



ROOOH: the Missing Piece of the Puzzle for OH measurements in low NO Environments

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Abstract. Field campaigns have been carried out with the FAGE technique in remote biogenic environments in the last decade to quantify the *in situ* concentrations of OH, the main oxidant in the atmosphere. These data have revealed concentrations of OH radicals up to a factor of 10 higher than predicted by models, whereby the disagreement increases with decreasing NO concentration. This was interpreted as a major lack in our understanding of the chemistry of biogenic VOCs, particularly isoprene, which are dominant in remote pristine conditions. But interferences in these measurements of unknown origin have also been discovered for some FAGE instruments. We present in this paper convincing experimental and modeling evidence that the disagreement between model and measurement is due to interference by the unexpected decomposition of a new class of molecule, ROOOH, in the FAGE instruments. Including ROOOH reflects the missing piece of the puzzle in our understanding of OH in the atmosphere.

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1 Introduction

OH radicals are the most important oxidant in the atmosphere, and the detailed understanding of their formation and reactivity is key for the understanding of the overall chemistry. Upon reaction with Volatile Organic Compounds (VOCs, such as methane and isoprene), OH oxidation leads to the production of organic peroxy radicals (RO₂) who play a crucial role in the chemistry of tropospheric ozone and secondary organic aerosol (Monks et al., 2015). The concentration of OH radicals has been measured for several decades now (Holland et al., 2003; Creasey et al., 1997; Brune et al., 1995), and comparison of OH concentration profiles with model outputs is taken as a good indicator on the degree of understanding of the chemistry going on. Good agreement is often obtained between measurements and models for polluted environments (where levels of nitrogen oxides (NO_x=NO+NO₂) are in excess of 500 pmol/mol, or ppt), however remote and clean

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environments show much less good agreement (Stone et al., 2012). Several field campaigns in remote environments, dominated by natural biogenic emissions, have been carried out during the last decade (Whalley et al., 2011;Lelieveld et al., 2008;Hofzumahaus et al., 2009), and a very poor agreement has been found, with measured OH concentrations exceeding model predictions by up to a factor of 10. These findings have been interpreted to reflect a lack in our understanding of the oxidation mechanism of biogenic VOCs under low NO_x conditions and have triggered a large number of studies aiming at improving the atmospheric oxidation mechanism of biogenic VOCs (Peeters et al., 2009;Crounse et al., 2011;Paulot et al., 2009;Archibald et al., 2010). Improvements have been made especially in the oxidation mechanism of isoprene (Wennberg et al., 2018), and new reaction pathways leading to OH recycling have been found. However, none of these new chemical pathways has led to a sufficiently significant increase in modelled OH concentration to bring models into reasonable agreement with measurements (Rohrer et al., 2014).

An alternative explanation for the unexpectedly high OH concentrations measured in biogenic, low NO environments is that the measurements suffer from an unidentified interference. Indeed, all of these measurements have been carried out using a technique named FAGE (Fluorescence Assay by Gas Expansion). Briefly, ambient air is rapidly expanded into a low pressure volume, where OH radicals are excited by 308 nm light and the resulting fluorescence is detected (Heard and Pilling, 2003). Calibration of the fluorescence signal allows the determination of absolute concentrations (Dusanter et al., 2008). Interferences can arise from different sources such as photolysis of suitable precursors by the fluorescence excitation laser (*e.g.*, O₃), the presence of fluorescing species other than OH or the decomposition of labile species during the gas expansion into the FAGE cell (Ren et al., 2004). The first source of interference can, in principle, be identified by varying the excitation laser energy: “real” OH only needs one photon to fluoresce, whilst other species need two (one for generating OH radicals by photolysis, another for their excitation). Therefore the fluorescence intensity would not vary linearly with the excitation energy. Even though in practice this method is highly uncertain, given the generally low OH concentrations (and the resulting low S/N ratio) and the high temporal variability of OH radical concentration, the high OH concentrations observed in the different field campaigns seems to arise from “real” OH and not from the photolysis of other species. The second source of interference can be identified by regularly measuring the fluorescence signal with the excitation laser wavelengths slightly tuned off the OH line. This procedure is always adopted during measurements as it enables to account for stray light reaching the detector from the excitation laser or the sun.

