1	Supporting Information
2	Soil emissions, soil air dynamics and model simulation of gaseous
3	mercury in subtropical forest
4	
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26 Supporting Text:

27 Site description

28 The mean annual precipitation, temperature and daily relative humidity at the TFP are 1230 29 mm, 18.2 °C and 95%, respectively. The ecosystem type at the TFP study site is a Masson Pine 30 dominated forest, with some associated ever-green broad-leaved species. Trees were planted in the 31 1960s. The soil is typically mountain yellow earth (corresponding to a Haplic Acrisol in FAO). The 32 soil is acidic, with a pH of 3.79. From previous studies, the mean Hg concentrations in precipitation, throughfall, litterfall and organic soils were 55.3 ng L⁻¹, 98.9 ng L⁻¹, 104.8 \pm 18.6 ng g⁻¹ and 191 \pm 33 65 ng g⁻¹, respectively, with an annual Hg input of 291.2 μ g m⁻² yr⁻¹.(Zhou et al., 2016;Zhou et al., 34 2015) 35

36

37 Dynamic Flux Chamber design and measurement

Semi-cylindrical quartz glass and open-bottom DFCs (4.71 L) were used during the sampling campaign. The area of the DFCs over the soil surface was 20×30 cm, with six inlet holes (1 cm diameter). At the outlet of the chamber, an orifice was connected to two exits; one to a regulated suction pump with a flow rate of 10 L min⁻¹, and the other to a gold cartridge for trapping outlet TGM. On the two opposite sections of the chamber, a gold trap was placed in the inlet to trap TGM entering from outside air.

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5 *Quality Assurance and Quality Control*

46 Gold cartridges used to sample pore TGM were collected at same time as soil TGM flux 47 measurements. All the cartridges were brought back to the laboratory at the TFP Forest Station for 48 quantitative analysis using a cold vapor atomic fluorescence spectroscopy (CVAFS) detector 49 (Brooks Rand III). For all Hg analysis, quality assurance and quality control (QA-QC) measures 50 included the system blanks, gold cartridge recovery, recovery and duplicates before and after the campaign in each season. A quantitative volume of saturated Hg air at a known temperature was 51 52 injected to calibrate the recovery of all gold cartridges' before and after the campaigns in each season. 53 The recoveries of gold cartridges before and after the operation ranged from 98.8 to 103.2% and 54 96.3 to 102.5%, respectively. The standard deviation of the recoveries was 2.6%. Before and after 55 field sampling, the collection efficiency of the gold quartz cartridges was detected by connecting two cartridges in sequence and sampling the ambient air for 24 h in laboratory. For all cartridges, 56 less than 1% Hg was detected on the second cartridges compared to the first cartridge, indicating 57 58 that more than 99% of TGM was absorbed by the gold cartridges during the field operation. Blanks 59 of the soil TGM flux sampling system were routinely measured by placing the DFC on a quartz 60 glass surface in the five plots. The sampling time of the blank measurement was same to the soil-61 air TGM flux measurement, which were collected twice a day: every morning (about 8:00) and S2

afternoon (about 17:00) representing night (17:00–8:00 of next day) and day (8:00–17:00) emissions, respectively. The averaged blank was 0.13 ± 0.21 ng m⁻² h⁻¹ (n=10), which was subtracted from the soil-air TGM flux in each season. The calibration curve was developed using Hg saturated air and had to have a correlation coefficient greater than 0.99 before the samples analysis could proceed.

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68 Environmental measurements

69 Daily meteorological parameters were collected and averaged over 5-min intervals. Daily air temperature and solar radiation were monitored using a TP 101 digital thermometer and a GLZ-C 70 71 photo synthetically radiometer (TOP Ltd. China), respectively, during diurnal measurements. 72 Percent moisture was monitored with Time Domain Reflectometry (TDR) Hydra Probe II 73 (SDI-12/RS485) and a Stevens water cable tester (USA). Measurements were taken at the same 74 time with gold trap collection. Solar radiation was collected with a weather station (Davis Wireless 75 Vantage VUE 06250 Weather Station, Davis Instruments, Hayward, CA) located in the TFP Forest 76 Station about 500 m away from the sub-catchment.

