

**Interactive comment on “Ozone source apportionment during peak summer events over southwestern Europe” by Maria Teresa Pay et al.**

**Anonymous Referee #2**

Received and published: 16 October 2018

We thank the Referee #2 for his/her thorough and constructive comments and suggestions, which have contributed to improve the quality of our paper. All his/her comments have been implemented and commented accordingly in the revised version of the manuscript.

Please, find below the item-by-item response. For more details on the review process, we have uploaded the manuscript with track-changes.

**1. Reviewer #2: In the introduction, the authors provide a good background of the ozone related issues in the Iberian Peninsula including an overview of meteorological conditions typically associated to high ground-level O<sub>3</sub> concentrations along with a number of relevant references. They also briefly discuss the trends and justify the need for their research. The topic is timely and interesting not only from the scientific point of view but also considering the legal implications and the need to identify potential interventions that may help alleviate O<sub>3</sub> pollution in the Mediterranean Basin. Given the limitations of brute force methods for source apportionment studies of secondary pollutants, the authors apply the Integrated Source Apportionment Method (ISAM) implemented within the Community Multiscale Air Quality (CMAQ) with 4x4 km<sup>2</sup> resolution for the whole Iberian Peninsula during a 10-day specific episode in summer 2012. The rationale and approach is clear but it would be interesting to explicitly state the main purpose of the study since this is relevant to understand whether the experiment design is appropriate and what are the limitations that can be expected from potential conclusions.**

**Authors:** We thank the Reviewer for their assessment of the scope and methodology of the manuscript. We have rewritten some parts of the abstract and the introduction to clarify the main propose of the study as follows:

In the abstract (*Page 1 – Line 19-21*):

*“The main goal of this study is to provide a first quantitative estimation of the contribution of the main anthropogenic activity sectors to peak O<sub>3</sub> events in Spain relative to the contribution of imported (regional and hemispheric) O<sub>3</sub>. We also assess the potential of our source apportionment method to improve O<sub>3</sub> modelling”*

In the introduction (*Page 4 – Line 25-34*):

*“The integrated source apportionment tools combined with high-resolution emission and meteorological models can help unravelling the sources responsible for peak summer events of O<sub>3</sub> in the Western Mediterranean Basin. Quantifying the contribution of emission sources during acute O<sub>3</sub> episodes is a prerequisite for the design of future mitigation strategies in the region. In this framework, the main goal of this study is to provide a quantitative estimation of the contribution of the main anthropogenic activity sectors compared to the imported concentration (regional and hemispheric) to peak O<sub>3</sub> events in Spain. We also assess the potential of our source apportionment method to improve O<sub>3</sub> modelling. Our study applies for the first time a countrywide O<sub>3</sub> source apportionment at high resolution over the Iberian Peninsula during the period between July 21<sup>st</sup> and 31<sup>st</sup>, 2012. We use the CMAQ-ISAM within the CALIOPE air quality forecast system for Spain ([www.bsc.es/caliope](http://www.bsc.es/caliope)), which runs at a horizontal resolution of 4x4 km<sup>2</sup> over the IP. The system is fed by the HERMESv2.0 emission model, which provides disaggregated emissions based on local information and state-of-the-art bottom-up approaches for the most polluting sectors.”*

**2. Reviewer #2: P3.Line 18: when discussing the possible reasons for the observed increase of O<sub>3</sub> in some urban areas in the Iberian Peninsula, the authors assume a VOC-limited situation. Reductions on NO<sub>x</sub> emissions have necessarily reduced NO titration but I suggest them to remove that assumption regarding the VOC-NO<sub>x</sub> regime because it may be an oversimplification and not necessarily true in all cases/seasons.**

**Authors:** We agree with the reviewer. We have removed the assumption regarding the VOC-NO<sub>x</sub> regime as follows:

Page 3–Line 17-19: “*The reasons behind the urban O<sub>3</sub> upward trend are not clear yet due to the complex VOC-NO<sub>x</sub> regime; part of the O<sub>3</sub> increase may have resulted from the reduction of NO emissions relative to NO<sub>2</sub> and therefore to a lower NO titration effect in VOC-limited situations.*”

**3. Reviewer #2: P4.Line 29: the authors claim that the period analysed (between July 21st and 31st) is representative of typical summer synoptic conditions in that particular region. That is quite a strong statement and it would require substantial discussion and evidence to demonstrate to what extent that is true. Regardless of that, although the period may be characteristic in terms of synoptic conditions, O<sub>3</sub> dynamics as previously stated, is strongly conditioned by both long-range transport and local conditions, including emissions of O<sub>3</sub> precursors, initial chemical conditions, etc. Although the paper constitutes a valuable contribution to the understanding of O<sub>3</sub> pollution in the Iberian Peninsula, I think that it cannot be assumed that the outcomes of the study may provide a source apportionment comprehensive description. Consequently, the insight to support the design abatement policies is limited and caution should be used to avoid extracting incorrect conclusions.**

**Authors:** Our characterization of the study period (July 21<sup>st</sup> and 31<sup>st</sup>, 2012) is based on the circulation type classification performed in Valverde et al. (2014, *Circulation-type classification derived on a climatic basis to study air quality dynamics over the Iberian Peninsula. Int. J. Climatol.* 35 (8)). Specifically, the classification in Valverde et al. (2014) is designed to study air quality dynamics over the Iberian Peninsula using an objective synoptic classification method over the present climate (1983–2012). According to the classification in Valverde et al. (2014), our study episode starts with an Iberian Thermal Low (ITL) (21-25<sup>th</sup> July, 2012), followed by a northwestern advection from the Atlantic (NWad) (26-29<sup>th</sup> July, 2012) and finishing with another ITL (30-31<sup>st</sup> July, 2012). ITL and NWad are circulation types that typically affect the Iberian Peninsula, which represent the 44% of the days in the IP both taking place in summer and alternate each other (Valverde et al., 2014).

We have provided a summary of these evidences to support the representativeness of the episode in the revised version of the manuscript as follows:

Page 9–Line 25: “*Our characterization of the study period is based on the circulation type classification performed in Valverde et al. (2014), who developed an objective synoptic classification method over the period 1983–2012, specifically designed to study air quality dynamics over the IP. [...]. According to the circulation type classification in Valverde et al. (2014), the selected episode started with the development of the ITL (July 21st-25th), followed by a NWad-venting period (July 26th-29th) and ended with the development of another ITL (July 30th-31st).*”

On the other hand, we want to remark that our study is a first quantitative O<sub>3</sub> source apportionment study, and the representativeness of our results is limited because they are focused in just one episode. We have clarified this limitation in the manuscript in the Section 4 as follows:

Page 20–Line 1: *“Our study has provided a first estimation of the main sources responsible for high O<sub>3</sub> concentration in the Western Mediterranean Basin during the period July 21st -31st, 2012.”*

Page 22–Line 22-23: *“[...] future studies should preferentially cover multiple summer periods in order to improve representativeness.”*

Furthermore, we want to highlight that the main goal of this study is not the design abatement policies, but it is to provide a first quantitative estimation of the contribution of the main anthropogenic activity sectors to peak O<sub>3</sub> events in Spain relative to the contribution of imported O<sub>3</sub>. Actually, source apportionment techniques alone cannot be used to the design abatement policies. Subsequent source sensitivity analyses tailoring the identified main contribution sources could predict how O<sub>3</sub> will respond to reductions in precursor emissions. Both, source apportionment and source sensitivity are complementary and essential studies to define the most efficient O<sub>3</sub> abatement strategies in the Western Mediterranean Basin. Therefore, this study has provide a perspective about the potential use of source apportionment methods for regulatory studies in non-attainment regions as a prerequisite for the design of future mitigation strategies. We have added some remarks about this point as follows:

*Page 4 – Line 28-30 (Section 1. Introduction): “In this framework, the main goal of this study is to provide a first quantitative estimation of the contribution of the main anthropogenic activity sectors compared to the imported concentration (regional and hemispheric) to peak O<sub>3</sub> events in Spain.”*

*Page 22–Line 23-27 (Section 4. Discussion and conclusions): “We note that our results cannot predict whether emission abatement will have either a positive or negative effect in O<sub>3</sub> changes due to the non-linearity of the O<sub>3</sub> generation process. Subsequent source sensitivity analyses tailoring the identified main contribution sources could predict how O<sub>3</sub> will respond to reductions in precursor emissions, which are essential to define the most efficient O<sub>3</sub> abatement strategies in the Western Mediterranean Basin.”*

*Page 23-Line 4-7 (Section 4. Discussion and conclusions): “This work has quantified the local and imported contributions to O<sub>3</sub> during an episode in a particular area in southwestern Europe. In addition, we have provided a perspective about the potential use of source apportionment method for regulatory studies in non-attainment regions. Further O<sub>3</sub> source apportionment studies targeting other nonattainment regions in Europe are necessary prior to design local mitigation measures that complement national and European-wide abatement efforts.”*

**4. Reviewer #2: An illustrative description of the CALIOPE and HERMES systems is provided in section 2.1. along with a number of relevant references. The authors state, however that emissions are based on 2009 since that is the most recent year with updates information on local emission activities. That statement is hard to understand and it may deserve further clarification. In addition, this section would benefit from a more consistent discussion on how this methodological choice may impact the results and to what extent potential inconsistencies with meteorology and boundary conditions for that specific modelling period are compatible with the research specific**

**aim (which is not clearly identified). This issue may be acceptable to gain a general understanding of O<sub>3</sub> contribution over a long period of time but for a short, high-O<sub>3</sub> episode this may be a potential flaw that should be carefully addressed.**

**Authors:** Our methodological choice has been to use a detailed bottom-up emission inventory instead of a typical top-down regional emission inventory. Bottom-up emissions, estimated using source-specific emission factors and activity statistics, accurately characterise pollutant sources and allow obtaining more realistic results than the ones reported by top-down or regional emission inventories. However, they require very large efforts to be compiled and consequently the updating processes cannot be implemented year-to-year.

In HERMESv2 emissions are based on 2009 data, which was the closest year with updated information on local emission activities in HERMES at the time this work started.

To understand the impact of the use of 2009 data to study year 2012 we revised the EMEP Centre on Emission Inventories and Projections (EMEP-CEIP), which collects and reviews the national emission inventories from Parties to the Convention on Long-range Transboundary Air Pollution (CLRTAP). Between 2009 and 2012 total NO<sub>x</sub> and NMVOC emissions in Spain decreased by -10.6% and -10.7%, respectively (EMEP CEIP, 2019). For NO<sub>x</sub>, around 80% of this reduction is linked to a reduction of road transport emissions, whereas in the case of NMVOC ~50% of the reduction is due to a decrease of industrial emissions. NO<sub>x</sub> emissions from shipping in Europe have also decreased in the period 2009-2012 by 15%.

For our modelling study, we consider these differences as small and acceptable, and not creating any major inconsistency. The difference of 10-15 % in emissions for certain precursors between 2009 and 2012 is within the typically larger ranges of uncertainty in emission inventories. We also note that all our results are thoroughly evaluated and critically assessed using observations.

In any case, we have followed the reviewer's suggestion, and we have discussed in the manuscript the potential impact of these differences when the contribution of each emission sector is analysed:

*Page 17–Line 32-33: “[...] This factor, added to the 15% decrease of NO<sub>x</sub> shipping emissions observed in Europe between 2009 (HERMESv2.0 base year) and 2012 (EMEP CEIP, 2019) could explain the discrepancies observed.”*

*Page 18–Line 12-14: “[...] it has been estimated that between 2009 and 2012 energy production in coal-fired power plants increased from 13.1% to 19.4% (UNESA, 2012), which implied an increase of NO<sub>x</sub> emissions from the power industry sector of around 19.5% (EMEP CEIP, 2019).”*

The Section 4 of the revised version of the manuscript includes now a comment on the methodological implication of using 2009 emissions for O<sub>3</sub> source apportionment studies in an episode in 2012 as follows:

*Page 5-Line 31-32: “HERMESv2.0 is currently based on 2009 data, which is the closest year with updated information on local emission activities in HERMES at the time this work started.”*

*Page 21-Line 31: “Our methodological choice has been to use a detailed bottom-up emission inventory instead of a typical top-down regional emission inventory. Bottom-up emissions, estimated using source-specific emission factors and activity statistics, accurately characterise pollutant sources and allow obtaining more realistic results than the ones reported by top-down or regional emission inventories. To understand*

*the impact of the use of 2009 data to study year 2012, we revised the EMEP Centre on Emission Inventories and Projections (EMEP-CEIP), which collects and reviews the national emission inventories from Parties to the Convention on Long-range Transboundary Air Pollution. Between 2009 and 2012, total NO<sub>x</sub> and NMVOC emissions in Spain decreased by -10.6% and -10.7%, respectively (EMEP CEIP, 2019). For NO<sub>x</sub>, around 80% of this reduction is linked to a reduction of road transport emissions, whereas in the case of NMVOC ~50% of the reduction is due to a decrease of industrial emissions. For our modelling study, we consider these differences as small and acceptable, and not creating any major inconsistency. The difference of 10-15 % in emissions for certain precursors between 2009 and 2012 is within the typically larger ranges of uncertainty in emission inventories.”*

Reference:

EMEP CEIP, 2019. Officially reported emission data. Available at: [http://www.ceip.at/ms/ceip\\_home1/ceip\\_home/data\\_viewers/official\\_tableau/](http://www.ceip.at/ms/ceip_home1/ceip_home/data_viewers/official_tableau/) (last access February 2019)

**5. Reviewer #2: VOC emissions are particularly relevant input for this analysis. However, our current understanding of VOC emission is limited, especially in urban areas (Lewis, 2018) which makes it difficult to accurately apportion contributions to tropospheric O<sub>3</sub>.**

**Authors:** For the estimation of the main VOC contributors in Spanish urban areas, namely the road transport (SNAP07) and the use of solvents (SNAP06), the HERMESv2.0 emission model uses a combination of bottom-up approaches and downscaling methodologies. The sectors road transport and solvents together account for more than 80% of total VOC emissions in Barcelona and Madrid cities (Soret et al., 2014).

For traffic, HERMESv2.0 estimates VOC emissions according to the Tier 3 method described in the EMEP/EEA guidelines, which is fully incorporated in COPERT 4. Speciation factors to map total VOC to the CB05 chemical mechanism species are also obtained from the EMEP/EEA guidelines.

For the solvent sector, emissions are estimated performing a downscaling methodology of the original Spanish National Emission Inventory due to the lack of specific information on activity data and emission factors. The Spanish inventory, developed by the Spanish Ministry of the Agriculture, Food and Environment (MAPAMA, personal communication), represents the official Spanish contribution to the EMEP emission inventory. It reports total annual emissions of primary pollutants by NUTS 2 level and SNAP elemental activity and it is based on the EMEP/EEA guidelines combined with local activity data (see the corresponding Inventory Informative Report, IIR, for more information: [http://www.ceip.at/ms/ceip\\_home1/ceip\\_home/status\\_reporting/2018\\_submissions/](http://www.ceip.at/ms/ceip_home1/ceip_home/status_reporting/2018_submissions/)).

HERMESv2 assigns a specific spatial proxy, temporal and speciation profile to each pollutant activity after defining it as point, lineal or area source. The speciation treatment of the original emissions is done using the profiles reported by the SPECIATE database (<https://www.epa.gov/air-emissions-modeling/speciate-version-45-through-40>). More details on the methods used to estimate VOC emissions from these two sectors can be found in Guevara et al. (2013).

Despite the efforts in providing detailed emission input data, it is true that there are still a lot of uncertainties and room for improvement in the estimation of urban VOC emission

inventories, as stated recently by several works (Liu et al., 2017; McDonald et al., 2018; Lewis, 2018). Most of these uncertainties are due to the lack of in-situ observational data, which is key for the development of local speciation profiles and for the evaluation of modelled concentrations of VOC. In order to overcome this problem, continuous monitoring of urban VOC should be performed in Spanish cities, following the example of other regions in which O<sub>3</sub> is also a major problem such as Mexico City (Jaime-Palomera et al., 2016). On the other hand, the use of satellite observations of formaldehyde (HCHO) columns to constrain urban VOC emissions could be also pointed out as a future task to improve the representativeness of urban emission inventories (Zhu et al., 2014).

Urban VOC emissions are particularly relevant emissions for this analysis, and uncertainties in the estimation of urban VOC emission inventories makes also uncertain their contribution to tropospheric O<sub>3</sub> source apportionment studies.

In the reviewed version of the manuscript, we have added a comment on the urban VOC emission uncertainty as follows:

Page 6-Line 13-20: *“Urban VOC emissions could be a relevant source for O<sub>3</sub> concentration. Over Spanish urban areas, HERMESv2.0 estimates VOC emissions from road transport and the use of solvents (Fig. 1) following bottom-up approaches (Guevara et al., 2013). However, uncertainties in the estimation of urban VOC emission inventories, as stated recently by several works (Pan et al., 2015; Liu et al., 2017; McDonald et al., 2018; Lewis, 2018) makes uncertain the urban VOC contribution to tropospheric O<sub>3</sub> concentrations. In order to overcome this problem, continuous monitoring of urban VOC should be performed in Spanish cities, following the example of other regions in which O<sub>3</sub> is also a major problem such as Mexico City (Jaimes-Palomera et al., 2016). In addition, the use of formaldehyde satellite observations to constrain urban VOC emissions could be also pointed out as a future task to improve the representativeness of urban emission inventories (Zhu et al., 2014)”*

#### References:

Guevara, M., Martínez, F., Arévalo, G., Gassó, S., and Baldasano, J.M.: An improved system for modelling Spanish emissions: HERMESv2.0, *Atmos. Environ.*, 81, 209–221, doi:10.1016/j.atmosenv.2013.08.053, 2013.

Jaimes-Palomera, M., Retama, A., Elias-Castro, G., Neria-Hernández, A., Rivera-Hernández, O., and Velasco, E.: Non-methane hydrocarbons in the atmosphere of Mexico City: Results of the 2012 ozone-season campaign, *Atmos. Environ.*, 132, 258–275, <https://doi.org/10.1016/j.atmosenv.2016.02.047>, 2016

Lewis, A. C.: The changing face of urban air pollution. *Science*, 359, 744-745. doi:10.1126/science.aar4925, 2018.

Liu, H., Man, H., Cui, H., Wang, Y., Deng, F., Wang, Y., Yang, X., Xiao, Q., Zhang, Q., Ding, Y., and He, K.: An updated emission inventory of vehicular VOCs and IVOCs in China, *Atmos. Chem. Phys.*, 17, 12709-12724, <https://doi.org/10.5194/acp-17-12709-2017>, 2017

McDonald, B. C., de Gouw, J. A., Gilman, J. B., Jathar, S. H., Akherati, A., Cappa, C. D., Jimenez, J. L., Lee-Taylor, J., Hayes, P. L., McKeen, S. A., Cui, Y. Y., Kim, S.-W., Gentner, D. R., Isaacman-VanWertz, G., Goldstein, A. H., Harley, R. A., Frost, G. J., Roberts, J. M., Ryerson, T. B., and Trainer, M.: Volatile chemical products emerging as largest petrochemical source of urban organic emissions, *Science (New York, N.Y.)*, 359, 760–764, 2018

Soret, A., Guevara, M., Baldasano, J.M., 2014. The potential impacts of electric vehicles on air quality in the urban areas of Barcelona and Madrid (Spain). *Atmospheric Environment*, 99, 51–63.

Zhu, L., Jacob, D. J., Mickley, L. J., Marais, E. A., Cohan, D. S., Yoshida, Y., Duncan, B. N., González Abad, G., and Chance, K. V.: Anthropogenic emissions of highly reactive volatile organic compounds in eastern Texas inferred from oversampling of satellite (OMI) measurements of HCHO columns, *Environ. Res. Lett.*, 9, 114004, doi:10.1088/1748-9326/9/11/114004, 2014.

**6. Reviewer #2: It is also widely accepted that biogenic VOCs play a major role on atmospheric photochemistry. For this study, the authors rely on the Model of Emissions of Gases and Aerosols from Nature (MEGAN) version 2.04 (P5.Line 30). This version was revised and extended through version 2.1 (Guenther et al., 2012) that also includes some code fixes. I'd strongly suggest the authors to perform a sensitivity run to understand whether using an outdated version of MEGAN may introduce relevant biases into their simulation. Species tagging and emission categories selection described in section 2.3 seem sensible although VOC emission shares should be reviewed taking into account the previous comment.**

**Authors:** Although we use the MEGANv2.0.4 model, we have used the most updated emission factors (version 2011) from the MEGANv2.1 model (<http://lar.wsu.edu/megan/guides.html>). In the Section 2 of the supplement we discuss the behaviour of our biogenic emissions using this configuration. Figure S2 shows the isoprene concentration at the Montseny station during the DAURE experimental campaign (Seco et al., 2011; <http://cires.colorado.edu/jimenez-group/wiki/index.php/DAURE>). This evaluation indicates that modelled isoprene concentrations with updated emission factors are in reasonably good agreement with observations.

We have improved the description of the upgraded MEGAN version used in this study as follows:

*Page 6 - Line 9-11: "In this study, we have updated MEGANv2.0.4 with emission factors from last MEGANv2.1 (<http://lar.wsu.edu/megan/guides.html>). In Sect. 2 of the supplement, we provide a comparison with measurements from the DAURE campaign (Pandolfi et al., 2014) showing the reasonably good behaviour of our modelled isoprene"*

References:

Pandolfi, M., Querol, X., Alastuey, A., Jimenez, J.L., Jorba, O., Day, D., Ortega, A., Cubison, M.J., Comerón, A., Sicard, M., Mohr, C., Prévot, A.S.H., Minguillón, M.C., Pey, J., Baldasano, J.M., Burkhardt, J.F., Seco, R., Peñuelas, J., van Drooge, B.L., Artiñano, B., Di Marco, C., Nemitz, E., Schallhart, S., Metzger, A., Hansel, A., Lorente, J., Ng, S., Jayne, J., Szidat, S.: Effects of sources and meteorology on particulate matter in the Western Mediterranean Basing: An overview of the DAURE campaign. *J. Geophys. Res. Atmos.*, 119, 4978-5010, 2014

We are currently working on upgrading our modelling system with MEGANv3 (<https://bai.ess.uci.edu/megan/versions/megan3>).

References:

Seco, R., Peñuelas, J., Filella, I., Llusà, J., Molowny-Horas, R., Schallhart, S., Metzger, A., Müller, M., and Hansel, A.: Contrasting winter and summer VOC mixing ratios at a forest site

in the Western Mediterranean Basin: the effect of local biogenic emissions, Atmos. Chem. Phys., 11, 13161-13179, doi:10.5194/acp-11-13161-2011, 2011.

**7. Reviewer #2: I'm also concerned about a potential double counting and/or erroneous spatio-temporal allocation of VOCs emissions from agriculture since plant functional types considered in MEGAN include crops. From previous literature, a share of 70% of total VOCs from SNAP 11 (nature) seems too high and may support that shortcoming. Please, double check this potential issue since it may bring about a considerable bias the outcomes of the study.**

**Authors:** We are not double counting crops emissions. Emissions from agriculture (SNAP10) only include VOC from manure management and field burning of agricultural residues. We only include VOC emissions from cultivated crops estimated by MEGAN.

We have added a clarification in the revised version of the manuscript as follows:

Page 6 - Line 7-9: *"Note that we configured MEGAN to compute VOC emissions from cultivated crops; the agriculture emission module in HERMESv2.0 estimates the VOC from manure management and field burning of agricultural residues."*

**8. Reviewer #2: The study makes advantage of a remarkably dense network of monitoring sites in the area of study to assess model performance through the computation of a series of common statistics (appendix B). The results however are difficult to interpret in their current form. Please, see corresponding suggestion in the results section.**

**Authors:** The presentation of the evaluation has been improved following the reviewer's suggestions in the results below. See Table 2 and Appendix B in the revised version of the manuscript.

**9. Reviewer #2: The authors discuss that the episode at hand concentrated an important percentage of exceedances in Spain (if I understood correctly; the first paragraph may be reviewed for the sake of clarity). However, the inspection of panel a) in Fig 2. Does not seem to indicate that this episode was particularly severe since the distribution of MDA8 is not dissimilar to those of previous years, even when they reflect the concentrations over a 6 month period. In addition, the outliers apparently have moderated values. It would be interesting to make the point for such comparison. I'm not completely sure that it is sound to identify a high pollution episode at national level since O3 largely depends on regional features. If O3 levels were actually high all over the modelling domain, the influence of exported ozone (influenced by synoptic conditions) may be too high and thus, the representativeness of the results and the potential implications policy-wise, rather limited. Please, reflect on that.**

**Authors:** Figure 2a supports the reviewer's comment that the selected episode is not the most severe between 2000-2012 that affected all of Spain. However, it comprises a period with high MDA8 O<sub>3</sub> concentrations measured at rural background stations, actually the 75<sup>th</sup> percentile of those values were above the Target Value, similar to the particularly severe summer of 2003 (Solberg et al., 2008).

This episode is also interesting because it affected Europe (EEA, 2013), for which 33% and 12% of the total number of exceedances were observed for the information threshold and the Target Value in 2012, respectively. The O<sub>3</sub> regional context of the episode allows us to study the influence of the imported O<sub>3</sub> to Spain.