The third source of interference, the generation of OH radicals during the expansion into the FAGE cell, is more difficult to identify because only one photon is needed and hence the interfering species would appear as “real” OH. Following the large disagreements between measurements and models, the group of W. Brune has conceived a method to quantify such possible interferences (Mao et al., 2012): a pre-injector device is installed just above the inlet into the FAGE cell, which injects regularly into the airflow a high concentration of a species rapidly reacting with OH radicals. This way all ambient OH radicals are scavenged before entering the FAGE cell, and any remaining signal can be identified as interference. The difference between the signal with and without the scavenger allows the quantification of the real ambient OH. This technique was used for the first time in 2012 in a forest in California (Mao et al., 2012) and led to the identification of a large



fluorescence signal following scavenging of all ambient OH radicals, corresponding to up to 50% of the total OH concentration. The OH concentrations obtained with the scavenger agreed well with models, while the OH concentrations obtained without the scavenger exceeded modeled concentrations by up to a factor of 3. Other groups have also developed a pre-injector system in the following years (Griffith *et al.*, 2016; Novelli *et al.*, 2014a; Tan *et al.*, 2017). Using this system, Novelli *et al.* (Novelli *et al.*, 2014a) have observed strong interferences in their FAGE system during three field campaigns in remote biogenic environments in Germany, Finland and Spain, while Griffith *et al.* (Griffith *et al.*, 2016) was able to account for the observations through known interferences by O₃ photolysis. Tan *et al.* (Tan *et al.*, 2017) have observed a small unexplained OH concentration using a prototype pre-injector device during a field campaign in rural China. However, technical difficulties with the prototype made it uncertain to draw final conclusions about the origin of this unexplained OH signal.

Novelli *et al.* proposed that ozonolysis of alkenes, leading to the formation of Criegee intermediates and the subsequent decomposition of these Criegee intermediates within the FAGE cell, was responsible for the interference (Novelli *et al.*, 2017). However, Rickly *et al.* (Rickly and Stevens, 2018) and Fuchs *et al.* (Fuchs *et al.*, 2016) could not confirm this source: even though they detected internally formed OH when mixing high concentrations of O₃ and alkenes in the laboratory, when they extrapolated their results to ambient conditions they found that the possible interference generated this way would be well below the detection limit of the FAGE. Chamber studies were carried out at the SAPHIR chamber in Jülich (Fuchs *et al.*, 2012), simulating remote forest conditions (*i.e.*, high biogenic VOC and low NO concentrations). OH concentrations were measured simultaneously by FAGE and by absolute DOAS absorption. No sizeable interference was detected in these experiments, even though the same group had previously observed unexpected high OH concentrations in the Pearl River delta in China (Hofzumahaus *et al.*, 2009; Rohrer *et al.*, 2014), exceeding modelled concentrations by up to a factor of 8.

Following several years of interference studies in various environments, recent work from W. Brune's group (Feiner *et al.*, 2016) concluded that the interference observed in their FAGE system (a) was due to a rather long-lived species because the interference persists into the evening, (b) it had been observed in different environments dominated by MBO, terpenes or isoprene, hence it must originate from a class of species rather than from only one species such as isoprene, (c) it strongly increased with increasing O(¹D), hence it must somehow be linked to photochemistry and (d) the species responsible for this interference was linked to a low NO_x oxidation pathway, because the extent of the interference steeply decreased with increasing NO concentration.

