

Reply to Reviewer Report 1

In the following the comments of the reviewer are presented (black) alongside with our replies (in blue) and changes made to the manuscript (in red).

General statement: The authors present an analysis of net O₃ production calculated from ship-borne trace gas measurements obtained during the AQABA campaign, mostly in July and August 2017. In the Oman Gulf, the Northern Red Sea, and the Arabian Gulf the authors found the highest values ranging from 14 ppb day⁻¹ to 28 ppb day⁻¹. Based on HCHO/NO₂ ratios, in most areas O₃ formation was NO_x limited apart from the Northern Red Sea, which was located in the VOC-NO_x limitation transition zone. The Arabian Sea and Arabian Gulf areas showed maximum HCHO/NO₂ values, clearly indicating NO_x limitation. This paper shows some interesting data from an area, which often lacks robust air quality data, but which is also an area with significant (and increasing) anthropogenic emissions, mostly related to oil and gas exploration. The paper is well-written and deserves publication. However, I was hoping the authors could address some of my concerns.

Dear reviewer, thank you very much for reviewing our manuscript and for the insightful comments. Below we provide detailed responses to your comments. Please find revised graphs at the end of the document, which were compiled based on your comments.

Comment 1: Page 1 L22-24: According to Fig 10 net O₃ formation actually mostly occurs in NO_x limitation regimes, not in the transition regime, as the authors mention here. Also, it seems Fig 10 does not support the findings by Pfannerstill et al. (2019) as stated by the authors on page 25 (L509-511). All, but one median, is above the threshold of HCHO/NO₂ > 2. I also doubt that a HCHO/NO₂ median of 2.2 (for OG) would signify a tendency towards VOC limitation (L511-512).

Indeed a median HCHO/NO₂-ratio of 2.2 does not fall within plain VOC limitation as deduced by Duncan et al. (2010), which needs to be re-written. Although Pfannerstill et al. (2019) highlight the scatter in the attribution of ozone production to NO_x- and VOC-limitation, their study shows that ozone production over the Arabian Sea over the Arabian Gulf is rather NO_x-limited, whereas the Gulf of Suez is characterized by a strong VOC-limitation (page 11516 and page 11517, Figure 9 [Pfannerstill et al., 2019]). We have revised Page 1 L21-24 to: Constrained by HCHO/NO₂-ratios, our photochemistry calculations show that net ozone production in the MBL around the Arabian Peninsula occurs mostly in NO_x-limitation regimes with a significant share of ozone production occurring in the transition regime between NO_x- and VOC-limitation over the Mediterranean and more significantly over the Northern Red Sea and Oman Gulf.

Comment 2: Page 3, L58-59: The reference Zhou et al (2014) shows up a couple of times in this paper. While I agree that it makes sense to compare the authors' Middle East study area with the Houston area (e.g. some similar emissions; similar latitude) there are better publications for the Houston case which include direct measurements of OH, HO₂, and also O₃ production itself among many other measurements (e.g. Chen et al., 2010; Mao et al, 2010; Ren et al. 2013). Also, I doubt that the term marine environment shows up in Zhou et al.

We have removed the term "marine environment" in the context of Zhou et al. (2014) and we have revised the manuscript as suggested on Page 2, L55-56 to: Measurements performed in the Houston Ship Channel revealed NOPR of the order of several tens of ppb h⁻¹ (Chen et al., 2010; Mao et al., 2010; Ren et al., 2013).

Comment 3: Page 3, L80-86: This explanation is confusing. It seems there is a main sampling inlet (not sure what instruments were connected), but for HCHO and NO_x sampling was done via a 1/2" PFA tubing (not sure how long the tube was), and for the other trace gas measurements (what were those?) it was done in a different way. May be a chart showing the experimental design would be helpful.

50 Please note that air was drawn from the stainless steel common inlet into each measurement container via
bypass systems. NO_x (CLD measurements) and HCHO measurements were located in one lab container (both
sampling air from the same bypass) and NO₂ (CRDS measurements) and O₃ in another lab container. We have
revised Page 3, L78-84 to: A 6 m high, 20 cm diameter cylindrical stainless steel common inlet was installed on
the front deck of the vessel to sample air at a total mass flow rate of 10,000 SLM. NO and NO₂
chemiluminescence measurements were obtained at a total bypass flow rate of 28.5 SLM sampling air from the
common inlet with a residence time in the tubing of ~3 s. HCHO, NO₂ cavity ring-down spectroscopy and O₃
55 measurements were obtained with similar bypass systems sampling air from the common inlet. H₂O vapor was
measured on the top of the ship mast in the front.

Comment 4: Page 6, L144: What was the flow of the calibration gas into the zero air?
60 The NO calibration flow into the zero air was 4.5 sccm. We have revised the manuscript. Now it says on Page 6,
L 140-142: Zero air measurements and NO calibrations were performed with a total mass flow of 3.44 SLM
achieving an overflow of 0.44 SLM to guarantee ambient air free standard measurements. The calibration gas
was added at 4.5 sccm to the zero air flow.

65 Comment 5: Page 6, L157: What do the authors mean by "notably high"?
Alkenes can react with ozone to produce a chemiluminescent signal which will bias NO measurement obtained
by chemiluminescence. To subtract such interferences, the CLD has been equipped with a prechamber to which
the sampled air can be directed during prechamber measurements. As the reaction of ozone with alkenes will be
70 much slower than the reaction of NO and with ozone, the CLD will be measuring only the signal from the reaction
of alkenes during prechamber measurements. It should be noted that the signal from the reaction of alkenes with
ozone observed during a prechamber measurement will also slightly decrease. Regions where alkenes are
strongly varying in time and magnitude might be plagued by a measurement offset. We have revised Page 6,
L148-150: However, in regions where alkene concentrations are strongly varying in time and magnitude, the CLD
75 is prone to enhanced backgrounds due to the interference of alkenes with ozone in the instrument.

Comment 6: Page 11, L259: I guess the authors mean Lu et al. (2010) here.
Thank you very much for noticing this typo. Now it says Lu et al. (2010) instead of Lu et al. (2016) on page 11,
80 L260.

Comment 7: Page 12, L276-277: The loss mechanism through H₂O is important. Also, it seems to vary a lot.
Some parts of the ship cruising legs might have already been exposed high humidity due to the Indian monsoon
85 system. It would be good to see the absolute humidity variation along the legs similar to Figs 3 and S4.
Figures showing the absolute variation along both legs have been added to the supplement. On P13 L328 now it
says: Supplementary Figure S5 shows the variation of the absolute humidity around the Arabian Peninsula.

90 Comment 8: Page 12, L281-282: At least, the authors want to include an estimate for the potential contribution of
halogens.
Based on oxidative pairs, Bourtsoukidis et al. found that the majority of the samples they collected during AQABA
were characterized by a OH/Cl-ratio of ~ 200:1 (Bourtsoukidis et al., 2019, Non-methane hydrocarbon (C₂-C₈)
sources and sinks around the Arabian Peninsula, doi:10.5194/acp-19-7209-2019). Measured daytime OH
95 concentrations were of the order of $5 \cdot 10^6$ molec cm⁻³, hence the Cl radical concentration can be estimated at
 $2.5 \cdot 10^4$ molec cm⁻³. Incorporating an ozone loss due to the reaction of O₃ with Cl (at Cl concentrations of $2.5 \cdot$
 10^4 molec cm⁻³) into the NOPR (Eq. 7) would decrease the diurnal net ozone production rates by roughly 0.2

ppbv day⁻¹ over the Arabian Sea and at most 0.6 ppbv day⁻¹ over the other regions. We have revised the manuscript as follows on page 12, L285-289: Based on oxidative pairs, Bourtsoukidis et al. (2019) have classified the majority of their samples collected during AQABA by an OH/CI-ratio of 200:1. As measured daytime OH concentrations were of the order of 5·10⁶ molecule cm⁻³, the estimate would yield a CI concentration of 2.5·10⁴ molecule cm⁻³, which would decrease the estimated diurnal net ozone production rates by roughly 0.2 ppbv day⁻¹ over the Arabian Sea and at most 0.6 ppbv day⁻¹ over the other regions, which does not substantially alter the here presented findings. The noontime chemical ozone loss rate can be summarized by

Comment 9: Page 12, Eq. 7 and Fig 9: It would be nice to see a break-down of the different terms in Eq. 7 for different legs as shown in Fig. 9 to evaluate what processes might be most relevant/different in those different legs.

We have included regional box-whisker plots (according to Figure 9) of the four terms of Eq. 7 in the supplements. We have revised the manuscript on Page 21, L460-461: A break-down of the different terms of Eq. 7 in the six regions is included in the supplementary Figures S10 – S13.

Comment 10: Page 13, L310: Authors mention NO_x values of several hundred ppbs. Where do they show up in Figs. 3, S2, and S3? What were the megacities along the cruising legs? I could think about Cairo, but according to Fig 3 NO_x values do not show extremely high values.

Values of up to several hundred of ppbv NO_x were observed when measuring own ship stack, stack of bypassing ships or when being at anchor in the direct vicinity of Jeddah or Kuwait City. Data measured during these time periods have been removed from the final data set as contamination by the ship exhaust itself could not be excluded for these period. Megacities along the cruising legs generally include Cairo, Kuwait City, megacities in the UAE and Jeddah (Saudi Arabia).

Comment 11: Page 13, L311-312: During the first leg very high O₃ values are found in the Arabian Gulf and potentially in the area of the Suez Canal. In the second leg those high O₃ values are pretty much gone. I doubt emissions have changed. I also doubt that weather conditions have changed drastically. What were the reasons for those distinct changes?

Based on back trajectories, previous studies (Pfannerstill et al., 2019) also already highlighted different air mass origins during the two legs for the air sampled e.g. over the Arabian Gulf. While during the first leg northwestern wind from Kuwait/Iraq was encountered, northeastern winds from Iran were encountered during the second leg over the Arabian Gulf. For the area of Suez, data coverage during the first leg (due to instrumental mal function and applied stack filter) is not as exhaustive as during the second leg. We would classify these changes of concentrations in the Suez region as insignificant (see also Figure S2 and the data coverage for the Suez region before July 06th 2017). We have revised the manuscript on Page 16, L351-355 as follows: However, a significantly larger whisker-interval of observed ozone of 31.4 ppbv over the Gulf of Oman indicates increasing amounts of pollution and advection from the Arabian Gulf where extreme events of ozone were observed several times during the campaign with maximum mixing ratios of up to 170 ppbv when wind was coming from Kuwait/Iraq. Please note that during the second leg wind was coming from Iran (Pfannerstill et al., 2019).

Comment 12: Page 16, L334-355: This section should include some more explanations: it seems there is a huge variation in NO_x and O₃ in AG (also a huge variation in NOPR as shown in Fig 9). What is the major driver of this: point sources from ships? Why are the highest NO_x values in OG and why are some of the lowest O₃ values found in OG? Why would you consider air masses over the Mediterranean as photochemically aged air masses due to the small whisker-interval, while the whisker-plots for AS and OG show pretty much the same with, but at much lower absolute O₃ ranges. There are no emission sources in that area of the Mediterranean?

Reasons for large variations of NO_x and O_3 over the AG are point sources (ship, oil and gas processing) as well as a change in the general wind direction observed during both legs. NO_x are high over the OG due to the magnitude of emissions from vessels. O_3 is generally very low over the OG because it has been partly characterized as VOC-limited and high NO_x values may contribute to net ozone destruction. Air over the Mediterranean has previously been characterized as photochemically processed emissions from Eastern Europe (Turkey, Greece) (Destroff et al., 2017; Pfannerstill et al., 2019). We have included Destroff et al., “Volatile organic compounds (VOCs) in photochemically aged air from the eastern and western Mediterranean, 2017, doi:10.5194/acp-17-9547-2017 as a reference. Page 16, L348-364 following now say (please note that only the underlined sentences have been changed): The low O_3 mixing ratios over the Arabian Sea was accompanied by the smallest variability (whisker-interval: 15.1 ppb_v). Although observing highest NO_x over the Oman Gulf, O_3 observed over the Oman Gulf was amongst the lowest detected throughout the whole campaign, which can be partly explained by the fact that high NO_x lead to low ozone production or even net ozone destruction. However, a significantly larger whisker-interval of observed ozone of 31.4 ppb_v over the Gulf of Oman indicates increasing amounts of pollution and advection from the Arabian Gulf where extreme events of ozone were observed several times during the campaign with maximum mixing ratios of up to 170 ppb_v when wind was coming from Kuwait/Iraq. Please note that during the second leg wind was coming from Iran (Pfannerstill et al., 2019). The whisker-interval over the Arabian Gulf was 100.9 ppb_v, more than six times higher than that over the Arabian Sea. Reasons for large variations of both NO_x and O_3 over the Arabian Gulf were a multitude of point sources as well as a change in the observed wind direction with air masses coming from Iraq/Kuwait area during the first leg and air masses coming from Iran during the second leg (Pfannerstill et al., 2019). Over the Mediterranean, the Northern Red Sea and the Southern Red Sea, median ozone was 61.5 ppb_v, 64.2 ppb_v and 46.9 ppb_v, respectively. The whisker-interval over the Northern Red Sea and the Southern Red Sea were 44.2 ppb_v and 31.6 ppb_v, respectively. Air masses over the Mediterranean were characterized as photochemically aged due to their impact by northerly winds (Etesians) which bring processed/oxidated air from eastern Europe (Turkey, Greece) to the Mediterranean area (Destroff et al., 2017; Pfannerstill et al., 2019). This photochemical ageing/oxidation over the Mediterranean leads to a rather small whisker-interval of 18.7 ppb_v in ozone.

Comment 13: Page 17, L367: As I understand it Velchev et al. (2011) show O_3 data from the Western Mediterranean. A reference looking closer to the area the authors studied would be Kouvarakis et al (2002). We have now replaced (Velchev et al., 2011) by (Kouvarakis et al., 2002) on page 17, L382.

Comment 14: Page 17, L369: Edwards et al. (2014) is not a good reference here. Edwards et al report wintertime O_3 in cold-pool conditions, i.e. extremely low boundary layer heights. The meteorological conditions reported in Edwards et al are pretty different from the ones observed in the Middle East during summertime. Also, there is no word on the impact of narrow shipping lanes in Kleinman et al (2005) and Zhou et al (2014). O_3 in Houston is predominantly driven by emissions from all kinds of petrochemical industries (including refineries, but no oil exploration) located in the Ship Channel area.

We agree that Edwards et al., 2014 is not the best reference in here, which has been removed in this context. Also Kleinman et al. (2005) and Zhou et al. (2014) have been removed in the context of shipping lanes. Instead we have included Mazzuca et al., 2016 “Ozone production and its sensitivity to NO_x and VOCs: results from the DISCOVER-AQ field experiment, Houston 2013”, doi:10.5194/acp-16-14463-2016. On page 17, L382-384 now it says: The latter are consistent with O_3 mixing ratios reported from regions influenced by oil and gas processing (Pfannerstill et al., 2019) and shipping lanes such as the Houston Ship Channel (Mazzuca et al., 2016).

Comment 15: Page 25, L516-517: Actually, Figure 10 shows that in almost all areas O_3 formation is NO_x limited. However, the authors say that this is typical for photochemically aged air masses over the Mediterranean. As

200 already mentioned further above, why do the authors explicitly consider the Mediterranean area having aged air masses? It is even more surprising as the results for the Mediterranean area in Figure 10 indicate that the Box-Whisker plot stretches into the transition between NO_x and VOC limitation.

205 The Mediterranean area is considered having aged air masses as it is impacted by processed emissions from eastern Europe (Destroff et al., 2017; Pfannerstill et al., 2019). We have removed the statement that photochemical ageing of air masses over the Mediterranean leads to NO_x-limitation, however it should be noted that on 29 August 2017 we were lying at anchor in front of Malta with a magnitude of pollution sources nearby.

210 This day is characterized by a low HCHO/NO₂-ratio and explains why the Box-Whisker-Plot for the Mediterranean stretches into the transition between NO_x- and VOC-limitation. Page 25, L534-538 now says: Ozone production over the Mediterranean was classified as rather NO_x-limited, however partly being in the transition regime between NO_x- and VOC-limitation, which can be explained by measurements obtained on 29 August 2017 when laying at anchor in front of Malta with a multitude of (NO_x)-emissions from nearby situated vessels. Average noontime NO_x on that particular day was about three times as large as the regional average noontime NO_x observed over the whole Mediterranean area.

215 Comment 16: Page 25. L517-519: Why would higher NO_x lead to higher O₃ pollution? For instance, according to Figure 4, OG has the highest NO_x values, but also pretty low O₃ values. With regard to NOPR, the Box-Whisker plot for OG shows positive, but also large negative values. In any case NOPR values are significantly lower than for AG, for instance.

220 Indeed, the passage needs more characterization. We have re-written the passage. Now it says on P25 L539-544: Note that a further increase in NO_x-emissions from increased shipping in the Arabian Gulf may initially lead to higher ozone production. However, a further increase in NO_x might eventually lead to a change from NO_x- to VOC-sensitivity and a decrease in ozone production for this region, as observed for the Oman Gulf (median HCHO/NO₂-ratio of 2.2 and average O₃ of 34 ppb.). See supplementary Table ST9 for detailed statistics on regional HCHO/NO₂-ratios.

225 Comment 17: Table 1: This table can go into the supplement.

230 We have moved the table into the supplement and numbered the tables in the manuscript and supplement accordingly. In the manuscript (Line 99-101) now it says: See supplementary Table ST1 for the range of latitudinal and longitudinal coordinates of the different regions and supplementary Table ST2 for a detailed day to day description of the route.

235 Comment 18: Figure 7: The legend mentions "Measurements", the figure captions says "estimated". From Eq 3 I understand that RO₂ was neither measured nor estimated, but calculated. Also, what would be the interpretation of the negative RO₂ concentrations (blue Box- Whisker plots) when calculated from Eq 3?

The legends (Figure 7) now says Estimated based on measured data and EMAC. Also the legend for Figure 9 now says Estimated based on measured data and Estimated based on simulated data.

240 Comment 19: Figures S2 and S3: It would be nice to see the time series of OH and HO₂ here as well.

We have added the time series of OH and preliminary HO₂ to Figures S2 and S3.

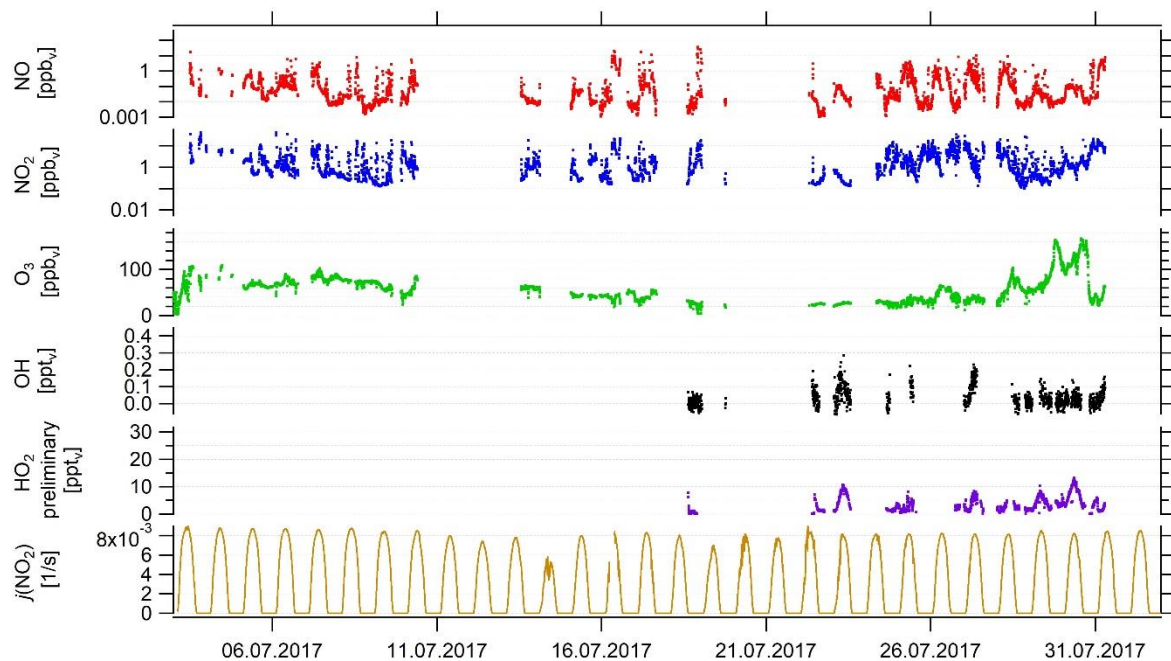


Figure S2: Timeline of NO, NO₂ (both CLD), O₃, OH, HO₂ preliminary and $j(\text{NO}_2)$ data during the first leg. See Table ST1 for additional information on the ship cruise. Note that HO₂ data are preliminary.

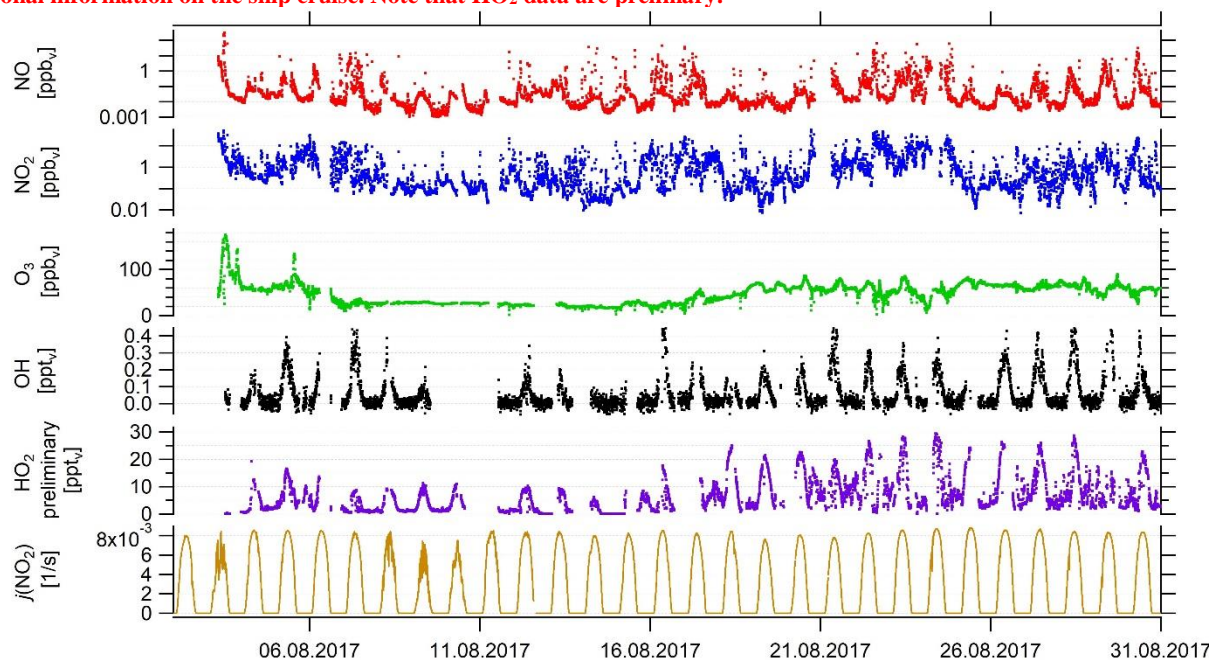
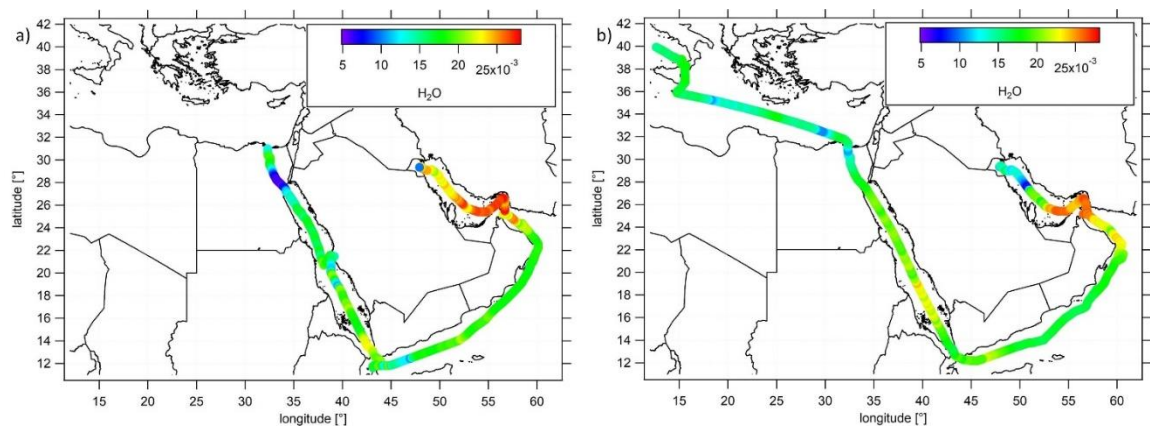
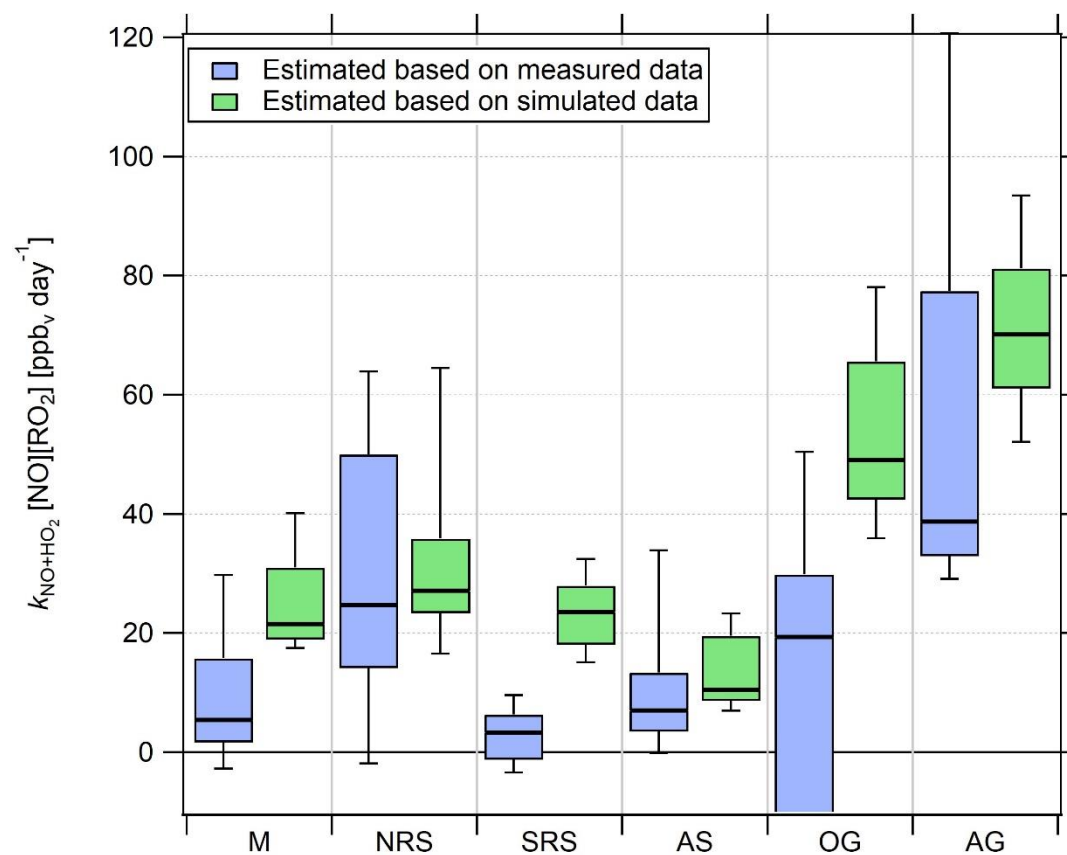


Figure S3: Timeline of NO, NO₂ (both CLD), O₃, OH, HO₂ preliminary and $j(\text{NO}_2)$ data during the second leg. See Table ST1 for additional information on the ship cruise. Note that HO₂ data are preliminary.



