- 1 Enhanced heterogeneous uptake of sulfur dioxide on
- 2 mineral particles through modification of iron speciation
- 3 during simulated cloud processing
- 4 Zhenzhen Wang<sup>1</sup>, Tao Wang<sup>1</sup>, Hongbo Fu<sup>1, 2, 3</sup>, Liwu Zhang<sup>1</sup>, Mingjin Tang<sup>4</sup>,
- 5 Christian George<sup>5</sup>, Vicki H. Grassian<sup>6</sup>, and Jianmin Chen<sup>1</sup>
- 6 <sup>1</sup>Shanghai Key Laboratory of Atmospheric Particle Pollution and Prevention,
- 7 Department of Environmental Science & Engineering. Institute of Atmospheric
- 8 Sciences, Fudan University, Shanghai, 200433, China
- 9 <sup>2</sup>Shanghai Institute of Pollution Control and Ecological Security, Shanghai 200092,
- 10 China
- 11 <sup>3</sup>Collaborative Innovation Center of Atmospheric Environment and Equipment
- 12 Technology (CICAEET), Nanjing University of Information Science and Technology,
- Nanjing 210044, China
- 14 <sup>4</sup>State Key Laboratory of Organic Geochemistry and Guangdong Key Laboratory of
- 15 Environmental Protection and Resources Utilization, Guangzhou Institute of
- 16 Geochemistry, Chinese Academy of Sciences, Guangzhou 510640, China
- <sup>5</sup>University of Lyon, Université Claude Bernard Lyon 1, CNRS, IRCELYON,
- 18 F-69626, Villeurbanne, France
- 19 <sup>6</sup>Departments of Chemistry and Biochemistry, University of California, San Diego, La
- 20 Jolla, California 92093, United States
- 21 Correspondence to: Hongbo Fu (fuhb@fudan.edu.cn); Jianmin Chen
- 22 (jmchen@fudan.edu.cn)

**Abstract.** Iron-containing mineral aerosols play a key role in the oxidation of sulfur species in the atmosphere. Simulated cloud processing (CP) of typical mineral particles, such as illite (IMt-2), nontronite (NAu-2), smectite (SWy-2) and Arizona test dust (ATD) is shown here to modify sulfur dioxide (SO<sub>2</sub>) uptake onto mineral surfaces. Heterogeneous oxidation of SO<sub>2</sub> on particle surfaces was firstly investigated using an in situ DRIFTS apparatus. Our results showed that the BET surface area normalized uptake coefficients (y<sub>BET</sub>) of SO<sub>2</sub> on the IMt-2, NAu-2, SWy-2 and ATD samples after CP were 2.2, 4.1, 1.5 and 1.4 times higher than the corresponding ones before CP, respectively. The DRIFTS results suggested that CP increased the amounts of reactive sites (e.g., surface OH groups) on the particle surfaces and thus enhanced the uptake of SO<sub>2</sub>. TEM showed that the particles broke up into smaller pieces after CP, and thus produced more active sites. The "free-Fe" measurements confirmed that more reactive Fe species were present after CP, which could enhance the SO<sub>2</sub> uptake more effectively. Mössbauer spectroscopy further revealed that the formed Fe phase were amorphous Fe(III) and nanosized ferrihydrite hybridized with Al/Si, which were possibly transformed from the Fe in the aluminosilicate lattice. The modification of Fe speciation was driven by the pH-dependent fluctuation coupling with Fe dissolution-precipitation repeatedly during the experiment. Considering both the enhanced SO<sub>2</sub> uptake and subsequent promotion of iron dissolution along with more active Fe formation, which in turn lead to more SO<sub>2</sub> uptake, it was proposed that there may be a positive feedback between SO<sub>2</sub> uptake and iron mobilized on particle surfaces during CP, thereby affecting climate and biogeochemical cycles. This self-amplifying mechanism generated on the particle surfaces may also serve as the basis of high sulfate loading in severe fog-haze events observed recently in China.

24

25

26

27

28

29

30

31

32

33

34

35

36

37

38

39

40

41

42

43

44

45

46

#### 1 Introduction

49

50

51

52

53

54

55

56

57

58

59

60

61

62

63

64

65

66

67

68

69

70

71

72

73

Mineral dust is a major fraction of global atmospheric aerosol budget, with an estimated annual emission flux of 1000 to 3000 Tg into the atmosphere (Jickells et al., 2005; Andreae and Rosenfeld, 2008). Mineral dust aerosol mainly consists of quartz, feldspars, carbonates (calcite, dolomite), and clay minerals (illite, kaolinite, chlorite, montmorillonite), the exact composition varies with source (Claquin et al., 1999; Formenti et al., 2008; Journet et al., 2008). A long range transport would result in a decrease of quartz relative to the clay fraction because of the more rapid removal of quartz, hence clay is an important component of mineral dusts (Mahowald et al., 2005; Journet et al., 2008). During the long-range transport, mineral dust provides a reactive surface for heterogeneous chemistry (Zhang et al., 2006; George et al., 2015; Huang et al., 2015). Heterogeneous reactions of atmospheric trace gases on mineral dust particles are of great significance as these reactions alter the chemical balance of the atmosphere and modify the properties of individual particles (Usher et al., 2003; Wu et al., 2011; Huang et al., 2015). SO<sub>2</sub> is an important trace gas, which is released mainly by fossil fuel combustion and volcanic emission. The heterogeneous conversion of SO<sub>2</sub> on mineral dust surfaces leads to the formation of sulfuric acid and sulfate aerosols, resulting in a significant cooling effect on the global climate by scattering solar radiation and acting as cloud condensation nuclei (CCN) to affect climate indirectly (Lelieveld and Heintzenberg, 1992; Usher et al., 2003; Kolb et al., 2010). In addition, sulfate containing particles play a significant role in the haze formation in China in recent years (Sun et al., 2014; Wang et al., 2014; Yang et al., 2017). SO<sub>2</sub> can be gaseous oxidized to sulfate by OH radical, and be aqueous oxidation in cloud and fog droplets by ozone and hydrogen peroxide (Luria and Sievering, 1991), or through heterogeneous processes that occur

on aerosol particle surfaces (Usher et al., 2003; Ullerstam et al., 2003). However, the high sulfate levels measured in recent field observations cannot be explained by current atmospheric models (Kerminen et al., 2000; Wang et al., 2003; Cheng et al., 2016), leading to a large gap between the modeled and field-observed sulfate concentrations using known oxidation pathways (Herman, 1991; Kasibhatla et al., 1997; Barrie et al., 2016). Overall, on a global scale, atmospheric SO<sub>2</sub> concentration were typically overestimated, while sulfate tended to be underestimated, suggesting missing sulfate production pathways (Harris et al., 2013; Kong et al., 2014). It has been suggested that the heterogeneous conversion of SO<sub>2</sub> could make an important contribution to the atmospheric sulfate loading. Laboratory studies typically focus on SO<sub>2</sub> uptake onto a variety of metal oxides and mineral particles (Goodman et al., 2001; Usher et al., 2002; Zhao et al., 2015; Yang et al., 2016), and have confirmed that its conversion rate on the surface of Fe (hydr)oxides was faster compared to other metal oxides investigated, in good agreement with the field-measurement (Usher et al., 2002; Zhang et al., 2006). Atmospheric Fe is emitted from both anthropogenic (primarily biomass burning, coal and oil combustion) and natural (mineral dust and volcanic ash) sources, with the mineral dust source dominant globally (Siefert et al., 1998; Luo et al., 2008; Ito et al., 2019). It has been established that an important in cloud S (IV) oxidation pathway is catalyzed by natural transition metal ions, especially Fe hosted within mineral particles (Alexander et al., 2009; Harris et al., 2013). Another important consideration for heterogeneous chemistry of mineral dust aerosol, is how mineral dust particles change in the atmosphere. During long-range transport, mineral particles often undergo chemical ageing by atmospheric processes (Mahowald et al., 2005; Baker and Croot, 2010; Shi et al., 2011). Cloud processing

74

75

76

77

78

79

80

81

82

83

84

85

86

87

88

89

90

91

92

93

94

95

96

97

involves cloud water condensation and evaporation on the particle surfaces, along with drastic liquid water content and pH fluctuations (Mackie, 2005; Shi et al., 2011; Rubasinghege et al., 2016). During CP, the high relative humidity (RH) results in high aerosol water content and relatively high pH (Behra et al., 1989; Baker and Croot, 2010; Shi et al., 2011). While water evaporation from cloud droplets to wet aerosol at higher temperature, the particles only contain a concentrated aqueous aerosol solution, in which the pH can be lower than 2 (Zhu et al., 1993; Meskhidze, 2003; Shi et al., 2015). Therefore, there is a highly acidic film (e.g., pH = 2) in the "wet aerosol" phase versus a less acidic droplet (near-neutral, 5-6) in the "cloud droplet" phase within clouds (Shi et al., 2015). During its lifetime, a typical aerosol particle may experience several cloud cycles involving large pH variations before being removed from the atmosphere as rain or through dry deposition (Pruppacher and Jaenicke, 1995; Maters et al., 2016). Herein, the simulated CP experiment was conducted by changing pH between 2 and 5-6, in accordance with the previous studies (Spokes et al., 1994; Mackie, 2005; Shi et al., 2009). It was well documented that pH is especially important for Fe mobilization (Zhu et al., 1993; Desboeufs et al., 2001; Deguillaume et al., 2010; Maters et al., 2016). The fluctuating pH during CP will impact and change the Fe speciation and morphology in dust particles (Zhuang et al., 1992; Wurzler et al., 2000; Shi et al., 2009; Kadar et al., 2014). The low pH will increase Fe solubility and bioavailability of dust during transport, thereby providing Fe external input to the open ocean surface to promote marine prime productivity (Spokes et al., 1994; Desboeufs et al., 2001). It has been found that Fe-rich nanoparticle aggregates were formed from Saharan soil and goethite upon simulated CP conditions, in good agreement with their field-measurements from the wet-deposited Saharan dusts collected from the western

99

100

101

102

103

104

105

106

107

108

109

110

111

112

113

114

115

116

117

118

119

120

121

122

Mediterranean (Shi et al., 2009). Fe nanoparticles are more chemically reactive (Wurzler et al., 2000; Desboeufs et al., 2001), possibly lead to a remarkable difference in heterogeneous chemistry. However, little is known about the influence of CP on SO<sub>2</sub> uptake onto particle surfaces up to now.

