

1 **XCO₂ in an emission hot-spot region: the COCCON Paris campaign 2015**

2

3 F.R. Vogel^{1,a}, M. Frey², J. Staufe^{3,a}, F. Hase², G. Broquet⁴, I. Xueref-Remy^{5,a}, F. Chevallier⁴,
4 P. Ciais⁴, M. K. Sha^{6,b}, P. Chelin⁷, P. Jeseck⁸, C. Janssen, Y. V. Te⁸, J. Groß², T.
5 Blumenstock², Q. Tu² and J. Orphal²

6

7 1. Climate Research Division, Environment and Climate Change Canada, Toronto, Canada

8 2. Karlsruhe Institute of Technology (KIT), Institute of Meteorology and Climate Research
9 (IMK), Karlsruhe, Germany

10 3. Thales, Labège, France

11 4. Laboratoire des Sciences du Climat et de l'Environnement (LSCE), IPSL, CEA-CNRS-
12 UVSQ, Université Paris-Saclay, Gif-Sur-Yvette, France

13 5. Observatoire de Haute Provence, OSU Pytheas, Saint-Michel l'Observatoire, France

14 6. Royal Belgian Institute for Space Aeronomy, Brussels, Belgium

15 7. Laboratoire Inter-Universitaire des Systèmes Atmosphériques (LISA), (CNRS UMR 7583,
16 Université Paris Est Créteil, Université Paris Diderot, Institut Pierre Simon Laplace), Créteil,
17 France.

18 8. Laboratoire d'Études du Rayonnement et de la Matière en Astrophysique et Atmosphères
19 (LERMA), IPSL, Sorbonne Universités, (CNRS, PSL Research University, Observatoire de
20 Paris), Paris, France

21

22 Previously at

23 a. Laboratoire des Sciences du Climat et de l'Environnement (LSCE), IPSL, CEA-CNRS-
24 UVSQ, Université Paris-Saclay, Gif-Sur-Yvette, France

25 b. Karlsruhe Institute of Technology (KIT), Institute of Meteorology and Climate Research
26 (IMK), Karlsruhe, Germany

27

28 *Correspondence to: Felix R. Vogel (Felix.Vogel@canada.ca)*

29

30 **Abstract.** Providing timely information on urban Greenhouse-Gas (GHG) emissions and their
31 trends to stakeholders relies on reliable measurements of atmospheric concentrations and the
32 understanding of how local emissions and atmospheric transport influence these
33 observations.

34 Portable Fourier Transform Infra-Red (FTIR) spectrometers were deployed at 5 stations in the
35 Paris metropolitan area to provide column-averaged concentrations of CO₂ (XCO₂) during a
36 field campaign in spring of 2015, as part of the Collaborative Carbon Column Observing
37 Network (COCCON). Here, we describe and analyze the variations of XCO₂ observed at

38 different sites and how they changed over time. We find that observations upwind and
39 downwind of the city centre differ significantly in their XCO₂ concentrations, while the overall
40 variability of the daily cycle is similar, i.e., increasing during night-time with a strong decrease
41 (typically 2-3 ppm) during the afternoon.

42 An atmospheric transport model framework (CHIMERE-CAMS) was used to simulate XCO₂
43 and predict the same behaviour seen in the observations, which supports key findings, e.g.
44 that even in a densely populated region like Paris (over 12 Million people), biospheric uptake
45 of CO₂ can be of major influence on daily XCO₂ variations. Despite a general offset between
46 modelled and observed XCO₂, the model correctly predicts the impact of the meteorological
47 parameters (e.g. wind direction and speed) on the concentration gradients between different
48 stations. Looking at the local gradients of XCO₂ for upwind and downwind station pairs, is
49 found to be less sensitive to changes in XCO₂ boundary conditions and biogenic fluxes within
50 the domain and we find the model-data agreement further improves. Our modelling framework
51 indicates that the local XCO₂ gradient between the stations is dominated by the fossil fuel CO₂
52 signal of the Paris metropolitan area. This further highlights the potential usefulness of XCO₂
53 observations to help optimise future urban GHG emission estimates.

54

55 **1 Introduction**

56 Atmospheric background concentrations of CO₂ measured since 1958 in Mauna Loa, USA,
57 have passed the symbolic milestone of 400 ppm (monthly mean) as of 2013 [Jones 2013].
58 Properly quantifying fossil fuel CO₂ emissions (FFCO₂) can contribute to define effective
59 climate mitigation strategies. Focussing our attention on cities is a critical part of this
60 endeavour as emissions from urban areas are currently estimated to represent from 53 % to
61 87 % of global FFCO₂, depending on the accounting method considered, and are predicted to
62 increase further [IPCC-WG3 2014, IEA 2008, Dhakal 2009]. As stated in the IPCC 5th
63 assessment report, “*current and future urbanisations trends are significantly different from the*
64 *past*” and “*no single factor explains variations in per-capita emissions across cities and there*
65 *are significant differences in per capita greenhouse gas (GHG) emissions between cities*
66 *within a single country*” [IPCC-WG3 2014]. Therefore, findings in one city can often not be
67 simply extrapolated to other urban regions. Furthermore, the large uncertainty of the global
68 contribution of urban areas to CO₂ emissions today and in the future is why a new generation
69 of city-scale observing and modelling systems are needed.

70 In recent years, more and more atmospheric networks have emerged that observe GHG
71 concentrations using the atmosphere as a large-scale integrator, for example in Paris, France
72 (e.g., Bréon et al. 2015, Xueref-Remy et al, 2018), Indianapolis, USA (e.g. Turnbull et al. 2015,
73 Lauvaux et al. 2016), Salt Lake City, USA (Strong et al. 2011, Mitchell et al. 2018), Heidelberg,
74 Germany (e.g. Levin et al. 2011, Vogel et al. 2013) and Toronto, Canada (e.g. Vogel et al.

75 2012). The air measured at in-situ ground-based stations is considered to be representative
76 of surface CO₂ fluxes of a larger surrounding area (1 km²-10000 km²), i.e. the emissions of
77 the Greater Paris Area dominate the airshed of the Ile-de-France (ca. 12'000 km²) (Staufer et
78 al. 2016). If CO₂ measurements are performed both up-wind and downwind of a city, the
79 concentration gradient between the two locations is influenced by the local net emission
80 strength between both sites and atmospheric mixing [Xueref-Remy et al, 2018, Bréon et al.
81 2015, Turnbull et al. 2015]. To derive quantitative flux estimates, measured concentration data
82 are typically assimilated into numerical atmospheric transport models which calculate the
83 impact of atmospheric mixing on concentration gradients for a given flux space-time
84 distribution. Such a data assimilation framework implemented for Paris with three atmospheric
85 CO₂ measurement sites [Xueref-Remy et al, 2018] previously allowed deriving quantitative
86 estimates of monthly emissions and their uncertainties over one year [Staufer et al. 2016].
87 Space-borne measurements of the column-average dry air mole fraction of CO₂ (XCO₂) are
88 increasingly considered for the monitoring of urban CO₂. This potential was shown with OCO-
89 2 and GOSAT XCO₂ measurements, even though the spatial coverage and temporal sampling
90 frequency of these two instruments were not optimized for FFCO₂ [Kort et al., 2012,
91 Janardanan et al. 2016, Schwandner et al. 2017], while other space-borne sensors dedicated
92 to FFCO₂ and with an imaging capability are in preparation [O'Brien et al, 2016, Broquet et al.
93 2017]. Important challenges of satellite measurements are that they are not as accurate as in-
94 situ ones, having larger systematic errors, while the XCO₂ gradients in the column are typically
95 7-8 times smaller than in the boundary layer. Another difficulty of space-borne imagery with
96 passive instruments is that they will only sample city XCO₂ plumes during clear sky conditions
97 for geostationary satellites and with an additional constraint to observations at around mid-
98 day for low-earth orbiting satellites.

99 The recent development of a robust portable ground-based FTIR (Fourier Transform InfraRed)
100 spectrometer as described in Gisi et al. [2012] and Hase et al. [2015] (EM27/SUN, Bruker
101 Optik, Germany) greatly facilitates the measurement of XCO₂ from the surface, with better
102 accuracy than from space and with the possibility of continuous daytime observation during
103 clear sky conditions. Typical compatibility (uncorrected bias) of the EM27/SUN retrievals of
104 the different instruments in a local network is better than 0.01 % (i.e. ` 0.04 ppm) after a careful
105 calibration procedure and a harmonized processing scheme for all spectrometers [Frey et al.
106 2015]. The Collaborative Carbon Column Observing Network (COCCON) [Frey et al. 2018]
107 intends to offer such a framework for operating the EM27/SUN. This type of spectrometer
108 therefore represents a remarkable opportunity to document XCO₂ variability in cities as a direct
109 way to estimate FFCO₂ [Hase et al. 2015] or in preparation of satellite missions.

110 When future low-Earth-orbit operational satellites with imaging passive spectrometers of
111 suitable capabilities to invert FFCO₂ will sample different cities, this will likely be limited to clear

112 sky conditions and at a time of the day close to local noon. Increasing the density of the
113 COCCON network stations around cities will allow to evaluate those XCO₂ measurements and
114 to monitor XCO₂ during the early morning and afternoon periods which will not be sampled
115 with satellites low-earth orbit satellite., From geostationary orbit, which can also have other
116 benefits, those time-periods can however be observed and could be compared to ground-
117 based measurements [e.g. Butz et al., 2015, O'Brien et al. 2016].

118 This study focuses on the measurements of XCO₂ from ground based EM27/SUN
119 spectrometers deployed within the Paris metropolitan area during a field campaign in the
120 spring of 2015, and modelling results. This campaign can be seen as a demonstration of the
121 COCCON network concept applied to the quantification of an urban FFCO₂ source. Several
122 spectrometers were operated by different research groups, while closely following the
123 common procedures suggested by Frey et al. [2015]. The paper is organised as follows. After
124 the instrumental and modelling setup descriptions of section 2, the observations of the field
125 campaign and the modelling results will be presented in section 3. Results are discussed in
126 section 4 together with the study conclusions.

127

128 **2 Methods and materials**

129 **2.1 Description of study area and field campaign design**

130 During the COCCON field campaign (April 28th to May 13th, 2015) five portable FTIR
131 spectrometers (EM27/SUN, Bruker Optik, Karlsruhe, Germany) were deployed in the Parisian
132 region (administratively known as *Île-de-France*) and within the city of Paris. The campaign
133 was conducted in early spring as the cloud cover is typically low in April and May and the time
134 between sunrise and sunset is more than 14 hours.

135 The Paris metropolitan area houses over 12 million people, with about 2.2 million inhabiting
136 the city of Paris. This urban region is the most densely populated in France with ~1000
137 inhabitants/km² and over 21000 inhabitants/km² for the city of Paris itself [INSEE 2016 -
138 <https://www.insee.fr/fr/statistiques>]. The estimated CO₂ emissions from the metropolitan
139 region are 39 Mt/year, according to the air quality association (AIRPARIF), that monitors the
140 airshed of Greater Paris. On-road traffic emissions, residential and the tertiary (i.e.
141 commercial) sector are the main sources (accounting for over 75 %), and minor contributions
142 from other sectors such as industrial sources and airports [<https://www.airparif.asso.fr/en/>,
143 AIRPARIF 2016]. It was crucial to understand the spatial distribution of these CO₂ sources to
144 optimally deploy the COCCON spectrometers. To this end a 1 km emission model for France
145 by IER (Institut fuer Energiewirtschaft und Rationelle Energieanwendung, University of
146 Stuttgart, Germany) was used as a starting point [Latoska 2009]. This emission inventory is
147 based on the available activity data such as, e.g., traffic counts, housing statistics, or energy
148 use, and the temporal disaggregation was implemented according to Vogel et al. [2013]. In

149 brief, the total emissions of the IER model were re-scaled to match the temporal factors for
150 the different emission sectors according to known national temporal emission profiles.

151 To quantify the impact of urban emissions on XCO₂, the FTIR instruments were deployed
152 along the dominant wind directions in this region in spring, i.e., southwesterly [Staufer et al
153 2016], in order to maximize the likelihood to capture upwind and downwind air masses (see
154 Figure 1). The two southwesterly sites (GIF and RES) are located in a less densely populated
155 area, where emissions are typically lower than in the city centre, where the station JUS is
156 located. The data in Fig. 1 show that the densest FFCO₂ emission area extends northwards
157 and eastwards. The two Northwesterly sites (PIS and MIT) were placed downwind of this area.
158 All instruments were operated manually and typically started operations around 7-8 am local
159 time from which they continuously observe XCO₂ until 5-6 pm.

