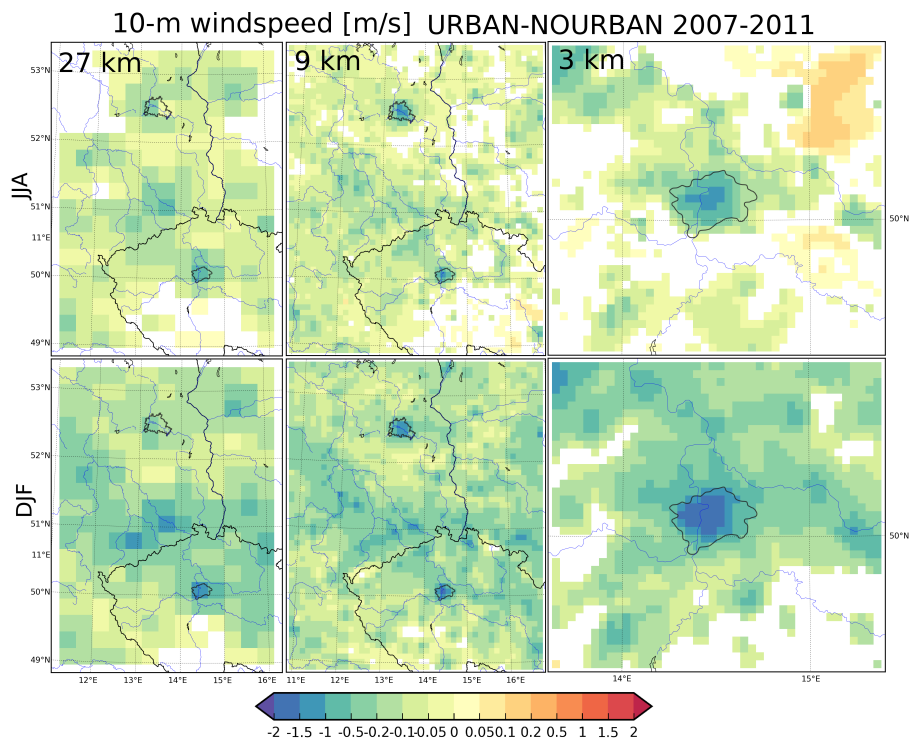
Figure 11. Impact of urban surfaces on 10 m wind speed in ms^{-1} for JJA (upper) and DJF (lower row) for the three resolutions (27, 9 and 3 km). Shaded areas represent statistically significant changes over the 98% threshold using two tailed t-test. The geographic location of Berlin and Prague is indicated by their administrative boundary.
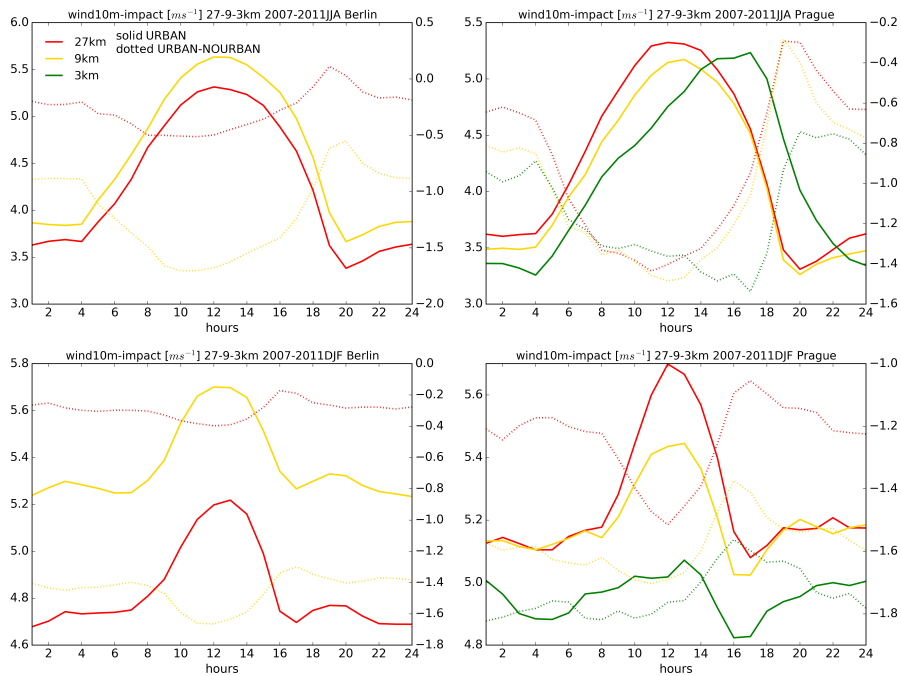
Figure 12. Impact of urban surfaces on the diurnal cycle of the 10 m wind speed in ms^{-1} for JJA (upper) and DJF (lower row) for the three resolutions (27 km – red, 9 km – orange and 3 km – dark green). Solid lines are the absolute values from the URBAN model experiment, dashed lines represent the urban impact.
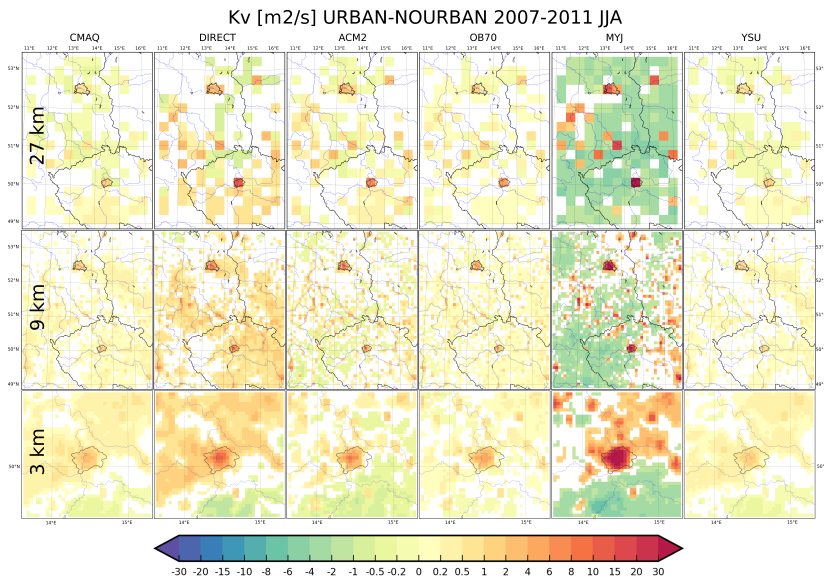
Figure 13. Impact of urban surfaces on the vertical eddy diffusion coefficient at 65 m in $\mathrm{m^2s^{-1}}$ for the 27 km (upper), 9 km (middle) and 3 km (lower row) for each Kv methods (CMAQ, DIRECT, ACM2, OB70, MYJ and YSU, from left to right). Shaded areas represent statistically significant changes over the 98% threshold using two tailed t-test. The geographic location of Berlin and Prague is indicated by their administrative boundary.
