1	Anthropogenic Fine Particulate Matter Pollution Will Be Exacerbated in Eastern							
2	China Due to 21 <sup>st</sup> -Century GHG Warming							
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#### 19 Abstract

China has experienced a substantial increase in severe haze events over the past 20 21 several decades, which is primarily attributed to the increased pollutant emissions 22 caused by its rapid economic development. The climate changes observed under the 23 warming scenarios, especially those induced by increases in greenhouse gases (GHG), are also conducive to the increase in air pollution. However, how the air pollution 24 changes in response to the GHG warming has not been thoroughly elucidated to date. 25 We investigate this change using the century-long large ensemble simulations with the 26 27 Community Earth System Model 1 (CESM1) with the fixed anthropogenic emissions at the year 2005. Our results show that although the aerosol emission is assumed to be 28 29 a constant throughout the experiment, anthropogenic air pollution presents positive responses to the GHG-induced warming. The anthropogenic PM<sub>2.5</sub> concentration is 30 estimated to increase averaged over eastern China at the end of this century, but 31 32 varying from regions, with an increase over northwestern part of eastern China and a 33 decrease over southeastern part. Similar changes can be observed for the light air 34 pollution days. However, the severe air pollution days is reported to increase across eastern China at the end of this century, particularly around the Jing-Jin-Ji region. 35 Further research indicates that the increased stagnation days and the decreased light 36 precipitation days are the possible causes of the increase in PM<sub>2.5</sub> concentration, as 37 well as the anthropogenic air pollution days. Estimation shows that the effect of 38 39 climate change induced by the GHG warming can account for 11%-28% of the changes in anthropogenic air pollution days over eastern China. Therefore, in the 40

- 41 future, more stringent regulations on regional air pollution emissions are needed to
- 42 balance the effect from climate change.

## 44 **1. Introduction**

The extraordinarily rapid development of China has caused extremely high 45 46 aerosol loading and gaseous pollutant emissions that have enveloped most regions across China in the recent decades. The increased pollutant emissions, particularly for 47 the particulate matter finer than 2.5 µm in aerodynamic diameter (PM<sub>2.5</sub>), generally 48 result in severe haze events and present a major threat to public health (Gao et al., 49 2017; Tang et al., 2017; Wang, 2018), crop production (Tie et al., 2016), and regional 50 climates (Cao et al., 2016). For example, the annual averaged PM<sub>2.5</sub> in Beijing 51 exceeded 75  $\mu$ g/m<sup>3</sup> during 2009-2016 (Fig. 1b), which more than three times the 52 recommended 24-hour standard (25  $\mu$ g/m<sup>3</sup>) of the World Health Organization (WHO). 53 This degeneration of the air pollution across China, which is similar to that in Beijing, 54 55 is primarily caused by the integrated effects of high emissions and poor ventilation (Chen and Wang, 2015; Zhang et al., 2016a). Many efforts are thus underway to 56 reduce emissions that cause severe haze pollutions. However, the question remains of 57 58 whether climate change will offset or facilitate these efforts.

Recent studies have documented that the exacerbation of air quality over eastern China was partly modulated by meteorological conditions and climate variability that are generally conducive to the severe haze occurrences (Li et al., 2018; Liao and Chang, 2014; Wang and Chen, 2016; Yang et al., 2016; Zhang et al., 2014; Zhang et al., 2016b). Specifically, Wang *et al.* (2015) revealed that the shrinking Arctic sea ice favors less cyclone activity and a more stable atmosphere conducive to haze formation, which can explain approximately 45%-67% of the interannual to

interdecadal variability of winter haze days over eastern China. Besides Arctic sea ice, 66 other decadal variability and changes, including weak East Asian winter monsoon 67 (Jeong et al., 2017; Li et al., 2016; Yin et al., 2015), strong El Niño-Southern 68 oscillation (Gao and Li, 2015; Zhao et al., 2018), high Pacific decadal oscillation 69 70 (Zhao et al., 2016), and high Arctic oscillation (Cai et al., 2017), may have contributed. In addition, the increasing winter haze days over eastern China may also be linked to 71 the low boundary layer height (Huang et al., 2018; Wang et al., 2018), weakened 72 northerly winds (Yang et al., 2017a), decreased relative humidity (Ding and Liu, 73 74 2014), and increased sea surface temperature (Xiao et al., 2015; Yin and Wang, 2016; Yin et al., 2017). 75