160

161 **2.2 Instrumentation, calibration, and data processing**

162 The EM27/SUN is a portable FTIR spectrometer which has been described in detail in, e.g.,
163 Gisi et al. [2012] and Frey et al. [2015]. Here, only a short overview is given. The centre piece
164 of the instrument is a Michelson interferometer which splits up the incoming solar radiation
165 into two beams. After inserting a path difference between the beams, the partial beams are
166 recombined. The modulated signal is detected by an InGaAs detector covering the spectral
167 domain from 5000 to 11000 cm⁻¹ and is called an interferogram. As the EM27/SUN analyzes
168 solar radiation, it can only operate in daylight sunny conditions. A Fourier transform of the
169 interferogram generates the spectrum and a DC correction is applied to remove the
170 background signal and only keep the AC signal (see Keppel-Aleks et al. [2007]). A numerical
171 fitting procedure (PROFFIT code) [Schneider and Hase et al., 2009] then retrieves column
172 abundances of the concentrations of the observed gases from the spectrum. The single-
173 channel EM27/SUN is able to measure total columns of O₂, CO₂, CH₄ and H₂O. The ratio over
174 the observed O₂ column, assumed to be known and constant, delivers the column-averaged
175 trace gas concentrations of XCO₂, XCH₄ in μmol / mol dry air, with a temporal resolution of
176 one minute. XCO₂ is the dry air mole fraction of CO₂, defined as $XCO_2 = \text{Column}[CO_2] /$
177 $\text{Column}[\text{Dry Air}]$. Applying the ratio over the observed oxygen (O₂) column reduces the effect
178 of various possible systematic errors; see Wunch et al. (2011).

179 In order to correctly quantify small differences in XCO₂ columns between Paris city upstream
180 and downstream locations, measurements were performed with the five FTIR instruments side
181 by side before and after the campaign, as we expect small calibration differences between the
182 different instruments due to slightly different alignment for each individual spectrometer. These
183 differences are constant over time and can be easily accounted for by applying a calibration
184 factor for each instrument. Previous studies showed that the instrument specific corrections
185 are well below 0.1 % for XCO₂ [Frey et al. 2015, Chen et al. 2016] and are stable for individual

186 devices. The 1-sigma precision for XCO₂ is in the order of 0.01 % - 0.02 % (< 0.08 ppm) e. g.
187 [Gisi et al. 2012, Chen et al. 2016, Hedelius et al. 2016, Klappenbach et al. 2015]. The
188 calibration measurements for this campaign were performed in Karlsruhe w.r.t. the Total
189 Carbon Column Observing Network (TCCON) [Wunch et al. 2011] spectrometer at the
190 Karlsruhe Institute of Technology (KIT), Germany for 7 days before the Paris campaign
191 between April 9th and 23rd, and after the campaign on May 18th until 21st.

192 Figure S1 (left panel) shows the XCO₂ time series of the calibration campaign, where small
193 offsets between the instruments raw data are visible. As these offsets are constant over time,
194 a calibration factor for each instrument can be easily applied; actually these are the calibration
195 factors previously found for the Berlin campaign [Frey et al. 2015]. These factors are given in
196 Table 1, where all EM27/SUN instruments are scaled to match instrument No. 1. The
197 calibrated XCO₂ values for April 15th are shown in Fig. S1 None of the five instruments that
198 participated in the Berlin campaign show any significant drift; in other words, the calibration
199 factors found one year before were still applicable. This is a good demonstration of the
200 instrument stability stated in section 2.2, especially as several instruments (Nos. 1, 3, 5) were
201 used in another campaign in Northern Germany in the meantime. The EM27/SUN XCO₂
202 measurements can also be made traceable to the WMO international scale for in-situ
203 measurements by comparison with measurements of a collocated TCCON spectrometer
204 which are calibrated against in-situ standards by aircraft and aircore measurements [Wunch
205 et al. 2010, Messerschmidt et al. 2012] performed using the WMO scale.

206 During the campaign and for the calibration measurements we recorded double-sided
207 interferograms with 0.5 cm⁻¹ spectral resolution. Each measurement of 58 s duration consisted
208 of 10 scans using a scanner velocity of 10 kHz. For precise timekeeping, we used GPS
209 sensors for each spectrometer.

210 In-situ surface pressure data used for the analysis of the calibration measurements performed
211 at KIT have been recorded at the co-located meteorological tall tower. During the campaign,
212 a MHD-382SD data-logger recorded local pressure, temperature and relative humidity at each
213 station. The analysis of the trace gases from the measured spectra for the calibration
214 measurements has been performed as described by Frey et al. [2015]. For the campaign
215 measurements we assume a common vertical pressure-temperature profile for all sites,
216 provided by the model, so that the surface pressure at each spectrometer only differs due to
217 different site altitudes. The 3-hourly temperature profile from the European Centre for Medium-
218 Range Weather Forecasts (ECMWF) operational analyses interpolated for site JUS located in
219 the centre of the array was used for the spectra analysis at all sites. The individual ground-
220 pressure was derived from site altitudes and pressure measurements performed at each site.
221 Before and after the Paris campaign, side by side comparison measurements were performed
222 with all 5 EM27/SUN spectrometers and the TCCON spectrometer operated in Karlsruhe at

223 KIT. All spectrometers were placed on the top of the IMK office building North of Karlsruhe.
224 The altitude is 133 m above sea level (a.s.l.), coordinates are 49.09° N and 8.43° E. The
225 processing of the Paris raw observations (measured interferograms) were performed as
226 described by Gisi et al. [2012] and Frey et al. [2015] for the Berlin campaign: spectra were
227 generated applying a DC correction, a Norton-Beer medium apodization function and a
228 spectral resampling of the sampling grid resulting from the FFT on a minimally sampled
229 spectral grid. PROFFWD was used as the radiative transfer model and PROFFIT as the
230 retrieval code.

231

232 **2.3 Atmospheric transport modelling framework**

233 We used the chemistry transport model CHIMERE (Menut et al., 2013) to simulate CO₂
234 concentrations in the Paris area. More specifically, we used the CHIMERE configuration over
235 which the inversion system of Bréon et al. [2015] and Staufer et al. [2016] was built to derive
236 monthly to 6-hour mean estimates of the CO₂ Paris emissions. Its horizontal grid, and thus its
237 domain and its spatial resolution, are illustrated in Figure S2. It has a 2 × 2 km² spatial
238 resolution for the Paris region, and 2 × 10 km² and 10 × 10 km² spatial resolutions for the
239 surroundings. It has 20 vertical hybrid pressure-sigma (terrain- following) layers that range
240 from the surface to the mid-troposphere, up to 500 hPa. It is driven by operational
241 meteorological analyses of the ECMWF Integrated Forecasting System, available at an
242 approximately 15 × 15 km² spatial resolution and 3 h temporal resolution.

243 In this study the CO₂ simulations are based on a forward run over April 25th - May 12th 2015
244 with this model configuration; we do not assimilate atmospheric CO₂ data and so no inversion
245 for surface fluxes was conducted. In the Paris area (the Île-de-France administrative region),
246 hourly anthropogenic emissions are given by the IER inventory, see section 2.1. The
247 anthropogenic emissions in the rest of the domain are prescribed from the EDGAR V4.2
248 database for the year 2010 at 0.1° resolution [Olivier and Janssens-Maenhout et al., 2012]. In
249 the whole simulation domain, the natural fluxes (the Net Ecosystem Exchange: NEE) are
250 prescribed using simulations of C-TESSSEL, which is the land-surface component of the
251 ECMWF forecasting system [Boussetta et al., 2013], at a 3 hourly and 15 × 15 km² resolution.
252 Finally, the CO₂ boundary conditions at the lateral and top boundaries of the simulation domain
253 and the simulation CO₂ initial conditions on April 25th 2015 are prescribed using the CO₂
254 forecast issued by the Copernicus Atmosphere Monitoring Service (CAMS,
255 <http://atmosphere.copernicus.eu/>) at a ~15 km global resolution [Agustí-Panareda et al.,
256 2014].

257 The CHIMERE transport model is used to simulate the XCO₂ data. However, since the model
258 does not cover the atmosphere up to its top, the CO₂ fields from CHIMERE are complemented

259 with that of the CAMS CO₂ forecasts from 500 hPa to the top of the atmosphere to derive total
 260 column concentrations. The derivation of modelled XCO₂ at the sites, involves obtaining a
 261 kernel-smoothed CO₂ profile of CHIMERE and CAMS and vertical integration of these
 262 smoothed profiles, weighted by the pressure at the horizontal location of the sites.

263 The parametrisation used to smooth modelled CO₂ profiles approximates the sensitivity of the
 264 EM27/sun CO₂ retrieval is a function of pressure and sun elevation. Between 1000 hPa and
 265 480 hPa, a linear dependency of the instrument averaging kernels on solar zenith angle (θ) is
 266 assumed with boundary values following Frey et al. [2015]:

267

268 (1a) $k(480 \text{ hPa}) = 1.125$

269 (1b) $k(1000 \text{ hPa}) = 1.0 + 0.45 s^3$

270

271 where $s = \theta/90^\circ$. k . Approximate averaging kernels are obtained by linear interpolation to the
 272 pressure levels of CHIMERE and CAMS, respectively. If $p > 1000$ hPa, k is linearly
 273 extrapolated. Above 480 hPa ($p < 480$ hPa), the averaging kernels can be approximated by

274

275 (2) $k(u, s) = 1.125 - 0.6 u^3 - 0.4 u s^3$

276

277 where u is $(480 \text{ hPa} - p) / 480$. The kernel-smoothed CO₂ profile, $CO_{2_model}^s$, is obtained by

278

279 (3) $CO_{2_model}^s = \mathbf{K} CO_{2_model} + (\mathbf{I} - \mathbf{K})CO_2^a$

280

281

282 where CO_{2_model} is the modelled CO₂ profile by CHIMERE or CAMS, \mathbf{I} the identity matrix and
 283 \mathbf{K} is a diagonal matrix containing the averaging kernels k . The a priori CO₂ profile, CO_2^a , is
 284 provided by the Whole Atmosphere Community Climate Model (WACCM) model (version 6)
 285 and interpolated to the pressure levels of CHIMERE and CAMS. $CO_{2_model}^s$ is the appropriate
 286 CO₂ profile to calculate modelled XCO₂ at the location of the sites.

287

288 For a given site, the simulated XCO₂ data are thus computed from the vertical profile of this
 289 site as:

290

291 (3) $XCO_{2_CHIMERE} = \frac{1}{p_{surf}} \int_{p_{surf}}^{p_{top_CHIM}} CO_{2_CHIM}^s dp + \int_{p_{top_CHIM}}^{P=0mbar} CO_{2_CAMs}^s dp$

292

293 where p_{surf} is the surface pressure, $p_{top_CHIM} = 500$ hPa the pressure corresponding to the top
 294 boundary of the CHIMERE model, and $CO_{2_CHIM}^s$ and $CO_{2_CAMs}^s$ are the smoothed CO₂

295 concentrations of CHIMERE and CAMS respectively. For comparison we also calculated
296 XCO₂ at a lower spatial resolution with the CAMS data alone as:

297

$$298 \quad (4) \quad XCO_{2_CAMS} = \int_{P_{surf}}^{P=0mbar} CO_{2_CAMS}^s dP$$

299

300 **3 Results and discussion**

301 **3.1 Observations**

302 **3.1.1 Meteorological conditions and data coverage/instrument performance.**

303 During the measurement campaign (April 28th until May 13th, 2015), meteorological conditions
304 were a major limitation for the availability of XCO₂ observations. Useful EM27/SUN
305 measurements require direct sunlight and low wind speeds typically yield higher local XCO₂.
306 Most of the time during the campaign conditions were partly cloudy and turbid, and so
307 successful measurements at high solar zenith angle (SZA) were rare. Therefore, the data
308 coverage between April 28th and May 3rd is limited (see Table 2). As is typical for spring periods
309 in Paris, the temperature and the wind direction vary and display less synoptic variations than
310 in winter. The dominant wind directions were mostly northeasterly at the beginning of the
311 campaign and mostly southeasterly during the second half of the campaign. We find that the
312 wind speeds during daytime nearly always surpass 3 m s⁻¹, which has been identified by Breon
313 et al. [2015] and Staufer et al. [2016] as the cut-off wind speed above which the atmospheric
314 transport model CHIMERE performs best in modelling CO₂ concentration gradients in the
315 mixed layer.

