# <u>Aerosol Mass yields of selected Biogenic Volatile Organic Compounds – a</u> <u>theoretical study with near explicit gas-phase chemistry</u>

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#### 8 Abstract

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In this study we modeled secondary organic aerosols (SOA) mass loadings from the oxidation (by  $O_3$ , 9 OH and  $NO_3$ ) of five representative Biogenic Volatile Organic compounds (BVOCs): isoprene, endocyclic 10 bond containing monoterpenes ( $\alpha$ -pinene and limonene), exocyclic double bond compound ( $\beta$ -pinene) and a 11 sesquiterpene (β-caryophyllene). The simulations were designed to replicate idealized smog chamber and 12 oxidative flow reactors (OFR). The master chemical mechanism (MCM) together with the peroxy radical 13 autoxidation mechanism (PRAM), were used to simulate the gas-phase chemistry. The aim of this study was 14 to compare the potency of MCM and MCM+PRAM in predicting SOA formation. SOA yields were in good 15 agreement with experimental values for chamber simulations when MCM+PRAM was applied, while a 16 standalone MCM under-predicted the SOA yields. Compared to experimental yields, the OFR simulations 17 using MCM+PRAM yields were in good agreement for BVOCs oxidized by both O<sub>3</sub> and OH. On the other 18 hand, a standalone MCM under-predicted the SOA mass yields. SOA yields increased with decreasing 19 temperatures and NO concentrations and vice-versa. This highlights the limitations posed when using fixed 20 SOA yields in a majority of global and regional models. Few compounds that play a crucial role (>95% of 21 mass load) in contributing to SOA mass increase (using MCM+PRAM) are identified. The results further 22 emphasized that incorporating PRAM in conjunction with MCM does improve SOA mass yields estimation. 23

# 24 1. Introduction

Atmospheric secondary organic aerosols, formed from gas to particle phase conversion of the oxidation products of volatile organic compounds (VOC) significantly impact the organic aerosol mass loadings (Griffin, 1999; Kanakidou et al., 2005). However, the scale of SOA contribution to the aerosol particle mass is still subject to high uncertainties (Hao et al., 2011,Glasius and Goldstein, 2016). The elevated aerosol particle concentrations are shown to have inimical effects on health (Miller et al., 2007), and a varying degree of influence on the climate by forming cloud condensation nuclei (CCN), altering the cloud properties

and radiative balance (Rosenfeld et al., 2014; Schmale et al., 2018). Therefore, it is acutely necessary to 31 understand the role and contributions of SOA to the particle loading in the atmosphere. Biogenic VOCs from 32 forest are estimated to contribute to about 90% of VOCs emissions globally (Guenther et al., 1995, 1999 and 33 2000). The most important BVOCs for SOA formation are isoprene ( $C_5H_8$ ), monoterpenes ( $C_{10}H_{16}$ ) and 34 35 sesquiterpenes (C<sub>15</sub>H<sub>24</sub>). These compounds are all alkenes containing at least one carbon-carbon double bond, enabling them to undergo oxidation by the dominant atmospheric oxidants: the hydroxyl radical (OH), ozone 36 (O<sub>3</sub>) and the nitrate radical (NO<sub>3</sub>). For some of the terpenes, initial oxidation steps can lead to formation of 37 highly oxygenated organic molecules (HOM). These HOMs generally have low volatilities and can condense 38 nearly irreversibly, thereby producing SOA (Ehn et al., 2014). HOMs, detected in both the ambient 39 atmosphere and chamber experiments (Ehn et al., 2012) are formed by autoxidation (Berndt et al., 2016; 40 Crounse and Nielsen, 2013) wherein peroxy radicals (RO<sub>2</sub>) undergo subsequent intramolecular H-shifts 41 accompanied by rapid reactions with O<sub>2</sub>. Autoxidation hence results in compounds containing multiple 42 functional groups such as hydroxyls, peroxides and carbonyls (Bianchi et al., 2017, Bianchi et al., 2019). 43

A majority of chamber and flow-tube experiments have focused on HOM formation from the 44 oxidation of various VOCs and their contribution to SOA mass loadings (Ehn et al., 2014; Kristensen et al., 45 2017). Oxidation of isoprene (Liu et al., 2016), endocyclic monoterpenes containing reactive double bonds 46 47 such as α-pinene and limonene (Zhao et al., 2015), or exocyclic double bond containing compounds such as  $\beta$ -pinene (Jokinen et al., 2015) and sesquiterpenes such as  $\beta$ -caryophyllene (Chen et al., 2012) have been 48 investigated. The SOA forming potential of various BVOCs depends on the isomeric structures (Friedman 49 50 and Farmer, 2018; Keywood et al., 2004). Ozonolysis of compounds containing reactive endocyclic bonds such as  $\alpha$ -pinene produce higher SOA mass yields of 41% in comparison to those with exocyclic bonds ( $\beta$ -51 pinene), which produce mass yields of 17 % (Lee et al., 2006a). One explanation for this dependence on the 52 isomeric structure is attributed to the formation of HOMs (Ehn et al., 2014). Another important factor 53 54 influencing HOM formation is the initial oxidant, as pointed out by Zhao and co-workers (2015). They showed that the SOA formation by OH oxidation of α-pinene and limonene were lower when compared to 55 their SOA formed by ozonolysis. Further they measured lower H/C ratio for SOA produced by monoterpene 56 ozonolysis (experiments were carried out in dark with CO as OH scavenger), in comparison to OH oxidation 57 of  $\alpha$ -pinene and limonene, while O/C ratio were similar for both oxidation cases. This was attributed to the 58 formation of  $RO_2$  radicals (monoterpenes  $+O_3$ ) which undergo internal hydrogen shifts and subsequently react 59 with another RO<sub>2</sub> radical, to form compounds containing carbonyl groups while losing hydrogen atoms in the 60 process. A similar analysis was conducted by Draper et al. (2015), who showed that an increase in  $NO_2$ 61 concentration reduced α-pinene ozonolysis SOA mass yields, while no appreciable reduction in mass yields 62

are reported for  $\beta$ -pinene and  $\Delta^3$ - carene ozonolysis. On the other hand, the mass yields from limonene ozonolysis increased with increasing NO<sub>2</sub> concentrations (Draper et al., 2015). This disparity in mass yields for different BVOCs in the presence of NO<sub>2</sub> is possibly caused by the formation of high molecular weight oligomers (or lack of in case of  $\alpha$ -pinene) through oxidation with NO<sub>3</sub> that contribute to SOA mass loadings (Draper et al., 2015).