250 **Figure S5: Ship cruises with color-scaled absolute humidity a) during the first and b) the second leg.**



255 **Figure S10: Comparison of the regional, absolute contribution of $k_{\text{NO}+\text{HO}_2}[\text{NO}][\text{RO}_2]$ to NOPR in the six different regions investigated during AQABA. The horizontal black bar indicates the median value, the box the 25- and 75-percentiles and the whiskers the 10- and 90-percentiles.**

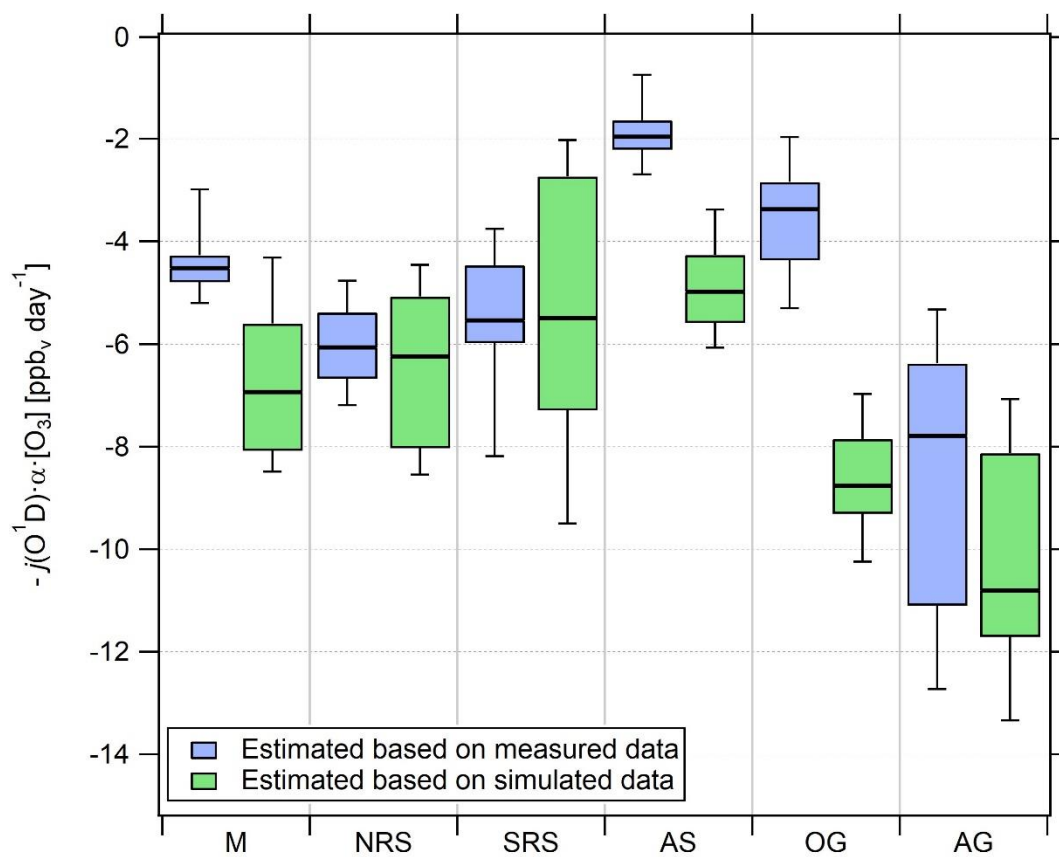


Figure S11: Comparison of the regional, absolute contribution of $-j(\text{O}^1\text{D}) \cdot \alpha \cdot [\text{O}_3]$ to NOPR in the six different regions investigated during AQABA. The horizontal black bar indicates the median value, the box the 25- and 75-percentiles and the whiskers the 10- and 90-percentiles.

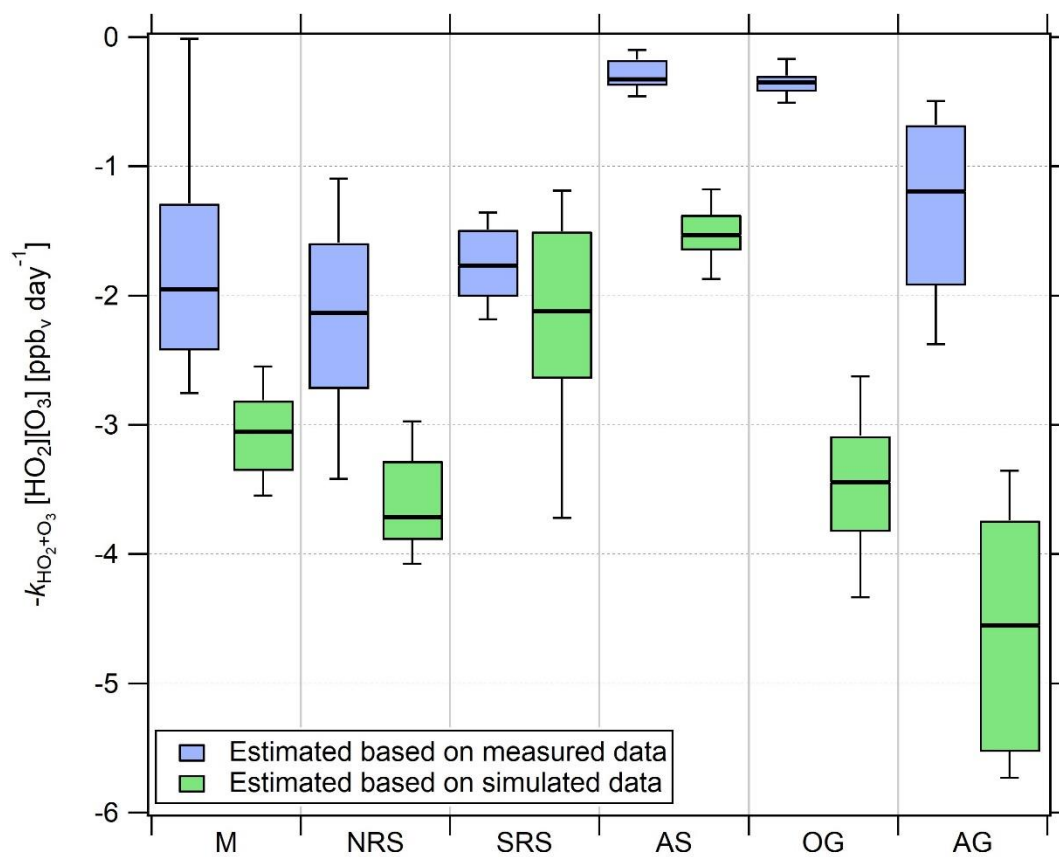


Figure S12: Comparison of the regional, absolute contribution of $-k_{\text{HO}_2+\text{O}_3}[\text{HO}_2][\text{O}_3]$ to NOPR in the six different regions investigated during AQABA. The horizontal black bar indicates the median value, the box the 25- and 75-percentiles and the whiskers the 10- and 90-percentiles.

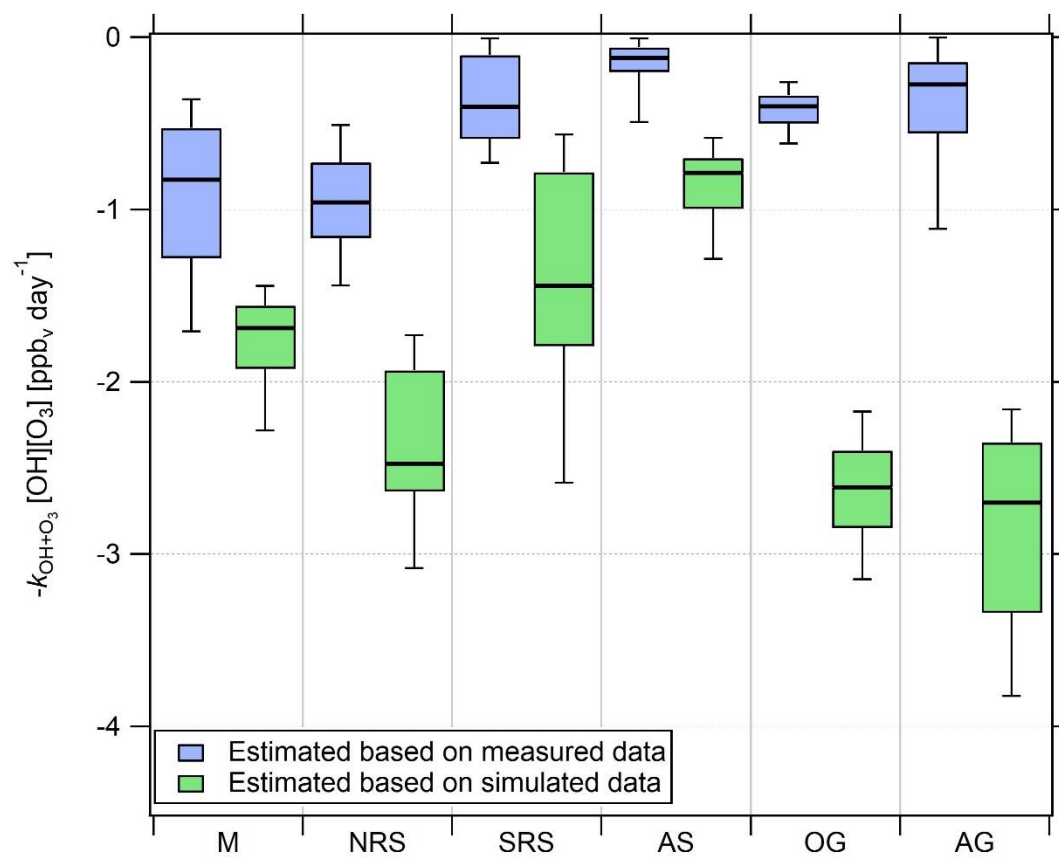


Figure S13: Comparison of the regional, absolute contribution of $-k_{\text{OH}+\text{O}_3}[\text{OH}][\text{O}_3]$ to NOPR in the six different regions investigated during AQABA. The horizontal black bar indicates the median value, the box the 25- and 75-percentiles and the whiskers the 10- and 90-percentiles.

Reply to Reviewer Report 2

In the following the comments of the reviewer are presented (black) alongside with our replies (in blue) and changes made to the manuscript (in red).

280 General statement: The authors have a very interesting data set. I'm not familiar with the chemistry of this region but I assume there are few high quality NO_x measurements and perhaps no radical measurements. Analysis is through the use of deviations from the Leighton
photostationary state to get peroxy radical concentrations which are then used in conjunction
with measured NO to obtain ozone production rates. Observed OH and HO₂ are presented only via a color-
285 coded ship track. As far as I can tell, the only use of the observed OH and HO₂ is to determine loss rates of ozone which when added to the PSS ozone production rate yields a net ozone production rate. The ratio of HCHO to NO_x (both observed quantities) is used as an indicator ratio to predict regions in which O₃ production is either NO_x or VOC sensitive.

290 Comparisons are made with a Eulerian model for NO_x, ozone, and RO₂ mixing ratios as well as for net O₃ production rate. This article contains the quantum of information to warrant publication. I can only guess that either they want to write up the radical and HCHO measurements in a separate study or that they are uninterruptable for one reason or another.

295 I found the article difficult to follow in places. It is my belief that the authors have underestimated the uncertainty of the instruments used to determine peroxy radicals. The authors need to re-examine their error propagation formula. I am calling this a major revision as it affects the most prominent results in the paper. In practice it can be done in an afternoon. They could also compare the PSS value of RO₂ with measured HO₂ (which is stated to be preliminary due to an up to 20% interference by RO₂). This is a reality test.

300 Dear reviewer, thank you very much for reviewing our manuscript and for the insightful comments. Below we provide detailed responses to your comments. Please note that PSS RO₂ has been compared to HO₂ in Sect. 3.2, P 21 L442.

305 Comment 1: line 124 "UV-induced positive bias in the NO₂ measurements due to photolysis of HONO, BrONO₂, NO₃ and ClNO₂ to produce NO was characterized ahead of the campaign to be 7.7 %, 7.2 %, 5.6 % and 1.5 % of the respective ambient concentration of HONO, BrONO₂, NO₃ and ClNO₂ respectively,"
(Italics mine). No where in the paper is it mentioned that these species are measured. Have interferences been determined based on model-calculated or typical concentrations? Or do these figures represent the percent
310 interference if the interferent has the same concentration as NO?

We have used absorption cross sections from the MPI-Mainz UV/VIS Spectral Atlas of Gaseous Molecules (Keller-Rudek et al., 2013) for the estimation of the parallel interference. The particular photolysis rate of NO₂ and the regarded NO_y-compound was calculated by (Eq. 1)

$$315 \quad j = \int F(\lambda, T) \cdot \sigma(\lambda, T) \cdot \varphi(\lambda) d\lambda, \quad (1)$$

where $F(\lambda, T)$ is the spectral emission of the UV LEDs, $\sigma(\lambda, T)$ the absorption cross section and $\varphi(\lambda)$ the quantum yield. For the calculation of the photolysis rate of the particular NO_y-compound the respective quantum yield was conservatively estimated to be equal 1, yielding an upper limit for the interference. On Page 5, L 120
320 now it says: The UV-induced positive bias in the NO₂-measurement due to photolysis of BrONO₂, HONO, NO₃ and ClNO₂ to produce NO was estimated at 6.1 %, 2.8 %, 2.7 % and 1.2 %, respectively, based on the absorption cross sections from the MPI-Mainz UV/VIS Spectral Atlas of Gaseous Molecules (Keller-Rudek et al., 2013). These values represent upper limits for the interference of the respective NO_y compound as the respective molecular quantum yield was estimated conservatively at 1. Note that the values represent percent interferences

325 if the interferent had the same concentration as NO₂. Due to small daytime concentrations of these molecules in
the MBL, a UV-induced bias was neglected for the observations in this study.
Please note that the values have been revised to 6.1 %, 2.8%, 2.7 % and 1.2 % for BrONO₂, HONO, NO₃ and
ClNO₂, respectively. See further changes.

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Comment 2: line 181 NO₂ was further measured by cavity ring-down spectroscopy (Sobanski et al., 2016) and
used for correcting the instrumental background of the CLD NO₂ data, as described above (the correction was
taken as the ultimate absolute measurement uncertainty in the CLD NO₂ data).

335 The uncertainty of the NO₂ chemiluminescent measurement is not equal to that of the cavity ring-down
instrument. The chemiluminescent NO₂ is the difference between two measurements, one of which gets divided
by 0.294 to take into account photolysis efficiency. In order to add errors in quadrature, I need to know the NO to
NO₂ ratio. I'm guessing that the relative uncertainty of NO₂ will be at least twice that of NO. The NO₂
chemiluminescent measurement is affected adversely by the relatively low photolysis efficiency. That accounts
for random errors in NO₂. Comparison with the cavity ringdown may take care of the NO₂ instrumental
340 background but how does it take care of the random errors?

Random errors are taken up by the calculation of a one-minute average before the instrumental background was
estimated from the difference in the two NO₂ measurements. Note that random errors in the data are expected to
be reduced as our final data analysis is based on a five minute average. NO₂ is calculated by

345
$$\text{NO}_2 = \frac{\text{NO}_c - \text{NO}}{K_e},$$

where NO_c is the signal of the channel equipped with the photolytic converter, NO the signal of the NO-channel
and K_e the NO₂ conversion efficiency. We agree that the relative uncertainty in the NO₂ has to be estimated by
means of the largest error possible from the relative uncertainties of NO (6 %), NO_c (6 %) and the conversion
350 efficiency K_e (3 %) rather than only from the uncertainty of the NO_c-signal and the uncertainty of the conversion
efficiency. The revised TMU for NO₂ follows as:

$$\Delta\text{NO}_2 = \sqrt{\Delta\text{NO}^2 + \Delta\text{NO}_c^2 + \Delta K_e^2} = \sqrt{6\%^2 + 6\%^2 + 3\%^2} = 9\%.$$

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We also agree that the absolute error of a single NO₂ data point might exceed 9 % when NO₂ is calculated from
two high numbers (e.g. if we assume 100 ppb_v NO and 101 ppb_v NO_c, NO₂ would be ~ 4 ppb_v), the error in NO₂
would be underestimated by 9 %. However the average and the median relative uncertainty of each data point of
the campaign at 5 minute integration time are 13.6 % and 11.8 %, respectively, which is indeed slightly higher
360 than the statistical estimate of 9 %. However the above average of 13.6 % yields a standard deviation of 6.9 %,
which yields that both the average and the median relative uncertainty of NO₂ are significantly not different from
the statistical estimate of 9 %. As we finally calculate median values for days and regions, it is more practicable
to estimate a statistical error for the NO₂ measurement at 9 %. The manuscript has been explicitly revised on
Page 7, L 158-160. Now it says: The TMU in NO₂ has been estimated by means of the largest error possible from

365 error propagation at

$$\Delta\text{NO}_2 = \sqrt{\Delta\text{NO}^2 + \Delta\text{NO}_c^2 + \Delta K_e^2} = \sqrt{6\%^2 + 6\%^2 + 3\%^2} = 9\%$$

at a confidence level of 1σ and an integration time of five minutes.

Also the TMU of NO₂ in Table 1 has been corrected to 9 %.

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Line 205 Total Measurement Uncertainty. All of these numbers appear very optimistic. In a previous comment, I
gave my reasons why the TMU of NO₂ appeared low. I do not know why the TMU of HO₂ is not larger than OH.
There are many sources of uncertainty in the conversion of HO₂ to OH which is the quantity actually measured.

375 The authors state a 20% bias due to RO₂ chemistry. I do not know to what extent quantification of HO₂ is made easier than that of OH because there is more HO₂ than OH and hence a larger signal.

380 The 1 sigma accuracy of OH is 20 % and the average precision is $3.4 \cdot 10^5 \text{ molec cm}^{-3}$. The accuracy of HO₂ is 20 % and the largest uncertainty due to interference by contribution of RO₂ is 7 % or 3 pptv, whichever is higher. The average precision of HO₂ is 3 pptv. The uncertainty in OH is estimated as the 1 sigma accuracy whereas the uncertainty in HO₂ is estimated at $\sqrt{(20 \%)^2 + (7 \%)^2} \approx 21 \%$. Note that HO₂ is calculated as the difference from two signals: $\text{HO}_2 = (\text{HO}_2 + \text{OH}) - \text{OH}$, whereas the signal (HO₂+OH) is about two order of magnitude larger than the signal from OH, which means that the uncertainty in OH can be neglected in the determination of the uncertainty in HO₂. On page 8, L 190-193 now it says: HO₂ data used in this study is still preliminary due to not yet corrected interference of organic peroxy radicals RO₂. The largest uncertainty due to interference by contribution of RO₂ is 7 % or 3 pptv, whichever is higher. The 1 sigma accuracy of both OH and HO₂ is 20 %. The uncertainty in the OH data is here estimated as the 1 sigma accuracy of the data set at 20 %, whereas the uncertainty in HO₂ is estimated at $\sqrt{(20 \%)^2 + (7 \%)^2} \approx 21 \%$. The measurement uncertainty for HO₂ has been corrected to 21 % in Table 1.

390 Comment 3: Line 272 “In low NO_x environments (< 100 pptv) previous studies have indicated that further NO oxidizing trace gases such as peroxy radicals (HO₂, RO₂) and halogen monoxides (XO) may result in a deviation from unity (Nakamura et al., 2003; Hosaynali Beygi et al., 2011; Reed et al., 2016).”

395 Deviations from the Leighton relation can also be important at higher NO. Departures of the Leighton ratio from unity depend (primarily) on the competition between HO₂+NO and O₃+NO. In polluted environments, HO₂ concentrations can increase, remain steady, or decrease only slowly as NO is increased.

400 Several studies address deviations from the Leighton Ratio. However in high NO_x regimes gaps between measurement and estimate can be closed by incorporating higher oxidations (HO₂, halogen monoxides). However, in low NO_x regimes, even addressing these further oxidants does not close the gap between observation and estimate (see for instance Hosaynali Beygi et al., 2011). We have added the following sentence Page 10, L 238: Deviations from expected NO/NO₂-ratios at low NO_x generally refer to missing oxidants converting NO to NO₂ (Hosaynali Beygi et al., 2011; Reed et al., 2016) or to a measurement error due to an instrumental background or a positive interference from thermal labile NO_x reservoir species (Reed et al., 2016; Silvern et al., 2018).

405 Line 250 and following. Does Fig. S1 show the ratio of actinic flux in the 4 hour window centered around noon to the total measured actinic flux? Or is Fig. S1 a ratio obtained by fitting a Gaussian, between zeroes in the AM and PM?

410 You are right, Figure S1 shows the ratio of the actinic flux in the 4 hour window (centered to noon) to the total measured actinic flux of that particular day. The Gaussian Fit was only used to estimate noontime of each particular day. We have revised caption of Figure S1. Now the first sentence of the caption of supplementary Figure S1 says: Ratio of the noontime actinic flux ($\pm 2\text{h}$ around noon) with regard to the total actinic flux $j(\text{NO}_2)$ of that particular day.

Comment 4: Lines 266-270, typos for O singlet D. Elsewhere, the “1” is in its proper place as a superscript.