316 Despite some periods with unfavourable conditions, more than 10,000 spectra were retrieved
317 among the five deployed instruments. The quality of the spectra for each day was rated
318 according to the overall data availability and consistent with Hase et al. (2015). The best
319 measurement conditions prevailed for the period between May 7th and May 12th.

320

321 **3.1.2 Observations of XCO₂ in Paris**

322 The observed XCO₂ in the Paris region for all sites (10415 observations) ranges from 397.27
323 to 404.66 ppm with a mean of 401.26 ppm (a median of 401.15 ppm). The strong atmospheric
324 variability of XCO₂ across Paris and within the campaign period is reflected in the standard
325 deviation of 1.04 ppm for 1-minute averages. We find that all sites exhibit very similar diurnal
326 behaviours with a clear decrease of XCO₂ during daytime and a noticeable day-to-day
327 variability as seen in Figure 2. This is to be expected as they are all subject to very similar
328 atmospheric transport in the boundary layer height and to similar large-scale influences, i.e.,
329 surrounding with stronger natural fluxes or air mass exchange with other regions at synoptic
330 time scales. However, observed XCO₂ concentrations at the downwind sites for our network

331 remain clearly higher from sites that are upwind of Paris (see Figure 2). The shifting dominant
332 wind conditions also explain why the site RES and GIF are lowest in the beginning of the
333 campaign and higher on May 12th and 13th after meteorological conditions changed. This
334 indicates that the influence of urban emissions is detectable with this network configuration
335 under favourable meteorological conditions. By comparing the different daily variations in Fig.
336 3, it is apparent that the day-to-day variations observed at the two southwesterly (typically
337 upwind) sites GIF and RES are approximately 1 ppm, with both sites exhibiting similar diurnal
338 variations throughout the campaign period. This can be expected as their close vicinity would
339 suggest that they are sensitive to emissions from similar areas and to concentrations of air
340 masses arriving from the southwest.

341 The typical decrease in XCO₂ found over the course of a day is about 2 to 3 ppm. This
342 decrease could be driven by (natural) sinks of CO₂, which can be expected to be very strong
343 as our campaign took place after the start of the growing season in Europe for most of southern
344 and central Europe [Roetzer and Chmielewski 2001].

345 The observations at the site located in Paris (JUS) displays similarly low day-to-day variations
346 and a clear decrease in XCO₂ over the course of the day. The latter feature indicates that even
347 in the dense city centre, XCO₂ is primarily representative of a large footprint like in other areas
348 of the globe [Keppel-Aleks 2011] and supports the findings of Belikov et al. (2017) concerning
349 the footprints for the Paris and Orleans TCCON sites, Thus, our total column observations are
350 less critically affected by local emissions than in-situ measurements [Breon et al. 2015,
351 Ammoura et al. 2016]. It is also apparent that the decrease in XCO₂ (the slope) during the
352 afternoon for April 28th and 29th as well as May 7th and 10th is noticeably smaller than at other
353 days during this campaign. As XCO₂ is not sensitive to vertical mixing, this has to be caused
354 by different CO₂ sources and sinks acting upon the total column arriving at JUS.

355 The two (typically downwind) sites PIS and MIT northeast of Paris show a markedly larger
356 day-to-day spread in their general XCO₂ levels as well as strongly changing slopes for the
357 diurnal XCO₂ decrease. For these sites the exact wind direction is critical as they can be
358 downwind of the city centre that has a much higher emission density or less dense suburbs
359 (see Fig. 1).

360

361 **3.1.3 Gradients in observed XCO₂**

362 In order to focus more on the impact of local emissions on atmospheric conditions and less on
363 that of of CO₂ fluxes from outside of our urban domain in our analysis of XCO₂, we choose to
364 study the spatial gradients (Δ) between different sites. Fundamentally, this approach assumes
365 that regional and large-scale fluxes have a similar impact on XCO₂ for the sites within our
366 network, due to the close proximity of sites and the smoothing of remote emission signals due
367 to atmospheric transport by the time the air-mass arrives in our domain. Ideal conditions were

368 sampled during May 7th, with predominantly southwesterly winds, and on May 10th with
369 southerly winds. We can see in Fig. 4 that all sites were, on average, elevated compared to
370 RES, chosen as reference here as it was upwind of Paris during those days. The hodographs
371 for both days also indicate that the wind fields were consistent across Paris (see Figure S3).
372 The observations from GIF showed only minimal differences with RES, while the rest of the
373 sites (PIS, JUS and MIT) had Δ values of 1 to 1.5 ppm. During southwesterly winds, MIT is
374 downwind of the densest part of the Paris urban area, and JUS is impacted by emissions of
375 neighborhoods to the southwest. The site of PIS is still noticeably influenced by the city centre
376 but, as can be seen in Fig. 1, we likely do not catch the plume of the most intense emissions
377 but rather from the suburbs. On May 10th, with its dominant southerly winds, the situation was
378 markedly different. While GIF was still only slightly elevated, the XCO₂ enhancement at MIT
379 was significantly lower and quite similar to JUS for large parts of the day. The highest Δ XCO₂
380 can be observed at PIS, again typically ranging from 1 to 1.5 ppm. As seen in Fig. 1, PIS is
381 then directly downwind of the densest emission area, while MIT is only exposed to CO₂
382 emissions from the eastern outskirts of Paris.

383 It is also important to note that the impact of the local biosphere that is assumed to cause the
384 strong decrease in XCO₂ during the day is not seen on both days for these spatial gradients.
385 For a more comprehensive interpretation of these observations the use of a transport model
386 (as described in section 2.3) is necessary.

387

388 **3.2 Modelling**

389 **3.2.1 Model performance**

390 Before interpreting the modelled XCO₂ we need to evaluate the performance of the chosen
391 atmospheric transport model framework as described in section 2.3. Comparing it to
392 meteorological observations (wind speed and wind direction) at GIF, we find that CHIMERE
393 predicts these variables well throughout the duration of the campaign (see Figure S4).
394 Changes in wind speed direction and speed are reproduced with a slight overestimation at low
395 wind-speeds (>1m/s). Besides the meteorological forcing, the model performance can also be
396 expected to depend on the chosen model resolution. Therefore, we compared XCO₂ at JUS
397 calculated based on the coarser resolution atmospheric transport and flux framework CAMS
398 (15 km), and the higher resolution emission modelling input for the framework based on
399 CHIMERE (2 km) for the inner domain and on CAMS boundary conditions (see Fig. S2). We
400 find that the coarser model displays similar inter-daily variations, but that the high-resolution
401 model modifies the modelling results on shorter time-scales. We find that the afternoon XCO₂
402 decreases are often more pronounced in CHIMERE. Only the high-resolution will be
403 considered and referred to in the following. The impact of using different flux maps (fossil fuel
404 CO₂) on the modelled XCO₂ can unfortunately not be explicitly investigated here as only one

405 high-resolution (1 km) emission product available for fossil fuel CO₂ was available for this
406 study region (see section 2.3) and other global emission products are usually not intended for
407 urban-scale studies.

408

409 **3.2.2 Modelled XCO₂ and its components**

410 The modelled XCO₂ for the five sites (Fig. 5) co-evolves over the period of the campaign with
411 occurrences of significant differences. This was already seen with the measurements, but the
412 model allows looking at the full time series. The model reveals clear daily cycles of XCO₂, with
413 an accumulation during night-time and a decrease during daytime. Despite a good general
414 agreement of modelled XCO₂ at all sites for, e.g., the timing of daily minima and their synoptic
415 changes, differences in XCO₂ are observed between the sites for many days. Typically the
416 northeasterly sites (PIS, MIT) show an enhancement in modelled XCO₂ compared to the
417 southwesterly sites (GIF, RES).

418 To understand the synoptic and diurnal variations of the modelled XCO₂, we analyzed the
419 contribution of different sources (and sinks) of CO₂, namely the net ecosystem exchange
420 (NEE), the fossil fuel CO₂ emissions (FFCO₂), and the boundary conditions (BC), i.e., the
421 variations of CO₂ not caused by fluxes within our domain (the example of JUS is given in Fig.
422 6). The day-to-day variability of modelled XCO₂ is dominated by changing boundary conditions
423 and coincides with synoptic weather changes. As the CO₂ emitted from the different sources
424 is transported in the model as independent tracers, the strong daily decrease in XCO₂ can be
425 directly linked to NEE, which leads to a decrease of ~1 ppm (but up to 4 ppm) during the day,
426 but can also cause positive enhancements during nighttime driven by biogenic respiration.
427 The XCO₂ from fossil fuel emissions causes significant enhancements compared to the
428 background, but is often compensated by NEE. During short periods, fossil fuel emissions can
429 however lead to enhancements of up to 4 ppm.

430

431 **3.2.3 Modelled ΔXCO₂ gradients and its components**

432 To be able to assess the impact of local sources and reduce the influence of NEE and BC on
433 the modelled signals, we analyse the XCO₂ gradient (i.e. station-to-station difference) with
434 RES being taken as reference. In Fig. 7 we compare Δ, in the top panel, and its components,
435 i.e. fossil fuel CO₂, biogenic CO₂ and CO₂ transported across the boundary of the domain
436 (boundary conditions: BC), along a south-north direction. For the modelled Δ we can see that
437 MIT shows a positive value during the campaign period whenever the predominant wind
438 direction was southwesterly. We also find that Δ between JUS and RES was both negative
439 and positive during the campaign, and predominantly negative between MIT and JUS. When
440 split into FFCO₂, BC and NEE components, we can clearly see that the total Δ is dominated

441 by FF causing XCO₂ offsets of up to 4 ppm, but more typically 1 ppm gradients are observed.
442 Gradients can also change rapidly (within a few hours) if the wind direction changes, for
443 example on May 1st and May 12th. This highlights the fact that, during such conditions, we
444 cannot assume a simple upwind-downwind interpretation of our sites. As expected, the
445 contributions from BC and NEE are generally greatly reduced when analysing Δ XCO₂. The
446 most important impact of NEE on the XCO₂ gradients of -1ppm and +1ppm can be seen on
447 May 8th and May 11th, respectively. This means that, despite greatly reducing the impact of
448 NEE on average, the contribution of NEE cannot be fully ignored. BC is an overall negligible
449 contribution to Δ XCO₂, even though it reaches -0.4 ppm on May 11th.

450

451 **3.3 Model data and observations comparison**

452 **3.3.1 XCO₂**

453 A comparison of modelled and observed XCO₂ is of course limited to the relatively short
454 periods when observations are available. Over these periods we can see a general issue in
455 reproducing the general XCO₂ for each day in the model as observed XCO₂ is significantly
456 lower revealing a fairly stable bias between 1 to 2 ppm. As our CO₂ boundary conditions were
457 from a forecast product, this is not unexpected, as already small issues in estimating carbon
458 uptake (or emissions) at the European scale can have such an impact on the boundary
459 conditions. However, we observe that the main features, like daily cycles and synoptic
460 changes of the modelled and observed XCO₂ are comparable as seen in Figure 8. The
461 daytime variations are well reproduced by the model and the general relative concentrations
462 between sites are preserved, e.g., the highest values for XCO₂ at MIT are on May 9th and
463 highest XCO₂ for PIS are later on May 10th and May 11th. We also see that the timing of the
464 daily minima is not fully covered in the observed data as it typically happens after sunset and
465 cessation of biosphere uptake. To reduce the impact of uncertainties of the boundary
466 conditions on our analysis a gradient approach was tested.

