



1 Contributions of Nordic anthropogenic emissions on air pollution and 2 premature mortality over the Nordic region and the Arctic

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4 Ulas Im^{1,2}, Jesper H. Christensen^{1,2}, Ole-Kenneth Nielsen^{1,2}, Maria Sand³, Risto Makkonen^{4,5},
5 Camilla Geels^{1,2}, Camilla Anderson⁶, Jaakko Kukkonen⁴, Susana Lopez-Aparicio⁷, Jørgen Brandt^{1,2}
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7
8 1 Aarhus University, Department of Environmental Science, Atmospheric Modelling Section,
9 Frederiksborgvej 399, Roskilde, Denmark.

10 2 Interdisciplinary Center for Climate Change (iCLIMATE), Frederiksborgvej 399, Roskilde,
11 Denmark.

12 3 Center for International Climate Research, Postboks 1129 Blindern, 0318 Oslo, Norway.

13 4 Finnish Meteorological Institute, Erik Palmenin aukio 1, P.O.Box 503, FI-00101, Helsinki,
14 Finland.

15 5 University of Helsinki, Institute for Atmospheric and Earth System Research, P.O. Box 64,
16 00014, Helsinki, Finland.

17 6 Swedish Meteorological and Hydrological Institute, SE-60176 Norrköping, Sweden.

18 7 NILU - Norwegian Institute for Air Research, Instituttveien 18, P.O. Box 100, 2027 Kjeller,
19 Norway.

20 21 Abstract

22
23 This modelling study presents the sectoral contributions of anthropogenic emissions in the four
24 Nordic countries; Denmark, Finland, Norway and Sweden, on air pollution levels and the associated
25 health impacts and costs over the Nordic and the Arctic region for the year 2015. The Danish
26 Eulerian Hemispheric Model (DEHM) has been used on a 50 km resolution over Europe in tagged
27 mode in order to calculate the response of a 30% reduction of each emission sector in each Nordic
28 country individually. The emission sectors considered in the study were energy production,
29 residential/commercial heating, industry, traffic, off-road mobile sources, and waste
30 management/agriculture. In total, 28 simulations were carried out. Following the air pollution
31 modelling, the Economic Valuation of Air Pollution (EVA) model has been used to calculate the
32 associated premature mortality and their costs. Results showed that more than 80% of PM_{2.5}
33 concentrations in the considered four Nordic countries were transported from outside these four
34 countries. The leading emission sector in each country was found to be non-industrial combustion
35 (contributing by more than 60% to the total PM_{2.5} mass), except for Sweden, where industry
36 contributed to PM_{2.5} with a comparable amount as non-industrial combustion. In addition to
37 residential combustion, the next most important source categories were industry, agriculture and
38 traffic. The main chemical constituent of PM_{2.5} concentrations was organic carbon in all countries,
39 which suggested that residential wood burning was the dominant national source of pollution in the
40 Nordic countries. We have estimated the total number of premature mortality cases due to air
41 pollution to be around 4 000 in Denmark and Sweden and around 2 000 in Finland and Norway.
42 These premature mortality cases led to a total cost of 7 billion Euros in the selected Nordic
43 countries. The assessment of the related premature mortality and associated cost estimates
44 suggested that residential combustion, together with industry and traffic, will be the main sectors to
45 be targeted in emission mitigation strategies in the future.
46
47



48 1. Introduction

49

50 Air pollution is the world's largest single environmental health risk (WHO, 2014), estimated to be
51 responsible for 3.7 million premature deaths in 2012 from urban and rural sources worldwide. In
52 Europe, recent results (Andersson et al., 2009; Brandt et al., 2013a; 2013b; Geels et al., 2015; Im et
53 al., 2018a; Liang et al., 2018; Solazzo et al., 2018) show that outdoor air pollution causes ~500 000
54 premature deaths in Europe. Brandt et al. (2013a) calculated that due to exposure to ambient air
55 pollution, there were around 3.500 premature deaths in 2011 in Denmark alone. Lehtomäki et al.
56 (2018) have recently evaluated that ambient air pollution caused approximately 2000 premature
57 deaths in Finland in 2015. Other studies have made assessments for some of the Nordic countries
58 (Denmark, Sweden and Finland) with estimates ranging from 6500 to 9500 for the year 2000 (Geels
59 et al., 2014; Watkiss et al., 2005, Karvosenoja et al., 2010, respectively). Kukkonen et al. (2018)
60 and Forsberg et al. (2015) have concluded that long-range transported fine particulate matter
61 dominates the health effects in the Scandinavian countries, with largest contribution to long-term
62 effects in Sweden originate from south-western Europe, while the largest contribution to short-term
63 exposure originates from south-eastern Europe (Jönsson et al. 2013).

64

65 Air pollution is a transboundary problem covering global, regional, national and local sources,
66 leading to large geographic variability and therefore to large differences in the geographical
67 distribution of human exposure to air pollution (Im et al., 2018a,b). In the Nordic countries, there
68 are large spatial differences in air pollution levels because of long-range transported and polluted
69 air masses especially from the south and east as well as due to the degree of urbanization. There are
70 also local differences depending on wind direction and distance from local emission sources such as
71 road transport, power plants and industry (Brandt et al., 2013a). Furthermore, the widespread use of
72 domestic wood stoves in the Nordic countries represents a special challenge for exposure to air
73 pollution, where e.g. more than a third of the health impacts from Danish emissions are due to
74 smoke from wood stoves. International ship traffic is also a significant source of air pollution and
75 health impacts in highly trafficked areas of the Baltic and North Seas (Brandt et al., 2013b;
76 Jalkanen et al., 2016, Johansson et al., 2017). Based on simulations for the period 1997-2003,
77 Andersson et al. (2009) calculated that Sweden contributed to 1.4% of the European primary $PM_{2.5}$
78 ($PPM_{2.5}$) concentrations while Denmark, Finland and Norway were responsible for 4% of European
79 $PPM_{2.5}$. Contribution to secondary inorganic $PM_{2.5}$ (SIA) levels were much smaller (0.5% from
80 Sweden and 1.4% from Denmark, Finland and Norway). They also calculated a death rate increase
81 of 2 and 3% due to exposure to $PPM_{2.5}$ and SIA, respectively, in Europe due to emissions from
82 Denmark, Finland, Norway and Sweden.

83

84 The external (or indirect) costs to society related to health impacts from air pollution are substantial.
85 In the whole of Europe, the total external costs have been estimated to be approx. 800 billion Euros
86 per year and in Denmark alone the external costs are nearly 4 billion Euro per year (Brandt et al.,
87 2013a). In a more recent study, Im et al. (2018a), using a multi-model ensemble of 14 chemistry
88 transport models (CTM), estimated that ambient air pollution in Europe in 2010 was responsible for
89 $414\ 000 \pm 100\ 000$ premature deaths, leading to a cost of 300 billion Euros. The study also showed
90 that a 20% decrease of anthropogenic emissions in Europe source could avoid 47 000 premature
91 deaths in Europe, while a similar reduction in the U.S. would avoid around 1 000 premature deaths
92 in Europe due to long-range transport.

93

94 The Nordic countries are generally characterized with low air pollution levels compared to the rest
95 of Europe. However, there are still large impacts of air pollution on human health and climate in the



96 region itself (Brandt et al., 2013a; Forsberg et al., 2015), as well as over the Arctic (Sand et al.,
97 2015). The Task Force on Short Lived Climate Forcers of the Arctic Council reported that black
98 carbon (BC) emission sources within Arctic Council nations generally have a greater impact on
99 climate change per unit of emissions compared to sources outside of the Arctic (Arctic Council,
100 2011). The report also states that measures aimed at decreasing these emissions will have positive
101 health effects for communities exposed to air pollution. In a recent study, Sand et al. (2015) showed
102 that although the largest Arctic warming source is from Asian emissions, the Arctic is most
103 sensitive, per unit mass emitted, to SLCFs emissions from a small number of activities within the
104 Arctic nations themselves.

105
106 The aim of the study is to quantify the contributions of the main emission sectors in each of the
107 Nordic countries to air pollutant levels and their impacts on premature mortality and associated
108 costs in the Nordic region and the Arctic, in order to identify the emission sectors that should be
109 targeted for mitigation to decrease the air pollution and exposure levels in the Nordic countries. In
110 order to achieve this, we have coupled the Danish Eulerian Hemispheric Model (DEHM) to the
111 Economic Valuation of Air Pollution (EVA) model and conducted a number of perturbation
112 simulations targeting different emission sectors in the four Nordic countries; Denmark, Finland,
113 Norway and Sweden, for the year 2015. The models and perturbation simulations are described in
114 Section 2, the model evaluation against surface measurements in the Nordic countries are presented
115 in Section 3.1, the contributions of sectoral emissions on the air pollution levels in the Nordic
116 region and the Arctic are presented in Section 3.2., and the health impacts and associated costs are
117 presented in Section 3.3. Conclusions are given in Section 4.

