## **1** Increased inorganic aerosol fraction contributes to air pollution and

# 2 haze in China

Yonghong Wang<sup>1,2</sup>, Yuesi Wang<sup>1,6,8</sup>, Lili Wang<sup>1</sup>, Tuukka Petäjä<sup>2,3</sup>, Qiaozhi Zha<sup>2</sup>,Chongshui Gong<sup>1,4</sup>,Sixuan Li<sup>7</sup>, Yuepeng Pan<sup>1</sup>, Bo Hu<sup>1</sup>, Jinyuan Xin<sup>1</sup> and Markku Kulmala<sup>2,3,5</sup>

- <sup>3</sup> <sup>1</sup>State Key Laboratory of Atmospheric Boundary Layer Physics and Atmospheric Chemistry
- 4 (LAPC), Institute of Atmospheric Physics, Chinese Academy of Sciences, Beijing 100029,
- 5 China
- <sup>6</sup> <sup>2</sup>Institute for Atmospheric and Earth System Research / Physics, Faculty of Science, P.O.Box
- 7 64, 00014 University of Helsinki, Helsinki, Finland
- 8 <sup>3</sup>Joint international research Laboratory of Atmospheric and Earth SysTem sciences
- 9 (JirLATEST), Nanjing University, Nanjing, China
- <sup>4</sup>Institute of Arid meteorology, China Meteorological Administration, Lanzhou 730000,

11 China

- 12 <sup>5</sup>Aerosol and Haze Laboratory, Beijing Advanced Innovation Center for Soft Matter Science
- 13 and Engineering, Beijing University of Chemical Technology (BUCT), Beijing, China
- 14 <sup>6</sup>Centre for Excellence in Atmospheric Urban Environment, Institute of Urban Environment,
- 15 Chinese Academy of Science, Xiamen, Fujian 361021, China
- 16 <sup>7</sup>State Key Laboratory of Numerical Modeling for Atmospheric Sciences and Geophysical
- 17 Fluid Dynamics (LASG), Institute of Atmospheric Physics, Chinese Academy of Sciences,
- 18 Beijing 100029, China
- <sup>8</sup>University of Chinese Academy of Sciences, Beijing 100049, China
- 20 Revised to: Atmospheric Chemistry and Physics

#### Corresponding authors: Y.S. Wang, L.L. Wang and M. Kulmala

# 22 E-mail: wys@mail.iap.ac.cn;wll@mail.iap.ac.cn; markku.kulmala@helsinki.fi

# 23 Abstract

24 The detailed formation mechanism of increased number of haze events in China is 25 still not very clear. Here, we found that reduced surface visibility from 1980-2010 and 26 an increase in satellite derived columnar concentrations of inorganic precursor from 27 2002 to 2012 are connected with each other. Typically higher inorganic mass fractions 28 lead to increased aerosol water uptake and light scattering ability in elevated relative 29 humidity. Satellite observation of aerosol precursors of NO<sub>2</sub> and SO<sub>2</sub> showed increased 30 concentrations during study period. Our in-situ measurement of aerosol chemical 31 composition in Beijing also confirmed increased contribution of inorganic aerosol 32 fraction as a function of increased particle pollution level. Our investigations 33 demonstrate that the increased inorganic fraction in the aerosol particles is a key 34 component in the frequently occurring haze days during studying period, and 35 particularly the reduction of nitrate, sulfate and their precursor gases would contribute 36 towards better visibility in China.

# 37 Introduction

As one of the most heavily polluted regions in the world, China has suffered from air pollution for decades (Hao et al., 2007; Zhang et al., 2015). Aerosol particles, as major air pollutant, have significant effects on human health (Lelieveld et al., 2015). The general public and the central government of China have realized the severe situation and have taken some actions to improve the air quality nationwide in the recent years. For example, the state council published a plan for air pollution control, in 44 September of 2013, aim to reduce PM<sub>2.5</sub> concentrations by 10%~25% in different 45 regions of China. The successful implementation requires a sufficient knowledge of 46 haze formation mechanism (Kulmala, 2015) and comprehensive observation network 47 (Kulmala, 2018). Our understanding on haze events with high PM<sub>2.5</sub> concentrations in 48 China is still limited due to the spatial-temporal variation of aerosol properties and 49 limited observation information (Wang et al., 2016). Recent studies found that 50 secondary aerosol components are important during the intense haze events in Beijing, 51 Xi'an, Chengdu and Guangzhou during January of 2013, and the reduction of aerosol 52 precursors is a key step to reduce particle pollution (Guo et al., 2014; Huang et al., 53 2014). The analysis of longer time series data from Nanjing shows that secondary 54 particles are typically dominating even the number concentrations in polluted 55 conditions (Kulmala., 2016). A recent study have suggested significantly decreased 56 trends of PM<sub>2.5</sub> and SO<sub>2</sub> in China from 2015-2017 by analyzing data sets from Ministry 57 of Ecology and Environment of China (Silver et al., 2018). The column NO<sub>2</sub> 58 concentration obtained from OMI showed increased trend during 2005-2011, while a 59 decreasing trend during 2012-2015 (Itahashi et al., 2016). The SO<sub>2</sub> concentration has 60 decreased around 50% from 2012-2015 in North China Plain due to economic 61 slowdown and governments efforts to restrain emissions from power and industrial 62 sectors (Krotkov et al., 2016). However, the most abundant mass fractions of 63 atmospheric aerosol are inorganic and organic components, which have large spatio-64 temporal variation (Jimenez et al., 2009). Identifying the most abundant as well as 65 critical aerosol species that contribute to the haze formation in a longtime perspective 66 is important to draw up effective plans for the air pollution control.