In the revised version of the manuscript, we have provided a more clear description of the relevance of the episode as follows:

Page 9 - Line 4-23: *“Our first estimation of the origin of peak O<sub>3</sub> events in Spain focuses on the episode July 21<sup>st</sup> -31<sup>st</sup>, 2012. Figure 2a illustrates the relevance of the episode showing the observed MDA8 O<sub>3</sub> concentrations trends at the Spanish EIONET stations during the (extended) summers (i.e., from April to September) from 2000 to 2012, together with the concentrations recorded during the episode. Although the selected episode is not the most severe between 2000 and 2012 at national scale, it comprises a period with high MDA8 O<sub>3</sub> concentrations measured at rural background stations, actually the 75<sup>th</sup> percentile of those values were above the Target Value, similar to the particularly severe summer of 2003 (Solberg et al., 2008).*

*This episode is also interesting because it was widespread and affected big parts of Europe (EEA, 2013). Only during this period 33% and 12% of the total number of exceedances for the information threshold and the Target Value in 2012, respectively, were measured. The O<sub>3</sub> regional context of the episode allows us to study the influence of the imported O<sub>3</sub> to Spain.*

*The maps of the 90<sup>th</sup> percentile of the measured MDA8 O<sub>3</sub> concentrations over Spain (Fig. 2b) shows high concentration spots all over the domain. The exceedances of the Target Value were found in the surroundings of large urban areas (Madrid, Barcelona, Valencia, Seville) and along Spanish valleys (i.e., Ebro Valley, Guadalquivir Valley).*

*There were more than 100 exceedances of the O<sub>3</sub> Target Value in most of the days during the episode, with relative maxima on July 25<sup>th</sup>, 28<sup>th</sup> and 31<sup>st</sup> attributed to the change in the synoptic conditions (Fig. S3). Figure 3 shows the meteorological patterns (temperature at 2m, wind at 10m, precipitation, mean sea level pressure and geopotential height at 500 hPa) modelled by WRF-ARW during the three distinctive days over the outer EU12 domain.”*

We agree with the reviewer that other O<sub>3</sub> episodes should be studied to extract statistically robust conclusions on the main source contribution to O<sub>3</sub> events in Spain. To remark that this work is a first quantitative O<sub>3</sub> source apportionment study, and the representativeness of our results is limited because they are focused in just one episode, we have added some comments (Section 4) as follows:

Page 20–Line 1-2: *“Our study has provided a first estimation of the main sources responsible for high O<sub>3</sub> concentration in the Western Mediterranean Basin during the period July 21<sup>st</sup> -31<sup>st</sup>, 2012*

Page 22–Line 24-25: *“[...] future studies should preferentially cover multiple summer periods in order to improve representativeness.”*

Page 22–Line 22-27: *“For regulatory applications, further source apportionment studies should target not only emissions from activity sectors, but also the source regions where the emission abatement strategies should be applied. In addition, future studies should preferentially cover multiple summer periods in order to improve representativeness. We note that our results cannot predict whether emission abatement will have either a positive or a negative effect in O<sub>3</sub> changes due to the*

*non-linearity of the O<sub>3</sub> generation process. Subsequent source sensitivity analyses tailoring the identified main contribution sources could predict how O<sub>3</sub> will respond to reductions in precursor emissions, which are essential to define the most efficient O<sub>3</sub> abatement strategies in the Western Mediterranean Basin.”*

**10. Reviewer #2:** They also claim that the episode affected the central and north of the IP, but Fig. 2c shows high concentration spots all over the domain (or maybe the colour scale is not clear enough). It is also hard to see what the influence of Madrid and Barcelona plumes is (something consistent, to my understanding, with dominant stagnant conditions). Please, try to clarify.

**Authors:** Following the review’s recommendation, in the revised version of the manuscript we have improved the colour scale in the aforementioned Figure (now Fig. 2b) to better distinguish the variability of the O<sub>3</sub> concentrations. In addition, we have clarified the explanation about the spatial coverage of the episode as follows:

Page 9 - Line 16-18: *“The maps of the 90<sup>th</sup> percentile of the measured MDA8 O<sub>3</sub> concentrations over Spain (Fig. 2b) show high concentration spots all over the domain. The exceedances of the Target Value were found in the surroundings of large urban areas (Madrid, Barcelona, Valencia, Seville) and along Spanish valleys (i.e., Ebro Valley, Guadalquivir Valley).”*

**11. Reviewer #2:** Fig 3 illustrates the meteorological conditions through some WRF-ARW outputs. I understand that a thorough model evaluation is not the main purpose of the paper but a minimal check (also through a statistic evaluation) of the credibility of the meteorological simulation (mainly wind fields) may substantially help the authors to gain a better insight of their results. For instance, that would help them to contrast hypotheses such as the one made in P10.Line 25. and following, where they attribute O<sub>3</sub> nighttime overestimation to the underestimation of vertical mixing during nighttime stable conditions. In that case, a comparison of observed (where available) and modelled PBLH may be useful.

**Authors:** According to the reviewer’s comment, we have included in Sect. 4 of the supplement an evaluation of wind speed (WS) and direction (WD) at 10 m and temperature at 2 m (T2M) using METeorological Aerodrome Report stations (METAR).

For the selected episode, there were 50 METAR stations located at airports (see location in Fig. S5). Table S2 shows the scores following the methodology explained in “Section 2.4 Evaluation method” for concentrations.

The modelled T2M shows the best behaviour when compared with observations ( $r=0.91$ ) (Table S2). The model slightly underestimates T2M ( $-0.2$  °C), especially for maximum and minimum temperatures ( $1.0$ °C and  $0.4$  °C for p25 and p75, respectively) (Fig. S5). The model reproduces the WS ( $r=0.42-0.70$ ) with an overestimation of  $\sim 0.3$  ms<sup>-1</sup> on average. The overestimation is particularly marked during nighttime (Fig. S5), coincident with low-level wind speeds. These biases may contribute to the underestimation of surface concentrations of O<sub>3</sub> precursors. The wind direction shows a lower correlation coefficient ( $0.1, 0.43$ ).

We did not evaluate the PBL height in this study, but Bank et al. (2012) used daytime radiosounding and PBL height estimations from backscatter lidar to perform a comprehensive evaluation of PBL parametrization from WRF in the North-East Iberian Peninsula. This study

found that there is a systematic underestimation of PBL height simulated by WRF. These results are consistent with Vautard et al. (2012), who found that models generally underpredict PBL heights at nighttime.

Overall, nighttime meteorology remains a challenge for meteorological models. The nighttime systematic overestimation of wind and underestimation of PBL height is a potential source of large error compensation for the modelling of NO<sub>2</sub> and O<sub>3</sub> nighttime concentrations.

The revised version of the manuscript includes a discussion of the meteorological evaluation in the evaluation section as follows:

Page 12 – 18-22: *“Section S4 in the supplement discusses the meteorological evaluation results and their impact on the pollutant concentrations. Not surprisingly, temperature shows the best behaviour when compared with observations (Table S2). The modelled wind speed is overestimated, particularly during nighttime (Fig. S5), coincident with low-level wind speed. The nighttime overestimation of wind is a source of error in modelled NO<sub>2</sub> and O<sub>3</sub> nighttime concentrations (Vautard et al., 2012; Bessagnet et al., 2016)”*

References:

Bessagnet, B., Pirovano, G., Mircea, M., Cuvelier, C., Aulinger, A., Calori, G., Ciarelli, G., Manders, A., Stern, R., Tsyro, S., García Vivanco, M., Thunis, P., Pay, M.-T., Colette, A., Couvidat, F., Meleux, F., Rouil, L., Ung, A., Aksoyoglu, S., Baldasano, J. M., Bieser, J., Briganti, G., Cappelletti, A., D’Isidoro, M., Fignardi, S., Kranenburg, R., Silibello, C., Carnevale, C., Aas, W., Dupont, J.-C., Fagerli, H., Gonzalez, L., Menut, L., Prévôt, A. S. H., Roberts, P., and White, L.: Presentation of the EURODELTA III intercomparison exercise – evaluation of the chemistry transport models’ performance on criteria pollutants and joint analysis with meteorology, *Atmos. Chem. Phys.*, 16, 12667–12701, doi:10.5194/acp-16-12667-2016, 2016.

Vautard, R., Moran, M. D., Solazzo, E., Gilliam, R. C., Matthias, V., Bianconi, R., Chemel, C., Ferreira, J., Geyer, B., Hansen, A.B., Jericevic, A., Prank, M., Segers, A., Silver, J.D., Werhahn, J., Wolke, R., Rao, S.T., Galmarini, S.: Evaluation of the meteorological forcing used for the Air Quality Model Evaluation International Initiative (AQMEII) air quality simulations. *Atmospheric Environment*, 53, 15-37, 2012.

**12. Reviewer #2: As for the discussion of the general meteorological conditions, I’m not sure that the attempt to discriminate between ITL and NWadv situations is nor illustrative or needed. The discussion is hard to follow and the application of deterministic CTM may make that effort redundant.**

**Authors:** Several studies in the Iberian Peninsula (IP) have addressed the causes of O<sub>3</sub> episodes looking at the circulation of air masses (Millán, 2014, and references therein). Specifically, recently, some studies found relevant to discriminate by synoptic situations for studying the phenomenology of summer O<sub>3</sub> over Catalunya (Querol et al., 2017), Madrid (Querol et al., 2018) and overall in whole Spain (Valverde et al., 2016).

As these authors claim, we believe that distinguishing by synoptic conditions is relevant to understand the O<sub>3</sub> origin in Spain. Stagnant conditions are characterized by weak synoptic winds, intense solar radiation, and the development of the Iberian Thermal Low, which forces the convergence of surface winds from the coast towards the central IP during the day and enhancing mesoscale process, which favours the accumulation of pollutants. In contrast,

northwestern advections transport air masses from the Atlantic towards the north and west of the IP, favouring the contribution of imported O<sub>3</sub> concentrations.

In addition, distinguishing by synoptic conditions allows assessing the performance of the deterministic model (WRF and CMAQ) to reproduce the synoptic transport and chemistry. For example the time series of source apportionment results in the northwest of the Iberian Peninsula (Fig. 9) show that:

Page 18 – line 16-22: *“The time series show that the model reproduces reasonably well the observed O<sub>3</sub> variability under different synoptic conditions. O<sub>3</sub> reaches the highest concentration (~100/150 µgm<sup>-3</sup> in urban/rural areas) under stagnant conditions (July 24<sup>th</sup>-27<sup>th</sup>) when the contribution of anthropogenic sources from all activity sectors is the highest (60-70%). O<sub>3</sub> concentrations decrease down to ~70 µgm<sup>-3</sup> under NW advective conditions (e.g., July 28<sup>th</sup>-30<sup>th</sup>) when the imported O<sub>3</sub> shows the highest contribution (80-90%). Saavedra et al. (2012) found that stationary anticyclones over the NWIP play an important role in the occurrence of high O<sub>3</sub> concentrations. Our results show that under these stagnant conditions O<sub>3</sub> concentrations are due largely to in situ production (photochemistry) from on-road traffic, shipping, power plants, and industry in almost the same proportion.”*

#### References:

Millán, M.M.: Extreme hydrometeorological events and climate change predictions in Europe, J. Hydrol., 518, 206-224, doi: 10.1016/j.jhydrol.2013.12.041, 2014.

Querol, X., Gangoiti, G., Mantilla, E., Alastuey, A., Minguillón, M.C., Amato, F., Reche, C., Viana, M., Moreno, T., Karanasiou, A., Rivas, I., Pérez, N., Ripoll, A., Brines, M., Ealo, M., Pandolfi, M., Lee, H.-K., Eun, H.-R., Park, Y.-H., Escudero, M., Beddows, D., Harrison, R.H., Bertrand, A., Marchand, N., Lyasota, A., Codina, B., Olid, M., Udina, M., Jiménez-Estevé, B., Soler, R.M., Alonso, L., Millán-M., and Ahn, K.-H.: Phenomenology of high ozone episodes in NE Spain, Atmos. Chem. Phys., 17, 2817-2838, doi: 10.5194/acp-17-2817-2017, 2017.

Querol, X., Alastuey, A., Gangoiti, G., Perez, N., Lee, H. K., Eun, H. R., Park, Y., Mantilla, E., Escudero, M., Titos, G., Alonso, L., Temime-Roussel, B., Marchand, N., Moreta, J. R., Revuelta, M. A., Salvador, P., Artíñano, B., García dos Santos, S., Anguas, M., Notario, A., Saiz-Lopez, A., Harrison, R. M., Millán, M., and Ahn, K.-H.: Phenomenology of summer ozone episodes over the Madrid Metropolitan Area, central Spain, Atmos. Chem. Phys., 18, 6511-6533, <https://doi.org/10.5194/acp-18-6511-2018>, 2018.

Valverde, V., Pay, M.T., and Baldasano, J.M.: Ozone attributed to Madrid and Barcelona on-road transport emissions: characterization of plume dynamics over the Iberian Peninsula, Sci. Total Environ. 543, 670–682, doi: 10.1016/j.scitotenv.2015.11.070, 2016.

**13. Reviewer #2: The results of the statistical evaluation for CMAQ outputs are summarized in Table 2. Some suggestions for this table:**

- Please include in the caption whether the exceedance column refers to observed or modelled values (the missing one may be also included either way)
- Two or three decimal points for r may be used
- MNB (%) may be more illustrative than MB (that can be derived directly from MM – MO)

- It may be misleading to pool together the statistics for different types of monitoring stations. As the authors discuss, it is arguable that outputs from a 4x4 km<sup>2</sup> model exercise should be compared against observations at traffic locations.
- It is unclear why some monitoring sites wouldn't fit into any of the categories considered. Please elaborate and state the rationale to include them in the analysis. It may be interesting to put these results into perspective by comparing them with those from other modelling exercises based on similar model suites in the IP or elsewhere (besides referring to previous applications of CALIOPE itself).

**Authors:** The revised version of the manuscript includes all the reviewer's suggestions regarding Table 2.

As the reviewer indicates, it is arguable that outputs from a 4x4 km<sup>2</sup> model exercise should be compared against observations at traffic locations. Actually, traffic station may not be representative of a 4x4 km<sup>2</sup> grid. Despite this limitation, we included traffic stations in our analysis discussing the model limitations as follows:

Page 11 – Line 27-28: *“Underestimation of NO<sub>2</sub> traffic peaks is a common problem in Eulerian mesoscale models (Pay et al., 2014), as emission heterogeneity is lost in the grid cell-averaging process, which is especially critical in urban areas”.*

The stations without a category corresponded to suburban background (SB) stations. The revised version of the manuscript includes now the SB category in both the discussion and Table 2. Note that now all the stations fit any of the five categories (i.e., IN, TR, UB, SB and RB), so the exceedance/station numbers in the IN/TR/UB/SB/RB rows do sum up to the numbers in the “ALL” rows.

It is difficult to compare the present evaluation results with other modelling studies because of the different period, domain, resolution, model setup, etc. However, we agree with the reviewer that it may be interesting to put these evaluation results into perspective. In this sense, we have added the following paragraph in the revised version of the manuscript:

Page 12 – Line 8-16: *“The comparison with previous CALIOPE studies (Baldasano et al., 2011; Pay et al., 2014) indicates that  $r$  is in the same range for O<sub>3</sub> (0.6-0.7) and NO<sub>2</sub> (0.4-0.5) at individual stations; the same applies to RMSE (15-29 µgO<sub>3</sub> m<sup>-3</sup> and 10-20 µgNO<sub>2</sub> m<sup>-3</sup>). Modelled O<sub>3</sub> shows higher performance at traffic stations in large cities, since stations influenced by road transport emissions (i.e., high-NO<sub>x</sub> environments) are better characterized with a more pronounced daily variability (Baldasano et al., 2011). At European scale, several model intercomparisons (Giornado et al. 2015; Bessagnet et al., 2016) indicate that O<sub>3</sub> concentrations in summer agree with the surface observations with  $r$  between 0.5 and 0.6. NO<sub>2</sub> hourly variability is overall underestimated due to uncertainties in the emission and meteorological modelling and model resolution. These studies highlight the limitations of models to simulate meteorological variables that affect the NO<sub>2</sub> hourly variability, and therefore the model performance for O<sub>3</sub> in high-NO<sub>x</sub> environments and downwind.”*

References:

Baldasano, J.M., Pay, M.T., Jorba, O., Gassó, S., Jiménez-Guerrero, P., 2011. An annual assessment of air quality with the CALIOPE modeling system over Spain. *Sci Total Environ*, 409, 2163-2178. Basart, S., Pérez, C., Nickovic, S., Cuevas, E., Baldasano, J.M., 2012. Development and evaluation of the BSC-DREAM8b dust regional model over Northern Africa,

the Mediterranean and the Middle East. *Tellus B* 64, 18539. <http://dx.doi.org/10.3402/tellusb.v64i0.18539>.

Bessagnet, B., Pirovano, G., Mircea, M., Cuvelier, C., Aulinger, A., Calori, G., Ciarelli, G., Manders, A., Stern, R., Tsyro, S., García Vivanco, M., Thunis, P., Pay, M.-T., Colette, A., Couvidat, F., Meleux, F., Rouïl, L., Ung, A., Aksoyoglu, S., Baldasano, J. M., Bieser, J., Briganti, G., Cappelletti, A., D'Isidoro, M., Finessi, S., Kranenburg, R., Silibello, C., Carnevale, C., Aas, W., Dupont, J.-C., Fagerli, H., Gonzalez, L., Menut, L., Prévôt, A. S. H., Roberts, P., and White, L.: Presentation of the EURODELTA III intercomparison exercise – evaluation of the chemistry transport models' performance on criteria pollutants and joint analysis with meteorology, *Atmos. Chem. Phys.*, 16, 12667–12701, doi:10.5194/acp-16-12667-2016, 2016

Giordano, L., Brunner, D., Flemming, J., et al., 2015. Assessment of the MACC reanalysis and its influence as chemical boundary conditions for regional air quality modeling in AQMEII-2. *Atmos. Environ.*, 115, 371-388.

Pay, M.T., Martínez, F., Guevara, M., Baldasano, J.M., 2014. Air quality at kilometre scale grid over Spanish complex terrains. *Geosci. Model Dev.* 7, 1979–1999. <http://dx.doi.org/10.5194/gmd-7-1979-2014>.

**14. Reviewer #2: P11.L10-14. The description of the performance-based categories is hard to follow and it is already condensed in Fig. 5. Please, simplify or simply remove that passage. I'd suggest the authors to re-compute statistics and assessment with the alternative BVOC emissions model run mentioned earlier.**

**Authors:** We agree with the reviewer. In the revised version of the manuscript, we have removed that passage on the description of the model performance based on bias categories as it is condensed in Fig. 5.

As explained before, we did not perform the BVOC emission sensitivity test with MEGAN comparing v2.0.4 and v2.1, because evaluation of MEGAN with updated emission factors indicates that modelled isoprene concentrations are in reasonably good agreement with observations. We foresee this BVOC emission sensitivity test and its effect on O<sub>3</sub> during the upgrade of our modelling system with MEGANv3 (<https://bai.ess.uci.edu/megan/versions/megan3>).

**15. Reviewer #2: If mention to specific cities or areas is made (e.g. Ebro Valley, Lleida Plain), please identify them in any of the maps in the manuscript.**

**Authors:** We have improved Figure 2 to identify the cities and areas that appear throughout the manuscript.

**16. Reviewer #2: The information summarized in Fig. 7 is interesting although the concept of “receptor regions” is unclear. It seems reasonable in terms of geographical location but differences regarding contributions from different sectors are not evident, especially if the results are put into the perspective of the typical uncertainties of**

**modelling exercises that can be inferred from the model evaluation previously presented.**

**Authors:** The O<sub>3</sub> receptors are defined as air quality stations located in regions with similar meteorological and O<sub>3</sub> patterns, main source contributors and geographical patterns. The O<sub>3</sub> receptor regions defined in this work are consistent with Diéguez et al. (2014) and Querol et al. (2016), who proposed a similar regionalization based only on observations from air quality stations.

Page 14 – Line 4-6: *“We have identified ten O<sub>3</sub> receptor regions with similar characteristics in terms of meteorological and geographical patterns, O<sub>3</sub> dynamics and main source contributors (Figs. 4 and 6). The receptor regions defined in our work are consistent with Diéguez et al. (2014) and Querol et al. (2016), who proposed a similar regionalization based on observations from air quality stations.”*

Fig. 7 shows the contribution from different sectors by O<sub>3</sub> receptor region ordered by decreasing concentration of imported O<sub>3</sub>. Differences between sectors are more evident in the normalized contribution (Fig. 7b).

We agree with the referee that the results of the source apportionment have an associated uncertainty. However, this uncertainty cannot be precisely quantified because of the lack of apportioned O<sub>3</sub> observations. We note that in sect 3.4 we extensively evaluate and discuss our results using O<sub>3</sub>, NO<sub>2</sub> and wind speed and direction observations. We show that for some cases we can clearly identify a particular sector to be responsible for the O<sub>3</sub> mismatch.

**17. Reviewer #2: It is interesting noting a relatively large contribution from the SNAP 8 sector. The share of mobile sources is particularly important in the SIP area, which would be consistent with the discussion regarding the influence of shipping. However, other areas such as GV or even CIP present a non-negligible contribution. Could the authors elaborate on that?**

**Authors:** The SNAP8 sector accounts for international shipping, airport service and agricultural machinery. The O<sub>3</sub> contribution from non-road transport in the central of the Iberian Peninsula may arise from the international shipping routes, Madrid’s airports and the agricultural machinery operating in the surrounding rural areas. The current study has not distinguished these subsectors but it maybe useful for future source apportionment studies. We have clarified the definition of the SNAP8 sector and its contribution in the CIP as follows:

Table 1: *“SNAP8: Non-road transport (international shipping, airport and agricultural machinery)”*

Page 16- Line 18-20: *“The O<sub>3</sub> contribution from non-road transport in this region may arise from the Madrid’s airports and the agricultural machinery operating in the surrounding rural areas mainly.”*

Page 20 – Line 28-29: *“The non-road transport sector (including international shipping, airport and agricultural machinery) is as significant as road transport inland (10-19% of the daily mean O<sub>3</sub> concentration during the peaks).”*

**18. Reviewer #2: Maybe the discussion in section 3.4 is too profuse and should be substantially shortened. I encourage the authors to summary here their findings and**

provide the region-by-region discussion as supplementary material, including Fig. 8 and Fig. 9. Oppositely, the rationale for the station sub-set selection may deserve further explanation. In general, the section is abundant in hypotheses and subjective interpretation that are not clearly supported by evidence. Personally, I don't think this contribution really benefits from such approach. The paper may be restricted to a more solid and consistent analysis and discussion of the findings from the application of CMAQ-ISAM. That is novel and interesting enough and further attempts to relate the results with detailed regional dynamics and atmospheric patterns may very well be addressed in future specific studies (using more specific methods and data, e.g. better resolved emission inventories).

**Authors:** We have summarized that part following the reviewer's suggestion. Please, have a look at the revised version of the manuscript.

However, we have kept the region-by-region discussion because the main purpose of the present study is the estimation of the contribution of the anthropogenic activity sectors and the imported concentration to peak O<sub>3</sub> events in Spain. In addition, this region-by-region discussion has provided a perspective about the potential use of source apportionment analysis for improving the O<sub>3</sub> modelling and designing future mitigation strategies at regions with a high on-road traffic contribution (i.e., CIP and NEIP in Fig. 8) and a high contribution from industry and energy production (i.e., NWIP and Guadalquivir Valley in Fig. 9).

**19. Reviewer #2:** The authors claim that the modelling exercise presented allowed an in-depth evaluation of the modelling system applied. This relates to my last comment regarding the results section. The paper may lack a well-defined objective and presents a huge amount of information without a clear purpose. For example, if the main interest was to assess the modelling system capabilities and identify options for improvement, the results and analysis should gravitate towards a more detailed statistical analysis within a better defined methodological framework. I acknowledge a valuable study but I think the authors should revise their manuscript under a clearly defined scientific question avoiding an excessive spread in their discussion that may lead to inconsequential or cursory analyses and reflect that also in this section. As for the discussion on model uncertainty I find particularly important to take into account the observations regarding biogenic emissions, although the mismatch between emissions and meteorological conditions may also hinder the discriminating power of the results.

**Authors:** We have rewritten the objective in the abstract and the introduction to clarify the objective of the study as follows:

In the abstract (*Page 1 – Line 19-21*):

*“The main goal of this study is to provide a first quantitative estimation of the contribution of the main anthropogenic activity sectors to peak O<sub>3</sub> events in Spain relative to the contribution of imported (regional and hemispheric) O<sub>3</sub>. We also assess the potential of our source apportionment method to improve O<sub>3</sub> modelling”*

In the introduction (*Page 4 – Line 25-30*):

*“The integrated source apportionment tools combined with high-resolution emission and meteorological models can help unravelling the sources responsible for peak summer events of O<sub>3</sub> in the Western Mediterranean Basin. Quantifying the contribution of emission sources during acute O<sub>3</sub> episodes is a prerequisite for the design of future mitigation strategies in the region. In this framework, the main goal of this study is to provide a first quantitative estimation of the contribution of the main anthropogenic*

*activity sectors compared to the imported concentration (regional and hemispheric) to peak O<sub>3</sub> events in Spain. We also assess the potential of our source apportionment method to improve O<sub>3</sub> modelling”*

Under clearly defined scientific objectives, we have substantially shortened Section 4 to avoid an excessive spread in the discussion. A new version of Section 4 is available in the revised manuscript.

As suggested by the reviewer we have included a comment on the uncertainty of biogenic emissions and emissions reference year as follows:

*Page 22 – Line 10-12: “Another relevant and uncertain source for O<sub>3</sub> concentration is the urban VOC emissions. Future research works should be devoted to continuous monitoring of urban VOC and take advantage of satellite observations to improve speciation and spatial variability of urban VOC emissions.”*

**20. Reviewer #2: In any case, caution should apply since the timespan of the period analysed makes it difficult to extract general conclusions. This is particularly important for the regulatory implications that may be derived from this study. As the authors conclude, I find reasonable to base recommendations for abatement strategies in more specific, regional scale, detailed analyses. Consequently, I'd keep such conclusions to a minimum in this contribution.**

**Authors:** As discussed before, we agree that the representativeness of our results is limited because they are focused in just one episode. Future studies should preferentially cover multiple summer periods in order to improve representativeness.

Although, the main goal of this study is not the design abatement policies, these source apportionment results has provide a perspective about the potential use of these methods for regulatory studies in non-attainment regions as a prerequisite for the design of future mitigation strategies. We have added a short comment on this in the discussion section:

*Page 23-Line 4-7: “This work has quantified the local and imported contributions to O<sub>3</sub> during an episode in a particular area in southwestern Europe. In addition, we have provided a perspective about the potential use of source apportionment method for regulatory studies in non-attainment regions. Further O<sub>3</sub> source apportionment studies targeting other nonattainment regions in Europe are necessary prior to design local mitigation measures that complement national and European-wide abatement efforts.”*

**21. Reviewer #2: Please revise equations 1 to 4 for a better readability**

**Authors:** We have increased the size of the font in the equation for better readability.

**22. Reviewer #2: P7.Line 4: SNAP3 and 4**

**Authors:** We note that SNAP34 is not a typo error. We have defined sector SNAP34 all together as mentioned in Table 1 as manufacturing industries. We follow the same reporting approach as the one proposed by the TNO\_MACC emission inventories, in which SNAP 3

and SNAP 4 emissions are merged to SNAP 34. This does not mean that we deal with SNAP3 and SNAP4 emissions all together. Emissions from each point source are estimated individually and applying specific activity and emission factors, as well as speciation and temporal profiles. It is just a matter of reporting format.