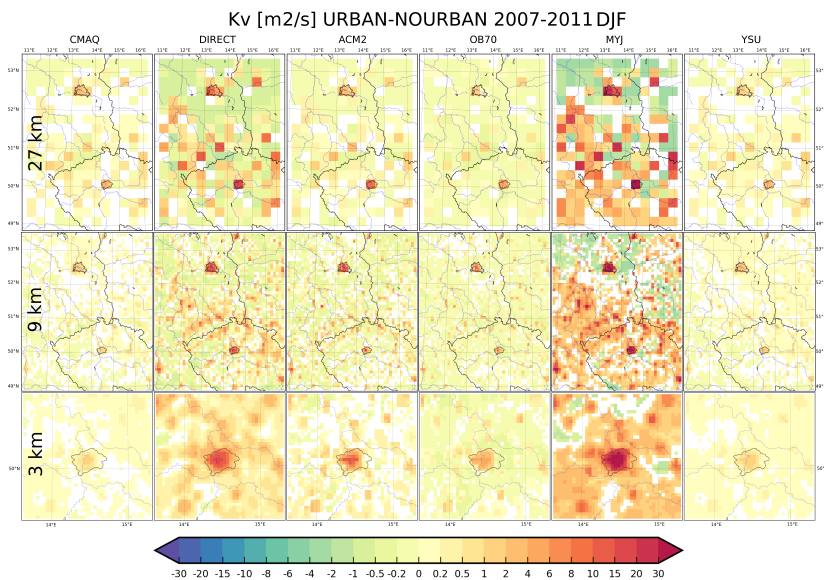
Figure 14. Same as Fig. 13 but for DJF.
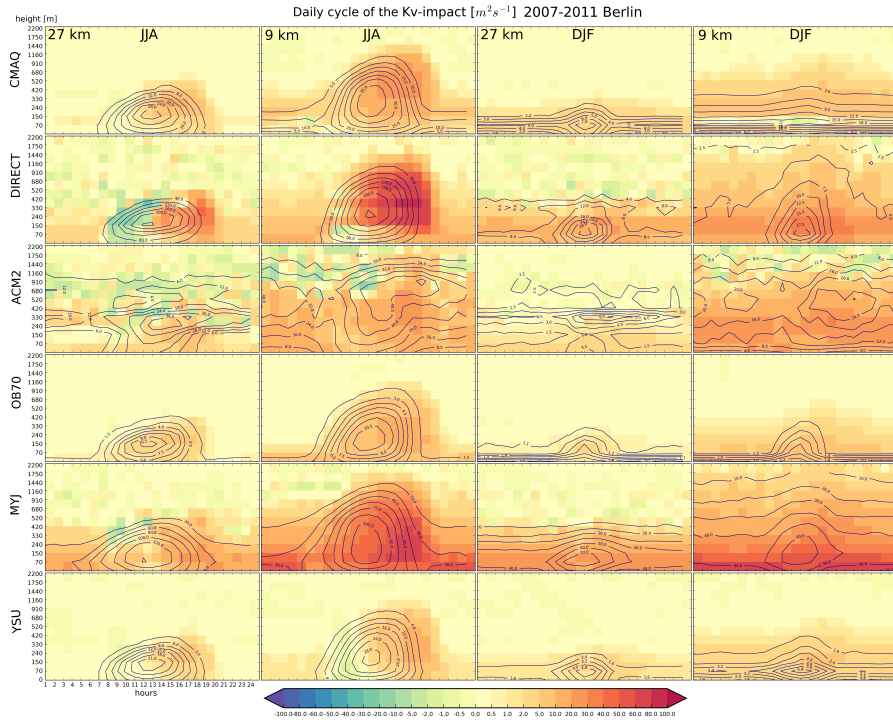
Figure 15. Impact of urban surfaces on the diurnal cycle of the vertical eddy diffusion profile in $\mathrm{m^2 s^{-1}}$ over Berlin for the 27 km (1st and 3rd column) and 9 km (2nd and 4th column) domains for both seasons (columns 1-2 JJA, columns 3-4 DJF). Individual rows represent different Kv methods (CMAQ, DIRECT, ACM2, OB70, MYJ and YSU, from top to bottom). Contour lines denote the absolute Kv values from the URBAN model experiment.
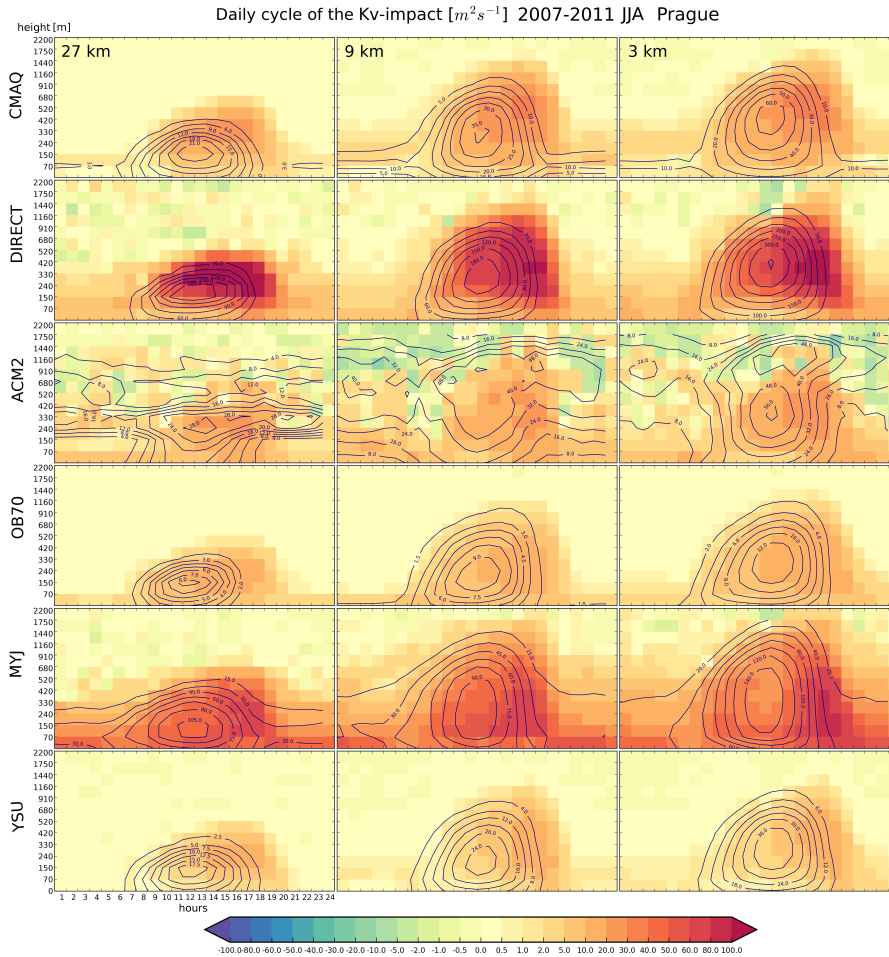
Figure 16. Impact of urban surfaces on the JJA diurnal cycle of the vertical eddy diffusion profile in $\mathrm{m^2s^{-1}}$ over Prague for the 27, 9 and 3 km domains (columns from left to right)4th column) domains. Individual rows represent different Kv methods (CMAQ, DIRECT, ACM2, OB70, MYJ and YSU, from top to bottom). Contour lines denote the absolute Kv values from the URBAN model experiment.
