

1 Comments

2 We thank the reviewers and the co-editor for their valuable comments. The replies are in blue.

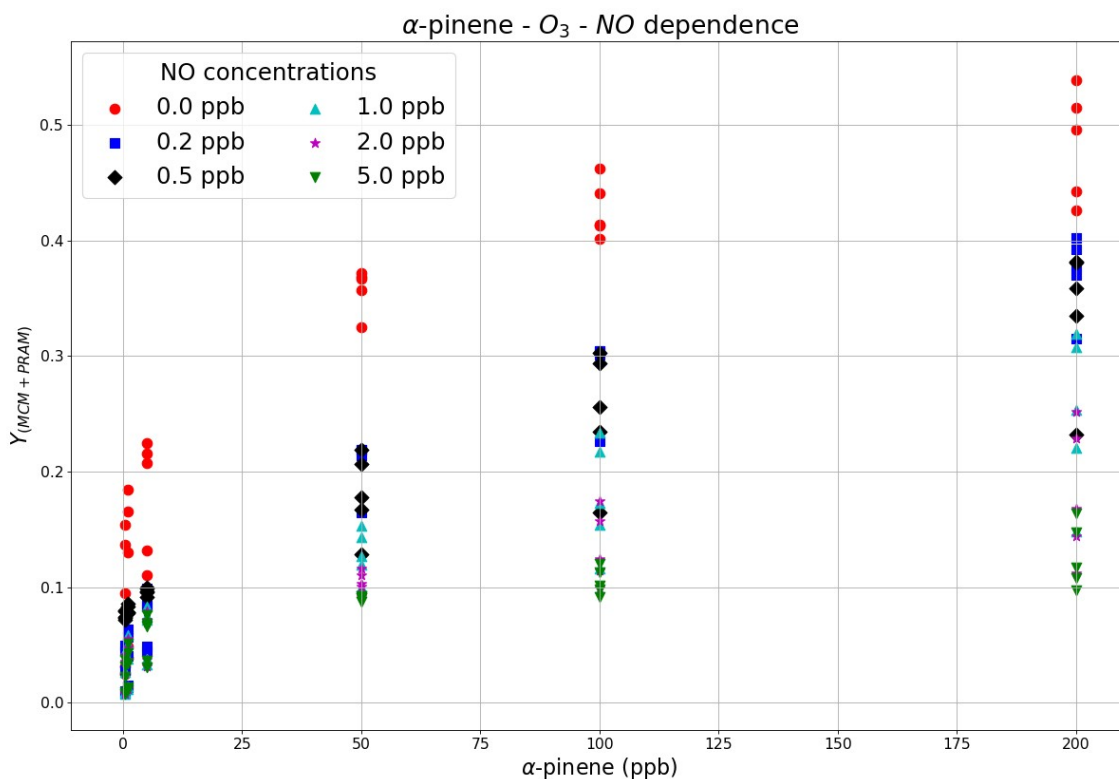
3 **General comment:**

4 R1. Many figures show yields vs SOA mass loading. I assume that the SOA mass loading corresponds to the
5 resulting SOA using different concentrations of precursors (alpha-pinene etc), and not to seed aerosol loading.
6 – Is this correct?

7 **Reply.** Yes. The SOA mass loading corresponds to the SOA mass formed when the different concentrations
8 of precursor gas is oxidized and not the seed aerosol loadings.

9 R2. Could you add to the figures the information what precursor concentrations have been used to obtain the
10 SOA mass yields? For example, I assume that the different ‘branches’ in Figure 6 correspond to different
11 initial alpha-pinene concentrations (0.5 – 200 ppb)? Would the figure look the same if you plotted yield vs
12 initial precursor concentration? – I’m not suggesting presenting all figures like this, but I’m just trying to
13 understand it.

14 **Reply.** The information of initial concentrations of precursor gas has been added to the figure 6 caption. The
15 different branches in Figure 6 correspond to different NO concentrations which is varied between 0-5 ppb. So
16 each branch represents mass loadings for varying α -pinene and O₃ concentrations but constant NO
17 concentrations. For example, the black diamonds correspond to mass yields from ozonolysis of different
18 concentrations of α -pinene (0.5-200ppb) and constant NO concentration of 0.5ppb. If the figure is plotted
19 with initial precursor concentration on the x-axes we get the following plot. For each concentration of α -
20 pinene (eg 5ppb) we have 5 yield points corresponding to 5 values of O₃ concentrations (1.0, 5.0,10.0,
21 50.0,100.0 in the unit of 10¹¹ #cm⁻³).



34 **Figure C1.** Concentration vs Mass yields

35 R3. Please use consistent descriptions on the y-axes (either Y or 'mass yields')

36 [Reply. Done.](#) The figures descriptions on the y axes are now consistent (Figure 6 and 8).

37 l. 65: Define MW

38 [Reply. Done.](#)

39 l. 113: Why is Table 1c cited before the other Table parts? – I think it is rather unusual anyway to have
40 multiple Table parts (a, b, c), but I leave it up to you to or to the copy editing service to change it or not.

41 [Reply. Changed.](#) The Table 1c is now cited later in a more appropriate context in the text. We follows the
42 citing sequence of Table 1a,b and c.

43 l. 209: The x-axis of the figure seems to only show a range of < 100 (~800?).

44 [Reply. Yes, That was a typo in the text.](#)

45 l. 232: replace 'value' by 'yield'

46 [Reply. Done.](#)

47 l. 252: 'ozonolysis' misspelled

48 [Reply. Corrected](#)

49 l. 268: Add the value measured by Lee et al (0.58?)

50 [Reply. Done.](#)

51 l. 281: Do you mean 'Figure 3'? (Fig. 2 shows ozone oxidation.)

52 [Reply. Yes. Corrected.](#)

53 l. 313: Shouldn't it just read '... from application of MCM' as PRAM is not available for NO₃ reactions? – If
54 so, please also change the header of the figure.

55 [Reply. Done.](#)

56 l. 400: Shouldn't that be 99% (as in the SI, you list 98.59%)?

57 [Reply. Yes. Changed.](#)

58 l. 567: Please replace by citation of ACP (not ACPD) paper.

59 [Reply. Done.](#)

60 l. 622: Reference is incomplete

61 [Reply. Changed.](#)

62 l. 644: Update reference.

63 [Reply. Done.](#)

64 l. 685: Incomplete.

65 [Reply. Changed.](#)

66 **Supplement material**

67 Table 1s a- c: Add to the table captions that it is for the oxidation of alpha-pinene by ozone

68 [Reply. Done.](#)

69 l. 28, 34 and 41: Should be called 'Figure 4s'

70 Reply. Yes. Changed.

71 l. 36: The reference to Figure 9 seems wrong here. Do you mean Figure 4s?

72 Reply. Yes. Changed.

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91 **Aerosol Mass yields of selected Biogenic Volatile Organic Compounds – a**
92 **theoretical study with near explicit gas-phase chemistry**

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97

98 **Abstract**

99 In this study we modeled secondary organic aerosols (SOA) mass loadings from the oxidation (by O₃,
100 OH and NO₃) of five representative Biogenic Volatile Organic compounds (BVOCs): isoprene, endocyclic
101 bond containing monoterpenes (α -pinene and limonene), exocyclic double bond compound (β -pinene) and a
102 sesquiterpene (β -caryophyllene). The simulations were designed to replicate idealized smog chamber and
103 oxidative flow reactors (OFR). The master chemical mechanism (MCM) together with the peroxy radical
104 autoxidation mechanism (PRAM), were used to simulate the gas-phase chemistry. The aim of this study was
105 to compare the potency of MCM and MCM+PRAM in predicting SOA formation. SOA yields were in good
106 agreement with experimental values for chamber simulations when MCM+PRAM was applied, while a
107 standalone MCM under-predicted the SOA yields. Compared to experimental yields, the OFR simulations
108 using MCM+PRAM yields were in good agreement for BVOCs oxidized by both O₃ and OH. On the other
109 hand, a standalone MCM under-predicted the SOA mass yields. SOA yields increased with decreasing
110 temperatures and NO concentrations and vice-versa. This highlights the limitations posed when using fixed
111 SOA yields in a majority of global and regional models. Few compounds that play a crucial role (>95% of
112 mass load) in contributing to SOA mass increase (using MCM+PRAM) are identified. The results further
113 emphasized that incorporating PRAM in conjunction with MCM does improve SOA mass yields estimation.

114 **1. Introduction**

115 Atmospheric secondary organic aerosols, formed from gas to particle phase conversion of the
116 oxidation products of volatile organic compounds (VOC) significantly impact the organic aerosol mass
117 loadings (Griffin, 1999; Kanakidou et al., 2005). However, the scale of SOA contribution to the aerosol
118 particle mass is still subject to high uncertainties (Hao et al., 2011, Glasius and Goldstein, 2016). The elevated
119 aerosol particle concentrations are shown to have inimical effects on health (Miller et al., 2007), and a varying
120 degree of influence on the climate by forming cloud condensation nuclei (CCN), altering the cloud properties

121 and radiative balance (Rosenfeld et al., 2014; Schmale et al., 2018). Therefore, it is acutely necessary to
122 understand the role and contributions of SOA to the particle loading in the atmosphere. Biogenic VOCs from
123 forest are estimated to contribute to about 90% of VOCs emissions globally (Guenther et al., 1995, 1999 and
124 2000). The most important BVOCs for SOA formation are isoprene (C_5H_8), monoterpenes ($C_{10}H_{16}$) and
125 sesquiterpenes ($C_{15}H_{24}$). These compounds are all alkenes containing at least one carbon-carbon double bond,
126 enabling them to undergo oxidation by the dominant atmospheric oxidants: the hydroxyl radical (OH), ozone
127 (O_3) and the nitrate radical (NO_3). For some of the terpenes, initial oxidation steps can lead to formation of
128 highly oxygenated organic molecules (HOM). These HOMs generally have low volatilities and can condense
129 nearly irreversibly, thereby producing SOA (Ehn et al., 2014). HOMs, detected in both the ambient
130 atmosphere and chamber experiments (Ehn et al., 2012) are formed by autoxidation (Berndt et al., 2016;
131 Crouse and Nielsen, 2013) wherein peroxy radicals (RO_2) undergo subsequent intramolecular H-shifts
132 accompanied by rapid reactions with O_2 . Autoxidation hence results in compounds containing multiple
133 functional groups such as hydroxyls, peroxides and carbonyls (Bianchi et al., 2017, Bianchi et al., 2019).