In this study, we employed four typical Fe-containing mineral samples as

surrogates to perform simulated CP experiments. The SO<sub>2</sub> uptakes on the mineral particles before and after CP were compared using *in situ* diffuse reflectance infrared Fourier transform spectroscopy (DRIFTS). Transmission electron microscopy (TEM) was applied to observe the morphological and mineralogical change of mineral particles. The Fe speciation modification during simulated CP was further monitored by the dissolved Fe measurement, the "free-Fe" analysis and Mössbauer spectroscopic characterization.

#### 2 Material and methods

#### 2.1 Mineral particles

The standard mineral samples of IMt-2, NAu-2 and SWy-2 were purchased from the Source Clay Minerals Repository (Purdue University, West Lafayette, IN). ATD was purchased from Powder Technology Inc. (Burnsville, MN, USA). The mineral samples were coarsely ground using a mortar and pestle before being more finely ground using an All-dimensional Planetary Ball Mill QM-QX (Nanjing University Instrument Plant) and were sieved to particle diameters ( $D_p$ ) < 45  $\mu$ m prior to analysis. The Brunauer-Emmett-Teller specific surface areas ( $S_{BET}$ ) of the samples were measured with a Quantachrome Nova 1200 BET apparatus. Total iron content (Fe<sub>T</sub>) of the samples were determined using an inductively coupled plasma atomic emission spectroscopy (ICP-AES, Jobin Yvon Ultima). The chemical compositions of the particles were analyzed by X-ray fluorescence spectrometry (XRF, PANalytical Axios

149 Advanced).

## 2.2 Cloud processing simulation experiment

The simulated CP experiments were conducted at a constant temperature (298  $\pm$  1K) using a Pyrex glass vessel with a water jacket. The suspensions contained a mineral particle loading of 1 g L<sup>-1</sup> were subjected to acidic (pH = 2  $\pm$  0.1, 24 h) and near-neutral pH (pH = 5-6, 24 h) cycles for 1-3 times according to the previous methods (Spokes et al., 1994; Mackie, 2005; Shi et al., 2009). Suspension pH was adjusted by adding dilute H<sub>2</sub>SO<sub>4</sub> or NH<sub>4</sub>OH. The CaCO<sub>3</sub> equivalent alkalinity of the dust was determined in accordance with APHA method 2320B so that acid additions to control pH could be adjusted accordingly (Mackie, 2005). The amount of acid or alkali added to achieve these pH cycles was less than 1% of the total volume of the suspensions. The experiments were performed under a constant stirring (about 50 rpm) in the dark for 144 h. At the end of the CP experiment, the suspensions were filtered through 0.2  $\mu$ m PTFE filters (Millipore). The filter residue was air-dried, and was further applied to the DRIFTS experiment, as well as TEM observation, "free-Fe" measurement and Mössbauer spectroscopic characterization.

#### 2.3 SO<sub>2</sub> uptake on the mineral particles

The SO<sub>2</sub> uptake on the particle surfaces before and after CP was investigated by a Shimadzu Tracer-100 FTIR spectrometer equipped with a high-sensitivity mercury cadmium telluride (MCT) detector and a diffuse reflectance accessory. A temperature controller was fitted to the DRIFTS chamber to ensure constant reaction temperature (298 K). Weighted sample was placed into a ceramic crucible (0.35 mm depth, 5 mm i. d.) in the chamber. Mass flow controllers (Beijing Sevenstar electronics Co., LTD) were used to adjust the reactant gases to a flux with expected concentration and relative humidity. The sample was firstly pretreated in a 100 mL min<sup>-1</sup> flow of

synthetic air (21% O<sub>2</sub> and 79% N<sub>2</sub>) for 1 h to blow off water and impurities on particle surface. When the background spectrum of the fresh sample reached steady state, the reactant gas of SO<sub>2</sub> (5.0 ppm) along with synthetic air was introduced into the chamber at a total flow rate of 120 mL min<sup>-1</sup> for 45 min, during which the IR spectrum was recorded automatically every 5 min at a resolution of 4 cm<sup>-1</sup> for 100 scans in the spectral range of 900 to 4000 cm<sup>-1</sup>. Atmospheric moisture was simulated with a RH level around 40 % by guiding one high-pure air flux through water. The humidity value was monitored using a hygrometer.

The sulfate products were analyzed by ion chromatography (IC) after the DRIFTS experiments. The particles were extracted with 5 ml ultrapure water by ultrasonic extractor. After 10 min, the extracted solution was passed through a 0.22 μm PTFE membrane filter and the leaching solution was analyzed using a Metrohm 883 Basic IC equipped with an A5-250 column. A weak base eluent (3.2 mmol L<sup>-1</sup> Na<sub>2</sub>CO<sub>3</sub> plus 1.0 mmol L<sup>-1</sup> NaHCO<sub>3</sub>) was used for anion detection at a flow rate of 0.70 ml min<sup>-1</sup>. To discriminate the adsorbed sulfate during simulated CP experiment and the sulfate ions generated from the heterogeneous reaction, the adsorbed sulfate on the particles during simulated CP experiment were initially measured as blank. The heterogeneous uptake of SO<sub>2</sub> was calculated by subtracting the blank value from the total sulfate ions.

The reactive uptake coefficient ( $\gamma$ ) was defined as the rate of sulfate formation on the surface (d[SO<sub>4</sub><sup>2-</sup>]/dt, ions s<sup>-1</sup>) divided by collision frequency (Z, molecules s<sup>-1</sup>) (Usher et al., 2003; Ullerstam et al., 2003; Kong et al., 2014; Huang et al., 2015).

$$\gamma = \frac{d[SO_4^{2-}]/dt}{Z},\tag{1}$$

$$Z = \frac{1}{4} \times A_s \times [SO_2] \times v, \tag{2}$$

$$v = \sqrt{\frac{8RT}{\pi M_{SO_2}}},\tag{3}$$

Where,  $A_s$  is the effective sample surface of the samples,  $m^2$ ; v is the mean molecular velocity of SO<sub>2</sub>, m s<sup>-1</sup>; R is the gas constant, J mol K<sup>-1</sup>; T is the absolute temperature, K; and  $M_{SO_2}$  is the molecular weight of SO<sub>2</sub>, kg mol<sup>-1</sup>.

A conversion factor was obtained by a calibration plot with numbers of  $SO_4^{2^-}$  analyzed by ion chromatography (IC, Metrohm 883 Basic, Switzerland) versus the integrated areas of sulfate products from DRIFTS spectra. The residual sulfate during simulated CP experiments were deducted as background. The calculated conversion factor of  $SO_4^{2^-}$  is  $1.170 \times 10^{15}$  (ions · integrated units<sup>-1</sup>). Integrated areas for the total sulfur-containing products were calculated to show the maximal sulfate formation rates. The reactive uptake coefficient for  $SO_2$  was determined to be  $\gamma_{BET}$  and  $\gamma_{geo}$  using the BET area ( $A_{BET} = \text{mass} \times S_{BET}$ ) and geometric area ( $A_{geo} = \text{mass} \times S_{geo}$ ) as the reactive area, respectively.

## 2.4 Morphological and mineralogical characterization of the Fe speciation

A FEI TECNAI G2 S-TWIN F20 TEM equipped with an Oxford energy-dispersive X-ray spectrometer (EDX) was used to analyze the morphological and chemical composition of individual particles before and after CP. Suspensions (0.2 g L<sup>-1</sup>) of each particle were prepared in methanol and sonicated for at least 1 h. A drop of this suspension was then applied to a carbon-coated Cu TEM grid (400 mesh; EMS) and allowed to air-dry. The operation was conducted in bright field mode at 120 kV. The Fe content of the typical individual mineral particle were calculated from the values of 50 typical particles. To obviously observe the morphological changes, high-resolution TEM (HRTEM) images were also collected to observe nanoscale structural features, e.g., surface roughness and lattice fringes.

The content of "free-Fe" in the mineral particles was determined by a citrate-buffered-dithionite (CBD) sequential Fe extractions method according to the literature (Lafon et al., 2004; Shi et al., 2009). Simply, 30 mg of the dust samples were treated for 24 h with a 10 mL ascorbate solution (pH = 7.5) to extract chemically highly labile Fe phases (FeA), mainly composed of amorphous, nanoparticle and/or poorly crystalline ferrihydrite. The solutions were filtered through 0.2 µm polycarbonate filters. The dust particles collected on the filters were subsequently treated for 2 h with a 10 mL sodium dithionite solution (pH = 4.8) to extract crystalline Fe (oxyhydr) oxides (Fe<sub>D</sub>), which are mainly goethite and hematite. After each reaction step, the dissolved Fe concentrations (FeA and FeD) in the filtrates were determined using ICP-AES. The sum of these two pools (Fe<sub>A</sub> + Fe<sub>D</sub>) was defined as the "free-Fe" fraction (Shi et al., 2011). The other fraction was donated as the "structural-Fe" in aluminosilicate crystals, which could be calculated from the difference between the Fe<sub>T</sub> and "free-Fe" fractions (Lafon et al., 2004). The Mössbauer spectroscopic analysis performed in transmission geometry with a constant acceleration was used to inspect the chemical valence and the surrounding structure of Fe in the particles before and after CP. <sup>57</sup>Co was used as the Mössbauer source, and a 1 mm thick Na(TI) scintillator coupled to a EMI9750B photoelectric multiplier was used as the detector (Cwiertny et al., 2008). The measurement was carried out at room temperature (RT) with a duration of 24 hours for one sample (around 1.5×10<sup>6</sup> counts per channel). Experimental data were fitted by a least-squares fitting-program. The isomer shift values were calibrated against a spectrum for  $\alpha$ -Fe metal foil. During the simulated CP experiment, the total dissolved iron (Fe<sub>s</sub>) and the dissolved Fe(II) in the suspensions were measured colorimetrically by the Ferrozine method, as

222

223

224

225

226

227

228

229

230

231

232

233

234

235

236

237

238

239

240

241

242

243

244

245

described in previous studies (Viollier et al., 2000; Cwiertny et al., 2008). For Fe(II) analysis, 200 mL of a 5 mM 1, 10-phenanthroline solution and 200 mL of an ammonium acetate buffer were added into 1 mL of sample. To avoid possible interference from Fe(III), which can also form a complex with 1,10-phenanthroline when present at high concentrations, 50 mL of 0.43 M ammonium fluoride was added to the sample prior to 1,10-phenanthroline. The mixture was allowed to sit in the dark for 30 min prior to ultraviolet-visible spectroscopy (UV-Vis) analysis, during which time a reddish-orange color developed if Fe(II) was present. Fe<sub>s</sub> was determined via the same protocol, except that 20 mL of 1.5 M hydroquinone, which reduces Fe(III) to Fe(II), was added to the sample rather than ammonium fluoride. Absorbance measured at 510 nm was converted to concentrations using aqueous standards prepared from anhydrous beads of ferrous chloride. Standards were prepared in each acid used in dissolution studies, and no matrix effects were observed. These conditions resulted in a detection limit of 1  $\mu$ M. The concentration of dissolved Fe(III) was calculated from the difference in experimentally measured concentrations of total dissolved iron and dissolved Fe(II). Additionally, the dissolved Fe(III) could precipitate out as the pH increased, and then the Fe mineraology of the deposit was also observed. NAu-2 released about 300 µM of dissolved Fe at pH 2. The dissolving solution (200 mL) was sampled after filtration (0.2 µm polycarbonate filter). The clear solution was subjected to changing acidity from pH 2 to 5 by the stepwise addition of dilute NH<sub>4</sub>OH. The precipitated particles were separated out by 0.2 µm filters and were used in TEM and Mössbauer analysis. Size distributions for the Fe-bearing particles formed in the suspensions were determined by a Horiba LB-500 light scattering microscopy within the size range of 3-6000 nm.