77 For each DFC sampling location, bulk soil samples were collected from the DFC footprints 78 (0-5 cm) in each month after the end of the measurement period. Soil samples were dried and 79 homogenized, and completely ground to a fine powder in a pre-cleaned stainless-steel blender. The total Hg concentration in the soil samples was determined using a DMA-80 direct Hg analyzer 80 81 (Milestone Ltd., Italy). SOM content in soils was determined using the sequential loss on ignition 82 (LOI) method.(Zhou et al., 2013) A homogenized soil sample (WS) was dried at 105 °C for about 83 12-24 h to obtain the dry weight of the samples (DW_{105}). The heated dry sample was then burned at 550 °C for 4 h and the weight of the sample after heating at 550 °C was DW₅₅₀. Thus, the TOM 84 85 concentration (LOI₅₅₀) was calculated according to the following formula:

86 $LOI_{550}=100(DW_{105}-DW_{550})/WS.$

87

88 Statistical analysis

Mean pore TGM concentrations and soil TGM fluxes were compared among the five plots. Separate one-way ANOVAs were used to determine if the differences in Hg concentrations and fluxes were evident among the depths and plots. All differences in means were significant at the p=0.05 level and all means are reported with \pm one standard deviation from the mean. The correlation was analyzed by Pearson's Correlation Tests using SPSS software (SPSS Inc. 16.0) and correlation coefficient and p values are presented and significantly correlated at the level of 0.05.

D1-4-	T		Date of flux mea	asurement		Da	SOM (0-	Area			
Plots	Locations	Spring	Summer	Autumn	Winter	Spring	Summer	Autumn	Winter	5, %)	(%)
Plot A	Top of the hillslope of the coniferous forest	5 Mar-7 Apr	17 -19 Jun; 1-31 Jul; 10-24 Aug	3 Nov-6 Dec	24 Dec- 14 Jan	5 Mar-7 Apr	17 -19 Jun; 21- 31 Jul; 10-24 Aug	3 Nov-6 Dec	24 Dec- 14 Jan	13.6	42.4
Plot B	Middle of the hill slo pe of the coniferous f orest	5 Mar-7 Apr	17 -19 Jun; 1-31 Jul; 10-24 Aug	3 Nov-6 Dec	24 Dec- 14 Jan		17 -19 Jun; 21- 31 Jul; 10-24 Aug	3 Nov-6 Dec	24 Dec- 14 Jan	16.3	42.4
Plot C	Wetland	5 Mar-7 Apr	1-31 Jul; 10-24 Aug	3 Nov-6 Dec	31 Dec- 14 Jan					4.9	2.9
Plot D	Broad-leaved forest	5 Mar-7 Apr	17 -19 Jun; 1-31 Jul; 10-24 Aug	3 Nov-6 Dec	24 Dec- 14 Jan	5 Mar-7 Apr	17 -19 Jun; 21- 31 Jul; 10-24 Aug	3 Nov-6 Dec	24 Dec- 14 Jan	8.8	10
Plot E	Open field (deserted a gricultural land)	22 Mar- 7 Apr	17 -19 Jun; 1-31 Jul; 10-24 Aug	3-23 Nov	30 Dec- 14 Jan		17 -19 Jun; 21- 31 Jul; 10-24 Aug	3-23 Nov	30 Dec- 14 Jan	4.1	2.3

Table S1. Characteristics and detail of measurements at five plots in the forested sub-catchments.

