



1 **Characterization of biogenic primary and secondary**
2 **organic aerosols in the marine atmosphere over the East**
3 **China Sea**

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25



26 Abstract

27 Molecular composition and abundance of sugars and secondary organic aerosols (SOA) from
28 biogenic sources over the East China Sea were investigated based on gas
29 chromatography/mass spectrometry. Biogenic SOA tracers and sugars exhibit higher levels in
30 the samples affected by continental air masses, suggesting the terrestrial outflows of organic
31 matter to the East China Sea. Glucose was the dominant sugar species ($0.31\text{--}209\text{ ng m}^{-3}$, 18.8
32 ng m^{-3}), followed by mannitol – a fungal spore tracer. All sugar compounds showed higher
33 concentrations in the nighttime than in the daytime. 3-Methyl-1,2,3-butanetricarboxylic acid,
34 one high-oxidation tracer of monoterpene SOA, was found to be the most abundant species
35 among measured biogenic SOA markers, suggesting the input of aged organic aerosols
36 through long-range transport. Fungal-spore-derived organic carbon (OC) was the biggest
37 contributor to total OC ($0.03\text{--}19.8\%$, 3.1%), followed by sesquiterpene-derived secondary OC
38 (SOC), biomass-burning-derived OC, monoterpene- and isoprene-derived SOC. Larger
39 carbon percentages of biogenic OCs and SOCs in total OC presented in the terrestrially
40 influenced aerosols indicate significant contributions of continental aerosols through long-
41 range transport. Positive matrix factorization results illustrate that the secondary sulfate
42 ammonia, secondary nitrate and biogenic SOA, sea salt and coal combustion were the main
43 sources controlling the marine aerosols over the East China Sea, again highlighting the
44 importance of Asian continent as a natural emitter of biogenic organic aerosols together with
45 anthropogenic aerosols over the coastal marine atmosphere.

46

47 1 Introduction

48 Oceans cover more than 70% of the Earth's surface and marine aerosols contribute
49 significantly to the global aerosol load (O'Dowd et al., 2004), playing an important role in the
50 albedo, atmospheric processes, atmospheric chemistry, climate, and biogeochemical cycling
51 of nutrients (O'Dowd and de Leeuw, 2007; Shi et al., 2011). Such influences depend on the
52 chemical composition and concentrations of marine aerosols. In recent years, significant
53 abundances of organic matter in marine aerosol and their importance to the cloud
54 condensation nuclei (CCN) formation as well as their direct and indirect radiative effects have
55 been reported (Tervahattu et al., 2002; O'Dowd et al., 2004; Facchini et al., 2008; Sciare et al.,
56 2009; Pringle et al., 2010; Bougiatioti et al., 2011; Ovadnevaite et al., 2011a; Ovadnevaite et



57 al., 2011b). However, information about marine organic aerosol remains poor owing to
58 various emission sources, complex formation mechanisms, and limited field measurements
59 regarding their chemical composition and concentrations (Cavalli et al., 2004; O'Dowd and de
60 Leeuw, 2007). Therefore, it is necessary to investigate the loadings, molecular compositions
61 and sources of marine organic aerosols, especially in coastal and offshore regions where the
62 land/ocean interaction is active.