134 A majority of chamber and flow-tube experiments have focused on HOM formation from the
135 oxidation of various VOCs and their contribution to SOA mass loadings (Ehn et al., 2014; Kristensen et al.,
136 2017). Oxidation of isoprene (Liu et al., 2016), endocyclic monoterpenes containing reactive double bonds
137 such as α -pinene and limonene (Zhao et al., 2015), or exocyclic double bond containing compounds such as
138 β -pinene (Jokinen et al., 2015) and sesquiterpenes such as β -caryophyllene (Chen et al., 2012) have been
139 investigated. The SOA forming potential of various BVOCs depends on the isomeric structures (Friedman
140 and Farmer, 2018; Keywood et al., 2004). Ozonolysis of compounds containing reactive endocyclic bonds
141 such as α -pinene produce higher SOA mass yields of 41% in comparison to those with exocyclic bonds (β -
142 pinene), which produce mass yields of 17% (Lee et al., 2006a). One explanation for this dependence on the
143 isomeric structure is attributed to the formation of HOMs (Ehn et al., 2014). Another important factor
144 influencing HOM formation is the initial oxidant, as pointed out by Zhao and co-workers (2015). They
145 showed that the SOA formation by OH oxidation of α -pinene and limonene were lower when compared to
146 their SOA formed by ozonolysis. Further they measured lower H/C ratio for SOA produced by monoterpene
147 ozonolysis (experiments were carried out in dark with CO as OH scavenger), in comparison to OH oxidation
148 of α -pinene and limonene, while O/C ratio were similar for both oxidation cases. This was attributed to the
149 formation of RO_2 radicals (monoterpenes + O_3) which undergo internal hydrogen shifts and subsequently react
150 with another RO_2 radical, to form compounds containing carbonyl groups while losing hydrogen atoms in the
151 process. A similar analysis was conducted by Draper et al. (2015), who showed that an increase in NO_2
152 concentration reduced α -pinene ozonolysis SOA mass yields, while no appreciable reduction in mass yields

153 are reported for β -pinene and Δ^3 -carene ozonolysis. On the other hand, the mass yields from limonene
154 ozonolysis increased with increasing NO_2 concentrations (Draper et al., 2015). This disparity in mass yields
155 for different BVOCs in the presence of NO_2 is possibly caused by the formation of high molecular weight
156 oligomers (or lack of in case of α -pinene) through oxidation with NO_3 that contribute to SOA mass loadings
157 (Draper et al., 2015).

158 Due to computational limitations, many regional and canopy scale atmospheric chemistry models
159 generally use isoprene and/or a representative monoterpene (generally α -pinene), to model SOA yields
160 (Friedman and Farmer, 2018). The SOA yields of different monoterpenes vary with structure, NO_x and
161 temperature (Friedman and Farmer, 2018; Kristensen et al., 2017; Presto et al., 2005). This poses a limitation
162 on using representative monoterpene fixed SOA yields in many of the global models and increases
163 uncertainties in predicting cloud condensation nuclei concentrations, cloud droplet number concentrations and
164 radiative balance due to aerosol loadings.

165 This work aims to investigate the SOA mass loading from the oxidation products of BVOCs with the
166 atmospheric oxidants OH, O_3 and NO_3 with a specific focus on the BVOCs isoprene, α -pinene, β -pinene,
167 limonene and β -caryophyllene. Further we study the effect of varying temperature (258.15 K – 313.15 K) and
168 NO concentrations (0 - 5 ppb) on α -pinene oxidation mass yields. We use the master chemical mechanism
169 (MCMv3.3.1) (Jenkin et al., 1997, 2012 and 2015; Saunders et al., 2003), a near explicit gas-phase chemical
170 mechanism together with peroxy radical autoxidation mechanism (PRAM, Roldin et al., 2019) (PRAM +
171 MCM). The aim is to understand the importance and contribution of peroxy radical autoxidation products to
172 the SOA mass yields from terpenes.

173

174 2. Model description

175 2.1 Malte Box

176 MALTE (Model to predict new Aerosol formation in Lower TropospherE) is a one-dimensional
177 model consisting of modules calculating boundary layer meteorology, emissions of BVOCs, gas-phase
178 chemistry and aerosol dynamics with the aim to simulate particle distribution and growth in the lower
179 troposphere (Boy et al., 2006). In this study, a zero-dimensional version, MALTE-Box is applied to simulate
180 an ideal chamber and flow-tube environment (i.e. no wall losses effects are considered in this study). For the
181 simulations performed in this study the emission module was switched off while only employing the gas-
182 phase chemistry and aerosol dynamics module.

183 Kinetic preprocessor (KPP) is used to generate a system of coupled differential equations to solve the
184 gas-phase chemistry schemes (Damian et al., 2002). The peroxy radical autoxidation mechanism (PRAM),
185 (Roldin et al., 2019, Qi et al., 2018, Öström et al., 2017), formulated based on the oxidation of monoterpenes
186 as described by Ehn et al. (2014) was incorporated alongside MCMv3.3.1. PRAM explicitly describes the
187 formation and evolution of peroxy radicals (RO_2) from the ozonolysis and OH oxidation of monoterpenes,
188 driven by subsequent H-shifts and O_2 additions. The current version of PRAM based on experimental and
189 theoretical studies, considers HOM autoxidation for a fraction of the peroxy radicals formed during the
190 ozonolysis of α -pinene and limonene and OH oxidation of α -pinene, β -pinene and limonene. This is achieved
191 by assigning species specific molar yields for the formation of first RO_2 , which subsequently initiates the
192 autoxidation chain (Roldin et al., 2019). Currently, in PRAM a maximum first generation RO_2 yield of 9% for
193 α -pinene ozonolysis, 21.9 % for limonene ozonolysis, 2.5 % for α -pinene+OH, and 1% for both
194 limonene+OH and β -pinene+OH first generation products are allowed to initiate autoxidation (Roldin et al.,
195 2019). For β -pinene ozonolysis the molar yield of RO_2 is minor (<0.1 %) (Roldin et al., 2019, Ehn et al. 2014)
196 and hence not considered in this work. The above mentioned RO_2 molar yields used in this work are close to
197 the experimental values obtained in both smog chamber and flow tube experiments. Ehn et al. (2014)
198 measured an RO_2 yield of $\sim 7\%$ for α -pinene ozonolysis and $\sim 17\%$ for limonene ozonolysis, whereas Jokinen
199 et al. (2015) measured 0.58 % and 0.93 % for OH oxidation of β -pinene and limonene respectively. The
200 autoxidation is terminated by bimolecular reactions, wherein the RO_2 formed reacts with NO, HO_2 or other
201 peroxy radicals, thereby forming alkoxy radicals, closed shell monomers or dimers (Roldin et al., 2019). The
202 PRAM considers temperature dependent autoxidation reaction rates, which is important when investigating
203 the SOA mass yields at varying temperatures (Table 1e). The temperature dependence in PRAM is based on
204 quantum chemical calculations wherein the autoxidation rates correspond to an activation energy of 24
205 kcal/mol. The activation energies vary for autoxidation of different RO_2 from α -pinene ozonolysis between 22
206 and 29 kcal/mol (Rissanen et al., 2015), leading to varying autoxidation rates at different temperatures
207 (Roldin et al., 2019). It should be noted that the temperature dependence in PRAM is a first of its kind but
208 needs further evaluation using recent measurements of HOM formation at different temperatures (e.g.
209 Quéléver et al., 2019).

210 The aerosol dynamics are simulated using the University of Helsinki Multicomponent Aerosol model
211 (UHMA) originally from Korhonen et al. (2004). The model has undergone significant development since
212 then to allow simulation with all the compounds from MCM. It now supports an unlimited number of
213 condensing vapors and solves condensation using the analytical predictor of condensation method from
214 Jacobson (1997). The condensation algorithm considers both, the Kelvin effect and Raoult's law. The

215 processes included in the model are nucleation, condensation, evaporation, coagulation and deposition. The
216 discretization of the size distribution and the time evolution is modeled with the moving section approach,
217 with optional redistribution to a fixed grid. In this work, the redistribution is active to make the coagulation
218 more accurate, since it requires that grid points are available near the size of the coagulated particles. In this
219 study nucleation and deposition are not active, and hence are not considered. A total of 100 size bins ranging
220 from 1nm to 20 μm with the fixed grid was applied for this study.

221 A group contribution method based on Nannoolal et al. (2008) using the UManSysProp online system
222 (Topping et al., 2016) was used to estimate the pure liquid saturation vapor pressures (p_0) of the organic
223 compounds in MCMv3.3.1. For the PRAM species, p_0 were estimated using the functional group method
224 SIMPOL (Pankow and Asher, 2008; see Roldin et al., 2019 for details). Temperature was used as an input to
225 estimate p_0 for both the group contribution methods.

226 2.2 Simulations

227 The simulations performed in this study are aimed to closely resemble an idealized smog chamber
228 (batch mode setup) and an Oxidative Flow Reactor (OFR) without interactions between the gas phase and the
229 system walls. For the chamber runs, the VOC and oxidants were introduced at the beginning (time, $t=0$ sec),
230 set to certain concentrations (Table 1a) and then allowed to react. Both chamber and OFR simulations are
231 performed using ammonium sulfate seed particles which are introduced at time $t=0$. The condensation sink
232 (CS) was inferred from the size distribution of seed particles used in the model. The CS for the chamber and
233 OFR simulations was set to 0.00067 s^{-1} and 0.067 s^{-1} respectively. SOA mass yields obtained using an OFR
234 are sensitive to short residence time used, hence the seed particle surface area should be chosen in order to
235 overcome the mass yield underestimation (Ahlberg et al., 2019). CS sensitivity runs (Supplement Figure S1)
236 were performed for α -pinene- O_3 to determine the CS for which there are no appreciable change in mass yields
237 with increasing particle surface.

238 The simulation for the chamber setup is run for a maximum time of 24 hours and ends when either of
239 the 2 criteria are satisfied: (1) the simulation time reaches the 24-hour mark or (2) 90 % of the initial
240 precursor VOC has reacted away. In the latter case the simulation is continued for an additional 2 hours to
241 ensure enough time for the vapors to condense onto the seed particles. By contrast, the OFR runs were
242 simulated for a maximum residence time of 100 seconds, ensuring all initial precursor vapors were oxidized.
243 Seed particles were also added in the OFR simulations. The oxidant concentrations used for the OFR
244 simulations (Table 1b) are significantly higher in comparison to the simulated chamber runs (~ 2 orders of
245 magnitude larger). The time step for the chamber and flow-tube simulations are set to $t=10 \text{ s}$ and $t=0.1 \text{ s}$

246 respectively. The runs performed were oxidant specific (i.e. VOCs would be oxidized by only one specific
 247 oxidant at any given time). For the O₃ specific simulations no OH could form in both, OFR and chamber
 248 setups, thus enabling oxidation of O₃ to be the only pathway.

249 The simulations were performed at atmospheric relevant NO_x (NO_x = NO +NO₂) concentrations,
 250 corresponding to [NO]=0.5 ppb and [NO₂] = 2.0 ppb conditions with the relative humidity (RH) set to 60 %
 251 and temperature to 293.15 K. The RH value considered in this study is based on previous published
 252 experimental studies performed at ~60 % in both smog chamber (Bruns et al., 2015a; Ehn et al., 2014;
 253 Stirnweis et al., 2017) and OFR (Ahlberg et al. et al., 2018). α-pinene ozonolysis runs were performed at four
 254 different temperatures: 258.15 K, 278.15 K, 303.15K and 313.15 K, respectively (Table 1c) . SOA mass
 255 yields are expected to increase with decreasing temperature (Saathoff and Naumann, 2009). A similar
 256 temperature dependence was observed by Kristensen et al. (2017) who observed SOA mass yield from α-
 257 pinene ozonolysis at ~ 40 % and ~20 % at 258 K and 293 K respectively. Analogous to analyzing the effect of
 258 varying temperature on SOA yields, we study the variation in α-pinene ozonolysis SOA mass yields by
 259 varying the NO_x concentrations (Table 1c) . SOA yields for α-pinene ozonolysis at high NO_x conditions
 260 should be suppressed (Ng and Chhabra, 2007), which could be due to the production of relatively, volatile
 261 organic nitrates under high NO_x conditions as compared to less volatile products during low NO_x conditions
 262 (Presto et al., 2005).

263 Furthermore, two different chemistry schemes were applied for the simulations. One scheme consisted
 264 of only the MCM chemistry mechanism and the second included the MCM+PRAM chemistry mechanism.
 265 Table 1a shows the concentrations of different BVOCs and Table 1b shows the oxidants concentrations used
 266 for the simulations.

