Response to comments of Reviewer 1

We thank the reviewer for the comments and suggestions on the manuscript. Our response (in blue) and the corresponding edits (in red) are shown below.

General comments

This work presents the evaluation of the halogen (Cl, Br, I) chemistry scheme from Sarwar et al [2015] within the CMAQ model on a regional scale (12km) for the month of July over Europe. The authors present an evaluation of halogen chemistry's impacts on air-quality metrics by comparing two runs, with chlorine ("CHL") or all halogens (HAL), against a model run without halogen chemistry ("BASE"). The focus is on how halogen chemistry impacts ozone in the European summer.

1. The authors argue the novel aspect of this study derives from the high resolution (12km) and focus on air quality impacts over Europe. However, this novelty is some-what challenged by the extensive referencing of a study in another model that considered also considered halogen chemistry within a nested model over Europe at ~ 25km during the summer too [Sherwen et al 2017]. Furthermore, this work seems to omit reference to the follow-on work using the same nested model ~25km for other seasons [Sommariva et al 2018] or using the model in a global for an entire year at 12km [Hu et al 2018]. It is not clear what the authors are suggesting is the main differences in between the nested global vs. nested hemispheric approaches, apart from the resolution.

Response: We have revised the introduction to describe the novelty of the present study, including (1) using the latest version of the CMAQ model (Sarwar et al., 2019) to investigate the halogen impact on air quality in Europe, (2) the implication for regulation, and (3) the highest spatial resolution used to investigate halogen impact on air quality over Europe.

The revised text in the introduction:

Only a few regional modeling studies have explored the combined influence of the halogen chemistry on air quality. The first modeling study with combined halogen (Cl, Br, and I) chemistry was conducted by Sarwar et al. (2015) who used a hemispheric version of the Community Multiscale Air Quality (CMAQ) model (Ching and Byun, 1999; Byun and

Schere, 2006; Mathur et al., 2017) and reported a decrease of surface O_3 by ~15% to ~48% over the Northern Hemisphere by Br and I. Gantt et al. (2017) then utilized the CMAQ model to explore the role of halogen chemistry at a regional scale over the continental United States (US). While these studies focused on Northern Hemisphere and the continental US, Muñiz-Unamunzaga et al. (2018) applied the full-halogen chemistry version of CMAQ with a resolution of 4 km and reported up to 5 ppbv decrease of O_3 in the city of Los Angeles, California, US. Sherwen et al. (2017) used a global model, GEOS-Chem, in a regional configuration (with a grid size of 0.25° × 0.315°, ~25km x ~25km) and predicted a large decrease of O_3, on average -13.5 pptv (25%) and as much as -28.9 pptv (45%) in Europe. Sarwar et al. (2019) further updated the halogen chemistry in CMAQ model and reported a reduction of -3 to -12 ppbv of annually average O_3 over seawater and -3 to -6 ppbv over coastal and -3 ppbv over inland area by Br and I. These previous regional studies using various models (or versions of models) in different areas reported a large range of the halogen impact on O_3 highlighting the uncertainty in this research field.

The regulation of air quality and the control of air pollutants emission in Europe started in the early 1970s and over forty years of effort has successfully improved air quality throughout Europe (EEA, 2018a). Nonetheless, poor air quality persist in major cities like Madrid, Paris, and London (EEA, 2018a); this shows the need for continiued air quality management and effective policy. Because the influence of halogens on air quality is uncertain and potentially has an impact on air quality management decisions, we have conducted regional simulations using the latest version of the CMAQ model implemented with comprehensive halogen sources and chemistry (Sarwar et al., 2019) to examine the overall effect of halogen species on air pollution over Europe. Considering that the grid size has a noticeable impact on air quality model predictions (Sommariva et al., 2018), we used a CMAQ model domain with 12 km horizontal resolution (higher than the previous studies on halogen impact covering Europe) to simulate the levels of halogen species over Europe, examine the effect on the oxidation capacity and the concentration of air pollutants, and explore the potential implications for air quality policy related to NO_2 and O_3.

2. Furthermore when considering the period of study: Halogen chemistry can be very seasonally dependent in Europe (e.g. ClNO2 - See Sommariva et al [2018]). So considering

the existing literature on modelling halogens, further model runs should at least be presented for a winter month to give an equivalent novel value on understanding the air-quality impacts.

Response: The present study focuses on the halogen impact on air quality during summer season. We have revised the title.

Impact of halogen chemistry on summertime air quality in coastal and continental Europe: application of CMAQ model and implication for regulation

3. Some opportunities to make this work more novel seem to have been missed. For instance, coarse two bin comparisons ("coastal" vs. "inland") are made with regional ozone and NO2 observations instead of comparisons that show if the model captures chemical/physical processing (e.g. diel plots by the hour). Oxidants are analysed as "snapshots" of peak or average concentrations, rather than considered in terms of a given hour. Considering the diel cycle of oxidants by hour over Europe (e.g. contribution of Cl in the morning vs. OH at midday, then NO3) and how halogens effect this would really elevate this manuscript.

Response: We have added diurnal plots of O_3, NO_2, OH, HO_2, NO_3, and Cl and the corresponding text in the revised manuscript.

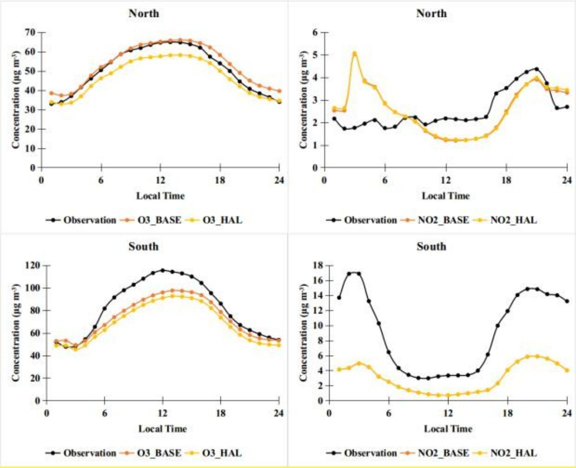

Figure S2. Diurnal variation of observation and simulation (BASE and HAL) of O_3 and NO_2 in northern and southern Europe.

The BASE simulation under-predicts O_3 compared to observations both at coastal and continental stations (Table 1), possibly due to the uncertainty of VOC emission inventory (Sherwen et al., 2017) and the underestimated NO_x (Table 1). The HAL simulation slightly improves the correlation coefficient of O_3 but decreases the average level of O_3 compared to the BASE case. Diurnal variation plots (Fig. S2) suggests that both BASE and HAL simulations produces the temporal patterns of O_3 and NO_2.

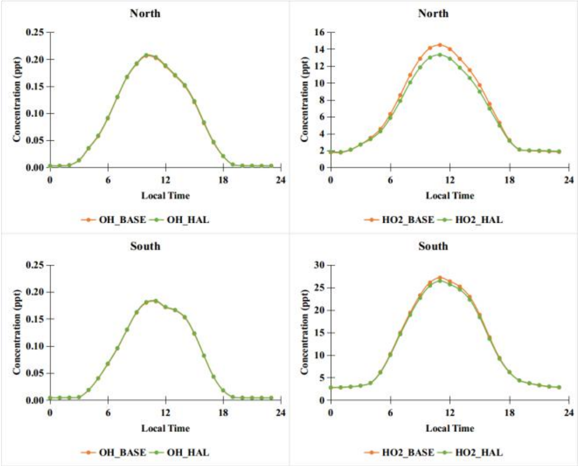

Figure S3. Diurnal variation of simulated (BASE and HAL) OH and HO_2 over northern and southern Europe.

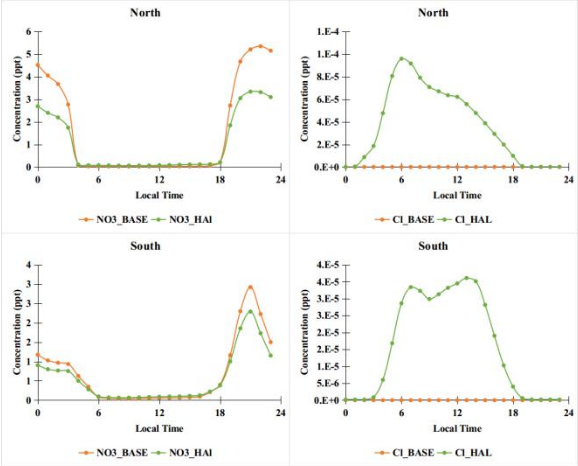

Figure S4. Diurnal variation of simulated (BASE and HAL) NO_3 and Cl over northern and southern Europe.

We also examine the diurnal variations of the four radicals in the BASE and HAL scenarios (Fig. S3 and S4). Halogens have small effect on the diurnal pattern of OH. HO_2 is reduced by halogens especially in the mid-day. NO_3 radical is strongly decreased throughout the night after the addition of halogens. Cl atom is released by the halogen chemistry evidently in the early morning. The significant effects of halogen chemistry on the diurnal variation of OH, HO_2, NO_3, and Cl radicals highlight the role of halogen chemistry in regulating the atmospheric oxidation capacity throughout the day with the highest effect on Cl in the early morning, maximum effects on OH and HO_2 in daytime, and largest effect on NO_3 at night.

4. The description of the model setup needs to be clearer. From reading the manuscript, I think I correctly understand that the configuration uses a regional grid for Europe and then sets "boundary conditions" outside this grid from two additional hemispheric runs. Please add more information to the model configuration section and clarify this. Writing is often verbose and would often benefit from re-wording to make the manuscript more concise. The introduction especially would benefit for re-writing for flow and clarity of the various different model studies cited. References (e.g. for observations) could be more up to date.

Response: In the revised version of the manuscript, we used the latest version of CMAQ model incorporated with up-to-date halogen chemistry (Sarwar et al., 2019) to re-run all simulation cases.

The boundary condition setting is also modified.

Boundary conditions for the model were derived from the hemispheric CMAQ simulations (Mathur et al., 2017). Three different annual simulations were conducted using the hemispheric CMAQ model for 2016: the first simulation used the Carbon Bond chemical mechanism without any halogen chemistry, the second simulation used the Carbon Bond chemical mechanism and the chlorine chemistry, and the third simulation used the Carbon Bond chemical mechanism and the full halogen chemistry. Results from the corresponding hemispheric CMAQ simulation were used to generate boundary conditions for the BASE, CL, and HAL simulations. Therefore, the difference between CL and BASE simulations represents the impact of the chlorine chemistry on air quality and the difference between HAL and BASE simulations represents the effect of halogen chemistry on air quality.

The spin-up time has also been changed.

The study was completed for the month of July 2016 with a spin-up period of 30 days.

Therefore, all figures, tables, numbers in the text have been modified using the new simulation results. Please note that the conclusions are not changed using the new model, new boundary condition setting, and new spin-up time.

We have also added more reference, e.g. Sommariva et al., (2018) and Bannan et al., (2017), in the halogen simulation (section 3.2).

Table 3. The comparison of observed and simulated halogen species

Location	Species	Observation *	Simulation #
Hessen, Germany [a]	$ClNO_2$	800.0	273.4
London, United Kingdom [b]	$ClNO_2$	724.0	801.5
Weybourne, United Kingdom [c]	$ClNO_2$	65	373
Weybourne, United Kingdom [d]	$ClNO_2$	946	373
Weybourne, United Kingdom [e]	$ClNO_2$	1100 (summer) 75.6 (autumn) 733 (winter)	373
Leicester, United Kingdom [e]	$ClNO_2$	274 (spring) 74.2 (summer) 248 (winter)	274
Penlee Point , United Kingdom [e]	$ClNO_2$	922	319
Mace Head, Ireland [f]	BrO	6.5	10.1
Brittany, France [g]	BrO	7.5	0.4
Dead Sea [h]	BrO	100.0	0.2
Mace Head, Ireland [i]	IO	4.0~50.0	3.9
Brittany, France [j]	IO	7.7~30.0	1.1
Dagebull, Germany [k]	IO	2.0	9.0
Atlantic Ocean [l] (Prados-Roman et al., 2015)	IO	0.4 to 0.5 (daytime average)	0.4 to 2.0 (daytime average)

*: Maximum value (pptv).
#: Maximum value (pptv) from the HAL simulation.
a: Phillips et al. 2012.
b: Bannan et al., 2015.
c: Banna et al., 2017.
d: Sherwen et al., 2017.
e: Sommariva et al., 2018.
f: Saiz-Lopez et al., 2004.
g: Mahajan et al. 2009.
h: Matveev et al., 2001; Holla et al., 2015.
i: Allan et al., 2000; Commane et al., 2011.
j: Britter et al., 2005; Furneaux et al., 2010.

k: Peters et al., 2005.
l: Prados-Roman et al., 2015

Several field campaigns have been conducted in Weybourne in the past few years to measure $ClNO_2$. Sherwen et al. (2017) reported a peak concentration of 946 pptv. Bannan et al. (2019) reported a peak value of 65 pptv, and Sommariva et al. (2018) reported a peak value of 1100 pptv in summer, 75.6 pptv in autumn and 733 pptv in winter. CMAQ simulated a maximum of 373 pptv at that location while GEOS-Chem predicted 458 pptv. Sommariva et al. (2018) also reported measurements of $ClNO_2$ at Leicester with a maximum value of 274 pptv in spring, 74.2 pptv in summer, 248 pptv in winter, and that at Penlee Point a peak value of 922 pptv. CMAQ predicted a maximum of 274 pptv at Leicester and 319 pptv at Penlee Point. Eger et al. (2019) conducted shipborne observation of $ClNO_2$ in the Mediterranean Sea and reported up to 600 pptv $ClNO_2$ during their campaign, which is similar to the prediction of the present study.

Specific comments

5. Page 1 - Lines 15-35 (Abstract) Please add headline numbers for changes seen in this manuscript to abstract.

What is the net effect of the domain on OH/HO2? The effect of halogens on OH /HO2 in box models and global models has been discussed at length by Stone et al. [2018].

Response: We have added the predicted values in the abstract.

Combined halogen chemistry induces complex effects on OH (between -0.023 pptv and 0.030 pptv) and HO_2 (in the range of -3.7 to 0.73 pptv), significantly reduces the concentrations of NO_3 (as much as 20 pptv) and O_3 (as much as 10 ppbv), and decreases NO_2 in the highly polluted regions (as much as 1.7 ppbv) and increases NO_2 (up to 0.20 ppbv) in other areas.