Global warming generally presents an adverse impact on the haze pollution 76 77 across China. Simulations of the dynamic downscaling by the regional climate model RegCM4 under the RCP4.5 (Representative Concentration Pathway) scenarios have 78 shown that the air environment carrying capacity tends to decrease, and the weak 79 ventilation days tend to increase, in the 21<sup>st</sup> century across China, suggesting an 80 increase in the haze pollution potential compared to the current state (Han et al., 2017). 81 82 Furthermore, Cai et al. (2017) projected that the days conducive to severe haze pollution in Beijing would increase by 50% at the end of the 21<sup>st</sup> century (2050-2099) 83 under the RCP8.5 scenarios compared to the historical period. 84

These qualitative estimations of the haze pollution response to climate changes generally derived from the *potential* changes of the corresponding meteorological conditions indirectly. No studies to date quantitatively assessed the simulated PM

directly. How the fine particulate matter pollution changes in response to the global 88 warming in China has not been thoroughly elucidated to date. This study particularly 89 90 focuses on the anthropogenic  $PM_{2,5}$  loading and its response to the future warming. In this study, the large ensemble simulations from the Community Earth System Model 91 Version 1 (CESM1) throughout the 21st century that are induced by increasing 92 93 greenhouse gases (GHG) emissions along the trajectory RCP8.5 but retaining the emissions of aerosols and/or their precursors fixed at the year of 2005 level 94 (RCP8.5 FixAerosol2005; Xu and Lamarque, 2018) will be utilized. 95

### 96 **2. Data and methods**

#### 97 2.1 PM<sub>2.5</sub> observational datasets

98 Surface hourly PM<sub>2.5</sub> concentration data released since 2013 are taken from the website of the Ministry of Environmental Protection (http://106.37.208.233:20035), 99 100 which covers 1602 sites across China. The duration of available datasets varies across 101 sites because of the gradual development of the monitoring network in recent years. In our study region of eastern China (east to 100 °E), there are 1263 sites remaining after 102 103 the sites with missing values were removed during 2015-2017. Additionally, surface daily PM<sub>2.5</sub> concentrations for the Beijing, Shanghai, Guangzhou, and Chengdu cities 104 that had relatively longer monitoring times are also collected from the U.S. Beijing 105 Embassy (http://www.stateair.net/web/historical/1/1.html). 106

#### 107 2.2 CESM1 model simulations

108 The CESM1 is an Earth system model involving the atmosphere, land, ocean,

and sea-ice components with a nominal 1 ° by 1 ° horizontal resolution (Hurrell et al., 109 2013). The RCP8.5\_FixAerosol2005 simulations are forced by the RCP8.5 scenario, 110 111 but all emissions of sulfate (SO<sub>4</sub>), black carbon (BC) and primary organic matter (POM), and secondary organic aerosols (SOA; or their precursors) and atmospheric 112 oxidants are fixed at the present-day level (2005). These simulations include 16 113 114 ensemble members, differing solely in their atmospheric initial conditions with a tiny random temperature difference (order of  $10^{-14}$  °C; Kay et al., 2015). For comparison, 115 the CESM1 large ensemble consists of 35-member simulations that forced by the 116 117 RCP8.5 scenario are also employed here. Using these relatively large ensembles can substantially reduce the contribution of natural variability of the climate system to the 118 119 result estimation (Xu and Lamarque, 2018).

120 For the aerosol emission in the RCP scenarios database, just its decadal change is considered rather than the emission at a single year (Lamarque et al., 2011). Here, the 121 years of 2006-2015 considered reference period 122 are as the in the 123 RCP8.5 FixAerosol2005 simulations. The differences of the mean climates from the reference period are largely due to the increase in GHG emissions and are not 124 125 attributed to the decline in aerosol emissions, as specified in RCP8.5. The changes of anthropogenic PM<sub>2.5</sub> loadings and anthropogenic air pollution days in our study are 126 thus only a result of the GHG-induced climate change, rather than changes in aerosol 127 emission. Note that just four species of PM2.5 components that show a substantial 128 129 threat to public health are considered here for analysis, including SO4, BC, POM, and SOA from the CESM1 simulations. 130

