China's Clean Air Action has suppressed unfavorable influences of climate on wintertime PM$_{2.5}$ concentrations in Beijing since 2002

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Abstract

Severe wintertime $PM_{2.5}$ pollution in Beijing has been receiving increasing worldwide attention, yet the decadal variations remain relatively unexplored. Combining field measurements and model simulations, we quantified the relative influences of anthropogenic emissions and meteorological conditions on $PM_{2.5}$ concentrations in Beijing over winters of 2002-2016. Between the winters of 2011 and 2016, stringent emission control measures resulted in a 21% decrease in mean mass concentrations of $PM_{2.5}$ in Beijing, with 7 fewer haze days per winter on average. Given the overestimation of $PM_{2.5}$ by model, the effectiveness of stringent emission control measures might have been slightly overstated. With fixed emissions, meteorological conditions over the study period would have led to an increase of haze in Beijing, but the strict emission control measures have suppressed the unfavorable influences of recent climate. The unfavorable meteorological conditions are attributed to the weakening of the East Asia Winter Monsoon associated particularly with an increase in pressure associated with the Aleutian low.

1 Introduction

In recent years, persistent and severe haze episodes with high $PM_{2.5}$ concentrations occur frequently in China, attracting worldwide attention (*Cheng et al., 2016; Gao et al., 2016*). High aerosol concentrations during haze have been reported to cause traffic jams and flight cancelations (*Wu et al., 2005*), and have been linked to health damages (*Dockery et al., 1993*), climate change (*Ramanathan and Carmichael, 2008*) and ecosystem degradation (*Chameides et al., 1999*). The annual mean $PM_{2.5}$ concentrations in Beijing exceeded 90 μg/m^3 in 2013, nearly twice the China's National Ambient Air Quality Standard (NAAQS) of 35 μg/m^3 (*MEP, 2012*). January 2013 was reported as the haziest month over the past 60 years in Beijing, with maximum hourly and daily mean $PM_{2.5}$ concentrations exceeding 1000 μg/m^3 and 500 μg/m^3, respectively (*Wang et al., 2014*).

Since then, the State Council of China issued the Air Pollution Prevention and Control Action Plan (APPCAP, denoted as the Clean Air Action hereafter), which describes explicitly the pollution control measures, and proposed specific goals for concentrations by 2017 (*China State Council, 2013*). This action has been considered the most stringent air pollution control policy in Chinese history. The Jing-Jin-Ji, Yangtze River Delta, and Pearl River Delta regions were required to reduce annual mean $PM_{2.5}$ concentrations by 15-25% compared with the concentrations in 2013, and the annual mean concentrations of $PM_{2.5}$ in Beijing should not exceed 60 μg/m^3 (*Cheng et al., 2018; China State Council, 2013*). Specific control measures included eliminating small coal-fired boilers, phasing out small, high-emitting factories, installing control facilities for emissions of VOCs (volatile organic compounds), and replacing residential coal burning with electricity and natural gas among others. (*Zheng et al., 2018*). With these rigorous control measures, China has made impressive progress, with annual mean $PM_{2.5}$ concentrations reduced in major metropolitan regions by 28-40% between 2013 and 2017 (*Zheng et al., 2018*).

A number of studies have used visibility as a surrogate to indicate trends of haze pollution in China over the past several decades (*Che et al., 2009; Chen and Wang, 2015; Ding and Liu, 2014; Wang and Chen, 2016*). *Chen and Wang (2015)* reported that haze days increased rapidly in the 1970s and remained relatively stable up to present. However, *Che et al. (2009)* illustrated

