



# Why the Indo-Gangetic Plain is the region with the largest NH<sub>3</sub> column in the globe during summertime?

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**Abstract.** Satellite observations show a global maximum in ammonia (NH<sub>3</sub>) over the Indo-Gangetic Plain (IGP), with a peak in summer. However, it has never been explained explicitly. In this study, we investigated the causes of high NH<sub>3</sub> loading over the IGP in summer using WRF-Chem (Weather Research and Forecasting model coupled to chemistry). IGP has relatively high NH<sub>3</sub> emission fluxes (0.4 t km<sup>-2</sup> month<sup>-1</sup>) due to intensive agricultural activities and high air temperature in summer. Additionally, low sulfur dioxide (SO<sub>2</sub>) and nitrogen oxides (NO<sub>x</sub>) emissions and high air temperature limit the gas-to-particle conversion of NH<sub>3</sub>, particularly for ammonium nitrate formation. Moreover, the barrier effects of the Himalayas in combination with the surface convergence weaken the horizontal diffusion of NH<sub>3</sub>. The high NH<sub>3</sub> loading over the IGP mainly results from the low gas-to-particle partitioning of NH<sub>3</sub> caused by low SO<sub>2</sub> and NO<sub>x</sub> emissions. It contrasts to those in the North China Plain, where high SO<sub>2</sub> and NO<sub>x</sub> emissions promote the conversion of gaseous NH<sub>3</sub> into particulate ammonium.

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## 1 Introduction

Ammonia (NH<sub>3</sub>) has multiple environmental implications. As the only alkaline gas in the atmosphere, it reacts with sulfuric acid (H<sub>2</sub>SO<sub>4</sub>) or nitric acid (HNO<sub>3</sub>) to produce ammonium (NH<sub>4</sub><sup>+</sup>) containing aerosols (Seinfeld and Pandis, 2006), which can affect Earth's radiative balance (Abbatt et al., 2006; Adams et al., 2001) and endanger public health (Pope et al., 2002; Stokstad, 2014). In addition, NH<sub>3</sub> is the main form of reactive nitrogen in the environment (Reis et al., 2009), the deposition of ammonia and ammonium can cause acidification of terrestrial ecosystems and eutrophication of water bodies (Paerl et al., 2014).

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Satellite observations (Van Damme et al., 2018; Warner et al., 2016) and ground-based measurements (Carmichael et al., 2003) have revealed that the Indo-Gangetic Plain (IGP) has the global maximum NH<sub>3</sub> loading, particularly during summer. Previous studies have suggested that the high NH<sub>3</sub> loading over the IGP during summer is caused by high NH<sub>3</sub> emissions from intensive agricultural activities (Clarisse et al., 2009; Van Damme et al., 2015b). Interestingly, satellite measurements show that the total columns of NH<sub>3</sub> over the IGP are much higher than those over the North China Plain (NCP), which has higher NH<sub>3</sub> emissions fluxes ([www.meicmodel.org/dataset-mix](http://www.meicmodel.org/dataset-mix)). Therefore, emissions alone might not be enough to explain the high NH<sub>3</sub> loading over the IGP.

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Apart from dry deposition and wet removal by precipitation, another main sink for  $\text{NH}_3$  is scavenging by acidic species to form particulate  $\text{NH}_4^+$ .  $\text{H}_2\text{SO}_4$  and  $\text{HNO}_3$  resulting from the oxidation of sulfur dioxide ( $\text{SO}_2$ ) and nitrogen oxides ( $\text{NO}_x$ ) are major acidic species in the atmosphere. Previous studies have confirmed that reduced  $\text{SO}_2$  and  $\text{NO}_x$  emissions are key factors driving the increase in  $\text{NH}_3$  concentration (Liu et al., 2018; Yu et al., 2018; Warner et al., 2017). In addition, meteorological conditions (including wind speed, precipitation, relative humidity and air temperature) also influence  $\text{NH}_3$  loading through various chemical and physical processes. These factors may be causing the high  $\text{NH}_3$  loading over the IGP, but these assumptions have not been verified in a modeling study.

In this study, we use a regional air quality model to investigate the causes of high  $\text{NH}_3$  loading over the IGP during summer. This is the first study to analyze the causes of high  $\text{NH}_3$  loading over the IGP considering all possible factors. The remainder of this paper is organized as follows. The air quality model and observational data are described in sect. 2. Section 3 analyzes the influences of several factors (including emissions, chemical conversion, and meteorological conditions) on  $\text{NH}_3$  loading. Among them,  $\text{SO}_2$  and  $\text{NO}_x$  emissions over the IGP are compared to those over the NCP to clearly illustrate their impacts on  $\text{NH}_3$  loading. Section 4 provides concluding remarks.

## 2 Methods

### 2.1 WRF-Chem model and emissions inventory

WRF-Chem (Fast et al., 2006; Grell et al., 2005) version 3.6.1 was applied to investigate the cause of the high  $\text{NH}_3$  loading over the IGP during summer. The simulation was performed on a domain with 30 km horizontal resolution covering the northern part of India and parts of Pakistan, Nepal, China, and Bangladesh with  $120 \times 90$  grid cells. There were 23 vertical levels from the surface to the top pressure of 50 hPa. June to August was considered summer. The simulations were conducted from 25 May to 31 August 2010 and the first 7 days (25-31 May) were treated as the spin-up period. The initial meteorological and boundary conditions were obtained from the National Centers for Environmental Prediction Final Analysis with a 6 h temporal resolution. The detailed model configurations are described in Table S1.

