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24 <u>ABSTRACT</u>

25 Hygroscopicity plays crucial roles in determining aerosol optical properties and aging processes 26 in the atmosphere. We investigated submicron aerosol hygroscopicity and composition by 27 connecting an aerosol time-of-flight mass spectrometer (ATOFMS) in series to a hygroscopic tandem differential mobility analyzer (HTDMA), to characterize hygroscopicity and 28 29 composition of ambient aerosols in Shanghai, China. The HTDMA-ATOFMS data suggested 30 that particle types, including biomass burning, EC, Dust/Ash, organics particles, cooking 31 particles and sea salt, were shown to have distinct hygroscopicity distributions. Peak intensities 32 in particle spectra were found nonlinearly correlated with hygroscopicity and the correlations were variant with particle types. Based on the measured hygroscopicity-composition relations, 33 34 we developed a statistical method to estimate ambient particle hygroscopicity just from their 35 mass spectra. The method was applied to another ambient ATOFMS dataset sampled during September 12nd to 28th, 2012 in Shanghai. The estimated hygroscopicity suggested that ambient 36 37 particles were present in three apparent hygroscopicity modes, whose growth factors peaked at 38 1.05, 1.42 and 1.60 (85% RH), respectively. The estimated GF were divided into four bins as 39 <1.1, 1.1-1.3, 1.3-1.5 and >1.5 to represent the nearly-hydrophobic (NH), less-hygroscopic 40 (LH), more-hygroscopic (MH) and sea salt (SS) mode. Number contributions of particle types 41 to hygroscopicity modes showed consistent results with the HTDAM-ATOFMS experiment. 42 Based on the combined information on particle composition, hygroscopicity, airmass back 43 trajectories and ambient pollutants concentrations, we inferred that the NH, LH, MH, SS modes 44 were characterized by POA/EC, SOA, SIA and salts compositions, respectively. The proposed 45 method would provide additional information to the study of particle mixing states, source 46 identification and visibility degradation.

48 <u>1. INTRODUCTION</u>

49 Atmospheric particles have critical impacts on climate and the environment. They affect climate 50 by directly interacting with sunlight and changing the energy balance of earth's atmosphere (Facchini et al., 1999;Lohmann and Feichter, 2005). Aerosol particles also act as cloud 51 52 condensation nuclei or ice nuclei and impact cloud formation (Lohmann and Feichter, 2005). 53 Aerosol particles provide surfaces for heterogeneous reactions to occur and act as the sink for 54 many atmospheric reactions (Gard et al., 1998; Qiu and Zhang, 2013), which are of significance 55 to air quality, visibility and human health. The climate-relevant and other properties of aerosols are largely determined by their hygroscopicity. In atmospheric conditions the hygroscopic 56 57 growth transform particles into micro droplets and their optical effects are altered importantly 58 (Cheng et al., 2008;Qu et al., 2015), which further impacts particle aging processes and 59 visibility degradation (Qu et al., 2015;Liu et al., 2012;Qiu and Zhang, 2013;Chen et al., 2012).

Atmospheric particles are a mixture of a complicated variety of chemical compounds. The bulk 60 61 chemical composition of particulate matter (PM) usually refers to its dry composition. However, 62 in ambient conditions the particulate water is also an important constitution of PM which has 63 not been accounted for in conventional gravitational analysis. The mass of particulate water may be much larger than the total mass of dried PM at elevated RHs (Swietlicki et al., 2008). 64 The factors that affect water contents in particles include the particle hygroscopicity, the particle 65 size distributions and ambient RH. The hygroscopicity parameter determines the particle's 66 ability to growth in their sizes in humidity environment, which is directly relates to particle 67 composition and size. To accurately predict particulate water content, a detailed knowledge on 68 particle hygroscopicity and composition is needed (Gysel et al., 2007;Sjogren et al., 69 70 2008;Laborde et al., 2013;Healy et al., 2014).

Aerosol hygroscopicity can be quantitatively measured by techniques such as the Hygroscopic Tandem Differential Mobility Analyzer (HTDMA) (Swietlicki et al., 2008). Previous studies measured aerosol hygroscopicity and chemical composition simultaneously by deploying HTDMA and chemical composition measurements in parallel (Gysel et al., 2007;Sjogren et al., 2008;Laborde et al., 2013). The measured hygroscopicity was compared with the reconstructed values using the mixing rules of variant compositions (Gysel et al., 2007). The hygroscopicity 77 reconstructed in this way is representing the averaged hygroscopicity of ensembled particles 78 and therefore could not reflect the mixing states of particles (Healy et al., 2014). However, the 79 HTDMA measurements suggested that several hygroscopicity modes exist simultaneously, 80 which evidenced the external mixing state of atmospheric particles. HTDMA hygroscopicity modes were generally fell into four categories: nearly hydrophobic (NH), less-hygroscopic 81 82 (LH), more-hygroscopic (MH), and sea-salt mode, with their center GFs (90% RH) lie in 1.0-1.11, 1.11-1.33, >1.33 and >1.8 ranges, respectively (Swietlicki et al., 2008). To investigate the 83 84 chemical nature of these hygroscopicity modes, it is better to connect HTDMA and composition 85 measurement techniques in tandem, since more direct connection between hygroscopicity and 86 composition could be established in this way (Buzorius et al., 2002;Zelenyuk et al., 2008;Herich et al., 2008;Laborde et al., 2013). For composition measurement techniques, single 87 particle mass spectrometers are preferred since they are sensitive to analyze particles of low 88 89 concentrations in the HTDMA outflow (Herich et al., 2008;Herich et al., 2009), and that the 90 particle mixing state information is preserved during analysis (Healy et al., 2014).

91 Only a few studies have reported simultaneous characterization of hygroscopicity and 92 composition using the tandem method (Herich et al., 2008;Herich et al., 2009;Buzorius et al., 93 2002;Zelenyuk et al., 2008). Zelenyuk et al. connected a single particle mass spectrometer 94 SPLAT with HTDMA to demonstrate the capability of this system to derive quantitative information on aerosol hygroscopicity, composition, and effective density (Zelenvuk et al., 95 96 2008). Herich et al. firstly applied the tandem HTDMA-ATOFMS system to characterize 97 particle composition of different hygroscopicity (Herich et al., 2008;Herich et al., 2009). A large 98 portion of the less hygroscopic modes were found to be contributed by organics and combustion 99 species both in the urban and remote site, while the sulfates and nitrates were present in almost 100 all particles independent of hygroscopicity. Similar findings were also observed in our 101 preliminary characterization using HTDMA-ATOFMS in Shanghai city (Wang et al., 2014), 102 except higher nitrate and sulfate intensities were found in hygroscopic particles in our study. However, the preliminary dataset was not sufficiently large since only a few GF were 103 104 characterized in that study (GF 1.05-1.1, 1.3, 1.4 and 1.5 at 85% RH). The primary objective of the present study is to establish more complete connections between hygroscopicity and 105

single particle signatures, which could be further utilized to predict hygroscopicity of ambient
particles. We conducted a comprehensive HTDMA-ATOFMS experiment with the particle GF
varied in a more complete range (0.9~1.7, 85% RH), which accounted for the main number
fraction of atmospheric particles in urban atmosphere (Liu et al., 2014;Liu et al., 2011;Ye et al.,
2013). Based on the HTDMA-ATOFMS data, we further developed and tested a method to
estimate the hygroscopicity of ambient particles analyzed by ATOFMS.