that there was a decreasing haze trend in winter for many cities over the interval 1981-2005. *Wang et al. (2019)* argued that visibility is impacted significantly by meteorological factors, especially relative humidity, and thus visibility does not accordingly reflect the real changes in air pollution. The influences of anthropogenic emissions and meteorological conditions on air quality over shorter periods have been investigated extensively (Gao et al., 2017a; Y. Gao et al., 2011; Liu et al., 2017; Wang et al., 2016; Xing et al., 2011). Due to the lack of long-term measurements of $PM_{2.5}$ in China, limited studies have been conducted exploring the roles of anthropogenic emissions and meteorological conditions for the long-term variations of $PM_{2.5}$ in China. *Yang et al. (2016)* used the GEOS-Chem model to simulate $PM_{2.5}$ in China between 1985 and 2005, and concluded that the increase of winter $PM_{2.5}$ was dominated over this period by the increase in anthropogenic emissions. They found that weakening of winds was the dominant meteorological factor. The variability of winds in North China has also been linked to air quality over the Yangtze River Delta, which is a downwind region of North China (*Kang et al., 2019*). While *Yang et al. (2016)* has explored the relative roles of emissions and meteorology, no model validation of $PM_{2.5}$ was provided, and the recent decades were not covered in the study period. Long-term measurements of $PM_{2.5}$ in Beijing reveal a slight decreasing trend of annual mean concentration over 2004-2012 (*Liu et al., 2015*). With the increase in availability of recent measurements, a closer reading of long-term variations is needed to better define the relative roles of anthropogenic emissions and meteorology.

In this study, we present a comprehensive analysis of the decadal trend of wintertime $PM_{2.5}$ in Beijing based on regional meteorology-chemistry modeling, a new decadal emission inventory, and long-term observations of $PM_{2.5}$, including their composition. We address the following questions: (1) the influences of decadal changes in anthropogenic emissions and meteorology on the variations of winter $PM_{2.5}$ in Beijing; and (2) the key driving factors for the decadal variation of meteorology. The descriptions of model, emissions, and numerical experiments are presented in Sect. 2. The two questions highlighted above are addressed in detail in Sect. 3; Sect. 4 provides an overall summary of the study.

2 Model Description and Configurations

2.1 Meteorology-chemistry modeling

We used the WRF-Chem (Weather Research and Forecasting model coupled with chemistry) model version 3.6.1 to simulate meteorology and emissions, transport, mixing, and the chemical transformation of trace gases and aerosols. Two nested domains were applied with the outer domain covering East Asia and part of Southeast Asia, with the inner domain focusing on North China (Fig. 1). Horizontal resolutions of 81km and 27km were configured respectively for these two domains, and the model accounted for 27 vertical layers extending from the surface to 50 hPa. The gas phase chemical mechanism CBMZ (*Zaveri and Peters, 1999*) coupled with the 8-bin sectional MOSAIC model with aqueous chemistry (*Zaveri et al., 2008*) was adopted. The model treats all the important aerosol species, including sulfate, nitrate, chloride, ammonium, sodium, black carbon (BC), primary organic and inorganic material. The Fast-j radiation scheme (*Wild et al., 2000*) was selected to calculate photolysis rates. These configurations have been shown in previous studies (*Gao et al., 2016a, 2016b, 2017b*) to be capable of reproducing winter haze episodes in North China.

2.2 Emissions

The monthly Multi-resolution Emission Inventory for China (MEIC, http://www.meicmodel. org/) covering the years 2002-2017 (*Zheng et al., 2018*) was used for anthropogenic emissions. This inventory considers emissions of sulfur dioxide (SO_2), nitrogen oxides (NO_x), carbon monoxide (CO), non-methane volatile organic compounds (NMVOC), ammonia (NH_3), black carbon (BC), organic carbon (OC), $PM_{2.5}$, PM_{10}, and carbon dioxide (CO_2) associated with power generation, and the industrial, residential, transportation, and agricultural sectors. The trends of wintertime emissions of these species in Beijing-Tianjin-Hebei region over the interval 2002-2016 are displayed in Fig. S1. Emissions of SO_2 have decreased continuously since 2004, while NO_x emissions have declined since 2011. Emissions of all involved species have decreased rapidly since 2012. Biogenic emissions were calculated online using the MEGAN model (*Guenther et al., 2006*). Emissions from biomass burning were taken from the GFED v3 dataset (*Randerson et al., 2015*). Dust emissions and online sea-salt emissions were also calculated online.