Anthropogenic emissions were obtained from the MIX inventory (Li et al., 2017), an Asian anthropogenic emissions inventory that harmonizes several local inventories using a mosaic approach. MIX uses Regional Emissions Inventory in Asia (REAS2, version 2) (Kurokawa et al., 2013) for  $\text{NH}_3$  emissions in India.

### 2.2 Observational dataset

Atmospheric total columns of  $\text{NH}_3$  were derived from measurements of an Infrared Atmospheric Sounding Interferometer (IASI) on board MetOp-A (<https://iasi.aeris-data.fr/NH3/>). MetOp-A was launched in 2006 in a Sun-synchronous orbit with a mean local solar overpass time of 9:30 a.m. and 9:30 p.m. Only the daytime measurements have been used here, because the nighttime measurements had larger relative errors caused by the general lower thermal contrast for the nighttime overpass (Van Damme et al., 2014). It has been found that the IASI samples at the overpass time could represent the entire day, and



65 IASI NH<sub>3</sub> observations are in fair agreement with ground-based measurements (Dammers et al., 2016; Van Damme et al.,  
2015a). This work used the ANNI-NH<sub>3</sub>-v2.2R-I retrieval product, which relied on ERA-Interim reanalysis for its  
meteorological inputs (Van Damme et al., 2017). The seasonal mean NH<sub>3</sub> column concentrations over East Asia on a 0.25° ×  
0.25° grid in summer 2010 have been determined based on the relative error weighting mean method (Van Damme et al.,  
2014).

70 Meteorological data at 38 sites over northern India obtained from the National Climate Data Center  
(<https://gis.ncdc.noaa.gov/maps/ncei/cdo/hourly>) were used to evaluate the accuracy of meteorological simulations. The  
evaluated variables included hourly wind speed at 10 m (WS10), wind direction at 10 m (WD10), relative humidity at 2 m  
(RH2) and temperature at 2 m (T2). The statistical parameters included the mean bias (MB), normalized mean bias (NMB),  
root mean square error (RMSE) and correlation coefficient (R). In addition, air temperature at 21 sites over the NCP obtained  
75 from the National Climate Data Center were also used in this work.

### 2.3 ISORROPIA-II thermodynamic model

The thermodynamic equilibrium model, ISORROPIA-II (Fountoukis and Nenes, 2007), treating the thermodynamics of NH<sub>4</sub><sup>+</sup>-  
SO<sub>4</sub><sup>2-</sup>-NO<sub>3</sub><sup>-</sup>-K<sup>+</sup>-Ca<sup>2+</sup>-Mg<sup>2+</sup>-Na<sup>+</sup>-Cl<sup>-</sup>-H<sub>2</sub>O aerosol system, was used to investigate the influence of air temperature on the NH<sub>3</sub>  
total columns. In this study, ISORROPIA-II was run in the “forward mode” and assuming particles are “metastable” with no  
80 solid precipitates. As inputs of ISORROPIA-II, the outputs (water-soluble ions, gas species, T and RH) of WRF-Chem were  
first averaged over the IGP and then averaged for summer 2010. Using ISORROPIA-II, we simulated 20 scenarios. In these  
cases, air temperature of each layer increased or decreased by 10 °C synchronously, with the interval of 1 °C. Meanwhile, the  
other input parameters remained the same. Then, we calculated the columnar  $\epsilon(\text{NH}_4^+)$  (partitioning ratios of NH<sub>4</sub><sup>+</sup> to total  
ammonia (TA, TA = NH<sub>3</sub> + NH<sub>4</sub><sup>+</sup>)) in each case. The columnar  $\epsilon(\text{NH}_4^+)$  is the sum of the  $\epsilon(\text{NH}_4^+)$  of each vertical level, but  
85 each weighted by the thickness of the layer and mass concentration of TA.

## 3 Results

### 3.1 High NH<sub>3</sub> emissions

IGP is a vast stretch of fertile alluvial plain spanning the banks of the Indus and Ganges Rivers and their tributaries. The main  
part of the IGP is located in India. The estimated NH<sub>3</sub> emissions in India in 2010 was 9.9 Tg, which is comparable to that in  
90 China (9.8 Tg) and accounts for about 34 % of total NH<sub>3</sub> emissions in Asia (Li et al., 2017). Agriculture is the largest NH<sub>3</sub>  
emitter in India, accounting for about 76 % of the total NH<sub>3</sub> emissions (Li et al., 2017). Agricultural NH<sub>3</sub> emissions mainly  
originate from animal husbandry and fertilizer application (Bouwman et al., 1997; Streets et al., 2003). India is the second  
largest N-fertilizer consumer (after China) and consumes 16.5 Tg N-fertilizers (16 % of the world’s total) (FAOSTAT, 2010).  
In addition, there are an estimated 302 million cattle and buffalo in India (19 % of world’s total), which is more than any other



95 country (FAOSTAT, 2010). It is estimated that cattle and buffalo account for about 80 % of NH<sub>3</sub> emissions among livestock in India (Aneja et al., 2012).