247

248

249

250

251

252

253

254

255

256

257

258

259

260

261

262

263

264

265

266

267

268

269

270

#### 272 3 Results and discussion

273

#### 3.1 Morphological change of the mineral particles after CP.

274 The characteristic results are shown in Table S1 and Table S2. The samples exhibited  $S_{\rm BET}$  in the range from  $4.3\pm0.3$  to  $22.6\pm2.3$  m<sup>2</sup>/g. The Fe<sub>T</sub> content were 275 276  $5.45 \pm 0.34\%$ ,  $26.30 \pm 0.57\%$ ,  $2.36 \pm 0.56\%$  and  $1.48 \pm 0.56\%$ , for IMt-2, NAu-2, 277 SWy-2 and ATD, respectively. The proportions of Fe<sub>2</sub>O<sub>3</sub> in IMt-2, NAu-2, SWy-2 278 and ATD were 7.95%, 39.03%, 5.55% and 2.57%, respectively. 279 Figure 1 shows the TEM images of the mineral particles before and after CP. As 280 shown in Figure 1 a, c, e and g, the IMt-2, NAu-2, SWy-2 and ATD samples before 281 CP primarily consisted of laminar aluminosilicate with irregular shape and rough 282 morphologies mainly at the micrometer scale, all of which were characterized by various fractions of Fe (1.5%-26.2%), along with minor Mg (0.1%-16.5%), K 283 284 (0.0%-7.8%) and Ca (0.0%-1.1%). The Fe within the aluminosilicates of the particles 285 was evenly distributed. Besides, some Fe-rich crystal with several hundreds of 286 nanometers in size were found to attach onto the ATD particles, which were identified 287 as α-Fe<sub>2</sub>O<sub>3</sub> (PDF: 33-664) from the typical d-spacing analysis of HRTEM (Janney et al., 2000). 288 289 After the simulated CP, all of the processed mineral particles showed much smaller 290 size than the ones before CP. For example, the typical IMt-2 and NAu-2 particles after 291 CP (Figure 1 b and d) were  $\leq 1 \mu m$  in size. Under the TEM, the average Fe content of 292 the individual IMt-2 and SWy-2 particles (Figure 1 b and f) decreased from 5.5% (± 293 1.9%; n = 50) to 4.1% (± 1.6%; n = 50) and from 2.4% (± 0.6%; n = 50) to 2.1% (± 294 0.5%; n = 50), respectively. In addition, the IMt-2 particles after CP showed a 295 heterogeneous distribution of the Fe on the basis of the EDX data. Most of the 296 aluminosilicate in IMt-2 after CP hosted lower Fe content (4.1%), whereas a few of the Fe-rich particles with less Si/Al were observed with irregular shapes at the nanoscale. The TEM images of the NAu-2 and ATD particles after CP (Figure 1 h) showed some pseudohexagonal nanoparticles with around 5 nm in diameter. Based on the EDX and SAED analysis, these nanoparticles were Fe-rich and the *d*-spacings was at about 1.5-2.5 Å, all of which were identified to be 2-line ferrihydrite (Janney et al., 2000; Shi et al., 2009).

The TEM observation suggested that CP induced the disintegration of mineral particles and thus produced enhanced surface area, resulting in more active sites available on the particle surfaces for SO<sub>2</sub> uptake. Results of TEM also showed that CP influenced the Fe mineralogy, and lead to the Fe-rich nanoparticle formation, which could partly explain the higher SO<sub>2</sub> uptake on the mineral particles after CP.

## Figure 1

## 3.2 Effect of simulated CP on heterogeneous transformation of SO<sub>2</sub>

The *in situ* DRIFTS spectra on the IMt-2, NAu-2, SWy-2 and ATD samples before and after CP exposed to SO<sub>2</sub> as a function of time are shown in Figure 2. For the IMt-2 sample before CP (Figure 2a and b), the intensities of the broad peaks from 3600 to 3000 cm<sup>-1</sup> and a weak peak at 1650 cm<sup>-1</sup> increased with time. The band between 3600 and 3000 cm<sup>-1</sup> was attributed to the vibrations of hydrogen-bonded hydroxyl species (Zhao et al., 2015), while the absorption peak at 1650 cm<sup>-1</sup> was mainly associated to H<sub>2</sub>O produced from the reaction between SO<sub>2</sub> and surface hydroxyls (Nanayakkara et al., 2012; Cheng et al., 2016). A weak vibration was observed at around 1100 cm<sup>-1</sup>, which might be attributed to free sulfate anions on the particle surface (Ullerstam et al., 2003; Nanayakkara et al., 2012; Yang et al., 2016). Previous studies established that various types of surface OH groups are the key reactive sites for sulfite/sulfate and bisulfite/bisulfate formation on mineral oxides

(Faust et al., 1989; Usher et al., 2003; Ullerstam et al., 2003), because of the complexes formed between sulfite/sulfate species and the surface OH. Generally, the SO<sub>2</sub> adsorption grow in intensity with decreasing OH stretching and H<sub>2</sub>O banding (Zhang et al., 2006). However, the OH peaks herein were not observed to decrease with prolonged time, because the losses of H<sub>2</sub>O and OH groups on the particle surfaces were replenished by maintaining the constant RH in this study. When the same set of experiments were carried out using the IMt-2 sample after CP (Figure 2b), the intensities of the prominent peaks were significantly higher than those on the IMt-2 sample before CP. Four new bands were readily observed at 1167, 1100, 1088 and 1077 cm<sup>-1</sup>. The new bands were easily assigned to the stretching motion of surface-coordinated sulfate species (1167 cm<sup>-1</sup>), i.e., bidentate surface sulfate complexes, free sulfate ion (1100 cm<sup>-1</sup>), and sulfite/bisulfite species (1088 and 1077 cm<sup>-1</sup>) (Peak et al., 1999; Ullerstam et al., 2003; Yang et al., 2016). These new bands remained when an argon blow-off process was carried out, suggesting that the surface-adsorbed sulfite/sulfate species between 1250 and 1000 cm<sup>-1</sup> was chemisorbed (Zhang et al., 2006). Upon adsorption of SO<sub>2</sub> on the surface of the NAu-2 sample before CP (Figure 2c and d), the broad band from 3600 to 2800 cm<sup>-1</sup> and the peaks at 1580 and 1675 cm<sup>-1</sup> increased drastically with time. These absorbance bands were all attributed to the surface hydroxyl species (OH) and H<sub>2</sub>O. No peaks were observed over the range of 1000 to 1250 cm<sup>-1</sup>, suggesting that the sulfite/sulfate products were not formed newly on the surface of the NAu-2 sample before CP. Upon adsorption of SO<sub>2</sub> on the surface of the NAu-2 sample after CP (Figure 2d), the new bands at 3661 and 3450 cm<sup>-1</sup>, the broad band between 3400 and 2700 cm<sup>-1</sup>, and the broad band centered at 2131 cm<sup>-1</sup>, were observed as the exposure time increased. In detail, the band at 3661 cm<sup>-1</sup> could

322

323

324

325

326

327

328

329

330

331

332

333

334

335

336

337

338

339

340

341

342

343

344

345

be assigned to stretching vibration modes of isolated or bridged surface hydroxyl groups bonded to the surface iron ions embedded in the octahedral and tetrahedral sites (Faust et al., 1989; Nanayakkara et al., 2012; Zhao et al., 2015). The peaks at around 3450 cm<sup>-1</sup>, 2131 cm<sup>-1</sup> and the band between 3400 and 2700 cm<sup>-1</sup> were all attributed to surface OH groups (Ma et al., 2010; Zhao et al., 2017). These new bands generated on the processed NAu-2 particles suggested that CP changed the location of diverse OH groups on the particle surfaces. Over the range of 1250-1000 cm<sup>-1</sup>, the new bands centered at 1170 cm<sup>-1</sup> was assigned to the asymmetric stretching of sulfate (Kong et al., 2014; Yang et al., 2015).

The spectra of the SWy-2 samples before and after CP (Figure 2e and f) showed a similar spectral character with those of the NAu-2 samples. The spectra for the ATD samples before and after CP (Figure 2g and h) were roughly the same as the ones for IMt-2. All of the results demonstrated that the characteristic peaks for the active OH sites and the sulfite/sulfate products on the mineral particles after CP were significantly higher than those on the ones before CP, indicating the higher hygroscopicity and more SO<sub>2</sub> uptake on the particles after CP. The data shown herein confirmed that CP could potentially promote the transformation of SO<sub>2</sub> on the particle surfaces.

#### Figure 2

## 3.3 Uptake coefficient of SO<sub>2</sub> on the mineral particles before and after CP

The areas of the bands (from 1250 to 1000 cm<sup>-1</sup>) attributed to the sulfite/sulfate products as a function of time are shown in Figure 3. It was evident that the peak areas of the products on the mineral particles after CP were generally greater than the ones before CP. The reaction on the sample surfaces was practically saturated to SO<sub>2</sub> uptake within 15 min, except for the NAu-2 and IMt-2 samples after CP. As for all of

the sample, the saturation coverages of the sulfite/sulfate products after CP were obviously greater than the corresponding values before CP, suggesting that CP favored the sulfate formation on the mineral surfaces due to improving active site number, as expected previously.