63 In general, sources of organic compounds in marine aerosols comprise natural and
64 anthropogenic emissions. For example, surface-active organic matter of biogenic origin in the
65 ocean (e.g. bacteria, viruses and detritus) can be transferred to the marine atmosphere by
66 bubble-bursting processes (Gershey, 1983; Mochida et al., 2002; Sciare et al., 2009; Gantt and
67 Meskhidze, 2013). Terrestrial air masses also bring plentiful organic matter derived from
68 natural and/or anthropogenic activities to the oceanic atmosphere via long-range atmospheric
69 transport (Zhou et al., 1990; Hawkins et al., 2010; Uematsu et al., 2010; Srinivas et al., 2011;
70 Kumar et al., 2012; Kang et al., 2017). Sugars, important water-soluble organic constituents
71 of atmospheric particulate matter, are substantial in both continental (Pashynska et al., 2002;
72 Iinuma et al., 2007; Yttri et al., 2007; Fu et al., 2008; Jia and Fraser, 2011; Cong et al., 2015)
73 and marine aerosols (Simoneit et al., 2004b; Fu et al., 2011). Due to their ubiquity and
74 abundance, sugars can be used to elucidate sources and transport of atmospheric organic
75 aerosols. Levoglucosan along with its two isomers, mannosan and galactosan, as the primary
76 thermal alteration products by the pyrolysis of cellulose and hemicellulose, has been
77 recognized as specific tracers for biomass burning (Simoneit, 2002). Trehalose is a recognized
78 fungal carbohydrate, which can be indicative of soil dust (Feng and Simpson, 2007; Fu et al.,
79 2012). Arabitol and mannitol are tracers for airborne fungal spores (Bauer et al., 2008).
80 Sucrose could serve as a marker for airborne pollen grains (Pacini, 2000; Fu et al., 2012),
81 while fructose and glucose can be emitted from plant pollen, fruits and detritus (Speranza et
82 al., 1997; Baker et al., 1998; Pacini, 2000). These sugar compounds have been detected in
83 some marine aerosols and effectively used as key tracers to assess contributions of different
84 emission sources (Simoneit et al., 2004b). Despite their importance, the knowledge about
85 molecular characterization of sugars in marine aerosols is still rare because of various sources
86 and inconvenience of sampling.

87 Apart from primary sources, atmospheric volatile organic compounds (VOCs) emitted from
88 ocean and/or continents also have significant impacts on marine aerosols. VOCs can react
89 with oxidants, such as ozone (O₃), nitrogen oxides (NO_x) and OH radicals in the atmosphere,



90 producing secondary organic aerosols (SOA) (Atkinson and Arey, 2003; Claeys et al., 2004a;
91 Jaoui et al., 2005; Meskhidze and Nenes, 2006; Claeys et al., 2007; Hallquist et al., 2009;
92 Gantt and Meskhidze, 2013; Hu et al., 2013). SOAs contribute substantially to atmospheric
93 particulate OM and strongly affects the regional and global air quality, climate and human
94 health (Zhu et al., 2016; Chen et al., 2017). Recent studies have revealed that SOA is an
95 important or even the dominant contributor to PM_{2.5} during heavy haze events in China (Cao
96 et al., 2017). Due to poor understanding of sources and formation mechanisms of SOA,
97 accurately measuring and modelling SOA concentrations remains a big issue (Zheng et al.,
98 2017). It was reported that emissions of biogenic VOCs (BVOCs) were one order of
99 magnitude larger than those of anthropogenic VOCs on a global scale (Guenther et al., 2006).
100 BVOCs emitted from terrestrial vegetation include isoprene, monoterpenes and
101 sesquiterpenes as well as other compounds (Guenther et al., 2006; Goldstein and Galbally,
102 2007), which could exert large impacts on marine aerosols through continental aerosols'
103 outflow. On the other hand, marine phytoplankton and seaweeds can release isoprene and
104 other BVOCs as well (Yokouchi et al., 1999; Shaw et al., 2010), especially during the
105 phytoplankton blooms (Hu et al., 2013). Marine-derived SOAs by oxidation of
106 phytoplankton-produced isoprene have been shown to remarkably influence the chemical
107 composition of marine CNN and cloud droplet number, forming an indirect climatic effect
108 (Meskhidze and Nenes, 2006; Gantt et al., 2009; Bikkina et al., 2014). In addition to isoprene,
109 the photooxidation products of oceanic monoterpenes (e.g. α - and β -pinene) were established
110 as well (Yassaa et al., 2008; Gantt et al., 2009; Fu et al., 2011). However, to date, the accurate
111 molecular characterisation and spatial distribution of biogenic SOAs (BSOA) in the marine
112 atmosphere are not well known due to their chemical complexity and process nonlinearity
113 (Chen et al., 2017). Consequently, there remains an urgent need for studies about marine
114 BSOA on a regional and/or global scale currently.