2.3 Numerical experiments

Meteorological initial and boundary conditions were obtained from the National Centers for Environmental Prediction (NCEP) Final Analysis (FNL) dataset. Chemical initial and boundary conditions were taken from the climatological data provided by the NOAA Aeronomy Lab Regional Oxidant Model (NALROM). Wintertime periods defined as the last month of the year and the following two months of the next year were simulated for years 2002-2016. For example, the winter of 2002 includes December of 2002, January of 2003 and February of 2003. Nudging (assimilation) of the analyses was applied to produce realistic meteorological simulations. The simulations were conducted month by month (15 years $\times$ 3 month/year = 45 months). To overcome the impacts of initial conditions, five more days were simulated for each month and discarded as spin-up. Two sets of simulations were performed to elucidate the relative roles of changes in anthropogenic emissions and meteorological conditions: (1) CTL simulation, simulations of winter periods from 2002 to 2016 with varying meteorological conditions and anthropogenic emissions; and (2) MET simulation, simulations of winter periods from 2002 to 2016 with varying meteorological conditions only, with anthropogenic emissions fixed at levels that applied in 2002. The MET simulation can be used to answer the question: what the climate would have done to the variations of $PM_{2.5}$ if emissions were not changing? The CTL simulation contains the information of changes in both meteorological conditions and emissions.

3 Results and Discussion

3.1 Model evaluation

Model evaluation was conducted in terms of both $PM_{2.5}$ concentrations and $PM_{2.5}$ chemical compositions, using measurements from urban Beijing (location is marked with red dot in Fig. 1). $PM_{2.5}$ was measured using the Two tapered Element Oscillating Microbalances (TEOM) system at the Institute of Atmospheric Physics site. More descriptions of the observations were archived in *Liu et al. (2015)*. Fig. 2 displays the variations of simulated and observed daily mean $PM_{2.5}$ concentrations over winters from 2002 to 2016. Temporal variations of simulated and observed $PM_{2.5}$ concentrations are generally consistent, with correlation coefficients

ranging from 0.75 to 0.83 (Fig. 2). Notably, the model overestimates $PM_{2.5}$ concentrations for all periods. However, the overestimations decline gradually over time. Values of the mean bias decrease from 54.1 $\mu g/m^3$ over the 2004-2006 period to 15.9 $\mu g/m^3$ for recent winters. The broad ranges of errors in different periods reflect the changing uncertainty of emission inventories for different periods. In early times, documentations of emission sources were not as comprehensive as those available for recent years, leading to larger errors in early inventories.

According to results reported by *Zheng et al. (2018)*, SO_2 emissions decreased by 59%, NO_x emissions decreased by 21%, BC emissions decreased by 28%, and OC emissions decreased by 32% for China between 2013 and 2017. These remarkable changes in emissions are expected to lead to notable changes in both the abundance and composition of $PM_{2.5}$. As shown in Fig. 3, sulfate and OC exhibit the largest declines, ammonium and BC show slight decreases, while nitrate concentrations remain relatively stable over 2013-2017. These measured trends are captured generally well by the model, except that sulfate is still underestimated, and BC is overestimated in Beijing. The underestimate of sulfate by models and the overestimate of BC in Beijing have been well documented in previous studies (*Cheng et al., 206; Gao et al., 2016a, 2018a; Song et al., 2018*), attributed to missing reaction pathways and aging/deposition treatments in models (*Song et al., 2019*). Several heterogeneous reaction pathways for sulfate formation have been proposed, including the oxidation of SO_2 by NO_2, transition-meta-catalyzed O_2, or H_2O_2 in aerosol water, and by NO_2 or O_2 on aerosol surfaces (*Cheng et al., 2016; He et al., 2014; Hung et al., 2018; Li et al., 2018*). More recently, *Song et al. (2019)* proposed that the heterogeneous production of hydroxymethanesulfonate (HMS) from the reaction of SO_2 and formaldehyde could be an important chemical mechanism for wintertime haze in China. With the rapid declines in sulfate in Beijing, the relative importance of nitrate in $PM_{2.5}$ is enhanced, a circumstance worthy of special attention for future pollution control policy.