## Figure 3

The maximum uptake coefficients ( $\gamma_{geo}$  and  $\gamma_{BET}$ ) for SO<sub>2</sub> uptake on the samples were estimated on the basis of the sulfate formation rates in the initial 15 min. The values on the mineral samples before and after CP are shown in Table 1. The  $\gamma_{geo}$  values of SO<sub>2</sub> on the IMt-2, NAu-2, SWy-2 and ATD samples before CP were  $1.03\times10^{-7}$ ,  $0.30\times10^{-7}$ ,  $1.72\times10^{-7}$  and  $1.37\times10^{-7}$ , respectively, which were in the order of SWy-2, ATD, IMt-2 and NAu-2. The  $\gamma_{geo}$  values of SO<sub>2</sub> on the IMt-2, NAu-2, SWy-2 and ATD samples after CP were 4.7, 19.4, 2.7 and 2.0 times higher than the values before CP, respectively, suggesting that the SO<sub>2</sub> uptake on the mineral particles significantly increased after CP.

#### Table 1

 $A_{\rm BET}$  was more appropriate to represent the effective area, because the reactant may diffuse into tiny holes of the entire sample. The  $\gamma_{\rm BET}$  values of SO<sub>2</sub> on the IMt-2, NAu-2, SWy-2 and ATD samples before CP were  $2.62\times10^{-12}$ ,  $0.75\times10^{-12}$ ,  $3.70\times10^{-12}$  and  $1.61\times10^{-11}$ , respectively, which were in the order of ATD, SWy-2, IMt-2 and NAu-2. It was noteworthy that the  $S_{\rm BET}$  of samples increased after CP, as shown in Table 1. The  $\gamma_{\rm BET}$  values of SO<sub>2</sub> on the IMt-2, NAu-2, SWy-2 and ATD after CP were 2.2, 4.1, 1.5 and 1.4 times higher than the values before CP, respectively. The discrepancies in the  $\gamma_{\rm BET}$  value confirmed that the higher sulfate formation rates of the particles after CP was not only due to the increased surface area of the particles, but also resulting from the chemical modification on the particle surfaces.

The estimated uptake coefficients were several orders of magnitude lower than the results from Ullerstam et al. (2003) and Usher et al. (2003), which could be partly explained by the difference in the preparation of mineral dust samples, or the difference between diverse experimental structures such as the DRIFTS and Knudsen cell in kinetics discussion. In this study, mineral dust particles were in a highly accumulative state in the sample support of Knudsen cell. The many layers of particles in the latter study will hinder the diffusion of gas into the underlayer particles, resulting in the underestimate of  $\gamma_{BET}$ . However, the values herein were comparable to those obtained by the similar DRIFTS setup (Fu et al., 2007), indicating the reliability of our measurements.

- In addition, the formation rate of sulfate appeared a linear increasing trend as a function of pH cycles. Specifically, the increasing amount of sulfate ions for the
- IMt-2, NAu-2, SWy-2 and ATD samples after each pH cycle during CP were 7.0×10<sup>10</sup>,
- $1.0 \times 10^{11}$ ,  $5.0 \times 10^{10}$ ,  $3.0 \times 10^{10}$ , in the order of NAu-2 > IMt-2 > SWy-2 > ATD (Figure
- 411 4). The multiples factors for  $\gamma_{\rm BET}$  ( $\gamma_{\rm geo}$ ) were coincided with the total Fe content of
- 412 these samples: NAu-2 (26.30%) > IMt-2 (5.45%) > SWy-2 (2.36%) > ATD (1.48%).
- We thus supposed that the SO<sub>2</sub> uptake on these dust samples was closely related to the
- 414 Fe hosted in the particles.

#### **Figure 4**

## 3.4 Fe speciation analysis before and after CP.

The fractions of "free-Fe" (including Fe<sub>A</sub> and Fe<sub>D</sub>) and "structural-Fe" in the mineral particles before and after CP were determined by the CBD extraction (Figure 5). In terms of total Fe, the amorphous Fe (Fe<sub>A</sub>) (e.g., nanoparticulate and poorly crystalline ferrihydrite) contents of the IMt-2, NAu-2, SWy-2 and ATD samples before CP were 0.7%, 0.5%, 0.7% and 3.8%, respectively. The crystalline Fe

- 422 (oxyhydr)oxides (Fe<sub>D</sub>) (e.g., α-FeOOH and α-Fe<sub>2</sub>O<sub>3</sub>) contents of the IMt-2, NAu-2,
- 423 SWy-2 and ATD samples before CP were 7.2%, 2.3%, 4.5% and 35.5%, respectively.
- As a result, the fractions of "structural-Fe" before CP were 92.1%, 97.2%, 94.8% and
- 425 60.7%, respectively, for IMt-2, NAu-2, SWy-2 and ATD.
- After CP, the Fe<sub>A</sub> contents of the IMt-2, NAu-2, SWy-2 and ATD samples reached
- 427 1.8%, 1.2%, 1.7% and 24.2%, respectively, which increased by 2.6, 2.4, 2.4 and 6.4
- 428 times as compared to the ones before CP. The crystalline Fe (oxyhydr)oxides (Fe<sub>D</sub>)
- contents of the samples after CP were not significantly changed as compared to the
- ones before CP; whereas the content of "structural-Fe" in the Al-Si crystals of the
- 431 IMt-2, NAu-2, SWy-2 and ATD samples after CP decreased by various degrees, to
- 432 91.1%, 96.1%, 93.2% and 42.5%, respectively. Previous research had indicated that
- 433 FeA increased as a result of the simulated CP (Shi et al., 2009). Herein, we further
- proposed that the increased fractions of Fe<sub>A</sub> could be mostly transformed from the
- "structural-Fe" in the aluminosilicate phase of the particles during CP, which is in
- good agreement with the TEM observation. For example, the Fe<sub>A</sub> in the ATD samples
- increased from 3.8% to 24.2% after CP, accompanied by a sharp decrease of the
- 438 structural-Fe content from 60.7% to 42.5%.

## 439 **Figure 5**

- The Mössbauer spectra and their fitted results are shown in Figure 6. The
- 441 corresponding hyperfine parameters estimated from the best fitted spectra are
- presented in Table S3. The central doublet with isomer shift (IS) of 0.37 mm s<sup>-1</sup> and
- quadrupole shift (QS) of 0.72 mm s<sup>-1</sup> were typical for high-spin Fe(III) in octahedral
- symmetry (Eyre and Dickson, 1995), while the other one with IS of 1.12 mm s<sup>-1</sup> and
- QS of 2.65 mm s<sup>-1</sup> was characteristic of high spin Fe(II) (Hofstetter et al., 2003;
- Kopcewicz et al., 2015). The two doublet components of the IMt-2, NAu-2, SWy-2

and ATD samples before CP were all attributed to different fractions of Fe(III) and Fe(II) in the aluminosilicate crystals, respectively. Before CP, the Fe(II) fraction in the IMt-2, NAu-2, SWy-2 and ATD samples were 34.0 %, 12.9 %, 18.3 % and 29.0 %, respectively (Figure 6 a, c, e and g). Furthermore, the spectra of the ATD sample before CP showed not only two central quadrupole doublets, but also one MHS sextet with IS of 0.39 mm  $s^{-1},\,QS$  of -0.13 mm  $s^{-1}$  and  $H_{\rm f}$  of 51.1 T. The MHS sextet, which shared 31.8 % of the whole area, could be ascribed to α-Fe<sub>2</sub>O<sub>3</sub> (Kopcewicz and Kopcewicz et al., 1991), in agreement with the TEM analysis and "free-Fe" measurement as mentioned previously. After CP, the Fe(II) content of the samples decreased to 31.5 %, 11.6 %, 17.1% and 10.9%, respectively, for IMt-2, NAu-2, SWy-2 and ATD (Figure 6 b, d, f and h). It was supposed that the Fe(II) release is more energetically favorable than one of Fe(III) due to the bond strength. As to the ATD sample after CP (Figure 6 h), not only did the Fe(II) fraction decrease from 29.0% to 10.9%, but also the Fe(III) fraction in the aluminosilicates decreased from 39.0% to 33.0%. Meanwhile, the α-Fe<sub>2</sub>O<sub>3</sub> fraction was not significantly changed (31.8% vs. 32.3%). As discussed previously, the Fe mobilization was dependent on the specific chemical bonds. The FeD phase in α-Fe<sub>2</sub>O<sub>3</sub> with the strong Fe-O bond was less liable than that embedded in the aluminosilicate lattice (Strehlau et al., 2017). It was well documented that the Fe replacing alkaline elements as the interlayer ions was easy to be mobilized than the Fe bound by covalent bonds in the aluminosilicate matrix (Luo et al., 2005; Cwiertny et al., 2008; Journet et al., 2008). Therefore, the Fe in the aluminosilicate fraction of the mineral particles exhibited varied iron solubility. Particularly, a new quadrupole doublet with IS of 0.67 mm s<sup>-1</sup> and QS of 1.21 mm

447

448

449

450

451

452

453

454

455

456

457

458

459

460

461

462

463

464

465

466

467

468

469

470

471

s<sup>-1</sup> was observed in the spectra of the ATD sample after CP (Figure 6 h), which shared

23.8% of the total area, and was possibly indicative of the Fe(III) oxide hybridized in the aluminosilicate matrix (Kopcewicz and Kopcewicz, 1991). The "free-Fe" measurement have indicated that the Fe<sub>A</sub> fraction of ATD increased by 20.4% after CP, so that this Fe phase was most likely to be amorphous Fe(III) hybridized with Al/Si. In the terms of the other samples after CP, the magnetic signal of the newly formed Fe(III) phase was not detected. It was probably due to the newly formed Fe fractions were not available at sufficiently high level to be clearly resolved by the Mössbauer spectroscopy, and/or the slight signal drift and the poor signal to noise ratio made an unambiguous identification difficult. Herein, the newly formed amorphous Fe(III) phase was supposed to be a reactive Fe-bearing component, of which may contribute significantly to the SO<sub>2</sub> uptake even at a low level.

## Figure 6

## 3.5 The dissolution-precipitation cycle of the mineral Fe during CP

During the simulated CP experiments, the concentrations of total dissolved Fe (Fe<sub>s</sub>), dissolved Fe(II) and Fe(III) released from the particles as a function of time are shown in Figure 7. Similar dissolution trends were observed for all of the samples. One can see that the suspensions at pH 2 induced a rapid increase of Fe<sub>s</sub>. Once increasing the pH from 2 to 5 resulted in a rapid and almost complete removal of Fe<sub>s</sub>. In fact, only a rather small fraction of the Fe in dusts could be dissolved at pH above 4 (Zuo and Hoigne, 1992). The dissolved Fe precipitated rapidly as insoluble deposit at pH 5. When the suspension pH was again reduced to 2, a steep increase in the Fe<sub>s</sub> concentration was measured once again. The fast Fe release was due to the redissolution of the Fe-rich precipitates, which was proposed to be reactive Fe phases (Shi et al., 2009; Shi et al., 2015). Such highly soluble Fe-bearing precipitates have been observed under the TEM, as well as the "free Fe" measurement and Mössbauer

497 characterization.