#### 131 **2.3 Definition of the fraction of attributable risk**

The influences of the GHG-induced climate changes on the anthropogenic air 132 133 pollutions in China are investigated using the metric of the fraction of attributable risk (FAR), which has been widely used for attribute analyses of climate extreme changes 134 (Chen and Sun, 2017; Stott et al., 2004). FAR is defined as the  $1-P_0/P_1$ , where  $P_0$  is 135 the probability of exceeding a certain threshold during the reference period and  $P_1$  is 136 the probability exceeding the same threshold during a given period. FAR thus presents 137 the quantitative estimations of effects of the GHG-induced climate changes on the 138 139 anthropogenic air pollutions.

#### 140 **2.4 Definition of stagnation days**

The changes of the stagnation days that were induced by the increase of GHG 141 142 emissions are also evaluated in our study to explore the possible impact of climate 143 change on the anthropogenic air pollutions. The day is considered to be stagnant when the daily mean near-surface wind speed is less than 3.2 m/s, the daily mean 500-hPa 144 145 wind speed is less than 13 m/s, and the daily accumulated precipitation is less than 1 mm (Horton et al., 2012). Early studies have suggested that this air stagnation 146 definition might not be applicable for China to represent the air pollution condition 147 under the seasonal scales (Feng et al., 2018; Wang et al., 2018). However, the annual 148 mean stagnation generally presents good agreement with that of air pollution across 149 China (Huang et al., 2017; 2018). The changes in the annual mean states of air 150 stagnations over China at the end of 21<sup>st</sup> century will thus be discussed in the 151 following. 152

## 153 **3. Results**

#### 154 **3.1 Observational changes in PM<sub>2.5</sub> pollutions**

The days of severe haze pollution increased over the past several decades across 155 eastern China, particularly for the episodes of January 2013, December 2015, and 156 December 2016, when several severe haze alerts were reached. High PM<sub>2.5</sub> loading 157 was centralized over the Jing-Jin-Ji (JJJ) region, Shangdong, and Henan provinces, as 158 well as the Sichuan Basin (SCB, Fig. 1a). The annual mean PM<sub>2.5</sub> mass concentrations 159 for most sites over these regions exceed 75  $\mu$ g/m<sup>3</sup>. According to the statistics, there 160 are approximately 95% sites where the annual mean PM<sub>2.5</sub> concentration exceeded the 161 WHO recommended 24-hour standard (25  $\mu$ g/m<sup>3</sup>) across eastern China, and there are 162 65 sites centralized by Beijing, where the annual mean PM<sub>2.5</sub> concentration was larger 163 than 75  $\mu$ g/m<sup>3</sup>, which would present the possibility of exposing people to serious 164 165 health hazards (World Health Organization, 2014).

Regarding the four economic zones of Beijing, Shanghai, Guangzhou, and 166 167 Chengdu cities over China, serious PM<sub>2.5</sub> pollution can be expected in recent years, especially for the Beijing and Chengdu regions (Fig. 1). Taking Beijing as an example, 168 the annual mean  $PM_{2.5}$  concentration was stably exceeding 100 µg/m<sup>3</sup>, and more than 169 a half of the year had experienced severe air pollution (> 75  $\mu$ g/m<sup>3</sup>) before 2013. 170 Since 2013, China's State Council released its Air Pollution Prevention and Control 171 Action Plan, which requires the key regions, including the JJJ, the Yangtze River 172 Delta (YRD), and the Pearl River Delta (PRD) to reduce their atmospheric levels of 173 PM<sub>2.5</sub> by 25%, 20%, and 15%, respectively, by the end of year 2017 (State Council, 174

175 2013). Effort is obvious, and the  $PM_{2.5}$  loading and the air pollution days present 176 sharp decreases in recent years. However, the strict emission policies substantially 177 cost the economic development, which cannot meet the current requirement of the 178 rapid development of China. Thus, scientifically quantifying the roles of 179 anthropogenic emissions and climate changes shows great importance for seeking the 180 balance between socioeconomic development and emission reduction.

