

# ***Interactive comment on “Numerical analysis of the impact of agricultural emissions on PM<sub>2.5</sub> in China using a high-resolution ammonia emissions inventory” by Xiao Han et al.***

**Xiao Han et al.**

zhlyun@126.com

Received and published: 7 May 2020

China is one of the largest agricultural countries in the world. The NH<sub>3</sub> emissions from agricultural activities in China, such as fertilizer and husbandry, farmland ecosystems, livestock waste, crop residue burning and fuel wood combustion, significantly affect regional air quality and horizontal visibility by contribution to secondary inorganic aerosols. In the manuscript, the air quality modeling system RAMS-CMAQ (regional atmospheric modeling system-community multiscale air quality), coupled with the ISAM (integrated source apportionment method) module is applied to capture the contribution of NH<sub>3</sub> emitted from total agriculture (Tagr) in China. It explores that the annual

Printer-friendly version

Discussion paper



average contribution of Tagr NH<sub>3</sub> to PM<sub>2.5</sub> mass burden in China was 14-18%. Specific to the PM<sub>2.5</sub> components, Tagr NH<sub>3</sub> provided a major contribution to ammonium formation (87.6%) but a tiny contribution to sulfate (2.2%). Though the Tagr NH<sub>3</sub> only contributed 10.1% of nitrate under current emissions scenarios, the reduction of nitrate could reach 98.8% upon removal of the Tagr NH<sub>3</sub> emissions. The results are meaningful, but the explanation for these phenomenon was not enough. I recommend the manuscript to be accepted after some minor revisions, and detail some issues below. Major points:

1. The most important gas in this manuscript was NH<sub>3</sub>, but there are no NH<sub>3</sub> in Figure 2 in comparing between the modeled and observed results.

R: Thanks for this comment. However, NH<sub>3</sub> is not included in the conventional observation species in China at present. Therefore, it is hard to collect available observation data of NH<sub>3</sub> mass concentration for model evaluation directly. Most of the available information was derived from the published research paper. In Han et al. (2017; Modeling dry deposition of reactive nitrogen in China with RAMS-CMAQ. Atmos. Environ.), the simulated NH<sub>3</sub> by RAMS-CMAQ has been compared with the observations from many studies in detail, including the multi-year observation results with Nationwide Nitrogen Deposition Monitoring Network and the seasonal variation characteristics from Pan et al. (2012; Wet and dry deposition of atmospheric nitrogen at ten sites in Northern China. Atmos. Chem. Phys.). In this paper, we also compare the simulation results with the value and seasonal variation at several stations from Pan et al. (2018) and Zhang et al. (2018) (Line 200-211). We kindly hope these content could reflect the reasonability of modeled NH<sub>3</sub>.

2. Why is the NH<sub>3</sub> contribution to nitrate small under "rich NH<sub>3</sub>" conditions and large in "poor NH<sub>3</sub>" environments? What is the internal logical relationship?

R: Thanks for this comment. In fact, the detail discuss about "rich NH<sub>3</sub>" and "poor NH<sub>3</sub>" can be found in Wang et al. (2011; Impact Assessment of Ammonia Emissions on

[Printer-friendly version](#)[Discussion paper](#)

Inorganic Aerosols in East China Using Response Surface Modeling Technique). The results of RSM (Response Surface Modeling) in their study shows that the change of NO<sub>3</sub>- mass concentration is very sensitive to the emission level of NH<sub>4</sub><sup>+</sup> and performs as nonlinear relationship. The reduction of NH<sub>3</sub> emissions can play a significant role in reducing the mass concentration of NO<sub>3</sub>- under NH<sub>3</sub>-poor condition. However, there will be excess NH<sub>3</sub> in the atmosphere under NH<sub>3</sub>-rich condition, and these excess NH<sub>3</sub> could neutralizes more nitric acid even in the case of emission reduction. Thus, the effect of emission reduction is not significant under NH<sub>3</sub>-rich condition. In addition, the SO<sub>2</sub> will compete for NH<sub>3</sub> and prevent the generation of NH<sub>4</sub>NO<sub>3</sub> under NH<sub>3</sub>-poor condition because the reaction between H<sub>2</sub>SO<sub>4</sub> and NH<sub>3</sub> takes precedence over the one between HNO<sub>3</sub> and NH<sub>3</sub>. Oppositely, SO<sub>2</sub> should be benefit for the formation of NO<sub>3</sub>- (especially in summer) under NH<sub>3</sub>-rich condition according to the calculation of Wang et al. (2011). This should be a reason why the effect of NH<sub>3</sub> emission control is not obvious in the case of NH<sub>3</sub>-rich condition as well.

3. The study period is January, April, July, and October, but only the modeled and observed results in January and July are compared in Figure A1, A2, A3 and A4.

R: Thanks for this comment. We added the comparison of meteorological factors in April and October. Please check if it is appropriate.