118

119 2. Materials and methods

120

121 2.1. DEHM

122

123 DEHM model was originally developed mainly to study the transport of SO₂ and SO₄ to the Arctic
124 (Christensen 1997), but has been extended to different applications during the last decades. It has
125 been documented extensively in Brandt et al. (2012) and evaluated in several intercomparison
126 studies (e.g. Solazzo et al., 2012 a,b; Solazzo et al., 2017; Im et al., 2018a,b) and recently joined the
127 suit of operational models in the Copernicus Atmospheric Monitoring System (CAMS) to provide
128 regional forecasts of air pollution over Europe. The DEHM model uses a 150 km×150 km spatial
129 resolution over the Northern Hemisphere, then nests to 50 km×50 km resolution over Europe,
130 extending up to 100 hPa through 29 vertical levels, with the first layer height of approximately 20
131 m. The meteorological fields were simulated by the Weather Research and Forecast Model (WRF,
132 Skamarock et al., 2008) setup with identical domains and resolution. The gas-phase chemistry
133 module includes 58 chemical species, 9 primary particles and 122 chemical reactions (Brandt et al.,
134 2012). The model also describes atmospheric transport and chemistry of lead, mercury, CO₂, as
135 well as POPs. Secondary organic aerosols (SOA) are calculated using the Volatility Base System
136 (VBS: Bergstrom et al., 2012).

137

138 In the current study, the DEHM model used anthropogenic emissions from the EDGAR-HTAP
139 database and biogenic emissions are calculated online based on the MEGAN model. The total
140 emission per country for the different pollutants are presented in Table 1. The sectoral distributions
141 of emissions in each country are presented in Fig. 1. As seen in the Table 2, most SNAP (Selected
142 Nomenclature for Air Pollutants; CEIP, 2019) sectors are considered individually, while some are
143 merged in order to reduce the computational costs. All sectors in relation to industrial activities



144 (combustion, processes, solvent use and extraction and transport of fossil fuels) are merged into an
 145 “Industry” source sector, while waste management and agriculture sectors were lumped into
 146 “Others” source sector.

147
 148 As seen in Fig.1, non-industrial combustion (orange bars), where residential combustion dominates,
 149 stands out as a major source contributing to CO and PM emissions while industry (grey bars) (Table
 150 2) is the largest source of NMVOCs, NOx and SOx. Traffic (yellow bars) also contributes
 151 significantly to CO and NOx. The largest source of NH3 is from agriculture and waste
 152 management, as seen in the ‘Other’ (green bars) (Table 2).

153

154 Table 1. Total pollutant emissions in the Nordic countries (in Gg) in 2015.

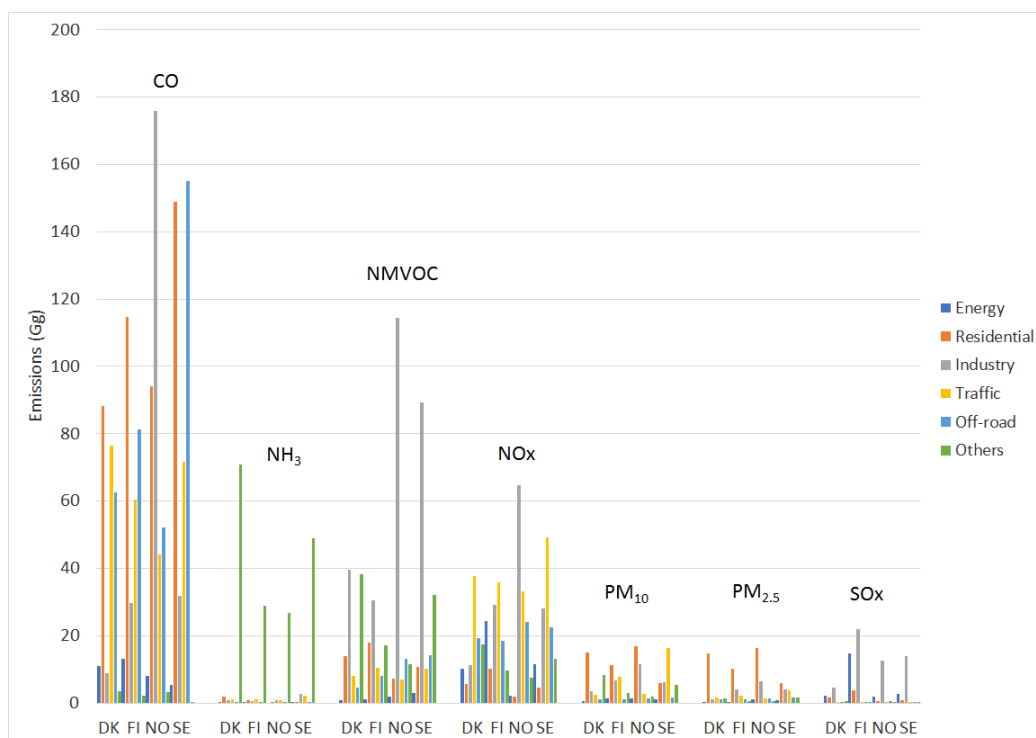
155

	CO	NH ₃	NMVOC	NO _x	SO ₂	PM ₁₀	PM _{2.5}
DK	251	75	106	102	9	31	20
FI	302	31	85	128	41	31	19
NO	378	28	155	133	16	35	27
SE	413	54	159	129	18	37	18

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160

161 Fig.1. Sectoral emissions of major pollutants in the Nordic countries.



162

163 2.1.1. Tagging Method

164

165 Tagging method keeps track of contributions to the concentration field from a particular emission
166 source or sector, as explained in detail in Brandt et al. (2013a). Tagging involves modelling the
167 background concentrations and the δ -concentrations (the contributions from a specific emission
168 source or sector to the overall air pollution levels) in parallel, where special treatment is required for
169 the non-linear process of atmospheric chemistry, since the δ -concentrations are strongly influenced
170 by the background concentrations in such processes. Although this treatment involves taking the
171 difference of two concentration fields, it does not magnify the spurious oscillations (the Gibbs
172 phenomenon), which are primarily generated in the advection step. The non-linear effects can be
173 accounted for in the δ -concentrations without losing track of the contributions arising from the
174 specific emission source or sector.

175

176 2.1.2. Model evaluation

177

178 Surface concentrations modelled by the DEHM model were evaluated against data at selected urban
179 background and regional or global monitoring stations in each Nordic country. The statistical
180 comparisons included using correlation coefficient (r), mean bias (MB) and normalized mean error
181 (NME) and root mean square error ($RMSE$), shown in the supplementary material. The station
182 information is provided in Table S1.

183

184 The Danish Air Quality Monitoring Programme consists of an urban monitoring network that
185 includes stations in the four largest Danish cities (Aalborg, Aarhus, Copenhagen and Odense) and
186 two background stations in rural areas (Keldsnor and Risø). The design of the Danish air quality
187 monitoring network is based on location of one or more pairs of stations in each of the four cities
188 (Ellermann et al., 2015). In each city one of the stations is located at a pollution hot spot close (at
189 the sidewalk) to a street lane with a high traffic density. The other station is located within a few
190 hundred meters from the street station. It is placed so that it is not influenced by emissions from a
191 single or a few streets or other nearby sources and hence is representative for the urban background
192 pollution. In most cases the background stations are placed on rooftops. In addition, rural stations
193 monitor the pollution outside city areas.

194

195 The measurement data for Finland represents regional and urban background levels. Data from the
196 global and regional background stations are reported to European Monitoring and Evaluation
197 Programme (EMEP) under the CLRTAP (Convention on Long-range Transboundary Air Pollution,
198 <http://www.unece.org/env/lrtap>), and are available at <http://ebas.nilu.no>. The data for the urban
199 background stations are reported at the 'Air Quality in Finland' web pages by the Finnish
200 Meteorological Institute (<https://en.ilmatieteenlaitos.fi/air-quality>).