NH<sub>3</sub> emissions over the IGP was 4.3 Tg in 2010 (estimated using the MIX database), which was mainly attributed to intensive agricultural practices. The IGP is known as the food bowl of India spreading across the states of Punjab, Haryana, Uttar Pradesh, Bihar, and West Bengal (blue quadrangle in Fig. 1). The total number of cattle and buffalo in the five states was estimated to  
100 be 103 million (34 % of the national total) in 2012 (19th Livestock Census 2012, 2012). The total amount of N-fertilizer applied in the five states was estimated to be 6.9 Tg (42 % of the national total) in 2010 (Agricultural Statistics At a Glance 2012, 2012a). NH<sub>3</sub> emissions over the IGP in summer are very high with a regional mean NH<sub>3</sub> emissions flux of 0.4 t km<sup>-2</sup> month<sup>-1</sup> (estimated using MIX database for 2010). This is consistent with satellite observations, which also show a summer peak of NH<sub>3</sub> columns over the IGP (Van Damme et al., 2015b). The summer peak of NH<sub>3</sub> emissions over the IGP might be  
105 the joint result of intensive N-fertilizer applications and high temperature. Summer is the sowing season in the IGP, when a large amount of N-fertilizer applied to the cropland as base fertilizer (Agricultural Statistics At a Glance 2012, 2012a). In addition, the summer air temperature is very high over the IGP with an observed regional mean value of 30.9 °C in summer 2010. Ammonia emissions increase exponentially with temperature (Riddick et al., 2016). The high application rate of N-fertilizer and high air temperature could cause high NH<sub>3</sub> emissions, resulting in the high NH<sub>3</sub> columns.

110 The spatial distribution of seasonal mean NH<sub>3</sub> total columns over East Asia during summer 2010 is shown in Fig. 1. The NH<sub>3</sub> columns over the IGP ( $7.6 \times 10^{16}$  molecules cm<sup>-2</sup>) were about twice as large as what was observed over the NCP ( $4.1 \times 10^{16}$  molecules cm<sup>-2</sup>) in summer 2010. The NCP is also a large agricultural region (Huang et al., 2012). The regional mean NH<sub>3</sub> emissions flux over the NCP was 0.7 t km<sup>-2</sup> month<sup>-1</sup> in summer 2010 (estimated using the MIX database), which was about 1.8 times that of the IGP. The IGP has much higher NH<sub>3</sub> total columns (Fig. 1) compared to the NCP, but lower NH<sub>3</sub> emissions  
115 fluxes (Fig. S1a). Therefore, other factors might lead to the high NH<sub>3</sub> loading over the IGP in summer besides high NH<sub>3</sub> emissions.

### 3.2 Low gas-to-particle conversion of NH<sub>3</sub>

The emissions fluxes of SO<sub>2</sub> and NO<sub>x</sub> (both are 0.3 t km<sup>-2</sup> month<sup>-1</sup>) over the IGP are only about one-fourth of that over the NCP (1.1 and 1.3 t km<sup>-2</sup> month<sup>-1</sup>) (Table 1 and Fig. S1). The relatively low SO<sub>2</sub> and NO<sub>x</sub> emissions could be an important  
120 factor causing the high NH<sub>3</sub> columns over the IGP. In this study, we used the molar ratio ( $R_{emis}$ ) of NH<sub>3</sub> emissions fluxes ( $E_A$ ) to the sum of twice the SO<sub>2</sub> emissions fluxes ( $E_S$ ) and NO<sub>x</sub> emissions fluxes ( $E_N$ ) to roughly represent the richness of NH<sub>3</sub> in the atmosphere, given by Eq. (1):

$$R_{emis} = \frac{E_A}{2 \times E_S + E_N} \quad (1)$$

The calculated  $R_{emis}$  in the IGP was 1.35, which was about 2.6 times as large as that in the NCP (0.51). We performed  
125 simulations for a base case and a ‘increased SO<sub>2</sub>/NO<sub>x</sub> emissions’ case to investigate the impact of SO<sub>2</sub> and NO<sub>x</sub> emissions on



NH<sub>3</sub> loading. In the increased SO<sub>2</sub>/NO<sub>x</sub> emissions case, the emissions of SO<sub>2</sub> and NO<sub>x</sub> increased 2.6 times to make R<sub>emis</sub> of the IGP equal to that of the NCP.

The simulated NH<sub>3</sub> columns in the base case are shown in Fig. 2a. The regional mean NH<sub>3</sub> total column over the IGP from the base case was  $8.8 \times 10^{16}$  molecules cm<sup>-2</sup>. It is noted that the IASI NH<sub>3</sub> columns cannot be quantitatively compared to modeled  
130 NH<sub>3</sub> columns as the IASI NH<sub>3</sub> products do not provide information on the vertical sensitivity to properly weight the model values. Nonetheless, the spatial distribution of the NH<sub>3</sub> columns in the base run was consistent with the satellite observations, both of which showed the highest values in northwestern IGP (Figs. 1 and 2a). The statistical performance of the meteorological predictions at 38 sites over Northern India are presented in Table S2. The predicted T2 matched well with the observations with a correlation coefficient of 0.8 and an NMB of 4.2 %. The predicted RH2 was slightly underestimated with an NMB of –  
135 13.4 % and a correlation coefficient of 0.8. The predicted WS10 agreed reasonably well with the observations with an NMB of –5.3 %. In addition, the simulated WD10 matched well with the observations, and both the predicted and observed dominant wind direction was SSE. The good agreement between the simulation and the observations confirms the reliability of the meteorological prediction over the simulation domain.