In this work we present convincing experimental and modelling evidence that this sought-after species is the product of the reaction between RO₂ radicals and OH radicals. In recent works it has been shown that this reaction is fast (Assaf *et al.*, 2017b; Assaf *et al.*, 2016) and could be competitive to other sinks for RO₂ radicals (Fittschen *et al.*, 2014; Archibald *et al.*, 2009), *i.e.* it becomes increasingly important with decreasing NO concentration. *Ab-initio* calculations (Müller *et al.*, 2016; Liu *et al.*, 2017; Assaf *et al.*, 2018b) have shown that the initial reaction product is a trioxide, ROOOH, obtained from the recombination of RO₂ and OH. The formation of this adduct is exothermic by around 120 kJ mol⁻¹ compared to the initial reaction partners and by around 110 kJ mol⁻¹ compared to the major decomposition products, RO + HO₂, largely independent



of the size of the alkyl moiety of the RO_2 . While for the smallest RO_2 radical, CH_3O_2 , stabilization of CH_3OOOH is not the major fate of the initial adduct (Assaf et al., 2017a; Müller et al., 2016) and the major products are $\text{CH}_3\text{O} + \text{HO}_2$, the HO_2 yield has been found to decrease with increasing size of the alkyl group and it is expected that for C_4 peroxy radicals the stabilization of the initially formed ROOOH is the major product (Assaf et al., 2018a). For RO_2 radicals obtained from an initial attack of OH radicals on biogenic VOCs, it can thus be expected that the major reaction product will also be the corresponding trioxides and that sizeable concentrations of this new class of species can accumulate and thus be present in low NO environments.

2. Results and Discussion

In the first part the experimental evidence for the interference generated in the FAGE by the presence of ROOOH molecules will be presented. In the second part, model calculations are used in order to estimate of the steady state concentration of ROOOH molecules that can possibly build up in different environments.

2.1 Experiments

With the goal of forming sizeable amounts of trioxide (ROOOH), experiments have been carried out in a pump-probe FAGE instrument of the University Lille (UL-FAGE), describes already in detail in earlier publications (Fuchs et al., 2017; Hansen et al., 2015; Parker et al., 2011). Briefly, a gas mixture containing isoprene (or C_4H_{10} or CH_4 , see Supplementary material) and $\text{O}_3/\text{H}_2\text{O}$ is photolysed at 266 nm at a repetition rate of 2 Hz. The photolysed mixture is expanded into the FAGE cell, and the OH concentration is monitored by time-resolved Laser Induced Fluorescence (LIF). The excitation laser operates at 5 kHz, hence the OH profiles are obtained with a time resolution of 200 ms. The residence time of the gas mixture in the photolysis cell is around 20 s, therefore the mixture is photolysed around 40 times before it reaches the FAGE inlet. Experiments start with a fresh mixture (*i.e.*, with the photolysis laser covered) and 40 decays are then registered every 0.5 s for 20 s. After 40 photolysis pulses the laser is covered again for 2 minutes to allow the mixture to completely refresh and, in order to improve the S/N ratio, a new series of measurements is started. After 20 series, the signals are averaged so that one OH decay profile is obtained for each sequential photolysis pulse. An example is shown in **Figure 1** where, for clarity, only one in every 10th decay profile is plotted.

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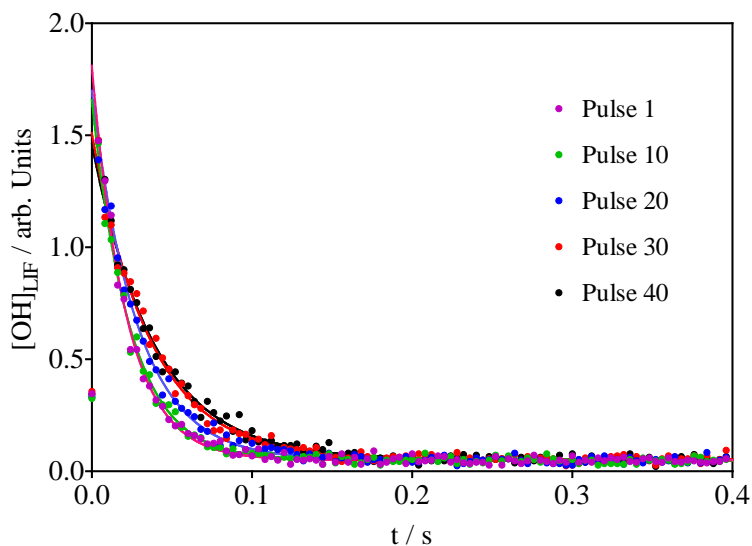