**23. Reviewer #2: P8.Lines 12-13: the brackets are not needed: “In Spain, around 60% of the annual exceedances also occurred during this period. (As shown in Querol et al. (2016) July is typically the month with the highest number of O<sub>3</sub> exceedances in Spain.) The”**

**Authors:** We have removed this statement for simplicity.

**24. Reviewer #2: P11.Line 11: “for 93% of the stations” instead of “for the 93% of stations”**

**Authors:** We have amended this issue in the revised version of the manuscript.

**25. Reviewer #2: P12.Line 2: “extremely low winds”?**

**Authors:** We have rewritten this statement as follows “stagnant conditions”.

**26. Reviewer #2: P20.Line 8: “O3at” is missing a space**

**Authors:** We have amended this issue in the revised version of the manuscript.

**27. Reviewer #2: P34.Line 5 (Fig. 2 caption): I guess the authors mean “Number of stations” instead of “Number of days”**

**Authors:** we have rewritten the caption of Figure 2 following the reviewer’s suggestions as follows:

Figure 2-caption: “*Number of the Spanish EIONET stations days exceeding the O<sub>3</sub> Target Value (120 µg/m<sup>3</sup>) per episode day*”

**Interactive comment on “Ozone source apportionment during peak summer events over southwestern Europe” by Maria Teresa Pay et al.**

**Anonymous Referee #3**

Received and published: 29 December 2018

**1. Reviewer #3: The paper gives important contribution to address the source apportionment study regarding ozone episode occurred in Spain. The paper is well structured and presents a complete analysis of the modelling results. However, there are some major points that should be addressed before recommended for publication. Besides that, English should be revised along the manuscript, there is some inconsistencies and grammatical errors. See below major and minor comments.**

**Authors:** We thank the reviewer #3 for the comments and suggestions for improvement. We have corrected errors and omissions, and introduced as much as possible the reviewer's suggestions.

Please, find below the item-by-item response. For more details on the review process, we have uploaded the manuscript with track-changes.

**2. Reviewer #3: Abstract; Line 15 (Page 4/Line 2): there is recently studies that showed that source-apportionment methods are not adequate to investigate plans and mitigation measures, in particular for non-linear pollutants like ozone, and that for that purpose “scenario analysis” based on “brute-force” are recommended. Authors should revise the text along the manuscript where it is mentioned the purpose of “designing plans”, which should not be the final objective of this source-apportionment study. (see Clappier, A., Belis, C., Pernigotti, D., Thunis, P. Source apportionment and sensitivity analysis: two methodologies with two different purposes. *Geosci. Model Dev. Discuss.* 10, 4245-4256 (2017))**

**Authors:** We totally agree with the reviewer's point of view and with the content of Clappier et al., (2017). The main goal of this study is not the design abatement policies, but it is to provide a first quantitative estimation of the contribution of the main anthropogenic activity sectors to peak O<sub>3</sub> events in Spain relative to the contribution of imported O<sub>3</sub>.

Actually, source apportionment techniques alone cannot be used to the design abatement policies. Subsequent source sensitivity analyses tailoring the identified main contribution sources could predict how O<sub>3</sub> will respond to reductions in precursor emissions. Both, source apportionment and source sensitivity are complementary and essential studies to define the most efficient O<sub>3</sub> abatement strategies in the Western Mediterranean Basin. The manuscript highlights now this idea in different sections that we recap as follows:

*Page 4 – Line 26-30: “Quantifying the contribution of emission sources during acute O<sub>3</sub> episodes is a prerequisite for the design of future mitigation strategies in the region. In this framework, the main goal of this study is to provide a first quantitative estimation of the contribution of the main anthropogenic activity sectors compared to the imported concentration (regional and hemispheric) to peak O<sub>3</sub> events in Spain.”*

*Page 22–Line 23-27: “We note that our results cannot predict whether emission abatement will have either a positive or negative effect in O<sub>3</sub> changes due to the non-linearity of the O<sub>3</sub> generation process. Subsequent source sensitivity analyses tailoring the identified main contribution sources could predict how O<sub>3</sub> will respond to reductions in precursor emissions, which are essential to define the most efficient O<sub>3</sub> abatement strategies in the Western Mediterranean Basin.”*

*Page 23-Line 4-7: "This work has quantified the local and imported contributions to O<sub>3</sub> during an episode in a particular area in southwestern Europe. In addition, we have provided a perspective about the potential use of source apportionment method for regulatory studies in non-attainment regions. Further O<sub>3</sub> source apportionment studies targeting other nonattainment regions in Europe are necessary prior to design local mitigation measures that complement national and European-wide abatement efforts."*

In order to be clear in the source apportionment applications and limitations of certain methods, we have added in the revised version of the manuscript a comment taking into account the recent findings in Clappier et al. (2017) as follows:

*Page 4-Line 7-11: "Brute force is simple to implement, as it does not require additional coding in the CTM. However, as it quantifies the contribution of each source based on its absence, it does not reproduce actual atmospheric conditions, and therefore it is susceptible to inaccuracies in the prediction of O<sub>3</sub> peaks under non-linear regimes (Cohan and Napelenok, 2011). Actually, brute force is not suitable to retrieve source contribution when the relationship between emissions and concentration is non-linear, but it is useful for analysing the concentration responses to emission abatement scenarios (Clappier et al., 2017)."*

**3. Reviewer #3: Page 4/Line 7-9: please review this sentence according to what has been commented before.**

**Authors:** The statement in Page 4/Line 7-9 has been expanded following the reviewer's suggestion as shown in the previous authors' answer (see Page 4-Line 9-13).

**4. Reviewer #3: Page 5/Line 26: the authors should comment about the representativeness of the 2009 emissions to the 2012 SA study presented. From 2009 to 2012 several changes happened in society and economy which was reflected in emission data.**

**Authors:** Our methodological choice has been to use a detailed bottom-up emission inventory instead of a typical top-down regional emission inventory. Bottom-up emissions, estimated using source-specific emission factors and activity statistics, accurately characterise pollutant sources and allow obtaining more realistic results than the ones reported by top-down or regional emission inventories. However, they require very large efforts to be compiled and consequently the updating processes cannot be implemented year-to-year.

In HERMESv2 emissions are based on 2009 data, which was the closest year with updated information on local emission activities at the time this work started.

To understand the impact of the use of 2009 data to study year 2012 we revised the EMEP Centre on Emission Inventories and Projections (EMEP-CEIP), which collects and reviews the national emission inventories from Parties to the Convention on Long-range Transboundary Air Pollution (CLRTAP). Between 2009 and 2012 total NO<sub>x</sub> and NMVOC emissions in Spain decreased by -10.6% and -10.7%, respectively (EMEP CEIP, 2019). For NO<sub>x</sub>, around 80% of this reduction is linked to a reduction of road transport emissions, whereas in the case of NMVOC ~50% of the reduction is due to a decrease of industrial emissions. NO<sub>x</sub> emissions from shipping in Europe have also decreased in the period 2009-2012 by 15%.

For our modelling study, we consider these differences as small and acceptable, and not creating any major inconsistency. The difference of 10-15 % in emissions for certain precursors between 2009 and 2012 is within the typically larger ranges of uncertainty in emission inventories. We also note that all our results are thoroughly evaluated and critically assessed using observations.

In any case, we have followed the reviewer's suggestion, and we have discussed in the manuscript the potential impact of these differences when the contribution of each emission sector is analysed:

*Page 17–Line 32-33: “[...] This factor, added to the 15% decrease of NO<sub>x</sub> shipping emissions observed in Europe between 2009 (HERMESv2.0 base year) and 2012 (EMEP CEIP, 2019) could explain the discrepancies observed.”*

*Page 18–Line 12-14: “[...] it has been estimated that between 2009 and 2012 energy production in coal-fired power plants increased from 13.1% to 19.4% (UNESA, 2012), which implied an increase of NO<sub>x</sub> emissions from the power industry sector of around 19.5% (EMEP CEIP, 2019).”*

The Section 4 of the revised version of the manuscript includes now a comment on the methodological implication of using 2009 emissions for O<sub>3</sub> source apportionment studies in an episode in 2012 as follows:

*Page 5-Line 31-32: “HERMESv2.0 is currently based on 2009 data, which is the closest year with updated information on local emission activities at the time this work started.”*

*Page 21-Line 31: “Our methodological choice has been to use a detailed bottom-up emission inventory instead of a typical top-down regional emission inventory. Bottom-up emissions, estimated using source-specific emission factors and activity statistics, accurately characterise pollutant sources and allow obtaining more realistic results than the ones reported by top-down or regional emission inventories. To understand the impact of the use of 2009 data to study year 2012, we revised the EMEP Centre on Emission Inventories and Projections (EMEP-CEIP), which collects and reviews the national emission inventories from Parties to the Convention on Long-range Transboundary Air Pollution. Between 2009 and 2012, total NO<sub>x</sub> and NMVOC emissions in Spain decreased by -10.6% and -10.7%, respectively (EMEP CEIP, 2019). For NO<sub>x</sub>, around 80% of this reduction is linked to a reduction of road transport emissions, whereas in the case of NMVOC ~50% of the reduction is due to a decrease of industrial emissions. For our modelling study, we consider these differences as small and acceptable, and not creating any major inconsistency. The difference of 10-15 % in emissions for certain precursors between 2009 and 2012 is within the typically larger ranges of uncertainty in emission inventories.”*

Reference:

EMEP CEIP, 2019. Officially reported emission data. Available at: [http://www.ceip.at/ms/ceip\\_home1/ceip\\_home/data\\_viewers/official\\_tableau/](http://www.ceip.at/ms/ceip_home1/ceip_home/data_viewers/official_tableau/) (last access February 2019)

**5. Reviewer #3: Page 7, Line 9: SNAP2 activity can be a particular important source for ozone precursors. Authors should comment about that when they mentioned that SNAP2 is aggregated with other activity sectors.**

**Authors:** One limitation of the current version of CMAQ-ISAMv5.0.2 is that the number of tagged sources increases the computational time. In the current version of CMAQv5.0.2 the increase of the computational resources does not decrease the computational time. A more computationally efficient version of the ISAM will be released with the final version of CMAQv5.3 in Spring 2019. Based on that limitation, we configured our first study tagging the energy, industrial, road transport and non-road transport sectors (Fig. 1b), which account for 92% of the total NO<sub>x</sub> emissions in Spain. The remaining emission sectors are lumped in the OTHER tag. This selection criterion is explained in the manuscript as follows:

*Page 7- Line 20-23: “The selected (tagged) SNAP categories in this study are the energy, industrial, road transport and non-road transport sectors (Fig. 1b), which account for the 92% of the total NO<sub>x</sub> emissions in Spain. An additional tracer (OTHR) gathers the remaining emission categories that were not explicitly tracked (i.e., SNAP2, 5, 6, 9, 10 and 11).”*

We have seen in other studies performed over different domains (e.g., Portugal) that SNAP2 can be an important contributor to O<sub>3</sub> precursors. We have mentioned this fact in the manuscript as follows:

*From Page 13- Line 33 to Page 14-Line 2: “The high OTHR concentration around the biggest cities in Portugal may be related to precursors emitted by the residential sector (SNAP2 and 9) and biogenic emissions, as found in other source apportionment studies over Portugal (Borrego et al., 2016; Karamchandani et al., 2017).”*

**6. Reviewer #3: Page 7, Lines 25-30: in the scope of FAIRMODE – Forum for Air Quality Modelling in Europe – tools were developed, namely the DELTA-Tool, to evaluate air quality models and conclude about their suitability to be used for legislation purposes. The authors should consider the application of this tool to evaluate model performance instead of calculating the traditional statistical indicators. In any case, authors should justify why they decided not using this tool.**

**Authors:** The evaluation with the Delta tool has been taken into account in previous evaluation studies of the CALIOPE system, although it has not been shown in the manuscript. We have added this comment to complement this information:

*Page 11-Line 2-6: “CALIOPE has been evaluated in detail elsewhere (Pay et al., 2014 and references therein). Furthermore, the system has been evaluated using the tool developed by the Forum for Air Quality Modelling in Europe, named DELTA-Tool (Thunis and Cuvelier, 2014) to support and harmonize the model evaluation in the frame of the Air Quality Directive. Valverde et al. (2016a; 2016b) used the DELTA-Tool v4.0 and showed that the CALIOPE system accomplishes the quality objectives as defined in the Air Quality Directive for 78% to 91% of the NO<sub>2</sub> and O<sub>3</sub> monitoring stations in summer conditions in 2012.”*

References:

Thunis, P., Cuvelier, C., 2014. DELTA Version 4.0. Joint Research Center, Ispra ([http://aqm.jrc.ec.europa.eu/DELTA/assessment/data/DELTA\\_UserGuide\\_V4\\_0.pdf](http://aqm.jrc.ec.europa.eu/DELTA/assessment/data/DELTA_UserGuide_V4_0.pdf)).

Valverde, V., Pay, M.T., and Baldasano, J.M.: Ozone attributed to Madrid and Barcelona on-road transport emissions: characterization of plume dynamics over the Iberian Peninsula, Sci. Total Environ. 543, 670–682, doi: 10.1016/j.scitotenv.2015.11.070, 2016a.

Valverde, V., Pay, M.T., and Baldasano, J.M.: A model-based analysis of SO and NO dynamics from coal-fired power plants under representative synoptic circulation types over the Iberian Peninsula, *Sci. Total Environ.*, 541, 701-713, doi: 10.1016/j.scitotenv.2015.09.111, 2016b.

**7. Review #3: Page 13, Lines 16-17: this information (model performance in terms of O<sub>3</sub> peaks) should be presented and discussed in the model validation section.**

**Authors:** The information provided in these lines is not a dedicated evaluation analysis on O<sub>3</sub> peaks. It is devoted to provide a general perspective on source apportionment results shown in Figure 7. In order to make it clear, we have rewritten this sentence as follows:

Page 14-Line 15-16: *“Figure 7 indicates that during exceedances of the MDA8 target value there is a good agreement ( $r = 0.79$ ) between the sum of apportioned O<sub>3</sub> and the observed concentrations over the receptor regions.”*

**8. Reviewer #3: Page 13, Line 22: this sentence should be completed with information about the area where this impact (up to 8%) is verified.**

**Authors:** The reviewer is right. We have rewritten the sentence as follows:

Page 14 – Line 20-21: *“Shipping emissions in the MED region contributed up to 8% of the total O<sub>3</sub>”.*

**9. Reviewer #3: Page 15, Lines 4-5: The authors should quantify how “model reproduces reasonably well”**

**Authors:** We agree with the reviewer and we have quantified the model performance at this station as follows:

Page 15-Line 33: *“At the urban station, the model reproduces the O<sub>3</sub> traffic cycle ( $r = 0.66$  and  $MB=22.5 \mu\text{gm}^{-3}$ ) featuring the typical low O<sub>3</sub> concentrations ( $< 40 \mu\text{gm}^{-3}$ ) in the early morning and in the afternoon due to O<sub>3</sub> titration (Fig. 8a).”*

**10. Reviewer #3: Page 15, Lines 9-10: Please clarify the sentence “The NO<sub>2</sub> overestimation correlates with the highest road transport contribution”. Page 15, Lines 11-12: Please explain why: “The results point towards a poor representation of the vertical mixing during the stagnant conditions”**

**Authors:** In order to clarify this sentence, we have rewritten this paragraph as follows:

Page 16-Line 2-5: *“O<sub>3</sub> was overestimated (MB type D) during daytime peaks due to the overestimation of the NO<sub>2</sub> morning peaks during stagnant conditions, coincident with the highest road transport contribution for both pollutants. The results point towards a poor representation of the meteorological condition in the city during the stagnant conditions as shown in the meteorological evaluation (Sect. 4 in the supplement)”*

**11. Reviewer #3: Page 16, Lines 4-10: this should be placed in the model validation section.**

**Authors:** We agree with the reviewer that evaluation results should be addressed in the dedicated section. However, the interpretation of the source apportionment results benefits from model evaluation, and at the same time the source apportionment results support enhanced model evaluation. We have added a comment on this in the Section 2.4 Evaluation method to clarify this issue as follow:

Page 8-Line 24-27: *“Evaluation results are discussed together with the source apportionment results. On the one side, the interpretation of the source apportionment results benefits from model evaluation. On the other side, the source apportionment results support enhanced model evaluation as it allows identifying potential errors in emission estimates for specific sectors and/or in the chemical boundary conditions.”*

**12. Reviewer #3: Page 16, Lines 13-14: how can the authors conclude that the model is able to reproduce all these different processes? Can the authors support better this statement?**

**Authors:** As shown in Fig. 8d, modelled O<sub>3</sub> peaks (> 120 µg m<sup>-3</sup>) are in a good agreement with observations (Fig. 8d), which suggests that overall the model reproduces the main transport paths, photochemical processes, and relative contributions from different sources. We have rewritten this statement as follows:

Page 16-Line 32: *“At the rural station, modelled O<sub>3</sub> peaks (> 120 µg m<sup>-3</sup>) are in a good agreement with observations (Fig. 8d), which suggests that overall the model reproduces the main transport paths, photochemical processes, and relative contributions from different sources.”*

**13. Reviewer #3: Page 18, Line 4: since only one rural station is analysed, the authors should not generalize as “In rural background areas”**

**Authors:** The reviewer is right; we have rewritten the sentence accordingly:

Page 18-Line 8-9: *“Despite the O<sub>3</sub> biases during stagnant conditions, the modelled O<sub>3</sub> concentration is in general agreement with observations at the rural background station (Fig. 9d)”*

**14. Reviewer #3: Page 18, Lines 30-33 to Page 19: the authors analyse the vertical profile in a single point, but this will be not representativeness of the all study domain. Authors should change the text according to this limitation and comment it, or increase the number of points analysed.**

**Authors:** The Figure 10 does not show a vertical profile, but it shows the O<sub>3</sub> distribution over a vertical cross section crossing the IP from the west to the east at the centre of the domain (approximately at a latitude of 40.38° N). Although Figure 10 is not representative of the whole

domain of study, it helps to understand the vertical variability of both pollutants with the PBL dynamics as schematized by Millán et al. (1996). We have clarified this point in the manuscript as follows:

Page 19-Line 1-3: *“Figure 10 shows the vertical cross-sections at 6, 12, and 18 UTC for O<sub>3</sub> and NO<sub>2</sub> at a constant latitude (40.38° N) on July 25th, 28th and 30th. It helps to understand the vertical variability of both pollutants according to the dynamics of the PBL as schematized by Millán et al. (1996).”*

**15. Reviewer #3: Figure 2: please review the figure caption “Number of days exceeding the O<sub>3</sub> target value (120 ug.m-3) by each day of the episode”**

**Authors:** We have rewritten the caption of Figure 2 following the reviewer’s suggestions as follows:

Figure 2-caption: *“Number of stations exceeding the O<sub>3</sub> Target Value (120 µg/m<sup>3</sup>) per episode day”*

Note that in the reviewed version this figure corresponds to the Figure S3 in the Supplement.

**16. Review #3: Page 1/Line 20: write 4x4 km<sup>2</sup> instead of 4x4 km (please correct this along the manuscript)**

**Authors:** We have amended that.

**17. Reviewer #3: Authors should refer the modelling system (CALIOPE) in the abstract.**

**Authors:** The suggestion of the reviewer has been included in the new version of the manuscript as follows:

*Abstract (Page 1-Line 21-24): “Our study applies and thoroughly evaluates a countrywide O<sub>3</sub> source apportionment method implemented in the CALIOPE air quality forecast system for Spain at high resolution (4 x 4 km<sup>2</sup>) to understand and quantify the origin of peak O<sub>3</sub> events over a 10-day period covering the most frequent synoptic summer conditions in the Iberian Peninsula.”*

**18. Reviewer #3: Page 3, Line 4: The following reference should be added, since it is the biggest ozone episode occurred in IP region: “Monteiro A., Gama C., Candido M., Ribeiro I., Lopes M. (2016) Investigating ozone high levels and the role of sea breeze on its transport. Atmospheric Pollution Research 7, 339-347.**

**Authors:** We have added Monteiro et al. (2016) to the list of studies as an example of a large ozone episode occurred in IP region.

**19. Reviewer #3: Page 7, Line 23: please indicate how many stations measure both O<sub>3</sub> and NO<sub>2</sub> pollutants.**

**Authors:** We have mentioned this information latter on when we introduce the NO<sub>2</sub>/O<sub>3</sub> source apportionment time series:

Page 8-Line 20-22: *“This work will only discuss in detail the source apportionment plots at key O<sub>3</sub> receptor regions, given the high number of stations (260) that simultaneously measure O<sub>3</sub> and NO<sub>2</sub>.”*

**20. Reviewer #3: Page 10, Lines 17, 24, 29: please add “average” when mentioning “hourly O<sub>3</sub>” (the values presented are an average of different locations and not an “hourly O<sub>3</sub> data”**

**Authors:** The suggestion of adding “average” when mentioning “hourly concentration” has been implemented in the whole manuscript.

**21. Reviewer #3:- Page 12, Lines 2-7: the following reference should be added to support this part: Borrego C., Monteiro A., Martins H., Ferreira J., Fernandes A.P., Rafael S., Miranda A.I., Guevara M., Baldasano J.M. (2016). Air quality plan for ozone: a case-study for North Portugal. Air Quality, Atmosphere & Health 9 (5), 447–460.**

**Authors:** The reviewed version of the manuscript includes the reference Borrego et al., (2016) over northern Portugal to support the fact that, under stagnant conditions, imported O<sub>3</sub> is depleted and O<sub>3</sub> photochemical production is enhanced around the largest industrial/urban areas.

Page 13-Line 4-6: *“In a source attribution study over northern Portugal, Borrego et al. (2016) also found a reduction of imported O<sub>3</sub> and the subsequent O<sub>3</sub> formation by local sources under similar meteorological conditions.”*

**22. Reviewer #3: Page 14, Line 30: please review the English.**

**Authors:** The reviewed version of the manuscript has rewritten this paragraph as follows:

Page 15-Line 28-29: *“The following sections analyse the source apportionment results at key regions (see Fig. 5) with a high on-road traffic contribution (i.e., CIP and NEIP) and a high contribution from industry and energy production (i.e., NWIP and Guadalquivir Valley).”*

**23. Reviewer #3: Page 17, Lines 19-24: authors should consult and use the following reference that compares the different shipping emission inventories mentioned: Russo M.A., Leitão J., Gama C., Ferreira J., Borrego C., Monteiro A. (2018) Shipping emissions over Europe: a state-of-the-art and comparative analysis. Atmospheric Environment 177, 187–194.**

**Authors:** The reviewed version of the manuscript include the reference Russo et al. (2018) to discuss the uncertainties on shipping emissions over Europe:

Page 17-Line 28-32: “A recent review on the state-of-the-art of marine traffic emissions (Russo et al., 2018) indicates that STEAM appears as the most reliable and detailed emissions inventory since it is based on Automatic Identification System data and specific vessel information, with a resolution of  $2.5 \times 2.5 \text{ km}^2$  (Jalkanen et al., 2016). A comparative analysis indicates that EMEP gridded inventories are overestimated, in particular over hotspots in the Mediterranean shipping routes, and underestimated in secondary routes.”

References:

Jalkanen, J.-P., Johansson, L., and Kukkonen, J.: A comprehensive inventory of ship traffic exhaust emissions in the European sea areas in 2011, *Atmos. Chem. Phys.*, 16, 71-84, <https://doi.org/10.5194/acp-16-71-2016>, 2016.

Russo, M. A., Leitão, J., Gama, C., Ferreira, J., & Monteiro, A.: Shipping emissions over Europe: A state-of-the-art and comparative analysis. *Atmospheric Environment*, 177, 187-194, 2018.

**24. Reviewer #3: Page 18, Line 22: please replace “These O3 ...” by “The results presented before ...”**

**Authors:** We have amended that the suggestion.

**25. Reviewer #3: Figure 3: Please review the units used along the manuscript, like “m/s”**

**Authors:** We have harmonize the use of units as  $\text{ms}^{-1}$ .

# Ozone source apportionment during peak summer events over southwestern Europe

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**Abstract.** It is well established that in Europe, high O<sub>3</sub> concentrations are most pronounced in southern/Mediterranean countries due to the more favorable climatological conditions for its formation. However, the contribution of the different sources of precursors to O<sub>3</sub> formation within each country relative to the imported (regional and hemispheric) O<sub>3</sub> is poorly quantified. This lack of quantitative knowledge prevents local authorities from effectively designing plans that reduce the exceedances of the O<sub>3</sub> Target Value set by the European Air Quality Directive. O<sub>3</sub> source attribution is a challenge because the concentration at each location and time results not only from local biogenic and anthropogenic precursors, but also from the transport of O<sub>3</sub> and precursors from neighbouring regions, O<sub>3</sub> regional and hemispheric transport and stratospheric O<sub>3</sub> injections. [The main goal of this study is to provide a first quantitative estimation of the contribution of the main anthropogenic activity sectors to peak O<sub>3</sub> events in Spain relative to the contribution of imported \(regional and hemispheric\) O<sub>3</sub>. We also assess the potential of our source apportionment method to improve O<sub>3</sub>.](#) Our study applies and thoroughly evaluates a countrywide O<sub>3</sub> source apportionment method implemented in [a Chemical Transport Model \(CTM\) the CALIOPE air quality forecast system for Spain](#) at high resolution (4 x 4 km<sup>2</sup>) [over the Iberian Peninsula \(IP\) to understand and quantify the origin of peak O<sub>3</sub> events](#) over a 10-day period [covering the most a frequent synoptic characterized by typical summer conditions in the Iberian Peninsula \(IP\).](#) The method tags both O<sub>3</sub> and its gas precursor emissions from source sectors within one simulation and each tagged species is subject to the typical physical processes (advection, vertical mixing, deposition, emission and chemistry) as the actual conditions remain unperturbed. We quantify the individual contributions of the largest NO<sub>x</sub> local sources to high O<sub>3</sub> concentrations compared to the contribution of imported O<sub>3</sub>. We show for the first time that imported O<sub>3</sub> is the largest input to the ground-level O<sub>3</sub> concentration in the IP, accounting for 46% to 68 % of the daily mean O<sub>3</sub> concentration during exceedances of the European Target Value. The hourly imported O<sub>3</sub> increases during typical northwestern advections (70-90%, 60-80 µg/m<sup>3</sup>), and decreases during typical stagnant conditions (30-40%, 30-60 µg/m<sup>3</sup>) due to the local NO titration [effect](#). During stagnant conditions, the [local](#) anthropogenic precursors control the O<sub>3</sub> peaks in areas downwind of the main urban and industrial regions (up to 40% in hourly peaks). We also show that ground-

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level O<sub>3</sub> concentrations are strongly affected by vertical mixing of O<sub>3</sub>-rich layers present in the free troposphere, which result from local/regional layering and accumulation, and continental/hemispheric transport. Indeed, vertical mixing largely explains the presence of imported O<sub>3</sub> at ground level in the [Iberian Peninsula](#). Our results demonstrate the need for detailed quantification of the local and remote contributions to high O<sub>3</sub> concentrations for local O<sub>3</sub> management, being the O<sub>3</sub> source apportionment an essential analysis prior to the design of O<sub>3</sub> mitigation plans in any non-attainment area. ~~To Achieving~~ the European O<sub>3</sub> objectives in ~~the southern Europe and after the estimated relative importance of the imported O<sub>3</sub> to local generation from anthropogenic precursors, requires not only strong ad hoc~~ local actions ~~should be complemented bybut also~~ decided national and European-wide strategies.