267 **Table 1a.** Concentrations of different BVOCs

α-pinene (ppb)	β-pinene (ppb)	Isoprene (ppb)	Limonene (ppb)	B-caryophyllene (ppb)
0.5, 1.0, 5.0, 50.0, 100.0, 200.0	0.5, 1.0, 5.0, 50.0, 100.0, 200.0	5.0, 50.0, 100.0, 200.0	1.0, 5.0, 50.0, 100.0, 200.0	0.5, 1.0, 2.0, 5.0, 10.0

268

269 **Table 1b.** Concentrations of different oxidants for chamber and flow-tube runs

OH (* 10 ⁶ #/cm ³) - chamber	O ₃ (* 10 ¹¹ #/cm ³) - chamber	NO ₃ (* 10 ⁷ #/cm ³) - chamber
OH (* 10 ⁸ #/cm ³) - OFR	O ₃ (* 10 ¹³ #/cm ³) - OFR	NO ₃ (* 10 ⁹ #/cm ³) - OFR

2.0, 5.0 ,10.0, 50.0,100.0	1.0, 5.0 ,10.0, 50.0,100.0	1.0, 5.0 ,10.0, 50.0,100.0
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270

271 **Table 1c.** NO concentrations and temperatures used for α -pinene ozonolysis

NO (ppb)	0.5 (default), 0, 0.2, 1, 2, 5
Temperature (K)	293.15 (default), 258.15, 278.15, 303.15, 313.15

272

273 **2.3 Mass Yields**

274 The SOA mass yields (Y) are determined by calculating the ratio of the amount of SOA or mass
275 concentration of organic aerosol formed (C_{OA}) to the amount of VOC (ΔVOC) reacted:

276

$$Y = \frac{C_{OA}}{\Delta VOC} \quad (1)$$

277

278 A volatility basis set is fit to the data to obtain the volatility distribution. In this study equilibrium
279 partitioning was only assumed for deriving the volatility distribution based on the model simulations.

280 Following Donahue et al. (2006), the SOA is assumed to be in equilibrium with the gas-phase and using the
281 effective saturation concentration C_i^* spaced logarithmically. The individual product partitioning to the

282 particle phase can be estimated using

282

$$E_i = \left(1 + \frac{C_i^*}{C_{OA}} \right)^{-1} \quad (2)$$

283

284 Where E_i is the fraction of species in the condensed particle phase. The above equation determines the
285 fraction of species in the particle phase as well as in the gas phase. For example, if we assume $C_{OA} = 10 \mu\text{g m}^{-3}$

286 a species with $C^* = 10 \mu\text{g m}^{-3}$ will partition 50 % to condensed phase and the rest 50% will reside in the gas
287 phase. The fidelity of this equilibrium partitioning enables the parameterization of product vapors in volatility

288 C^* bins that are near the C_{OA} concentrations (Henry et al., 2012).

287

288 **3. Results and Discussion**

289

290 SOA mass yields were simulated for the oxidation of various biogenic volatile organic compounds
291 (isoprene, α -pinene, limonene and β -caryophyllene, β -pinene) by dominant atmospheric oxidants OH, O_3 and
292 NO_3 . The following section examines the comparison between the yields derived using MCM+PRAM and a
standalone MCM for chamber and flow-tube experiments.

293

294 3.1 BVOCs – O₃ chamber and flow-tube simulations

295 In Fig. 1 panel A indicates the SOA mass yields derived on applying a coupled MCM+PRAM mechanism to
296 ozonolysis of α -pinene and limonene (PRAM is only available for ozonolysis of α -pinene and limonene) and
297 the lower panel B shows ratio of yields obtained by MCM and coupled MCM+PRAM.

298 The abscissa, depicted on a log scale, considers the entire range of SOA mass loadings from 1-~~1150~~
299 **800** $\mu\text{g m}^{-3}$. Each data point is representative of simulated SOA mass yields resulting from variable BVOC
300 loading. The resulting mass yields for α -pinene in the range shown in Table 2a. are consistent with the yields
301 found in various smog chamber experiments. The mass yields derived using MCM+PRAM for α -pinene
302 ozonolysis are in good agreement with the experimental yields measured for similar mass loadings by
303 Kristensen et al. (2017) and Pathak et al. (2007). The standalone MCM, on the other hand, severely under-
304 predicts the mass yields for α -pinene ozonolysis. The MCM+PRAM also shows better agreement with
305 experiments when estimating the lower range mass yields for SOA mass loadings of $< 15 \mu\text{g m}^{-3}$. This is
306 supported by the values obtained by Shilling et al. (2008), where the authors measured a 0.09 yield from α -
307 pinene ozonolysis for SOA mass loading of $10.6 \mu\text{g m}^{-3}$. Limonene ozonolysis mass yields using
308 MCM+PRAM in comparison to standalone MCM, are much closer to the values given by Waring (2016).

309 The formation of HOM from β -pinene ozonolysis is low (Ehn et al., 2014; Jokinen et al., 2015) and
310 hence not considered in PRAM. The peroxy radical autoxidation mechanism for β -caryophyllene ozonolysis
311 has not yet been developed and therefore, not considered in PRAM. When comparing the measured mass
312 yield values for β -caryophyllene (Chen et al. 2012) and β -pinene ozonolysis (Griffin (1999) and Pathak et al.
313 (2008)) to the modeled values using the MCM scheme, it is evident that the MCM scheme drastically under-
314 predicts the SOA mass yields (Fig. 2).

315 Today oxidation flow reactor (OFR) experiments are complementing the traditional batch mode smog
316 chamber experiments. The OFR generally exhibits lower mass yields compared to the smog chamber
317 experiments at ranges of equivalent oxidant exposure (Lambe et al., 2015). We modeled flow-tube simulation
318 after the potential aerosol mass (PAM) OFR, where the residence time is in the order of a few to several
319 minutes (Lambe et al., 2011). The model simulations are performed with a maximum residence time of 100
320 seconds with O₃ exposures ranging from $1.0 \times 10^{15} - 1.0 \times 10^{17}$ molecules $\text{cm}^{-3} \text{s}$ (residence time * [O₃]).
321 Kang and Root (2007) measured a ~~value~~ **yield** of 0.2 for ozonolysis of α -pinene for an initial precursor VOC
322 concentration of 100 ppbv, while we obtain ~ 0.25 (MCM+PRAM) for the similar initial precursor
323 concentrations. The OFR yields for β -pinene (MCM-only) are significantly lower (0.02) than the values

324 measured by Kang and Root (2007) wherein they measured a yield of 0.49 for similar initial precursor
 325 concentrations. Addition of seed particles promotes condensation, leading to increased SOA yields (Lambe et
 326 al., 2015) which was confirmed by Ahlberg et al, (2019). Kang and Root (2007) found that using seed
 327 particles, the yield from α -pinene ozonolysis increased by a factor of ~ 1.4 which can explain our yields for α -
 328 pinene ozonolysis simulations. The mass spectra plot (Figure S2) shows that PRAM contributes the majority
 329 of dimers to the particle phase, while MCM dominate monomer contribution. Another interesting facet of
 330 Figure S2 are the different condensing compounds in both OFR and chamber simulations. The higher absolute
 331 RO_2 concentrations in the OFR simulations explain the lower concentration of HOM monomers and dimers
 332 relative to the chamber simulations, i.e. the high RO_2 concentrations in the OFR cause termination of the
 333 peroxy radical autoxidation chain before the RO_2 become highly oxygenated, thereby influencing SOA yields.
 334 Hence, this should be taken into account when using yields from OFR as inputs to regional and global
 335 models.

Table 2a. Mass yields for BVOCs ozonolysis at 293 K for different range of mass loadings using a chamber[†]
 setup. The values in parenthesis in the column ‘Experimental yields’ indicates the corresponding experimental
 mass loadings in unit of $\mu\text{g m}^{-3}$.

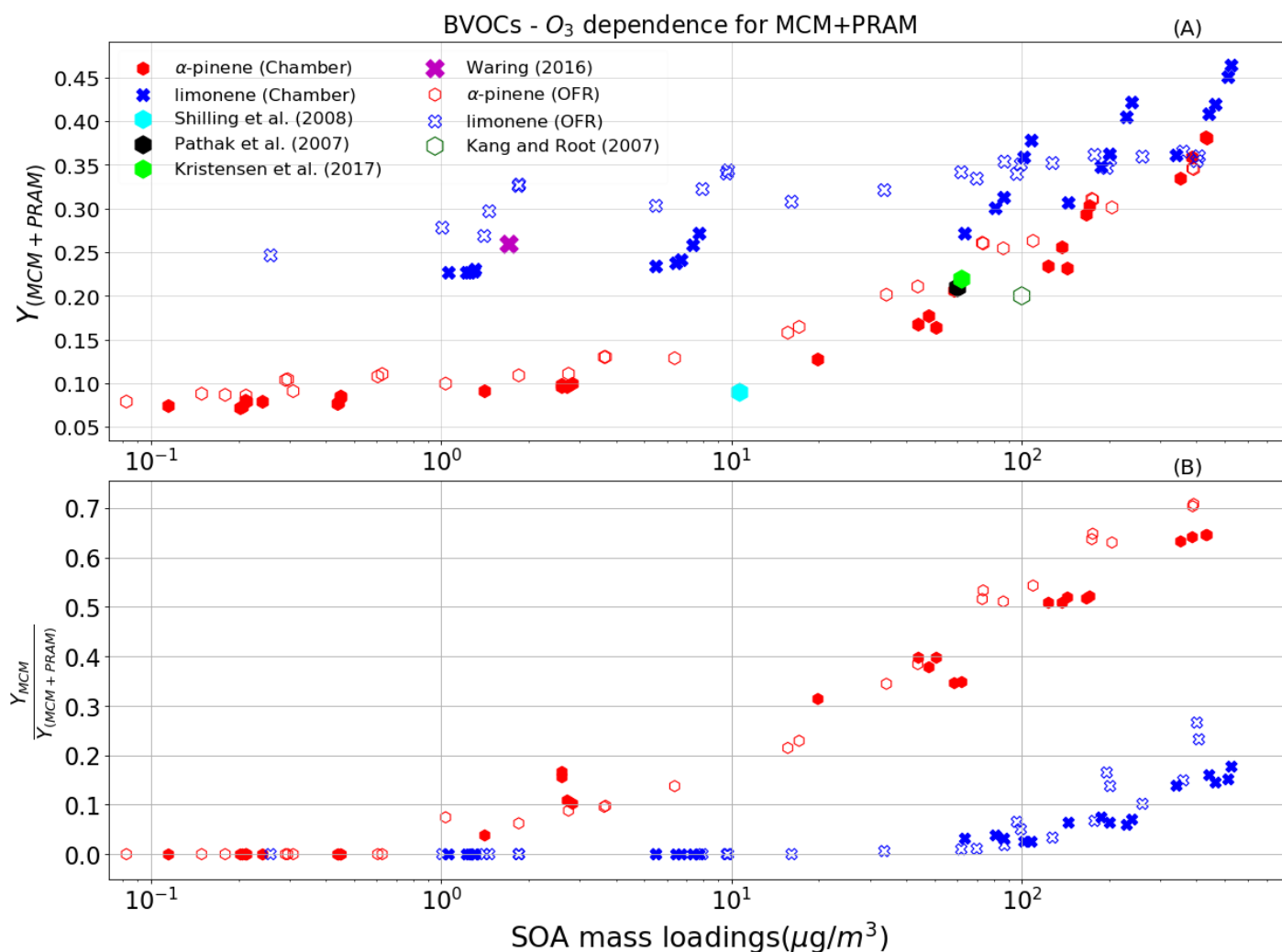
SOA mass loading ($\mu\text{g m}^{-3}$)	MCM + PRAM mass yields range	MCM mass yields range	BVOC	Experimental yields	References
0– 15 [†]	0.07– 0.08	0.00 – 0.06	α -pinene	0.09 (10.6)	Shilling et al. (2008)
16 - 60 [†]	0.12 – 0.20	0.06 – 0.11	α -pinene	0.16 – 0.21 (15 -60)	Pathak et al. (2007)
61 – 200 [†]	0.22 – 0.30	0.12 – 0.15	α -pinene	0.22 (62)	Kristensen et al. (2017)
1.1– 550 [†]	0.24 -0.48	0.007-0.06	limonene	0.26 (1.7)	Waring (2016)
0 - 100 [†]	0 – 0.09 ^{!!}	0 – 0.09	β -pinene	0.03-0.22 (7.2 - 100)	Griffin (1999)
0 -10 [†]	0 – 0.01 ^{!!}	0 – 0.01	β -caryophyllene	0.13 (1.8)	Chen et al. (2012)

336 ^{!!}indicates that no PRAM mechanism available yet i.e the yields are same as the MCM yields.