#### 181 **3.2 Simulated changes in anthropogenic PM<sub>2.5</sub> pollutions**

A strong spatial correlation (0.69) is found for the annual mean PM<sub>2.5</sub> 182 concentration between the site observation and median ensemble of CESM1 183 simulations over eastern China (Fig. S1). The high concentrations across eastern 184 China, including the regions centralized by Beijing and Chengdu, are reasonably 185 186 reproduced. However, a negative bias is obvious. Early studies (Li et al., 2016; Yang et al., 2017b; c) have documented that this low bias of aerosol concentration simulated 187 by models is much more complicated in China and the causes mainly involve the 188 189 uncertainties from aerosol emission amount, emission injection height, lack of nitrate, 190 aerosol treatment in model as well as the coarse model resolution.

The median ensemble-mean change of the  $PM_{2.5}$  surface concentration presents strong regional dependence across China with significantly decreasing trends over the southeastern part of eastern China and significantly increasing trends over the other regions throughout the 21<sup>st</sup> century (Fig. S2), even though the emissions are constant throughout the experiment. The regional differences in the total  $PM_{2.5}$  changes are mainly due to SO<sub>4</sub>, which can account for approximately 50% of the total  $PM_{2.5}$  mass

(Xu and Lamarque, 2018). The species of BC and POM are reported to significantly 197 increase in the 21<sup>st</sup> century across eastern China, although the aerosol emissions were 198 fixed at the level in 2005. Figure 2 presents the simulated PM<sub>2.5</sub> loadings from the 199 CESM1 model, in terms of column burden and surface concentration, are significantly 200 increasing throughout the 21<sup>st</sup> century. The increase in the total PM<sub>2.5</sub> is 201 approximately 8% for the column burden and 2% for the surface concentration at the 202 end of the 21<sup>st</sup> century (2090-2099) with respect to the current state (2006-2015). 203 These increasing trends of PM<sub>2.5</sub> loadings are mainly due to the significant increases 204 205 of the major PM<sub>2.5</sub> species, except for SOA, in which the surface concentration presents a slight decrease. Furthermore, the increases of all major PM<sub>2.5</sub> species in 206 terms of column burden (BC: 11%, SO<sub>4</sub>: 6%, SOA: 11%, and POM: 11%) show 207 208 stronger than the surface concentration (BC: 4%, SO<sub>4</sub>: 2%, SOA: -1%, and POM: 4%). 209

For comparison, we also evaluated the future changes of PM<sub>2.5</sub> concentrations 210 and the associated species along the RCP8.5 forcing trajectory from the large 211 ensemble simulations of CESM1 (Figure not shown). Different from changes of 212 aerosol concentrations under the fixed aerosol simulations, the PM<sub>2.5</sub> concentrations 213 and the associated species present uniformly decreasing trends across eastern China 214 from the simulations along the RCP8.5 forcing. The decreasing trends in the RCP8.5 215 simulations are mainly attributed to the prescribed decrease of aerosol forcing in the 216 future in RCP database (Xu and Lin, 2017). The climate change induced by the 217 GHG-warming might exacerbate the air pollution, but the impacts cannot compensate 218

the prescribed decreasing trend of aerosol concentration.

As mentioned above, the PM<sub>2.5</sub> surface concentration in the two economic zones 220 221 of YRD and PRD present a negative response to the GHG-induced warming, while the corresponding column burden shows significantly increasing trends (Fig. S3). The 222 223 decreases of the surface concentration over these two zones are primarily contributed by the changes of SO<sub>4</sub> and SOA, while there are no obvious trends for BC and POM 224 (Figs. S4-S7). The robust response of the increased surface wind speed and decreased 225 226 upper-level wind speed to GHG warming can be partly responsible for the changes of 227 the major PM<sub>2.5</sub> species in these two zones, which will be further discussed. Over the zones of JJJ and SCB, both the PM<sub>2.5</sub> concentrations and the associated major PM<sub>2.5</sub> 228 species present the significantly rising trends throughout the 21<sup>st</sup> century. For the 229 230 surface concentration, PM<sub>2.5</sub> is reported to increase by 3% and 4% in the regions of JJJ and SCB, respectively, at the end of the 21<sup>st</sup> century. The BC is reported to 231 increase by 4% and 8% for JJJ and SCB, respectively. The other species, such as SO<sub>4</sub> 232 and POM, increase by 4% and 4%, respectively, in the JJJ regions and by 2% and 9%, 233 respectively, in SCB regions. Relatively stronger responses can be seen in changes of 234 the column burden for all major species (Figs. S4-S7). The increased concentrations 235 of  $PM_{25}$  species finally result in significantly increasing trends of the total  $PM_{25}$ 236 loading over these two regions, which will present a more direct effect on human 237 health. 238