112 <u>2. EXPERIMENTAL SECTION</u>

113 **2.1. HTDMA**

114 The custom-built HTDMA (Ye et al., 2009) consists of two DMAs and a humidifier connected in series (Figure 1). Aerosol was dried before entering HTDMA (RH ~10%) by a diffusional 115 116 silica gel tube. The dried aerosol reached charge equilibrium in a Kr-85 neutralizer. The DMA1 (Model 3081, TSI Inc.) selected particles by electrical mobility size as D_{dry} . The monodisperse 117 particles from DMA1 grew in a Nafion humidifier (RH=85%). The sizes of humidified particles 118 $D_{\rm RH}$ was determined by the second DMA connected by a CPC to measure their concentrations. 119 120 The sheath flow rate in DMA2 (3 l/min) was regulated by mass flow controller. The RH of the 121 DMA2 sheath flow was managed to match the humidifier (85% RH). The HTDMA were 122 installed in thermostatic chamber in which temperature was controlled to 25 ± 0.1 °C. The total 123 aerosol flow was 0.4 L/min (the sum flow rate of the CPC, 0.3 L/min and the ATOFMS, 0.1 124 L/min). Aerosol residence time in humidifier was ~10 s. PSL spheres of known size and (NH₄)₂SO₄ salt were used to calibrate the HTDMA. The HTDMA uncertainty in GF 125 126 determination is ± 0.05 (Ye et al., 2009;Swietlicki et al., 2008).

127 2.2. ATOFMS

The schematic of ATOFMS (Model 3800-100, TSI. Inc) is illustrated in Figure 1. Particles were drawn into ATOFMS through a 0.1mm orifice and focused into narrow beam through successive expansions and contractions in the aerodynamic focusing Lens (AFL). Particles leaving the AFL obtain velocities depending on their aerodynamic sizes. In the ATOFMS sizing region particles pass through two orthogonally oriented continuous lasers (Nd: YAG, 532 nm) and laser light was scattered. The scattered light generates pulses in two photomultiplier tubes (PMT) and the signal delay between the two pulses is used to calculate particle velocity. Particle
velocity was also used to trigger ionization laser (Nd: YAG, 266 nm) at exact time to ionize
particles. The negative and positive ions generated from particles are recorded by a dual polar
time-of-flight mass spectrometer. More details of ATOFMS were described elsewhere (Su et
al., 2004).

139 The ATOFMS data was analyzed within the YAADA toolkit (http://www.yaada.org/). Particles 140 showing similar composition were classified by the adaptive resonance theory-based clustering 141 algorithm (ART-2a) (Song et al., 1999). The ART-2a algorithm parameters were set to: vigilance 142 factor = 0.85, learning rate = 0.05 and number of iterations = 20. The clusters generated by the ART-2a were manully regrouped into major types by considering their common composition 143 144 patterns. The obtained particle types were labelled by refering to previous single particle characterization studies (Spencer et al., 2006;Silva et al., 1999;Sullivan et al., 2007;Gaston et 145 146 al., 2011;Qin et al., 2012).

147 **2.3. Sampling description**

148 The HTDMA-ATOFMS characterization was carried out at the building of department of environmental science and technology in Fudan university (31°18'N, 121°29'E) from Feb-26 to 149 Mar-7, 2014. Aerosol inlet was installed at the building roof about 6 m above the ground. The 150 151 Fudan campus was influenced by local emissions sources from transportation, residential, 152 business and cooking activities from surrounding areas which can be viewed as urban environment. A period of ambient ATOFMS data, which persisted from Sep-12 to Sep-28, 2012, 153 was recorded at the same site in Fudan campus. Ambient air quality data of pollutants 154 155 concentrations (PM_{2.5}, O₃, and SO₂) in Shanghai city were provided by Shanghai Environmental 156 Monitoring Center (SEMC).

The sampling procedure was similar to our previous study (Wang et al., 2014). The typical HTDMA GF distributions in this site showed two separated hygroscopicity modes. In 85% RH condition, the two modes were present with respective center GF of 1.05 and 1.45, which were conventionally classified as Near-Hydrophobic (NH) and More-Hygroscopic (MH) modes, respectively (Swietlicki et al., 2008). These modes were normally present elsewhere in China and other areas (Liu et al., 2011;Swietlicki et al., 2008). The HTDMA data suggest that the
majority of particles (>97 %) were of GF in 0.9-1.7 range. Consequently, particles in this GF
range were characterized by HTDMA-ATOFMS with a GF step of 0.1.

To characterize the desired GFs, the two DMAs in HTDMA were set at certain diameters D_{dry} 165 166 and $D_{\rm RH}$ according to GF= $D_{\rm RH}/D_{\rm dry}$. The HTDMA-ATOFMS system was kept sampling until a sufficient number of particles (> 200) were analyzed by ATOFMS for each GF setting (Table 167 168 1). We fixed the DMA1 (D_{dry}) diameter to 250 nm, while the DMA2 diameter (D_{RH}) was set as 169 shown in Table 1. The number of particle spectra in ATOFMS was affected by ambient particle 170 concentrations of certain GF. Since particle concentrations in the downstream of HTDMA were 171 very low, longer sampling were maintained to record sufficient number of spectra in ATOFMS 172 (See the CPC concentrations in Figure S1). The ATOFMS instrument used in this study has size detection range of 100-3000 nm. Considering this, the detection efficiency for 250 nm is 173 174 expected to be low, as 250 nm is on the lower end of ATOFMS detection range. However, in 175 HTDMA-ATOFMS experiment we selected 250 nm particles, because the concentrations of 176 larger particles were found to decrease further in SMPS size distributions. Therefore, the 177 selection of D_{drv} as 250 nm is a compromise between detection efficiency and particle concentrations (Wang et al., 2014:Herich et al., 2008). With the measured ATOFMS particle 178 179 numbers and CPC concentrations, the detection efficiency of ATOFMS were calculated to be $\sim 1.6 \times 10^{-2}$ at the dry size. In Figure S2 we presented ATOFMS detection efficiencies together 180 181 with the particle hit rate (hit particles/total sized particles) at different growth factors. Generally, 182 the detection efficiencies suggested variations at different GFs. We found higher detection 183 efficiencies in moderate GF range (1.2-1.4) and higher hit rates in GF<1.3 range, which is probably caused by variations of compositions with GF (Hatch et al., 2014). 184

185 **2.4. Estimation of ambient particle hygroscopicity**

Particles in HTDMA-ATOFMS dataset is comparable to ambient ATOFMS particles. Particle types typically present in HTDMA-ATOFMS study were also preset in ambient ATOFMS studies. Therefore, it is possible to assign similar GF to ambient particles if they have similar composition. The estimation method was firstly performed by evaluating spectra similarities between ATOFMS and HTDMA-ATOFMS dataset (dot products of normalized spectra). The