Due to computational limitations, many regional and canopy scale atmospheric chemistry models
generally use isoprene and/or a representative monoterpene (generally α-pinene), to model SOA yields
(Friedman and Farmer, 2018). The SOA yields of different monoterpenes vary with structure, NO<sub>x</sub> and
temperature (Friedman and Farmer, 2018; Kristensen et al., 2017; Presto et al., 2005). This poses a limitation
on using representative monoterpene fixed SOA yields in many of the global models and increases
uncertainties in predicting cloud condensation nuclei concentrations, cloud droplet number concentrations and
radiative balance due to aerosol loadings.

75 This work aims to investigate the SOA mass loading from the oxidation products of BVOCs with the atmospheric oxidants OH, O<sub>3</sub> and NO<sub>3</sub> with a specific focus on the BVOCs isoprene,  $\alpha$ -pinene,  $\beta$ -pinene, 76 limonene and β-caryophyllene. Further we study the effect of varying temperature (258.15 K – 313.15 K) and 77 NO concentrations (0 - 5 ppb) on  $\alpha$ -pinene oxidation mass yields. We use the master chemical mechanism 78 79 (MCMv3.3.1) (Jenkin et al., 1997, 2012 and 2015; Saunders et al., 2003), a near explicit gas-phase chemical mechanism together with peroxy radical autoxidation mechanism (PRAM, Roldin et al., 2019) (PRAM + 80 MCM). The aim is to understand the importance and contribution of peroxy radical autoxidation products to 81 the SOA mass yields from terpenes. 82

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#### 84 2. Model description

#### 85 2.1 **Malte Box**

MALTE (Model to predict new Aerosol formation in Lower TropospherE) is a one-dimensional model consisting of modules calculating boundary layer meteorology, emissions of BVOCs, gas-phase chemistry and aerosol dynamics with the aim to simulate particle distribution and growth in the lower troposphere (Boy et al., 2006). In this study, a zero-dimensional version, MALTE-Box is applied to simulate an ideal chamber and flow-tube environment (i.e. no wall losses effects are considered in this study). For the simulations performed in this study the emission module was switched off while only employing the gasphase chemistry and aerosol dynamics module.

Kinetic preprocessor (KPP) is used to generate a system of coupled differential equations to solve the 93 gas-phase chemistry schemes (Damian et al., 2002). The peroxy radical autoxidation mechanism (PRAM), 94 (Roldin et al., 2019, Qi et al., 2018, Öström et al., 2017), formulated based on the oxidation of monoterpenes 95 as described by Ehn et al. (2014) was incorporated alongside MCMv3.3.1. PRAM explicitly describes the 96 97 formation and evolution of peroxy radicals (RO<sub>2</sub>) from the ozonolysis and OH oxidation of monoterpenes, driven by subsequent H-shifts and O<sub>2</sub> additions. The current version of PRAM based on experimental and 98 theoretical studies, considers HOM autoxidation for a fraction of the peroxy radicals formed during the 99 ozonolysis of  $\alpha$ -pinene and limonene and OH oxidation of  $\alpha$ -pinene,  $\beta$ -pinene and limonene. This is achieved 100 by assigning species specific molar yields for the formation of first RO<sub>2</sub>, which subsequently initiates the 101 autoxidation chain (Roldin et al., 2019). Currently, in PRAM a maximum first generation RO<sub>2</sub> yield of 9% for 102  $\alpha$ -pinene ozonolysis, 21.9 % for limonene ozonolysis, 2.5 % for  $\alpha$ -pinene+OH, and 1% for both 103 limonene+OH and  $\beta$ -pinene+OH first generation products are allowed to initiate autoxidation (Roldin et al., 104 2019). For  $\beta$ -pinene ozonolysis the molar yield of RO<sub>2</sub> is minor (<0.1 %) (Roldin et al., 2019, Ehn et al. 2014) 105 106 and hence not considered in this work. The above mentioned RO<sub>2</sub> molar yields used in this work are close to the experimental values obtained in both smog chamber and flow tube experiments. Ehn et al. (2014) 107 measured an RO<sub>2</sub> yield of ~7% for  $\alpha$ -pinene ozonolysis and ~17% for limonene ozonolysis, whereas Jokinen 108 et al. (2015) measured 0.58 % and 0.93 % for OH oxidation of β-pinene and limonene respectively. The 109 autoxidation is terminated by bimolecular reactions, wherein the RO<sub>2</sub> formed reacts with NO, HO<sub>2</sub> or other 110 peroxy radicals, thereby forming alkoxy radicals, closed shell monomers or dimers (Roldin et al., 2019). The 111 PRAM considers temperature dependent autoxidation reaction rates, which is important when investigating 112 the SOA mass yields at varying temperatures (Table 1c). The temperature dependence in PRAM is based on 113 quantum chemical calculations wherein the autoxidation rates correspond to an activation energy of 24 114 kcal/mol. The activation energies vary for autoxidation of different RO<sub>2</sub> from *α*-pinene ozonolysis between 22 115 and 29 kcal/mol (Rissanen et al., 2015), leading to varying autoxidation rates at different temperatures 116 (Roldin et al., 2019). It should be noted that the temperature dependence in PRAM is a first of its kind but 117 needs further evaluation using recent measurements of HOM formation at different temperatures (e.g. 118 Quéléver et al., 2019). 119

The aerosol dynamics are simulated using the University of Helsinki Multicomponent Aerosol model (UHMA) originally from Korhonen et al. (2004). The model has undergone significant development since then to allow simulation with all the compounds from MCM. It now supports an unlimited number of condensing vapors and solves condensation using the analytical predictor of condensation method from Jacobson (1997). The condensation algorithm considers both, the Kelvin effect and Raoult's law. The processes included in the model are nucleation, condensation, evaporation, coagulation and deposition. The discretization of the size distribution and the time evolution is modeled with the moving section approach, with optional redistribution to a fixed grid. In this work, the redistribution is active to make the coagulation more accurate, since it requires that grid points are available near the size of the coagulated particles. In this study nucleation and deposition are not active, and hence are not considered. A total of 100 size bins ranging from 1nm to 20 µm with the fixed grid was applied for this study.