115 In this study, we investigated molecular compositions, abundances and spatial distributions of
116 sugars and BSOA tracers in marine aerosols over the East China Sea (ECS), which is located
117 between the east coast of Asian continent and the western North Pacific. The ECS is
118 susceptible to the outflow of continental OM from natural and anthropogenic activities in the
119 mainland. The contributions of secondary organic carbon (SOC) from isoprene, monoterpene
120 and sesquiterpene to the total OC were estimated. Positive matrix factorization (PMF) was
121 also used to estimate the contributions of sources identified by biomass burning tracers (e.g.



122 levoglucosan), primary saccharides including fungal spore and pollen tracers, BSOA tracers
123 from oxidation of BVOCs, inorganic ions and some other reliable source markers.

124

125 **2 Materials and methods**

126 **2.1 Aerosol sampling**

127 Marine total suspended particles (TSP) were collected during May 18 to June 12, 2014. The
128 sampling was performed using a high-volume air sampler (Kimoto, Japan) at an airflow rate
129 of $0.8 \text{ m}^3 \text{ min}^{-1}$ on board of the KEXUE-1 Research Vessel during a National Natural Science
130 Foundation of China (NSFC) sharing cruise. Figure 1 shows the cruise tracks and
131 concentrations of chlorophyll-a over the ECS. Detailed sampling information and map of
132 sampling sites are described elsewhere (Kang et al., 2017).

133 **2.2 Organic species analysis**

134 Filter aliquots were ultrasonically extracted for 10 min three times using
135 dichloromethane/methanol mixture (2:1, v/v). The solvent extracts were filtered through
136 quartz wool packed in a Pasteur pipette and concentrated with a rotary evaporator, and then
137 blown down to dryness with pure nitrogen gas. The extracts were reacted with 50 μL of N,O-
138 bis-(trimethylsilyl)trifluoroacetamide (BSTFA) containing 1% trimethylsilyl chloride and 10
139 μL of pyridine at 70°C for 3 hour in order to convert COOH and OH to the corresponding
140 trimethylsilyl (TMS) esters and ethers. After the reaction, 140 μL internal standard (C_{13} *n*-
141 alkane, $1.43 \text{ ng } \mu\text{L}^{-1}$) was added to the derivatives before injection into gas
142 chromatography/mass spectrometry (GC/MS).

143 **2.3 Gas chromatography/mass spectrometry**

144 Two organic species (i.e. sugars and SOA tracers) were determined on an Agilent model 7890
145 GC coupled to an Agilent model 5975C mass-selective detector (MSD). The GC instrument
146 was equipped with a split/splitless injector and a DB-5ms fused silica capillary column (30 m
147 $\times 0.25 \text{ mm i.d.}$, $0.25 \text{ } \mu\text{m}$ film thickness) with the GC oven temperature programmed from
148 50°C (2 min) to 120°C at $15^\circ\text{C min}^{-1}$, and then to 300°C at 5°C min^{-1} with final isotherm
149 hold at 300°C for 16 min. Helium was used as carrier gas. The GC injector temperature was
150 maintained at 280°C . The mass spectrometer was operated in the Electron Ionization (EI)



151 mode at 70 eV and scanned in the m/z range of 50 to 650 Da. Data were processed using
152 ChemStation software. Each compound was determined through comparing mass spectra with
153 those of authentic standards or literature data. GC/MS response factors were obtained with
154 authentic standards or surrogate standards. Recoveries of the standards that were spiked onto
155 pre-combusted quartz filters and measured as the samples ($n = 3$) were generally better than
156 80%. A field blank filter was treated as the real samples for quality assurance. The present
157 data were corrected with field blanks but not for recoveries.