## 1 Introduction

10 Tropospheric ozone (O<sub>3</sub>) is an air pollutant of major public concern as it harms human health (WHO, 2013) and sensitive vegetation (Booker et al., 2009), and contributes to climate change (Jacob and Winner, 2009). O<sub>3</sub> is formed in the atmosphere through nonlinear photochemical reactions among carbon monoxide (CO), peroxy radicals generated by the photochemical oxidation of volatile organic compounds (VOC), and nitrogen oxides (NO<sub>x</sub>) (Crutzen, 1973). Therefore, meteorological stagnation, high temperature, and low precipitation enhance tropospheric O<sub>3</sub> formation (Demuzere et al.,  
15 2009; Otero et al., 2016). Atmospheric circulation also controls the short and long-range transport of O<sub>3</sub> affecting its lifetime in the atmosphere (Monks et al., 2015). For example, the transport of precursors emitted in urban and industrialized areas may cause O<sub>3</sub> production downwind (Holloway et al., 2003).

20 According to the European Environmental Agency (EEA) around 95-98% of the population in Europe during 2013-2015 were exposed to O<sub>3</sub> concentrations that exceeded the guidelines of the World Health Organization (WHO) (EEA, 2017). These guidelines establish a maximum daily 8-hour averaged (MDA8) O<sub>3</sub> concentration of 100 µgm<sup>-3</sup> never to be exceeded. The European Air Quality Directive (2008/50/EC) is less restrictive as it sets an O<sub>3</sub> Target Value of 120 µgm<sup>-3</sup> for the MDA8 concentration, which can be exceeded up to 25 days per calendar year averaged over three years.

25 Southern European countries around the Mediterranean Basin are particularly exposed to exceedances of the O<sub>3</sub> Target Value in summer due to the influence of frequent anticyclonic and clear-sky conditions that favour photochemical O<sub>3</sub> formation in the troposphere (EEA, 2017). In addition, its geographic location also makes the Basin a receptor of the long-range transport of pollution from Europe, Asia and even North America (Lelieveld et al., 2002; Gerasopoulos, 2005). The importance of long-range transport on surface O<sub>3</sub> has been studied in the Mediterranean Basin, indicating that the emission  
30 sources within the Basin have a dominating influence on surface O<sub>3</sub>, while remote sources are more important than local sources for O<sub>3</sub> mixing ratios at higher altitudes (Richards et al., 2013; Safieddine et al., 2014). Recent studies suggest that the upper O<sub>3</sub>-rich air masses could increase the surface O<sub>3</sub> concentration in the Mediterranean Basin (Kalabokas et al., 2017;

Querol et al., 2018). Further detailed and quantitative studies on the mechanism linking upper O<sub>3</sub>-rich layer with increases of the ground-level O<sub>3</sub> concentration in episodes need further clarification particularly regarding the contribution of O<sub>3</sub> transported at regional and hemispheric scales.

5 Several studies in the Iberian Peninsula (IP) have addressed the causes of O<sub>3</sub> episodes looking at the circulation of air masses (Millán, 2014, and references therein). In the Atlantic region, the blocking anticyclones over Western Europe favour the inter-regional transport of O<sub>3</sub> in the area and its accumulation for several days during the most severe episodes (Alonso et al., 2000; Gangoiti et al., 2002, 2006; Valdenebro et al., 2011; Saavedra et al., 2012; Monteiro et al., 2016). On the other hand, in the Mediterranean coast, the typical summer synoptic meteorological conditions with a lack of strong synoptic advection, 10 combined with the orographic characteristics and the sea and land breezes, favour episodes where high levels of O<sub>3</sub> are accumulated by recirculation of air masses loaded with O<sub>3</sub> precursors (Millán et al., 1997 and 2000; Toll and Baldasano, 2000; Gangoiti et al., 2001; Pérez et al., 2004; Jiménez et al., 2006; Gonçalves et al., 2009; Millán, 2014, Querol et al., 2017; Querol et al., 2018). The coupling between synoptic and mesoscale processes governing the levels of O<sub>3</sub> in the Western Mediterranean Basin need further research in order to understand the O<sub>3</sub> intercontinental contribution. Furthermore, from our 15 understanding there is a lack of research quantifying the contribution of the activity sources to the O<sub>3</sub> local formation during peak events in this region.

O<sub>3</sub> analyses in the Western Mediterranean Basins show that regional background O<sub>3</sub> levels have remained high without significant changes (EEA, 2016; EMEP-CCC, 2016; Querol et al., 2016). However, they have increased at traffic and urban 20 background sites (EEA, 2016; Querol et al., 2016; Sicard et al., 2016; Saiz-Lopez et al., 2017). ~~The reasons behind the urban O<sub>3</sub> upward trend are not clear yet due to the complex VOC-NO<sub>x</sub> regime; part of the O<sub>3</sub> increase may have resulted from the reduction of NO emissions relative to NO<sub>2</sub> and therefore to a lower NO titration effect in VOC-limited situations. The reasons behind the urban O<sub>2</sub> upward trend are not clear yet; it might be a result of the reduction of NO emissions relative to NO<sub>2</sub> and therefore a lower NO titration effect or the fact that urban O<sub>2</sub> formation is VOC-limited, and a reduction in NO<sub>x</sub> emissions might enhance O<sub>2</sub> formation.~~ The most intense O<sub>3</sub> events in the last decade, measured by the number of exceedances of the O<sub>3</sub> Target Value are recorded over areas downwind of large urban and industrial hot spots (Monterio et al, 2012; Querol et al., 2016; EEA, 2016). Overall, the number of these type O<sub>3</sub> events occur in June-July and during summer 25 heatwaves (i.e., 2003 and 2015).

30 According to the European Air Quality Directive, in zones exceeding the O<sub>3</sub> Target Value, Member States must develop plans to attain compliance by reducing the emission of O<sub>3</sub> precursors. Abatement of tropospheric O<sub>3</sub> concentration in the Western Mediterranean Basin has been insufficient so far (Querol et al., 2018). Effective planning requires an accurate quantitative knowledge of the sources of these precursors and their respective contributions to the exceedances of the O<sub>3</sub> Target Value (Querol et al., 2016; Borrego et al., 2015). However, source attribution of surface O<sub>3</sub> concentration remains a

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challenge, because the concentration at each location and time results not only from local biogenic and anthropogenic precursors, but also from the transport of O<sub>3</sub> and its precursors from neighbouring regions, O<sub>3</sub> hemispheric transport (UNECE, 2010), and stratospheric O<sub>3</sub> injections (Monks et al., 2015).

- 5 At present, there are no methods based on observations that distinguish the origin of O<sub>3</sub>. Despite their inherent uncertainties, Chemical Transport Models (CTMs) allow apportioning the contribution of any source (by sector and/or region) to O<sub>3</sub> concentrations. The most widely used approach is the “brute force” method, which consists on running an ensemble of simulations zeroing out the sources one by one and then comparing them with a baseline simulation that accounts for all of the sources. Several O<sub>3</sub> source apportionment studies at European scale have applied the brute force method to quantify the contribution of one or two emission sectors. For example, road transport emissions with the EMEP model (Reis et al., 2000), biogenic and anthropogenic emissions with the Polyphemus model (Sartelet et al. 2012), transport-related emissions including road transport, shipping, and aviation with the WRF-CMAQ model (TRANSPHORM, 2014), and ship emissions with CAMx (Aksoyuglu et al., 2016). ~~This approach~~ Brute force is simple to implement, as it does not require additional coding in the CTM. However, as it quantifies the contribution of each source based on its absence, it does not reproduce actual atmospheric conditions, and therefore it is susceptible to inaccuracies in the prediction of O<sub>3</sub> peaks under non-linear regimes (Cohan and Napelenok, 2011). Actually, brute force is not suitable to retrieve source contributions when the relationship between emissions and concentration is non-linear. ~~Despite these limitations, brute force but it~~ is useful for analysing the concentration responses to emission abatement scenarios (Clappier et al., 2017).
- 20 Recently, CTMs include algorithms that tag multiple pollutants by source (region and/or sector) all the way through the pollutant’s lifetime, from emission to deposition. This integrated source apportionment approach has several advantages. First, it allows identifying the main sources contributing to high O<sub>3</sub> levels under actual atmospheric conditions, which is a preliminary step towards designing refined and efficient emission abatement scenarios. Second, as we show below, it supports enhanced model evaluation and therefore potential model improvements by identifying problems in emission estimates (sectors or regions) or chemical boundary conditions. The Integrated Source Apportionment Method (ISAM) within the Community Multiscale Air Quality (CMAQ) model has shown promising results for O<sub>3</sub> tagging, exhibiting less noise in locations where brute force results are demonstrably inaccurate (Kwok et al., 2013, 2015). Recent ISAM experiments have quantified that the contribution of traffic in the cities of Madrid and Barcelona to the daily O<sub>3</sub> peaks downwind of the urban areas is particularly significant (up to 80-100 µgm<sup>-3</sup>) (Valverde et al., 2016a). O<sub>3</sub> tagging methods are also included in other regional and global models applied over Europe (Karamchandani et al., 2017; Butler et al., 2018).

The integrated source apportionment tools combined with high-resolution emission and meteorological models can help unravelling the sources responsible for peak summer events of O<sub>3</sub> in the Western Mediterranean Basin. Quantifying the contribution of emission sources during acute O<sub>3</sub> episodes is a prerequisite for the design of future mitigation strategies in the

region. In this framework, the main goal of this study is to provide a first quantitative estimation of the contribution of the main anthropogenic activity sectors compared to the imported concentration (regional and hemispheric) to peak O<sub>3</sub> events in Spain. We also assess the potential of our source apportionment method to improve O<sub>3</sub> modelling. Our study applies for the first time a countrywide O<sub>3</sub> source apportionment at high resolution over the (IP) Iberian Peninsula to investigate the local sources responsible for high O<sub>3</sub> concentration compared to the imported (regional and hemispheric) O<sub>3</sub> during a typical summer episode. Our analysis focuses on over the period between July 21<sup>st</sup> and 31<sup>st</sup>, 2012, which is representative of the typical summer synoptic conditions in the region. We use the CMAQ-ISAM within the CALIOPE air quality forecast system for Spain ([www.bsc.es/calioppe](http://www.bsc.es/calioppe)), which runs at 4-km horizontal resolution of 4x4 km<sup>2</sup> over the IP. The system is fed by the HERMESv2.0 emission model, which provides disaggregated emissions based on local information and state-of-the-art bottom-up approaches for the most polluting sectors.

The paper is organized as follows. In Section 2 we introduce the CALIOPE system, the set-up of ISAM and the HERMESv2.0 emission model for O<sub>3</sub> source apportionment studies, and the methodology used to quantify evaluate the model. In Section 3 we demonstrate the representativeness of the selected episode, we evaluate the model, and we provide an analysis of the source-sector contribution to the Spanish O<sub>3</sub> under the different synoptic patterns occurring during the study period. In Section 4, we discuss our findings, the regulatory implications, and future research.

## 2 Methodology

### 2.1 Air quality model

We used the CALIOPE air quality modelling system ([www.bsc.es/calioppe](http://www.bsc.es/calioppe)) to simulate the O<sub>3</sub> dynamics over the IP during the selected episode. CALIOPE is described elsewhere (Baldasano et al., 2008; Pay et al., 2014; Valverde et al., 2016a; and reference therein). The system consists of the HERMESv2.0 emission model (Guevara et al, 2013), the WRF-ARWv3.6 meteorological model (Skamarock and Klemp, 2008), the CMAQ v5.0.2 chemical transport model (Byun and Schere, 2006) and the BSC-DREAM8bv2 mineral dust model (Basart et al., 2012). CALIOPE first runs over Europe at 12-km resolution (12x12 km<sup>2</sup>, named EU12 domain) and then over the IP at 4-km resolution (4x4 km<sup>2</sup>, named IP4 domain) (Fig. S1). In the present work, the system is configured with 38 sigma layers up to 50 hPa, both for WRF and CMAQ. The planetary boundary layer is characterized with approximately 11 layers, where the bottom layer's depth is ~39 m. The EU12 domain uses meteorological initial and boundary conditions from the Final Analyses provided by the National Centers of Environmental Prediction (FNL/NCEP) at 0.5° by 0.5°. The first 12 h of each meteorological run are treated as cold start, and the next 23 h are provided to the chemical transport model. Boundary conditions for reactive gases and aerosols come from the global MOZART-4/GEOS-5 model at 1.9° by 2.5° (Emmons et al., 2010). CMAQ uses the CB05 gas-phase mechanism with active chlorine chemistry, an updated toluene mechanism (CB05TUCL; Whitten et al., 2010; Sarwar et al., 2012), and the sixth generation CMAQ aerosol mechanism including sea salt, aqueous/cloud chemistry and the ISORROPIA II

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thermodynamic equilibrium module (AERO6; Reff et al., 2009; Appel et al., 2013). **Table S1** depicts the remaining CALIOPE configuration options.

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For the IP4 domain, HERMESv2.0 estimates emissions for Spain with a temporal and spatial resolution of 1 h and up to 1 km by 1 km, according to the Selected Nomenclature for Air Pollution (SNAP), which are then aggregated to ~~4 km~~ 4-km resolution (Guevara et al., 2013). HERMESv2.0 is suitable for source apportionment studies thanks to its level of detail in the calculation of the emission fluxes by source (Guevara et al., 2014). HERMESv2.0 is currently based on 2009 data, which is the ~~most recent~~ closest year with updated information on local emission activities in HERMES at the time this work started. For neighbouring countries and international shipping activities, HERMESv2.0 uses the annual gridded national emission inventory provided by the European Monitoring and Evaluation Programme (EMEP) disaggregated to 4-km resolution using a SNAP-sector-dependent spatial, temporal and speciation treatment (Ferreira et al., 2013). ~~The VOC and NO<sub>x</sub> emissions from vegetation, which are critical in the formation of O<sub>3</sub>, are estimated with the Model of Emissions of Gas and Aerosols from Nature (MEGANv2.0.4) (Guenther et al., 2006) using temperature and solar radiation from the WRF model. HERMESv2.0 is suitable for source apportionment studies thanks to its level of detail in the calculation of the emission fluxes by source (Guevara et al., 2014).~~

HERMESv2.0 integrates the Model of Emissions of Gas and Aerosols from Nature (MEGANv2.0.4; Guenther et al., 2006) to estimate VOCs and NO<sub>x</sub> emissions from vegetation, which play a major role on O<sub>3</sub> photochemistry, using temperature and solar radiation from the WRF model. Note that we configured MEGAN to compute VOC emissions from cultivated crops; the agriculture emission module in HERMESv2.0 estimates the VOC from manure management and field burning of agricultural residues. In this study, we have updated MEGANv2.0.4 with emission factors from ~~last~~ MEGANv2.1 (<http://lar.wsu.edu/megan/guides.html>). In Sect. 2 of the supplement, we provide a comparison with measurements from the DAURE campaign (Pandolfi et al., 2014) showing the reasonably good behaviour of our modelled isoprene.

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Urban VOC emissions could be a relevant source ~~offer~~ O<sub>3</sub> concentration. Over Spanish urban areas, HERMESv2.0 estimates VOC emissions from road transport and the use of solvents (Fig. 1) following bottom-up approaches (Guevara et al., 2013). However, uncertainties in the estimation of urban VOC emission inventories, as stated recently by several works (Pan et al., 2015; Liu et al., 2017; McDonald et al., 2018; Lewis, 2018) makes uncertain the urban VOC contribution to tropospheric O<sub>3</sub> concentrations. In order to overcome this problem, continuous monitoring of urban VOC should be performed in Spanish cities, following the example of other regions in which O<sub>3</sub> is also a major problem such as Mexico City (Jaimes-Palomera et al., 2016). In addition, the use of formaldehyde satellite observations to constrain urban VOC emissions could be also pointed out as a future task to improve the representativeness of urban emission inventories (Zhu et al., 2014).

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## 2.2 Ozone source apportionment method

We applied ISAM to quantify contributions from different SNAP categories to the surface O<sub>3</sub> over the IP. The ISAM O<sub>3</sub> tagging method is a mass balance technique that tags both O<sub>3</sub> and its gas precursor emissions (NO<sub>x</sub> and VOC) from each source sector within one simulation (Kwok et al., 2013, 2015). Each tagged species undertakes typical physical processes (advection, vertical mixing, deposition, emission and chemistry) without perturbing the actual conditions. The O<sub>3</sub> rate of change for each tag in any grid cell is calculated as follows (Eq. 1):

$$\frac{dC_{tag}}{dt} = P_{tag} - D \frac{C_{tag}}{\sum_{tag} C} \quad (1)$$

Where C<sub>tag</sub> represents the O<sub>3</sub> concentration related to a tagged source of interest, P<sub>tag</sub> is the chemical production rate of O<sub>3</sub> formed by the precursors emitted for each tag, and D is the total chemical destruction rate of O<sub>3</sub> in this grid cell. Different ratios of NO<sub>x</sub>/VOC cause the formation of O<sub>3</sub> in each grid cell, which is controlled either by NO<sub>x</sub>- or VOC-limited conditions. ISAM uses the ratio H<sub>2</sub>O<sub>2</sub>/HNO<sub>3</sub> to determine whether O<sub>3</sub> is NO<sub>x</sub>- or VOC-sensitive (above or below 0.35, respectively) (Zhang et al., 2009). The bulk O<sub>3</sub> concentration in each model grid cell (P<sub>bulk</sub>) is equal to the sum of O<sub>3</sub> tracers that were produced in either NO<sub>x</sub> or VOC-sensitive conditions (Eq. 2),

$$P_{bulk} = \sum_{tag} P_{tag} = \sum_{tag} P_{tag}^N + \sum_{tag} P_{tag}^V \quad (2)$$

where P<sub>tag</sub><sup>N</sup> and P<sub>tag</sub><sup>V</sup> are the O<sub>3</sub> produced under NO<sub>x</sub>- and VOC-limited conditions, respectively according Eqs. 3 and 4:

$$P_{tag}^{N,new} = P_{tag}^{N,old} + P_{bulk}^{new} \frac{\sum_x NO_{x,tag}}{\sum_{tag} \sum_x NO_{x,tag}} \quad (3)$$

$$P_{tag}^{V,new} = P_{tag}^{V,old} + P_{bulk}^{new} \frac{\sum_y VOC_{y,tag} \times MIR_y}{\sum_{tag} \sum_y VOC_{y,tag} \times MIR_y} \quad (4)$$

NO<sub>x,tag</sub> and VOC<sub>j,tag</sub> are the concentrations of the x nitrogen and y VOC species in CB05 that participate in the photochemical O<sub>3</sub> formation for each source sector tag and grid cell. MIR<sub>y</sub> is the maximum incremental reactivity factor of each y species of VOC emitted by each source sector tag, corresponding to the O<sub>3</sub> generating potential of each single VOC species (Carter, 1994).

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### 2.3 Ozone tagged species

Table 1 summarizes the O<sub>3</sub> tagged sources in the present study and Fig. 1a depicts the HERMESv2.0's estimates of the contribution by each SNAP category to the total emissions of O<sub>3</sub> precursors in Spain. The largest NO<sub>x</sub> sources are road transport (SNAP7; 42%), non-road transport (SNAP8; 19%), manufacturing industries (SNAP34; 16%), and energy production (SNAP1; 16%). VOC are dominated by biogenic sources (SNAP11; 70%) and to a lesser extent by the agricultural sector (SNAP10; 11%), solvent and other product uses (SNAP6; 9%) and road transport (SNAP7; <7%). The selected (tagged) SNAP categories in this study are the energy, industrial, road transport and non-road transport sectors (Fig. 1b), which account for the 92% of the total NO<sub>x</sub> emissions in Spain. An additional tracer (OTHR) gathers the remaining emission categories that were not explicitly tracked (i.e., SNAP2, 5, 6, 9, 10 and 11).

In addition to the selected sources, we tracked the contributions of the chemical boundary conditions (BCON) and the initial conditions (ICON). BCON represents both the O<sub>3</sub> directly transported through the IP4 domain boundaries and the formation of O<sub>3</sub> resulting from precursors that are also transported through the boundaries. BCON O<sub>3</sub> comes from the EU12 parent domain, which includes the O<sub>3</sub> produced in Europe and the O<sub>3</sub> transported at global scale (both tropospheric and stratospheric O<sub>3</sub>) provided by the MOZART-4/GEOS model (Fig. S1). Hereinafter, we name BCON O<sub>3</sub> as the imported O<sub>3</sub> to the IP4 domain. Tagging the O<sub>3</sub> initial allows quantifying the number of spin-up days to minimize the impact of model initialization. For the present run, we required 6 days of spin-up to set the contribution of initial conditions to less than 1% of the net hourly O<sub>3</sub> concentration over 95% of the available O<sub>3</sub> stations.

### 2.4 Evaluation method

We evaluate the simulated concentrations against air quality measurements from the Spanish monitoring stations that are part of the European Environment Information and Observation Network (EIONET; <https://www.eionet.europa.eu/>). The EIONET network provides a relatively dense geographical coverage of the Spanish territory. During the July 21<sup>st</sup>-31<sup>st</sup> episode we used the measurements from 347 stations for O<sub>3</sub> and 357 stations for NO<sub>2</sub> with a temporal coverage above 85% on an hourly basis. Fig. S2 shows the distribution of the stations for O<sub>3</sub> and NO<sub>2</sub>.

The evaluation based on discrete statistics includes the correlation coefficient (r), Mean Bias (MB), [Normalized Mean Bias \(NMB\)](#) and the Root Mean Square Error (RMSE) ([Appendix B](#)). We used the package "openair" (Carslaw and Ropkins, 2012) for R (v3.3.2; R Core Team, 2016) to compute the statistics. We calculate statistics on an hourly basis for O<sub>3</sub> and NO<sub>2</sub>, as well as for the regulatory MDA8 in the case of O<sub>3</sub>. The evaluation also takes into account the station type, following the categories established by the EEA (i.e., rural background, [suburban background](#), urban background, industrial and traffic).

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There are no direct evaluation methods for apportioned pollutants. Instead, we designed a diagnostic plot for source apportionment analysis at each individual receptor, including a time series of measured and observed O<sub>3</sub> and NO<sub>2</sub> concentrations together with the simulated tagged sources. In addition, this plot includes the simulated wind speed and direction. These plots are helpful as they compare the modelled O<sub>3</sub> and NO<sub>2</sub> with the observations, while highlighting the sources and circulation patterns at least partly responsible for the model behaviour. This work will only discuss in detail the source apportionment plots at key O<sub>3</sub> receptor regions, given the high number of stations (260) that simultaneously measure O<sub>3</sub> and NO<sub>2</sub>.

Evaluation results will be discussed together with the source apportionment results. On the one side, the interpretation of the source apportionment results benefits from model evaluation. On the other side, the source apportionment results support enhanced model evaluation as it allows identifying potential errors in emission estimates for specific sectors and/or in the chemical boundary conditions.

### 3 Results

#### 3.1 Description of the ozone episode

In 2012, the largest O<sub>3</sub> episode in Europe occurred between July 21<sup>st</sup> and 31<sup>st</sup>. That period alone comprised 33% of the total number of exceedances of the information threshold, and 12% of the number of exceedances of the O<sub>3</sub> Target Value in 2012 (EEA, 2013). In Spain, around 60% of the annual exceedances also occurred during this period. (As shown in Querol et al. (2016) July is typically the month with the highest number of O<sub>3</sub> exceedances in Spain.) The episode affected the central and north of the IP (Fig. 2c), in particular the 90<sup>th</sup> percentile of the MDA8 O<sub>3</sub> concentration exceeded the Target Value at the surroundings of the Madrid Metropolitan Area (MMA), and an area located north of the Barcelona Metropolitan Area (BMA), which correspond to hotspots and the tail end of large urban and/or industrial plumes.

Figure 2a shows the observed MDA8 O<sub>3</sub> concentrations trends at the Spanish EIONET stations during the (extended) summers (i.e., from April to September) from 2000 to 2012, together with the values recorded during the selected episode. We have categorized the MDA8 O<sub>3</sub> concentrations by station type. Rural background stations show the highest MDA8 O<sub>3</sub> concentrations during the episode followed by industrial, traffic and urban stations, which is consistent with the observed previous summer patterns. More specifically, the episode shares a similar 75<sup>th</sup> percentile (75p) of the MDA8 O<sub>3</sub> concentrations at rural background stations above the Target Value with the particularly severe summer of 2003 (Solberg et al., 2008). Note that the MDA8 O<sub>3</sub> concentrations at rural and urban background sites barely changed between 2000 and 2012, which suggests that the benefits from European emission controls may have been significantly counterbalanced by increasing background O<sub>3</sub> (Monks et al., 2015). In contrast, the MDA8 O<sub>3</sub> concentrations at TR and IN sites slightly

increased, likely because of a lower O<sub>2</sub> titration due to the preferential abatement of NO vs. NO<sub>2</sub>. However, we need more research on O<sub>3</sub> source apportionment to confirm this hypothesis (Querol et al., 2016; Sieard, et al., 2016).

Our first estimation of the origin of peak O<sub>3</sub> events in Spain focuses on the episode July 21st -31st, 2012. Figure 2a illustrates the relevance of the episode showing the observed MDA8 O<sub>3</sub> concentrations trends at the Spanish EIONET stations during the (extended) summers (i.e., from April to September) from 2000 to 2012, together with the concentrations recorded during the episode. Although the selected episode is not the most severe between 2000 and 2012 at national scale, it comprises a period with high MDA8 O<sub>3</sub> concentrations measured at rural background stations, actually the 75<sup>th</sup> percentile of those values ~~were~~ above the Target Value, similar to the particularly severe summer of 2003 (Solberg et al., 2008).

This episode is also interesting because it was widespread and affected big parts of Europe (EEA, 2013). Only during this period 33% and 12% of the total number of exceedances for the information threshold and the Target Value in 2012, respectively, were measured. The O<sub>3</sub> regional context of the episode allows us to study the influence of the imported O<sub>3</sub> to Spain.

The maps of the 90th percentile of the measured MDA8 O<sub>3</sub> concentrations over Spain (Fig. 2b) show high concentration spots all over the domain. The exceedances of the Target Value were found in the surroundings of large urban areas (Madrid, Barcelona, Valencia, Seville) and along Spanish valleys (i.e., Ebro Valley, Guadalquivir Valley).

