

1 Increased inorganic aerosol fraction contributes to air pollution and

2 haze in China

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22 **Abstract**

23 The detailed formation mechanism of increased number of haze events in China is
24 still not very clear. Here, we found that reduced surface visibility and an increase in
25 satellite derived columnar concentration of inorganic precursor concentrations are
26 connected with each other. Typically higher inorganic mass fractions lead to increased
27 aerosol water uptake and light scattering ability in elevated relative humidity. Satellite
28 observation of aerosol precursors of NO_2 and SO_2 showed increased concentrations
29 during study period. Our in-situ measurement of aerosol chemical composition in
30 Beijing also confirmed increased contribution of inorganic aerosol fraction as a function
31 of increased particle pollution level. Our investigations demonstrate that the increased
32 inorganic fraction in the aerosol particles is a key component in the frequently occurring
33 haze days during studying period, and particularly the reduction of nitrate, sulfate and
34 their precursor gases would contribute towards better air quality in China.

35 **Introduction**

36 As one of the most heavily polluted regions in the world, China has suffered from
37 air pollution for decades (Hao et al., 2007; Zhang et al., 2015). Aerosol particles, as
38 major air pollutant, have significant effects on human health (Lelieveld et al., 2015).
39 The general public and the central government in China have realized the severe
40 situation and have taken some measures to improve the air quality nationwide in the
41 recent years. For example, the state council published a plan for air pollution control,
42 in September of 2013, aim to reduce $\text{PM}_{2.5}$ concentrations by 10%~25% in different
43 regions of China. The successful implementation requires a sufficient knowledge of

44 haze formation mechanism (Kulmala, 2015) and comprehensive observation network
45 (Kulmala, 2018). Our understanding on haze events with high PM_{2.5} concentrations in
46 China is still limited due to the spatial-temporal variation of aerosol properties and
47 limited observation information (Wang et al., 2016). Recent studies found that
48 secondary aerosol components are important during the intense haze events in Beijing,
49 Xi'an, Chengdu and Guangzhou during January of 2013, and the reduction of aerosol
50 precursors is a key step to reduce particle pollution (Guo et al., 2014; Huang et al.,
51 2014). The analysis of longer time series data from Nanjing shows that secondary
52 particles are typically dominating even the number concentrations in polluted
53 conditions (Kulmala M., 2016). The most abundant mass fractions of atmospheric
54 aerosol are inorganic and organic components, which have large spatio-temporal
55 variation (Jimenez et al., 2009). Identifying the most abundant as well as critical aerosol
56 species that contribute to the haze formation in a longtime perspective is important to
57 draw up effective plans for the air pollution control.

58 Here, a comprehensive data sets were used to reveal that an increasing trend of
59 inorganic components in atmospheric aerosol may be a pivotal factor, at least, which
60 leads to frequently occurred haze events in China from 1980-2010. We suggests that
61 the controlling of inorganic aerosol components of nitrate, sulfate and their precursors
62 should be of a high priority due to their strong water uptake abilities and therefore, light
63 scattering ability in high RH conditions.

64 **2.Methodology**

65 The daily averaged visibility and relative humidity data in 262 sites of China are
66 obtained from the Integrated Surface Dataset (ISD) from National Oceanic and
67 Atmospheric Administration National Climate Data Center of the USA

68 (https://www.ncdc.noaa.gov/isd). The visibility observations were made three times a
69 day at 8-hour intervals begins at 00:00 by well trained technicians. They measured
70 visual range using distinctive markers, such as tall buildings, mountains and towers,
71 to which the distance from the meteorological monitoring stations are known. First,
72 we quantified the importance of relative humidity to visibility as the hygroscopic
73 inorganic compounds typically grow in size in high humidity (Swietlicki et al., 2008).