98	Table S2. A summary of empirica	al models in the literature for soil-air Hg fluxes.
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Parameters	Soil type	Equations	References
Temperature	Forest lake and soil	F = EXP (-E/RT),	Xiao et al.(Xiao
		E is the apparent activation energy; R is the gas constant.	et al., 1991)
Temperature	Forest, open and agricultural	$\operatorname{Ln} F = E/RT + \beta_1,$	Carpi and
	fields.	E is the apparent activation energy; R is the gas constant.	Lindberg(Carpi
			and Lindberg,
			1997)
Temperature	Forest soil	$\operatorname{Log} F = \beta_2 T + \beta_{3,}$	Xu et al.(Xu et
			al., 1999)
Temperature, soil Hg content	Bare soil	$\ln F = -\gamma/T + \beta_4 \ln Sc + \beta_5$	Gbor et al.(Gbor et
		γ is related to the apparent activation energy.	al., 2006)
Solar radiation, soil Hg content	Forest soil and artificially	$\ln F = \beta_6 L + \beta_7 \ln S c + \beta_8$	Gbor et al.(Gbor et
	shaded background soil		al., 2006)
Temperature, solar radiation	Forest soil during leaf-on	$F = \beta_9 L + \beta_{10} \operatorname{EXP}(\beta_{11} T)$	Choi and
	period		Thomas(Choi and
			Holsen, 2009)
Temperature	Forest soil during leaf-off	$F = \beta_{12} + [\beta I_3 \text{ EXP}(\beta_{14}T - 1)]/\beta_{15}$	Choi and
	period		Thomas(Choi and
			Holsen, 2009)
Temperature, solar radiation, soil	Laboratory study on	$F = S_c \times [\beta_{16} + \beta_{17}T + \beta_{18}W + \beta_{19}L + \beta_{20}(T \times L) + \beta_{21}(T \times W) + \beta_{22}(W)$	Lin et al. ^{(Lin et al.,}
moisture, Hg content		$(\times L) + \beta_{23}T^2 + \beta_{24}W^2 + \beta_{25}L^2]$	2010)
Temperature, solar radiation, Hg	Bare soil and soil under the	$F = (10^{-3} \times S_c) \times [\beta_{26} + \beta_{27}T + \beta_{28}L + \beta_{29}(T \times L) + \beta_{30}T^2 + \beta_{31}L^2]$	Kikuchi et
content	leaf canopy		al. (Kikuchi et al., 2013)

Temperature, solar radiation, soil	Forest soil and bare soil	$F = (a \times S_c) \times [\partial_0 + \partial_1 T + \partial_2 W + \partial_3 L + \partial_4 C_a + \partial_5 (T \times W) + \partial_6 (T \times L) +$	This study					
moisture, Hg content, atmospheric		$\partial_7(T \times C_a) + \partial_8(W \times L) + \partial_9(W \times Ca) + \partial_{10}(L \times C_a) + \partial_{11}T^2 + \partial_{12}W^2 +$						
TGM		$\partial_{I3}L^2 + \partial_{I4}C_a{}^2$]						

	Masson pine		Camphor		Whole sub-
Coefficients	forest	Wetland	forest	Bare soil	catchment
∂_0	1.18	1.98	7.05×10^{-1}	1.13×10	1.39
∂_1	1.80×10^{-1}	1.39×10^{-1}	3.06×10 ⁻²	-1.41	1.28×10^{-1}
∂_2	-1.13×10	-1.10	-9.24	4.48×10	-9.55
∂_3	1.05×10^{-2}	-2.37×10^{-2}	2.34×10^{-2}	2.75×10 ⁻³	1.06×10^{-2}
∂_4	-117×10^{-1}	-1.10	4.28×10^{-1}	-4.20×10^{-1}	-9.82×10^{-2}
∂_5	-1.27×10^{-1}	6.81×10^{-3}			-1.08×10^{-1}
∂_6	-2.43×10^{-4}	-1.13×10^{-4}	-5.11×10^{-4}	1.44×10^{-3}	-2.27×10^{-4}
∂_7					
∂_8				-4.91×10^{-2}	-1.13×10 ⁻³
∂_9	-2.19×10^{-4}	2.55×10^{-1}	-2.61		-2.54×10^{-1}
∂_{10}				3.43×10 ⁻²	7.90×10 ⁻⁷
∂_{11}	-2.47×10^{-3}	-1.00×10^{-3}	4.19×10 ⁻⁴	3.78×10^{-2}	-1.22×10^{-3}
∂_{12}	1.71×10		23.2×10	-1.05×10^{2}	1.44×10
<i>∂</i> 13	-4.88×10^{-5}	2.26×10^{-4}	-4.77×10^{-5}	-1.57×10^{-5}	-4.00×10^{-5}
∂_{14}		2.23×10^{-2}	3.33×10 ⁻²		4.85×10^{-3}