191 ATOFMS is known to have higher detection efficiencies toward some metals (such as Na, K, 192 Fe), resulting inappropriately large peaks in particle spectra. We solved the bias by taking the 193 0.5 power treatment to peaks intensities (Rehbein et al., 2012). In this treatment the larger peaks were suppressed in some degree while smaller peaks increased their weigh relatively. The 0.5 194 power treatment to peaks intensity was applied because it offered better results in the estimation 195 196 of hygroscopicity than without it, as discussed in the supplemental information (Figure S3). In 197 the second step we searched matched particles from the HTDMA-ATOFMS dataset showing 198 the best similarities with the ambient particle (dot products in 95-100% range of the maximum dot product). In this study we set a threshold similarity (0.7 dot product) in matching particles, 199 200 as was required in ART-2a algorithm (Song et al., 1999). Ambient particles with matching dot 201 products <0.7 were excluded from analysis of the estimated GF. The similarity data suggests 202 that 96.2% of the matching similarities are > 0.7 and 79% of them are > 0.8 (Figure S4). Since 203 each of the matched particles in HTDMA-ATOFMS dataset was associated with a GF, we obtained a collection of the matching GFs (0.9 -1.7 in 0.1 step). The estimated GF of the 204 ambient particle was determined to be the weighted average of the matched GFs, with the 205 206 weights being the number percentage of matched particles in each GF bins:

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$$GF_{pred} = \frac{\sum GF_i \cdot F_i}{\sum F_i}$$

where: GF_{pred} = the estimated GF of ambient particle, GF_i = GF value from 0.9 to 1.7 interspaced by 0.1, F_i = number percentages of the matched particles in each GF bin.

210 The estimation process relied on statistical approach in estimating the most probable 211 hygroscopicity for ambient particles, rather than by inferring particle compositions of single 212 particles (Healy et al., 2014). The latter method derived quantitative concentrations of various 213 compositions from peak intensities, which were then applied to predict particle hygroscopicity 214 using the Zdanovskii-Stokes-Robinson mixing rules. Some assumptions including material densities were needed in that method. As a comparison, we inferred hygroscopicity by matching 215 216 particles with HTDMA-ATOFMS particles of known hygroscopicity. Therefore, the estimated GFs were derived from the measured GF and the assumptions of composition densities and 217 detection sensitivities in ATOFMS are obviated. 218

219 The uncertainties in the GF prediction in this method were estimated. The uncertainties in 220 eventual GF may stem from the intrinsic uncertainties in HTDMA-ATOFMS techniques. For 221 the estimation algorithm itself, only few parameters exist that are capable to affect the estimated 222 GF. With the 0.5 power treatment to peak intensities, the only parameter that could influence the estimated GF would be the matching criteria of particles. We have adjusted the matching 223 criteria of 95-100% maximum dot products to 90-100% and 98-100% and the variations in 224 particle GFs were inspected (Figure S5). Based on the variations of the obtained GF, we 225 226 estimated that the uncertainty in GF estimation is within ± 0.15 .

227 3. RESULTS AND DISCUSSIONS

228 **3.1. Single particle composition and hygroscopicity in HTDMA-ATOFMS experiment**

229 **3.1.1 Hygroscopicity distribution of different particle types**

230 The particles in the HTDMA-ATOFMS dataset were classified into major types based on their 231 mass spectra. The ART-2a algorithm was applied to particle clustering and then similar clusters 232 were combined. The majority of particles were eventually classified into 9 types including 233 Biomass, Fresh EC, Aged EC, Dust/Ash, HMOC, Amine-rich, Ammonium/OC, Cooking and 234 Sea salt. In Figure 2 we present the average spectra of each type and their numbers detected in 235 each GF bin. Since the total number of detected particles in GF bins were not equal (Table 1), 236 we also present the particle numbers normalized by the total numbers to indicate their detection 237 probability in each GF bin (Figure. 2). The hygroscopicity of particle types showed different 238 distribution patterns with GF. The hygroscopicity characters of Biomass, EC, Dust/ash, HMOC 239 and Amine-rich types were described previously and their hygroscopicity have shown 240 consistent trend with the previous characterization (Wang et al., 2014).

241 The Biomass particles produced characteristic peaks of -26(CN), -42(CNO), -59(C₂H₃O₂), -

242 $73(C_3H_5O_2)$ and dominant peak at 39K and related peaks at $113(K_2Cl)$ or $213(K_3SO_4)$ (Silva et

243 al., 1999;Zauscher et al., 2013;Pratt and Prather, 2009). Biomass particles displayed low

244 hygroscopicity since majority of them were present in GF \leq 1.2 range, with the peak detection

probability at GF 1.1 (Figure 2 (a)). The hygroscopicity of Biomass particles is consistent with

other HTDMA measurement of biomass particles (Rissler et al., 2006;Laborde et al., 2013).

The hygroscopicity of biomass particles in ambient environment were similarly detected in the HTDMA-SP2, which suggested the center GF of 1.1~1.2 of biomass particles (at 90% RH),

corresponding to the GF of 1.06-1.13 at 85% RH (Laborde et al., 2013).

250 The EC particles were detected by a series of elemental carbon peaks at C_n (n=1, 2, 3...) in the 251 negative and positive spectra (Ault et al., 2010;Spencer et al., 2006;Toner et al., 2008). EC 252 particles distributed broadly from nearly-hydrophobic (NH) mode to more-hygroscopic (MH) 253 mode. However, the mass spectra of hydrophobic and hygroscopic EC particles were different 254 in their mass spectra. As shown in Figure 2 (b-c), The mass spectra of hygroscopic EC particles produced stronger secondary peaks (-62NO3⁻, -97HSO4⁻, 18NH4⁺) than hydrophobic EC 255 particles, consistent with the significant fractions of secondary matters in hygroscopic EC 256 257 (Laborde et al., 2013). According to their hygroscopicity distributions, the general EC type was 258 divided into Fresh EC and aged EC type. The mass spectra of Fresh EC suggested they were 259 freshly emitted without significant secondary coatings (Weingartner et al., 1997;Laborde et al., 260 2013;Herich et al., 2009). Peak intensity trends of 62NO₃⁻, -97HSO₄⁻, 18NH₄⁺ and other related 261 peaks at different GF were summarized for EC particles (Figure S6). Based on the statistics of 262 peak intensities, we found clear increasing trends of secondary peak intensities in GF 0.9-1.2 263 range, but not in all GFs (Laborde et al., 2013;Herich et al., 2008).

Dust/ash type particles produced inorganic peaks of salts and metals (Gaston et al., 2017;Ault 264 265 et al., 2011;Sullivan et al., 2007). Most of dust/ash particles were detected in hygroscopic range 266 (GF>1.3). Most of Dust/ash particles were internally mixed with nitrate. Within the general Dust/ash type there are many sub-clusters according to specific association of metal peaks in 267 268 particle spectra. Some of the clusters showed characteristic hygroscopicity distributions which 269 offered values in the source apportionment of these particles. As an illustration, we presented 270 the mass spectra and hygroscopicity distribution of the Al-Si cluster in Figure 3. The mass 271 spectra of Al-Si particles showed stronger aluminum (27Al⁺) and silicate (-76SiO₃⁻) peaks in 272 their positive and negative spectra, respectively. Particle number distribution of Al-Si particles 273 suggested that they were detected with the highest probability at GF 1.1. In the preliminary 274 study we identified the similar Al-Si particles exclusively in NH mode (Wang et al., 2014). 275 Based on their hygroscopicity distribution, we assumed the Al-Si particles are soil dusts

according to their reported low hygroscopicity (Koehler et al., 2009).