Figure 1: OH concentration time profiles following the photolysis of 600 ppb O₃ (leading to initial OH concentrations of around $1.4 \times 10^{10} \text{ cm}^{-3}$) in the presence of $3 \times 10^{11} \text{ cm}^{-3}$ isoprene. For clarity, only every 10th photolysis shot is shown. Time resolution was decreased from 200 μs to 4 ms by averaging 20 data points for clarity only: data analysis is carried out with raw data.

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The initial isoprene concentration ($3 \times 10^{11} \text{ cm}^{-3}$ in **Figure 1**) was chosen to make the reaction of RO₂ with OH compete efficiently with that of isoprene with OH after several photolysis pulses: with initial OH concentrations of $1.4 \times 10^{10} \text{ cm}^{-3}$ (obtained from calibration in separate experiments, see Supplementary data), the isoprene concentration decreases with each photolysis shot, while the RO₂ radical concentration increases. It can thus be expected that the concentration of ROOOH

10 increases with every photolysis pulse.

A single-exponential decay was fitted to the OH profiles from **Figure 1** and the resulting pseudo-first order decay rates are shown as blue dots in **Figure 2**. It can be seen that the decay rate decreases with increasing number of photolysis pulses. This is expected due to the ongoing transformation of reactive isoprene (and RO₂ radicals) into less reactive species. The decrease of $\sim 20 \text{ s}^{-1}$ corresponds to a decrease in isoprene concentration of around $2 \times 10^{11} \text{ cm}^{-3}$, in good agreement with

15 predictions of a kinetic model (see Supplementary data). The OH LIF signal at long reaction times, obtained as the plateau of the single-exponential fit (red dots in **Figure 2**), increases with increasing number of photolysis pulses ($m = (1.2 \pm 0.2) \times 10^{-4}$ arb. Units / photolysis pulse). This can be interpreted as interference due to decomposition of the increased concentration of ROOOH within the FAGE.

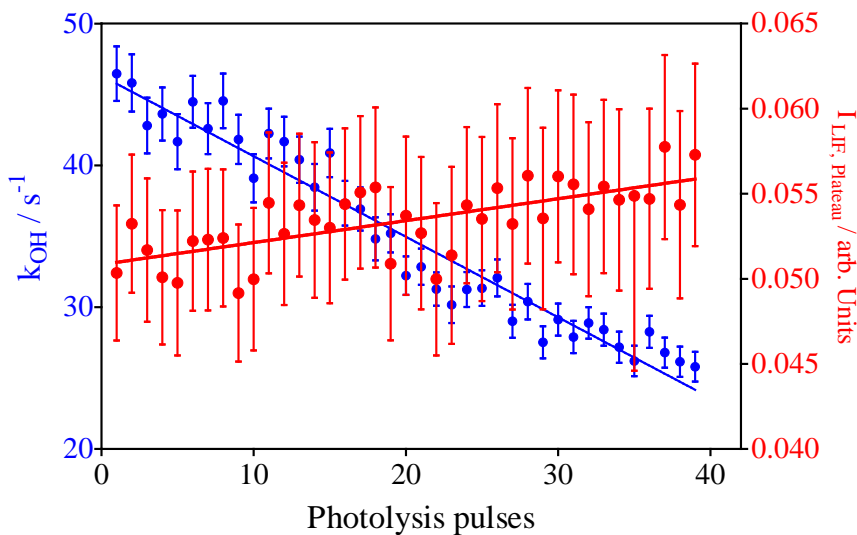


Figure 2: Results of fitting a mono-exponential decay to the raw signal of the experiments shown in Figure 1. Blue dots: OH decay rates from the mono-exponential fit (left y-axis). Red dots: fluorescence signal after reaction of all OH radicals obtained as plateau of the single-exponential fit (right y-axis). Error bars show 95% confidence interval such as obtained from the fit of the raw data.