101 Table S3. Coefficients of the empirical models to predicting soil-air Hg fluxes using the annual

U_0	1.10	1.90	7.03~10	1.13^10	1.39
∂_I	1.80×10^{-1}	1.39×10^{-1}	3.06×10^{-2}	-1.41	1.28×10^{-1}
∂_2	-1.13×10	-1.10	-9.24	4.48×10	-9.55
∂_3	1.05×10^{-2}	-2.37×10^{-2}	2.34×10^{-2}	2.75×10^{-3}	1.06×10^{-2}
∂_4	-117×10^{-1}	-1.10	4.28×10^{-1}	-4.20×10^{-1}	-9.82×10^{-2}
∂_5	-1.27×10^{-1}	6.81×10 ⁻³			-1.08×10^{-1}
∂_6	-2.43×10^{-4}	-1.13×10^{-4}	-5.11×10^{-4}	1.44×10^{-3}	-2.27×10^{-4}
∂_7					
∂_8				-4.91×10^{-2}	-1.13×10^{-3}
∂_9	-2.19×10^{-4}	2.55×10^{-1}	-2.61		-2.54×10^{-1}
∂_{10}				3.43×10^{-2}	7.90×10^{-7}
∂_{11}	-2.47×10^{-3}	-1.00×10^{-3}	4.19×10^{-4}	3.78×10^{-2}	-1.22×10^{-3}
∂_{12}	1.71×10		23.2×10	-1.05×10^{2}	1.44×10
∂_{13}	-4.88×10^{-5}	2.26×10 ⁻⁴	-4.77×10^{-5}	-1.57×10^{-5}	-4.00×10^{-5}
∂_{14}		2.23×10^{-2}	3.33×10^{-2}		4.85×10 ⁻³
Notes: Both fi	rst-order and secon	d-order terms	of the four fac	tors investiga	ted show net positiv

data and corresponding environmental factors of the different ecosystems. 102

103 ve 104 effects on the measured Hg fluxes within the data ranges of the regression analyses, and the corresponding equation is $F = (a \times S_c) \times [\partial_0 + \partial_1 T + \partial_2 W + \partial_3 L + \partial_4 C_a + \partial_5 (T \times W) + \partial_6 (T \times L) + \partial_7$ 105 $(T \times C_a) + \partial_8 (W \times L) + \partial_9 (W \times Ca) + \partial_{10} (L \times C_a) + \partial_{11} T^2 + \partial_{12} W^2 + \partial_{13} L^2 + \partial_{14} C_a^2].$ 106