A group contribution method based on Nannoolal et al. (2008) using the UManSysProp online system (Topping et al., 2016) was used to estimate the pure liquid saturation vapor pressures (p<sub>0</sub>) of the organic compounds in MCMv3.3.1. For the PRAM species, p<sub>0</sub> were estimated using the functional group method SIMPOL (Pankow and Asher, 2008; see Roldin et al., 2019 for details). Temperature was used as an input to estimate p<sub>0</sub> for both the group contribution methods.

#### 136 2.2 Simulations

The simulations performed in this study are aimed to closely resemble an idealized smog chamber 137 (batch mode setup) and an Oxidative Flow Reactor (OFR) without interactions between the gas phase and the 138 system walls. For the chamber runs, the VOC and oxidants were introduced at the beginning (time, t=0 sec), 139 set to certain concentrations (Table 1a) and then allowed to react. Both chamber and OFR simulations are 140 performed using ammonium sulfate seed particles which are introduced at time t=0. The condensation sink 141 (CS) was inferred from the size distribution of seed particles used in the model. The CS for the chamber and 142 OFR simulations was set to 0.00067 s<sup>-1</sup> and 0.067 s<sup>-1</sup> respectively. SOA mass yields obtained using an OFR 143 are sensitive to short residence time used, hence the seed particle surface area should be chosen in order to 144 overcome the mass yield underestimation (Ahlberg et al., 2019). CS sensitivity runs (Supplement Figure S1) 145 were performed for α-pinene-O<sub>3</sub> to determine the CS for which there are no appreciable change in mass yields 146 with increasing particle surface. 147

148 The simulation for the chamber setup is run for a maximum time of 24 hours and ends when either of the 2 criteria are satisfied: (1) the simulation time reaches the 24-hour mark or (2) 90 % of the initial 149 precursor VOC has reacted away. In the latter case the simulation is continued for an additional 2 hours to 150 ensure enough time for the vapors to condense onto the seed particles. By contrast, the OFR runs were 151 simulated for a maximum residence time of 100 seconds, ensuring all initial precursor vapors were oxidized. 152 Seed particles were also added in the OFR simulations. The oxidant concentrations used for the OFR 153 simulations (Table 1b) are significantly higher in comparison to the simulated chamber runs (~2 orders of 154 magnitude larger). The time step for the chamber and flow-tube simulations are set to t=10 s and t =0.1 s 155

respectively. The runs performed were oxidant specific (i.e. VOCs would be oxidized by only one specific oxidant at any given time). For the  $O_3$  specific simulations no OH could form in both, OFR and chamber setups, thus enabling oxidation of  $O_3$  to be the only pathway.

The simulations were performed at atmospheric relevant  $NO_x$  ( $NO_x = NO + NO_2$ ) concentrations, 159 corresponding to [NO]=0.5 ppb and  $[NO_2]=2.0$  ppb conditions with the relative humidity (RH) set to 60 % 160 and temperature to 293.15 K. The RH value considered in this study is based on previous published 161 162 experimental studies performed at ~60 % in both smog chamber (Bruns et al., 2015a; Ehn et al., 2014; Stirnweis et al., 2017) and OFR (Ahlberg et al. et al., 2018).  $\alpha$ -pinene ozonolysis runs were performed at four 163 different temperatures: 258.15 K, 278.15 K, 303.15K and 313.15 K, respectively (Table 1c) . SOA mass 164 yields are expected to increase with decreasing temperature (Saathoff and Naumann, 2009). A similar 165 166 temperature dependence was observed by Kristensen et al. (2017) who observed SOA mass yield from  $\alpha$ pinene ozonolysis at ~ 40 % and ~20 % at 258 K and 293 K respectively. Analogous to analyzing the effect of 167 varying temperature on SOA yields, we study the variation in  $\alpha$ -pinene ozonolysis SOA mass yields by 168 varying the NO<sub>x</sub> concentrations (Table 1c). SOA yields for  $\alpha$ -pinene ozonolysis at high NO<sub>x</sub> conditions 169 170 should be suppressed (Ng and Chhabra, 2007), which could be due to the production of relatively, volatile organic nitrates under high NO<sub>x</sub> conditions as compared to less volatile products during low NO<sub>x</sub> conditions 171 (Presto et al., 2005). 172

Furthermore, two different chemistry schemes were applied for the simulations. One scheme consisted of only the MCM chemistry mechanism and the second included the MCM+PRAM chemistry mechanism. Table 1a shows the concentrations of different BVOCs and Table 1b shows the oxidants concentrations used for the simulations.

<b>α-pinene (ppb)</b>	β-pinene (ppb)	Isoprene (ppb)	Limonene (ppb)	B-caryophyllene (ppb)
0.5, 1.0, 5.0,	0.5, 1.0, 5.0,	5.0, 50.0, 100.0,	1.0, 5.0, 50.0,	0.5, 1.0, 2.0, 5.0, 10.0
50.0, 100.0,	50.0, 100.0,	200.0	100.0, 200.0	
200.0	200.0			

**Table 1a.** Concentrations of different BVOCs

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179 Table 1b. Concentrations of different oxidants for chamber and flow-tube runs

OH (* 10 <sup>6</sup> #/cm <sup>3</sup> ) - chamber	O <sub>3</sub> (* 10 <sup>11</sup> #/cm <sup>3</sup> ) - chamber	<b>NO</b> <sub>3</sub> (* 10 <sup>7</sup> #/cm <sup>3</sup> ) - chamber
OH (* 10 <sup>8</sup> #/cm <sup>3</sup> ) - OFR	O <sub>3</sub> (* 10 <sup>13</sup> #/cm <sup>3</sup> ) - OFR	NO <sub>3</sub> (* 10 <sup>9</sup> #/cm <sup>3</sup> ) - OFR

2.0, 5.0 ,10.0, 50.0,100.0	1.0, 5.0 ,10.0, 50.0,100.0	1.0, 5.0 ,10.0, 50.0,100.0

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**Table 1c.** NO concentrations and temperatures used for α-pinene ozonolysis

NO	(ppb)	0.5 (default), 0, 0.2, 1, 2, 5
Ten	nperature (K)	293.15 (default), 258.15, 278.15, 303.15, 313.15

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# 183 2.3 Mass Yields

184 The SOA mass yields (Y) are determined by calculating the ratio of the amount of SOA or mass 185 concentration of organic aerosol formed ( $C_{OA}$ ) to the amount of VOC ( $\Delta$ VOC) reacted:

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$$Y = \frac{C_{OA}}{\Delta VOC}$$
(1)