Here, a comprehensive data sets were used to reveal that an increasing trend ofinorganic components in atmospheric aerosol may be a pivotal factor, at least, which

leads to frequently occurred haze events in China from 1980-2010. We suggests that
the controlling of inorganic aerosol components of nitrate, sulfate and their precursors
should be of a high priority due to their strong water uptake abilities and therefore, light
scattering ability in high RH conditions.

# 73 **2.Methodology**

74 The daily averaged visibility and relative humidity data in 262 sites of China are 75 obtained from the Integrated Surface Dataset (ISD) from National Oceanic and 76 Atmospheric Administration National Climate Data Center of the USA from 1980-77 2010 (https://www.ncdc.noaa.gov/isd). The visibility observations were made three 78 times a day at 8-hour intervals begins at 00:00 by well trained technicians. They 79 measured visual range using distinctive markers, such as tall buildings, mountains and 80 towers, to which the distance from the meteorological monitoring stations are known. 81 We quantified the importance of relative humidity to visibility as the hygroscopic 82 inorganic compounds typically grow in size in high humidity (Swietlicki et al., 2008). 83 Aerosol size growth and composition change in high humidity condition are highly 84 related light scattering ability (Zhang et al., 2015). Studies always use f (RH), a 85 parameter which is defined as the ratio of light scattering coefficient under high RH 86 with that under low RH. f(RH) is a unitless number, usually ranges from one to two. 87 At ambient RH around 80%, a higher f(RH) value usually correspondes to higher 88 inorganic aerosol fraction, while a lower value usually corespondes to high organic 89 fraction. The reason is that inorganic aerosol compounds of nitrate, sulfate and 90 ammonium have more strong water uptake ability than organic comopounds. In 91 addition, the high humidity condition in ambient prefers the formation of inroganici 92 aerosol from precusors of NO<sub>2</sub> and SO<sub>2</sub> (Wang et al., 2014). In this study, for a given

site and given year, we defined a *f* (RH)-like parameter, Ri, using the observed annual
visibility (V) as a ratio (R<sub>i</sub>) between visibility values from the surface observation
stations, when the daily average RH was below 40% for more than 20 days. In the
corresponding high-humidity cases daily RH was between 80%~90% for more than
20 days each year at a given observation site:

98

99 
$$R_i = \frac{V_{dry}}{V_{wet}}.$$

We use this ratio to infer long trend of aerosol hygroscopicity information. In
addition, we calculate anomaly (A) from the ratio for a given year *i* as a difference from
the 30-year (R<sub>30y</sub>) from 1980 to 2010:

$$A = R_i - R_{30y}$$

Our spatial focus is placed on North China Plain, Yangtze River Plain and Sichuan
Basin due to frequent haze events (Zhang et al., 2012). The stations in Pearl River delta
region and other Southern China stations were not included due to limited days with
the daily average RH below 40%.

108 The atmospheric column amount of NO<sub>2</sub> and SO<sub>2</sub> data are obtained from 2002-2012 109 and 2004-2012, respectively, from SCIAMACHY (Scanning Imaging Absorption 110 spectrometer for Atmospheric CHartographY) satellite products. SCIAMACHY is an 111 atmospheric sensor aboard the European satellite ENVISAT. It was launched in March 2002 as 112 a joint project of Germany, the Netherlands and Belgium. It measures atmospheric absorption 113 in spectral bands from the ultraviolet to the near infrared (240 nm - 2380 nm), allows to retrieve 114 atmospheric column concentrations of O<sub>3</sub>, BrO, OClO, ClO, SO<sub>2</sub>, H<sub>2</sub>CO, NO, NO<sub>2</sub>, NO<sub>3</sub>, CO, 115 CO<sub>2</sub>, CH<sub>4</sub>, H<sub>2</sub>O, N<sub>2</sub>O, aerosols, radiation and cloud properties (Boersma et al., 2004). Aerosol

