



# 1 **Consequences of dynamic and timing properties of** 2 **new aerosol particle formation and consecutive growth events**

3 Imre Salma and Zoltán Németh

4 Institute of Chemistry, Eötvös University, H-1518 Budapest, P.O. Box 32, Hungary

5 Correspondence to: Imre Salma ([salma@chem.elte.hu](mailto:salma@chem.elte.hu))

6 **Abstract.** Dynamic properties, i.e. particle formation rate  $J_6$  and particle diameter growth rate  
7  $GR_{10}$ , and timing properties, i.e. starting time ( $t_1$ ) and duration time interval ( $\Delta t$ ) of 247  
8 quantifiable (class 1A) atmospheric new particle formation (NPF) and consecutive particle  
9 diameter growth events identified in the city centre and near-city background of Budapest over  
10 6 full measurement years together with related gas-phase  $H_2SO_4$  proxy, condensation sink (CS)  
11 of vapours, basic meteorological data and concentrations of criteria pollutant gases were  
12 derived, evaluated, discussed and interpreted. In the city centre, nucleation ordinarily starts at  
13 09:15 UTC+1, and it is maintained for approximately 3 h. The NPF and growth events produce  
14 4.6 aerosol particles with a diameter of 6 nm in 1 cm<sup>3</sup> of air in 1 s, and cause the particles with  
15 a diameter of 10 nm to grow with a typical rate of 7.3 nm h<sup>-1</sup>. Nucleation starts approximately  
16 1 h earlier in the near-city background, it shows substantially smaller  $J_6$  (with a median of 2.0  
17 cm<sup>-3</sup> s<sup>-1</sup>) and  $GR_{10}$  values (with a median of 5.0 nm h<sup>-1</sup>), while the duration of nucleation is  
18 similar to that in the centre. Monthly distributions of the dynamic properties and daily  
19 maximum  $H_2SO_4$  proxy do not follow the mean monthly pattern of the event occurrence  
20 frequency. The factors that control the event occurrence and that govern the intensity of particle  
21 formation and growth are not directly linked. Condensing atmospheric chemical species and/or  
22 their processes in the city centre seem to contribute equally to new particle formation and  
23 particle growth. In the near-city background, however, chemical compounds available and their  
24 processes power particle growth more than particle formation. There is a minimum growth rate  
25 of approximately 1.8 nm h<sup>-1</sup> that is required for nucleated particles to reach the lower end of  
26 the diameter interval measured (6 nm) under the actual/local conditions. Monthly distributions  
27 and relationships among the properties mentioned provided several indirect evidence that  
28 chemical species other than  $H_2SO_4$  largely influence the particle growth and possibly  
29 atmospheric NPF process as well. The  $J_6$ ,  $GR_{10}$  and  $\Delta t$  can be described by log-normal  
30 distribution. Most of the extreme dynamic properties could not be explained by  $H_2SO_4$  proxy,  
31 CS, meteorological data or pollutant gas concentrations. Approximately 40% of the NPF and



32 growth events exhibited broad beginning, which can be an urban feature. For 9% of all cases,  
33 it was feasible to calculate 2 separate sets of dynamic properties. The later onset frequently  
34 shows more intensive particle formation and growth than the first onset by a typical factor of  
35 approximately 1.4. The first event is of regional type, while the second event, superimposed on  
36 the first, is often associated with sub-regional, thus urban NPF and growth process.

37

## 38 **1 Introduction**

39

40 Molecules and molecular fragments in the air collide randomly, and can form electrically  
41 neutral or charged clusters. Most clusters decompose shortly. Chemical stabilising interactions  
42 among certain components within a cluster can enhance its lifetime, during which it can grow  
43 further by additional molecular collisions through some distinguishable size regimes (Kulmala  
44 et al., 2014). If the diameter of these clusters reaches a critical value of  $1.5\pm 0.3$  nm (Kulmala  
45 et al., 2013), they become thermodynamically stable, and their further growth turns into a  
46 spontaneous process. Supersaturation is a necessary atmospheric condition for this principal  
47 conversion. It is virtually a phase transition, which takes place in a dispersed manner in the  
48 atmosphere, so it generates an aerosol system. The newly formed particles grow further by  
49 condensation to larger particles in most cases due to the existing supersaturation. Photochemical  
50 oxidation products such as  $\text{H}_2\text{SO}_4$  (Sipilä et al., 2010), extremely low-volatile organic  
51 compounds (ELVOCs, Ehn et al., 2014; Jokinen et al., 2015) and highly oxygenated molecules  
52 (HOMs, Bianchi et al., 2016; Kirkby et al., 2016; Tröstl et al., 2016) together with  $\text{H}_2\text{O}$  vapour,  
53  $\text{NH}_3$  (Kirkby et al., 2011), amines (Almeida et al., 2013), other oxidation products of volatile  
54 organic compounds (VOCs; Metzger et al., 2010; Schobesberger et al., 2013; Riccobono et al.,  
55 2014) and  $\text{NO}_2$  can play an important role in both the formation and growth. The VOCs include  
56 compounds of anthropogenic and biogenic origin mainly isoprenoids such as  $\alpha$ -pinene (Kirkby  
57 et al., 2016). In some specific coastal regions, iodine oxides produced from marine biota are  
58 involved (O'Dowd et al., 2002). Atmospheric concentration of these key compounds at a level  
59 that is smaller typically by 12–14 orders of magnitude than the concentration of air molecules  
60 is already sufficient for the phenomenon (Kulmala et al., 2014). Relative importance of the  
61 organics increases with particle size (Riipinen et al., 2011; Ehn et al., 2014), and their  
62 supersaturation is maintained by fast gas-phase autooxidation chemical reactions of VOCs  
63 (Crouse et al., 2013). The overall phenomenon is ordinarily confined in time for 1 day or so,  
64 and, therefore, it can be regarded as an event in time, and is referred as new aerosol particle  
65 formation (NPF) and consecutive particle growth event.



66

67 Such events appear to take place almost everywhere in the world and anytime (Kulmala et al.,  
68 2004; Nieminen et al., 2018). Their occurrence frequency and, more importantly, their  
69 contribution to particle number concentrations were found to be substantial or determinant in  
70 the global troposphere (Spracklen et al., 2006; Kulmala et al., 2014). Moreover, their  
71 contribution to the number of cloud condensation nuclei (CCN) can be 50% or even more  
72 (Makkonen et al., 2009; Merikanto et al., 2009; Sihto et al., 2011), which links the events to  
73 climate system, and emphasizes their global relevance (Kerminen et al., 2012; Carslaw et al.,  
74 2013; Makkonen et al., 2012; Gordon et al., 2016). New particle formation and growth events  
75 were proved to be common in polluted air of large cities as well with a typical relative  
76 occurrence frequency between 10% and 30% (Woo et al., 2001; Baltensperger et al., 2002;  
77 Alam, et al., 2003; Wehner et al., 2004; Salma et al., 2011; Dall'Osto et al., 2013; Xiao et al.,  
78 2015; Zhang et al., 2015; Kulmala et al., 2017, Nieminen et al., 2018). The coupling and  
79 relationships between regional and urban (sub-regional) NPF was demonstrated at least under  
80 favourable orographic conditions (Salma et al., 2016b). New particle formation can increase  
81 the existing particle number concentrations in city centres by a factor of approximately 2 on  
82 nucleation days, while it can produce approximately 28% of ultrafine (UF) particles on a longer  
83 (e.g. annual) time scale (Salma et al., 2017). Particle concentrations from NPF are also  
84 important when compared to the (primary) particles emitted by their dominant source in cities,  
85 namely by road vehicles with internal combustion engines (Paasonen et al., 2016). These jointly  
86 imply that particles from NPF and growth events in cities can influence not only the urban  
87 climate but can contribute to the public's excess health risk from particle number exposures  
88 (Oberdörster et al., 2005; Braakhuis et al., 2014; Salma et al., 2015), and, furthermore, could  
89 be linked to the role of human actions in all these effects.

90

91 In spite of these potentials, conclusive interpretation of the data obtained and results derived  
92 specifically for cities remained hindered so far. Several-year long semi-continuous critically  
93 evaluated and coherent data sets are required for this purpose, which have been generated  
94 gradually. As part of this international progress, investigations dedicated to urban NPF and  
95 growth events in Budapest have been going on since November 2008. Measurements for 5 full  
96 years were realised in the city centre at a fixed location, 1 full year was devoted to  
97 measurements in a near-city background environment, and some other measurements were  
98 accomplished in different urban microenvironments for time intervals of a few months. The  
99 main objectives of this study are to determine, present and analyse the dynamic properties, i.e.



100 particle formation rate and particle diameter growth rate, timing properties, i.e. starting time  
101 and duration time interval of nucleation process of NPF and growth events together with the  
102 major sources and sink of condensing vapours, basic meteorological data and criteria pollutant  
103 gases for 6 years, to investigate and interpret their relationships, to discuss their monthly  
104 distributions, to evaluate and detect some of their features specific for urban atmospheric  
105 environments, and to demonstrate some specific urban influence on the determination of these  
106 properties. These quantities and relationships are of basic importance in many atmospheric  
107 processes for several reasons, and some of them are also discussed.

108

## 109 **2 Experimental methods**

110

111 The measurements took place at two urban locations in Budapest, Hungary. Most measurements  
112 were realised at the Budapest platform for Aerosol Research and Training (BpART) facility (N  
113 47° 28' 29.9", E 19° 3' 44.6", 115 m above mean sea level (a.s.l.; Salma et al., 2016a). This site  
114 represents a well-mixed, average atmospheric environment for the city centre. The other  
115 location was situated at the NW border of Budapest in a wooded area of the Konkoly  
116 Astronomical Observatory of the Hungarian Academy of Sciences (N 47° 30' 00.0", E 18° 57'  
117 46.8", 478 m a.s.l.). This site characterises the air masses entering the city since the prevailing  
118 wind direction in the area is NW. The measurements were accomplished for 6 full-year long  
119 time intervals, i.e. from 03–11–2008 to 02–11–2009, from 19–01–2012 to 18–01–2013, from  
120 13–11–2013 to 12–11–2014, from 13–11–2014 to 12–11–2015, from 13–11–2015 to 12–11–  
121 2016 and from 28–01–2017 to 27–01–2018. In the measurement year 2012–2013, the  
122 instruments were set up in the near-city background, while in all other years, they were installed  
123 in the city centre. Local time (LT=UTC+1 or daylight saving time, UTC+2) was chosen as the  
124 time base of the data unless otherwise indicated because the daily activity time pattern of  
125 inhabitants substantially influences the atmospheric processes in cities (Salma et al., 2014).

126

127 The main measuring system was a flow-switching type differential mobility particle sizer  
128 (DMPS). It consists of a radioactive ( $^{60}\text{Ni}$ ) bipolar charger, a Nafion semi-permeable membrane  
129 dryer, a 28-cm long Vienna-type differential mobility analyser and a butanol-based  
130 condensation particle counter (TSI, model CPC3775). The sample flow was  $2.0 \text{ L min}^{-1}$  in the  
131 high-flow mode, and  $0.31 \text{ L min}^{-1}$  in the low-flow mode with sheath air flow rates 10 times  
132 larger than for the sample flows. The DMPS measures particle number concentrations in an  
133 electrical mobility diameter range from 6 to 1000 nm in the dry state of particles (with a relative



134 humidity of  $RH < 30\%$ ) in 30 channels, which finally yields 27 channels after averaging 3  
135 overlapping channels when joining the data for the 2 flow modes. The time resolution of the  
136 measurements was approximately 10 min till 18–01–2013, and 8 min from 13–11–2013 (after  
137 a planned update of the DMPS system). There was no upper size cut-off inlet applied to the  
138 sampling line, and a weather shield and insect net were only attached. The sampling inlet was  
139 installed at a height of 12.5 m above the street level in the city centre, and of approximately 1.7  
140 m in the near-city background. The measurements were performed according to the  
141 international technical standard (Wiedensohler et al., 2012). The availability of the DMPS data  
142 over 1-year long time intervals are summarised in Table 1. Synoptic meteorological data for air  
143 temperature ( $T$ ), RH, wind speed (WS) and wind direction (WD) were obtained from a  
144 measurement station of the Hungarian Meteorological Service (HMS, no. 12843) by  
145 standardised methods with a time resolution of 1 h. Global solar radiation (GRad) data were  
146 measured by the HMS at a distance of 10 km in E direction with a time resolution of 1 h.  
147 Meteorological data were available in  $>90\%$  of the possible cases in each year. Concentrations  
148 of  $SO_2$ ,  $O_3$ ,  $NO_x$  and CO were obtained from measurement stations of the National Air Quality  
149 Network in Budapest (in a distance of 4.5 km from the urban site, and of 6.9 km from the near-  
150 city background site) located in the upwind prevailing direction from the measurement sites.  
151 They are measured by UV fluorescence (Ysselbach 43C), UV absorption (Ysselbach 49C),  
152 chemiluminescence (Thermo 42C) and IR absorption methods (Thermo 48i), respectively with  
153 a time resolution of 1 h. The concentration data were available in  $>85\%$  of the yearly time  
154 intervals. It is worth mentioning that concentration of  $SO_2$  in the Budapest area is ordinarily  
155 distributed without larger spatial (and temporal) differences (Salma et al., 2011). For the present  
156 study, this was actually proved by evaluating the concentration ratios from two different  
157 municipal station which were located in 2 different directions with an angle between them of  
158  $60^\circ$  in the closest distance from the BpART. The mean  $SO_2$  concentration ratio and standard  
159 deviation (SD) for the 2 stations were  $81 \pm 20\%$  over the 5-year long measurement time interval,  
160 which appears to lay within the experimental uncertainty. The assumption can be justified  
161 indirectly by a conclusion on the monthly distribution of  $SO_2$  concentration in Sect. 4.2.