201

202 The measurement dataset from Norway is from the national monitoring program of air pollutants
203 financed by the Norwegian Environment Agency (Aas et al, 2018), and also reported to European
204 Monitoring and Evaluation Programme (EMEP) under the CLRTAP (Convention on Long-range
205 Transboundary Air Pollution, <http://www.unece.org/env/lrtap>). The data is openly available at
206 <http://ebas.nilu.no>. The data from the city background stations is reported to EEA (European
207 Environmental Agency, <http://www.eea.europa.eu/>) as required in the EU air quality directive (EU,
208 2008) and it is available at <http://www.luftkvalitet.info>.

209



210 The measurement dataset for Sweden were extracted from the openly available Shair data base
211 (<http://shair.smhi.se/portal/concentrations-in-air>), which includes most national environmental data
212 and is financed by the Swedish Environmental Agency. The observation sites used here were
213 carefully selected to represent urban background at rooftop level, rural or regional background, and
214 to have known good quality.

215 216 2.2. EVA

217
218 The EVA system (Brandt et al., 2013a,b; Geels et al., 2015; Im et al., 2018) is based on the impact-
219 pathway chain method (Friedrich and Bickel, 2001), and it calculates health impacts of ambient air
220 pollution due to exposure to surface concentrations of O₃, CO, SO₂ and PM_{2.5}, and the associated
221 external costs. The EVA system requires gridded concentrations along with gridded population
222 data, exposure-response functions (ERFs) for health impacts, and economic valuation functions of
223 the impacts from air pollution. The EVA system can estimate various health impacts, including
224 different morbidity outcomes as well as acute and chronic mortality, related to short term (acute)
225 exposure to O₃, CO, and SO₂, and long term (chronic) exposure to PM_{2.5}. EVA calculates and uses
226 the annual mean concentrations of CO, SO₂, and PM_{2.5}, while for O₃, it uses the SOMO35 metric
227 that is defined as the annual sum of the daily maximum of 8-hour running average over 35 ppb,
228 following WHO (2013) and EEA (2017). In addition, EVA uses population densities over fixed age
229 intervals, corresponding to babies, children, adults and elderlies.

230
231 Exposure response functions (ERF) for all-cause chronic mortality due to PM_{2.5} are based on Pope
232 et al., 2002; Krewski et al., 2009; WHO, 2013). Following Pope et al. (2002), the relative risk (RR)
233 is 1.062 (1.040-1.083) on 95% confidence interval. The counterfactual PM_{2.5} concentration is
234 assumed to be 0 µg m⁻³ following the EEA methodology, meaning that the impacts have been
235 estimated for the full range of modelled concentrations. Regarding short-term exposure to O₃, EVA
236 uses the ERF recommended by the CAFE Programme (Hurley et al., 2005) and WHO (2013) that
237 uses the daily maximum of 8-hour mean O₃ concentrations. The ERFs used in EVA to calculate
238 mortality are presented in Table S2. For the valuation of the health impacts, a value of EUR 1.5
239 million was applied for preventing an acute death, following expert panel advice (EC, 2001), while
240 for the valuation of a life year, a value of EUR 57 500 per year of life lost (YOLL) were applied
241 (Alberini et al., 2006). More details can be found in Im et al. (2018a).

242 243 2.3. Scenarios (response and contribution)

244
245 We have applied a 30% reduction on land-based anthropogenic emissions from each of the
246 continental Nordic countries, which include Denmark, Finland, Norway and Sweden. The
247 perturbations are applied based on the SNAP sectors. Each simulation perturbed a SNAP sector
248 from an individual Nordic country, which are listed in Table 2.

249
250 DEHM model has been run on “tagged” mode, explained in section 2.1., so each simulation
251 included a “perturbed” and “non-perturbed” concentration, which we used to calculate the response
252 to the 30% reduction in the particular country and sector. These responses are then converted to
253 contributions by assuming a linear extrapolation to 100%. We have also simulated a 100%
254 reduction scenario to all sectors per country (“All” in Table 2) to see the impact of a 100%
255 reduction and how it compares to the scaled 30% response at each country.

256
257



258 Table 2. Source sectors used in the perturbation scenarios.

259

Source Sectors	SNAP Code
Combustion in energy and transformation industries	1
Non-industrial Combustion	2
Industry	3,4,5,6
Road transport	7
Other mobile sources and machinery	8
Others (waste and agriculture)	9,10
All	1,2,3,4,5,6,7,8,9,10

260

261

262 3. Results and Discussion

263

264 3.1. Evaluation

265

266 Surface ozone and PM_{2.5} concentrations calculated by the DEHM model have been evaluated using
 267 surface observations from the Nordic countries, described in 2.1.2. The comparison of the mean of
 268 all observed concentrations in each country and the corresponding modelled concentrations are
 269 presented in Table 3 while Figs. 2 and 3 present Taylor diagrams for each station in each Nordic
 270 country. As seen in Table 3, O₃ levels are well reproduced by the DEHM model over all countries
 271 ($r > 0.7$), however with a slight overestimation of ~10% over Denmark, Finland and Sweden, and
 272 ~30% over Norway. The monthly variations of PM_{2.5} levels, averaged over all stations in each
 273 Nordic country are well reproduced for Denmark and Norway ($r \sim 0.7$), moderately over Sweden
 274 and poorly ($r \sim 0$) over Finland (Table 3).

275

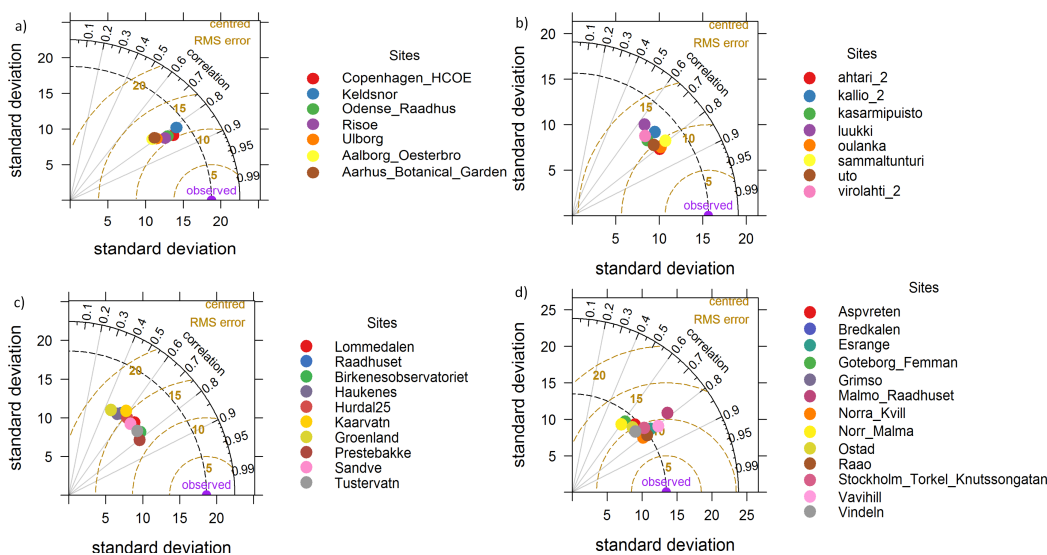
276 Table 3. Model evaluation for the daily mean concentrations of O₃ and PM_{2.5} for all the selected
277 stations in the Nordic countries.