The spatial distribution of the NH<sub>3</sub> total column in the increased emissions case is shown in Fig. 2b. The NH<sub>3</sub> total columns  
140 significantly decreased over the entire IGP, with a regional mean value of  $2.5 \times 10^{16}$  molecules cm<sup>-2</sup> (a 72.2 % decrease compared to the base case). The surface  $\epsilon(\text{NH}_4^+)$  in the base case and the increased SO<sub>2</sub>/NO<sub>x</sub> emissions case are shown in Fig. 2 (panels c and d, respectively). The surface  $\epsilon(\text{NH}_4^+)$  in the base case was fairly low with a regional mean value of 0.3 over the IGP, close to the observations in Delhi (0.4 in summer 2010) (Singh and Kulshrestha, 2012). In the increased SO<sub>2</sub>/NO<sub>x</sub> emissions case, the regional mean surface  $\epsilon(\text{NH}_4^+)$  increased to 0.6 over the IGP. Significant increases in surface SO<sub>4</sub><sup>2-</sup> and  
145 NO<sub>3</sub><sup>-</sup> concentrations were also found (Fig. S2). The regional mean surface SO<sub>4</sub><sup>2-</sup> and NO<sub>3</sub><sup>-</sup> concentrations increased from 9.7 to 24.9 and from 7.2 to 20.0 μg m<sup>-3</sup>, respectively. Additionally, the regional mean columnar  $\epsilon(\text{NH}_4^+)$  over the IGP is 0.56 in the base case and increases to 0.87 in the increased SO<sub>2</sub>/NO<sub>x</sub> emissions case. This suggests that the increased SO<sub>2</sub> and NO<sub>x</sub> emissions enhanced the formation of acidic species and promoted the conversion of NH<sub>3</sub> into NH<sub>4</sub><sup>+</sup>. The effectively reduced NH<sub>3</sub> total columns in the increased SO<sub>2</sub>/NO<sub>x</sub> emissions case indicate that low SO<sub>2</sub> and NO<sub>x</sub> emissions could be the major cause  
150 of the high NH<sub>3</sub> loading over the IGP.

Besides the amount of acidic species, air temperature is also an important factor affecting the thermodynamic equilibrium of NH<sub>3</sub> between the gas phase and the particle phase. Higher air temperature limits the gas-to-particle conversion of NH<sub>3</sub> and enhances volatilization of NH<sub>4</sub>NO<sub>3</sub> (Seinfeld and Pandis, 2006). The observed regional mean air temperature over the IGP in summer 2010 was 30.9 °C, about 4.9 °C higher than the NCP (26.0 °C). Thermodynamic calculation was conducted using  
155 ISORROPIA-II based on WRF-Chem outputs (The detailed descriptions are in Sect. 2.3). As shown in Fig. 3, the columnar  $\epsilon(\text{NH}_4^+)$  increases as temperature decreases. A 10°C decrease in temperature results in a 0.07 increase in  $\epsilon(\text{NH}_4^+)$  and a consequent 17 % decrease in NH<sub>3</sub> total columns. Additionally, a 10°C increase in temperature results in a 0.08 decrease in  $\epsilon(\text{NH}_4^+)$  and a consequent 20 % increase in NH<sub>3</sub> total columns. If the temperature over the IGP drops to the temperature typical of the NCP (a 4.9 °C decrease), the NH<sub>3</sub> total columns over the IGP will only decrease by 10 %. In contrast, if the SO<sub>2</sub>/NO<sub>x</sub>



160 emissions over the IGP increase to make the  $R_{emis}$  of the IGP equal to that of the NCP, the  $NH_3$  column over the IGP will decrease by 72.2 %. Therefore, the low  $SO_2/NO_x$  emissions have a greater effect on causing high  $NH_3$  columns over the IGP than the high air temperature.

### 3.2 Weak horizontal diffusion of $NH_3$

The IGP is located in the northern part of the Indian subcontinent and is surrounded by unique topography with the Himalayan range to the north and the Sulaiman range to the west. Weather on the Indian subcontinent is controlled by the low-level Indian monsoon regime from June through September (Lawrence and Lelieveld, 2010). Fig. 4a shows the spatial distributions of surface wind flow and wind speed in summer 2010. The dominant wind direction is southwest over the Indian peninsula and southeast over the IGP. Air mainly flows from the west coast of India and the south coast of Bengal. Surface wind speed is high on the west coast of India ( $>5 \text{ m s}^{-1}$ ) and on the south coast of Bengal ( $>4 \text{ m s}^{-1}$ ) but decreases from the coast inland. Mountains serve as barriers to the airflow on the surface of the Earth (Barry, 2008). Chow et al. (2013) reported that when stably stratified airflow encounters an extra-tropical mountain barrier, it is forced to rise and cool adiabatically. Consequently, higher pressure along the slope could be created, which could decelerate and block the flow. After a while, geostrophic adjustment occurs. As a result, the airflow turns left (right) in the northern (southern) hemisphere, and a barrier jet blowing parallel to the barrier is formed. As shown in Fig. 4a, the southerly airflow from the Bay of Bengal turns left when approaching the Himalayas, and then an easterly barrier jet parallel to the Himalayas is formed. The southwesterly airflow from the west coast of India also turns left when approaching the Himalayas. Similarly, wind flow at 850 hPa (Fig. 4b) also shows left-turning airflow near the Himalayas. The left-turning airflow indicates that the barrier effect of the Himalayas limits the northward movement of polluted air. Both satellite observations and the model simulation show that the high  $NH_3$  columns over the IGP are effectively cut off by mountains to the north (Fig. 2a and Fig. S3).

180 As shown in Fig. 4b, an area of low geopotential height extends from Pakistan to east India following the IGP. This elongated region of low pressure is known as the monsoon trough (Bohlinger et al., 2017). It causes wind to converge over this region. The convergence of horizontal wind can be observed from wind flow at both the surface and at 850 hPa. The prevailing wind directions south and east of the IGP are southwest and southeast, respectively. As a result of convergence of horizontal winds, an area of low wind speed forms and covers most of the IGP. The regional mean surface wind speed over the IGP is  $<3 \text{ m s}^{-1}$ .  
185 The weak wind speed in association with the convergence weakens the horizontal advection of  $NH_3$  and results in the accumulation of  $NH_3$  over the IGP.