162

### 163 3 Data treatment

164

165 The measured DMPS data were evaluated according to the procedure protocol recommended  
166 by Kulmala et al. (2012) with some refinements that are related mainly to urban features (see  
167 Sect. 3.1). Particle number concentrations in the diameter ranges from 6 to 1000 nm ( $N$ ), from



168 6 to 25 nm ( $N_{6-25}$ ) and from 6 to 100 nm ( $N_{6-100}$  or UF particles) were calculated from the  
169 measured and inverted DMPS concentrations. Particle number size distribution surface plots  
170 showing jointly the variation in particle diameter and particle number concentration density in  
171 time were also generated. Identification and classification of NPF and growth events was  
172 accomplished on these surface plots (Dal Maso et al., 2005; refined in Németh et al., 2018 for  
173 urban sites) on a daily basis into the following main classes: NPF event days, non-event days,  
174 days with undefined character, and days with missing data (for more than 4 h during the  
175 midday). Relative occurrence frequency of events was determined for each month and year as  
176 the ratio of the number of event days to the total number of relevant (i.e. all–missing) days. A  
177 subset of events with uninterrupted evolution in time, which are called quantifiable (class 1A)  
178 events, were further distinguished because the time evolution of their size distribution functions  
179 were utilised to determine the dynamic and timing properties with good accuracy and reliability.

180

### 181 **3.1 Dynamic and timing properties**

182

183 Growth rate (GR) of nucleation-mode particles was calculated by mode-fitting method  
184 (Kulmala et al., 2012). Particle number median mobility diameter (NMMD) of the nucleation  
185 mode were obtained from fitting the individual size distributions by DoFit algorithm (Hussein  
186 et al., 2004). The growth rate was determined as the slope of the linear line fitted to the time  
187 series of the NMMD data within a time interval around a diameter  $d$ , where the dependence  
188 could be satisfactorily approximated by linear fit. Since the nucleation mode was mostly  
189 estimated by  $N_{6-25}$  in the calculations of the formation rate (see below), and since the median  
190 of the related diameter interval (from 6 to 25 nm) is close to  $d=10$  nm, GRs for particles with a  
191 diameter of 10 nm were determined ( $GR_{10}$ ). This type of GR can be interpreted as an average  
192 GR as far as the given particle diameter range is concerned, but it actually expresses the  
193 beginning of the growth process only, which may have considerable effects on the formation  
194 rate calculations in specific cases (see later).

195

196 Time evolution of an aerosol population is described by the general dynamic equation which  
197 was rearranged, simplified and approximated by several quantities (Kulmala et al., 2001; Dal  
198 Maso et al., 2002; Kulmala et al., 2012; Cai and Jiang, 2017) to express the formation rate  $J_6$   
199 of particles with the smallest detected diameter of  $d_{\min}=6$  nm in a form actually utilised in the  
200 present evaluation as

201



$$J_6 = \frac{dN_{6-25}}{dt} + \text{CoagS}_{10} N_{6-25} + \frac{\text{GR}_{10}}{(25-6)} N_{6-25} - \frac{dN_{A_i, <25}}{dt}. \quad (1)$$

203

204 The first term on the right side of Eq. 1 expresses the concentration increment. The particle  
205 number concentration in the size range from 6 to 25 nm ( $N_{6-25}$ ) was selected to approximate the  
206 nucleation-mode particles  $N_{\text{nuc}} \approx N_{6-25}$ . This is a usual and reasonable choice because it was  
207 proved to be advantageous and effective way in handling fluctuating data sets since  $N_{6-25}$  often  
208 exhibits less sensitivity and smaller scatter in time than the fitted area of the nucleation mode.  
209 It is implicitly assumed that the intensity of the NPF is constant for a certain time interval, and,  
210 therefore,  $dN_{6-25}/dt$  can be determined as the slope of the linear function of  $N_{6-25}$  versus time  $t$   
211 within an interval where the dependence could be satisfactorily approximated by linear fit. A  
212 limitation of the relatively wide size range (6–25 nm) selected can be manifested by  
213 disturbances from primary particles in particular in urban environments. This is taken into  
214 account by the last term of Eq. 1, and is discussed later. The second term on the right side of  
215 Eq. 1 expresses the loss of particles due to coagulation scavenging (due to pre-existing larger  
216 particles). The coagulation scavenging efficiency for particles with a diameter of 10 nm  
217 ( $\text{CoagS}_{10}$ ) was selected to approximate the mean coagulation efficiency of nucleation-mode  
218 particles ( $\text{CoagS}_{\text{nuc}}$ ). This diameter was chosen by considering the median of the related  
219 diameter range, which was discussed above for GR. The coagulation efficiency was calculated  
220 from classical aerosol mechanics with adopting a mass accommodation coefficient of 1 and  
221 utilizing the Fuchs' transition-regime correction factor separately for all the size channels  
222 belonging to the selected size range (Kulmala et al., 2001; Dal Maso et al., 2005; Kulmala et  
223 al., 2013) by using the computation scripts of the University of Helsinki. Self-coagulation  
224 within the nucleation mode was neglected due to its limited concentration. Hygroscopic growth  
225 of particles was not considered since this depends on chemical composition of particles, which  
226 is unknown but expected to change substantially in time. The third term on the right side of Eq.  
227 1 expresses the particle growth out of the considered size range by condensation of vapours.  
228 The  $\text{GR}_{10}$  was selected to approximate a representative value at the median of the particle  
229 diameter range considered (Vuollekoski et al., 2012). It is implicitly assumed that  $\text{GR}_{10}$  can be  
230 regarded to be constant over the time interval of interest. Nevertheless, the diameter growth of  
231 nucleation-mode particles in time is occasionally strongly limited (Fig. S1a). In these specific  
232 cases, the mean relative area of the nucleation mode below 25 nm was determined by fitting the  
233 individual size distributions around the time of the maximum nucleation-mode NMMD, and  
234 the ratios were averaged. The correction in form of the mean relative area was adopted as a





235 multiplication factor for the growth out term in Eq. 1. On a very few days, the particle growth  
236 was followed by a decrease in nucleation-mode NMMD (Salma et al., 2016a). In these cases,  
237 the shrinkage rate (with a formal  $GR_{10} < 0$ ) was derived and adopted. The fourth term on the  
238 right side of Eq. 1 expresses the contribution of (high-temperature) emission sources, usually  
239 vehicular road traffic (Paasonen et al., 2016; Salma et al., 2017) to the  $N_{6-25}$ , which can  
240 provisionally disturb the assumption of  $N_{nuc} \approx N_{6-25}$ . A typical example of such a situation is  
241 shown in Fig. S1b from 10:09 to 12:23 LT. In these specific cases, the contribution of primary  
242 emissions was estimated from the slope of the time series of the fitted peak area of the Aitken  
243 mode below  $d < 25$  nm ( $N_{Ai, < 25}$ ) in the time region of interest. Reliable separation of the  
244 nucleation and Aitken modes from each other was hindered or was not possible for a few  
245 individual size distributions due to overlapping modes and scatter in the concentration data, and  
246 these individual Aitken-mode areas were excluded from or skipped in the time series. Relative  
247 contributions of the concentration increment, coagulation loss and growth out from the diameter  
248 interval – which were significant in all calculations – to  $J_6$  are decreasing in this order with  
249 mean values of 71%, 17% and 12%, respectively (Table S1).

250

251 The formation and growth rates for the measurement years of 2008–2009 and 2012–2013 were  
252 obtained earlier by a slightly different way and neglecting the urban features (Salma et al., 2011,  
253 2016b). To obtain consistent data sets, the dynamic properties for these 2 years were re-  
254 evaluated now by adopting the refinements described above and implementing the experience  
255 gained over the years. The mean new-to-old rate ratios with SDs for the  $GR_{10}$  and  $J_6$  were  
256  $1.06 \pm 0.32$  and  $1.23 \pm 0.37$ , respectively in the city centre (2008–2009) and  $1.04 \pm 0.21$  and  
257  $1.20 \pm 0.35$ , respectively in the near-city background (2012–2013). The re-calculation yielded  
258 non-negligible improvements for the mean formation rates. Some individual formation and  
259 growth rates, in particular the smaller values, were substantially impacted.

260

261 The assumption and estimations above usually represent a reasonable approximation to reality.  
262 The  $N_{6-25}$  is derived from the experimental data in a straightforward way, the  $GR_{10}$  and the  
263 corrections for primary particles and limited particle growth depend on the quality of the size  
264 distribution fitting program as well, while the  $CoagS_{10}$  is determined by using a theoretical  
265 model. The resulting accuracies of the dynamic properties, in particular of  $J_6$ , look rather  
266 complicated, and depend on the spatial heterogeneity in the air masses measured particularly  
267 for the observations performed at a fixed site, size and time resolution of the concentrations  
268 measured, diameter range of the size distributions, fluctuations in the experimental data,





269 selection of the particle diameter channels (interval), choice of the time interval of interest (for  
270 linear fits), sensitivity of the models to the uncertainties (Vuollekoski et al., 2012), and also on  
271 the extent of the validity of the assumptions applied – particularly under highly polluted  
272 conditions (Cai and Jiang, 2017). The situation is further complicated with the fact that the  
273 dynamic (and also the timing) properties are connected to each other. Finally, it is important to  
274 recognise that some NPF and growth curves on the surface plots have rather broad starting time  
275 interval (Fig. S1a and S1c). They occur in a considerable abundance in cities, e.g. in 40% of all  
276 quantifiable events in Budapest (Sect. 4.4). This may yield badly defined or composite dynamic  
277 properties, whose uncertainty can have principle limitations overwhelming the experimental  
278 and model uncertainties.

279

280 Timing properties of NPF and growth events are increasingly recognised, and they can provide  
281 very valuable information even if they are estimated indirectly from the observed diameter  
282 interval  $>1.5$  nm (Sect. 1). The earliest estimated time of the beginning of a nucleation ( $t_1$ ) and  
283 the latest estimated time of the beginning of a nucleation ( $t_2$ ) were derived by a comparative  
284 method (Németh and Salma, 2014) based on the variation in the content of the first size channel  
285 of the DMPS system. Both time parameters includes a time shift that accounts for the particle  
286 growth from the stable neutral cluster mode at approximately 2 nm to the smallest detectable  
287 diameter limit of the DMPS systems (6 nm in our case) by adopting the GR value in the size  
288 window nearest to it in size space. The difference  $\Delta t = t_2 - t_1$  was considered as the duration time  
289 interval of the nucleation process. It represents the time interval during which new aerosol  
290 particles are generated in the air. The uncertainty of the timing parameters is regarded to be  
291 around 30 min under ordinary NPF and growth situations.

292

### 293 3.2 Sources and sink

294

295 Gas-phase  $\text{H}_2\text{SO}_4$  was not measured in the campaigns, and its relative effect was estimated by  
296 its proximity measure (proxy) containing both the major source and sink terms under steady-  
297 state conditions according to Petäjä et al. (2009). It was calculated for radiations  $>10$   $\text{W m}^{-2}$ . In  
298 principle, it is possible to convert the  $\text{H}_2\text{SO}_4$  proxy values to  $\text{H}_2\text{SO}_4$  concentrations by an  
299 empirical scaling factor of  $k = 1.4 \times 10^{-7} \times \text{GRad}^{-0.70}$ , where GRad is expressed in a unit of  $\text{W m}^{-2}$   
300 (Petäjä et al., 2009). The factor was, however, derived for a remote boreal site, and, therefore,  
301 we usually prefer not to perform the conversion since urban areas are expected to differ from  
302 the boreal regions, and adopting the factor could distort the dynamic relationships or time trends



303 investigated. The conversion was applied only to estimate the order of average H<sub>2</sub>SO<sub>4</sub>  
304 atmospheric concentration levels. The results derived by utilising the proxy are subject to larger  
305 uncertainties than in other cases because of these limitations, but they may indicated gross  
306 tendencies.

307

308 Condensation sink for vapour molecules onto the surface of existing aerosol particles was  
309 computed for discrete size distributions as described in earlier papers (Kulmala et al., 2001; Dal  
310 Maso et al., 2002, 2005) and summarised by Kulmala et al. (2013). The equilibrium vapour  
311 pressure of the condensing species was assumed implicitly to be negligible at the surface of the  
312 particles, so similar to sulfuric acid. Dry particle diameters were considered in the calculations.