415 Thank you for noticing. Now it says on Page 11, Line 269: $\text{O}(^1\text{D}) + \text{H}_2\text{O} \rightarrow 2\text{OH}$

Comment 5: Line 285. Am I correct that insofar as ozone production is concerned the only use of the HO₂ and OH measurements is their contribution to the loss rate of ozone (and hence their effect on net ozone production)?

420 HO_x data have been used to calculate net ozone production rates, however HO₂ has also been used as a reality test: The peroxy radical estimate based on Eq. 3 was compared to measured HO₂ in Section 3.2 (Page 21, L 442).

425 Comment 6: Line 373. “We find that the median NO_x(model)/NO_x(measurement)-ratio throughout the whole campaign is 0.91, indicating that the model underestimates NO_x by roughly 10 %.” The median of what? Could you please specify what items you are taking the median of; i.e., what are the data points. Ratios by Region? Days? Individual data points. I may have missed it; how long are data points? Equal to the 5 minute instrument averaging time?

430 The median NO_x(model)/NO_x(measurement) has been calculated as the median of five minute individual data points of the whole data set of the campaign. We have cleared out the misunderstanding in the text, now it says on Page 17, L386-387 following: We find that the median NO_x(model)/NO_x(measurement)-ratio of all five minute averaged data points of the whole campaign is 0.91, indicating that the model underestimates NO_x by roughly 10 %.

435

Comment 7: Line 391. “Noontime RO₂ was estimated based on Eq. 3. As the steady state assumption will not hold for air masses originating from fresh emissions (times to acquire steady state estimated from the inverse sum of the loss and production terms for NO₂ typically ranged from 1-2 minutes during AQABA) and for fast changes in the actinic flux,” What I think you want to say is mis-stated. As written, it says: We can’t use samples that had fresh emissions, so we used samples taken when actinic flux was slowly changing. Was the data screened to eliminate time periods in which NO_x (or less likely O₃) was rapidly varying? From the looks of the actinic flux plot you did not have many clouds giving rapid variations in jNO₂. I would be surprised if the time window around noon could not have been wider. How much does jNO₂ change between, say noon-3 hours and noon – 175 minutes and what change in HO₂ does that produce?

445 Please note that before performing photochemistry calculation, the whole data set was filtered based on a stack filter which was established based on parameters such as wind direction, wind speed, variability in NO data, O₃, SO₂ (see Lines 209-211). However rapid changes in the actinic flux j(NO₂) were observed e.g. after sunrise and before sunset. For ± 2h around noon the variation of the actinic flux (within the 4 hour period) was about 7 %, whereas the variation of the actinic flux within a 6-hour window (± 3h around noon) would be at least 20 % and assumptions that the actinic flux does not change within this time frame would not have been supportable. Differences in the actinic flux between noon-3 hours and noon-175 minutes would already be about 2 % of the maximum noontime value (and about the same change in HO₂). Differences in the actinic flux between noon-2 hours and noon-115 minutes would also be about 1-2 % of the maximum noontime value (slightly less than for 6 hours), but however expecting to decrease to zero towards noontime. Constant RO₂ and NOPR can be rather assumed for a 4-hour window around noon than for a 6-hour window around noon.

460 Comment 8: Figure 6. Obtaining peroxy radical concentrations from photostationary state calculations is not easy. Line 413 “the total uncertainty in the RO₂ estimates is estimated at 14%. This is a way too optimistic estimate of the uncertainty of RO₂. Instrument precision is too high and no account (except for NO₂) is taken of biases. Even still, the PSS is a difference in two numbers, often of comparable magnitude. I’m not certain the error propagation was correctly performed. I would like to see the formula that represents “errors added in quadrature”. Compare those results with a simple Monte Carlo calculation that can be done on a spreadsheet. And keep in mind that the result will not take into account errors in rate constants.

465 We agree that obtaining RO₂ from PSS calculations is not trivial and very much depends on the errors of the used quantities. The TMU in RO₂ was estimated based on Gaussian error propagation by means of the estimate of the largest error possible. RO₂ is calculated by

470
$$[RO_2] = \frac{j(NO_2) \cdot [NO_2] - k_{NO+O_3} \cdot [NO][O_3]}{k_{NO+HO_2} \cdot [NO]}.$$

The relative uncertainty in the RO₂ data has to be calculated from the relative uncertainty of NO, NO₂, O₃ and $j(\text{NO}_2)$ at roughly 6 %, 9 % (instead of 7 %), 2 % and 10 %, respectively, which yields a relative uncertainty of 15 % for RO₂.

$$\Delta \text{RO}_2 = \sqrt{\Delta \text{NO}^2 + \Delta \text{NO}_2^2 + \Delta \text{O}_3^2 + \Delta j(\text{NO}_2)^2} = \sqrt{6\%^2 + 9\%^2 + 2\%^2 + 10\%^2} \approx 15 \%$$

The relative uncertainty in RO₂ has been revised in the manuscript, now it says on Page 20 line 425-428: Based on the total measurement uncertainties of the measured quantities in Eq. 3, the uncertainty in RO₂ is estimated by means of the largest error possible at 15 %.

$$\Delta \text{RO}_2 = \sqrt{\Delta \text{NO}^2 + \Delta \text{NO}_2^2 + \Delta \text{O}_3^2 + \Delta j(\text{NO}_2)^2} = \sqrt{6\%^2 + 9\%^2 + 2\%^2 + 10\%^2} \approx 15 \% \quad (9)$$

Note that our calculation assumes that errors in the used rate coefficients are negligible.

Comment 9: Line 416. "Negative values for all regions are regularly found in the vicinity of fresh emissions and air masses not in photochemical equilibrium" That might be the explanation. It would be useful to quantify this point. HO₂ and RO₂ concentrations will be low in a high NO_x environment. The negative values may reflect the measurement accuracy needed to distinguish, for example, – 5 ppt RO₂ from zero.

Note that the uncertainty of the best estimates of RO₂ for each day/each region is estimated at 15 %. A reason why negative values were regularly found in the vicinity of fresh emissions is that the assumption of photostationary steady state (PSS) might not be fulfilled for fresh emissions. Time to acquire PSS was estimated as the inverse sum of the loss and production terms of NO₂ and was of the order of 1-2 minutes during AQABA, depending on the particular conditions (temperature, radiation, ozone concentration). The assumption that the regarded air masses were in PSS might not be true in the vicinity of fresh emissions (with transport time from source to instrument of less than 2 minutes). For these air masses production of NO₂ (oxidation of NO) prevailed over photolysis of NO₂ leading to $j(\text{NO}_2) \cdot [\text{NO}_2] - k_{\text{NO}+\text{O}_3} \cdot [\text{NO}][\text{O}_3] < 0$. Negative values might hence simply reflect that the regarded air mass has not reached PSS.

Comment 10: Figure 7. Here the RO₂ data looks much better. There are differences mentioned in the text, but for this reader could you please provide a concise reason why Fig. 7. looks so much better than Fig. 6. Is it the data groupings? I assume that in both Fig. 6 and 7. the blue RO₂ data is from Eq. 3. I am not totally positive because you were measuring HO₂ and some fraction of RO₂.

Figure 7 has been processed based on five minute data points in each region whereas Figure 6 has been processed for single days which might be more biased by single data outliers. Grouping into regions rather than for single days allows for a statistically more established estimate.

Comment 11: Line 421. As peroxy radicals are short-lived molecules generated from the oxidation of VOCs, enhanced RO₂ concentrations observed over the Arabian Gulf are most likely due to high VOC observed over the Arabian Gulf are most likely due to high VOC emissions from intense oil and gas activities in the region. High HO₂ can also occur in aged air masses in which NO_x and VOCs have reacted away but still have significant O₃ and (perhaps) HCHO. Photolysis could then yield peroxy radicals.

We have added the comment to the manuscript on Page 21, L. 438-439. However high HO₂ and RO₂ can also occur in aged air masses with low NO_x and VOCs but still significant O₃ (and perhaps HCHO whose photolysis would then yield peroxy radicals).

Comment 12: Line 440. Regarding the extrapolated actinic flux curve to get a daily ozone production. $j(1OD)$ decreases early and late in the day faster than $J(NO_2)$. Not sure how much difference it makes. I would be very leery of this extrapolation over land; I'm assuming that you are far enough away that varying traffic and boundary layer heights are not a concern.

Except for $j(O^1D) \cdot \alpha \cdot [O_3]$ all other terms will scale with $j(NO_2)$. However errors due to a different diurnal profiles of $j(O^1D)$ will remain insignificant with respect to the relative uncertainty of NOPR, as NOPR are mainly determined by $k \cdot [NO] \cdot [RO_2]$ and loss reactions due to OH and HO_2 .

Comment 13: Line 455 "the uncertainty of the regional NOPR is 40 % which has been estimated by error propagation" I don't disagree with this value. Merely surprised at its magnitude compared to a 15% uncertainty for RO_2 from Eq. 3.

The uncertainty in NOPR was estimated by means of the largest error possible and is mainly driven by the high uncertainty in OH and HO_2 (20 % and 21 %, respectively) and by the uncertainty in RO_2 (15 %). Diurnal net ozone production rates were estimated by scaling the median hourly noontime value to a diurnal value (by multiplying by the hourly value by 4 due to a 4-hour window around noon and by further dividing by 0.461 (noontime fraction ($\pm 2h$) of $J(NO_2)$). Strictly speaking, the relative uncertainty (6 %) in the division by 0.461, referred to as Δs (see Supplements Figure 1), also needs to be accounted for when estimating the uncertainty in NOPR. Also the error in α (Eq. 5) is rather determined by the error in H_2O (5 %) than from its relative variation over the campaign (relative variation ~ 21 %). This yields a measurement uncertainty for NOPR of

$$\Delta NOPR = \sqrt{\Delta NO^2 + \Delta NO_2^2 + \Delta O_3^2 + \Delta j(NO_2)^2 + \Delta j(O^1D)^2 + \Delta \alpha^2 + \Delta RO_2^2 + \Delta OH^2 + \Delta HO_2^2 + \Delta s^2}$$

$$= \sqrt{6\%^2 + 9\%^2 + 2\%^2 + 10\%^2 + 10\%^2 + 5\%^2 + 15\%^2 + 20\%^2 + 21\%^2 + 6\%^2} \approx 38 \%$$

We have corrected the error in NOPR from 40 % to 38 % in the manuscript and specified error calculation in detail also for NOPR in the manuscript.

On Page 12, L274 now it says: The error in α is mainly determined by the error of H_2O at 5 %.

On Page 21, L453 now it says: These noontime values are scaled to diurnal production rates (Figure 8). As photochemical net ozone production is in good approximation linear with actinic flux $j(NO_2)$ and as on average (46.1 ± 2.8) % of the total $j(NO_2)$ occurred $\pm 2h$ around noon, the median noontime NOPR estimate was multiplied by $4/0.461 \approx 8.68$ to obtain a diurnal value. The error in the total actinic flux located $\pm 2h$ around noon is estimated from the standard deviation of the best estimate of 0.461 at $\Delta s \approx 6$ %.

On Page 23, L 472 now it says: Based on the total measurement uncertainties of the measured quantities in Eq. 3, the systematic error in NOPR is estimated from error propagation by means of the largest error possible at 38 %.

$$\Delta NOPR = \sqrt{\Delta NO^2 + \Delta NO_2^2 + \Delta O_3^2 + \Delta j(NO_2)^2 + \Delta j(O^1D)^2 + \Delta \alpha^2 + \Delta RO_2^2 + \Delta OH^2 + \Delta HO_2^2 + \Delta s^2}$$

$$= \sqrt{6\%^2 + 9\%^2 + 2\%^2 + 10\%^2 + 10\%^2 + 5\%^2 + 15\%^2 + 20\%^2 + 21\%^2 + 6\%^2} \approx 38 \%$$

Comment 14: Line 473. "Although EMAC predicts high ozone levels over the Arabian Sea, it also reports the lowest NOPR in this region. Deviations between model-calculated estimate and the estimate based on measured

560 tracer data over the Mediterranean and over the Southern Red Sea could be linked to NO_x being overestimated in the model in these regions.” I’m not following. There is a low net ozone production rate. which to me implies that the model has a too low NO concentration, but you say that the model overpredicts NO_x.
As demonstrated in Figure 4, the model overpredicts the NO_x measurement over the Mediterranean and the Southern Red Sea. As these regions are classified as rather NO_x sensitive regimes (see Figure 10), higher NO_x (in the model) will likewise imply higher NOPR (under stagnant chemistry, i.e. NO_x sensitivity). The text is not fully clear and will be rewritten on P. 23 L492-495 as: Although EMAC predicts high ozone levels over the Arabian Sea, it also reports the lowest NOPR in this region. On the other side, the large overestimation of the model-calculated estimate NOPR against the one based on measured tracer data over the Mediterranean and over the Southern Red Sea could be linked to NO_x being overestimated in the model in these regions.

570

Comment 15: Section 3.4 VOC and NO_x sensitivity. Makes sense.
We have changed the title of chapter 3.4 to VOC- and NO_x-sensitivity.

575

Further changes

We have further made the following changes in the manuscript:

- 580 Change 1: There has been an updated version of the preliminary HO₂ data, which were used to revise NOPR in the different regions yielding 16 ppb_v day⁻¹ over the Oman Gulf (instead of 14 ppb_v day⁻¹), 32 ppb_v day⁻¹ over the Arabian Gulf (instead of 28 ppb_v day⁻¹) and -1 ppb_v day⁻¹ over the Mediterranean (instead of -2 ppb_v day⁻¹). The values have been corrected in the manuscript and the supplement, e.g. in Line 17 now it says: Net ozone production rates (NOPR) were greatest with 16 ppb_v day⁻¹ over both the Gulf of Oman and the Northern Red Sea and with 32 ppb_v day⁻¹ over the Arabian Gulf.
- 585 In the manuscript in Line 490 and Line 569 the values have also been revised. We have also revised the median noontime RO₂(measurement estimate)/HO₂(estimate)-ratio to 1.88 in Line 450-451: We find that the median noontime RO₂(measurement estimate)/HO₂(estimate)-ratio throughout the whole campaign is 1.88.
A new file with updated HO₂ data has been uploaded to zenodo and a new DOI has been generated. The data is now available at <https://doi.org/10.5281/zenodo.3693988>.
- 590 Change 2: Each end of the photolytic converter features 200 UV LED units instead of 50 UV LED units. In Line 122, 123 now it says: The photolytic NO₂ converter features a set of 200 UV LED units attached to each end of the converter.
- Change 3: The calculation of the UV-induced positive bias was revised, which reduced the estimated maximum interferences of BrONO₂, HONO, NO₃ and ClNO₂ to 6.1 %, 2.8 %, 2.7 % and 1.2 %, respectively (instead of 7.7 %, 7.2 %, 5.6 % and 1.5 % for HONO, BrONO₂, NO₃ and ClNO₂, respectively). Now it says in Line 124-127: The UV-induced positive bias in the NO₂-measurement due to photolysis of BrONO₂, HONO, NO₃ and ClNO₂ to produce NO was estimated at 6.1 %, 2.8 %, 2.7 % and 1.2 %, respectively, based on the absorption cross sections from the MPI-Mainz UV/VIS Spectral Atlas of Gaseous Molecules (Keller-Rudek et al., 2013).
- 600 Change 4: In the manuscript it says “supplementary Figure” and “supplementary Table” instead of “supplements Figure” and “supplements Table”, respectively. Also we now write “Figure” and “Table” instead of “Fig.” and “Tab.”, respectively, and we have revised the used date format in the manuscript to, e.g. 21 July 2017 instead of July 21st 2017.
- 605 Change 5: Table 1 was moved into the supplement. Figure S5 was also added to the supplement. All other figures and tables have been numbered accordingly in the manuscript and supplement.
- Change 6: In the caption of supplementary Figures S6-S9 now it says “The error bars represent the 25-75-percentile variation” instead of “The error bars are represented by the 25-75-percentile variation”
- 610 Changes made to the final manuscript and supplements are written in red.

Net ozone production and its relationship to NO_x and VOCs in the marine boundary layer around the Arabian Peninsula

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Abstract. Strongly enhanced tropospheric ozone mixing ratios have been reported in the Arabian Basin, a region with intense solar radiation and high concentrations of ozone precursors such as nitrogen oxides and volatile organic compounds. To analyze photochemical ozone production in the marine boundary layer (MBL) around the Arabian Peninsula, we use ship-borne observations of NO, NO₂, O₃, OH, HO₂, HCHO, actinic flux, water vapor, pressure and temperature obtained during the summer 2017 Air Quality and Climate in the Arabian Basin (AQABA) campaign, compare them to simulation results of the ECHAM-MESSy atmospheric chemistry (EMAC) general circulation model. Net ozone production rates (NOPR) were greatest with 16 ppb_v day⁻¹ over both the Gulf of Oman and the Northern Red Sea and with 32 ppb_v day⁻¹ over the Arabian Gulf. NOPR over the Mediterranean, the Southern Red Sea and the Arabian Sea did not significantly deviate from zero; however, results for the Arabian Sea indicate weak net ozone production of 5 ppb_v day⁻¹, and net ozone destruction over the Mediterranean and the Southern Red Sea with -1 ppb_v day⁻¹ and -4 ppb_v day⁻¹, respectively. Constrained by HCHO/NO₂-ratios, our photochemistry calculations show that net ozone production in the MBL around the Arabian Peninsula occurs mostly in NO_x-limitation regimes with a significant share of ozone production occurring in the transition regime between NO_x- and VOC-limitation over the Mediterranean and more significantly over the Northern Red Sea and Oman Gulf.

1 Introduction

Revenues from exploitation of the great oil reserves in the states of and around the Arabian Peninsula have propelled remarkable economic development associated with industrialization and urbanization. Strong population growth and anthropogenic emissions of gases and particulates in the last few decades have resulted in the Middle East becoming a hotspot for air pollution and associated health effects, while it is also one of the regions worldwide where climate change is particularly rapid (Lelieveld et al., 2016a). Unique meteorological conditions such as intense solar radiation, high temperatures and aridity, as well as strong anthropogenic emissions of volatile organic compounds (VOCs) and NO_x ($= \text{NO} + \text{NO}_2$) by on- and off-shore petrochemical industries, dense ship traffic, fossil energy production for air conditioning and desalination, and urban development are expected to further intensify in the future and contribute to photochemical ozone production (Lelieveld et al., 2009; Krotkov et al., 2016; Pfannerstill et al., 2019). Understanding the sources and sinks of NO_x and other ozone precursors on and around the Arabian Peninsula is therefore of major importance for atmospheric chemistry studies, including the investigation of net ozone production rates (NOPR) (Monks et al., 2015; Reed et al., 2016; Bozem et al., 2017).

NO_x plays a central role in atmospheric photochemistry (Nakamura et al., 2003; Tuzson et al., 2013; Reed et al., 2016). It is the primary precursor for tropospheric ozone (O_3), secondary organic aerosols and photochemical smog in urban areas (Hollaway et al., 2012; Javed et al., 2019). Main ground-based sources of NO and NO_2 are fossil fuel combustion and to a lesser extent bacterial processes in soils, and both lightning and aircraft emissions in the upper troposphere (Nakamura et al., 2003; Miyazaki et al., 2017; Javed et al., 2019). Transport of NO_x in the atmosphere is relatively limited due to its short lifetime of a few hours (Reed et al., 2016). It is removed from the troposphere mainly by conversion to HNO_3 (via reaction with OH) during the day, or the formation of N_2O_5 (in the reaction of NO_2 with NO_3 at night-time), which also leads to formation of nitric acid by heterogeneous hydrolysis on aerosol surfaces (Crutzen, 1973; Liu et al., 2016; Reed et al., 2016). Ultimately, the deposition of HNO_3 constitutes the major loss process of NO_x from the atmosphere. Ozone is a secondary pollutant that is photochemically formed in the troposphere from its precursors NO_x and VOCs (Bozem et al., 2017; Jaffe et al., 2018). It is an important greenhouse gas, an atmospheric oxidant and the most important primary precursor for OH (Lelieveld et al., 2004; Monks et al., 2015; Bozem et al., 2017). O_3 in the planetary boundary layer causes health damage, notably respiratory diseases, and reduces crop yields (Monks et al., 2015; Jaffe et al., 2018).

NO_x and O_3 mixing ratios in the troposphere vary from less than 20 ppt_v and 10 ppb_v, respectively, for pristine conditions such as the remote marine boundary layer (MBL) up to mixing ratios of several hundreds of ppb_v in regions with heavy automobile traffic and in international shipping lanes (for NO_x) and downwind of urbanized areas (for O_3) (Reed et al., 2016; Jaffe et al., 2018). Low NO_x environments such as the clean MBL and the lower free troposphere are considered net ozone destruction regimes whereas the upper troposphere and areas with anthropogenic emissions of ozone precursors are regions

of net ozone production (Klonecki and Levy, 1997; Bozem et al., 2017). Measurements performed in the the Houston Ship Channel revealed NOPR of the order of several tens of ppb h⁻¹ (Chen et al., 2010; Mao et al., 2010; Ren et al., 2013).

60 In the last decade much effort has been successfully devoted to the mitigation of NO_x emissions over Europe and America, and levels of reactive nitrogen trace gases have decreased (Miyazaki et al., 2017). But in Asia, India and the Middle East, NO_x emissions have substantially increased during the last decade so that the global NO_x burden has essentially remained constant (Miyazaki et al., 2017). NO_x emissions by ocean-going vessels have attracted considerable attention as they are reported to account for 15 % of the global NO_x emission burden (Celik et al., 2019). Model calculations suggest that the
65 Arabian Gulf, with an estimated annual NO_x emission density of about one ton km⁻² from ship traffic, is among the regions with highest NO_x emission densities worldwide (Johansson et al., 2017). Although NO_x emissions in the Red Sea and Arabian Sea areas were reported to be three and five times smaller than for the Arabian Gulf, respectively, these values are still 50-100 times larger than the emission density reported for the South Pacific Ocean, for example (Johansson et al., 2017).

In the present study, we characterize photochemical NOPR in the MBL around the Arabian Peninsula. In Sect. 2, the
70 campaign, instrument description, data processing and a description of the methods used in this study is presented. In Sect. 3, mixing ratios of nitrogen oxides and ozone around the Arabian Peninsula are reported. Based on concurrent measurements of HO_x, actinic flux, temperature and pressure, noontime RO₂ mixing ratios are estimated and used to calculate NOPR in the different regions around the Arabian Peninsula. Observation-based analysis of HCHO/NO₂-ratios will be used to distinguish between NO_x- or VOC-limited chemistry in the particular regions. A comparison of the results with data retrieved from the
75 3D global circulation model EMAC is also included.

2 Experimental

2.1 AQABA campaign

The AQABA ship campaign (Air Quality and Climate in the Arabian Basin) investigated the chemical composition of the MBL around the Arabian Peninsula. From late June to early September 2017, the *Kommandor Iona* Research and Survey
80 Vessel sailed from Toulon (France) to Kuwait and back in order to perform gas-phase and particle measurements in the region. The gas-phase and aerosol measurement instrumentation was housed in five laboratory containers on the front deck. A 6 m high, 20 cm diameter cylindrical stainless steel common inlet was installed on the front deck of the vessel to sample air at a total mass flow rate of 10,000 SLM. NO and NO₂ chemiluminescence measurements were obtained at a total bypass flow rate of 28.5 SLM sampling air from the common inlet with a residence time in the tubing of ~3 s. HCHO, NO₂ cavity
85 ring-down spectroscopy and O₃ measurements were obtained with similar bypass systems sampling air from the common inlet. H₂O vapor was measured on the top of the ship mast in the front. The OH and HO₂ detection units were placed on the prow to allow for inlets with residence times less than 10 ms.

The *Kommandor Iona* left Malta in late June 2017 traversing the Mediterranean Basin, the Suez Canal and the Northern Red Sea. A 3 day stop over at KAUST University (Saudi Arabia) was made from 11 July 2017 to 13 July 2017 before passing the Southern Red Sea area. On 17 July 2017, we briefly stopped at Djibouti port before passing the Gulf of Aden, the Arabian Sea and the Gulf of Oman. Kuwait at the northern end of the Arabian Gulf marked the turning point of the ship cruise where, during a second 3-day stop-over, scientific staff was exchanged. The *Kommandor Iona* started the second leg on 03 August 2017 arriving in Toulon (France) in early September 2017 without any further stops. Figure 1 shows the ship's route subdivided into six different regimes.