Figure 2b shows the number of observed exceedances of the O<sub>3</sub> Target Value per day during the selected episode. There were more than 100 exceedances of the O<sub>3</sub> Target Value in most of the days during the episode, with relative maxima on July 25<sup>th</sup>, 28<sup>th</sup> and 31<sup>st</sup> attributed to the change in the synoptic conditions (Fig. S3). Figure 3 shows the meteorological patterns (2m temperature, 10m wind, precipitation, mean sea level pressure and geopotential height at 500 hPa) modelled by WRF-ARW during the three distinctive days over the outer EU12 domain.

Our characterization of the study period is based on the circulation type classification ~~performed~~ proposed in Valverde et al. (2014), who developed an objective synoptic classification method over the period 1983–2012, specifically designed to study air quality dynamics over the IP. Stagnant conditions and northwestern advections are the most frequent summer synoptic circulation patterns over the IP, occurring ~44% of the days in a year (Jorba et al., 2004; Valverde et al., 2014). Stagnant conditions are characterized by reduced surface pressure gradients and weak synoptic winds, intense solar radiation, and the development of the Iberian Thermal Low (ITL). The ITL forces the convergence of surface winds from the coastal areas towards the central plateau enhancing sea breezes and mountain-valley winds and subsidence over the Western Mediterranean Basin as described by (Millán et al., 1997, 2000; Millán 2014). In contrast, northwestern advections (NWad) transport air masses from the Atlantic towards the north and west of the IP and they are characterized by atmospheric

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instability and intense ventilation. Periods of accumulation and venting of pollutants follow the same sequence of pressure ridging and throughing respectively, of the lower and middle troposphere of the IP during the warm season (Querol et al., 2017; 2018). [According to the circulation type classification in Valverde et al. \(2014\)](#), the selected episode started with the development ~~of an accumulation-the~~ ITL ~~period~~ (July 21<sup>st</sup>-25<sup>th</sup>), followed by a NWad-venting period (July 26<sup>th</sup>-29<sup>th</sup>) and ended with the development of another ITL (July 30<sup>th</sup>-31<sup>st</sup>).

Figure 4 shows the 90<sup>th</sup> percentile (90p) of the simulated hourly O<sub>3</sub> and NO<sub>2</sub> concentrations corresponding to the three distinctive days with the relative maxima of exceedances. In the northern Spanish Mediterranean areas, intense O<sub>3</sub> episodes often affect the plains and valleys located ~~60 km north of the to the north of the BMA~~ [Barcelona Metropolitan Area \(BMA\)](#) in summer (Toll and Baldasano, 2000; Gonçalves et al., 2009; Valverde et al., 2016a; Querol et al., 2017). High NO<sub>x</sub> concentrations from the BMA combined with high ~~summer~~ biogenic VOC levels are driven inland by [mesoscale processes \(sea breezes and mountain-valley winds\)](#) ~~sea breezes and mountain valley winds, channelled by north-south valleys towards an intra-mountain plains located 60 km north of the BMA~~. This happened on July 31<sup>st</sup> when the highest 90p of the hourly O<sub>3</sub> concentrations (160-180 µgm<sup>-3</sup>) in Spain occurred over ~~the N-north and NW~~ [northwest of the](#) BMA. Occasionally, as it happened on July 25<sup>th</sup>, anticyclonic winds over the western Mediterranean Sea deflect the sea-breeze flow enriched with precursors from the BMA towards the Gulf of Lion where it reaches the highest 90p of the hourly O<sub>3</sub> concentrations (160-180 µgm<sup>-3</sup>) in the IP Mediterranean region. Eastern Spanish Mediterranean areas show similar O<sub>3</sub> dynamics, with inland regions depicting the highest O<sub>3</sub> peaks (140-160 µgm<sup>-3</sup>) when ~~low-pressure gradients~~ [stagnant conditions](#) cover central and eastern IP.

In the centre of the IP, intense O<sub>3</sub> episodes occurred during the development of the ITL, where the affected area depends on the synoptic conditions [\(Querol et al., 2018\)](#). Under the absence of synoptic forcing (e.g., July 25<sup>th</sup>), the MMA had the highest 90p of the hourly O<sub>3</sub> concentrations (~140-160 µgm<sup>-3</sup>). In contrast, when mountain valley winds are reinforced with synoptic westerlies (e.g., July 31<sup>st</sup>) (Fig. 3) the urban NO<sub>x</sub> plume is channelled along the mountain ranges in Madrid towards the northeast and the highest 90p of the hourly O<sub>3</sub> concentrations are found along the valley (~140-160 µgm<sup>-3</sup>).

In the ~~north~~ and ~~northeast~~ of the IP, the 90p of the hourly O<sub>3</sub> concentrations show a significant increase when the blocking anticyclone over Western Europe is combined with the development of the ITL (e.g. July 25<sup>th</sup>). The stagnant conditions favour the accumulation of O<sub>3</sub> precursors around main cities and industrial areas and enhance the local O<sub>3</sub> formation.

The NWad pattern (e.g., July 28<sup>th</sup>) significantly decreases the 90p of the hourly O<sub>3</sub> concentrations in the ~~central~~ and ~~north~~ of the IP. The northwesterly winds decrease the temperature and therefore the O<sub>3</sub> formation. As consequence, O<sub>3</sub> levels are reduced in the plumes from the BMA and the MMA, although they are still significant in the latter. Overall, the

90p of the hourly O<sub>3</sub> concentrations during the NWadv pattern were ~100 µgm<sup>-3</sup> in most background areas. In contrast, during the ITL it was above 120 µgm<sup>-3</sup>.

### 3.2 Statistical evaluation

CALIOPE has been evaluated in detail elsewhere (Pay et al., 2014 and references therein). ~~Furthermore, the system has been evaluated using the DELTA-Tool (Thunis and Cuvelier, 2014) tool developed by the Forum for Air Quality Modelling in Europe, named DELTA-Tool (Thunis and Cuvelier, 2014) to support and harmonize the model evaluation in the frame of the Air Quality Directive. Valverde et al. (2015; 2016) used the DELTA-Tool v4.0 and showed that the CALIOPE system accomplishes the quality objectives as defined in the Air Quality Directive for 78% of the NO<sub>2</sub> and 91% of the NO<sub>2</sub> and O<sub>3</sub> monitoring stations in summer conditions during the summer of 2012.~~ Here, we evaluate the updated version of CALIOPE using ISAM to quantify the system's ability to reproduce O<sub>3</sub> and NO<sub>2</sub> concentrations during the selected episode. Table 2 compiles the quartiles of the statistics calculated by station type. ~~(Note that there are stations not fitting any of the four categories (i.e., IN, TR, UB and RB), so the exceedance/station numbers in the IN/TR/UB/RB rows do not sum up to the numbers in the "ALL" rows.)~~

The model slightly overestimates the average hourly and MDA8 O<sub>3</sub> concentrations with MB of +12 µgm<sup>-3</sup> and +6 µgm<sup>-3</sup>, respectively. The *r* is above 0.6 in more than 50% of the stations and above 0.7 in 25% of them. The MB for average hourly and MDA8 O<sub>3</sub> concentrations are lower at RB stations (±4 µgm<sup>-3</sup>) than at IN, TR and UB stations (between +6 and +16 µgm<sup>-3</sup>) in 50% of the stations. As expected, the highest number of exceedances of the O<sub>3</sub> Target Value was recorded at RB stations (260 exceedances) followed by IN stations (204 exceedances).

At RB stations, average hourly O<sub>3</sub> is overestimated (+4 µgm<sup>-3</sup>) and MDA8 O<sub>3</sub> is underestimated (-4 µgm<sup>-3</sup>), which indicates that nighttime O<sub>3</sub> is overestimated. The nighttime overestimation is a common feature of CTMs and it is typically attributed to the underestimation of vertical mixing during nighttime stable conditions and to underestimation of the O<sub>3</sub> titration by NO (Bessagnet et al., 2016; Sharma et al., 2017).

CALIOPE underestimates the average hourly NO<sub>2</sub> concentrations with -7 µgm<sup>-3</sup> at TR stations and -2 µgm<sup>-3</sup> at RB stations. This partly explains the high overestimation of the average hourly and MDA8 O<sub>3</sub> concentration at TR and UB stations, as well as the systematic overestimation of average hourly O<sub>3</sub> concentration at nighttime (due to a lack of O<sub>3</sub> titration by NO). The average hourly NO<sub>2</sub> concentration at TR stations feature the highest *r* (with 25% of stations above 0.6), which proves the reasonably accurate representation of temporal emission in urban areas by the HERMESv2.0 (Guevara et al, 2014; Baldasano et al., 2011). In contrast, the RMSE is highest at TR stations, which results from the underestimation of NO<sub>2</sub> peaks during traffic rush hours. Underestimation of NO<sub>2</sub> traffic peaks is a common problem in Eulerian mesoscale models (Pay et al., 2014), as emission heterogeneity is lost in the grid cell-averaging process, which is especially critical in urban

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areas. Next generation microscale models will potentially solve this problem (Lateb et al., 2016). Besides the dilution of the emission in the grid, meteorology also may play an important role in the low performance of NO<sub>2</sub> and O<sub>3</sub> in hotspot areas. Several inter-comparison studies (e.g., EURODELTA and AQMEII) agree on the limitations of models to simulate meteorological variables that affect the average hourly NO<sub>2</sub> temporal variability, which controls model performance for O<sub>3</sub> in high NO<sub>x</sub> environments and areas downwind (Bessagnet et al., 2016; Solazzo et al., 2017).

Figure 5 classifies the average hourly and MDA8 O<sub>3</sub> concentrations at the air quality stations into four MB categories that account for the 93% of the stations. There are two categories with the lowest hourly O<sub>3</sub>-MB ( $\pm 10 \mu\text{g m}^{-3}$ ): type A with a MDA8 O<sub>3</sub>-MB between  $-40$  and  $-10 \mu\text{g m}^{-3}$  and type B with MDA8 O<sub>3</sub>-MB of  $\pm 10 \mu\text{g m}^{-3}$ . There are two categories with the highest hourly O<sub>3</sub>-MB (between  $10$  and  $40 \mu\text{g m}^{-3}$ ): type C with a MDA8 O<sub>3</sub>-MB of  $\pm 10 \mu\text{g m}^{-3}$  and the type D with MDA8 O<sub>3</sub>-MB between  $10$  and  $40 \mu\text{g m}^{-3}$ . In addition, Fig. S2 depicts the MB, RMSE and  $r$  for hourly and MDA8 O<sub>3</sub>, and hourly NO<sub>2</sub> concentrations. The best performances for O<sub>3</sub> (type B) are found at the 28% of the stations, located in the surroundings of the MMA, the BMA and most of the northern Mediterranean stations, which is consistent with the highest  $r$  ( $0.6 < r < 0.9$ ) found in the center and north of the IP (Fig. S3, Fig. S4). The highest O<sub>3</sub> overestimations (type D) are present at 36% of the stations, mainly located in highly industrialized areas in Spain (Guadalquivir Basin Valley, Strait of Gibraltar, Valencia) and inside the MMA. Next sections analyze the origin of these O<sub>3</sub> biases using the source apportionment time series.

The comparison with previous CALIOPE studies (Baldasano et al., 2011; Pay et al., 2014) indicates that  $r$  is in the same range for O<sub>3</sub> (0.6-0.7) and NO<sub>2</sub> (0.4-0.5) at individual stations; the same applies to the RMSE ( $15-29 \mu\text{g O}_3 \text{ m}^{-3}$  and  $10-20 \mu\text{g NO}_2 \text{ m}^{-3}$ ). Modelled O<sub>3</sub> shows higher performance at traffic stations in large cities, since stations influenced by road transport emissions (i.e., high-NO<sub>x</sub> environments) are better characterized with a more pronounced daily variability (Baldasano et al., 2011). At European scale, several model intercomparisons (Giornado et al. 2015; Bessagnet et al., 2016) indicate that O<sub>3</sub> concentrations in summer agree with the surface observations with  $r$  between 0.5 and 0.6. NO<sub>2</sub> hourly variability is overall underestimated due to uncertainties in the emission estimates, and meteorological modelling input and model resolution. These studies highlight the limitations of models to simulate meteorological variables that affect the NO<sub>2</sub> hourly variability, and therefore the model performance for O<sub>3</sub> in high-NO<sub>x</sub> environments and downwind. The comparison with previous CALIOPE studies (Baldasano et al., 2011; Pay et al., 2014) indicates that  $r$  is in the same range for O<sub>3</sub> (0.6-0.7) and NO<sub>2</sub> (0.4-0.5) at individual stations; the same applies to RMSE ( $15-29 \mu\text{g O}_3 \text{ m}^{-3}$  and  $10-20 \mu\text{g NO}_2 \text{ m}^{-3}$ ). Modelled O<sub>3</sub> shows higher performance at traffic stations in large cities, since stations influenced by road transport emissions (i.e., high-NO<sub>x</sub> environments) are better characterized with a more pronounced daily variability (Baldasano et al., 2011).

Section S4 in the supplement discusses the meteorological evaluation results and their impact on the pollutant concentrations. Not surprisingly, temperature shows the best behaviour when compared with observations (Table S2). The modelled wind speed are overestimated, particularly during nighttime (Fig. S5), coincident with low level wind speed. The

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nighttime overestimation of wind is a source of ~~large error compensation for the modelling error of~~ simulated  $\text{NO}_2$  and  $\text{O}_3$  nighttime concentrations (Vautard et al., 2012; Bessagnet et al., 2016).

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### 3.3 Source-sector ozone contributions during peak episodes

Figure 6 shows the 90p of the average hourly  $\text{O}_3$  concentration over the IP tagged by source type (Table 1) for different days (July 25<sup>th</sup>, 28<sup>th</sup> and 31<sup>st</sup>). (~~Fig. S3-Fig. S6~~ in the ~~supplementary material~~ shows similar plots for  $\text{NO}_2$ .) The imported  $\text{O}_3$  is by far the largest contributor showing a 90p ranging from 70 to 120  $\mu\text{g m}^{-3}$  in the east/north/centre of the IP on July 25<sup>th</sup> /28<sup>th</sup> /31<sup>st</sup>, respectively (~~Fig. 6~~). The imported-  $\text{O}_3$  enters the study domain through the IP4 domain boundaries and it can only be transported, scavenged deposited or depleted by  $\text{O}_3$  precursors. Therefore, areas with low imported  $\text{O}_3$  concentrations ( $< 50 \mu\text{g m}^{-3}$ ) are good indicators of (1) the accumulation of specific  $\text{O}_3$  precursors that deplete imported  $\text{O}_3$ , and (2) the subsequent  $\text{O}_3$  photochemical production that occur mostly under stagnant conditions and around the largest industrial/urban areas. The 90p of the hourly imported  $\text{O}_3$  concentration shows the lowest values in two different conditions and regions. ~~First, on~~ July 25<sup>th</sup> in the northwestern IP and Portugal, ~~extremely low winds~~stagnant conditions allow the accumulation of local pollutants that titrate imported  $\text{O}_3$  concentrations down to 30-70  $\mu\text{g m}^{-3}$ ; ~~at~~At the same time  $\text{O}_3$  is locally produced due to traffic emission downwind of major northern cities (La Coruña, Gijón, Bilbao) (60-120  $\mu\text{g m}^{-3}$ ), shipping activities (up to 40  $\mu\text{g m}^{-3}$ ) and the generation of energy and industrial processes (10-20  $\mu\text{g m}^{-3}$ ). ~~Second, on~~ July 31<sup>st</sup> in the northeast of ~~the~~ IP, when the pollutants transported from the Gulf of Lion and Catalonia towards the Mediterranean act as a sink of imported  $\text{O}_3$  reducing its concentration down to 60  $\mu\text{g m}^{-3}$ . As a result, there is a local  $\text{O}_3$  formation up to 120-160  $\mu\text{g m}^{-3}$  along the Ebro Valley and the Lleida Plain. ~~Borrego et al. (2016) in a source attribution study over northern Portugal Borrego et al. (2016) also found thea~~ reduction of imported  $\text{O}_3$  and the subsequent  $\text{O}_3$  formation by local sources under similar meteorological conditions.

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After the imported  $\text{O}_3$ , the largest contributor to  $\text{O}_3$  is the road transport sector. Downwind of the major urban areas in Spain (i.e., Madrid, Barcelona, Bilbao, ~~Sevilla~~Seville, Valencia, ~~Gijon~~, Pontevedra), on-road traffic contributed to the 90p of the hourly  $\text{O}_3$  concentrations as much as 60-120  $\mu\text{g m}^{-3}$ , and affected different areas depending on the synoptic/mesoscale regimes (Fig. 6). In the ~~north~~east of the BMA, the 90p of the hourly  $\text{O}_3$  concentration from the road transport sector reaches its maximum when the stagnant conditions affects the centre and eastern IP (e.g., July 31<sup>st</sup>). As noted above, mesoscale winds carry traffic  $\text{O}_3$  precursors from the BMA inland, channelled by north-south valleys towards the intra-mountain plain in the north. Over the MMA, the 90p of the hourly  $\text{O}_3$  concentration from the road transport sector showed the maximum when the ITL was combined with the synoptic westerlies (e.g., July 31<sup>st</sup>), carrying high  $\text{O}_3$  as far as to the Ebro Valley, as shown in Fig 4.

Regarding the contribution from the non-road transport sector, the Atlantic regions of the IP show the highest 90p of the average hourly  $\text{O}_3$  concentration (25-40  $\mu\text{g m}^{-3}$ ) on July 25<sup>th</sup>. The stagnant condition favoured the accumulation of precursors

from the Atlantic shipping route and the formation of O<sub>3</sub> within the region. The Spanish Mediterranean region shows the highest 90p of the [average](#) hourly O<sub>3</sub> concentrations from the non-road transport sector in front of the southeastern coasts of the IP (~180 µgm<sup>-3</sup>) when the westerlies in the Strait of Gibraltar injected precursors from international shipping into the Mediterranean Basin (e.g., July 28<sup>th</sup> and 31<sup>st</sup>). Note that during days with high 90p of the [average](#) hourly O<sub>3</sub> concentration from non-road transport, the imported O<sub>3</sub> concentration shows the lowest 90p due to the NO titration effect over emission areas.

The elevated point source emission sectors (i.e., energy and industry) contributed less to O<sub>3</sub> than the traffic sector, but their contributions were significant reaching 15-25 µgm<sup>-3</sup> of the 90p of the [average](#) hourly O<sub>3</sub> concentrations (Fig. 6). The north and northeastern of the IP, the Mediterranean coast, and the Guadalquivir Valley are the most affected regions under stagnant conditions.

The contribution of the remaining sectors (OTHR) to the 90p of the [average](#) hourly O<sub>3</sub> concentrations was similar to that of the elevated point sources (15-25 µgm<sup>-3</sup>), but it reached up to 30 µgm<sup>-3</sup> in areas downwind of Oporto and Lisbon (Fig. 6). OTHR includes the formation of O<sub>3</sub> from the remaining anthropogenic and biogenic sources (accounting for less than 8% of total NO<sub>x</sub> emissions, but 93% of total VOC). The high OTHR concentration around the biggest cities in Portugal may be related to precursors emitted by the residential sector (SNAP2 and 9) and biogenic emissions, as found in other source apportionment studies over Portugal (Borrego et al., 2016; Karamchandani et al., 2017).

### 3.4 Regionalization of source-sector contributions

~~The analysis of the source-sector contributions to O<sub>3</sub> allowed~~ We have identified ~~ten~~ O<sub>3</sub> receptor regions with similar characteristics in terms of meteorological and ~~O<sub>3</sub>-geographical patterns, O<sub>3</sub> dynamics and main source contributors~~ ~~main source contributors and geographical patterns~~ (Figs. 4 and 6). ~~The receptor regions defined in our work are consistent with Diéguez et al. (2014) and Querol et al. (2016), who proposed a similar regionalization based on observations from air quality stations.~~ Figure 7c shows the location of the air quality stations belonging to each receptor region corresponding to the ~~central of the~~ IP (CIP), the east ~~of the~~ IP (EIP), the Ebro Valley (EV), the Guadalquivir Valley (GV), the Mediterranean Sea (MED), the northeast ~~of the~~ IP (NEIP), the north ~~of the~~ IP (NIP), the northwest ~~of the~~ IP (NWIP), the south ~~of the~~ IP (SIP) and the west ~~of the~~ IP (WIP). ~~The O<sub>3</sub> receptor regions are consistent with Diéguez et al. (2014) and Querol et al. (2016), who proposed a similar regionalization based on observations from air quality stations.~~

Figure 7a shows the absolute O<sub>3</sub> contribution of each tagged source at air quality stations by region along with the modelled and observed daily mean concentration during exceedances of 120 µgm<sup>-3</sup> of the observed MDA8 ozone. ~~Note that differences between sectors are more evident in the when normalized contribution~~ (Fig. 7b). (Table S3 compiles de numerical values of Fig. 7.) ~~Figure 7 indicates that during exceedances of the MDA8 target value there is a good agreement~~

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( $r = 0.79$ ) between the sum of apportioned  $O_3$  and the observed concentrations over the receptor regions. Fig. 7b shows the respective fractional mean  $O_3$  contribution. (Table S2 compiles de numerical values of Fig. 7.) We have excluded the ICON  $O_3$  in Table S2 because its contribution is negligible after six days of spin-up. The spatial  $r$  between modelled and observed daily mean  $O_3$  concentration during exceedances of the MDA8 target value is 0.79, which indicates that CALIOPE is able to reproduce reasonably well the spatial variability of  $O_3$  during peak episodes in Spain.

The MED region represented by stations in the Balearic Islands shows the highest imported  $O_3$  contribution (76%) because it is relatively far away from important anthropogenic  $NO_x$ +VOC sources in the IP. Under the ITL influence (July 25<sup>th</sup> and 31<sup>st</sup>), MED received air masses enriched with on-road traffic precursors from southern France and the NEIP, which enhanced  $O_3$  production up to 7%. Shipping emissions in the MEDMediterranean region contributed up to 8% of the total  $O_3$ .

After MED, there is a cluster of regions along the Spanish Mediterranean coast (i.e., NEIP, EV and EIP) showing imported  $O_3$  contributions between 60 and 68 % of the daily mean  $O_3$  under exceedances. This is explained by their proximity to the eastern boundary and the frequent mesoscale phenomena enhancing the recirculation and accumulation of imported  $O_3$  along the Spanish Mediterranean coasts. The contribution of road and non-road transport is similar (~11-16%) because these regions have both important roads and maritime trade routes. Note that the SIP region, which is also located in the Spanish Mediterranean coast, shows a daily mean imported  $O_3$  concentration lower than the regions along the Spanish Mediterranean coast other Mediterranean regions (~57%) and the highest non-road transport contribution in Spain (19%). The main sink of imported  $O_3$  are precursors resulting from dense shipping traffic through the Strait of Gibraltar, which have substantial impact in the  $O_3$  production downwind (either in the Alboran Sea or the Gulf of Cadiz).

The regions including the largest metropolitan areas in Spain are the CIP (Madrid) and the NEIP (Barcelona). Both regions show an imported  $O_3$  contribution of ~60% and a similar contribution from the road transport sector (18 and 16%, respectively). However, the NEIP shows a slightly higher contribution from non-road transport (13 vs 10%) due to the influence of international shipping near coastal areas.

The northern and northwestern regions of the IP (NIP and NWIP) had relatively lower imported  $O_3$  contributions (56-59%). The contribution of non-road transport was ~10-12%, slightly lower than in the Mediterranean coast, and that of road transport was also significant (~14-15%). The contribution from the industrial sector was one of the highest in the country (~5%) due probably related to the influence of the large industrial facilities located in several areas of Galicia, Cantabria, Asturias, and the Basque Country the north of Spain. The contribution from the energy sector in the NWIP region was the highest in Spain (~5%) due to emissions from large coal-fired power plants located in the area, including As Pontes, Meirama, Aboño, Guardo, and Compostilla.

The WIP had the lowest daily mean O<sub>3</sub> concentrations during days exceeding the O<sub>3</sub> Target Value (93.5 µgm<sup>-3</sup>) and a high imported O<sub>3</sub> contribution (~63%). NO<sub>x</sub> emissions in the WIP region are moderate (Fig. 4), which could explain the low daily mean O<sub>3</sub> concentration. There is a significant contribution from traffic (14% for road transport and 11% for non-road transport) and industrial and energetic sectors (7%) to the daily mean O<sub>3</sub> concentrations. These anthropogenic contributions suggest that O<sub>3</sub> in the WIP ~~may be produced~~<sup>is produced</sup> by precursors transported from the surrounding cities (Oporto, Lisbon and Madrid) and the highly industrialized areas in the NWIP and the NIP (Fig. 6).

The Guadalquivir Valley had the lowest imported O<sub>3</sub> contribution of the IP (~46%) and the highest daily O<sub>3</sub> concentration during days of exceedance. ~~The relatively low imported O<sub>3</sub> contribution resulted from the titration effect driven by the high NO<sub>x</sub> emissions from industrial and urban hotspots in the region.~~ The on-road traffic was the highest anthropogenic contributor to O<sub>3</sub> (~18%) due to the emissions from three major cities (Sevilla, Huelva, and Cordoba). Although O<sub>3</sub> in Huelva may be overestimated (as discussed later), shipping is the second most important contributor to O<sub>3</sub> in the Guadalquivir Valley (~17%) ~~due probably link to~~ the important fluvial transport along the Guadalquivir River. (The Guadalquivir River is one of the most important routes for merchandise transport in Europe.) In fact, the non-road transport sector is the highest contributor (~17-19%) in Southern Spain, both in the Guadalquivir Valley and the SIP, also due to the dense maritime routes across the Strait of Gibraltar.

~~The following sections analyse the source apportionment results eight stations located in key regions (see Fig. 5). We analyse two regions with high on-road traffic contribution (CIP and NEIP), and two regions with high contribution from industry and energy production (NWIP and Guadalquivir Valley).~~

~~The following sections analyse the source apportionment results at key regions with a high on-road traffic contribution (i.e., CIP and NEIP) and a high contribution from industry and energy production (i.e., NWIP and Guadalquivir Valley).~~

#### 3.4.1 The centre of the Central Iberian Peninsula

Figure 8 show the source apportionment time series of the average hourly O<sub>3</sub> and NO<sub>2</sub> concentrations at two stations ~~in the CIP~~, an urban station in Madrid (station 1 in Fig. 8a), and another one located in Guadalajara (station 2 in Fig. 8a), which is a medium size city affected by Madrid's urban plume. ~~The model reproduces reasonably well the main processes and source contributions, with a significant imported O<sub>3</sub> contribution in both stations.~~

At the urban station, the model reproduces the O<sub>3</sub> traffic cycle (r = 0.66 and MB=22.5 µgm<sup>-3</sup>) featuring the typical low O<sub>3</sub> concentrations (< 40 µgm<sup>-3</sup>) in the early morning and in the afternoon due to O<sub>3</sub> titration (Fig. 8a). However, O<sub>3</sub> was overestimated (MB type D) during daytime peaks due to the overestimation of the NO<sub>2</sub> morning peaks during stagnant conditions, coincident with the highest road transport contribution for both pollutants. The results point towards a poor

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representation of the meteorological conditions in the city during the stagnant conditions as shown in the meteorological evaluation (Sect. 4 in the supplement).