278

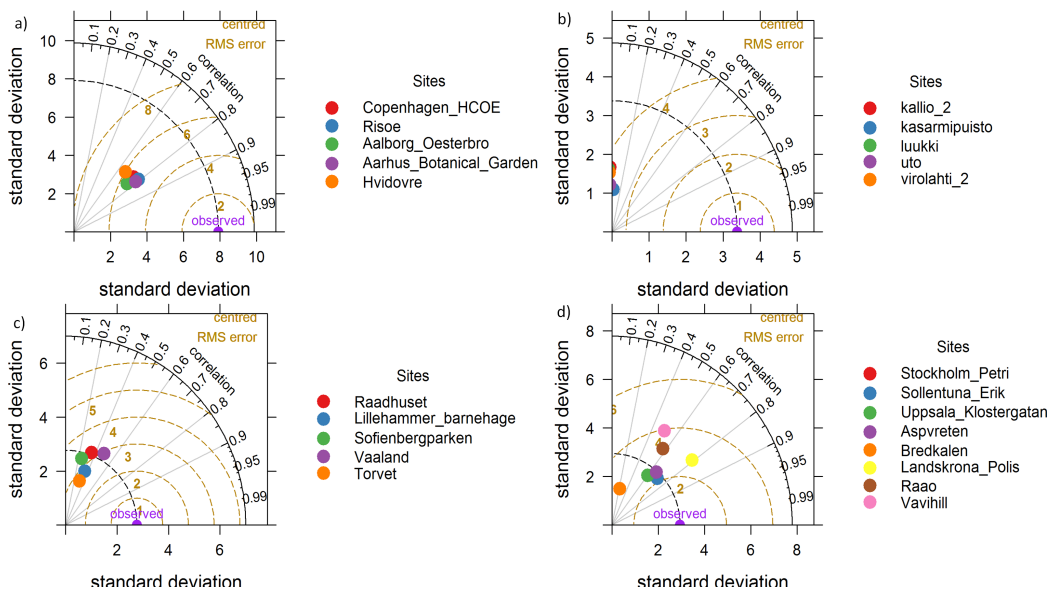
	O ₃				PM _{2.5}			
	R	MB ($\mu\text{g m}^{-3}$)	NME (%)	RMSE ($\mu\text{g m}^{-3}$)	R	MB ($\mu\text{g m}^{-3}$)	NME (%)	RMSE ($\mu\text{g m}^{-3}$)
Denmark	0.81	5.67	0.16	11.60	0.75	-3.41	0.36	6.22
Finland	0.74	4.77	0.19	12.44	-0.03	-0.80	0.52	3.83
Norway	0.64	12.02	0.32	18.31	0.35	-2.56	0.47	4.52
Sweden	0.74	7.00	0.19	13.25	0.59	0.33	0.50	3.23

279

280



281
282 Fig. 2. Taylor diagrams for daily mean O₃ for all stations in a) Denmark, b) Finland, c) Norway and
283 d) Sweden.
284
285
286



287
288 Fig. 3. Taylor diagrams for daily mean PM_{2.5} for all stations in a) Denmark, b) Finland, c) Norway
289 and d) Sweden.
290
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292

3.2. Sectoral contributions to surface concentrations



293

294 *3.2.1. Nordic countries*

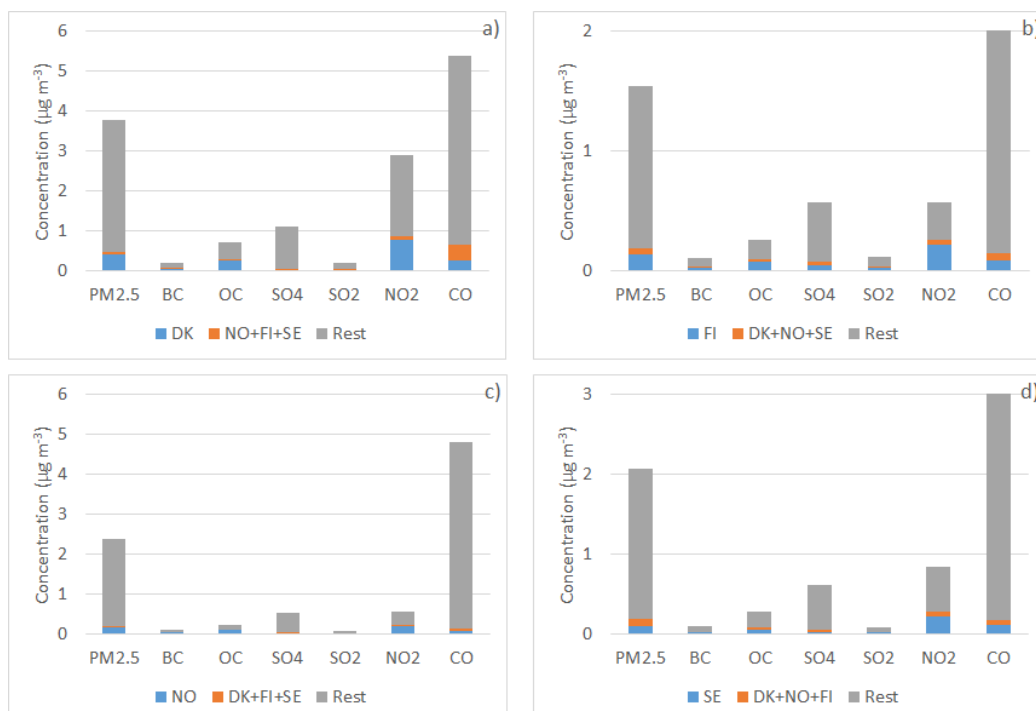
295

296 In general, the long-term transport of air pollutants from one country to another is dependent on the
 297 global and regional atmospheric circulation and on the relative geographic positions of the
 298 countries. Nordic countries are influenced by substantial long-range transported contributions of air
 299 pollution especially from the central, western and central eastern parts of Europe. In the region
 300 containing the continental Nordic countries, the prevailing atmospheric flow directions near the
 301 ground surface are from the west, south-west and south. Caused by the atmospheric circulation
 302 patterns, it is therefore to be expected that, e.g., the emissions in Denmark will have a relatively
 303 larger influence on the pollution levels in the other Nordic countries than those in Finland.

304

305 Fig. 4 compares the contribution of the total contribution of each Nordic country on the surface
 306 concentrations over the country itself, with contributions from rest of the Scandinavian countries
 307 and rest of the world. The figure clearly shows that over 90% or more of PM_{2.5} surface levels are
 308 coming outside each country. Similar high contributions for other species including CO also shows
 309 that Scandinavian countries are exposed to airmasses coming from rest of the world while local
 310 pollution is low. The figure also shows that PM_{2.5} levels are generally low in the Scandinavian
 311 countries, with annual means of 2–4 $\mu\text{g m}^{-3}$ (highest in Denmark and lowest in Finland). Similar to
 312 PM_{2.5}, annual mean surface O₃ levels are also low ($\sim 30 \mu\text{g m}^{-3}$).

313

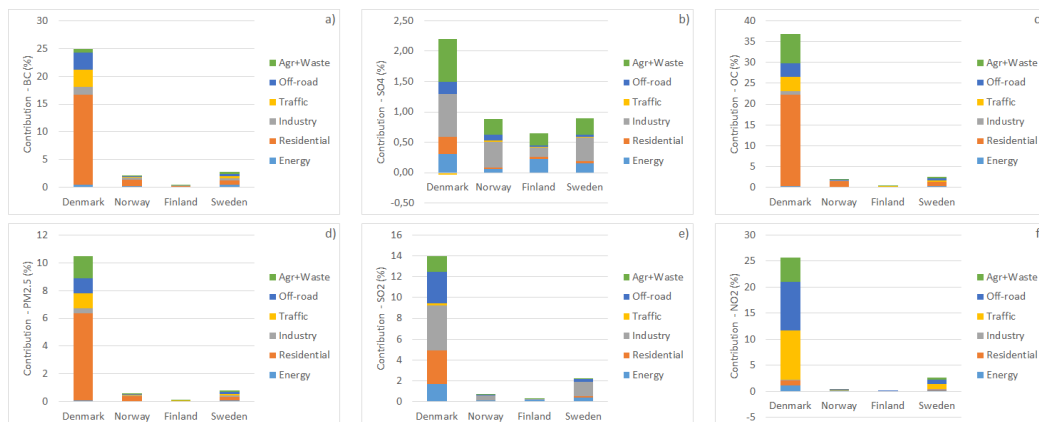


314

315 Fig. 4. Contribution of national, Scandinavian and other sources on the surface levels of major air
 316 pollutants over a) Denmark, b) Finland, c) Norway and d) Sweden. Note that CO concentrations are
 317 divided by 20 to scale with other pollutants.