The net effect in the domain of halogens on OH and HO2 is added in the revised manuscript.

The net change of OH due to halogen was near zero.

The overall difference of HO_2 because of halogens was -0.59 pptv in the European domain.

The reference of Stone et al. (2018) is added in the discussion of the halogen impact on HO_x.

Another GEOS-Chem study, however, predicted an increase of OH over the Mediterranean Sea (Stone et al., 2018).

6. Page 2 - Line 35 - 41 Earlier references should be used here to give fair credit to the original work on this, instead of recent reviews.

It would be cleaner and more instructive to the reader to include the following sentence, rather than just citing a recent review.

“The chemistry of halogens in the troposphere has been described in detail in recent reviews (Saiz-Lopez and von Glasow, 2012; Simpson et al., 2015), so we just briefly outline it here.”

Additionally, the effect of halogen nitrate hydrolysis on decreasing ozone production should be mentioned here as it has been shown to have a larger impact on ozone than increased loss [Schmidt et al 2016].

Response: We have added the suggested sentence in the introduction.

The chemistry of halogens in the troposphere has been described in detail in recent reviews and references therein (Saiz-Lopez and von Glasow, 2012; Simpson et al., 2015), so it is just briefly outlined here.

We have also added the effect of the hydrolysis of halogen nitrate on the O_3 level in the introduction.

indirectly decreasing O_3 production by reducing NO_2 (R2 and R3),

$$XO + NO_2 \rightarrow XONO_2 \qquad (R2)$$

$$XONO_2 + H_2O\ (l) \rightarrow HOBr + HNO_3(l) \qquad (R3)$$

7. Page 2 - Line 46 “methane ch4” should read “methane (CH4)”

Response: Revised.

8. Page 2 - Line 52-55 “Evaluation of the complex role of halogen chemistry in air quality requires the employment of advanced, high-resolution chemical transport models”

Has this argument really been evidenced here? Could not the processed discussed be captured by existing coarse resolution approaches? Are the uncertainties on halogen modelling substantially small enough that horizontal resolution of models is the main limitation for the fidelity of simulation to observations and capturing chemical and physical processes?

Response: Please refer to response 1.

9. Page 2 Line 56-61 Should not the faraday discussions paper you mentioned earlier be included here as you say in the manuscript it covered ClNO2 too? Other model studies have looked at this too and should be included here also. Adding “e.g. X et al., Y et al.” to illustrate the reader that a couple of examples have been given would also be appropriate. There are other points in the manuscript where this would be appropriate too.

Response: Added more reference to the previous studies on simulating $ClNO_2$, including Sherwen et al. (2017).

The chemistry of chlorine, mainly that of $ClNO_2$, has been reported to increase the oxidation capacity and the formation of O_3 in recent studies (Sarwar et al., 2012, 2014; Li et al., 2016; Sherwen et al., 2017; Sommariva et al., 2018).

10. Page 3 - 61-71 It is hard for the reader to follow the way in which the previous work is being discussed. The authors have referred to another nested regional study in GEOSChem in the final intro paragraph [Sherwen et al 2017], but then did not include it in the discussion of existing regional modelling work here.

Response: Added Sherwen et al. (2017) in the description of regional model simulation.

Sherwen et al. (2017) used a global model, GEOS-Chem, in a regional configuration (with a grid size of 0.25° × 0.315°, ~25km × ~25km) and predicted a large decrease of O_3, on average 13.5 pptv (25%) and as much as 28.9 pptv (45%) in Europe.

11. Page 3 - Line 71 It is arguable that 4km is a higher resolution. However, is a 4km horizontal resolution accepted to be sufficient to represent the processes going on in a city? I would suggest updating as follows:

From “halogen sources on air quality at a city scale (4 km resolution) in Los Angeles, California, US.” To “halogen sources on air quality at a resolution of 4 km in the city of Los Angeles (California, US). “

Response: Revised.

12. Page 3 - Line 74 A URL in brackets is not an appropriate format here. If a URL must be used please include a reference to the EEA and a data accessed for the data. Referencing an EEA report would be preferable.

Response: Revised.

13. Page 3 - Line 72 to 82 It is hard for the reader to follow the semantic. The previous study is also a domain based study over Europe using boundary conditions from a model with a larger (global) domain with offline meteorology. Are the authors arguing that the main difference between the 12x12km CMAQ approach presented here and the existing work, at the coarser resolution (25x25km), is just resolution? Both models use similar halogen chemistry are nested within larger domains, correct? Why is so much change in simulation skill expected to be seen between 12x12 km and 25x25km? Or would a higher resolution, say 4x4 km [MunÌCˇ iz-Unamunzaga et al., 2018], be required to notable gains in capturing processes or differences? This leads to a more philosophical question: is it a model resolution or processes holding back science currently? Would not other uncertainties in halogen chemistry be greater than the difference caused by a change in resolution too? (e.g. emissions developed for coarser resolutions or new developments in the representation of halogen chemistry in models - Xuan et al [2019])

Response: Please refer to response 1.

14. Page 3 - Line 84 Is “instrumented” the appropriate word here? “Including” would be a better choice or “which includes”.

Response: “instrumented” has been changed to “implemented”.

15. Page 3 - Line 85 Why is “state-of-the-art” used here? It seems unnecessarily verbose, especially when referring to a paper that is at least four years old in a fast-moving part of the literature.

Response: In the revised version of manuscript, we use the updated halogen chemistry in CMAQ (Sarwar et al., 2019). We have revised the paragraph as follows:

The regulation of air quality and the control of air pollutants emission in Europe started in the early 1970s and over forty years of effort has successfully improved air quality throughout Europe (EEA, 2018a). Nonetheless, poor air quality persist in major cities like Madrid, Paris, and London (EEA, 2018a); this shows the need for continiued air quality management and effective policy. Because the influence of halogens on air quality is uncertain and potentially has an impact on air quality management decisions, we have conducted regional simulations using the latest version of the CMAQ model implemented with comprehensive halogen sources and chemistry (Sarwar et al., 2019) to examine the overall effect of halogen species on air pollution over Europe. Considering that the grid size has a noticeable impact on air quality model predictions (Sommariva et al., 2018), we used a CMAQ model domain with 12 km horizontal resolution (higher than the previous studies on halogen impact covering Europe) to simulate the levels of halogen species over Europe, examine the effect on the oxidation capacity and the concentration of air pollutants, and explore the potential implications for air quality policy related to NO_2 and O_3.

16. Page 4 - Line 90-96 The meteorology is offline? Or is CMAQ being run in coupled mode? Please explicitly state if the meteorology is offline.

Response: The CMAQ model was run offline.

The meteorological inputs for the CMAQ model were obtained from the Weather Research and Forecasting model (WRF 3.7.1) (Skamarock and Klemp, 2008; Borge et al., 2008a) as an offline input.

17. Page 4 - Line 96 Why is the word “besides” used in this sentence. Please rephrase for clarity.

Response: Removed.

18. Page 4 - Line 111 How has coastal been defined here? More broadly, is 12km sufficient resolution to capture “coastal” effects?

Response: Previous studies (Sarwar et al., 2015; Sherwen et al., 2016; Sarwar et al., 2019) suggest that halogen chemistry affects O_3 not only over marine environments but also over inland locations far away from marine environments. Thus, we use coastal area to include greater land area adjacent to marine environments and employ all monitoring stations within 24-km from the coastline to examine the impact of the halogen chemistry impact on coastal areas as well as over inland areas.

19. Page 6 - Line 131 -132 Is this 12x12km domain nested within another domain not computed online? Or is the model run hemispherically at 12x12km here for all simulations? Please make this clearer.

Response: In the present study, the CMAQ model is run with one domain with a spatial resolution of 12x12 km. The hemispherical simulation is only used as the boundary condition.

The CMAQ model is applied over a domain covering the entirety of Europe (Fig. 1) with 12 km horizontal resolution.

20. Page 6 - line 143 The name "CHL" may make many readers from the Earth sciences community think of chlorophyll. I would suggest using "Cl" instead to make the paper more accessible to a broader readership.

Response: Revised.

21. Page 6 - Line 139 -145 Only 7 days spin up was used for a hemispheric simulation? Or is this for the European grid? Is this sufficient to ensure initial conditions are "washed" through the grid? Is there a reference showing this is sufficient? What spin up was used for "the hemispheric CMAQ simulations"/hemispheric grid? What initial conditions were used? This needs to be clearer. Global studies on halogens in CMAQ, GEOS-chem and CAM-Chem have highlighted the importance of changes in background concentrations. The relative contribution of boundary and local effects seems to be a core focus of the Sherwen et al [2017] manuscript, but not really given much attention or discussion here.

Response: The hemispheric simulation of CMAQ, which is used for the boundary condition, is described in Mathur et al. (2017).

In the revised version, we have used a spin-up period of 30 days.

In the present study, we focus on the overall impact of halogens on air quality in Europe, instead of the relative contribution of boundary and local effects.

22. Page 6 - Line 152-160 The model is being run on a hemispheric grid of 12x12, but just analysed for the domain in Figure 1? The "Boundary conditions" are on the same grid, but global? and these are provided to the outside of the modelling domain? This needs to be explained more clearly.

Response: Please refer to response 19 and 21.

23. Page 6 - Line 152 "Boundary conditions for the model were derived from the hemispheric CMAQ simulations."

Which hemispheric CMAQ simulations? Where have these been described?

Response: Please refer to response 21.

24. Page 7 - Line 161-163 Please include a table that shows which boundary conditions and chemistry used in each domain for each model experiment.

“the difference between CHL and BASE simulations represents the impact of the chlorine chemistry on air quality”

Don’t both of these simulations have the same boundary conditions (“Carbon Bond chemical mechanism and the chlorine chemistry”)? Surely then the most this difference can show is the contribution of chlorine chemistry locally. A preferable approach would be to use a “BASE” set of boundary conditions without halogen chemistry for the “BASE” simulation.

“the difference between HAL and BASE simulations represents the effect of halogen chemistry on air quality.”

Again: according to the text, “BASE” includes the effects of chlorine globally as it includes chlorine in boundary conditions (“Carbon Bond chemical mechanism and the chlorine chemistry”). Therefore “BASE” - “HAL” is giving the effects of halogens minus the global effect of chlorine.

Although locally chlorine can provide an oxidant effect and lead to ozone formation, it also can act as a sink for ozone through the loss of chlorine nitrates on a global scale. This has been discussed in some of the global modelling papers cited here. What are the global effects of chlorine in this model? Can global effects be excluded here?

Response: Please refer to response 4.

25. Page 7 - line 169 Please give a justification for the use for “within 24 km from the coast” definition. Coastal processes (notably halogens) can be confined to a very small area (e.g. macroalgae in tidal zones).

Response: Please refer to response 18.

26. Page 7 - Line 169-171 The level of the evaluation presented needs to be increased to be in line with the high-resolution output the authors argue is notable here.

Please plot ozone and NO2 as a diel (24-hour) comparison compared with the model, preferably showing a few regions in Europe. Regions like the coastal Mediterranean should not just be lumped together with the coastal Scandinavian, as they have very different seasonal characteristics. Simply using a table to do a very coarse two bin comparison does not provide much insight and it is not really in line with the current level analysis presented in the literature (e.g. Schnell et al. 2015), instead, it smooths out the extra information gained.

Response: Please refer to response 3.

27. Page 7 - Line 179-181 Please provide references for the proposed explanations for model bias or expand this discussion.

Response: We have added one reference, Jung et al., 2017, for the explanation of the NO_2 simulation.

Jung, J., Lee, J., Kim, B. and Oh, S.: Seasonal variations in the NO2 artifact from chemiluminescence measurements with a molybdenum converter at a suburban site in Korea (downwind of the Asian continental outflow) during 2015–2016. Atmos. Environ., 165, 290-300, 2017.

28. Page 7 - Line 184-186 Only a single table of comparisons has been provided to back up this statement. More evidence is needed. Please provide diel plots of core species (e.g. CO, NO2, O3).

Response: Please refer to response 3.

29. Page 7 - Line 182 How does this comparison look on a diel basis? Is there an offset or difference in the diel cycle at certain times of the day? Does the model capture the diel cycle before or after or is there a structural issue in the model (e.g. caused by emissions or boundary layer mixing?)?

Response: Please refer to response 3.

30. Page 11 - Line 223-255 The coastal influence or inland nature of observations should be made clearer. At the resolution presented here (12x12km) it would not be fair to expect the model to reproduce many of these observations (e.g those influenced by local emissions from the tidal zone).

Response: Please refer to response 18.

31. Page 11 - Table 1 This table of observations does not seem in line with current literature (e.g. observations for ClNO2 are available across seasons in Northern Europe [Sommeriva et al 2018]). Please expand these comparisons.

Response: Please refer to response 4.

32. Page 11 - Line 249 The largest IO dataset has been published since Saiz-Lopez and von Glasow [2012] by Prados-Roman [2015]. Did this not this start and end in the Mediterranean? How does the model compare against this? What other more recent datasets are there?

Response: We have added the comparison between the CMAQ modeling results with the observations reported in Prados-Roman et al. (2015).

Prados-Roman et al. (2015) reported the level of IO during a ship-based campaign in the range of <0.4 to >1.4 pptv (daytime average) around the globe and 0.4 to 0.5 pptv (daytime average) in the south of Spain and the west of Africa (over the Atlantic), and the present study predicted 0.4 to 2.0 pptv (daytime average) of IO in that area.

33. Page 12 - Line 261-264 Please see earlier comments about the inclusion of chlorine in boundary conditions for both “BASE” and “CHL” simulations.

Response: Please refer to response 4.

34. Page 13 - Line 267 Please give OH units in the more commonly used units of molecules cm-3 or at least show this in brackets throughout the text. Please do the same for Cl (atoms cm-3).

Response: Revised.

35. Page 13 - Line 290 What about the resolution difference could cause this? What other differences could explain this? Are there any differences in the chemistry between the Sarwar et al (2015) and Sherwen et al (2017) Mechanism?

Response: We have revised the discussion.

Another GEOS-Chem study, however, predicted an increase of OH over the Mediterranean Sea (Stone et al., 2018). The discrepancy among the previous studies and between those works and the present one is difficult to deduce and requires further investigation. Several possible causes could lead to different simulated levels of halogens and their impact on oxidants, including the different mechanism of producing and recycling the halogen species (Sarwar et al., 2019), spatial resolution (Sommariva et al., 2018), emission inventory (Wang et al., 2019), and different spatio-temporal scale of interest (Stone et al., 2018).

