1	Enhanced heterogeneous uptake of sulfur dioxide on
2	mineral particles through modification of iron speciation
3	during simulated cloud processing
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24	Abstract. Iron-containing mineral aerosols play a key role in the oxidation of sulfur
25	species in the atmosphere. Simulated cloud processing (CP) of typical mineral
26	particles, such as illite (IMt-2), nontronite (NAu-2), smectite (SWy-2) and Arizona
27	test dust (ATD) is shown here to modify sulfur dioxide (SO2) uptake onto mineral
28	surfaces. Heterogeneous oxidation of SO2 on particle surfaces was firstly investigated
29	using an in situ DRIFTS apparatus. Our results showed that the BET surface area
30	normalized uptake coefficients (γ_{BET}) of SO ₂ on the IMt-2, NAu-2, SWy-2 and ATD
31	samples after CP were 2.2, 4.1, 1.5 and 1.4 times higher than the corresponding ones
32	before CP, respectively. The DRIFTS results suggested that CP increased the amounts
33	of reactive sites (e.g., surface OH groups) on the particle surfaces and thus enhanced
34	the uptake of SO ₂ . TEM showed that the particles broke up into smaller pieces after
35	CP, and thus produced more active sites. The "free-Fe" measurements confirmed that
36	more reactive Fe species were present after CP, which could enhance the SO ₂ uptake
37	more effectively. Mössbauer spectroscopy further revealed that the formed Fe phase
38	were amorphous Fe(III) and nanosized ferrihydrite hybridized with Al/Si, which were
39	possibly transformed from the Fe in the aluminosilicate lattice. The modification of
40	Fe speciation was driven by the pH-dependent fluctuation coupling with Fe
41	dissolution-precipitation repeatedly during the experiment. Considering both the
42	enhanced SO ₂ uptake and subsequent promotion of iron dissolution along with more
43	active Fe formation, which in turn lead to more SO ₂ uptake, it was proposed that there
44	may be a positive feedback between SO2 uptake and iron mobilized on particle
45	surfaces during CP, thereby affecting climate and biogeochemical cycles. This
46	self-amplifying mechanism generated on the particle surfaces may also serve as the
47	basis of high sulfate loading in severe fog-haze events observed recently in China.

49 1 Introduction

Mineral dust is a major fraction of global atmospheric aerosol budget, with an 50 51 estimated annual emission flux of 1000 to 3000 Tg into the atmosphere (Jickells et al., 52 2005; Andreae and Rosenfeld, 2008). Mineral dust aerosol mainly consists of quartz, 53 feldspars, carbonates (calcite, dolomite), and clay minerals (illite, kaolinite, chlorite, 54 montmorillonite), the exact composition varies with source (Claquin et al., 1999; 55 Formenti et al., 2008; Journet et al., 2008). A long range transport would result in a 56 decrease of quartz relative to the clay fraction because of the more rapid removal of 57 quartz, hence clay is an important component of mineral dusts (Mahowald et al., 2005; 58 Journet et al., 2008). During the long-range transport, mineral dust provides a reactive 59 surface for heterogeneous chemistry (Zhang et al., 2006; George et al., 2015; Huang 60 et al., 2015). Heterogeneous reactions of atmospheric trace gases on mineral dust 61 particles are of great significance as these reactions alter the chemical balance of the 62 atmosphere and modify the properties of individual particles (Usher et al., 2003; Wu 63 et al., 2011; Huang et al., 2015).

64 SO₂ is an important trace gas, which is released mainly by fossil fuel combustion and volcanic emission. The heterogeneous conversion of SO₂ on mineral dust surfaces 65 66 leads to the formation of sulfuric acid and sulfate aerosols, resulting in a significant 67 cooling effect on the global climate by scattering solar radiation and acting as cloud 68 condensation nuclei (CCN) to affect climate indirectly (Lelieveld and Heintzenberg, 69 1992; Usher et al., 2003; Kolb et al., 2010). In addition, sulfate containing particles 70 play a significant role in the haze formation in China in recent years (Sun et al., 2014; 71 Wang et al., 2014; Yang et al., 2017). SO₂ can be gaseous oxidized to sulfate by OH 72 radical, and be aqueous oxidation in cloud and fog droplets by ozone and hydrogen 73 peroxide (Luria and Sievering, 1991), or through heterogeneous processes that occur 74 on aerosol particle surfaces (Usher et al., 2003; Ullerstam et al., 2003). However, the 75 high sulfate levels measured in recent field observations cannot be explained by current atmospheric models (Kerminen et al., 2000; Wang et al., 2003; Cheng et al., 76 77 2016), leading to a large gap between the modeled and field-observed sulfate 78 concentrations using known oxidation pathways (Herman, 1991; Kasibhatla et al., 79 1997; Barrie et al., 2016). Overall, on a global scale, atmospheric SO₂ concentration were typically overestimated, while sulfate tended to be underestimated, suggesting 80 81 missing sulfate production pathways (Harris et al., 2013; Kong et al., 2014).

82 It has been suggested that the heterogeneous conversion of SO₂ could make an 83 important contribution to the atmospheric sulfate loading. Laboratory studies typically 84 focus on SO₂ uptake onto a variety of metal oxides and mineral particles (Goodman et 85 al., 2001; Usher et al., 2002; Zhao et al., 2015; Yang et al., 2016), and have confirmed 86 that its conversion rate on the surface of Fe (hydr)oxides was faster compared to other 87 metal oxides investigated, in good agreement with the field-measurement (Usher et al., 88 2002; Zhang et al., 2006). Atmospheric Fe is emitted from both anthropogenic 89 (primarily biomass burning, coal and oil combustion) and natural (mineral dust and 90 volcanic ash) sources, with the mineral dust source dominant globally (Siefert et al., 91 1998; Luo et al., 2008; Ito et al., 2019). It has been established that an important in 92 cloud S (IV) oxidation pathway is catalyzed by natural transition metal ions, 93 especially Fe hosted within mineral particles (Alexander et al, 2009; Harris et al., 94 2013).

Another important consideration for heterogeneous chemistry of mineral dust aerosol, is how mineral dust particles change in the atmosphere. During long-range transport, mineral particles often undergo chemical ageing by atmospheric processes (Mahowald et al., 2005; Baker and Croot, 2010; Shi et al., 2011). Cloud processing

99 involves cloud water condensation and evaporation on the particle surfaces, along 100 with drastic liquid water content and pH fluctuations (Mackie, 2005; Shi et al., 2011; 101 Rubasinghege et al., 2016). During CP, the high relative humidity (RH) results in high 102 aerosol water content and relatively high pH (Behra et al., 1989; Baker and Croot, 103 2010; Shi et al., 2011). While water evaporation from cloud droplets to wet aerosol at 104 higher temperature, the particles only contain a concentrated aqueous aerosol solution, 105 in which the pH can be lower than 2 (Zhu et al., 1993; Meskhidze, 2003; Shi et al., 106 2015). Therefore, there is a highly acidic film (e.g., pH = 2) in the "wet aerosol" phase 107 versus a less acidic droplet (near-neutral, 5-6) in the "cloud droplet" phase within 108 clouds (Shi et al., 2015). During its lifetime, a typical aerosol particle may experience 109 several cloud cycles involving large pH variations before being removed from the 110 atmosphere as rain or through dry deposition (Pruppacher and Jaenicke, 1995; Maters 111 et al., 2016). Herein, the simulated CP experiment was conducted by changing pH between 2 and 5-6, in accordance with the previous studies (Spokes et al., 1994; 112 113 Mackie, 2005; Shi et al., 2009).

