



1   **Predictions of the glass transition temperature and viscosity of**  
2   **organic aerosols by volatility distributions**

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5   **Ying Li<sup>1,\*</sup>, Douglas A. Day<sup>2,3</sup>, Harald Stark<sup>2,3,4</sup>, Jose L. Jimenez<sup>2,3</sup> and**  
6   **Manabu Shiraiwa<sup>1,\*</sup>**

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9   [1] Department of Chemistry, University of California, Irvine, CA 92697-2025, USA

10   [2] Cooperative Institute for Research in Environmental Sciences (CIRES), University  
11   of Colorado, Boulder, CO 80309, USA

12   [3] Department of Chemistry, University of Colorado, Boulder, CO 80309, USA

13   [4] Aerodyne Research Inc., Billerica, Massachusetts 01821, USA

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15   \*Correspondence to: Ying Li (yingl47@uci.edu) or Manabu Shiraiwa  
16   (m.shiraiwa@uci.edu)

17



18    **Abstract:**

19    Volatility and viscosity are important properties of organic aerosols (OA), affecting  
20    aerosol processes such as formation, evolution and partitioning of OA. Volatility  
21    distributions of ambient OA particles have often been measured, while viscosity  
22    measurements are scarce. We have previously developed a method to estimate the glass  
23    transition temperature ( $T_g$ ) of an organic compound containing carbon, hydrogen, and  
24    oxygen. Based on analysis of over 2300 organic compounds including oxygenated  
25    organic compounds as well as nitrogen- and sulfur-containing organic compounds, we  
26    extend this method to include nitrogen- and sulfur-containing compounds based on  
27    elemental composition. In addition, parameterizations are developed to predict  $T_g$  as a  
28    function of volatility and the atomic oxygen-to-carbon ratio based on a negative  
29    correlation between  $T_g$  and volatility. The prediction method of  $T_g$  and viscosity is  
30    applied to ambient observations of volatility distributions at eleven field sites. The  
31    predicted  $T_g$  varies mainly from 290 K to 339 K and the predicted viscosities are  
32    consistent with the results of ambient particle phase state measurements in the  
33    southeastern US and the Amazonian rain forest. Reducing the uncertainties in measured  
34    volatility distributions would be helpful to improve predictions of viscosity especially  
35    at low relative humidity. We also predict the  $T_g$  of OA components identified via  
36    positive matrix factorization of aerosol mass spectrometer data. The predicted viscosity  
37    of oxidized OA is consistent with previously reported viscosity of SOA derived from  
38     $\alpha$ -pinene, toluene, isoprene epoxydiol (IEPOX), and of diesel fuel. Comparison of the  
39    predicted viscosity based on the observed volatility distributions with the viscosity  
40    simulated by a chemical transport model implies that missing low volatility compounds  
41    in a global model can lead to underestimation of OA viscosity at some sites. The  
42    relation between volatility and viscosity can be applied in the molecular corridor or  
43    volatility basis set approaches to improve OA simulations in chemical transport models  
44    by consideration of effects of particle viscosity in OA formation and evolution.

45



46 **1. Introduction**

47       Organic aerosols (OA) contribute substantially to the mass loadings of  
48 atmospheric fine particulate matter (Hallquist et al., 2009; Jimenez et al., 2009). OA  
49 formed from various anthropogenic or biogenic precursors have complex  
50 physicochemical properties (Goldstein and Galbally, 2007; Nizkorodov et al., 2011;  
51 Ditto et al., 2018), which makes predictions of their role in air quality, climate and  
52 public health challenging (Kanakidou et al., 2005; Shrivastava et al., 2017). Volatility  
53 and viscosity are important properties of OA, both of which affect important aerosol  
54 processes such as gas–particle partitioning, new particle formation and evolution of size  
55 distribution, heterogeneous reactions, and cloud condensation and ice nucleation  
56 pathways of OA, as summarized in recent review articles (Krieger et al., 2012; Bilde et  
57 al., 2015; Poschl and Shiraiwa, 2015; Knopf et al., 2018; Reid et al., 2018).

58       Recent measurements show that OA can exist in liquid (low dynamic  
59 viscosity  $\eta$ ;  $\eta < 10^2$  Pa s), semi-solid ( $10^2 \leq \eta \leq 10^{12}$  Pa s), and amorphous solid ( $\eta >$   
60  $10^{12}$  Pa s) states (Reid et al., 2018). Even though there are several particle bounce  
61 measurements to infer ambient OA phase state, there are limited ambient measurements  
62 of particle phase state or viscosity depending on temperature ( $T$ ), relative humidity  
63 (RH), and particle composition (Virtanen et al., 2010; O'Brien et al., 2014; Bateman et  
64 al., 2016; Pajunoja et al., 2016; Bateman et al., 2017; Liu et al., 2017; Ditto et al., 2019;  
65 Slade et al., 2019). Viscosity can be directly converted to bulk diffusivity in organic  
66 molecules using the Stokes–Einstein equation, which has been shown to work well for  
67 organic molecules diffusing through low viscous materials (Price et al., 2016;  
68 Chenyakin et al., 2017). This relation is inapplicable for predicting the bulk diffusivity  
69 of water and small molecules and it may also underestimate the diffusivity of organic  
70 molecules in a highly viscous matrix, which can be corrected using a fractional Stokes–  
71 Einstein equation (Price et al., 2016; Evoy et al., 2019).

72       Chemical composition of OA is complex and molecular specificity is often  
73 unavailable, which makes viscosity predictions of ambient OA challenging (Bosse,



74 2005; Song et al., 2016b; Rovelli et al., 2019). Viscosity can be related to the glass  
75 transition temperature ( $T_g$ ), at which phase transition between amorphous solid and  
76 semi-solid states occurs (Koop et al., 2011). Ambient temperature varies through 100  
77 K throughout the troposphere, greatly influencing the viscosity of the mixture. When  
78 the ambient temperature is below  $T_g$ , an amorphous particle behaves as a solid, while a  
79 particle would be semi-solid or liquid when the ambient temperature is above  $T_g$ . OA  
80 particles contain a number of organic compounds and also a variable amount of liquid  
81 water depending on RH, which can act as a plasticizer to reduce  $T_g$ : these mixture  
82 effects can be estimated using the Gordon-Taylor relation (Mikhailov et al., 2009; Koop  
83 et al., 2011; Dette et al., 2014). In addition, ambient OA may often be internally mixed  
84 with inorganic species such as sulfate and nitrate, which would further lower  $T_g$  and  
85 viscosity if they are well-mixed in one phase; when the phase separation occurs, the  
86 inorganic-rich and organic-rich phases may undergo glass transition at different  
87 temperatures (Dette and Koop, 2015).

88 We have recently developed a set of semi-empirical parameterizations using  
89 molar mass ( $M$ ) and atomic O:C ratio (Shiraiwa and Li et al., 2017) or elemental  
90 composition (DeRieux and Li et al., 2018) to predict  $T_g$  for compounds comprised of  
91 carbon, hydrogen, and oxygen (CHO compounds). These parameterizations have been  
92 applied to high-resolution mass spectrometry measurements to estimate viscosity of  
93 organic aerosols (DeRieux and Li et al., 2018; Schum et al., 2018; Ditto et al., 2019)  
94 and coupled into a thermodynamic model (Gervasi et al., 2019). Note that heteroatoms  
95 and the effects of molecular structure and functional groups on  $T_g$  are not considered in  
96 parameterizations of Shiraiwa and Li et al. (2017) and DeRieux and Li et al. (2018).

97 Viscosity of pure compounds has been found to be inversely correlated with  
98 vapor pressure (Thomas et al., 1979). The molecular corridor (Shiraiwa et al., 2014; Li  
99 et al., 2016) based analysis of hundreds of SOA components has shown that compounds  
100 with lower pure compound saturation mass concentration ( $C^0$ ) have higher  $T_g$  (Shiraiwa  
et al., 2017). Rothfuss & Petters (2017) found that there is a similar trend between the



102 sensitivity of viscosity to functional group addition and the sensitivity of vapor pressure  
103 to functional group addition. Measurements of the evaporation kinetics of maleic acid  
104 showed that decreasing particle viscosity leads to a suppression in the effective vapor  
105 pressure of maleic acid (Marshall et al., 2018). Champion et al. (2019) found secondary  
106 organic aerosols (SOA) with higher condensed-phase fractions of extremely low  
107 (ELVOC) and low volatile organic compounds (LVOC) showed an increased viscosity.  
108 Zhang et al. (2019) measured  $T_g$  of isoprene SOA components including isoprene  
109 hydroxy hydroperoxide (ISOPPOOH), isoprene-derived epoxydiols (IEPOX), 2-  
110 methyltetrols, and 2-methyltetrol sulfates (2-MT-OS), observing a tight correlation  
111 between  $T_g$  and vapor pressure.