36. Page 15 - Line 305 “Our study, along with the previous work, highlights the vital role of halogen chemistry in the nighttime chemistry.” Which previous work? How is this chemistry constrained by lab work/observations?

Is it based on theoretical calculations? If So, then the uncertainty on this should be highlighted here.

Response: The sentence has been removed.

37. page 15 Line 307 - 309 Cly is in the boundary conditions for both “BASE” and “CHL”, correct? How much Cly is transported into the domain?

Response: Please refer to the response 4.

38. Page 15 - 312-318 How do these values compare against known constraints on tropospheric chlorine (e.g. Gromov et al [2018])? Are there any explanations for the differences? (e.g. It is worth noting too that Hossani et al [2016] used unrealistic anthropogenic chloride emissions - see Xuan et al [2019] for details on this)

Response: The suggested reference, Gromov et al., 2018, is a study on the role of chlorine on methane in the southern hemisphere, which is not relevant to the present study. Their reported range of the concentration of Cl, 9 × 10^3 to 2.8 × 10^4 atom cm^{-3}, is within the range in the review we cited (Saiz-Lopez and von Glasow, 2012), 10^3 to 10^5 atom cm^{-3}.

We have added the following sentence to acknowledge the work by Wang et al. (2019).

In their study, Hossaini et al. (2016) used the Reactive Emission Inventory of Chlorine (Keene et al., 1999) which Wang et al. (2019) reported to be unrealistic for present day applications.

39. page 15 Line 320 "The current study and the previous works simulated a broad range of the surface Cl concentrations although they were all within the scope of the reported observed (observation- based calculation) values of 103 to 105 atom cm-3 (4.0 x10-5 to 4.0 x10-3 pptv) according to the review of Saiz-Lopez and von Glasow (2012)."

Why is a review that is > seven years old being used as the basis for comparison? There have been a large amount of Cly measurements since then (e.g. Gromov et al [2018], Haskins et al [2018] etc ...) and more work to constrain tropospheric Cl [Gromov et al 2018].

Response: For the suggested reference, Gromov et al. (2018), please refer to the response 38.

The other reference, Haskins et al. (2018), is a study on the measurements of inorganic chlorine species in US. They focused on the partitioning of the gaseous and particulate chlorine and provided constraints on the total atmospheric inorganic chlorine, not the chlorine atom.

40. page 22 Fig 8 AOT40 is calculated over a growing season. Why is this shown for a single month in Fig. 8 and discussed in the text? It would be more appropriate to give output in units of exceedances for a given month as this usage of “AOT40” could mislead readers. What over relevant ozone thresholds are there? What about the particulate matter if the oxidants have changed (e.g. PM10, PM2.5)?

Response: We have revised the manuscript. We used the threshold of 40 ppbv as in AOT40.

Here we use the simulation results (BASE and HAL) in July to calculate the accumulated O3 and the difference between two scenarios. We note that the accumulated O_3 is noticeably reduced (>15% along the coast) after the addition of halogens.

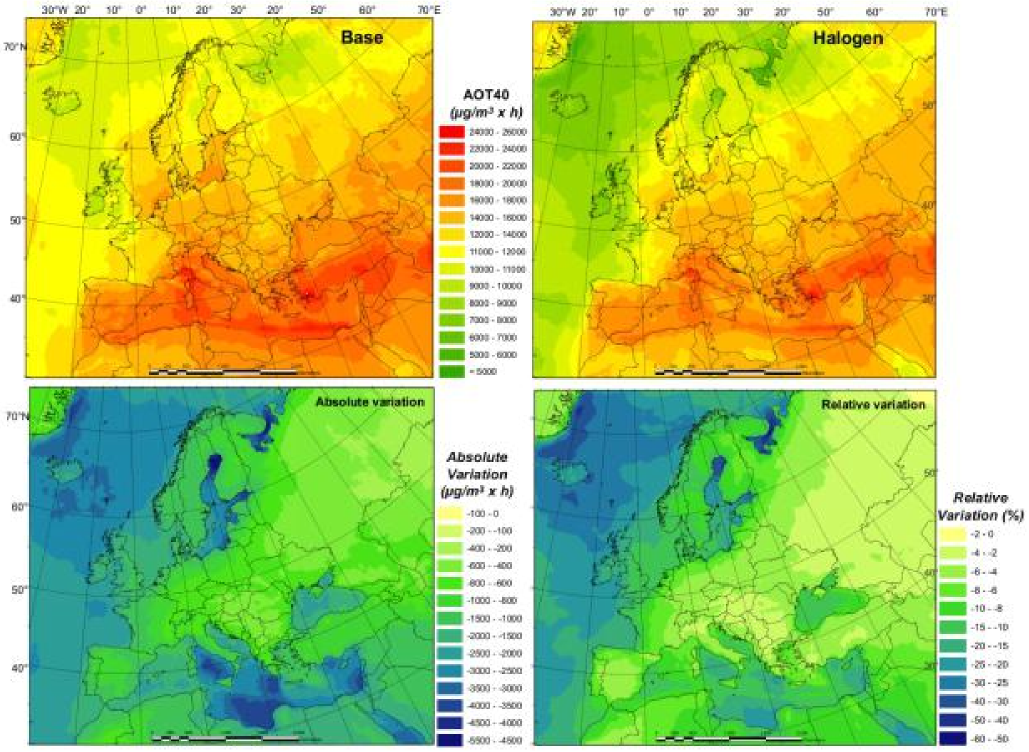

Figure 8. Accumulated O_3 in July in the BASE and HAL simulations, and absolute and relative changes between the two simulations.”

We investigated the halogen impact on oxidants, O_3, and NO_2 in the present study. The impact on aerosols is not the focus.

41. page 23 Lines 430-432 “These models do not include the comprehensive halogen chemistry, potentially leading to [an] unrealistic simulation of O3 concentration responsiveness to the predicted NOx and/or VOCs emission changes in Europe.”

As only a coarse comparison is provided for ozone (table 1) and no diel cycles are shown it is hard to see if halogens are aiding the capture of processes seen within the observations. This weakens the argument that halogen processes are needed to gain a “realistic” simulation in air quality models. Please back up this claim with figures.

Response: Please refer to response 3. We have also modified this sentence:

The models which do not include the comprehensive halogen chemistry can potentially lead to different O_3 concentration responses to NO_x and/or VOC emission changes in Europe.

42. page 24 Lines 460-465 The final paragraph of the conclusion comes across as vague. Please provide a few examples of uncertainties that are more specific than just “chemical mechanism”.

Response: We have added a few examples of uncertainties in the final paragraph.

Although the incorporation of the halogen chemistry may improve the capabilities of 3D Eulerian chemical transport models, we acknowledge that large uncertainties still exist in the assessment of halogen chemistry impact due to emission inventories (e.g., chlorine emission inventory; Wang et al., 2019), model configuration (e.g., grid size; Sommariva et al., 2018), chemical mechanism (e.g., photolysis rate of iodine oxides, recycling rate of halogen species on aerosol; Simpson et al., 2015), etc.

Response to comments of Review 2

We thank the reviewer for the comments and suggestions on the manuscript. Our response (in blue) and the corresponding edits (in red) are shown below.

General comments

This paper presents a model study of the effects of halogen chemistry on the air quality of Europe. This study provides an interesting overview of the impact of halogens on ozone and other pollutants, a research question that is still open. The paper is well written and presented and I have only some minor comments (see below). Overall I think it is suitable for publication in ACP.

1. In Section 3.2 the CMAQ results are compared to the observations and to the GEOSChem results from Sherwen et al. (2017). First of all, there are other observations of ClNO2 in Europe besides those in Table 3. In fact some of these are mentioned in the Sherwen paper itself (as well as in Sommariva et al., 2018) and in Bannan et al. (2017). These measurements should be included in the discussion. Second, using the maximum observed concentration is not a good metric to assess the agreement with the model. For example, the observations in Phillips et al. (2012) show quite a range of peak nocturnal concentrations of ClNO2. I would also argue that GEOS-Chem shows better agreement with the measurements than CMAQ, especially wrt ClNO2 (lines 231-233). The discussion of the model-measurements comparison is better when dealing with iodine and bromine species, but please revise Section 3.2 to be more accurate.

Response: The observational results of $ClNO_2$ reported in Sherwen et al. (2017), Sommariva et al. (2018) and Bannan et al. (2017) have been included in Table 3 of the revised manuscript.

Table 3. The comparison of observed and simulated halogen species

Location	Species	Observation [*]	Simulation [#]
Hessen, Germany [a]	$ClNO_2$	800.0	273.4
London, United Kingdom [b]	$ClNO_2$	724.0	801.5
Weybourne, United Kingdom [c]	$ClNO_2$	65	373
Weybourne, United Kingdom [d]	$ClNO_2$	946	373
Weybourne, United Kingdom [e]	$ClNO_2$	1100 (summer) 75.6 (autumn) 733 (winter)	373
Leicester, United Kingdom [e]	$ClNO_2$	274 (spring) 74.2 (summer) 248 (winter)	274
Penlee Point , United Kingdom [e]	$ClNO_2$	922	319

Mace Head, Ireland [f]	BrO	6.5	10.1
Brittany, France [g]	BrO	7.5	0.4
Dead Sea [h]	BrO	100.0	0.2
Mace Head, Ireland [i]	IO	4.0~50.0	3.9
Brittany, France [j]	IO	7.7~30.0	1.1
Dagebull, Germany [k]	IO	2.0	9.0
Atlantic Ocean [l] (Prados-Roman et al., 2015)	IO	0.4 to 0.5 (daytime average)	0.4 to 2.0 (daytime average)

*: Maximum value (pptv).
#: Maximum value (pptv) from the HAL simulation.
a: Phillips et al. 2012.
b: Bannan et al., 2015.
c: Banna et al., 2017.
d: Sherwen et al., 2017.
e: Sommariva et al., 2018.
f: Saiz-Lopez et al., 2004.
g: Mahajan et al. 2009.
h: Matveev et al., 2001; Holla et al., 2015.
i: Allan et al., 2000; Commane et al., 2011.

Several field campaigns have been conducted in Weybourne in the past few years to measure $ClNO_2$. Sherwen et al. (2017) reported a peak concentration of 946 pptv. Bannan et al. (2019) reported a peak value of 65 pptv, and Sommariva et al. (2018) reported a peak value of 1100 pptv in summer, 75.6 pptv in autumn and 733 pptv in winter. CMAQ simulated a maximum of 373 pptv at that location while GEOS-Chem predicted 458 pptv. Sommariva et al. (2018) also reported measurements of $ClNO_2$ at Leicester with a maximum value of 274 pptv in spring, 74.2 pptv in summer, 248 pptv in winter, and that at Penlee Point a peak value of 922 pptv. CMAQ predicted a maximum of 274 pptv at Leicester and 319 pptv at Penlee Point. Eger et al. (2019) conducted shipborne observation of $ClNO_2$ in the Mediterranean Sea and reported up to 600 pptv $ClNO_2$ during their campaign, which is similar to the prediction of the present study.

We agree with the reviewer that the peak concentration of $ClNO_2$ (normally around the time of sunrise) at one location could have a large range, implying the large day-to-day variation of the level of $ClNO_2$ precursors (NO_x, O_3, Cl^-). The capability of CMAQ model to reproduce the maximum level of $ClNO_2$ at several locations throughout Europe (United Kingdom, Germany, Mediterranean Sea etc.) represents the ability of CMAQ to satisfactorily simulate emission, transport, and chemical transformation processes related to $ClNO_2$. Considering that the present study was not designed to reproduce the level of $ClNO_2$ in a certain campaign, we think that the current validation metric adequate enough to show that the CMAQ model and the current setting can be used to investigate the halogen impact on the air quality in Europe.

We have removed the sentence that compares the performance of CMAQ with that of GEOS-Chem with regard to $ClNO_2$.

2. Figure 2 is interesting in the sense that it shows some different results from the corresponding figure 5 in the Sherwen paper especially when it comes to BrO. It looks like CMAQ is calculating lower concentrations than GEOS-Chem both for Cl and for HCl, which deserves some comment. It would also be good to include some of the European observations of HCl in this discussion. I realize that a comparison between CMAQ and GEOS-Chem is beyond the scope of this paper, but the differences in the geographical distributions of some species (and related impacts on O3 and other species) are sometimes striking and require at least a brief comment.

Response: CMAQ predicted lower level of BrO compared to GEOS-Chem. We have added a sentence to acknowledge that.

The predicted BrO levels over Europe are low (average value ~0.17 pptv) with the largest predicted value occurring within the Arctic circle while GEOS-Chem predicted >1.0 pptv level of BrO in Mediterranean Sea (Sherwen et al., 2017).

For the modeled level of Cl, CMAQ had a similar distribution and magnitude compared with the GEOS-Chem model, although CMAQ predicted a slightly higher maxima value (7.0×10^{-4} pptv, or ~1.75×10^{4} atom cm^{-3}) than the GEOS-Chem model (1.4×10^{4} atom cm^{-3}).

We have also added some discussion on the prediction of HCl level.

The observed level of HCl in Europe is in the range of <100 pptv to 5000 pptv (Hossaini et al., 2016 and the reference therein). The CMAQ model predicted a monthly average concentration of HCl between 6.3 and 1249 pptv, which is similar to the observation range. GEOS-Chem (Sherwen et al., 2017) predicted a maximum of 12 pptv for HCl, which is significantly lower than the available measurements in Europe.

In the original version, we compared the present study with the GEOS-Chem work in the simulation of halogen species ($ClNO_2$, BrO, and IO) oxidants (e.g., OH), and pollutants (e.g. O_3) and we also noted the difference between the two models. In the revised version, we have added some more discussions on the difference of the level and distribution of some species of the two models, e.g., BrO and HCl.

3. line 688: correct typo in name.