467

468 **3.3.2 Δ XCO₂**

469 Due to the prevailing southeasterly wind conditions, we can compare XCO₂ at the typical
470 downwind sites (PIS, MIT) relative to the mostly upwind sites (RES, GIF) and expect elevated
471 XCO₂ downwind. Furthermore, we can expect to see negative gradients for opposing wind
472 conditions, i.e. northwesterly. For other wind conditions, the concentration difference is not
473 determined by emissions between the station pairs, but rather by the areas upwind of the sites,
474 (see Figure 1). We find that the Δ XCO₂ of PIS relative to RES generally falls along the 1:1 line
475 with a slope of 1.07 ± 0.09 with a Pearson's R of 0.8. Negative Δ XCO₂ values, seen in Fig. 9,
476 are associated with meteorological conditions when winds come from northerly directions,

477 i.e., the roles of normal upwind and downwind sites are reversed. For wind perpendicular to
478 the direct line of sight for (PIS, RES) the concentration enhancements are small and harder to
479 interpret and no slope was calculated. The gradient of XCO₂ MIT relative to RES has a
480 significantly lower range for modelled XCO₂ while the observed range of XCO₂ is similar to
481 PIS. The slope of observed to modelled Δ XCO₂ for upwind-downwind (or downwind-upwind
482 conditions) is 1.72 ± 0.06 with a Pearson's R of 0.96. This points to a significant
483 underestimation of the impact of urban sources on the MIT-RES gradient, which is especially
484 visible in the more negative Δ XCO₂ during northerly wind conditions. This could indicate that
485 the spatial distribution of our emissions prior should be improved, i.e., emissions in the eastern
486 outskirts/suburbs are likely underestimated in the IER emissions model. The low modelled
487 Δ XCO₂ could also be due to overestimated horizontal dispersion in the model, which seems
488 less likely. Again the model does not predict concentration differences well for perpendicular
489 wind conditions. When comparing the mean modelled daily cycle of the days with south-
490 westerly wind conditions and when observations exist with the mean diurnal cycle for all days
491 within the field campaign period when MIT and PIS can be considered downwind of RES, we
492 find that the days with observations do not significantly differ from those without observations
493 (see Fig. 10). An investigation of typical diurnal variations of modelled Δ XCO₂ can only be
494 performed to a limited degree with the observational data available for suitable wind
495 conditions. Within the large uncertainties, the modelled and observed Δ XCO₂ agree
496 throughout the day. When analysing the modelled Δ XCO₂ components we also find that the
497 observed daytime increases of Δ XCO₂ are driven by CO₂ added by urban FF CO₂ burning
498 and that the impact of FF is significantly higher at PIS (up to 1 ppm) than at MIT site (0.5 ppm)
499 in the model, when both sites are downwind of Parisian emissions. Our observations indicate
500 that both sites have strong diurnal variations. Given that the most important biogenic sinks, in
501 our domain, can be expected to be found in the rural parts surrounding Paris we would expect
502 the biogenic contribution to be similar at both sites (as predicted by the model). This would
503 further point towards, that the impact of FF emissions on the MIT site is larger than predicted
504 by our modelling framework.

505 Different Δ XCO₂ diurnal variations can be found for other upwind-downwind site pairs, but they
506 are all systematically driven by the locally-added CO₂ from FFCO₂.

507

508 **5 Conclusion and Outlook**

509 For the two-week field campaign we demonstrated the ability of a network of five EM27/SUN
510 spectrometers, placed in the outskirts of Paris, to track the XCO₂ changes due to the urban
511 plume of the city. However, we also found that XCO₂ cannot be simply interpreted in the
512 context of local emissions as, even in such a densely populated area, XCO₂ is still significantly

513 influenced by natural CO₂ uptake during the growing season. Understanding the area
514 influencing XCO₂ and/or the use of suitable atmospheric transport models seems
515 indispensable to correctly interpret atmospheric XCO₂ variations. Using a gradient approach,
516 i.e., analysing the difference between XCO₂ measured at upwind and downwind stations,
517 greatly reduced the impact of CO₂ boundary condition, that reflect fluxes outside the domain
518 and biogenic fluxes within the domain. Overall, the XCO₂ variability modelled using our
519 ECMWF-CHIMERE system with IER (1 x 1 km²) emissions data was found to be comparable
520 with the observed variability and diurnal evolution of XCO₂, despite a higher background for
521 modelled XCO₂. Our modelling framework, run at a 2 x 2 km² resolution over Paris also
522 predicts that biogenic fluxes and boundary conditions (i.e. the influence of CO₂ being
523 transported into our domain) have only very small impact on ΔXCO₂ only noticeably impacting
524 it during a few situations, specifically when meteorological conditions changes made the
525 concept of 'upwind' and 'downwind' not applicable. When comparing modelled and measured
526 ΔXCO₂ we find strong correlations (Pearson's R) of 0.8 and 0.96 for PIS-RES and MIT-RES,
527 respectively. The offset between model and observations also diminished for ΔXCO₂ and the
528 slope found between observed and modeled PIS-RES gradient is statistically in accordance
529 with a 1:1 relationship (1.07±0.09). However, the slope of the MIT-RES XCO₂ gradient of
530 1.72±0.06 suggests that the emission model could potentially be improved, as it seems
531 unlikely that the general atmospheric transport in the model is the key issue as both site pairs
532 would be subject to very similar winds. Another potential source of error that needs to be
533 investigated is if such an underestimation of ΔXCO₂ could be caused by the limited model
534 resolution.. It also seems rather likely that a 2x2km² model would cause a general spreading
535 of point source emissions and not systematically underestimate emissions impacts from less
536 densely populated, parts of Île-de-France. The data also confirm previous results by models
537 that XCO₂ gradients caused by a megacity do not exceed 2 ppm, which supports the previous
538 requirement for satellite observations of less than 1 ppm precision on individual soundings,
539 and biases lower than 0.5 ppm (Ciais et al. 2015). The gradients are mainly caused by the
540 transport of FFCO₂ emissions but, interestingly, during specific episodes, a noticeable
541 contribution comes from biogenic fluxes, suggesting that these fluxes cannot always be
542 neglected even when using gradients.

543 Unfortunately, the duration of the campaign was relatively short, so that an in-depth analysis
544 of mean daily cycles or the impact of ambient conditions (traffic conditions, temperature, solar
545 insolation, etc.) on the observed gradient and underlying fluxes could not be investigated here.
546 Hence, future studies in Paris and elsewhere should aim to perform longer-term observations
547 during different seasons, which will allow better understanding changes in biogenic and
548 anthropogenic CO₂ fluxes. A remotely-controllable shelter for the EM27/SUN instrument is

549 currently under development [Heinle and Chen, 2017]. This will considerably facilitate the
550 establishment of permanent spectrometer arrays around cities and other sources of interest.
551 Nevertheless, our study already indicates that such observations of urban XCO₂ and ΔXCO₂
552 contain original information to understand local sources and sinks and that the modelling
553 framework used here is a step forward to support their detailed interpretation in the future. An
554 improved model will also be able to adjust or better model the background conditions and
555 potentially use this type of observations to estimate local CO₂ fluxes using a Bayesian
556 inversion scheme similar to the existing system based on in-situ observations for Paris [Staufer
557 et al. 2016].

558 We expect that the previous successful collaboration in the framework of the Paris campaign
559 will mark the permanent implementation of COCCON as a common framework for a French-
560 Canadian-German collaboration on the EM27/SUN instrument. The acquisition of additional
561 spectrometers is planned by several partners.

562

563 **Author contribution**

564 FRV, MF, FH, IXR, MKS, PCh, PJ, YVT, CJ, TB, QT and JO, supported the field campaign
565 and contributed data to this study.

566 MF, FH, FRV, JS, GB and PCi planned the fieldwork and modelling activities for this study.

567 JS, GB, FC, and FRV performed the CHIMERE modelling, provided modelling data input
568 and/or analysed the output data.

569 MF, FH and FRV processed and analysed the EM27Sun data.

570 FRV, MF, JS, FH and PCi wrote sections of the manuscript and created figures and tables.

571 All authors reviewed, edited and approved the manuscript.

572

573 **Acknowledgement**

574 All authors would like to thank the two anonymous reviewers for their comments that helped
575 to significantly improve this manuscript. ECCC would like to thank Ray Nasser (CRD) and
576 Yves Rochon (AQRD) for their internal review. The authors from LSCE acknowledge the
577 support of the SATINV group of Frederic Chevallier. The authors from KIT acknowledge
578 support from the Helmholtz Research Infrastructure ACROSS. The authors from LISA
579 acknowledge support from the OSU-EFLUVE (Observatoire des Sciences de l'Univers-
580 Enveloppes Fluides de la Ville à l'Exobiologie).

581

582 **References**

583 Agustí-Panareda, A., Massart, S., Chevallier, F., Bousetta, S., Balsamo, G., Beljaars, A., Ciais, P.,
584 Deutscher, N.M., Engelen, R., Jones, L. and Kivi, R., 2014. Forecasting global atmospheric CO₂.
585 Atmospheric Chemistry and Physics, 14(21), pp.11959-11983.

586 AIRPARIF, 2016. Inventaire régional des émissions en Île-de-France Année de référence 2012 –
587 éléments synthétiques, Edition Mai 2016, Paris, France. Last access Dec. 14th, 2017, available at:
588 https://www.airparif.asso.fr/_pdf/publications/inventaire-emissions-idf-2012-150121.pdf
589 Ammoura, L., Xueref-Remy, I., Vogel, F., Gros, V., Baudic, A., Bonsang, B., Delmotte, M., Té, Y. and
590 Chevallier, F., 2016. Exploiting stagnant conditions to derive robust emission ratio estimates for CO₂,
591 CO and volatile organic compounds in Paris. *Atmospheric Chemistry and Physics*, 16(24), pp.15653-
592 15664.

593 Belikov, D., Maksyutov, S., Ganshin, A., Zhuravlev, R., Deutscher, N.M., Wunch, D., Feist, D.G.,
594 Morino, I., Parker, R.J., Strong, K. and Yoshida, Y., 2017. Study of the footprints of short-term
595 variation in XCO₂ observed by TCCON sites using NIES and FLEXPART atmospheric transport
596 models.

597 Bréon, F.M., Broquet, G., Puygrenier, V., Chevallier, F., Xueref-Remy, I., Ramonet, M., Dieudonné, E.,
598 Lopez, M., Schmidt, M., Perrussel, O. and Ciais, P., 2015. An attempt at estimating Paris area CO₂
599 emissions from atmospheric concentration measurements. *Atmospheric Chemistry and Physics*, 15(4),
600 pp.1707-1724. <https://doi.org/10.5194/acp-15-1707-2015>, 2015.

601 Broquet, G., Bréon, F.M., Renault, E., Buchwitz, M., Reuter, M., Bovensmann, H., Chevallier, F., Wu,
602 L. and Ciais, P., 2018. The potential of satellite spectro-imagery for monitoring CO₂ emissions from
603 large cities. *Atmospheric Measurement Techniques*, 11(2), pp.681-708.

604 Boussetta, S., Balsamo, G., Beljaars, A., Panareda, A.A., Calvet, J.C., Jacobs, C., Hurk, B., Viterbo,
605 P., Lafont, S., Dutra, E. and Jarlan, L., 2013. Natural land carbon dioxide exchanges in the ECMWF
606 Integrated Forecasting System: Implementation and offline validation. *Journal of Geophysical*
607 *Research: Atmospheres*, 118(12), pp.5923-5946.

608 Butz, A., Orphal, J., Checa-Garcia, R., Friedl-Vallon, F., von Clarmann, T., Bovensmann, H.,
609 Hasekamp, O., Landgraf, H., Knigge, T., Weise, D., Sqalli-Houssini, O., and D. Kemper, *Geostationary*
610 *Emission Explorer for Europe (G3E): mission concept and initial performance assessment*, *Atmos.*
611 *Meas. Tech.*, 8, 4719-4734, 2015

612 Chen, J., Viatte, C., Hedelius, J. K., Jones, T., Franklin, J. E., Parker, H., Gottlieb, E. W., Wennberg, P.
613 O., Dubey, M. K., and Wofsy, S. C., 2016. Differential column measurements using compact solar-
614 tracking spectrometers, *Atmos. Chem. Phys.*, 16, 8479-8498, [https://doi.org/10.5194/acp-16-8479-](https://doi.org/10.5194/acp-16-8479-2016)
615 2016, 2016.