112 Based on the above evidence showing a close relation between volatility and  
113 viscosity, in this study we develop the parameterizations predicting  $T_g$  as a function of  
114  $C^0$  based on data from over 2000 compounds. Functional group contribution approaches  
115 are often used to predict  $C^0$  (Capouet and Müller, 2006; Pankow and Asher, 2008;  
116 Compernolle et al., 2011; O'Meara et al., 2014), thereby using volatility to predict  $T_g$   
117 would include the molecular structure effect indirectly. The developed  
118 parameterizations are applied to field observations of volatility distributions to predict  
119 viscosity of ambient OA.

120

## 121 **2. Parameterizations**

122 Figure 1a shows a dependence of  $T_g$  on  $C^0$  of 2325 organic compounds  
123 compiled in previous studies with available measured otherwise estimated  $T_g$  and  $C^0$   
124 (Koop et al., 2011; Li et al., 2016; Rothfuss and Petters, 2017; Lessmeier et al., 2018).  
125 In this study “estimated  $T_g$ ” represents the  $T_g$  estimated from the melting temperature  
126 ( $T_m$ ) applying the Boyer-Kauzmann rule of  $T_g = g \cdot T_m$  with  $g = 0.7$  as validated in a  
127 previous study (Koop et al., 2011). The 2325 compounds are classified into four classes  
128 based on chemical composition: CH, CHO, CHON, and CHOS, with the number of  
129 433, 783, 274, and 835, respectively. Detailed information of this dataset is described



130 in the Supplement. Figure 1a shows compounds with lower  $C^0$  have higher  $T_g$  and the  
131  $T_g$  appears to level at around 420 K at  $C^0 < \sim 10^{-10} \mu\text{g m}^{-3}$ . The dependence of  $T_g$  on the  
132 atomic O:C ratio is weaker (Fig. 1a), in agreement with previous studies (Koop et al.,  
133 2011; Shiraiwa et al., 2017). Based on the trend shown in Fig. 1a, we develop a  
134 parameterization (Eq. 1) to predict  $T_g$  as a function of  $C^0$  and the atomic O:C ratio,  
135 which are the parameters used in the two-dimensional VBS (2D-VBS) framework  
136 (Donahue et al., 2011):

137

138 
$$T_g = 289.10 - 16.50 \times \log_{10}(C^0) - 0.29 \times [\log_{10}(C^0)]^2 + 3.23 \times \log_{10}(C^0) (\text{O:C}) \quad (1)$$

139

140 The coefficients in Eq. (1) are obtained by fitting the  $T_g$  of 2325 compounds in Fig. 1a  
141 with multi-linear least squares analysis with 68% prediction and confidence intervals.  
142 Figure 1b shows that the  $T_g$  of those 2325 compounds are predicted well by Eq. (1) as  
143 indicated by a high correlation coefficient ( $R$ ) of 0.92. The average absolute value of  
144 the relative error (AAVRE, Aiken et al., 2007) is 12%.  $T_g$  of individual compounds can  
145 be predicted within  $\pm 31$  K as indicated by the prediction band (dotted lines in Fig. 1b);  
146 however, this uncertainty may be much smaller for multicomponent SOA mixtures  
147 under ideal mixing conditions as indicated in the confidence band (dashed lines, almost  
148 overlapping with the 1:1 line) (Shiraiwa and Li et al., 2017; DeRieux and Li et al., 2018;  
149 Song et al., 2019).

150 Equation (1) is also evaluated with a different set of organic compounds as a  
151 test dataset, including 654 CHO compounds and 212 CHON compounds found in SOA  
152 oxidation products (Shiraiwa et al., 2014). The values of their  $T_g$  are predicted by Eq.  
153 (1) with saturation mass concentrations estimated using the EVAPORATION model  
154 (Compernolle et al., 2011). The predicted  $T_g$  by Eq. (1) are compared with the  $T_g$   
155 estimated from the  $T_m$  applying the Boyer-Kauzmann rule.  $T_m$  values are adopted from  
156 the estimations of the Estimation Programs Interface (EPI) Suite software (version 4.1)  
157 developed by the US Environmental Protection Agency (US EPA, 2015). As shown in



158 Fig. 1(c), Eq. (1) also presents a good performance for predicting the  $T_g$  of these SOA  
159 components with  $R = 0.96$  and AAVRE = 6 %. Note that Eq. (1) was developed based  
160 on the training dataset containing a large number of compounds with measured  $C^0$   
161 (Table S1 in Supplement) and aimed to be applied in the 2D-VBS framework to predict  
162 the viscosity of SOA mixtures. For pure organic compounds with known molecular  
163 structure, viscosity can be predicted by group contribution approaches (Cao et al., 1993;  
164 Bosse, 2005; Song et al., 2016b; Gervasi et al., 2019; Rovelli et al., 2019).

165 We also develop a parameterization (Eq. 2) predicting  $T_g$  as a function of  $C^0$   
166 solely, which could be applied to the information available with the one-dimensional  
167 VBS (1D-VBS) framework (Donahue et al., 2006), and can be used when the O:C ratio  
168 is not available in measurements.

169

170 
$$T_g = 288.70 - 15.33 \times \log_{10}(C^0) - 0.33 \times [\log_{10}(C^0)]^2 \quad (2)$$

171

172 The coefficients in Eq. (2) are obtained following the procedures developing Eq. (1)  
173 and the same training dataset is used. Figure S1 shows that Eq. (2) gives very similar  
174 predictions as Eq. (1) particularly for the compounds with low O:C ratio. As Eq. (1)  
175 and (2) are developed based on the compounds with their  $C^0$  higher than  $\sim 10^{-20} \mu\text{g m}^{-3}$ ,  
176 Eqs. (1–2) may not be applicable for compounds with  $C^0 < \sim 10^{-20} \mu\text{g m}^{-3}$  (Fig. 1a).

177 We recently developed a parameterization (Eq. 3) predicting  $T_g$  as a function  
178 of the number of carbon ( $n_C$ ), hydrogen ( $n_H$ ), and oxygen ( $n_O$ ) atoms (DeRieux and Li  
179 et al., 2018), similar to the formulation used to predict  $C^0$  (Donahue et al., 2011; Li et  
180 al., 2016).

181

182 
$$T_g = (n_C^0 + \ln(n_C)) b_C + \ln(n_H) b_H + \ln(n_C) \ln(n_H) b_{CH} + \ln(n_O) b_O + \ln(n_C) \ln(n_O) b_{CO} \quad (3)$$

183

184 Values of the coefficients [ $n_C^0$ ,  $b_C$ ,  $b_H$ ,  $b_{CH}$ ,  $b_O$ , and  $b_{CO}$ ] are [1.96, 61.99, -113.33, 28.74,  
185 0, 0] for CH compounds and [12.13, 10.95, -41.82, 21.61, 118.96, -24.38] for CHO



186 compounds. We broaden the parameterizations for CH and CHO compounds (Eq. 3) to  
187 the following equations applicable to CHON (Eq. 4) and CHOS compounds (Eq. 5):  
188

$$189 T_g = (n_C^0 + \ln(n_C)) b_C + \ln(n_O) b_O + \ln(n_N) b_N + \ln(n_C) \ln(n_O) b_{CO} + \ln(n_C) \ln(n_N) b_{CN} + \ln(n_O) \\ 190 \ln(n_N) b_{ON} \quad (4)$$

$$191 T_g = (n_C^0 + \ln(n_C)) b_C + \ln(n_O) b_O + \ln(n_S) b_S + \ln(n_C) \ln(n_O) b_{CO} + \ln(n_C) \ln(n_S) b_{CS} + \ln(n_O) \\ 192 \ln(n_S) b_{OS} \quad (5)$$