The ventilation rate ( $V_r$ ) of the four edges of the IGP was used to illustrate the accumulation of an air mass over the IGP (Fig. 4d). The  $V_r$  of one edge is defined as the product of sectional area to the transport wind, given by Eq. (2):

$$V_r = AU_T \quad (2)$$



190 The sectional area  $A$  can be expressed as  $A = ZL$ , where  $Z$  is the mean boundary layer height along the edge and  $L$  is the length

$$U_T = \frac{1}{m} \sum_{j=1}^m \left( \frac{1}{n} \sum_{i=1}^n U_{ij} \right)$$

of the edge. The transport wind  $U_T$  is given by  $m$  and  $n$  are the number of locations along the edge and vertical levels within the ABL where the winds are measured or predicted.  $U_{ij}$  is the wind speed perpendicular to the cross-section at each height and location along the edge. The ventilation rates of the four edges of the IGP were calculated using the WRF-Chem simulation results. The total  $V_r$  of the inflow from the southern and eastern edges ( $3.1 \times 10^9 \text{ m}^3 \text{ s}^{-1}$ ) was 64 %  
195 higher than the total  $V_r$  of the outflow from the western and northern edges ( $1.9 \times 10^9 \text{ m}^3 \text{ s}^{-1}$ ). The strong inflow and weak outflow indicate accumulation of the air mass over the IGP in summer. Therefore, outward transport of  $\text{NH}_3$  from the IGP through horizontal advection could be weak.

Interestingly, both relative humidity and precipitation are high over the IGP (Figs. S4), with regional mean values of 63 % and 660 mm during summer 2010. The high relative humidity and precipitation suggest strong gas-to-particle conversion and wet  
200 scavenging of  $\text{NH}_3$  (Seinfeld & Pandis, 2006). The observed high  $\text{NH}_3$  loading under such a wet condition further indicates the effectiveness of other factors leading to high  $\text{NH}_3$  loading. The meteorological conditions in the northwest IGP are characterized by higher air temperature, lower humidity, and lower rainfall compared to the southeast IGP (Figs. 4c and S4), all of which are conducive to the increase of  $\text{NH}_3$ . Consistently,  $\text{NH}_3$  total columns decrease from northwest to southeast along the IGP as revealed by both the satellite measurements and model simulations (Figs. 1 and 2a). However, emission fluxes of  
205  $\text{NH}_3$  over the northwest IGP are also obviously higher than the southeast IGP (Fig. S1). To exclude the impact of emissions on the spatial distributions of  $\text{NH}_3$ , simulations for a “homogeneous emissions” case was performed by using WRF-Chem, where emissions of all primary pollutants over the IGP were set to their regional mean values. As shown in Fig. 5,  $\text{NH}_3$  total columns in the homogeneous emissions case still appear to decrease from northwest to southeast along the IGP. It is indicated that the meteorological factors (atmospheric diffusion, temperature, relative humidity, and precipitation) are important causes  
210 of the higher  $\text{NH}_3$  loadings over the northwest IGP than the southeast IGP.

#### 4 Conclusions

Satellite observations have revealed that the IGP has the global maximum  $\text{NH}_3$  loading with a peak in summer. Our study reveals that the high  $\text{NH}_3$  loading over the IGP appears to be the joint result of high  $\text{NH}_3$  emissions, weak chemical loss, and weak horizontal diffusion. Intensive agricultural activities in combination with high temperature resulted in relatively high  
215  $\text{NH}_3$  emissions over the IGP, with a regional mean  $\text{NH}_3$  emissions flux of  $0.4 \text{ t km}^{-2} \text{ month}^{-1}$ . The low  $\text{SO}_2$  and  $\text{NO}_x$  emissions and high temperature limited the conversion of  $\text{NH}_3$  to  $\text{NH}_4^+$ . The low chemical loss of gaseous  $\text{NH}_3$  is a key reason for the high  $\text{NH}_3$  loading over the IGP. In addition, orographic and meteorological conditions also play important roles in  $\text{NH}_3$  accumulation over the IGP. The barrier effects of the Himalayas limit the northward movement of monsoon air in summer.



220 The low wind speed ( $<3 \text{ m s}^{-1}$ ) in association with the surface convergence over the IGP weakens horizontal diffusion, which is conducive to the accumulation of  $\text{NH}_3$  over the IGP.

### Data availability

The IASI data used in this study was provided by the AERIS data infrastructure (<https://iasi.aeris-data.fr/NH3/>). The meteorological data used in this study was obtained from the National Climate Data Center integrated surface database (<https://gis.ncdc.noaa.gov/maps/ncei/cdo/hourly>). The anthropogenic emissions are available from MIX inventory  
225 ([www.meicmodel.org/dataset-mix](http://www.meicmodel.org/dataset-mix)).

### Author contribution

Y.S initiated the investigation. T.W performed the modelling analyses. T.W, Y.S, Z.X and T.Z wrote and edited the manuscript. M.L, T.X, W.L, L.Y, X.C, H.Z and L.K contributed to discussions of the results and the manuscript.