Table 2b. Mass yields for BVOCs ozonolysis at 293 K for different range of mass loadings using an OFR^{||} setup.

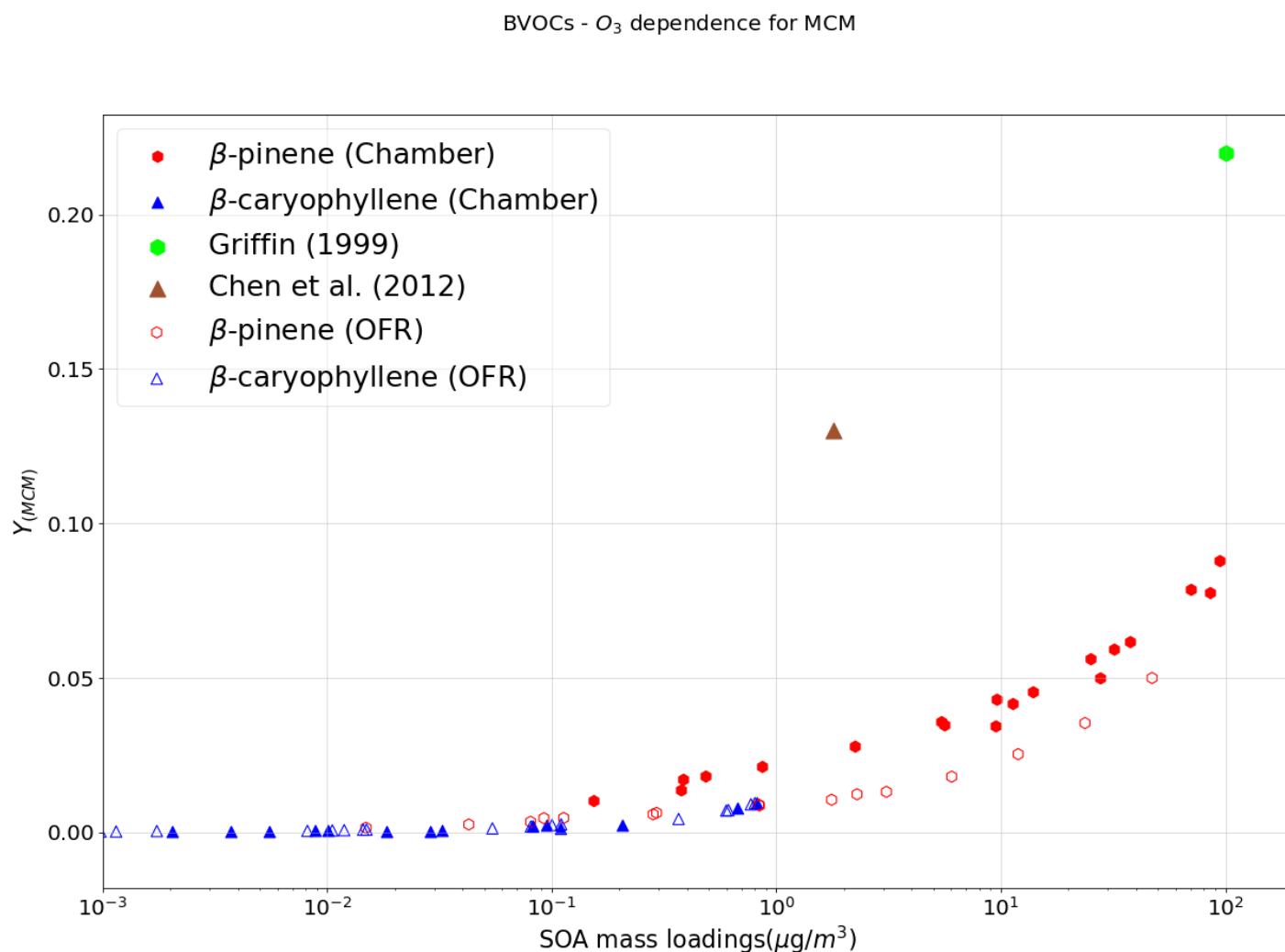
SOA mass loading (ppb)	MCM + PRAM mass yields range	MCM mass yields range	BVOC	Experimental yields	References
0-100	0.07-0.25	0-0.13	α -pinene	0.2 (100)	Kang and Root (2007)
0-156	0 – 0.02	0 – 0.02	β -pinene	0.49 (156)	Kang and Root (2007)

337 ^{||}indicates that no PRAM mechanism available yet i.e the yields are same as the MCM yields.



338 **Figure 1.** The mass yields from the ozonolysis of BVOCs α -pinene (red heptagon) and limonene (blue crosses) modelled after
 339 chamber (filled symbols) and flow-tube settings (open symbols). The figure shows a comparison of SOA mass yields obtained from
 340 simulations with MCM + PRAM (panel A) and ratio of yields from MCM and MCM+PRAM (panel B). Currently PRAM is

341 available for **ozonolysis** of limonene and α -pinene. The clumps are a result of SOA mass yields for the oxidation of specific
 342 oxidant concentration with varying BVOC concentration



344 **Figure 2.** The mass yields from the ozonolysis of BVOCs β -pinene and β -caryophyllene modelled after chamber (filled symbols)
 345 and flow-tube (open symbols) settings. The figure shows a comparison of SOA mass yields obtained from simulations with only
 346 MCM as currently there is no PRAM available for these compounds. The experimental values are provided for comparison.

347 3.2 BVOCs – OH chamber and flow-tube simulations

348 The mass yields obtained by MCM+PRAM for α -pinene – OH oxidation are close to the measured
 349 values (Kristensen et al., 2017), while using only MCM under-predicts the mass yields (Figure 3 , panel A
 350 and B, and Table 3). The maximum SOA mass yield for OH oxidation of α -pinene is lower than the yield
 351 from ozonolysis which is suspected to arise due to the formation of more volatile oxidation products produced
 352 during OH oxidation (Bonn and Moortgat, 2002; Kristensen et al., 2014). The OH oxidation of β -pinene
 353 results in mass yields similar to the measurements obtained by Lee et al. (2006b) for similar mass loadings.

354 The β -pinene SOA yields are comparatively well represented by MCM+PRAM in comparison to the
355 standalone MCM. On the other hand, the limonene mass yields are under-predicted by MCM+PRAM for
356 similar mass loadings. Yields for limonene SOA mass loadings of $350 \mu\text{g m}^{-3}$ are around 0.31 which is lower
357 than the experimental values of 0.58, measured by Lee et al. (2006b).

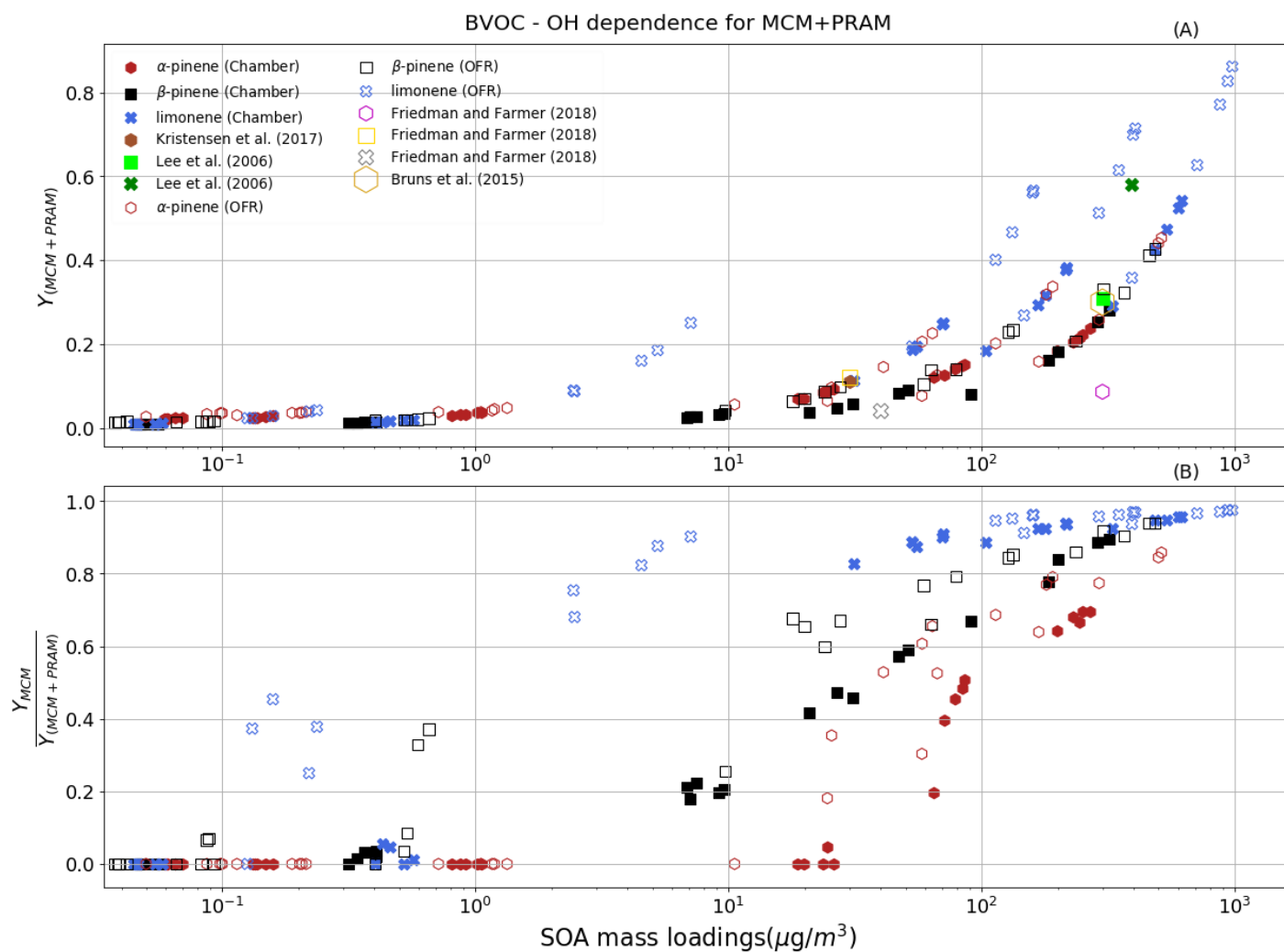
358 For β -caryophyllene, the modeled values are in good agreement with experimental measured yields in
359 the range of mass loadings provided by Griffin (1999) and Tasoglou and Pandis (2015). Currently there are no
360 experiments providing HOM yields from OH oxidation of β -caryophyllene, and hence, those species are not
361 included in PRAM. The simulation results for yields from OH oxidation of β -caryophyllene, indicate that the
362 MCM scheme is able to reproduce the experimental values (Fig. 4). Only MCM was used for modeling the
363 mass yields for OH oxidation of isoprene due to current lack of PRAM mechanism for isoprene. The mass
364 yields derived from OH oxidation of isoprene vary from 0.01 - 0.31 covering a range of mass loadings from
365 $0.003 - 132 \mu\text{g m}^{-3}$. At low mass loadings $< 10 \mu\text{g m}^{-3}$ the maximum yield obtained is ~ 0.06 , which is a factor
366 of 3 greater than the experimental results obtained by (Lee et al., 2006b) where they measured yield of 0.02.
367 The mass yields are in good agreement with the experimental results from Liu et al. (2016), wherein they
368 measured a yield of 0.13 for $22 \mu\text{g m}^{-3}$ (Table 3).

