



# 1 Increased inorganic aerosol fraction contributes to air pollution and

## 2 haze in China

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## 22 Abstract

23 The detailed formation mechanism of increased number of haze events in China is 24 still not very clear. Here, we found that reduced surface visibility and an increase in 25 satellite derived columnar concentration of inorganic precursor concentrations are 26 connected with each other. Typically higher inorganic mass fractions lead to increased 27 aerosol water uptake and light scattering ability in elevated relative humidity. Satellite 28 observation of aerosol precursors of NO2 and SO2 showed increased concentrations 29 during study period. Our in-situ measurement of aerosol chemical composition in 30 Beijing also confirmed increased contribution of inorganic aerosol fraction as a function 31 of increased particle pollution level. Our investigations demonstrate that the increased 32 inorganic fraction in the aerosol particles is a key component in the frequently occurring 33 haze days during studying period, and particularly the reduction of nitrate, sulfate and 34 their precursor gases would contribute towards better air quality in China.

## 35 Introduction

36 As one of the most heavily polluted regions in the world, China has suffered from 37 air pollution for decades (Hao et al., 2007; Zhang et al., 2015). Aerosol particles, as 38 major air pollutant, have significant effects on human health (Lelieveld et al., 2015). 39 The general public and the central government in China have realized the severe 40 situation and have taken some measures to improve the air quality nationwide in the 41 recent years. For example, the state council published a plan for air pollution control, 42 in September of 2013, aim to reduce  $PM_{2.5}$  concentrations by 10%~25% in different 43 regions of China. The successful implementation requires a sufficient knowledge of





haze formation mechanism (Kulmala, 2015) and comprehensive observation network 44 45 (Kulmala, 2018). Our understanding on haze events with high  $PM_{2.5}$  concentrations in 46 China is still limited due to the spatial-temporal variation of aerosol properties and 47 limited observation information (Wang et al., 2016). Recent studies found that 48 secondary aerosol components are important during the intense haze events in Beijing, 49 Xi'an, Chengdu and Guangzhou during January of 2013, and the reduction of aerosol 50 precursors is a key step to reduce particle pollution (Guo et al., 2014; Huang et al., 51 2014). The analysis of longer time series data from Nanjing shows that secondary 52 particles are typically dominating even the number concentrations in polluted 53 conditions (Kulmala M., 2016). The most abundant mass fractions of atmospheric 54 aerosol are inorganic and organic components, which have large spatio-temporal 55 variation (Jimenez et al., 2009). Identifying the most abundant as well as critical aerosol 56 species that contribute to the haze formation in a longtime perspective is important to 57 draw up effective plans for the air pollution control.

Here, we used comprehensive data sets to reveal that an increasing trend of inorganic components in atmospheric aerosol may be a pivotal factor, at least, which leads to frequently occurred haze events in China from 1980-2010. As a result, the controlling of inorganic aerosol components of nitrate, sulfate and their precursors should be of a high priority.

## 63 2.Methodology

- 64 The daily averaged visibility and relative humidity data in 262 sites of China are
- 65 obtained from the Integrated Surface Dataset (ISD) from National Oceanic and
- 66 Atmospheric Administration National Climate Data Center of the USA
- 67 (https://www.ncdc.noaa.gov/isd). The visibility observations were made three times a





- 68 day at 8-hour intervals begins at 00:00 by well trained technicians. They measured
- 69 visual range using distinctive markers, such as tall buildings, mountains and towers,
- 70 to which the distance from the meteorological monitoring stations are known. First,
- 71 we quantified the importance of relative humidity to visibility as the hygroscopic
- 72 inorganic compounds typically grow in size in high humidity(Swietlicki et al., 2008).
- 73 In this study, for a given site and given year, we used the observed annual visibility
- (V) as a ratio  $(R_i)$  between visibility values from the surface observation stations,
- vhen the daily average RH was below 40% for more than 20 days. In the
- 76 corresponding high-humidity cases daily RH was between 80%~90% for more than
- 77 20 days each year at a given observation site:
- 78

$$R_i = \frac{V_{dry}}{V_{wet}}.$$

We use this ratio to infer long trend of aerosol hygroscopicity information. In
addition, we calculate anomaly (A) from the ratio for a given year *i* as a difference from
the 30-year (R<sub>30y</sub>) from 1980 to 2010:

$$A = R_i - R_{30\gamma}.$$

Our spatial focus is placed on North China Plain, Yangtze River Plain and Sichuan Basin due to frequent haze events (Zhang et al., 2012). The stations in Pearl River delta region and other Southern China stations were not included due to limited days with the daily average RH below 40%.