The spectra of HMOC particles showed obvious organic peaks in higher m/z range (>150). Some HMOC particles produce obvious polycyclic aromatic hydrocarbons (PAH) peaks in positive spectra and high mass signals in negative spectra. The mass spectra of HMOC suggested they were generated from combustion including traffic emissions (Dall'Osto et al., 2013;Toner et al., 2008). The majority of HMOC particles displayed low hygroscopicity (GF<1.2) (Wang et al., 2014;Herich et al., 2008).

283 Amine-rich type particles produced amine peaks at $+59(C_3H_9N)$, $+86(C_5H_{12}N)$ and 284 +101(C₆H₁₅N) (Angelino et al., 2001;Pratt et al., 2009). Particulate amine formation was 285 favored in low temperatures and higher humidity conditions (Huang et al., 2012;Zhang et al., 286 2012). The elevated amine particle fractions may be related to the low temperature (6 °C) and 287 high humidity (78% RH) condition during this experiment. Both the preliminary and present 288 study identified the hydrophilicity of Amine-rich particles, with the highest number 289 contributions to GF>1.5 range (Wang et al., 2014). Short alkyl chain aliphatic amines are known to have relatively high vapor pressures and basic in nature, their presence in particles indicates 290 291 they are most likely occur in the form of aminium salts, whose formation is greatly favored in the presence of particulate water (Angelino et al., 2001;Chen et al., 2019). Mass spectra of 292 Amine-rich particles suggest that 77% of them were internally mixed with sulfate or nitrate. 293

294 With the expanded GF range and sampling durations, we identified other particle types of 295 specific hygroscopic patterns, including Ammonium/OC, Cooking and Sea salt particles in this study. The Ammonium/OC particles demonstrated some similarities with biomass particles. 296 There was predominant potassium peak ³⁹K and many organic peaks in the positive mass 297 298 spectra, as shown in Figure 2. Additionally, stronger 18NH₄⁺ peaks for ammonium and sulfate 299 (-97HSO₄) were also present in these particles. The typical -26CN⁻ and -42CNO⁻ peaks 300 observed for biomass particles were absent or very weak, suggesting the composition differences between Ammonium/OC and biomass particles (Silva et al., 1999;Zauscher et al., 301 302 2013;Pratt and Prather, 2009). The hygroscopicity of the Ammonium/OC particles was unique 303 since they have the largest contributions to moderate GF range (GF 1.1-1.3), with the maximum 304 contribution found at GF=1.2. The GF of Ammonium/OC particles suggests that they can be

305 categorized as LH mode (Swietlicki et al., 2008). A prior ATOFMS study identified that 306 Ammonium/OC particles were from agricultural sources, and found most of them were present 307 in higher photochemical oxidation periods (Qin et al., 2012), consistent with the prominent 308 secondary peaks of ammonium found in this study. It is likely the organics in this type is secondary since the GF 1.2 is close to the hygroscopicity of SOA (GF=1.24 at 90% RH) (Gysel 309 et al., 2007;Sjogren et al., 2008). We inferred that ammonium was not contributing major 310 fractions to Ammonium/OC particles, since ammonium salts was very hydrophilic while 311 312 Ammonium/OC demonstrate only moderate hygroscopicity.

313 Cooking is an important source of primary organic aerosol (POA) in urban regions (Crippa et al., 2013; Dall'Osto and Harrison, 2012). Zhang et al. estimated that up to 35% of POA are 314 315 attributed to cooking aerosol during meal hours (Zhang et al., 2007). Cooking particles around the site was likely to be significant considering that the Fudan campus is located in a heavily 316 317 populated area. The ATOFMS characterization of cooking particles have been performed 318 previously (Dall'Osto et al., 2013). The marker peaks at -255(C₁₆H₃₂O₂, palmitic) and -281 319 (C₁₇H₃₄O₂, oleic acid) in the negative spectra were used to identify Cooking particles (Dall'Osto 320 and Harrison, 2012; Silva, 2000). As shown in Figure 2(h), Cooking particles demonstrated very 321 low hygroscopicity and were detected exclusively in GF<1.1 range. It is noted that GF 0.9 does 322 not necessarily indicate a particle shrinkage in 85% RH. Cooking particles might become more 323 spherical in elevated RH, resulting in smaller mobility diameters. This phenomenon was 324 observed for other organic particles (Shi et al., 2012;Pratt and Prather, 2009). The low 325 hygroscopicity of cooking particles is consistent with the enriched organic as indicated by the 326 fatty acids (-171, -255, -279, -281) and HOA (+55, +57) peaks in the spectra. The detection of 327 cooking particles in NH mode complemented to the conclusion that combustion processes are 328 mainly responsible for NH particles (Swietlicki et al., 2008;Laborde et al., 2013;Herich et al., 2009). 329

With particles of higher GF being analyzed in this experiment, we also identified the Sea salt particle which constitutes an important particle type in ambient air in coastal areas (Herich et al., 2009;Gard et al., 1998). Sea salt mass spectra contain dominant sodium peak 23Na⁺ and other sodium cluster peaks at $62Na_2O^+$, $63Na_2OH^+$, $81Na_2Cl^+$ (Gaston et al., 2017). The 334 hydrophilicity of Sea salt is clear that they were mostly detected in the largest GF bins (>1.5), 335 with their number fractions increased from GF 1.5 to 1.7 (Figure. 2). HTDMA studies in marine 336 environment shown that sea salt particles constitute a separated hygroscopicity mode of the 337 largest GF (Swietlicki et al., 2008), which is generally consistent with the observed GF range in the experiment. However, the observed sea salt particle hygroscopicity is somewhat different 338 339 from the HTDMA-ATOFMS characterization in a subarctic region, where sea salt particles were found mainly detected in GF 1.3-1.5 range at 82% RH (Herich et al., 2009). We inferred 340 341 that sea salt hygroscopicity properties are variant with locations and other factors (organics in 342 seawater, marine microbiological conditions, aging) should be considered (Facchini et al., 343 2008;Randles et al., 2004).

344 3.1.2 Peak intensity variations with GF

Apart from particle number distributions, the HTDMA-ATOFMS dataset provided another aspect of information regarding peak intensities with GF. In this study, we used relative peak intensities (peak areas normalized by the total areas in spectrum) to investigate its relation to GF. Generally, the responses of peak intensity to GF variation were found to be nonlinear, since they were correlated only within specific GF ranges. A simple trend applicable to whole GF range was not observed.