114 It was well documented that pH is especially important for Fe mobilization (Zhu et 115 al., 1993; Desboeufs et al., 2001; Deguillaume et al., 2010; Maters et al., 2016). The 116 fluctuating pH during CP will impact and change the Fe speciation and morphology in 117 dust particles (Zhuang et al., 1992; Wurzler et al., 2000; Shi et al., 2009; Kadar et al., 118 2014). The low pH will increase Fe solubility and bioavailability of dust during 119 transport, thereby providing Fe external input to the open ocean surface to promote 120 marine prime productivity (Spokes et al., 1994; Desboeufs et al., 2001). It has been 121 found that Fe-rich nanoparticle aggregates were formed from Saharan soil and 122 goethite upon simulated CP conditions, in good agreement with their 123 field-measurements from the wet-deposited Saharan dusts collected from the western

Mediterranean (Shi et al., 2009). Fe nanoparticles are more chemically reactive (Wurzler et al., 2000; Desboeufs et al., 2001), possibly lead to a remarkable difference in heterogeneous chemistry. However, little is known about the influence of CP on SO₂ uptake onto particle surfaces up to now.

128 In this study, we employed four typical Fe-containing mineral samples as 129 surrogates to perform simulated CP experiments. The SO₂ uptakes on the mineral 130 particles before and after CP were compared using in situ diffuse reflectance infrared 131 Fourier transform spectroscopy (DRIFTS). Transmission electron microscopy (TEM) 132 was applied to observe the morphological and mineralogical change of mineral 133 particles. The Fe speciation modification during simulated CP was further monitored 134 by the dissolved Fe measurement, the "free-Fe" analysis and Mössbauer spectroscopic 135 characterization.

136 2 Materials and methods

137 **2.1 Mineral particles**

138 The standard mineral samples of IMt-2, NAu-2 and SWy-2 were purchased from 139 the Source Clay Minerals Repository (Purdue University, West Lafayette, IN). ATD 140 was purchased from Powder Technology Inc. (Burnsville, MN, USA). The mineral 141 samples were coarsely ground using a mortar and pestle before being more finely 142 ground using an All-dimensional Planetary Ball Mill QM-QX (Nanjing University Instrument Plant) and were sieved to particle diameters $(D_p) < 45 \ \mu m$ prior to analysis. 143 144 The Brunauer-Emmett-Teller specific surface areas (S_{BET}) of the samples were 145 measured with a Quantachrome Nova 1200 BET apparatus. Total iron content (Fe_T) of the samples were determined using an inductively coupled plasma atomic emission 146 147 spectroscopy (ICP-AES, Jobin Yvon Ultima). The chemical compositions of the 148 particles were analyzed by X-ray fluorescence spectrometry (XRF, PANalytical Axios

149 Advanced).

150 **2.2 Cloud processing simulation experiment**

The simulated CP experiments were conducted at a constant temperature $(298 \pm 1K)$ 151 152 using a Pyrex glass vessel with a water jacket. The suspensions contained a mineral particle loading of 1 g L⁻¹ were subjected to acidic (pH = 2 ± 0.1 , 24 h) and 153 near-neutral pH (pH = 5-6, 24 h) cycles for 1-3 times according to the previous 154 155 methods (Spokes et al., 1994; Mackie, 2005; Shi et al., 2009). Suspension pH was 156 adjusted by adding dilute H₂SO₄ or NH₄OH. The CaCO₃ equivalent alkalinity of the 157 dust was determined in accordance with APHA method 2320B so that acid additions 158 to control pH could be adjusted accordingly (Mackie, 2005). The amount of acid or 159 alkali added to achieve these pH cycles was less than 1% of the total volume of the 160 suspensions. The experiments were performed under a constant stirring (about 50 161 rpm) in the dark for 144 h. At the end of the CP experiment, the suspensions were 162 filtered through 0.2 µm PTFE filters (Millipore). The filter residue was air-dried, and was further applied to the DRIFTS experiment, as well as TEM observation, "free-Fe" 163 164 measurement and Mössbauer spectroscopic characterization.

165 2

2.3 SO₂ uptake on the mineral particles

166 The SO₂ uptake on the particle surfaces before and after CP was investigated by a 167 Shimadzu Tracer-100 FTIR spectrometer equipped with a high-sensitivity mercury 168 cadmium telluride (MCT) detector and a diffuse reflectance accessory. A temperature 169 controller was fitted to the DRIFTS chamber to ensure constant reaction temperature 170 (298 K). Weighted sample was placed into a ceramic crucible (0.35 mm depth, 5 mm i. d.) in the chamber. Mass flow controllers (Beijing Sevenstar electronics Co., LTD) 171 172 were used to adjust the reactant gases to a flux with expected concentration and relative humidity. The sample was firstly pretreated in a 100 mL min⁻¹ flow of 173

174 synthetic air (21% O₂ and 79% N₂) for 1 h to blow off water and impurities on particle surface. When the background spectrum of the fresh sample reached steady 175 state, the reactant gas of SO₂ (5.0 ppm) along with synthetic air was introduced into 176 the chamber at a total flow rate of 120 mL min⁻¹ for 45 min, during which the IR 177 spectrum was recorded automatically every 5 min at a resolution of 4 cm⁻¹ for 100 178 179 scans in the spectral range of 900 to 4000 cm⁻¹. Atmospheric moisture was simulated 180 with a RH level around 40 % by guiding one high-pure air flux through water. The 181 humidity value was monitored using a hygrometer.

182 The sulfate products were analyzed by ion chromatography (IC) after the DRIFTS 183 experiments. The particles were extracted with 5 ml ultrapure water by ultrasonic 184 extractor. After 10 min, the extracted solution was passed through a 0.22 μ m PTFE 185 membrane filter and the leaching solution was analyzed using a Metrohm 883 Basic 186 IC equipped with an A5-250 column. A weak base eluent (3.2 mmol L⁻¹ Na₂CO₃ plus 1.0 mmol L⁻¹ NaHCO₃) was used for anion detection at a flow rate of 0.70 ml min⁻¹. 187 188 To discriminate the adsorbed sulfate during simulated CP experiment and the sulfate 189 ions generated from the heterogeneous reaction, the adsorbed sulfate on the particles 190 during simulated CP experiment were initially measured as blank. The heterogeneous 191 uptake of SO₂ was calculated by subtracting the blank value from the total sulfate 192 ions.

193 The reactive uptake coefficient (γ) was defined as the rate of sulfate formation on 194 the surface (d[SO₄²⁻]/dt, ions s⁻¹) divided by collision frequency (*Z*, molecules s⁻¹) 195 (Usher et al., 2003; Ullerstam et al., 2003; Kong et al., 2014; Huang et al., 2015).

196
$$\gamma = \frac{d[\mathrm{SO}_4^{2-}]/dt}{Z},\tag{1}$$

197
$$Z = \frac{1}{4} \times A_s \times [SO_2] \times v, \qquad (2)$$

198
$$v = \sqrt{\frac{8RT}{\pi M_{SO_2}}},\tag{3}$$

199 Where, A_s is the effective sample surface of the samples, m²; v is the mean 200 molecular velocity of SO₂, m s⁻¹; R is the gas constant, J mol K⁻¹; T is the absolute 201 temperature, K; and M_{SO_2} is the molecular weight of SO₂, kg mol⁻¹.