239 The increase in  $PM_{2.5}$  surface concentration throughout the  $21^{st}$  century 240 substantially leads to the significant increase of the light anthropogenic  $PM_{2.5}$ 

pollution days ( $PM_{2.5} > 25 \mu g/m^3$ ) across the northwestern part of eastern China (Fig. 241 3). Due to the decrease of  $PM_{2.5}$  concentration over the southeastern part of eastern 242 243 China, the light anthropogenic air pollution days can be expected to decrease in this region. Estimation shows that the number of the light air pollution days would be 244 decreased by approximately 10 days at the end of the 21<sup>st</sup> century with respect to the 245 early period of this century in the region. However, the annual mean light air pollution 246 days is reported to increase averaged over the eastern China at the end of this century 247 despite the aerosol emission is constant throughout the experiment. In contrast to the 248 light air pollution days, the severe anthropogenic air pollution days ( $PM_{2.5} > 75 \mu g/m^3$ ) 249 show a positive response to the GHG-induced warming across eastern China, 250 particularly for the regions around JJJ in which the high PM<sub>2.5</sub> concentration was 251 252 localized (Fig. 3). The severe air pollution days is estimated to increase by more than 2 days at the end of this century when compared to the early period over this region. 253 Considering the underestimation in aerosol concentration by CESM1 model in China, 254 255 the percentile threshold metric is also applied here to estimate the future changes in light (90th) and severe (99th) air pollution days. Similar results can be obtained (Fig. 256 S8). 257

#### 258 **3.3 Attributable changes due to GHG warming**

Although the aerosol emission was constant throughout the experiment, our study reveals that the  $PM_{2.5}$  loadings and their associated pollution days still present increases throughout the  $21^{st}$  century, primarily resulting from the impact of climate change induced by GHG warming. One may ask how large a contribution the climate change exerts on the changes in anthropogenic air pollution. To quantitatively address
this issue, the framework of the "Fraction of Attributable Risk (FAR)" that has been
widely used for attribute analyses of climate extreme changes (Chen and Sun, 2017;
Stott et al., 2004) is employed in this study.

Figure 4 shows the percentage changes of the anthropogenic air pollution days 267 throughout the 21<sup>st</sup> century over eastern China and their associated FAR variations. 268 The regional averaged anthropogenic air pollution days present an obvious increase in 269 the 21<sup>st</sup> century as addressed above. Correspondingly, synchronous increasing trends 270 can be found in FAR for both light and severe anthropogenic air pollution days. For 271 the light pollution days, FAR is estimated to be 28% at the end of the 21<sup>st</sup> century, 272 implying that approximately 28% of the pollution days are contributed by the climate 273 274 change that was induced by GHG warming. For the severe pollution days, FAR shows a relatively smaller value of approximately 11%. Furthermore, the high FAR values 275 are mainly located over the regions of high PM<sub>2.5</sub> loadings concentrated over eastern 276 China, suggesting considerably stronger effects of climate changes in these regions. 277 Note that the FAR values estimated in this research may be underestimated because 278 the GHG-induced warming impact was involved in the selected reference period that 279 resulted in the overestimation of the probability of anthropogenic air pollution days. 280

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#### 3.4 Effects of the changes in meteorological conditions

We further examined the changes of meteorological conditions induced by the GHG warming that alternatively exerted effects on air pollution. Our results show that the local boundary layer height presents as higher under the warming scenario (Fig. 5a), which benefits the vertical transport of the air pollutant.