351 We presented the statistics of peak intensity of nitrate (46NO₂⁻, 62NO₃⁻, 125H(NO₃)₂⁻) and 352 sulfate $(80SO_3^-, 97HSO_4^-)$ which were known to be critical to particle hygroscopicity (Figure 353 4). As previously observed, the nitrate and sulfate peaks were present in the majority of particles 354 in all GF bins (Herich et al., 2009;Herich et al., 2008;Wang et al., 2014). However, peak 355 intensities of nitrate and sulfate were indeed stronger in hygroscopic particles than hydrophobic 356 particles. In Figure 4 we observed positive correlation between nitrate and sulfate intensities 357 and GF in GF <1.2 range, suggesting contribution of nitrate and sulfate to particle 358 hygroscopicity in low GF range (Figure 4). However, in higher GF range (GF 1.3-1.5), nitrate 359 and sulfate peaks seem to reach a plateau with unclear dependence on GF. Nitrate and sulfate 360 were known to contribute large fractions of particle mass in MH particles (Swietlicki et al., 361 2008;Laborde et al., 2013;Liu et al., 2014). The unclear trend of nitrate and sulfate with GF 362 seem to suggest that nitrate and sulfate were in stable ratios since nitrate and sulfate peaks were

363 dominating peak areas in negative spectra. For particles of even higher GF, differences were 364 observed between GF 1.3-1.5 and GF 1.5-1.7 range in that stronger nitrate and weaker sulfate 365 peaks were detected in the GF 1.3-1.5 range. Particles classification suggests that this general characteristic is also at variance for different particle types. The same statistics for EC and 366 Dust/ash particles were presented in Figure 4. Compared with EC particles, smaller sulfate and 367 368 stronger nitrate peaks were found in Dust/ash spectra, and the observed trend in total particles 369 were less obvious in Dust/ash. These facts highlight the nonlinearity between peak intensities 370 and GF and that particle types should also be considered in describing peak intensities.

371 The analysis of peak intensities with GF can disclose some atmospheric processes happened on 372 aerosol. We take the Sea salt as an illustration. Sea salt particles were known to react with 373 atmospheric nitric acid, with NaCl in fresh sea salt be transformed into NaNO₃ in the reacted sea salt (Gard et al., 1998). This composition transformation is indicated in corresponding 374 375 changes of NaCl and NaNO₃ peak intensities in particle spectra. The unreacted sea salt particles 376 tend to produce larger peaks of Na₂Cl⁺ and NaCl₂⁻ in spectra (Gaston et al., 2011;Prather et al., 377 2013). In particle spectra of reacted sea salt, the NaCl peaks (Na₂Cl⁺, NaCl₂⁻) decrease while 378 NaNO₃ peaks $(Na_2NO_3^+, Na(NO_3)_2^-)$ increase. We presented peak intensities of sea salt in GF 379 1.5-1.7 range where sea salt particles were detected with largest numbers (Figure 5). We found 380 that the positive correlation between NaCl peak intensity and GF, and the negative correlation 381 for NaNO₃ peaks. Therefore, the HTDMA-ATOFMS data supported that reacted sea salt have 382 reduced hygroscopicity (Herich et al., 2009;Gaston et al., 2018). Laboratory HTDMA study 383 suggested that NaCl and NaNO₃ have deliquesced at 85% RH and that the NaNO₃ (GF ~1.8) is 384 less hygroscopic than NaCl (GF ~2.2) (Hu et al., 2010). The reduced hygroscopicity of sea salt 385 is in line with the GF of sodium salts. However, the sea salt hygroscopicity (GF 1.5-1.7) was 386 smaller than pure NaNO₃ salt (supposing fully reacted), suggested that the chemical 387 transformation alone is not sufficient to account for the observed hygroscopicity of sea salt. We hypothesize that other compositions as organics were mixed into sea salt and contributed to the 388 reduction of sea salt hygroscopicity (Gaston et al., 2011;Randles et al., 2004;Facchini et al., 389 2008). 390

391 3.2 Predictability of hygroscopicity from particle mass spectra

The GF of a particle can be estimated based on HTDMA-ATOFMS data for two reasons. First, different particle types had distinct GF distributions. Second, particles in different GF bins had different mass spectra. The GF estimation from particle spectra requires that the HTDMA-ATOFMS data is capable to represent the major particle types normally presented in atmosphere, which is evidenced in the preceding discussions. In another aspect, the GF prediction from mass spectra also demands that HTDMA-ATOFMS data are sensitive to reflect the composition differences with GF variations.

399 To test the sensitivity of HTDMA-ATOFMS data, we evaluated the average spectral similarities 400 between each pair of GF groups. The average similarities were calculated from the similarities between every possible pairs of particles from the two GF groups. The self-comparing of 401 402 particles within the same GF group were excluded. As shown in Figure S7, we observed a 403 general trend that particles in the same GF bins tend to produce the highest similarities. As the 404 GF differences increase, the mass spectra similarity between two GF bins tended to decrease. 405 This result is an evidence that the particles with different GFs are more likely to have 406 discriminable mass spectra, which suggests that the HTDMA-ATOFMS dataset are capable to 407 estimate hygroscopicity just from particle mass spectra.

408 **3.3 Estimated hygroscopicity of ambient particles**

409 A case study of the hygroscopicity estimation were carried out based on a period of ambient 410 ATOFMS measurement. The ATOFMS data was collected at the same Fudan site from Sep-12 to Sep-28, 2012. During this period the ATOFMS recorded 538,983 mass spectra of individual 411 412 particles. With the described estimation method, the GF value (corresponding to 85% RH) was 413 generated for each particle based on individual particle mass spectra. A fraction of the estimated 414 GF (4%) were excluded from analysis since their maximum similarities failed to exceed the threshold value (dot product > 0.7) between ATOFMS and HTDMA-ATOFMS particles. 415 Particle mixing states in this period were analyzed by clustering particles using ART-2a 416 algorithm (Song et al., 1999). After merging the clusters of similar composition and temporal 417 trends, the majority of particles were finally grouped into the same general types as discussed 418 in HTDMA-ATOFMS dataset (Fresh EC, Aged EC, Dust/Ash, HMOC, Amine-rich, 419

420 Ammonium/OC, Cooking and Sea salt), which account for 90.8% of the total analyzed particles421 in this period.

422 **3.3.1 Hygroscopicity modes and contributions of particle types**

The estimation method determined that the GF of ATOFMS particles were restricted within the 423 424 GF range in HTDAM-ATOFMS dataset (0.9-1.7). Within this GF range, the ATOFMS particle 425 GF distribution suggested several hygroscopicity modes similar to the HTDMA measurement. 426 As shown in Figure 6, three hygroscopicity modes were clear in the GF-number distributions, 427 with particle GF centered at about 1.05, 1.42, and 1.6, respectively (85% RH). Prior HTDMA 428 studies suggested the regular presence of the nearly-hydrophobic mode with center GF in 429 1.05~1.1 range in Shanghai area (Ye et al., 2011), consistent with the GF 1.05 mode in this 430 study. The second mode at GF 1.42 mode in ATOFMS particles corresponds to the MH mode 431 (GF 1.43~1.47) in Shanghai and other sites using HTDMA (Ye et al., 2013;Ye et al., 2011;Liu et al., 2011). The sea salt mode in HTDMA GF distribution is not always clear because of the 432 larger size of sea salt particles. However, the sea salt particles were readily detected by 433 434 ATOFMS because of the detection range of ATOFMS. The particles in GF 1.6 mode contained rich sodium content and their mass spectra suggest typical sea salt peaks (Figure S2). In marine 435 436 areas the sea salt particles were found to constitute hygroscopicity mode of the largest GF (about 2.0 at 90% RH, corresponding to 1.76 at 85% RH) (Swietlicki et al., 2008). 437