116 chemical composition from GEOS (Goddard Earth Observing System)-Chem chemical 117 transport model combined with satellite AOD products in China during 1998-2012 is 118 used. The model utilizes assimilated meteorology data and regional emission 119 inventories with a horizontal resolution of  $2^{\circ} \times 2^{\circ}$  with 47 vertical levels from surface to 120 80 km. The PM<sub>2.5</sub> concentration was retrieved from AOD of satellite and the 121 relationship between PM<sub>2.5</sub> and AOD in GEOS-Chem. The detailed information about 122 the model can be found in (Boys et al., 2014). Aerosol chemical composition of organic, 123 sulfate, nitrate, ammonium and chloride were measured with a high-resolution-time of 124 flight-aerosol mass spectrometers during an intensive campaign in urban Beijing from 125 November of 2010 to January of 2011 (DeCarlo et al., 2006). Detailed information of 126 data analysis, collection efficiencies (CE) and relative ionization efficiencies are 127 presented in Zhang et al. (2014).

128

# **3. Results and discussion**

### 129 3.1 Decreasing trend in visibility in high relative humidity conditions

130 According to the geographical division, our study sites are mainly in North China 131 Plain (NCP), Sichuan Basin (SCB) and Yangtze River Plain (YRP) as showed in Figure 132 1. The average visibility in low RH in NCP, SCB, YRP and China are 18.2 km, 21.4 133 km, 19.5 km and 23.3 km, while the values in high RH conditions are 10.6 km, 13.7 134 km, 13.7 km and 17.4 km, respectively. In general, visibility in low RH condition has 135 fluctuated trend, particularly in Northern China Plain, Sichuan Basin and Yangtze river 136 Plain region, whereas visibility in high RH conditions showed decreasing trend as 137 shown in Figure S1 (a) and (b). The average ratio of visibility in low RH to that in high 138 RH from 1980-2010 is presented in Figure 1. The maximum ratios were identified in 139 eastern China and in some western Chinese cities. Three heavily polluted regions,

140 Northern China Plain, Sichuan Basin and Yangtze river Plain were identified based on 141 values of high R<sub>i</sub>, which are also constant with aerosol mass concentrations and haze 142 day distributions (van Donkelaar et al., 2010; Xin et al., 2015). That is, the higher ratios 143 occurred in more severe air pollution areas, like North China Plain, Sichuan Basin and 144 the city of Urumqi, where the contribution of hygroscopic aerosol is more pronounced 145 in comparison with non-hygroscopic dust particles. The average R<sub>i</sub> during 1980-1984 146 in Northern China Plain, Sichuan Basin and Yangtze river Plain are 1.62, 1.41, 1.29 147 and 1.31, respectively, contrasting with the values of 1.98, 1.81, 1.70 and 1.52 during 148 2006-2010. The increments are 22.3%, 27.3%, 31% and 16%, respectively. It is worth 149 noting that the R<sub>i</sub> in Yangtze river Plain region exhibits the most increment, which 150 implies the increased emissions with rapid economic growth. Long time trends of this 151 ratio in a specific site can reveal the variation of aerosol inorganic fraction and organic 152 fraction due to their different hygroscopicity and water uptake capacity and associated 153 light extinction ability. That is, the mass fractions and concentrations of sulfate, nitrate 154 and ammonium may have increased over study period as they dominate water uptake 155 ability compared with other components (e.g., organic, black carbon, dust and metal 156 elements, see Table S1) in the atmospheric aerosol (Wang et al., 2015). For the selected 157 regions, we have calculated the anomaly as a regional average as shown in Figure 2. 158 The ratio showed increasing trends over three regions of China and the maximum trends 159 occurred in North China Plain with the value of 0.0168 per year, which indicate an 160 increase of hygroscopic aerosol in these regions during the 30-year observation period.

161 To corroborate our results, Yang et al. (2011) showed an increasing fraction of 162 inorganic components by 20% in Beijing from 1998 to 2008 using in-situ offline aerosol 163 chemical composition measurement, especially in summer, while the fractions of 164 hydrophobic components such as organic and black carbon decreased in the aerosol phase. A study by Boys et al. (2014) revealed that increasing fraction of secondary inorganic aerosol is dominated in the increased mass concentration of PM<sub>2.5</sub> in China from 1998-2012 using GEOS-Chem model combined satellite results. By using observed meteorology data sets, Fu et al. (2014) revealed that the number of haze days have significantly increased in the past three decades over North China Plain due to the increase in hygroscopic inorganic aerosol composition.