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Additional experiments have been carried out with the goal to test if the observed increase in LIF signal at long reaction times is due to a 1- or 2-photon process. Experiments have been carried out with either decreased excitation laser energy (factor 2 lower) or with decreased excitation laser repetition rate (1 kHz instead of 5 kHz). Because photolytically generated interferences need two photons for generating one fluorescence photon, it would be expected that at lower photolysis energy the residual LIF signal would decrease with decreasing excitation laser energy: this was not the case, the same increase of residual LIF signal (relative to the LIF signal at time 0 after the photolysis pulse) was observed for both energies. The same is valid for changing the repetition rate of the excitation laser: photolytically generated interferences appear because the air mass within the excitation volume is not completely renewed between two excitation laser pulses (200 μ s at 5 kHz), and thus OH radicals generated with one pulse can be excited with the following pulse. Hence, such interference would be expected to decrease with decreasing repetition rate: this also was not the case in our experiments, the increase in the fluorescence signal was of the same order of magnitude for both repetition rates (see Supplementary data for full details).

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To further support the hypothesis that the observed increase in residual LIF signal is due to an interference generated by the decomposition of ROOOH species, additional experiments have been carried out with C₄H₁₀ and CH₄ instead of isoprene. C₄H₁₀ has been chosen because Assaf et al. (Assaf et al., 2018a) have shown experimentally that the HO₂ yield for the reaction of the corresponding RO₂ radical with OH is very low and *ab-initio* and RRKM calculations support the hypothesis that the major reaction product with increasing alkyl size of the RO₂ radical becomes the corresponding trioxide. For the reaction of the corresponding isoprene peroxy radical with OH such direct evidence is currently not available, and it could be imagined that the OH radicals would rather add to the remaining double bond rather than to the peroxy site. Note however,

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that the major conclusion from the above experiments (the product of the reaction between the isoprene-peroxy radical with OH generates an interference in the UL-FAGE) would still be the same. Using C_4H_{10} as peroxy radical precursor, the same increase in residual LIF signal with increasing number of photolysis pulses is observed (see supplementary data for more details). Experiments with CH_4 have been carried out because it is known that the HO_2 yield in the reaction of CH_3O_2 with OH is very high, and that the yield of stabilized CH_3OOOH is expected to be very low (Assaf et al., 2018a; Assaf et al., 2017a). Therefore, no increase in the residual OH signal would be expected with increasing number of photolysis pulses. And this is what has been observed in our experiments (see Supplementary data for details).

Additional experiments have been carried out with much higher isoprene or butane concentration such that the OH reactivity with the hydrocarbon always remains the major reaction path and that the reaction of OH with the accumulated RO_2 never becomes competitive. Under these conditions, negligible $ROOOH$ will be formed, while still comparable concentrations of RO_2 radicals are generated, and with this, comparable concentrations of the products of their cross reaction or reaction with HO_2 . In such conditions no increase in residual OH signal is observed, neither for isoprene nor for butane. This observation further supports the hypothesis that the observed increase in residual LIF signal is due to decomposition of $ROOOH$ and not to any other reaction product (see Supplementary data for details).

The increase in residual LIF signal in **Figure 2** over the 40 photolysis pulses is around 0.005 arb. units. This can be compared with the raw OH decays shown in **Figure 1**: the initial LIF signal of ≈ 1.7 arb. units corresponds to an OH concentration of $1.4 \times 10^{10} \text{ cm}^{-3}$. Therefore, the increase in residual signal corresponds to an equivalent OH concentration of $\approx 4 \times 10^7 \text{ cm}^{-3}$. The concentration of $ROOOH$ accumulated after 40 photolysis pulses was estimated to be $[ROOOH] \approx 1 \times 10^{11} \text{ cm}^{-3}$ using a simple model (see supplementary data), i.e. a fraction of $\approx 4 \times 10^{-4}$ of $ROOOH$ decomposed to OH radicals during the expansion within the UL-FAGE. It can hence be concluded that in the UL-FAGE an interference signal corresponding to $[OH] = 1 \times 10^6 \text{ cm}^{-3}$ (order of magnitude of the disagreement between model and measurements) could be generated by less than 100 ppt of $ROOOH$.