616 Ciais, P., Crisp, D., Denier van der Gon, H., Engelen, R., Heimann, M., Janssens-Maenhout, G.,
617 Rayner, P. and Scholze, M., 2015. Towards a European Operational Observing System to Monitor
618 Fossil CO₂ Emissions. Final Report from the Expert Group, European Commission, October 2015.
619 Available at http://edgar.jrc.ec.europa.eu/news_docs/CO2_report_22-10-2015.pdf Accessed February
620 6th, 2018

621 Dhakal, S., 2009. Urban energy use and carbon emissions from cities in China and policy implications,
622 *Energy Policy* 37:4208-4219

623 Frey, M., F. Hase, T. Blumenstock, J. Groß, M. Kiel, G. Mengistu Tsidu, K. Schäfer, M. Kumar Sha, and
624 J. Orphal, 2015. Calibration and instrumental line shape characterization of a set of portable FTIR
625 spectrometers for detecting greenhouse gas emissions, *Atmos. Meas. Tech.*, 8, 3047-3057,
626 [doi:10.5194/amt-8-3047-2015](https://doi.org/10.5194/amt-8-3047-2015)

627 Frey, M., Sha, M.K., Hase, F., Kiel, M., Blumenstock, T., Harig, G., Surawicz, G., Deutscher, N.M.,
628 Shiomi, K., Franklin, J., Bösch, H., Chen, J., Grutter, M., Ohyama, H., Sun, Y., Butz, A., Mengistu Tsidu,
629 G., Ene, D., Wunch, D., Song, C.Z., Garcia, O., Ramonet, M., Vogel, F., and J. Orphal, Building the
630 COllaborative Carbon Column Observing Network (COCCON): Long term stability and ensemble
631 performance of the EM27/SUN Fourier transform spectrometer, *Atmos. Meas. Tech. Diss*, submitted,
632 2018

633 Gisi, M., F. Hase, S. Dohe, T. Blumenstock, A. Simon, and A. Keens, 2012. XCO₂-measurements with
634 a tabletop FTS using solar absorption spectroscopy, *Atmos. Meas. Tech.*, 5, 2969-2980,
635 doi:10.5194/amt-5-2969-2012

636 Hase, F., M. Frey, T. Blumenstock, J. Groß, M. Kiel, R. Kohlhepp, G. Mengistu Tsidu, K. Schäfer, M. K.
637 Sha, and J. Orphal, 2015. Application of portable FTIR spectrometers for detecting greenhouse gas
638 emissions of the major city Berlin, *Atmos. Meas. Tech.*, 8, 3059-3068, doi:10.5194/amt-8-3059-2015

639 Hase, F., M. Frey, M. Kiel, T. Blumenstock, R. Harig, A. Keens, and J. Orphal, 2016. Addition of a
640 channel for XCO observations to a portable FTIR spectrometer for greenhouse gas measurements,
641 *Atmos. Meas. Tech.*, 9, 2303-2313, doi:10.5194/amt-9-2303-2016

642 Hedelius, J. K., Viatte, C., Wunch, D., Roehl, C. M., Toon, G. C., Chen, J., Jones, T., Wofsy, S. C.,
643 Franklin, J. E., Parker, H., Dubey, M. K., and Wennberg, P. O., 2016. Assessment of errors and biases
644 in retrievals of X_{CO₂}, X_{CH₄}, X_{CO}, and X_{N₂O} from a 0.5 cm⁻¹ resolution solar-viewing spectrometer, *Atmos.*
645 *Meas. Tech.*, 9, 3527-3546, <https://doi.org/10.5194/amt-9-3527-2016>

646 Heinle, L. and Chen, J., 2017 in review. Automated Enclosure and Protection System for Compact
647 Solar-Tracking Spectrometers, *Atmos. Meas. Tech. Discuss.*, <https://doi.org/10.5194/amt-2017-292>

648 IEA, International Energy Agency, 2008, *World Energy Outlook*, IEA Publications, Paris, France ISBN:
649 978926404560-6

650 IPCC-WG1, *Climate Change 2013: The Physical Science Basis*. Contribution of Working Group I to the
651 Fifth Assessment Report of the Intergovernmental Panel on Climate Change [Stocker, T.F., D. Qin, G.-
652 K. Plattner, M. Tignor, S.K. Allen, J. Boschung, A. Nauels, Y. Xia, V. Bex and P.M. Midgley (eds.)].
653 Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 1535 pp

654 IPCC-WG3, *Climate Change 2014: Mitigation of Climate Change*. Contribution of Working Group III to
655 the Fifth Assessment Report of the Intergovernmental Panel on Climate Change [Edenhofer, O., R.
656 Pichs-Madruga, Y. Sokona, E. Farahani, S. Kadner, K. Seyboth, A. Adler, I. Baum, S. Brunner, P.
657 Eickemeier, B. Kriemann, J. Savolainen, S. Schlömer, C. von Stechow, T. Zwickel and J.C. Minx (eds.)].
658 Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA

659 Janardanan, R., S. Maksyutov, T. Oda, M. Saito, J. W. Kaiser, A. Ganshin, A. Stohl, T. Matsunaga, Y.
660 Yoshida, and T. Yokota (2016), Comparing GOSAT observations of localized CO₂ enhancements by
661 large emitters with inventory-based estimates, *Geophys. Res. Lett.*, 43, 3486–3493,
662 doi:10.1002/2016GL067843.

663 Jones, N., 2013, Troubling milestone for CO₂. *Nature Geoscience* 6, no. 8, 589-589.

664 Keppel-Aleks, G., Toon, G.C., Wennberg, P.O. and Deutscher, N.M., 2007. Reducing the impact of
665 source brightness fluctuations on spectra obtained by Fourier-transform spectrometry. *Applied optics*,
666 46(21), pp.4774-4779.

667 Keppel-Aleks, G., P. O. Wennberg, and T. Schneider (2011), Sources of variations in total column
668 carbon dioxide, *Atmospheric Chemistry and Physics*, 11(8), 3581-3593, doi:10.5194/acp-11-3581-2011
669 Klappenbach, F., Bertleff, M., Kostinek, J., Hase, F., Blumenstock, T., Agusti-Panareda, A., Razingger,
670 M., and Butz, A., 2015. Accurate mobile remote sensing of XCO₂ and XCH₄ latitudinal transects from
671 aboard a research vessel, *Atmos. Meas. Tech.*, 8, 5023-5038, <https://doi.org/10.5194/amt-8-5023-2015>
672 Kort, E. A., C. Frankenberg, C. E. Miller, and T. Oda (2012), Space-based observations of megacity
673 carbon dioxide, *Geophys. Res. Lett.*, 39, L17806, doi:10.1029/2012GL052738.

674 Lauvaux, T., Miles, N.L., Deng, A., Richardson, S.J., Cambaliza, M.O., Davis, K.J., Gaudet, B., Gurney,
675 K.R., Huang, J., O'Keefe, D. and Song, Y., 2016. High-resolution atmospheric inversion of urban CO₂
676 emissions during the dormant season of the Indianapolis Flux Experiment (INFLUX). *Journal of*
677 *Geophysical Research: Atmospheres*, 121(10), pp.5213-5236.

678 Latoska, A., 2009. Erstellung eines räumlich hoch aufgelösten Emissionsinventar von Luftschadstoffen
679 am Beispiel von Frankreich im Jahr 2005, Master's thesis, Institut für Energiewirtschaft und Rationelle
680 Energieanwendung, Universität Stuttgart, Stuttgart, Germany

681 Levin, I., Hammer, S., Eichelmann, E. and Vogel, F.R., 2011. Verification of greenhouse gas emission
682 reductions: the prospect of atmospheric monitoring in polluted areas. *Philosophical Transactions of the*
683 *Royal Society of London A: Mathematical, Physical and Engineering Sciences*, 369(1943), pp.1906-
684 1924.

685 Messerschmidt, J., Geibel, M. C., Blumenstock, T., Chen, H., Deutscher, N. M., Engel, A., Feist, D. G.,
686 Gerbig, C., Gisi, M., Hase, F., Katrynski, K., Kolle, O., Lavrič, J. V., Notholt, J., Palm, M., Ramonet, M.,
687 Rettinger, M., Schmidt, M., Sussmann, R., Toon, G. C., Truong, F., Warneke, T., Wennberg, P. O.,
688 Wunch, D., and Xueref-Remy, I., 2011. Calibration of TCCON column-averaged CO₂: the first aircraft
689 campaign over European TCCON sites, *Atmos. Chem. Phys.*, 11, 10765-10777,
690 <https://doi.org/10.5194/acp-11-10765-2011>

691 Mitchell, L.E., Lin, J.C., Bowling, D.R., Pataki, D.E., Strong, C., Schauer, A.J., Bares, R., Bush, S.E.,
692 Stephens, B.B., Mendoza, D. and Mallia, D., 2018. Long-term urban carbon dioxide observations reveal
693 spatial and temporal dynamics related to urban characteristics and growth. *Proceedings of the National*
694 *Academy of Sciences*, 115(12), pp.2912-2917.

695 Nassar, R., Hill, T.G., McLinden, C.A., Wunch, D., Jones, D.B.A. and D. Crisp, 2017. Quantifying CO₂
696 emissions from individual power plants from space, *JGR*, 44, 19, 10045-1053.

697 Nassar, R., Napier-Linton, L., Gurney, K.R., Andres, R.J., Oda, T., Vogel, F.R. and Deng, F., 2013.
698 Improving the temporal and spatial distribution of CO₂ emissions from global fossil fuel emission data
699 sets. *Journal of Geophysical Research: Atmospheres*, 118(2), pp.917-933.

700 O'Brien, D.M., Polonsky, I.N., Utembe, S.R. and Rayner, P.J., 2016. Potential of a geostationary
701 geoCARB mission to estimate surface emissions of CO₂, CH₄ and CO in a polluted urban environment:
702 case study Shanghai. *Atmospheric Measurement Techniques*, 9(9), p.4633.

703 Olivier, J. and G. Janssens-Maenhout, CO₂ Emissions from Fuel Combustion -- 2012 Edition, IEA CO₂
704 report 2012, Part III, Greenhouse-Gas Emissions, ISBN 978-92-64-17475-7

705 Rötzer, T., and F-M. Chmielewski. 2001. Phenological maps of Europe., *Climate research* 18.3, 249-
706 257.

707 Schwandner, F.M., Gunson, M.R., Miller, C.E., Carn, S.A., Eldering, A., Krings, T., Verhulst, K.R.,
708 Schimel, D.S., Nguyen, H.M., Crisp, D. and O'dell, C.W., 2017. Spaceborne detection of localized
709 carbon dioxide sources. *Science*, 358(6360), p.eaam5782.

710 Schneider, M. and Hase, F.: Ground-based FTIR water vapour profile analyses, *Atmos. Meas. Tech.*,
711 2, 609–619, doi:10.5194/amt-2-609-2009, 2009.

712 Staufer, J., Broquet, G., Bréon, F.M., Puygrenier, V., Chevallier, F., Xueref-Rémy, I., Dieudonné, E.,
713 Lopez, M., Schmidt, M., Ramonet, M. and Perrussel, O., 2016. The first 1-year-long estimate of the
714 Paris region fossil fuel CO₂ emissions based on atmospheric inversion. *Atmospheric Chemistry and*
715 *Physics*, 16(22), pp.14703-14726.

716 Strong, C., Stwertka, C., Bowling, D.R., Stephens, B.B. and Ehleringer, J.R., 2011. Urban carbon
717 dioxide cycles within the Salt Lake Valley: A multiple-box model validated by observations. *Journal of*
718 *Geophysical Research: Atmospheres*, 116(D15).

719 Turnbull, J.C., Sweeney, C., Karion, A., Newberger, T., Lehman, S.J., Tans, P.P., Davis, K.J., Lauvaux,
720 T., Miles, N.L., Richardson, S.J. and Cambaliza, M.O., 2015. Toward quantification and source sector
721 identification of fossil fuel CO₂ emissions from an urban area: Results from the INFLUX experiment.
722 *Journal of Geophysical Research: Atmospheres*, 120(1), pp.292-312.