#### Figure 7

498

499

500

501

502

503

504

505

506

507

508

509

510

511

512

513

514

515

516

517

518

519

520

For each pH cycle during the simulated CP experiment, the overall changes of total released Fe concentrations were reproducible. The Fe ion on the particle surfaces would experience a continuous dissolution-precipitation-redissolution-reprecipitation process when the pH cycles between pH 2 and pH 5 (cloud-aerosol modes). During this process, the Fe(II) fraction would be transformed to Fe(III). The results shown herein suggested that CP could significantly modify Fe partitioning between dissolved and particulate phases in the real atmosphere. Not only did the increase of specific surface area contribute to the enhanced sulfate formation, but also the highly reactive Fe on the particle surfaces yielded during CP were also responsible for the higher SO<sub>2</sub> uptake on the particles after CP. When investigating the NAu-2 sample, once the pH of the clear solution increased from 2 to 5-6, the Fe-bearing nanoparticles separated out from the solution rapidly and precipitate out slowly. It developed an initial yellow color and then an orange colored suspension. The TEM images of the precipitated particles are shown in Figure 8. The particles could be categorized into two different types. One type of particle could be characterized as hundreds of nanometers in size, with low Fe but high Si/Al content. The other type displayed particle sizes nearly 1 micrometer, and were Fe-rich but contained a smaller amount of Si/Al components. These bigger particles were ambiguously identified as Na<sub>0.42</sub>Fe<sub>3</sub>Al<sub>6</sub>B<sub>309</sub>Si<sub>6</sub>O<sub>18</sub>(OH)<sub>3.65</sub> (PDF: 89-6506) on the basis of the EDX data and SAED analysis. It is likely that the Al/Si elements also precipitated out along with the Fe.

#### Figure 8

The Mössbauer spectra of the precipitated Fe-rich particles are shown in Figure 9. Two central doublets were distinguished, with one (48.4%) of IS = 0.45 mm s<sup>-1</sup>, QS =  $0.75 \text{ mm s}^{-1}$ , and the other (51.6%) of IS = 0.24 mm s<sup>-1</sup>, QS = 0.76 mm s<sup>-1</sup>. Both of the two doublet components could be attributed to the Fe(III) fraction in the aluminosilicates (Kopcewicz et al., 2015). The results were in good agreement with the TEM observation, which showed that most of these Fe particles were mostly present as the Fe(III) hybridized with Al/Si. The particle size distributions in the suspensions were also determined by dynamic light scattering, as shown in Figure 10. When pH was lower than 2.0, the particles seemed to stabilize below 10 nm in size. These Fe colloids were thought to be a source of soluble Fe (Janney et al., 2000). Once pH increased, the size of precipitated particles quickly increased, even to micro-scale, and the suspension was featured with a polydispersed size distribution. Conclusively, the precipitated Fe were mainly Fe(III) with weak crystal structure and/or ferrihydrite nanoparticle hybridized with Al/Si, which were possibly transformed from the Fe hosted in the aluminosilicate matrix of the particles. The particle surfaces after CP were coated by these reactive Fe to provide more surface

## Figure 9

521

522

523

524

525

526

527

528

529

530

531

532

533

534

535

536

537

538

540

541

542

543

544

545

#### **Figure 10**

## 4 Conclusion and implication

OH species, resulting in enhanced SO<sub>2</sub> uptake.

Transition metal ions, especially Fe(III), could catalyze SO<sub>2</sub> oxidation rapidly in cloud drops (Harris et al., 2013). This study further confirmed that SO<sub>2</sub> uptake on the mineral particles could be greatly enhanced by CP, possibly more than described previously. The higher uptake coefficient of the particles after CP was not only due to increased surface area, but also resulted from the chemical modification of the particle

surfaces. The "free-Fe" and Mössbauer analysis suggested that CP triggered newly formation of amorphous Fe particles on the surfaces, of which were mostly transformed from the Fe hosted in the aluminosilicate matrix. TEM showed that the amorphous Fe(III) and/or ferrihydrite nanoparticle were hybridized with Al/Si. In general, the acidity fluctuation during CP enables the dissolution-precipitation cycles of mineral Fe to yielded more reactive Fe, resulting in more SO<sub>2</sub> uptake on the particle surfaces. More SO<sub>2</sub> adsorption further increases the surface acidity of dust particles, in turn leading to higher Fe solubility; again, more sulfate formation. It was thus proposed that there is a positive feedback relative to SO<sub>2</sub> update and iron mobilized from mineral particles during CP, therefore enhanced sulfate formation greatly.

Our results also serve to explain high sulfate loading in fog-haze episodes of China.

It has been recommended that sulfate contributed significantly to the explosive growth of fine particles, thus exacerbating severe fog-haze development (Kasibhatla et al., 1997; Nie et al., 2014; Barrie et al., 2016). Haze and fog within an episode was often found to transform each other at a short time due to the diurnal variation of RH, whereby the haze-fog transition was probably analogous to the aerosol-cloud interaction. Water content of aerosol or fog drops was regulated by RH, and thus allowed the particle acidity fluctuation. Although the aerosol acidity could not be accurately determined from field measurements or calculated using the thermodynamic model, we recognized that the large pH fluctuations between the haze-fog modes could significantly modify the microphysical properties of mineral particles, and triggered formation of reactive Fe particles and thus accelerated sulfate formation via a self-amplifying process, contributing to explosive growth of fine particles at the initial stage of fog-haze events. The data presented herein also

highlight that CP provide more bioavailable iron from mineral particle than one expected previously, of which is a key speciation to promote oceanic primary productivity. Results of this study could partly explain the missing source of sulfate and improve agreement between models and field observations.

571

572

573

574

575

576

577

578

579

580

581

582

583

584

585

586

587

588

589

590

591

592

Additionally, previous studies indicated that Fe in pyrogenic aerosols was always presented as liable Fe, such as ferric sulfate and aggregated nanocrystals of magnetite (Fe<sub>3</sub>O<sub>4</sub>) (Fu et al., 2012), and displayed higher Fe solubility compared to dust (Desboeufs et al., 2005; Sedwick et al., 2007; Ito et al., 2019). Alexander et al. demonstrated that the sulfate formed through metal catalysis was highest over the polluted industrial regions of northern Eurasia, suggesting that the increasing importance of the metal-catalyzed S(IV) oxidation pathway due to anthropogenic emissions (Alexander et al, 2009). With the rapid development of industry and agriculture, the pyrogenic Fe-containing aerosols are indispensable contributors to the atmospheric Fe load in China. Thus, the acidic solution at pH 2 and high sulfate loading of fine particles in severe fog-haze events of China might be more relevant to Fe-containing combustion aerosols than mineral dust. Based on the current findings, not only the potential influences of cloud liquid water content, light, and organic ligands, but also the solubility and speciation of Fe in pyrogenic aerosols will be considered during the simulated CP experiments in the future. A more detailed understanding of the iron-sulfur cycle during CP is therefore critical to estimate accurately the contribution of CP to global sulfate loading and its impact on the climate.

Author contributions. Z.W., H.F. and J.C. designed the experiments, Z.W., T.W., H.F. and L.Z. performed the laboratory experiments. H.F., J.C., L.Z. and V.G. contributed reagents/analytic tools. C.G., V.G. and M.T. gave some valuable suggestions in designing the experiments. Z.W., T.W. and H.F. analyzed data. Z.W. and H.F. wrote the manuscript, with inputs from all coauthors.

Competing interests. The authors declare no conflict of interest.

Acknowledgments. This work was supported by National Natural Science Foundation of China (Nos. 91744205, 21777025, 21577022, 21177026), National Key R&D Program of China (2016YFC0202700), and Opening Project of Shanghai Key Laboratory of Atmospheric Particle Pollution and Prevention.

#### References

Alexander, B., Park, R. J., Jacob, D. J., and Gong, S.: Transition metal-catalyzed oxidation of atmospheric sulfur: Global implications for the sulfur budget, J. Geophys. Res., 114, D02309, 2009.

Andreae, M. O., and Rosenfeld, D.: Aerosol-cloud-precipitation interactions. Part 1. The nature and sources of cloud-active aerosols, Earth Sci. Rev., 89, 13-41, 2008.

Baker, A. R., and Croot, P. L.: Atmospheric and marine controls on aerosol iron solubility in seawater, Mar. Chem., 120, 4-13, 2010.

Barrie, L. A., Yi, Y., Leaitch, W. R., Lohmann, U., Kasibhatla, P., Roelofs, G. J., Wilson, J.,
McGovern, F., Benkovitz, C., Méliéres, M. A., Law, K., Prospero, J., Kritz, M., Bergmann,
D., Bridgeman, C., Chin, M., Christensen, J., Easter, R., Feichter, J., Land, C., Jeuken, A.,

- Kjellström, E., Koch, D., and Rasch, P.: A comparison of large-scale atmospheric sulphate aerosol models (COSAM): Overview and highlights, Tellus B, 53, 615-645, 2016.
- Behra, P., Sigg, L., and Stumm, W.: Dominating influence of NH<sub>3</sub> on the oxidation of aqueous SO<sub>2</sub>: the coupling of NH<sub>3</sub> and SO<sub>2</sub> in atmospheric water, Atmos. Environ., 23, 2691-2707, 1989.
- Cheng, Y. F., Zheng, G. J., Wei, C., Mu, Q., Zheng, B., Wang, Z. B., Gao, M., Zhang, Q., He,K. B., Carmichael, G., Poschl, U., and Su, H.: Reactive nitrogen chemistry in aerosol wateras a source of sulfate during haze events in China, Sci. Adv., 2, 1-11, 2016.
- Claquin, T., Schulz, M., and Balkanski, Y. J.: Modeling the mineralogy of atmospheric dust sources, J. Geophys. Res.-Atmos., 104, 22243-22256, 1999.
- Cwiertny, D. M., Baltrusaitis, J., Hunter, G. J., Laskin, A., Scherer, M. M., and Grassian, V. H.: Characterization and acid-mobilization study of iron-containing mineral dust source materials, J. Geophys. Res.-Atmos., 113, D05202, 2008.
- Deguillaume, L., Desboeufs, K. V., Leriche, M., Long, Y., and Chaumerliac, N.: Effect of iron dissolution on cloud chemistry: from laboratory measurements to model results, Atmos. Pollut. Res., 1, 220-228, 2010.
- Desboeufs, K. V., Losno, R., and Colin, J. L.: Factors influencing aerosol solubility during cloud processes, Atmos. Environ., 35, 3529-3537, 2001.
- Desboeufs, K. V., Sofikitis, A., Losno, R., Colin, J. L., and Ausset, P.: Dissolution and solubility of trace metals from natural and anthropogenic aerosol particulate matter, Chemosphere, 58, 195-203, 2005.