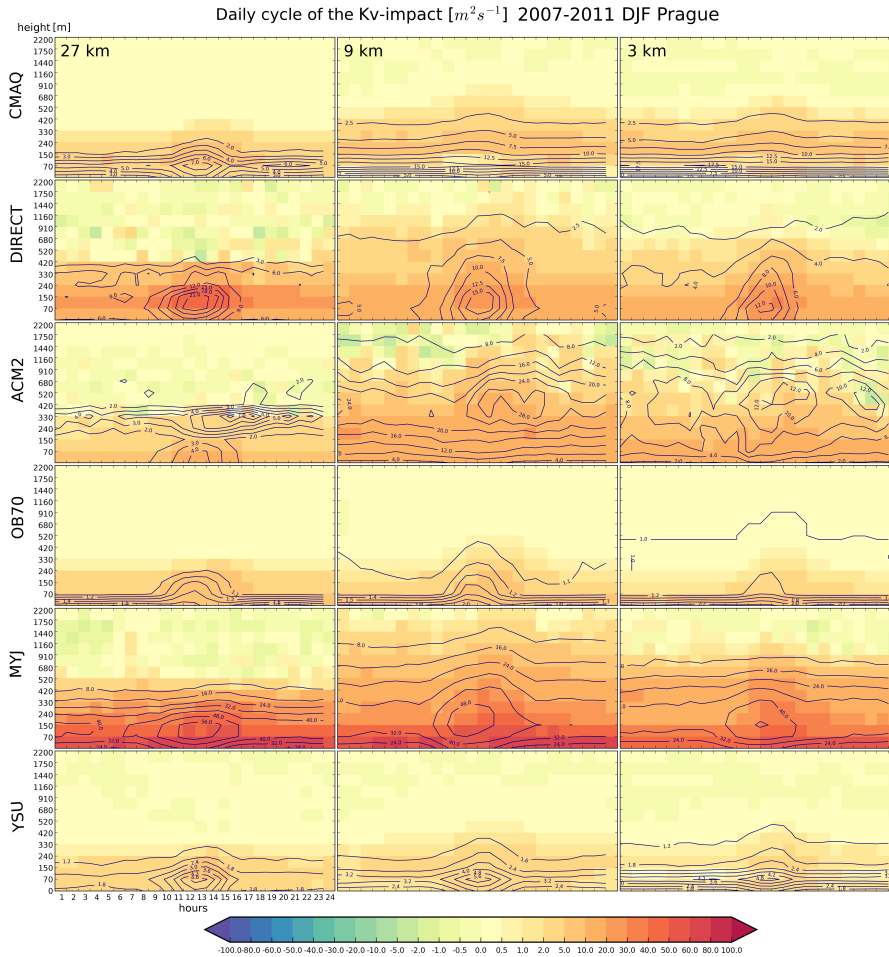
Figure 17. Impact of urban surfaces on the DJF diurnal cycle of the vertical eddy diffusion profile in m^{2}s^{-1} over Prague for the 27, 9 and 3 km domains (columns from left to right)4th column) domains. Individual rows represent different Kv methods (CMAQ, DIRECT, ACM2, OB70, MYJ and YSU, from top to bottom). Contour lines denote the absolute Kv values from the URBAN model experiment.
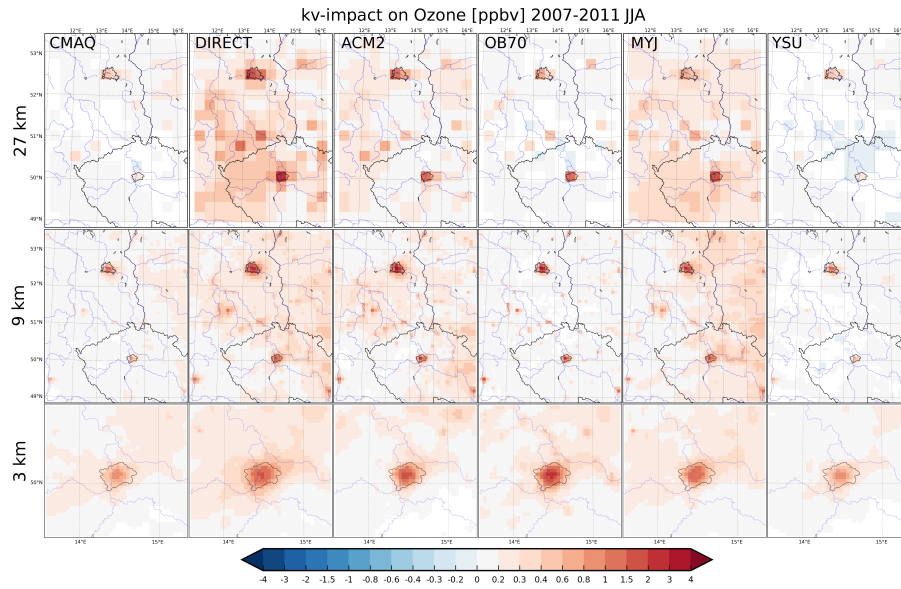
Figure 18. The impact of urbanization induced Kv enhancement (i.e. increase of vertical eddy diffusivity) on surface O_3 concentrations in ppbv for the 27, 9 and 3 km resolutions (top to bottom) for JJA for the six Kv methods (CMAQ, DIRECT, ACM2, OB70, MYJ and YSU). The geographic location of Berlin and Prague is indicated by their administrative boundary. Shaded areas represent statistically significant changes on the 98% level using two tailed t-test.