A conversion factor was obtained by a calibration plot with numbers of SO4²⁻ 202 203 analyzed by ion chromatography (IC, Metrohm 883 Basic, Switzerland) versus the 204 integrated areas of sulfate products from DRIFTS spectra. The residual sulfate during 205 simulated CP experiments were deducted as background. The calculated conversion factor of SO_4^{2-} is 1.170×10^{15} (ions \cdot integrated units⁻¹). Integrated areas for the total 206 207 sulfur-containing products were calculated to show the maximal sulfate formation rates. The reactive uptake coefficient for SO₂ was determined to be γ_{BET} and γ_{geo} using 208 the BET area ($A_{BET} = mass \times S_{BET}$) and geometric area ($A_{geo} = mass \times S_{geo}$) as the 209 210 reactive area, respectively.

211 **2.4 Morphological and mineralogical characterization of the Fe speciation**

212 A FEI TECNAI G2 S-TWIN F20 TEM equipped with an Oxford energy-dispersive 213 X-ray spectrometer (EDX) was used to analyze the morphological and chemical 214 composition of individual particles before and after CP. Suspensions (0.2 g L^{-1}) of 215 each particle were prepared in methanol and sonicated for at least 1 h. A drop of this 216 suspension was then applied to a carbon-coated Cu TEM grid (400 mesh; EMS) and 217 allowed to air-dry. The operation was conducted in bright field mode at 120 kV. The 218 Fe content of the typical individual mineral particle were calculated from the values 219 of 50 typical particles. To obviously observe the morphological changes, 220 high-resolution TEM (HRTEM) images were also collected to observe nanoscale structural features, e.g., surface roughness and lattice fringes. 221

222 The content of "free-Fe" in the mineral particles was determined by a 223 citrate-buffered-dithionite (CBD) sequential Fe extractions method according to the 224 literature (Lafon et al., 2004; Shi et al., 2009). Simply, 30 mg of the dust samples 225 were treated for 24 h with a 10 mL ascorbate solution (pH = 7.5) to extract chemically 226 highly labile Fe phases (Fe_A), mainly composed of amorphous, nanoparticle and/or 227 poorly crystalline ferrihydrite. The solutions were filtered through 0.2 μ m 228 polycarbonate filters. The dust particles collected on the filters were subsequently treated for 2 h with a 10 mL sodium dithionite solution (pH = 4.8) to extract 229 230 crystalline Fe (oxyhydr) oxides (Fe_D), which are mainly goethite and hematite. After 231 each reaction step, the dissolved Fe concentrations (Fe_A and Fe_D) in the filtrates were 232 determined using ICP-AES. The sum of these two pools (Fe_A + Fe_D) was defined as the "free-Fe" fraction (Shi et al., 2011). The other fraction was donated as the 233 234 "structural-Fe" in aluminosilicate crystals, which could be calculated from the 235 difference between the Fe_T and "free-Fe" fractions (Lafon et al., 2004).

236 The Mössbauer spectroscopic analysis performed in transmission geometry with a 237 constant acceleration was used to inspect the chemical valence and the surrounding structure of Fe in the particles before and after CP. ⁵⁷Co was used as the Mössbauer 238 239 source, and a 1 mm thick Na(TI) scintillator coupled to a EMI9750B photoelectric 240 multiplier was used as the detector (Cwiertny et al., 2008). The measurement was carried out at room temperature (RT) with a duration of 24 hours for one sample 241 (around 1.5×10^6 counts per channel). Experimental data were fitted by a least-squares 242 243 fitting-program. The isomer shift values were calibrated against a spectrum for α -Fe metal foil. 244

During the simulated CP experiment, the total dissolved iron (Fe_s) and the dissolved Fe(II) in the suspensions were measured colorimetrically by the Ferrozine method, as 247 described in previous studies (Viollier et al., 2000; Cwiertny et al., 2008). For Fe(II) analysis, 200 mL of a 5 mM 1, 10-phenanthroline solution and 200 mL of an 248 249 ammonium acetate buffer were added into 1 mL of sample. To avoid possible 250 interference from Fe(III), which can also form a complex with 1,10-phenanthroline 251 when present at high concentrations, 50 mL of 0.43 M ammonium fluoride was added 252 to the sample prior to 1,10-phenanthroline. The mixture was allowed to sit in the dark 253 for 30 min prior to ultraviolet-visible spectroscopy (UV-Vis) analysis, during which 254 time a reddish-orange color developed if Fe(II) was present. Fe_s was determined via 255 the same protocol, except that 20 mL of 1.5 M hydroquinone, which reduces Fe(III) to 256 Fe(II), was added to the sample rather than ammonium fluoride. Absorbance 257 measured at 510 nm was converted to concentrations using aqueous standards 258 prepared from anhydrous beads of ferrous chloride. Standards were prepared in each 259 acid used in dissolution studies, and no matrix effects were observed. These conditions resulted in a detection limit of 1 μ M. The concentration of dissolved 260 261 Fe(III) was calculated from the difference in experimentally measured concentrations 262 of total dissolved iron and dissolved Fe(II).

263 Additionally, the dissolved Fe(III) could precipitate out as the pH increased, and 264 then the Fe mineraology of the deposit was also observed. NAu-2 released about 265 300 µM of dissolved Fe at pH 2. The dissolving solution (200 mL) was sampled 266 after filtration (0.2 μ m polycarbonate filter). The clear solution was subjected to 267 changing acidity from pH 2 to 5 by the stepwise addition of dilute NH₄OH. The 268 precipitated particles were separated out by 0.2 μ m filters and were used in TEM and Mössbauer analysis. Size distributions for the Fe-bearing particles formed in 269 270 the suspensions were determined by a Horiba LB-500 light scattering microscopy 271 within the size range of 3-6000 nm.

272 **3 Results and discussion**

273 **3.1 Morphological change of the mineral particles after CP.**

274 The characteristic results are shown in Table S1 and Table S2. The samples 275 exhibited S_{BET} in the range from 4.3 ± 0.3 to 22.6 ± 2.3 m²/g. The Fe_T content were

276 5.45 \pm 0.34%, 26.30 \pm 0.57%, 2.36 \pm 0.56% and 1.48 \pm 0.56%, for IMt-2, NAu-2,

277 SWy-2 and ATD, respectively. The proportions of Fe₂O₃ in IMt-2, NAu-2, SWy-2

and ATD were 7.95%, 39.03%, 5.55% and 2.57%, respectively.

279 Figure 1 shows the TEM images of the mineral particles before and after CP. As 280 shown in Figure 1 a, c, e and g, the IMt-2, NAu-2, SWy-2 and ATD samples before 281 CP primarily consisted of laminar aluminosilicate with irregular shape and rough 282 morphologies mainly at the micrometer scale, all of which were characterized by various fractions of Fe (1.5%-26.2%), along with minor Mg (0.1%-16.5%), K 283 284 (0.0%-7.8%) and Ca (0.0%-1.1%). The Fe within the aluminosilicates of the particles 285 was evenly distributed. Besides, some Fe-rich crystal with several hundreds of 286 nanometers in size were found to attach onto the ATD particles, which were identified 287 as α -Fe₂O₃ (PDF: 33-664) from the typical *d*-spacing analysis of HRTEM (Janney et al., 2000). 288

289 After the simulated CP, all of the processed mineral particles showed much smaller 290 size than the ones before CP. For example, the typical IMt-2 and NAu-2 particles after 291 CP (Figure 1 b and d) were $< 1 \mu m$ in size. Under the TEM, the average Fe content of 292 the individual IMt-2 and SWy-2 particles (Figure 1 b and f) decreased from 5.5% (\pm 293 1.9%; n = 50 to 4.1% (± 1.6%; n = 50) and from 2.4% (± 0.6%; n = 50) to 2.1% (± 294 0.5%; n = 50), respectively. In addition, the IMt-2 particles after CP showed a 295 heterogeneous distribution of the Fe on the basis of the EDX data. Most of the 296 aluminosilicate in IMt-2 after CP hosted lower Fe content (4.1%), whereas a few of the Fe-rich particles with less Si/Al were observed with irregular shapes at the nanoscale. The TEM images of the NAu-2 and ATD particles after CP (Figure 1 h) showed some pseudohexagonal nanoparticles with around 5 nm in diameter. Based on the EDX and SAED analysis, these nanoparticles were Fe-rich and the *d*-spacings was at about 1.5-2.5 Å, all of which were identified to be 2-line ferrihydrite (Janney et al., 2000; Shi et al., 2009).