74 **Aerosol size growth and composition change in high humidity condition are highly**
75 **related light scattering ability (Zhang et al., 2015). Studies always use f (RH), a**
76 **parameter which is defined as the ratio of light scattering coefficient under high RH**
77 **with that under low RH. f (RH) is a unitless number, usually ranges from one to two.**
78 **At ambient RH around 80%, a higher f (RH) value usually corresponds to higher**
79 **inorganic aerosol fraction, while a lower value usually corresponds to high organic**
80 **fraction. The reason is that inorganic aerosol compounds of nitrate, sulfate and**
81 **ammonium have more strong water uptake ability than organic compounds. In**
82 **addition, the high humidity condition in ambient prefers the formation of inorganic**
83 **aerosol from precursors of NO_2 and SO_2 (Wang et al., 2014). In this study, for a given**
84 **site and given year, we defined a f (RH)-like parameter, R_i , using the observed annual**
85 **visibility (V) as a ratio (R_i) between visibility values from the surface observation**
86 **stations, when the daily average RH was below 40% for more than 20 days. In the**
87 **corresponding high-humidity cases daily RH was between 80%~90% for more than**
88 **20 days each year at a given observation site:**

89

90
$$R_i = \frac{V_{dry}}{V_{wet}}$$

91 We use this ratio to infer long trend of aerosol hygroscopicity information. In
92 addition, we calculate anomaly (A) from the ratio for a given year i as a difference from
93 the 30-year (R_{30y}) from 1980 to 2010:

94
$$A = R_i - R_{30y}.$$

95 Our spatial focus is placed on North China Plain, Yangtze River Plain and Sichuan
96 Basin due to frequent haze events (Zhang et al., 2012). The stations in Pearl River delta
97 region and other Southern China stations were not included due to limited days with
98 the daily average RH below 40%.

99 The atmospheric column amount of NO_2 and SO_2 data are obtained from
100 SCIAMACHY (Scanning Imaging Absorption spectrometer for Atmospheric
101 CHartographY) satellite products. Modeled aerosol chemical composition from GEOS
102 (Goddard Earth Observing System) - Chem chemical transport model in China during
103 1998-2012 is used. The model utilizes assimilated meteorology data and regional
104 emission inventories with a horizontal resolution of $2^\circ \times 2^\circ$ with 47 vertical levels from
105 surface to 80 km. The detailed information about the model can be found in (Boys et
106 al., 2014). Aerosol chemical composition of organic, sulfate, nitrate, ammonium and
107 chloride were measured with a high-resolution-time of flight-aerosol mass
108 spectrometers (DeCarlo et al., 2006). Detailed information of data analysis, collection
109 efficiencies (CE) and relative ionization efficiencies are presented in Zhang et al. (2014).

110 **3. Results and discussion**

111 **3.1 Decreasing trend in visibility in high relative humidity conditions**

112 According to the geographical division, our study sites are mainly in North China
113 Plain (NCP), Sichuan Basin (SCB) and Yangtze River Plain (YRP) as showed in Figure

114 1. The average visibility in low RH in NCP, SCB, YRP and China are 18.2 km, 21.4
115 km, 19.5 km and 23.3 km, while in high RH conditions are 10.6 km, 13.7 km, 13.7 km
116 and 17.4 km, respectively. In general, visibility in low RH condition has fluctuated
117 trend, particularly in Northern China Plain, Sichuan Basin and Yangtze river Plain
118 region, whereas visibility in high RH conditions showed decreasing trend as shown in
119 Figure S1 (a) and (b) . The average ratio of visibility in low RH to that in high RH from
120 1980-2010 is presented in Figure 1. The maximum ratios were identified in eastern
121 China and in some western Chinese cities. Three heavily polluted regions, Northern
122 China Plain, Sichuan Basin and Yangtze river Plain were identified based on values of
123 high R_i , which are also constant with aerosol mass concentrations and haze day
124 distributions (van Donkelaar et al., 2010; Xin et al., 2015). That is, the higher ratios
125 occurred in more severe air pollution areas, like North China Plain, Sichuan Basin and
126 the city of Urumqi, where the contribution of hygroscopic aerosol is more pronounced
127 in comparison with non-hygroscopic dust particles. The average R_i during 1980-1984
128 in Northern China Plain, Sichuan Basin and Yangtze river Plain are 1.62, 1.41, 1.29
129 and 1.31, respectively, contrasting with the values of 1.98, 1.81, 1.70 and 1.52 during
130 2006-2010. The increments are 22.3%, 27.3%, 31% and 16%, respectively. It is worth
131 noting that the R_i in Yangtze river Plain region exhibits the most increment, which
132 implies the increased emissions with rapid economic growth. Long time trends of this
133 ratio in a specific site can reveal the variation of aerosol inorganic fraction and organic
134 fraction due to their different hygroscopicity and water uptake capacity and associated
135 light extinction ability. That is, the mass fractions and concentrations of sulfate, nitrate
136 and ammonium may have increased over study period as they dominate water uptake
137 ability compared with other components (e.g., organic, black carbon, dust and metal
138 elements, see Table S1) in the atmospheric aerosol (Wang et al., 2015). For the selected