88 The atmospheric column amount of NO<sub>2</sub> and SO<sub>2</sub> data are obtained from
89 SCIAMACHY (Scanning Imaging Absorption spectrometer for Atmospheric
90 CHartographY) satellite products. Modeled aerosol chemical composition from GEOS





91 (Goddard Earth Observing System) - Chem chemical transport model in China during 92 1998-2012 is used. The model utilizes assimilated meteorology data and regional 93 emission inventories with a horizontal resolution of  $2^{\circ} \times 2^{\circ}$  with 47 vertical levels from 94 surface to 80 km. The detailed information about the model can be found in (Boys et 95 al., 2014). Aerosol chemical composition of organic, sulfate, nitrate, ammonium and 96 chloride were measured with a high-resolution-time of flight-aerosol mass 97 spectrometers (DeCarlo et al., 2006). Detailed information of data analysis, collection 98 efficiencies (CE) and relative ionization efficiencies are presented in Zhang et al. (2014).

## 99 3. Results and discussion

#### 100 3.1 Decreasing trend in visibility in high relative humidity conditions

101 According to the geographical division, our study sites are mainly in North China 102 Plain (NCP), Sichuan Basin (SCB) and Yangtze River Plain (YRP) as showed in Figure 103 1. Looking at the time series as a whole, the visibility in dry (RH <40%) condition did 104 not have a clear trend in Yangze River Plain and Sichuan Bsain region, whereas 105 visibility in high RH (80%~90%) conditions showed a decreasing trend. The average 106 ratio of visibility in low RH to that in high RH from 1980-2010 is presented in Figure 107 1. The maximum ratios were identified in eastern China and in some western Chinese 108 cities. Overall, the enhancement factor due to hygroscopicity is within expected values 109 both from modeling and observation results (van Donkelaar et al., 2010; Xin et al., 110 2015). The higher ratios occurred in more severe air pollution areas, like North China 111 Plain, Sichuan Basin and the city of Urumqi, where the contribution of hygroscopic 112 aerosol is more pronounced in comparison with non-hygroscopic dust particles. Long 113 time trends of this ratio in a specific site can reveal the variation of aerosol inorganic 114 fraction and organic fraction due to their different hygroscopicity and water uptake





115 capacity and associated light extinction ability. That is, the mass fractions and 116 concentrations of sulfate, nitrate and ammonium may have increased over study period as they dominate water uptake ability compared with other components (e.g., organic, 117 118 black carbon, dust and metal elements) in the atmospheric aerosol (Wang et al., 2015). 119 For the selected regions, we have calculated the anomaly as a regional average as shown 120 in Figure 2. The ratio showed increasing trends over three regions of China and the 121 maximum trends occurred in North China Plain with the value of 0.0168 per year, 122 which indicate an increase of hygroscopic aerosol in these regions during the 30-year 123 observation period.

124 To corroborate our results, Yang et al. (2011) showed an increasing fraction of inorganic components by 20% in Beijing from 1998 to 2008, especially in summer, 125 126 while the fractions of hydrophobic components such as organic and black carbon 127 decreased in the aerosol phase. A recent study by Boys et al. (2014) revealed that 128 increasing fraction of secondary inorganic aerosol is dominated in the increased mass 129 concentration of PM<sub>2.5</sub> in China from 1998-2012. Due to the increase in hygroscopic 130 inorganic aerosol composition, the number of haze days have significantly increased in 131 the past three decades over North China Plain (Fu et al., 2014).