158 **2.4 Positive matrix factorization (PMF) analysis**

159 To further investigate the potential sources of marine organic aerosols, the positive matrix
160 factorization (PMF) analysis are used in this study. For this analysis, the measured ambient
161 concentrations less than or equal to the method detection limit (MDL) were replaced by
162 MDL/2 and associated uncertainties were $(5/6)*MDL$. For the concentrations greater than the
163 MDL, the calculation of uncertainty is based on the following equation:

$$164 \quad \text{Uncertainty} = \sqrt{(\text{error fraction} * \text{concentration})^2 + (0.5 * MDL)^2} \quad (1)$$

165 where the error fraction is a user-provided estimation of the analytical uncertainty of the
166 measured concentration or flux. In the present study, the error fraction was set as 0.2 for all
167 species for receptor-oriented source apportionment analyses (Han et al., 2017).

168

169 **3 Results and discussion**

170 Concentrations of sugars and biogenic SOA tracers in marine aerosols over the ECS were
171 presented in Table 1. Total abundance of the quantified sugar compounds ranged from 1.8–
172 950 ng m^{-3} with an average of 81.5 ng m^{-3} much lower than those in urban atmosphere in
173 China (Wang et al., 2006), and the nighttime aerosols contained more sugars (average 90.1 ng
174 m^{-3}) than the daytime ones (72.6 ng m^{-3}). Individual sugar compounds also showed higher
175 nighttime concentrations, but still lower than those in urban aerosols (Wang et al., 2011). The
176 total biogenic SOA tracers were in the range of 1.1–135 ng m^{-3} (average 22.9 ng m^{-3}) with
177 lower nighttime abundance (22.2 ng m^{-3}) than daytime (23.6 ng m^{-3}). In contrast to
178 monoterpene and sesquiterpene SOA tracers, the isoprene SOA tracers presented higher levels
179 in the nighttime aerosols (9.6 ng m^{-3}) than those in the daytime samples (7.1 ng m^{-3}).



180 Generally, higher values of these organics were observed in coastal regions compared with
181 those far away from mainland (Fig. 2, Fig. 4 and Fig. S1), similar to the spatial pattern of
182 lipids, PAHs and phthalates in our previous report (Kang et al., 2017), suggesting the impact
183 from the outflow of continental OM on the basis of back trajectory analysis (Fig. S2-S6).

184 **3.1 Sugars and sugar-alcohols**

185 Sugars, a major class of water-soluble organic constituents in the atmosphere, have been
186 reported to be ubiquitous in marine aerosols (Simoneit et al., 2004b; Burshtein et al., 2011; Fu
187 et al., 2011; Chen et al., 2013). Most of them can serve as tracers of biological primary
188 aerosol particles and biomass burning (Simoneit, 2002; Bauer et al., 2008). Eleven sugar
189 compounds including anhydrosugars (levoglucosan, galactosan, mannosan), sugars (fructose,
190 glucose, sucrose and trehalose) and sugar alcohols (erythritol, arabitol, mannitol and inositol)
191 were measured in marine aerosols collected over the ECS. The concentrations of sugars are
192 shown in Table 1 and Fig. 2. Glucose ($0.31\text{--}209\text{ ng m}^{-3}$, mean 18.8 ng m^{-3}) and mannitol
193 ($0.03\text{--}169\text{ ng m}^{-3}$, 16.3 ng m^{-3}) were the dominant species, followed by sucrose ($0.09\text{--}216\text{ ng}$
194 m^{-3} , 11.7 ng m^{-3}), trehalose ($0.08\text{--}96.0\text{ ng m}^{-3}$, 9.4 ng m^{-3}), fructose ($0.09\text{--}106\text{ ng m}^{-3}$, 9.1 ng
195 m^{-3}), and levoglucosan ($0.09\text{--}64.3\text{ ng m}^{-3}$, 7.3 ng m^{-3}) (Fig. 3). Chen et al. (2013) also
196 reported that glucose and mannitol are the major sugar compounds detected in marine
197 aerosols collected at Chichi-Jima Island during spring and summer. Moreover, nighttime
198 concentrations of all sugars were generally higher than the daytime ones.