Response: Corrected. (Dub has been modified to be Dubé)

Impact of halogen chemistry on summertime air quality in coastal and continental Europe: application of CMAQ model and implication for regulation

Qinyi Li[1], Rafael Borge[2], Golam Sarwar[3], David de la Paz[2], Brett Gantt[4], Jessica Domingo[2], Carlos A. Cuevas[1], and Alfonso Saiz-Lopez[1]*

1 Department of Atmospheric Chemistry and Climate, Institute of Physical Chemistry Rocasolano, CSIC, Madrid 28006, Spain
2 Environmental Modelling Laboratory, Department of Chemical & Environmental Engineering, Universidad Politécnica de Madrid (UPM), Madrid, Spain
3 National Exposure Research Laboratory, Environmental Protection Agency, Research Triangle Park, NC 27711, United States
4 Office of Air Quality Planning and Standards, Environmental Protection Agency, Research Triangle Park, NC 27711, United States

*Correspondence to: Alfonso Saiz-Lopez (a.saiz@csic.es)

Abstract:

Halogen (Cl, Br, and I) chemistry has been reported to influence the formation of secondary air pollutants. Previous studies mostly focused on the impact of chlorine species on air quality over large spatial scales. Very little attention has been paid to the effect of the combined halogen chemistry on air quality over Europe and its implications for control policy. In the present study, we utilize a widely-used regional model, the Community Multiscale Air Quality Modeling System (CMAQ), incorporated with the latest halogen sources and chemistry, to simulate the abundance of halogen species over Europe and to examine the role of halogens in the formation of secondary air pollution. The results suggest that the CMAQ model is able to reproduce the level of O_3, NO_2, and halogen species over Europe. Chlorine chemistry slightly increases the levels of OH, HO_2, NO_3, O_3, and NO_2 and substantially enhances the level of Cl radical. Combined halogen chemistry induces complex effects on OH (ranging from -0.023 pptv to 0.030 pptv) and HO_2 (in the range of -3.7 to 0.73 pptv), significantly reduces the concentrations of NO_3 (as much as 20 pptv) and O_3 (as much as 10 ppbv), and decreases NO_2 in the highly polluted regions (as much as 1.7 ppbv) and increases NO_2 (up to 0.20 ppbv) in other areas. The maximum effects of halogen chemistry occur over oceanic and coastal regions, but some noticeable impacts also occur over continental Europe. Halogen chemistry affects the number of days exceeding the European Union target threshold for the protection of human being and vegetation from ambient O_3. In light of the significant impact of halogen chemistry on air quality, we recommend that halogen chemistry be considered for inclusion in air quality policy assessments, particularly in coastal cities.

1. Introduction

Halogen (Cl, Br, and I) species and related processes have been known to deplete stratospheric ozone (O_3) for several decades (Molina and Rowland, 1974; Farman et al., 1985). In the troposphere, the chemistry of halogens in the troposphere has been described in detail in recent reviews and references therein (Saiz-Lopez and von Glasow, 2012; Simpson et al., 2015), so it is just briefly outlined here. Halogen species affect the concentration of air pollutants, e.g., directly destroying O_3 (R1), indirectly decreasing O_3 production by reducing NO_2 (R2 and R3), influencing the NO/NO_2 ratio (R2 and R4) and the HO_2/OH ratio (R5 and R6). The budgets of NO_x ($NO+NO_2$) and HO_x ($OH+HO_2$) also affect the formation of O_3 (e.g., Sillman, 1999; Li et al., 2018).

$$X\ (Cl, Br, I) + O_3 \rightarrow XO + O_2 \tag{R1}$$

$$XO + NO_2 \rightarrow XONO_2 \tag{R2}$$

$$XONO_2 + H_2O\ (l) \rightarrow HOX + HNO_3(l) \tag{R3}$$

$$XO + NO \rightarrow X + NO_2 \tag{R4}$$

$$XO + HO_2 \rightarrow HOX + O_2 \tag{R5}$$

$$HOX + h\nu \rightarrow OH + X \tag{R6}$$

Chlorine radical (Cl) initiates the oxidation of hydrocarbons (R7), including methane (CH_4) and non-methane volatile organic compounds (NMVOC), in a similar way to OH radical, reducing the lifetime of CH_4 and NMVOC and leading to the formation of O_3 in the presence of NO_x (Thornton et al., 2010).

$$RH + Cl \rightarrow HCl + RO_2 \tag{R7}$$

The combined effect of halogen chemistry on air quality, therefore, is complicated and depends heavily on local conditions, e.g., atmospheric compositions, oxidative capacity, etc. (Sherwen et al., 2016; Muñiz-Unamunzaga et al., 2018). Evaluation of the complex role of halogen chemistry in air quality requires the employment of advanced, high-resolution chemical transport models.

A number of modeling studies have been conducted to investigate the impact of individual halogen species on air quality. The chemistry of chlorine, mainly that of nitryl chloride ($ClNO_2$), has been reported to increase the oxidation capacity and the formation of O_3 in recent studies

(Sarwar et al., 2012, 2014; Li et al., 2016; Sherwen et al., 2017; Sommariva et al., 2018). Bromine (Fernandez et al., 2014) and iodine (Saiz-Lopez et al., 2014) chemistry are reported to decrease the concentration of O_3 over the oceanic and terrestrial regions.

Only a few regional modeling studies have explored the combined influence of the halogen chemistry on air quality. The first modeling study with combined halogen (Cl, Br, and I) chemistry was conducted by Sarwar et al. (2015) who used a hemispheric version of the Community Multiscale Air Quality (CMAQ) model (Ching and Byun, 1999; Byun and Schere, 2006; Mathur et al., 2017) and reported a decrease of surface O_3 by ~15% to ~48% over the Northern Hemisphere by Br and I. Gantt et al. (2017) then utilized the CMAQ model to explore the role of halogen chemistry at a regional scale over the continental United States (US). While these studies focused on Northern Hemisphere and the continental US, Muñiz-Unamunzaga et al. (2018) applied the full-halogen chemistry version of CMAQ with a resolution of 4 km and reported up to 5 ppbv decrease of O_3 in the city of Los Angeles, California, US. Sherwen et al. (2017) used a global model, GEOS-Chem, in a regional configuration (with a grid size of 0.25° × 0.315°, ~25km x ~25km) and predicted a large decrease of O_3, on average 13.5 pptv (25%) and as much as 28.9 pptv (45%) in Europe. Sarwar et al. (2019) further updated the halogen chemistry in CMAQ model and reported a reduction of 3 to 12 ppbv of annually average O_3 over seawater and 3 to 6 ppbv over coastal and 3 ppbv over inland area by Br and I. These previous regional studies using various models (or versions of models) in different areas reported a large range of the halogen impact on O_3 highlighting the uncertainty in this research field.

The regulation of air quality and the control of air pollutants emission in Europe started in the early 1970s and over forty years of effort has successfully improved air quality throughout Europe (EEA, 2018a). Nonetheless, poor air quality persist in major cities like Madrid, Paris, and London (EEA, 2018a); this shows the need for continiued air quality management and effective policy. Because the influence of halogens on air quality is uncertain and potentially has an impact on air quality management decisions, we have conducted regional simulations using the latest version of the CMAQ model implemented with comprehensive halogen sources and chemistry (Sarwar et al., 2019) to examine the overall effect of halogen species on air pollution over Europe. Considering that the grid size has a noticeable impact on air quality model predictions (Sommariva et al., 2018), we used a CMAQ model domain with 12 km horizontal resolution (higher than the previous studies on halogen impact covering Europe) to simulate the levels of halogen species over Europe, examine the effect on the oxidation capacity and the concentration of air pollutants, and explore the potential implications for air quality policy related to NO_2 and O_3.

2. Method and Materials

2.1 Data

The meteorological inputs for the CMAQ model were obtained from the Weather Research and Forecasting model (WRF 3.7.1) (Skamarock and Klemp, 2008; Borge et al., 2008a) as an offline input. The WRF model was initialized from global reanalyses from the National Centers for Environmental Prediction (NCEP) Global Forecast System (GFS) with a spatial resolution of 1° × 1° and a temporal resolution of 6 h (available online at http://rda.ucar.edu/datasets/ds083.2/) which was updated daily from NCEP global analyses with 0.5° resolution (available online at http://www.nco.ncep.noaa.gov/pmb/products/sst/). NCEP's ADP global upper-air (NCAR archive ds351.0) and global surface observations (NCAR archive ds461.0) were used to drive the simulation with a Newtonian relaxation technique in the WRF model.

Anthropogenic emissions for the year 2016 were taken from the 0.1° × 0.1° gridded EMEP inventory (EMEP/CEIP, 2014). It should be noted that no anthropogenic chlorine sources are included in our emission inventory. The temporal profiles and vertical distribution needed to resolve the emissions were those used in the EuroDelta experiment (van Loon et al., 2007). Biogenic emissions were estimated using the Model of Emissions of Gases and Aerosols from Nature (MEGANv2.10) (Guenther et al., 2012). All emissions were gridded to our model domain, temporally allocated and chemically speciated using the Sparse Matrix Operator Kernel Emissions (SMOKE) model, version 3.6.5 (UNC, 2015; Borge et al., 2008b).

In addition, we used measurement data of NO_2 and O_3 from over 400 background stations (traffic and industrial stations are not included) across Europe from database *AirBase* (public air quality database system of the European Environment Agency, 2018) to compare the results of our simulation with observations (Fig. 1). Among these stations, 315 are located in inland areas (208 for NO_2 and 266 for O_3), and 125 are located in the coastal areas (83 for NO_2 and 93 for O_3).

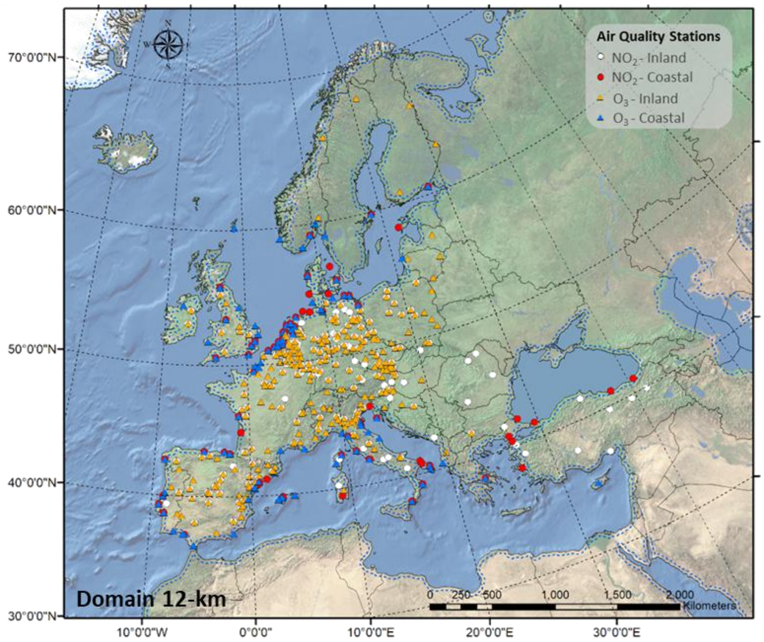

Figure 1. Geographic representation of the modeling domain and air quality stations used for model evaluation

2.2 Modeling system

The CMAQ model is widely used and includes comprehensive representations of many essential atmospheric processes. The skill of the model in reproducing observed air quality has been demonstrated in many previous studies (Foley et al., 2010; Appel et al., 2013, 2017; Mathur et al., 2017), including applications over Europe (Borge et al., 2008a; Appel et al., 2012; Solazzo et al., 2017). CMAQ version 5.2 (www.epa.gov/cmaq; doi:10.5281/zenodo.1167892) containing the Carbon Bond chemical mechanism (Appel et al., 2017) with halogen chemistry was used in this study. We updated the chlorine chemistry of Sarwar et al. (2012) and implemented it in CMAQ version 5.2 (Table S1). The heterogeneous hydrolysis of dinitrogen pentoxide (N_2O_5) can produce $ClNO_2$ and nitric acid (HNO_3) in the presence of particulate chloride. In the absence of particulate chloride, heterogeneous hydrolysis of N_2O_5 produces only HNO_3. Sarwar et al. (2015) implemented the initial bromine and iodine chemistry in CMAQ which has recently been updated by Sarwar et al. (2019). The updates include revising the gas-phase bromine and iodine reactions, incorporating several heterogeneous reactions of bromine and iodine species, incorporating several aqueous-phase bromine reactions, and revising the iodine and bromine emissions. We combine the updated chlorine, bromine, and iodine chemistry with the Carbon Bond chemical mechanism and use it in this study.

2.3 Simulation setup

A detailed description of physics and other model options (Table S2) can be found in (de la Paz et al., 2016). The CMAQ model is applied over a domain covering the entirety of Europe (Fig. 1) with 12 km horizontal resolution. The vertical extent of the model extended from the surface to 100 mbar and contained 35 layers with an average surface layer thickness of approximately 20 m. The CMAQ chemical transport model is configured to use the Piecewise Parabolic Method to describe advection processes, the Asymmetric Convective Model (version 2) to describe vertical diffusion processes, and the multiscale method to describe horizontal diffusion processes. Gas-phase chemistry, aqueous chemistry, aerosol processes, and dry and wet deposition were also included. The Rosenbrock solver was used for gas-phase chemistry.

The study was completed for the month of July 2016 with a spin-up period of 30 days. We performed three simulations to isolate the effect of halogen chemistry on air quality (in brackets the name of the scenario used hereafter):

(1) Base model without halogen chemistry (BASE),

(2) BASE with chlorine chemistry (CL), and

(3) CL with Br and I chemistry (HAL).

The BASE model simulation includes the Carbon Bond 2015 chemical mechanism but does not contain any halogen chemistry, and only the HNO_3 is produced from the heterogeneous hydrolysis of N_2O_5. The CL simulation contains the Carbon Bond chemical mechanism with chlorine chemistry and considers $ClNO_2$ and HNO_3 production from the heterogeneous uptake of N_2O_5 on the aerosol surface. The HAL simulation contains the Carbon Bond chemical mechanism with full halogen chemistry and produces $ClNO_2$ and HNO_3 from the heterogeneous uptake of N_2O_5 on the aerosol surface.

Boundary conditions for the model were derived from the hemispheric CMAQ simulations (Mathur et al., 2017). Three different annual simulations were conducted using the hemispheric CMAQ model for 2016: the first simulation used the Carbon Bond chemical mechanism without any halogen chemistry, the second simulation used the Carbon Bond chemical mechanism and the chlorine chemistry, and the third simulation used the Carbon Bond chemical mechanism and the full halogen chemistry. Results from the corresponding hemispheric CMAQ simulation were used to generate boundary conditions for the BASE, CL, and HAL simulations. Therefore, the difference between CL and BASE simulations represents the impact of the chlorine chemistry on air quality and the difference between HAL and BASE simulations represents the effect of halogen chemistry on air quality.