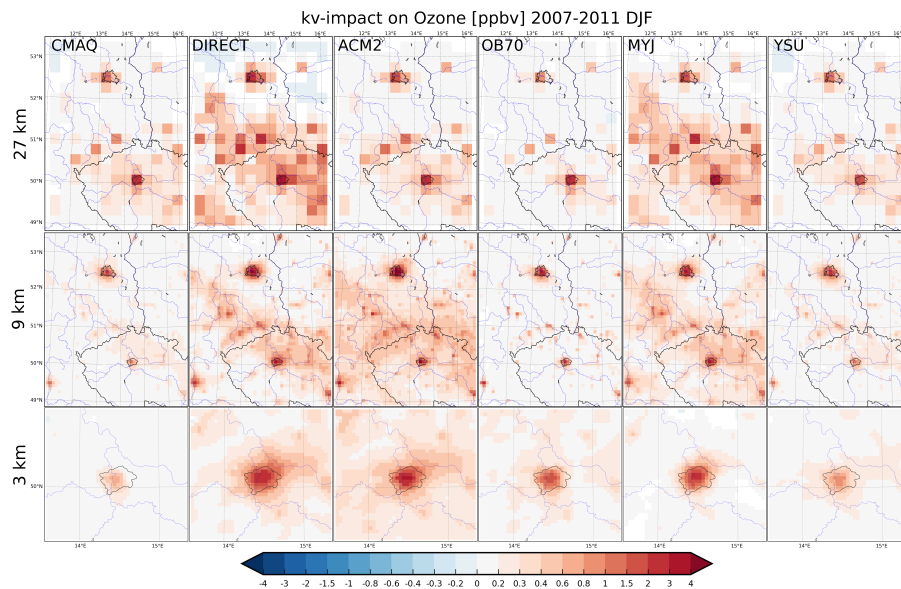
Figure 19. Same as Fig. 18 but for DJF.
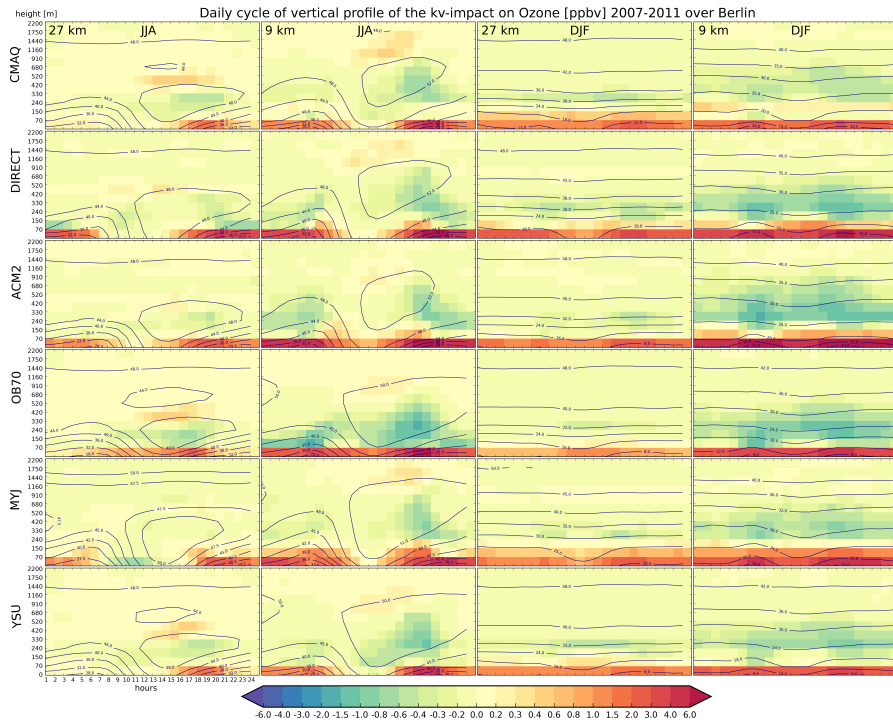
Figure 20. The impact of the urbanization induced Kv enhancement (i.e. increase of vertical eddy diffusivity) on the diurnal cycle of the O_3 profile over Berlin for JJA (columns 1–2) and DJF (columns 3–4) evaluated over the 27 km (columns 1 and 3) and 9 km (columns 2 and 4) domains. Rows correspond to individual Kv-methods. Colors mean the difference, contours stand for the absolute concentrations (from the total-impact). Units in ppbv.
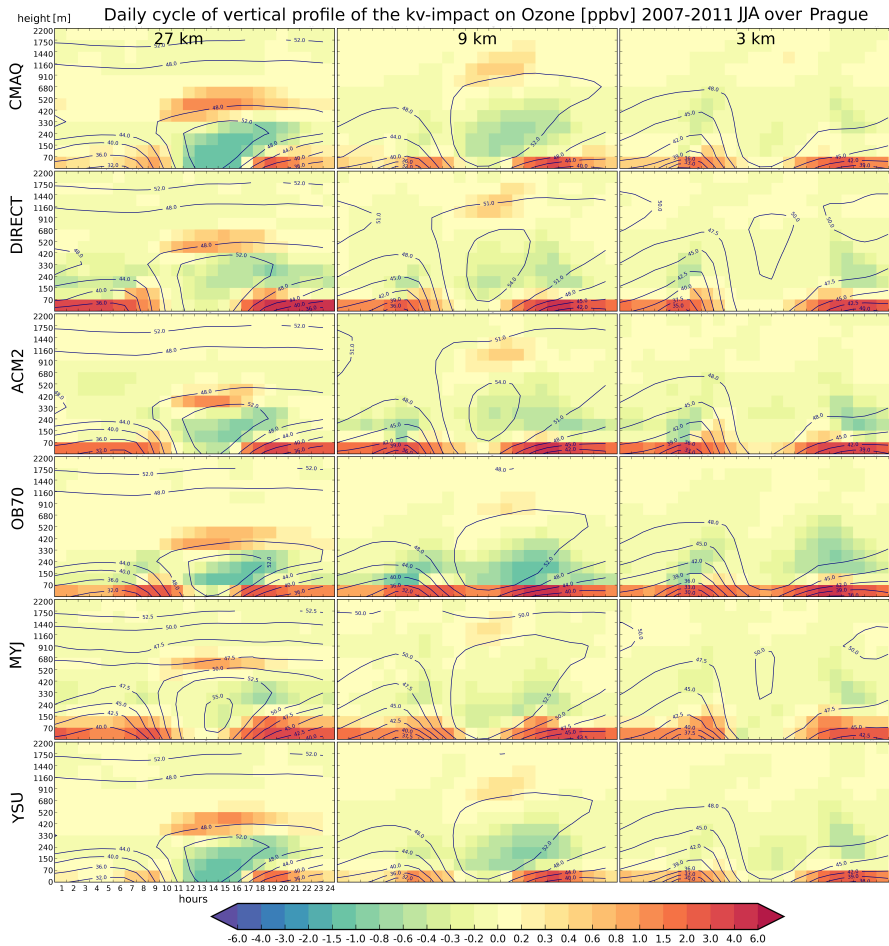
Figure 21. The impact of the urbanization induced Kv enhancement (i.e. increase of vertical eddy diffusivity) on the diurnal cycle of the O_3 profile over Prague for JJA. Columns represent the three resolutions (27km, 9km and 3 km, from left to right). Rows correspond to individual Kv-methods. Colors mean the difference, contours stand for the absolute concentrations (from the total-impact). Units in ppbv.