193  
194 Values of the coefficients [ $n_C^0$ ,  $b_C$ ,  $b_O$ ,  $b_N$ ,  $b_{CO}$ ,  $b_{CN}$  and  $b_{ON}$ ] in Eq. (4) are [5.34, 31.53,  
195 -7.06, 134.96, 6.54, -34.36, -15.35] and [ $n_C^0$ ,  $b_C$ ,  $b_O$ ,  $b_S$ ,  $b_{CO}$ ,  $b_{CS}$  and  $b_{OS}$ ] in Eq. (5) are  
196 [1.12, 68.41, 64.95, 35.77, -12.32, -9.85, 13.80], respectively. These values are obtained  
197 by fitting the  $T_g$  of CHON and CHOS compounds included in the training dataset (Fig.  
198 1a, Table S1) with multi-linear least squares analysis. In the training dataset there are  
199 35 CHON compounds with measured  $T_g$  available (Koop et al., 2011) and  $T_g$  of other  
200 239 CHON compounds are estimated by the Boyer-Kauzmann rule using their  
201 measured  $T_m$  (Supplement). Figure 2(a) shows a fair agreement between the predicted  
202  $T_g$  using Eq. (4) and the measured otherwise estimated  $T_g$  with  $R$  of 0.55 and relatively  
203 large AAVRE of 16 %. Figure 2 (b) shows a better prediction performance with  $R$  of  
204 0.83 and AAVRE of 9 % for 212 CHON compounds included in the test dataset with  
205 estimated  $T_g$  by the Boyer-Kauzmann rule using the EPI-estimated  $T_m$ . Measured  $T_g$   
206 values for CHOS compounds are sparse (Zhang et al., 2019) and the  $T_g$  of the 835  
207 CHOS compounds shown in Fig. 2c are estimated by the Boyer-Kauzmann rule using  
208 the EPI-estimated  $T_m$ . Figure 2 (c) shows that Eq. 5 performs well for those CHOS  
209 compounds with  $R$  of 0.87 and AAVRE of 8 %.

210 Comparison of  $T_g$  predicted by Eqs. (4, 5) with  $T_g$  predicted by Eq. (1) shows  
211 that the agreement between the two sets of parameterizations for nitrogen- and sulfur-  
212 containing compounds is not as good as that for CHO compounds (Fig. S2), indicating  
213 that there are limitations of predicting  $T_g$  by the elemental composition for nitrogen-



214 and sulfur-containing compounds with complex elemental composition and molecular  
215 structures. As volatility depends significantly on functional groups contained in a  
216 molecule (Pankow and Asher, 2008; Compernolle et al., 2011), predicting  $T_g$  by  
217 volatility (Eq. 1) indirectly incorporates the effects of molecular structure. Considering  
218 Eqs. (4, 5) are useful in high-resolution mass spectrometry measurements (DeRieux  
219 and Li et al., 2018; Schum et al., 2018), future experiments measuring more  $T_g$  data for  
220 nitrogen- and sulfur-containing organics would help improve the  $T_g$  parameterizations.

221 Recently Zhang et al. (2019) developed a semi-empirical parameterization  
222 (Eq. 6) using vapor pressure ( $p_0$  in atm) to predict  $T_g$  based on measured  $T_g$  of 11 SOA  
223 compounds:

224

$$225 \quad T_g = 480.1 - \frac{54395}{(\log_{10}(p_0) - 1.7929)^2 + 116.49} \quad (6)$$

226

227  $p_0$  can be converted to  $C^0$  via  $C^0 = (10^6 M p_0)/(RT)$ , where  $R$  is the ideal gas constant ( $R$   
228  $= 8.2 \times 10^{-5}$  m<sup>3</sup> atm mol<sup>-1</sup> K<sup>-1</sup>),  $M$  is the molar mass (g mol<sup>-1</sup>), and  $T$  is the temperature  
229 (K). Figure 3 compares the measured  $T_g$  (Koop et al., 2011; Rothfuss and Petters, 2017;  
230 Lessmeier et al., 2018; Zhang et al., 2019) to  $T_g$  predicted by (a)  $C^0$  and the atomic O:C  
231 (Eq. 1), (b) elemental composition (Eqs. 3-5), and (c) Eq. (6) by Zhang et al. (2019).  
232 While all three methods perform reasonably well, the predictions using elemental  
233 composition (Eqs. 3-5) show better performance (Fig. 3b) with  $R$  of 0.93 and AAVRE  
234 of 11 %, respectively.

235 The prediction performance is influenced by the training dataset used  
236 developing parameterizations of  $T_g$ . The compounds shown in Fig. 3 contain mostly  
237 carboxylic acid and hydroxyl functional groups (Koop et al., 2011; Rothfuss and  
238 Petters, 2017) and are included in the training dataset used developing Eq. (1) and Eqs.  
239 (3-5). The training dataset used in Zhang et al. (2019) included 11 organic compounds,  
240 and their parameterization predicted  $T_g$  of isoprene SOA very well (Zhang et al., 2019),  
241 while underpredicting some low- $T_g$  compounds (Fig. 3c). For compounds with their



242 measured  $T_g$  higher than 200 K, predictions by Zhang et al. (2019) show good  
243 performance and are consistent with the predictions given by Eq. (1). Predicted  $T_g$  of  
244 2-MT-OS using the three methods are 297 K (Eq. 1), 275 K (Eq. 5) and 280 K (Eq. 6),  
245 comparable with the measured  $T_g$  of  $276 \pm 15$  K (Zhang et al., 2019).

246 Note that predictions using elemental composition (Eq. 3) overestimate the  $T_g$   
247 of phthalate compounds (the star markers in Fig. 3). For instance, the observed  $T_g$  of  
248 dioctyl phthalate is 194 K (Zhang et al., 2018), while the prediction is higher than 300  
249 K (Fig. 3b). The reason is that ester is not an effective functional group to increase  
250 viscosity compared to carboxylic acid and hydroxyl (Rothfuss and Petters, 2017).  
251 Parameterizations using volatility (Eqs. 1 and 6) improve the predicted  $T_g$  of phthalate  
252 compounds (Fig. 3a, c). Figure A1 shows comparing to the predictions using Eq. (6)  
253 provided in Zhang et al. (2019), predictions by  $C^0$  and the atomic O:C (Eq. 1) and  
254 elemental composition (Eq. 3) agree better with the  $T_g$  estimated from the Boyer-  
255 Kauzmann rule. Future experiments measuring more  $T_g$  of SOA components would  
256 help verify the  $T_g$  predictions by different parameterizations.

257

### 258 **3. Application in field observations**

259  $T_g$  of ambient OA are estimated by applying the developed parameterizations  
260 to field observations of volatility distributions. The term volatility refers to the effective  
261 saturation mass concentration ( $C^*$ ) and we assume ideal thermodynamic mixing in  
262 which case  $C^*$  is equal to  $C^0$  (Donahue et al., 2011).  $T_g$  are then placed into the 2D-VBS  
263 framework of O:C ratio vs. volatility as shown in Fig. 4. The isolines in Fig. 4 show the  
264 predicted  $T_g$  by Eq. (1), showing that  $T_g$  would be below  $\sim 250$  K for intermediate  
265 volatility organic compounds (IVOC;  $300 < C^0 < 3 \times 10^6 \mu\text{g m}^{-3}$ ), from  $\sim 260$  K to 290  
266 K for semi-volatile organic compounds (SVOC;  $0.3 < C^0 < 300 \mu\text{g m}^{-3}$ ), and higher  
267 than 300 K for low-volatile organic compounds (LVOC;  $3 \times 10^{-4} < C^0 < 0.3 \mu\text{g m}^{-3}$ ) and  
268 extremely low-volatile organic compounds (ELVOC;  $C^0 < 3 \times 10^{-4} \mu\text{g m}^{-3}$ ).  $T_g$  increases  
269 as the O:C ratio increases for SVOC and IVOC, consistent with previous studies (Koop



270 et al., 2011; Saukko et al., 2012; Berkemeier et al., 2014).  $T_g$  slightly decreases as the  
271 O:C ratio increases for LVOC and ELVOC compounds, which might be due to the  
272 uncertainties in Eq. (1) which is derived from a dataset containing fewer LVOC and  
273 ELVOC compounds as shown in Fig. 1a, which exhibits lower  $T_g$  with higher O:C.