139 regions, we have calculated the anomaly as a regional average as shown in Figure 2.
140 The ratio showed increasing trends over three regions of China and the maximum
141 trends occurred in North China Plain with the value of 0.0168 per year, which indicate
142 an increase of hygroscopic aerosol in these regions during the 30-year observation
143 period.

144 To corroborate our results, Yang et al. (2011) showed an increasing fraction of
145 inorganic components by 20% in Beijing from 1998 to 2008 **using in-situ offline**
146 **measurement**, especially in summer, while the fractions of hydrophobic components
147 such as organic and black carbon decreased in the aerosol phase. A recent study by
148 Boys et al. (2014) revealed that increasing fraction of secondary inorganic aerosol is
149 dominated in the increased mass concentration of PM_{2.5} in China from 1998-2012 **using**
150 **GEOS-Chem model results**. By using observed meteorology data sets, Fu et al. (2014)
151 **revealed that the number of haze days have significantly increased in the past three**
152 **decades over North China Plain due to the increase in hygroscopic inorganic aerosol**
153 **composition**.

154 **3.2 Enhanced emissions of inorganic aerosol precursors**

155 The longterm trends of aerosol precursors and their spatial variability can improve
156 our understanding of the trends in aerosol chemical composition. Figure 3 and Figure
157 4 show atmospheric column trends of NO₂ and SO₂ observed from SCIAMACHY. The
158 column NO₂ level can be a good proxy for vehicle and coal burning emission associated
159 with oil and coal consumption (Richter et al., 2005). The column amount of NO₂
160 showed pronounced increasing trends in three regions of China, particularly in Northern
161 China with the value of 0.14×10^{15} molecule/cm²/year from 2002 to 2011. This is
162 probably associated with the increase in power plant and on-road vehicle emissions

163 (Wu et al., 2012). The average NO_2 concentration in Northern China increased by more
164 than two-fold, while in the Yangze River Plain region experienced a considerable
165 smaller trend in NO_2 , with the value of 9.7×10^{15} molecule/cm² in 2010 and 6.4×10^{15}
166 molecule/cm² in 2002. As a whole, the column NO_2 concentration in China doubled
167 from 2002 to 2010, with the values of 1.4×10^{15} molecule/cm² in 2002 and 2.8×10^{15}
168 molecule/cm² in 2010, respectively.