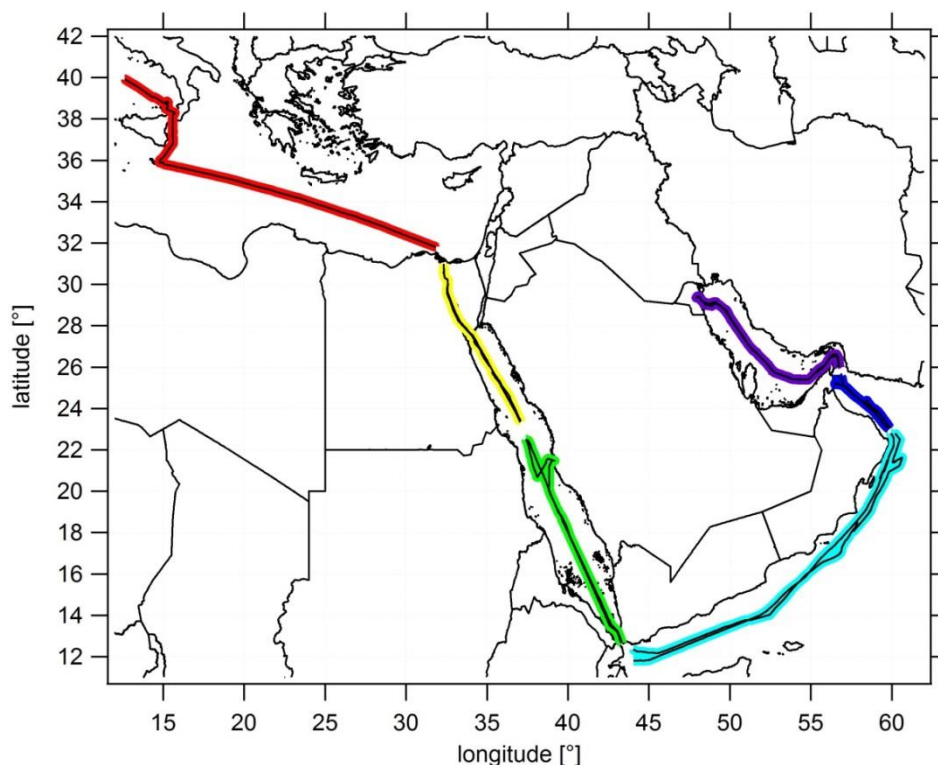


Figure 1: Ship cruises during both legs and color-coded subdivision into six different regimes. The following abbreviations will be used: AG for Arabian Gulf (purple), OG for Oman Gulf (dark blue), AS for Arabian Sea (blue), SRS for Southern Red Sea (green), NRS for Northern Red Sea (yellow), M for Mediterranean (red).

To enhance the statistical significance of our results and due to comparable signatures of the NO_x and O_3 measurements in the northern part of the Red Sea, the Suez Gulf and the Suez Canal, we have combined these regions which are represented by the 'Northern Red Sea' (NRS). For the same reasons we have merged the Gulf of Aden with the Arabian Sea (AS). See supplementary Table ST1 for the range of latitudinal and longitudinal coordinates of the different regions and supplementary Table ST2 for a detailed day to day description of the route.

2.2 Measurements of nitrogen oxides during AQABA

Chemiluminescent detection of NO and NO₂ is a widely applied method to quantify mixing ratios from the ppm_v down to the low ppt_v range (Nakamura et al., 2003; Pollack et al., 2011; Hosaynali Beygi et al., 2011; Reed et al., 2016). During AQABA we deployed a compact, robust and commercially available two-channel chemiluminescence instrument CLD 790 SR (ECO Physics AG, Dürnten, Switzerland) that has been optimized for in situ field measurements during the last decade (Hosaynali Beygi et al., 2011). The measurement principle of the CLD is based on the addition of O₃ to NO to produce stoichiometric quantities of excited state NO₂^{*} that will emit an infrared photon ($\lambda > 600$ nm) forming the chemiluminescent detection principle for NO (Drummond et al., 1985; Reed et al., 2016). Both channels feature an identical layout and were operated at a mass flow of 1.5 SLM during AQABA. One channel of the CLD (NO_c-channel) has additionally been equipped with a LED solid state photolytic converter (Droplet Measurement Techniques, Boulder, Colorado) installed upstream of the O₃ addition to selectively photolyze NO₂ to NO, which is subsequently measured. In this section, we will concentrate on modifications made prior to the campaign and especially on operational conditions of the photolytic converter during the campaign. Further details on the measurement principle are described elsewhere (Pollack et al., 2011; Hosaynali Beygi et al., 2011; Reed et al., 2016).

During AQABA, the cylindrical photolytic converter (length 14 cm, volume ~ 0.079 l) was operated at a constant pressure of 95 hPa yielding a residence time of ~ 0.3 s. The photolytic NO₂ converter features a set of 200 UV LED units attached to each end of the converter. The emission profile of the UV LED units was characterized in laboratory measurements to peak at 398 nm with a Full Width at Half Maximum (FWHM) of 16 nm. The UV-induced positive bias in the NO₂-measurement due to photolysis of BrONO₂, HONO, NO₃ and ClNO₂ to produce NO was estimated at 6.1 %, 2.8 %, 2.7 % and 1.2 %, respectively, based on the absorption cross sections from the MPI-Mainz UV/VIS Spectral Atlas of Gaseous Molecules (Keller-Rudek et al., 2013). These values represent upper limits for the interference of the respective NO_y compound as the respective molecular quantum yield was estimated conservatively at 1. Note that the values represent percent interferences if the interferent had the same concentration as NO₂. Due to small daytime concentrations of these molecules in the MBL, a UV-induced bias was neglected for the observations in this study. To limit wall loss of NO₂, the inner cavity surface is made of PTFE (polytetrafluoroethylene), which may potentially provide a reservoir (via surface adsorption) for NO_y that can thermally dissociate to increase the background signal of the NO₂ measurement (Reed et al., 2016). The conversion efficiency K_e of the photolytic NO₂ conversion was estimated by gas phase titration (SYCOS K-GPT-DLR, ansyco, Karlsruhe, Germany) several times before, during and after the campaign at (29.4 ± 0.9) % allowing the calculation of NO₂ concentrations by $[\text{NO}_2] = \frac{[\text{NO}_c] - [\text{NO}]}{K_e}$. To avoid chemical interferences due to adding ozone in excess during a gas phase titration, a small but not vanishing amount of NO has always been left unoxidized during gas phase titrations.

During AQABA, regular dry zero-air measurements as well as NO and NO₂ calibrations were performed autonomously over a 10 minute period every 6 hours to accurately quantify the instrumental background and to correct for sensitivity drifts. An autonomous cycle of ‘2 min zero air measurements – 2 min NO calibration – 2 min zero air measurement – 2 min NO₂ calibration – 2 min zero air measurement’ was implemented. Continuous flows NO and NO₂ calibration gases were added to the synthetic airflow or directed to a pump by switching solenoid valves. The NO calibration standard (1.954 ± 0.039 ppm_v NO in N₂, Air Liquide, Germany) used during the campaign was compared to a primary standard (5.004 ± 0.025) ppm_v (NPL, Teddington, UK) after the campaign yielding an effective NO mixing ratio of (2.060 ± 0.057) ppm_v in the NO calibration gas. Zero air measurements and NO calibrations were performed with a total flow of 3.44 SLM achieving an overflow of 0.44 SLM to guarantee ambient air free standard measurements. The calibration gas was added at 4.5 sccm to the zero air flow. During AQABA, NO calibrations at 2.5 ppb_v were achieved. During the first leg of the campaign, zero air was sampled from a bottle (Westfalen AG, Germany), whereas during the second leg zero air was generated from a zero air generator (Air Purifier CAP 180, acuraLine). Zero air measurements generated with the zero air generator were statistically not significantly different from those achieved by a bottle. To correctly account for the photomultiplier background and chemical interferences due to reactions of ozone with ambient alkenes additional pre-chamber measurements were performed every 5 minutes as well as at the beginning of zero air measurements and calibrations for 25 s each. This correction is removing a large fraction of the interference signal from alkenes. However, in regions where alkene concentrations are strongly varying in time and magnitude, the CLD is prone to enhanced backgrounds due to the interference of alkenes with ozone in the instrument. A schematic setup of the two-channel CLD instrument is given in Figure 2.

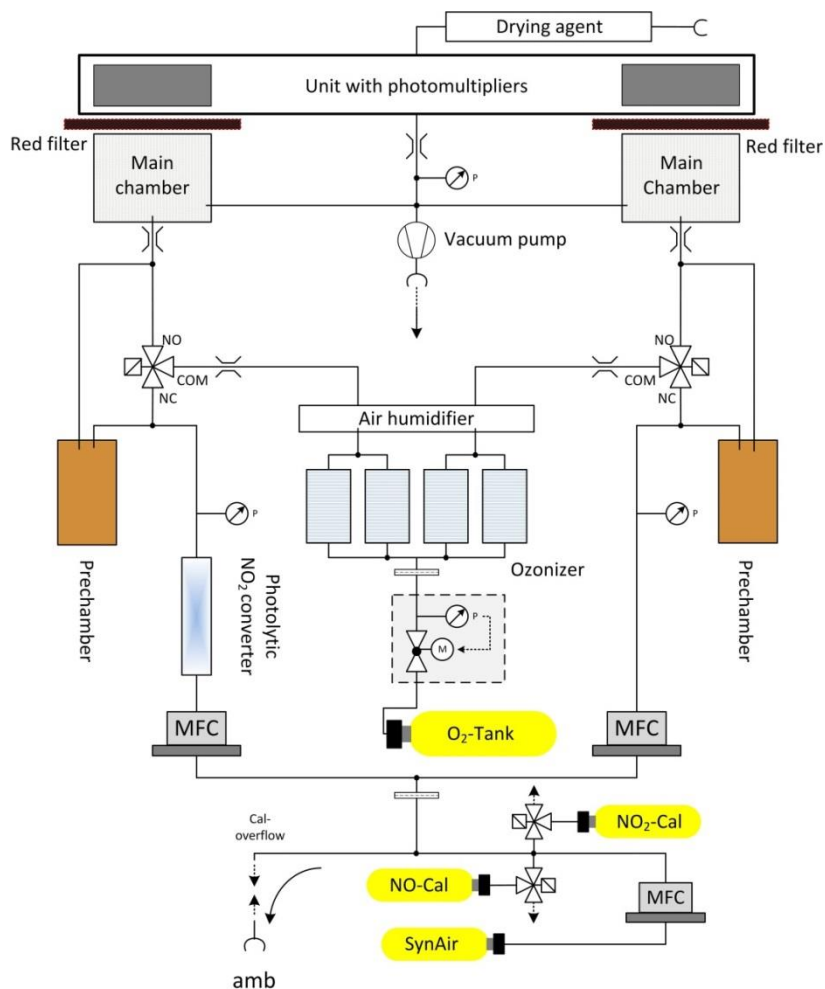


Figure 2: Schematic setup of the two channel CLD instrument in the configuration used during AQABA. NO and NO₂ calibration gases were running continuously and were added to the zero airflow by switching the respective solenoid valves.

The total measurement uncertainty (TMU) in the NO data is 5.5 % at a 5 min integration time and a confidence level of 1 σ . The limit of detection in the NO channel was estimated as the full width at half maximum of the frequency distribution of all zero air measurements obtained during the campaign to be 9 ppt_v at a 5 min integration time and a confidence level of 1 σ . The TMU in NO₂ has been estimated by means of the largest error possible from error propagation at

$$\Delta\text{NO}_2 = \sqrt{\Delta\text{NO}^2 + \Delta\text{NO}_c^2 + \Delta K_c^2} = \sqrt{6\%^2 + 6\%^2 + 3\%^2} = 9\%$$

at a confidence level of 1 σ and an integration time of five minutes. As the zero air measurements in the NO₂ channel produced an increased background affected by memory effects after exposure to high NO_x levels e.g. during measurements of stack emissions, the NO₂ raw data were initially processed without converter background subtraction. As we therefore expect the CLD NO₂ data to be offset due to not being initially background corrected, the converter background was estimated at 112 ppt_v from the centre of a Gaussian fit representing the difference of 1-minute averaged CLD NO₂ and

concurrent cavity ring-down spectroscopy (CRDS) NO₂ measurements for data points below 10 ppb_v. Setting the threshold for calculating the difference of the two concurrent data sets to 10 ppb_v is somewhat arbitrary, however, changing this limit to 5 ppb_v or 20 ppb_v does not significantly vary the estimated offset of the CLD NO₂ data. The offset correction of 112 ppt_v was taken as the ultimate absolute measurement uncertainty of the CLD NO₂ measurement. Further corrections of to the final CLD data include residence time corrections as well as corrections for NO and O₃ losses and the subsequent formation of NO₂ in the sampling line (Ryerson et al., 2000). Both NO and NO₂ CLD data have also been corrected for nonlinearities for concentrations higher than 55 ppb_v, as experienced during probing of stack emissions.

2.3 Further measurements used in this study

An extensive set of concurrent measurements providing mixing ratios of O₃, NO₂, HCHO, OH, HO₂, absolute humidity and actinic flux, temperature and pressure data obtained during AQABA was used in this study. Ozone was measured with an absorption photometer (Model 202 Ozone Monitor, 2B Technologies, Boulder, Colorado) based on the well-established absorption of the mercury line in the Hartley band at 254 nm (Viallon et al., 2015). Eliminating water and particle interferences during sampling was achieved via sampling through a nafion tube and a Teflon filter. The ozone monitor was zeroed ten times during the campaign. NO₂ was further measured by cavity ring-down spectroscopy (Sobanski et al., 2016) and used for correcting the instrumental background of the CLD NO₂ data, as described above (the correction was taken as the ultimate absolute measurement uncertainty in the CLD NO₂ data). Note that in this study we will use the NO₂ CLD data rather than the NO₂ CRDS data as the temporal coverage of the CLD NO₂ data over the course of the campaign is about 60 % compared to about 35 % for the cavity ring-down measurement. Formaldehyde (HCHO) was measured with an Aerolaser 4021 (AERO-LASER GmbH, Garmisch-Partenkirchen, Germany), which is a fully automatized monitor based on the Hantzsch technique (Kormann et al., 2003). H₂O measurements were obtained using a cavity ring-down spectroscopy monitor (PICARRO G2401, Santa Clara, California) supervised by Laboratoire des Sciences du Climat et de l'Environnement (LSCE) (Kwok et al., 2015). Measurements of OH and HO₂ were performed with the custom-built **HydrOxyl Radical measurement Unit** based on fluorescence **Spectroscopy** (HORUS) instrument based on laser-induced fluorescence (LIF) spectroscopy of the OH molecule and NO titration of HO₂ to OH followed by LIF spectroscopy detection of the OH molecule (Martinez et al., 2010; Regelin et al., 2013). **HO₂ data used in this study is still preliminary due to not yet corrected interference of organic peroxy radicals RO₂. The largest uncertainty due to interference by contribution of RO₂ is 7 % or 3 ppt_v whichever is higher. The 1 sigma accuracy of both OH and HO₂ is 20 %. The uncertainty in the OH data is here estimated as the 1 sigma accuracy of the data set at 20 %, whereas the uncertainty in HO₂ is estimated at**
 $\sqrt{20\%^2 + 7\%^2} \approx 21\%$. Wavelength resolved down-welling actinic flux was measured with a spectral radiometer (model CCD Spectroradiometer 85237). The *j*-values for NO₂ and O₃ were not corrected for upwelling UV radiation and were estimated to have a ~ 10 % measurement uncertainty (Meusel et al., 2016). The radiometer was installed 10 m above sea level, respectively 5 m above the front deck surface. Decreases in sensitivity due to sensor contamination with e.g. sea-spray

were corrected with a linear interpolation between two (daily) cleaning events. Temperature and pressure measurements were performed with the Shipborne European Common Automatic Weather Station (EUCAWS), a weather station specifically designed for ships. The weather station incorporates sensors, processing units, satellite positioning and communication systems in one device and is implemented and coordinated by the European National Meteorological Service EUMETNET. Table 1 lists the measurement methods and the TMU for each observation.

Table 1: List of observations and gas phase measurements during AQABA. The TMU at a confidence level of 1σ and at the particular temporal resolution as well as a reference of the measurement operability are given.

Molecule	Method	TMU	References
NO	chemiluminescence	6 %	Hosaynali Beygi et al., 2011
NO ₂	photolysis-chemiluminescence	9 %	Hosaynali Beygi et al., 2011
NO ₂	cavity ring-down spectroscopy	7 %	Sobanski et al., 2016
O ₃	UV absorbance	2 %	Viallon et al., 2015
OH	LIF	20 %	Martinez et al., 2010
HO ₂	NO titration / LIF	21 %	Martinez et al., 2010
HCHO	Hantzsch technique	13 %	Kormann et al., 2003
H ₂ O	cavity ring-down spectroscopy	5 %	Kwok et al., 2015
actinic flux	spectral radiometer	10 %	Meusel et al., 2016

The *Kommandor Iona* Research and Survey Vessel sailed whenever possible with the wind coming from the bow to avoid contamination by stack emissions. However, based on the relative wind direction, the variability in NO as well as the temporal evolution of NO_x, SO₂, and O₃ sections of data in which the air mass was contaminated by the ship's stack were identified. All data used here to calculate RO₂ and NOPR have been filtered to remove contaminated air masses. Altogether, 21 % of the sampling time was potentially contaminated by the ship exhaust of the KI of which 87 % occurred on the first leg. During the second leg the ship sailed against the wind and most of the data was free of stack contamination. Our analysis is based on a 5-minute running mean for each data set, whereby only averages that have been calculated at a temporal coverage greater than 30 % have been used. A time series of the NO, NO₂ (both CLD), O₃, OH, HO₂ preliminary and $j(\text{NO}_2)$ measurements is given in the supplementary Figures S2 and S3.

NO and NO₂ were measured from 03 July 2017 to 31 August 2017, O₃ was measured from 22 June 2017 to 01 September 2017, HCHO from 01 July 2017 to 31 August 2017 and OH and HO₂ from 18 July 2017 to 31 August 2017. For the analysis of peroxy radicals RO₂ and NOPR around the Arabian Peninsula we have removed data measured during the stop-overs in Jeddah (11 July to 13 July), Kuwait (31 July to 03 August) and during bunkering at Fujairah City (06 August, 07:00 – 15:00

225 UTC). Due to HO_x data being available from 18 July 2017 onward, we have limited the net ozone production analysis to the period after this date.

2.4 Methods

230 The so-called NO_x-O₃-null cycle represents a rapid daytime cycling between NO, NO₂ and O₃. Solar UV radiation photolyzes NO₂ to NO and O(³P) (R1) which will reform O₃ in the subsequent reaction with molecular oxygen O₂ (R2) (Leighton, 1961). NO and O₃ react to form NO₂ and O₂ (R3). R1, R2 and R3 constitute a so called null cycle which establishes photostationary steady state (PSS) for both NO_x and O₃ in mid latitudes during noon time on a time scale of ~100 s (Thornton et al., 2002; Mannschreck et al., 2004).



Under the assumption of PSS, the Leighton Ratio φ is unity (Leighton, 1961)

$$\varphi = \frac{j(\text{NO}_2) \cdot [\text{NO}_2]}{k_{\text{NO}+\text{O}_3} \cdot [\text{NO}][\text{O}_3]} = 1 \quad (1)$$

240 with $j(\text{NO}_2)$ being the NO₂ photolysis rate [s⁻¹]. In low NO_x environments (< 100 ppt_v) previous studies have indicated that further NO oxidizing trace gases such as peroxy radicals (HO₂, RO₂) and halogen monoxides (XO) may result in a deviation from unity (Nakamura et al., 2003; Hosaynali Beygi et al., 2011; Reed et al., 2016).



245 Deviations from expected NO/NO₂-ratios at low NO_x generally refer to missing oxidants converting NO to NO₂ (Hosaynali Beygi et al., 2011; Reed et al., 2016) or to a measurement error due to an instrumental background or a positive interference from thermal labile NO_x reservoir species (Reed et al., 2016; Silvern et al., 2018). In the present study we include HO₂ and R_iO₂ into the production term for NO₂.

$$j(\text{NO}_2) \cdot [\text{NO}_2] = k_{\text{NO}+\text{O}_3} \cdot [\text{NO}][\text{O}_3] + k_{\text{NO}+\text{HO}_2} \cdot [\text{NO}][\text{HO}_2] + [\text{NO}] \cdot \sum_i k_{\text{NO}+\text{R}_i\text{O}_2} \cdot [\text{R}_i\text{O}_2] \quad (2)$$

250 Assuming that the temperature-dependent rate coefficient for the reaction of each particular peroxy radical R_iO_2 with NO equals the rate k_{NO+HO_2} for Reaction R4 (Hauglustaine et al., 1996; Cantrell et al., 1997; Thornton et al., 2002), we can combine HO_2 and the sum of all organic peroxy radicals R_iO_2 to the entity RO_2 that can be estimated using the steady state equation

$$[RO_2] = \frac{j(NO_2) \cdot [NO_2] - k_{NO+O_3} \cdot [NO][O_3]}{k_{NO+HO_2} \cdot [NO]}. \quad (3)$$

255 However, the steady state assumption is not valid if the sampled air parcel is affected by fresh emissions or fast changes in the actinic flux (Thornton et al., 2002). After sampling a fresh emission e.g. a ship plume, for which NO_x went up typically to values of several tens of ppb_v with simultaneous titration in O_3 , we assume that PSS is re-established on a time scale of 2 minutes (Thornton et al., 2002; Mannschreck et al., 2004). To best approximate PSS in our analysis we have restricted the estimation of RO_2 on time frames ± 2 h around noontime for which we expect the smallest relative changes in the actinic flux. Noontime for each day was determined as the centre of a Gaussian fit that was applied to the actinic flux data. We applied a Gaussian Fit to the actinic flux data as this fitting method is sufficient to estimate the centre of the diurnal actinic flux. To further limit the effect of periods for which PSS is not fulfilled, we use the median instead of the average that is often disproportionately biased by strong NO_x sources nearby. See [supplementary Tables ST3, ST5 and ST7](#) for detailed statistics and a further motivation on regional averages and median values. See [supplementary Figure S1](#) for a detailed illustration of the calculation of the fraction of the noontime integral.

A further part of the analysis will be the investigation of NOPR. Ozone production is initiated by reactions that produce HO_x , for which primary production is from the photolysis of ozone, formaldehyde, nitrous acid (HONO) and hydrogen peroxide (H_2O_2) (Thornton et al., 2002; [Lu et al., 2010](#); Hens et al., 2014; Mallik et al., 2018). The production of ozone can be approximated by the rate of oxidation of NO with RO_2 ($HO_2 + \sum_i R_iO_2$) to form NO_2 that will rapidly form O_3 (R1-R2) (Bozem et al., 2017). For RO_2 we use the result from Eq. 3 that incorporates HO_2 and the sum of all further peroxy radicals $\sum_i R_iO_2$ (Parrish et al., 1986; Thornton et al., 2002).

$$P(O_3) = k_{NO+HO_2} \cdot [NO][RO_2] \quad (4)$$

Photochemical O_3 loss is mainly due to photolysis ($\lambda < 340$ nm) in the presence of water [vapor](#) and the reactions of ozone with OH and HO_2 (Bozem et al., 2017).



α , the fraction of $O(^1D)$ that reacts with H_2O

$$\alpha = \frac{k_{O(^1D)+H_2O}[H_2O]}{k_{O(^1D)+H_2O}[H_2O] + k_{O(^1D)+M}[M]} \quad (5)$$

was (10.6 ± 2.2) % during AQABA with a quasi linear dependence on water concentrations. The error in α is mainly determined by the error of H_2O at 5 %. Furthermore, ozone is lost due to reactions with alkenes (R12) and halogen radicals (R13).