274

### 275 **3.1 Southern Oxidant and Aerosol Study (SOAS)**

276 The volatility of ambient OA was measured during the SOAS campaign  
277 which took place in the southeastern United States (Centreville, Alabama) in summer  
278 2013 (Carlton et al., 2018).  $T_g$  in each volatility bin ( $T_{g,i}$ ) is calculated by Eq. (1) under  
279 the assumption of ideal thermodynamic mixing and the measured O:C is used (Xu et  
280 al., 2015). The glass transition temperatures of organic aerosols under dry conditions  
281 ( $T_{g,org}$ ) are calculated by the Gordon-Taylor equation assuming the Gordon-Taylor  
282 constant ( $k_{GT}$ ) of 1:  $T_{g,org} = \sum_i w_i T_{g,i}$  (Dette et al., 2014), where  $w_i$  is the mass  
283 fraction in the particle phase for each volatility bin (Table S2). To place  $T_{g,org}$  into the  
284 2D-VBS framework, the average  $\log_{10}(C^*)$  is calculated by  $\sum_i w_i \log_{10}(C_i^*)$   
285 (Kostenidou et al., 2018).

286 Figure 4 shows that  $T_{g,org}$  of total OA (TOA) ranges from 232 K to 334 K,  
287 depending on volatility distributions measured by different methods. Stark et al. (2017)  
288 used three methods to derive volatility distributions applying the measurements of  
289 organic acids (which comprised half of the total OA; Yatavelli et al., 2015) from a high-  
290 resolution chemical ionization time-of-flight mass spectrometer (HRToF-CIMS)  
291 equipped with a filter inlet for gases and aerosols (FIGAERO, Lopez-Hilfiker et al.,  
292 2014; Thomson et al., 2017). The “Formulas” method used the SIMPOL group  
293 contribution method (Pankow and Asher, 2008) to calculate vapor pressures from the  
294 composition of the identified ions. Many of the detected species can be thermal  
295 decomposition products rather than actual SOA molecules (Stark et al., 2015; Stark et  
296 al., 2017), which can lead to overestimations of volatilities mainly in the IVOC range,  
297 resulting in the unlikely low  $T_{g,org}$  of 232 K. The “Partitioning” method used the



298 measured particle-phase mass fractions of each species to estimate  $C^*$  based on the  
299 partitioning theory (Pankow, 1994). The estimated  $C^*$  is distributed mainly in the  
300 SVOC range (Stark et al., 2017), leading to a  $T_{g,\text{org}}$  of 279 K (Fig. 4). This value is very  
301 close to the  $T_{g,\text{org}}$  (281 K) simulated by a global chemical transport model EMAC-  
302 ORACLE (Shiraiwa et al., 2017). However, Stark et al. (2017) note that the  
303 partitioning-based volatility distribution is likely too high due to an artifact of signal-  
304 to-noise limitations, confining the  $C^*$  characterizable by the partitioning method to a  
305 relatively narrow range centered around the ambient OA concentration (by definition,  
306 the semi-volatile range). In the thermogram method,  $C^*$  at 298 K is estimated from the  
307 desorption temperature after calibration with known species (Faulhaber et al., 2009).  
308 This method results in 93% of OA mass distributed in the LVOC and ELVOC (Stark  
309 et al., 2017), and a high  $T_{g,\text{org}}$  of 330 K is predicted (Fig. 4). These analyses indicate  
310 that the volatility distributions derived from different methods, even when based on the  
311 same measurements, significantly affect the predicted  $T_{g,\text{org}}$ .

312 Among these three methods using CIMS measurements of organic acids,  
313 Stark et al. (2017) note that the result from the thermogram method is more consistent  
314 with those measured by an aerosol mass spectrometer (AMS) with a thermo denuder  
315 (TD), which also applied the thermogram method to estimate the  $C^*$  distributions (Hu  
316 et al., 2016). Saha et al. (2017) applied an evaporation kinetic model (Lee et al., 2011)  
317 based on the VBS approach to extract the  $C^*$  distributions, and the effects of enthalpy  
318 of vaporization ( $\Delta H_{\text{vap}}$ ) and accommodation coefficient ( $\alpha$ ) are considered, resulting in  
319 the estimated  $T_{g,\text{org}}$  of 313 K. This study retrieved  $\alpha \sim 0.5$ , which is consistent with recent  
320 observations (Krechmer et al., 2017; Liu et al., 2019). In summary,  $T_{g,\text{org}}$  during the  
321 SOAS campaign span the range of 313 – 330 K.

322 Figure 4 also shows  $T_{g,\text{org}}$  of OA components identified via positive matrix  
323 factorization (PMF) of the aerosol mass spectrometer data. Isoprene-epoxydiols-  
324 derived SOA (IEPOX-SOA) has  $T_{g,\text{org}}$  of 345 K with very low volatility with the  
325 average  $C^*$  lower than  $10^{-4} \mu\text{g m}^{-3}$  (Hu et al., 2016; Lopez-Hilfiker et al., 2016; D'Ambro



326 et al., 2019), which may be due to substantial formation of organosulfates and other  
327 oligomers (Lin et al., 2012; Hu et al., 2015; Riva et al., 2019). The predicted  $T_{g,\text{org}}$  of  
328 IEPOX-SOA is higher than previously reported  $T_{g,\text{org}}$  of 263 - 293 K for monoterpene-  
329 derived ( $\alpha$ -pinene,  $\Delta^3$ -carene, myrcene, limonene and ocimene) SOA (Petters et al.,  
330 2019).

331 The phase state of aerosol particles strongly depends on their water content  
332 (Mikhailov et al., 2009; Koop et al., 2011). Under humid conditions,  $T_g$  of organic-  
333 water mixtures at given RH can be estimated considering hygroscopic growth  
334 combined with the Gordon-Taylor equation. The effective hygroscopicity parameter ( $\kappa$ )  
335 (Petters and Kreidenweis, 2007) of TOA during the SOAS is measured as 0.14 (Cerully  
336 et al., 2015). The Gordon-Taylor constant for organic-water mixtures is suggested to be  
337 2.5 (Zobrist et al., 2008; Koop et al., 2011). Using the  $T_g$  of organic-water mixtures, the  
338 temperature-dependence of viscosity can be calculated applying the Vogel-Tamman-  
339 Fulcher (VTF) equation (Angell, 1991; Rothfuss and Petters, 2017; DeRieux and Li et  
340 al., 2018). Figure 5(a) shows the predicted viscosity of total OA at different RH.  $T$  is  
341 adopted as 298 K, the average value during the SOAS campaign (Hu et al., 2016). The  
342 characteristic timescale of mass transport and mixing by molecular diffusion ( $\tau_{\text{mix}}$ ) is  
343 also calculated:  $\tau_{\text{mix}} = d_p^2 / (4\pi^2 D_b)$  (Seinfeld and Pandis, 2006), where  $d_p$  is the particle  
344 diameter and the bulk diffusion coefficient  $D_b$  is calculated from the predicted viscosity  
345 by the fractional Stokes-Einstein relation (Evoy et al., 2019). We assume the radius of  
346 the diffusing molecule of  $10^{-10}$  m and the particle diameter of 200 nm (Shiraiwa et al.,  
347 2011). Note that these estimated timescales represent rough estimations, as molecular  
348 interactions in complex mixtures are not considered.

349 The viscosity of TOA at RH of 83% (average RH during SOAS) is predicted  
350 to be less than  $10^2$  Pa s with  $\tau_{\text{mix}}$  less than 1 s, which is consistent with the particle  
351 bounce measurements which suggested organic-dominated particles were mostly liquid  
352 during the SOAS campaign (Pajunoja et al., 2016). When RH was below ~50% in the  
353 sampling inlet, the particles were found to adopt a semi-solid state (Pajunoja et al.,



354 2016), which agrees with the predicted viscosity of  $10^7 - 10^{11}$  Pa s and  $\tau_{\text{mix}}$  can be higher  
355 than 1 hour at RH of 50% (Fig. 5a).