#### 132 3.2 Enhanced emissions of inorganic aerosol precursors

The longterm trends of aerosol precursors and their spatial variability can improve our understanding of the trends in aerosol chemical composition. Figure 3 and Figure 4 show atmospheric column trends of NO<sub>2</sub> and SO<sub>2</sub> observed from SCIAMACHY. The column NO<sub>2</sub> level can be a good proxy for vehicle and coal burning emission associated with oil and coal consumption (Richter et al., 2005). The column amount of NO<sub>2</sub> showed pronounced increasing trends in three regions of China, particularly in Northern





139 China with the value of 0.14  $\times 10^{15}$  molecule/cm<sup>2</sup>/year from 2002 to 2011. This is 140 probably associated with the increase in power plant and on-road vehicle emissions 141 (Wu et al., 2012). The average NO<sub>2</sub> concentration in Northern China increased by more 142 than two-fold, while in the Yangze River Plain region experienced a considerable smaller trend in NO<sub>2</sub>, with the value of 9.7  $\times 10^{15}$  molecule/cm<sup>2</sup> in 2010 and 6.4  $\times 10^{15}$ 143 144 molecule/cm<sup>2</sup> in 2002. As a whole, the column NO<sub>2</sub> concentration in China doubled from 2002 to 2010, with the values of  $1.4 \times 10^{15}$  molecule/cm<sup>2</sup> in 2002 and  $2.8 \times 10^{15}$ 145 146 molecule/cm<sup>2</sup> in 2010, respectively.

147 Figure 4 depicts trend in SO<sub>2</sub> concentration in four regions of China from 2004 to 148 2010. The SO<sub>2</sub> concentration showed and increasing trends in North China Plain, 149 Sichuan Basin and Yangze River Plain, but nearly a constant trend in China as a whole. 150 A decreasing trend was observed during the year of 2008 and 2009, especially in 151 Northern China Plain. This may be due to a combination of Chinese economic downturn and emission reduction during the Olympic games (Lin and McElroy, 2011) 152 153 (Wang et al., 2010). Anyway, the aerosol precursor of  $NO_2$  showed the most increasing 154 trend in China, consistent with the trend of increased aerosol concentration (Xing et al., 155 2015) . Figure S1 shows the annual trends of aerosol inorganic fraction in PM2.5 mass 156 concentration from 1998-2012 with GEOS-Chem model in China. The results indicate 157 that North China Plain area suffered the most from heavily pollution, consistent with 158 our surface observations as well as earlier work (Xin et al., 2015). Aerosol concentrations 159 have increased considereably from 1980 to 2010. The modeling results show that concurrently 160 the fraction of inorganic fractio'n has increased more rapidly. Consequently the water uptake 161 of the aerosol have increased leading to reduced visibility, which is consistent with our ground-162 based observations.





#### 163 3.3 Validation of increased inorganic aerosol components with elevated air

#### 164 pollution level with in-situ measurement

165 To validate our hypothesis that the increased inorganic components contribute to 166 visibility degradation, we used four months of aerosol chemical composition and 167 visibility data from urban Beijing from November of 2010 to February of 2011. As 168 shown in Figure 5, we divided the visibility values into four bins, which corresponds to 169 clean time to heavy pollution time and to conditions in between. The inorganic aerosol 170 precursors of SO<sub>2</sub> and NO<sub>x</sub> doubled as the visibility decreased from more than 10 km 171 (clean time) to less than 2 km (heavily polluted time). At the same time, the mass concentration of nitrate, sulfate and ammonium components increased to 13.5 µg m<sup>-3</sup>, 172 15.5  $\mu$ g m<sup>-3</sup> and 10.6  $\mu$ g m<sup>-3</sup>, respectively. Meanwhile, the mass fraction of these 173 174 inorganics increased from 11% to 13%, from 14% to 21% and from 10% to 14%, respectively. At the same time, the mass concentration and fraction of organic 175 components decreased from 12.2  $\mu$ g m<sup>-3</sup> to 33.4  $\mu$ g m<sup>-3</sup> and 60% to 46%, respectively. 176 177 This direct observation shows that the contribution of inorganic components increased 178 during this campaign. It is plausible that the increased concentration of SO<sub>2</sub> and NO<sub>x</sub> 179 are highly associated with this giving rise to the long-term trends observed in Figure 2 180 (Pan et al., 2016; Wang et al., 2014).