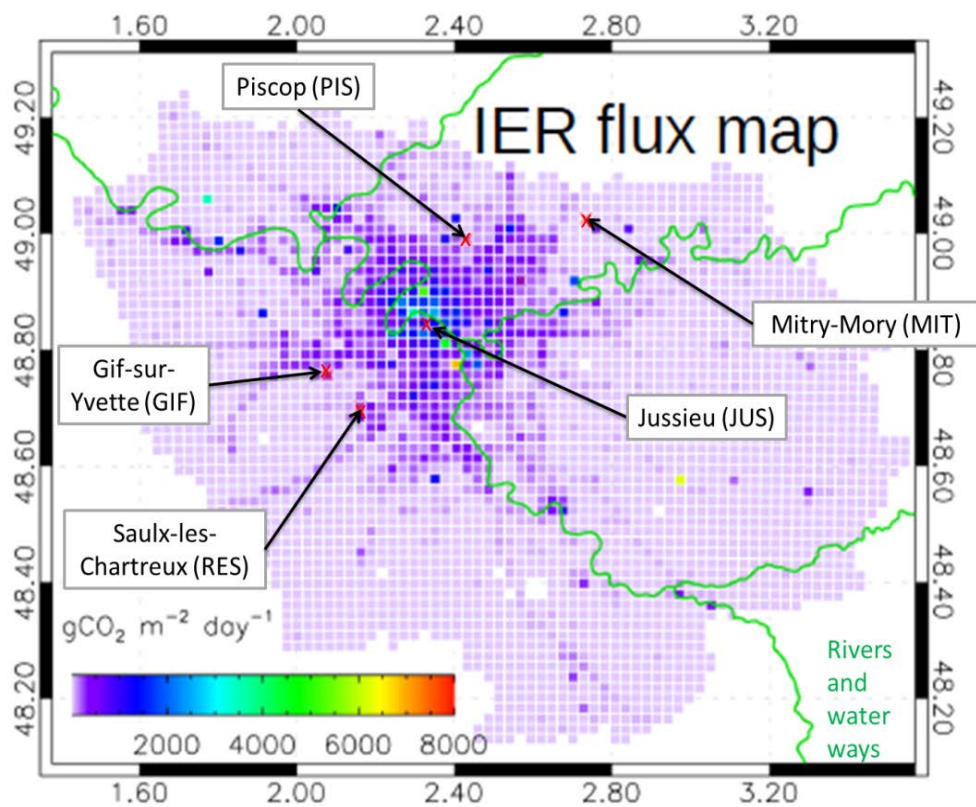
723 Vogel, F.R., Ishizawa, M., Chan, E., Chan, D., Hammer, S., Levin, I. and Worthy, D.E.J., 2012. Regional
724 non-CO₂ greenhouse gas fluxes inferred from atmospheric measurements in Ontario, Canada. *Journal*
725 *of Integrative Environmental Sciences*, 9(sup1), pp.41-55.

726 Vogel, F.R., Thiruchittampalam, B., Theloke, J., Kretschmer, R., Gerbig, C., Hammer, S. and Levin, I.,
727 2013. Can we evaluate a fine-grained emission model using high-resolution atmospheric transport
728 modelling and regional fossil fuel CO₂ observations?. *Tellus B: Chemical and Physical Meteorology*,
729 65(1), p.18681.

730 Wunch, D., Toon, G.C., Blavier, J.F.L., Washenfelder, R.A., Notholt, J., Connor, B.J., Griffith, D.W.T.,
731 Sherlock, V. and Wennberg, P.O., 2011. The total carbon column observing network, *Philos. T. Roy.*
732 *Soc. A*, 369, 2087–2112.

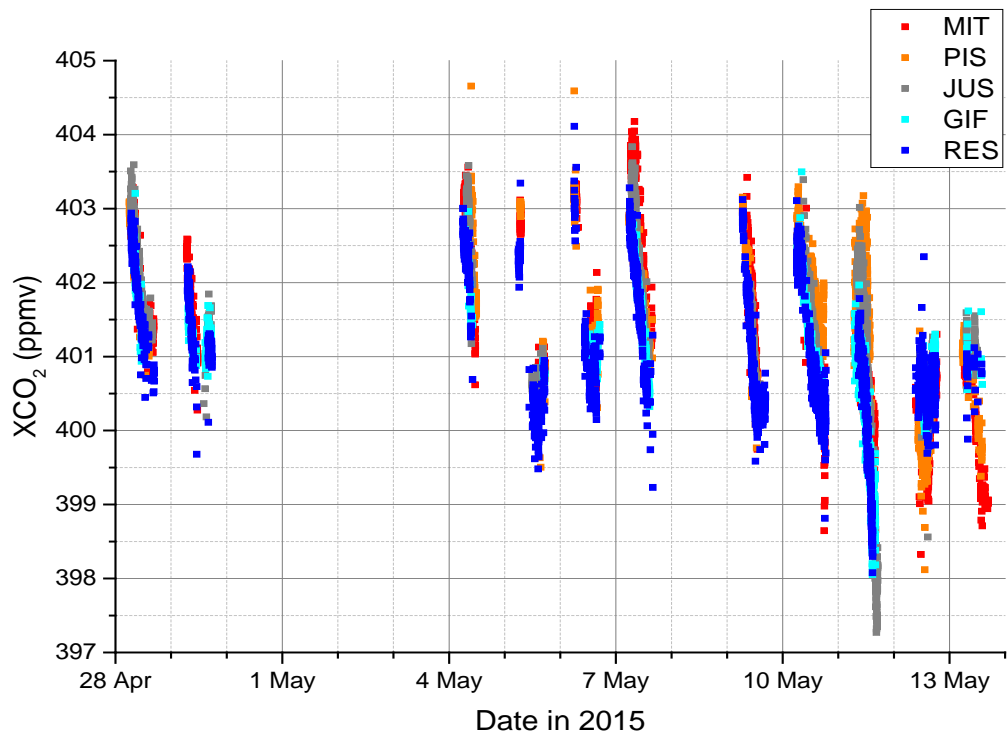
733 Wunch, D., Toon, C., Wennberg, O., Wofsy, C., Stephens, B., Fischer, L., Uchino, O., Abshire, B.,
734 Bernath, P., Biraud, C. and Blavier, F., 2010. Calibration of the total carbon column observing network
735 using aircraft profile data. *Atmospheric Measurement Techniques*, 3(5), pp.1351-1362.

736 Xueref-Remy, I., Dieudonné, E., Vuillemin, C., Lopez, M., Lac, C., Schmidt, M., Delmotte, M., Chevallier,
737 F., Ravetta, F., Perrussel, O., Ciais, P., Bréon, F.-M., Broquet, G., Ramonet, M., Spain, T. G., and
738 Ampe, C.: Diurnal, synoptic and seasonal variability of atmospheric CO₂ in the Paris megacity area,
739 *Atmospheric Chemistry and Physics*, 18(5), pp.3335-3362.



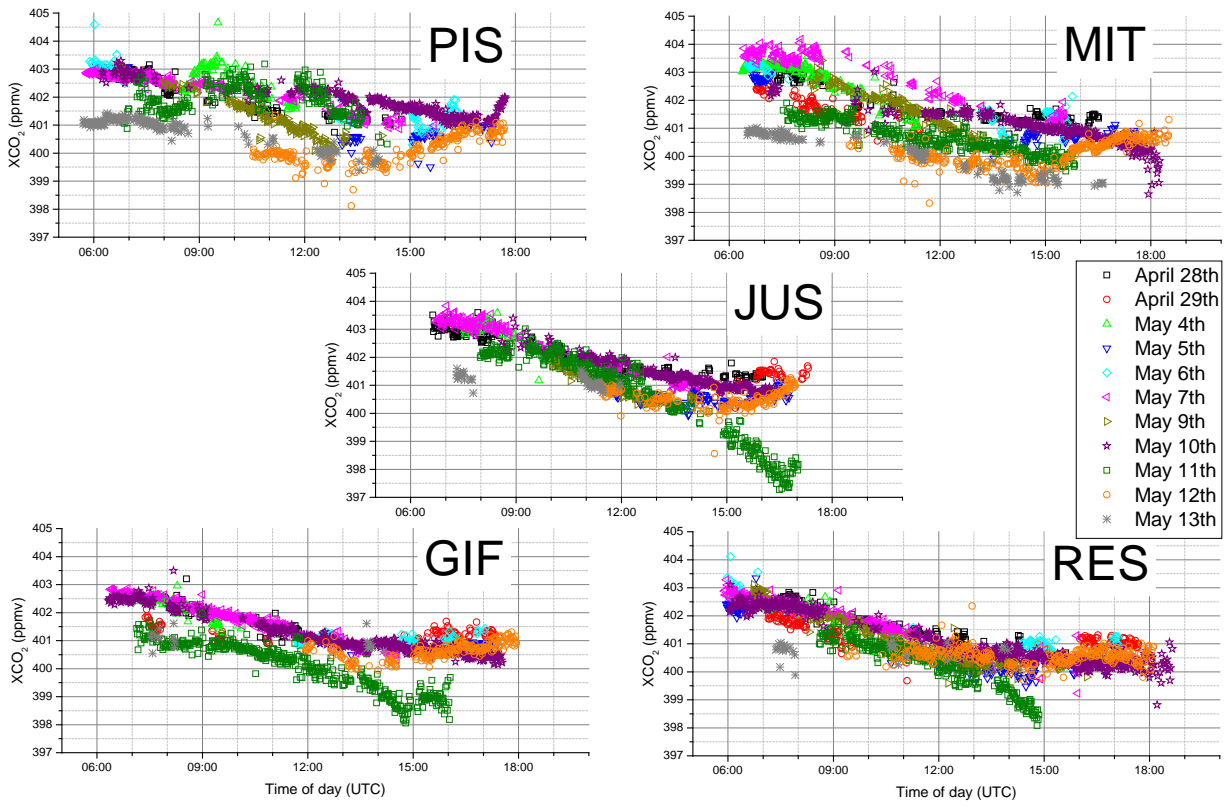
740
741
742

Figure 1. CO₂ emissions in the Île-de-France region according to the IER emission inventory. Measurement sites are indicated by red crosses.



743
 744
 745

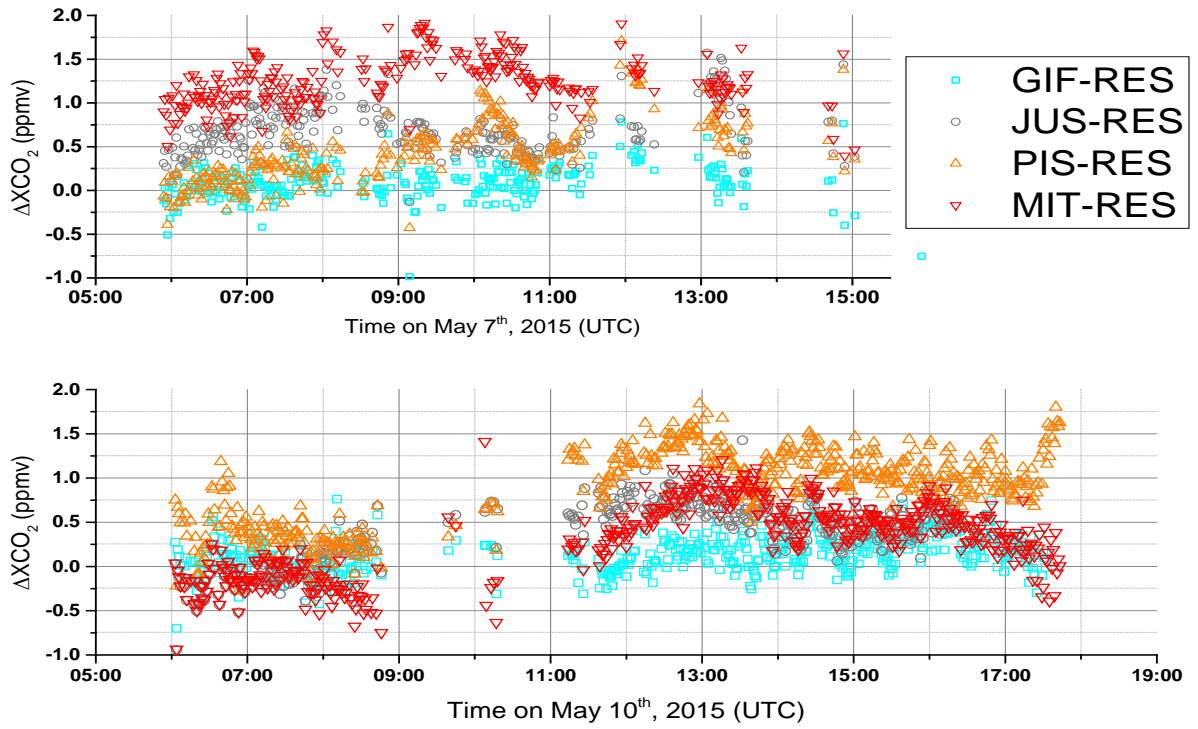
Figure 2. Time series of observed XCO₂ in the Parisian region for all five sites (all valid data of 1 minute averages).