313

#### 314 **4 Results and discussion**

315

316 Annual median particle number concentrations based on the individual data in the near-city  
317 background in 2012–2013, and in the city centre for the separate measurement years of 2008–  
318 2009, 2013–2014, 2014–2015, 2015–2016 and 2017–2018 were  $3.4 \times 10^3$ , and  $11.5 \times 10^3$ ,  
319  $9.7 \times 10^3$ ,  $9.3 \times 10^3$ ,  $7.5 \times 10^3$  and  $10.6 \times 10^3$  cm<sup>-3</sup>, respectively, which indicate some overall  
320 decreasing trend in the city centre. There was no obvious pattern in the distribution of monthly  
321 median particle number concentrations. The mean UF/N ratio with SD for the same  
322 measurement time intervals were  $67 \pm 14\%$ , and  $79 \pm 6\%$ ,  $75 \pm 10\%$ ,  $75 \pm 11\%$ ,  $76 \pm 11\%$  and  
323  $80 \pm 10\%$ , respectively. The values correspond to ordinary urban atmospheric environments in  
324 Europe (Putaud et al., 2010). An overview on the number of classified days separately for the  
325 1-year long measurement time intervals is given in Table 1. The availability of the daily size  
326 distribution surface plots with respect to all days ensures that the data are representative on  
327 yearly and monthly time scales, except for 08 and 09–2015, when there were missing days in  
328 larger ratios. The number of quantifiable event days (248 cases) is also considerable, which  
329 establishes to arrive at firm conclusion on the NPF and growth events as well.

330



331 **Table 1.** Number of days with new aerosol particle formation and growth event, quantifiable (class 1A)  
 332 event days, non-event days, undefined days, missing days and the coverage of relevant days with respect  
 333 to all days in the near-city background and city centre separately for the 1-year long measurement time  
 334 intervals.  
 335

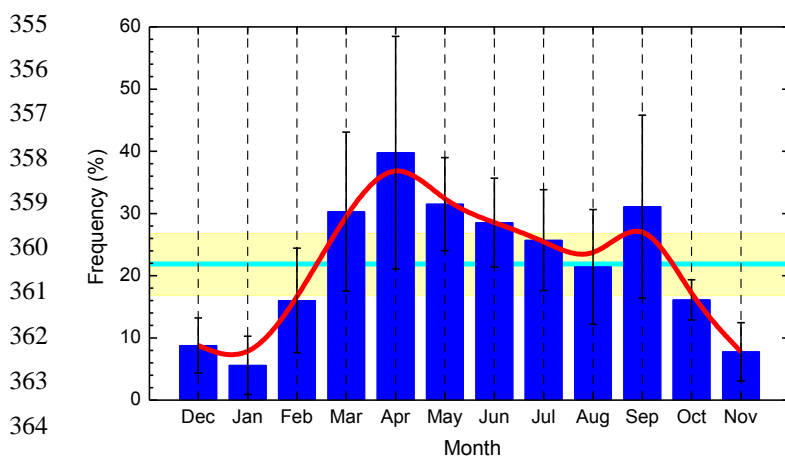
Environment	Background		Centre			
	Time interval	2012–2013	2008–2009	2013–2014	2014–2015	2015–2016
Event days	96	83	72	81	35	83
Quantifiable days	43	31	48	56	18	52
Undefined days	19	34	24	25	8	23
Non-event days	231	229	267	240	226	257
Coverage (%)	95	95	99	95	73	99
Missing days	20	19	2	19	97	2

336

337

338 It was previously shown that the NPF and growth events observed in the city centre and its  
 339 background ordinarily happen above a larger territory or region (Salma et al., 2011; Németh  
 340 and Salma, 2014), and they are linked to each other (Salma et al., 2016b). From the point of the  
 341 occurrence frequency distribution, they could, therefore, be evaluated jointly in the first  
 342 approximation. An overall monthly mean relative occurrence frequency of nucleation days  
 343 derived for all 6 measurement years is shown in Fig. 1. The annual mean frequency with SD  
 344 was  $22 \pm 5\%$ , which is considerable and is in line with other urban sites (Sect. 1). The monthly  
 345 mean frequency has a temporal variation, which can be characterised by a noteworthy pattern.  
 346 The mean monthly dependency exhibits an absolute and a local minimum in January (5.6%)  
 347 and August (21%), respectively, and an absolute and a local maximum in April (40%), and  
 348 September (31%), respectively. Nevertheless, the SDs of the monthly (and annual) means  
 349 indicate prominent variability from year to year. The pattern can be related to multivariate  
 350 relationships and complex interplay among the influencing factors, which include the air  
 351 temperature (January is the coldest month, while August is the warmest month in the Carpathian  
 352 Basin) and enhanced emission of biogenic VOCs in springtime (March–April) and early autumn  
 353 (September) as well (Salma et al., 2016b).

354



**Figure 1.** Monthly mean relative occurrence frequency of new aerosol particle formation and consecutive particle diameter growth events with respect to the number of relevant days for the joint 6-year long data set. The error bars show  $\pm 1$  standard deviation, the horizontal line in cyan indicates the overall annual mean frequency, the yellow bands represent  $\pm 1$  standard deviation of the annual mean, and the smooth curve in red serves to guide the eye.

The properties and variables studied were derived in full time resolution. They were averaged in several ways for different conditions and for various purposes to obtain typical average descriptive characteristics. In 1 case (31–08–2016), the NPF and growth event could reliably be identified, while the measured absolute particle number concentrations could not be validated due to some experimental troubles, and, therefore, it was left out from the further calculations. Similarly, there were 1 and 4 NPF and growth events with unusually/extraordinarily large dynamic properties in the measurement years 2014–2015 and 2017–2018, respectively. More specifically, 5 individual  $J_6$  data when expressed in a unit of  $\text{cm}^{-3} \text{s}^{-1}$  and 1 individual  $\text{GR}_{10}$  data when given in  $\text{nm h}^{-1}$  were  $>20$  (Table 3). These extremes were left out from the overview statistics to maintain the representativity (they could be influenced by some unknown extra or very local sources) and to fulfil better the basic requirements of correlation analysis. If an event showed a double beginning then the dynamic properties for the first onset were considered in the basic overview since this onset is of regional relevance (Salma et al., 2016b). The extreme dynamic properties and the characteristics for the second onsets were, however, evaluated separately and are discussed in detail and interpreted in Sect. 4.4.

388

389 **4.1 Ranges and averages**

390

391 Ranges and averages with SDs of formation rate  $J_6$ , growth rate  $GR_{10}$ , starting time of  
392 nucleation ( $t_1$ ) and duration time interval of nucleation ( $\Delta t$ ) are summarised in Table 2 for  
393 separate measurement years and for the joint 5-year long city centre data set. In the city centre,  
394 nucleation generally starts at 09:15 UTC+1, and it is typically maintained for approximately 3  
395 h. The NPF and growth events ordinarily produce 5.6 new aerosol particles with a diameter of  
396 6 nm in 1 cm<sup>3</sup> of air in 1 s, and cause the particles with a diameter of 10 nm to grow with a  
397 typical rate of 7.6 nm h<sup>-1</sup>. The statistics for  $J_6$  and  $GR_{10}$  are based on 199 and 203 events,  
398 respectively. The corresponding data for the separate years show considerable variability  
399 without obvious trends or tendencies. The differences between the years can likely be related  
400 to changes in actual atmospheric chemical and physical situations and conditions, and to the  
401 resulting modifications in the sensitive balance and delicate coupling among them from a year  
402 to year. Spread of the individual data for  $GR_{10}$  is smaller than for  $J_6$ ; the relative SDs for the  
403 joint 5-year long city centre data set were 38% and 68%, respectively, while the (external)  
404 relative SDs calculated from the annual mean values were 4.2% and 14.0%, respectively. The  
405 dynamic properties and  $t_1$  data tend to be smaller in the near-city background than in the city  
406 centre. In general, nucleation starts 1 h earlier in the background, and the events typically show  
407 significantly smaller  $J_6$  (with a median of 2.0 cm<sup>-3</sup> s<sup>-1</sup>) and  $GR_{10}$  (with a median of 5.0 nm h<sup>-1</sup>).  
408 Duration of the nucleation is very similar to that in the city centre. All starting times of  
409 nucleation were larger than (in a few cases, very close to) the time of the sunrise. This implies  
410 that no nocturnal NPF and growth event has been identified in Budapest so far. The particle  
411 growth process (the so-called banana curve) could be traced usually for a longer time interval  
412 (up to 1.5 d) in the background than in the centre. These conclusions are in line with the ideas  
413 on atmospheric nucleation and consecutive particle growth process (e.g. Kulmala et al., 2014;  
414 Zhang et al., 2015), and they also confirm some our earlier findings with respect to Budapest  
415 and its regional background within the Carpathian Basin achieved with shorter, 2-year long data  
416 sets (Salma et al., 2016b).

417



418 **Table 2.** Ranges, averages and standard deviations of aerosol particle formation rate  $J_6$ , particle diameter  
 419 growth rate  $GR_{10}$ , starting time ( $t_1$ ) and duration time interval ( $\Delta t=t_2-t_1$ ) of nucleation process of  
 420 quantifiable (class 1A) new particle formation and growth events in the near-city background and city  
 421 centre separately for the 1-year long measurement time intervals and for the joint 5-year long city centre  
 422 data set.  
 423

Environment	Background		Centre				
	2012– 2013	2008– 2009	2013– 2014	2014– 2015	2015– 2016	2017– 2018	All 5 years
Formation rate $J_6$ ( $\text{cm}^{-3} \text{s}^{-1}$ )							
Minimum	0.48	1.47	1.13	0.81	1.19	1.60	0.81
Median	2.0	4.2	3.5	4.4	4.6	6.3	4.6
Maximum	5.6	15.9	17.8	18.0	15.3	17.3	18.0
Mean	2.2	4.7	5.2	5.6	5.0	6.6	5.6
St. deviation	1.3	2.6	3.7	4.2	3.7	3.3	3.8
Growth rate $GR_{10}$ ( $\text{nm h}^{-1}$ )							
Minimum	3.0	3.7	3.1	2.8	3.2	3.3	2.8
Median	5.0	7.6	6.6	6.5	8.0	7.5	7.3
Maximum	9.8	17.4	19.0	18.0	15.5	19.8	19.8
Mean	5.2	7.8	7.2	7.3	7.7	8.0	7.6
St. deviation	1.4	2.6	2.8	3.2	3.0	2.8	2.9
Starting time, $t_1$ (HH:mm UTC+1)							
Minimum	05:51	07:14	06:44	05:48	07:31	05:57	05:48
Median	08:19	09:26	09:22	08:48	09:45	09:18	09:15
Maximum	11:09	11:38	12:21	11:23	12:45	12:15	12:45
Mean	08:17	09:27	09:25	08:49	10:02	09:24	09:19
St. deviation	01:11	01:05	01:26	01:22	01:23	01:36	01:26
Duration time, $\Delta t$ (HH:mm)							
Minimum	01:23	00:52	00:42	00:31	01:03	01:26	00:31
Median	03:16	02:36	02:04	03:53	02:31	03:49	02:57
Maximum	06:44	06:04	05:34	07:46	06:05	07:55	07:55
Mean	03:30	02:44	02:14	03:52	02:58	03:57	03:18
St. deviation	01:40	01:11	01:01	01:40	01:47	01:39	01:40

424  
 425 Ranges and averages with SDs of some related atmospheric properties, namely of mean CS  
 426 averaged for the time interval from  $t_1$  to  $t_2$ , daily maximum gas-phase  $\text{H}_2\text{SO}_4$  proxy, daily mean  
 427  $T$  and RH (Table S2), and of daily median concentrations of  $\text{SO}_2$  (as the major precursor of gas-  
 428 phase  $\text{H}_2\text{SO}_4$ ),  $\text{O}_3$  (as an indicator of photochemical activity),  $\text{NO}_x$  and CO gases (as indicators  
 429 of anthropogenic combustion activities and road vehicle emissions) (Table S3) were also  
 430 derived for quantifiable NPF and growth event days, and are further evaluated. The annual  
 431 mean CS values exhibited decreasing tendency in the city centre over the years (as can be



432 expected from the particle number concentrations as well). The individual values remained  
433 below approximately  $20 \times 10^{-3} \text{ s}^{-1}$ , which agrees well with the results of our earlier study (Salma  
434 et al., 2016b) according to which the CS suppresses NPF above this level in the Carpathian  
435 Basin. Maximum  $\text{H}_2\text{SO}_4$  proxy values reached substantially higher levels (by a factor of  
436 approximately 2) in the near-city background than in the city centre due mainly to the  
437 differences in the CS and also  $[\text{SO}_2]$ . The differences between the 2 sets of proxy data (between  
438 the 2 sites) are particularly evident when considering their smallest values. The largest  
439 variability in the annual average values were observed for the proxy. A maximum concentration  
440 of  $\text{H}_2\text{SO}_4$  molecules of  $10^6$ – $10^7 \text{ cm}^{-3}$  were estimated by adopting the scaling factor (Sect. 3.2).  
441 The dynamic properties seem to be not very sensitive to air  $T$ ; which displayed quite similar  
442 and comparable values over the years and at both sites. This conclusion is likely linked to  
443 several temperature-dependent environmental processes simultaneously present in the area  
444 which balance the effect of  $T$ . Some events happened at daily mean temperatures below zero.  
445 The daily mean RH and its SD for the city centre and near-city background were  $54 \pm 11\%$  and  
446  $64 \pm 12\%$ , respectively. There were events that occurred at RHs as high as 90%. Relationships  
447 of the dynamic properties with  $T$  and RH are also obscured with strong seasonal cycle of these  
448 meteorological data and with the fact that air masses arriving to the receptor site in different  
449 trajectories are often characterised by the distinct levels of meteorological data. As far as the  
450 pollutant gases are concerned,  $\text{SO}_2$  showed somewhat smaller daily median values, and  $\text{O}_3$   
451 exhibited substantially smaller levels on event days in the city centre than in the near-city  
452 background, while concentrations of  $\text{NO}_x$  and CO were obviously larger in the city than in its  
453 close background. The differences can primarily be explained by the intensity and spatial  
454 distribution of their major sources and atmospheric chemical reactions, and the joined  
455 concentration data resembles typical situations without photochemical smog episodes in cities.  
456 There was no obvious decrease in  $\text{SO}_2$  concentration during these years in contrast with an  
457 earlier decreasing trend from mid-1980s till about 2000. No evident or sensitive effect of  
458 atmospheric gases on the dynamic or timing properties could be deduced from the averaged  
459 data. This can probably be explained by a dedicated balance between the intensifying and  
460 suppressing effects, which were averaged out on a yearly time scale. Relationships on shorter  
461 scales are further investigated and discussed in more detail in the following sections.