356 Figure 5(b) shows diurnal variations of predicted viscosity of total OA using  
357 measured  $T$  and RH during the SOAS campaign (Hu et al., 2016). During 10:00 – 20:00  
358 when  $\text{RH} < 70\%$  and  $T > 298$  K, three simulations using different  $T_{g,\text{org}}$  values predict  
359 that total OA occur as semi-solid with the predicted viscosity of  $10^2 - 10^7$  Pa s and the  
360 mixing times less than 1 hour. Particles are predicted to have low viscosity of  $< 1$  Pa s  
361 adopting a liquid phase during nighttime. The lowest viscosity occurs around 5:00 –  
362 6:00 with  $\text{RH} > 95\%$ . Here we did not consider the effects of the diurnal variations of  
363 volatility distributions, as they did not vary dramatically over the campaign period  
364 (Saha et al., 2017). Besides  $T$  and RH, diurnal variation of ambient aerosol phase state  
365 also depends on particle chemical composition and mixing states. Organic particles in  
366 Amazon were found to be more viscous at night than the daytime due to the influence  
367 of biomass burning that may form non-liquid particles (Bateman et al., 2017). Particles  
368 in a mixed forest in northern Michigan were also found more viscous at night despite  
369 higher RH than the daytime, due to the formation of high molar mass organic  
370 compounds and smaller inorganic sulfate mass fractions (Slade et al., 2019). Phase state  
371 measurements during daytime and nighttime at Atlanta suggested that the ambient  
372 particle phase state was influenced by OA composition, the presence of inorganic ions,  
373 aerosol liquid water and particle mixing state (Ditto et al., 2019).

374

### 375 **3.2 $T_{g,\text{org}}$ at 11 global sites**

376 Figure 6 summarizes  $T_{g,\text{org}}$  at 11 sites where the measured volatility  
377 distributions with volatility bins of four or more are available (Table S2). We did not  
378 include the data with narrower volatility ranges which may not correctly characterize  
379 the properties of ambient SOA (Bilde et al., 2015), and thus may not be appropriate for  
380 estimating fundamental particle properties and volatility distributions without  
381 considering realistically low  $C^*$  bins would result in too low  $T_g$ . Note that a narrow VBS



382 may still be useful for efficiency in 3-dimentional chemical transport models for SOA  
383 evaporation and condensation under the narrow range of ambient temperature  
384 variations (Kostenidou et al., 2018).  $T_{g,org}$  is color-coded in the 2D-VBS framework of  
385 O:C vs.  $\log_{10}C^*$  in the panel (a), whereas the panel (b) shows  $T_{g,org}$  vs.  $\log_{10}C^*$  with  
386 markers color-coded with the O:C ratio.  $T_{g,org}$  of total OA (TOA) varies from 290 K to  
387 339 K. The lower  $T_{g,org}$  occurs at Beijing, China in June 2018 (Xu et al., 2019). OA in  
388 Beijing was found to be overall more volatile with the particle-phase semi-volatile  
389 fraction of 63%, leading to a lower  $T_{g,org}$ . The predicted  $T_{g,org}$  of total OA at numerous  
390 other sites range between 300 K and 320 K, including Paris (Paciga et al., 2016),  
391 Mexico city (Cappa and Jimenez, 2010), Centreville (Hu et al., 2016; Saha et al., 2017;  
392 Stark et al., 2017), Raleigh (Saha et al., 2017), and Durham (Saha et al., 2018) in  
393 southeastern US. The  $T_{g,org}$  (316 K) is higher at 220 m from a highway in Durham than  
394 the  $T_{g,org}$  (309 K) at 10 m downwind distance due to the dilution and mixing of traffic-  
395 sourced particles with background air and evaporation of semi-volatile species during  
396 downwind transport (Saha et al., 2018).  $T_{g,org}$  is predicted to be higher (>320 K) at the  
397 sites in Athens, Pasadena, Colorado Rocky Mountain and Amazon (Fig. 6).

398 At the same site in Mexico city or Athens, Figure 6 shows that the  $T_{g,org}$  is  
399 lowest for the BBOA factor (287 K – 305 K) followed by HOA (290 K – 331 K) and  
400 OOA (> 350 K for the MO-OOA in Mexico city and Paris), which follows the trend of  
401 the decrease in the average volatility of these OA factors. Cooking OA (COA) has  
402 comparable  $T_{g,org}$  (290 K – 324 K) with HOA. Marine OA (MOA) was characterized  
403 with high volatility containing around 60% SVOC, leading to the  $T_{g,org}$  of 295 K in  
404 summer in Paris.

405 Figure 7 shows the OA viscosity variation of OA components against RH.  
406 The hygroscopic growth is considered based on hygroscopicity ( $\kappa$ ), which is estimated  
407 as a function of the O:C ratio (Lambe et al., 2011) when  $\kappa$  was not measured (Table  
408 S2). Figure 7a shows OA factors with low O:C ratio, i.e., HOA, COA and BBOA, occur  
409 as liquid only when RH is very high (> ~80 %), as their  $\kappa$  is estimated to be low (<



410 0.08). HOA, COA and BBOA in Athens and COA in Paris can undergo glass transition  
411 (i.e., viscosity reaches  $10^{12}$  Pa s) at RH between 25 % and 68 %. OA factors with higher  
412 O:C ratio and hygroscopicity including LO-OOA, MO-OOA, IEPOX SOA and  
413 isoprene-OA tend to become liquid (viscosity  $< 10^2$  Pa s) at intermediate RH (Fig. 7b).

414 The predicted ambient viscosity is compared with the experimental observed  
415 viscosity of SOA formed from isoprene (Song et al., 2015),  $\alpha$ -pinene (Abramson et al.,  
416 2013; Renbaum-Wolff et al., 2013; Kidd et al., 2014; Pajunoja et al., 2014; Bateman et  
417 al., 2015; Zhang et al., 2015; Grayson et al., 2016; Petters et al., 2019), toluene (Song  
418 et al., 2016a) and diesel fuel (Song et al., 2019). The majority of experimental values  
419 are well bounded by the predicted viscosity of OOA, represented by the pink shaded  
420 area. One exception is the measured viscosity of isoprene SOA is lower than the  
421 predicted viscosity of IEPOX SOA at low RH (<30 %). One possible reason is that the  
422 isoprene SOA in experiments was formed with high oxidant concentrations with short  
423 reaction time in an oxidation flow reactor in the absence of inorganic seed particles  
424 (OFR) (Song et al., 2015). In ambient environments heterogeneous reactions with  
425 acidic sulfate particles forming oligomers are suggested to be an important pathway  
426 (Surratt et al., 2010; Lin et al., 2013; Hu et al., 2015, 2016). These particle-phase  
427 organosulfates may contribute to a higher viscosity, as indicated by the predicted  
428 viscosity of IEPOX-derived organosulfate mixtures with their  $T_{g,org}$  estimated to be 313  
429 K (Riva et al., 2019). Another reason could be the mass concentrations ( $100 \sim 1000 \mu\text{g}$   
430  $\text{m}^{-3}$ ) of isoprene SOA generated in the OFR (Song et al., 2015) are much higher than  
431 the ambient OA concentrations ( $5 \mu\text{g m}^{-3}$ , Stark et al., 2017) observed during the SOAS  
432 campaign. Higher mass concentrations can lead to lower viscosity, as more semi-  
433 volatile compounds can partition into the particle phase (Pankow, 1994; Donahue et al.,  
434 2006; Grayson et al., 2016; Jain et al., 2018; Champion et al., 2019).  
435

436 **4. Implications on SOA simulations in chemical transport models**



437 Shiraiwa et al. (2017) simulated the global distribution of annual averages of  
438 SOA phase state using the chemical transport model EMAC (Jöckel et al., 2006)  
439 coupled with the organic aerosol module ORACLE (Tsimpidi et al., 2014). ORACLE  
440 uses the 1D-VBS framework with four  $C^*$  bins ( $1, 10, 10^2$ , and  $10^3 \mu\text{g m}^{-3}$ ). To estimate  
441  $T_g$  the values of molar mass and O:C ratio were assigned for each volatility bin based  
442 on molecular corridors (Shiraiwa et al., 2014). Note that the molar mass assigned for  
443 the volatility bin of  $1 \mu\text{g m}^{-3}$  was assumed to have relatively high molar mass to partially  
444 compensate for the fact that ORACLE does not consider lower volatility bins with  
445 higher molar mass. Global distributions of  $T_g / T$  presented in Shiraiwa et al. (2017) is  
446 converted to viscosity using the VTF equation, as shown in Fig. 8. Figure 8 also  
447 includes the viscosity of total OA at 11 sites by applying the model simulated 5 years'  
448 average  $T$  and RH with  $\kappa$  assumed to be 0.1 (Pringle et al., 2010). It shows the predicted  
449 viscosity at 11 sites (markers) generally agree well with the global simulations: the  
450 amorphous solid or semi-solid phase occurs over relatively dry lands, including western  
451 US, Mexico city, and Beijing; the lower viscosity occurs in southeastern US and Paris.