746

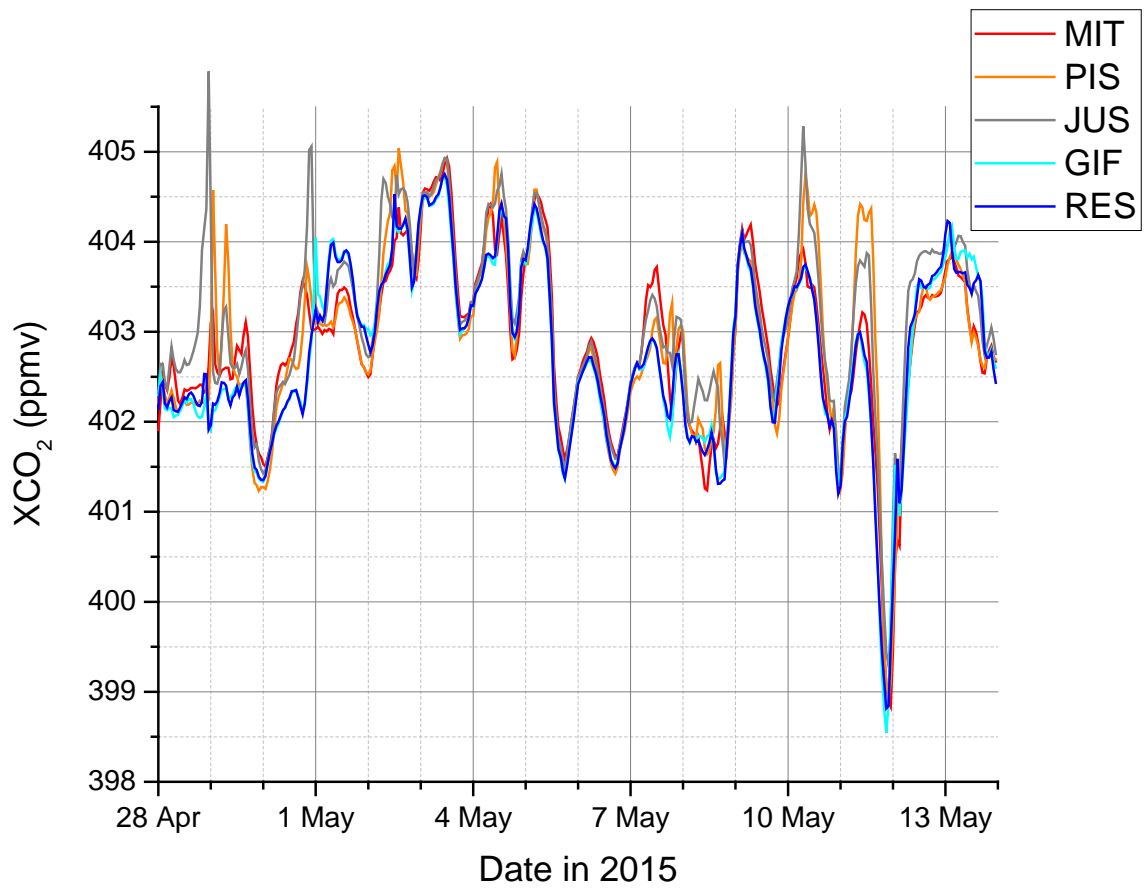
747

Figure 3. Time series of observed XCO₂ in the Parisian region sorted by station.

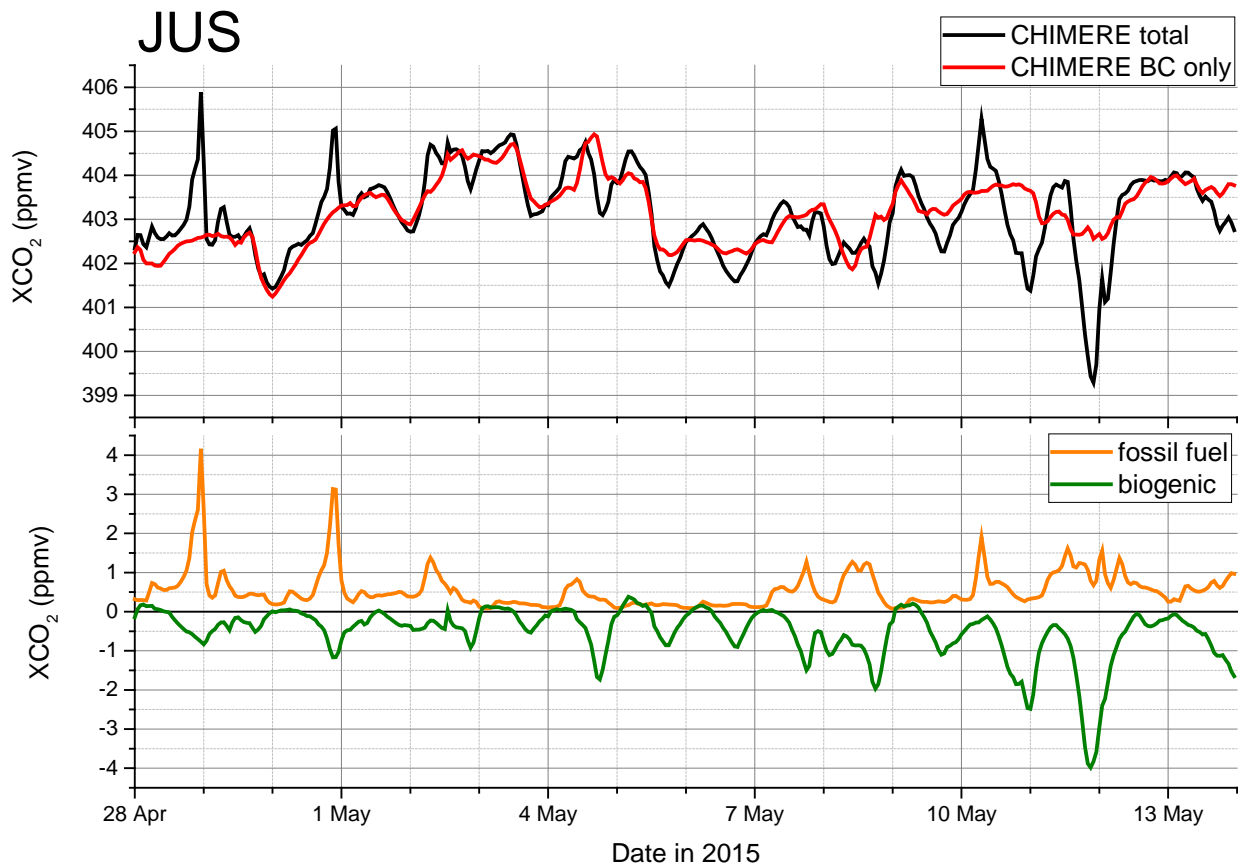


748
749
750

Figure 4. Observed spatial gradients of XCO₂ for May 7th (southwesterly winds) and May 10th (southerly winds).

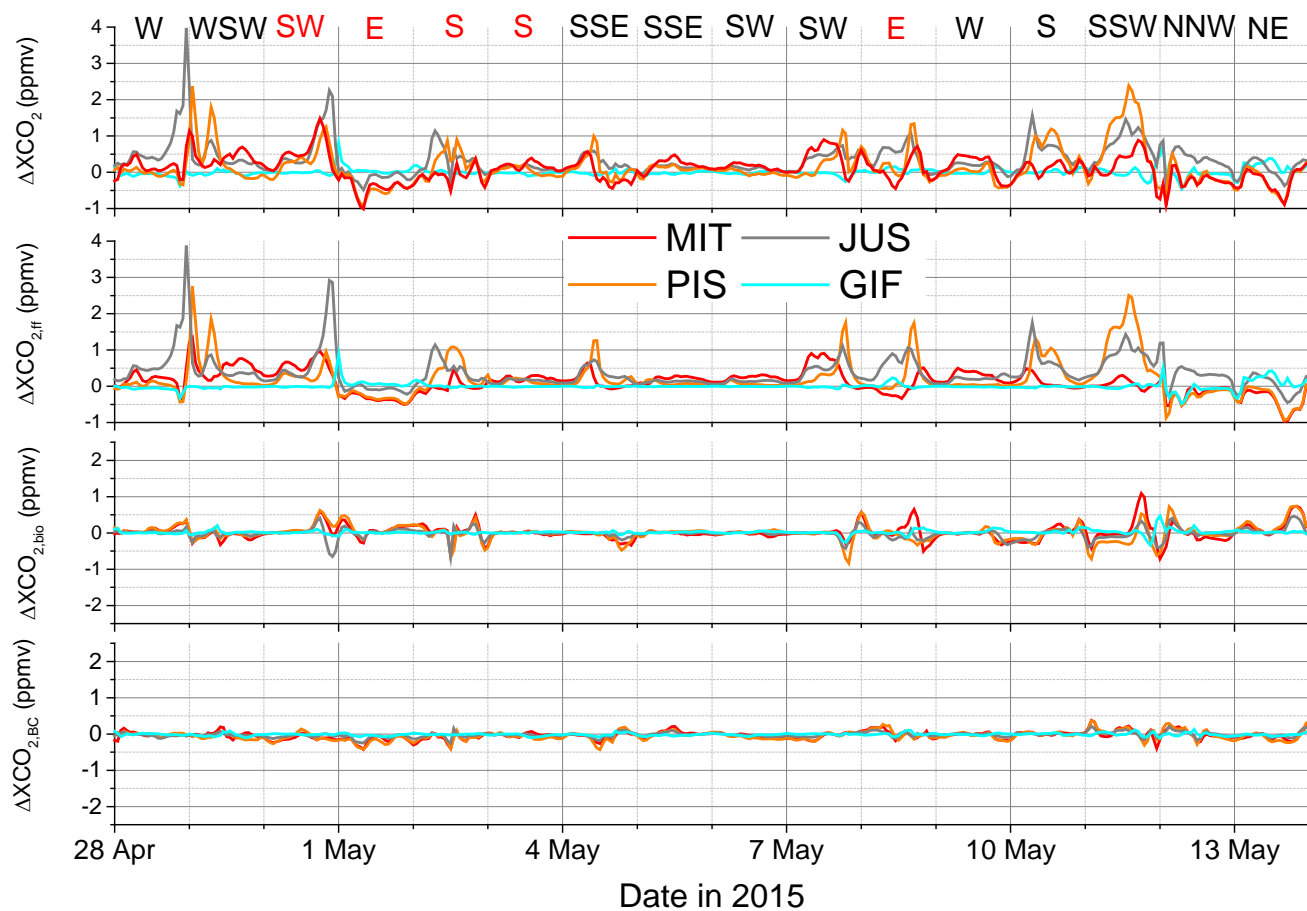


751
752 **Figure 5. Modelled XCO₂ for all stations.**



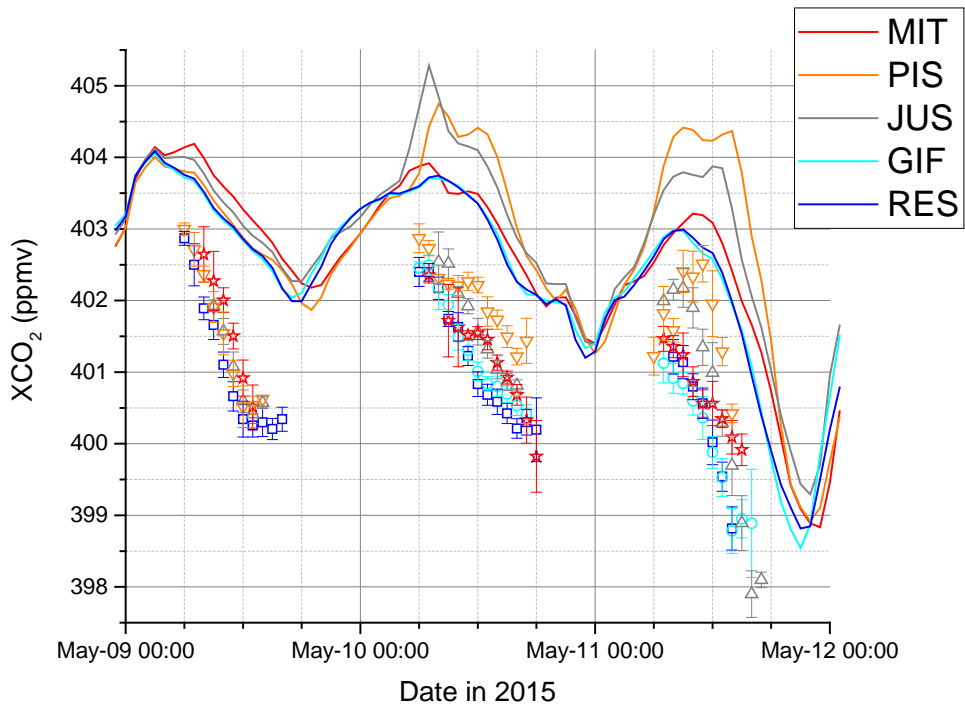
753
754
755
756
757

Figure 6: Time series of XCO₂ and related fluxes for JUS. The top panel provides a comparison of modelled total XCO₂ and XCO₂ variations due to changes in boundary conditions (BC only). The lower panel shows the contribution of the different flux components, namely fossil fuel CO₂ emissions and biogenic fluxes.