3. Results and Discussions

3.1 Evaluation of model performance

The performance of the CMAQ model in simulating air quality over Europe is evaluated using observation data collected from >400 measurement stations. We separate the stations into coastal (within 24 km from the coast) and continental stations (Fig. 1). Table 1 presents the statistics of the model performance for O_3 and NO_2 for BASE and HAL simulations.

The BASE and HAL simulations generally reproduce the concentration levels and the temporal variations of O_3 and NO_2 both at coastal and continental stations. The correlation coefficients between simulations and observations (Fig. S1) show that CMAQ satisfactorily reproduces the variation of O_3 and NO_2 over most of Europe especially the coastal regions (> 0.6 for O_3 and > 0.4 for NO_2). The BASE simulation under-predicts O_3 compared to observations both at coastal and continental stations (Table 1), possibly due to the uncertainty of NMVOC emission inventory (Sherwen et al., 2017) and the underestimated NO_x (Table 1). The HAL simulation slightly improves the correlation coefficient of O_3 but decreases the average level of O_3 compared to the BASE case. Diurnal variation plots (Fig. S2) suggest that both BASE and HAL simulations can reproduce the temporal patterns of O_3.

The BASE simulation under-predicts NO_2 compared to observations both at coastal and continental stations (Table 1). Such an under-estimation of NO_2 can occur for many reasons including (1) positive artifacts of NO_2 monitors (Jung et al., 2017), (2) under-estimation of NO_x in the emission inventory, and (3) rapid transformation of NO_2 into HNO_3 in the model compared to the real atmosphere. Model performance is reasonable as the NO_2 underestimation is relatively small. The HAL case predicted very similar NO_2 concentrations (Table 1).

Overall, the evaluation of CMAQ results over Europe demonstrates that the model is capable of reproducing the levels of atmospheric chemical species and can be used to investigate the impact of halogen chemistry on air quality over Europe. It also suggests that the incorporation of halogen chemistry changes the model performance for O_3 concentrations by a small margin without a noticeable impact on the model performance for NO_2.

Table 1. Statistical summary of model performance

Statistics	**O_3 (µg/m³)**				**NO_2 (µg/m³)**			
	Coastal		Inland		Coastal		Inland	
	Base	HAL	Base	HAL	Base	HAL	Base	HAL
MB	-0.9	-6.8	-2.3	-6.6	-3.7	-3.7	-3.1	-3.1
ME	16.9	17.5	19.7	20.2	5.8	5.8	5.7	5.6
RMSE	22.6	23.1	25.8	26.4	7.7	7.7	7.2	7.2
r	0.65	0.67	0.60	0.61	0.44	0.44	0.42	0.42
IOA	0.64	0.62	0.62	0.61	0.31	0.31	0.18	0.19

Note: MB = Mean Bias, RMSE = Root Mean Square Error, r = correlation coefficient, IOA = Index of Agreement.

3.2 Simulated halogen species

Average surface concentrations of the inorganic halogen species predicted in the HAL simulation over the ocean are summarized in Table 2. HCl is the dominant chlorine species with an average level of 247.9 pptv representing over 96% to the total inorganic chlorine (Cl_y) while the average $ClNO_2$ and HOCl is 4.9 pptv (1.9%) and 3.8 pptv (1.5%), respectively, and the remaining species contributing less than 1%. The Br_y species are relatively evenly partitioned, with HOBr (1.06 pptv, 27.0%), BrCl (0.71 pptv, 18.2%), $BrNO_3$ (0.67 pptv, 16.9), HBr (0.66

pptv, 16.8%), and Br_2 (0.33 pptv, 8.4%) being the abundant species while the remaining species contribute <5%. HOI (5.1 pptv, 52.0%), INO_3 (2.8 pptv, 28.7%), IO (1.0 pptv, 10.3%) contribute over 90% of I_y over the ocean, while the remaining species contribute ~10%. The predicted average concentrations of the critical halogen radicals, Cl, BrO, and IO, are $2.0x10^{-4}$ pptv, 0.03 pptv, and 0.4 pptv, respectively, over the ocean in the Europe.

Table 2. Simulated average concentrations of inorganic halogen species over the ocean

Species	Concentration (ppt)	Percentage (%)	**Species**	Concentration (ppt)	Percentage (%)	**Species**	Concentration (ppt)	Percentage (%)
HCl	247.9	96.1	**HOBr**	1.06	27.0	**HOI**	5.1	52.0
$\mathbf{ClNO_2}$	4.9	1.9	**BrCl**	0.71	18.2	$\mathbf{INO_3}$	2.8	28.7
HOCl	3.8	1.5	$\mathbf{BrNO_3}$	0.67	16.9	**IO**	1.0	10.3
$\mathbf{ClNO_3}$	1.2	0.5	**HBr**	0.66	16.8	**I**	0.4	4.0
ClO	0.25	0.1	**BrO**	0.38	9.7	$\mathbf{I_2O_3}$	0.3	3.1
$\mathbf{Cl_2}$	0.02	0.008	$\mathbf{Br_2}$	0.33	8.4	**HI**	0.1	1.0
Cl	0.0002	0.0001	$\mathbf{BrNO_2}$	0.09	2.4	$\mathbf{I_2}$	0.05	0.5
-			**Br**	0.03	0.7	**INO**	0.03	0.3
-			-			$\mathbf{I_2O_2}$	0.01	0.1
-			-			$\mathbf{INO_2}$	0.01	0.1
-			-			$\mathbf{I_2O_4}$	0.004	<0.1
Total Cl	258.1	100	**Total Br**	3.9	100	**Total I**	9.8	100

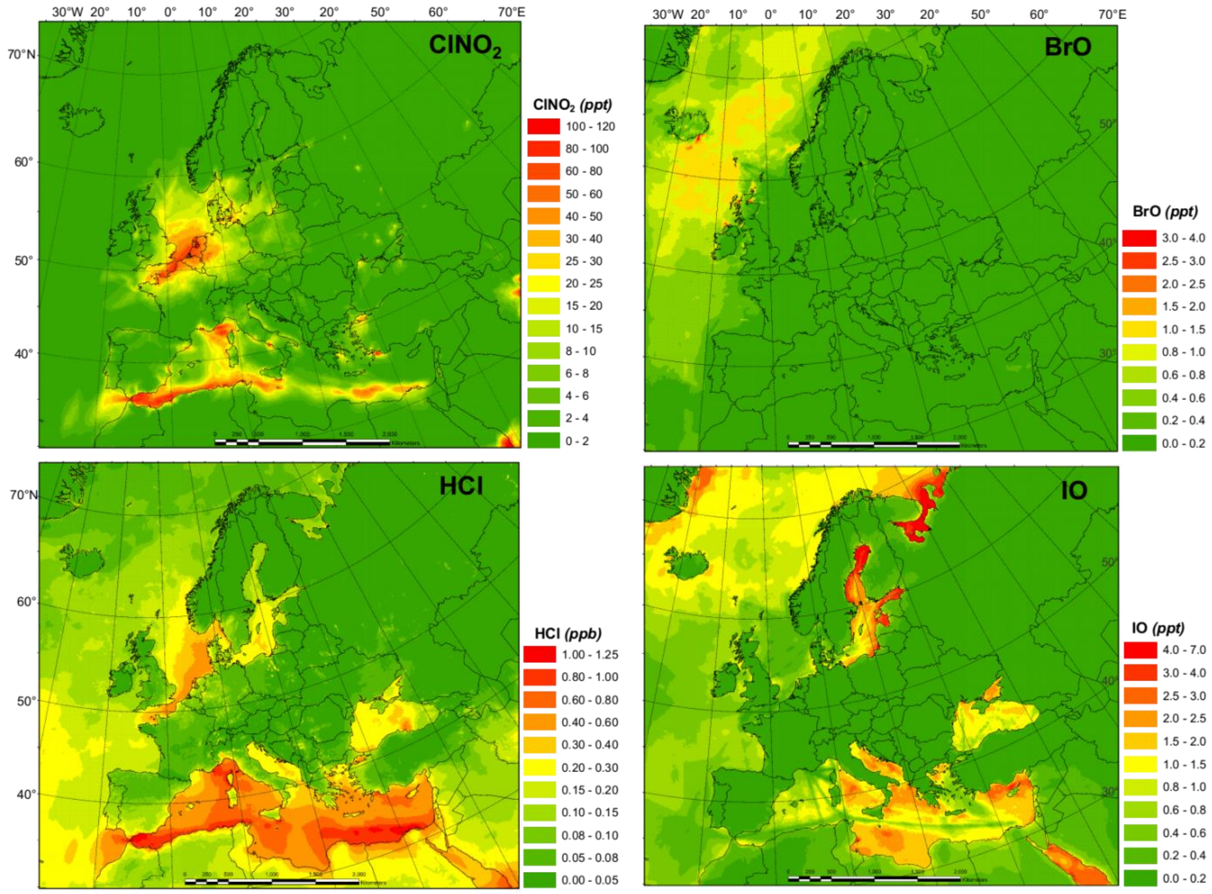

Figure 2. Monthly average ClNO$_2$, HCl, BrO, and IO concentration in the HAL simulation.

The spatial distributions of key halogen species are shown in Fig 2. The HAL simulation with full halogen chemistry simulates generally higher ClNO$_2$ levels (with the highest monthly average value of 113.0 pptv) along the coast of the Mediterranean Sea and the North Sea with some influence into continental Europe, especially in Germany. The simulated HCl shows a similar pattern to that of ClNO$_2$ but with much higher concentration (>10 times higher). The predicted BrO levels over Europe are low (average value ~0.17 pptv) with the largest predicted value occurring within the Arctic circle while GEOS-Chem predicted >1.0 pptv level of BrO in Mediterranean Sea (Sherwen et al., 2017). The predicted IO peaks over the Mediterranean region with a maximum value of 6.9 pptv.

Direct measurements of halogen species are very scarce and not available for the period covered in the present study (July 2016). Since a direct comparison is not possible, here we present a comparison of the simulated concentrations with observations from previous studies (Table 3), to provide an approximate assessment of the representation of halogen species in the HAL simulation of the CMAQ model over Europe.

Table 3. The comparison of observed and simulated halogen species

Location	Species	Observation *	Simulation #
Hessen, Germany [a]	$ClNO_2$	800.0	273.4
London, United Kingdom [b]	$ClNO_2$	724.0	801.5
Weybourne, United Kingdom [c]	$ClNO_2$	65	373
Weybourne, United Kingdom [d]	$ClNO_2$	946	373
Weybourne, United Kingdom [e]	$ClNO_2$	1100 (summer) 75.6 (autumn) 733 (winter)	373
Leicester, United Kingdom [e]	$ClNO_2$	274 (spring) 74.2 (summer) 248 (winter)	274
Penlee Point , United Kingdom [e]	$ClNO_2$	922	319
Mace Head, Ireland [f]	BrO	6.5	10.1
Brittany, France [g]	BrO	7.5	0.4
Dead Sea [h]	BrO	100.0	0.2
Mace Head, Ireland [i]	IO	4.0~50.0	3.9
Brittany, France [j]	IO	7.7~30.0	1.1
Dagebull, Germany [k]	IO	2.0	9.0
Atlantic Ocean [l] (Prados-Roman et al., 2015)	IO	0.4 to 0.5 (daytime average)	0.4 to 2.0 (daytime average)

*: Maximum value (pptv).
#: Maximum value (pptv) from the HAL simulation.
a: Phillips et al. 2012.
b: Bannan et al., 2015.
c: Banna et al., 2017.
d: Sherwen et al., 2017.
e: Sommariva et al., 2018.
f: Saiz-Lopez et al., 2004.
g: Mahajan et al. 2009.
h: Matveev et al., 2001; Holla et al., 2015.
i: Allan et al., 2000; Commane et al., 2011.
j: Britter et al., 2005; Furneaux et al., 2010.
k: Peters et al., 2005.
l: Prados-Roman et al., 2015

Numerous $ClNO_2$ measurements have been reported around the globe which show that $ClNO_2$ is ubiquitous in the boundary layer with maximum values ranging from hundreds to thousands pptv in polluted coastal (Osthoff et al., 2008; Wang et al., 2016) and continental regions (Tham et al., 2016; Thornton et al., 2010). A few campaigns have been conducted in Europe. Phillips et al. (2012) reported a maximum value of 800 pptv for $ClNO_2$ in Hessen, Germany where CMAQ predicts a concentration of 273 pptv. Bannan et al. (2015) observed a peak value of 724 pptv in London where CMAQ predicts a concentration of 802 pptv. Simulations with the GEOS-Chem model (Sherwen et al., 2017) reported maximum values of 110 pptv and 140 pptv at Hessen and London, respectively. Several field campaigns have been conducted in Weybourne in the past few years to measure $ClNO_2$. Sherwen et al. (2017) reported a peak concentration of 946 pptv. Bannan et al. (2019) reported a peak value of 65 pptv, and Sommariva et al. (2018) reported a peak value of 1100 pptv in summer, 75.6 pptv in autumn and 733 pptv in winter. CMAQ simulated a maximum of 373 pptv at that location while GEOS-Chem predicted 458 pptv. Sommariva et al. (2018) also reported measurements of $ClNO_2$ at Leicester with a maximum value of 274 pptv in spring, 74.2 pptv in summer, 248 pptv in winter, and that at Penlee Point a peak value of 922 pptv. CMAQ predicted a maximum of 274 pptv at Leicester and 319 pptv at Penlee Point. Eger et al. (2019) conducted shipborne observation of $ClNO_2$ in the Mediterranean Sea and reported up to 600 pptv $ClNO_2$ during their campaign, which is similar to the prediction of the present study.

The observed levels of HCl in Europe range from <100 pptv to 5000 pptv (Hossaini et al., 2016 and the reference therein). The CMAQ model predicted monthly average concentrations of HCl 6.3 to 1249 pptv, which is similar to the observation ranges. GEOS-Chem (Sherwen et al., 2017) predicted a maximum of 12 pptv for HCl, which is significantly lower than the measurements in Europe.