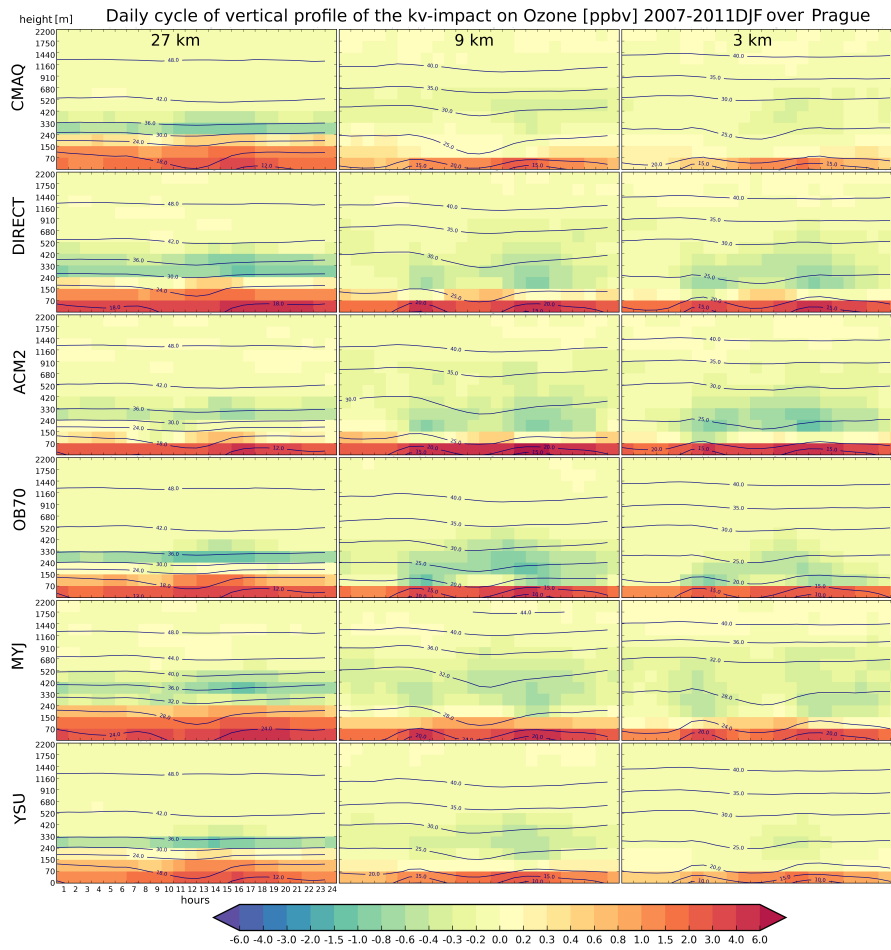
Figure 22. Same as Fig. 21 but for DJF.
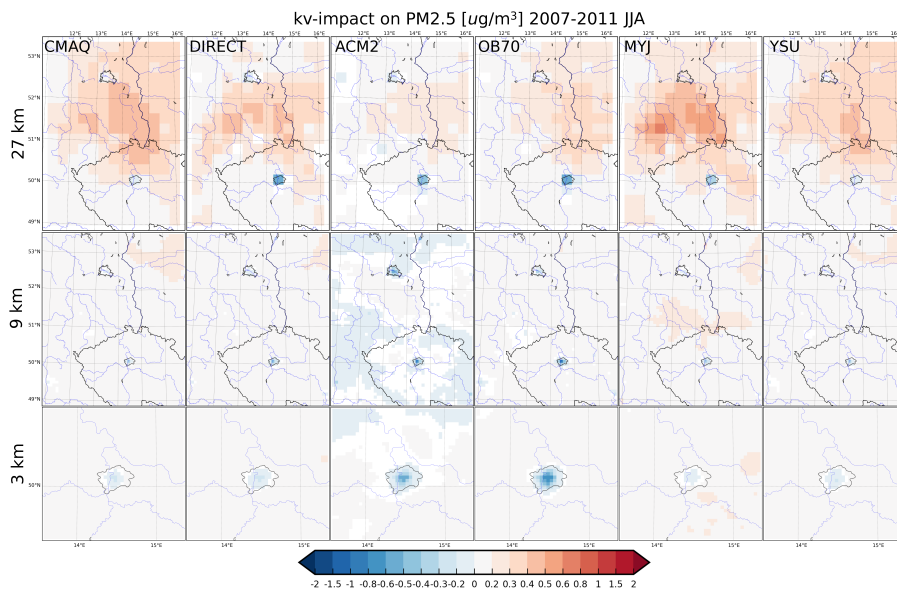
Figure 23. The impact of urbanization induced Kv enhancement (i.e. increase of vertical eddy diffusivity) on surface PM2.5 concentrations in μgm^{-3} for the 27, 9 and 3 km resolutions (top to bottom) for JJA for the six Kv methods (CMAQ, DIRECT, ACM2, OB70, MYJ and YSU). The geographic location of Berlin and Prague is indicated by their administrative boundary. Shaded areas represent statistically significant changes on the 98% level using two tailed t-test.