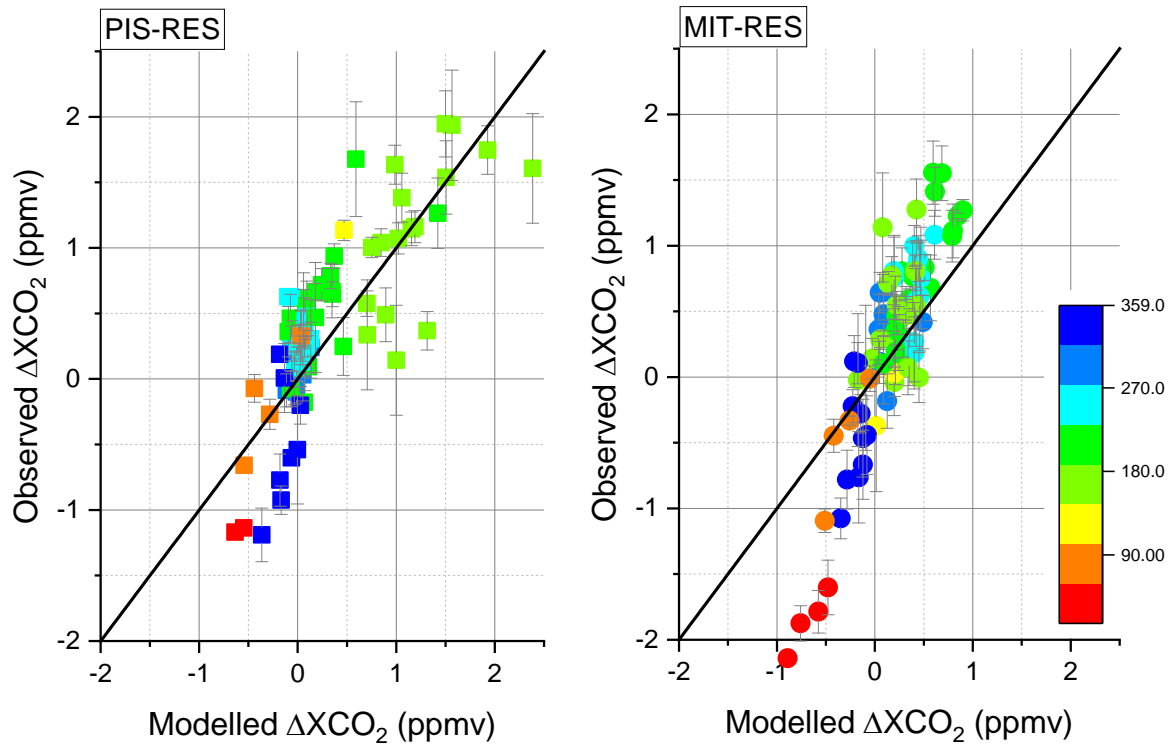
758
759
760
761
762
763
764

Figure 7. Modelled XCO₂ gradients for each station relative to RES are given in the top panel with its contributing components in the panels below. Total ΔXCO_2 (top), the fossil fuel contribution $\Delta\text{XCO}_{2,\text{ff}}$ (second from top), the biogenic contribution $\Delta\text{XCO}_{2,\text{bio}}$ (third from top) and the influence of the boundary conditions, $\Delta\text{XCO}_{2,\text{BC}}$ (bottom), The dominant wind conditions for each day given at the top of the figure and days without observations due to precipitation are in red.



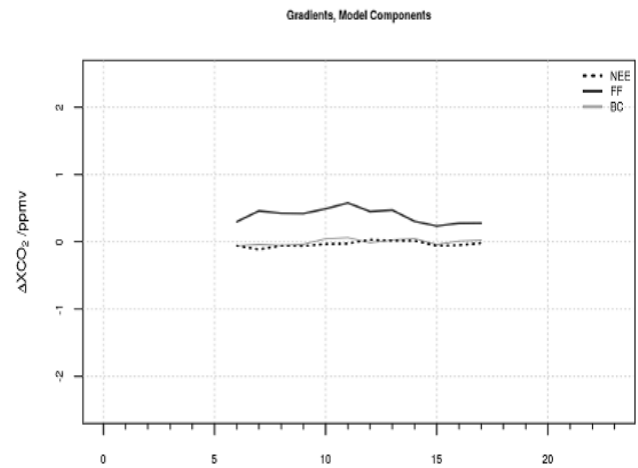
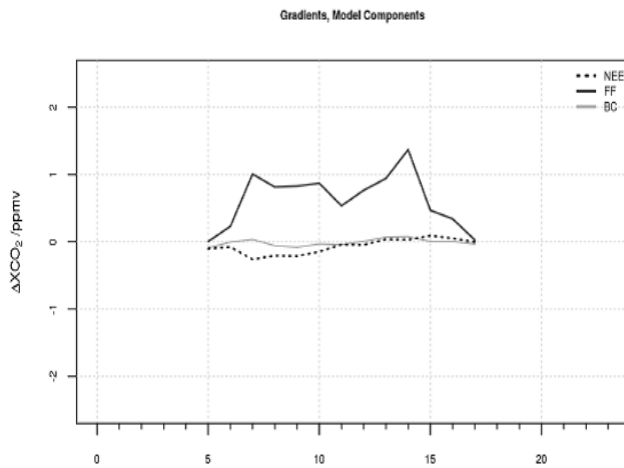
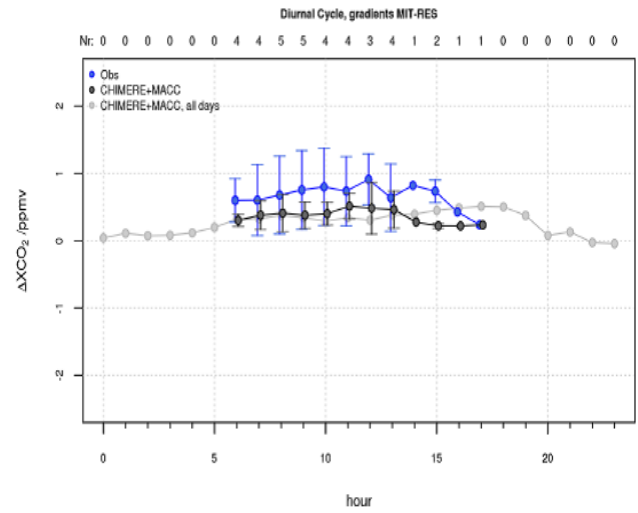
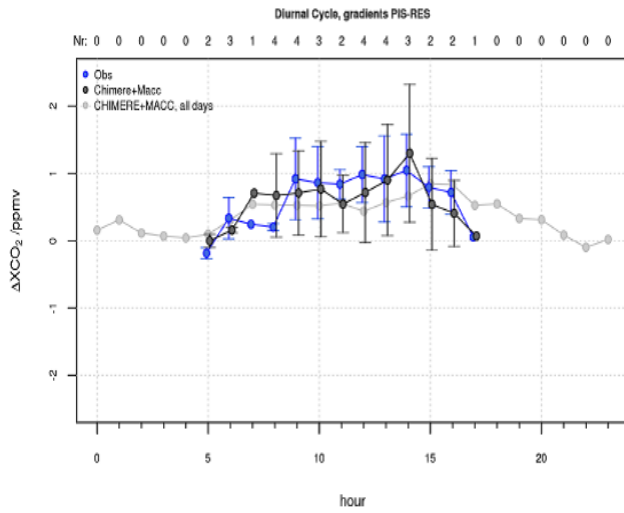
765
 766
 767

Figure 8. Comparison of modelled (solid lines) and observed hourly averaged XCO₂ (symbols) with standard deviations as error bars.



768
 769
 770
 771
 772

Figure 9. Comparison of modelled and observed hourly averaged ΔXCO_2 for gradients between PIS and RES (left) and MIT and RES (right), with standard deviations of the minute values of the hourly mean as vertical bars and the points color coded by wind direction from 0 to 359 degrees.



773
774
775
776
777
778
779
780

Figure 10. Comparison of modelled (black) and observed mean daily cycle (blue) of hourly averaged ΔXCO_2 of PIS (top left) and of MIT (top right) during the campaign when RES can be considered as upwind site. Labels on top of the upper figures denote the number of days contributing to the mean. The mean daily cycle for all days within the campaign period when PIS and MIT are downwind of RES is given in light grey. The modelled contribution of different CO_2 sources/sinks to the mean daily cycle for days with observations for the two sites is given in the bottom panels.

Instrument	XCO ₂ factor Berlin	XCO ₂ factor before Paris	XCO ₂ factor after Paris
1	1.0000 (0.0003)	1.0000 (0.0003)	1.0000 (0.0003)
2	0.9992 (0.0003)	0.9991 (0.0003)	0.9992 (0.0003)
3	1.0002 (0.0003)	1.0001 (0.0004)	1.0000 (0.0005)
4	0.9999 (0.0003)	1.0000 (0.0004)	1.0000 (0.0004)
5	0.9996 (0.0003)	0.9995 (0.0003)	0.9995 (0.0003)

781 Table 1. Normalisation factors for the five EM27/SUN instruments derived during
782 measurements before and after the Paris field campaign. Values in parentheses are
783 standard deviations. Measurements of instrument 1 were arbitrarily chosen as
784 reference from which the others were scaled. The calibration factors from a previous
785 field campaign in Berlin [Hase et al. 2015] are also shown. Calibration factors between
786 the two field campaigns agree well within 0.02 % (~0.08 ppm) for all instruments.

Date	No. of observations					Quality	Wind speed (m s ⁻¹)	Wind direction
	MIT	GIF	PIS	RES	JUS			
28 Apr 2015 (Tu)	179	102	178	199	234	++	4	W
29 Apr 2015 (We)	110	124	0	161	53	+	5	SW-W
04 Mai 2015 (Mo)	194	85	96	163	83	+	6	S-SE
05 Mai 2015 (Tu)	77	27	85	185	92	+	8	S-SW
06 Mai 2015 (We)	81	88	87	139	0	+	8	SW
07 Mai 2015 (Th)	169	313	252	286	238	+++	3	SW
09 Mai 2015 (Sa)	179	0	181	289	149	++	6	W
10 Mai 2015 (Su)	325	478	362	542	282	++++	3	S
11 Mai 2015 (Mo)	410	431	251	298	413	++++	3	SSW
12 Mai 2015 (Tu)	324	222	230	326	203	+++	4	NNW
13 Mai 2015 (We)	159	18	182	28	56	+	4	NE

787 **Table 2. Summary of all measurement days with the number of observations at each of**
788 **the sites, Mitry Mory (MIT) , Gif Sur Yvette (GIF), Piscop (PIS), Saulx-les-Chartreux**
789 **(RES), Jussieu (JUS), the overall quality ranking of each day according to the number**
790 **of available observations and temporal coverage (with classification from poor to great:**
791 **+, ++, +++, +++++), and the ground-level wind speed and direction.**