However, a robust negative response of the horizontal advection to the 286 GHG-induced warming across eastern China can be found in the troposphere (Fig. 5b, 287 c), facilitating air pollutant accumulation. The change of surface wind speed in 288 response to the GHG warming is highly similar with the variation of PM<sub>2.5</sub> surface 289 concentration, with wind speed increasing in the southeastern part of eastern China 290 and decreasing in the northwestern part. Variations of surface wind speeds are thus 291 mainly responsible for the changes of PM<sub>2.5</sub> surface concentration over eastern China. 292 293 Different responses can be found for the tropospheric upper-level wind speeds, which are reported to substantially decrease. These decreases would directly result in 294 significant increases of the stagnation days over eastern China, particularly over the 295 northern region and SCB (Fig. 6). The decreasing trend of wind speed in the 21<sup>st</sup> 296 century across China not only exists in CEMS1 model, but also happens in the other 297 global climate models that participated in Coupled Model Intercomparison Project 298 Phase 3 (CMIP3) and CMIP5 (Jiang et al., 2010a; McInnes et al., 2011), as well as in 299 regional climate models (Jiang et al., 2010b). 300

In response to the GHG-induced warming, the stagnation days over eastern China are estimated to increase by 6% at the end of 21<sup>st</sup> century with respect to the current period. For the specific economic zones, the stagnation days over the SCB and JJJ regions show considerably stronger rising trends, while relatively weaker increases are observed over the YRD and PRD regions. The number of stagnation days is estimated to increase by 13% and 6% at the end of the 21<sup>st</sup> century for the SCB and JJJ regions, respectively. Briefly, though the atmospheric stratification appears to be considerably more unstable in response to the GHG warming, the weakened horizontal advection would substantially increase the stagnation days over eastern China, which provides a beneficial background for the air pollutant accumulation and further increases the occurrence probability of the anthropogenic air pollution events.

Early studies have documented a significant increase in total precipitation across 312 China due to the GHG-induced warming (Chen, 2013; Li et al., 2018; Wang et al., 313 314 2012), which seems to represent a conflict with the increase of the anthropogenic air 315 pollution days. To resolve this issue, the precipitation changes in terms of light precipitation days (daily accumulated precipitation < 10 mm) and heavy precipitation 316 days (> 10 mm) are further examined (Fig. 5d, e). Clearly, the heavy precipitation 317 318 days present an increase, while the light precipitation days show a decrease, across eastern China in response to the warming. Though the precipitation shifts toward 319 heavy precipitation events, its cleansing impact on air pollutants has not increased 320 321 because an increase in heavy precipitation days appears to be insufficient to further enhance the wet removal ability (Xu and Lamarque, 2018). In contrast, the decrease in 322 light precipitation days substantially weakens the wet deposition of air pollutants, 323 leading to the increase of the PM<sub>2.5</sub> loading, as well as anthropogenic air pollution 324 days. The future changes of precipitation days present much robust. Both the 325 increasing trends of heavy precipitation days and the decreasing trends of light 326 precipitation days are also obvious across China simulated by the CMIP5 models 327 (Chen and Sun, 2013; 2018), as well as the regional climate models (Gao et al., 2012). 328

## 329 **4. Conclusions**

The world is predicted to experience increased disasters, such as heat waves, 330 331 flash floods, and storms, due to the continuous global warming induced by the GHG increase. The research question we aim to address in this study is how the GHG 332 warming would affect the anthropogenic PM<sub>2.5</sub> pollutions across China. Our 333 evaluations show that the anthropogenic PM<sub>2.5</sub> loadings, as well as the anthropogenic 334 PM<sub>2.5</sub> pollution days, would increase under the global warming conditions even the 335 aerosol emissions fixed at current levels. More stringent regulations are thus 336 337 suggested for regional aerosol emissions to maintain the air quality standard as the 338 current state.

The climate changes induced by GHG warming exert their effects on the 339 340 anthropogenic air pollutions across eastern China via two ways that are of interest in this study. First, the weakened tropospheric wind speed induced by the GHG warming 341 would result in a decrease of the horizontal advection and lead to an increase in the 342 number of stagnation days, facilitating the local accumulation of air pollutants. 343 Second, the number of light precipitation days would decrease due to GHG-induced 344 warming, although the total precipitation would clearly increase across China. This 345 shift toward more no-rainfall days would further weaken the wet deposition of PM<sub>2.5</sub> 346 pollutants. Thus, the increased stagnation days and decreased light precipitation days 347 provide a beneficial background for the occurrence of anthropogenic air pollution. Of 348 349 course, under the warming scenarios, a large discrepancy exists among the different meteorological processes that benefit the air pollutions at the current state, leading to 350