3.2 Influences of Anthropogenic Emissions and Meteorological Conditions on Haze in Beijing

Fig. 4 illustrates the wintertime mean concentrations of $PM_{2.5}$ and the numbers of haze days

from the CTL and MET simulations. Haze days are defined as occasions with daily mean concentrations exceeding 150 $\mu g/m^3$. With fixed anthropogenic emissions, wintertime averaged concentrations of $PM_{2.5}$ would have increased at a rate of 2.1 $\mu g/m^3$/year in Beijing (Fig. 4a). Due to the implementation of China's Clean Air Action, the wintertime averaged concentrations of $PM_{2.5}$ have been declining since 2012. Over 2002-2016 winters, mean concentrations of $PM_{2.5}$ in Beijing decreased at a rate of 1.4 $\mu g/m^3$/year. Compared to concentrations in the MET simulation, the mean mass concentrations of $PM_{2.5}$ decreased by 21% in Beijing over the winters of 2011-2016.

The effectiveness of China's Clean Air Action has been highlighted also in the decline in the number of haze days in Beijing. In the MET simulation, the total number of haze days over 2011-2016 winters amounted to 157 days, reduced by 44 days as a result of the emission controls implemented over this period (Fig. 4b). On average, China's Clean Air Action resulted in 7 fewer haze days per winter over 2011-2016. Over the entire study period 2002-2016, China's Clean Air Action altered the direction of changes in wintertime haze days, with rates changing from 0.8 day/year to -0.3 day/year. The differences in both mean concentrations of $PM_{2.5}$ and numbers of haze days from the two sets of simulations underscore the impressive success of China's Clean Air Action. Given the more serious overestimate of $PM_{2.5}$ in early periods (Fig. 2), the declining rates inferred from the CTL simulation might have been slightly overstated. With unchanged anthropogenic emissions, the increasing trend of haze pollution that would have occurred for Beijing highlights the unfavorable influences of recent changes in local meteorology. We explored if different thresholds of haze days would change the findings and found the variations are similar when a threshold of 75 $\mu g/m^3$ was used (Fig. S2). Due to the overestimation of number of haze days by the model in early years, the declining rates of number of haze days inferred from the CTL simulation might have been overstated. As seen from Fig. S3, there is no notable declining trend in number of haze days inferred from observations over 2005-2016. However, it is consistent that both model and observations indicate rapid declines in number of haze days (-4.8 days and -3.0 days per winter, respectively). With fixed emissions, the MET simulation suggests that unfavorable climate conditions would have led to more haze days, emphasizing the significance of emission control in recent years. Using similar approach, *Cheng et al. (2019)* found that meteorological conditions explain 12.1%

of the improved PM$_{2.5}$ air quality during 2013-2017, while large portions of the improvement are dominated by local (65.4%) and regional (22.5%) emission reductions. Zhou et al. (2019) concluded that emission reductions play a more important than meteorological conditions in determining the declines in PM2.5 over 2011/2012-2017/2018. The current study examines longer term trend since 2002, but the more favorable meteorological conditions mentioned in *Cheng et al. (2019)* are illustrated also in Fig. 4. Our findings highlight also the significance of emission reductions, especially after 2013, while the long-term trend of meteorological conditions since 2002 differs from it during 2013-2017.

3.3 Significance of Different Meteorological Variables

To identify the key meteorological variables for the unfavorable influences on air quality, we applied the stepwise linear regression model (SLR) to determine the relative significance of multiple meteorological variables in terms of their contributions to the variations of PM$_{2.5}$ concentrations. In a SLR model, the selected predictors are included in the regression equation one by one. The predictor that contributes the most to the model is included first, and the process is continued if the additional predictor can statistically improve the regression (*Bendel and Afifi, 1977*). Thus, the SLR model is widely used to select meaningful predictors. In this study, boundary layer heights (BLH), precipitation (PREC), near surface relative humidity (RH2), near surface temperature (T2) and near surface wind speeds (WS10) were selected as predictors for the SLR model. These variables were extracted from the WRF-Chem meteorological simulations with analyses nudging applied. Table 1 summarizes the p values for each predictor. RH2 and WS10 were selected as the most significant predictors for wintertime PM$_{2.5}$ in Beijing (p values < 0.05). The influence of WS10 is greater than that for RH2. *Shen et al. (2018)* reported also that RH and meridional wind speeds drive stagnation and chemical production of PM$_{2.5}$ in Beijing.