The TEM observation suggested that CP induced the disintegration of mineral particles and thus produced enhanced surface area, resulting in more active sites available on the particle surfaces for SO_2 uptake. Results of TEM also showed that CP influenced the Fe mineralogy, and lead to the Fe-rich nanoparticle formation, which could partly explain the higher SO_2 uptake on the mineral particles after CP.

308 Figure 1

309 **3.2 Effect of simulated CP on heterogeneous transformation of SO**₂

310 The in situ DRIFTS spectra on the IMt-2, NAu-2, SWy-2 and ATD samples before 311 and after CP exposed to SO₂ as a function of time are shown in Figure 2. For the 312 IMt-2 sample before CP (Figure 2a and b), the intensities of the broad peaks from 3600 to 3000 cm⁻¹ and a weak peak at 1650 cm⁻¹ increased with time. The band 313 between 3600 and 3000 cm⁻¹ was attributed to the vibrations of hydrogen-bonded 314 hydroxyl species (Zhao et al., 2015), while the absorption peak at 1650 cm⁻¹ was 315 mainly associated to H₂O produced from the reaction between SO₂ and surface 316 317 hydroxyls (Nanayakkara et al., 2012; Cheng et al., 2016). A weak vibration was observed at around 1100 cm⁻¹, which might be attributed to free sulfate anions on the 318 319 particle surface (Ullerstam et al., 2003; Nanayakkara et al., 2012; Yang et al., 2016). 320 Previous studies established that various types of surface OH groups are the key 321 reactive sites for sulfite/sulfate and bisulfite/bisulfate formation on mineral oxides

322 (Faust et al., 1989; Usher et al., 2003; Ullerstam et al., 2003), because of the 323 complexes formed between sulfite/sulfate species and the surface OH. Generally, the 324 SO₂ adsorption grow in intensity with decreasing OH stretching and H₂O banding 325 (Zhang et al., 2006). However, the OH peaks herein were not observed to decrease 326 with prolonged time, because the losses of H₂O and OH groups on the particle 327 surfaces were replenished by maintaining the constant RH in this study.

328 When the same set of experiments were carried out using the IMt-2 sample after CP 329 (Figure 2b), the intensities of the prominent peaks were significantly higher than those 330 on the IMt-2 sample before CP. Four new bands were readily observed at 1167, 1100, 1088 and 1077 cm⁻¹. The new bands were easily assigned to the stretching motion of 331 332 surface-coordinated sulfate species (1167 cm⁻¹), i.e., bidentate surface sulfate complexes, free sulfate ion (1100 cm⁻¹), and sulfite/bisulfite species (1088 and 1077 333 334 cm⁻¹) (Peak et al., 1999; Ullerstam et al., 2003; Yang et al., 2016). These new bands 335 remained when an argon blow-off process was carried out, suggesting that the surface-adsorbed sulfite/sulfate species between 1250 and 1000 cm⁻¹ was 336 337 chemisorbed (Zhang et al., 2006).

338 Upon adsorption of SO₂ on the surface of the NAu-2 sample before CP (Figure 2c and d), the broad band from 3600 to 2800 cm⁻¹ and the peaks at 1580 and 1675 cm⁻¹ 339 340 increased drastically with time. These absorbance bands were all attributed to the 341 surface hydroxyl species (OH) and H₂O. No peaks were observed over the range of 1000 to 1250 cm⁻¹, suggesting that the sulfite/sulfate products were not formed newly 342 343 on the surface of the NAu-2 sample before CP. Upon adsorption of SO₂ on the surface of the NAu-2 sample after CP (Figure 2d), the new bands at 3661 and 3450 cm⁻¹, the 344 broad band between 3400 and 2700 cm⁻¹, and the broad band centered at 2131 cm⁻¹, 345 were observed as the exposure time increased. In detail, the band at 3661 cm⁻¹ could 346

347 be assigned to stretching vibration modes of isolated or bridged surface hydroxyl groups bonded to the surface iron ions embedded in the octahedral and tetrahedral 348 349 sites (Faust et al., 1989; Nanayakkara et al., 2012; Zhao et al., 2015). The peaks at around 3450 cm⁻¹, 2131 cm⁻¹ and the band between 3400 and 2700 cm⁻¹ were all 350 351 attributed to surface OH groups (Ma et al., 2010; Zhao et al., 2017). These new bands generated on the processed NAu-2 particles suggested that CP changed the location of 352 353 diverse OH groups on the particle surfaces. Over the range of 1250-1000 cm⁻¹, the new bands centered at 1170 cm⁻¹ was assigned to the asymmetric stretching of sulfate 354 355 (Kong et al., 2014; Yang et al., 2015).

356 The spectra of the SWy-2 samples before and after CP (Figure 2e and f) showed a 357 similar spectral character with those of the NAu-2 samples. The spectra for the ATD 358 samples before and after CP (Figure 2g and h) were roughly the same as the ones for 359 IMt-2. All of the results demonstrated that the characteristic peaks for the active OH sites and the sulfite/sulfate products on the mineral particles after CP were 360 361 significantly higher than those on the ones before CP, indicating the higher 362 hygroscopicity and more SO₂ uptake on the particles after CP. The data shown herein 363 confirmed that CP could potentially promote the transformation of SO₂ on the particle 364 surfaces.

Figure 2 365

366 **3.3 Uptake coefficient of SO₂ on the mineral particles before and after CP**

The areas of the bands (from 1250 to 1000 cm⁻¹) attributed to the sulfite/sulfate products as a function of time are shown in Figure 3. It was evident that the peak areas of the products on the mineral particles after CP were generally greater than the ones before CP. The reaction on the sample surfaces was practically saturated to SO_2 uptake within 15 min, except for the NAu-2 and IMt-2 samples after CP. As for all of the sample, the saturation coverages of the sulfite/sulfate products after CP were obviously greater than the corresponding values before CP, suggesting that CP favored the sulfate formation on the mineral surfaces due to improving active site number, as expected previously.

Figure 3 Figure 3

377 The maximum uptake coefficients (γ_{geo} and γ_{BET}) for SO₂ uptake on the samples were estimated on the basis of the sulfate formation rates in the initial 15 min. The 378 values on the mineral samples before and after CP are shown in Table 1. The γ_{geo} 379 380 values of SO₂ on the IMt-2, NAu-2, SWy-2 and ATD samples before CP were 1.03×10^{-7} , 0.30×10^{-7} , 1.72×10^{-7} and 1.37×10^{-7} , respectively, which were in the order 381 of SWy-2, ATD, IMt-2 and NAu-2. The γ_{geo} values of SO_2 on the IMt-2, NAu-2, 382 383 SWy-2 and ATD samples after CP were 4.7, 19.4, 2.7 and 2.0 times higher than the values before CP, respectively, suggesting that the SO₂ uptake on the mineral particles 384 385 significantly increased after CP.