169 Figure 4 depicts trend in SO_2 concentration in four regions of China from 2004 to
170 2010. The SO_2 concentration showed an increasing trends in North China Plain,
171 Sichuan Basin and Yangze River Plain, **but increased mostly in China from 2004 to**
172 **2012**. A decreasing trend was observed during the year of 2008 and 2009, especially in
173 Northern China Plain. This may be due to a combination of Chinese economic
174 downturn and emission reduction during the Olympic games (Lin and McElroy, 2011)
175 (Wang et al., 2010). **Anyway, as an important aerosol precursor, NO_2 showed the**
176 **most increasing trend in China, consistent with the trend of increased aerosol**
177 **concentration by modeling result (Xing et al., 2015)**. Figure S3 shows the annual trends
178 of aerosol inorganic fraction in $\text{PM}_{2.5}$ mass concentration from 1998-2012 with GEOS-
179 Chem model **combined with satellite** results in China. The results indicate that North
180 China Plain area suffered the most from heavily pollution, consistent with our surface
181 observations as well as earlier work (Xin et al., 2015). Aerosol concentrations have
182 increased considerably from 1980 to 2010. The modeling combined with satellite results by
183 **Boys et al. (2014)** show that concurrently the fraction of inorganic fraction has increased more
184 rapidly. **Consequently, the water uptake of the aerosol have increased leading to reduced**
185 **visibility as we suggested, which is consistent with ground-based observations (Yang et al.,**
186 **2011)**.

187 **3.3 Validation of increased inorganic aerosol components with elevated air**

188 **pollution level with in-situ measurement**

189 To validate our hypothesis that the increased inorganic components contribute to
190 visibility degradation, we used four months of aerosol chemical composition and
191 visibility data from urban Beijing from November of 2010 to February of 2011. As
192 shown in Figure 5, we divided the visibility values into four bins, which corresponds to
193 clean time to heavy pollution time and to conditions in between. The inorganic aerosol
194 precursors of SO_2 and NO_2 nearly doubled as the visibility decreased from more than
195 10 km (clean time) to less than 2 km (heavily polluted time). At the same time, the mass
196 concentration of nitrate, sulfate and ammonium components increased to $13.5 \mu\text{g m}^{-3}$,
197 $15.5 \mu\text{g m}^{-3}$ and $10.6 \mu\text{g m}^{-3}$, respectively. Meanwhile, the mass fraction of these
198 inorganics increased from 11.3% to 17.3%, from 13.0% to 19.9% and from 9.6% to
199 13.6%, respectively. At the same time, the mass concentration and fraction of organic
200 components changed from $12.2 \mu\text{g m}^{-3}$ to $33.4 \mu\text{g m}^{-3}$ and 60% to 46%, respectively.

201 We also investigated the relationship between relative humidity (RH) and volume
202 fractions of ammonium sulfate, ammonium nitrate and organic aerosols as shown in Figure 6.
203 The results indicated that ammonium nitrate increased most significantly in elevated RH. On
204 the contrary, ammonium sulfate, as another inorganic compound, showed only a moderate
205 positive correlation with RH and a decrease in the volume fraction was observed in RH values
206 larger than 75%. This might be associated with liquid phase oxidation of SO_2 under high RH
207 condition, to sulfate aerosol. Increasing RH may provide more atmospheric oxidants and
208 reaction media for the aqueous-phase oxidation (Zhang et al., 2015). The volume fraction of
209 organic aerosol showed negative correlation with increasing RH, as presented in Figure 6 (c),
210 which was maybe due to an faster increasing volume fraction of inorganic aerosol than organic
211 aerosol.

212 This direct observation shows that the contribution of inorganic components increased
213 during this campaign. It is plausible that the increased concentration of SO₂ and NO₂
214 are highly associated with this giving rise to the long-term trends observed in Figure 2
215 (Pan et al., 2016; Wang et al., 2014).

216 **4. Conclusion and implication for atmospheric air pollution control**

217 Atmospheric pollution and associated haze events has a dramatic effect on climate
218 change, human health and visibility degradation (Ding et al., 2013; Petäjä et al., 2016;
219 Wang et al., 2015; Zhang et al., 2015). Here, longterm visibility measurements
220 combined with satellite data sets, in-situ measurements and model results revealed that
221 increased fractions of inorganic aerosol components in the particle matter are crucial in
222 contributing to more haze events. In this way, aerosol hygroscopic growth has occurred
223 in lower relative humidity conditions than before due to more ammonium nitrate
224 aerosol, and the light scattering ability of atmospheric aerosol enhanced as shown in
225 Figure 7. Another mechanism is that high concentration of NO_x can promote the
226 conversion of SO₂ to form sulfate aerosol via aqueous phase oxidation during intensive
227 pollution periods (He et al., 2014; Wang et al., 2016). Considering the vast energy
228 consumption in the future decades and the sources of inorganic components in
229 atmospheric aerosol, we denmonstrate that the reduction nitrate, sulfate, ammonium
230 and their precursors maybe more critical in China.