199 Figure 2 presents the temporal variations in sugar compounds determined in the marine
200 aerosol samples collected over the ECS. The concentrations of total sugars were characterized
201 by higher levels in the regions close to continent and/or influenced by land air masses,
202 indicating substantial influence of continental outflow compared to marine air over the
203 pelagic ocean.

204 **3.1.1 Anhydrosugars**

205 Levoglucosan, a thermal degradation production of cellulose (Simoneit, 2002) and a specific
206 indicator for biomass burning (Simoneit et al., 1999), can largely modify the chemical
207 composition of atmospheric aerosols on a regional to global scale. Due to its water solubility,
208 levoglucosan contribute to water-soluble organic carbon in aerosols, significantly enhancing
209 the hygroscopic properties of atmospheric aerosols (Mochida and Kawamura, 2004). The
210 abundance of levoglucosan detected in the present study ($0.09\text{--}41.2\text{ ng m}^{-3}$, mean 6.1 ng m^{-3})



211 and 0.10–64.3 ng m⁻³, mean 8.4 ng m⁻³ for day and night, respectively) was not the highest
212 among measured sugar species (Fig. 3 and Table 1). The average concentration of
213 levoglucosan is close to those reported at Gosan, Jeju Island in summer (mean 8.0 ng m⁻³) (Fu
214 et al., 2012). During the whole sampling period, higher concentrations of levoglucosan were
215 found in the samples under the effect of continental air masses based on five-day backward
216 trajectories (Fig. S2-S6), consistent with our previous report about lipids, PAHs and
217 phthalates (Kang et al., 2017). Mannosan and galactosan, isomers of levoglucosan, are
218 produced by pyrolysis of cellulose and hemicellulose and can also act as biomass burning
219 tracers (Simoneit, 2002; Fabbri et al., 2009). Mannosan and galactosan were detected in the
220 samples with similar variation trends to levoglucosan (Table 1 and Fig. 2), further indicating
221 the strong impact of continental biomass burning activities on the marine atmosphere. As
222 shown in Table 1 and Fig. 3, the nighttime average concentrations of all three anhydrosugars
223 were higher than the daytime ones, probably attributable to enhanced biomass combustion
224 and lower height of planetary boundary layer (PBL) as well as the land breeze during the
225 night.

226 In addition, hard wood tend to contain higher levels of cellulose than hemicellulose, therefore
227 the mass concentration ratios of levoglucosan to mannosan (L/M) can be utilized as a
228 diagnostic parameter for diverse biomass burning substrates (Zhu et al., 2015). Previous
229 studies found that L/M ratios from softwood are in a range of 3–10 and those from hardwood
230 are 15–25, while the ratios from crop residues are even higher (25–40) (Zhu et al., 2015). In
231 this study, the L/M ratios range from 0.71–32 with an average of 4.2, and higher values were
232 observed in the terrestrially influenced aerosols, especially coastal areas near Fujian province
233 (Fig. S7). The lower L/M ratios in our study (average 4.2) suggest that the aerosols were
234 mainly associated with the burning of softwood, consistent with the lower L/M ratios (2.1–4.8)
235 observed on Okinawa Island in May–June (Zhu et al., 2015) and Mt. Fuji (4.6–7.6, mean 5.5)
236 (Fu et al., 2014). It is worth noting that a couple of samples near Fujian Province in Southeast
237 China were characterized by a higher L/M ratio (>20) (Fig. S7), agreeing well with the values
238 of straw burning smokes (Chen et al., 2013). Zhu et al. (2015) also reported combustion of
239 agricultural residues, peat and wood could contribute to high L/M ratios. Therefore, these
240 higher values probably indicate emissions from terrestrial burning of straw residues, which
241 could affect the chemical components of aerosols in the western North Pacific via long-range
242 atmospheric transport as confirmed by backward trajectories (Fig. S5).