Table 1

 A_{BET} was more appropriate to represent the effective area, because the reactant may 387 diffuse into tiny holes of the entire sample. The γ_{BET} values of SO₂ on the IMt-2, 388 NAu-2, SWy-2 and ATD samples before CP were 2.62×10⁻¹², 0.75×10⁻¹², 3.70×10⁻¹² 389 and 1.61×10⁻¹¹, respectively, which were in the order of ATD, SWy-2, IMt-2 and 390 391 NAu-2. It was noteworthy that the S_{BET} of samples increased after CP, as shown in 392 Table 1. The γ_{BET} values of SO₂ on the IMt-2, NAu-2, SWy-2 and ATD after CP were 393 2.2, 4.1, 1.5 and 1.4 times higher than the values before CP, respectively. The 394 discrepancies in the γ_{BET} value confirmed that the higher sulfate formation rates of the 395 particles after CP was not only due to the increased surface area of the particles, but 396 also resulting from the chemical modification on the particle surfaces.

397 The estimated uptake coefficients were several orders of magnitude lower than the 398 results from Ullerstam et al. (2003) and Usher et al. (2003), which could be partly 399 explained by the difference in the preparation of mineral dust samples, or the 400 difference between diverse experimental structures such as the DRIFTS and Knudsen 401 cell in kinetics discussion. In this study, mineral dust particles were in a highly 402 accumulative state in the sample support of Knudsen cell. The many layers of 403 particles in the latter study will hinder the diffusion of gas into the underlayer 404 particles, resulting in the underestimate of γ_{BET} . However, the values herein were 405 comparable to those obtained by the similar DRIFTS setup (Fu et al., 2007), 406 indicating the reliability of our measurements.

407 In addition, the formation rate of sulfate appeared a linear increasing trend as a 408 function of pH cycles. Specifically, the increasing amount of sulfate ions for the 409 IMt-2, NAu-2, SWy-2 and ATD samples after each pH cycle during CP were 7.0×10^{10} , 1.0×10^{11} , 5.0×10^{10} , 3.0×10^{10} , in the order of NAu-2 > IMt-2 > SWy-2 > ATD (Figure 410 411 4). The multiples factors for γ_{BET} (γ_{geo}) were coincided with the total Fe content of 412 these samples: NAu-2 (26.30%) > IMt-2 (5.45%) > SWy-2 (2.36%) > ATD (1.48%). 413 We thus supposed that the SO₂ uptake on these dust samples was closely related to the 414 Fe hosted in the particles.

415 **Figure 4**

416 **3.4 Fe speciation analysis before and after CP.**

The fractions of "free-Fe" (including Fe_A and Fe_D) and "structural-Fe" in the mineral particles before and after CP were determined by the CBD extraction (Figure 5). In terms of total Fe, the amorphous Fe (Fe_A) (e.g., nanoparticulate and poorly crystalline ferrihydrite) contents of the IMt-2, NAu-2, SWy-2 and ATD samples before CP were 0.7%, 0.5%, 0.7% and 3.8%, respectively. The crystalline Fe 422 (oxyhydr)oxides (Fe_D) (e.g., α-FeOOH and α-Fe₂O₃) contents of the IMt-2, NAu-2,
423 SWy-2 and ATD samples before CP were 7.2%, 2.3%, 4.5% and 35.5%, respectively.

424 As a result, the fractions of "structural-Fe" before CP were 92.1%, 97.2%, 94.8% and

425 60.7%, respectively, for IMt-2, NAu-2, SWy-2 and ATD.

426 After CP, the Fe_A contents of the IMt-2, NAu-2, SWy-2 and ATD samples reached 1.8%, 1.2%, 1.7% and 24.2%, respectively, which increased by 2.6, 2.4, 2.4 and 6.4 427 428 times as compared to the ones before CP. The crystalline Fe (oxyhydr)oxides (Fe_D) 429 contents of the samples after CP were not significantly changed as compared to the 430 ones before CP; whereas the content of "structural-Fe" in the Al-Si crystals of the 431 IMt-2, NAu-2, SWy-2 and ATD samples after CP decreased by various degrees, to 432 91.1%, 96.1%, 93.2% and 42.5%, respectively. Previous research had indicated that 433 Fe_A increased as a result of the simulated CP (Shi et al., 2009). Herein, we further 434 proposed that the increased fractions of Fe_A could be mostly transformed from the 435 "structural-Fe" in the aluminosilicate phase of the particles during CP, which is in 436 good agreement with the TEM observation. For example, the Fe_A in the ATD samples increased from 3.8% to 24.2% after CP, accompanied by a sharp decrease of the 437 438 structural-Fe content from 60.7% to 42.5%.

439 **Figure 5**

The Mössbauer spectra and their fitted results are shown in Figure 6. The corresponding hyperfine parameters estimated from the best fitted spectra are presented in Table S3. The central doublet with isomer shift (IS) of 0.37 mm s⁻¹ and quadrupole shift (QS) of 0.72 mm s⁻¹ were typical for high-spin Fe(III) in octahedral symmetry (Eyre and Dickson, 1995), while the other one with IS of 1.12 mm s⁻¹ and QS of 2.65 mm s⁻¹ was characteristic of high spin Fe(II) (Hofstetter et al., 2003; Kopcewicz et al., 2015). The two doublet components of the IMt-2, NAu-2, SWy-2 447 and ATD samples before CP were all attributed to different fractions of Fe(III) and 448 Fe(II) in the aluminosilicate crystals, respectively. Before CP, the Fe(II) fraction in the 449 IMt-2, NAu-2, SWy-2 and ATD samples were 34.0 %, 12.9 %, 18.3 % and 29.0 %, respectively (Figure 6 a, c, e and g). Furthermore, the spectra of the ATD sample 450 451 before CP showed not only two central quadrupole doublets, but also one MHS sextet with IS of 0.39 mm s $^{-1}$, QS of –0.13 mm s $^{-1}$ and $\rm H_{f}$ of 51.1 T. The MHS sextet, which 452 453 shared 31.8 % of the whole area, could be ascribed to α -Fe₂O₃ (Kopcewicz and 454 Kopcewicz et al., 1991), in agreement with the TEM analysis and "free-Fe" 455 measurement as mentioned previously.

456 After CP, the Fe(II) content of the samples decreased to 31.5 %, 11.6 %, 17.1% and 457 10.9%, respectively, for IMt-2, NAu-2, SWy-2 and ATD (Figure 6 b, d, f and h). It 458 was supposed that the Fe(II) release is more energetically favorable than one of Fe(III) 459 due to the bond strength. As to the ATD sample after CP (Figure 6 h), not only did the 460 Fe(II) fraction decrease from 29.0% to 10.9%, but also the Fe(III) fraction in the aluminosilicates decreased from 39.0% to 33.0%. Meanwhile, the α -Fe₂O₃ fraction 461 was not significantly changed (31.8% vs. 32.3%). As discussed previously, the Fe 462 463 mobilization was dependent on the specific chemical bonds. The Fe_D phase in α-Fe₂O₃ with the strong Fe-O bond was less liable than that embedded in the 464 aluminosilicate lattice (Strehlau et al., 2017). It was well documented that the Fe 465 466 replacing alkaline elements as the interlayer ions was easy to be mobilized than the Fe 467 bound by covalent bonds in the aluminosilicate matrix (Luo et al., 2005; Cwiertny et 468 al., 2008; Journet et al., 2008). Therefore, the Fe in the aluminosilicate fraction of the 469 mineral particles exhibited varied iron solubility.