369 The OFR simulations results for the OH oxidation of BVOCs with an equivalent exposure range from
370 $2.0 \times 10^{10} - 2.0 \times 10^{12}$ molecules $\text{cm}^{-3} \text{s}$, is shown in Fig. 3. Our yields for α -pinene agree well with the yields
371 obtained by Bruns et al. (2015) where they measured yield of ~ 0.3 for mass loading of $\sim 300 \mu\text{g m}^{-3}$ at
372 equivalent OH exposures. Friedman and Farmer (2018) found mass yields of 0 - 0.086 for α -pinene
373 (ammonium sulfate seeded experiment), 0- 0.12 for β -pinene (no seed particles) and 0-0.04 for limonene (no
374 seed particles), by varying the OH exposures between $4.7 \times 10^{10} - 7.4 \times 10^{11}$ molecules $\text{cm}^{-3} \text{s}$. Our simulated
375 yields for OH oxidation of α -pinene, β -pinene and limonene suggest higher mass yields for α -pinene and
376 limonene at equivalent mass loadings, while mass yields for β -pinene are in good agreement with the
377 experimental yields. Friedman and Farmer (2018) suggest that the reason for this underestimation in mass
378 yields could arise due to the exclusion of large particle sizes in the experiments and propose that these yields
379 could represent lower bounds.

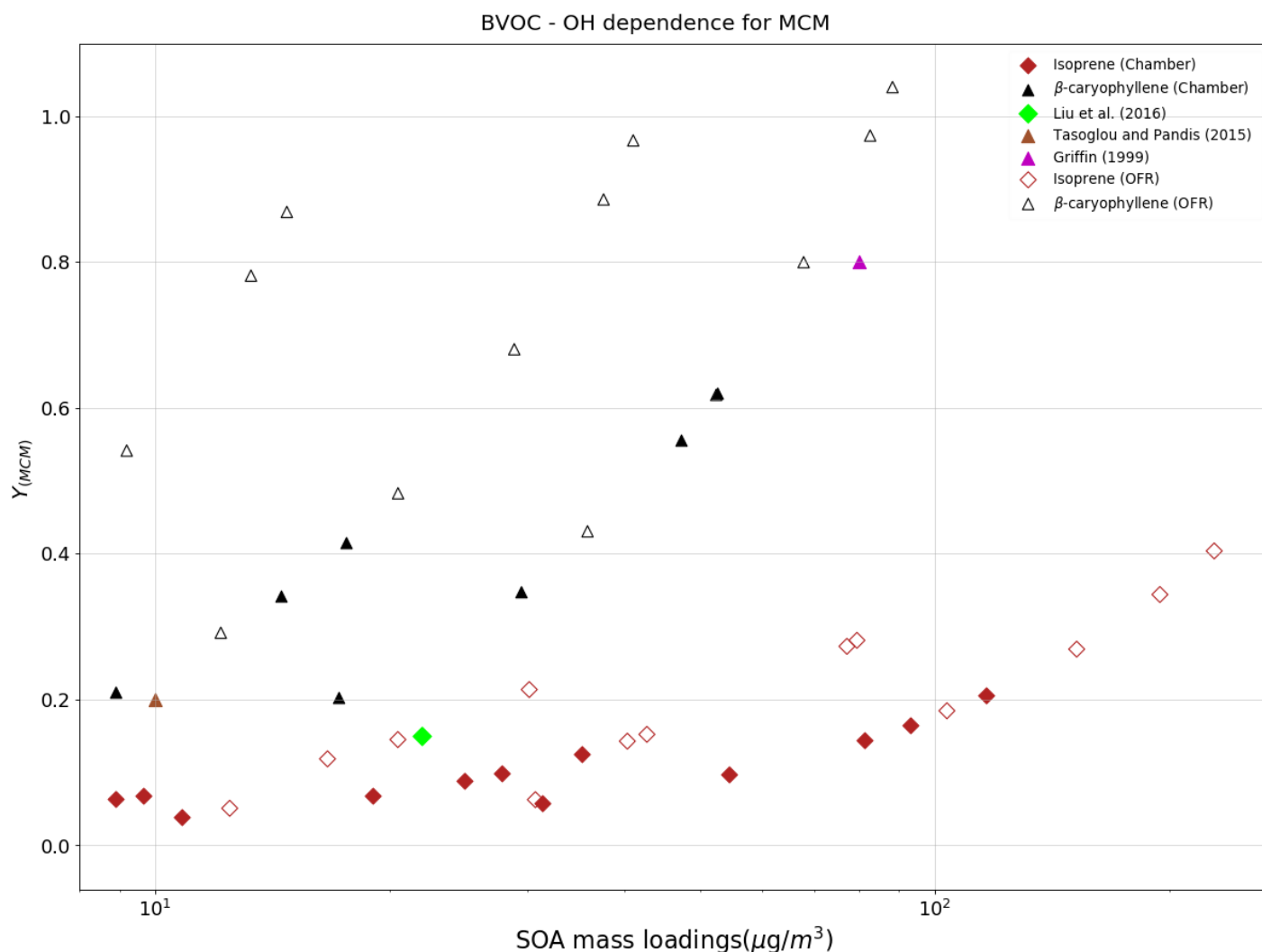
380 **Table 3.** Mass yields for OH oxidation of BVOCs at 293 K for different range of mass loadings using a
381 chamber[†] and OFR^{||} setup.

SOA mass loading ($\mu\text{g m}^{-3}$)	MCM + PRAM mass yields	MCM mass yields	BVOC	Experimental yields	References
300 [†]	0.28	0.25	β -pinene	0.31 (293)	Lee et al. (2006b)
350 [†]	0.31	0.06 – 0.11	limonene	0.58 (394)	Lee et al. (2006b)
30	0.09	0.004	α -pinene	0.11 (30)	Kristensen et al., 2017
< 10 [†]	0.21 ^{!!}	0.21	β -caryophyllene	0.2 (8.8)	Tasoglou and Pandis (2015)
20 – 80 [†]	0.3 – 0.7 ^{!!}	0.3 – 0.7		0.37 – 0.79 (17-82)	Griffin (1999)
22 [†]	0.1 ^{!!}	0.1	Isoprene	0.13 (22)	Liu et al. (2016)
<10	0.06 ^{!!}	0.06		0.02 (9)	Lee et al. (2006b)
0-300 ^{!!}	0.05 – 0.31	0 – 0.2	α -pinene	0 – 0.086 (0-300) 0.3 (300)	Friedman and Farmer (2018) Bruns et al. (2015)
0-30 ^{!!}	0-0.1	0-0.01	β -pinene	0 – 0.12 (30)	Friedman and Farmer (2018)
0-40 ^{!!}	0.-0.19	0-0.17	limonene	0.0 – 0.04 (35)	Friedman and Farmer (2018)

382 ^{!!}indicates that no PRAM mechanism available yet i.e the yields are same as the MCM yields.



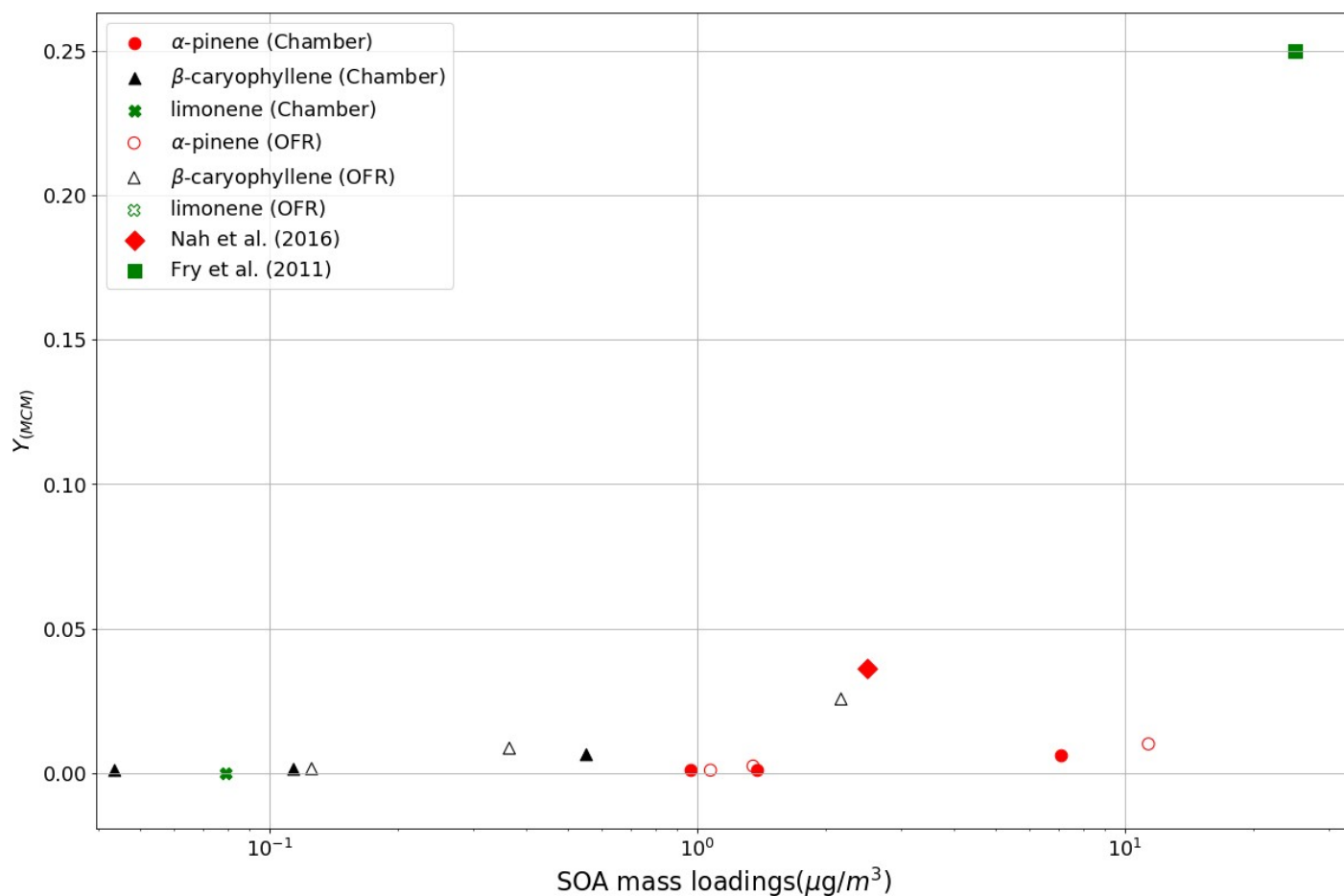
383 **Figure 3.** The mass yields from OH oxidation of BVOCs α -pinene (red heptagons), β -pinene (black squares) and limonene (blue
 384 crosses) modeled after chamber (filled symbols) and flow-tube settings (open symbols). The figure shows a comparison of SOA
 385 mass yields obtained from application of MCM+PRAM (panel A) and ratio of yields from MCM and couple MCM+PRAM (panel
 386 B). Currently PRAM is available for OH oxidation of limonene and α -pinene and β -pinene.