- Eyre, J. K., and Dickson, D. P. E.: Mössbauer spectroscopy analysis of iron-containing minerals in the Chinese loess, J. Geophys. Res., 100, 17925-17930, 1995.
- Faust, B. C., Hoffmann, M. R., and Bahnemann, D. W.: Photocatalytic oxidation of sulfur dioxide in aqueous suspensions of α-Fe<sub>2</sub>O<sub>3</sub>, J. Phys. Chem., 93, 6371-6381, 1989.
- Fu, H. B., Wang, X., Wu, H. B., Yin, Y., and Chen, J. M.: Heterogeneous uptake and oxidation of SO<sub>2</sub> on iron oxides, J. Phys. Chem. C, 111, 6077-6085, 2007.
- Fu, H. B., Lin, J., Shang, G. F., Dong, W. B., Grassian, V. H., Carmichael, G. R., Li, Y., and Chen, J. M.: Solubility of iron from combustion source particles in acidic media linked to iron speciation, Environ. Sci. Technol., 46, 11119-11127, 2012.
- Formenti, P., Rajot, J. L., Desboeufs, K., Caquineau, S., Chevaillier, S., Nava, S., Gaudichet, A., Journet, E., Triquet, S., Alfaro, S., Chiari, M., Haywood, J., Coe, H., and Highwood, E.: Regional variability of the composition of mineral dust from western Africa: Results from the AMMA SOP0/DABEX and DODO field campaigns, J. Geophys. Res., 113, D00C13, 2008.
- George, C., Ammann, M., D'Anna, B., Donaldson D. J., and Nizkorodov, S. A.: Heterogeneous Photochemistry in the Atmosphere, Chem. Rev., 115, 4218-4258, 2015.
- Goodman, A. L., Li, P., Usher, C. R., and Grassian, V. H.: Heterogeneous uptake of sulfur dioxide on aluminum and magnesium oxide particles, J. Phys. Chem. A, 105, 6109-6120, 2001.
- Harris, E., Sinha, B., Foley, S., Crowley, J. N., Borrmann, S., and Hoppe, P.: Sulfur isotope fractionation during heterogeneous oxidation of SO<sub>2</sub> on mineral dust, Atmos. Chem. Phys.,

- 12, 4867-4884, 2012.
- Harris, E., Sinha, B., van Pinxteren, D., Tilgner, A., Fomba, K. W., Schneider, J., Roth, A., Gnauk, T., Fahlbusch, B., Mertes, S., Lee, T., Collett, J., Foley, S., Borrmann, S., Hoppe, P., and Herrmann, H.: Enhanced role of transition metal ion catalysis during in-cloud oxidation of SO<sub>2</sub>, Science, 340, 727-730, 2013.
- Herman, L. M. S.: Heterogeneous and homogeneous oxidation of SO<sub>2</sub> in the remote marine atmosphere, Atmos. Environ., 25, 1489-1496, 1991.
- Huang, L., Zhao, Y., Li, H., and Chen, Z.: Kinetics of heterogeneous reaction of sulfur dioxide on authentic mineral dust: Effects of relative humidity and hydrogen peroxide, Environ. Sci. Technol., 49, 10797-17805, 2015.
- Hofstetter, T. B., Schwarzenbach, R. P., and Haderlein, S. B.: Reactivity of Fe(II) species associated with clay minerals, Environ. Sci. Technol., 37, 519-528, 2003.
- Ito, A., Myriokefalitakis, S., Kanakidou, M., Mahowald, N., Scanza, R., Hamilton, D., Baker, A., Jickells, T., Sarin, M., Bikkina, S., Gao, Y., Shelley, R., Buck, C., Landing, W., Bowie, A., Perron, M., Guieu, C., Meskhidze, N., Johnson, M., Feng, Y., Kok, J., Nenes, A. and Duce, R.: Pyrogenic iron: The missing link to high iron solubility in aerosols, Sci. Adv., 5(5), eaau7671, 2019.
- Janney, D. E., Cowley, J. M., and Buseck, P. R.: Transmission electron microscopy of synthetic 2-and 6-line ferrihydrite, Clay Clay Miner., 48, 111-119, 2000.
- Jickells, T. D., An, Z. S., Andersen, K. K., Baker, A. R., Bergametti, G., Brooks, N., Cao, J. J., Boyd, P. W., Duce, R. A., Hunter, K. A., Kawahata, H., Kubilay, N., laRoche, J., Liss, P. S.,

- Mahowald, N., Prospero, J. M., Ridgwell, A. J., Tegen, I., and Torres, R.: Global iron connections between desert dust, ocean biogeochemistry, and climate, Science, 308, 67-71, 2005.
- Journet, E., Desboeufs, K. V., Caquineau, S., and Colin, J.-L.: Mineralogy as a critical factor of dust iron solubility, Geophys. Res. Lett., 35, L07805, 2008.
- Kadar, E., Fisher, A., Stolpe, B., Calabrese, S., Lead, J., Valsami-Jones, E., and Shi, Z.: Colloidal stability of nanoparticles derived from simulated cloud-processed mineral dusts, Sci. Total. Environ., 466-467, 864-870, 2014.
- Kasibhatla, P., Chameides, W. L., and John, J. S.: A three-dimensional global model investigation of seasonal variations in the atmospheric burden of anthropogenic sulfate aerosols, J. Geophys. Res.-Atmos., 102, 3737-3759, 1997.
- Kerminen, V. M., Pirjola, L., Boy, M., Eskola, A., Teinila, K., Laakso, L., Asmi, A., Hienola, J., Lauri, A., Vainio, V., Lehtinen, K., and Kulmala, M.: Interaction between SO<sub>2</sub> and submicron atmospheric particles, Atmos. Res., 54, 41-57, 2000.
- Kolb, C. E., Cox, R. A., Abbatt, J. P. D., Ammann, M., Davis, E. J., Donaldson, D. J., Garrett,
  B. C., George, C., Griffiths, P. T., Hanson, D. R., Kulmala, M., McFiggans, G., Pöschl, U.,
  Riipinen, I., Rossi, M. J., Rudich, Y., Wagner, P. E., Winkler, P. M., Worsnop, D. R., and O'
  Dowd, C. D.: An overview of current issues in the uptake of atmospheric trace gases by
  aerosols and clouds, Atmos. Chem. Phys., 10, 10561-10605, 2010.
- Kong, L. D., Zhao, X., Sun, Z. Y., Yang, Y. W., Fu, H. B., Zhang, S. C., Cheng, T. T., Yang, X., Wang, L., and Chen, J. M.: The effects of nitrate on the heterogeneous uptake of sulfur

- dioxide on hematite, Atmos. Chem. Phys., 14, 9451-9467, 2014.
- Kopcewicz, B., and Kopcewicz, M.: Mössbauer study of iron-containing atmospheric aerosols, Struct. Chem., 2, 303-312, 1991.
- Kopcewicz, B., Kopcewicz, M., and Pietruczuk, A.: The Mössbauer study of atmospheric iron-containing aerosol in the coarse and PM<sub>2.5</sub> fractions measured in rural site, Chemosphere, 131, 9-16, 2015.
- Lafon, S., Rajot, J.-L., Alfaro, S. C., and Gaudichet, A.: Quantification of iron oxides in desert aerosol, Atmos. Environ., 38, 1211-1218, 2004.
- Lelieveld, J., and Heintzenberg, J.: Sulfate cooling effect on climate through in-cloud oxidation of anthropogenic SO<sub>2</sub>, Science, 258, 117-120, 1992.
- Luo, C., Mahowald, N. M., Meskhidze, N., Chen, Y., Siefert, R. L., Baker, A. R., and Johansen, A. M.: Estimation of iron solubility from observations and a global aerosol model, J. Geophys. Res., 110, D23, 2005.
- Luo, C., Mahowald, N., Bond, T., Chuang, P. Y., Artaxo, P., Siefert, R., Chen, Y., and Schauer,J.: Combustion iron distribution and deposition, Global Biogeochem. Cycles, 22, GB1012,2008.
- Luria, M., and Sievering, H.: Heterogeneous and homogeneous oxidation of SO<sub>2</sub> in the remote marine atmosphere, Atmos. Environ., 25, 1489-1496, 1991.
- Ma, Q., He, H., and Liu, Y.: In situ DRIFTS study of hygroscopic behavior of mineral aerosol,J. Environ. Sci., 22, 555-560, 2010.
- Mackie, D. S.: Simulating the cloud processing of iron in Australian dust: pH and dust

- concentration, Geophys. Res. Lett., 32, L06809, 2005.
- Mahowald, N. M., Baker, A. R., Bergametti, G., Brooks, N., Duce, R. A., Jickells, T. D., Kubilay, N., Prospero, J. M., and Tegen, I.: Atmospheric global dust cycle and iron inputs to the ocean. Global Biogeochem. Cycles, 19, GB4025, 2005.
- Maters, E. C., Delmelle, P., and Bonneville, S.: Atmospheric processing of volcanic glass: Effects on iron solubility and redox speciation. Environ. Sci. Technol., 50 (10), 5033-5040, 2016.
- Meskhidze, N.: Iron mobilization in mineral dust: Can anthropogenic SO<sub>2</sub> emissions affect ocean productivity? Geophys. Res. Lett., 30, 1-2, 2003.
- Nanayakkara, C. E., Pettibone, J., and Grassian, V. H.: Sulfur dioxide adsorption and photooxidation on isotopically-labeled titanium dioxide nanoparticle surfaces: Roles of surface hydroxyl groups and adsorbed water in the formation and stability of adsorbed sulfite and sulfate, Phys. Chem. Chem. Phys., 14, 6957-6966, 2012.
- Nie, W., Ding, A. J., Wang, T., Kerminen, V., George, C., Xue, L. K., Wang, W. X., Zhang, Q. Z., Petäjä, T., Qi, X. M., Gao, X. M., Wang, X. F., Yang, X. Q., Fu, C. B., and Kulmala, M.: Polluted dust promotes new particle formation and growth, Sci. rep., 4, 6634, 2014.
- Peak, D., Ford, R. G., and Sparks, D. L.: An in situ ATR-FTIR investigation of sulfate bonding mechanisms on goethite, J. Colloid Interface Sci., 218, 289-299, 1999.
- Pruppacher, H. R., Jaenicke, R.: The processing of water-vapor and aerosols by atmospheric clouds, a global estimate, Atmos. Res., 38, 283-295, 1995.
- Rubasinghege, G., Lentz, R. W., Scherer, M. M., and Grassian, V. H.: Simulated atmospheric