470 Particularly, a new quadrupole doublet with IS of 0.67 mm s⁻¹ and QS of 1.21 mm 471 s⁻¹ was observed in the spectra of the ATD sample after CP (Figure 6 h), which shared 472 23.8% of the total area, and was possibly indicative of the Fe(III) oxide hybridized in 473 the aluminosilicate matrix (Kopcewicz and Kopcewicz, 1991). The "free-Fe" measurement have indicated that the Fe_A fraction of ATD increased by 20.4% after CP, 474 475 so that this Fe phase was most likely to be amorphous Fe(III) hybridized with Al/Si. 476 In the terms of the other samples after CP, the magnetic signal of the newly formed 477 Fe(III) phase was not detected. It was probably due to the newly formed Fe fractions were not available at sufficiently high level to be clearly resolved by the Mössbauer 478 479 spectroscopy, and/or the slight signal drift and the poor signal to noise ratio made an 480 unambiguous identification difficult. Herein, the newly formed amorphous Fe(III) 481 phase was supposed to be a reactive Fe-bearing component, of which may contribute 482 significantly to the SO₂ uptake even at a low level.

483 Figure 6

484 **3.5** The dissolution-precipitation cycle of the mineral Fe during CP

485 During the simulated CP experiments, the concentrations of total dissolved Fe (Fe_s), 486 dissolved Fe(II) and Fe(III) released from the particles as a function of time are shown 487 in Figure 7. Similar dissolution trends were observed for all of the samples. One can 488 see that the suspensions at pH 2 induced a rapid increase of Fe_s. Once increasing the 489 pH from 2 to 5 resulted in a rapid and almost complete removal of Fes. In fact, only a 490 rather small fraction of the Fe in dusts could be dissolved at pH above 4 (Zuo and 491 Hoigne, 1992). The dissolved Fe precipitated rapidly as insoluble deposit at pH 5. 492 When the suspension pH was again reduced to 2, a steep increase in the Fe_s 493 concentration was measured once again. The fast Fe release was due to the 494 redissolution of the Fe-rich precipitates, which was proposed to be reactive Fe phases 495 (Shi et al., 2009; Shi et al., 2015). Such highly soluble Fe-bearing precipitates have 496 been observed under the TEM, as well as the "free Fe" measurement and Mössbauer

497 characterization.

498 **Figure 7**

499 For each pH cycle during the simulated CP experiment, the overall changes of total 500 released Fe concentrations were reproducible. The Fe ion on the particle surfaces 501 would experience a continuous dissolution-precipitation-redissolution-reprecipitation 502 process when the pH cycles between pH 2 and pH 5 (cloud-aerosol modes). During 503 this process, the Fe(II) fraction would be transformed to Fe(III). The results shown 504 herein suggested that CP could significantly modify Fe partitioning between dissolved 505 and particulate phases in the real atmosphere. Not only did the increase of specific 506 surface area contribute to the enhanced sulfate formation, but also the highly reactive 507 Fe on the particle surfaces yielded during CP were also responsible for the higher SO₂ 508 uptake on the particles after CP.

509 When investigating the NAu-2 sample, once the pH of the clear solution increased 510 from 2 to 5-6, the Fe-bearing nanoparticles separated out from the solution rapidly 511 and precipitate out slowly. It developed an initial yellow color and then an orange 512 colored suspension. The TEM images of the precipitated particles are shown in Figure 513 8. The particles could be categorized into two different types. One type of particle 514 could be characterized as hundreds of nanometers in size, with low Fe but high Si/Al 515 content. The other type displayed particle sizes nearly 1 micrometer, and were Fe-rich 516 but contained a smaller amount of Si/Al components. These bigger particles were 517 ambiguously identified as Na_{0.42}Fe₃Al₆B₃₀₉Si₆O₁₈(OH)_{3.65} (PDF: 89-6506) on the 518 basis of the EDX data and SAED analysis. It is likely that the Al/Si elements also 519 precipitated out along with the Fe.

520 **Figure 8**

21

521 The Mössbauer spectra of the precipitated Fe-rich particles are shown in Figure 9. Two central doublets were distinguished, with one (48.4%) of IS = 0.45 mm s⁻¹, OS = 522 0.75 mm s⁻¹, and the other (51.6%) of IS = 0.24 mm s⁻¹, QS = 0.76 mm s⁻¹. Both of 523 524 the two doublet components could be attributed to the Fe(III) fraction in the 525 aluminosilicates (Kopcewicz et al., 2015). The results were in good agreement with the TEM observation, which showed that most of these Fe particles were mostly 526 527 present as the Fe(III) hybridized with Al/Si. The particle size distributions in the 528 suspensions were also determined by dynamic light scattering, as shown in Figure 10. 529 When pH was lower than 2.0, the particles seemed to stabilize below 10 nm in size. 530 These Fe colloids were thought to be a source of soluble Fe (Janney et al., 2000). 531 Once pH increased, the size of precipitated particles quickly increased, even to 532 micro-scale, and the suspension was featured with a polydispersed size distribution.

533 Conclusively, the precipitated Fe were mainly Fe(III) with weak crystal structure 534 and/or ferrihydrite nanoparticle hybridized with Al/Si, which were possibly 535 transformed from the Fe hosted in the aluminosilicate matrix of the particles. The 536 particle surfaces after CP were coated by these reactive Fe to provide more surface 537 OH species, resulting in enhanced SO₂ uptake.

538 Figure 9

539 **Figure 10**

540 4 Conclusion and implication

Transition metal ions, especially Fe(III), could catalyze SO₂ oxidation rapidly in cloud drops (Harris et al., 2013). This study further confirmed that SO₂ uptake on the mineral particles could be greatly enhanced by CP, possibly more than described previously. The higher uptake coefficient of the particles after CP was not only due to increased surface area, but also resulted from the chemical modification of the particle

surfaces. The "free-Fe" and Mössbauer analysis suggested that CP triggered newly 546 formation of amorphous Fe particles on the surfaces, of which were mostly 547 548 transformed from the Fe hosted in the aluminosilicate matrix. TEM showed that the 549 amorphous Fe(III) and/or ferrihydrite nanoparticle were hybridized with Al/Si. In 550 general, the acidity fluctuation during CP enables the dissolution-precipitation cycles 551 of mineral Fe to yielded more reactive Fe, resulting in more SO₂ uptake on the 552 particle surfaces. More SO₂ adsorption further increases the surface acidity of dust 553 particles, in turn leading to higher Fe solubility; again, more sulfate formation. It was 554 thus proposed that there is a positive feedback relative to SO₂ update and iron 555 mobilized from mineral particles during CP, therefore enhanced sulfate formation 556 greatly.

557 Our results also serve to explain high sulfate loading in fog-haze episodes of China. It has been recommended that sulfate contributed significantly to the explosive 558 559 growth of fine particles, thus exacerbating severe fog-haze development (Kasibhatla 560 et al., 1997; Nie et al., 2014; Barrie et al., 2016). Haze and fog within an episode was 561 often found to transform each other at a short time due to the diurnal variation of RH, 562 whereby the haze-fog transition was probably analogous to the aerosol-cloud 563 interaction. Water content of aerosol or fog drops was regulated by RH, and thus 564 allowed the particle acidity fluctuation. Although the aerosol acidity could not be accurately determined from field measurements or calculated using the 565 566 thermodynamic model, we recognized that the large pH fluctuations between the 567 haze-fog modes could significantly modify the microphysical properties of mineral particles, and triggered formation of reactive Fe particles and thus accelerated sulfate 568 569 formation via a self-amplifying process, contributing to explosive growth of fine 570 particles at the initial stage of fog-haze events. The data presented herein also 571 highlight that CP provide more bioavailable iron from mineral particle than one 572 expected previously, of which is a key speciation to promote oceanic primary 573 productivity. Results of this study could partly explain the missing source of sulfate 574 and improve agreement between models and field observations.