The ATOFMS measured particle aerodynamic diameters simultaneously for individual particles. 438 Together with the estimated GF, we inspected particle number distribution as a bivariate 439 440 function of the estimated GF and diameter (aerodynamic diameter, d_{va}) (Figure 7). The 441 hygroscopicity modes were clearer in the GF- d_{va} diagram, which suggest the increasing trend 442 of particle diameter with increasing GF, a very consistent result with HTDMA studies (Ye et 443 al., 2011;Ye et al., 2013). Healy et.al. applied a different method to estimate particle 444 hygroscopicity from single particle data using ZSR mixing rule (Healy et al., 2014). The particle 445 aerodynamic diameter d_{va} was transformed to equivalent mobility diameter d_m by assuming a 446 particle density in that study. Although the methods are different, the identified hygroscopicity modes were similar between the two studies, except the sea salt mode which was not found by 447 Healy et al. (Healy et al., 2014). 448

Based on the GF of hygroscopicity modes, prior HTDMA studies conventionally classified the 449 450 observed modes into categories as nearly-hydrophobic (NH), less-hygroscopic (LH), more-451 hygroscopic (MH) and sea salt (SS) modes respectively (Swietlicki et al., 2008;Liu et al., 452 2011;Sjogren et al., 2008). However, the chemical nature of these hygroscopicity modes was not clear since the HTDAM technique is based on particle numbers and the particle composition 453 information was not obtainable. With the ATOFMS single particle data, particle composition 454 and hygroscopicity was connected directly. To facilitate comparison, we similarly divide the 455 456 estimated GF into four bins (<1.1, 1.1-1.3, 1.3-1.5 and >1.5) to roughly represent the NH, LH, 457 MH, SS particles according to the conventional classification of hygroscopicity modes (Swietlicki et al., 2008). As shown in Figure 6, particle types were distributed differently in GF 458 modes. For example, the organic particles including HMOC, Biomass and Freshly emitted EC 459 particles were mainly enriched in NH mode, which suggests directly that combustion sources 460 461 are mainly responsible for NH particles in the ambient (Herich et al., 2008;Herich et al., 462 2009;Ye et al., 2011).

463 In Table 2 we made the statistics on average number contributions of particle types to the NH, 464 LH, MH and SS mode. The presented statistics were based on the temporal contributions of 465 each particle types in daily resolution. It is noted that particle number contributions presented 466 in Table 2 may be different from HTDMA-ATOFMS dataset (Figure 2). For example, the 467 Cooking particles contributions to NH mode was significantly lower in the ATOFMS dataset 468 (3%) than HTDMA-ATOFMS dataset (19%). This result is understandable because particle 469 concentrations are variant with particle size and HTDMA-ATOFMS only analyzed a narrow 470 size bin from the total particle size distribution. For each hygroscopicity mode, there were 471 multiple particle types contributing significant number fractions, suggesting that even within 472 the same hygroscopicity mode there were still some heterogeneity in particle composition. 473 Particles in the same hygroscopicity mode may share some common features in compositions but their differences are distinguishable in single particle data. In another respect, the 474 475 contributions of each type also suggest the existence of a predominant type that accounts for major fractions in respective modes compared with other types, such as Aged EC in MH mode, 476 Ammonium/OC in LH mode. The comparison between Table 2 and Figure 2 suggests that, 477

478 although their absolute contributions may be different, the hygroscopicity patterns of particle 479 types in the two datasets are in good agreement. Based on this fact, we concluded that the 480 composition-hygroscopicity connections contained in HTDMA-ATOFMS dataset was 481 successfully reflected into the predicted GF.

482

3.3.2. Temporal variations of estimated hygroscopicity

483 The temporal variation of particle estimated GF from Sep-12 to Sep-28 was illustrated in Figure 484 8. Four distinct periods (P1-P4) were identified based on their different hygroscopicity 485 distributions. Generally, the P1 and P3 periods were characterized by elevated MH mode which 486 dominated the ATOFMS particles numbers, while in P2 and P4 the MH particles decreased 487 significantly and sea salt mode was pronounced. Back trajectories during P1-P4 were analyzed 488 using HYSPLIT mode (Draxler, R. R. and Rolph, G. D., 2003) to inspect the airmass that 489 influenced the sampling site (Figure S9). The 24-hour back trajectories suggests that the airmass in P1 period mainly circulated in local regions from northwest direction to Shanghai. The local 490 491 circulations brought regional aerosol pollution to the sampling site, resulted in elevated 492 concentrations of particles, especially the MH particles. During P2, the airmass originated from 493 the ocean in northeast direction with less continental influence. The cleaner air from the ocean almost wiped out the accumulated particles observed in P1 and the concentrations of sea salt 494 particles increased. In the majority of time during P3, the airmass stayed over continental areas. 495 496 The MH particles dominated particle numbers in this period and the sea salt mode were barely present. During Sep-18 to Sep-20 in P3, the LH particles showed increased concentrations and 497 gradually decreased after Sep-20. Similar to P1, the origin of airmass in P4 shifted to the ocean 498 499 in eastern directions and SS mode emerged again. Both the particle spectra and the back 500 trajectories supported that the GF mode of 1.6 can be mainly attributed to sea salt particles.

Except meteorological conditions, other differences exist in the temporal trends for each hygroscopicity modes. Generally, the NH mode showed relatively stable trends irrespective of different periods of P1-P4, as indicated in Figure 8. Closer inspection of NH particles suggests a notable feature that obvious sharp spikes were present in NH particle temporal concentrations (Figure 8). This character is typical for particles from local emission sources, with undissipated plumes at the time of detection. With the combined information from particle composition in Table 2, we inferred that the NH particles were fresh emitted particles from local sources with high organic or elemental carbon content (Laborde et al., 2013;Herich et al., 2008;Weingartner et al., 1997). The mass spectra of NH particles indicated low nitrate and sulfate signals, suggesting that secondary matters have not accumulated significantly on these particles, consistent with negligible coating thickness on NH particles (Laborde et al., 2013). We tend to ascribe the organics in NH particles to be primary organic carbons (POA) considering their relatively fresh emission state (Sjogren et al., 2008;Liu et al., 2011;Gysel et al., 2007).

514 Some characters of LH particles were noticed. Similar to the particles in MH range, the LH 515 particles mainly presented in continent influenced periods (P1 and P3) (Figure 8). However, the 516 temporal concentrations trends suggested differences between LH and MH particles. For 517 example, MH mode dominated particle numbers in the entire P3, while the LH particles were only pronounced from Sep-17 to Sep-21, with peak concentrations observed on Sep-19. The 518 519 particle contributions showed that Ammonium/OC is the main contributor to NH particles 520 (Table 2). This conclusion is also correct in temporal basis since the Ammonium/OC 521 contributions to LH range were always much larger than other types of particles in all the 522 studied period. We investigated the nature of the Ammonium/OC particles by comparing its 523 concentration with ambient pollutants levels. As shown in Figure 9, the number fractions of 524 Ammonium/OC particles showed strong connections to ambient O₃ concentrations. During 525 Sep-17 to Sep-21 there were daily oscillations of O₃ levels, which were followed by the same 526 pattern of Ammonium/OC particles with lags of several hours. The maximum O₃ concentrations 527 was found on Sep-19 (219 µgm⁻³) in the period, the same day when the highest Ammonium/OC 528 particle contribution was observed. HTDMA studies suggested that LH mode became 529 pronounced in new particle formation (NPF) periods with high atmospheric reactivity 530 (Swietlicki et al., 2008). Based on these facts, we think the Ammonium/OC particles were 531 related to the oxidation processes of organics vapors by oxidants such as O₃ (Varutbangkul et 532 al., 2006). The moderate hygroscopicity of Ammonium/OC agrees with the hygroscopicity of ambient secondary organic aerosol (SOA) (Gysel et al., 2007;Sjogren et al., 2008). During 533 534 periods of higher Ammonium/OC contributions (Sep-17 to Sep-21), increased SO₂ levels were also encountered, which coincided with high sulfate signals in mass spectra of Ammonium/OC 535

(Figure 2). Compared with the organic compositions, the sulfate was inferred to have minorcontributions to mass fractions because of the moderate hygroscopicity of Ammonium/OC.