The impacts of RH on aerosol composition and processes in winter were examined using measurements in Beijing, and the largest impacts were found for the growth of sulfate and organic aerosols associated with coal combustion (*Sun et al., 2013; Wang et al., 2017*). Although RH has been shown to be a good predictor for PM$_{2.5}$, it decreases slightly (-5.3 %/decade) between 2002 and 2016 period (Fig. 5a), which contradicts the predicted

increasing trend of PM$_{2.5}$ under fixed emissions (Fig. 4a). RH2 has been found to explain the interannual variability of PM$_{2.5}$, but not the increasing trend. Thus, the variability of RH2 is unlikely to be the driver of enhanced PM$_{2.5}$ under changing conditions of climate. As suggested

280   by Ding et al. (2014), the decrease of RH was partly caused by the increase of surface temperature.

Between 2002 and 2016, simulated wintertime WS10 in Beijing declined gradually at a rate of 0.3 m/s/decade (Fig. 5b), in agreement with the declining trends inferred from observations (Fig. S4). Weaker wintertime near surface wind speeds are associated with enhanced PM$_{2.5}$

285   concentrations and increasing numbers of haze days since 2002. In winter, the North China plain features northwesterly winds associated with the East Asian Winter Monsoon (EAWM), properties of which depend largely on the development of both the Siberian high and the Aleutian low (*Jhun and Lee, 2004*). We calculated the intensity of the EAWM index using the pressure difference between area mean sea level pressure (SLP, data were taken from MERRA-

290   2 reanalysis) over the region 90º E-110º E, 40º N-50º N and the area mean SLP over the region 120º -170º E, 40º-60º N. These two regions represent the central focal areas of the Siberian high and the Aleutian low, respectively. Fig. 5(c) indicates that the variations of EAWM intensity has declined gradually over time, consistent with the variations of wind speeds in Beijing. The weakening of EAWM intensity is due partially to the increasing pressure in the

295   regions of the Aleutian low. As shown in Fig. 6, there is no significant change in SLP over the regions of the Siberian high, but SLP of the Aleutian low intensity decreased significantly between 2002 and 2016 (more than 50 Pa/year). Although no significant trend was observed for the strength of the Siberian high, the changes in the position of the Siberian high has been linked to wintertime air quality in China over the past decades (*Jia et al., 2015*).

300   *Yin et al. (2015)* found a significant negative correlation between winter haze and the East Asia Winter Monsoon from 1986 to 2010. Deterioration of air quality in China has been linked also to a weaker East Asia Summer Monsoon (*Chin et al., 2012*). The slacking of winds is observed not only for China (*Sherman et al., 2017*), but also for other countries including India (*Gao et al., 2018b*). Analyses using climate projections suggest that wind speeds in continental regions

305   in the Northern Hemisphere will continue to decline under a warming climate (*Karnauskas et al., 2018*), imposing greater pressure on measures for future control of air pollution.

4 Summary

Combining field measurements and model simulations, we quantified the relative influences on $PM_{2.5}$ concentrations in Beijing of anthropogenic emissions and meteorological conditions over winters of 2002-2016. China's Clean Air Action has been effective in reducing both mass concentrations of $PM_{2.5}$ and the number of haze days. With fixed emissions, meteorological conditions over the study period should have resulted in an increase in haze in Beijing, but the strict emission control measures implemented by the government have suppressed the unfavorable influences associated with recent climate. Using a statistical method, we concluded that RH2 and WS10 offer useful predictors for wintertime $PM_{2.5}$ in Beijing, with the variations of WS10 in particularly good agreement with the increasing trend of $PM_{2.5}$ concentrations with fixed emissions. The increasing trend of $PM_{2.5}$ under unfavorable meteorological conditions was attributed further to the weakening of Aleutian low and the EAWM. The variations of $PM_{2.5}$ concentrations were investigated in this study for the urban Beijing region, so were the changes in wind speeds. We do not exclude the possibility that the trend may have been influenced by other factors, including for example increases in surface roughness (*Vautard et al, 2010*).

4 figures are listed in the supplement.

Author contribution

M.G. and M.B.M designed the study; M.G. performed model simulations and analyzed the data with the help from S.S., P.S., J.X., J.K.J. and G.R.C.; B.Z. and Q. Z. provided the emission inventory; Z.W., Y.W., Z.L., D.J., J.Y.X., and C.L. provided measurements. M.G. and M.B.M. wrote the paper with inputs from all other authors.