231

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238 Atmospheric Sciences and the National Natural Science Foundation of China (41605119).

239

240 **Competing financial interests**

241 The authors declare no competing financial interests.

242 **Author contributions**

243 Y.H.W had the original idea. L.L.W and C.S.G provided and processed satellite and
244 visibility data. Y.S.W provided measurements of aerosol chemical composition
245 data. Y.H.W, Y.S.W, L.L.W, T.P and M.K. interpreted the data and write the paper.

246 All the authors commented on the paper.

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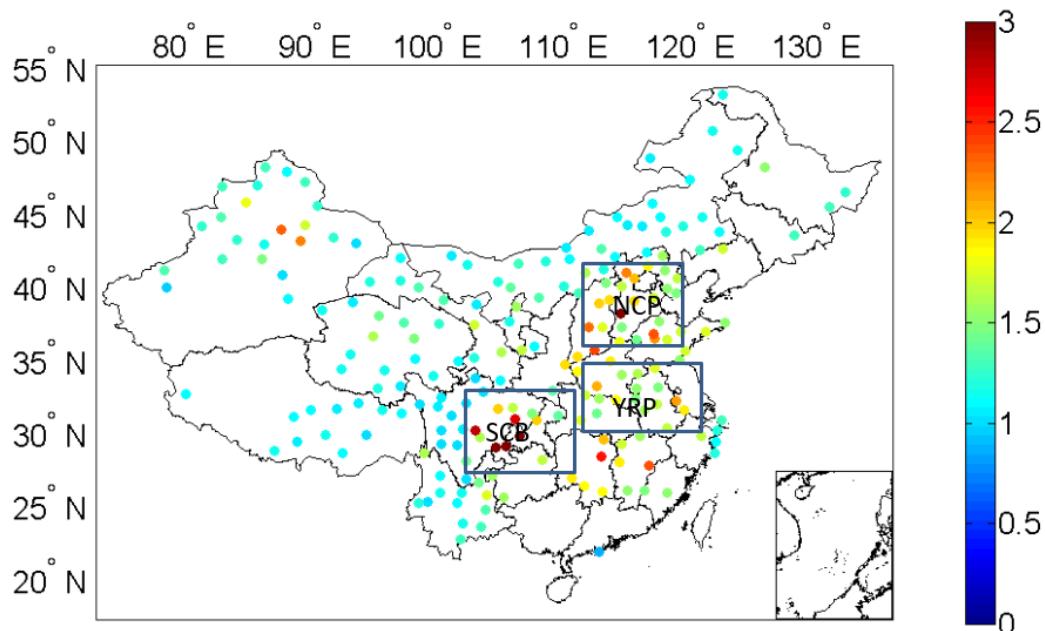
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321

322 **Figure caption**



323

324 **Figure 1.** The distribution of the average surface visibility ratio in dry and wet
325 conditions based on observations at 262 surface observation sites in China. The
326 aerosol in the industrialized regions of China in the East are more hygroscopic than
327 aerosol particles in the west of China.