538 Particles in MH mode dominated the particle numbers for the majority of time in P1 and P3 539 (60%) compared with the average fraction of 25% in P2 and P4. Inspection of the temporal 540 trends of MH particles also suggest some diurnal variations with higher concentrations in 541 nighttime (Figure 8). As illustrated in Figure 4 and Figure S6, mass spectra of MH particles 542 were dominated by sulfate, nitrate peaks, suggested that MH particles were mixed with 543 significant fraction of secondary inorganic matters (SIA). The coating thickness of the 544 secondary matters was determined by HTDAM-SP2 at different GF (Laborde et al., 2013). The coating thickness of secondary coating was measured to 40-80 nm ($D_{dry}=265 \text{ nm}$) in GF 1.2-545 1.7 range, being equivalent to 55-76% of hygroscopic particle volume was attributed to 546 secondary matters (Laborde et al., 2013;Healy et al., 2014). 547

548 3.3.3 Peak intensity variations with estimated GF

549 Particle hygroscopicity and peak intensities in particle mass spectra were correlated to show 550 their connections. The correlation was illustrated similarly to the analysis of the HTDMA-ATOFMS dataset, as shown in Figure 10. In addition to the statistics on peak intensities of 551 552 different GF, the number distributions ATOFMS particles with GF and peak intensities were 553 presented for nitrate $(46NO_2^-, 62NO_3^-, 125H(NO_3)_2^-)$ and sulfate peaks $(80SO_3^-, 97HSO_4^-)$ in 554 the lower panels in Figure 10. The general trends of peak intensities with GF in HTDMA-ATOFMS dataset was preserved in estimated GF of ambient particles. The trends of nitrate and 555 556 sulfate peak intensities showed increases from NH to LH range and remained constant in the 557 MH mode. Similar to HTDMA-ATOFMS particles, stronger nitrate peaks were detected in SS 558 particles compared with the MH particles, while an opposite trend was observed for sulfate 559 intensities. These results highlight the nonlinearity between GF and peak intensities of 560 ATOFMS particles.

The nonlinearity of peak intensities with GF was also suggested by the different particle types presented in ATOFMS data. As shown in the lower panels in Figure 10, the distribution of ATOFMS individual particles showed enrichment in different areas in the GF-peak intensity

564 diagram, suggesting the presence of particle groups of different compositions. To illustrate this 565 character, we selected two areas with clear particle enrichments in GF-peak intensity diagram 566 and their particle composition were analyzed (denoted as A and B in lower left panel in Figure 567 10). Obviously, particles in area A produced much larger nitrate signals than particles in area B. Particle numbers in the two areas suggested that Dust/ash dominated particles in area A (59%) 568 while in area B the Dust/ash only accounted for 14% of particles (Figure S10). As a contrast, 569 particles in area B were dominated by Aged EC type (53%) followed by 25% Dust/ash. Table 570 571 2 suggests that Aged EC and Dust/ash are the major types presented in the same MH mode. However, peak intensity responses to GF were indeed different for different particle types, 572 suggesting the importance of particle types in describing peak intensities. 573

574 The particle distribution with sulfate intensities showed similar enrichment patterns to nitrate (lower right panel in Figure 11). Sulfate peak intensities were found to correlated with 575 576 hygroscopicity in GF <1.2 range but in MH range no correlated with GF was observed. We note 577 that except the larger peaks of nitrate and sulfate, some smaller peaks were also found to 578 correlated with GF within specific particle type. We correlated peak intensities of Na₂Cl⁺ and 579 NO_3^- with the estimated GF of sea salt particles (Figure S11). The Na₂Cl⁺ peaks were positively 580 correlated with GF while the nitrate peaks were negatively correlated with GF. The observed 581 correlation in sea salt particles are consistent with discussed trends in the HTDMA-ATOFMS 582 dataset. These results demonstrate that the GF estimation method have effectively reflected the 583 minor changes in particle mass spectra into the estimated hygroscopicity.

584 **3.3.4** Comparing the estimated hygroscopicity with visibility

585 Particle optical properties were closely connected to hygroscopicity (Liu et al., 2012;Qu et al., 586 2015; Chen et al., 2012). The hygroscopic growth increases particle volumes and cross sections 587 and is contributing to the visibility degradation. With the estimated hygroscopicity of ATOFMS particles, we correlated atmospheric visibility with particle concentrations to study their 588 contributions to the visibility variations. The ATOFMS particle volume concentrations were 589 590 calculated for hygroscopicity modes of NH, LH, MH and SS based on ATOFMS particle 591 diameter and numbers. The particle volume concentrations was used because hygroscopic 592 growth change particle sizes rather than numbers (Chen et al., 2012). The visibility data was

593 obtained from (https://www.wunderground.com/) logged in the Hongqiao airport (31°12' N, 594 121°20' E) and Pudong airport (31°9.3' N, 121°49' E) during the study period (see the map in 595 Figure S12). The temporal variations of visibility in two sites correlated strongly (Figure S12), 596 despite the 45 kilometers distance between two airports. The Fudan site is located roughly between the two airports, and the two sets of visibility data were averaged to represent the study 597 598 site. In P2 and P4 the site was under the influences from ocean, resulting visibilities larger than 10 km (Figure 8). Apart from ATOFMS particles, contemporary PM_{2.5} volume concentrations 599 600 were also correlated with visibility. The PM2.5 volume concentrations were derived from PM2.5 mass concentrations using particle density (1.4 gcm⁻³). A strong correlation between ATOFMS 601 602 particle numbers and PM_{2.5} was found ($R^2=0.80$).

603 An exponential relation between visibility and PM concentrations was found by the previous study (Qu et al., 2015). After applying the exponential fitting to the visibility and particle 604 volume concentrations, we found a moderate correlation for ATOFMS particles ($R^2=0.45$) and 605 606 better correlations for PM_{2.5} concentrations (R^2 =0.64) (Figure S13). However, the fitting errors 607 were clearly dependent on ambient RH, with larger errors in higher humidity, indicating that 608 hygroscopicity might affect visibility degradations, which were consistent with other studies 609 (Chen et al., 2012;Liu et al., 2012). To further examine the effect of particle hygroscopicity on visibility, we derived particle volumes in different RH using estimated κ values (Petters and 610 611 Kreidenweis, 2007). The κ values were calculated using the GF of individual particles at 85% 612 RH for ATOFMS data and the average GF of 1.36 for PM_{2.5} volume concentrations. With hygroscopicity being considered, we found notable improvements of the correlations between 613 614 PM concentrations and visibility, with the improved correlation observed for PM_{2.5} 615 concentrations ($R^2=0.82$) after applying correction for hygroscopicity (Figure 11). However, 616 this improvement was barely the case for NH particles, probably due to the negligible 617 hygroscopic growth. For the ATOFMS particles in different GF modes, we found the highest R^2 (0.65) for the MH particles. The correlation between SS particles and visibility was distorted 618 619 due to the visibility reached its limit (10 km) when sea salt mode was pronounced (Figure 8). The R^2 between MH particles and visibility suggests that the variation of MH particles 620 accounted for the major part of visibility changes (65%) during this period, which coincided 621

with the major contribution of nitrate and sulfate to light extinction (61%) in eastern China
areas (Qu et al., 2015). These results indicate the importance of discriminating particles by
hygroscopicity in explaining the measured visibility.