A volatility basis set is fit to the data to obtain the volatility distribution. In this study equilibrium partitioning was only assumed for deriving the volatility distribution based on the model simulations. Following Donahue et al. (2006), the SOA is assumed to be in equilibrium with the gas-phase and using the effective saturation concentration  $C_i^*$  spaced logarithmically. The individual product partitioning to the particle phase can be estimated using

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$$E_i = \left(1 + \frac{C_i^*}{C_{OA}}\right)^{-1}$$
 (2)

Where  $E_i$  is the fraction of species in the condensed particle phase. The above equation determines the fraction of species in the particle phase as well as in the gas phase. For example, if we assume  $C_{OA} = 10 \ \mu g \ m^{-3}$ a species with  $C^* = 10 \ \mu g \ m^{-3}$  will partition 50 % to condensed phase and the rest 50% will reside in the gas phase. The fidelity of this equilibrium partitioning enables the parameterization of product vapors in volatility  $C^*$  bins that are near the  $C_{OA}$  concentrations (Henry et al., 2012).

# 198 3. Results and Discussion

SOA mass yields were simulated for the oxidation of various biogenic volatile organic compounds (isoprene,  $\alpha$ -pinene, limonene and  $\beta$ -caryophyllene,  $\beta$ -pinene) by dominant atmospheric oxidants OH, O<sub>3</sub> and NO<sub>3</sub>. The following section examines the comparison between the yields derived using MCM+PRAM and a standalone MCM for chamber and flow-tube experiments. 203

#### **3.1 BVOCs – O<sub>3</sub> chamber and flow-tube simulations**

In Fig. 1 panel A indicates the SOA mass yields derived on applying a coupled MCM+PRAM mechanism to
 ozonolysis of α-pinene and limonene (PRAM is only available for ozonolysis of α-pinene and limonene) and
 the lower panel B shows ratio of yields obtained by MCM and coupled MCM+PRAM.

The abscissa, depicted on a log scale, considers the entire range of SOA mass loadings from 1-1150 208 800 µg m<sup>-3</sup>. Each data point is representative of simulated SOA mass yields resulting from variable BVOC 209 loading. The resulting mass yields for  $\alpha$ -pinene in the range shown in Table 2a. are consistent with the yields 210 found in various smog chamber experiments. The mass yields derived using MCM+PRAM for α-pinene 211 ozonolysis are in good agreement with the experimental yields measured for similar mass loadings by 212 Kristensen et al. (2017) and Pathak et al. (2007). The standalone MCM, on the other hand, severely under-213 predicts the mass yields for α-pinene ozonolysis. The MCM+PRAM also shows better agreement with 214 experiments when estimating the lower range mass yields for SOA mass loadings of  $< 15 \ \mu g \ m^{-3}$ . This is 215 supported by the values obtained by Shilling et al. (2008), where the authors measured a 0.09 yield from  $\alpha$ -216 pinene ozonolysis for SOA mass loading of 10.6 µg m<sup>-3</sup>. Limonene ozonolysis mass yields using 217 MCM+PRAM in comparison to standalone MCM, are much closer to the values given by Waring (2016). 218

The formation of HOM from  $\beta$ -pinene ozonolysis is low (Ehn et al., 2014; Jokinen et al., 2015) and hence not considered in PRAM. The peroxy radical autoxidation mechanism for  $\beta$ -caryophyllene ozonolysis has not yet been developed and therefore, not considered in PRAM. When comparing the measured mass yield values for  $\beta$ -caryophyllene (Chen et al. 2012) and  $\beta$ -pinene ozonolysis (Griffin (1999) and Pathak et al. (2008) ) to the modeled values using the MCM scheme, it is evident that the MCM scheme drastically underpredicts the SOA mass yields (Fig. 2).

225 Today oxidation flow reactor (OFR) experiments are complementing the traditional batch mode smog chamber experiments. The OFR generally exhibits lower mass yields compared to the smog chamber 226 experiments at ranges of equivalent oxidant exposure (Lambe et al., 2015). We modeled flow-tube simulation 227 after the potential aerosol mass (PAM) OFR, where the residence time is in the order of a few to several 228 minutes (Lambe et al., 2011). The model simulations are performed with a maximum residence time of 100 229 seconds with O<sub>3</sub> exposures ranging from 1.0 x  $10^{15} - 1.0 \times 10^{17}$  molecules cm<sup>-3</sup> s (residence time \* [O<sub>3</sub>]). 230 Kang and Root (2007) measured a value yield of 0.2 for ozonolysis of  $\alpha$ -pinene for an initial precursor VOC 231 concentration of 100 ppbv, while we obtain ~0.25 (MCM+PRAM) for the similar initial precursor 232 concentrations. The OFR yields for  $\beta$ -pinene (MCM-only) are significantly lower (0.02) than the values 233

234 measured by Kang and Root (2007) wherein they measured a yield of 0.49 for similar initial precursor concentrations. Addition of seed particles promotes condensation, leading to increased SOA yields (Lambe et 235 al., 2015) which was confirmed by Ahlberg et al, (2019). Kang and Root (2007) found that using seed 236 particles, the yield from  $\alpha$ -pinene ozonolysis increased by a factor of ~1.4 which can explain our yields for  $\alpha$ -237 238 pinene ozonolysis simulations. The mass spectra plot (Figure S2) shows that PRAM contributes the majority of dimers to the particle phase, while MCM dominate monomer contribution. Another interesting facet of 239 Figure S2 are the different condensing compounds in both OFR and chamber simulations. The higher absolute 240 RO<sub>2</sub> concentrations in the OFR simulations explain the lower concentration of HOM monomers and dimers 241 relative to the chamber simulations, i.e. the high RO<sub>2</sub> concentrations in the OFR cause termination of the 242 peroxy radical autoxidation chain before the RO<sub>2</sub> become highly oxygenated, thereby influencing SOA yields. 243 Hence, this should be taken into account when using yields from OFR as inputs to regional and global 244 models. 245

**Table 2a.** Mass yields for BVOCs ozonolysis at 293 K for different range of mass loadings using a chamber<sup>†</sup> setup. The values in parenthesis in the column 'Experimental yields' indicates the corresponding experimental mass loadings in unit of μg m<sup>-3</sup>.