452 The global simulations show that the particles are liquid in the Amazon,  
453 while they occur as semi-solid in our predictions based on measured volatility  
454 distributions. The reason of this disagreement may be mainly due to the substantial  
455 fraction of low volatility compounds observed in ambient measurements largely  
456 missing from global simulations. Hu et al. (2016) observed that 90% of OA have  
457 volatilities lower than  $1 \mu\text{g m}^{-3}$ , which is the lowest  $C^*$  bin in the global simulations.  
458 The ambient phase state measurements show that for background conditions of the  
459 Amazonian tropical forest, particles are mostly liquid, while with the anthropogenic  
460 influence of biomass burning, they occur as a nonliquid phase (Bateman et al., 2016;  
461 Bateman et al., 2017). The volatility distributions were measured in the dry season  
462 heavily influenced by biomass burning (Hu et al., 2016), which can lead to the higher  
463 predicted viscosity. A similar case is observed in Athens, that our predictions indicate



464 the glassy phase state while the global model predicts the occurrence of a semi-solid  
465 phase.

466 Most of the current chemical transport models treat particles as liquid or  
467 homogeneously well-mixed without considering particle-phase diffusion limitations  
468 (Pankow, 1994), which can lead to bias in simulations of SOA mass concentrations and  
469 evolution of size distributions (Shiraiwa and Seinfeld, 2012; Zaveri et al., 2018). The  
470 SOA simulations applying the VBS framework have not included the effects of  
471 viscosity on SOA formation and evolution. When the gas-particle partitioning is in the  
472 bulk diffusion-limited regime, kinetic treatments of SOA partitioning may need to be  
473 applied (Perraud et al., 2012; Liu et al., 2016; Yli-Juuti et al., 2017; Li and Shiraiwa,  
474 2019). Some chamber experiments probing the mixing timescales of SOA particles did  
475 not observe significant kinetic limitations at moderate and high RH under room  
476 temperature (Loza et al., 2013; Ye et al., 2018), warranting further investigations on the  
477 degree of kinetic limitations in ambient tropospheric conditions. The method developed  
478 in this study can be applied in the VBS or the molecular corridor-based approach to  
479 improve OA simulations in chemical transport models (Pye et al., 2017; Schmedding  
480 et al., 2019). Several important aspects should be further explored in future studies. The  
481 interplay of diffusion limitations and phase separation impacts heterogeneous and  
482 multiphase chemistry (Vander Wall et al., 2018; DeRieux et al., 2019; Zhou et al., 2019)  
483 and gas-particle partitioning (Zuend and Seinfeld, 2012; Freedman, 2017; Pye et al.,  
484 2017; Gorkowski et al., 2019a). The particle morphology and the degree of non-ideal  
485 mixing and liquid-liquid phase separation can evolve upon atmospheric aging  
486 (Gorkowski et al., 2019b). These aspects may also need to be considered for better  
487 representation of organic aerosols in future studies.

488

489 **Author contribution.**

490 YL, JLJ and MS designed the research. YL developed the parameterizations. DAD, HS  
491 and JLJ provided measured volatility distributions for the SOAS campaign. YL and MS



492 wrote the manuscript. All authors discussed the results and contributed to manuscript  
493 editing.

494

495 **Data availability.** The data used in this study is available in the supplement.

496

497 **Competing interests.** The authors declare that they have no conflict of interest.

498

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506

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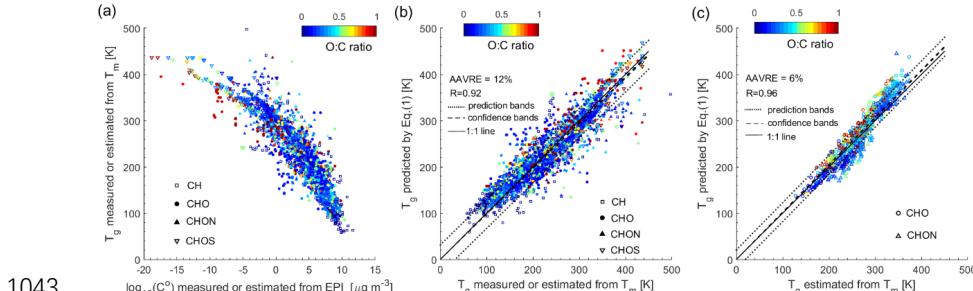
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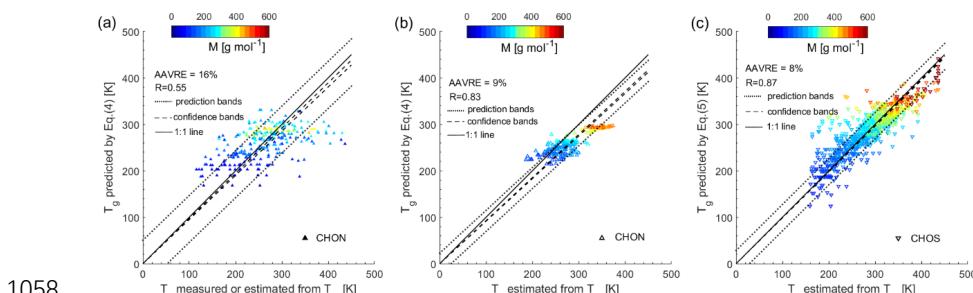
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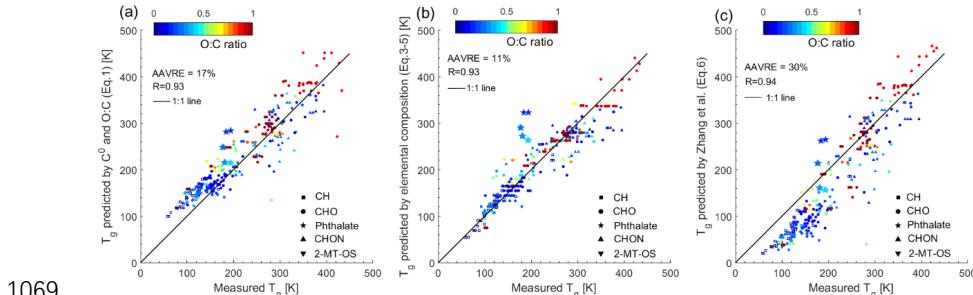
1044 **Figure 1.** (a)  $T_g$  of organic compounds as measured and estimated by the melting  
 1045 temperature ( $T_m$ ) with the Boyer-Kauzmann rule plotted against pure compound  
 1046 saturation mass concentration ( $C^0$ ) as measured and estimated from the EPI suite (see  
 1047 supplement for details of the dataset). (b) Predicted  $T_g$  for compounds shown in (a)  
 1048 using a parameterization (Eq. 1) developed in this study compared to measured and  
 1049 estimated  $T_g$  from  $T_m$  with the Boyer-Kauzmann rule. (c) Predicted  $T_g$  for SOA  
 1050 components (Shiraiwa et al., 2014) using Eq. (1) plotted against estimated  $T_g$  from  $T_m$   
 1051 with the Boyer-Kauzmann rule. The markers are color-coded by the atomic O:C ratio.  
 1052 Organic compounds include CH compounds (squares), CHO compounds (circles),  
 1053 CHON compounds (triangles), and CHOS compounds (inverted triangles). The  
 1054 correlation coefficient ( $R$ ) and the average absolute value of the relative error (AAVRE)  
 1055 are shown. The solid line shows the 1:1 line and the dashed and dotted lines show 68%  
 1056 confidence and prediction bands, respectively.