- processing of iron oxyhydroxide minerals at low pH: roles of particle size and acid anion in iron dissolution, Proc. Natl. Acad. Sci. U. S. A., 107, 6628-6633, 2010.
- Sedwick, P. N., Sholkovitz, E. R., and Church, T. M.: Impact of anthropogenic combustion emissions on the fractional solubility of aerosol iron: Evidence from the Sargasso Sea, Geochem. Geophys. Geosyst., 8, Q10Q06, 2007.
- Shi, Z., Bonneville, S., Krom, M. D., Carslaw, K. S., Jickells, T. D., Baker, A. R., and Benning, L. G.: Iron dissolution kinetics of mineral dust at low pH during simulated atmospheric processing, Atmos. Chem. Phys., 11, 995-1007, 2011.
- Shi, Z., Krom, M. D., Bonneville, S., Baker, A. R., Jickells, T. D., and Benning, L. G.: Formation of iron nanoparticles and increase in iron reactivity in mineral dust during simulated cloud processing, Environ. Sci. Technol., 43, 6592-6596, 2009.
- Shi, Z., Krom, M. D., Bonneville, S., Baker, A. R., Bristow, C., Drake, N., Mann, G., Carslaw, K., McQuaid, J. B., Jickells, T., and Benning, L. G.: Influence of chemical weathering and aging of iron oxides on the potential iron solubility of Saharan dust during simulated atmospheric processing, Global Biogeochem. Cycles, 25, GB2010, 2011.
- Shi, Z., Krom, M. D., Bonneville, S., and Benning, L. G.: Atmospheric processing outside clouds increases soluble iron in mineral dust, Environ. Sci. Technol., 49, 1472-1477, 2015.
- Siefert, R. L., Johansen, A. M., Hoffmann, M. R., and Pehkonen, S. O.: Measurements of trace metal (Fe, Cu, Mn, Cr) oxidation states in fog and stratus clouds, J. Air Waste Manage., 48(2), 128–143, 1998.
- Spokes, L. J., Jickells, T. D., and Lim, B.: Solubilization of aerosol trace-metals by cloud

- processing a laboratory study, Geochim. Cosmochim. Acta., 58, 3281-3287, 1994.
- Strehlau, J. H., Schultz, J. D., Vindedahl, A. M., Arnold, W. A., and Penn, R. L.: Effect of nonreactive kaolinite on 4-chloronitrobenzene reduction by Fe(II) in goethite-kaolinite heterogeneous suspensions, Environ. Sci.: Nano, 4, 325-334, 2017.
- Sun, Y. L., Jiang, Q., Wang, Z. F., Fu, P. Q., Li, J., Yang, T., and Yin, Y.: Investigation of the sources and evolution processes of severe haze pollution in Beijing in January 2013, J. Geophys. Res.-Atmos., 119, 4380-4398, 2014.
- Ullerstam, M., Johnson, M. S., Vogt, R., and Ljungstrom, E.: DRIFTS and Knudsen cell study of the heterogeneous reactivity of SO<sub>2</sub> and NO<sub>2</sub> on mineral dust, Atmos. Chem. Phys., 3, 2043-2051, 2003.
- Usher, C. R., Al-Hosney, H., Carlos-Cuellar, S., and Grassian, V. H.: A laboratory study of the heterogeneous uptake and oxidation of sulfur dioxide on mineral dust particles, J. Geophys. Res.-Atmos., 107, 161-169, 2002.
- Usher, C. R., Michel, A. E., and Grassian, V. H.: Reactions on mineral dust, Chem. rev., 103, 4883-4940, 2003.
- Viollier, E., Inglett, P. W., Hunter, K., Roychoudhury, A. N., and Van Cappellen, P.: The ferrozine method revisited: Fe(II)/Fe(III) determination in natural waters, Appl. Geochem., 15, 785-790, 2000.
- Wang, G., Wang, H., Yu, Y., Gao, S., Feng, J., Gao, S., and Wang, L.: Chemical characterization of water-soluble components of PM<sub>10</sub> and PM<sub>2.5</sub> atmospheric aerosols in five locations of Nanjing, China, Atmos. Environ., 37, 2893-2902, 2003.

- Wang, Y., Zhang, Q., Jiang, J., Zhou, W., Wang, B., He, K., Duan, F., Zhang, Q., Philip, S., and Xie, Y.: Enhanced sulfate formation during China's severe winter haze episode in January 2013 missing from current models, J. Geophys. Res.-Atmos., 119, 10425-10440, 2014.
- Wu, L. Y., Tong, S. R., Wang, W. G., and Ge, M. F.: Effects of temperature on the heterogeneous oxidation of sulfur dioxide by ozone on calcium carbonate, Atmos. Chem. Phys., 11, 6593-6605, 2011.
- Wurzler, S., Reisin, T. G., and Levin, Z.: Modification of mineral dust particles by cloud processing and subsequent effects on drop size distributions, J. Geophys. Res.-Atmos., 105, 4501-4512, 2000.
- Yang, W., He, H., Ma, Q., Ma, J., Liu, Y., Liu, P., and Mu, Y.: Synergistic formation of sulfate and ammonium resulting from reaction between SO<sub>2</sub> and NH<sub>3</sub> on typical mineral dust, Phys. Chem. Chem. Phys., 18, 956-964, 2016.
- Yang, W., Zhang, J., Ma, Q., Zhao, Y., Liu, Y., and He, H.: Heterogeneous reaction of SO<sub>2</sub> on manganese oxides: the effect of crystal structure and relative humidity, Sci. rep., 7, 4550, 2017.
- Zhang, X., Zhuang, G., Chen, J., Wang, Y., Wang, X., An, Z., and Zhang, P.: Heterogeneous reactions of sulfur dioxide on typical mineral particles, J. Phys. Chem. B, 110, 12588-12596, 2006.
- Zhao, X., Kong, L., Sun, Z., Ding, X., Cheng, T., Yang, X., and Chen, J.: Interactions between heterogeneous uptake and adsorption of sulfur dioxide and acetaldehyde on hematite, J.

- Phys. Chem. A, 119, 4001-4008, 2015.
- Zhao, Y., Liu, Y., Ma, J., Ma, Q., and He, H.: Heterogeneous reaction of SO<sub>2</sub> with soot: The roles of relative humidity and surface composition of soot in surface sulfate formation, Atmos. Environ., 152, 465-476, 2017.
- Zhu, X., Prospero, J. M., Savoie, D. L., Millero, F. J., Zika, R. G., and Saltzman, E. S.: Photoreduction of iron(III) in marine mineral aerosol solutions, J. Geophys. Res.-Atmos., 98, 9039-9046, 1993.
- Zhuang, G. S., Yi, Z., Duce, R. A., and Brown, P. R.: Link between iron and sulfur cycles suggested by detection of Fe(II) in remote marine aerosols, Nature, 355 (6360), 537-539, 1992.
- Zuo, Y. G., and Hoigne, J.: Formation of hydrogen-peroxide and depletion of oxalic-acid in atmospheric water by photolysis of iron(III) oxalato complexes, Environ. Sci. Technol., 26, 1014-1022, 1992.

# Captions of Figures and Tables

Figure 1. Comparison of morphologies and chemical properties for samples collected before and after CP using TEM. The dotted circles indicate the positions of the electron beam for the HRTEM images and SAED patterns. Elements of the detected parts of individual particles are also presented. Square brackets indicate mass percent of iron. The iron species were identified by the Miller indices and the SAED patterns. (a) IMt-2 particles characterized by high fractions of Al and Si, along with other crustal elements including Mg, K and Fe. (b) IMt-2 particles after CP were almost all less than 1 µm in size. Some Fe-rich particles with less Si and Al were observed on nanoscale dimension. (c) NAu-2 particles with high Fe/Si-ratios, contain Mg, Al and Ca elements. (d) NAu-2 particles after CP were much smaller than the ones before CP. Some ferrihydrite clusters were observed and were attached on the surface of the NAu-2 particles after CP. (e) Typical SWy-2 particles were Al/Si-rich, containing Fe, Mg and Ca elements. (f) TEM images of the SWy-2 particles after CP appeared smaller than the particles before CP. (g) The Si/Al-rich crystal in ATD particles was aluminosilicate with low content of Fe, and a typical of the α-Fe<sub>2</sub>O<sub>3</sub> particles (PDF: 33-664) was found to attach onto the aluminosilicate surface. (h) The pseudohexagonal nanoparticles were observed to on the surface of α-Fe<sub>2</sub>O<sub>3</sub> crystal among the ATD particles. The SAED lattice constant of these nanoparticles were found to be very close to that of 2-line ferrihydrite.

**Figure 2.** Comparison of the DRIFT spectra of mineral dust samples upon exposure to SO<sub>2</sub> for 45 min before and after CP. Data for IMt-2 (a and b), NAu-2 (c and d), SWy-2 (e and f), ATD (g and h), are shown in the ranges of 4000 to 1250 cm<sup>-1</sup> and 1250 to 1000 cm<sup>-1</sup>, respectively.

**Figure 3.** Comparison of the integrated areas on DRIFTS spectra in the range of 1250-1000 cm<sup>-1</sup> for the sulfate species formed on the samples before and after CP.

**Table 1.** Sulfate formation rates and uptake coefficients of SO<sub>2</sub> on particle samples before and after CP.

Figure 4. Comparison of the sulfate formation rates as a function of pH cycle.

**Figure 5.** The fractions of "free-Fe" (Fe<sub>A</sub> and Fe<sub>D</sub>) and "structural-Fe" were measured by the chemical CBD extractions for the samples before and after CP. Results are present as relative percentage of Fe<sub>T</sub>.

**Figure 6.** Mössbauer spectroscopy measured for samples. IMt-2 before and after CP (a and b), NAu-2 before and after CP (c and d), SWy-2 before and after CP (e and f), ATD before and after CP (g and h). Experimental data were fit using a least-squares fitting-program. The IS values were relative to  $\alpha$ -Fe at RT. Prominent spectral features associated with different iron species are indicated.

**Figure 7.** The concentrations of Fe<sub>s</sub>, dissolved Fe(II) and Fe(III) in the suspensions measured over 144 h in the solution cycled between pH 2 and pH 5 for IMt-2 (a), NAu-2 (b), SWy-2 (c) and ATD (d), respectively.