462

#### 463 4.2 Monthly distributions

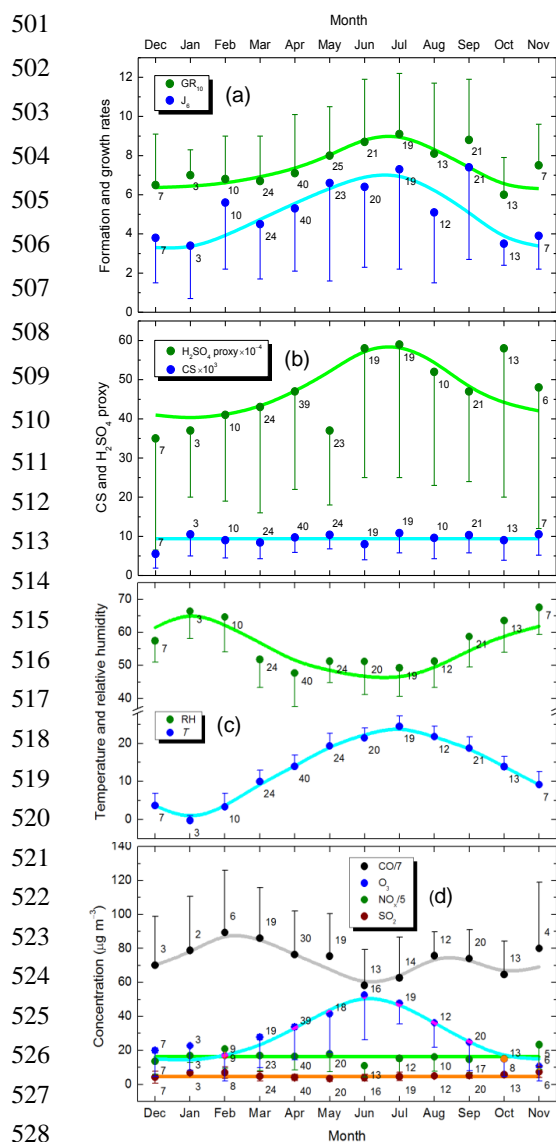
464

465 Distributions of the monthly mean  $J_6$ ,  $\text{GR}_{10}$ , daily maximum gas-phase  $\text{H}_2\text{SO}_4$  proxy, mean CS,  
466 daily mean air  $T$  and RH, and daily median  $\text{SO}_2$ ,  $\text{O}_3$ ,  $\text{NO}_x$  and CO concentrations for quantifiable





467 NPF and growth events for the joint 5-year long city centre data sets are shown in Fig. 2. The  
468 distributions – eminently for  $J_6$ ,  $GR_{10}$ ,  $H_2SO_4$  proxy and  $SO_2$  – do not follow the monthly pattern  
469 of the event occurrence frequency at all (cf. Fig. 1). Instead, the  $J_6$ ,  $GR_{10}$  and  $H_2SO_4$  proxy tend  
470 to exhibit larger values in summer months, and they temporal changes over the other months  
471 are smooth and do not show distinctive features. The elevations are substantial; the estimated  
472 maximum level was larger than the baseline by a factor of 2.1 for the  $J_6$ , and by a factor of  
473 approximately 1.4 for the  $GR_{10}$  and  $H_2SO_4$  proxy. The result is in line with other observations  
474 according to which the GR exhibits almost exclusively a summer maximum (Nieminen et al.,  
475 2018). The intensity of solar radiation at the surface, its seasonal cycling, concentration of  
476 atmospheric condensing vapours in different months, biogenic processes and the fact that rate  
477 coefficient of many thermal chemical/physicochemical processes in the nature (including GR,  
478 Paasonen et al., 2018) increases with  $T$  could play an important role in explained the  
479 distributions of the dynamic properties, although a more comprehensive study involving  
480 chemicals and their photochemistry is required for more detailed explanation. The properties  
481 are biased and influenced jointly by the intensity of the solar electromagnetic radiation.  
482 Nevertheless, the misalignment between the monthly distributions of occurrence frequency and  
483 all the other properties indicates that 1) the event occurrence or its basic causes are not linked  
484 with the dynamic properties in a straightforward or linear manner, and 2) gas-phase  $H_2SO_4$  does  
485 not seem to be the controlling factor of NPF occurrence in the Carpathian Basin including  
486 Budapest. The mean  $J_6$ ,  $GR_{10}$  and  $H_2SO_4$  proxy data still have considerable uncertainty, which  
487 makes their interpretation not yet completely conclusive. The uncertainties are influenced by  
488 the inherent fluctuations in the data sets, number of the individual data available for different  
489 properties and months, variations in other or unknown relevant environmental conditions, and  
490 by their variability from a year to year. The resulting uncertainties are expected to decrease with  
491 the length of the available data sets, which emphasized the need for continuation of the  
492 measurements. The monthly distributions of CS, and  $SO_2$  and  $NO_x$  concentrations could be  
493 represented by constant values of the overall means and SDs of  $(9.4 \pm 4.3) \times 10^{-3} \text{ s}^{-1}$ ,  $4.7 \pm 2.1 \mu\text{g}$   
494  $\text{m}^{-3}$  and  $81 \pm 38 \mu\text{g m}^{-3}$ , respectively with an acceptable accuracy. This suggests that CS,  $SO_2$   
495 and  $NO_x$  in Budapest do not critically or substantially affect either the dynamic properties (or  
496 the event occurrence). Monthly distributions of air  $T$  and  $O_3$  concentration showed a maximum  
497 over summer months, while RH reflected the  $T$  tendency. Distribution of CO was more  
498 changing and less certain than for the other gases, and, therefore, its interpretation is  
499 encumbered so far. However, it doesn't seem to obviously affect the dynamic properties.  
500



**Figure 2.** Distribution of monthly mean aerosol particle formation rate  $J_6$  in a unit of  $\text{cm}^{-3} \text{s}^{-1}$  and particle diameter growth rate  $\text{GR}_{10}$  in a unit of  $\text{nm h}^{-1}$  (a), mean condensation sink for vapours (CS) in a unit of  $\text{s}^{-1}$  averaged over the nucleation time interval ( $t_1, t_2$ ) and daily maximum gas-phase  $\text{H}_2\text{SO}_4$  proxy in a unit of  $\mu\text{g m}^{-5} \text{W s}$  (b), daily mean air temperature ( $T$ ) in a unit of  $^\circ\text{C}$  and daily mean relative humidity (RH) in % (c), and daily median concentrations of  $\text{SO}_2$ ,  $\text{O}_3$ ,  $\text{NO}_x$  and CO for quantifiable (class 1A) new particle formation and growth events in the city centre for the joint 5-year long time interval. The error bars are shown for one side for clarity, and indicate 1 standard deviation. Number of the individual data averaged in each month is displayed next to the symbols. The horizontal lines indicate the overall mean. The nonlinear curves assist to guide the eye.

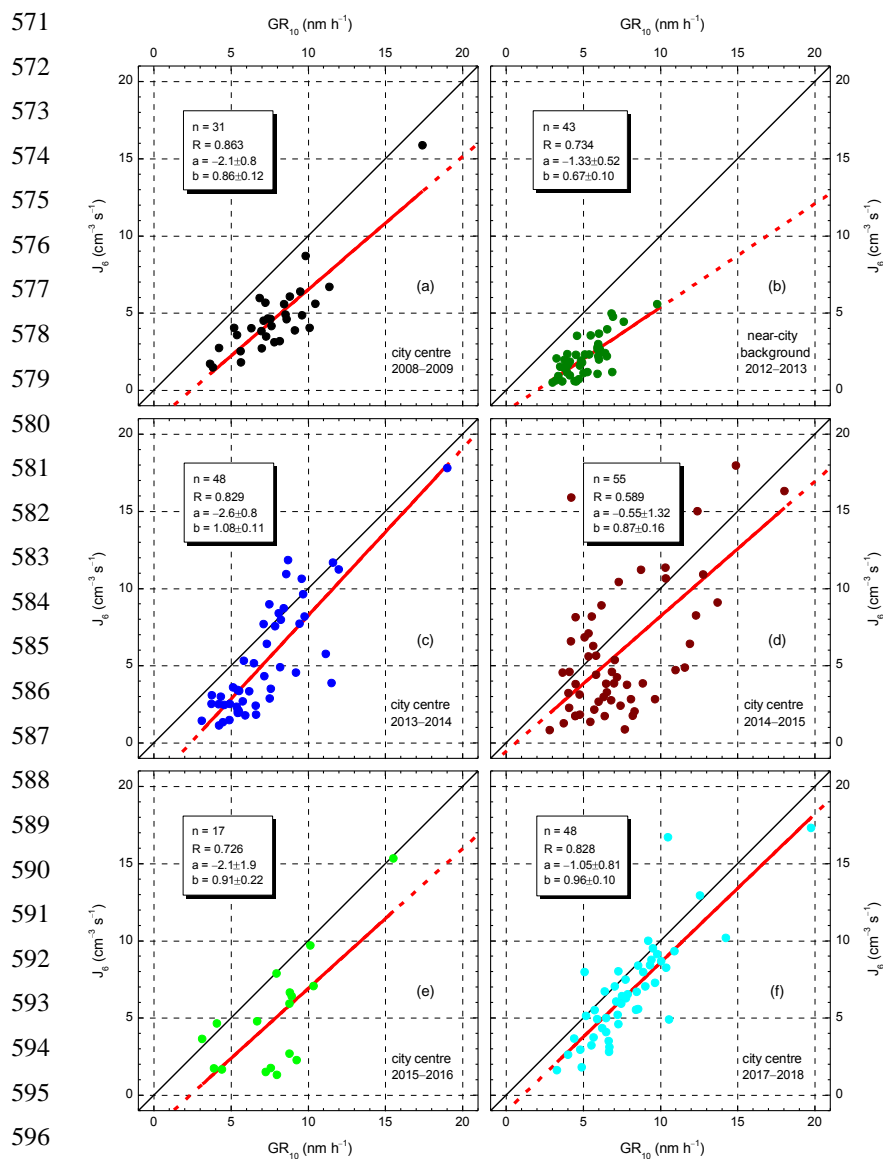


### 538 4.3 Relationships

539

540 Pearson's coefficients of correlation ( $R$ ) between  $J_6$  and  $GR_{10}$  revealed significant linear  
541 relationship between them for all annual data sets (the mean  $R$  and SD were  $0.768 \pm 0.099$ ,  
542 number of data pairs  $n=243$ ). It has to be noted that  $J_6$  and  $GR_{10}$  are not completely independent  
543 variables (see Eq. 1 and Table S1). Scatter plots between  $J_6$  and  $GR_{10}$  for the separate 1-year  
544 long measurement time intervals are shown in Fig. 3. For the city centre, the regression lines  
545 follow the line of equality in all 5 years. The mean slope ( $b$ ) with SD for the joint 5-year long  
546 city centre data set was  $b=0.94 \pm 0.07$  expressed formally in a unit of  $\text{cm}^{-3} \text{s}^{-1} \text{nm}^{-1} \text{h}$ . At the  
547 same time, the regression line for the near-city background deviated significantly from the  
548  $J_6=GR_{10}$  line, with a  $b=0.67 \pm 0.10 \text{ cm}^{-3} \text{ s}^{-1} \text{ nm}^{-1} \text{ h}$ . This implies that the relevant chemical  
549 species and/or their processes in the air of the city centre contribute equally to the formation of  
550 6-nm particles and to their growth process. In the near-city background, however, the chemical  
551 compounds available and their processes power particle growth more than new particle  
552 formation. This may be related to the differences between the city and its close environments  
553 as far as the atmospheric composition, chemistry and physics, and other delicate conditions are  
554 concerned (Paasonen et al., 2018). The narrower range and smaller number of individual  
555 dynamic properties available in the near-city background relative to those in the city centre may  
556 represent some weakness in the explanation. The intercepts ( $a$ ) of the regression lines were,  
557 however, identical for all data sets within their uncertainty interval. The mean intercept and SD  
558 were estimated to be  $-1.7 \pm 0.8 \text{ cm}^{-3} \text{ s}^{-1}$ . This finding is interpreted as the existence of a  
559 minimum GR or more exactly of a minimally required GR that corresponds to  $J_6=0$ . Particles  
560 that exhibit at least this level of GR can escape coagulation mainly with larger particles, and  
561 reach the diameter of 6 nm detectable in the present study by condensational growth. The  
562 minimal GR was derived as  $GR_{\min} = -a/b$ , and its mean and SD are  $1.8 \pm 1.0 \text{ nm h}^{-1}$  for the  
563 conditions ordinarily present in the Budapest air. Nucleation processes which are initiated under  
564 circumstances that cause the newly formed particle with a diameter of 10 nm to grow with a  
565 rate  $< GR_{\min}$  are normally not observed. Anyway, these are expected to be weak phenomena due  
566 to the relationship between  $GR_{10}$  and  $J_6$ , while the events close to it can also be masked by  
567 fluctuating experimental data. Their identification and evaluation can be made feasible by  
568 decreasing the lower measurement diameter limit of DMPS systems down to 3 nm, or by  
569 different instruments such as particle size magnifier or neutral cluster and air ions spectrometer.