We find that the loss rate is dominated by the photolysis of ozone with subsequent reaction of $O(^1D)$ with H_2O , was 60 – 80 % of the total loss rate, followed by the reaction of O_3 with HO_2 , which makes up 10 – 30 % (note that the uncertainty in HO_2 radical concentrations mentioned above has no significant influence on the total O_3 loss rate, due to its small contribution). The remaining fraction (10-30 %) is due to the reaction of O_3 with OH . The reaction of ozone with ethene is on average $0.005 - 0.01$ ppbv h^{-1} and therefore generally less than 2 % of the total ozone loss rate (Bourtsoukidis et al., 2019). The reaction of O_3 with all alkenes will hence be neglected. Halogen radicals were not measured during AQABA and will not be incorporated into our study. Based on oxidative pairs, Bourtsoukidis et al. (2019) have classified the majority of their samples collected during AQABA by an OH/Cl -ratio of 200:1. As measured daytime OH concentrations were of the order of $5 \cdot 10^6$ molecule cm^{-3} , the estimate would yield a Cl concentration of $2.5 \cdot 10^4$ molecule cm^{-3} , which would decrease the estimated diurnal net ozone production rates by roughly 0.2 ppbv day^{-1} over the Arabian Sea and at most 0.6 ppbv day^{-1} over the other regions, which does not substantially alter the here presented results. The noontime chemical ozone loss rate can be summarized by

$$L(O_3) = [O_3] \cdot (\alpha \cdot j(O^1D) + k_{OH+O_3} \cdot [OH] + k_{HO_2+O_3} \cdot [HO_2]). \quad (6)$$

NOPR presented in this study is finally calculated as the difference of Eq. 4 and Eq. 6.

$$NOPR = k_{NO+RO_2}[NO][RO_2] - [O_3] \cdot (\alpha \cdot j(O^1D) + k_{OH+O_3}[OH] + k_{HO_2+O_3}[HO_2]). \quad (7)$$

Under the assumption of constant chemical composition for a given day, the NOPR is expected to have a diel cycle following the measured actinic flux. Hence integrating the estimated NOPR over the course of a day based on the particular fractional noontime integral of $j(NO_2)$ will yield a diurnal value for NOPR. A detailed calculation of the diurnal fractional integrals is given in the supplementary Figure S1. Note that all reaction rate constants used are from the IUPAC Task Force on Atmospheric Chemistry Chemical Kinetic Data Evaluation (Atkinson et al., 2004). Indications whether a chemical regime is NO_x -limited or VOC-limited can be derived from the ratio of $HCHO$ to NO_2 . Former studies have derived $HCHO/NO_2$ -

ratios from satellite measurements to establish whether ozone production is NO_x-limited or VOCs-limited. The results indicate NO_x-limitation for HCHO/NO₂ > 2 and prevailing VOC-limitation for HCHO/NO₂ < 1 (Duncan et al., 2010).

2.5 ECHAM/MESSy Atmospheric Chemistry (EMAC) model

EMAC is a 3D general circulation model that includes a variety of sub-models to describe numerous processes in the troposphere, their interaction with oceans and land surfaces and incorporates anthropogenic influences. Here we use the second development cycle of the Modular Earth Submodel System (MESSy2) (Jöckel et al., 2010) and ECHAM5 (Röckner et al., 2006) which is the fifth generation European Centre Hamburg general circulation model in the T106L31 resolution (corresponding to a quadratic grid of roughly 1.1° and 1.1°). The model has 31 vertical pressure levels and involves the complex organic chemistry mechanism MOM (Mainz Organic Mechanism) as presented by Sander et al. (2019) that includes further developments of the version used by Lelieveld et al. (2016b). Here we use the lowest pressure level in a terrain following coordinates (equivalent to the surface level) and simulations of NO, NO₂, O₃, OH, HO₂, *j*(NO₂) and *j*(O¹D). The sum of peroxy radicals was estimated as the sum of all radicals R_iO₂ with less than four carbon atoms. Net ozone production based on data retrieved from EMAC was estimated as

$$\text{NOPR} = [\text{NO}] \cdot (k_{\text{NO}+\text{HO}_2}[\text{HO}_2] + \sum_i k_{\text{NO}+\text{R}_i\text{O}_2}[\text{R}_i\text{O}_2]) - [\text{O}_3] \cdot (\alpha \cdot j(\text{O}^1\text{D}) + k_{\text{OH}+\text{O}_3}[\text{OH}] + k_{\text{HO}_2+\text{O}_3}[\text{HO}_2]). \quad (8)$$

A list of all included peroxy radicals R_iO₂ for the reaction with NO is given in the supplementary Table ST10.

3 Results and discussions

3.1 NO_x and O₃ in the MBL around the Arabian Peninsula

During AQABA NO_x mixing ratios varied over five orders of magnitude with lowest values of less than 50 ppt_v observed in relatively pristine regions and highest values of several hundred ppb_v found in the vicinity of megacities or nearby passing ships. Ozone mixing ratios ranged from values of less than 20 ppb_v, detected over the Arabian Sea, to more than 150 ppb_v during episodes of severe pollution. Figures 3a) and 3b) show distributions of NO_x measured during the first and second leg of the campaign (range from 0.1 ppb_v to 20 ppb_v) while Figure 3c) and 3d) show corresponding ozone mixing ratios covering a range from 20 ppb_v to 100 ppb_v, respectively. A classification of the different regions based on Box-Whisker-Plots, including the 25-75-percentile interval (box) and whiskers for the 10-90-percentile interval, is shown in Figure 4 and Figure 5 for NO_x and O₃, respectively. As average NO_x is often influenced by fresh, localized emissions, we have included the median (black bar) instead of the average in the Box-Whisker-Plot for NO_x, which is less sensitive to extreme values. For O₃, although the difference between median and mean is mostly negligible, we also use the median in Figure 5. NO_x and O₃ averages, medians, standard deviations, 1st and 3rd quantiles and the number of data points quantified per region are given in the supplementary Table ST3. See supplementary Figure S4 for OH and preliminary HO₂ mixing ratios around the Arabian Peninsula. Supplementary Figure S5 shows the variation of the absolute humidity around the Arabian Peninsula.

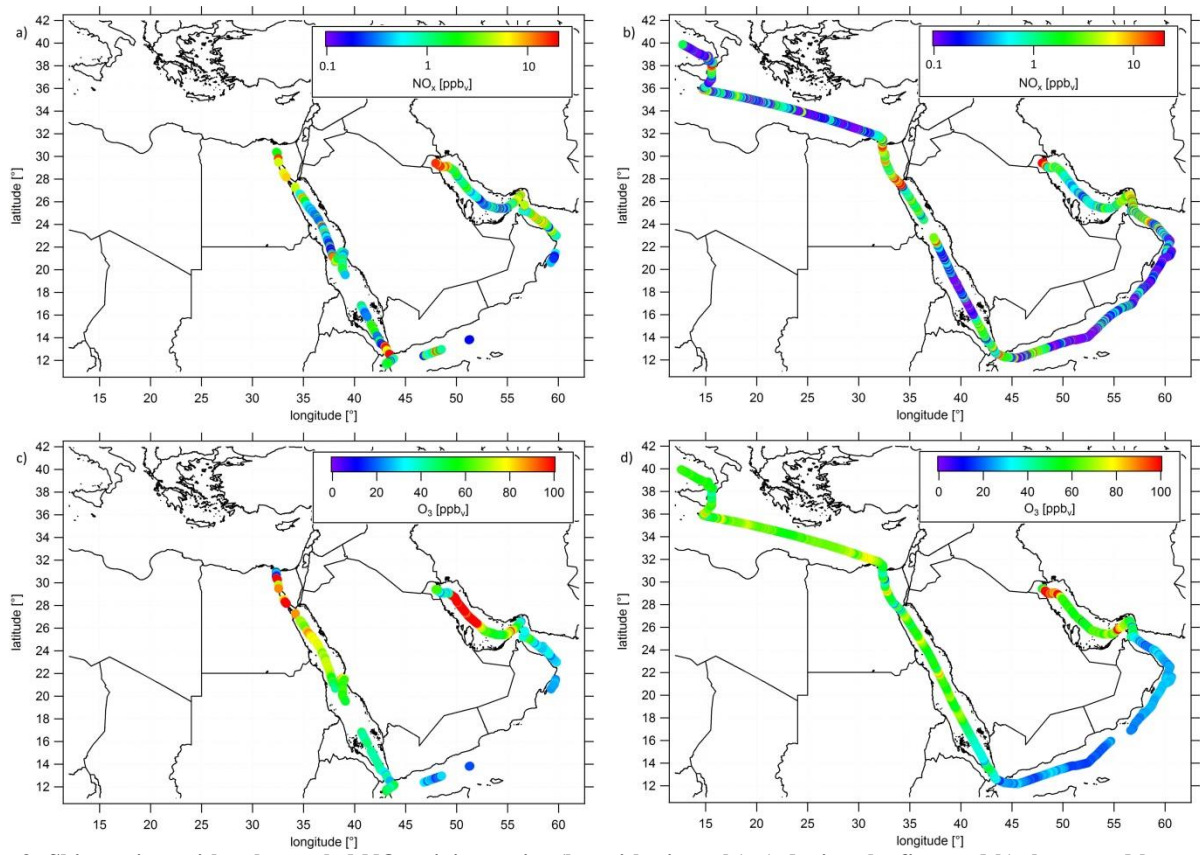


Figure 3: Ship cruises with color-scaled NO_x mixing ratios (logarithmic scale) a) during the first and b) the second leg and color-scaled O₃ mixing ratios (linear scale) c) during the first and d) during the second leg. Note that both NO_x and O₃ has been filtered for own stack contamination.

340

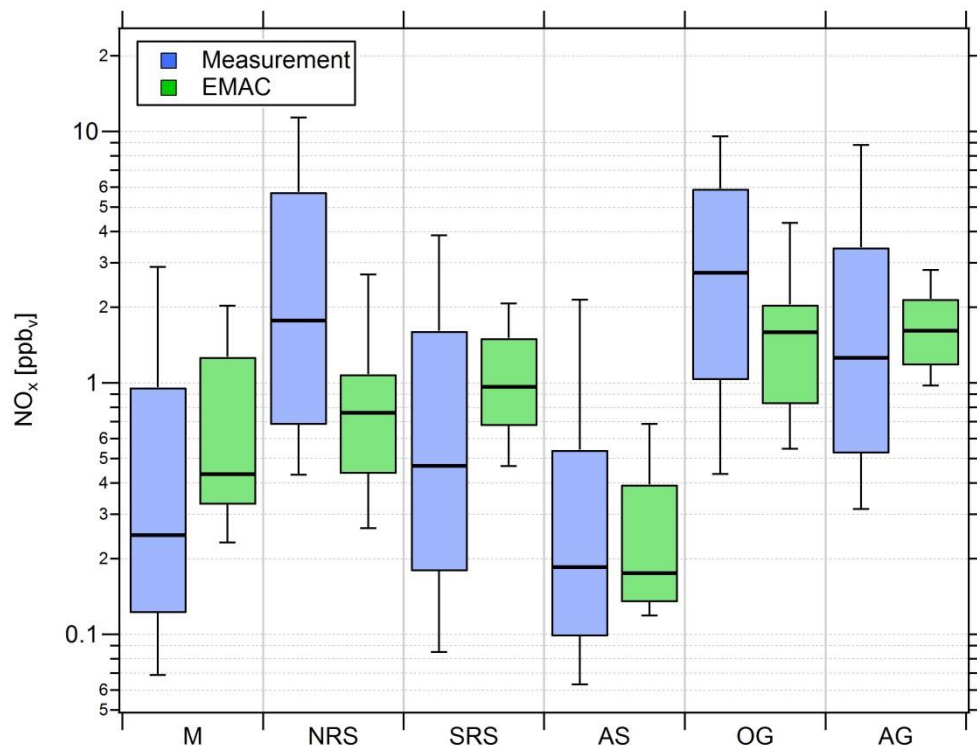


Figure 4: Comparison of measured (blue) and simulated (green) NO_x mixing ratios in the six different regions investigated during AQABA. The horizontal black bar indicates the median value, the box the 25- and 75-percentiles and the whiskers the 10- and 90-percentiles.

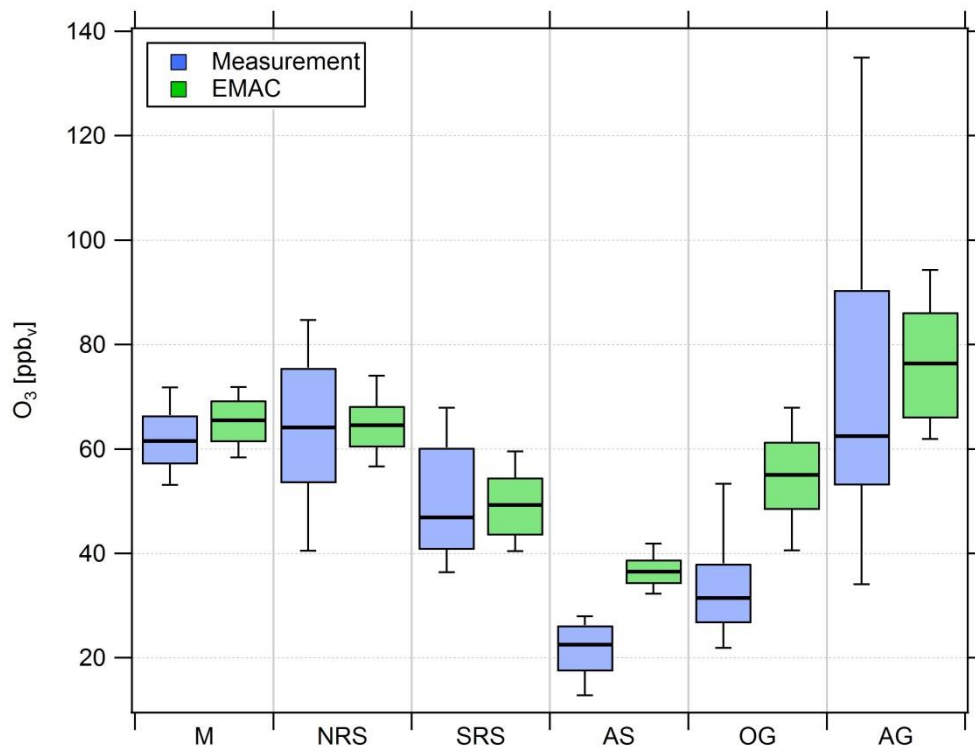


Figure 5: Comparison of measured (blue) and simulated (green) O_3 mixing ratios in the six different regions investigated during AQABA. The horizontal black bar indicates the median value, the box the 25- and 75-percentiles and the whiskers the 10- and 90-percentiles.

Overall, we find that NO_x mixing ratios over the Northern Red Sea, the Gulf of Oman and the Arabian Gulf are approximately one order of magnitude higher than in the other three regions (Southern Red Sea, Arabian Sea, Mediterranean). NO_x medians over the Arabian Gulf, the Northern Red Sea and the Gulf of Oman are 1.26 ppb_v, 1.76 ppb_v and 2.74 ppb_v, respectively. Lower median NO_x mixing ratios were measured over the Southern Red Sea (0.46 ppb_v), the Mediterranean (0.25 ppb_v) and the Arabian Sea (0.19 ppb_v). With respect to observed O_3 mixing ratios, the Arabian Sea is the only region representing remote MBL conditions with lowest median and average O_3 of 21.5 ppb_v and 22.5 ppb_v respectively, followed by the Gulf of Oman where median and mean O_3 were 31.5 ppb_v and 34 ppb_v, respectively. The low O_3 mixing ratios over the Arabian Sea were accompanied by the smallest variability (whisker-interval: 15.1 ppb_v). Although observing highest NO_x over the Oman Gulf, O_3 observed over the Oman Gulf was amongst the lowest detected throughout the whole campaign, which can be partly explained by the fact that high NO_x lead to low ozone production or even net ozone destruction. However, a significantly larger whisker-interval of observed ozone of 31.4 ppb_v over the Gulf of Oman indicates increasing amounts of pollution and advection from the Arabian Gulf where extreme events of ozone were observed several times during the campaign with maximum mixing ratios of up to 170 ppb_v when wind was coming from Kuwait/Iraq. Please note that during the second leg wind was coming from Iran (Pfannerstill et al., 2019). The whisker-interval over the Arabian

Gulf was 100.9 ppb_v, more than six times higher than that over the Arabian Sea. Reasons for large variations of both NO_x and O₃ over the Arabian Gulf were a multitude of point sources as well as a change in the observed wind direction with air masses coming from Iraq/Kuwait area during the first leg and air masses coming from Iran during the second leg (Pfannerstill et al., 2019). Over the Mediterranean, the Northern Red Sea and the Southern Red Sea, median ozone was 61.5 ppb_v, 64.2 ppb_v and 46.9 ppb_v, respectively. The whisker-intervals over the Northern Red Sea and the Southern Red Sea were 44.2 ppb_v and 31.6 ppb_v, respectively. Air masses over the Mediterranean were characterized as photochemically aged due to their impact by northerly winds (Etesians) which bring processed/oxidized air from eastern Europe (Turkey, Greece) to the Mediterranean area (Derstroff et al., 2017; Pfannerstill et al., 2019). This photochemical ageing/oxidation over the Mediterranean leads to a rather small whisker-interval of 18.7 ppb_v in ozone. In summary, median NO_x over the Oman Gulf was 56 % and 117 % higher than over the Northern Red Sea and the Arabian Gulf, respectively. However, the highest NO_x average was measured over the Northern Red Sea at 4.69 ppb_v, similar to the values observed over the Oman Gulf (4.16 ppb_v) and the Arabian Gulf (3.65 ppb_v). Note that highest NO_x mixing ratios over the Oman Gulf and over the Northern Red Sea are not always associated with high O₃ mixing ratios. We find that average ozone was highest over the Arabian Gulf with 74 ppb_v followed by the Northern Red Sea region (63.4 ppb_v). The average ozone mixing ratio over the Oman Gulf was 34 ppb_v, which corresponds to 46 % of the value observed over the Arabian Gulf. Photochemically aged air masses over the Mediterranean Basin show an ozone average of 61.6 ppb_v and air masses encountered over the Northern Red Sea (O₃ median of 64.2 ppb_v, O₃ average of 63.4 ppb_v) are comparable to the Arabian Gulf.

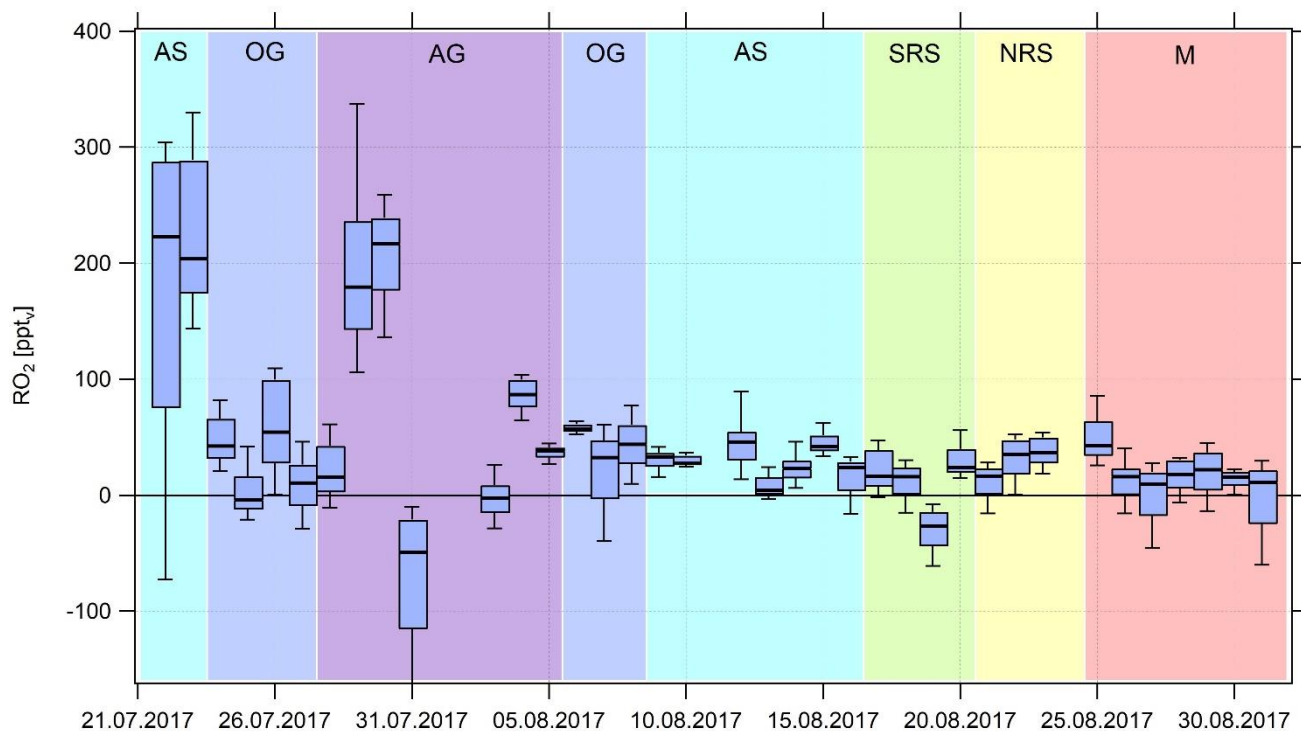
Due to a number of large pollution sources in the region around the Arabian Peninsula such as passing ships, highly urbanized areas as well as on- and off-shore petrochemical processing, NO_x levels were rarely as low as those found in remote locations such as over the South Atlantic (Fischer et al., 2015) where NO_x levels may be under 20 ppt_v. Apart for a few occasions where NO_x was below 50 ppt_v for short periods (Arabian Sea, the Southern Red Sea and the Mediterranean), NO_x levels during AQABA generally ranged from 100 ppt_v up to several ppb_v. The campaign NO_x median of 0.65 ppb_v and mean value of (2.51 ± 5.84) ppb_v is comparable to urban sites (Kleinman et al., 2005). A detailed emission density analysis performed by Johansson et al. (2017) shows that NO_x emissions on and around the Arabian Peninsula are amongst the highest worldwide, which could explain the rather high NO_x level in the MBL around the peninsula (Johansson et al., 2017; Pfannerstill et al., 2019). O₃ mixing ratios measured during AQABA were also very variable with O₃ mixing ratios ranging between less than 20 ppb_v in the remote MBL (Fischer et al., 2015) to 60-70 ppb_v in the Mediterranean (consistent with previous ship-based measurements in the region (Kouvarakis et al., 2002) and as high as 150 ppb_v measured over the Arabian Gulf region. The latter are consistent with O₃ mixing ratios reported from regions influenced by oil and gas processing (Pfannerstill et al., 2019) and shipping lanes such as the Houston Ship Channel (Mazzuca et al., 2016).

Figure 4 also shows that the general trend for NO_x mixing ratios in the different regions is widely reproduced by the EMAC model. We find that the median NO_x(model)/NO_x(measurement)-ratio of all five minute averaged data points of the whole

campaign is 0.91, indicating that the model underestimates NO_x by roughly 10 %. The average ratio and its standard deviation are significantly larger at 2.57 and 5.71, respectively, indicating that single modeled data points strongly exceed the measurements, especially during periods of low in situ NO_x (see supplementary Figure S6). Particularly over the Arabian Sea and the Southern Red Sea, the model generally simulates NO_x mixing ratios higher than 100 and 200 ppt_v, respectively while the measurements indicate mixing ratios of less than 50 ppt_v for certain periods. Furthermore, as expected, the model is not able to reproduce point sources such as passing ships for which we observe a significant underestimation of the measured NO_x . For ozone we find that the median $\text{O}_3(\text{model})/\text{O}_3(\text{measurement})$ -ratio throughout the campaign is 1.23, indicating that over the course of the campaign the model overestimates O_3 by about 23 %. This could partly be related to the same limitation, i.e. the inability of the model to resolve point sources in which O_3 is locally reduced due to titration by NO. While the model is in rather good agreement with the measurements over the Mediterranean, the Northern Red Sea and Southern Red Sea, large deviations are found over the Arabian Sea and the Oman Gulf, where the model overestimation with respect to the regional median is 63 % and 75 %, respectively. A possible explanation for the overestimation of both ozone and NO_x in pristine regions such as over the Arabian Sea and the Oman Gulf could be related to the model resolution of $1.1^\circ \times 1.1^\circ$. Interpolation of model simulations along the *Kommandor Iona* ship track close to the coast at this resolution will most likely incorporate contributions from nearby land areas, affected by anthropogenic emissions. See supplementary Table ST3 and Table ST4 for further information and Figure S6 and S7 for additional scatterplots of measured and simulated regional median NO_x and O_3 , respectively.

3.2 Estimation of RO_2 around the Arabian Peninsula

Noontime RO_2 was estimated based on Eq. 3. As the steady state assumption will not hold for air masses originating from fresh emissions (times to acquire steady state estimated from the inverse sum of the loss and production terms for NO_2 typically ranged from 1-2 minutes during AQABA) and for fast changes in the actinic flux, we have calculated Box-Whisker-Plots for ± 2 h around noontime for which we expect relatively minor changes in the actinic flux (Figure 6). The noontime of each day was approximated by applying a Gaussian fit routine to the measured $j(\text{NO}_2)$ values whereas $j(\text{NO}_2)$ values being less than 10^{-3} s^{-1} were neglected. Due to the availability of OH and HO_2 data from 18 July 2017 onwards, we have limited the analysis to this period. Note that there are no noontime RO_2 estimates from 18 July to 21 July due to contamination by the ship exhaust and on 24 August 2017 due to missing data. The black bar in Figure 6 indicates the median value, with the Box-interval marking the 25- and 75-percentile and the whisker showing the 10- and 90-percentile. Figure 7 shows summarized regional trends of the RO_2 estimates for measured and simulated data.