388 **Figure 4.** The mass yields from OH oxidation of BVOCs β -caryophyllene (black triangles) and isoprene (maroon diamonds)
 389 modeled after chamber (filled symbols) and flow-tube settings (open symbols). The figure shows a comparison of SOA mass yields
 390 obtained from application of MCM as currently there is no PRAM available for these compounds.

391 3.3 BVOC – NO₃ chamber and OFR simulations

392 Figure 5. shows the yields derived from the oxidation of BVOCs by NO₃. Currently, as no PRAM is
 393 available for NO₃ oxidation, Figure 5 represents SOA yields derived using MCM. Due to limited
 394 experimental constraints, PRAM presently does not consider autoxidation of RO₂ formed from NO₃ oxidation
 395 of VOCs, which could explain the huge discrepancy between the measured and simulated mass yields (Figure
 396 5). The yields obtained for oxidation of α -pinene (0.002-0.007) by NO₃ are low in comparison to those
 397 obtained by Nah et al. (2016), where they measured a yield of 0.036. Measured mass yields for limonene
 398 oxidation by NO₃ resulting in mass yields between 0.25-0.4 (Fry et al., 2011), whereas we obtain negligible
 399 (~ 0.0003) mass yields for the same.

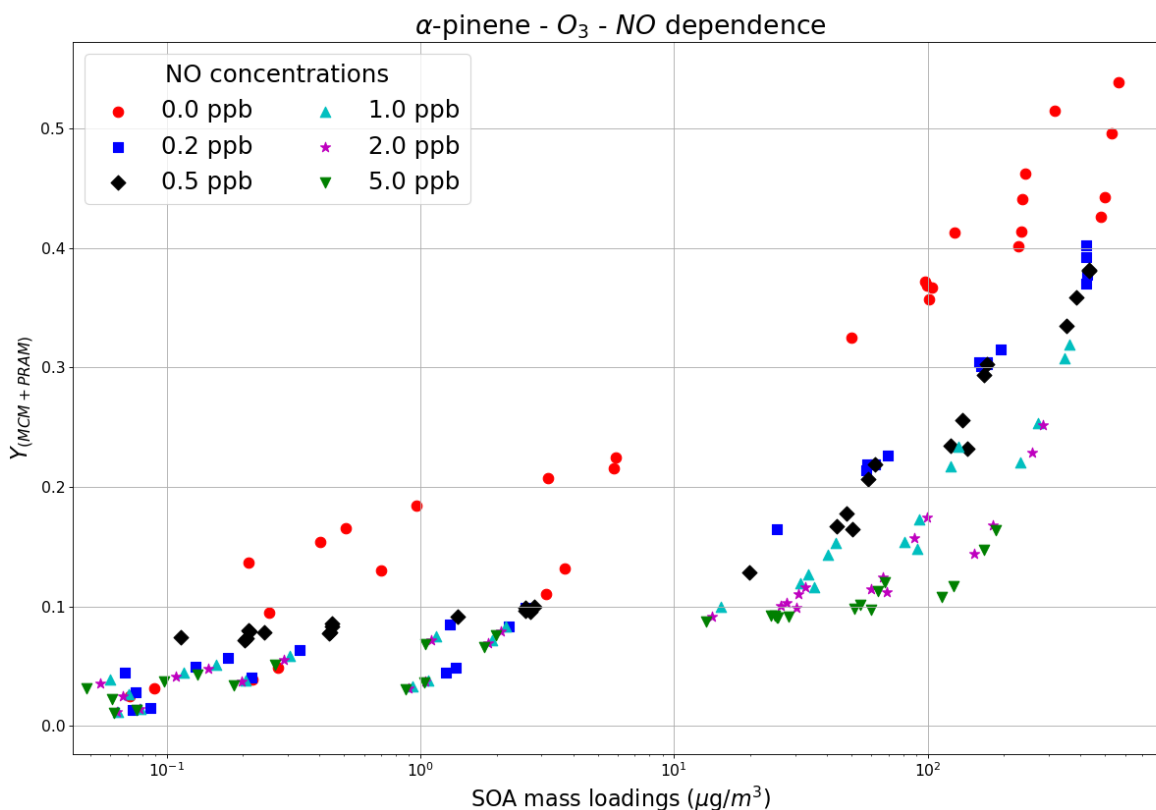
BVOC - NO_3 dependence for MCM

400 **Figure 5.** The mass yields from NO_3 oxidation of BVOCs modeled after chamber and flow-tube settings. The figure shows a
 401 comparison of SOA mass yields obtained from application of MCM. Appreciable mass yields were only obtained for α -pinene,
 402 limonene and β -caryophyllene.

403 3.4 NO_x dependence

404 Varying NO_x concentrations changes the fate of RO_2 radical formed during organic oxidations by
 405 altering HO_2/RO_2 ratio, thereby impacting the distribution of reaction products and aerosol formation (Presto
 406 et al., 2005; Zhao et al., 2018; Sarrafzadeh et al., 2016). We modeled the SOA mass yields for α -pinene - O_3
 407 setup with varying NO_x concentrations (NO was varied whereas NO_2 was kept constant for all the runs), for
 408 initial α -pinene mixing ratios in the range 0.5 - 200 ppb (Fig. 6). A maximum SOA yield value of 0.55 is
 409 obtained for a combination of the lowest value of NO (0 ppb, red circles). As the NO concentrations increase
 410 from 0.2 ppb (blue squares) to 5 ppb (green inverted triangles) the yields begin to decrease, and this pattern is
 411 observable and valid for all concentration ranges of reacted precursor VOC. The NO_x dependence of α -pinene

412 ozonolysis is consistent with the findings of Draper et al. (2015) and Presto et al. (2005) wherein they
 413 observed a trend of decreasing SOA mass yields for α -pinene ozonolysis with increasing NO_x concentrations.

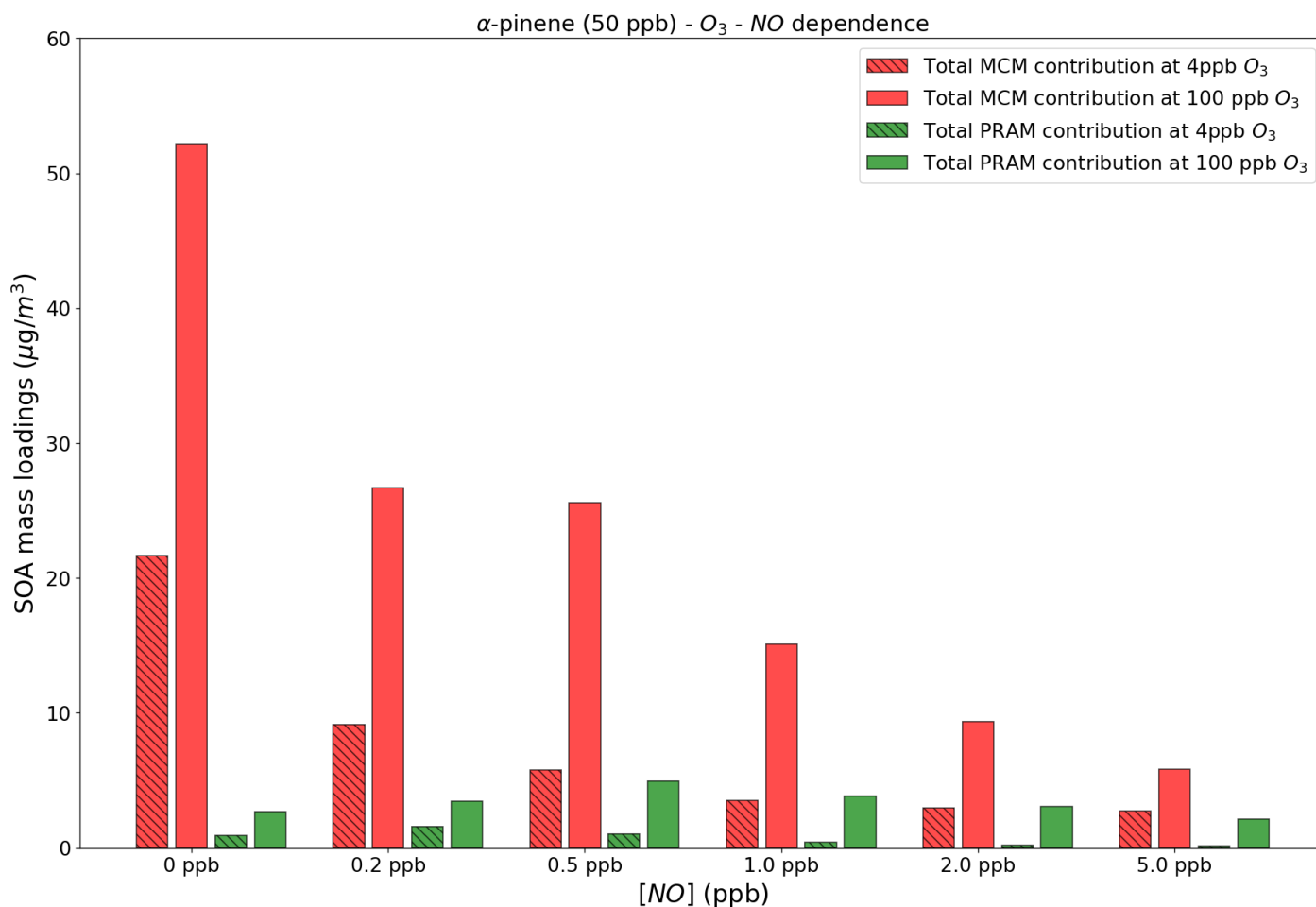


414 **Figure 6.** The SOA mass yields from O_3 oxidation of α -pinene modeled for different NO concentrations with the chamber setup.
 415 The model runs were performed using MCM+PRAM.

416 At low NO_x concentrations RO_2 radicals undergo rapid autoxidation until they react with HO_2 or RO_2
 417 resulting in production of low volatility hydro-peroxide products (Sarrafzadeh et al., 2016), closed shell
 418 monomers or dimers (Ehn et al., 2014; Roldin et al., 2019), which increase SOA mass. This contrasts with
 419 high NO_x conditions where the $\text{RO}_2 + \text{NO}$ reactions dominate over reactions with HO_2 or RO_2 , resulting in the
 420 formation of more volatile products such as aldehydes, ketones and organonitrates (Presto et al., 2005;
 421 Sarrafzadeh et al., 2016), and likely suppressing the autoxidation process leading to a decrease in SOA mass
 422 loadings (Ehn et al., 2014).