570



598 **Figure 3.** Scatter plots for aerosol particle formation rate  $J_6$  and consecutive particle diameter growth  
 599 rate  $GR_{10}$  in city centre (a and c–f) and near-city background (b) separately for the 1-year long  
 600 measurement time intervals. Number of data point ( $n$ ), their coefficient of correlation ( $R$ ) and the  
 601 intercept ( $a$ ) and slope ( $b$ ) of the regression line with standard deviations are also indicated. The lines in  
 602 black represent the line of equality  $J_6=GR_{10}$ , the solid lines in red show the regression lines, while the  
 603 dashed parts in red are extrapolated from the regression line.

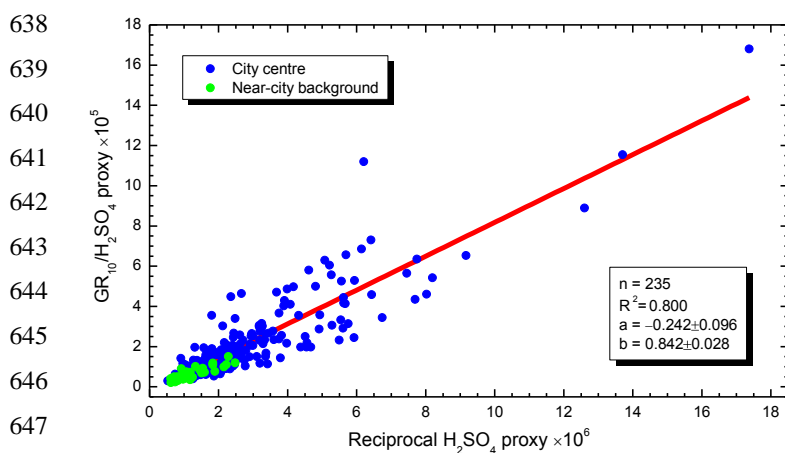


604 Correlations between individual H<sub>2</sub>SO<sub>4</sub> proxy values on one side and J<sub>6</sub> or GR<sub>10</sub> on the other  
605 side were not significant. This is consistent with the corresponding conclusion of Sect. 4.2 and  
606 with the earlier results according to which the mean contribution of H<sub>2</sub>SO<sub>4</sub> condensation to the  
607 particle GR was only 12.3% in Budapest (Salma et al., 2016b). The lack of correlation  
608 repeatedly suggests that other chemical species have larger influence on the formation and  
609 growth of new particles. Coefficients of correlation between CS on one side and J<sub>6</sub> or GR<sub>10</sub> on  
610 the other side for the joint city centre data sets were modest ( $R=0.41$  and  $0.32$ , respectively with  
611  $n=194$  and  $197$ , respectively). This is simply related to the fact that larger GR values are typical  
612 for polluted urban air (Kulmala et al., 2017) since particles capable of escaping coagulation  
613 scavenging need to grow faster in comparison to cleaner environments, and the enhanced  
614 requirements for the growth are linked to increased formation rates as well. As far as the  
615 pollutant gases are concerned, no correlation could be identified between J<sub>6</sub> or GR<sub>10</sub> on one side  
616 and the gas concentrations on the other side. The coefficients of correlation between CS and  
617 NO<sub>x</sub> or CO were modest ( $R=0.37$  and  $0.42$ , respectively with  $n=164$  and  $152$ , respectively),  
618 while correlation of NO<sub>x</sub> and CO on one side with WS was also modest but negative ( $R=-0.32$   
619 and  $-0.42$ , respectively with  $n=167$  and  $155$ , respectively). The former relationships can be  
620 explained by the fact that vehicular road traffic in cities is a considerable common sources of  
621 NO<sub>x</sub>, CO and primary particles (Paasonen et al., 2016), and the emitted particles largely  
622 contribute to CS levels, while the latter relationships are linked to the effect of large-scale air  
623 mass transport (often connected to high WSs) on urban air pollution or air quality.

624

625 Importance and contribution of condensing vapours other than H<sub>2</sub>SO<sub>4</sub> are further demonstrated  
626 in Fig. 4. The data for the city centre and near-city background were presented in a linearized  
627 form and separately for the 2 sites. Nevertheless, the fitting of the correlation line was  
628 accomplished for the joint 6-year long data set. It can be demonstrated in particular on non-  
629 linearized plot of the GR<sub>10</sub>/H<sub>2</sub>SO<sub>4</sub> proxy as function of H<sub>2</sub>SO<sub>4</sub> proxy (not shown here) that the  
630 2 data sets merge into each other without any relevant structure, and, therefore, that they can be  
631 regarded to be coherent. This approach seems sensible when considering also the limited  
632 accuracy of the values. The relationship between the 2 composite variables in Fig. 4 was  
633 significant ( $R^2=0.800$ ,  $p<0.05$ ). It can be interpreted as the increasing atmospheric concentration  
634 of gas-phase H<sub>2</sub>SO<sub>4</sub> can be related to larger contributions of other vapours than H<sub>2</sub>SO<sub>4</sub> to  
635 particle growth. The other or competing compounds may include oxidation products and their  
636 dimers from photooxidation of VOC precursors from both biogenic and anthropogenic sources.

637



649 **Figure 4.** Dependence of the growth rate  $GR_{10}$  (in a unit of  $\text{nm h}^{-1}$ ) of new particle formation and growth  
650 events normalised to the daily maximum gas-phase  $\text{H}_2\text{SO}_4$  proxy ( $\mu\text{g m}^{-5} \text{W s}$ ) on the reciprocal proxy  
651 value in the city centre and near-city background. The linear line in red represents the line fitted to the  
652 joint data set. Number of individual data considered ( $n$ ), their coefficient of determination ( $R^2$ ) and the  
653 intercept ( $a$ ) and slope ( $b$ ) of the fitted regression line with standard deviations are also shown.

654

#### 655 4.4 Extreme and multiple events

656

657 The joint 6-year long data sets of  $J_6$ ,  $GR_{10}$  and  $\Delta t$  containing all, 247 individual data each could  
658 be characterised by lognormal distribution function. This is demonstrated by log-probability  
659 graph for  $J_6$  in Fig. S2 as example. The coefficient of determination, median and geometric  
660 standard deviation for  $J_6$ ,  $GR_{10}$  and  $\Delta t$  data sets were 0.990,  $4.0 \text{ cm}^{-3}$  and 2.3; 0.993,  $6.8 \text{ nm h}^{-1}$   
661 and 1.46; and 0.998, 02:57 (0.123 d) and 1.74, respectively. One of the major properties of  
662 this distribution type is that it contains relatively large individual data with considerably high  
663 abundances. There were 5 individual  $J_6$  and 5 individual  $GR_{10}$  data above the 98% percentile  
664 of the data sets, which belonged to 9 separate NPF and growth events (days). Their  
665 specifications, properties and parameters are summarised in Table 3. All these events occurred  
666 in the city centre from April to September. Their number in the separate consecutive  
667 measurement years (Sect. 2) were 1, 0, 1, 2, 0 and 5, respectively. The medians of  $J_6$ ,  $GR_{10}$ , CS  
668 and air  $T$  for the subsets of these 9 extreme event days were larger by factors of 5.2, 2.4, 1.5  
669 and 1.4, respectively than for the city centre data, while the medians of the other properties and  
670 concentrations in these 2 respective data sets agreed within approximately 10%. There was a  
671 single event associated with an extreme  $\text{H}_2\text{SO}_4$  proxy (of  $23 \times 10^{-5} \mu\text{g m}^{-5} \text{W s}$ ) and relatively



672 low  $\text{NO}_x$  concentration ( $44 \mu\text{g m}^{-3}$ ), which indicate exceptionally favourable conditions for  
 673 NPF and growth. In addition to this case, there were only a few (1–2) days that were  
 674 characterised by an unusually large CS ( $23 \times 10^{-3} \text{ s}^{-1}$ ) – which could in turn be linked to higher  
 675 dynamic rates (Sect. 4.3) – or by somewhat larger  $\text{SO}_2$  ( $8.1 \mu\text{g m}^{-3}$ ) or lower  $\text{NO}_x$  concentration  
 676 ( $34 \mu\text{g m}^{-3}$ ). For all the other events, however, no simple or compound property of the  
 677 investigated variables could explain the extreme rates. Instead, they may be related to some  
 678 other chemical species and/or atmospheric processes, which were not including in the present  
 679 study. Since the extreme NPF and growth events usually resembled the time evolution for  
 680 regional events (well developed banana curves) – sometimes with multiple onsets –, the missing  
 681 atmospheric players in increased concentrations or their relevant processes are expected to  
 682 appear on a larger horizontal spatial scale.

683

684 **Table 3.** Date (in a format of dd–MM–yyyy), new particle formation rate  $J_6$  (in a unit of  $\text{cm}^{-3} \text{ s}^{-1}$ ),  
 685 particle diameter growth rate  $\text{GR}_{10}$  ( $\text{nm h}^{-1}$ ), starting time  $t_1$  of nucleation (HH:mm UTC+1), duration  
 686 time interval  $\Delta t = t_2 - t_1$  of nucleation (HH:mm), mean condensation sink CS during the nucleation process  
 687 ( $10^{-3} \text{ s}^{-1}$ ), daily maximum gas-phase  $\text{H}_2\text{SO}_4$  proxy ( $10^4 \mu\text{g m}^{-5} \text{ W s}$ ), daily mean air temperature  $T$  ( $^\circ\text{C}$ ),  
 688 daily mean relative humidity RH (%), daily median concentrations of  $\text{SO}_2$ ,  $\text{O}_3$ ,  $\text{NO}_x$  ( $\mu\text{g m}^{-3}$ ) and CO  
 689 ( $\text{mg m}^{-3}$ ) gases, and the type of the onset for extreme quantifiable (class 1A) new particle formation and  
 690 growth events. The cells in yellow indicate the values which are above the 98% percentile of the  
 691 corresponding data sets. N.a.: not available.