243 3.1.2 Sugar alcohols

244 Sugar alcohols detected in these samples consist of arabitol, mannitol, inositol and erythritol,
245 which had similar temporal patterns (Fig. 2e-h). Higher abundances of sugar alcohols
246 observed in the terrestrially influenced aerosol samples suggesting a significant contribution
247 from terrestrial source (Fig. S2-S6). While the marine fungi and algae, which can also release
248 fungal spores into the marine atmosphere via bubble bursting, contribute little to the sugar
249 alcohols in the coastal aerosols. Obviously, mannitol and arabitol were the most abundant
250 sugar alcohols detected in the present study (Fig. 3 and Table 1), ranging from 0.03–169 ng
251 m⁻³ (16.3 ng m⁻³) and 0.02–51.0 ng m⁻³ (5.1 ng m⁻³), respectively. A strong positive
252 correlation between arabitol and mannitol were found in marine aerosols ($p < 0.001$, $r = 0.996$,
253 $N = 51$), suggesting a similar origin. Mannitol and arabitol are very common in fungi, and are
254 the most frequently occurring sugar alcohols in plants (Burshtein et al., 2011), which can be
255 utilized to assess the contribution of fungal spores to the aerosol OC (Bauer et al., 2008). For
256 instance, mannitol is particularly abundant in algae (Burshtein et al., 2011). The peak
257 concentrations of sugar alcohols were found in the nighttime samples, which can be attributed
258 to the increased activities of yeasts and fungi at night (Graham et al., 2003). Moreover, the
259 maxima concentrations of these sugar polyols, especially arabitol and mannitol, were
260 observed in offshore regions in early summer (June) (Fig. 2e-h), likely due to the more active
261 microbial activities resulting from warmer temperature in June and more biota in coastal
262 regions (Fig. 1).

263 3.1.3 Sugars

264 Glucose, fructose, sucrose, and trehalose are the primary saccharides measured in marine
265 aerosols over the ECS. Glucose was the dominant sugar compound (0.31–209 ng m⁻³, 18.8 ng
266 m⁻³), followed by sucrose (0.09–216 ng m⁻³, 11.7 ng m⁻³) (Fig. 2i-l and Fig. 3). Glucose and
267 fructose originate from plant materials, such as pollen, fruits and their fragments (Graham et
268 al., 2003; Fu et al., 2012). Both glucose and fructose presented higher levels in the coastal
269 areas, demonstrating great contribution from terrestrial vegetation. Medeiros and Simoneit
270 (2007) reported that high abundance of glucose associated with lower molecular weight fatty
271 acids (mainly C₁₆) was attributed to the spring bloom of algae. In our study, C_{16:0} fatty acid
272 correlated well with glucose ($p = 0.001$, $r = 0.46$, $N = 51$), which suggests that marine sources
273 also contributed to the particulate glucose in the oceanic atmosphere. It was noted that
274 fructose was obviously related to glucose ($p < 0.001$, $r = 0.94$, $N = 51$), indicating they share