Data availability

The measurements and model simulations data can be accessed through contacting the corresponding authors.

Competing interests

340   The authors declare that they have no conflict of interests.

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Table 1. p-values for stepwise linear regression model for Beijing.

Meteorological Variables	p-values
BLH	0.38
PREC	0.64
RH2	0.02
T2	0.95
WS10	0.00

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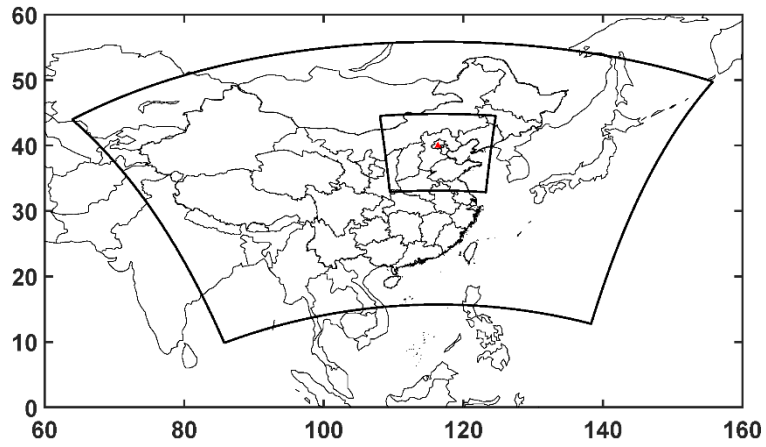
Fig.1. WRF-Chem modeling domain settings and locations of observations.

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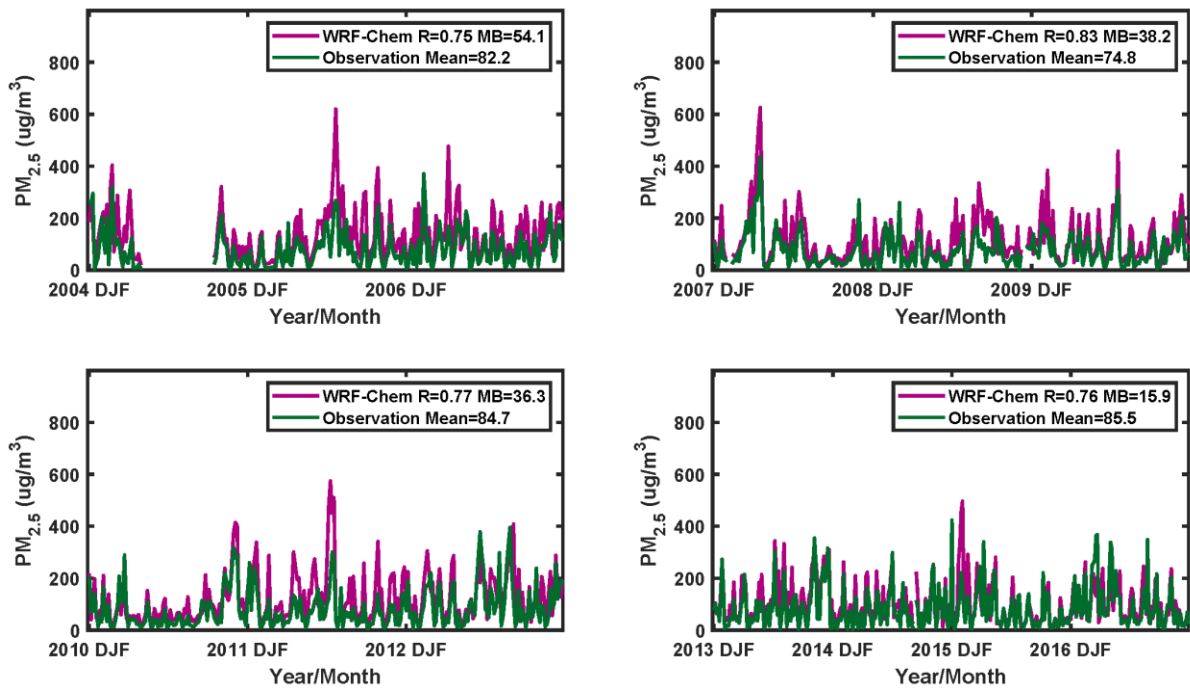
Fig. 2. Simulated and observed temporal variations of daily mean PM$_{2.5}$ concentrations in Beijing, with correlation coefficient and mean bias.