# 171 3.2 Enhanced emissions of inorganic aerosol precursors

172 The longterm trends of aerosol precursors and their spatial variability can improve 173 our understanding of the trends in aerosol chemical composition. Figure 3 and Figure 174 4 show atmospheric column trends of NO<sub>2</sub> and SO<sub>2</sub> observed from SCIAMACHY. The 175 column NO<sub>2</sub> level can be a good proxy for vehicle and coal burning emission associated with oil and coal consumption (Richter et al., 2005). The column amount of NO<sub>2</sub> 176 177 showed pronounced increasing trends in three regions of China, particularly in Northern China with the value of  $0.14 \times 10^{15}$  molecule/cm<sup>2</sup>/year from 2002 to 2011. This is 178 179 probably associated with the increase in power plant and on-road vehicle emissions 180 (Wu et al., 2012; Krotkov et al., 2016). The average NO<sub>2</sub> concentration in Northern 181 China increased by more than two-fold, while in the Yangze River Plain region experienced a considerable smaller trend in NO<sub>2</sub>, with the value of 9.7  $\times 10^{15}$ 182 molecule/cm<sup>2</sup> in 2010 and 6.4  $\times 10^{15}$  molecule/cm<sup>2</sup> in 2002. It is worting noting a 183 184 dectresed trend occurred during year 2008, which is mainly due to emission reduction 185 during the Olympic games and economic downturn (Lin and McElroy, 2011). As a 186 whole, the column  $NO_2$  concentration in China doubled from 2002 to 2010, with the values of  $1.4 \times 10^{15}$  molecule/cm<sup>2</sup> in 2002 and 2.8  $\times 10^{15}$  molecule/cm<sup>2</sup> in 2010, 187 188 respectively.

189 Figure 4 depicts trend in SO<sub>2</sub> concentration in four regions of China from 2004 to 190 2010. The SO<sub>2</sub> concentration showed an increasing trends in North China Plain, 191 Sichuan Basin and Yangze River Plain, but increased mostly in China from 2004 to 192 2012. A decreasing trend was observed during the year of 2008 and 2009, especially in 193 Northern China Plain. This may be due to a combination of Chinese economic 194 downturn and emission reduction during the Olympic games (Lin and McElroy, 2011) 195 (Wang et al., 2010). Anyway, as an important aerosol precursor, NO<sub>2</sub> showed the 196 most increasing trend in China from 2002-2012, consistent with the trend of increased 197 aerosol concentration by modeling result (Xing et al., 2015). Figure S3 shows the 198 annual trends of aerosol inorganic fraction in PM<sub>2.5</sub> mass concentration from 1998-2012 199 with GEOS-Chem model combined with satellite results in China. The results indicate 200 that North China Plain area suffered the most from heavily pollution, consistent with 201 our surface observations (Xin et al., 2015). Aerosol concentrations have increased 202 considereably from 1980 to 2010. The modeling combined with satellite results by Boys et al. 203 (2014) show that concurrently the fraction of inorganic fraction has increased more rapidly. 204 Consequently, the water uptake of the aerosol have increased leading to reduced visibility as 205 we suggested, which is consistent with ground-based observations (Yang et al., 2011).

# **3.3 Validation of increased inorganic aerosol components with elevated air**

## 207 pollution level with in-situ measurement

To validate our hypothesis that the increased inorganic components contribute to visibility degradation, we used four months of aerosol chemical composition and visibility data from urban Beijing from November of 2010 to February of 2011. As shown in Figure 5, we divided the visibility values into four bins, which corresponds to clean time to heavy pollution time and to conditions in between. The inorganic aerosol 213 precursors of SO<sub>2</sub> and NO<sub>2</sub> nearly doubled as the visibility decreased from more than 214 10 km (clean time) to less than 2 km (heavily polluted time). At the same time, the mass 215 concentrations of nitrate, sulfate and ammonium components increased to 13.5  $\mu$ g m<sup>-3</sup>, 216 15.5  $\mu$ g m<sup>-3</sup> and 10.6  $\mu$ g m<sup>-3</sup>, respectively. Meanwhile, the mass fraction of these 217 inorganics increased from 11.3% to 17.3%, from 13.0% to 19.9% and from 9.6% to 218 13.6%, respectively. At the same time, the mass concentration and fraction of organic 219 components changed from12.2  $\mu$ g m<sup>-3</sup> to 33.4  $\mu$ g m<sup>-3</sup> and 60% to 46%, respectively.

220 We also investigated the relationship between relative humidity (RH) and volume 221 fractions of ammonium sulfate, ammonium nitrate and organic aerosols as shown in Figure 6. 222 The results indicated that ammonium nitrate increased most significantly with elevated RH. On 223 the contrary, ammonium sulfate, as another inorganic compound, showed only a moderate 224 positive correlation with RH and a decrease in the volume fraction was observed in RH values 225 larger than 75%. This might be associated with liquid phase oxidation of SO<sub>2</sub> under high RH 226 condition, to sulfate aerosol. Increasing RH may provide more atmospheric oxidants and 227 reaction media for the aqueous-phase oxidation (Zhang et al., 2015). The volume fraction of 228 organic aerosol showed negative correlation with increasing RH, as presented in Figure 6 (c), 229 which was maybe due to a faster increasing volume fraction of inorganic aerosol than organic 230 aerosol.