275 similar sources. Sucrose, as a predominant sugar species in the phloem of plants as well as
276 developing flower buds, is reported to be the richest and dominant component of airborne
277 pollen grains (Graham et al., 2003). High abundance of sucrose along with fructose and
278 glucose observed in our study implies a large emission of airborne pollen grains into the
279 marine atmosphere over the ECS. Significantly, the concentration of glucose, fructose and
280 sucrose were highest at the beginning of June in early summer (Fig. 2i-l), probably
281 attributable to an enhanced pollen emission, because pollen counts tend to be highest in late
282 spring/early summer in temperate zones (Graham et al., 2003). Such peak concentrations in
283 early June are in accord with a previous study by Pashynska et al. (2002). Trehalose, a fungal
284 metabolite, is present in a variety of microorganisms (fungi, bacteria, yeast and algae), and a
285 few higher plants as well as invertebrates (Medeiros et al., 2006a). Thus, trehalose can be
286 used as a microbial biomarker and stress protectant. Furthermore, trehalose was reported to be
287 the most abundant sugar in soil (Medeiros et al., 2006b; Rogge et al., 2007; Jia and Fraser,
288 2011), thus the enrichment of trehalose in aerosols can be indicative of soil resuspension and
289 unpaved road dust (Simoneit et al., 2004a; Fu et al., 2012). The trehalose in our study was
290 positively correlated with non-sea-salt calcium (nss-Ca^{2+}) ($p < 0.01$, $r = 0.43$, $N = 51$), the
291 best tracer for soil dust (Virkkula et al., 2006), suggesting the atmospheric trehalose over the
292 ECS was mainly derived from resuspension of soil particles. Higher abundance of trehalose in
293 coastal regions (Fig. 2l), may be in connection with the outflow of Asian dust primarily
294 occurring in winter/spring (Fu et al., 2012), and inland soil resuspension/dust aerosols.

295 **3.2 Secondary organic aerosols**

296 The sum of all SOA tracers ranged from 1.1–135 ng m^{-3} (22.9 ng m^{-3}), which were higher
297 than previous report of marine aerosols (0.19–27 ng m^{-3} , 6.6 ng m^{-3}) (Fu et al., 2011), but
298 much lower than those in the continental sites (Fu et al., 2010; Ding et al., 2014). These
299 differences suggest that the major source of SOA tracers over the ECS is probably of
300 terrestrial origin. Specifically, the total concentrations of detected isoprene SOA-tracers
301 ranges from 0.15–64.0 ng m^{-3} (8.4 ng m^{-3}), comparable to those reported in marine aerosols
302 (mean 8.5 ng m^{-3}) (Hu et al., 2013), but lower than those of urban aerosols (Ding et al., 2014;
303 Ren et al., 2017); total monoterpene SOA-tracers in the range of 0.26–87.2 ng m^{-3} (11.6 ng m^{-3})
304 lower than those in urban aerosols as well (Ren et al., 2017); total sesquiterpene SOA-
305 tracers ranged between 0.16 and 17.2 ng m^{-3} (2.9 ng m^{-3}). Such phenomena that BSOA
306 derived from monoterpenes are more abundant than those from isoprene were also reported in



307 Chinese urban areas (Ren et al., 2017). On the whole, biogenic SOA tracers exhibited higher
308 loadings in the coastal areas than those remote sampling sites, further indicating that long-
309 range atmospheric transport of terrestrial aerosols has significant influence on the chemical
310 composition and abundance of SOA over oceans. Marine VOC spatial distributions are
311 expected to be linked to the distributions of photosynthetic pigments in seawater, such as
312 chlorophyll-a (Ooki et al., 2015). The higher concentrations of chlorophyll-a in the coastal
313 regions (Fig. 1) stand for higher biological activities and more emission of VOCs, which
314 agree well with higher SOA tracers over the coastal waters. Such spatial variations in
315 biogenic SOA tracers are in agreement with a previous report about marine organic aerosols
316 collected during a round-the-world cruise (Fu et al., 2011).

317 **3.2.1 Isoprene SOA tracers**

318 Isoprene is a reactive biogenic hydrocarbon and primarily originates from terrestrial
319 photosynthetic vegetation (e.g. trees and plants). Marine phytoplankton and seaweed can also
320 emit isoprene (Yokouchi et al., 1999; Shaw et al., 2010). Moreover, bacteria produce isoprene
321 as well, and the bacterial isoprene production is temperature-dependent (Kurihara et al., 2010).
322 In spite of much lower emission strength in the ocean region, more recent researches have
323 suggested that the oceanic source of isoprene significantly impact atmospheric chemistry and
324 cloud microphysical properties in the remote marine boundary layer because of its high
325 reactivity (about 1–2 hour lifetime) (Hackenberg et al., 2017).