SOA mass	MCM +	MCM mass	BVOC	Experimental	References
loading (µg	PRAM mass	yields range		yields	
m⁻³)	yields range				
0– 15 <sup>†</sup>	0.07-0.08	0.00 - 0.06	α-pinene	0.09 (10.6)	Shilling et al. (2008)
16 - 60 <sup>†</sup>	0.12 - 0.20	0.06 - 0.11	α-pinene	0.16 – 0.21 (15	Pathak et al. (2007)
				-60)	
$61 - 200^{\dagger}$	0.22 – 0.30	0.12 – 0.15	α-pinene	0.22 (62)	Kristensen et al. (2017)
$1.1 - 550^{\dagger}$	0.24 -0.48	0.007-0.06	limonene	0.26 (1.7)	Waring (2016)
0 - 100 <sup>†</sup>	$0 - 0.09^{!!}$	0-0.09	β-pinene	0.03-0.22 (7.2	Griffin (1999)
				- 100)	
0 -10 <sup>†</sup>	0-0.01"	0-0.01	β-caryophyllene	0.13 (1.8)	Chen et al. (2012)

<sup>246</sup> <sup>"</sup>indicates that no PRAM mechanism available yet i.e the yields are same as the MCM yields.

Table 2b. Mass yields for BVOCs ozonolysis at 293 K for different range of mass loadings using an OFR<sup>I</sup> setup.

SOA mass	MCM + PRAM	MCM mass	BVOC	Experimental	References
loading (ppb)	mass yields	yields range		yields	
	range				
0-100 <sup>  </sup>	0.07-0.25	0-0.13	α-pinene	0.2 (100)	Kang and Root (2007)
0-156 <sup>  </sup>	0-0.02"	0-0.02	β-pinene	0.49 (156)	Kang and Root (2007)



Figure 1. The mass yields from the ozonolysis of BVOCs α-pinene (red heptagon) and limonene (blue crosses) modelled after
chamber (filled symbols) and flow-tube settings (open symbols). The figure shows a comparison of SOA mass yields obtained from
simulations with MCM + PRAM (panel A) and ratio of yields from MCM and MCM+PRAM (panel B). Currently PRAM is

available for ozonolysis of limonene and α-pinene. The clumps are a result of SOA mass yields for the oxidation of specific
 oxidant concentration with varying BVOC concentration





Figure 2. The mass yields from the ozonolysis of BVOCs β-pinene and β-caryophyllene modelled after chamber (filled symbols)
and flow-tube (open symbols) settings. The figure shows a comparison of SOA mass yields obtained from simulations with only
MCM as currently there is no PRAM available for these compounds. The experimental values are provided for comparison.

257 3.2 BVOCs – OH chamber and flow-tube simulations

The mass yields obtained by MCM+PRAM for  $\alpha$ -pinene – OH oxidation are close to the measured values (Kristensen et al., 2017), while using only MCM under-predicts the mass yields (Figure 3, panel A and B, and Table 3). The maximum SOA mass yield for OH oxidation of  $\alpha$ -pinene is lower than the yield from ozonolysis which is suspected to arise due to the formation of more volatile oxidation products produced during OH oxidation (Bonn and Moortgat, 2002; Kristensen et al., 2014). The OH oxidation of  $\beta$ -pinene results in mass yields similar to the measurements obtained by Lee et al. (2006b) for similar mass loadings. The β-pinene SOA yields are comparatively well represented by MCM+PRAM in comparison to the
standalone MCM. On the other hand, the limonene mass yields are under-predicted by MCM+PRAM for
similar mass loadings. Yields for limonene SOA mass loadings of 350 µg m<sup>-3</sup> are around 0.31 which is lower
than the experimental values of 0.58, measured by Lee et al. (2006b).

For β-caryophyllene, the modeled values are in good agreement with experimental measured yields in 268 the range of mass loadings provided by Griffin (1999) and Tasoglou and Pandis (2015). Currently there are no 269 experiments providing HOM yields from OH oxidation of β-caryophyllene, and hence, those species are not 270 included in PRAM. The simulation results for yields from OH oxidation of β-caryophyllene, indicate that the 271 MCM scheme is able to reproduce the experimental values (Fig. 4). Only MCM was used for modeling the 272 mass yields for OH oxidation of isoprene due to current lack of PRAM mechanism for isoprene. The mass 273 yields derived from OH oxidation of isoprene vary from 0.01 - 0.31 covering a range of mass loadings from 274 0.003 - 132  $\mu$ g m<sup>-3</sup>. At low mass loadings < 10  $\mu$ g m<sup>-3</sup> the maximum yield obtained is ~0.06, which is a factor 275 of 3 greater than the experimental results obtained by (Lee et al., 2006b) where they measured yield of 0.02. 276 The mass yields are in good agreement with the experimental results from Liu et al. (2016), wherein they 277 measured a yield of 0.13 for 22  $\mu$ g m<sup>-3</sup> (Table 3). 278

The OFR simulations results for the OH oxidation of BVOCs with an equivalent exposure range from 279 2.0 x  $10^{10}$  – 2.0 x  $10^{12}$  molecules cm<sup>-3</sup> s, is shown in Fig. 3. Our yields for  $\alpha$ -pinene agree well with the yields 280 obtained by Bruns et al. (2015) where they measured yield of ~0.3 for mass loading of ~300  $\mu$ g m<sup>-3</sup> at 281 equivalent OH exposures. Friedman and Farmer (2018) found mass yields of 0 - 0.086 for  $\alpha$ -pinene 282 (ammonium sulfate seeded experiment), 0- 0.12 for β-pinene (no seed particles) and 0-0.04 for limonene (no 283 seed particles), by varying the OH exposures between  $4.7 \times 10^{10} - 7.4 \times 10^{11}$  molecules cm<sup>-3</sup> s. Our simulated 284 yields for OH oxidation of  $\alpha$ -pinene,  $\beta$ -pinene and limonene suggest higher mass yields for  $\alpha$ -pinene and 285 limonene at equivalent mass loadings, while mass yields for β-pinene are in good agreement with the 286 experimental yields. Friedman and Farmer (2018) suggest that the reason for this underestimation in mass 287 yields could arise due to the exclusion of large particle sizes in the experiments and propose that these yields 288 289 could represent lower bounds.

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Table 3. Mass yields for OH oxidation of BVOCs at 293 K for different range of mass loadings using a
chamber<sup>†</sup> and OFR<sup>||</sup> setup.