BrO measurements have been reported at ground-based sites and during the ship cruises which generally demonstrate a range of 0.5 to 2.0 pptv maximum values for land measurements and 3.0 to 3.6 pptv for ship measurements (Saiz-Lopez and von Glasow, 2012). BrO observations have been reported at several coastal sites in Europe. BrO level of up to 6.5 pptv (Saiz-Lopez et al., 2004) and 7.5 pptv (Mahajan et al., 2009) were reported in Mace Head and Brittany, respectively. CMAQ predicts 10.1 pptv and 0.4 pptv at those locations, which are lower than the measurements. Sherwen et al. (2017) also predicted small values with a maximum of 0.8 pptv in Mace Head and 0.5 pptv in Brittany. An extremely high level of BrO, ~100 pptv, was observed over the Dead Sea (Matveev et al., 2001; Holla et al., 2015). CMAQ is not able to reproduce such a high level of BrO due to the lower bromide content in typical ocean water (which was used in the present study for the Dead Sea) compared to the exceptionally high bromide content in Dead Sea (Tas et al., 2006; Sarwar et al., 2015).

Global measurements show that the IO levels observed by ground-based campaigns were generally between 0.2 and 2.4 pptv while those by ship measurement were ~3.5 pptv (Saiz-Lopez and von Glasow, 2012). Observations of IO have also been conducted in Europe. Maximum IO levels of 4.0~50.0 pptv were measured at Mace Head (Allan et al., 2000; Commane et al., 2011). CMAQ predicts a value of 3.9 pptv at Mace Head while GEOS-Chem predicted a value of 0.6 pptv (Sherwen et al., 2017). In Brittany, up to 7.7~30.0 pptv of IO were observed by Bitter et al. (2005) and Furneaux et al. (2010). CMAQ predicts 1.1 pptv of IO at Brittany and Sherwen et al. (2017) predicted 0.07 pptv. A maximum IO concentration of 2.0 pptv was reported in Dagebull (Peters et al., 2005), and CMAQ predicts 9.0 pptv at that site, while GEOS-Chem predicted 1.8 pptv (Sherwen et al., 2017). Prados-Roman et al. (2015) reported the level of IO during a ship-based campaign in the range of <0.4 to >1.4 pptv (daytime average) around the globe and 0.4 to 0.5 pptv (daytime average) in the south of Spain and the west of Africa (over the Atlantic), and the present study predicted 0.4 to 2.0 pptv (daytime average) of IO over those areas.

3.3 Influence of halogen chemistry on the atmospheric oxidation capacity

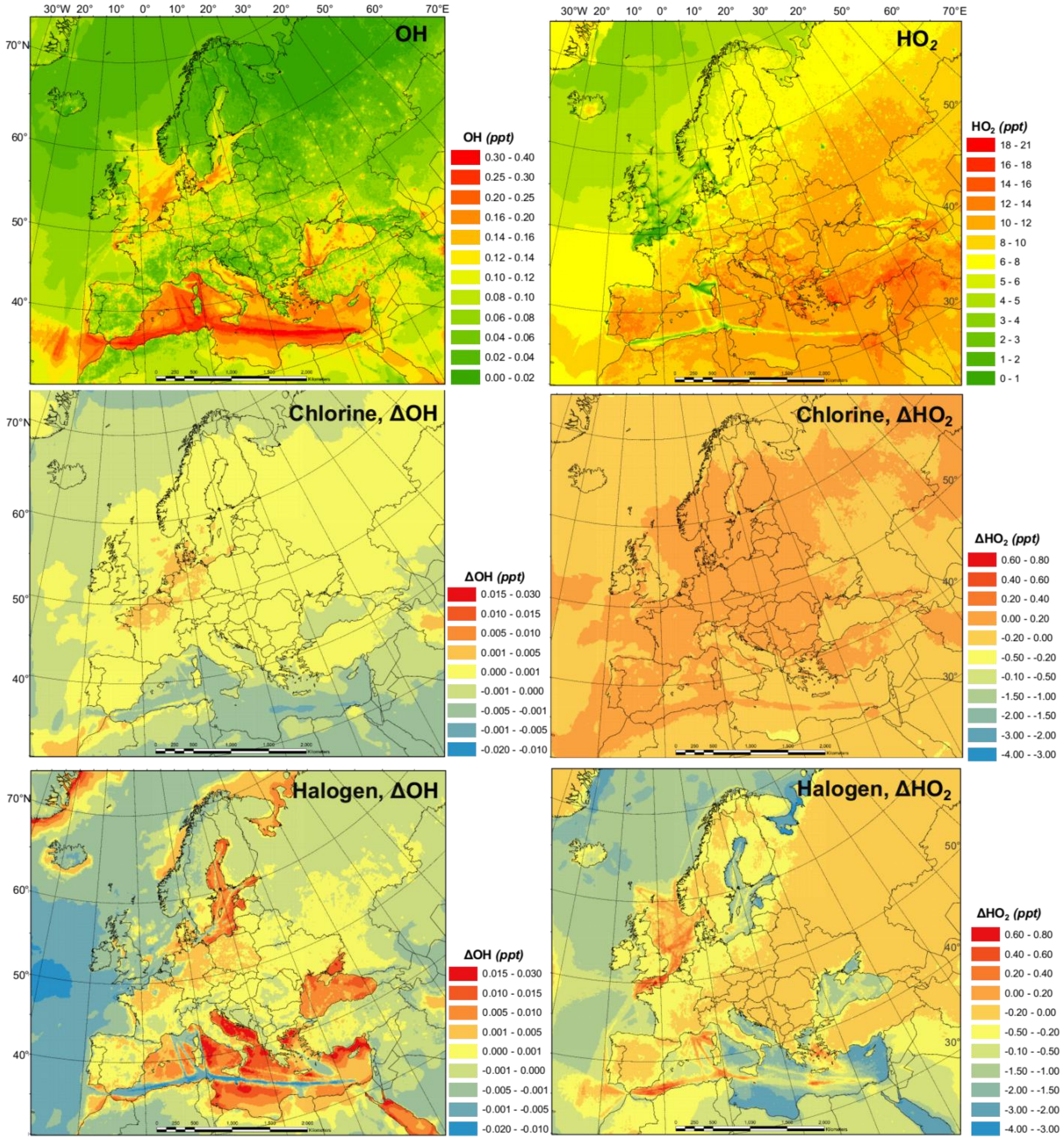

Figure 3. Monthly average OH and HO_2 concentration in the BASE simulation, and changes due to chlorine (CL) and full halogen chemistry (HAL).

Fig. 3 shows the monthly average concentrations of the OH and HO_2 radicals predicted by the BASE simulation and the impact of chlorine chemistry (CL-BASE), and the full halogen chemistry (HAL-BASE), on the simulated OH and HO_2 levels. In the BASE simulation, the highest OH concentration levels are predicted over the oceans especially along ship tracks, with a maximum value of 0.38 pptv (~9.5×10^6 molecule cm^{-3}). The chlorine chemistry slightly increased the OH level over most of the domain by up to 0.0041 pptv (1.0×10^5 molecule cm^{-3}). The impact of the halogen chemistry has competing effects on OH concentrations with a maximum increase of 0.030 pptv (7.5×10^5 molecule cm^{-3}) and a reduction of as much as 0.023 pptv (5.8×10^5 molecule cm^{-3}). The net change of OH due to halogen was near zero. The BASE simulation predicts the highest values of HO_2 over the Mediterranean Sea with a maximum value of 21 pptv. The chlorine chemistry increases the HO_2 level in the areas of elevated $ClNO_2$ predictions (Fig. 2 and Fig. 3). The further addition of halogen chemistry lowers the HO_2 as much as 3.7 pptv and increases HO_2 up to 0.73 pptv compared to the BASE simulation. The overall difference of HO_2 because of halogens was -0.59 pptv in the European domain. The effect of halogen chemistry on OH and HO_2 is the combined effect of the following four pathways: (1) conversion of HO_2 to OH via XO (R5 and R6), in which HO_2 decreases and OH increases; (2) reduction of O_3 (R1 and Figure 7) and the reduced production of OH by O_3 photolysis, in which both the OH and HO_2 decrease; (3) increase of NO_2 (R4 and Figure 7) and the enhanced consumption of OH by the reaction with NO_2, in which both OH and HO_2 decrease; and (4) increased oxidation of VOCs due to halogens in which both OH and HO_2 increase. Pathway (1) leads to the increase of OH and decrease of HO_2 over coastal area, Mediterranean Sea and Baltic Sea (except the ship tracks). Pathway (2) results in a decrease of OH and HO_2 over the remote ocean. Pathway (2), (3), and (4) results in a decrease of OH and an increase or decrease of HO_2 along the ship tracks in Mediterranean Sea.

Sarwar et al. (2015) reported a small overall decrease of OH (1%) and a significant decrease of HO_2 (11%) in the Northern Hemisphere due to the bromine and iodine chemistry. Their results suggest a considerable reduction of the HO_2/OH ratio which is consistent with the present study. Muñiz-Unamunzaga et al. (2018) found a slight increase of diurnal OH (1-2%) and a noticeable decrease of HO_2 (4%) leading to a decrease of HO_2/OH in Los Angeles, California. Sherwen et al. (2017) suggested that OH was reduced across their European domain due to the halogen chemistry, and concluded that the shift of HO_2 to OH by XO could not compensate for the decrease of OH due to the loss of O_3. Another GEOS-Chem study, however, predicted an increase of OH over the Mediterranean Sea (Stone et al., 2018). The discrepancy among the previous studies and between those works and the present one is difficult to deduce and requires further investigation. Several possible causes could lead to different simulated levels of halogens and their impact on oxidants, including the different mechanism of producing and recycling halogen species (Sarwar et al., 2019), spatial resolution (Sommariva et al., 2018), emission inventory (Wang et al., 2019), and different spatio-temporal scale of interest (Stone et al., 2018).

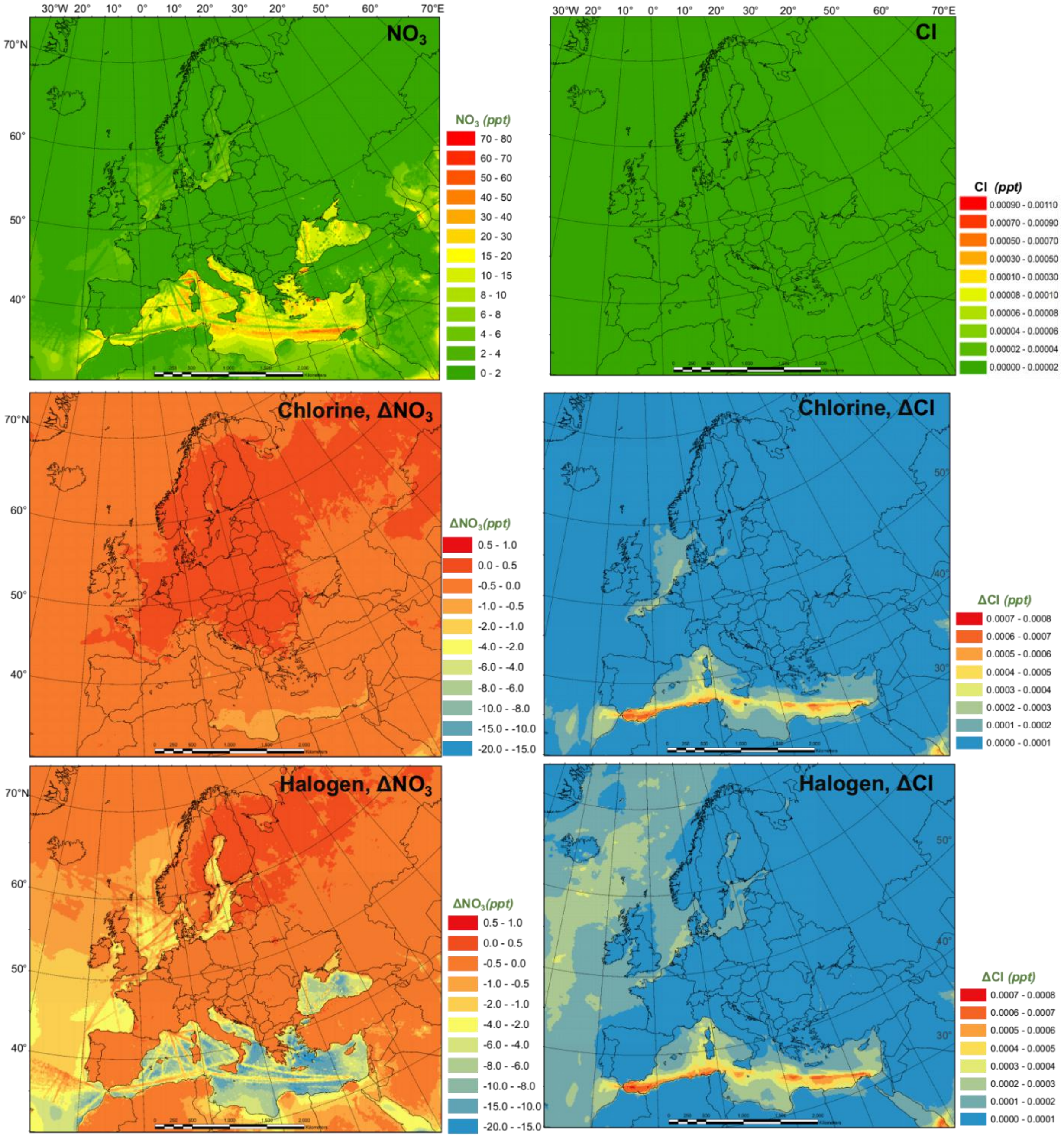

Figure 4. Monthly average NO_3 and Cl radical concentrations in the BASE simulation, and changes induced by chlorine (CL) and full halogen chemistry (HAL).

Fig. 4 presents the monthly average prediction of NO_3 and Cl radicals in the BASE scenario and the influence of chlorine (CL-BASE) and halogen chemistry (HAL-BASE) on the levels of NO_3 and Cl. The BASE simulation predicted relatively high NO_3 concentrations over the Mediterranean Sea along the busy shipping tracks. Although concentrations as high as 71 pptv are found, the majority of oceanic regions have concentrations in the range of 10 to 40 pptv and between 0 and 4 pptv over land. The chlorine chemistry slightly increases the NO_3 radical due to the increase of the O_3 (see section 3.4). In contrast, the halogen chemistry considerably reduces NO_3 concentrations by as much as 20 pptv at nighttime over the Mediterranean Sea. Muñiz-Unamunzaga et al. (2018) reported a 20-50% (2 to 4 pptv) decrease of NO_3 radical in Los Angeles, California when considering the halogen chemistry.