### Competing interests

230 The authors declare no competing interests.

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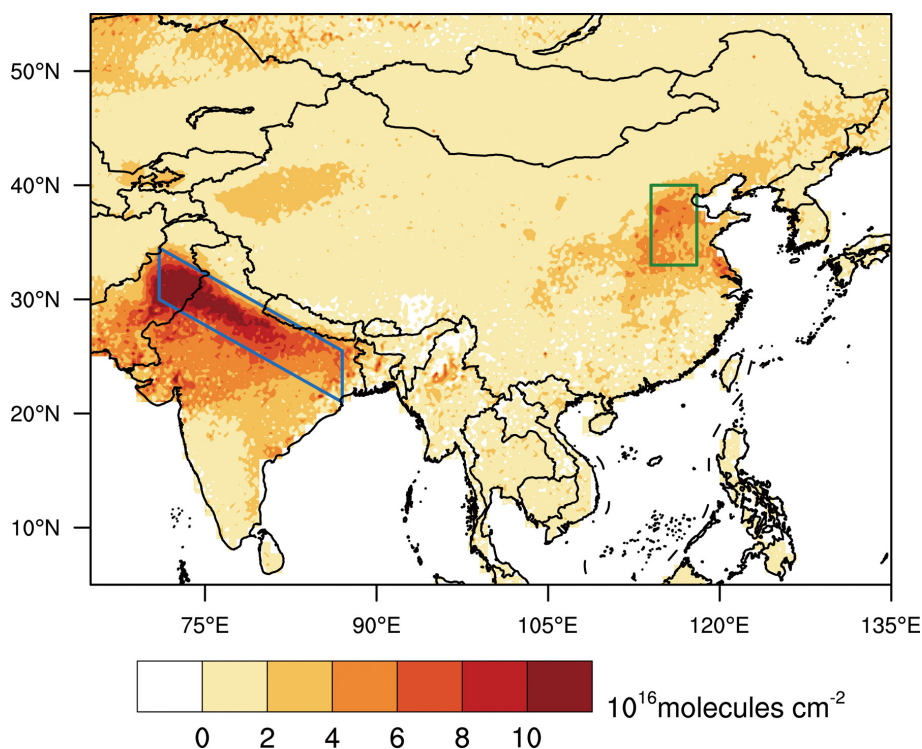


**Table 1.** Regional Mean  $\text{NH}_3$  Total Columns and Emissions Fluxes of  $\text{NH}_3$ ,  $\text{SO}_2$ , and  $\text{NO}_x$  of the IGP and the NCP in Summer 2010.

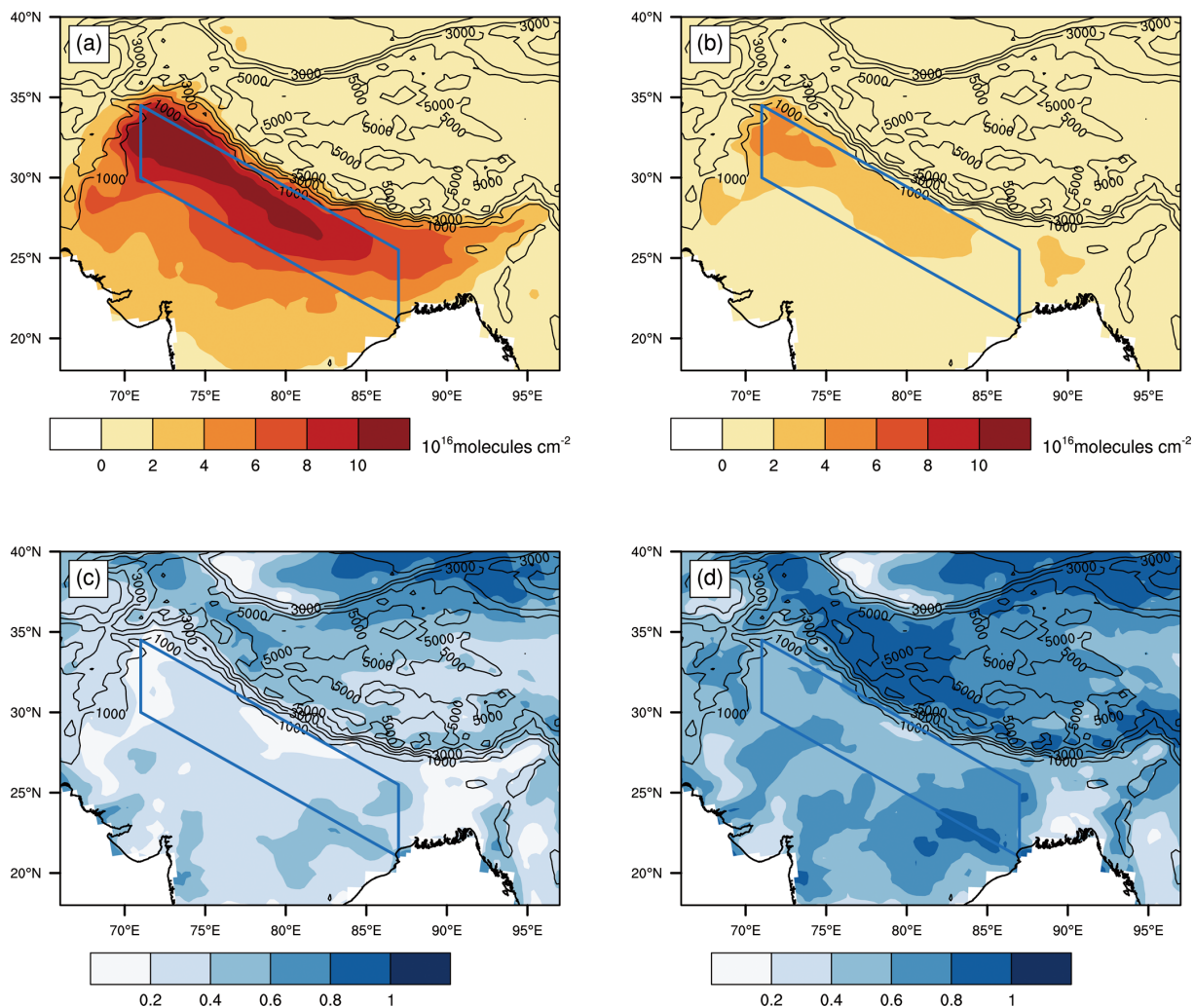
	$\text{NH}_3$ total columns <sup>a</sup> (molecules $\text{cm}^{-2}$ )	Emissions fluxes <sup>b</sup> ( $\text{t km}^{-2} \text{month}^{-1}$ )		
		$\text{NH}_3$	$\text{SO}_2$	$\text{NO}_x$
IGP	$7.6 \times 10^{16}$	0.4	0.3	0.3
NCP	$4.1 \times 10^{16}$	0.7	1.1	1.3