D	seaso		Plot A	A Plot B				Plot D)	Plot E			
Parameter	n	3	6	10	3	6	10	3	6	10	3	6	10
	Spr	0.37 a	0.54 a	0.48 a				0.43 a	0.09	0.13			
Soil	Sum	0.43 a	0.53 a	0.40 a	0.56 a	0.32 b	0.09	0.39 a	0.42 a	0.43 a	0.42 a	0.39 a	0.34
temperatur e	Fall	0.46 a	0.34 a	0.08	0.56 a	0.32 b	0.09	0.32 b	0.04	0.15	0.3 ^b	0.3 ^b	0.3 ^b
	Win	0.25	0.45 a	- 0.01	0.34 b	- 0.06	0.34 b	0.31 b	0.16	0.09	0.43 b	0.06	0
	Full year	0.85 a	0.89 a	0.83 a	0.86 a	0.74 a	0.82 a	0.84 a	0.75 a	0.67 a	0.70 a	0.67 a	0.62
	Spr	- 0.18	- 0.18	- 0.02				- 0.05	- 0.05	- 0.11			
	Sum	- 0.22	- 0.34 b	- 0.30 b	- 0.34 b	- 0.36 a	- 0.26	- 0.32 b	- 0.27 b	- 0.40 a	-0.1	- 0.02	-0.16
Soil moisture	Fall	- 0.03	0.09	- 0.07	- 0.34 b	- 0.36 a	- 0.26	0.03	- 0.06	- 0.01	- 0.43 a	- 0.28	-0.28
	Win	- 0.25	- 0.24	0.25	0.16	0.13	0.22	-0.1	0.12	0.05	- 0.28	- 0.23	- 0.38 b
	Full year	- 0.58 a	- 0.65 a	- 0.54 a	- 0.62 a	- 0.58 a	- 0.62 a	- 0.59 a	- 0.52 a	- 0.49 a	- 0.47 a	- 0.41 a	- 0.41
	Spr	0.59 a	0.41 a	0.26 b				0.49 a	0.24	0.34 a			
	Sum	0.42 a	0.45 a	0.36 a	0.42 a	0.26	0.18	0.51 a	0.52 a	0.27 b	0.71 a	0.59 a	0.64
Exchange fluxes	Fall	0.37 b	0.35 a	- 0.12	0.42 a	0.26	0.18	0.50 a	0.12	0.09	0.33 b	0.13	0.13
	Win	0.57 a	0.53 a	0.18	0.64 a	- 0.46 a	0.25	0.51 a	0.29	0.47 a	0.65 a	0.38 b	0.3
	Full year	0.64 a	0.65 a	0.55 a	0.72 a	0.64 a	0.67 a	0.63 a	0.51 a	0.36 a	0.77 a	0.69 a	0.65

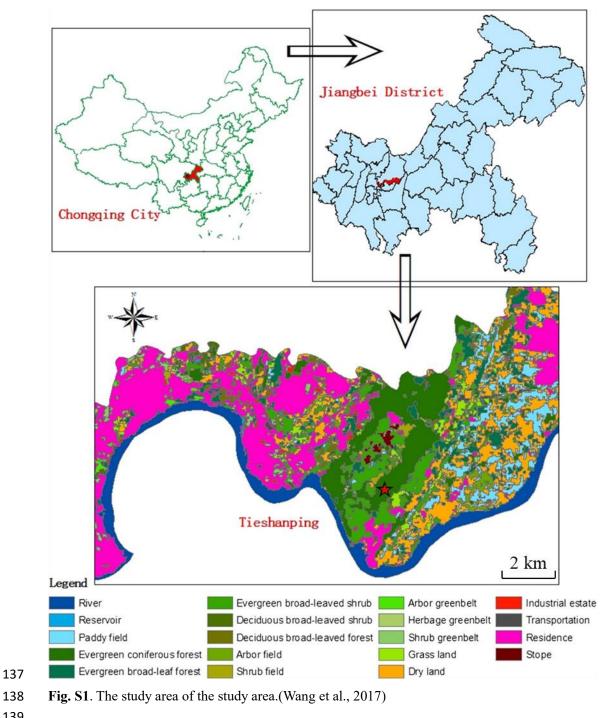
Table S4. Correlation coefficients of soil pore TGM concentrations in experimental plots with soil
 temperature, moisture and soil-air TGM fluxes for the four seasons and throughout the year.

^a Correlation is significant at the 0.01 level (two-tailed); ^b Correlation is significant at the 0.05 level (two-tailed). Note that soil temperature and moisture were directly determined in surface soil by Time Domain
Reflectometry (TDR) with a Stevens water cable tester, not in each layer.