625 <u>4. CONCLUSIONS</u>

626 The hygroscopicity and composition of submicron particles were characterized in a megacity 627 in eastern China. A single particle mass spectrometer was connected to the downflow of an 628 HTDMA to analyze particle composition of different hygroscopicity at 85% RH. Direct 629 connections between hygroscopicity and single particle signatures were established. The 630 HTDMA-ATOFMS dataset suggested that particle types were distributed differently in various hygroscopicity ranges. Generally, Fresh EC particles and organic particle types including 631 632 Biomass, Cooking and high molecular organic carbon (HMOC) were enriched in nearly-633 hydrophobic (NH) mode (GF<1.1). The majority of particles in NH mode suggested characters of freshly emitted particles from combustion sources. Particle types in the more-hygroscopic 634 635 (MH) range (GF1.3-1.5) include Aged EC, Amine-rich and Dust/ash particles. The mass spectra 636 of MH particles suggested that they were mixed with significant fractions of secondary matters. 637 The sea salt particles were detected in the higher GF range (GF 1.5-1.7), with increasing detection probability at larger GF. In the moderate hygroscopicity range (GF 1.1-1.3), the 638 Ammonium/OC particles was identified with peak detection probability at GF 1.2. 639

640 Single particle spectra peak intensities were correlated with GF to disclose their possible 641 connections. The peak intensities were nonlinearly correlated with GF. The peak intensities of nitrate and sulfate showed gradual increase from GF 0.9 to 1.2 while in larger GF range the 642 643 same trends were not observed. Peak intensities showed opposite directions for nitrate and 644 sulfate in GF 1.5-1.7 range with stronger nitrate peaks and smaller sulfate peaks. Except the 645 general nonlinearity of peak intensities is subjected to variations with different particle types. The peak intensity analysis suggested the reduced hygroscopicity of sea salt after atmospheric 646 647 aging.

Based on the established connections between composition and composition, we developed a
 method to estimate particle hygroscopicity from single particle mass spectra. The method was

650 tested in a period of ATOFMS data in Shanghai. The estimated GF of individual particles 651 suggested that ATOFMS particles were present in similar hygroscopicity modes as HTDMA measurement on ambient aerosol. We discriminated particles into NH, LH, MH and SS mode 652 653 by the GF of individual particles. Particle type contributions to the these hygroscopicity modes suggested consistent results with HTDMA-ATOFMS dataset. Based on the combined 654 information on particle composition, hygroscopicity and airmass back trajectories, we inferred 655 that the POA/EC, SOA, SIA and salts are the characteristic compositions for particles in the 656 657 NH, LH, MH, SS modes, respectively. The proposed method is a novel way of single particle 658 mass spectrometry data analysis, which would provide additional information to the study of 659 particle mixing states, source identification and visibility degradation.

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883 *TABLES*

Table 1. Statistics of the DRH, GF, sampling duration and the number of chemically analyzed particles by ATOFMS ($D_{dry} = 250$ nm, RH = 85%).

D _{RH} (nm)	225	250	275	300	325	350	375	400	425
Growth Factor	0.9	1.0	1.1	1.2	1.3	1.4	1.5	1.6	1.7
Duration (hours)	42	67	11	20	8	11	34	20	11
Number of particle	742	1665	709	1401	2330	4469	6399	723	262
spectra									

886

Table 2. Statistics on particle number contributions of ATOFMS particle types to different GF
modes. The statistics are the average contributions and variation ranges (in brackets) based on
temporal data in daily resolution.

Contribution (0/)	NH	LH	MH	SS	
Contribution (%)	(GF <1.1)	(GF 1.1-1.3)	(GF 1.3-1.5)	(GF >1.5)	
Fresh EC	14 (7-17)	2 (1-4)	0 (0-3)	0 (0-1)	
Cooking	3 (0-7)	1 (0-3)	0 (0-0)	0 (0-0)	
Biomass	18 (7-35)	9 (2-17)	0 (0-1)	0 (0-0)	
НМОС	40 (30-68)	8 (2-15)	0 (0-1)	0 (0-0)	
Ammonium/OC	11 (3-21)	32 (20-45)	2 (1-5)	0 (0-0)	
Aged EC	2 (1-5)	12 (5-20)	47 (15-72)	13 (1-42)	
Dust/ash	3 (1-6)	13 (9-22)	27 (13-53)	26 (4-44)	
Amine-rich	3 (1-5)	12 (2-41)	13 (4-39)	11 (1-44)	
Sea salt	0 (0-1)	0 (0-1)	1 (0-4)	40 (7-78)	

891 *FIGURES*



892

893 Figure 1. The schematic of HTDMA-ATOFMS characterization setup showing the major parts

894 of HTDMA (left), ATOFMS (right). The humidifier in HTDMA was maintained at 85% RH.



Figure 2. (Left panel) Averaged particle mass spectra of the major particle types detected in
HTDMA-ATOFMS characterization. Peaks of significance were labelled. Right panels show
the particle numbers of each type (left-axis) and their relative number fractions in total particles
(right-axis) as a function of GF.





901 Figure 3. Average mass spectra and hygroscopicity distribution of Al-Si particles.



Figure 4. Statistics of nitrate and sulfate peak intensities (minimum, 25th percentile, median,
75th percentile, maximum) with GF in HTDMA-ATOFMS experiment. The intensity statistics



Figure 5. Statistics of peak intensities in sea salt particle mass spectra in GF 1.5-1.7 range. The
statistics include minimum, 25th percentile, median, 75th percentile, maximum and outliers for
each GF bin.

910



912 Figure 6. Particle number distribution of different particle types as a function of estimated GF

⁹¹³ during Sep-12 to Sep-28, 2012.





Figure 7. The distribution of ATOFMS particles as a bivariate function of estimated GF andaerodynamic diameter.



918

919 Figure 8. Temporal variations of atmospheric visibility and number concentrations of nearly-920 hydrophobic (estimated GF < 1.1) and hygroscopic (GF > 1.1) particles (Upper). The lower 921 panel show the temporal variation of estimated GF from Sep-12 to Sep-28, 2012. In the lower-922 right panel is the particle number distribution with GF during the period.



Figure 9. Temporal traces of Ammonium/OC relative number concentrations, ambient O₃ and
SO₂ concentrations during Sep-12 to Sep-28, 2012



928

929 Figure 10. Statistics (minimum, 25th percentile, median, 75th percentile, maximum) on the nitrate
930 and sulfate peak intensities of ambient particles at different GFs (upper). The lower panels show
931 particle distributions as a bivariate function of estimated GF and peak intensities of nitrate and
932 sulfate.



934 Figure 11. The R-squares between visibility and PM volume concentrations with and without

935 considering particle hygroscopicity