At the urban station downwind (Fig. 8b) the modelled  $O_3$  shows a positive bias (MB type B) due to the underestimation of  $NO_2$  concentration. Note that the uncertainty of traffic- $NO_2$  traffic emissions in medium size cities is larger than in the largest urban areas (i.e. Madrid, Barcelona) because data is generally unavailable and emissions are estimated using emission factors that depend on population density (Baldasano et al., 2011).

Despite the limitation, the source apportionment results show that imported  $O_3$  is the main contributor in both stations, but  $O_3$  formation from traffic increases significantly during peaks in both stations. The highest  $O_3$  concentrations ( $\sim 160 \mu\text{g m}^{-3}$ ) are modelled when westerly winds channelled along the Tajo valley carry the polluted air masses in a NE direction through the Tajo Valley. This, which results in a  $O_3$  contribution of  $\sim 70 \mu\text{g m}^{-3}$  from the road transport sector in areas downwind (see wind vectors in Fig. 8b on July 28<sup>th</sup> and 31<sup>st</sup>). The  $O_3$  contribution from the industrial sector (whose precursors could come from facilities in the south of the MMA, Fig. 4) reinforces the  $O_3$  peaks up to  $\sim 10 \mu\text{g m}^{-3}$ , meanwhile the contribution from non-road transport increases systematically the background  $O_3$  concentration by  $\sim 15 \mu\text{g m}^{-3}$ . The  $O_3$  contribution from non-road transport in this region may arise mainly from the Madrid's airports and the agricultural machinery operating in the surrounding rural areas.

In Madrid, the model reproduces the  $O_3$ -traffic cycle featuring the typical low- $O_3$  concentrations ( $< 40 \mu\text{g m}^{-3}$ ) in the early morning and in the afternoon due to  $O_3$  titration (Fig. 8a). However,  $O_3$  was overestimated (MB type D) during daytime peaks due to the overestimation of the  $NO_2$  morning peaks during stagnant conditions. The  $NO_2$  overestimation correlates with the highest road transport contribution. The  $NO_2$  overestimation in Madrid is inconsistent with the overall underestimation of  $NO_2$  in urban areas due to the dilution effect. The results point towards a poor representation of the vertical mixing during the stagnant conditions. In urban areas downwind (Fig. 8b) the modelled  $O_3$  shows a positive bias (MB type B) due to the underestimation of  $NO_2$ . While traffic- $NO_2$  emissions in the largest urban areas (i.e. Madrid, Barcelona) are estimated at the road link level using specific data such as daily average traffic, mean speed circulation and real-world vehicle fleet composition profiles, in medium size cities this information is not available and traffic emissions are estimated using emission factors that depend on population density (Guevara et al., 2013), which implies having a larger uncertainty in the results.

The  $O_3$  contribution from the industrial sector (whose precursors could come from facilities in the southern MMA, Fig. 4) reinforces the  $O_3$  peaks up to  $\sim 10 \mu\text{g m}^{-3}$ , meanwhile the contribution from non-road transport increases systematically the

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background O<sub>3</sub> concentration by  $-15 \mu\text{g m}^{-3}$ . The O<sub>3</sub> contribution from non-road transport in this region may arise from the Atlantic shipping route, Madrid's airports and the agricultural machinery operating in the surrounding rural areas.

Figure 8a and 8b also shows that the urban O<sub>3</sub> plume presents different intensity and distribution depending on the meteorological pattern. Downwind of Madrid (Fig. 8b), stagnant conditions favour O<sub>3</sub> peaks exceeding  $120 \mu\text{g m}^{-3}$  and on-road transport contributes up to  $20\text{--}30 \mu\text{g m}^{-3}$  (e.g., July 25<sup>th</sup>). The highest O<sub>3</sub> concentrations ( $\sim 160 \mu\text{g m}^{-3}$ ) are modelled when westerly winds channelled along the Tajo valley carry the polluted air masses in a NE direction through the Tajo Valley, which results in a O<sub>3</sub> contribution of  $\sim 70 \mu\text{g m}^{-3}$  from road transport sector in areas downwind (see wind vectors in Fig. 8b on July 28<sup>th</sup> and 31<sup>st</sup>).

#### 3.4.2 The northeast of the North-Eastern Iberian Peninsula

Figure 8c and 8d show the source apportionment time series at two stations in the NEIP, an urban station in Barcelona (station 3) and a remote rural area downwind (station 4). ~~In Not surprisingly, at the urban station, NO<sub>2</sub> levels of up to  $100 \mu\text{g m}^{-3}$  affect O<sub>3</sub> concentrations by titration during traffic peaks. In contrast, the rural station downwind depicts a higher O<sub>3</sub> and lower NO<sub>2</sub> concentration than the urban station. Absolute O<sub>3</sub> biases in both stations are  $\sim 10 \mu\text{g m}^{-3}$  (MB type B).~~

~~Absolute O<sub>3</sub> biases in both stations are  $\sim 10 \mu\text{g m}^{-3}$  (MB type B). In the urban area, the modelled O<sub>3</sub> concentrations are in agreement with the observations. As at Madrid urban station, The model suggests that O<sub>3</sub> mostly results from import and from the NO titration effect due to local road transport and industrial sources (Fig. 8c). However, the O<sub>3</sub> diurnal cycle in the NEIP urban areas of the NEIP is less marked than in the CIP due to persistently high O<sub>3</sub> concentration at night ( $\sim 60 \mu\text{g m}^{-3}$ ). The breezes and mountain-valley winds contribute to the accumulation and recirculation of pollutants in this region. However, the model presents some under/overestimation at nighttime, which could result from the NO<sub>2</sub> underestimation and complex mesoscale phenomena not captured by the model. Note that the O<sub>3</sub> diurnal cycle in the NEIP urban areas is less marked than in the CIP due to persistently high O<sub>3</sub> concentration at night ( $\sim 60 \mu\text{g m}^{-3}$ ). The breezes and mountain-valley winds contribute to the accumulation and recirculation of pollutants in this region.~~

~~At the rural station, Despite of the slight overestimation of NO<sub>2</sub> concentration in rural areas, modelled O<sub>3</sub> peaks ( $> 120 \mu\text{g m}^{-3}$ ) are highly correlated in a good agreement with observations (Fig. 8d), which suggests that overall the model reproduces the main transport paths, photochemical processes, and relative contributions from different sources. Imported O<sub>3</sub> is one of the main contributors to ground-level O<sub>3</sub> (from  $40$  to  $100 \mu\text{g m}^{-3}$ ), but during peaks the on-road traffic contribution sharply increases up to  $80 \mu\text{g m}^{-3}$ .~~

The O<sub>3</sub> concentration from the road transport sector arriving at rural areas in the NEIP, mainly comes from the Barcelona and surroundings as a result of the afternoon sea breezes (see wind vectors in Fig. 8c and 8d). However, under specific

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meteorological patterns these winds also carry precursors from other cities located in the NW Mediterranean Basin. For example, on July 28<sup>th</sup> both the model and observations show that O<sub>3</sub> exceeded 120 µgm<sup>-3</sup> in both urban and rural areas and the model attributed ~60% of it to regional anthropogenic sources (non-imported O<sub>3</sub>). In the morning, two main meteorological phenomena favoured the transport of precursors towards the Western Mediterranean. First, the North African Thermal Low reinforced the sea breezes in the NEIP coast. Second, NW synoptic winds channelled between the Pyrenees and the French Central Massif towards the Gulf of Lion transported precursors from urban areas upwind (e.g., Marseille, Toulouse). In the afternoon, the sea breezes transported air masses enriched with O<sub>3</sub> and precursors inland. Other authors (Gangoiti et al., 2001) have hypothesized that high O<sub>3</sub> concentration in the Western Mediterranean Basin is influenced by transport from France via the Carcassonne gap. The present experiment cannot quantify the contribution of French cities to the O<sub>3</sub> concentration over the NEIP, but future studies could explicitly tag the emission from the French regions.

### 3.4.3 Guadalquivir Valley

The largest modelled O<sub>3</sub> overestimations are found in the Guadalquivir Valley (MB-type D). Precursors emitted along the Guadalquivir Valley arise from large industrial facilities and a coastal power plant (Fig. S3), a few densely populated cities (notably Sevilla and Cordoba with more than 700,000 and 320,000 inhabitants, respectively), and the navigable river. The airflow in the region is controlled by the Atlantic synoptic conditions, the development of breezes in the coast of the Gulf of Cadiz and the channelling through the Guadalquivir Valley (Adame et al., 2008).

We have selected two stations along the Guadalquivir Valley to evaluate O<sub>3</sub>, one in the urban area of Sevilla-Seville (station 5 in Fig. 9a), and one in a rural coastal area (station 6 in Fig. 9b). The contribution of anthropogenic sources to NO<sub>2</sub> at the coastal rural background station is low but still significant. The contribution of non-road transport is due the influence of one of the largest Spanish harbours. The contribution of the energy sector to the O<sub>3</sub> concentration is also noticed (e.g., July 25<sup>th</sup>). As expected, the urban station (Fig. 9a) shows a high NO<sub>2</sub> concentration dominated by on-road traffic. NO<sub>x</sub> from traffic is the main sink of O<sub>3</sub> in the city and the model reproduces the titration effect in agreement with observations (e.g., July 28<sup>th</sup>-31<sup>st</sup>).

High O<sub>3</sub> overestimations (10-30 µgm<sup>-3</sup>) at both stations (Fig. 9a and b) were detected during July 25<sup>th</sup>-28<sup>th</sup> which corresponds to intense and persistent SW winds transporting air masses from the Atlantic Sea along the Guadalquivir Valley, as shown by the wind vectors in Fig. 9a and b. Although the model overestimates O<sub>3</sub> concentrations, it reproduces the temporal variability.

Our results suggest that the non-road transport sector is a significant contributor along the Guadalquivir Valley during these days. The impact of shipping emission on O<sub>3</sub> in the Guadalquivir Valley region is mainly evidenced by the relative high NO<sub>x</sub> from ship exhaust (Fig. S3). NO<sub>2</sub> time series at the coastal station (Fig. 9b) indicates that the model overestimates NO<sub>2</sub> concentrations in hours where the NO<sub>2</sub> contribution from non-road transport is the highest. The unrealistic NO<sub>2</sub> peaks from non-road transport suggest that shipping emissions are overestimated in the HERMESv2.0 model, which HERMESv2.0 uses

the EMEP gridded emission ~~database-inventory~~ at 50 km x 50 km horizontal resolution to estimate shipping emissions and spatially distributes them to the 4-km model domain using the marine routes reported by Wang et al. (2008).

5 ~~A recent review on the state-of-the-art of marine traffic emissions (Russo et al., 2018) indicates that STEAM appears as the most reliable and detailed emissions inventory since it is based on Automatic Identification System data and specific vessel information, with a resolution of 2.5 x 2.5 km<sup>2</sup> (Jalkanen et al., 2016). A comparative analysis indicates that EMEP gridded inventories are overestimated, in particular over hotspots in the Mediterranean shipping routes, and underestimated in secondary routes.~~ Marine traffic emissions estimated with current state-of-the-art methods, such as the use of data provided by Automatic Identification System (AIS), show large discrepancies with the reported CEIP-EMEP gridded inventories both  
10 ~~in term of totals and spatial distribution (Jalkanen et al., 2012), suggesting that uncertainties associated to current official emissions may be large.~~ This factor, added to the 15% decrease of NO<sub>x</sub> shipping emissions observed in Europe between 2009 (HERMESv2.0 base year) and 2012 (EMEP CEIP, 2019~~http://www.ceip.at~~) ~~could-can at least partly~~ explain the discrepancies observed.

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15 ~~O<sub>3</sub> peaks are highly biased (by ~40-50 µg m<sup>-3</sup>) during stagnant conditions on July 28<sup>th</sup>-31<sup>st</sup>, especially in the urban station (station 5). During these days, measured NO<sub>2</sub> daily peaks increased up to 40 µg m<sup>-3</sup>. Although the model reproduces the variability, it overestimates the concentration due to an excessive on road transport contribution. As discussed for the CIP, under low dispersion conditions NO<sub>2</sub> concentrations are difficult to model due to complex mixing regimes.~~

#### 3.4.4 ~~The Northwestern~~ northwest of the Iberian Peninsula

20 Figures 9c and 9d show the source apportionment time series at one urban (station 7) and one rural (station 8) background station in the NWIP. The urban station (Fig. 9c) located in Santiago de Compostela, a medium size city with ~100,000 inhabitants, shows a high NO<sub>2</sub> concentration with a dominant contribution from the road transport sector. Traffic NO<sub>2</sub> is the main sink of urban O<sub>3</sub> via titration. Because NO<sub>2</sub> is underestimated, especially during stagnant conditions (July 24<sup>th</sup>-27<sup>th</sup>), O<sub>3</sub> concentrations are overestimated (MB category C).

25 ~~In rural background areas (Fig. 9d),~~ ~~Despite the O<sub>3</sub> biases during stagnant conditions,~~ the modelled O<sub>3</sub> concentration is in general agreement with observations ~~at the rural background station (Fig. 9d)(MB category B), but the model tends to underestimate the observed concentrations, particularly under weak winds.~~ NO<sub>2</sub> is likely ~~to be~~ underestimated due to missing traffic emissions. As noted previously, traffic emissions are poorly constrained in small and medium size cities, due to a lack  
30 of detailed information. There is also additional uncertainty in the precursors emitted from the large coal power plants and industries in the region (Valverde et al., 2016b). Our study uses emissions for 2009, and it has been estimated that between 2009 and 2012 energy production in coal-fired power plants increased from 13.1% to 19.4% (UNESA, 2012), ~~which. This implies~~ ~~implied~~ an increase ~~of in~~ NO<sub>x</sub> emissions from the power industry sector of around 19.5% (EMEP CEIP, 2019).

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~~Despite the O<sub>3</sub> biases during stagnant conditions, t~~The time series show that ~~the~~ model reproduces reasonably well the observed O<sub>3</sub> variability under different synoptic conditions. O<sub>3</sub> reaches the highest concentration (~100/150 µgm<sup>-3</sup> in urban/rural areas) under stagnant conditions (July 24<sup>th</sup>-27<sup>th</sup>) when the contribution of anthropogenic sources from all activity sectors is the highest (60-70%). O<sub>3</sub> concentrations decrease down to ~70 µgm<sup>-3</sup> under NW advective conditions (e.g., July 28<sup>th</sup>-30<sup>th</sup>) when the imported O<sub>3</sub> shows the highest contribution (80-90%). Saavedra et al. (2012) found that stationary anticyclones ~~over the NWIP~~~~over the northern~~ play an important role in the occurrence of high O<sub>3</sub> concentrations ~~over the NWIP~~. Our results show that under these stagnant conditions O<sub>3</sub> concentrations are due largely to in situ production (photochemistry) from on-road traffic, shipping, power plants, and industry in almost the same proportion.

### 3.5 Imported ozone

~~These O<sub>3</sub> source apportionment results~~~~The results presented before~~Our results indicate that imported O<sub>3</sub> represents the highest contribution to ground-level O<sub>3</sub> concentration in southwestern Europe. Imported O<sub>3</sub> enters the IP4 domain through the boundaries; it includes the contribution of O<sub>3</sub> from the EU12 domain, which in turn includes the contribution of hemispheric O<sub>3</sub> from the MOZART-4 global model. The imported O<sub>3</sub> is as large as the background O<sub>3</sub> regionally produced within the IP. ~~We~~~~Note~~ that the small biases at rural background stations obtained in the evaluation section indicate an overall high performance of the modelled background O<sub>3</sub> in the IP4 domain (Fig. S1). Given ~~its importance~~~~the important implications and robustness of this results~~, ~~this section is devoted to analyse the importance of the imported O<sub>3</sub> in southwestern Europe~~~~we further analyse this contribution below~~. In particular, we aim to understand its high contribution within the IP, ~~even~~, far from the model domain boundaries.

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Figure 10 shows the vertical cross-sections at 6, 12, and 18 UTC for O<sub>3</sub> and NO<sub>2</sub> at a constant latitude (40.38° N) on July 25<sup>th</sup>, 28<sup>th</sup> and 30<sup>th</sup>. ~~It helps to understand the vertical variability of both pollutants with the PBL dynamic as schematized by Millán et al. (1996).~~ The model predicts a pronounced O<sub>3</sub> vertical gradient above a height of 4 km above sea level (asl), showing that O<sub>3</sub> in the free troposphere is, to a large extent, imported to the IP4 domain (Safieddine et al., 2014). As expected, NO<sub>2</sub> mixing ratios show a negative gradient with altitude as it is mainly emitted at the surface. In the morning, the sun starts to heat up the ground, producing convective thermals and forcing the growth of the mixing layer. At noon, the mixing height reaches its maximum, being the highest in the CIP (2-4 km) and decreasing towards the coast (< 1 km) (Fig. 10). At the mixing layer top the O<sub>3</sub>-enriched air aloft is entrained into the mixing layer, mixing with O<sub>3</sub> and other pollutants produced locally within the mixing layer. When the mixing height decreases, O<sub>3</sub> is left in the free troposphere forming high O<sub>3</sub> residual layers (Gangoiti et al., 2001) that contribute to the regional transport. Over the following days, these residual layers (composed of imported O<sub>3</sub> and local O<sub>3</sub> produced within the domain over previous days) can be entrained by fumigation into the mixing layer to reach the surface. This fumigation effect, previously described in the eastern USA

(Zhang and Rao, 1999; Langford et al., 2015) and in the Western Mediterranean Basin (Kalabokas et al., 2017; Querol et al., 2018), leads to a rapid increase in O<sub>3</sub> concentrations at ground level. The accumulation and recirculation of air masses is intensified along the Eastern Mediterranean coast (Millán et al., 1996, 2000; Gangoiti et al., 2001; Querol et al., 2017) by the action of the breezes and mountain-valley winds. Furthermore, the small deposition velocity of O<sub>3</sub> over the sea and its high atmospheric lifetime in the free troposphere contributes to enrich the O<sub>3</sub> background concentration (c.a. several weeks, Monks et al., 2015; Seinfeld and Pandis, 2016).

The O<sub>3</sub> fumigation effect was particularly intense on July 30<sup>th</sup>-31<sup>st</sup>, when a high O<sub>3</sub> levels are found in the free troposphere compared to previous days (Fig. 10). The analysis of the O<sub>3</sub> concentration map with the imported O<sub>3</sub> contributions (Fig. 6) indicates that ground-based O<sub>3</sub> is neither advected nor titrated; therefore, it can only result from vertical mixing. The high O<sub>3</sub> mixing ratio in the free troposphere was mainly due to O<sub>3</sub> advection entering the Atlantic boundary driven by westerlies (Fig. 3). We hypothesize that two events may have contributed to the increase of the O<sub>3</sub> concentration by long-range transport. First, a low-pressure system in the British Islands on July 28<sup>th</sup> could have transported significant O<sub>3</sub> amounts from the stratosphere to the free troposphere (Fig. S4 Fig. S7). Second, O<sub>3</sub> episodes generated in mid-July over the Eastern USA predicted by the MOZART-4 model could have contributed to an increase in the O<sub>3</sub> transported from N America to Europe by the action of the prevailing westerlies (Fig. S7 Fig. S4) associated with cyclonic systems along the “warm conveyor belt” (Pausata, et al., 2012; Derwent et al., 2015).

#### 4 Discussion and conclusions

Our study has provided a first estimation of the main sources responsible for high O<sub>3</sub> concentration in the Western Mediterranean Basin during the episode period July 21<sup>st</sup> -31<sup>st</sup>, 2012. We used the Integrated Source Apportionment Method (ISAM) within the CALIOPE system to estimate the contribution of the main anthropogenic activity sectors compared to the imported concentrations to peak O<sub>3</sub> events in Spain compared to the imported O<sub>3</sub>. In addition, the use of ISAM has allowed an in-depth evaluation of the applied modelling system the model, and has provided a perspective on about its potential use for regulatory studies in non-attainment regions.

The results demonstrate that the O<sub>3</sub> problem over the Western Mediterranean Basin is local, regional, and hemispheric. Long-range transport of O<sub>3</sub> from beyond the IP domain is the main contributor to the ground-level daily mean O<sub>3</sub> concentration (~45%) during peak episodes. The imported O<sub>3</sub> contribution ranges from 40% during O<sub>3</sub> peaks to 80% at night or during well-ventilated conditions. The absolute imported O<sub>3</sub> is higher in the NEIP than in the central-CIP due to the recirculation and accumulation of pollutants along the Mediterranean. The high imported O<sub>3</sub> at the surface far away from the model boundaries is consistent with the high levels of O<sub>3</sub> in the free troposphere (resulting from local/regional layering and accumulation, and continental/hemispheric transport) along with intense vertical mixing during the day.

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Our results support the European Commission (EC, 2004) in pointing out that the effectiveness of abatement strategies for achieving compliance with the European air quality standards in southern Europe might be compromised by the long-range transport of O<sub>3</sub>. This is especially true in Mediterranean regions (i.e., NEIP, EV and EIP) where the contribution of imported O<sub>3</sub> is particularly dominant (60-68% of the daily mean O<sub>3</sub> concentration ~~in episodes~~) as a result of the accumulation and recirculation of pollutants over the Mediterranean Basin. In those areas, if the long-range transport of O<sub>3</sub> is not reduced, the mean background level will not decrease making it more vulnerable to exceedances of the O<sub>3</sub> target values by enhanced local production under stagnant conditions.

During high O<sub>3</sub> events, the imported O<sub>3</sub> is added to the formation from local and regional anthropogenic sectors. The road transport is an important contributor to the O<sub>3</sub> concentration in rural areas downwind of the large cities in Spain. It contributed up to 16-18% of the daily mean O<sub>3</sub> concentration under exceedances of the target value for human health protection, and up to 70 µg m<sup>-3</sup> on an hourly basis downwind of Barcelona and Madrid.

The non-road transport sector (including international shipping, airport and agricultural machinery) is as significant as road transport inland (10-19% of the daily mean O<sub>3</sub> concentration during the peaks). There is a high influence of international shipping (13%), affecting the coastal areas in the Mediterranean and the south of the IP (along the Strait of Gibraltar) with contributions of up to ~20 and ~30 µg/m<sup>3</sup>, respectively. Although the non-road transport contribution was found to be overestimated in coastal areas in the south of the IP in the present experiment, it cannot be neglected and actions controlling international shipping should be considered as important as those related with road transport especially in regions with big harbours (e.g., Huelva and Barcelona). Dalsøren et al. (2010) indicated that annual O<sub>3</sub> concentration is increasing yearly in a range of 0.5-2.5 ppb in areas impacted by shipping activities. Recent studies indicate that shipping emissions are projected to increase significantly due to increases in transportation demand and traffic. As the Strait of Gibraltar is the only shipping route connecting the Atlantic Ocean with the Mediterranean Basin, the regulation of these emissions is key in order to control O<sub>3</sub> exceedances in Spain and the Mediterranean Basin. Shipping emissions can be regulated by each country within 400 km of coastlines, but policy-induced controls for offshore emissions are very dependent on the success of adopted and proposed regulations within the International Maritime Organization.

The energy and industrial sectors contribute ~6-11% of the daily mean O<sub>3</sub> concentration during the peaks and over all the receptor regions. As they ~~usually~~ are usually injected in high altitudes, their contribution extends way beyond their surroundings. The energy combustion sector (3%) and industrial and non-industrial combustion sectors (3%) have a mean contribution of 2-4 µg m<sup>-3</sup>, going up to 4-6 µg m<sup>-3</sup> in NO<sub>x</sub>-limited areas (i.e., the western IP). In highly industrialized regions (i.e., Guadalquivir Basin and northwestern IP), abatement strategies affecting all sectors at regional scale could contribute to

decrease the local formation of O<sub>3</sub> as the regional/local anthropogenic contribution can be greater than 50% during several days.

In the Barcelona Metropolitan Area the contribution from energy and industrial sectors to the NO<sub>2</sub> concentrations can be in the same range than the contribution from road transport (~40-60%, Fig. 8c). In contrast, in areas downwind Barcelona the contribution from energy and industrial sectors to O<sub>3</sub> concentrations is relatively low compared with the contribution from road transport (Fig. 8d). The different contribution to O<sub>3</sub> concentration might ~~link with~~ be related to the different reactivity of VOC for O<sub>3</sub> formation. Each VOC emission source emits a different mix of VOC, which makes a different contribution to photochemical ozone formation. For example, in the UK, Derwent et al. (2007) showed a higher photochemical O<sub>3</sub> creation potential for road transport emissions than for production processes and combustion. Future national policy actions to control the emissions of VOC should tackle the sources that contribute more to photochemical O<sub>3</sub> formation.

The remaining sectors (i.e., SNAP 2, 5, 6, 9, and 11, see Table 1), are the fourth main contributors to the daily mean O<sub>3</sub> concentration during the days exceeding the target value (~2-8%). Future work should tag the biogenic sources as an individual sector as it is the main contributor to VOC emissions in Spain (i.e., ~70% in 2009 according to HERMESv2.0 model).

~~According to the model uncertainty, the~~ air quality and meteorological evaluation ~~of the model~~ indicate that uncertainties in our model are in the same range as the most recent inter-comparison studies using state-of-the-art air quality models. In addition, our model evaluation together with the source apportionment results has allowed a better understanding of the origin of model errors related to ~~the emission estimates~~. Our methodological choice has been to use a detailed bottom-up emission inventory instead of a typical top-down regional emission inventory. Bottom-up emissions, estimated using source-specific emission factors and activity statistics, accurately characterise pollutant sources and allow obtaining more realistic results than the ones reported by top-down or regional emission inventories. To understand the impact of the use of 2009 data to study year 2012, we revised the EMEP Centre on Emission Inventories and Projections (EMEP-CEIP), which collects and reviews the national emission inventories from Parties to the Convention on Long-range Transboundary Air Pollution. Between 2009 and 2012, total NO<sub>x</sub> and NMVOC emissions in Spain decreased by -10.6% and -10.7%, respectively (EMEP CEIP, 2019). For NO<sub>x</sub>, around 80% of this reduction is linked to a reduction of road transport emissions, whereas in the case of NMVOC ~50% of the reduction is due to a decrease ~~of~~ industrial emissions. For our modelling study, we consider these differences as small and acceptable, and not creating any major inconsistency. The difference of 10-15 % in emissions for certain precursors between 2009 and 2012 is within the typically larger ranges of uncertainty in emission inventories.