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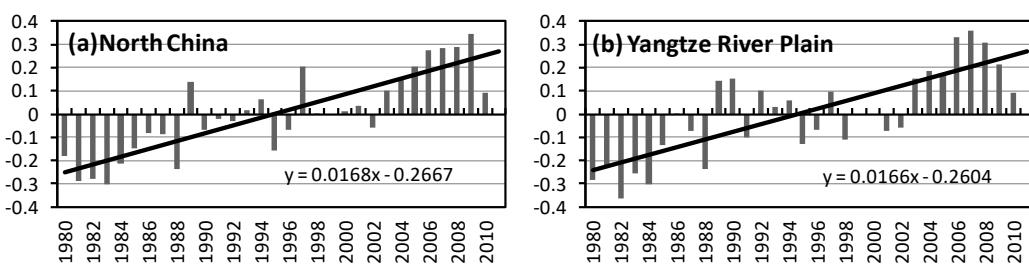
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338 **Figure 2.** Anomalies and trends of ratio of visibility in North China Plain, Yangtze
339 Plain, Sichuan Basin and in China as a whole. The relative contribution of
340 hygroscopic aerosols to the visibility reduction has increased from 1980 to 2010 in
341 China.

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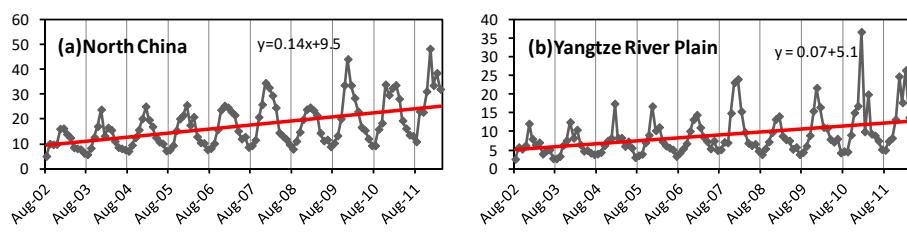
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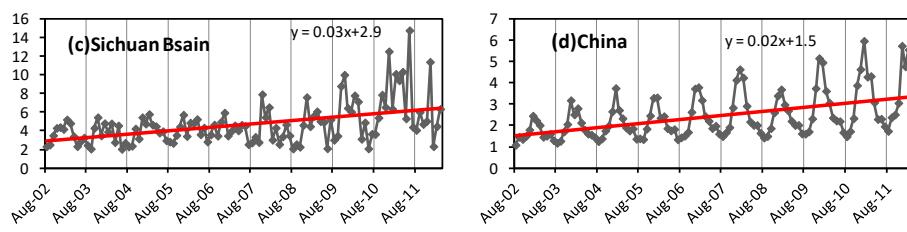
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Figure 3. Trends of NO₂ concentration over China from SCIAMACHY from the year

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2002 to 2012 (10^{15} mol/cm²)

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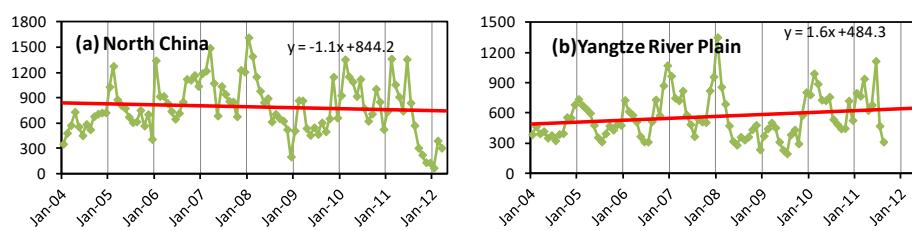
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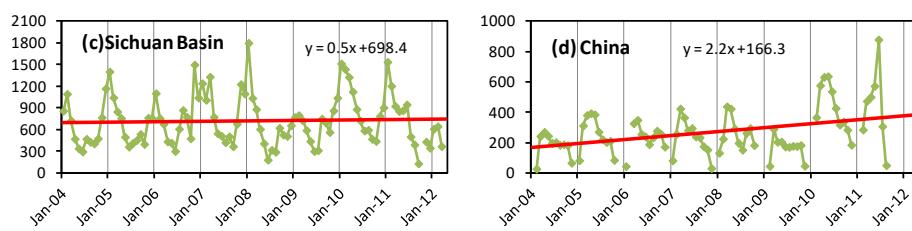
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Figure 4. Trends of SO₂ concentration over China from SCIAMACHY from the year of