351	the fuzzy recognition of air pollution change. For example, the boundary layer height
352	shows an increase in response to the GHG warming that may strengthen the vertical
353	dissipation of air pollutants. Thus, more studies are suggested in the future to further
354	understand the mechanisms governing air quality across China.
355	

# 357 Author contributions

358	H. P. Chen and H. J. Wang designed the research; H. P. Chen analyzed the data.
359	All the authors discussed the results and wrote the paper.
360	
361	Competing interests
362	The authors declare that they have no conflict of interest.
363	
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## 530 Figure captions

531 Figure 1. Observed PM<sub>2.5</sub> pollution conditions over eastern China during the past

532 **years.** (a) Annual averaged  $PM_{2.5}$  concentration ( $\mu g/m^3$ ) for the years of 2015-2017.

533 (b) Variations of annual averaged  $PM_{2.5}$  concentration (green bars) in Beijing city and

the corresponding number of the severe  $PM_{2.5}$  pollution days (red bars). The severe pollution days are defined as the daily averaged  $PM_{2.5}$  concentration exceeding 75  $\mu g/m^3$ . (c), (d), and (e) are similar to (b), but for the results of Shanghai, Guangzhou,

537 and Chengdu city, respectively.

Figure 2. Plots of future changes of the total  $PM_{2.5}$  as well as its associated species averaged over eastern China in terms of the surface concentration ( $\mu$ g/m<sup>3</sup>, right axis in red) and column burden (mg/m<sup>2</sup>, left axis in blue) from the simulations of the RCP8.5\_FixAerosol2005 experiment. (a) PM<sub>2.5</sub>, (b) BC, (c) SO<sub>4</sub>, (d) POM, and (e) SOA. Ensemble variance (1 sigma) for surface concentration is shown in red shadings.

Figure 3. Changes of the anthropogenic PM<sub>2.5</sub> pollution days across eastern 544 China from the RCP8.5\_FixAerosol2005 experiment. The top panel (a, b) shows 545 the changes of light air pollution days (> 25  $\mu$ g/m<sup>3</sup>) and the bottom panel (c, d) shows 546 the results of severe air pollution days (> 75  $\mu$ g/m<sup>3</sup>). The left panel (a, c) illustrates the 547 annual averaged severe pollution days in 2006-2015 and the right panel (b, d) shows 548 changes of the pollution days at the end of the 21<sup>st</sup> century with respect to 2006-2015. 549 Dots in (b) and (d) mean the changes are significant at the 95% confidence level using 550 Student T-test for all years and ensembles. Units: days. 551

Figure 4. Attributable changes of anthropogenic air pollution days to the increased greenhouse gases emissions. (a) Spatial distribution of FAR for the changes of severe  $PM_{2.5}$  pollutions (> 75 µg/m<sup>3</sup>) at the end of the 21<sup>st</sup> century over eastern China. (b) Regional averaged relative changes of air pollution days (left axis in red; > 25 µg/m<sup>3</sup>) and the corresponding variation of FAR (right axis in blue). Ensemble variance (1 sigma) for the relative changes of pollution days is shown in red shadings. (c) is similar to (b), but for the severe  $PM_{2.5}$  pollution days. Units: %.

Figure 5. Simulated changes in weather conditions of the air pollutions across 559 eastern China due to the GHG-induced warming. (a) Changes of the planetary 560 boundary layer height (PBLH) at the end of the 21<sup>st</sup> century with respect to the years 561 of 2006-2015 from the RCP8.5\_FixAerosol2005 experiment. (b) and (c) are similar to 562 563 (a) but for the wind speed at near-surface and 500-hPa levels, respectively. (d) Changes in the light precipitation days (daily accumulated precipitation < 10 mm) at 564 the end of the  $21^{st}$  century with respect to the current state. (e) is similar to (d) but for 565 the heavy precipitation days (> 10 mm). Dots in the figure mean the changes are 566 significant at the 95% confidence level using Student T-test for all years and 567 ensembles. Units: %. 568

Figure 6. Changes in the stagnant conditions across China due to the GHG-induced warming. (a) Distribution of the relative changes of the stagnation days at the end of the 21<sup>st</sup> century against the current state (2006-2015). Dots mean the changes are significant at the 95% confidence level using Student T-test for all years and ensembles. (b) Variations of the regional averaged stagnation days over