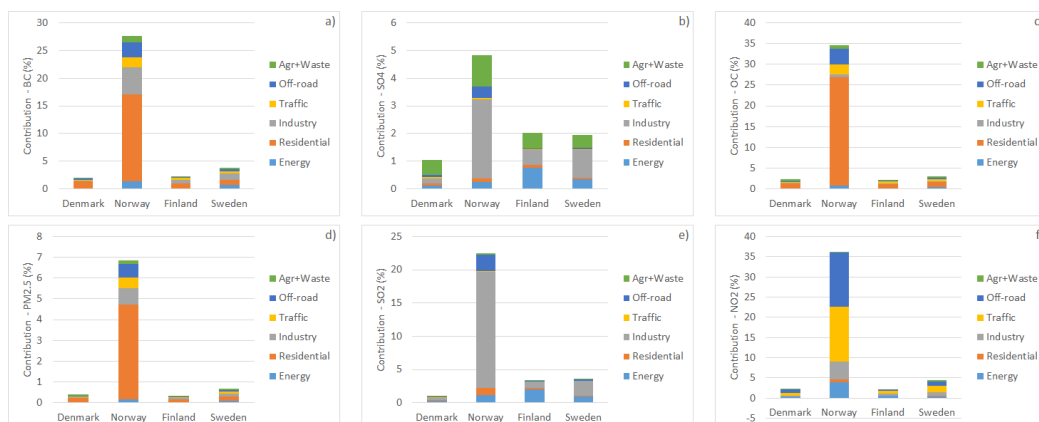


318 Danish emissions contribute to only $0.4 \mu\text{g m}^{-3}$ (10%) of the surface $\text{PM}_{2.5}$ concentrations over
319 Denmark ($3.8 \mu\text{g m}^{-3}$), while contributions to other Nordic countries are less than 1% (Fig. 5). Non-
320 industrial combustion (SNAP2), which is dominated by residential combustion, is responsible for
321 $0.24 \mu\text{g m}^{-3}$ (65%) of the Danish contribution to surface $\text{PM}_{2.5}$ concentrations over Denmark.
322 Residential combustion contributes to $0.16 \mu\text{g m}^{-3}$ (60%) of the Danish contribution to surface
323 organic carbon (OC) concentrations over the country, suggesting the importance of residential wood
324 burning for heating. Industry contributes to $7 \mu\text{g m}^{-3}$ (31%) of the Danish contribution to the
325 surface SO_2 concentrations over Denmark, while on-road and off-road transport contributes equally
326 to the Danish share of the in surface NO_2 concentrations by $0.55 \mu\text{g m}^{-3}$ (~72% together).
327 Agriculture and waste handling are important sources for surface SO_4 levels over Denmark as well
328 as over the other Nordic countries, via the formation of ammonium sulfate $((\text{NH}_4)_2\text{SO}_4)$ due to the
329 large ammonia (NH_3) emissions from these sectors. Among the other Nordic countries, Danish
330 emissions, in particular non-industrial combustion, have the largest contribution to the pollutant
331 levels over Sweden.
332



333 Fig. 5. Contributions of sectoral Danish emissions on surface a) BC, b) SO_4 , c) OC, d) $\text{PM}_{2.5}$, e) SO_2
334 and f) NO_2 over the Nordic countries.
335
336

337 Contributions of the Norwegian emissions over the Nordic countries are presented in Fig. 6. Similar
338 to the Danish emissions, Norwegian emissions contribute to $0.17 \mu\text{g m}^{-3}$ (7%) of the surface $\text{PM}_{2.5}$
339 concentrations over Norway, while contributions to other Nordic countries are below 1%, except for
340 NO_2 , where on-road transport emissions from Norway contributes to almost $0.08 \mu\text{g m}^{-3}$ (15%) of
341 the surface NO_2 levels over Finland. Non-industrial combustion is the main source of pollutant
342 levels, in particular for OC, where Norwegian emissions are responsible for $0.07 \mu\text{g m}^{-3}$ (77%) of
343 local contribution to the surface OC levels over Norway. Industry is a major source of surface SO_2
344 levels over Norway, contributing to $0.01 \mu\text{g m}^{-3}$ (76%) of the local contribution.
345



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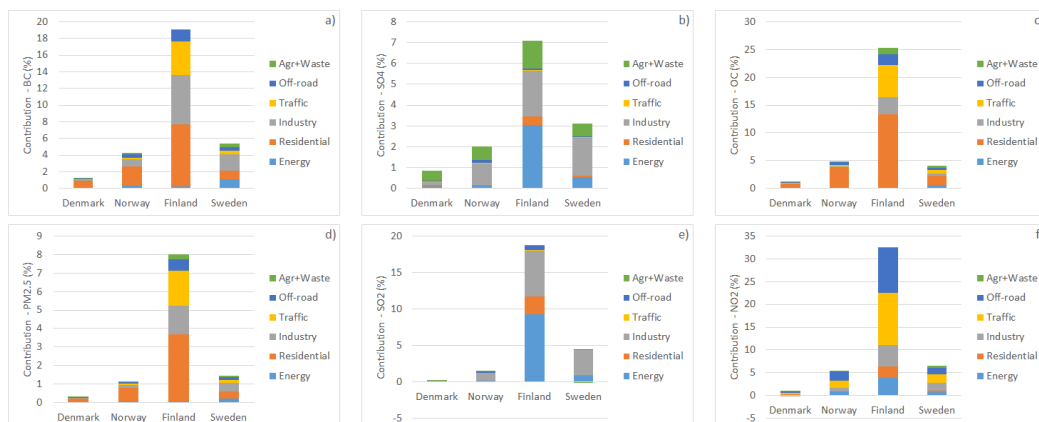
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348 Fig. 6. Contributions of sectoral Norwegian emissions on surface a) BC, b) SO₄, c) OC, d) PM_{2.5}, e)349 SO₂ and f) NO₂ over the Nordic countries.

350

351 Fig. 7 shows the contributions of Finnish emissions on the pollutant levels over the Nordic
 352 countries. Similar to Denmark and Norway, non-industrial combustion, which is dominated by
 353 residential combustion, is the major source of pollution over Finland, although contributions are
 354 lower compared to Denmark and Norway (0.06 $\mu\text{g m}^{-3}$ (58%) of PM_{2.5} and 0.04 $\mu\text{g m}^{-3}$ (66%) of
 355 OC). Another noticeable difference is that energy production is also an important contributor to
 356 surface SO₂ (0.01 $\mu\text{g m}^{-3}$: %51) and SO₄ (0.02 $\mu\text{g m}^{-3}$: 43%) levels over Finland. Finnish
 357 emissions, in particular industrial combustion, contribute largest to the air pollution over Sweden.

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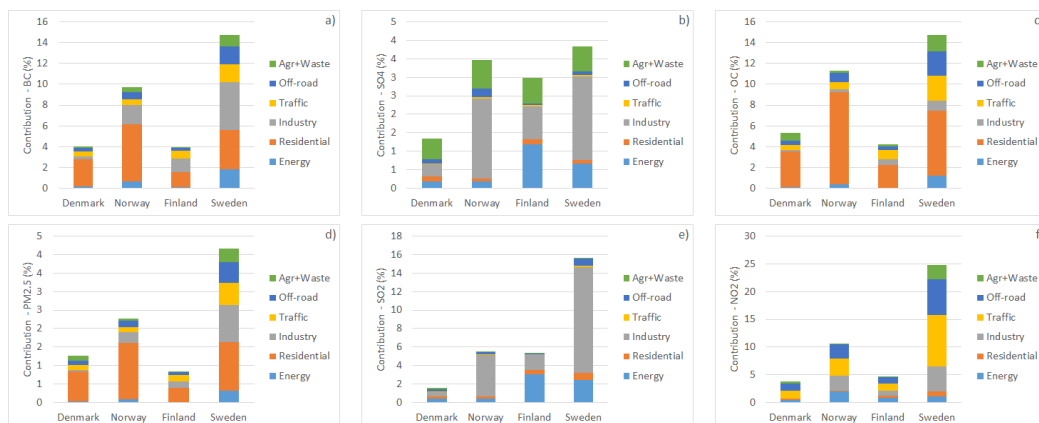
361 Fig. 7. Contributions of sectoral Finnish emissions on surface a) BC, b) SO₄, c) OC, d) PM_{2.5}, e)362 SO₂ and f) NO₂ over the Nordic countries.

363

364 Contributions from the Swedish emission sources to surface pollutant levels over the Nordic
 365 countries are presented in Fig. 8. Unlike other Nordic countries, Swedish emissions have larger
 366 contributions to pollution levels over the other Nordic countries, in particular over Norway. The
 367 figure also shows that Sweden does not experience as dominant contribution from non-industrial



368 combustion (28%) like the other Nordic countries show. Swedish emissions from SNAP2 are much
369 lower than for the rest of the Nordic countries (official emissions reported to the CLRTAP), most
370 probably due to lower emission factors. Residential combustion and industry contribute similarly to
371 the surface $PM_{2.5}$ levels. Industry also has a dominant contribution to surface SO_4 levels ($0.01 \mu g m^{-3}$:
372 51%), as well to SO_2 ($0.01 \mu g m^{-3}$: 74%) and BC ($0.004 \mu g m^{-3}$: 31%).
373



374

375

376 Fig 8. Contributions of sectoral Swedish emissions on surface a) BC, b) SO_4 , c) OC, d) $PM_{2.5}$, e)
377 SO_2 and f) NO_2 over the Nordic countries.