423 Figure 7 shows the absolute contributions to SOA mass loadings by PRAM and MCM compounds at
 424 two different O_3 concentrations of 4 and 100 ppb and varying NO concentrations. The figure shows that with
 425 an increase in NO concentrations more than 1 ppb the contribution of PRAM compounds to the particle phase

426 decreases at both 4 and 100 ppb of O₃ concentrations. In PRAM the RO₂ + NO reaction leads either to the
 427 formation of organonitrate HOM, closed shell monomers with carbonyl group or fragmentation products with
 428 higher volatility (Roldin et al., 2019). HOM Dimer formation is suppressed with increasing NO
 429 concentrations in PRAM (Roldin et al., 2019) which explains the lower contribution by PRAM compounds to
 430 SOA mass loadings with increasing NO. At NO concentrations <1ppb the PRAM contribution increases as,
 431 first generation RO₂ are capable of undergoing autoxidation forming highly oxygenated RO₂ which
 432 subsequently reacts with NO forming organic nitrates (Ehn et al., 2014). As NO concentrations exceed 1ppb
 433 the first generation RO₂ is scavenged by NO thereby reducing the concentration of organonitrate HOM (Ehn
 434 et al., 2014), possibly affecting SOA yields. The MCM contribution also decreases with increasing NO
 435 concentrations mostly due to the formation of more volatile organonitrates (Jenkin et al., 2019) .



436 **Figure 7.** Contribution to the SOA mass loadings by total PRAM and MCM compounds at different NO levels and O₃
 437 concentrations. For comparison we use 4 ppb and 100 ppb O₃ concentrations, respectively, at 50 ppb α-pinene.

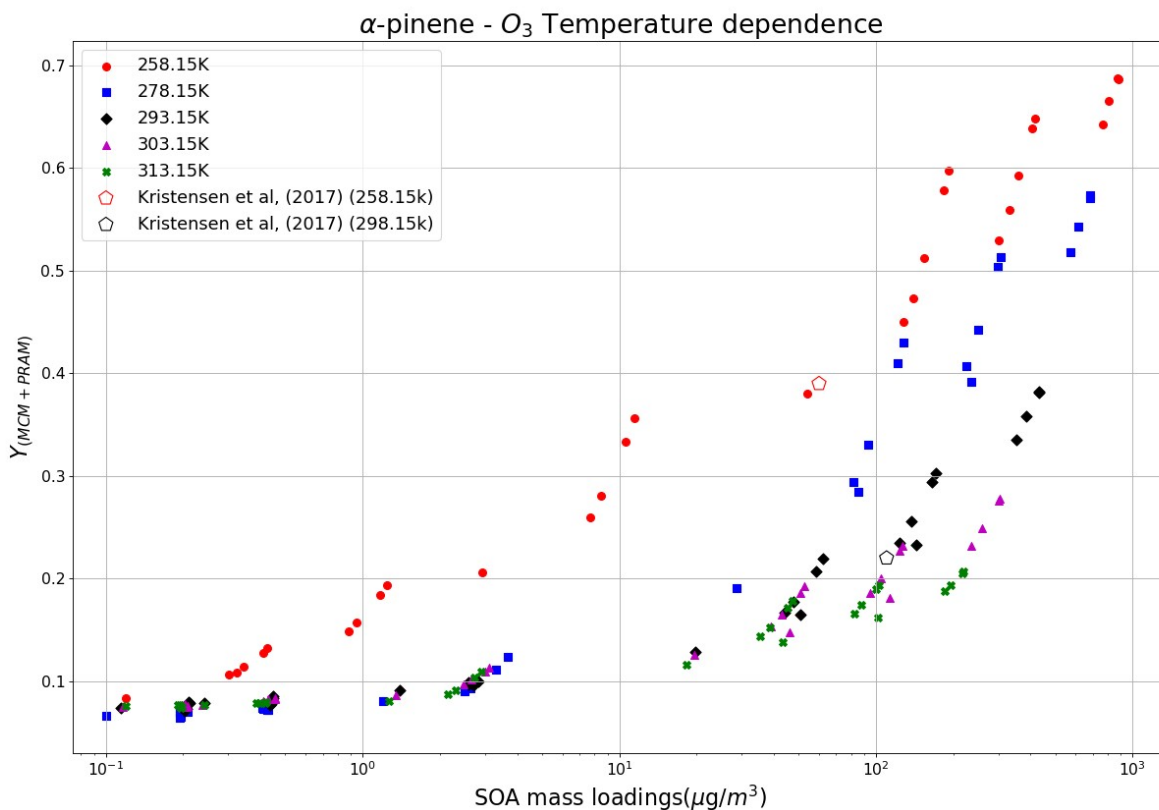
438

439

440 3.5 Temperature dependence

441 The formation of SOA from α -pinene ozonolysis in the temperature range of 258.15 - 313.15 K was
442 investigated in this study using MCM+PRAM. Strong dependence of SOA mass yield on temperature was
443 reported by Saathoff and Naumann, (2009) wherein they measured the decreasing mass yields from 0.42 at
444 273.15 K to 0.09 to 313.15 K for SOA loadings of 53 and 92 $\mu\text{g m}^{-3}$ respectively. Our results in Figure 8
445 show increasing SOA mass yields for α -pinene ozonolysis with decreasing temperature, which is attributed to
446 the augmented condensation of oxidation products termed as semi volatile organic compounds (SVOC)
447 (Kristensen et al., 2017) at lower temperatures.

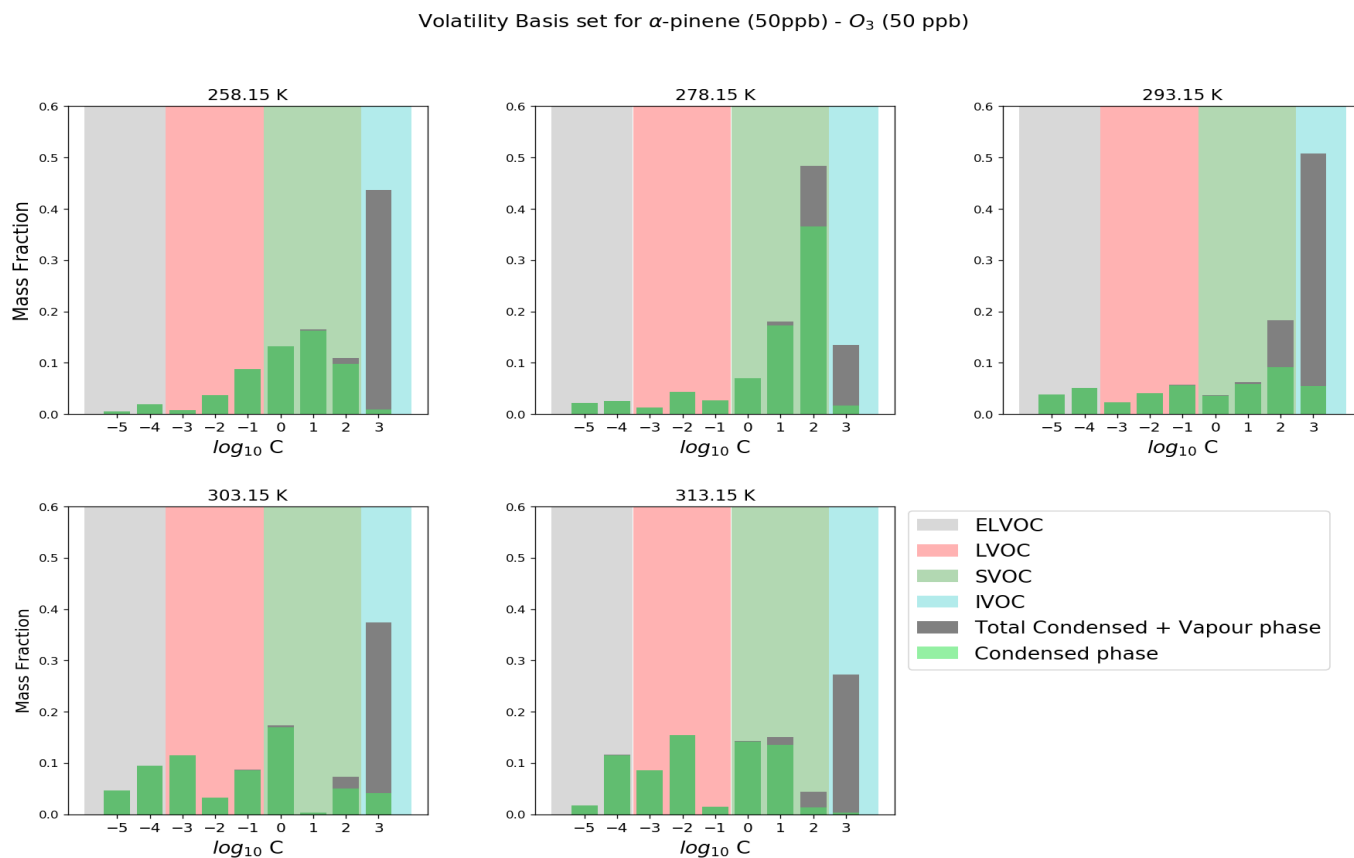
448 For α -pinene maximum mass loading $< 150 \mu\text{g m}^{-3}$ the mass yields reach a maximum value of 0.38 at
449 temperatures as low as 258.15 K and decrease to 0.27 for a temperature of 293.15 K and to 0.1 for the
450 temperature of 313.15 K. These yields are comparable to the results obtained by Kristensen et al.
451 (2017) where they measured yields of 0.39 for 258.15 K and 0.22 for 293.15 K for mass loading $< 150 \mu\text{g m}^{-3}$.
452 The results show a weak dependence of SOA mass yields on temperatures in the range of 278.15 K - 313.15
453 K at low SOA mass loadings which become more pronounced as the mass loadings increase. At the lowest
454 temperature of 258.15 K the mass yields are higher in comparison to other temperatures regardless the mass
455 loadings. These results are in good agreement with the findings by Pathak et al. (2007) where they found a
456 strong temperature dependence of SOA mass yields at lower temperature (0 – 15° C), which decreases as the
457 temperature increases. Furthermore, similar to the measurements made by Pathak et al. (2007), our
458 simulations were able to reproduce the experimental findings that show no appreciable differences in the SOA
459 mass yields for loadings below 1 $\mu\text{g m}^{-3}$ (initial mixing ratio of 1 ppb) for temperatures > 273.15 K.



460 **Figure 8.** Temperature dependence of SOA mass yields at different temperatures using the MCM+PRAM. The open pentagons
 461 represent measurement data from Kristensen et al. (2017) at 258.15 K and 298.15 K.