SOA mass	MCM +	MCM mass	BVOC	Experimental	References
loading (µg	PRAM mass	yields		yields	
m-3)	yields				
300 <sup>†</sup>	0.28	0.25	β-pinene	0.31 (293)	Lee et al. (2006b)
350 <sup>†</sup>	0.31	0.06 - 0.11	limonene	0.58 (394)	Lee et al. (2006b)
30	0.09	0.004	α-pinene	0.11 (30)	Kristensen et al., 2017
$< 10^{\dagger}$	0.21"	0.21	β-caryophyllene	0.2 (8.8)	Tasoglou and Pandis (2015)
$20 - 80^{\dagger}$	0.3 - 0.7"	0.3 - 0.7		0.37 – 0.79 (17-	Griffin (1999)
				82)	
22 <sup>†</sup>	0.1"	0.1	Isoprene	0.13 (22)	Liu et al. (2016)
<10	0.06"	0.06		0.02 (9)	Lee et al. (2006b)
0-300 <sup>  </sup>	0.05 – 0.31	0-0.2	α-pinene	0-0.086 (0-300)	Friedman and Farmer (2018)
				0.3 (300)	Bruns et al. (2015)
0-30 <sup>  </sup>	0-0.1	0-0.01	β-pinene	0 – 0.12 (30)	Friedman and Farmer
					(2018)
0-40 <sup>  </sup>	00.19	0-0.17	limonene	0.0 – 0.04 (35)	Friedman and Farmer (2018)

295 <sup>"</sup>indicates that no PRAM mechanism available yet i.e the yields are same as the MCM yields.



Figure 3. The mass yields from OH oxidation of BVOCs α-pinene (red heptagons), β-pinene (black squares) and limonene (blue
crosses) modeled after chamber (filled symbols) and flow-tube settings (open symbols). The figure shows a comparison of SOA
mass yields obtained from application of MCM+PRAM (panel A) and ratio of yields from MCM and couple MCM+PRAM (panel
B). Currently PRAM is available for OH oxidation of limonene and β-pinene.



Figure 4. The mass yields from OH oxidation of BVOCs β-caryophyllene (black triangles) and isoprene (maroon diamonds)
 modeled after chamber (filled symbols) and flow-tube settings (open symbols). The figure shows a comparison of SOA mass yields
 obtained from application of MCM as currently there is no PRAM available for these compounds.

# **304 3.3 BVOC – NO<sub>3</sub> chamber and OFR simulations**

Figure 5. shows the yields derived from the oxidation of BVOCs by NO<sub>3</sub>. Currently, as no PRAM is 305 available for NO<sub>3</sub> oxidation, Figure 5 represents SOA yields derived using MCM. Due to limited 306 experimental constraints, PRAM presently does not consider autoxidation of RO<sub>2</sub> formed from NO<sub>3</sub> oxidation 307 of VOCs, which could explain the huge discrepancy between the measured and simulated mass yields (Figure 308 5). The yields obtained for oxidation of  $\alpha$ -pinene (0.002-0.007) by NO<sub>3</sub> are low in comparison to those 309 obtained by Nah et al. (2016), where they measured a yield of 0.036. Measured mass yields for limonene 310 oxidation by NO<sub>3</sub> resulting in mass yields between 0.25-0.4 (Fry et al., 2011), whereas we obtain negligible 311 312  $(\sim 0.0003)$  mass yields for the same.

BVOC - NO3 dependence for MCM



Figure 5. The mass yields from NO<sub>3</sub> oxidation of BVOCs modeled after chamber and flow-tube settings. The figure shows a
comparison of SOA mass yields obtained from application of MCM. Appreciable mass yields were only obtained for α-pinene,
limonene and β-caryophyllene.

#### 316 **3.4 NO<sub>x</sub> dependence**

Varying NO<sub>x</sub> concentrations changes the fate of RO<sub>2</sub> radical formed during organic oxidations by 317 altering HO<sub>2</sub>/RO<sub>2</sub> ratio, thereby impacting the distribution of reaction products and aerosol formation (Presto 318 et al., 2005; Zhao et al., 2018; Sarrafzadeh et al., 2016). We modeled the SOA mass yields for  $\alpha$ -pinene - O<sub>3</sub> 319 setup with varying  $NO_x$  concentrations (NO was varied whereas  $NO_2$  was kept constant for all the runs), for 320 initial  $\alpha$ -pinene mixing ratios in the range 0.5 - 200 ppb (Fig. 6). A maximum SOA yield value of 0.55 is 321 obtained for a combination of the lowest value of NO (0 ppb, red circles). As the NO concentrations increase 322 from 0.2 ppb (blue squares) to 5 ppb (green inverted triangles) the yields begin to decrease, and this pattern is 323 observable and valid for all concentration ranges of reacted precursor VOC. The NO<sub>x</sub> dependence of α-pinene 324

ozonolysis is consistent with the findings of Draper et al. (2015) and Presto et al. (2005) wherein they observed a trend of decreasing SOA mass yields for  $\alpha$ -pinene ozonolysis with increasing NO<sub>x</sub> concentrations.



 $\alpha$ -pinene -  $O_3$  - NO dependence

Figure 6. The SOA mass yields from O<sub>3</sub> oxidation of α-pinene modeled for different NO concentrations with the chamber setup.
The model runs were performed using MCM+PRAM.

At low NO<sub>x</sub> concentrations RO<sub>2</sub> radicals undergo rapid autoxidation until they react with HO<sub>2</sub> or RO<sub>2</sub> resulting in production of low volatility hydro-peroxide products (Sarrafzadeh et al., 2016), closed shell monomers or dimers (Ehn et al., 2014; Roldin et al., 2019), which increase SOA mass. This contrasts with high NO<sub>x</sub> conditions where the RO<sub>2</sub>+NO reactions dominate over reactions with HO<sub>2</sub> or RO<sub>2</sub>, resulting in the formation of more volatile products such as aldehydes, ketones and organonitrates (Presto et al., 2005; Sarrafzadeh et al., 2016), and likely suppressing the autoxidation process leading to a decrease in SOA mass loadings (Ehn et al., 2014).