In order to estimate if $ROOOH$ concentrations in this range can possibly be accumulated in remote biogenic environments, calculations using global and box models have been performed.

2.2 Modeling Results

The global distribution of $ROOOH$ species produced by the $RO_2 + OH$ reaction was investigated using the Met Office's Unified Model with the United Kingdom Chemistry and Aerosols scheme (UM-UKCA), version 8.4 (Abraham et al., 2012). UM-UKCA is a global chemistry-climate model with a horizontal resolution of 1.875° in longitude \times 1.25° in latitude on 85 vertical levels from the surface up to a height of 85 km (in its N96-L85 configuration). The chemistry scheme and emissions used in the present study were described in detail in a recent work (Ferracci et al., 2018) and included isoprene oxidation (Archibald et al., 2010) and isoprene emissions.



Crucially, the model simulated the abundances of a number of peroxy radicals resulting from the oxidation of emitted VOCs: CH_3O_2 (methyl peroxy), $\text{CH}_3\text{CH}_2\text{O}_2$ (ethyl peroxy), $\text{CH}_3\text{CH}_2\text{CH}_2\text{O}_2$ (*n*-propyl peroxy), $(\text{CH}_3)_2\text{CHO}_2$ (*i*-propyl peroxy), $\text{CH}_3\text{C}(\text{O})\text{O}_2$ (acetyl peroxy), $\text{CH}_3\text{CH}_2\text{C}(\text{O})\text{O}_2$ (propionyl peroxy), $\text{CH}_3\text{C}(\text{O})\text{CH}_2\text{O}_2$ (propylldioxy peroxy). Peroxy radicals from the first oxidation of isoprene were lumped into one species, as those from the oxidation of isoprene oxidation products (methacrolein and methyl vinyl ketone). These were used, along with the modelled number densities of OH and a rate constant k_1 of $1.5 \times 10^{-10} \text{ cm}^3 \text{ s}^{-1}$ for all $\text{RO}_2 + \text{OH}$ reactions (consistent with laboratory studies (Faragó et al., 2015; Assaf et al., 2017b; Assaf et al., 2016)) to calculate the total rate of production of ROOOH species. The total atmospheric abundance of trioxide species, $[\text{ROOOH}]_{\text{ss}}$, was then calculated offline assuming steady state between the production and loss (L) processes of ROOOH, according to the equation:

$$[\text{ROOOH}]_{\text{ss}} = \frac{k_1[\text{OH}]\sum_{i=1}^n[\text{RO}_{2,i}]}{L} \quad (1)$$

where the sum is across all RO_2 radicals in the model excluding methyl peroxy radicals, for which it has been shown that the production of a trioxide species is only a minor product channel (Assaf et al., 2017a) while the trioxide yield is expected to be close to 1 for larger peroxy radicals (Assaf et al., 2018a).

Steady-state ROOOH abundances were calculated “offline” using the modelled abundances of hourly $[\text{OH}]$ and $[\text{RO}_2]$ along with a rate constant (Assaf et al., 2016; Assaf et al., 2017b) for ROOOH formation of $1.5 \times 10^{-10} \text{ cm}^3 \text{ s}^{-1}$. As the lifetime of these ROOOH species is not known, different removal rates were tested, ranging from 10^{-5} to 10^{-2} s^{-1} . The modelled $[\text{ROOOH}]$ followed a diurnal and seasonal cycle similar to that of its precursors (OH and RO_2). Therefore the highest $[\text{ROOOH}]$ values were found around midday-2pm in the summer months (JJA in the Northern Hemisphere, DJF in the Southern Hemisphere). The peak $[\text{ROOOH}]$ values shown in **Figure 3** and in **Figure S9** were determined by producing an average seasonal diurnal cycle for each model grid cell and then plotting only its peak $[\text{ROOOH}]$ value. **Figure 3** shows the average diurnal peak concentration of ROOOH in the Boreal (left) and Austral (right) summer obtained using a removal rate of 10^{-4} s^{-1} , leading to ROOOH lifetimes of around 3 hours. Peak concentrations of several 100 ppt are reached in this scenario, especially at tropical latitudes, which would lead to an interference in the UL-FAGE system of the order of $1 \times 10^6 \text{ cm}^{-3}$.