This direct observation shows that the contribution of inorganic components increased during this campaign. It is plausible that the increased concentration of SO<sub>2</sub> and NO<sub>2</sub> are highly associated with this giving rise to the long-term trends observed in Figure 2 (Pan et al., 2016; Wang et al., 2014).

## **4.** Conclusion and implication for atmospheric air pollution control

Atmospheric pollution and associated haze events has a dramatic effect on climatechange, human health and visibility degradation (Ding et al., 2013; Petäjä et al., 2016;

238 Wang et al., 2015; Zhang et al., 2015). Here, longterm visibility measurements combined with satellite data sets, in-situ measurements and model results revealed that 239 240 increased fractions of inorganic aerosol components in the particle matter are crucial in 241 contributing to more haze events from 1980-2010. In this way, aerosol hygroscopic 242 growth has occurred in lower relative humidity conditions than before due to more ammonium nitrate aerosol, and the light scattering ability of atmospheric aerosol 243 244 enhanced as shown in Figure 7. Another mechanism is that high concentration of NO<sub>x</sub> 245 can promote the conversion of SO<sub>2</sub> to form sulfate aerosol via aqueous phase oxidation 246 during intensive pollution periods (He et al., 2014; Wang et al., 2016). Considering the vast energy consumption in the future decades and the sources of inorganic components 247 248 in atmospheric aerosol, we demonstrate that the reduction nitrate, sulfate, ammonium 249 and their precursors should be continued to get better visibility in China.

### 250 Author contributions

251 Y.H.W had the original idea. L.L.W and C.S.G provided and processed satellite and

visibility data. Y.S.W provided measurements of aerosol chemical composition

253 data.Y.H.W, Y.S.W, L.L.W, T.P and M.K interpreted the data and write the paper.

All the authors commented on the paper.

255

256

#### 257 Acknowledgements

258 We acknowledge Dr B. Boys and Professor R. Martin of Dalhousie University for providing

- 259 GEOS-Chem model results in China. We acknowledge the free use of tropospheric NO<sub>2</sub> and SO<sub>2</sub>
- 260 column data from the SCIAMACHY sensor from www.temis.nl. This work was supported by the

- 261 Ministry of Science and Technology of China (No: 2017YFC0210000), the National Research
- 262 Program for key issues in air pollution control (DQGG0101), the National Natural Science
- 263 Foundation of China No.41775162 and Academy of Finland via Center of Excellence in
- 264 Atmospheric Sciences and the National Natural Science Foundation of China (41605119).

- 266 Competing financial interests
- 267 The authors declare no competing financial interests.

#### 268 References

269 Boersma, K. F., Eskes, H. J. and Brinksma, E. J.: Error analysis for tropospheric NO 2 retrieval

270 from space, Journal of Geophysical Research: Atmospheres, 109(D4), n/a-n/a,

- doi:10.1029/2003jd003962, 2004.
- Boys, B.L. et al. Fifteen-Year Global Time Series of Satellite-Derived Fine Particulate Matter.
  Environmental Science & Technology, 48(19): 11109-11118, 2014.
- 274 DeCarlo, P.F. et al. Field-Deployable, High-Resolution, Time-of-Flight Aerosol Mass

275 Spectrometer. Analytical Chemistry, 78(24): 8281-8289, 2006.

Ding, A.J. et al.. Intense atmospheric pollution modifies weather: a case of mixed biomass burning
with fossil fuel combustion pollution in eastern China. Atmos. Chem. Phys., 13(20):

**278** 10545-10554, 2013.

279 Fu, G.Q., Xu, W.Y., Yang, R.F., Li, J.B. and Zhao, C.S. The distribution and trends of fog and

haze in the North China Plain over the past 30 years. Atmos. Chem. Phys., 14(21): 11949-

- **281** 11958, 2014.
- Guo, S. et al. Elucidating severe urban haze formation in China. Proceedings of the National
  Academy of Sciences of the United States of America, 111(49): 17373-8, 2014.
- Hao, J., He, K., Duan, L., Li, J. and Wang, L. Air pollution and its control in China. Frontiers of
  Environmental Science & Engineering in China, 1(2): 129-142, 2007