4. The author thinks that the obvious deviation between the observed and modeled SO<sub>2</sub> in January may be a systemic underestimation due to the lack of emission intensity in this month. Did the lack of emission intensity only appear in SO<sub>2</sub>? Why are SO<sub>2</sub> and NO<sub>2</sub> underestimated and PM<sub>2.5</sub> overestimated?

R: Thanks for this comment. The monthly mean observation data were used in the submitted version. However, we would like to provide more details about the evaluation. Thus, the hourly observation data from the China National Environmental Monitoring Centre were collected and compared with simulation results. The scatter plots (Figure 2) were replaced and the comparison of SO<sub>2</sub>, NO<sub>2</sub> and PM<sub>2.5</sub> in January, April,

[Printer-friendly version](#)[Discussion paper](#)

July, and October at six sites were presented, and the statistical summary of the comparisons and related discussion were modified (Line 186-198). Please check if it is appropriate.

5. How much NH<sub>3</sub> is removed in Figure 7? And it's more intuitive to use a negative value for reduction.

R: Thanks for this comment. Here the emission of NH<sub>3</sub> from all agricultural sources were removed. For detail information, please see the percentage shown in Figure A6 which we added. In addition, the horizontal distribution of the PKU-NH<sub>3</sub> emission inventory can be viewed in Kang et al. (2016) (Kang et al., 2016: High-resolution ammonia emissions, High-resolution, ammonia 1980, 2012.). On the other hand, the Figure 7 was modified. Please check if it is appropriate.

6. Why do the trend of the decrease in ammonium mass concentration accelerate while NH<sub>3</sub> emissions is less than 20%?

R: Thanks for this comment. Here the simulation scenario was conducted for each emission reduction of 10% so that the acceleration should appear between 20% and 30%. In fact, it can be found that the accelerated decline mainly started when the emission reduction exceeds 50%. Therefore, we could deduce that the accelerated decline should be emerged gradually with NH<sub>3</sub> emission reduction. This feature indicates that the formation of NH<sub>4</sub><sup>+</sup> should be nonlinear with NH<sub>3</sub> emission intensity as well. The reason may also be related to the complex neutralization reaction among sulfate, nitrate and ammonium. The consumption of NH<sub>3</sub> should become more sufficient when the mass concentration of NH<sub>3</sub> is lower. Thus, the variation of ammonium is more sensitive under low NH<sub>3</sub> mass burden.

7. What is the horizontal distributions of the contribution percentage of NH<sub>3</sub> emissions to ammonium, nitrate and sulfate mass concentration, respectively? Which aerosol determines the horizontal distributions of SNA mass concentration? Why is the horizontal distributions of NH<sub>3</sub> emissions different with the horizontal distributions of the

[Printer-friendly version](#)[Discussion paper](#)

contribution percentage of NH<sub>3</sub> emissions to SNA mass concentration?

R: Thanks for this comment. The horizontal distributions of NH<sub>3</sub> emission contribution to sulfate, nitrate and ammonium is shown in Figure R1, and ammonium provided the major contribution to SNA (Table 4 also presented related information). In addition, Figure 6 shows the horizontal distributions of contribution percentage which may not follow the distribution pattern of mass concentration. For example, it can be seen that the agricultural NH<sub>3</sub> emission generally provided more than 90% contribution to ammonium over China in January as shown in Figure R1. Therefore, the contribution ratio should differ from the horizontal distribution pattern.

Minor points: 1. In Figure 6 and Figure 7, it should be the horizontal distributions in January, April, July, and October. 2. In Line 226, it should be "Since NH<sub>3</sub> concerns mainly with secondary inorganic aerosols (SNA): sulfate, nitrate, and ammonium formation". 3. In line 269, what is "TA NH<sub>3</sub> emission"? 4. In Line 833, should it be "The regional percent (%) of Tagr NH<sub>3</sub> contribution"?

R: Thanks for the comments. All error points were modified.

Please also note the supplement to this comment:

<https://www.atmos-chem-phys-discuss.net/acp-2019-1128/acp-2019-1128-AC1-supplement.pdf>

---

Interactive comment on Atmos. Chem. Phys. Discuss., <https://doi.org/10.5194/acp-2019-1128>, 2020.

Printer-friendly version

Discussion paper



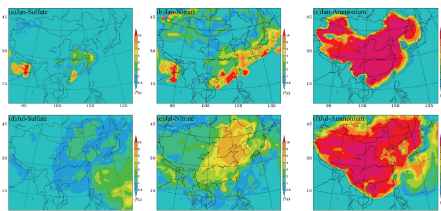


Figure R1 The horizontal distributions of the contribution percentage of  $\text{NH}_3$  emissions to sulfate, nitrate and ammonium mass burden (%) in January and July.

**Fig. 1.** Figure R1 The horizontal distributions of the contribution percentage of  $\text{NH}_3$  emissions to sulfate, nitrate and ammonium mass burden (%) in January and July.