Figure 8. TEM images of the newly formed particles in the precipitation experiment. Based on the TEM-EDX measurement and SAED analysis, these particles could be categorized into two different types, which were circled in Figure 8 a. The typical sizes of the first type were hundreds of nanometers. The enlarged images are displayed in Figure 8 b, c and d. The insert EDX data and SAED image confirmed that they were poor crystalline aluminosilicate with low Fe but high Si/Al content. The second type (Figure 8 e, f and g) were Fe-rich but with less amount of Si/Al, which were nearly 1 micrometer in size. Based on the EDX data and the SAED

analysis, these bigger particles were ambiguously identified as  $Na_{0.42}Fe_3Al_6B_{309}Si_6O_{18}(OH)_{3.65} \ (PDF: 89-6506).$ 

**Figure 9.** Mössbauer spectroscopy measured at RT for the neo-formed particles collected in the precipitation experiment.

**Figure 10.** During the precipitation experiment, the particle size distributions in the suspensions were determined by dynamic light scattering. The presented size distributions are characteristic of neo-formed nanoparticles or microparticles as the suspension pH raised from 1.0 to 3.8.

Figure 1

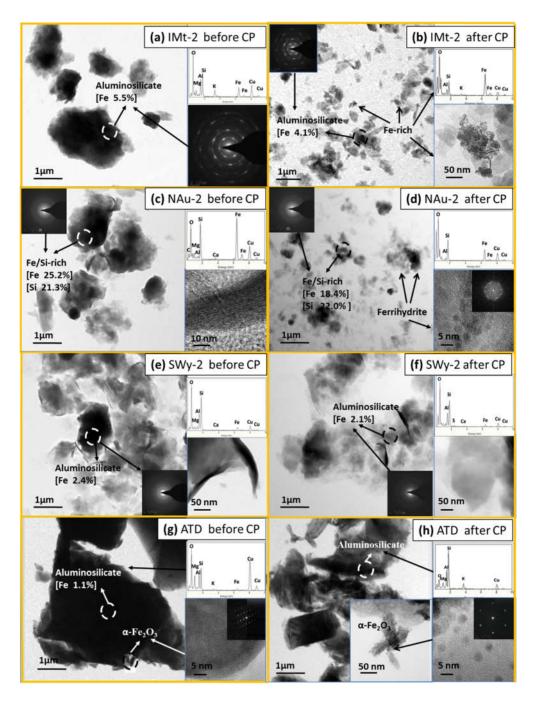


Figure 2

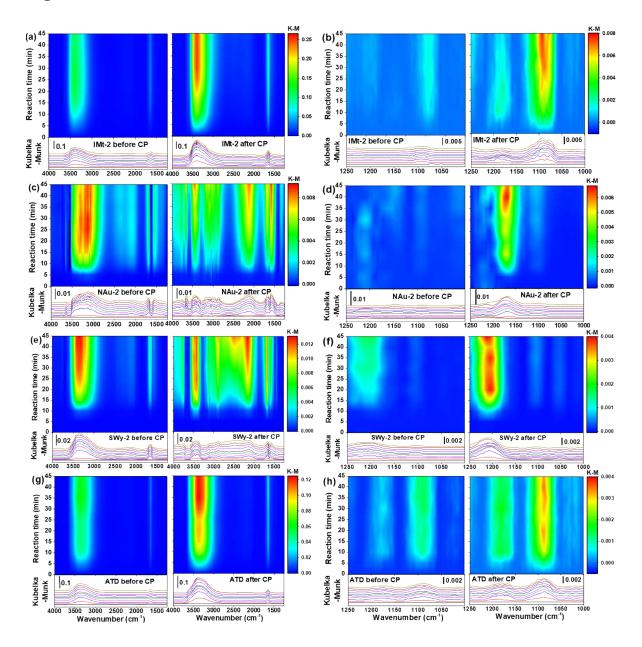


Figure 3

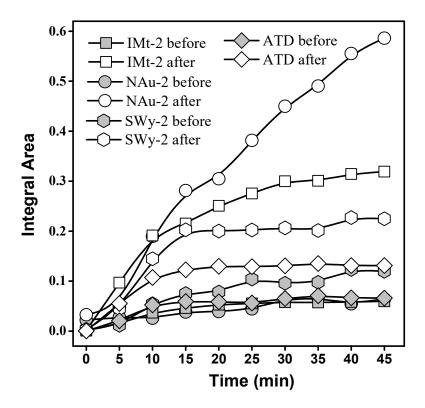


Table 1

Samples	$A_{ m BET}$ (m <sup>2</sup> )	Sulfate formation rate (ions s <sup>-1</sup> ) ( $\times 10^{10}$ )	$A_{\text{geometric}}$ (m <sup>2</sup> ) (×10 <sup>-5</sup> )	γ <sub>BET</sub> (×10 <sup>-12</sup> )	$\gamma_{ m geometric}$ ( $ imes 10^{-7}$ )
IMt-2 before CP	0.770	6.13	1.95	2.62	1.03
IMt-2 after CP	1.640	28.72	1.95	5.76	4.85
NAu-2 before CP	0.790	1.80	1.95	0.75	0.30
NAu-2 after CP	3.749	34.57	1.95	3.06	5.83
SWy-2 before CP	0.906	10.20	1.95	3.70	1.72
SWy-2 after CP	1.631	27.19	1.95	5.49	4.59
ATD before CP	0.166	8.11	1.95	16.05	1.37
ATD after CP	0.241	16.33	1.95	22.33	2.76

Figure 4

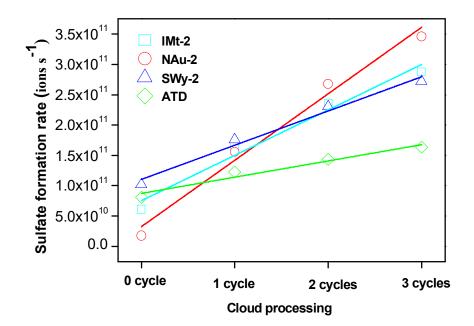


Figure 5

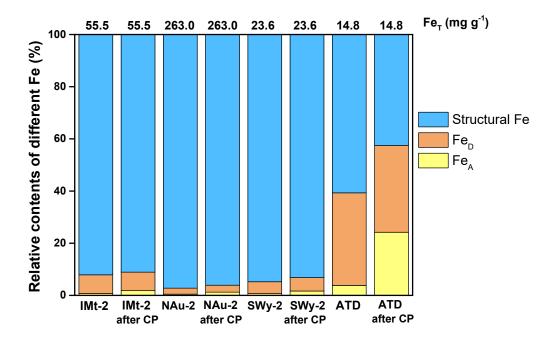


Figure 6

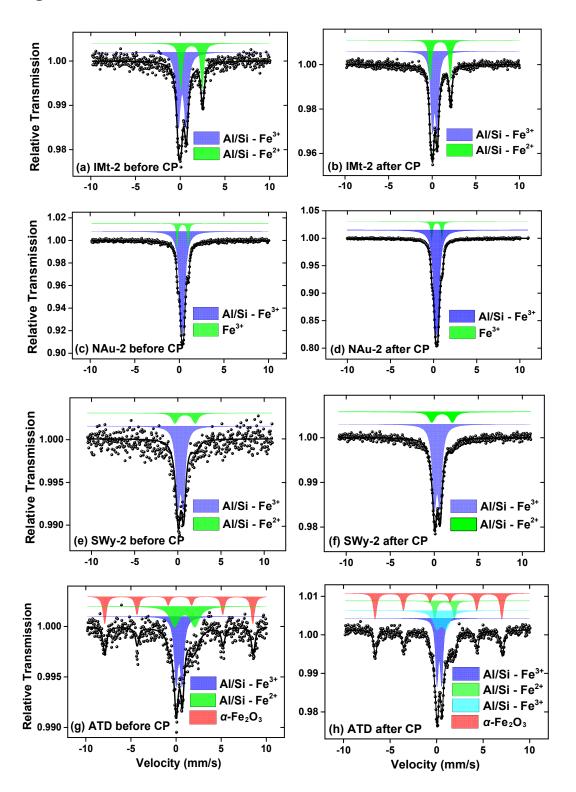


Figure 7

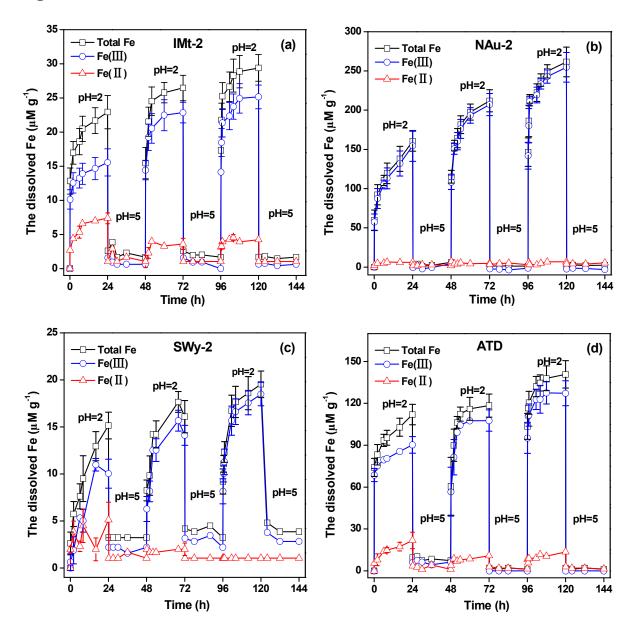


Figure 8

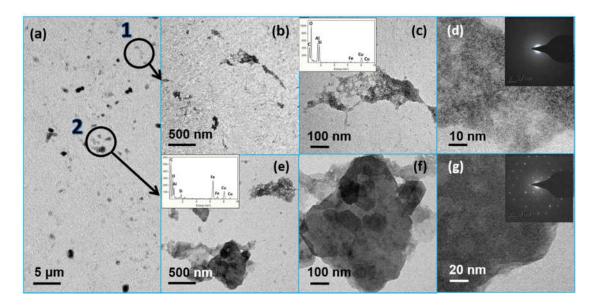


Figure 9

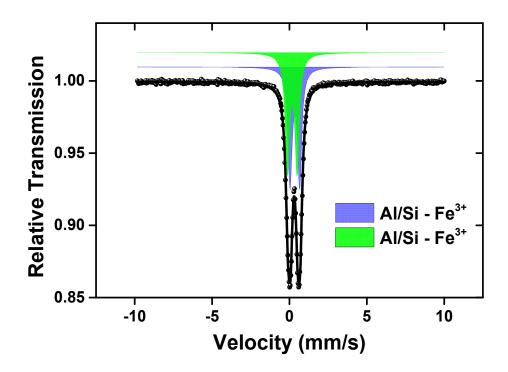


Figure 10