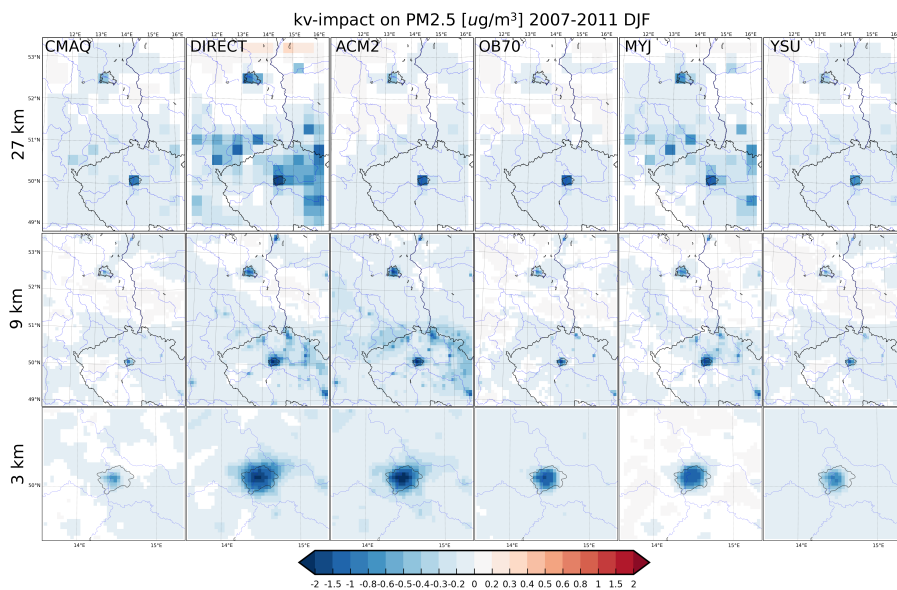
Figure 24. Same as Fig. 23 but for DJF.
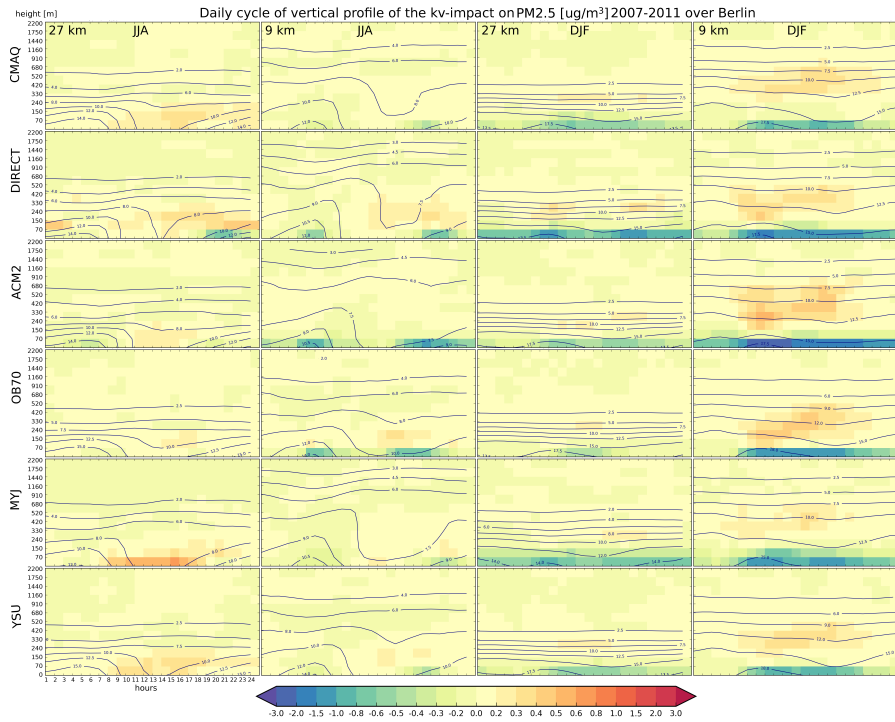
Figure 25. The impact of the urbanization induced Kv enhancement (i.e. increase of vertical eddy diffusivity) on the diurnal cycle of the PM2.5 profile over Berlin for JJA (columns 1–2) and DJF (columns 3–4) evaluated over the 27 km (columns 1 and 3) and 9 km (columns 2 and 4) domains. Rows correspond to individual Kv-methods. Colors mean the difference, contours stand for the absolute concentrations (from the total-impact). Units in $\mu\mathrm{gm}^{-3}$.
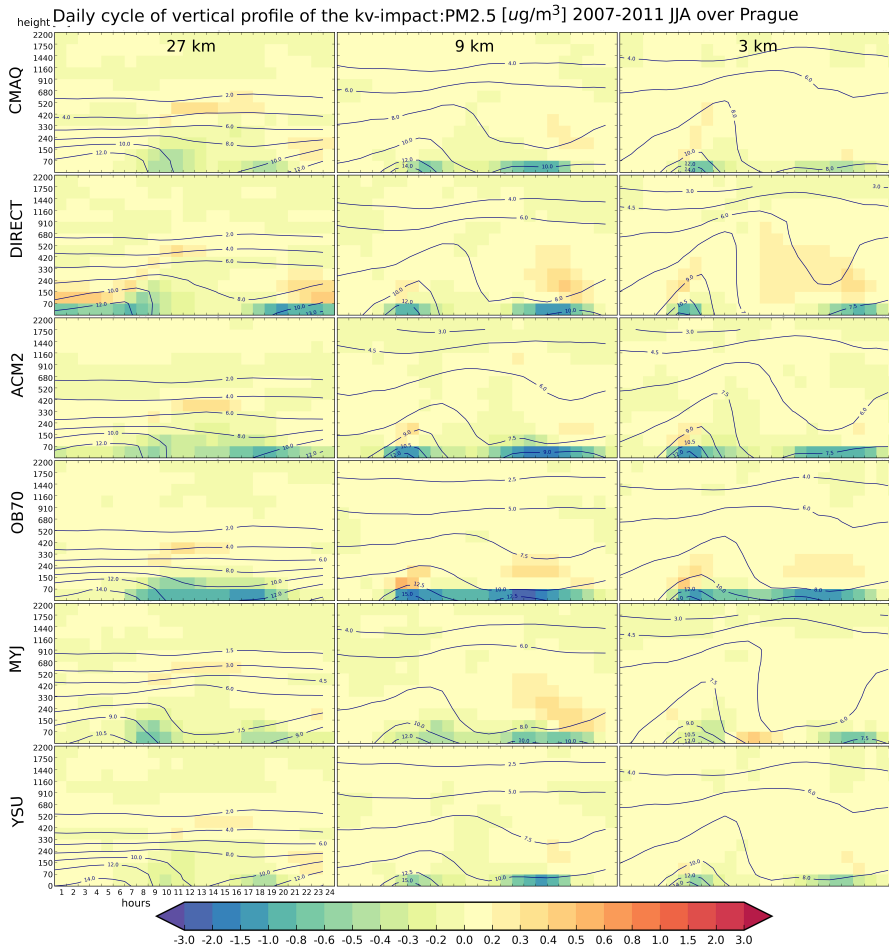
Figure 26. The impact of the urbanization induced Kv enhancement (i.e. increase of vertical eddy diffusivity) on the diurnal cycle of the PM2.5 profile over Prague for JJA. Columns represent the three resolutions (27km, 9km and 3 km, from left to right). Rows correspond to individual Kv-methods. Colors mean the difference, contours stand for the absolute concentrations (from the total-impact). Units in μgm^{-3}.