462 Figure 9 shows the volatility distribution of α -pinene ozonolysis derived SOA at different
 463 temperatures. The saturation vapor pressure limits for defining extremely low volatility (ELVOCs - grey
 464 shaded), low volatility (LVOCs - red shaded), semi volatile (SVOCs - green shaded) and intermediate
 465 volatility (IVOCs - cyan shaded) organic compounds used in the Volatility basis set (VBS) are set according
 466 to the values suggested in Donahue et al. (2012). In this work, we categorize compounds (ELVOCs, LVOCs,
 467 SVOCs and IVOCs) based on effective saturation vapor pressures (C^*) in the range of $\{10^{-5}$ to $10^3\}$ $\mu\text{g m}^{-3}$
 468 and temperature of 298 K (Donahue et al., 2009). At the lowest temperature of 258.15 K, the SVOCs
 469 contribution to the particle phase is dominant in comparison to LVOCs and ELVOCs, a trend which is
 470 subsequently reversed as the temperatures are increased. At 293.15 K a majority of SVOCs and IVOCs are in
 471 the gas phase while the contribution of LVOCs and ELVOCs to particle phases increases. These results are in
 472 good agreement with observations made by Kristensen et al. (2017) wherein they observed an increasing
 473 contribution of SVOCs at sub-zero temperatures of 258.15 K, which decrease the fraction of SOA formed

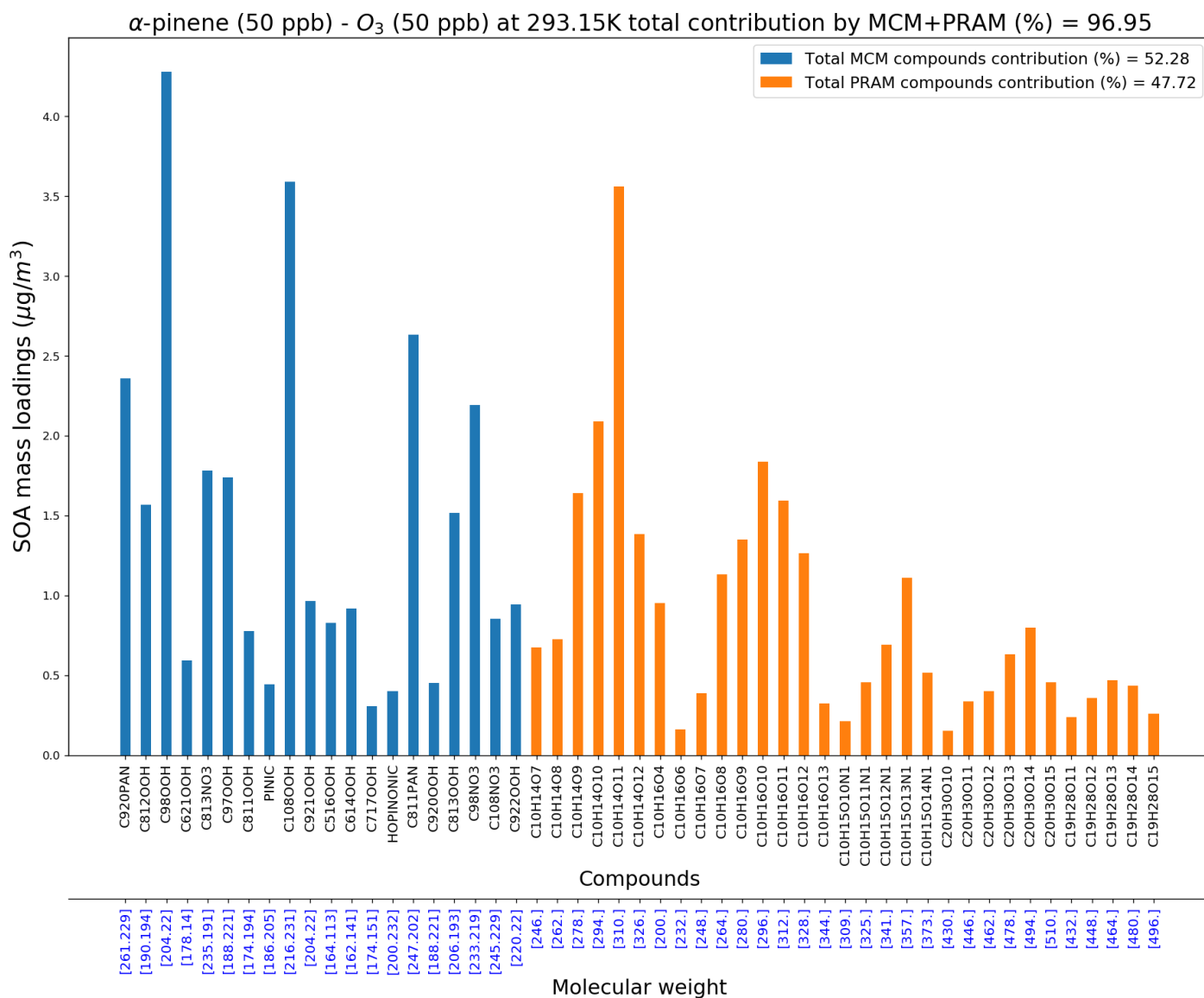
474 from ELVOCs. Again, it should be noted that the temperature dependence of peroxy radical autoxidation
 475 product formation still needs further validation based on recent experiments (e.g. Quéléver et al., 2019).
 476



477 **Figure 9.** Modeled volatility distribution of SOA at different temperatures. The volatility bins span a range of effective saturation
 478 vapor pressures $C = C^* = \{10^{-5} \text{ to } 10^3\} \mu\text{g m}^{-3}$. The VBS distribution is based on a reference temperature of 298 K.

479 3.6 Composition

480 MCM+PRAM can be used to narrow down and compile a list of compounds playing a pivotal role in
 481 contributing to SOA mass loadings and, also compare the relative importance of implementing PRAM
 482 alongside the MCM. Figure 10 shows the most important compounds from both the MCM and PRAM that
 483 together contribute to more than 95% of α -pinene ozonolysis SOA mass loading at 293.15 K.



484 **Figure 10.** MCM and PRAM Compounds contributing to > 95 % of SOA mass at 293 K and 50ppb O_3 and α -pinene
 485 concentrations.

486 Figure 10 shows that contribution to SOA mass loadings by PRAM compounds is ~48 % (of 97 %) while
 487 MCM compounds contribute ~52 % (of 97%). On lowering the temperature to 258K the relative contributions
 488 of PRAM drop to 15 % (of ~99 %), while MCM dominates by contributing ~85 % (of ~99 %) respectively
 489 (Figure S3a). The contribution of PRAM increases to ~64 % (of ~97 %) and MCM contribution drops to 36
 490 % (of ~97 %) at 313 K (Figure S3b). These results reflect the importance of PRAM as its contribution plays
 491 an increasingly dominant role with increasing temperatures and highlights the crucial few compounds that
 492 contribute to maximum SOA mass loadings for α -pinene ozonolysis. The list of abundant compounds which
 493 together add up to contribute more than 95 % of SOA mass loadings at 258 K, 293 K and 313 K are presented
 494 in the supplement Table 1s (a, b & c). At 258 K MCM compounds namely pinonic acid ($C_{10}H_{16}O_3$, 4.4 %),

495 C920PAN ($C_{10}H_{15}NO_7$, 9.3 %), C108NO3 ($C_{10}H_{15}NO_6$, 8.9 %), C811PAN ($C_9H_{13}NO_7$, 10.1 %), C717NO3
496 ($C_7H_9NO_6$, 11.3 %) contribute significantly to the total SOA mass loadings while PRAM compounds such as
497 $C_{10}H_{14}O_7$ (0.88 %), $C_{10}H_{16}O_4$ (1.3 %), $C_{10}H_{16}O_6$ (1.13 %) contribute significantly less. An increase in
498 temperature to 293 K results in an overall increase in contribution by PRAM compounds, with $C_{10}H_{14}O_{10}$ (3.6
499 %), $C_{10}H_{14}O_{11}$ (6.2 %), $C_{10}H_{16}O_{10}$ (3.2 %) playing an important role in contributing to the SOA mass loadings.
500 This trend of relative increase in the contribution by PRAM compounds over MCM compounds to SOA mass
501 loadings is also evident as the temperatures are further increased to 313 K, where the PRAM compounds
502 $C_{10}H_{14}O_{11}$ (18.3 %), $C_{10}H_{14}O_{12}$ (6 %) and $C_{10}H_{16}O_{12}$ (6.6 %) play a dominant role in increasing SOA mass
503 loadings.

504 4. Conclusions

505 We simulated SOA mass yields derived from the oxidation of various BVOCs (isoprene, α -pinene, β -
506 pinene, limonene and β -caryophyllene), by the oxidants O_3 , OH and NO_3 using the zero-dimensional model
507 MALTE-Box. The gas phase chemistry was simulated using the MCM in conjunction with PRAM. The aim
508 was to verify the efficacy of MCM+PRAM in simulating the SOA mass yields. Additional simulations were
509 performed to test the MCM+PRAM under varying temperature and NO concentrations. A few important
510 compounds playing a major role in increasing the SOA mass yields for α -pinene ozonolysis at different
511 temperatures are also highlighted.

512 The simulations were designed to resemble ideal smog chambers experiments and experiments in
513 oxidative flow reactors (OFR). No interactions between the gas phase and chamber walls were considered
514 during the simulations. For the smog chamber setting, the standalone MCM generally under-predicts the mass
515 yields obtained by the ozonolysis and OH oxidation of BVOCs. In contrast, the yields derived using
516 MCM+PRAM for the smog chamber setup is in good agreement with the experimental results. For an
517 idealized OFR setup, MCM+PRAM yields are in good agreement with experimental yields, while again the
518 MCM under-predicts the SOA yields. The relative contribution of HOM monomers and dimers to the particle
519 phase in OFR simulations is low when compared to the chamber simulations. This is due to higher RO_2
520 concentrations in OFR leading to termination of peroxy radical autoxidation, thereby affecting SOA yields.
521 This needs to be considered when applying yields based on OFR simulations in regional or global chemical
522 transport models

523 The model does not simulate appreciable SOA mass yields for oxidation of BVOCs with NO_3 , as
524 PRAM currently does not consider autoxidation of RO_2 formed from NO_3 oxidation of VOCs. This underlines
525 the need for developing a NO_3 oxidation scheme which can better constrain and predict SOA mass yields. In

526 accordance to the previous studies, the simulated SOA yields tend to decrease at higher temperatures. The
527 PRAM contribution to mass yields at low temperatures (258.15 K) is ~14 %, which is substantially lower than
528 that of MCM (~86 %). As the temperature is increased to 313.15 K, the contribution of PRAM to SOA mass
529 yields begins to dominate over MCM. This most likely is due to MCM producing more SVOCs (compounds
530 classified as SVOCs at 298 K), which show stronger contribution to particle phase at lower temperatures, due
531 to decrease in saturation vapor pressures with temperature. It should be noted that the present temperature
532 dependency of mass yields using PRAM are a first, and currently the best estimate in understanding the
533 influence of temperature on the peroxy radical autoxidation formation. The simulated SOA yields with
534 varying NO concentrations agree well with experimental results, i.e. SOA yields decrease with increasing NO
535 concentrations due to the formation of more volatile compounds such as organonitrates and ketones.

536 Using PRAM coupled with MCM helps us bridge the gap in understanding the role and contribution
537 of peroxy radical autoxidation to SOA formation. The variation of SOA yields for temperature and NO
538 concentrations, indicates the limitations of global and regional models in predicting e.g. cloud condensation
539 nuclei (CCN) effects using fixed SOA yields. The good agreement of modeled and experimental yields from
540 smog chambers, could further help us parameterize the SOA yields, that could be applied at a global and
541 regional model scale, to more accurately predict the direct and indirect impact of aerosol particles on e.g.
542 radiation balance by aerosol scattering/absorption and CCN concentrations. Furthermore, implementation of a
543 condensed PRAM version to regional and global models has been tested but still need further validation
544 (Roldin et al., 2019).

545 **Data availability**

546 The complete PRAM mechanism written in a format compatible with the Kinetic PreProcessor (KPP) together
547 with all species information can also be downloaded from <https://doi.org/10.1594/PANGAEA.905102>
548

549 **Author Contributions**

550 CX and MB served as the chief authors and editors of the paper. CX was performing the model simulations.
551 The study was designed by CX, MB and PR. All other co-authors contributed to the analysis and writing of
552 the paper.

553 **Acknowledgements**

554 The presented research has been funded by the Academy of Finland (Center of Excellence in Atmospheric
555 Sciences) grant no. 4100104 and the Swedish Research Council FORMAS, project no. 2018-01745. We
556 would also like to acknowledge the invaluable contribution of computational resources from CSC – IT Center
557 for Science, Finland.

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