575 Additionally, previous studies indicated that Fe in pyrogenic aerosols was always 576 presented as liable Fe, such as ferric sulfate and aggregated nanocrystals of magnetite 577 (Fe₃O₄) (Fu et al., 2012), and displayed higher Fe solubility compared to dust 578 (Desboeufs et al., 2005; Sedwick et al., 2007; Ito et al., 2019). Alexander et al. 579 demonstrated that the sulfate formed through metal catalysis was highest over the 580 polluted industrial regions of northern Eurasia, suggesting that the increasing 581 importance of the metal-catalyzed S(IV) oxidation pathway due to anthropogenic emissions (Alexander et al, 2009). With the rapid development of industry and 582 agriculture, the pyrogenic Fe-containing aerosols are indispensable contributors to the 583 584 atmospheric Fe load in China. Thus, the acidic solution at pH 2 and high sulfate 585 loading of fine particles in severe fog-haze events of China might be more relevant to 586 Fe-containing combustion aerosols than mineral dust. Based on the current findings, 587 not only the potential influences of cloud liquid water content, light, and organic 588 ligands, but also the solubility and speciation of Fe in pyrogenic aerosols will be 589 considered during the simulated CP experiments in the future. A more detailed 590 understanding of the iron-sulfur cycle during CP is therefore critical to estimate 591 accurately the contribution of CP to global sulfate loading and its impact on the 592 climate.

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Competing interests. The authors declare no conflict of interest.

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Captions of Figures and Tables

Figure 1. Comparison of morphologies and chemical properties for samples collected before and after CP using TEM. The dotted circles indicate the positions of the electron beam for the HRTEM images and SAED patterns. Elements of the detected parts of individual particles are also presented. Square brackets indicate mass percent of iron. The iron species were identified by the Miller indices and the SAED patterns. (a) IMt-2 particles characterized by high fractions of Al and Si, along with other crustal elements including Mg, K and Fe. (b) IMt-2 particles after CP were almost all less than 1µm in size. Some Fe-rich particles with less Si and Al were observed on nanoscale dimension. (c) NAu-2 particles with high Fe/Si-ratios, contain Mg, Al and Ca elements. (d) NAu-2 particles after CP were much smaller than the ones before CP. Some ferrihydrite clusters were observed and were attached on the surface of the NAu-2 particles after CP. (e) Typical SWy-2 particles were Al/Si-rich, containing Fe, Mg and Ca elements. (f) TEM images of the SWy-2 particles after CP appeared smaller than the particles before CP. (g) The Si/Al-rich crystal in ATD particles was aluminosilicate with low content of Fe, and a typical of the α -Fe₂O₃ particles (PDF: 33-664) was found to attach onto the aluminosilicate surface. (h) The pseudohexagonal nanoparticles were observed to on the surface of α-Fe₂O₃ crystal among the ATD particles. The SAED lattice constant of these nanoparticles were found to be very close to that of 2-line ferrihydrite.

Figure 2. Comparison of the DRIFT spectra of mineral dust samples upon exposure to SO₂ for 45 min before and after CP. Data for IMt-2 (a and b), NAu-2 (c and d), SWy-2 (e and f), ATD (g and h), are shown in the ranges of 4000 to 1250 cm⁻¹ and 1250 to 1000 cm⁻¹, respectively.

Figure 3. Comparison of the integrated areas on DRIFTS spectra in the range of $1250-1000 \text{ cm}^{-1}$ for the sulfate species formed on the samples before and after CP.

Table 1. Sulfate formation rates and uptake coefficients of SO₂ on particle samples before and after CP.

Figure 4. Comparison of the sulfate formation rates as a function of pH cycle.

Figure 5. The fractions of "free-Fe" (Fe_A and Fe_D) and "structural-Fe" were measured by the chemical CBD extractions for the samples before and after CP. Results are present as relative percentage of Fe_T .

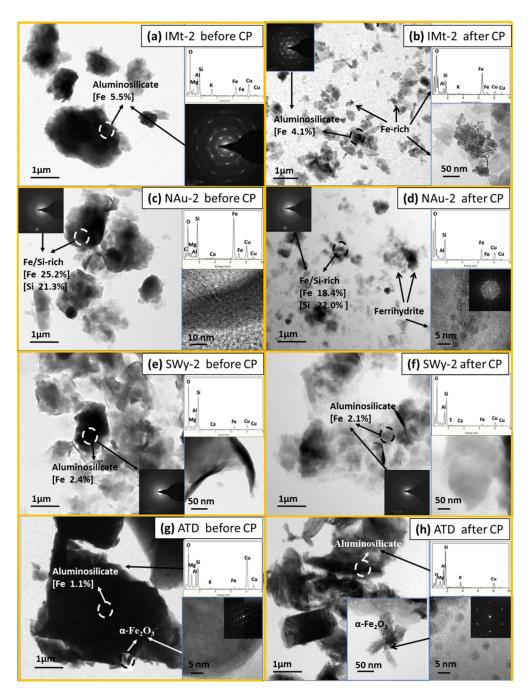
Figure 6. Mössbauer spectroscopy measured for samples. IMt-2 before and after CP (a and b), NAu-2 before and after CP (c and d), SWy-2 before and after CP (e and f), ATD before and after CP (g and h). Experimental data were fit using a least-squares fitting-program. The IS values were relative to α -Fe at RT. Prominent spectral features associated with different iron species are indicated.

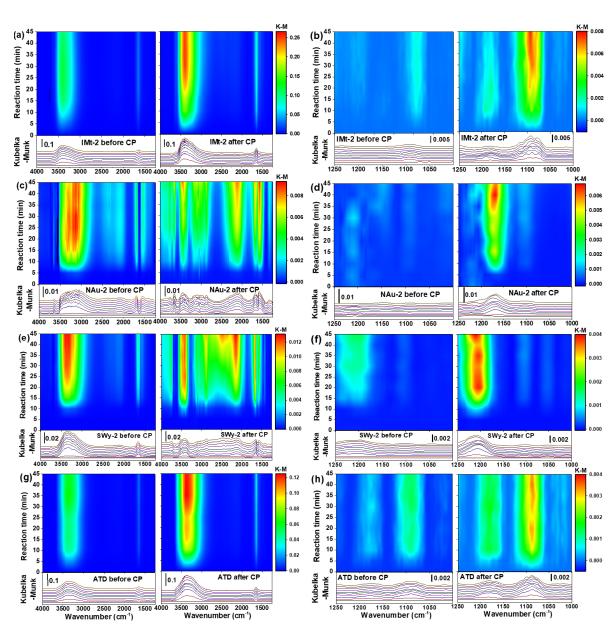
Figure 7. The concentrations of Fe_s , dissolved Fe(II) and Fe(III) in the suspensions measured over 144 h in the solution cycled between pH 2 and pH 5 for IMt-2 (a), NAu-2 (b), SWy-2 (c) and ATD (d), respectively.

Figure 8. TEM images of the newly formed particles in the precipitation experiment. Based on the TEM-EDX measurement and SAED analysis, these particles could be categorized into two different types, which were circled in Figure 8 a. The typical sizes of the first type were hundreds of nanometers. The enlarged images are displayed in Figure 8 b, c and d. The insert EDX data and SAED image confirmed that they were poor crystalline aluminosilicate with low Fe but high Si/Al content. The second type (Figure 8 e, f and g) were Fe-rich but with less amount of Si/Al, which were nearly 1 micrometer in size. Based on the EDX data and the SAED analysis, these bigger particles were ambiguously identified as Na_{0.42}Fe₃Al₆B₃₀₉Si₆O₁₈(OH)_{3.65} (PDF: 89-6506).