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## 182 4. Conclusion and implication for atmospheric air pollution control

Atmospheric pollution and associated haze events has a dramatic effect on climate
change, human health and visibility degradation (Ding et al., 2013; Petäjä et al., 2016;
Wang et al., 2015; Zhang et al., 2015). Here, longterm visibility measurements





186	combined with satellite data sets, in-situ measurements and model results revealed that
187	increased fractions of inorganic aerosol components in the particle matter are crucial in
188	contributing to more haze events. In this way, aerosol hygroscopic growth has
189	occurred in lower relative humidity conditions than before due to more ammonium
190	nitrate aerosol, and the light scattering ability of atmospheric aerosol enhanced as
191	shown in Figure 6. Another mechanism is that high concentration of $NO_x$ can promote
192	the conversion of SO <sub>2</sub> to form sulfate aerosol via aqueous phase oxidation during
193	intensive pollution periods (He et al., 2014; Wang et al., 2016). Considering the vast
194	energy consumption in the future decades and the sources of inorganic components in
195	atmospheric aerosol, we denmonstrate that the reduction nitrate, sulfate, ammonium
196	and their precursors maybe more critical in China.

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## 206 Competing financial interests

207 The authors declare no competing financial interests.





#### 208 Author contributions

- 209 Y.H.W had the original idea. L.L.W and C.S.G provided and processed satellite and
- 210 visibility data. Y.S.W provided measurements of aerosol chemical composition
- 211 data.Y.H.W, Y.S.W, L.L.W, T.P and M.K interpreted the data and write the paper.
- All the authors commented on the paper.

#### 213 References

214 Boys, B.L. et al. Fifteen-Year Global Time Series of Satellite-Derived Fine Particulate

215 Matter. Environmental Science & Technology, 48(19): 11109-11118, 2014.

- 216 DeCarlo, P.F. et al. Field-Deployable, High-Resolution, Time-of-Flight Aerosol Mass
- 217 Spectrometer. Analytical Chemistry, 78(24): 8281-8289, 2006.
- 218 Ding, A.J. et al.. Intense atmospheric pollution modifies weather: a case of mixed biomass
- 219 burning with fossil fuel combustion pollution in eastern China. Atmos. Chem. Phys.,
- **220** 13(20): 10545-10554, 2013.
- 221 Fu, G.Q., Xu, W.Y., Yang, R.F., Li, J.B. and Zhao, C.S. The distribution and trends of fog
- and haze in the North China Plain over the past 30 years. Atmos. Chem. Phys.,

**223** 14(21): 11949-11958, 2014.

224 Guo, S. et al. Elucidating severe urban haze formation in China. Proceedings of the National

Academy of Sciences of the United States of America, 111(49): 17373-8, 2014.

- 226 Hao, J., He, K., Duan, L., Li, J. and Wang, L. Air pollution and its control in China. Frontiers
- of Environmental Science & Engineering in China, 1(2): 129-142, 2007
- 228 He, H. et al.Mineral dust and NOx promote the conversion of SO2 to sulfate in heavy
- 229 pollution days. Sci Rep, 4: 4172, 2014.
- 230 Huang, R.J. et al. High secondary aerosol contribution to particulate pollution during haze
- events in China. Nature, 514(7521): 218-22, 2014.
- 232 Jimenez, J.L. et al. Evolution of Organic Aerosols in the Atmosphere. Science, 326(5959):
- **233** 1525-1529, 2009.





234	Kulmala, M Build a global Earth observatory.Nature, 553, 2018.
235	Kulmala, M China's choking cocktail. Nature, 526, 2015.
236	Kulmala M., L.K., Virkkula A., Petäjä T., Paasonen P., Kerminen VM., Nie W., Qi X., Shen
237	Y., Chi X. & Ding A. On the mode-segregated aerosol particle number concentration
238	load: contributions of primary and secondary particles in Hyytiälä and Nanjing.
239	Boreal Env. Res, 21: 319–331, 2016.
240	Lelieveld, J., Evans, J.S., Fnais, M., Giannadaki, D. and Pozzer, A. The contribution of
241	outdoor air pollution sources to premature mortality on a global scale. Nature,
242	525(7569): 367-71, 2015.
243	Lin, J.T. and McElroy, M.B. Detection from space of a reduction in anthropogenic emissions
244	of nitrogen oxides during the Chinese economic downturn. Atmos. Chem. Phys.,
245	11(15): 8171-8188, 2011.
246	Pan, Y. et al. Redefining the importance of nitrate during haze pollution to help optimize an
247	emission control strategy. Atmospheric Environment, 141: 197-202, 2016.
248	Peng, J. et al. Markedly enhanced absorption and direct radiative forcing of black carbon
249	under polluted urban environments. Proceedings of the National Academy of
250	Sciences, 113(16): 4266-4271, 2016.
251	Petäjä, T. et al. Enhanced air pollution via aerosol-boundary layer feedback in China.
252	Scientific Reports, 6: 18998, 2016.
253	Richter, A., Burrows, J.P., Nusz, H., Granier, C. and Niemeier, U. Increase in tropospheric
254	nitrogen dioxide over China observed from space. Nature, 437(7055): 129-132, 2005.
255	Swietlicki, E. et al. Hygroscopic properties of submicrometer atmospheric aerosol particles
256	measured with H-TDMA instruments in various environments-a review. Tellus B:
257	Chemical and Physical Meteorology, 60(3): 432-469, 2008.
258	van Donkelaar, A. et al.Global estimates of ambient fine particulate matter concentrations
259	from satellite-based aerosol optical depth: development and application.
260	Environmental health perspectives, 118(6): 847-855, 2010.