- **113 Figure Captions:**
- 114

Fig. S1. The study area of the study area.(Wang et al., 2017)

- Fig. S2. Correlation between the averaged solar radiation (8: 0-17:00) and daily soil-air Hg flux
 measured as the average of day and night values for the five plots.
- Fig. S3. Effects of precipitation events on soil-air TGM fluxes at the five plots for the four seasonsand annually.
- Fig. S4. Correlation between the soil Hg concentrations ($S_c \pm SD$) and soil-air Hg flux ($F \pm SD$) under the forest canopy. Standard deviations of soil Hg concentrations were obtained from Hg concentrations in the four seasons (n=12). Because fluxes are often controlled by solar
- radiation for bare soils, the correlation analysis above does not include the open field (plot E).
- Fig. S5. Correlation between the soil temperature and daily soil-air Hg flux measured as the averageof day and night values for the five plots.
- Fig. S6. Correlation between the soil moisture and daily soil-air Hg flux measured as the averageof day and night values for the five plots.
- Fig. S7. Correlation between the air TGM concentration and daily soil-air Hg flux measured as theaverage of day and night values for the five plots.
- Fig. S8. Scatterplots of model-predicted and DFC-measured fluxes between soil and air in the
 Masson pine (a), wetland (b), camphor (c) and open field (d) plots.
- Fig. S9. Hg (a) and TOM concentrations (b) with the soil depth at the collection depths of soil poreTGM.
- Fig. S10. Correlation between the gradient of TGM concentrations between soil pore (3 cm) andatmosphere values and soil-air TGM flux at the four plots.
- 136



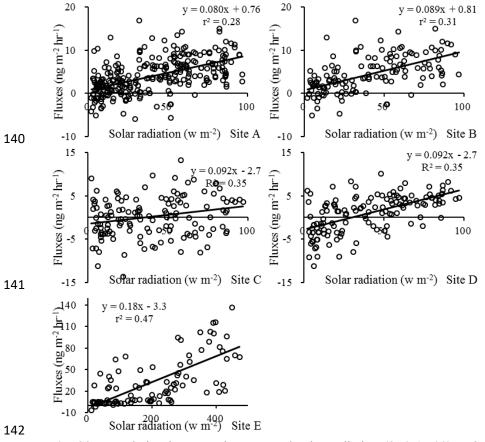
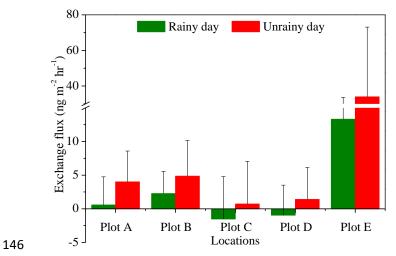
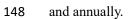


Fig. S2. Correlation between the averaged solar radiation (8: 0-17:00) and daily soil-air Hg flux

- 144 measured as the average of day and night values for the five plots.
- 145



147 Fig. S3. Effects of precipitation events on soil-air TGM fluxes at the five plots for the four seasons



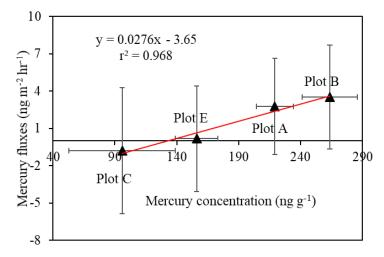
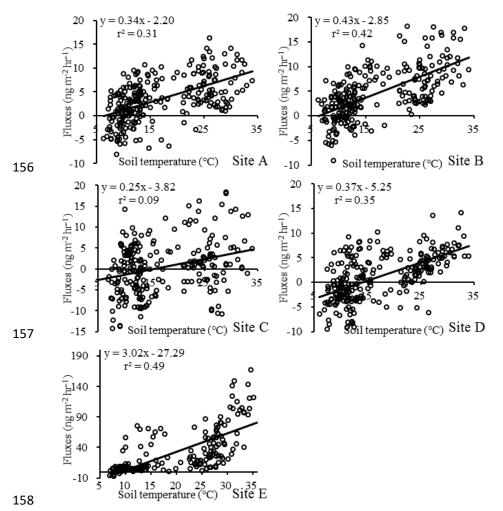


Fig. S4. Correlation between the soil Hg concentrations ($S_c \pm SD$) and soil-air Hg flux ($F \pm SD$) under the forest canopy. Standard deviations of soil Hg concentrations were obtained from Hg concentrations in the four seasons (n=12). Because fluxes are often controlled by solar radiation for bare soils, the correlation analysis above does not include the open field (plot E).