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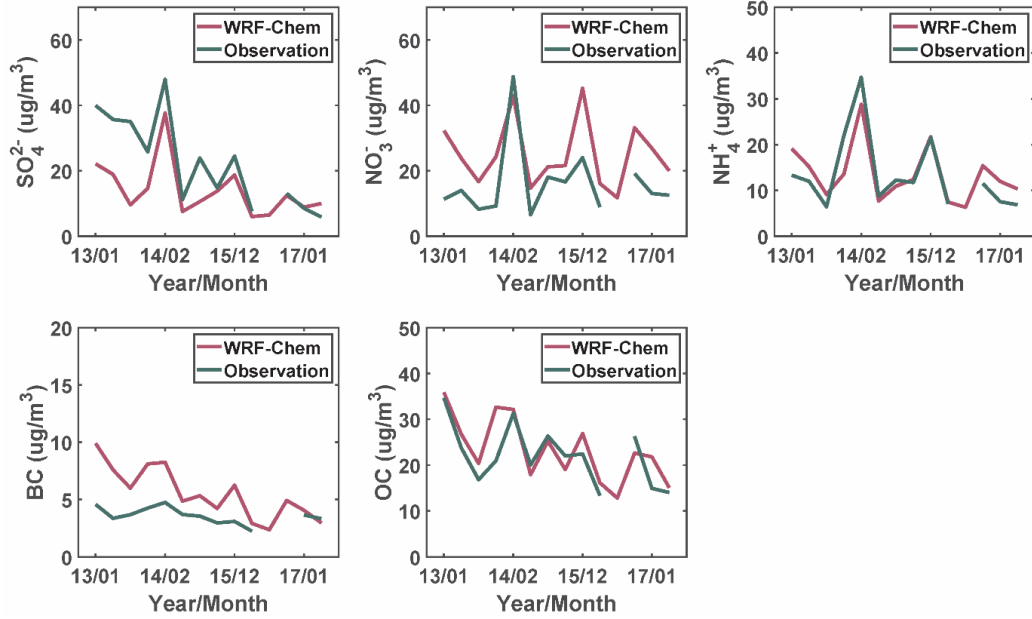
Fig. 3. Simulated and observed temporal variations of monthly mean concentrations of PM$_{2.5}$ chemical from 2013 to 2017.

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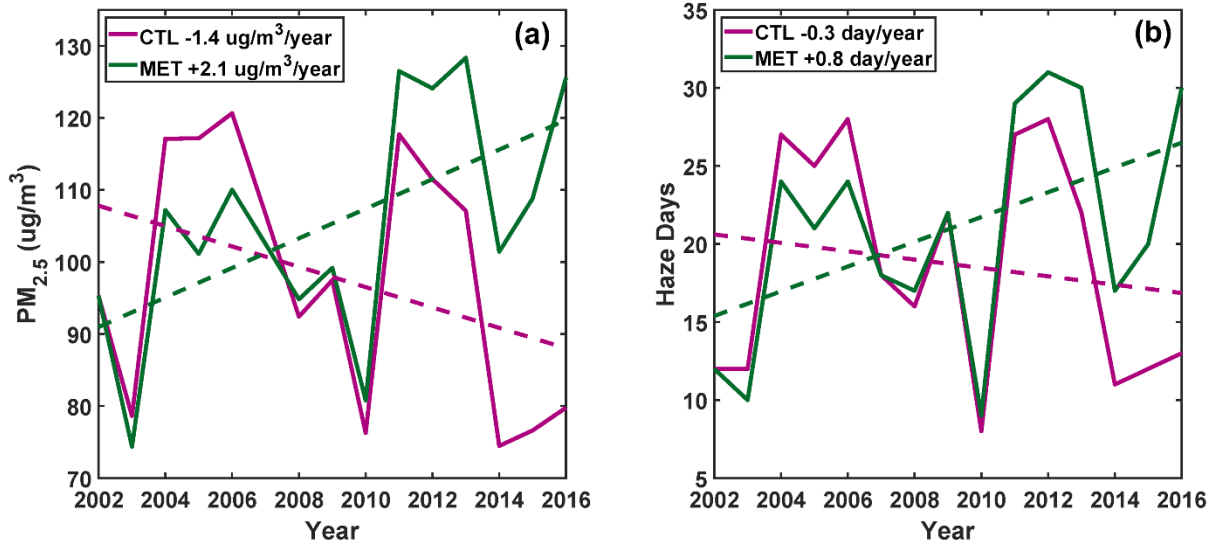
Fig. 4. Wintertime mean concentrations of PM$_{2.5}$ and number of haze days (defined with daily mean concentration above 150 µg/m^3) from the CTL and MET simulations.