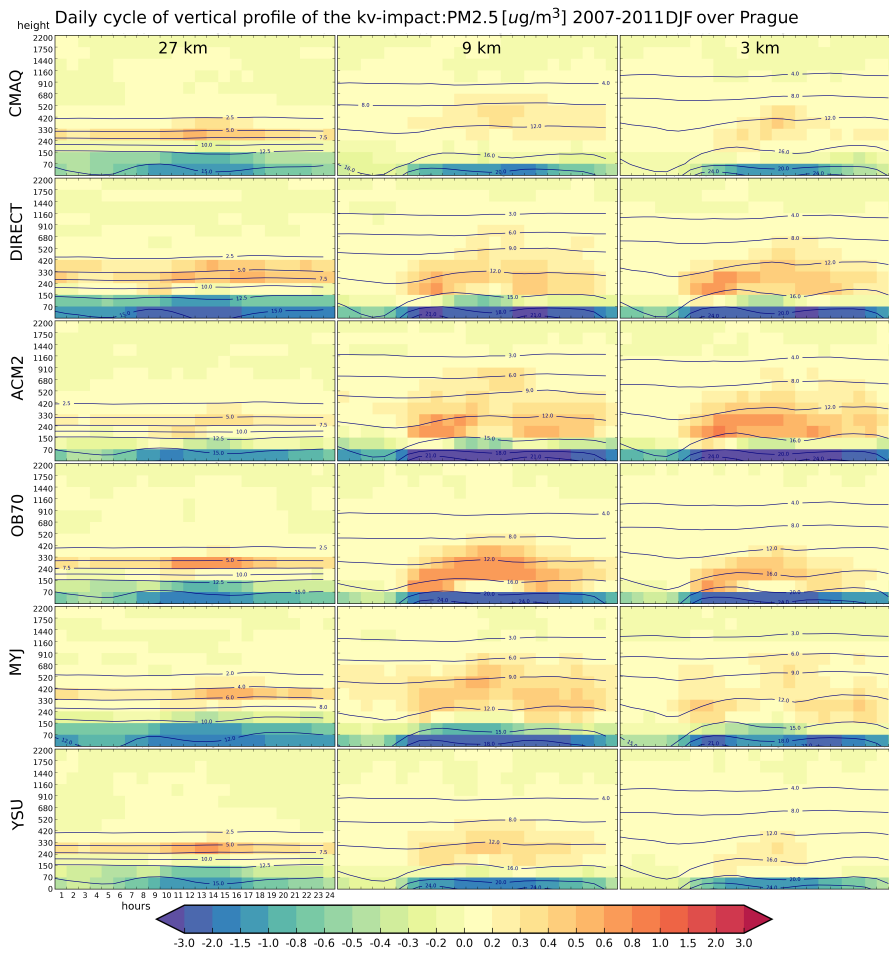
Figure 27. Same as Fig. 26 but for DJF.
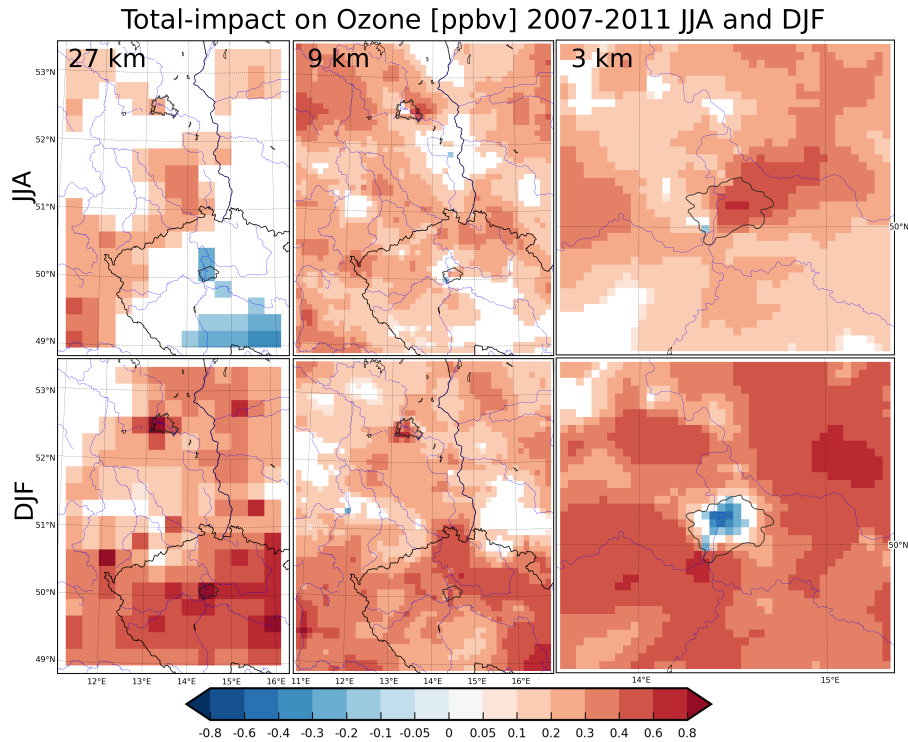
Figure 28. The total impact of urban meteorological changes (i.e. temperature, humidity, wind and vertical eddy diffusivity) on surface O_3 concentrations in ppbv for the 27, 9 and 3 km resolutions (left to right) for JJA (top) and DJF (bottom). Shaded areas represent statistically significant changes on the 98% threshold level using two tailed t-test. The default CMAQ Kv method was invoked in this calculation. The geographic location of Berlin and Prague is indicated by their administrative boundary.