- He, H. et al.Mineral dust and NOx promote the conversion of SO2 to sulfate in heavy pollution
  days. Sci Rep, 4: 4172, 2014.
- Huang, R.J. et al. High secondary aerosol contribution to particulate pollution during haze events
  in China. Nature, 514(7521): 218-22, 2014.
- 290 Itahashi, S., Muto, T., Irie, H., Uno, I. and Kurokawa, J.: Turnaround of Tropospheric Nitrogen
- 291 Dioxide Pollution Trends in China, Japan, and South Korea, Sola, 12(0), 170–174,
- doi:10.2151/sola.2016-035, 2016.
- Jimenez, J.L. et al. Evolution of Organic Aerosols in the Atmosphere. Science, 326(5959): 15251529, 2009.
- 295 Krotkov, N. A., McLinden, C. A., Li, C., Lamsal, L. N., Celarier, E. A., Marchenko, S. V.,
- 296 Swartz, W. H., Bucsela, E. J., Joiner, J., Duncan, B. N., Boersma, K. F., Veefkind, J. P.,
- 297 Levelt, P. F., Fioletov, V. E., Dickerson, R. R., He, H., Lu, Z. and Streets, D. G.: Aura
- 298 OMI observations of regional SO<sub&amp;gt;2&amp;lt;/sub&amp;gt; and
- 299 NO<sub&amp;gt;2&amp;lt;/sub&amp;gt; pollution changes from 2005 to 2015,
- **300** Atmospheric Chemistry and Physics, 16(7), 4605–4629, doi:10.5194/acp-16-4605-2016,
- 301 2016
- 302 Kulmala, M.. Build a global Earth observatory.Nature, 553, 2018.
- 303 Kulmala, M.. China's choking cocktail. Nature, 526, 2015.
- 304 Kulmala M., L.K., Virkkula A., Petäjä T., Paasonen P., Kerminen V.-M., Nie W., Qi X., Shen Y.,
- Chi X. & Ding A. On the mode-segregated aerosol particle number concentration load:
  contributions of primary and secondary particles in Hyytiälä and Nanjing. Boreal Env.
  Res, 21: 319–331, 2016.
- Lelieveld, J., Evans, J.S., Fnais, M., Giannadaki, D. and Pozzer, A. The contribution of outdoor air
  pollution sources to premature mortality on a global scale. Nature, 525(7569): 367-71,
  2015.
- Lin, J.T. and McElroy, M.B. Detection from space of a reduction in anthropogenic emissions of
  nitrogen oxides during the Chinese economic downturn. Atmos. Chem. Phys., 11(15):
  8171-8188, 2011.

314 Pan, Y. et al. Redefining the importance of nitrate during haze pollution to help optimize an 315 emission control strategy. Atmospheric Environment, 141: 197-202, 2016. 316 Peng, J. et al. Markedly enhanced absorption and direct radiative forcing of black carbon under 317 polluted urban environments. Proceedings of the National Academy of Sciences, 113(16): 318 4266-4271, 2016. 319 Petäjä, T. et al. Enhanced air pollution via aerosol-boundary layer feedback in China. Scientific 320 Reports, 6: 18998, 2016. 321 Richter, A., Burrows, J.P., Nusz, H., Granier, C. and Niemeier, U. Increase in tropospheric 322 nitrogen dioxide over China observed from space. Nature, 437(7055): 129-132, 2005. 323 Swietlicki, E. et al. Hygroscopic properties of submicrometer atmospheric aerosol particles 324 measured with H-TDMA instruments in various environments—a review. Tellus B: 325 Chemical and Physical Meteorology, 60(3): 432-469, 2008. 326 van Donkelaar, A. et al.Global estimates of ambient fine particulate matter concentrations from 327 satellite-based aerosol optical depth: development and application. Environmental health 328 perspectives, 118(6): 847-855, 2010. 329 Wang, G. et al.. Persistent sulfate formation from London Fog to Chinese haze. Proceedings of the 330 National Academy of Sciences, 113(48): 13630-13635, 2016. 331 Wang, S. et al. Quantifying the Air Pollutants Emission Reduction during the 2008 Olympic 332 Games in Beijing. Environmental Science & Technology, 44(7): 2490-2496, 2010. 333 Wang, Y. et al. Mechanism for the formation of the January 2013 heavy haze pollution episode 334 over central and eastern China. Science China Earth Sciences, 57(1): 14-25, 2014. 335 Wang, Y.H. et al. Aerosol physicochemical properties and implications for visibility during an 336 intense haze episode during winter in Beijing. Atmos. Chem. Phys., 15(6): 3205-3215, 337 2015. 338 Wu, Y. et al. The challenge to NOx emission control for heavy-duty diesel vehicles in China. 339 Atmos. Chem. Phys., 12(19): 9365-9379, 2012. 340 Xin, J. et al. The Campaign on Atmospheric Aerosol Research Network of China: CARE-China. 341 Bulletin of the American Meteorological Society, 96(7): 1137-1155. 2015.

- Xing, J. et al. Observations and modeling of air quality trends over 1990–2010 across the Northern
  Hemisphere: China, the United States and Europe. Atmos. Chem. Phys., 15(5): 27232747, 2015.
- Yang, F. et al. Characteristics of PM2.5 speciation in representative megacities and across China.
  Atmos. Chem. Phys., 11(11): 5207-5219, 2011.
- 347 Zhang, J.K. et al. Characterization of submicron aerosols during a month of serious pollution in
  348 Beijing, 2013. Atmos. Chem. Phys., 14(6): 2887-2903, 2014.
- Zhang, R. et al. Formation of urban fine particulate matter. Chem Rev, 115(10): 3803-55, 2015
- 350 Zhang, X.Y. et al. Atmospheric aerosol compositions in China: spatial/temporal variability,
- 351 chemical signature, regional haze distribution and comparisons with global aerosols.
- Atmos. Chem. Phys., 12(2): 779-799, 2012.