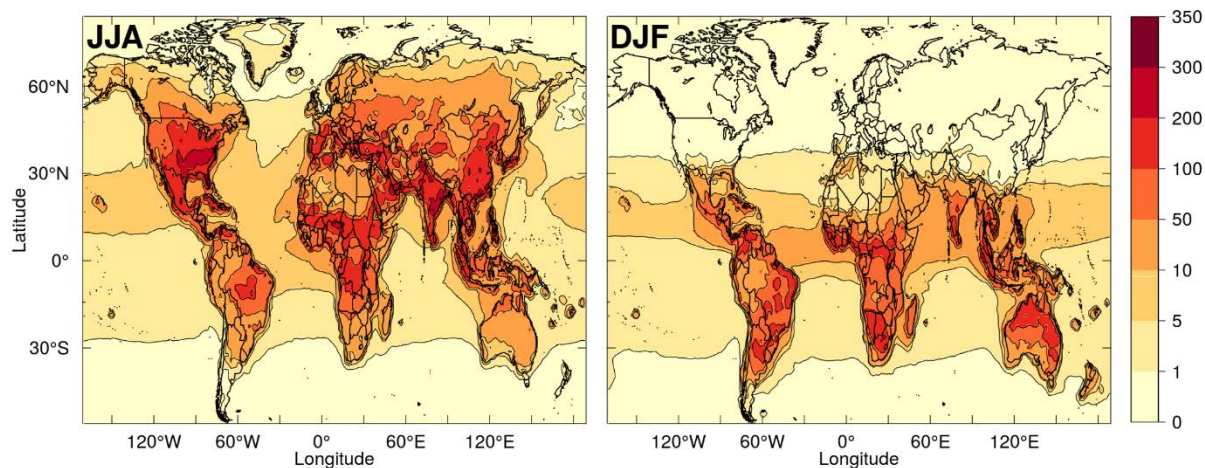


Figure 3: Modelled mean diurnal peak ROOOH volume mixing ratio (in ppt) during Northern (left) and Southern (right) summer months, using a combined removal rate for all ROOOH of 10^{-4} s^{-1} .

- 5 To confirm these global model results, a steady-state box model, constrained to observations made in the South East USA (Feiner et al., 2016), was developed. The results of the calculations with the steady-state model are shown in **Figure 4**, which highlights that at low levels of [NO] (< 200 ppt), typical in remote BVOC rich environments, levels of [ROOOH] are predicted to be on the order of 50-200 ppt. The two datasets plotted in **Figure 4** span a range of different NMVOC (isoprene) mixing ratios and highlight that ROOOH levels increase with increasing [VOC] and decreasing [NO], in agreement with the
- 10 global 3D modelling results shown in **Figure 3**.