In the BASE simulation, the Cl concentration was negligible because there was no relevant chlorine source incorporated in the CMAQ model. The CL simulation contains the production of $ClNO_2$ and its subsequent photolysis which increases the Cl concentration of as high as 7.0×10^{-4} pptv (~1.75×10^4 atom cm^{-3}). The HAL simulation predicted a very similar magnitude and spatial distribution of chlorine concentration. Sherwen et al. (2017) reported Cl concentrations less than 1.4×10^4 atom cm^{-3} (~5.6×10^{-4} pptv) over Europe, comparable to our prediction. Hossaini et al. (2016) reported more than 1.0×10^4 atom cm^{-3} (~4.0×10^{-4} pptv) of chlorine over Asia, Europe and North America, with a maximum of 8.5×10^4 atom cm^{-3} (~3.4×10^{-3} pptv), using a global chemical transport model (TOMCAT) that incorporated chlorine sources from sea salt dechlorination, coal and biomass burning, oxidation of natural and anthropogenic chlorocarbon, and heterogeneous reactions on sea salt and sulfate aerosol. In their study, Hossaini et al. (2016) used the Reactive Emission Inventory of Chlorine (Keene et al., 1999) which Wang et al. (2019) reported to be unrealistic for present day applications.

The current and the previous studies simulated a broad range of the surface Cl concentrations although they were all within the scope of the reported observed (observation-based calculation) values of 10^3 to 10^5 atom cm^{-3} (~ 4.0×10^{-5} to 4.0×10^{-3} pptv) according to the review of Saiz-Lopez and von Glasow (2012). In light of the considerable variation of observed and model predicted Cl level, further study may be needed to comprehensively evaluate the significant role of Cl in the troposphere.

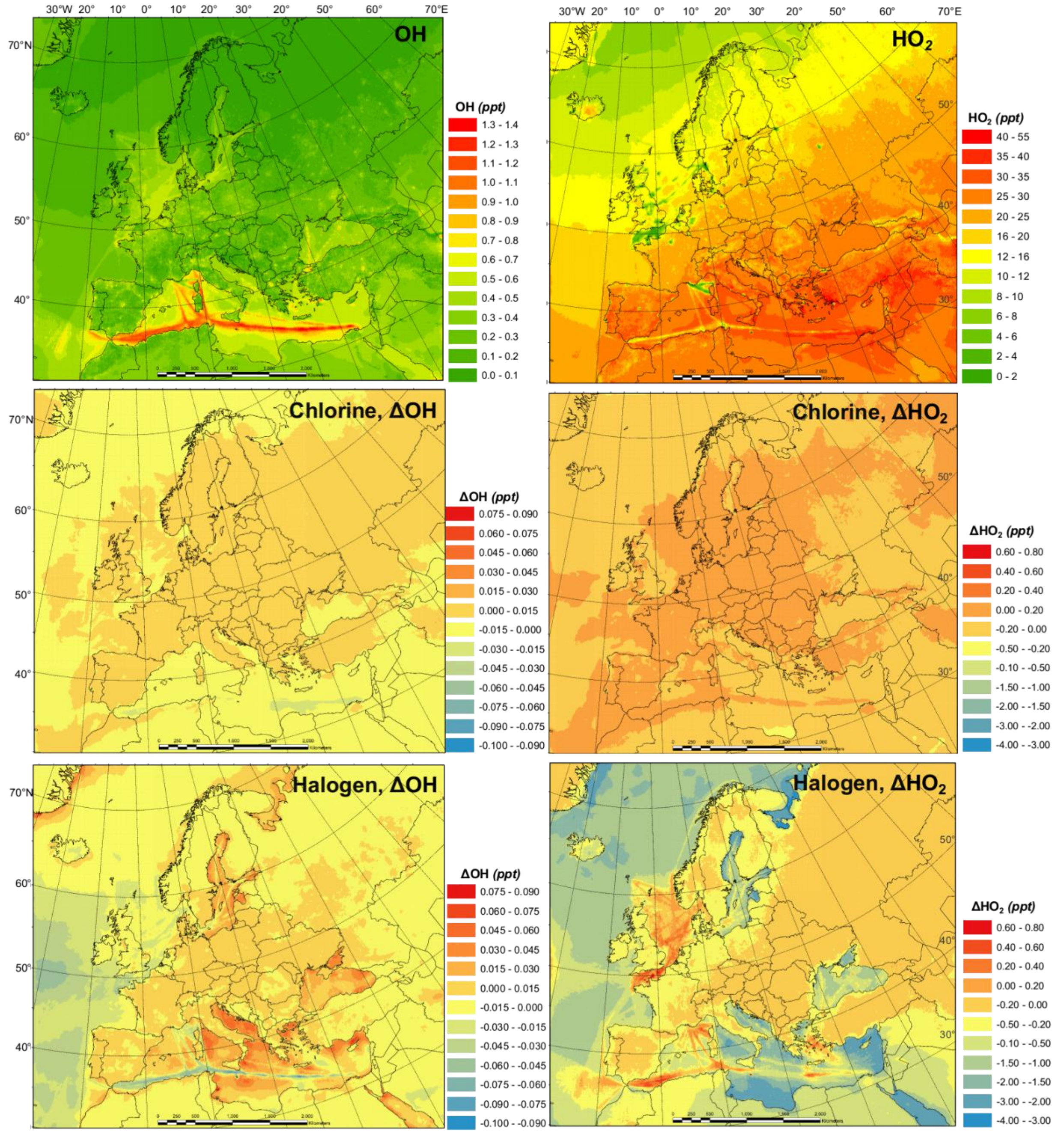

Figure 5. Monthly average of daily-maximum concentrations of OH and HO$_2$ in the BASE simulation, and changes due to chlorine (CL) and full halogen chemistry (HAL).

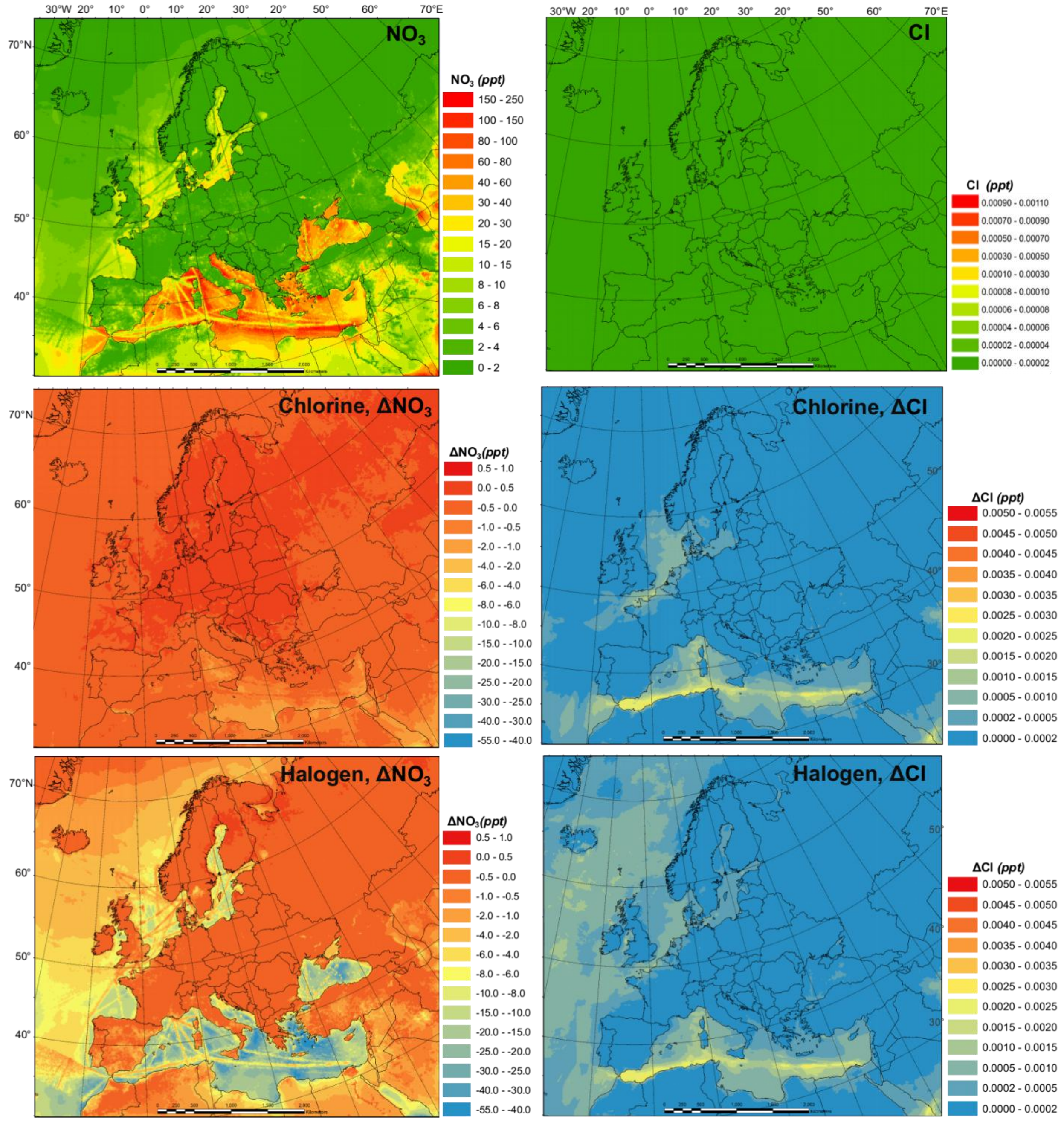

Figure 6. Monthly average of daily-maximum concentrations of NO_3 and Cl radical in the BASE simulation, and changes induced by chlorine (CL) and full halogen chemistry (HAL).

Fig. 5. and Fig. 6 demonstrate the monthly average of the daily-maximum concentrations of OH, HO_2, NO_3, and Cl in the BASE simulation and also the impact of chlorine (CL-BASE) and the halogen chemistry (HAL-BASE). The maximum values of OH, HO_2, and Cl were predicted during the daytime but they peak at different hours with Cl in the early morning and OH and HO_2 later in the day, while the highest levels of NO_3 radical were simulated during night-time.

The monthly average of daily-maxima of OH, HO_2, NO_3, and Cl (Fig. 5 and Fig. 6) have similar spatial pattern and higher concentrations (or changes of concentration) compared to those of the monthly averages (Fig. 3 and Fig. 4). The monthly average of daily-maximum OH (HO_2 and NO_3) radical is 3.7 (2.5 and 3.4) times of the monthly average OH (HO_2 and NO_3) concentration in the BASE simulations, while both the monthly average of daily-maximum and monthly average Cl shows negligible values. The monthly average of daily-maximum changes of OH (HO_2) concentration due to the chlorine and halogen chemistry has magnitude of -0.10 to 0.09 pptv (-4.0 to 0.8 pptv), which is wider than (the same as) that of the monthly averages, i.e., -0.020 to 0.03 pptv (-4.0 to 0.8 pptv). For the NO_3 (Cl) radicals, the magnitude of changes in monthly average is -20 to 1.0 pptv (0.0 to 0.0008 pptv) while that in monthly average of daily-maximums is -55 to 1.0 pptv (0.0 to 0.0055 pptv).

We also examine the diurnal variations of the four radicals in the BASE and HAL scenarios (Fig. S3 and S4). Halogens have small effect on the diurnal pattern of OH. HO_2 is reduced by halogens especially in the mid-day. NO_3 radical is strongly decreased throughout the night after the addition of halogens. Cl atom is released by the halogen chemistry in the early morning. The significant effects of halogen chemistry on the diurnal variation of OH, HO_2, NO_3, and Cl radicals highlight the role of halogen chemistry in regulating the atmospheric oxidation capacity throughout the day with the highest effect on Cl in the early morning, maximum effects on OH and HO_2 in daytime, and largest effect on NO_3 at night.

3.4 Impact of halogen chemistry on regulated gaseous air pollutants

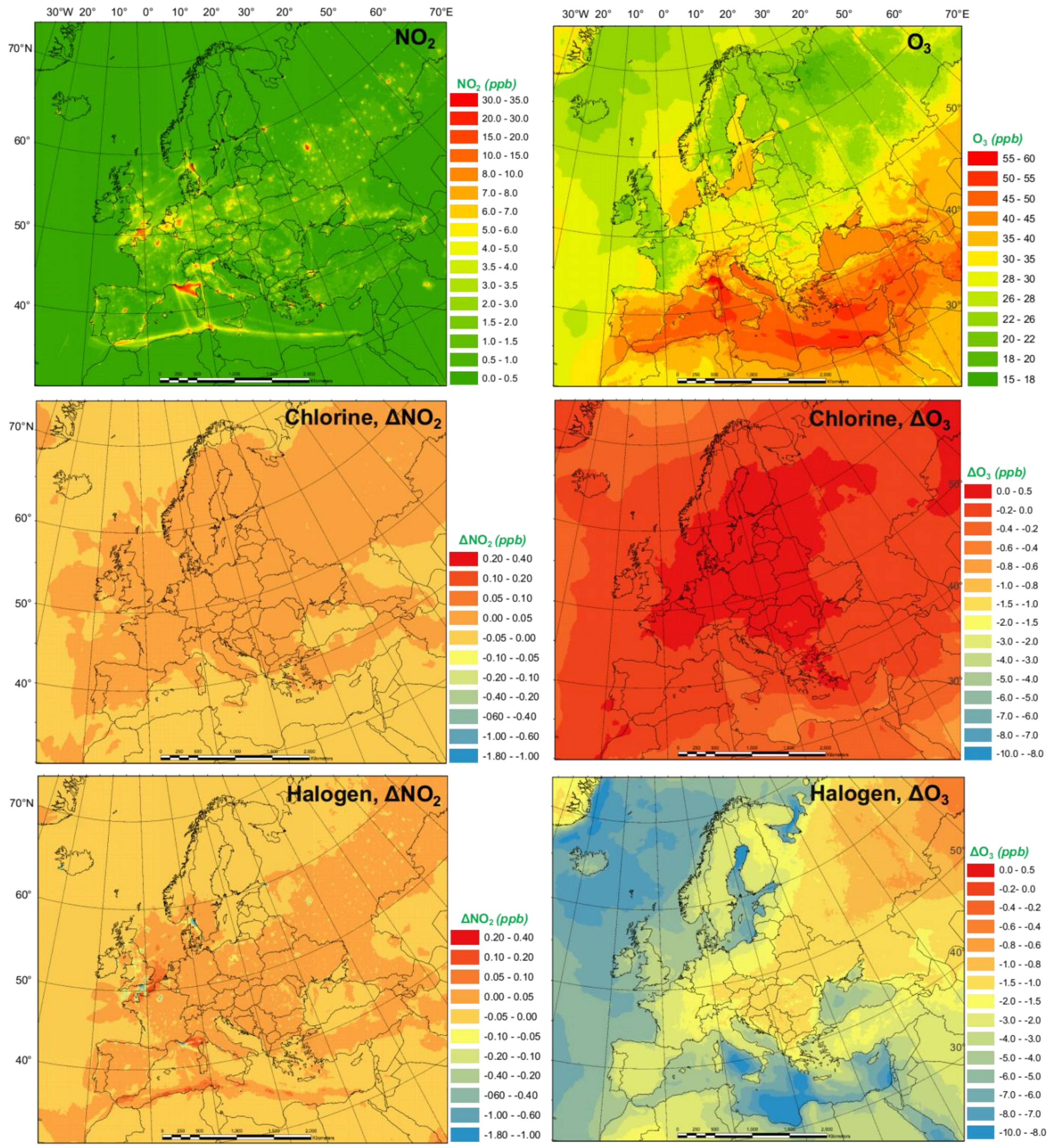

Figure 7. Monthly average NO$_2$ and O$_3$ concentration in the BASE simulation, and changes induced by chlorine (CL) and full halogen chemistry (HAL).