Figure 7 shows the absolute contributions to SOA mass loadings by PRAM and MCM compounds at two different O<sub>3</sub> concentrations of 4 and 100 ppb and varying NO concentrations. The figure shows that with an increase in NO concentrations more than 1 ppb the contribution of PRAM compounds to the particle phase 339 decreases at both 4 and 100 ppb of  $O_3$  concentrations. In PRAM the  $RO_2$  + NO reaction leads either to the formation of organonitrate HOM, closed shell monomers with carbonyl group or fragmentation products with 340 higher volatility (Roldin et al., 2019). HOM Dimer formation is suppressed with increasing NO 341 concentrations in PRAM (Roldin et al., 2019) which explains the lower contribution by PRAM compounds to 342 SOA mass loadings with increasing NO. At NO concentrations <1ppb the PRAM contribution increases as, 343 first generation RO<sub>2</sub> are capable of undergoing autoxidation forming highly oxygenated RO<sub>2</sub> which 344 subsequently reacts with NO forming organic nitrates (Ehn et al., 2014). As NO concentrations exceed 1ppb 345 the first generation RO<sub>2</sub> is scavenged by NO thereby reducing the concentration of organonitrate HOM (Ehn 346 et al., 2014), possibly affecting SOA yields. The MCM contribution also decreases with increasing NO 347 concentrations mostly due to the formation of more volatile organonitrates (Jenkin et al., 2019). 348

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Figure 7. Contribution to the SOA mass loadings by total PRAM and MCM compounds at different NO levels and O<sub>3</sub>
concentrations. For comparison we use 4 ppb and 100 ppb O<sub>3</sub> concentrations, respectively, at 50 ppb α-pinene.

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#### 353 **3.5 Temperature dependence**

The formation of SOA from  $\alpha$ -pinene ozonolysis in the temperature range of 258.15 - 313.15 K was investigated in this study using MCM+PRAM. Strong dependence of SOA mass yield on temperature was reported by Saathoff and Naumann, (2009) wherein they measured the decreasing mass yields from 0.42 at 273.15 K to 0.09 to 313.15 K for SOA loadings of 53 and 92 µg m<sup>-3</sup> respectively. Our results in Figure 8 show increasing SOA mass yields for  $\alpha$ -pinene ozonolysis with decreasing temperature, which is attributed to the augmented condensation of oxidation products termed as semi volatile organic compounds (SVOC) (Kristensen et al., 2017) at lower temperatures.

For  $\alpha$ -pinene maximum mass loading < 150 µg m<sup>-3</sup> the mass yields reach a maximum value of 0.38 at 361 temperatures as low as 258.15 K and decrease to 0.27 for a temperature of 293.15 K and to 0.1 for the 362 temperature of 313.15 K. These yields are comparable to the results obtained by Kristensen et al. 363 (2017) where they measured yields of 0.39 for 258.15 K and 0.22 for 293.15 K for mass loading  $<150 \mu g m^{-3}$ . 364 The results show a weak dependence of SOA mass yields on temperatures in the range of 278.15 K - 313.15 365 K at low SOA mass loadings which become more pronounced as the mass loadings increase. At the lowest 366 temperature of 258.15 K the mass yields are higher in comparison to other temperatures regardless the mass 367 loadings. These results are in good agreement with the findings by Pathak et al. (2007) where they found a 368 strong temperature dependence of SOA mass yields at lower temperature  $(0 - 15^{\circ} \text{ C})$ , which decreases as the 369 temperature increases. Furthermore, similar to the measurements made by Pathak et al. (2007), our 370 simulations were able to reproduce the experimental findings that show no appreciable differences in the SOA 371 mass yields for loadings below 1 µg m<sup>-3</sup> (initial mixing ratio of 1 ppb) for temperatures > 273.15 K. 372



 $\alpha$ -pinene -  $O_3$  Temperature dependence

Figure 8. Temperature dependence of SOA mass yields at different temperatures using the MCM+PRAM. The open pentagons
represent measurement data from Kristensen et al. (2017) at 258.15 K and 298.15 K.

Figure 9 shows the volatility distribution of  $\alpha$ -pinene ozonolysis derived SOA at different 375 temperatures. The saturation vapor pressure limits for defining extremely low volatility (ELVOCs - grey 376 shaded), low volatility (LVOCs - red shaded), semi volatile (SVOCs - green shaded) and intermediate 377 volatility (IVOCs - cyan shaded) organic compounds used in the Volatility basis set (VBS) are set according 378 to the values suggested in Donahue et al. (2012). In this work, we categorize compounds (ELVOCs, LVOCs, 379 SVOCs and IVOCs) based on effective saturation vapor pressures (C\*) in the range of  $\{10^{-5} \text{ to } 10^3\} \mu \text{g m}^{-3}$ 380 and temperature of 298 K (Donahue et al., 2009). At the lowest temperature of 258.15 K, the SVOCs 381 contribution to the particle phase is dominant in comparison to LVOCs and ELVOCs, a trend which is 382 subsequently reversed as the temperatures are increased. At 293.15 K a majority of SVOCs and IVOCs are in 383 the gas phase while the contribution of LVOCs and ELVOCs to particle phases increases. These results are in 384 385 good agreement with observations made by Kristensen et al. (2017) wherein they observed an increasing contribution of SVOCs at sub-zero temperatures of 258.15 K, which decrease the fraction of SOA formed 386

from ELVOCs. Again, it should be noted that the temperature dependence of peroxy radical autoxidation
product formation still needs further validation based on recent experiments (e.g. Quéléver et al., 2019).



Volatility Basis set for  $\alpha$ -pinene (50ppb) -  $O_3$  (50 ppb)

**Figure 9**. Modeled volatility distribution of SOA at different temperatures. The volatility bins span a range of effective saturation vapor pressures  $C = C^* = \{10^{-5} \text{ to } 10^3\} \ \mu \text{g m}^{-3}$ . The VBS distribution is based on a reference temperature of 298 K.

#### 392 **3.6 Composition**

MCM+PRAM can be used to narrow down and compile a list of compounds playing a pivotal role in
 contributing to SOA mass loadings and, also compare the relative importance of implementing PRAM
 alongside the MCM. Figure 10 shows the most important compounds from both the MCM and PRAM that
 together contribute to more than 95% of α-pinene ozonolysis SOA mass loading at 293.15 K.



**Figure 10**. MCM and PRAM Compounds contributing to > 95 % of SOA mass at 293 K and 50ppb  $O_3$  and  $\alpha$ -pinene concentrations.