# Figure caption

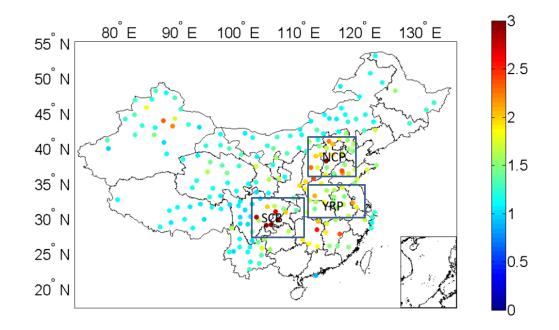
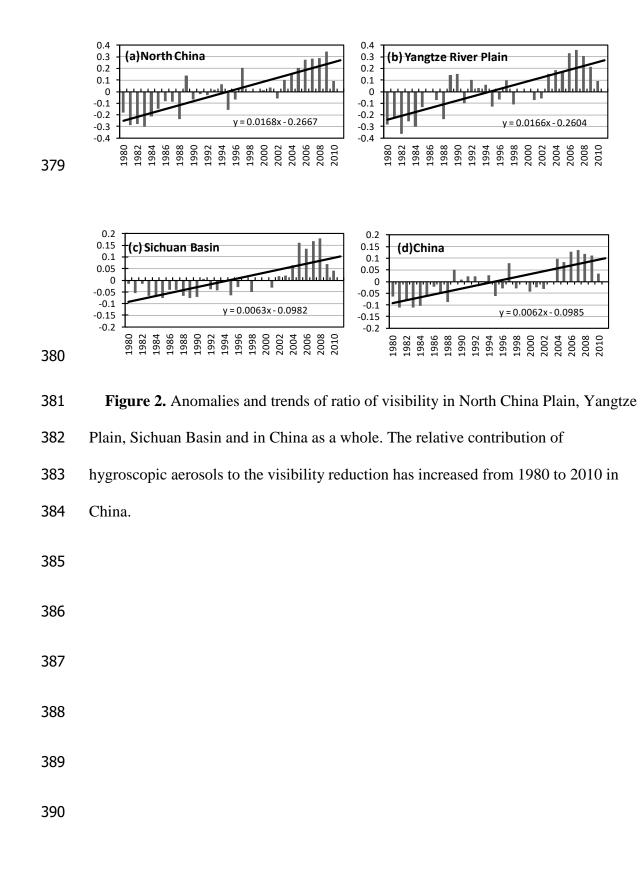
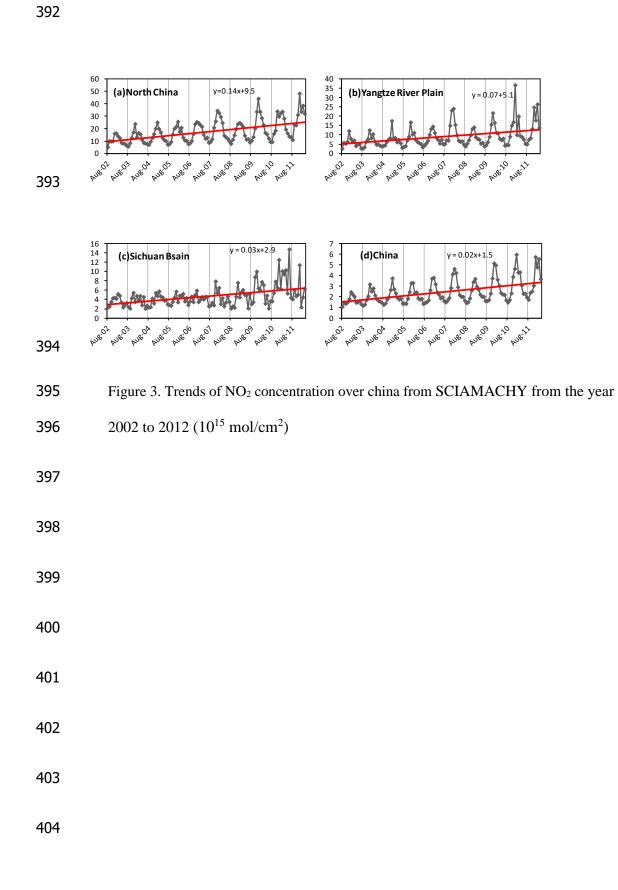