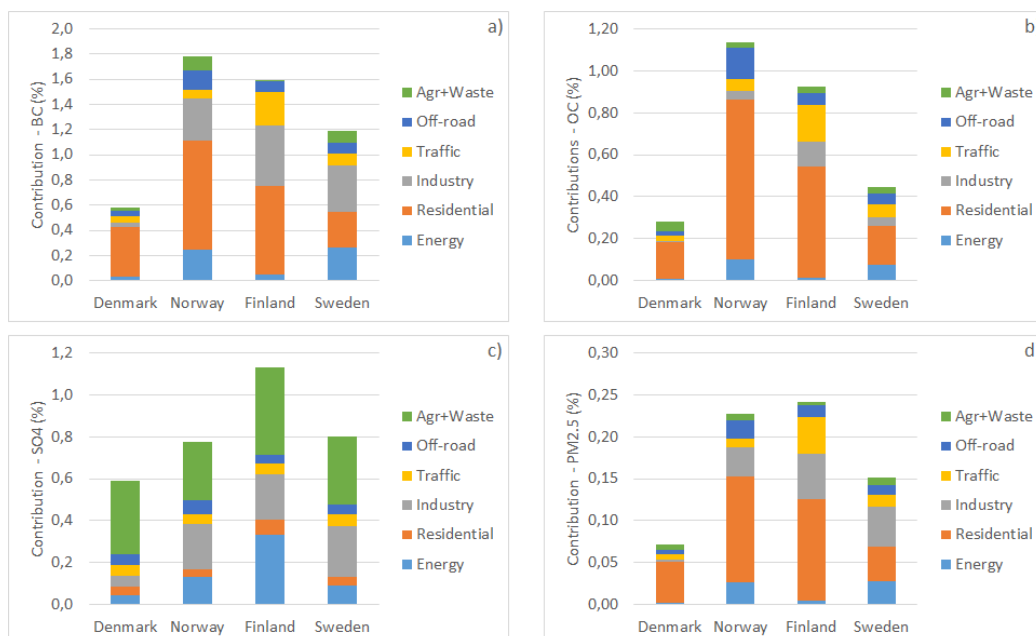
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3.2.2. Arctic

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381 The contributions of the emission sources in the different Nordic countries on the surface aerosol
382 concentrations over the Arctic region (defined as the area north of $67^\circ N$ latitude) are presented in
383 Fig. 9. Results show that overall, Norway has the largest contribution to surface aerosol levels over
384 the Arctic, while Denmark has the lowest contribution, although contributions are only a few
385 percent. Norwegian emissions, in particular non-industrial combustion, contributes to about 2% of
386 the surface BC levels over the Arctic. Non-industrial combustion in the Nordic countries is also the
387 largest contributor to Arctic BC levels, except for Sweden, where industry plays a more important
388 role. Non-industrial combustion is also the dominant contributor to OC levels over the Arctic.
389 Sulfate levels are largely influenced by the contributed from the agriculture and waste treatment
390 facilities over the Nordic countries. Contributions to Arctic $PM_{2.5}$ levels are similar to the
391 contributions to the BC levels.
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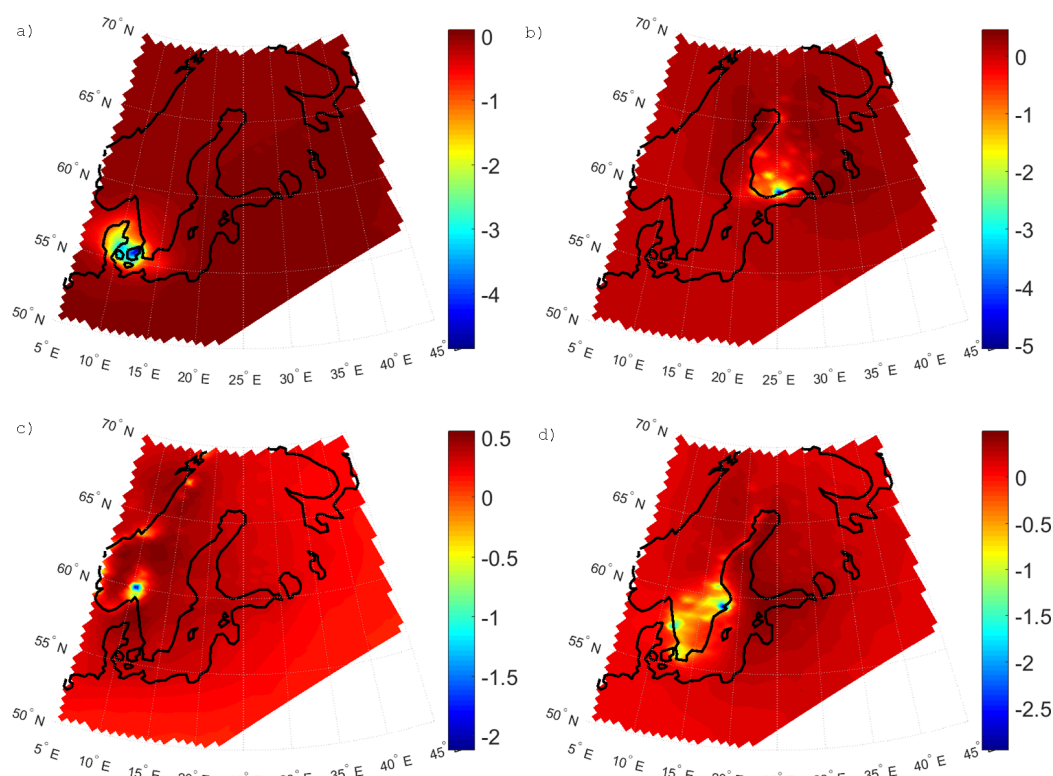


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Fig. 9. Contributions from a) Denmark, b) Norway, c) Finland and d) Sweden to the surface aerosol levels over the Arctic (north of 67°N).

3.2.3. Spatial distributions of contributions

The geographical distributions of total anthropogenic emissions from each Nordic country to surface PM_{2.5} and O₃ levels are calculated to investigate the extent of contributions from each Nordic country to its neighbours and to the Arctic. Fig. 10 shows the annual-mean relative contributions (%) of total land-based anthropogenic emissions to surface O₃ levels in the Nordic region from each country. The annual-mean contributions are very low, generally lower than 5% and are mainly calculated in the source country itself. Danish anthropogenic emissions (Fig. 10a) leads to a titration of around 4-5%, particularly over the Zealand region over the country where it leads to a very small O₃ increase (>1%) in the downwind towards south. The largest impact of Finnish emissions is around the Helsinki area, responsible for up to 5% of surface O₃ destruction over the area (Fig. 10b). Similar to Denmark, Finnish emissions also lead to an increase of surface O₃ levels by less than 1% over the downwind regions to the southeast and northwest. Impact of Norwegian emissions to surface O₃ levels (Fig. 10c) are largest (2%) over the Oslo area and the impact extents over the northern part of Oslo with a slightly larger contribution to O₃ levels compared to Denmark and Finland. The Swedish emissions have a larger geographical impact on the surface O₃ levels (Fig. 10d) over the country itself compared to the other Nordic countries but the magnitude is similar to the impact from the Norwegian emissions.