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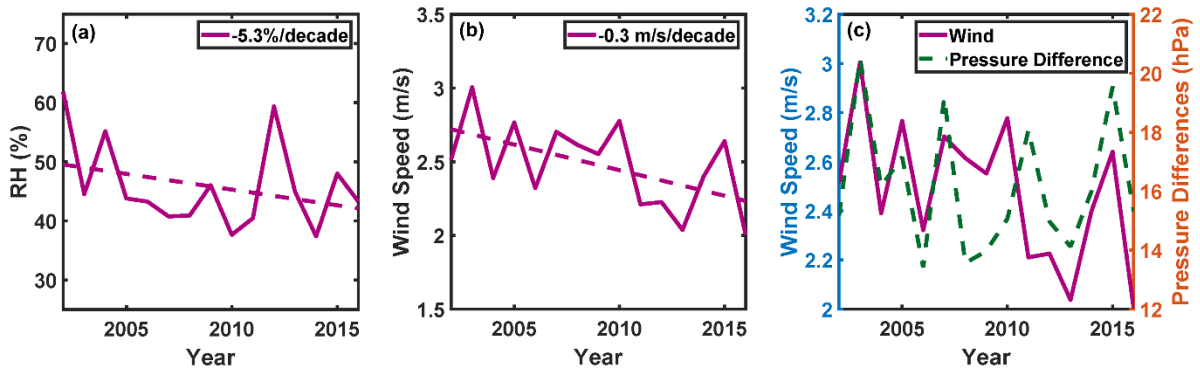
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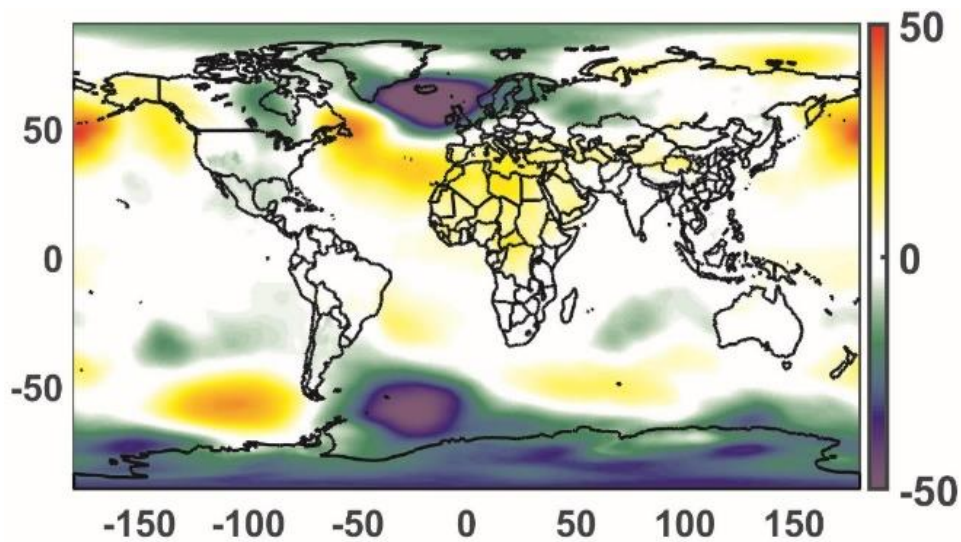
Fig. 5. Simulated winter mean RH, wind speeds in Beijing with declining rates, and pressure difference indicating the intensity of East Asia Winter Monsoon.

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Fig. 6. Trends of winter sea level pressure during 2002-2016 period (unit: Pa/year).

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