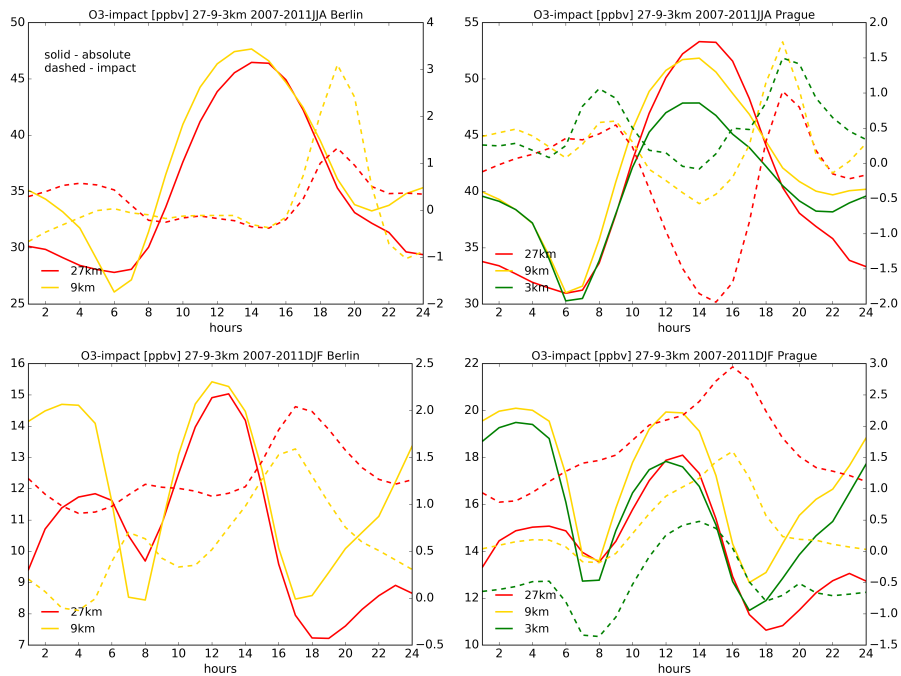
Figure 29. The diurnal cycle of the absolute ozone concentrations (solid) as well as the total-impact (i.e. the $t+q+uv+kv$-impact; dashed) above Berlin (left) and Prague (right) for JJA (top) and DJF (bottom) in ppbv. Red, orange and green colors stamd for 27km, 9km and 3 km resolutions, respectively.
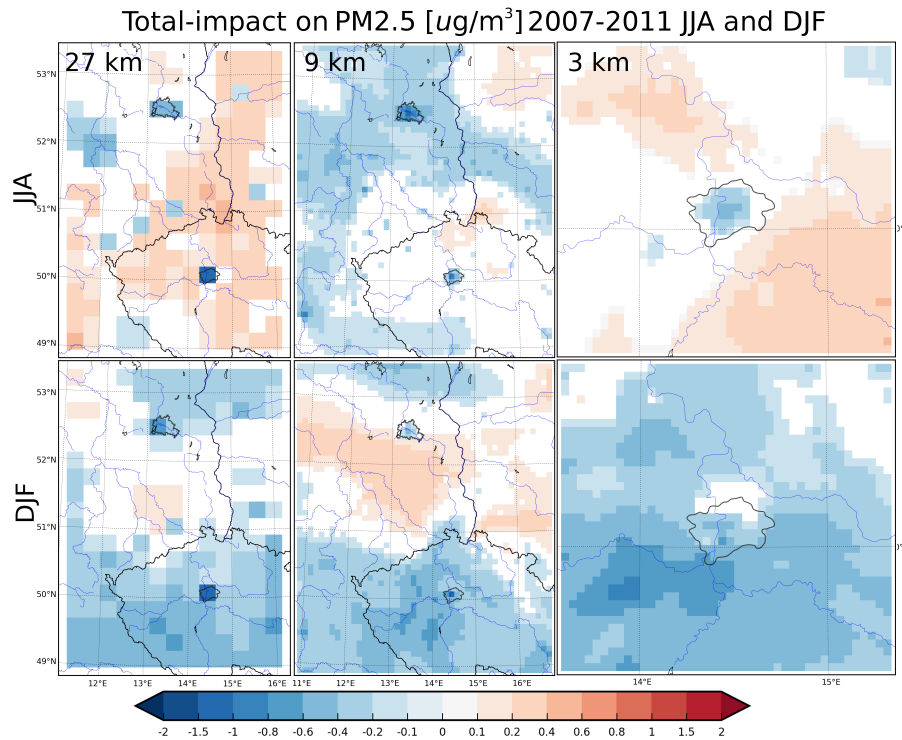
Figure 30. The total impact of urban meteorological changes (i.e. temperature, humidity, wind and vertical eddy diffusivity) on surface PM2.5 concentrations in ppbv for the 27, 9 and 3 km resolutions (left to right) for JJA (top) and DJF (bottom). Shaded areas represent statistically significant changes on the 98% threshold level using two tailed t-test. The default CMAQ Kv method was invoked in this calculation. The geographic location of Berlin and Prague is indicated by their administrative boundary.
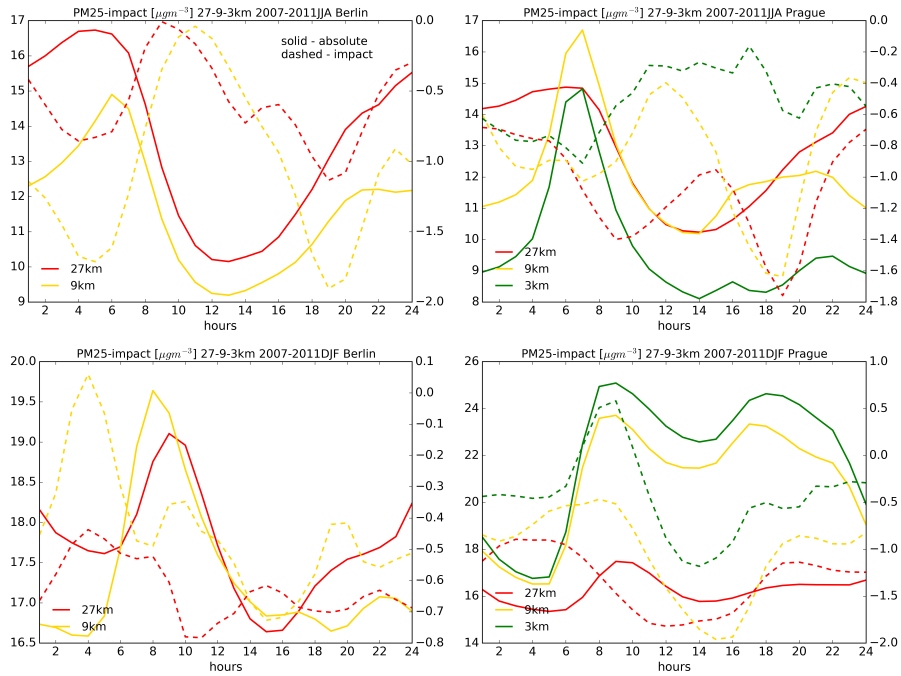
Figure 31. The diurnal cycle of the absolute PM2.5 concentrations (solid) as well as the total-impact (i.e. the $t+q+uv+kv$-impact; dashed) above Berlin (left) and Prague (right) for JJA (top) and DJF (bottom) in μgm^{-3}. Red, orange and green colors stamd for 27km, 9km and 3 km resolutions, respectively.