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2004 to 2012 (1000DU)

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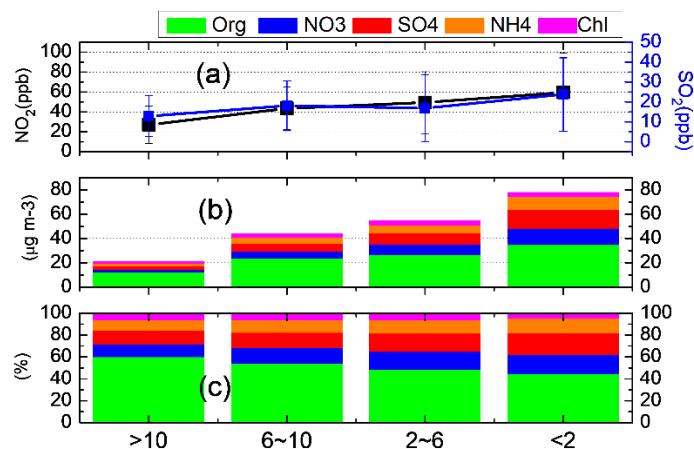
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379 **Figure 5:** Variation of (a) NO_2 , SO_2 , (b) chemical composition (c) mass fraction of
380 organic, nitrate, sulfate, ammonium and chloride with decreased visibility during the
381 intensive campaign in Beijing.

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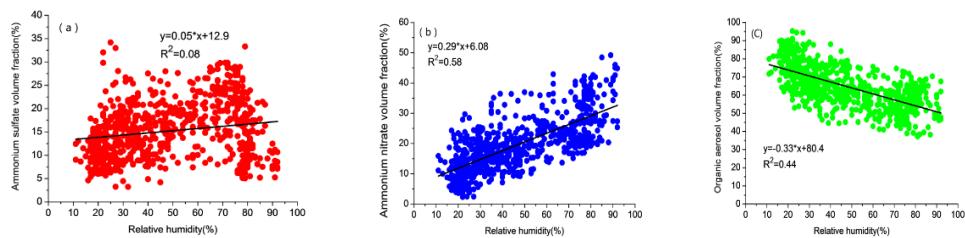
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388 Figure 6. Relation between relative humidity (RH) and volume fractions of (a) ammonium
389 sulfate (b) ammonium nitrate (c) organic aerosol.

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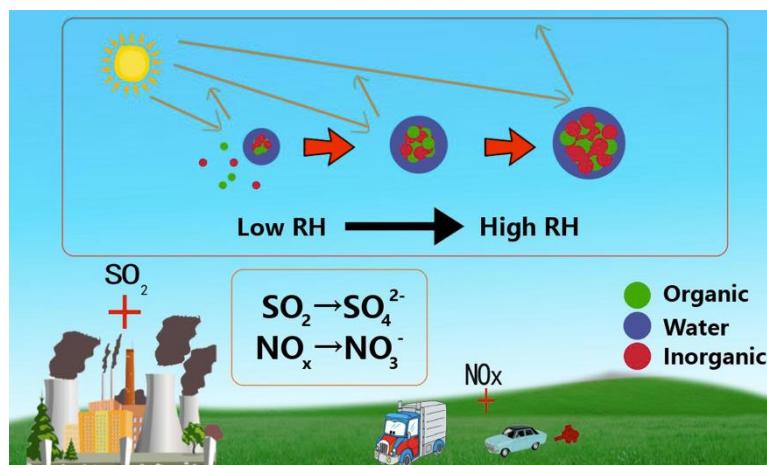
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403 Figure 7. A schematic picture illustrating the process of enhanced emission of
404 aerosol inorganic precursors and formation of aerosol inorganic components leading to
405 increased hygroscopicity and aerosol water uptake ability leading to considerable
406 visibility degradation in China. The plus symbols represents the strengthening of a
407 specific process.

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