692

Date/ property	15– 09– 2009	20– 04– 2014	19– 05– 2015	04– 07– 2015	28– 05– 2017	25– 06– 2017	02– 08– 2017	31– 08– 2017	09– 09– 2017
$J_6$	15.9	17.8	24	16.3	27	33	30	47	17.3
$\text{GR}_{10}$	17.4	19.0	12.2	18.0	9.2	17.0	11.8	21	19.8
$t_1$	10:20	08:52	08:52	09:38	06:34	10:18	07:39	10:06	11:38
$\Delta t$	01:23	01:42	03:57	02:06	07:15	02:46	06:58	06:19	02:06
Proxy	38	42	25	16	229	41	69	92	45
CS	13.4	8.9	13.7	11.9	6.9	10.5	23	18.2	15.5
$T$	20	13.0	22	26	20	24	29	23	19.1
RH	60	62	48	40	40	68	49	47	58
$\text{SO}_2$	6.1	2.5	4.4	2.3	3.4	3.1	5.6	8.1	6.6
$\text{O}_3$	16.3	43	n.a.	33	61	56	34	24	12.9
$\text{NO}_x$	69	34	174	70	44	66	n.a.	109	112
CO	0.42	n.a.	0.71	0.33	0.31	0.50	0.97	0.62	0.71
Onset	ordinary	double	broad	ordinary	broad	broad	broad	broad	ordinary

693





694 Each quantifiable NPF and growth event was labelled as ordinary or broad by visual inspection  
695 of its beginning part. If the width of the beginning was smaller than approximately 2 h or there  
696 was a determinant single growth curve (rib) on the size distribution surface plot then the onset  
697 was labelled as ordinary, otherwise as broad (Fig. S1a and S1c for broad onsets). Broad onsets  
698 can be generated by 1) long-lasting nucleation process, 2) disrupted and started over nucleation  
699 due to changing atmospheric and meteorological conditions or 3) multiple nucleation processes  
700 close to each other in time (Salma et al., 2016b). The broad onsets were specified as doublets  
701 if the nucleation mode could be separated into 2 submodes by size distribution fitting.  
702 Approximately 40% of all quantifiable events had a broad onset. This indicates that NPF and  
703 growth events with broad/multiple onsets are abundant in the urban environment. This could be  
704 an important difference from remote or clean atmospheres. For 9% of all cases, it was feasible  
705 to calculate 2 sets of dynamic properties for onsets 1 and 2, respectively with a reasonable  
706 accuracy. In the near-city background, the medians of  $J_6$  and  $GR_{10}$  for the onset 1 were similar  
707 or somewhat smaller than the corresponding medians for the whole near-city background data  
708 set, while for the onset 2, they were substantially larger, namely  $4.1 \text{ cm}^{-3} \text{ s}^{-1}$  and  $10.0 \text{ nm h}^{-1}$ ,  
709 respectively (cf. Table 2). Actually, the latter values were closer to the medians of the city centre  
710 than for the near-city background. The dynamic properties for the city centre for both the onset  
711 1 and onset 2 were somewhat larger than for the whole the city centre data set. Approximately  
712 75% of the doublets resulted in individual onset2/onset1 ratios larger than unity. Their overall  
713 median ratios for  $J_6$  and  $GR_{10}$  were similar and approximately 1.4, while for the near-city  
714 background, they were about 2. The results are in line with and confirm our earlier conclusion  
715 according to which the second onsets (if it is a new formation process and not just a started over  
716 event) often generate new particles more intensively than the first onsets (Salma et al., 2016b).  
717 These particles also grow faster. This can be explained by the fact that the first event is often  
718 or likely of regional scale since its dynamic properties resemble those of the regional  
719 background process (Yli-Juuti et al., 2009), while the later event can be characterised by values  
720 typical for the city centre (Salma et al., 2016b). The later event (or events) are mainly caused  
721 and governed by sub-regional processes. These findings are also coherent with a previous  
722 observation of NPF and growth events with multiple onsets in semi-clean savannah and  
723 industrial environments (Hirsikko et al., 2013), and they also fit well into the existing ideas on  
724 mixing regional and urban air parcels that exhibit different properties such as precursor  
725 concentrations,  $T$  and  $RH$  (Kulmala et al., 2017).

726

727 **5 Conclusions**

728

729 Dynamic and timing properties of 247 NPF and growth events were studied together with  
730 supporting aerosol properties, meteorological data and pollutant gas concentrations in near-city  
731 background and city centre of Budapest for 6 years. The results and conclusions derived here  
732 can form an important component that is based on atmospheric observations, and is to be  
733 combined with results from laboratory experiments and finally, with theoretical models to further  
734 improve our understanding on the NPF potentials to increase UF and CCN concentrations on  
735 various spatial and temporal scales. Magnitude of the particle number concentration level  
736 produced solely by NPF and growth (strength of the events) can roughly be estimated by  
737 considering the median  $J_6$ , median duration of the nucleation process (Table 2) and the typical  
738 coagulation loss of these particles (0.17; Sect. 3.1 and Table S1). In central Budapest, it yields  
739 a concentration of  $10^4 \text{ cm}^{-3}$ , which is comparable to the annual median concentrations (Sect.  
740 4). This simple example indicates that the phenomenon is not only relevant for aerosol load and  
741 climate issues on larger/regional or global spatial scales (which were first recognised) but it can  
742 affect the urban climate and the health risk of inhabitants of cities as well. This recognition led  
743 to emergence of the urban studies. At present, there is an evident and strong need for continuous  
744 urban NPF and growth studies on long term at fixed locations including relevant precursors and  
745 potential chemical participants of the phenomenon. They can also contribute to our general  
746 understanding of the nucleation phenomenon when their specialities and peculiarities are  
747 resolved and taken into account when dealing with its various implications.

748

749 The present research based on ambient atmospheric measurements provided several evidence  
750 that some important chemical players in the NPF and growth events are still missing.  
751 Considering the results and conclusions of cloud chamber experiments, these factors are  
752 expected to be related mainly to oxidation products of VOCs and/or their processes. Further  
753 dedicated research including sophisticated measurements, data evaluations and modelling  
754 studies is required to find and identify these species and their processes, and to account their  
755 multifactorial role in more detail. Such measurement campaign focusing on chemical  
756 composition of molecular clusters, precursors and nucleating vapours by applying recent  
757 expedient instruments in Budapest over the months of the highest expected event occurrence  
758 has been just realised within a frame of an international cooperation. Its perspective results can  
759 hopefully provide additional and valuable information for some of the conclusion based on  
760 indirect evidence so far, and can further clarify the overall picture on urban multicomponent  
761 nucleation and growth phenomenon.



762 **Data availability.** The observational data used in this paper are available on request from the  
763 corresponding author or at the website of the Budapest platform for Aerosol Research and Training  
764 (<http://salma.web.elte.hu/BpART>).

765 **Competing interest.** The authors declare that they have no conflict of interest.

766 **Acknowledgements.** The authors thank Markku Kulmala and his research team at the University of  
767 Helsinki for their cooperation. Financial support by the National Research, Development and Innovation  
768 Office, Hungary (contracts K116788 and PD124283); by the European Regional Development Fund  
769 and the Hungarian Government (GINOP-2.3.2-15-2016-00028) is gratefully acknowledged.

770

#### 771 **References**

- 772 Alam, A., Shi, J. P., and Harrison, R. M.: Observations of new particle formation in urban air, *J.*  
773 *Geophys. Res.*, 108 (D3), 4093, doi:10.1029/2001JD001417, 2003.
- 774 Almeida, J., Schobesberger, S., Kurten, A., Ortega, I. K., Kupiainen-Maatta, O., Praplan, A. P.,  
775 Adamov, A., Amorim, A., Bianchi, F., Breitenlechner, M., David, A., Dommen, J., Donahue, N.  
776 M., Downard, A., Dunne, E., Duplissy, J., Ehrhart, S., Flagan, R. C., Franchin, A., Guida, R.,  
777 Hakala, J., Hansel, A., Heinritzi, M., Henschel, H., Jokinen, T., Junninen, H., Kajos, M.,  
778 Kangasluoma, J., Keskinen, H., Kupc, A., Kurten, T., Kvashin, A. N., Laaksonen, A., Lehtipalo,  
779 K., Leiminger, M., Leppa, J., Loukonen, V., Makhmutov, V., Mathot, S., McGrath, M. J.,  
780 Nieminen, T., Olenius, T., Onnela, A., Petäjä, T., Riccobono, F., Riipinen, I., Rissanen, M., Rondo,  
781 L., Ruuskanen, T., Santos, F. D., Sarnela, N., Schallhart, S., Schnitzhofer, R., Seinfeld, J. H.,  
782 Simon, M., Sipilä, M., Stozhkov, Y., Stratmann, F., Tome, A., Tröstl, J., Tsagkogeorgas, G.,  
783 Vaattovaara, P., Viisanen, Y., Virtanen, A., Vrtala, A., Wagner, P. E., Weingartner, E., Wex, H.,  
784 Williamson, C., Wimmer, D., Ye, P. L., Yli-Juuti, T., Carslaw, K. S., Kulmala, M., Curtius, J.,  
785 Baltensperger, U., Worsnop, D. R., Vehkamäki, H., and Kirkby, J.: Molecular understanding of  
786 sulphuric acid–amine particle nucleation in the atmosphere, *Nature*, 502, 359–363, 2013.
- 787 Baltensperger, U., Streit, N., Weingartner, E., Nyeki, S., Prévôt, A. S. H., Van Dingenen, R., Virkkula,  
788 A., Putaud, J. P., Even, A., Brink, H., Blatter, A., Neftel, A., and Gaggeler, H. W.: Urban and rural  
789 aerosol characterization of summer smog events during the PIPAPO field campaign in Milan, Italy,  
790 *J. Geophys. Res.*, 107(D22), 8193, doi:10.1029/2001JD001292, 2002.
- 791 Bianchi, F., Tröstl, J., Junninen, H., Frege, C., Henne, S., Hoyle, C. R., Molteni, U., Herrmann, E.,  
792 Adamov, A., Bukowiecki, N., Chen, X., Duplissy, J., Gysel, M., Hutterli, M., Kangasluoma, J.,  
793 Kontkanen, J., Kürten, A., Manninen, H. E., Münch, S., Peräkylä, O., Petäjä, T., Rondo, L.,  
794 Williamson, C., Weingartner, E., Curtius, J., Worsnop, D. R., Kulmala, M., Dommen, J., and  
795 Baltensperger, U.: New particle formation in the free troposphere: A question of chemistry and  
796 timing, *Science*, 352, 1109–1112, <https://doi.org/10.1126/science.aad5456>, 2016.
- 797 Braakhuys, H. M., Park, M. V., Gosens, I., De Jong, W. H., and Cassee, F. R.: Physicochemical  
798 characteristics of nanomaterials that affect pulmonary inflammation, *Part. Fibre Toxicol.*, 11:18,  
799 doi: 10.1186/1743-8977-11-18, 2014.



- 800 Cai, R. and Jiang, J.: A new balance formula to estimate new particle formation rate: reevaluating the  
801 effect of coagulation scavenging, *Atmos. Chem. Phys.*, 17, 12659–12675, 2017.
- 802 Carslaw, K. S., Lee, L. A., Reddington, C. L., Pringle, K. J., Rap, A., Forster, P. M., Mann, G. W.,  
803 Spracklen, D. V., Woodhouse, M. T., Regayre, L. A., and Pierce, J. R.: Large contribution of  
804 natural aerosols to uncertainty in indirect forcing, *Nature*, 503, 67–71, 2013.
- 805 Crounse, J. D., Nielsen, L. B., Jørgensen, S., Kjaergaard, H. G., and Wennberg, P. O.: Autoxidation of  
806 organic compounds in the atmosphere, *J. Phys. Chem. Lett.*, 4, 20, 3513–3520, 2013.
- 807 Dal Maso, M., Kulmala, M., Lehtinen, K. E. J., Mäkelä, J. M., Aalto, P. P., and O’Dowd, C.:  
808 Condensation and coagulation sinks and formation of nucleation mode particles in coastal and  
809 boreal forest boundary layers, *J. Geophys. Res.*, 107(19D), 8097, 10.1029/2001jd001053, 2002.
- 810 Dal Maso, M., Kulmala, M., Riipinen, I., Wagner, R., Hussein, T., Aalto, P. P., and Lehtinen, K. E. J.:  
811 Formation and growth of fresh atmospheric aerosols: eight years of aerosol size distribution data  
812 from SMEAR II, Hyytiälä, Finland, *Boreal Environ. Res.*, 10, 323–336, 2005.
- 813 Dall’Osto, M., Querol, X., Alastuey, A., O’Dowd, C., Harrison, R. M., Wenger, J., and Gómez-  
814 Moreno, F. J.: On the spatial distribution and evolution of ultrafine particles in Barcelona, *Atmos.*  
815 *Chem. Phys.*, 13, 741–759, 2013.
- 816 Ehn, M., Thornton, J. A., Kleist, E., Sipilä, M., Junninen, H., Pullinen, I., Springer, M., Rubach, F.,  
817 Tillmann, R., Lee, B., Lopez-Hilfiker, F., Andres, S., Acir, I. H., Rissanen, M., Jokinen, T.,  
818 Schobesberger, S., Kangasluoma, J., Kontkanen, J., Nieminen, T., Kurten, T., Nielsen, L. B.,  
819 Jørgensen, S., Kjaergaard, H. G., Canagaratna, M., Dal Maso, M., Berndt, T., Petäjä, T., Wahner,  
820 A., Kerminen, V. M., Kulmala, M., Worsnop, D. R., Wildt, J., and Mentel, T. F.: A large source of  
821 low-volatility secondary organic aerosol, *Nature*, 506, 476–479, 2014.
- 822 Gordon, H., Sengupta, K., Rap, A., Duplissy, J., Frege, C., Williamson, C., Heinritzi, M., Simon, M.,  
823 Yan, C., Almeida, J., Tröstl, J., Nieminen, T., Ortega, I. K., Wagner, R., Dunne, E. M., Adamov,  
824 A., Amorim, A., Bernhammer, A. K., Bianchi, F., Breitenlechner, M., Brilke, S., Chen, X., Craven,  
825 J. S., Dias, A., Ehrhart, S., Fischer, L., Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Hakala, J.,  
826 Hoyle, C. R., Jokinen, T., Junninen, H., Kangasluoma, J., Kim, J., Kirkby, J., Krapf, M., Kürten,  
827 A., Laaksonen, A., Lehtipalo, K., Makhmutov, V., Mathot, S., Molteni, U., Monks, S. A., Onnela,  
828 A., Peräkylä, O., Piel, F., Petäjä, T., Praplan, A. P., Pringle, K. J., Richards, N. A. D., Rissanen, M.  
829 P., Rondo, L., Sarnela, N., Schobesberger, S., Scott, C. E., Seinfeld, J. H., Sharma, S., Sipilä, M.,  
830 Steiner, G., Stozhkov, Y., Stratmann, F., Tomé, A., Virtanen, A., Vogel, A. L., Wagner, A. C.,  
831 Wagner, P. E., Weingartner, E., Wimmer, D., Winkler, P. M., Ye, P., Zhang, X., Hansel, A.,  
832 Dommen, J., Donahue, N. M., Worsnop, D. R., Baltensperger, U., Kulmala, M., Curtius, J., and  
833 Carslaw, K. S.: Reduced anthropogenic aerosol radiative forcing caused by biogenic new particle  
834 formation, *Proc. Natl. Acad. Sci. U.S.A.*, 113, 12053–12058,  
835 <https://doi.org/10.1073/pnas.1602360113>, 2016.
- 836 Hirsikko, A., Vakkari, V., Tiitta, P., Hatakka, J., Kerminen, V.-M., Sundström, A.-M., Beukes, J. P.,  
837 Manninen, H. E., Kulmala, M., and Laakso, L.: Multiple daytime nucleation events in semi-clean  
838 savannah and industrial environments in South Africa: analysis based on observations, *Atmos.*  
839 *Chem. Phys.*, 13, 5523–5532, 2013.