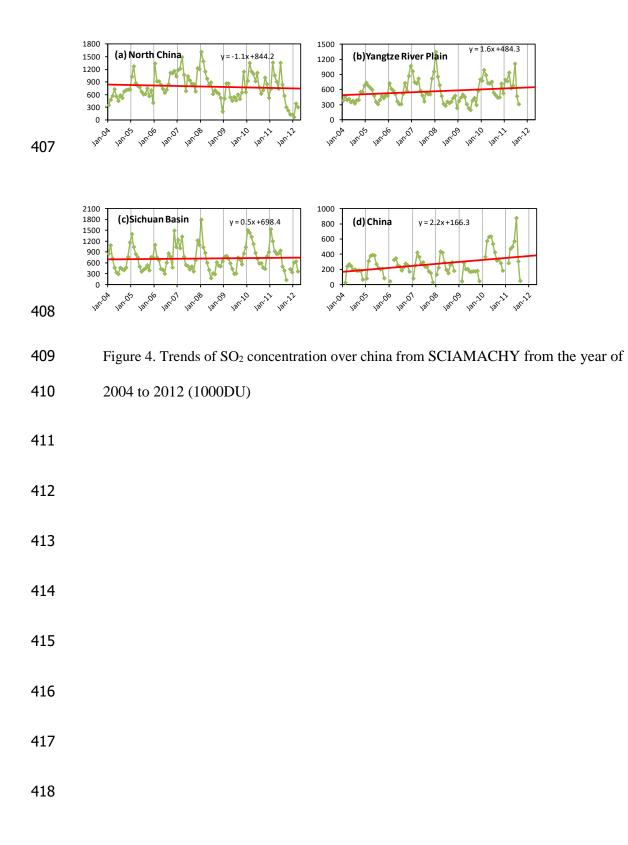


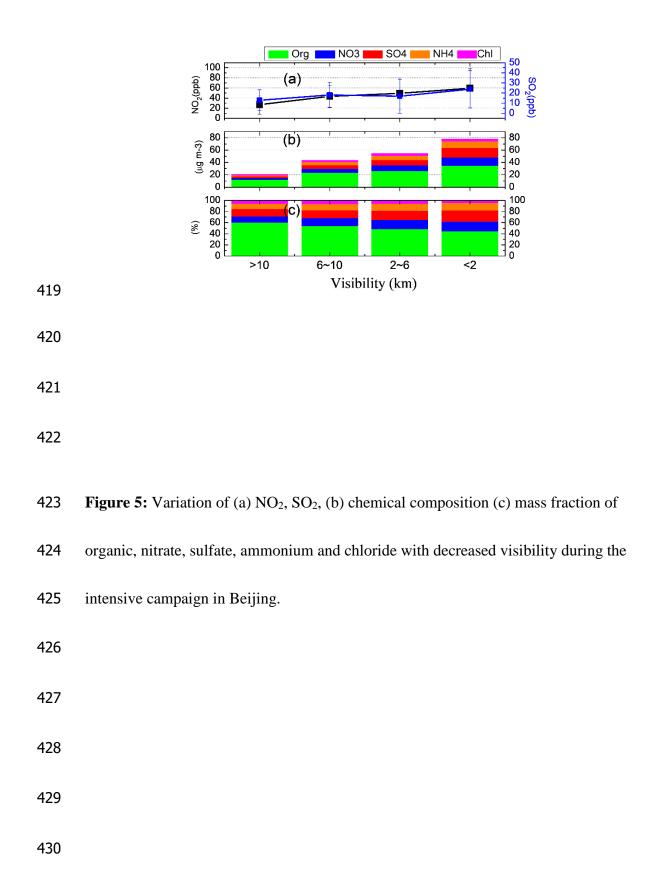
Figure 1. The distribution of the average surface visibility ratio in dry and wet
conditions based on observations at 262 surface observation sites in China. The
aerosol in the industrialized regions of China in the East are more hygroscopic than
aerosol particles in the west of China.

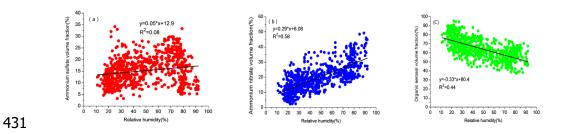












432 Figure 6. Relation between relative humidity (RH) and volume fractions of (a) ammonium

- 433 sulfate (b) ammonium nitrate (c) organic aerosol.
- 434
  435
  436
  437
  438
  439
  440
  441
  442
  443
  444
  445

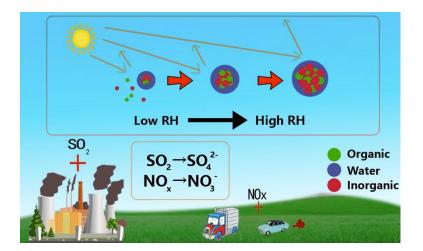


Figure 7. A schematic picture illustrating the process of enhanced emission of
aerosol inorganic precusors and formation of aerosol inorganic components leading to
increased hygroscopicity and aerosol water uptake ability leading to considerable
visibility degradation in China. The plus symbols represents the strengthening of a
specific process.