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159 Fig. S5. Correlation between the soil temperature and daily soil-air Hg flux measured as the average

- 160 of day and night values for the five plots.
- 161

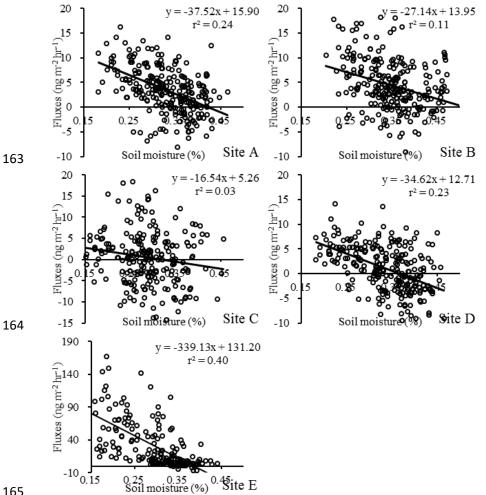
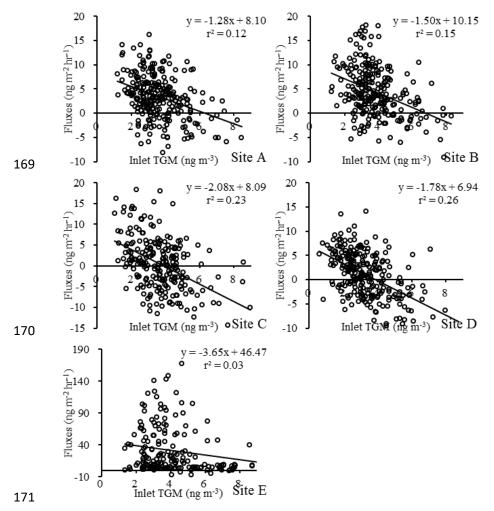


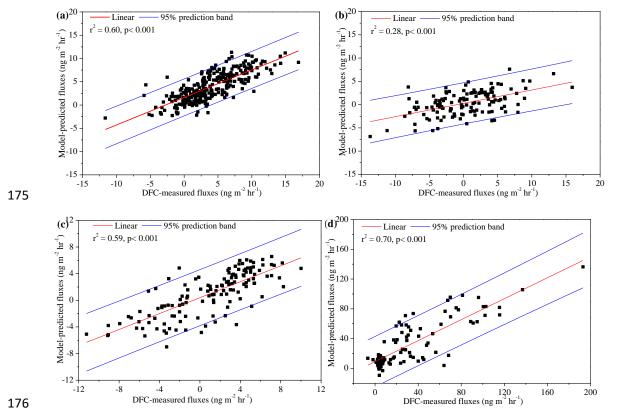
Fig. S6. Correlation between the soil moisture and daily soil-air Hg flux measured as the average

of day and night values for the five plots.

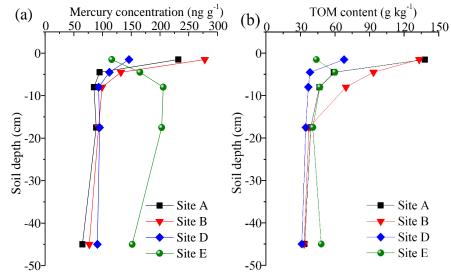


172 Fig. S7. Correlation between the air TGM concentration and daily soil-air Hg flux measured as the

- average of day and night values for the five plots.
- 174



177 Fig. S8. Scatterplots of model-predicted and DFC-measured fluxes between soil and air in the178 Masson pine (a), wetland (b), camphor (c) and open field (d) plots.