- 840 Hussein, T., Puustinen, A., Aalto, P. P., Mäkelä, J. M., Hämeri, K., and Kulmala, M.: Urban aerosol  
841 number size distributions, *Atmos. Chem. Phys.*, 4, 391–411, 2004.
- 842 Hussein, T., Martikainen, J., Junninen, H., Sogacheva, L., Wagner, R., Dal Maso, M., Riipinen, I.,  
843 Aalto, P. P., and Kulmala, M.: Observation of regional new particle formation in the urban  
844 atmosphere, *Tellus 60B*, 509–521, 2008.
- 845 Jokinen, T., Berndt, T., Makkonen, R., Kerminen, V.-M., Junninen, H., Paasonen, P., Stratmann, F.,  
846 Herrmann, H., Guenther, A. B., Worsnop, D. R., Kulmala, M., Ehn, M. and Sipilä, M.: Production  
847 of extremely low volatile organic compounds from biogenic emissions: Measured yields and  
848 atmospheric implications, *Proc. Natl. Acad. Sci. U.S.A.*, 112, 7123–7128, 2015.
- 849 Kerminen, V.-M., Paramonov, M., Anttila, T., Riipinen, I., Fountoukis, C., Korhonen, H., Asmi, E.,  
850 Laakso, L., Lihavainen, H., Swietlicki, E., Svenningsson, B., Asmi, A., Pandis, S. N., Kulmala, M.,  
851 and Petäjä, T.: Cloud condensation nuclei production associated with atmospheric nucleation: a  
852 synthesis based on existing literature and new results, *Atmos. Chem. Phys.*, 12, 12037–12059,  
853 2012.
- 854 Kirkby, J., Curtius, J., Almeida, J., Dunne, E., Duplissy, J., Ehrhart, S., Franchin, A., Gagné, S., Ickes,  
855 L., Kürten, A., Kupc, A., Metzger, A., Riccobono, F., Rondo, L., Schobesberger, S.,  
856 Tsagkogeorgas, G., Wimmer, D., Amorim, A., Bianchi, F., Breitenlechner, M., David, A.,  
857 Dommen, J., Downard, A., Ehn, M., Flagan, R. C., Haider, S., Hansel, A., Hauser, D., Jud, W.,  
858 Junninen, H., Kreissl, F., Kvashin, A., Laaksonen, A., Lehtipalo, K., Lima, J., Lovejoy, E. R.,  
859 Makhutov, V., Mathot, S., Mikkilä, J., Minginette, P., Mogo, S., Nieminen, T., Onnela, A., Pereira,  
860 A., Petäjä, T., Schnitzhofer, R., Seinfeld, J. H., Sipilä, M., Stozhkov, Y., Stratmann, F., Tome, A.,  
861 Vanhanen, J., Viisanen, Y., Vrtala, A., Wagner, P.E., Walther, H., Weingartner, E., Wex, H.,  
862 Winkler, P.M., Carslaw, K. S., Worsnop, D. R., Baltensperger, U., and Kulmala, M.: The role of  
863 sulfuric acid, ammonia and galactic cosmic rays in atmospheric aerosol nucleation, *Nature*, 476,  
864 429–433, 2011.
- 865 Kirkby, J., Duplissy, J., Sengupta, K., Frege, C., Gordon, H., Williamson, C., Heinritzi, M., Simon,  
866 M., Yan, C., Almeida, J., Tröstl, J., Nieminen, T., Ortega, I. K., Wagner, R., Adamov, A., Amorim,  
867 A., Bernhammer, A.-K., Bianchi, F., Breitenlechner, M., Brilke, S., Chen, X., Craven, J., Dias, A.,  
868 Ehrhart, S., Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Hakala, J., Hoyle, C. R., Jokinen, T.,  
869 Junninen, H., Kangasluoma, J., Kim, J., Krapf, M., Kürten, A., Laaksonen, A., Lehtipalo, K.,  
870 Makhmutov, V., Mathot, S., Molteni, U., Onnela, A., Peräkylä, O., Piel, F., Petäjä, T., Praplan, A.  
871 P., Pringle, K., Rap, A., Richards, N. A. D., Riipinen, I., Rissanen, M. P., Rondo, L., Sarnela, N.,  
872 Schobesberger, S., Scott, C. E., Seinfeld, J. H., Sipilä, M., Steiner, G., Stozhkov, Y., Stratmann, F.,  
873 Tomé, A., Virtanen, A., Vogel, A. L., Wagner, A., Wagner, P. E., Weingartner, E., Wimmer, D.,  
874 Winkler, P. M., Ye, P., Zhang, X., Hansel, A., Dommen, J., Donahue, N. M., Worsnop, D. R.,  
875 Baltensperger, U., Kulmala, M., Carslaw, K. S., and Curtius, J.: Ion-induced nucleation of pure  
876 biogenic particles, *Nature*, 533, 521–526, <https://doi.org/10.1038/nature17953>, 2016.
- 877 Kulmala, M., Dal Maso, M., Mäkelä, J. M., Pirjola, L., Väkevä, M., Aalto, P., Miikkulainen, P.,  
878 Hämeri, K., and O'Dowd, C. D.: On the formation, growth and composition of nucleation mode  
879 particles, *Tellus B53*, 479–490, 2001.



- 880 Kulmala, M., Vehkamäki, H., Petäjä, T., Dal Maso, M., Lauri, A., Kerminen, V.-M., Birmili, W., and  
881 McMurry, P.: Formation and growth rates of ultrafine atmospheric particles: a review of  
882 observations, *J. Aerosol Sci.*, 35, 143–176, 2004.
- 883 Kulmala, M., Petäjä, T., Nieminen, T., Sipilä, M., Manninen, H. E., Lehtipalo, K., Dal Maso, M.,  
884 Aalto, P. P., Junninen, H., Paasonen, P., Riipinen, I., Lehtinen, K. E. J., Laaksonen, A., and  
885 Kerminen, V.-M.: Measurement of the nucleation of atmospheric aerosol particles, *Nat. Protoc.*, 7,  
886 1651–1667, doi:10.1038/nprot.2012.091, 2012.
- 887 Kulmala, M., Kontkanen, J., Junninen, H., Lehtipalo, K., Manninen, H. E., Nieminen, T., Petäjä, T.,  
888 Sipilä, M., Schobesberger, S., Rantala, P., Franchin, A., Jokinen, T., Järvinen, E., Äijälä, M.,  
889 Kangasluoma, J., Hakala, J., Aalto, P.P., Paasonen, P., Mikkilä, J., Vanhanen, J., Aalto, J., Hakola,  
890 H., Makkonen, U., Ruuskanen, T., Mauldin, R. L. III, Duplissy, J., Vehkamäki, H., Bäck, J.,  
891 Kortelainen, A., Riipinen, I., Kurtén, T., Johnston, M. V., Smith, J. N., Ehn, M., Mentel, T. F.,  
892 Lehtinen, K. E. J., Laaksonen, A., Kerminen, V.-M., and Worsnop, D. R.: Direct observations of  
893 atmospheric aerosol nucleation, *Science*, 339, 943–946, 2013.
- 894 Kulmala, M., Petäjä, T., Ehn, M., Thornton, J., Sipilä, M., Worsnop, D. R., and Kerminen, V.-M.:  
895 Chemistry of atmospheric nucleation: On the recent advances on precursor characterization and  
896 atmospheric cluster composition in connection with atmospheric new particle formation, *Annu.*  
897 *Rev. Phys. Chem.*, 65, 21–37, 2014.
- 898 Kulmala, M., Kerminen, V. M., Petäjä, T., Ding, A. J., and Wang, L.: Atmospheric gas-to-particle  
899 conversion: why NPF events are observed in megacities, *Faraday Discuss.*,  
900 doi:10.1039/C6FD00257A, 2017.
- 901 Makkonen, R., Asmi, A., Korhonen, H., Kokkola, H., Järvenoja, S., Räisänen, P., Lehtinen, K. E. J.,  
902 Laaksonen, A., Kerminen, V.-M., Järvinen, H., Lohmann, U., Bennartz, R., Feichter, J., and  
903 Kulmala, M.: Sensitivity of aerosol concentrations and cloud properties to nucleation and  
904 secondary organic distribution in ECHAM5-HAM global circulation model, *Atmos. Chem. Phys.*,  
905 9, 1747–1766, 2009.
- 906 Makkonen, R., Asmi, A., Kerminen, V.-M., Boy, M., Arneth, A., Hari, P., and Kulmala, M.: Air  
907 pollution control and decreasing new particle formation lead to strong climate warming, *Atmos.*  
908 *Chem. Phys.*, 12, 1515–1524, 2012.
- 909 Merikanto, J., Spracklen, D. V., Mann, G. W., Pickering, S. J., and Carslaw, K. S.: Impact of  
910 nucleation on global CCN, *Atmos. Chem. Phys.*, 9, 8601–8616, 2009.
- 911 Metzger, A., Verheggen, B., Dommen, J., Duplissy, J., Prévôt, A. S. H., Weingartner, E., Riipinen, I.,  
912 Kulmala, M., Spracklen, D. V., Carslaw, K. S., and Baltensperger, U.: Evidence for the role of  
913 organics in aerosol particle formation under atmospheric conditions, *Proc. Natl. Acad. Sci. U. S.*  
914 *A.*, 107, 6646–6651, 2010.
- 915 Németh, Z. and Salma, I.: Spatial extension of nucleating air masses in the Carpathian Basin, *Atmos.*  
916 *Chem. Phys.*, 14, 8841–8848, 2014.
- 917 Németh, Z., Rosati, B., Ziková, N., Salma, I., Bozó, L., Dameto de España, C., Schwarz, J., Ždímal,  
918 V., and Wonaschütz, A.: Comparison of atmospheric new particle formation and growth events in  
919 three Central European cities, *Atmos. Environ.*, 178, 191–197, 2018.