325 <sup>a</sup> $\text{NH}_3$  total columns were derived from IASI measurements

<sup>b</sup>Emissions fluxes were estimated using the MIX database

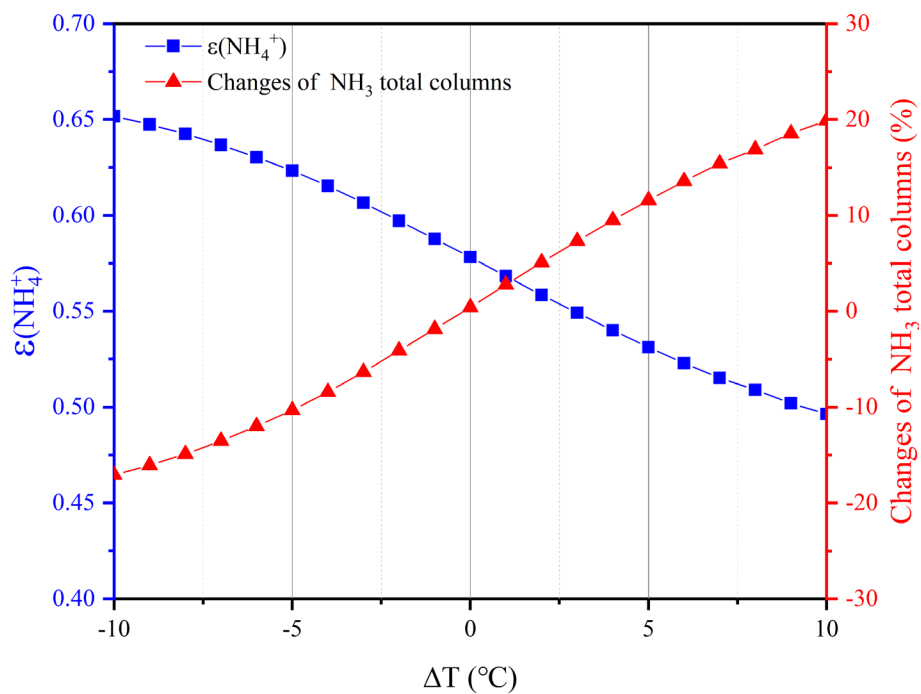


**Figure 1.** The spatial distribution of  $\text{NH}_3$  total columns over East Asia in summer 2010 retrieved from  
330 IASI measurements. The blue quadrangle represents the Indo-Gangetic Plain (IGP), and the green  
quadrangle represents the Northern China Plain (NCP).