The monthly average modeled NO$_2$ and O$_3$, two major gaseous air pollutants in Europe, and the effect of chlorine and halogen chemistry on the two regulated gaseous species were shown in Fig. 7. The BASE simulation produced many hot spots of NO$_2$ over Europe, in the vicinity of the major cities and the ship trajectories. The chlorine chemistry slightly increases the level of NO$_2$

(by up to 0.038 ppbv) in the majority of the domain since (1) the production and the subsequent photolysis of $ClNO_2$ recycles the NO_x extending its lifetime, which increases both NO_2 and NO (Fig. 7 and S5) and (2) the increased O_3 level (Fig. 7) enhances the transformation of NO to NO_2, which increases NO_2 and decreases NO (Fig. 7 and S2). Some grid cells show a decrease of NO_2 (as much as 0.09 ppbv) because the enhanced oxidative capacity (Section 3.3) promotes the cleansing of NO_2 via OH forming HNO_3. The full halogen chemistry enhances NO_2 (up to 0.20 ppbv) over the North Sea and the Mediterranean Sea and decreases NO_2 (as much as 1.7 ppbv) in the most polluted hot spots. The increase of NO_2 occurs by the reactions of XO with NO. Meanwhile, in the most polluted regions, the NO-NO_2 balance is predominantly controlled by the reactions of NO with HO_2 and O_3. With the decrease of HO_2 and O_3 due to the halogen chemistry, the transformation of NO to NO_2 is reduced which leads to decreasing NO_2 and increasing NO.

The monthly average O_3 concentration over Europe from the BASE simulation was relatively high (up to 57 ppbv) especially over the southern Europe, where higher temperature and more intensive radiation promote the formation of this secondary pollutant. Chlorine chemistry increases O_3 levels over the land of Europe with a maximum increment of 0.22 ppbv and decreases O_3 over the oceanic area as much as 0.76 ppbv. The full halogen chemistry decreases O_3 throughout the domain with a maximum reduction of 10 ppbv. On average, the halogen chemistry reduces O_3 concentration by more than 3.0 ppbv in coastal Europe and by over 2.0 ppbv over western and central Europe (nearly one thousand kilometers from the ocean). Our model simulation highlights the fact that halogen chemistry has a large impact on O_3 concentrations over the oceanic areas and a moderate impact on O_3 over coastal and continental regions of Europe.

Muñiz-Unamunzaga et al. (2018) reported a decrease of -2.0 ppbv O_3 in the inland areas of the western US (several hundreds of kilometers from the ocean) and a reduction of 2.5 to 5.0 ppbv O_3 in the coastal regions due to the full-halogen chemistry. Sarwar et al. (2015) suggested that the inclusion of halogen processes reduced O_3 concentrations by 2.0 to 4.0 ppbv over most of the terrestrial regions in the North Hemisphere, and over 6.0 ppbv in some coastal areas. Sherwen et al. (2017) used a revised version of GEOS-Chem (Sherwen et al., 2016) with halogen chemistry to show substantial reductions in O_3 over Europe with an average reduction of 13.5 ppbv in the domain and a maximum of 28.9 ppbv in some locations.

3.5 Implications for policy assessment

The current air quality management in Europe has two main objectives: (1) to protect human health and (2) to protect the environment. While many plans and measures have prioritized Particulate Matter (PM) or NO_2, policies to reduce O_3 concentrations are still needed (EEA, 2018a). The World Health Organization (WHO) Air Quality Guidelines value for O_3 (maximum daily 8-hour mean of 100 $\mu g \cdot m^{-3}$) was exceeded in 96% of all the reporting stations in Europe, although this is especially true for the areas near the Mediterranean Sea. According to the European Environment Agency latest report, 12% of the EU-28 urban population is exposed to O_3 concentrations above the European Union target value threshold (maximum daily 8-hour mean of 120 $\mu g \cdot m^{-3}$ not to be exceeded on more than 25 days/year, as set out by the Directive 2008/50/EC) in 2016. Apart from significant potential health effects (Jerrett et al., 2009; Malley et al., 2017), O_3 is also known to have a negative impact on vegetation (Mills et al., 2011). The target value for the protection of vegetation (18,000 $\mu g \cdot m^{-3} \cdot h$ accumulated over May to July), based on the Accumulated Ozone exposure over a Threshold of 40 ppbv index (AOT40 index), was exceeded in about 31 % of all agricultural land in all European countries. The critical level for this pollutant (10,000 $\mu g \cdot m^{-3} \cdot h$ accumulated over April to September) was exceeded in 60% of the total forest area of the continent in 2016 (EEA, 2018a). Here we use the simulation results (BASE and HAL) in July to calculate the accumulated O_3 and the difference between two scenarios. We note that the accumulated O_3 is considerably reduced (~15% along the coast) after the addition of halogens (Fig. 8).

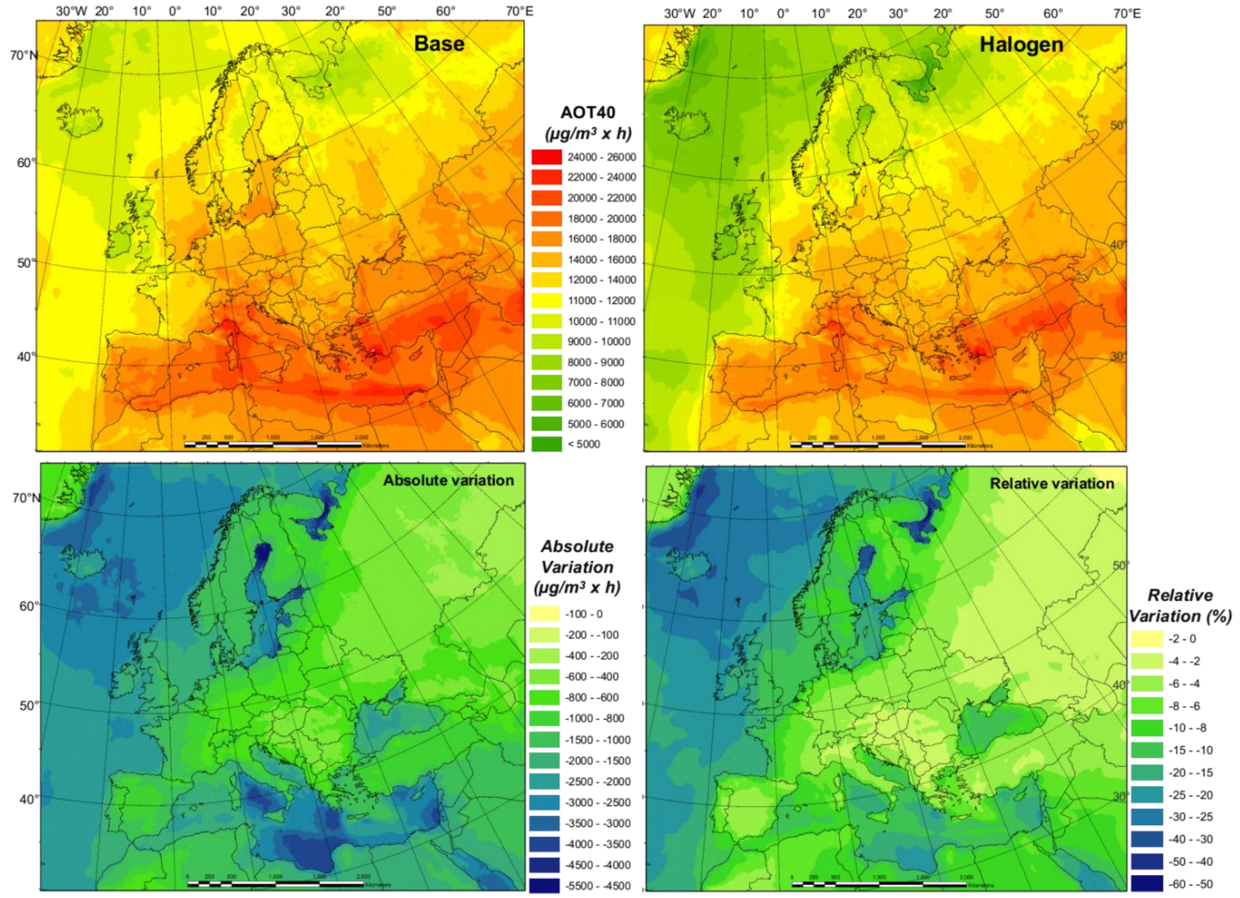

Figure 8. Accumulated O_3 in July in the BASE and HAL simulations, and absolute and relative changes between the two simulations.

We find that halogen chemistry strongly affects ambient O_3 concentration and may need to be considered in the formulation of plans and strategies for O_3 non-attainment areas. We see differences between BASE and HAL simulations (over land in July 2016) as high as 12% and 36% for the number of days with daily maximum 8 h O_3 over 120 μg·m^{-3} and the monthly average daily maximum 8 h O_3 level, respectively (Fig. S6 and Fig. S7). Furthermore, we notice strong regional differences, mainly between coastal and inland areas. The considerable effect of halogen chemistry on air quality implies the need to improve the robustness and accuracy of modeling tools to design customized policies to control O_3.

In Section 3.3 and 3.4, we have also discussed the effect of halogen chemistry on the partitioning of OH/HO_2 and NO/NO_2. The budgets of HO_x and NO_x are key parameters to accurately simulate the formation of O_3 and its response to the reductions of the precursors, namely NO_x and VOCs (e.g., Li et al., 2018). Air quality models are predominantly used to formulate air pollution control policy by examining the responses of O_3 levels to various reduction rates of NO_x and/or VOCs. The models which do not include the comprehensive halogen chemistry can potentially lead to different O_3 concentration responses to NO_x and/or VOC emission changes in Europe.

This study also demonstrates that chlorine chemistry affects the formation of O_3. The current policy is only designed to control the long-lived chlorinated species (Hossaini et al., 2015), but not reactive chlorine species, e.g., HCl, chloride, and short-lived chlorocarbons, from the coal burning, biomass burning, and industrial activities. The coal-fired power plants in EU (EEA, 2018b; Kuklinska et al., 2015) can potentially provide chlorine sources, making the implications of halogen chemistry even more relevant.

4. Conclusion

We applied the CMAQ model with comprehensive halogen chemistry (Cl, Br and I) to conduct high-resolution simulations for examining the impact of halogen chemistry on air quality over Europe.

The comparison of model results with observations from over 400 monitoring sites indicates that the CMAQ model is capable of reproducing the concentrations and temporal variations of air pollutants over Europe and can be employed to study the impact of halogen chemistry in Europe. The comparison of predicted halogen species concentrations with measurements suggests that the CMAQ model is able to predict observed levels of chlorine and iodine species although it underestimates bromine species.

The chlorine chemistry enhances the atmospheric oxidation capacity by significantly increasing the level of Cl radical and affects the levels of OH, HO_2, NO_3, O_3, and NO_2. The combined halogen chemistry marginally increases the level of OH and reduces HO_2, NO_3, and O_3. The impact of halogen chemistry on ambient concentration of NO_2 is smaller but non-negligible.

Halogen chemistry significantly influences the atmospheric oxidation capacity throughout the day by imposing the highest effect on Cl in the early morning, maximum effects on OH and HO_2

in daytime, and largest effect on NO_3 at night. Halogen chemistry can have a strong influence on atmospheric composition over oceanic and coastal regions but also some noticeable impacts over continental Europe. This study highlights the potential benefit of incorporating halogen chemistry into air quality models for policy development.

Although the incorporation of the halogen chemistry may improve the capabilities of 3D Eulerian chemical transport models, we acknowledge that large uncertainties still exist in the assessment of halogen chemistry impact due to emission inventories (e.g., chlorine emission inventory; Wang et al., 2019), model configuration (e.g., grid size; Sommariva et al., 2018), chemical mechanism (e.g., photolysis rate of iodine oxides, recycling rate of halogen species on aerosol; Simpson et al., 2015), etc. Further field, laboratory, and theoretical studies are needed to constraint modeling studies for evaluating the impacts of halogen chemistry on air quality and for assessing air quality policy implications.

Data availability. The data used and demonstrated in this study are available upon the request to the corresponding author.

Author contributions. AS-L designed research. GS and BG conducted the CMAQ modeling. RB, DP, JD, GS, BG, QL, and AS-L analyzed the results. QL, RB, GS, and AS-L wrote the paper with contributions from all authors.

Competing interests. The authors declare that they have no conflict of interest.

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DISCLAIMER

The views expressed in this paper are those of the authors and do not necessarily represent the views or policies of the U.S. EPA.

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