- 920 Nieminen, T., Kerminen, V.-M., Petäjä, T., Aalto, P. P., Arshinov, M., Asmi, E., Baltensperger, U.,  
921 Beddows, D. C. S., Beukes, J. P., Collins, D., Ding, A., Harrison, R. M., Henzing, B., Hooda, R.,  
922 Hu, M., Hörrak, U., Kivekäs, N., Komsaare, K., Krejčí, R., Kristensson, A., Laakso, L., Laaksonen,  
923 A., Leaitch, W. R., Lihavainen, H., Mihalopoulos, N., Németh, Z., Nie, W., O'Dowd, C., Salma, I.,  
924 Sellegri, K., Svenningsson, B., Swietlicki, E., Tunved, P., Ulevicius, V., Vakkari, V., Vana, M.,  
925 Wiedensohler, A., Wu, Z., Virtanen, A., and Kulmala, M.: Global analysis of continental boundary  
926 layer new particle formation based on long-term measurements, *Atmos. Chem. Phys. Discuss.*,  
927 <https://doi.org/10.5194/acp-2018-304>, in review, 2018.
- 928 Oberdörster, G., Oberdörster, E., and Oberdörster, J.: Nanotoxicology: an emerging discipline  
929 evolving from studies of ultrafine particles, *Environ. Health Perspect.*, 113, 823–839, 2005.
- 930 O'Dowd, C. D., Jimenez, J. L., Bahreini, R., Flagan, R. C., Seinfeld, J. H., Hämeri, K., Pirjola, L.,  
931 Kulmala, M., Jennings, S. G., and Hoffmann, Th.: Marine aerosol formation from biogenic iodine  
932 emissions, *Nature* 417, 632–636, 2002.
- 933 Paasonen, P., Kupiainen, K., Klimont, Z., Visschedijk, A., Denier van der Gon, H. A. C., and Amann,  
934 M.: Continental anthropogenic primary particle number emissions, *Atmos. Chem. Phys.*, 16, 6823–  
935 6840, 2016.
- 936 Paasonen, P., Peltola, M., Kontkanen, J., Junninen, H., Kerminen, V.-M., and Kulmala, M.:  
937 Comprehensive analysis of particle growth rates from nucleation mode to cloud condensation  
938 nuclei in Boreal forest, *Atmos. Chem. Phys. Discuss.*, <https://doi.org/10.5194/acp-2018-169>, in  
939 review, 2018.
- 940 Petäjä, T., Mauldin, III, R. L., Kosciuch, E., McGrath, J., Nieminen, T., Paasonen, P., Boy, M.,  
941 Adamov, A., Kotiaho, T., and Kulmala, M.: Sulfuric acid and OH concentrations in a boreal forest  
942 site, *Atmos. Chem. Phys.*, 9, 7435–7448, 2009.
- 943 Putaud, J.-P., Van Dingenen, R., Alastuey, A., Bauer, H., Birmili, W., Cyrus, J., Flentje, H., Fuzzi, S.,  
944 Gehrig, R., Hansson, H. C., Harrison, R. M., Herrmann, H., Hitznerberger, R., Hüglin, C., Jones,  
945 A.M., Kasper-Giebl, A., Kiss, G., Kousa, A., Kuhlbusch, T. A. J., Löschau, G., Maenhaut, W.,  
946 Molnár, A., Moreno, T., Pekkanen, J., Perrino, C., Pitz, M., Puxbaum, H., Querol, X., Rodriguez,  
947 S., Salma, I., Schwarz, J., Smolík, J., Schneider, J., Spindler, G., ten Brink, H., Turšič, J., Viana,  
948 M., Wiedensohler, and A., Raes, F.: A European Aerosol Phenomenology - 3: physical and  
949 chemical characteristics of particulate matter from 60 rural, urban, and kerbside sites across  
950 Europe, *Atmos. Environ.*, 44, 1308–1320, 2010.
- 951 Riccobono, F., Schobesberger, S., Scott, C., Dommen, J., Ortega, I., Rondo, L., Almeida, J., Amorim,  
952 A., Bianchi, F., Breitenlechner, M., David, A., Downard, A., Dunne, E., Duplissy, J., Ehrhart, S.,  
953 Flagan, R., Franchin, A., Hansel, A., Junninen, H., Kajos, M., Keskinen, H., Kupc, A., Kurten, A.,  
954 Kvashin, A., Laaksonen, A., Lehtipalo, K., Makhmutov, V., Mathot, S., Nieminen, T., Onnela, A.,  
955 Petäjä, T., Praplan, A., Santos, F., Schallhart, S., Seinfeld, J., Sipila, M., Spracklen, D., Stozhkov,  
956 Y., Stratmann, F., Tome, A., Tsagkogeorgas, G., Vaattovaara, P., Viisanen, Y., Vrtala, A., Wagner,  
957 P., Weingartner, E., Wex, H., Wimmer, D., Carslaw, K., Curtius, J., Donahue, N., Kirkby, J.,  
958 Kulmala, M., Worsnop, D., and Baltensperger, U.: Oxidation products of biogenic emissions  
959 contribute to nucleation of atmospheric particles, *Science*, 344, 717–721, 2014.





- 960 Riipinen, I., Pierce, J. R., Yli-Juuti, T., Nieminen, T., Häkkinen, S., Ehn, M., Junninen, H., Lehtipalo,  
961 K., Petäjä, T., Slowik, J., Chang, R., Shantz, N. C., Abbatt, J., Leaitch, W. R., Kerminen, V.-M.,  
962 Worsnop, D. R., Pandis, S. N., Donahue, N. M., and Kulmala, M.: Organic condensation: a vital  
963 link connecting aerosol formation to cloud condensation nuclei (CCN) concentrations, *Atmos.*  
964 *Chem. Phys.*, 11, 3865–3878, 2011.
- 965 Salma, I., Borsós, T., Weidinger, T., Aalto, P., Hussein, T., Dal Maso, M., and Kulmala, M.:  
966 Production, growth and properties of ultrafine atmospheric aerosol particles in an urban  
967 environment, *Atmos. Chem. Phys.*, 11, 1339–1353, 2011.
- 968 Salma, I., Borsós, T., Németh, Z., Weidinger, T., Aalto, T., and Kulmala, M.: Comparative study of  
969 ultrafine atmospheric aerosol within a city, *Atmos. Environ.*, 92, 154–161, 2014.
- 970 Salma, I., Füre, P., Németh, Z., Farkas, Á., Balásházy, I., Hofmann, W., and Farkas, Á.: Lung burden  
971 and deposition distribution of inhaled atmospheric urban ultrafine particles as the first step in their  
972 health risk assessment, *Atmos. Environ.*, 104, 39–49, 2015.
- 973 Salma, I., Németh, Z., Weidinger, T., Kovács, B., and Kristóf, G.: Measurement, growth types and  
974 shrinkage of newly formed aerosol particles at an urban research platform, *Atmos. Chem. Phys.*,  
975 16, 7837–7851, 2016a.
- 976 Salma, I., Németh, Z., Kerminen, V. M., Aalto, P., Nieminen, T., Weidinger, T., Molnár, Á., Imre, K.,  
977 and Kulmala, M.: Regional effect on urban atmospheric nucleation, *Atmos. Chem. Phys.*, 16,  
978 8715–8728, 2016b.
- 979 Salma, I., Varga, V., and Németh, Z.: Quantification of an atmospheric nucleation and growth process  
980 as a single source of aerosol particles in a city, *Atmos. Chem. Phys.*, 17, 15007–15017, 2017.
- 981 Schobesberger, S., Junninen, H., Bianchi, F., Lonn, G., Ehn, M., Lehtipalo, K., Dommen, J., Ehrhart,  
982 S., Ortega, I. K., Franchin, A., Nieminen, T., Riccobono, F., Hutterli, M., Duplissy, J., Almeida, J.,  
983 Amorim, A., Breitenlechner, M., Downard, A. J., Dunne, E. M., Flagan, R. C., Kajos, M.,  
984 Keskinen, H., Kirkby, J., Kupc, A., Kurten, A., Kurten, T., Laaksonen, A., Mathot, S., Onnela, A.,  
985 Praplan, A. P., Rondo, L., Santos, F. D., Schallhart, S., Schnitzhofer, R., Sipilä, M., Tome, A.,  
986 Tsagkogeorgas, G., Vehkamäki, H., Wimmer, D., Baltensperger, U., Carslaw, K. S., Curtius, J.,  
987 Hansel, A., Petäjä, T., Kulmala, M., Donahue, N. M., and Worsnop, D. R.: Molecular  
988 understanding of atmospheric particle formation from sulfuric acid and large oxidized organic  
989 molecules, *Proc. Natl. Acad. Sci. U.S.A.*, 110, 17223–17228, [10.1073/pnas.1306973110](https://doi.org/10.1073/pnas.1306973110), 2013.
- 990 Sihto, S.-L., Mikkilä, J., Vanhanen, J., Ehn, M., Liao, L., Lehtipalo, K., Aalto, P. P., Duplissy, J.,  
991 Petäjä, T., Kerminen, V.-M., Boy, M., and Kulmala, M.: Seasonal variation of CCN concentrations  
992 and aerosol activation properties in boreal forest, *Atmos. Chem. Phys.*, 11, 13269–13285, 2011.
- 993 Sipilä, M., Berndt, T., Petäjä, T., Brus, D., Vanhanen, J., Stratmann, F., Patokoski, J., Mauldin, R. L.  
994 3rd, Hyvärinen, A. P., Lihavainen, H., and Kulmala, M.: The role of sulfuric acid in atmospheric  
995 nucleation, *Science*, 327(5970), 1243–6. doi: [10.1126/science.1180315](https://doi.org/10.1126/science.1180315), 2010.
- 996 Spracklen, D. V., Carslaw, K. S., Merikanto, J., Mann, G. W., Reddington, C. L., Pickering, S., Ogren,  
997 J. A., Andrews, E., Baltensperger, U., Weingartner, E., Boy, M., Kulmala, M., Laakso, L.,  
998 Lihavainen, H., Kivekäs, N., Komppula, M., Mihalopoulos, N., Kouvarakis, G., Jennings, S. G.,  
999 O'Dowd, C., Birmili, W., Wiedensohler, A., Weller, R., Gras, J., Laj, P., Sellegri, K., Bonn, B.,



- 1000 Krejčí, R., Laaksonen, A., Hamed, A., Minikin, A., Harrison, R. M., Talbot, R., and Sun, J.: The  
1001 contribution of boundary layer nucleation events to total particle concentrations on regional and  
1002 global scales, *Atmos. Chem. Phys.*, 6, 5631–5648, 2006.
- 1003 Tröstl, J., Chuang, W. K., Gordon, H., Heinritzi, M., Yan, C., Molteni, U., Ahlm, L., Frege, C.,  
1004 Bianchi, F., Wagner, R., Simon, M., Lehtipalo, K., Williamson, C., Craven, J. S., Duplissy, J.,  
1005 Adamov, A., Almeida, J., Bernhammer, A. K., Breitenlechner, M., Brilke, S., Dias, A., Ehrhart, S.,  
1006 Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Gysel, M., Hansel, A., Hoyle, C. R., Jokinen, T.,  
1007 Junninen, H., Kangasluoma, J., Keskinen, H., Kim, J., Krapf, M., Kürten, A., Laaksonen, A.,  
1008 Lawler, M., Leiminger, M., Mathot, S., Möhler, O., Nieminen, T., Onnela, A., Petäjä, T., Piel, F.  
1009 M., Miettinen, P., Rissanen, M. P., Rondo, L., Sarnela, N., Schobesberger, S., Sengupta, K., Sipilä,  
1010 M., Smith, J. N., Steiner, G., Tomè, A., Virtanen, A., Wagner, A. C., Weingartner, E., Wimmer, D.,  
1011 Winkler, P. M., Ye, P. L., Carslaw, K. S., Curtius, J., Dommen, J., Kirkby, J., Kulmala, M.,  
1012 Riipinen, I., Worsnop, D. R., Donahue, N. M., and Baltensperger, U.: The role of low-volatility  
1013 organic compounds in initial particle growth in the atmosphere, *Nature*, 533, 527,  
1014 10.1038/nature18271, 2016.
- 1015 Vuollekoski, H., Sihto, S.-L., Kerminen, V.-M., Kulmala, M., and Lehtinen, K. E. J.: A numerical  
1016 comparison of different methods for determining the particle formation rate, *Atmos. Chem. Phys.*,  
1017 12, 2289–2295, 2012.
- 1018 Wehner, B., Wiedensohler, A., Tuch, T. M., Wu, Z. J., Hu, M., Slanina, J., and Kiang, C. S.:  
1019 Variability of the aerosol number size distribution in Beijing, China: new particle formation, dust  
1020 storms, and high continental background, *Geophys. Res. Lett.*, 31, L22108, 2004.
- 1021 Wiedensohler, A., Cheng, Y. F., Nowak, A., Wehner, B., Achtert, P., Berghof, M., Birmili, W., Wu, Z.  
1022 J., Hu, M., Zhu, T., Takegawa, N., Kita, K., Kondo, Y., Lou, S. R., Hofzumahaus, A., Holland, F.,  
1023 Wahner, A., Gunthe, S. S., Rose, D., Su, H., and Pöschl, U.: Mobility particle size spectrometers:  
1024 harmonization of technical standards and data structure to facilitate high quality long-term  
1025 observations of atmospheric particle number size distributions, *Atmos. Meas. Tech.*, 5, 657–685,  
1026 2012.
- 1027 Woo, K. S., Chen, D. R., Pui, D. Y. H., and McMurry, P. H.: Measurement of Atlanta aerosol size  
1028 distributions: observations of ultrafine particle events, *Aerosol Sci. Technol.*, 34, 75–87, 2001.
- 1029 Xiao, S., Wang, M. Y., Yao, L., Kulmala, M., Zhou, B., Yang, X., Chen, J. M., Wang, D. F., Fu, Q.  
1030 Y., Worsnop, D. R., and Wang, L.: Strong atmospheric new particle formation in winter in urban  
1031 Shanghai, China, *Atmos. Chem. Phys.*, 15, 1769–1781, 2015.
- 1032 Yli-Juuti, T., Riipinen, I., Aalto, P. P., Nieminen, T., Maenhaut, W., Janssens, I. A., Claeys, M.,  
1033 Salma, I., Ocskay, R., Hoffer, A., Imre, K., and Kulmala, M.: Characteristics of new particle  
1034 formation events and cluster ions at K-pusztá, Hungary. *Boreal Environ. Res.*, 14, 683–698, 2009.
- 1035 Zhang, R., Wang, G., Guo, S., Zamora, M. L., Ying, Q., Lin, Y., Wang, W., Hu, M., and Wang, Y.:  
1036 Formation of urban fine particulate matter, *Chem. Rev.*, 115, 3803–3855, 2015.