425 **Figure 6: Timeline of median RO₂ noontime estimates from 22 July to 31 August 2017. Due to contamination by the ship exhaust itself, there is no data from 18 July to 21 July 2017. See annotations for the classification of the different regions.**

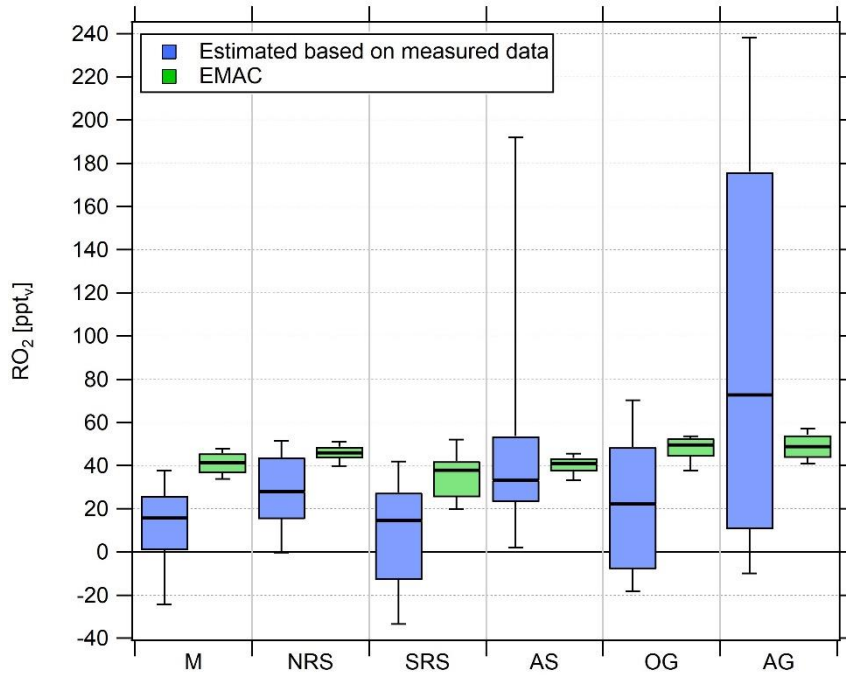


Figure 7: Comparison of Box-Whisker-Plots of the regional estimated noontime RO₂ median based on estimated and simulated data for the period from 18 July 2017 onwards.

We find median noontime RO₂ mixing ratios over the Mediterranean, the Northern Red Sea, the Southern Red Sea, the Arabian Sea and Oman Gulf of 16 ppt_v, 28 ppt_v, 15 ppt_v, 33 ppt_v and 22 ppt_v, respectively, with each respective 75-percentile RO₂ being equal or less than 54 ppt_v. Based on the total measurement uncertainties of the measured quantities in Eq. 3, the uncertainty in RO₂ is estimated by means of the largest error possible at 15 %.

$$\Delta RO_2 = \sqrt{\Delta NO^2 + \Delta NO_2^2 + \Delta O_3^2 + \Delta j(NO_2)^2} = \sqrt{6\%^2 + 9\%^2 + 2\%^2 + 10\%^2} \approx 15 \%$$

Note that our calculation assumes that errors in the used rate coefficients are negligible. Only over the Arabian Gulf, the RO₂ estimate yields a median noontime mixing ratio of 73 ppt_v accompanied by the largest variations in the box-interval of the whole campaign. While the box-interval of the RO₂ estimate in the other regions is 25-57 ppt_v, the box-interval over the Arabian Gulf is significantly higher at 165 ppt_v. Negative values for all regions are regularly found in the vicinity of fresh emissions and air masses not in photochemical equilibrium. The elevated 90-percentile over the Arabian Sea is due to high RO₂ estimates during the first leg on 22 and 23 July.

Estimated RO₂ mixing ratios based on measured tracer data are in general agreement with previous studies performed in marine boundary layer environments which report maximum mixing ratios between 30 and 55 ppt_v around noontime

445 (Hernandez et al., 2001). As peroxy radicals are short-lived molecules generated from the oxidation of VOCs, enhanced RO₂ concentrations observed over the Arabian Gulf are most likely due to high VOC emissions from intense oil and gas activities in the region (Bourtsoukidis et al., 2019; Pfannerstill et al., 2019). However high HO₂ and RO₂ can also occur in aged air masses with low NO_x and VOCs but still significant O₃ (and perhaps HCHO whose photolysis would then yield peroxy radicals). Bourtsoukidis et al. report that spatial volume mixing ratios of ethane and propane over the Arabian Gulf were
450 about a factor of 10-15 times higher than over the Arabian Sea and the Southern Red Sea (Bourtsoukidis et al., 2019). We find that the median noontime RO₂(measurement estimate)/HO₂(measurement)-ratio throughout the whole campaign is 1.88. Note that during single days, HO₂ may be higher than the RO₂ estimate, which is within the uncertainty of the RO₂ estimate.

EMAC modelled, median noontime RO₂ mixing ratios estimated as the sum of simulated HO₂ and all simulated peroxy radicals with less than four carbon molecules are 41 ppt_v, 46 ppt_v, 38 ppt_v, 41 ppt_v, 50 ppt_v and 49 ppt_v over the
455 Mediterranean, the Northern Red Sea, the Southern Red Sea, the Arabian Sea, the Oman Gulf and the Arabian Gulf, respectively. The observation based RO₂ estimate yields 16 ppt_v, 28 ppt_v, 15 ppt_v, 33 ppt_v, 22 ppt_v and 73 ppt_v respectively. We find that the median point by point RO₂(model)/RO₂(measurement estimate)-ratio from 18 July onward is 1.05 so that, on average, the model overestimates the measurement by 5 %. Please note that the observational variability is much higher than the modeled one and that the median of 1.05 is accompanied by a larger average (1.84) and a large variability (42.51).
460 See supplementary Table ST5 and ST6 for further information and Figure S8 for an additional scatterplot of measured and simulated regional median RO₂.

3.3 Net ozone production rates around the Arabian Peninsula

In the following, net ozone production rates (at noon) are calculated based on Eq. 7 for the different regions. These noontime values are scaled to diurnal production rates (Figure 8). As photochemical net ozone destruction is in good approximation
465 linear with actinic flux $j(\text{NO}_2)$ and as on average $(46.1 \pm 2.8) \%$ of the total $j(\text{NO}_2)$ occurred $\pm 2\text{h}$ around noon, the median noontime NOPR estimate was multiplied by $4/0.461 \approx 8.68$ to obtain a diurnal value. The error in the total actinic flux located $\pm 2\text{h}$ around noon is estimated from the standard deviation of the best estimate of 0.461 at $\Delta s \approx 6 \%$. Due to contamination by the own ship exhaust and due to the availability of OH and HO₂ data only from 18 July 2017 onwards, we have limited the analysis to the period from 22 July 2017 to 31 August 2017. A comparison of NOPR estimated based on
470 measured and simulated data for the different regions is shown in Figure 9. A break-down of the different terms of Eq. 7 in the six regions is included in the supplementary Figures S10-S13.

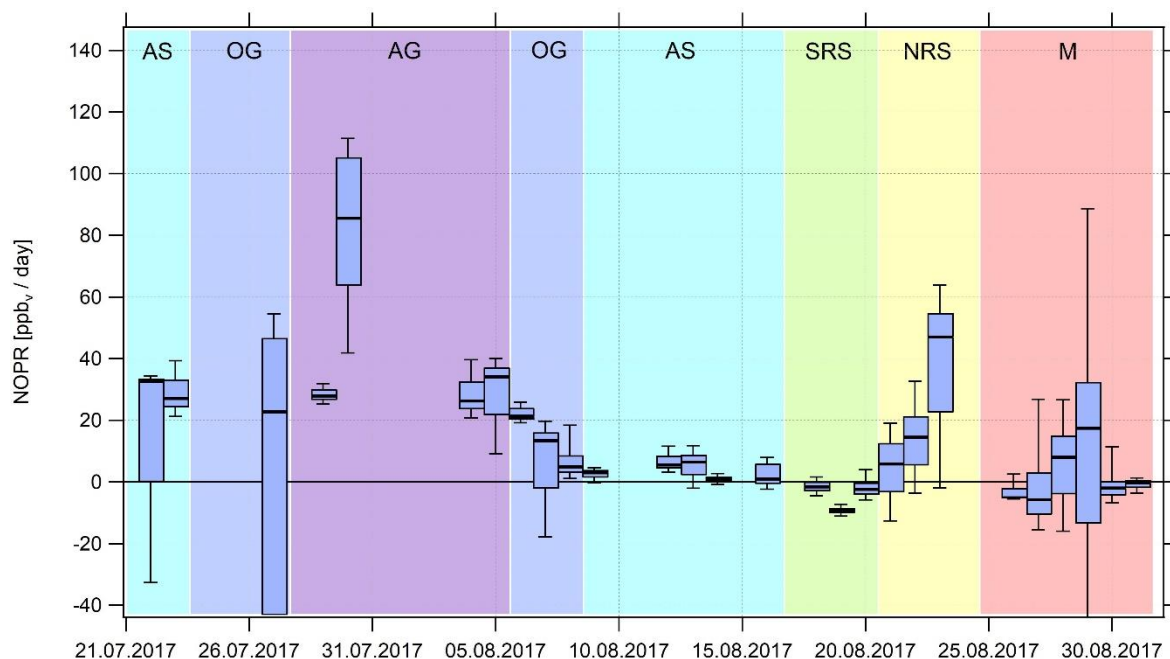


Figure 8: Timeline of the diurnal NOPR from **22 July** to **31 August** 2017. NOPR calculations are limited to the time period from **22 July** onwards due to missing HO_x data and contamination by the ship exhaust itself before this period. See annotations for the classification of the different regions.

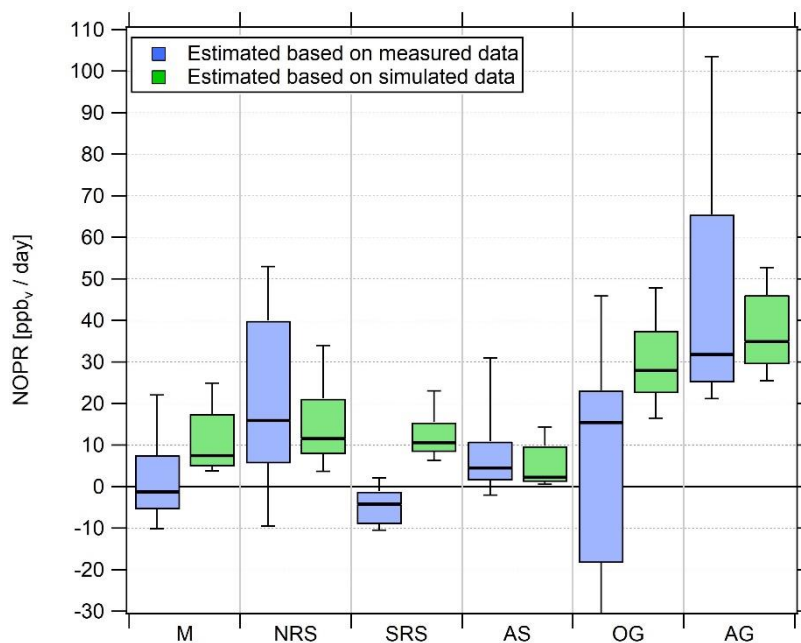


Figure 9: Diurnal net ozone production rates in the different regions. Related to the magnitude of pollution sources, the lower whisker of the NOPR estimate over the Oman Gulf is -324 ppb day⁻¹.

Over the Mediterranean and the Southern Red Sea, NOPR values do not significantly deviate from zero (production equals loss) within the atmospheric variability. Based on the total measurement uncertainties of the measured quantities in Eq. 3, the systematic error in NOPR is estimated from error propagation by means of the largest error possible at 38 %.

$$\Delta \text{NOPR} = \sqrt{\Delta \text{NO}^2 + \Delta \text{NO}_2^2 + \Delta \text{O}_3^2 + \Delta j(\text{NO}_2)^2 + \Delta j(\text{O}^1\text{D})^2 + \Delta \alpha^2 + \Delta \text{RO}_2^2 + \Delta \text{OH}^2 + \Delta \text{HO}_2^2 + \Delta s^2}$$

$$= \sqrt{6\%^2 + 9\%^2 + 2\%^2 + 10\%^2 + 10\%^2 + 5\%^2 + 15\%^2 + 20\%^2 + 21\%^2 + 6\%^2} \approx 38 \%$$

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The best estimate indicates slight net ozone destruction for the Mediterranean and Southern Red Sea (- 1 ppb day⁻¹) and (- 4 ppb day⁻¹) respectively, and slight net production for the Arabian Sea (5 ppb day⁻¹), which is significantly positive within the variability of the box-interval. Variations in NOPR calculated as the width of the 25-75-percentile-box yield comparable values of 9-11 ppb day⁻¹ for these three regions. Substantial net ozone production was inferred over the Oman Gulf, the Northern Red Sea, and the Arabian Gulf with the respective median values being 16 ppb day⁻¹, 16 ppb day⁻¹ and 32 ppb day⁻¹, respectively. Especially over the Red Sea we find a strong latitudinal gradient in net ozone production rates with higher values towards the northern end, while slight net ozone destruction of -4 ppb day⁻¹ is reported over the southern part.

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NOPR estimates for the Oman Gulf, the Northern Red Sea and the Arabian Gulf are comparable to results reported for dense traffic shipping routes such as the Houston Ship Channel with NOPR of a few tens of ppb h⁻¹ for periods of severe pollution (Zhou et al., 2014). Similar net ozone production rates have been reported for regions of Beijing in summer 2006 (Lu et al., 2010). For regions with low anthropogenic influence such as the Southern Red Sea and the Arabian Sea we estimate net ozone production that does not differ significantly from zero. This is due to the rather low NO_x mixing ratios in the clean marine boundary layer (Bozem et al., 2017). Note that we calculated net ozone destruction only for a few days over the Southern Red Sea and the Arabian Sea, indicating that the marine boundary layer around the Arabian Peninsula is rarely free from anthropogenic influence owing to the multitude of on- and off-shore anthropogenic activities.

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We find that model-calculated estimates of NOPR reproduce the trends observed for NOPR calculated from in situ measurements except over the Mediterranean and the Southern Red Sea. Although EMAC predicts high ozone levels over the Arabian Sea, it also reports the lowest NOPR in this region. On the other side, the large overestimation of the model-calculated estimate NOPR against the one based on measured tracer data over the Mediterranean and over the Southern Red Sea could be linked to NO_x being overestimated in the model in these regions. In the model, pollution emissions, especially over the Oman Gulf and the Arabian Gulf, seem to be averaged over a large (1.1° grid size) region. High background concentrations of ozone precursors hence contribute to net ozone production rates that compare to conditions observed in the Houston case (Zhou et al., 2014). Even in the more pristine regions such as over the Southern Red Sea and the Arabian Sea, the model is not able to reproduce net ozone destruction, which is consistent with the fact the ozone is generally too high and

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510 that NO_x levels below 0.1 ppb_v are not found in the model. See supplementary Table ST7 and ST8 for further information and supplementary Figure S9 for an additional scatterplot of measured and simulated regional NOPR.

3.4 VOC- and NO_x -sensitivity

Ozone is photochemically formed when the precursors NO_x and VOCs are abundant in the presence of sunlight (Bozem et al., 2017; Jaffe et al., 2018). In order to determine whether a chemical system is NO_x - or VOC-limited or in a transition
515 between those two regimes, one has to estimate the total amount of OH reactivity towards VOCs and towards NO_x . Therefore the VOC/ NO_x -ratio is an important indicator of the behavior of NO_x , VOCs and O_3 in a system. Since it is not feasible to precisely define all ambient VOCs (could be thousands), formaldehyde mixing ratios have been used as a proxy for the OH reactivity towards VOCs since it is a short-lived oxidation product of many VOCs that is often positively correlated with peroxy radicals (Sillman et al., 1995; Duncan et al., 2010). Sillman et al. first used afternoon concentrations
520 of indicator species such as HCHO and total reactive nitrogen (NO_y) to determine the sensitivity of ozone production to VOCs or NO_x (Sillman et al., 1995). Their approach was later successfully transferred to space-based satellite observations by using the ratio of tropospheric columns of HCHO and NO_2 to determine the sensitivity of ozone production (Martin et al., 2004). Here we use HCHO/ NO_2 -ratios (referred to as “Ratio”) deduced by Duncan et al. as indicators for the sensitivity of ozone production to NO_x - and VOC-limitations in megacities in the United States with large amounts of anthropogenic NO_x
525 and VOC emissions (Duncan et al., 2010). The Ratio is an indicator of surface photochemistry as most of the atmospheric column of HCHO and NO_2 is located in the planetary boundary layer (Duncan et al., 2010). Duncan et al. have derived NO_x -limited ozone production regimes for $\text{HCHO}/\text{NO}_2 > 2$ and VOC-limited ozone production for $\text{HCHO}/\text{NO}_2 < 1$ (Duncan et al., 2010). For $1 < \text{HCHO}/\text{NO}_2 < 2$ both NO_x and VOC emission reductions may lead to a reduction in ozone. Figure 10 shows the Box-Whisker-Plot classification of the HCHO/ NO_2 -ratio of the different regions during noontime.

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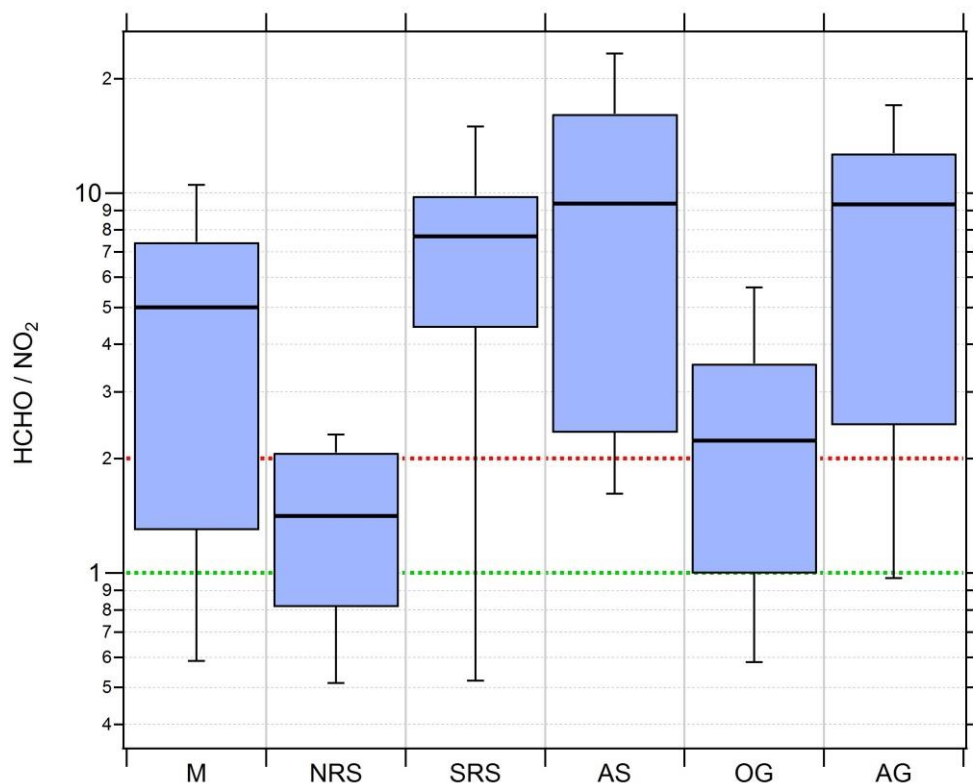


Figure 10: Box-Whisker-Plots of the HCHO/NO₂-ratio for the different regions with the black bar indicating the median value. Red (ratio = 2) and green (ratio = 1) lines indicate the limits for HCHO/NO₂ deduced by Duncan et al. (2010) for NO_x-limitation and VOC-limitation, respectively.

Median HCHO/NO₂-ratios of 5, 7.7, 9.4 and 9.3 over the Mediterranean, the Southern Red Sea, the Arabian Sea and the Arabian Gulf respectively indicate tendencies towards NO_x-limited regimes. In a previous study based on measured OH reactivity, Pfannerstill et al. classified these regions as being mostly in a transition between NO_x- and VOC-limitation, with a tendency towards NO_x-limitation (2019). Median HCHO/NO₂-ratios of 1.4 and 2.2 estimated over the Northern Red Sea and the Oman Gulf signify tendencies towards VOC-limitation. However, none of the medians of the six regions falls below the VOC-limit deduced by Duncan et al. (2010).

Over the Red Sea we find a latitudinal gradient in the HCHO/NO₂-ratio, similar to the gradients for NO_x and NOPR. Due to very low NO_x over the Southern Red Sea, O₃ production is NO_x-limited, changing into a more VOC-limited regime over the Northern Red Sea. **Ozone production over the Mediterranean was classified as rather NO_x-limited, however partly being in the transition regime between NO_x- and VOC-limitation., which can be explained by measurements obtained on 29 August 2017 when laying at anchor in front of Malta with a multitude of (NO_x)-emissions from nearby situated vessels. Average noontime NO_x on that particular day was about three times as large as the regional average noontime NO_x observed over the**

whole Mediterranean area. NO_x limitation is also inferred for the relatively clean Arabian Sea and the polluted Arabian Gulf atmosphere. Note that a further increase in NO_x -emissions from increased shipping in the Arabian Gulf may initially lead to higher ozone production. However, a further increase in NO_x might eventually lead to a change from NO_x - to VOC-sensitivity and a decrease in ozone production for this region, as observed for the Oman Gulf (median HCHO/NO_2 -ratio of 2.2 and average O_3 of 34 ppb_v). See supplementary Table ST9 for detailed statistics on regional HCHO/NO_2 -ratios.

4 Conclusion

In situ observations of NO , NO_2 , O_3 , HCHO , OH , HO_2 , absolute humidity, actinic flux, temperature and pressure were carried out in the marine boundary layer around the Arabian Peninsula during the AQABA ship campaign from late June to early September 2017. Concentration ranges of both NO_x and O_3 clearly showed anthropogenic influence in the MBL. NO_x was highest over the Arabian Gulf, the Northern Red Sea and the Oman Gulf. Lowest NO_x was observed over the Arabian Sea and over the Southern Red Sea during the second leg. O_3 mixing ratios were highest over the Arabian Gulf. We observed a latitudinal gradient in O_3 concentrations with higher values towards the northern part of the Red Sea. Although comparable O_3 averages were measured over the Northern Red Sea and over the Mediterranean, lower variability over the Mediterranean towards the end of August 2017 indicates photochemically more extensively aged air masses. The lowest regional O_3 mixing ratio average was detected over the Arabian Sea, which is broadly comparable to remote marine boundary layer conditions in the Northern Hemisphere.

Noontime RO_2 estimates based on deviations from the Leighton Ratio yield median values around the Arabian Peninsula amount to 15 – 33 ppt_v for all regions except over the Arabian Gulf where the median is 73 ppt_v . The uncertainty due to the missing up-welling actinic flux portion is expected to be insignificant. Furthermore, we estimated noontime and diurnal NOPR based on Eq. 6 and the integral over the actinic flux. Highest diurnal NOPR were observed over the Oman Gulf, the Northern Red Sea and the Arabian Gulf with median values of 16 $\text{ppb}_v \text{ day}^{-1}$, 16 $\text{ppb}_v \text{ day}^{-1}$ and 32 $\text{ppb}_v \text{ day}^{-1}$, respectively, which is in agreement with previous studies that predicted net photochemical O_3 formation conditions in the region. Net ozone destruction was only observed for a few days with clean conditions over the Arabian Sea and the Southern Red Sea. Based on HCHO/NO_2 -ratios our analysis suggests tendencies towards NO_x -limitation over the Mediterranean, the Southern Red Sea, the Arabian Sea and the Arabian Gulf and VOC-limitation over the Northern Red Sea and the Oman Gulf, which reproduces the trends observed by Pfannerstill et al. (2019).

NO_x results from the atmospheric chemistry – general circulation model EMAC underestimate the measurement data by 10 % whereas median modeled O_3 overestimates the measurement by 23 %, the latter being related to limitations in model resolution in coastal proximity and near shipping lanes. Although EMAC generally reproduces regional NO_x and O_3 medians, the scatter when comparing both data sets is large. NO_x is generally too low as it does not resolve local point

sources and too high for clean regions. Lowest NO_x of less than 0.1 ppb_v found in the in situ measurements is not reproduced
580 by the model as emissions are averaged over a large area (1.1°). Median noontime RO₂ retrieved from the EMAC model are
~ 5 % higher than RO₂ estimates based on measurement data, however, the RO₂ sum deduced from EMAC is sometimes
about a factor of 2 higher than the regional RO₂ estimate based on the Leighton Ratio and measured tracer data. NOPR
estimates based on modeled data reproduce the tendencies derived from the measurements very well. However, the model
does not reproduce observed net ozone destruction along some clean parts of the ship cruise.