181 Fig. S9. Hg (a) and TOM concentrations (b) with the soil depth at the collection depths of soil pore



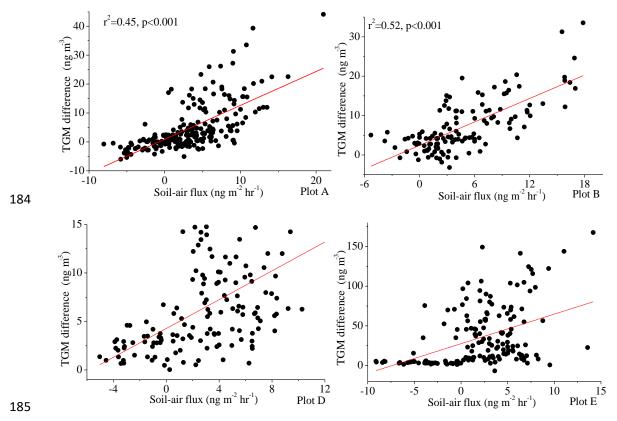


Fig. S10. Correlation between the gradient of TGM concentrations between soil pore (3 cm) andatmosphere values and soil-air TGM flux at the four plots.

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189 **References:**

- Carpi, A., and Lindberg, S. E.: Sunlight-mediated emission of elemental mercury from soil amended with
 municipal sewage sludge, Environmental Science & Technology, 31, 2085-2091, 1997.
- Choi, H.-D., and Holsen, T. M.: Gaseous mercury emissions from unsterilized and sterilized soils: The
 effect of temperature and UV radiation, Environmental pollution, 157, 1673-1678,
 10.1016/j.envpol.2008.12.014, 2009.
- Gbor, P. K., Wen, D., Meng, F., Yang, F., Zhang, B., and Sloan, J. J.: Improved model for mercury
 emission, transport and deposition, Atmospheric Environment, 40, 973-983, 2006.
- Kikuchi, T., Ikemoto, H., Takahashi, K., Hasome, H., and Ueda, H.: Parameterizing soil emission and atmospheric oxidation-reduction in a model of the global biogeochemical cycle of mercury, Environmental Science & Technology, 47, 12266-12274, 2013.
- Lin, C. J., Gustin, M. S., Singhasuk, P., Eckley, C., and Miller, M.: Empirical models for estimating
 mercury flux from soils, Environmental Science & Technology, 44, 8522-8528, 2010.
- Wang, J., Zhang, X., Wang, Z., and Kang, R.: A relative method for measuring nitric oxide (NO) fluxes
 from forest soils, Science of the Total Environment, 574, 544-552, 10.1016/j.scitotenv.2016.09.012,
 204 2017.
- Xiao, Z. F., Munthe, J., Schroeder, W. H., and Lindqvist, O.: Vertical fluxes of volatile mercury over
 forest soil and lake surfaces in Sweden, Tellus Series B-Chemical and Physical Meteorology, 43,
 267-279, 10.1034/j.1600-0889.1991.t01-1-00001.x, 1991.
- Xu, X., Yang, X., Miller, D. R., Helble, J. J., and Carley, R. J.: Formulation of bi-directional atmosphere surface exchanges of elemental mercury, Atmospheric Environment, 33, 4345-4355, 1999.
- Zhou, J., Feng, X., Liu, H., Zhang, H., Fu, X., Bao, Z., Wang, X., and Zhang, Y.: Examination of total
 mercury inputs by precipitation and litterfall in a remote upland forest of Southwestern China,
 Atmospheric Environment, 81, 364-372, 10.1016/j.atmosenv.2013.09.010, 2013.
- Zhou, J., Wang, Z., Zhang, X., and Chen, J.: Distribution and elevated soil pools of mercury in an acidic
 subtropical forest of southwestern China, Environmental pollution, 202, 187-195,
 10.1016/j.envpol.2015.03.021, 2015.
- Zhou, J., Wang, Z., Sun, T., Zhang, H., and Zhang, X.: Mercury in terrestrial forested systems with highly
 elevated mercury deposition in southwestern China: The risk to insects and potential release from
 wildfires, Environmental pollution, 212, 188-196, 10.1016/j.envpol.2016.01.003, 2016.
- 219