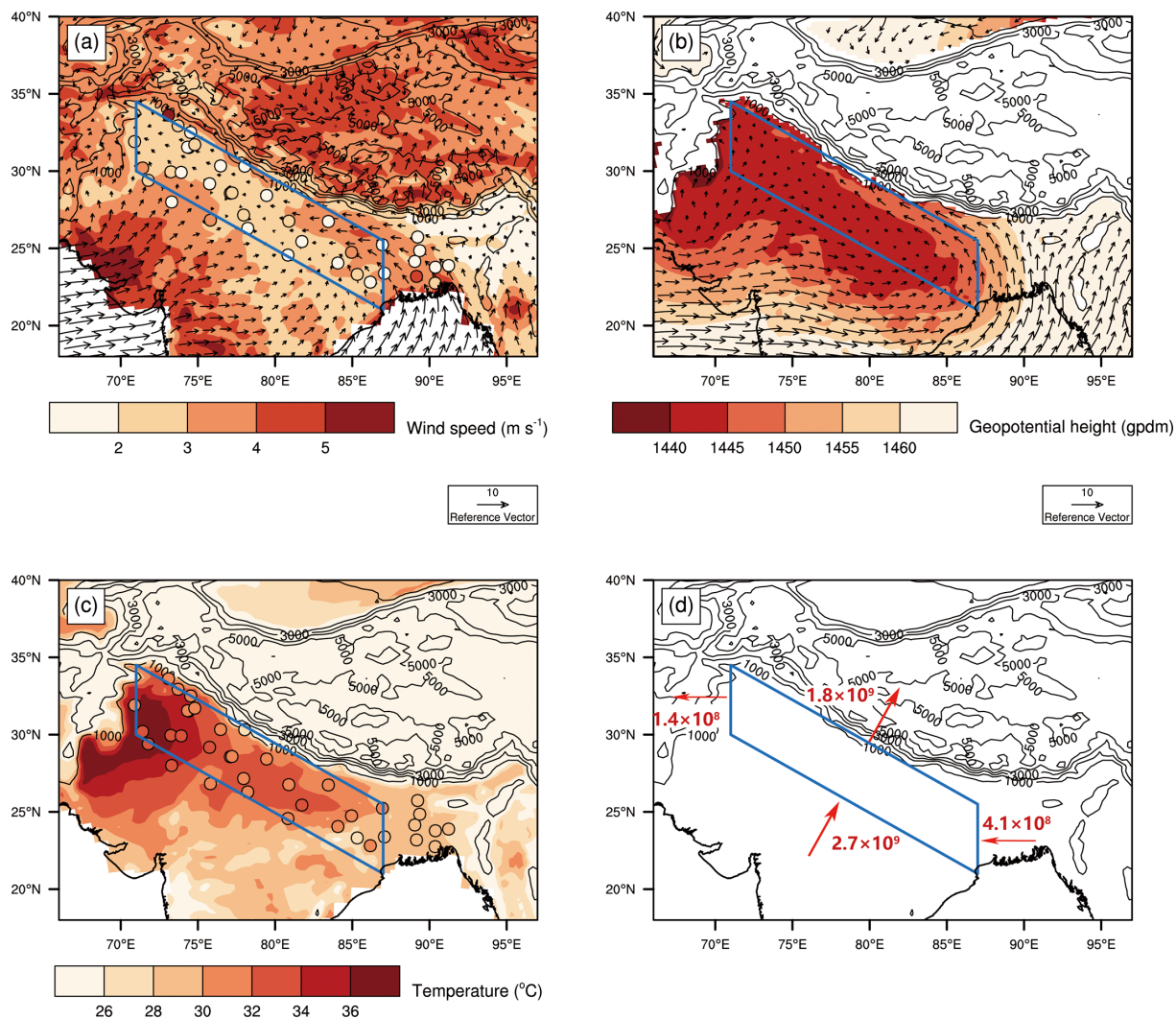


**Figure 2.** Spatial distributions of WRF-Chem predicted total columns of  $\text{NH}_3$  and surface  $\epsilon(\text{NH}_4^+)$  in summer 2010. (a) and (b) are total columns of  $\text{NH}_3$  in the base case and the increased  $\text{SO}_2/\text{NO}_x$  emissions case, respectively. (c) and (d) are surface  $\epsilon(\text{NH}_4^+)$  in the base case and the increased  $\text{SO}_2/\text{NO}_x$  emissions case, respectively.

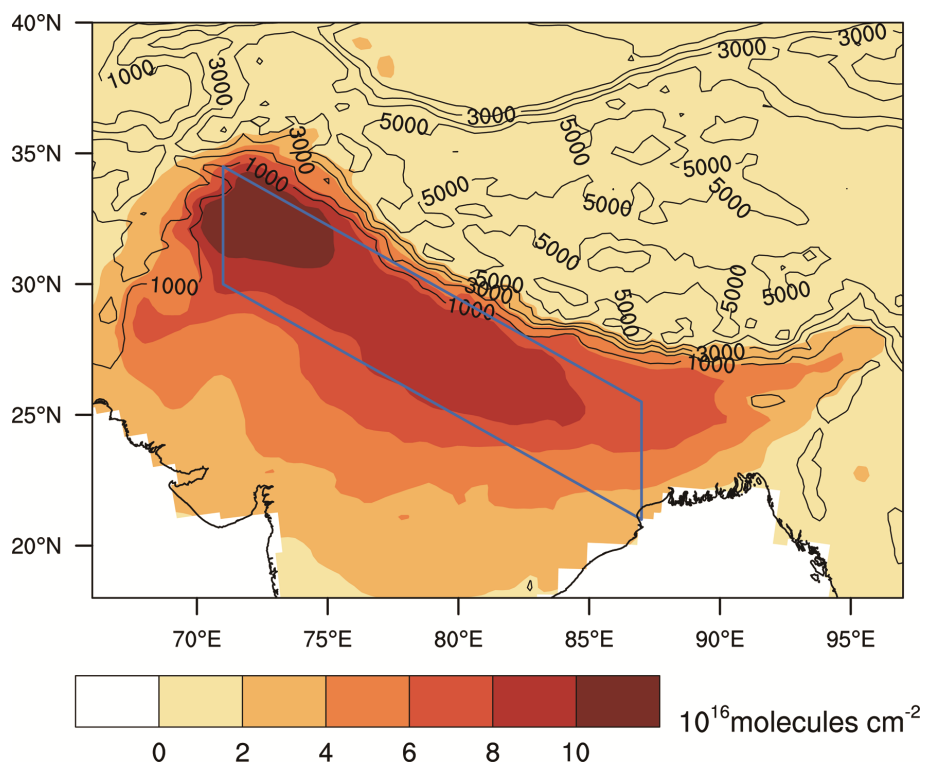
335



340 **Figure 3.** Columnar  $\epsilon(\text{NH}_4^+)$  and changes of  $\text{NH}_3$  total columns with the changes of temperatures predicted by ISORROPIA-II.



**Figure 4.** Spatial distributions of WRF-Chem predicted meteorological variables in summer 2010. (a) Wind flow and wind speed at 10 m. (b) Wind flow and geopotential height at 850 hPa. (c) Air temperature at 2 m. (d) Ventilation rate ( $\text{m}^3 \text{s}^{-1}$ ) of the four edges of the IGP. Circles in (a) and (c) show the observed wind speed at 10 m and air temperature at 2 m, respectively.



**Figure 5.** Spatial distributions of WRF-Chem predicted total columns of  $\text{NH}_3$  in summer 2010 in the homogeneous emissions case.