585 **Data availability**

Data used in this study is available to all scientists agreeing to the AQABA protocol at
<https://doi.org/10.5281/zenodo.3693988>.

Author contributions

IT, HF and JL designed the study. UP and IT performed the CLD NO and NO₂ measurements and processed the data. JC and
590 PE performed the O₃ measurements, JS performed the actinic flux measurements. JS performed cavity ring-down
spectroscopy measurements of NO₂. DD and BH performed the HCHO measurements. HH, MM, RR, ST performed the OH
and HO₂ measurements. J-DP was responsible for the H₂O measurements. Model simulations were made by AP. All authors
have contributed to writing this manuscript

Competing interests

595 The authors declare no conflict of interest.

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we would like to thank the whole AQABA community for a successful campaign.

Appendix: Acronyms and abbreviations

605 General

AQABA Air Quality and Climate in the Arabian Basin campaign
 Cyl The Cyprus Institute
 KAUST King Abdullah University of Science and Technology
 KISR Kuwait Institute for Scientific Research

610 Regions

AG Arabian Gulf
 AS Arabian Sea
 M Mediterranean Sea
 NRS Northern Red Sea
 615 OG Oman Gulf
 SRS Southern Red Sea

Scientific

CLD Chemiluminescence detector
 CRDS Cavity ring-down spectroscopy
 620 ECHAM5 Fifth generation European Centre Hamburg general circulation model
 EMAC ECHAM/MESSy Atmospheric Chemistry model
 FWHM Full width at half maximum
 GC-FID Gas chromatography – flame ionization detector
 HORUS HydOxyl Radical measurement Unit based on fluorescence Spectroscopy instrument
 625 HO_x OH + HO₂
 LED Light emitting diode
 LIF Laser induced fluorescence
 MBL Marine boundary layer
 MESSy Modular Earth Submodel System
 630 NOPR Net ozone production rate
 NO_x NO + NO₂
 PFA Perfluoroalkoxy
 PSS Photostationary steady state
 PTFE Polytetrafluoroethylene
 635 SLM Standard liter per minute
 STEAM3 Ship Traffic Emission Assessment Model 3
 TMU Total measurement uncertainty
 VOC Volatile organic compounds
 UV Ultraviolet

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Supplement of

Net ozone production and its relationship to NO_x and VOCs in the marine boundary layer around the Arabian Peninsula

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Figure S1 shows a day to day calculation of the correction factor to scale the fractional noontime integral to a diurnal value.

840 An average \pm standard deviation of (46.1 ± 2.8) % of the diurnal integral within ± 2 h around noontime was estimated.

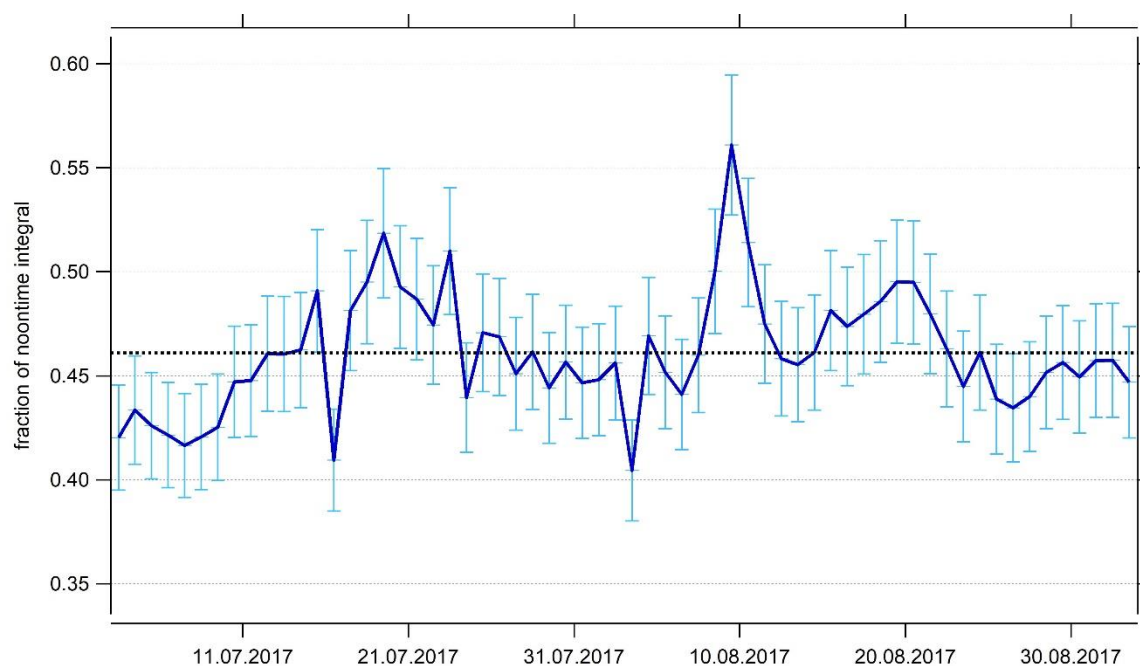


Figure S1: Ratio of the noontime actinic flux (± 2 h around noon) with regard to the total actinic flux $j(\text{NO}_2)$ of that particular day. Dashed line represents the campaign average of (46.1 ± 2.8) %. The errors bars are represented by the relative uncertainty (6 %) of the campaign average.

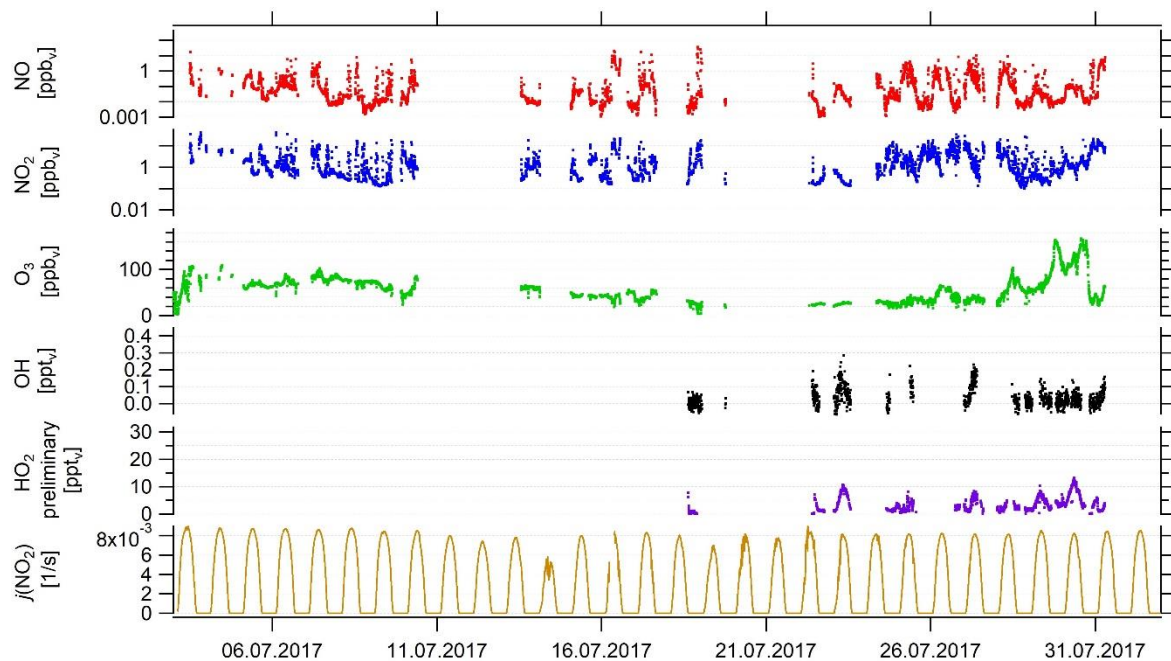


Figure S2: Timeline of NO, NO₂ (both CLD), O₃, OH, HO₂ preliminary and $j(\text{NO}_2)$ data during the first leg. See Table ST2 for additional information on the ship cruise. Note that HO₂ data are preliminary.

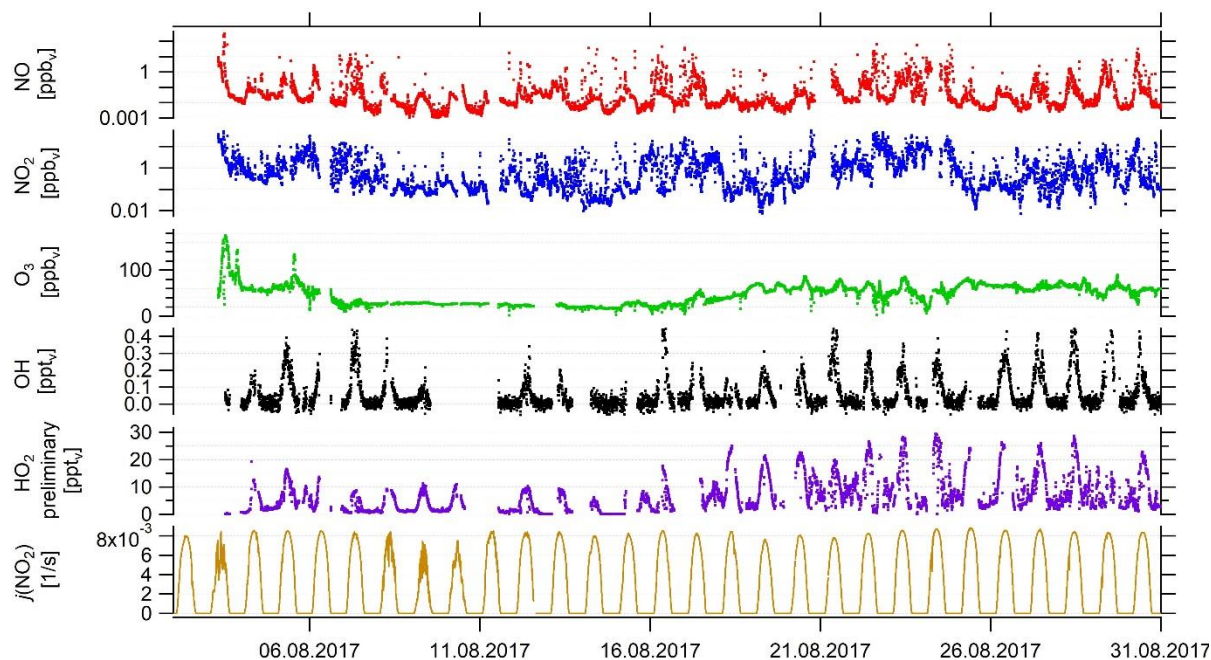


Figure S3: Timeline of NO, NO₂ (both CLD), O₃, OH, HO₂ preliminary and $j(\text{NO}_2)$ data during the second leg. See Table ST2 for additional information on the ship cruise. Note that HO₂ data are preliminary.

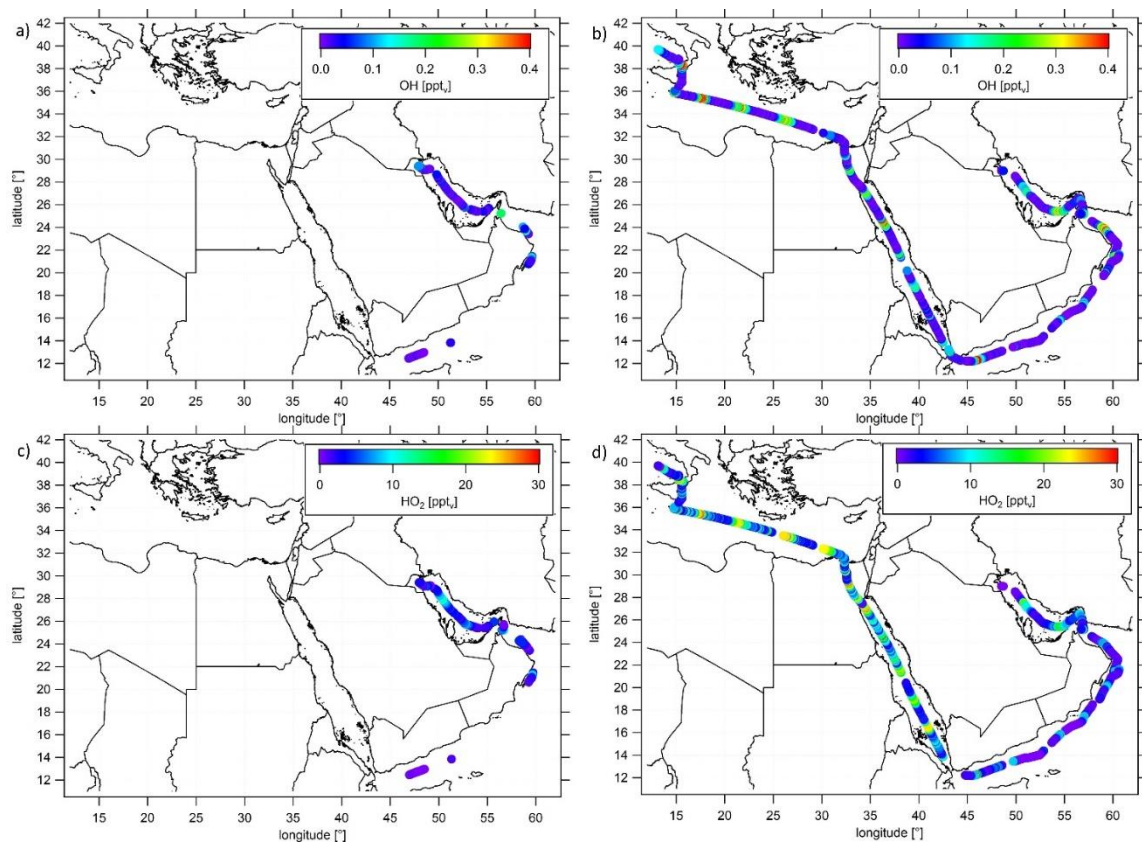


Figure S4: Ship cruises with color-scaled OH mixing ratios a) during the first and b) the second leg and color-scaled HO₂ mixing ratios (preliminary) c) during the first and d) the second leg. Note that OH and preliminary HO₂ data have been filtered for own stack contamination.

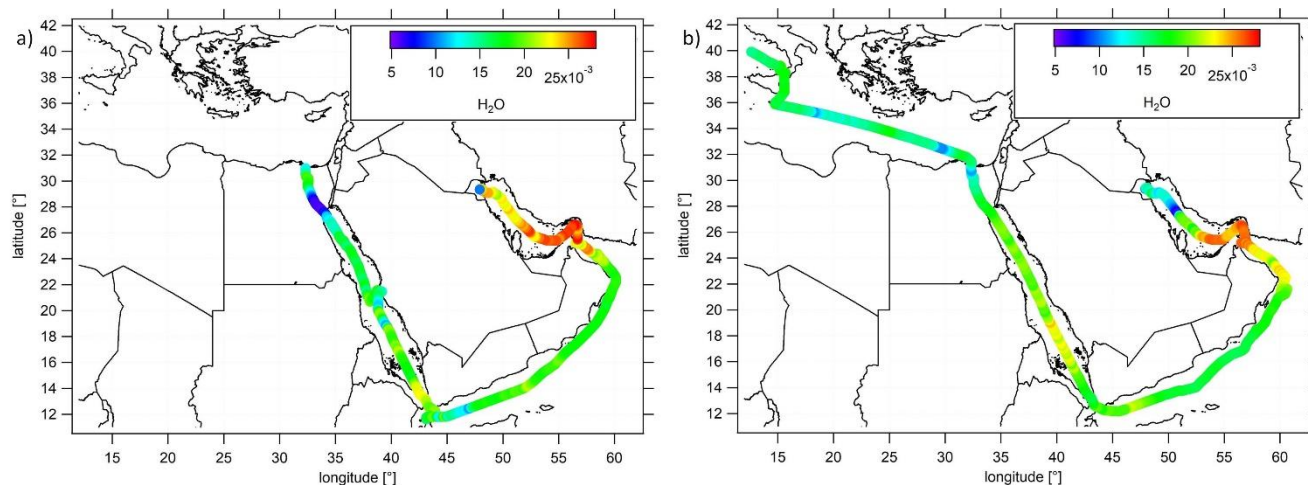


Figure S5: Ship cruises with color-scaled absolute humidity a) during the first and b) the second leg.

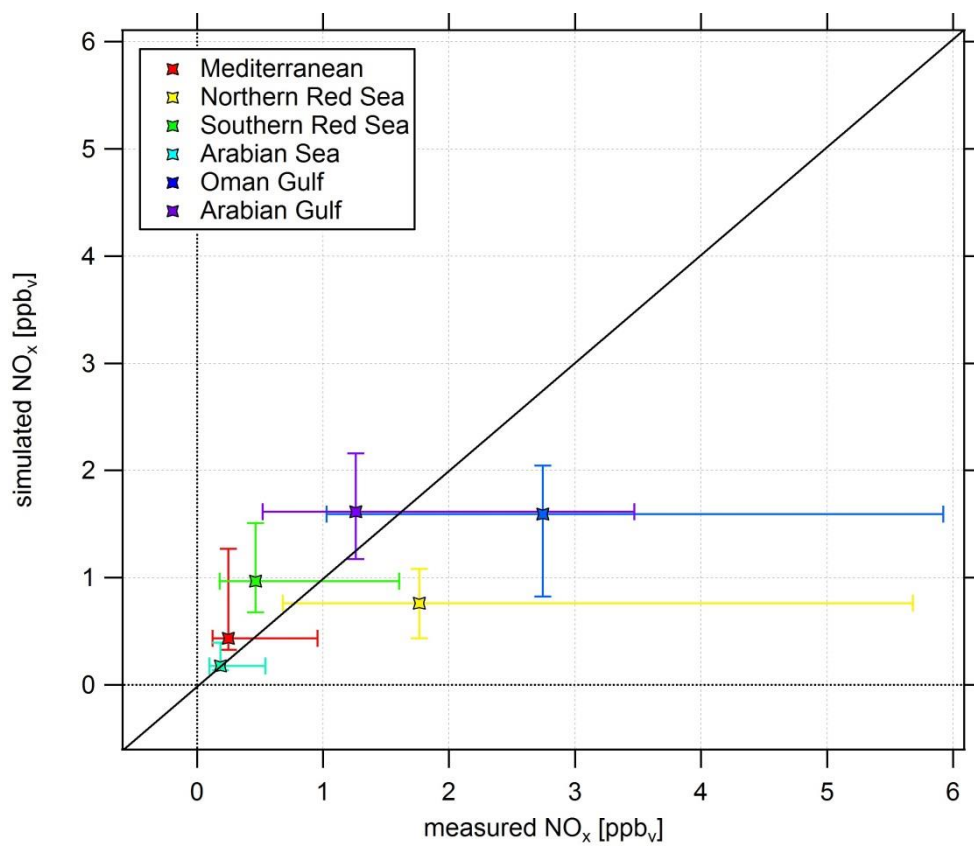
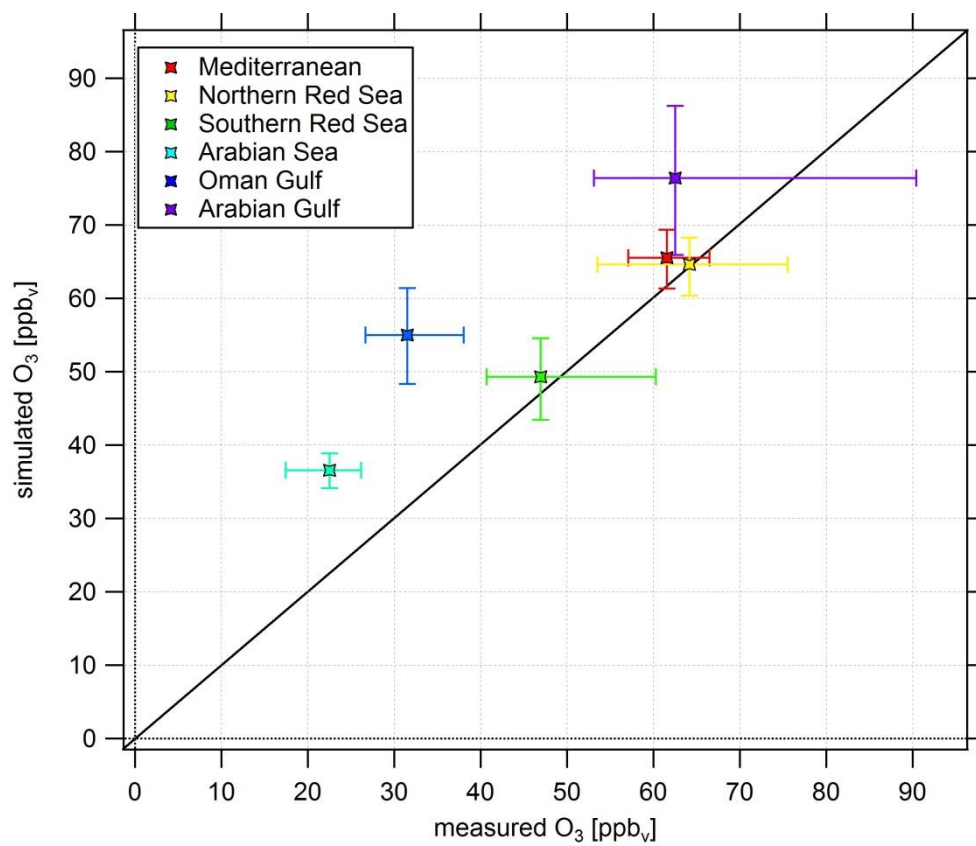


Figure S6: Scatterplot of simulated and measured regional NO_x median in ppbv. 1:1 line added for orientation. The error bars represent the 25-75-percentile variation.