261	Wang, G. et al Persistent sulfate formation from London Fog to Chinese haze. Proceedings
262	of the National Academy of Sciences, 113(48): 13630-13635, 2016.
263	Wang, S. et al. Quantifying the Air Pollutants Emission Reduction during the 2008 Olympic
264	Games in Beijing. Environmental Science & Technology, 44(7): 2490-2496, 2010.
265	Wang, Y. et al. Mechanism for the formation of the January 2013 heavy haze pollution
266	episode over central and eastern China. Science China Earth Sciences, 57(1): 14-25,
267	2014.
268	Wang, Y.H. et al. Aerosol physicochemical properties and implications for visibility during
269	an intense haze episode during winter in Beijing. Atmos. Chem. Phys., 15(6): 3205-
270	3215, 2015.
271	Wu, Y. et al. The challenge to NOx emission control for heavy-duty diesel vehicles in China.
272	Atmos. Chem. Phys., 12(19): 9365-9379, 2012.
273	Xin, J. et al. The Campaign on Atmospheric Aerosol Research Network of China: CARE-
274	China. Bulletin of the American Meteorological Society, 96(7): 1137-1155. 2015.
275	Xing, J. et al. Observations and modeling of air quality trends over 1990–2010 across the
276	Northern Hemisphere: China, the United States and Europe. Atmos. Chem. Phys.,
277	15(5): 2723-2747, 2015.
278	Yang, F. et al. Characteristics of PM2.5 speciation in representative megacities and across
279	China. Atmos. Chem. Phys., 11(11): 5207-5219, 2011.
280	Zhang, J.K. et al. Characterization of submicron aerosols during a month of serious pollution
281	in Beijing, 2013. Atmos. Chem. Phys., 14(6): 2887-2903, 2014.
282	Zhang, R. et al. Formation of urban fine particulate matter. Chem Rev, 115(10): 3803-55,
283	2015
284	Zhang, X.Y. et al. Atmospheric aerosol compositions in China: spatial/temporal variability,
285	chemical signature, regional haze distribution and comparisons with global aerosols.
286	Atmos. Chem. Phys., 12(2): 779-799, 2012.

















35 30 25 20 15 10 y=0.14x+9.5 40 30 20 10 (a)North China (b)Yangtze River Plain y = 0.07+5. 0 ĝ, 14 12 10 8 6 (d)China (c)Sichuan Bsain y = 0.02x+1.5 Figure 3. Trends of NO $_2$  concentration over china from SCIAMACHY from the year 2002 to 2012 (10<sup>15</sup> mol/cm<sup>2</sup>) 





(a) North China y = 1.6x + 484.3y = -1.1x +844.2 (b)Yangtze River Plain ŝ ~ ~ 1500 (c)Sichuan Basin y = 0.5x + 698.4(d) China y = 2.2x + 166.3 900 600 300 Figure 4. Trends of  $SO_2$  concentration over china from SCIAMACHY from the year of 2004 to 2012 (1000DU) 





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Figure 5: Variation of (a) NO<sub>x</sub>, SO<sub>2</sub>, (b) chemical composition (c) mass fraction of
organic, nitrate, sulfate, ammonium and chloride with decreased visibility during the
intensive campaign in Beijing.







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352 Figure 6. A schematic picture illustrating the process of enhanced emission of

aerosol inorganic precusors and formation of aerosol inorganic components leading to

354 increased hygroscopicity and aerosol water uptake ability leading to considerable

- 355 visibility degradation in China. The plus symbols represents the strengthening of a
- 356 specific process.

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