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Another relevant and uncertain source for  $O_3$  concentration is the urban-VOC emitted in urban areas emissions. Future research works should be devoted to the continuous monitoring of urban VOC and take advantage of satellite observations to improve speciation and spatial variability of urban VOC emissions.

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We have identified two sources of uncertainty in the modelling estimation of the imported  $O_3$  concentration. First, it depends on both the performance of the CALIOPE system over EU12, and on MOZART-4 at global scale. The small biases at rural background stations support an overall high performance baseline background  $O_3$  in the IP4 domain (Fig. S1). Second, our set-up involves some uncertainty in the estimation of the imported contribution of  $O_3$ . In reality, there is a fraction of the imported  $O_3$  that may have been generated within the IP4 domain before the period of simulation (including the spin-up). We have assumed that this fraction is negligible and future works should check to what extent this assumption is correct.

For regulatory applications, further source apportionment studies should target not only emissions from activity sectors, but also the source regions where the emission abatement strategies should be applied. In addition, future studies should preferentially cover multiple summer periods in order to improve representativeness. We note that our results cannot predict whether emission abatement will have either a positive or a negative effect in  $O_3$  changes due to the non-linearity of the  $O_3$  generation process. Subsequent source sensitivity analyses tailoring the identified main contribution sources could predict how  $O_3$  will respond to reductions in precursor emissions, which are essential to define the most efficient  $O_3$  abatement strategies in the Western Mediterranean Basin.

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Overall, we find that the imported  $O_3$  is the largest input to the ground-level  $O_3$  concentration in the IP during the studied episode. However, during stagnant conditions, the emission from local anthropogenic activities in the IP control the  $O_3$  peaks in areas downwind of the main urban and industrial regions. Furthermore, ground-level  $O_3$  concentrations are strongly affected by vertical downward mixing of  $O_3$ -rich layers in the free troposphere, which result from local/regional layering and accumulation, and continental/hemispheric transport. The importance of both imported and local contributions to the  $O_3$  peaks in the IP demonstrates the need for detailed quantification of both contributions to high  $O_3$  concentrations for local  $O_3$  management. Furthermore, the influence of local sources and topographical and meteorological conditions in the high  $O_3$  concentration indicate the importance of designing  $O_3$  abatement policies at local scale.

This work has identified and quantified the local and imported contributions to  $O_3$  concentration during an episode in a particular area in southwestern Europe. In addition, we have provided a perspective about the potential use of source apportionment method for regulatory studies in non-attainment regions. However, further  $O_3$  source apportionment studies targeting other non-attainment regions in Europe covering different episodes are necessary prior to defining local mitigation

~~measurements that complement national and European-wide abatement strategies to reach the efforts. European O<sub>3</sub> objectives.~~

We investigate the origin of ground-level O<sub>3</sub> concentration over southwestern Europe during the episode July 21<sup>st</sup>–31<sup>st</sup>, 2012. The period includes the two most frequent summer synoptic circulation patterns over the region: stagnant conditions with development of the Iberian Thermal Low (ITL) and North-Western Advection (NWAd). We used the Integrated Source Apportionment Method (ISAM) within the CALIOPE system to estimate the main sources responsible of the high O<sub>3</sub> concentrations in receptor regions over the IP compared to the imported (regional and hemispheric) O<sub>3</sub> contribution. In addition, the source apportionment method has allowed an in-depth evaluation of the modelling system applied. Below, we summarize the new findings from the source attribution study, we discuss the model uncertainties and we give a perspective on some of the regulatory implications, which may be applicable to other non-attainment regions.

#### 4.1 Source apportionment results

The results demonstrate that the O<sub>3</sub> problem over the Western Mediterranean Basin is local, regional, and hemispheric. Long-range transport of O<sub>3</sub> from beyond the IP domain is the main contributor to the ground-level daily mean O<sub>3</sub> concentration (~45%) during peak episodes. The imported O<sub>3</sub> contribution ranges from 40% during O<sub>3</sub> peaks to 80% at night or during well-ventilated conditions. The absolute imported O<sub>3</sub> is higher in NEIP than in the central IP due to the recirculation and accumulation of pollutants along the Mediterranean. The high imported O<sub>3</sub> at the surface far away from the model boundaries is consistent with the high levels of O<sub>3</sub> in the free troposphere (resulting from local/regional layering and accumulation, and continental/hemispheric transport) along with intense vertical mixing during the day.

During high O<sub>3</sub> events, the imported O<sub>3</sub> is added to the local and regional anthropogenic contributions. The road transport is an important contributor to the O<sub>3</sub> concentration in rural areas downwind of the large cities in Spain. It contributed up to 16–18% of the daily mean O<sub>3</sub> concentration under exceedances of the target value for human health protection, and up to 70  $\mu\text{g m}^{-3}$  on an hourly basis downwind of Barcelona and Madrid. The non-road transport sector is as significant as road transport inland (10–19% of the daily mean O<sub>3</sub> concentration during the peaks). There is a high influence of international shipping (13%), affecting the coastal areas in the Mediterranean and the southern IP (along the Strait of Gibraltar) with contributions of up to ~20 and ~30  $\mu\text{g m}^{-3}$ , respectively. The energy and industrial sectors contribute ~6–11% of the daily mean O<sub>3</sub> concentration during the peaks and over all the receptor regions. As they usually are injected in high altitudes, their contribution extend way beyond their surroundings. The energy combustion sector (3%) and industrial and non-industrial combustion sectors (3%) have a mean contribution of 2–4  $\mu\text{g m}^{-3}$ , going up to 4–6  $\mu\text{g m}^{-3}$  in NO<sub>x</sub>-limited areas (i.e., the western IP). The OTHR tag, which lumped the remaining sectors (i.e., SNAP 2, 5, 6, 9, and 11, see Table 1), is the fourth main contribution to the daily mean O<sub>3</sub> concentration during the days exceeding the target value (~2.8%). Future work

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should tag the biogenic sources as an individual sector as it is the main contributor to VOC emissions in Spain (i.e., ~70% in 2009 according to HERMESv2.0 model).

#### 4.2 Model uncertainty

The modelled concentrations present uncertainties that are in the same range as the most recent inter-comparison studies using state-of-the-art air quality models. Our results show a good agreement between the modelled and observed  $O_3$ , which provides confidence in our source apportionment results.

In addition, our model evaluation together with the source apportionment results has allowed a better understanding of the origin of model errors. Under stagnant conditions, the model shows lower performances: in coastal areas, this may be related to a poor representation of mesoscales processes such as sea breezes and recirculation; in Madrid,  $NO_2$  overestimations may be due to a poor representation of the vertical mixing. There is likely an overestimation of shipping emissions, and an underestimation of road transport emissions in medium-size cities.

Overall, the estimated imported  $O_3$  depends on both the performance of the CALIOPE system over EU12, and on MOZART 4 at global scale. The evaluation of model in the EU12 domain using the European EIONET stations (not shown here) showed good skills for  $O_3$  and its precursors. Furthermore, the source apportionment results have shown that  $O_3$  local production could be as large as the imported  $O_3$ , so the small biases at rural background stations indicate an overall high performance baseline background  $O_3$  in the IP4 domain (Fig. S1). This study has not evaluated MOZART 4, but there are many studies supporting its good performance. Several inter-comparison studies (e.g., EURODELTA and AQMEII) also found the importance of chemical boundary conditions to reproduce the  $O_3$  background concentrations (Giordano et al., 2015; Bessagnet et al., 2016; Solazzo et al., 2017).

Our set-up involves some uncertainty in the estimation of the imported contribution of  $O_3$ . In reality, there is a fraction of the imported  $O_3$  that may have been generated within the IP4 domain before the period of simulation (including the spin-up). We have assumed that this fraction is negligible and future works should check to what extent this assumption is correct.

#### 4.3 Regulatory implications

Our study has identified the main sources responsible of high  $O_3$  in receptor regions in the Western Mediterranean Basin, but cannot predict whether emission abatement will have either a positive or negative effect in  $O_3$  changes due to the non-linearity of the  $O_3$  generation process (Monks et al., 2015 and papers therein). Subsequent source sensitivity analysis tailoring the identified main contribution sources could predict how  $O_3$  will respond to reductions in precursor emissions. Future source sensitivity studies are essential to define the most efficient abatement strategies to reduce high  $O_3$  concentrations in the Western Mediterranean Basin. Complementarily, it would be important to know where the emission

abatement strategies should be applied. To answer this question further source apportionment studies should be conducted tagging both source activities and source regions.

5 This case study supports a relatively old conclusion by the European Commission (EC, 2004) pointing out that the effectiveness of abatement strategies for achieving compliance with the European air quality standards in southern Europe might be compromised by the long-range transport of O<sub>3</sub>. This is especially true in Mediterranean regions (i.e., NEIP, EV and EIP) where the contribution of imported O<sub>3</sub> is particularly dominant (60-68% of the daily mean O<sub>3</sub> concentration in episodes) as a result of the accumulation and recirculation of pollutants over the Mediterranean Basin. In those areas, if the long-range transport of O<sub>3</sub> is not reduced, the mean background level will not decrease making it more vulnerable to exceedances of the O<sub>3</sub> target values by enhanced local production under stagnant conditions.

10 Reduction of the background O<sub>3</sub> concentration can be achieved by decreasing the concentrations of O<sub>3</sub> and precursors aloft. Given the polluted dome aloft the Western Mediterranean Basin during a typical O<sub>3</sub> episode, which may result from local/regional layering and accumulation, and continental/hemispheric transport, it will necessary and important to implement emission reduction strategies on the hemispheric and continental scale. There is an urgent need of international collaboration at European and hemispheric level to reduce O<sub>3</sub> and its precursors. High background contribution to Iberian O<sub>3</sub> could represent a substantial future challenge to the attainment of O<sub>3</sub> limit values, which requires controlling hemispheric O<sub>3</sub> (EC, 2004; Fowler et al., 2013).

15 Besides the long range transport contribution, O<sub>3</sub> peaks during episodes result from the local/regional photochemical formation from the transport and industrial sectors, and therefore regional emission reduction strategies are also urgent. The contribution from road transport determines O<sub>3</sub> peaks in rural areas (up to 70 µg/m<sup>3</sup>) downwind of the main Spanish Metropolitan Areas (i.e., Madrid and Barcelona). To reduce O<sub>3</sub> peaks downwind of the Madrid city, a decrease of road transport emissions in the urban area would be effective. However, to reduce O<sub>3</sub> peaks downwind of the Barcelona city, the reduction of road traffic emission in the city should be complemented with abatement strategies in other Mediterranean cities, especially in Southern France.

20 In the BMA the contribution from energy and industrial sectors to NO<sub>2</sub> concentration can be in the same range than the contribution from road transport (~40-60%, Fig. 8c). In contrast, in areas downwind of the BMA the contribution from energy and industrial sectors to O<sub>3</sub> concentrations is relatively low compared with the contribution from road transport (Fig. 8d). The differences contribution to O<sub>3</sub> concentration might link with the different reactivity of VOC for O<sub>3</sub> formation. Each VOC emission source emits a different mix of VOC, which makes a different contribution to photochemical ozone formation. For example, in the UK, Derwent et al. (2007) showed a higher photochemical O<sub>3</sub> creation potential for road

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transport emissions than for production processes and combustion. Future national policy actions to control the emissions of VOC should tackle the sources that contribute more to photochemical O<sub>3</sub> formation.

5 Although the non road transport contribution was found to be overestimated in coastal areas in the south of the IP in the present experiment, it cannot be neglected and actions controlling international shipping should be considered as important as those related with road transport especially in regions with big harbours (e.g., Huelva and Barcelona). Dalsøren et al. (2010) indicated that annual O<sub>3</sub> concentration is increasing yearly in a range of 0.5–2.5 ppb in areas impacted by shipping activities. Recent studies indicate that shipping emissions are projected to increase significantly due to increases in transportation demand and traffic. As the Strait of Gibraltar is the only shipping route connecting the Atlantic Ocean with the  
10 Mediterranean Basin, the regulation of these emissions is key in order to control O<sub>3</sub> exceedances in Spain and the Mediterranean Basin. Shipping emissions can be regulated by each country within 400 km of coastlines, but policy-induced controls for offshore emissions are very dependent on the success of adopted and proposed regulations within the International Maritime Organization.

15 In highly industrialized regions (i.e., Guadalquivir Basin and northwestern IP), abatement strategies affecting all sectors at regional scale could contribute to decrease the local formation of O<sub>3</sub> as the regional/local anthropogenic contribution can be greater than 50% during several days.

20 Overall, we find that the imported O<sub>3</sub> is the largest input to the ground level O<sub>3</sub> concentration in the IP during episodes. In addition, during stagnant conditions, the emission from local anthropogenic activities in the IP control the O<sub>3</sub> peaks in areas downwind of the main urban and industrial regions. Furthermore, ground level O<sub>3</sub> concentrations are strongly affected by vertical downward mixing of O<sub>3</sub>-rich layers in the free troposphere, which result from local/regional layering and accumulation, and continental/hemispheric transport. The importance of both imported and local contributions to the O<sub>3</sub> peaks in the IP demonstrates the need for detailed quantification of both contributions to high O<sub>3</sub> concentrations for local O<sub>3</sub>  
25 management. Furthermore, the influence of local sources and topographical and meteorological conditions in the high O<sub>3</sub> concentration indicate the importance of designing O<sub>3</sub> abatement policies at local scale.

This work has identified local and imported contributions to O<sub>3</sub> concentration during an episode in a particular area in southwestern Europe. However, further detailed and longer O<sub>3</sub> source apportionment studies targeting other nonattainment  
30 regions in Europe are necessary prior to define local mitigation measurements that complement national and European wide strategies to reach the European O<sub>3</sub> objectives.

### Data availability

Air quality measurements are available at the EIONET database (<https://www.eea.europa.eu/data-and-maps/data/airbase-the-european-air-quality-database-7>). The CMAQ code is available at the <https://www.cmascenter.org/>. The WRF-ARW code is available at [http://www2.mmm.ucar.edu/wrf/users/download/get\\_source.html](http://www2.mmm.ucar.edu/wrf/users/download/get_source.html). Download Mozart-4 outputs are available at <http://www.acd.ucar.edu/wrf-chem/mozart.shtml>. The program that generates CMAQ boundary conditions from MOZART-4 output is available at: <http://www.camx.com/getmedia/a0c2d710-6187-47bb-af3f-bd9b4a4d19bb/mozart2camx-12jul17.tgz>. The HERMESv2 outputs are available upon request. The modelled data in this study are available upon request from the corresponding author.

## 10 Appendix A

Table 1A shows the acronyms used in the text.

BCON	Chemical boundary conditions to IP4, also named as imported O <sub>3</sub> contribution
BMA	Barcelona Metropolitan Area
CIP	Central Iberian Peninsula
CMAQ	Community Multiscale Air Quality model
EIP	Eastern Iberian Peninsula
EU12	European domain at 12-km horizontal resolution (12x12 km <sup>2</sup> )
EV	Ebro Valley
GV	Guadalquivir Valley
IP	Iberian Peninsula
IP4	Iberian Peninsula domain at 4-km horizontal resolution (4x4 km <sup>2</sup> )
ISAM	Integrated Source Apportionment Method
MDA8	Maximum daily 8-hour average concentration
MMA	Madrid Metropolitan Area
MED	Mediterranean Sea
NEIP	Northeast of the <del>em</del> Iberian Peninsula
NIP	North of the <del>em</del> Iberian Peninsula
NWIP	Northwest of the <del>em</del> Iberian Peninsula
SIP	South of the <del>em</del> Iberian Peninsula
SNAP1	Emission sector on combustion in energy
SNAP34	Emission sector on combustion and processes in industry
SNAP7	Emission sector on road transport, exhaust and non-exhaust
SNAP8	Emission sector on non-road transport (international shipping, airport and agricultural machinery )
WIP	West of the <del>em</del> Iberian Peninsula
WRF	Weather Research and Forecasting Model

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## Appendix B

Definition of the discrete statistics used in the evaluation: correlation coefficient ([r](#), Eq. B1), mean bias ([MB](#), Eq. [B1B2](#)), [normalized mean bias \(NMB, Eq. B3\)](#) and root mean squared error ([RMSE](#), Eq. [B3B4](#)). Where  $C_m(x, t)$  and  $C_o(x, t)$  are the modelled and observed concentrations at a location ( $x$ ) and time ( $t$ );  $N$  is the number of pairs of data.  $\overline{C_m}$  and  $\overline{C_o}$  are the modelled and observed mean concentrations over the whole period, respectively.

$$r = \frac{\sum_{i=1}^N (C_m(x, t) - \overline{C_m})(C_o(x, t) - \overline{C_o})}{\sqrt{\sum_{i=1}^N (C_m(x, t) - \overline{C_m})^2} \sqrt{\sum_{i=1}^N (C_o(x, t) - \overline{C_o})^2}} \quad (\text{B1})$$

$$MB = \frac{1}{N} \sum_{i=1}^N (C_m(x, t) - C_o(x, t)) \quad (\text{B2})$$

$$NMB = \frac{\sum_{i=1}^N (C_m(x, t) - C_o(x, t))}{\sum_{i=1}^N (C_o(x, t))} \times 100 = \left(\frac{C_m}{C_o} - 1\right) \times 100 \quad (\text{B2B3})$$

$$RMSE = \frac{1}{N} \sqrt{\sum_{i=1}^N (C_m(x, t) - C_o(x, t))^2} \quad (\text{B3B4})$$

### Author contribution

[MTP designed the research. SN supported on the ISAM set-up for this experiment. MTP performed the model simulations. All the authors analysed the data and discussed the results. MTP and CPGP wrote the paper.](#)

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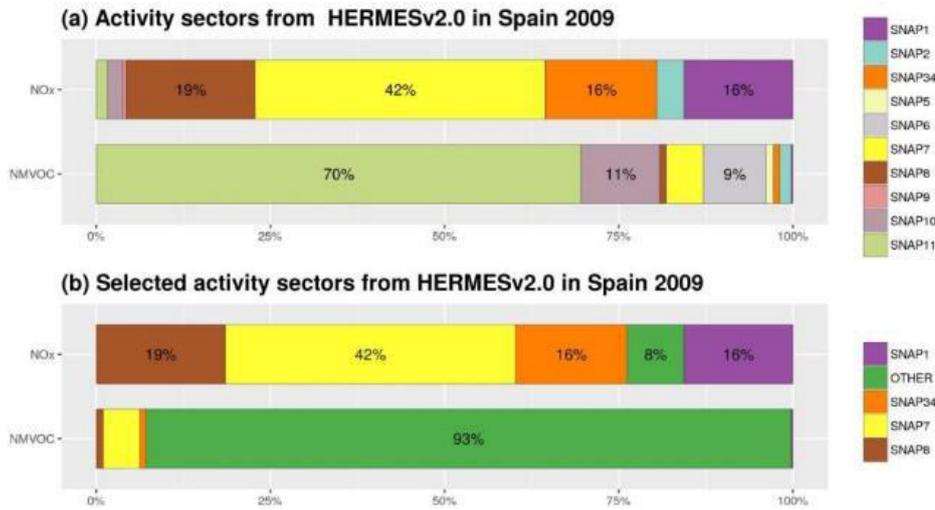


Figure 1: Percentage of the contribution of emissions to total annual emissions by SNAP sector calculated by HERMES for Spain 2009 (a) and for the selected SNAP sector accounting for more than 90% of NO<sub>x</sub> total emission to be tracked with ISAM (b). “OTHER” compiles the SNAP categories 2 (Residential combustion), 5 (Fugitive emissions from fuels), 6 (Solvent use), 9 (Waste management), 10 (Agriculture) and 11 (Other sources).

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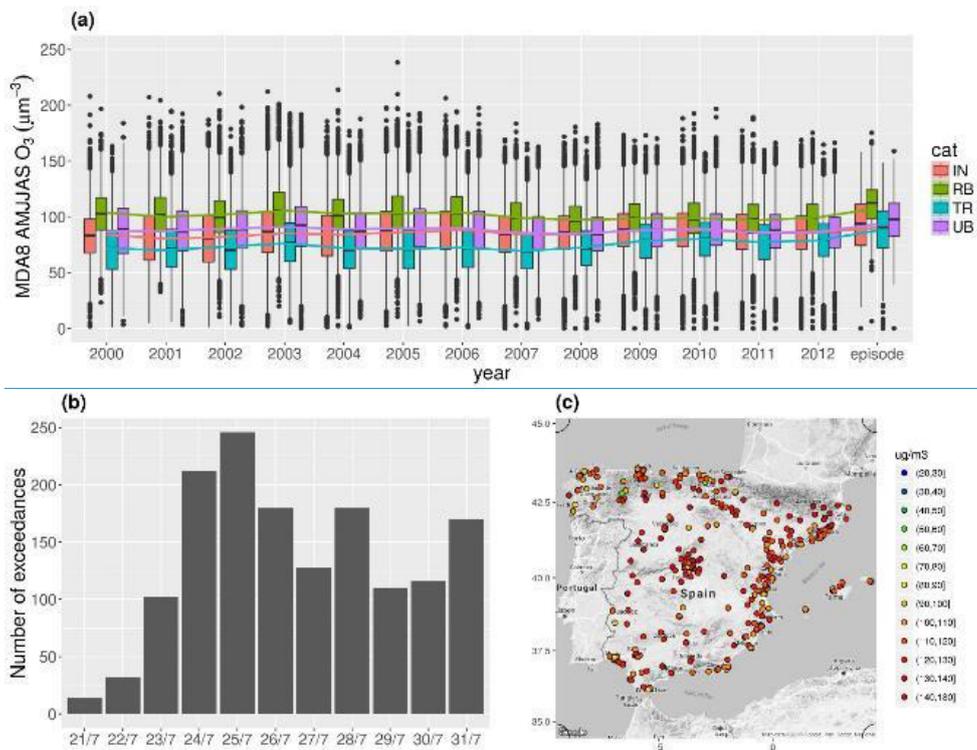
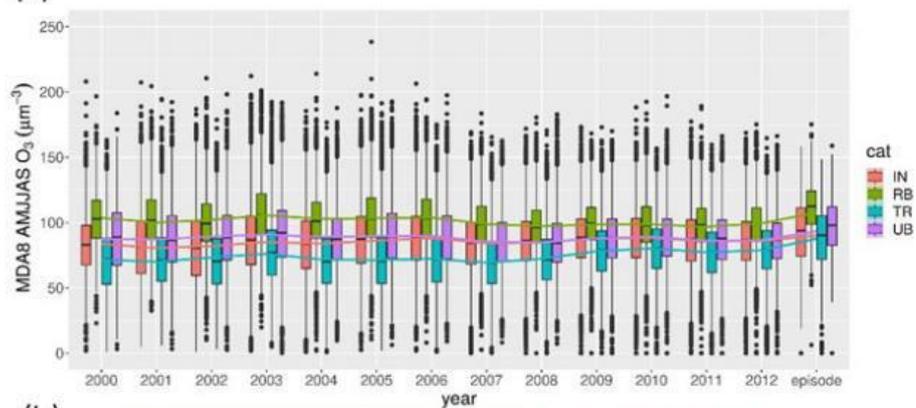


Figure 2: (a) Temporal distribution of the MDAS O<sub>3</sub> concentration during the extended summer (from April to September, AMJJAS) at the Spanish EIONET stations for the period 2000-2012 and the episode (from 21/07/2012 to 31/07/2012) by station type: IN (industrial), RB (rural background), TR (traffic) and UB (urban background). (b) Number of days exceeding the O<sub>3</sub> Target Value (120 µg/m<sup>3</sup>) by each day of the episode. (c) 90<sup>th</sup> percentile of the MDAS O<sub>3</sub> concentration at the Spanish EIONET stations during the

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episode:

(a)



(b)



Figure 2: (a) Temporal distribution of the MDA8 O<sub>3</sub> concentration during the extended summer (from April to September, AMJJAS) at the Spanish EIONET stations for the period 2000-2012 and the episode (from 21/07/2012 to 31/07/2012) by station type: IN (industrial), RB (rural background), TR (traffic) and UB (urban background). (b) 90<sup>th</sup> percentile of the MDA8 O<sub>3</sub> concentration at the Spanish EIONET stations during the episode. Numbers indicate the stations cited in section 3.4.

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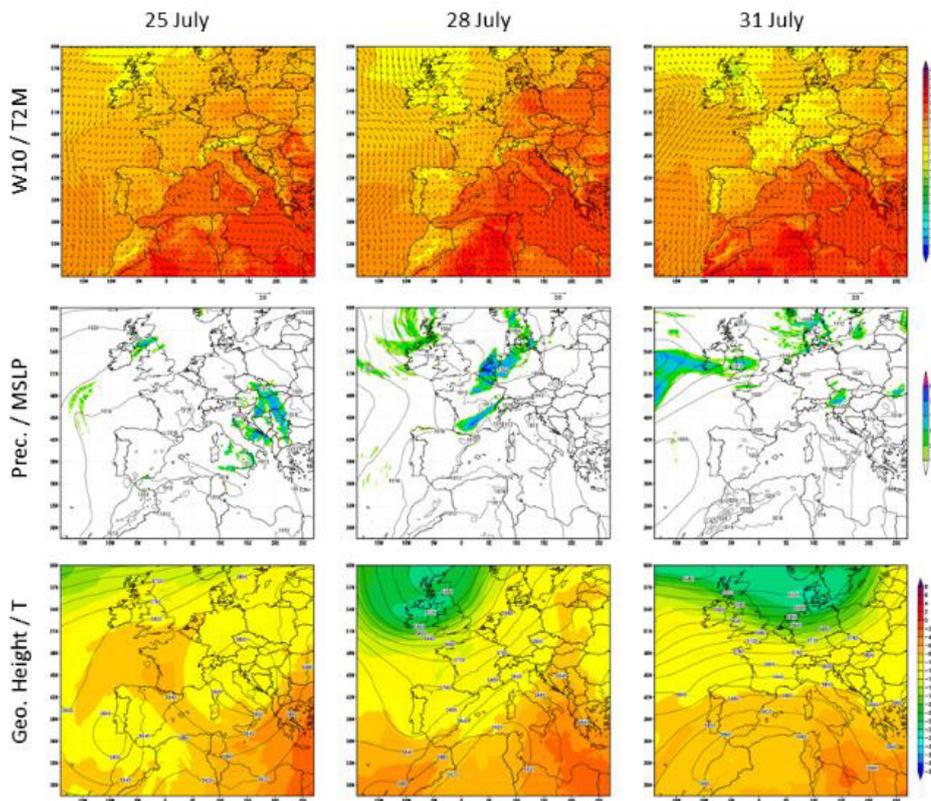


Figure 3: WRF-ARW meteorological fields at 6 UTC for July 25<sup>th</sup>, 28<sup>th</sup> and 31<sup>st</sup> in the EU12 domain: 10-m wind speed (W10,  $\text{ms}^{-1}$ ), 2-m temperature (T2M, C), 6 h accumulated precipitation (Prec., mm), mean sea level pressure (MSLP, hPa), 500-hPa geopotential height in contours (Geo. Height, m), and 500-hPa temperature in shaded colors (T, °C).