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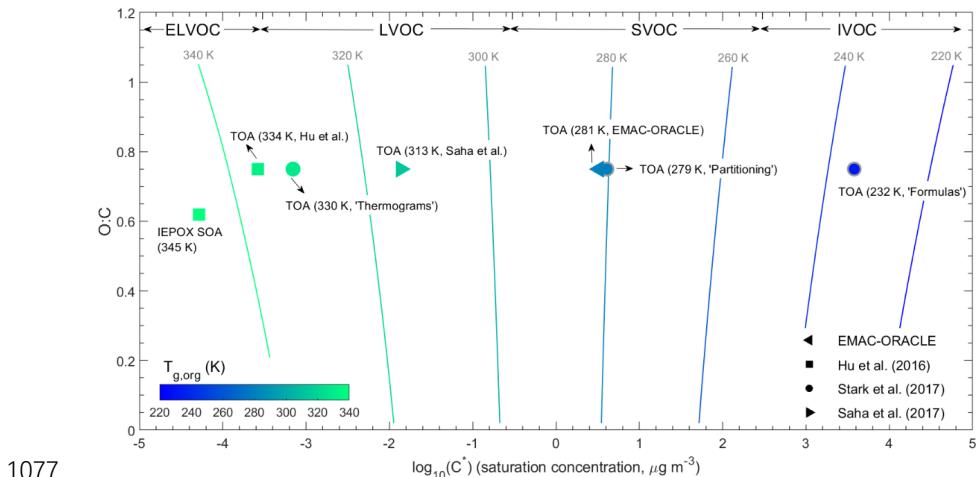
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1059 **Figure 2.** (a) Predicted  $T_g$  for CHON compounds using a parameterization (Eq. 4)  
 1060 developed in this study compared to measured or estimated  $T_g$  by the Boyer-Kauzmann  
 1061 rule using measured  $T_m$ . (b) Predicted  $T_g$  for CHON compounds using Eq. (4) compared  
 1062 to estimated  $T_g$  by the Boyer-Kauzmann rule with  $T_m$  estimated by the EPI suite. (c)  
 1063 Predicted  $T_g$  for CHOS compounds using Eq. (5) compared to estimated  $T_g$  by the  
 1064 Boyer-Kauzmann rule with  $T_m$  estimated by the EPI suite. The markers are color-coded  
 1065 by the molar mass. The solid line shows the 1:1 line and the dashed and dotted lines  
 1066 show 68% confidence and prediction bands, respectively. The correlation coefficient  
 1067 ( $R$ ) and the average absolute value of the relative error (AAVRE) are included in each  
 1068 figure legend.



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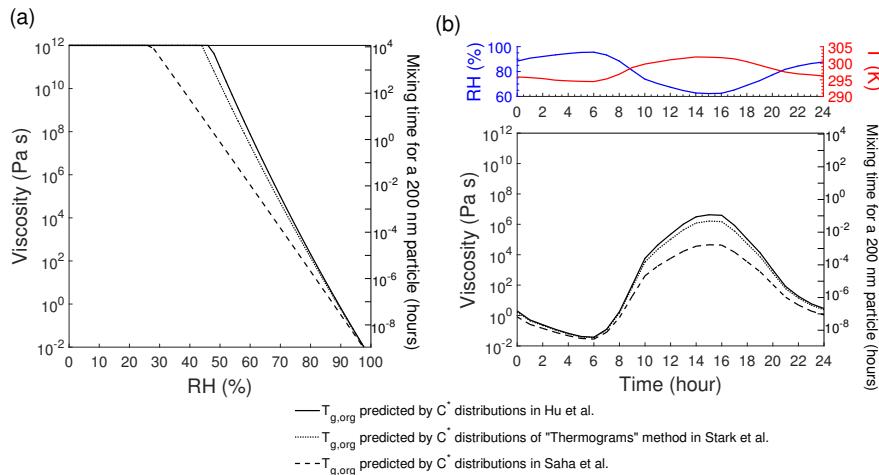
1070 **Figure 3.** Comparison between measured  $T_g$  and  $T_g$  predicted by (a)  $C^0$  and O:C (Eq.  
1071 1), (b) elemental composition (Eqs. 3-5), and (c) the parameterization (Eq. 6) in Zhang  
1072 et al. (2019). The markers are color-coded by the atomic O:C ratio. The solid line shows  
1073 the 1:1 line. The compounds include CH compounds (squares), CHO compounds  
1074 (circles), phthalates (stars), CHON compounds (triangles), and 2-methyltetrol sulfates  
1075 (2-MT-OS, inverted triangle). The correlation coefficient (R) and the average absolute  
1076 value of the relative error (AAVRE) are included in each figure legend.



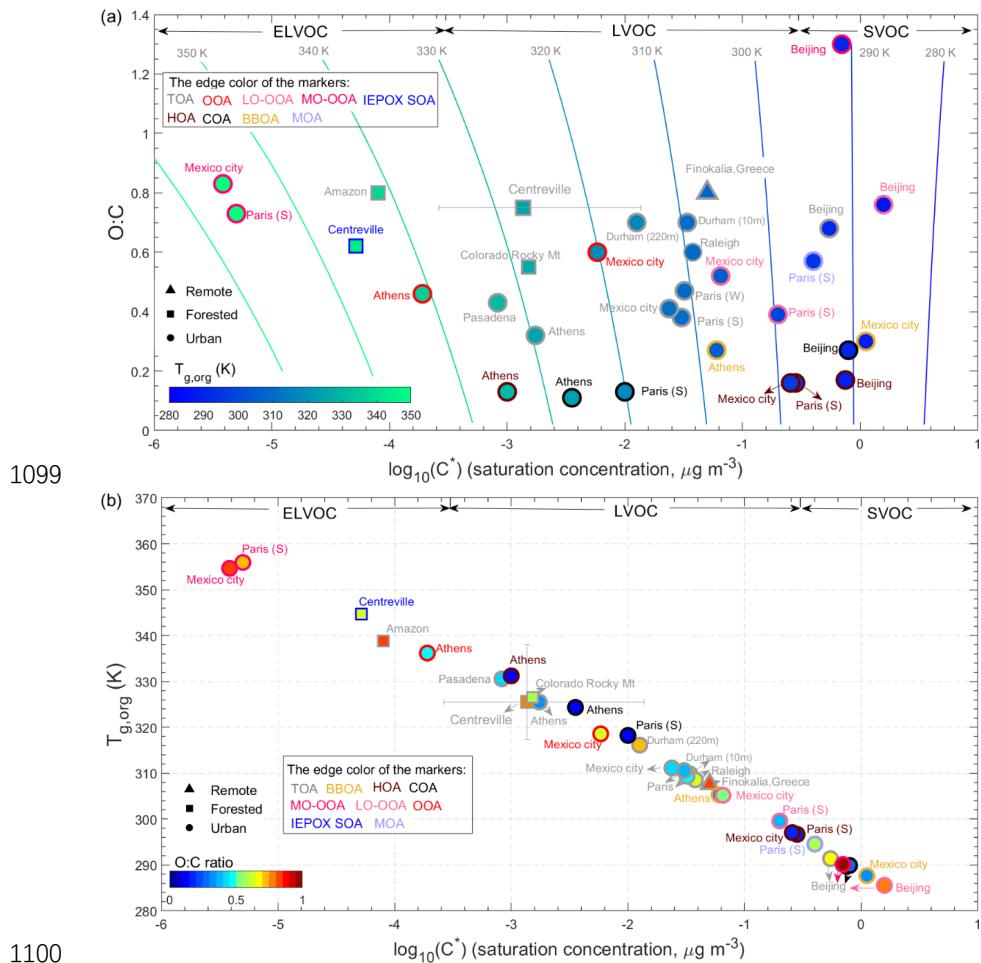
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1078 **Figure 4.** Predicted glass transition temperatures of organic aerosols under dry  
1079 conditions ( $T_{g,\text{org}}$ ) during the SOAS campaign, which are placed into the two-  
1080 dimensional VBS framework. The isopleths correspond to the  $T_{g,\text{org}}$  calculated using Eq  
1081 (1) with the effective saturation mass concentration ( $C^*$ ) and the O:C ratio defined in  
1082 the 2D-VBS. The markers represent the  $T_{g,\text{org}}$  of total OA (TOA) and IEPOX SOA  
1083 calculated from the volatility distributions simulated by a global chemical transport  
1084 model EMAC-ORACLE (Shiraiwa et al., 2017) or measured during the SOAS  
1085 campaign (Hu et al., 2016; Saha et al., 2017; Stark et al., 2017). Three methods  
1086 ('Formulas', 'Partitioning', and 'Thermograms') are applied in Stark et al. (2017) to  
1087 derive the  $C^*$  distributions, while the "Thermograms" method provides the most  
1088 credible volatility distributions compared to 'Formulas' and 'Partitioning' (edge lines  
1089 are in gray).