574	eastern China.	Ensemble	variance (	1 sigma)	is shown	in red	shadings.	(c), (d),	(e), and
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- 575 (f) are similar to (b), but for the results of four Chinese economic zones, i.e., JJJ, YRD,
- 576 PRD, and SCB. Units: %.
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## 579 Figures

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Figure 1. Observed  $PM_{2.5}$  pollution conditions over eastern China during the past years. (a) Annual averaged  $PM_{2.5}$  concentration (µg/m<sup>3</sup>) for the years of 2015-2017. (b) Variations of annual averaged  $PM_{2.5}$  concentration (green bars) in Beijing city and the corresponding number of the severe  $PM_{2.5}$  pollution days (red bars). The severe pollution days are defined as the daily averaged  $PM_{2.5}$  concentration exceeding 75 µg/m<sup>3</sup>. (c), (d), and (e) are similar to (b), but for the results of Shanghai, Guangzhou, and Chengdu city, respectively.

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Figure 2. Plots of future changes of the total  $PM_{2.5}$  as well as its associated species averaged over eastern China in terms of the surface concentration ( $\mu$ g/m<sup>3</sup>, right axis in red) and column burden (mg/m<sup>2</sup>, left axis in blue) from the simulations of the RCP8.5\_FixAerosol2005 experiment. (a) PM<sub>2.5</sub>, (b) BC, (c) SO<sub>4</sub>, (d) POM, and (e) SOA. Ensemble variance (1 sigma) for surface concentration is shown in red shadings.



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Figure 3. Changes of the anthropogenic PM<sub>2.5</sub> pollution days across eastern 600 China from the RCP8.5\_FixAerosol2005 experiment. The top panel (a, b) shows 601 the changes of light air pollution days (> 25  $\mu$ g/m<sup>3</sup>) and the bottom panel (c, d) shows 602 the results of severe air pollution days (> 75  $\mu$ g/m<sup>3</sup>). The left panel (a, c) illustrates the 603 annual averaged severe pollution days in 2006-2015 and the right panel (b, d) shows 604 changes of the pollution days at the end of the 21<sup>st</sup> century with respect to 2006-2015. 605 Dots in (b) and (d) mean the changes are significant at the 95% confidence level using 606 Student T-test for all years and ensembles. Units: days. 607



Figure 4. Attributable changes of anthropogenic air pollution days to the increased greenhouse gases emissions. (a) Spatial distribution of FAR for the changes of severe  $PM_{2.5}$  pollutions (> 75 µg/m<sup>3</sup>) at the end of the 21<sup>st</sup> century over eastern China. (b) Regional averaged relative changes of air pollution days (left axis in red; > 25 µg/m<sup>3</sup>) and the corresponding variation of FAR (right axis in blue). Ensemble variance (1 sigma) for the relative changes of pollution days is shown in red shadings. (c) is similar to (b), but for the severe  $PM_{2.5}$  pollution days. Units: %.

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Figure 5. Simulated changes in weather conditions of the air pollutions across 621 eastern China due to the GHG-induced warming. (a) Changes of the planetary 622 boundary layer height (PBLH) at the end of the 21<sup>st</sup> century with respect to the years 623 of 2006-2015 from the RCP8.5\_FixAerosol2005 experiment. (b) and (c) are similar to 624 (a) but for the wind speed at near-surface and 500-hPa levels, respectively. (d) 625 Changes in the light precipitation days (daily accumulated precipitation < 10 mm) at 626 the end of the  $21^{st}$  century with respect to the current state. (e) is similar to (d) but for 627 the heavy precipitation days (> 10 mm). Dots in the figure mean the changes are 628 significant at the 95% confidence level using Student T-test for all years and 629 ensembles. Units: %. 630

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Figure 6. Changes in the stagnant conditions across China due to the 633 GHG-induced warming. (a) Distribution of the relative changes of the stagnation 634 days at the end of the 21<sup>st</sup> century against the current state (2006-2015). Dots mean 635 the changes are significant at the 95% confidence level using Student T-test for all 636 637 years and ensembles. (b) Variations of the regional averaged stagnation days over eastern China. Ensemble variance (1 sigma) is shown in red shadings. (c), (d), (e), and 638 (f) are similar to (b), but for the results of four Chinese economic zones, i.e., JJJ, YRD, 639 640 PRD, and SCB. Units: %.

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