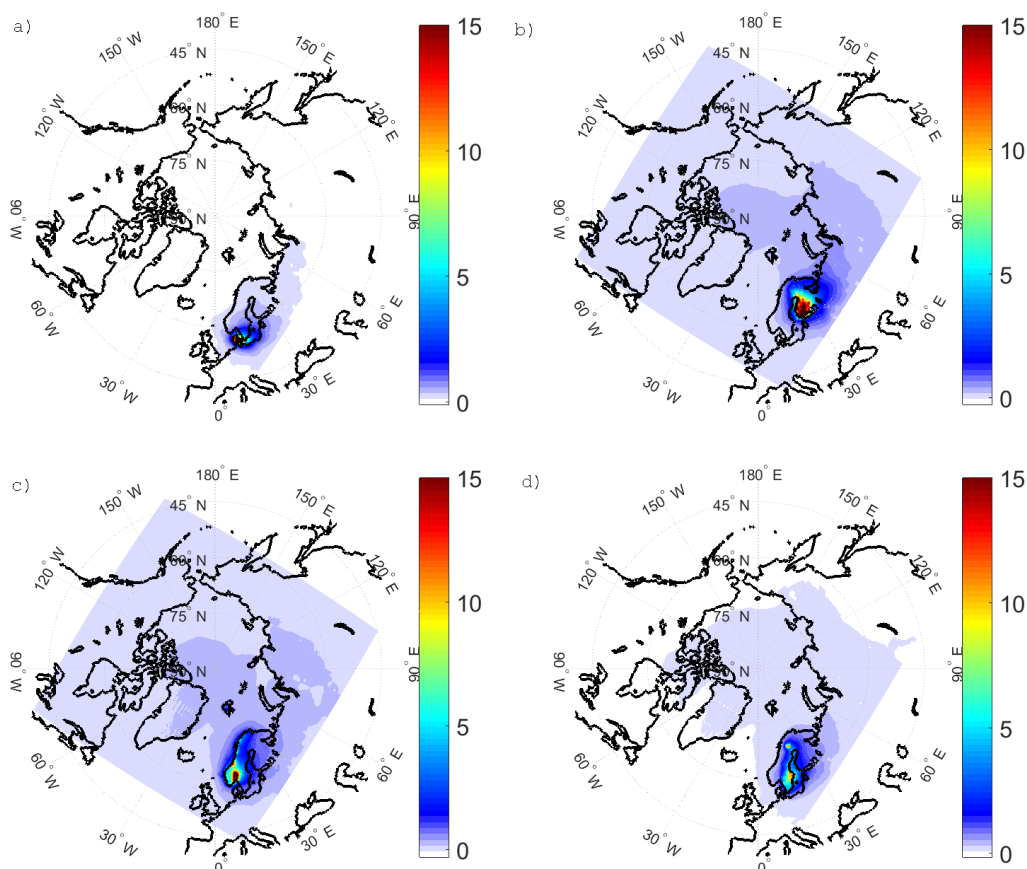
417
418 Fig. 10. Spatial distributions of annual-mean relative contributions (%) of total emissions from a)
419 Denmark, b) Finland, c) Norway, and d) Sweden to surface O₃ levels in the Nordic region.
420

421 Fig. 11 shows the annual-mean relative contributions of each Nordic country on the surface PM_{2.5}
422 levels in the entire model domain. Danish anthropogenic emissions are responsible for up to 20% of
423 surface PM_{2.5} levels over Denmark, with largest contributions over the Zealand region (Fig. 11a).
424 Danish land emissions also impact the surface PM_{2.5} levels over the southern part of Sweden and
425 Norway, by around 4% and 2%, respectively. The Finnish anthropogenic emissions have the largest
426 impact on surface PM_{2.5} levels over the southern part of the country, around the capital region by up
427 to 30% (Fig. 11b). Finnish emissions also have a small impact, lower than 3%, on the central part of
428 Sweden and northern parts of Norway. Norwegian anthropogenic emissions have largest
429 contributions to surface PM_{2.5} level around the capital region by up to 30%, while there is also a
430 significant impact on surface PM_{2.5} levels over Sweden by around 7% (Fig. 11c). Finally, Swedish
431 anthropogenic emissions have large contribution to surface PM_{2.5} levels over the Stockholm area by
432 around 15% and also contributes to PM_{2.5} levels over Finland, in particular over the southwestern
433 parts of Finland, by up to 5% (Fig. 11d).
434

435 Fig. 11 also shows the impact of anthropogenic emissions from each Nordic country to the surface
436 PM_{2.5} over the Arctic. Overall, the impacts are very small, around a few per cent, as seen in the
437 figure. The Danish emissions (Fig. 11a) have a more local contribution compared to other Nordic
438 countries and the impact does not reach above roughly 70 °N. The outflow from Finland, Norway



439 and Sweden can reach to the central Arctic ocean over to the northern parts of Greenland, however
440 contributions are around 1-2% (Figs. 11b-d).
441
442



443 Fig. 11. Spatial distributions of annual-mean relative contributions (%) of total emissions from a)
444 Denmark, b) Finland, c) Norway, and d) Sweden to surface $PM_{2.5}$ levels over the Nordic and the
445 Arctic regions (north of $67^\circ N$).
446

447 448 3.3. Contribution to premature mortality and costs 449

450 The number of acute and chronic premature mortality in the four selected Nordic countries and the
451 Arctic region (north of $67^\circ N$), along with the associated costs are presented in Table 4. As seen in
452 the Table, chronic mortality due to $PM_{2.5}$ is the major source for premature mortality, as EVA
453 calculates chronic mortality only due to exposure to $PM_{2.5}$ (see Table S2). The highest number of
454 cases is calculated for Sweden (~ 4 200 cases), followed by Denmark (~ 3 500 cases), Finland
455 (~ 1 800) and Norway (~ 1 700). These numbers lead to an associated cost of more than 2 billion
456 Euros in Sweden and Denmark and ~ 1 billion Euros in Finland and Norway. The number of
457 premature death cases are comparable with existing literature (e.g. Brandt et al., 2013a for



458 Denmark; Solazzo et al., 2018 for all four Nordic countries; EEA, 2017 for all four Nordic
 459 countries). In the Arctic region, the total number of premature mortality cases is calculated to be 94,
 460 93 of which are due to exposure to PM_{2.5} (chronic), leading to a cost of 58 million Euros.

461

462 Table 4. Acute and chronic premature death cases in the Nordic countries and the Arctic region
 463 (north of 67°N) in 2015 and the associated costs.

464

	Denmark	Finland	Norway	Sweden	Arctic
Premature Mortality (number of cases)					
Acute	19	18	6	25	1
Chronic	3 332	1 707	1 596	4 091	93
Total	3 351	1 725	1 602	4 116	94
Cost (million Euros)					
Acute	30	28	9	38	1
Chronic	2 031	1 040	973	2 494	57
Total	2 061	1 068	982	2 532	58

465

466 The EVA model has been used to calculate the contributions of Nordic emissions to the total
 467 premature mortality (acute + chronic) in the Nordic countries for the year 2015. Table 5 presents a
 468 source/receptor matrix of the contributions to premature mortality on the Nordic countries. Danish
 469 emissions contribute to ~400 premature deaths in Denmark, dominated by agriculture (33%), non-
 470 industrial combustion (31%) and traffic (18%). In Norway, the dominating sector contributing is
 471 non-industrial combustion, responsible for 48% of the ~200 premature deaths in Norway. In
 472 Finland, the total number of premature deaths in 2015 is calculated to be ~270, where non-industrial
 473 combustion and traffic are responsible for more than half. Finally, in Sweden, traffic and waste
 474 management/agriculture are responsible for 50% of the total premature death in Sweden (~330).

475

476 Table 5. Source/Receptor relationships of the contributions of anthropogenic emissions from the
 477 Nordic countries to the premature mortality in the Nordic area.

478

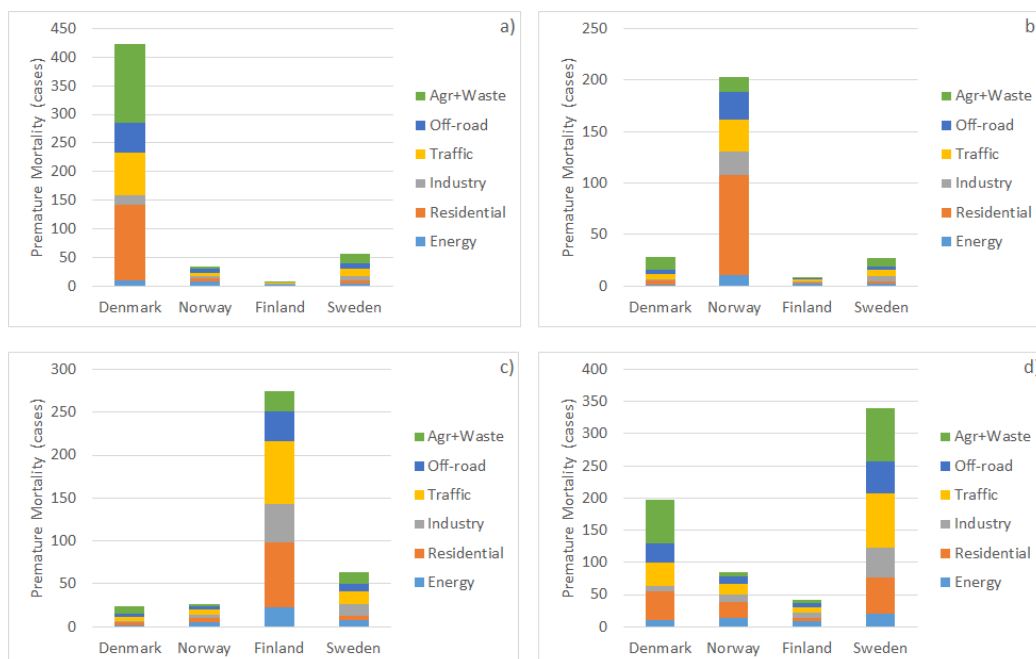
Source/Receptor	Denmark	Finland	Norway	Sweden
Denmark	422	24	29	198
Finland	8	274	9	42
Norway	33	26	203	86
Sweden	57	64	27	340

479

480

481 Fig. 12 shows the contributions of sectoral emissions from each Nordic country to the total
 482 premature death cases in 2015 in the different Nordic countries. Overall, Nordic countries
 483 contribute to low premature death cases in their Nordic neighbours (≤ 50). The largest
 484 transboundary contribution is calculated for the Danish emissions, dominated by agriculture, non-
 485 industrial combustion and traffic, contributing to ~200 premature death cases in Sweden.