Figure 9. Mössbauer spectroscopy measured at RT for the neo-formed particles collected in the precipitation experiment.

Figure 10. During the precipitation experiment, the particle size distributions in the suspensions were determined by dynamic light scattering. The presented size distributions are characteristic of neo-formed nanoparticles or microparticles as the suspension pH raised from 1.0 to 3.8.





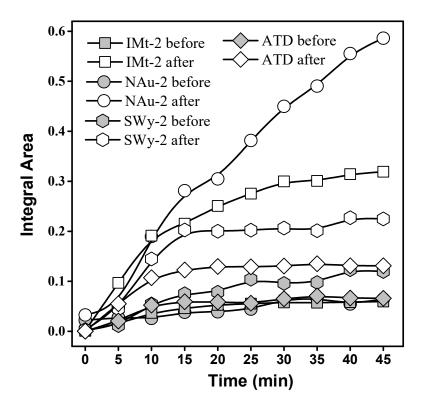
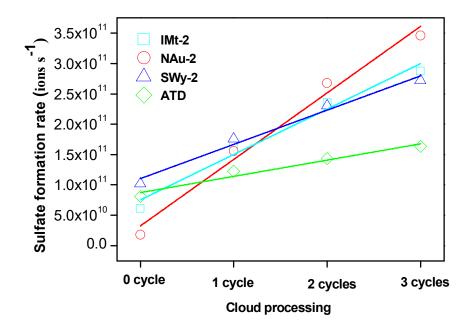
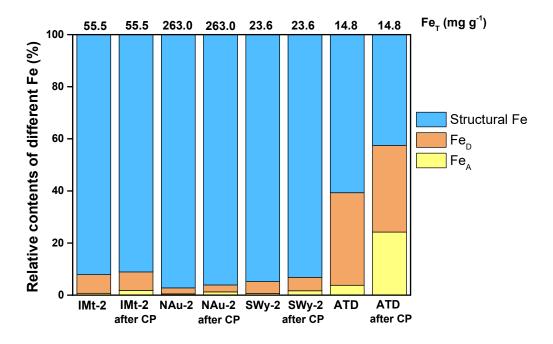


Table	1
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Samples	$A_{\rm BET}$ (m ²)	Sulfate formation rate (ions s ⁻¹) ($\times 10^{10}$)	$A_{\text{geometric}}$ (m ²) (×10 ⁻⁵)	γ _{BET} (×10 ⁻¹²)	$\gamma_{\text{geometric}}$ ($\times 10^{-7}$)
IMt-2 before CP	0.770	6.13	1.95	2.62	1.03
IMt-2 after CP	1.640	28.72	1.95	5.76	4.85
NAu-2 before CP	0.790	1.80	1.95	0.75	0.30
NAu-2 after CP	3.749	34.57	1.95	3.06	5.83
SWy-2 before CP	0.906	10.20	1.95	3.70	1.72
SWy-2 after CP	1.631	27.19	1.95	5.49	4.59
ATD before CP	0.166	8.11	1.95	16.05	1.37
ATD after CP	0.241	16.33	1.95	22.33	2.76





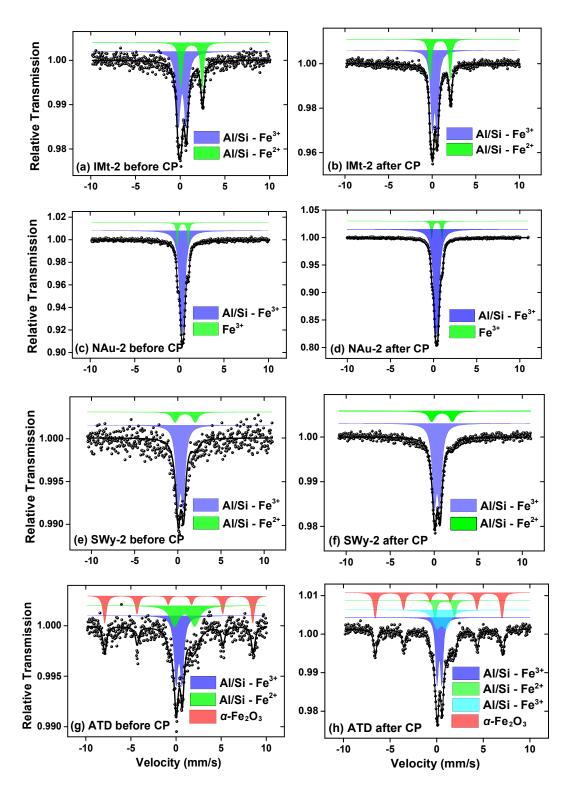
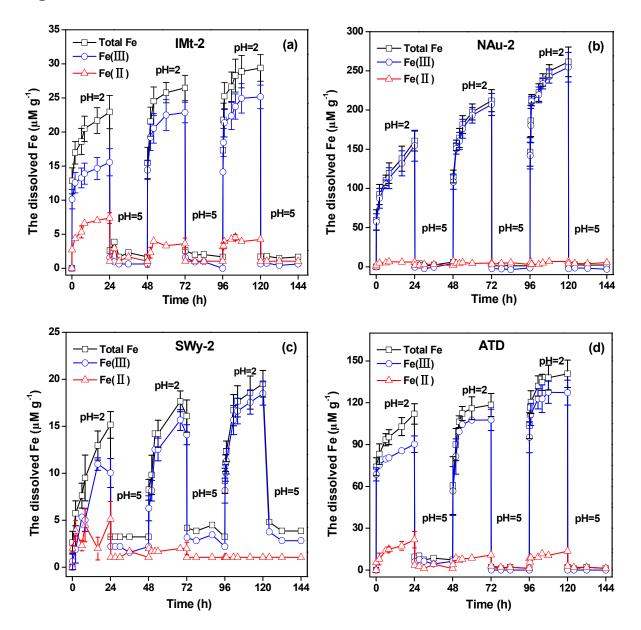


Figure 7



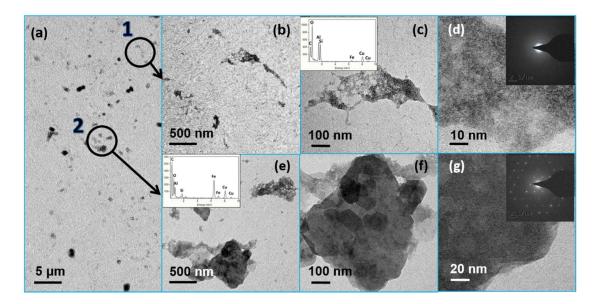


Figure 9

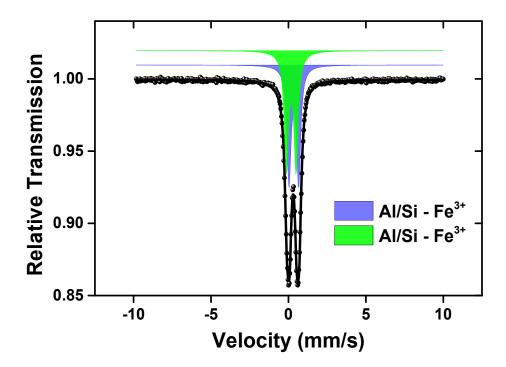


Figure 10