865 **Figure S7: Scatterplot of simulated and measured regional O₃ median in ppbv. 1:1 line added for orientation. The error bars represent the 25-75-percentile variation.**

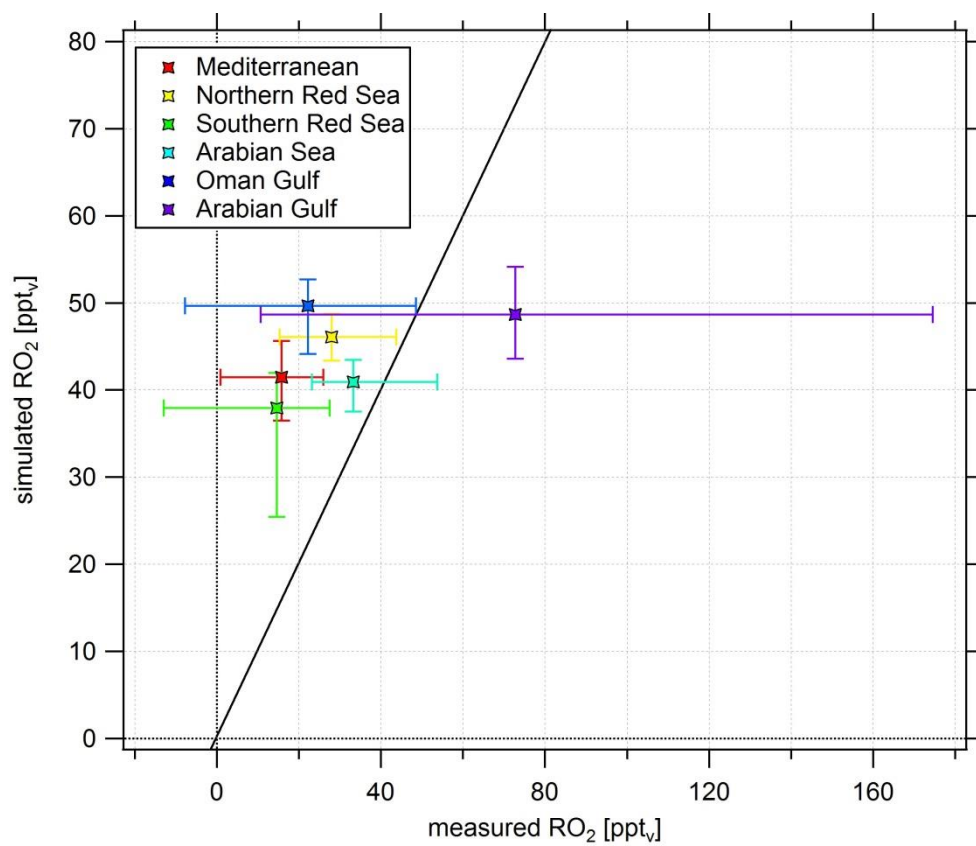
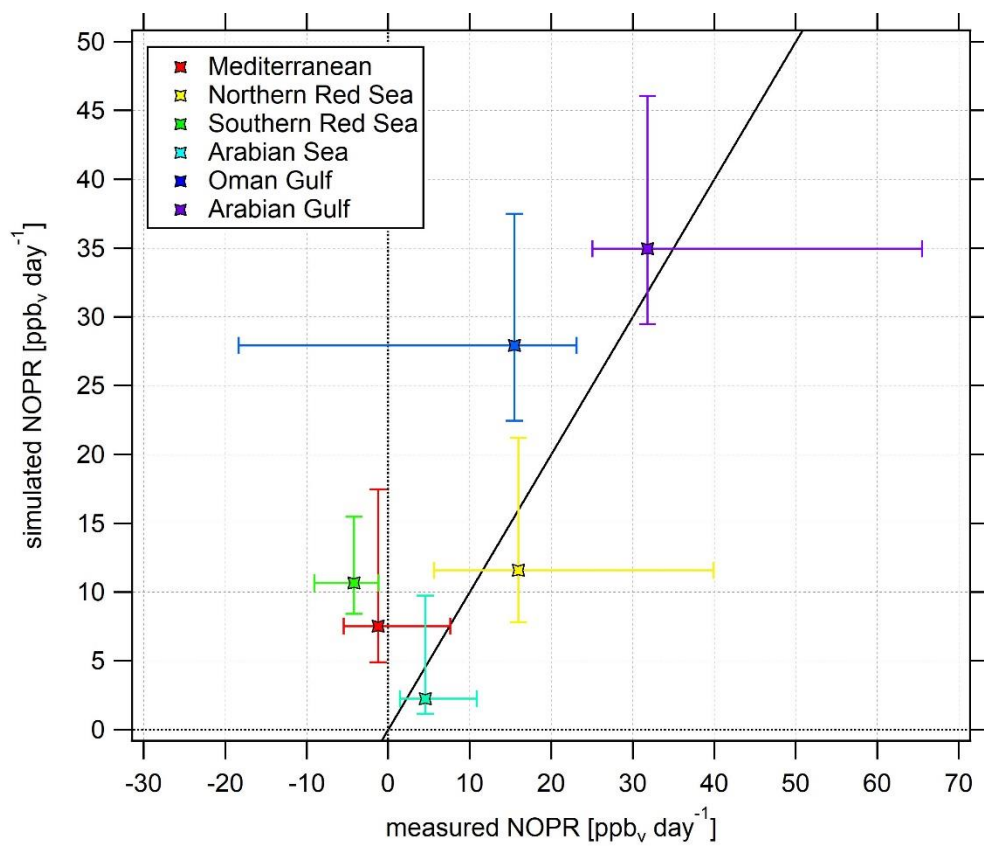
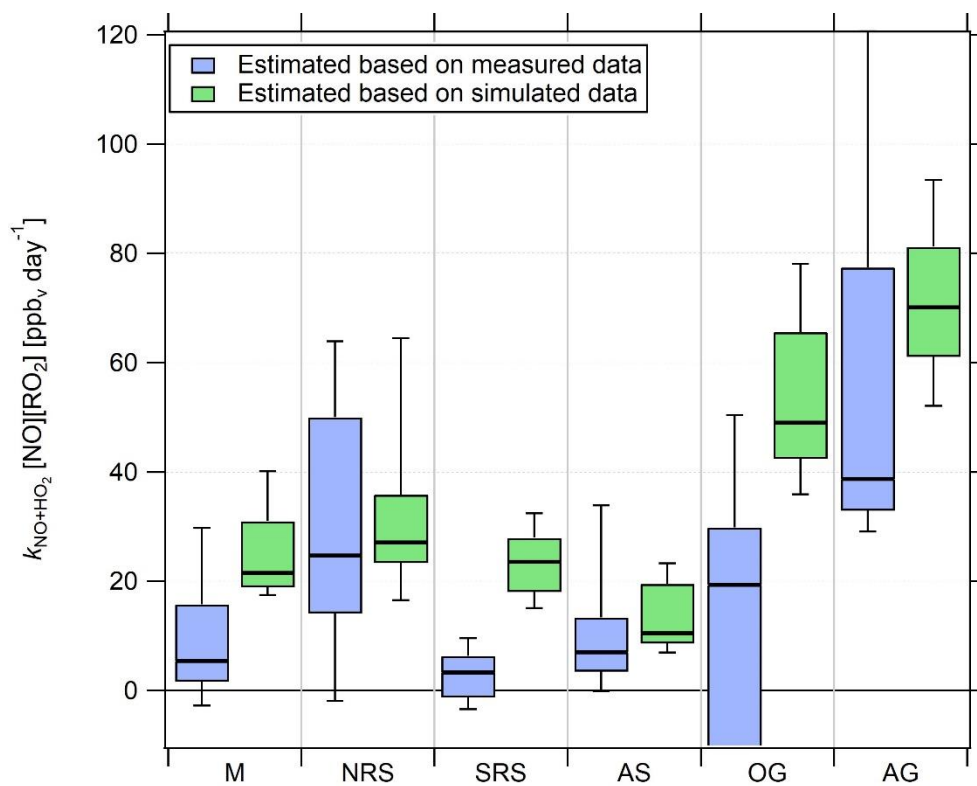


Figure S8: Scatterplot of simulated regional RO_2 median and regional RO_2 median estimated based on measured tracer data in ppt_v . 1:1 line added for orientation. The error bars represent the 25-75-percentile variation.



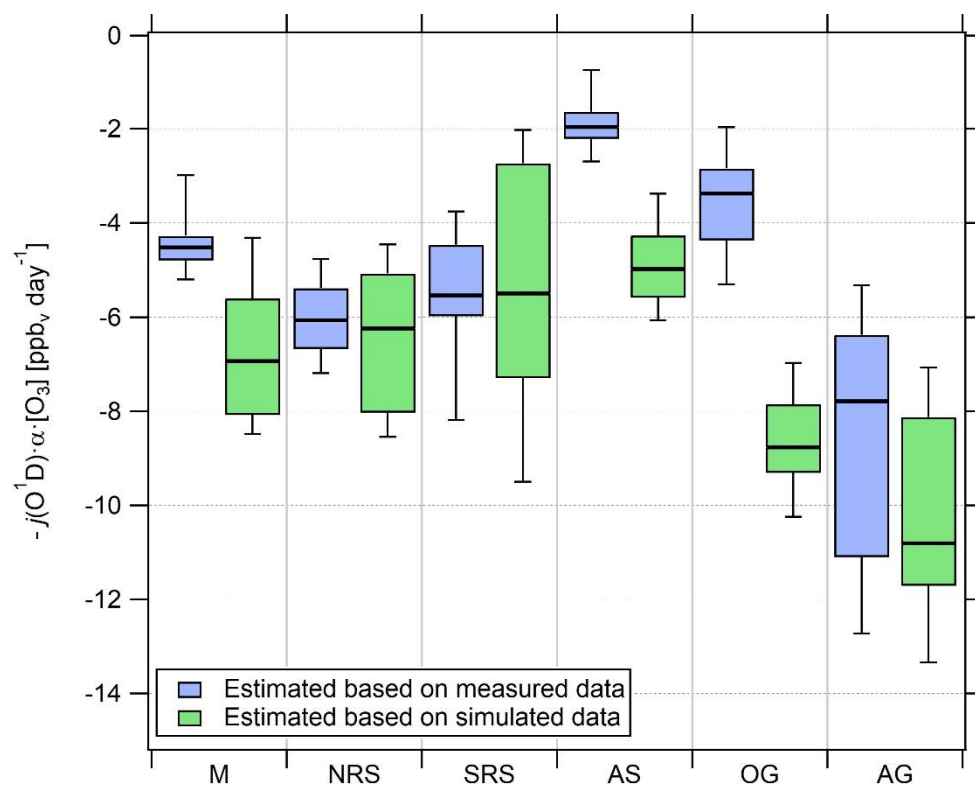
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Figure S9: Scatterplot of median NOPR in $\text{ppb}_v \text{ day}^{-1}$ estimated based on simulated and measured tracer data. 1:1 line added for orientation. The error bars represent the 25-75-percentile variation.

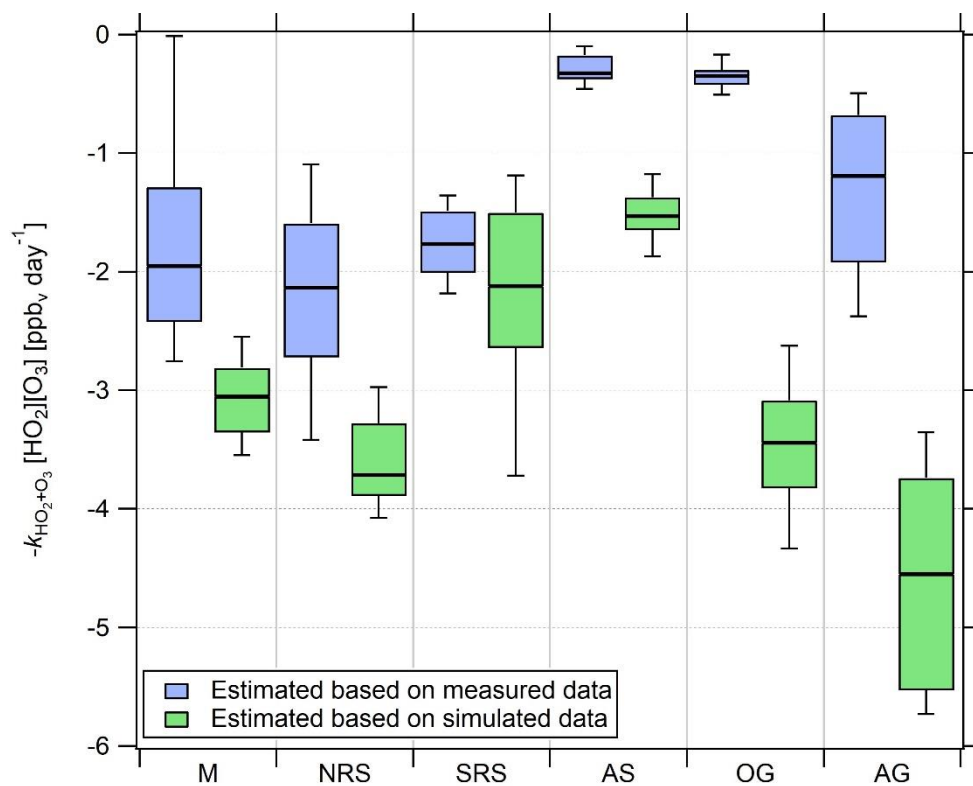


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Figure S10: Comparison of the regional, absolute contribution of $k_{\text{NO}+\text{HO}_2}[\text{NO}][\text{RO}_2]$ to NOPR in the six different regions investigated during AQABA. The horizontal black bar indicates the median value, the box the 25- and 75-percentiles and the whiskers the 10- and 90-percentiles.

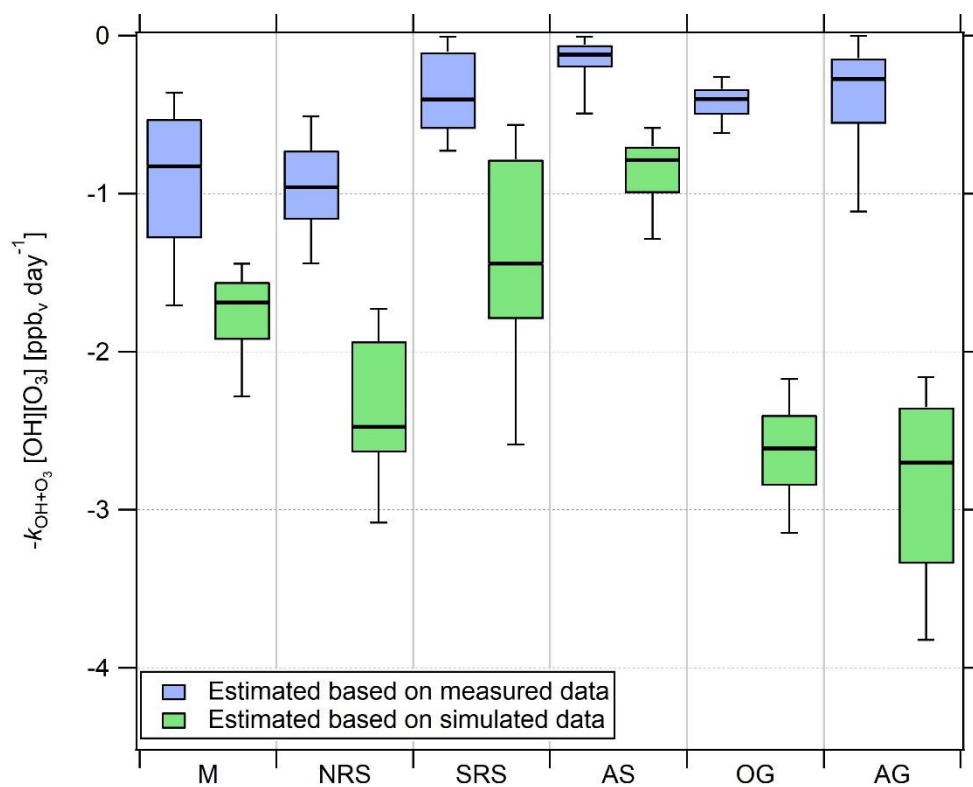


880 **Figure S11: Comparison of the regional, absolute contribution of $-j(\text{O}^1\text{D}) \cdot \alpha \cdot [\text{O}_3]$ to NOPR in the six different regions investigated during AQABA. The horizontal black bar indicates the median value, the box the 25- and 75-percentiles and the whiskers the 10- and 90-percentiles.**



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Figure S12: Comparison of the regional, absolute contribution of $-k_{\text{HO}_2+\text{O}_3}[\text{HO}_2][\text{O}_3]$ to NOPR in the six different regions investigated during AQABA. The horizontal black bar indicates the median value, the box the 25- and 75-percentiles and the whiskers the 10- and 90-percentiles.



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Figure S13: Comparison of the regional, absolute contribution of $-k_{\text{OH}+\text{O}_3}[\text{OH}][\text{O}_3]$ to NOPR in the six different regions investigated during AQABA. The horizontal black bar indicates the median value, the box the 25- and 75-percentiles and the whiskers the 10- and 90-percentiles.

895 **Table ST1: Range of latitudinal and longitudinal coordinates and dates during both legs of the different regions.**

region (abbreviation)	latitudinal range	longitudinal range	Date (1 st leg)	Date (2 nd leg)
Mediterranean (M)	31.810° N- 39.923° N	12.620° E- 31.850° E	---	25.08.2017 – 31.08.2017
Northern Red Sea (NRS)	23.343° N- 30.986° N	32.305° E- 37.085° E	03.07.2017 – 08.07.2017	21.08.2017 – 24.08.2017
Southern Red Sea (SRS)	12.672° N- 22.494° N	37.411° E- 43.327° E	09.07.2017 – 16.07.2017	17.08.2017 – 20.08.2017
Arabian Sea (AS)	11.797° N- 22.782° N	44.035° E- 60.636° E	18.07.2017 – 24.07.2017	07.08.2017 – 16.08.2017
Oman Gulf (OG)	23.050° N- 25.622° N	56.492° E- 59.913° E	24.07.2017 – 27.07.2017	05.08.2017 – 07.08.2017
Arabian Gulf (AG)	25.396° N- 29.425° N	47.920° E- 56.772° E	27.07.2017 – 31.07.2017	03.08.2017 – 05.08.2017

Table ST2: Overview of the time spent in the particular regions during AQABA. Red color indicates periods with KI at anchor that are not included in the data analysis. Data measured during bunkering at Fujairah City (06 August, 07:00-15:00 UTC) were also not included in the analysis.

Date	Region	Date	Region
03.07.2017	Northern Red Sea	02.08.2017	Kuwait port
04.07.2017	Northern Red Sea	03.08.2017	Kuwait port/Arabian Gulf
05.07.2017	Northern Red Sea	04.08.2017	Arabian Gulf
06.07.2017	Northern Red Sea	05.08.2017	Arabian Gulf/Oman Gulf
07.07.2017	Northern Red Sea	06.08.2017	Oman Gulf
08.07.2017	Northern Red Sea	07.08.2017	Oman Gulf/Arabian Sea
09.07.2017	Southern Red Sea	08.08.2017	Arabian Sea
10.07.2017	Southern Red Sea	09.08.2017	Arabian Sea
11.07.2017	Southern Red Sea/Jeddah port	10.08.2017	Arabian Sea
12.07.2017	Jeddah port	11.08.2017	Arabian Sea
13.07.2017	Jeddah port/Southern Red Sea	12.08.2017	Arabian Sea
14.07.2017	Southern Red Sea	13.08.2017	Arabian Sea
15.07.2017	Southern Red Sea	14.08.2017	Arabian Sea
16.07.2017	Southern Red Sea	15.08.2017	Arabian Sea
17.07.2017	Djibouti port	16.08.2017	Arabian Sea
18.07.2017	Arabian Sea	17.08.2017	Southern Red Sea
19.07.2017	Arabian Sea	18.08.2017	Southern Red Sea
20.07.2017	Arabian Sea	19.08.2017	Southern Red Sea
21.07.2017	Arabian Sea	20.08.2017	Southern Red Sea
22.07.2017	Arabian Sea	21.08.2017	Northern Red Sea
23.07.2017	Arabian Sea	22.08.2017	Northern Red Sea
24.07.2017	Arabian Sea/Oman Gulf	23.08.2017	Northern Red Sea
25.07.2017	Oman Gulf	24.08.2017	Northern Red Sea
26.07.2017	Oman Gulf	25.08.2017	Mediterranean
27.07.2017	Oman Gulf/Arabian Gulf	26.08.2017	Mediterranean
28.07.2017	Arabian Gulf	27.08.2017	Mediterranean
29.07.2017	Arabian Gulf	28.08.2017	Mediterranean
30.07.2017	Arabian Gulf	29.08.2017	Mediterranean
31.07.2017	Arabian Gulf/Kuwait port	30.08.2017	Mediterranean
01.08.2017	Kuwait port	31.08.2017	Mediterranean

Table ST3: Overview of measured NO_x (upper table) and measured O₃ (lower table) spatial volume mixing ratio average, standard deviation, 1st quantile, median, 3rd quantile (all in ppb_v) and number of considered data points.

NO _x (upper), O ₃ (lower)	Mediterranea n	Northern Red Sea	Southern Red Sea	Arabian Sea	Oman Gulf	Arabian Gulf
data points	1767	1694	1755	2656	1056	1539
average	1.24	4.69	1.62	0.95	4.16	3.65
stdev	3.34	7.9	3.7	3.15	4.33	9.24
1 st quantile	0.12	0.68	0.18	0.10	1.03	0.52
median	0.25	1.76	0.46	0.19	2.74	1.26
3 rd quantile	0.96	5.68	1.6	0.54	5.92	3.47
data points	2010	2717	2307	4130	1249	1809
average	61.56	63.39	50.35	21.53	34.04	73.99
stdev	8.25	18.45	12.96	6.8	11.27	35.68
1 st quantile	57.05	53.51	40.68	17.45	26.66	53.08
median	61.54	64.16	46.93	22.52	31.5	62.5
3 rd quantile	66.48	75.51	60.28	26.19	38.03	90.42

Table ST4: Overview of simulated NO_x (upper table) and simulated O₃ (lower table) spatial volume mixing ratio average, standard deviation, 1st quantile, median, 3rd quantile (all in ppb_v) and number of considered data points.

NO _x (upper), O ₃ (lower)	Mediterranea n	Northern Red Sea	Southern Red Sea	Arabian Sea	Oman Gulf	Arabian Gulf
data points	2012	2719	2310	4464	1253	1810
average	0.84	1.27	1.13	0.31	1.88	1.91
stdev	0.75	1.97	0.62	0.29	1.47	1.37
1 st quantile	0.33	0.43	0.67	0.14	0.82	1.17
median	0.43	0.76	0.97	0.18	1.59	1.61
3 rd quantile	1.27	1.08	1.51	0.39	2.05	2.16
data points	2012	2719	2310	4464	1253	1810
average	65.15	64.76	49.5	36.91	55	76.85
stdev	4.99	7.7	7.21	3.87	9.94	12.8
1 st quantile	61.35	60.34	43.47	34.12	48.37	65.88
median	65.53	64.6	49.33	36.54	55.01	76.39
3 rd quantile	69.33	68.25	54.58	38.85	61.42	86.22

Table ST5: Overview of noontime RO₂ spatial volume mixing ratio average, standard deviation, 1st quantile, median, 3rd quantile (all in ppt_v) estimated based on measured tracer data. Number of considered data points added in the first line.

RO ₂	Mediterranean	Northern Red Sea	Southern Red Sea	Arabian Sea	Oman Gulf	Arabian Gulf
data points	288	126	190	338	166	242
average	13	27	7	64	23	94
stdev	24	20	34	83	42	113
1 st quantile	1	15	-13	23	-8	11
median	16	28	15	33	22	73
3 rd quantile	26	44	27	54	49	176

910 **Table ST6: Overview of simulated noontime RO₂ spatial volume mixing ratio average, standard deviation, 1st quantile, median, 3rd quantile (all in ppt_v). Number of considered data points added in the first line.**

RO ₂	Mediterranean	Northern Red Sea	Southern Red Sea	Arabian Sea	Oman Gulf	Arabian Gulf
data points	336	192	192	720	203	293
average	41	46	36	40	48	49
stdev	5	4	11	5	6	7
1 st quantile	36	43	25	38	44	44
median	41	46	38	41	50	49
3 rd quantile	46	49	42	43	53	54

Table ST7: Overview of NOPR average, standard deviation, 1st quantile, median, 3rd quantile (all in ppbv day⁻¹) estimated based on measured tracer data. Number of considered data points added in the first line.

NOPR	Mediterranean	Northern Red Sea	Southern Red Sea	Arabian Sea	Oman Gulf	Arabian Gulf
data points	148	114	89	187	84	111
average	-3	16	-5	-67	-105	-24
stdev	43	39	12	576	362	449
1 st quantile	-5	6	-9	1	-18	25
median	-1	16	-4	5	16	32
3 rd quantile	8	40	-1	11	23	65

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Table ST8: Overview of NOPR average, standard deviation, 1st quantile, median, 3rd quantile (all in ppbv day⁻¹) estimated based on simulated tracer data. Number of considered data points added in the first line.

NOPR	Mediterranean	Northern Red Sea	Southern Red Sea	Arabian Sea	Oman Gulf	Arabian Gulf
data points	336	192	192	720	203	293
average	11	16	12	5	30	38
stdev	8	12	6	6	11	10
1 st quantile	5	8	8	1	22	30
median	8	12	11	2	28	35
3 rd quantile	17	21	15	10	37	46

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Table ST9: Overview of measured HCHO/NO₂-ratio average, standard deviation, 1st quantile, median, 3rd quantile and number of considered data points.

Ratio	Mediterranean	Northern Red Sea	Southern Red Sea	Arabian Sea	Oman Gulf	Arabian Gulf
data points	203	79	48	252	108	122
average	5.4	1.5	8.5	11.1	2.7	9
stdev	4.7	0.7	6.5	8.9	2.1	6.4
1 st quantile	1.3	0.8	4.4	2.3	1	2.5
median	5	1.4	7.7	9.4	2.2	9.3
3 rd quantile	7.4	2.1	9.8	16.1	3.6	12.7

925 **Table ST10: List of included peroxy radicals (with less than four carbon atoms) for the reaction with NO as recommended by Sander et al. (2019).**

Species
HO ₂
CH ₃ O ₂
C ₂ H ₅ O ₂
C ₂ H ₅ CO ₃
CH ₃ CO ₃
C3DIALO2 (C ₃ H ₃ O ₄)
CH ₃ CHOHO ₂
CH ₃ COCH ₂ O ₂
CH ₃ COCO ₃
CHOCOCH ₂ O ₂
CO ₂ H ₃ CO ₃
HCOCH ₂ CO ₃
HCOCH ₂ O ₂
HCOCO ₃
HCOCO ₂ HCO ₃
HOC ₂ H ₄ CO ₃
HOCH ₂ CH ₂ O ₂
HOCH ₂ CO ₃
HOCH ₂ COCH ₂ O ₂
HOCH ₂ O ₂
CH ₃ CHO ₂ CH ₂ OH
IC3H7O2 (isopropylperoxy radical)
NC3H7O2 (propylperoxy radical)
NCCH ₂ O ₂
NO ₃ CH ₂ CO ₃
CH ₃ CHO ₂ CH ₂ ONO ₂