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Fig. 12. Source contributions from the anthropogenic emissions of a) Denmark, b) Norway, c) Finland, and d) Sweden to total premature mortality (acute+chronic) in the Nordic countries.

Table 6 shows the cost of air pollution on human health in each of the Nordic countries in the source country and the neighbouring Nordic countries. Among the four Nordic countries, Denmark has the largest external costs due to air pollution, followed by Sweden, Finland and Norway, respectively. Following the mortality rates, Denmark, Finland and Norway have the largest cost contribution to Sweden, while Sweden contributes largest to Denmark.

Table 6. Contribution of costs (million €) of air pollution impacts on human health in the Nordic countries.

Source	Receptors			
	Denmark	Finland	Norway	Sweden
Denmark	260	14	17	121
Finland	5	172	5	26
Norway	20	16	125	52
Sweden	35	39	16	211

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Regarding the costs attributed to each of the source sectors, Fig.S1 summarizes the contributions per country. For Denmark, results suggest that non-industrial combustion and agriculture/waste management are the main sectors to be targeted to reduce the negative impacts of air pollution. In Norway, reduction of non-industrial combustion emissions alone can substantially reduce the costs of air pollution. In Finland, similar to Denmark and Norway, non-industrial combustion should be



507 targeted for developing emission reduction strategies, along with the traffic emissions, which
508 contribute as large as the residential combustion. Finally, in Sweden, traffic and agriculture/waste
509 management sectors should be targeted to reduce the adverse impacts of air pollution and their
510 associated costs.

511

512 4. Conclusions

513

514 The sectoral contributions of land-based anthropogenic emission sources in the four Nordic
515 countries; Denmark, Finland, Norway and Sweden, on air pollution levels and premature mortality
516 in these countries and over the Arctic have been estimated using the DEHM/EVA impact
517 assessment system for the year 2015. The chemistry and transport model, DEHM, was run with
518 tagging mode in order to calculate inline the sectoral contributions based on 30% reductions of each
519 sector separately. Using the modelled surface concentrations of O₃, SO₂ and PM_{2.5}, the EVA model
520 calculated the acute (O₃ and SO₂) and chronic (PM_{2.5}) premature mortality due to exposure to these
521 pollutants.

522

523 Results show that the Nordic countries are responsible for 5-10% of the regional background
524 surface PM_{2.5} concentrations in the countries itself. The non-industrial combustion (SNAP2), which
525 is dominated by the residential wood combustion, is responsible for 50% to 80% of the contribution
526 to surface PM_{2.5} in the Nordic countries. In Denmark, Finland and Norway, non-industrial
527 combustion contributes largely to surface OC (by 60% - 80%). In Sweden, SNAP2 is responsible
528 for 43% of the contribution to surface OC, while 43% comes from industrial activities. Similar to
529 OC, BC is also dominated by non-industrial combustion (by 50%-65%), except for Sweden, where
530 25% originates from non-industrial combustion and 31% from industrial activities. The dominant
531 source for surface SO₄ and SO₂ in all four Nordic countries is calculated to be industrial activities.
532 In Norway and Sweden, around 70% of SO₂ are coming from industrial activities, while in
533 Denmark and Finland, industrial activities are responsible for around 30% of SO₂. Off-road traffic
534 is responsible for 21% of SO₂, while energy production is responsible for 50% of SO₂ in Finland.
535 Industrial activities are also responsible for 60% of SO₄ in Norway and Sweden and 30% in
536 Denmark and Finland. The dominant source for NO₂ is calculated to be mobile sources, and the share
537 between on-road and off-road traffic varies depending on the country. Almost 35% of NO₂ comes
538 from on-road traffic in all four Nordic countries while off-road traffic contributes by 25% to 35%.

539

540 Norway has the largest contribution to aerosol levels over the Arctic, while Denmark has the lowest
541 contribution, although contributions are only a few percent. Non-industrial combustion in the
542 Nordic countries is also the largest contributor to Arctic OC and BC levels, except for Sweden,
543 where industry plays a more important role in relation to the Arctic levels. Agriculture and waste
544 treatment facilities over the Nordic countries are responsible contribute to the sulfate levels over the
545 Arctic.

546

547 Anthropogenic emissions lead to a titration of around 4-5%, particularly over the source countries
548 and lead to a very small surface O₃ increase (>1%) in the downwind regions. The largest impacts
549 are calculated to be around the capital regions. Danish emissions also impact the surface PM_{2.5}
550 levels over the southern part of Sweden and Norway, by around 3%. Finnish emissions also have a
551 small impact, lower than 3%, on the central part of Sweden and northern parts of Norway.
552 Norwegian anthropogenic emissions impacts PM_{2.5} levels over Sweden by around 7% while
553 Swedish anthropogenic emissions contribute to PM_{2.5} levels over the southwestern parts of Finland,
554 by up to 5%. It should be noted that these results are calculated for a specific year, 2015, therefore



555 transport from one country to others can significantly vary in different years due to meteorology, in
556 particular wind speed and direction.

557

558 The total number of premature mortality cases due to air pollution are calculated to be ~4 000 in
559 Denmark and Sweden and ~2 000 in Finland and Norway, leading to a total cost of 7 billion Euros
560 in the selected Nordic countries. The contributions of emission sectors to premature mortality in
561 each of the Nordic countries vary. Danish agriculture and industrial emissions contribute similarly
562 (by 33%) to ~400 premature mortality cases in Denmark. In Norway, non-industrial combustion,
563 dominated by residential wood combustion, is responsible for 48% of the ~200 premature deaths in
564 Norway. In Finland, non-industrial combustion and traffic are responsible for more than half of the
565 ~270 premature deaths in 2015. Finally, in Sweden, traffic and waste management/agriculture are
566 responsible for 50% of the total premature death in Sweden (~330). In Denmark, Finland and
567 Norway, non-industrial combustion is the main sectors to be targeted to reduce the negative impacts
568 of air pollution, while in Sweden, traffic and agriculture/waste management sectors should be
569 targeted to reduce the adverse impacts of air pollution and their associated costs. Overall, Nordic
570 countries contribute to low premature death cases in their Nordic neighbours (≤ 50). Among the four
571 Nordic countries, Denmark has the largest external costs due to air pollution, followed by Sweden,
572 Finland and Norway, respectively. Following the mortality rates, Denmark, Finland and Norway
573 have the largest cost contribution to Sweden, while Sweden contributes largest to Denmark.

574

575 Overall, results from the estimates of pollution export, premature mortality and associated costs
576 suggest that in the Nordic countries, non-industrial combustion, which is dominated by residential
577 wood combustion, together with industry and traffic are the main sectors to be targeted for emission
578 mitigation strategies. The contributions of emissions from Nordic countries to each other are small
579 ($\leq 10\%$), and to the Arctic (up to 2%), meaning that large reductions can be achieved only by
580 coordinated efforts to decrease emissions in the upwind countries.

581

582 **Author Contribution**

583

584 UI and JHC conducted the model simulations. JHC and OKN worked with the emissions input. MS
585 and RM contributed to the experimental design of the model simulations. UI, JK, CA and SL-A
586 extracted measurement data from Denmark, Finland, Sweden and Norway, respectively. CG and JB
587 contributed to premature mortality and cost calculations. All co-authors contributed to the
588 manuscript.

589

590 **Acknowledgements**

591

592 This study has been conducted under the FREYA project, funded by the Nordic Council of
593 Ministers, Climate and Air Pollution Group (grant agreement no. MST-227-00036). AU gratefully
594 acknowledges the NordicWelfare project funded by the NordForsk's Nordic Programme on Health
595 and Welfare (grant agreement no. 75007). The work has also been funded by the Academy of
596 Finland within the project GLORIOA and by the Research Council of Norway under the project
597 BlackArc (contract no 240921).

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