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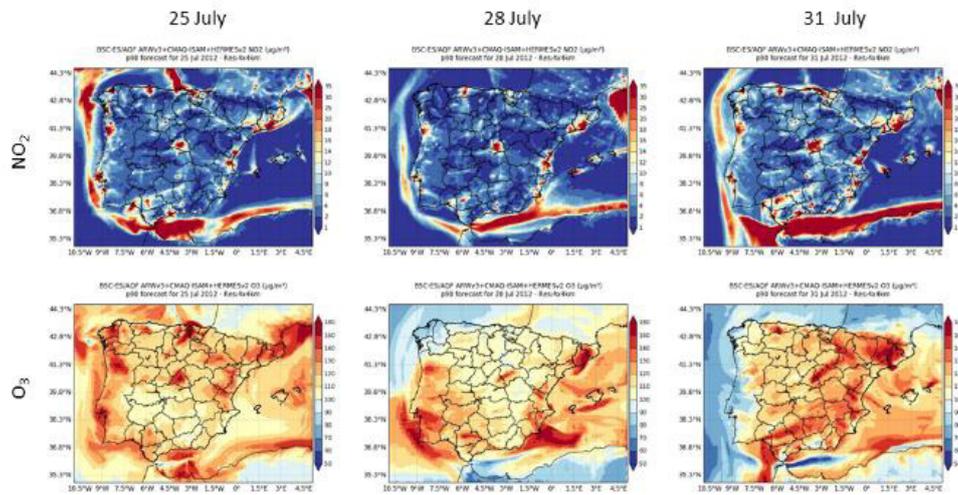
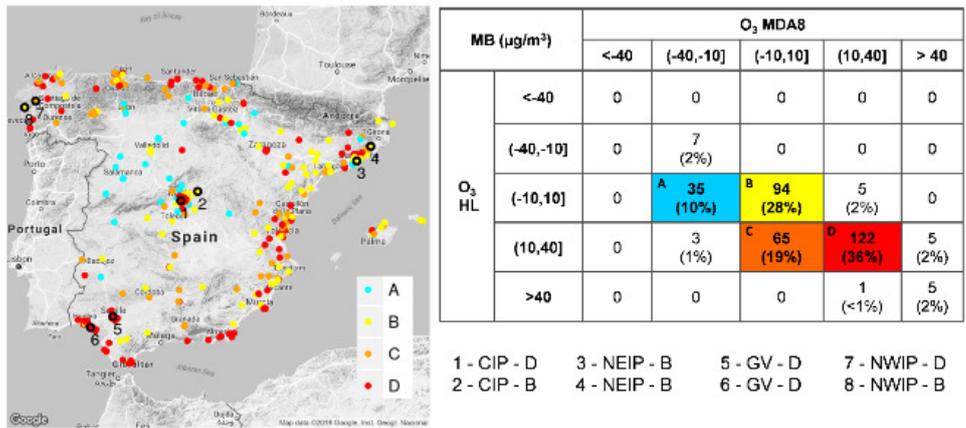


Figure 4: Ground-based concentration maps (in  $\mu\text{g m}^{-3}$ ) for  $\text{NO}_2$  (first row) and  $\text{O}_3$  (second row) corresponding to the 90<sup>th</sup> percentile of the average hourly concentrations on 25<sup>th</sup> (first column), 28<sup>th</sup> (second column) and 31<sup>st</sup> (third columns) July 2012.



5 Figure 5: Air quality stations classified by both mean bias (MB, in  $\mu\text{g m}^{-3}$ ) for average hourly and MDA8 O<sub>3</sub> at the Spanish EIONET stations and lumped by categories (A, B, C, and D). Numbered black circles indicate stations under study in Central IP (CIP; station 1 and 2), Northeastern IP (NEIP; station 3 and 4), Guadalquivir Valley (GV; station 5 and 6), and Northwestern IP (NWIP; station 7 and 8).

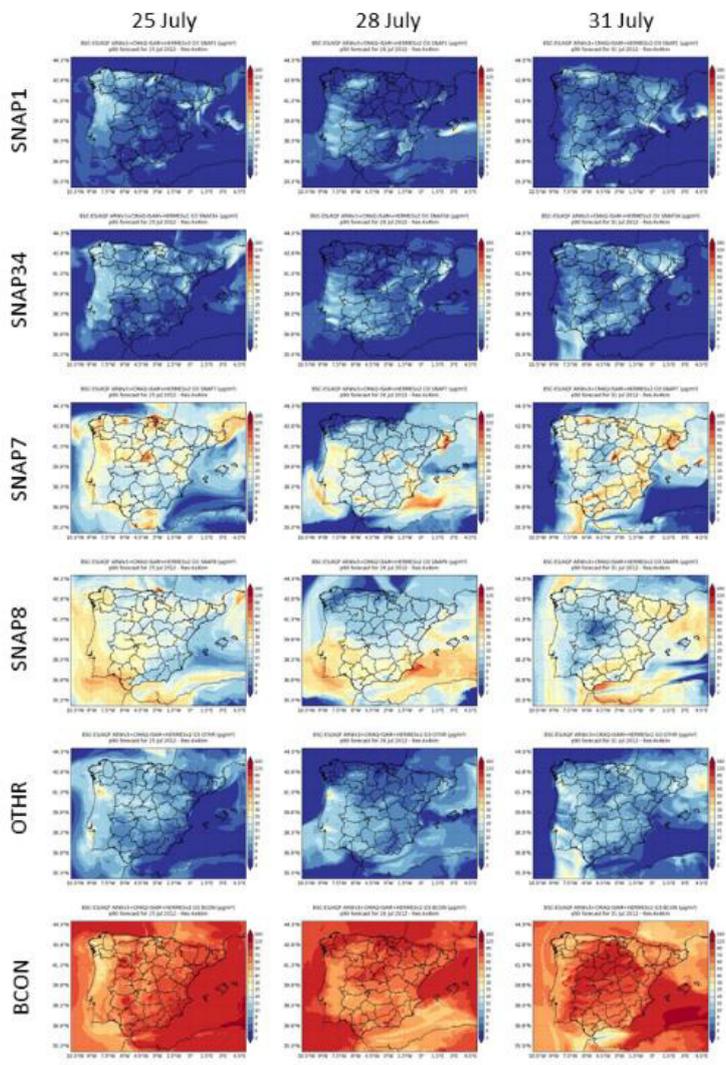
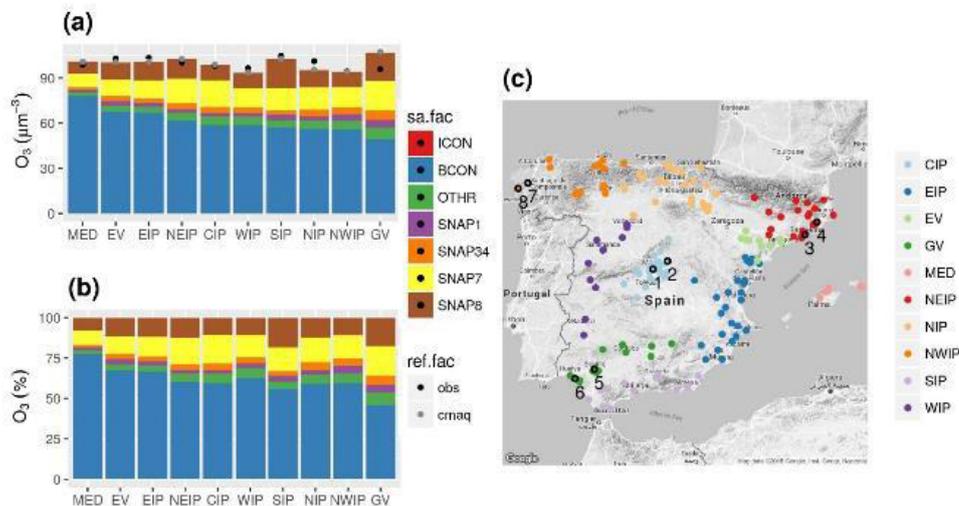


Figure 6: Tagged O<sub>3</sub> concentrations (in  $\mu\text{g}\text{m}^{-3}$ ) corresponding to the 90<sup>th</sup> percentile (90p) of the average hourly concentrations: SNAP1, SNAP34, SNAP7, SNAP8, OTHER, and BCON for July 25<sup>th</sup> (first column), 28<sup>th</sup> (second column) and 31<sup>st</sup> (third columns) in 2012.



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Figure 7: Daily mean contribution in  $\mu\text{g}\text{m}^{-3}$  (a) and in percentage (b) of tagged sources to O<sub>3</sub> during exceedances of the observed 120  $\mu\text{g}\text{m}^{-3}$  for MDA8 O<sub>3</sub> averaged by the identified receptor regions (c). Black and grey dots represent observed and modelled daily mean concentration during exceedances of 120  $\mu\text{g}\text{m}^{-3}$  of the observed MDA8 O<sub>3</sub>. Regions correspond to the Center of the IP (CIP), the Eastern of the IP (EIP), the Ebro Valley (EV), the Guadalquivir Valley (GV), the Mediterranean Sea (MED), the North-Eastern of the IP (NEIP), the Northern of the IP (NIP), the North-Western of the IP (NWIP), the Southern of the IP (SIP) and the Western of the IP (WIP). Numbered black circles indicate stations under study CI (1-2), NEIP (3-4), GV (5-6) and NWIP (7-8).

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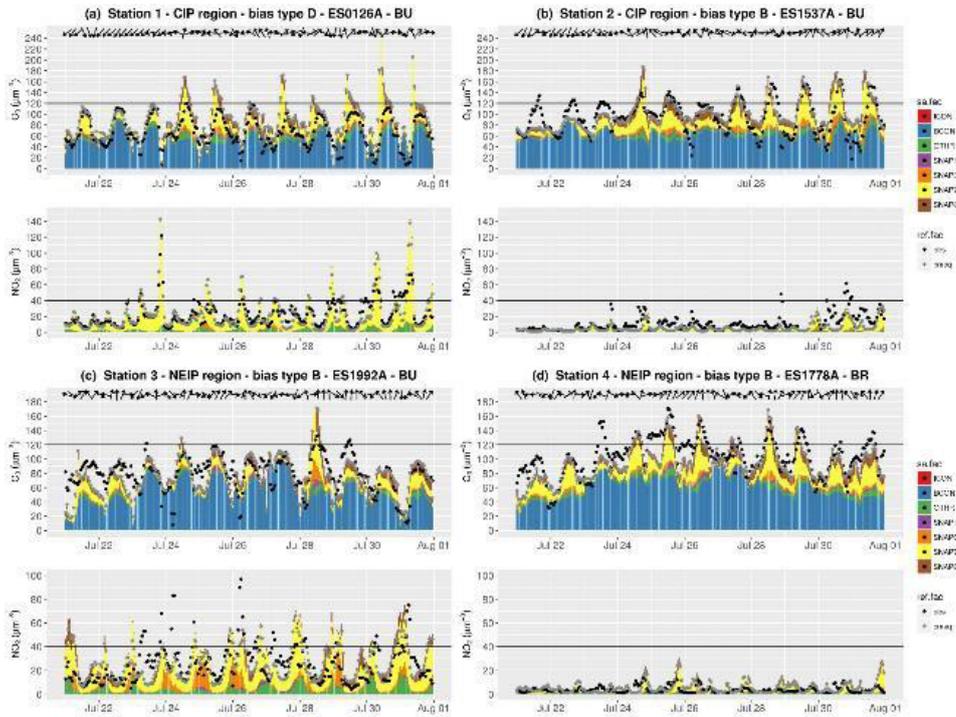


Figure 8: Source apportionment time series for  $O_3$  and  $NO_2$  concentrations (in  $\mu\text{g m}^{-3}$ ) in the episode at the selected stations in the Central-center of the IP (CIP) region (a,b), and in the North-Eastern of the IP (NEIP) (c,d). Color bars (sa.fac) indicate the  $O_3$  tags. Black and grey dots (ref.fac) indicate the observed and modelled concentrations, respectively. Black horizontal lines represent  $O_3$  target value ( $120 \mu\text{g m}^{-3}$ ) and  $NO_2$  limit value ( $40 \mu\text{g m}^{-3}$ ) as a reference. The location of the stations is shown in Fig. 7 by the corresponding numbers.

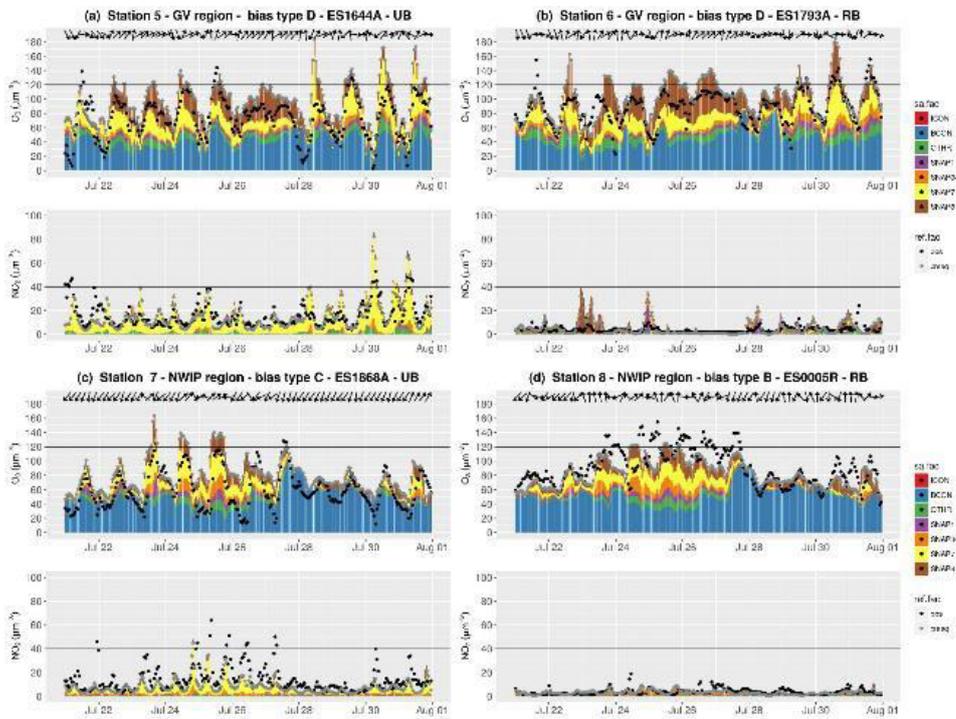


Figure 9: Source apportionment time series for  $O_3$  and  $NO_2$  concentrations (in  $\mu g m^{-3}$ ) in the episode at the selected stations in the Guadalquivir Valley (GV) (a,b), and in the northwest of the IP (NWIP) (c,d). Color bars (sa.fac) indicate the  $O_3$  tags. Black and grey dots (ref.fac) indicate the observed and modelled concentrations, respectively. Black horizontal lines represent  $O_3$  target value ( $120 \mu g m^{-3}$ ) and  $NO_2$  limit value ( $40 \mu g m^{-3}$ ) as a reference. The location of the stations is shown in Fig. 7 by the corresponding numbers.

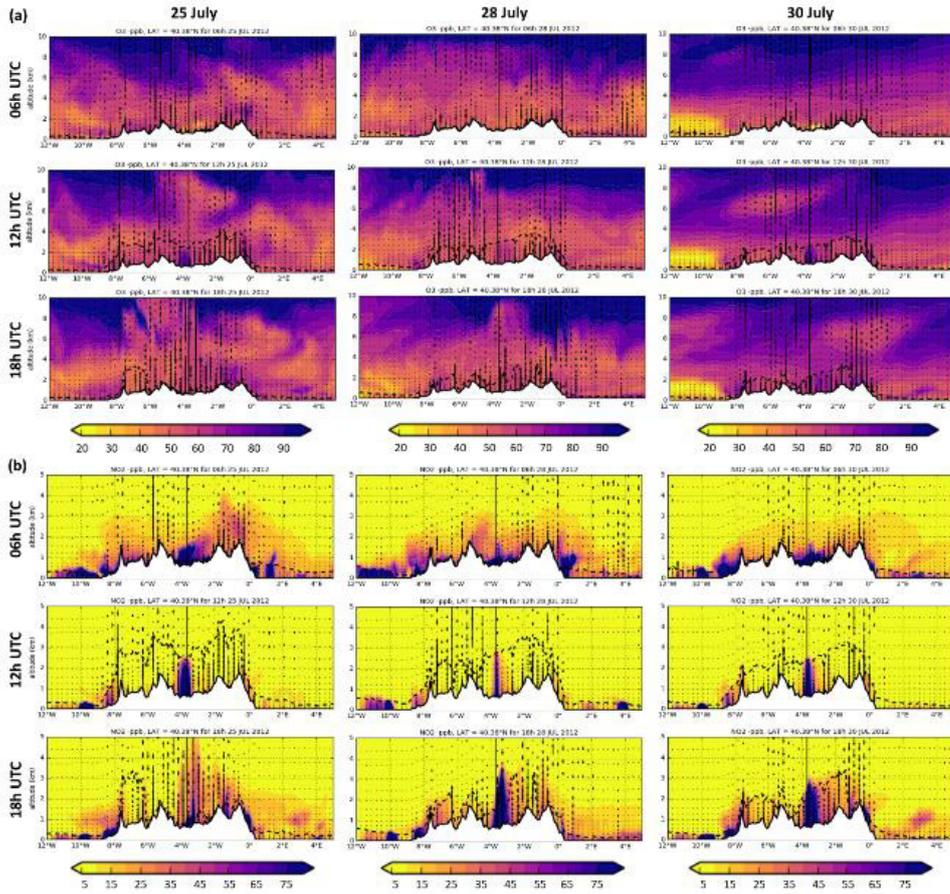


Figure 9: Cross section of modelled mixing ratios (in ppb) for O<sub>3</sub> (a) and NO<sub>2</sub> (b) at a constant latitude (Latitude = 40.38°, Madrid city) of the daily Iberian Thermal Low circulation, equivalent to the conceptual scheme of Millán et al. (1996) for July 25<sup>th</sup> (first column), 28<sup>th</sup> (second column) and 30<sup>th</sup> (third column) at 06, 12 and 18 UTC. Dot lines indicate the boundaries of the planetary boundary layer height. Vertical arrows indicate the vertical wind. Up arrows depict positive winds. Note the different scale for the y-axis between O<sub>3</sub> and NO<sub>2</sub>.

**Table 1: Description of the O<sub>3</sub> tagged sources in the present study.**

ISAM tag*	Emission by SNAP category	Description
SNAP1	SNAP1	SNAP1: Energy industry
SNAP34	SNAP34	SNAP34: <a href="#">Manufacturing Industries</a> (combustion and processes-)
SNAP7	SNAP7	SNAP7: Road transport, exhaust and non-exhaust
SNAP8	SNAP8	SNAP8: Non-road transport (international shipping, <a href="#">airport and agricultural machinery</a> )
OTHR	SNAP2 + SNAP5 + SNAP6 + SNAP9 + SNAP10 + SNAP11	SNAP2: residential and commercial/institutional combustion SNAP5: Fugitive emissions from fuels SNAP6: Product use including solvents SNAP9: waste management SNAP10: Agriculture SNAP11: Other sinks
BCON	-	Chemical boundary conditions to IP4 domain from the EU12 simulation which includes the contribution from Europe and international contribution from MOZART-4. O <sub>3</sub> external contribution
ICON	-	Initial chemical condition of the domain IP4

\*Each ISAM tag is applied to O<sub>3</sub> and its precursor species in the CB05 (NO<sub>x</sub> and VOCs). NO<sub>x</sub> species contributing to O<sub>3</sub> formation involve (9 species): NO, NO<sub>2</sub>, nitrogen trioxide (NO<sub>3</sub>), dinitrogen pentoxide (N<sub>2</sub>O<sub>5</sub>), nitrous acid (HONO), peroxyacyl nitrates (PAN), higher peroxyacyl nitrates (PANX), peroxyacetic acid (PNA), and organic nitrates (NTR). VOC species contributing to O<sub>3</sub> formation include (14 species): acetaldehyde (ALD2), higher aldehydes (ALDX), ethene (ETH), ethane (ETHA), ethanol (ETOH), formaldehyde (FORM), internal olefin (IOLE), isoprene (ISOP), methanol (MEOH), olefin (OLE), paraffin (PAR), monoterpene (TERP), toluene (TOL), and xylene (XYL).

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Table 2: Statistics for average hourly O<sub>3</sub>, MDA8 O<sub>3</sub>, and average hourly NO<sub>2</sub> concentrations in the episode as a function of the station type. Exceedances indicate the number of exceedances of the European Air Quality Directive Standards for hourly O<sub>3</sub> (180 µg<sup>m</sup><sup>-3</sup>), MDA8 O<sub>3</sub> (120 µg<sup>m</sup><sup>-3</sup>) and average hourly NO<sub>2</sub> (200 µg<sup>m</sup><sup>-3</sup>). N indicates the number of monitoring stations used in the statistics calculation. MO and MM depict the measured and modelled mean concentrations, respectively. Statistics are calculated by considering more than 75% of the hours in a day, as established by Directive 2008/50/EC. The statistics correspond to following quantiles 50<sup>th</sup> (25<sup>th</sup>, 75<sup>th</sup>) by station. Type indicates the station categories in the calculation of statistics: all the stations (ALL), industrial (IN), traffic (TR), urban background (UB), suburban background (SB) and rural background (RB) stations. Note that there are stations that do not fit in any of these four classifications.

Pollutant	Type	Exceedances:		N	MO (µg <sup>m</sup> <sup>-3</sup> )	MM (µg <sup>m</sup> <sup>-3</sup> )	MB (µg <sup>m</sup> <sup>-3</sup> )	NMB (%)	RMSE (µg <sup>m</sup> <sup>-3</sup> )		
		Observed/	Modelled								
Hourly O <sub>3</sub>	ALL	26/216		348	77.3 (66.7,86.5)	89.9 (83.5,95.6)	12.6 (4.4,20.2)	16.8 (5.1,29.2)	26.7 (21.5,32.1)	0.60.65 (0.60,57.0-70.72)	Formatted Table
	IN	5/23		106	74.1 (62.2,83.2)	82.2 (83.4,94.5)	14.1 (4.5,21.5)	19.8 (5.5,34.7)	26.8 (21.1,32.8)	0.70.66 (0.60,56.0-70.73)	
	TR	0/58		70	68.4 (57.1,76.7)	83.8 (74.4,89.2)	15.9 (8,21.8)	23.4 (10.1,38.5)	28.8 (24.5,33.9)	0.60.63 (0.50,54.0-70.70)	
	UB	0/78		56	74.8 (64.1,80.0)	89.4 (81.4,94.5)	15.6 (7.8,21.3)	22.1 (10.2,31.1)	28.2 (24.5,33.5)	0.70.65 (0.60,60.0-70.70)	
	SB	4/48		44	83.2 (78.1,87.2)	93.5 (88.2,97.3)	10 (4.6,14.6)	12.9 (5.4,17.5)	25.1 (21.4,29.3)	0.66 (0.60,0.72)	
	RB	17/9		66	91.2 (80.9,97.1)	96.2 (92.6,99.6)	4.5 (-3.3,14.7)	4.8 (-3.2,17.2)	21.2 (18.3,28.0)	0.70.67 (0.60,57.0-70.72)	Formatted Table
	MDA8 O <sub>3</sub>	ALL	751/822		348	101 (91.0,113.4)	106.2 (101.8,113.2)	5.7 (-3.7,16.9)	5.5 (-3.3,17.4)	17.9 (13.5,25.4)	0.60.64 (0.40,39.0-80.78)
	IN	204/187		106	96.7 (88.9,109.5)	104.8 (99.4,109.2)	5.7 (-2.9,16.5)	5.7 (-2.9,17.8)	17.0 (13.0,24.5)	0.70.68 (0.50,48.0-80.82)	
	TR	62/145		70	92.7 (83.6,99.9)	104.1 (98.3,110.2)	13.3 (4.6,23.8)	15.1 (4.6,27.2)	21.2 (15.2,31.2)	0.50.54 (0.30,27.0-80.76)	
	UB	86/174		56	100.0 (89.3,105.5)	107.6 (102.2,122.8)	14.1 (2.1,21.7)	14.8 (2.1,22.5)	22.6 (15.6,28)	0.60.59 (0.30,34.0-80.77)	
	SB	139/131		44	110.7 (99.8,118.5)	110.0 (105.5,119.2)	3.1 (-6.2,10.5)	2.9 (-5.4,10.1)	16.9 (13.6,22)	0.5 (0.31,0.67)	
	RB	260/185		66	113.1 (104.9,119.9)	108.6 (104.5,113.2)	-3.7 (-11.1,5.5)	-3.3 (-9.1,5.1)	15 (11.8,20.2)	0.70.72 (0.60,60.0-70.72)	Formatted Table
Hourly NO <sub>2</sub>	ALL	3/0		358	14.0 (8.4,21.2)	9.9 (4.4,16.2)	-4.1 (-8.4,-0.5)	-31.9 (-54.2,-4.9)	12.8 (8.2,17.6)	0.40.43 (0.30,29.0-60.55)	
	IN	0/0		120	11.5 (7.9,16.3)	8.8 (3.8,14.7)	-3.0 (-6.6,0.2)	-28.6 (-55.0,1.1)	11.0 (7.9,15.6)	0.40.40 (0.30,26.0-60.55)	Formatted Table
	TR	3/0		95	22.9 (17.7,28.7)	14.1 (8.7,21.6)	-7.4 (-13.9,-2.6)	-35.2 (-58.1,-14.3)	18.4 (14.1,23.1)	0.40.42 (0.30,28.0-60.57)	
	UB	0/0		63	17.0 (13,21.2)	13.2 (7.8,18.4)	-4.0 (-8,-0.3)	-26.9 (-47.4,-2)	14.1 (10.5,17.3)	0.50.49 (0.40,36.0-60.63)	
	SB	0/0		33	12 (8.2,14.6)	7.6 (4.8,11.1)	-2.8 (-6.5,0.8)	-34.4 (-48.6,7.5)	6.4 (7.9,13.6)	0.5 (0.41,0.65)	
RB	0/0		41	4.3 (3.5,9.3)	2.8 (1.6,4.0)	-2.0 (-5.3,-0.3)	-49.5 (-72.3,-9.7)	4.4 (2.7,7.1)	0.30.34 (0.20,24.0-70.72)	Formatted Table	