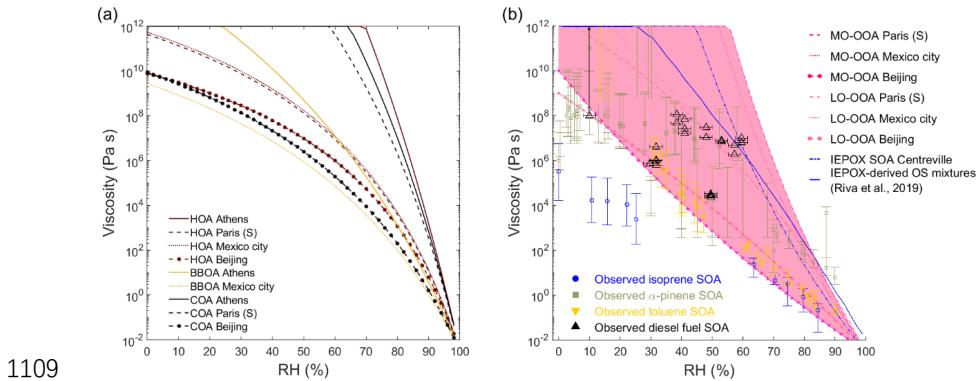
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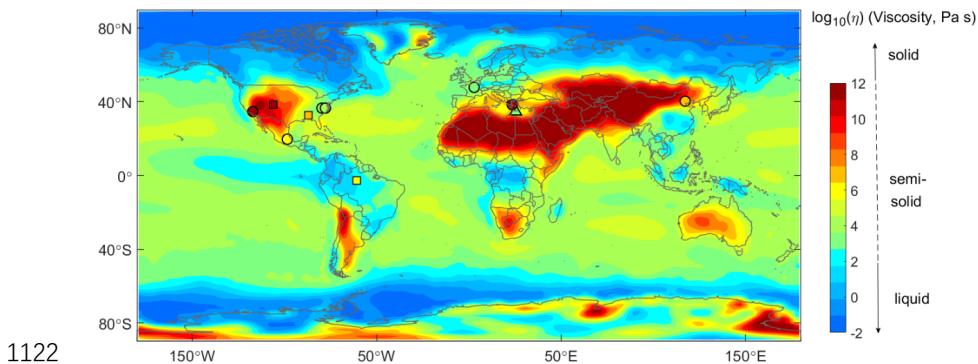
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1092 **Figure 5.** (a) Predicted viscosity of total OA measured during the SOAS campaign as  
1093 a function of RH. (b) Diurnal variations of viscosity of total OA predicted employing  
1094 the measured RH and  $T$  (Hu et al., 2016) during the SOAS campaign.  $T_{g,org}$  are  
1095 calculated using the volatility distributions measured in Hu et al., (2016), Saha et al.  
1096 (2017), and the “Thermograms” method in Stark et al. (2017). Characteristic mixing  
1097 timescales of organic molecules with the radius of  $10^{-10}$  m within 200 nm particles are  
1098 also shown in the right axis.



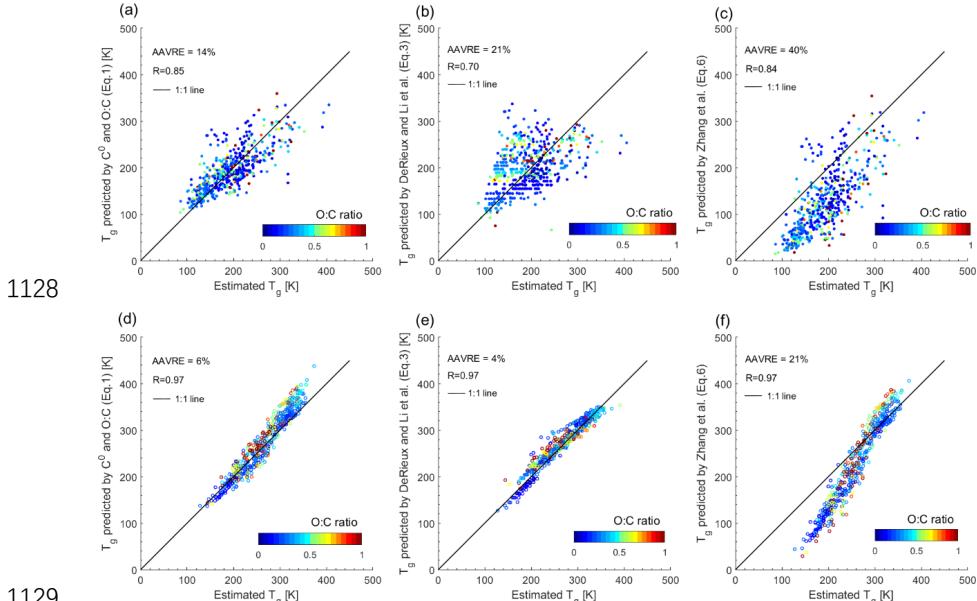
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 1100 **Figure 6.** Predicted glass transition temperatures of organic aerosols under dry  
 1101 conditions ( $T_{g,org}$ ) at 11 sites. The isopleths in (a) correspond to the  $T_{g,org}$  varied with the  
 1102 effective saturation mass concentration ( $C^*$ ) and the O:C ratio in the two-dimensional  
 1103 VBS (2D-VBS) framework. The markers represent the  $T_{g,org}$  of total OA (TOA) and its  
 1104 components calculated from measured volatility distributions (Table S2) and color-  
 1105 coded with  $T_{g,org}$  in (a) and the O:C ratio in (b). The edge color of the markers indicates  
 1106 the OA components identified via positive matrix factorization (PMF) of aerosol mass  
 1107 spectrometer data.  
 1108



1109  
1110 **Figure 7.** Predicted viscosity of (a) HOA, COA and BBOA and (b) LO-OOA, MO-  
1111 OOA, IEPOX SOA and Isoprene OA in different locations at 298 K as a function of  
1112 RH. Experimentally measured viscosity of laboratory-generated SOA formed from  
1113 isoprene (Song et al., 2015),  $\alpha$ -pinene (Abramson et al., 2013; Renbaum-Wolff et al.,  
1114 2013; Kidd et al., 2014; Pajunoja et al., 2014; Bateman et al., 2015; Zhang et al., 2015;  
1115 Grayson et al., 2016; Petters et al., 2019), toluene (Song et al., 2016), and diesel fuel  
1116 (Song et al., 2019) are also shown. Predicted viscosity of IEPOX-derived OS mixtures  
1117 (solid blue line) is from Riva et al. (2019). Note that in case these PMF factors are  
1118 internally mixed with other components, the predicted viscosity would not represent  
1119 real ambient complex organic mixtures.  
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1122  
1123 **Figure 8.** Comparison of global distributions of annual averages of SOA viscosity at  
1124 the surface simulated by a chemical transport model (Shiraiwa et al., 2017) with the  
1125 viscosity predicted by measured volatility distributions at 11 global sites as indicated  
1126 with markers (triangle, circle and square represent remote, urban and forested sites,  
1127 respectively, Table S2).



1130 **Figure A1.** Predicted  $T_g$  for CHO compounds by (a)  $C^0$  and O:C (Eq. 1), (b) elemental  
1131 composition (Eq. 3), and (c) the parameterization (Eq. 6) in Zhang et al. (2019) plotted  
1132 against estimated  $T_g$ .  $T_m$  and  $C^0$  in (a) – (c) are measured;  $T_m$  and  $C^0$  in (d) – (f) are  
1133 estimated by the EPI Suite and the EVAPORATION model, respectively (see  
1134 Supplement). Estimated  $T_g$  is then calculated from  $T_m$  applying the Boyer-Kauzmann  
1135 rule. The markers are color-coded by the atomic O:C ratio. The solid line shows the 1:1  
1136 rule. The correlation coefficient (R) and the average absolute value of the relative error  
1137 (AAVRE) are shown.