399 Figure 10 shows that contribution to SOA mass loadings by PRAM compounds is ~48 % (of 97 %) while MCM compounds contribute ~52 % (of 97%). On lowering the temperature to 258K the relative contributions 400 of PRAM drop to 15 % (of ~99 %), while MCM dominates by contributing ~85 % (of ~99 %) respectively 401 (Figure S3a). The contribution of PRAM increases to ~64 % (of ~97 %) and MCM contribution drops to 36 402 % (of ~97 %) at 313 K (Figure S3b). These results reflect the importance of PRAM as its contribution plays 403 an increasingly dominant role with increasing temperatures and highlights the crucial few compounds that 404 contribute to maximum SOA mass loadings for  $\alpha$ -pinene ozonolysis. The list of abundant compounds which 405 together add up to contribute more than 95 % of SOA mass loadings at 258 K, 293 K and 313 K are presented 406

in the supplement Table 1s (a, b & c). At 258 K MCM compounds namely pinonic acid (C<sub>10</sub>H<sub>16</sub>O<sub>3</sub>, 4.4 %),

C920PAN (C<sub>10</sub>H<sub>15</sub>NO<sub>7</sub>, 9.3 %), C108NO3 (C<sub>10</sub>H<sub>15</sub>NO<sub>6</sub>, 8.9 %), C811PAN (C<sub>9</sub>H<sub>13</sub>NO<sub>7</sub>, 10.1 %), C717NO3 408 (C<sub>7</sub>H<sub>9</sub>NO<sub>6</sub>, 11.3 %) contribute significantly to the total SOA mass loadings while PRAM compounds such as 409 C<sub>10</sub>H<sub>14</sub>O<sub>7</sub> (0.88 %), C<sub>10</sub>H<sub>16</sub>O<sub>4</sub> (1.3 %), C<sub>10</sub>H<sub>16</sub>O<sub>6</sub> (1.13 %) contribute significantly less. An increase in 410 temperature to 293 K results in an overall increase in contribution by PRAM compounds, with C<sub>10</sub>H<sub>14</sub>O<sub>10</sub> (3.6 411 412 %),  $C_{10}H_{14}O_{11}$  (6.2 %),  $C_{10}H_{16}O_{10}$  (3.2 %) playing an important role in contributing to the SOA mass loadings. This trend of relative increase in the contribution by PRAM compounds over MCM compounds to SOA mass 413 loadings is also evident as the temperatures are further increased to 313 K, where the PRAM compounds 414  $C_{10}H_{14}O_{11}$  (18.3 %),  $C_{10}H_{14}O_{12}$  (6 %) and  $C_{10}H_{16}O_{12}$  (6.6 %) play a dominant role in increasing SOA mass 415 loadings. 416

# 417 4. Conclusions

We simulated SOA mass yields derived from the oxidation of various BVOCs (isoprene, α-pinene, βpinene, limonene and β-caryophyllene), by the oxidants O<sub>3</sub>, OH and NO<sub>3</sub> using the zero-dimensional model
MALTE-Box. The gas phase chemistry was simulated using the MCM in conjunction with PRAM. The aim
was to verify the efficacy of MCM+PRAM in simulating the SOA mass yields. Additional simulations were
performed to test the MCM+PRAM under varying temperature and NO concentrations. A few important
compounds playing a major role in increasing the SOA mass yields for α-pinene ozonolysis at different
temperatures are also highlighted.

The simulations were designed to resemble ideal smog chambers experiments and experiments in 425 oxidative flow reactors (OFR). No interactions between the gas phase and chamber walls were considered 426 427 during the simulations. For the smog chamber setting, the standalone MCM generally under-predicts the mass yields obtained by the ozonolysis and OH oxidation of BVOCs. In contrast, the yields derived using 428 MCM+PRAM for the smog chamber setup is in good agreement with the experimental results. For an 429 idealized OFR setup, MCM+PRAM yields are in good agreement with experimental yields, while again the 430 431 MCM under-predicts the SOA yields. The relative contribution of HOM monomers and dimers to the particle phase in OFR simulations is low when compared to the chamber simulations. This is due to higher RO<sub>2</sub> 432 433 concentrations in OFR leading to termination of peroxy radical autoxidation, thereby affecting SOA yields. This needs to be considered when applying yields based on OFR simulations in regional or global chemical 434 435 transport models

The model does not simulate appreciable SOA mass yields for oxidation of BVOCs with NO<sub>3</sub>, as
PRAM currently does not consider autoxidation of RO<sub>2</sub> formed from NO<sub>3</sub> oxidation of VOCs. This underlines
the need for developing a NO<sub>3</sub> oxidation scheme which can better constrain and predict SOA mass yields. In

439 accordance to the previous studies, the simulated SOA yields tend to decrease at higher temperatures. The PRAM contribution to mass yields at low temperatures (258.15 K) is ~14 %, which is substantially lower than 440 that of MCM (~86 %). As the temperature is increased to 313.15 K, the contribution of PRAM to SOA mass 441 yields begins to dominate over MCM. This most likely is due to MCM producing more SVOCs (compounds 442 classified as SVOCs at 298 K), which show stronger contribution to particle phase at lower temperatures, due 443 to decrease in saturation vapor pressures with temperature. It should be noted that the present temperature 444 dependency of mass yields using PRAM are a first, and currently the best estimate in understanding the 445 influence of temperature on the peroxy radical autoxidation formation. The simulated SOA yields with 446 varying NO concentrations agree well with experimental results, i.e. SOA yields decrease with increasing NO 447 concentrations due to the formation of more volatile compounds such as organonitrates and ketones. 448

449 Using PRAM coupled with MCM helps us bridge the gap in understanding the role and contribution of peroxy radical autoxidation to SOA formation. The variation of SOA yields for temperature and NO 450 concentrations, indicates the limitations of global and regional models in predicting e.g. cloud condensation 451 nuclei (CCN) effects using fixed SOA yields. The good agreement of modeled and experimental yields from 452 453 smog chambers, could further help us parameterize the SOA yields, that could be applied at a global and regional model scale, to more accurately predict the direct and indirect impact of aerosol particles on e.g. 454 radiation balance by aerosol scattering/absorption and CCN concentrations. Furthermore, implementation of a 455 condensed PRAM version to regional and global models has been tested but still need further validation 456 (Roldin et al., 2019). 457

#### 458 Data availability

The complete PRAM mechanism written in a format compatible with the Kinetic PreProcessor (KPP) together with all species information can also be downloaded from <u>https://doi.org/10.1594/PANGAEA.905102</u>

462 **Author Contributions** 

CX and MB served as the chief authors and editors of the paper. CX was performing the model simulations.
The study was designed by CX, MB and PR. All other co-authors contributed to the analysis and writing of
the paper.

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