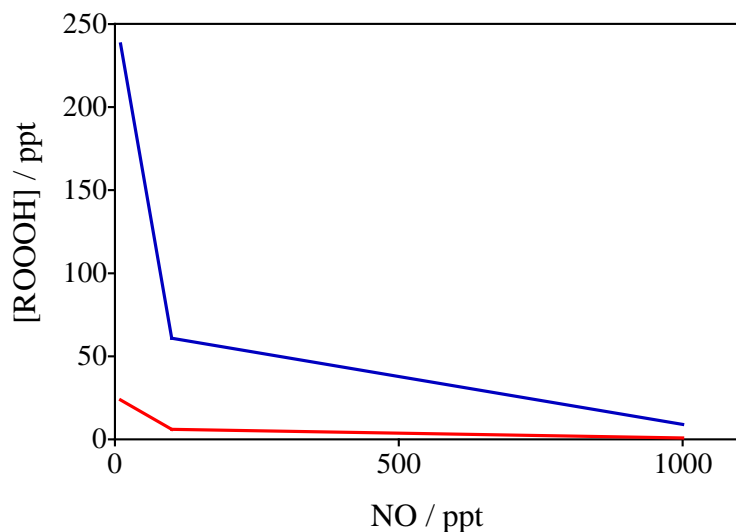




Figure 4: Variation in ROOOH as a function of NO (x-axis) and VOC reactivity (different colours) constrained by data from Feiner et al. (Feiner et al., 2016). Those data in red reflect a situation of VOC reactivity of 5 s^{-1} whilst the blue data reflect VOC reactivity of 24 s^{-1} (similar to that seen in regions like the Amazon).

5 3. Discussion

In this work we have shown that the product of the reaction of RO_2 radicals with OH radicals leads to an OH interference signal in the UL-FAGE instrument. If occurring also with a comparable intensity in other FAGE instruments, it can be high enough to explain numerous observations obtained with FAGE instruments from other research groups including:

- 10 (i) Underestimation by models of OH concentrations measured in remote, biogenic environments: the global model predicts ROOOH peak concentrations in remote environments that are probably high enough to explain, at least partially, the observed disagreement between model and measurements (Whalley et al., 2011; Lelieveld et al., 2008; Hofzumahaus et al., 2009; Tan et al., 2017).
- 15 (ii) Variability of interferences observed in field campaigns: The box model calculations have shown that the concentration of ROOOH species varies with NO, VOC concentration and $J(\text{O}^1\text{D})$ in the same way as the amplitude of the interference such as observed by the group of W. Brune (Feiner et al., 2016).
- 20 (iii) Interference observed from $\text{O}_3 + \text{alkenes}$: the tentative explanation of alkene ozonolysis being the source of internally formed OH radicals through decomposition of the stabilized Criegee intermediate (Novelli et al., 2017) is possibly also, at least partially, due to ROOOH formed in a secondary reaction from RO_2 and OH, both generated during the ozonolysis (Johnson and Marston, 2008) of the very high VOC and O_3 concentrations in laboratory experiments (Novelli et al., 2014b; Rickly and Stevens, 2018; Fuchs et al., 2016). Indeed, it is observed in these experiments that the interference scales with the $\text{O}_3 + \text{alkene}$ turnover rate, i.e. the time that ROOOH can accumulate.
- 25 (iv) Interferences observed in SAPHIR chamber: Fuchs *et al.* have carried out experiments under low NO conditions by comparing OH concentrations measured by FAGE and DOAS (Fuchs et al., 2012). Most of the time the agreement between both techniques was excellent, but on a few days towards the end of the campaign higher OH concentrations were measured by FAGE compared to DOAS. The NO concentrations on these days were lower, making the formation of ROOOH more likely, than on days with excellent agreement between FAGE and DOAS (Table 2 in Fuchs *et al.* (Fuchs et al., 2012)).
- 30 The results presented in this work thus propose a plausible solution to answer many open questions. Of course, the uncertainties are currently high on both, the observed FAGE interference per ROOOH molecule as well as the maximum ROOOH concentration that can accumulate in real environments. The first point could be improved through well-designed chamber studies under very low NO concentrations: such experiments have already been carried out (Nguyen et al., 2014)



- and a detailed analysis of the data might support the conclusions from this work. The second point is more difficult to ameliorate because the steady state ROOOH concentration directly scales with its removal rate, and currently nothing is known about the fate of ROOOH. Perhaps the table can be turned by using the evolution of the observed interferences to learn about the fate of ROOOH?
- 5 Nonetheless, even with current uncertainties the implications on our understanding of atmospheric oxidation chemistry are significant. We provide strong evidence for a plausible mechanism for how and why high OH levels in some environments are bolstered by a false signal, in a sense validating our current generation of models and reducing the need for speculative chemistry to explain the difference in simulated and observed OH in pristine environments. With further observations and model development, the outcome will be to improve our ability to predict the OH budget in pristine environments and the
- 10 impacts of changes on the global chemistry-climate system.

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