326 Six isoprene SOA tracers, including 2-methylglyceric acid, three C₅-alkene triols, and two 2-
327 methyltetrols (2-methylthreitol and 2-methylerythritol), were identified in the marine aerosols
328 over ECS. Isoprene SOA tracers showed diurnal variations with higher average
329 concentrations at night (Fig. 4a-c, Fig. 5, Fig. S8), consistent with the report by Fu et al.
330 (2010). The higher abundance during the night can be explained by the enhanced gas-to-
331 particle partition at cooler temperatures during the nighttime and/or increased input of
332 continent-originated isoprene-SOA into the oceanic atmosphere via land-sea breeze
333 circulations at night. However, T-test showed that the difference between daytime and
334 nighttime concentrations for 2-methyltetrols, C₅-alkene triols and 2-MGA was not that
335 significant ($p > 0.05$). As expected, much higher concentrations of isoprene tracers were
336 observed in coastal regions, where continental outflows exert larger effects from spring to
337 early summer. For the remote ocean, terrestrial sources have weak impacts because of the
338 short atmospheric lifetime of isoprene and the dilution effects during long-range atmospheric



339 transport. In addition to the effect of continental outflow, more nutrients in the coastal and
340 estuarine regions could be another factor responsible for the higher levels of isoprene SOA
341 tracers compared to the pelagic areas. Because nutrient-rich surface water can promote the
342 development of phytoplankton blooms and increases chlorophyll-a concentrations.
343 Chlorophyll-a is a measure of phytoplankton, or algal, biomass (Quinn et al., 2014) and
344 currently most widely used proxy for predicting isoprene concentrations in water
345 (Hackenberg et al., 2017). Numerous studies reported the positive relationship between
346 isoprene emission and chlorophyll-a in the surface seawater (Zhu et al., 2016; Hackenberg et
347 al., 2017). In the present study, the temporal and spatial distributions of chlorophyll-a at the
348 ECS surface during the whole sampling period are characterized by higher coastal levels, such
349 as waters near Zhejiang and Fujian provinces, but lower abundance in eastern Taiwan and the
350 remote sea (Fig. 1). Thus, high chlorophyll-a waters in coastal locations mean more isoprene
351 emissions than remote open waters. However, the isoprene in the remote ocean may mainly
352 originate in situ from biological production by marine biota at the ocean surface.

353 The low-NO_x products 2-methyltetrols with mass concentrations ranging from 0.03–41.9 ng
354 m⁻³ (4.8 ng m⁻³) were the major species among the isoprene SOA tracers, in line with early
355 report about summer aerosols in China (Ding et al., 2014). Specially, concentration ranges of
356 2-methyltetrols were 0.11–17.0 ng m⁻³ (3.8 ng m⁻³) during the daytime and 0.03–41.9 ng m⁻³
357 (5.7 ng m⁻³) at night with 2-methylerythritol being about 2.1-fold more abundant than 2-
358 methylthreitol. This ratio is similar to those calculated in previous studies (Ion et al., 2005;
359 Cahill et al., 2006; Fu et al., 2010). The atmospheric levels of 2-methyltetrols are comparable
360 to those reported in marine aerosols collected during a round-the-world cruise (0.07–15 ng m⁻³
361 ³, 2.4 ng m⁻³) (Fu et al., 2011). However, these values are small compared to those from
362 terrestrial emissions, such as mountain (Cahill et al., 2006; Fu et al., 2014) and forest aerosols
363 (Claeys et al., 2004a; Fu et al., 2010; Miyazaki et al., 2012). A significant positive
364 relationship between 2-methyltetrols and levoglucosan ($p < 0.001$, $r = 0.87$, $N = 51$) in our
365 study suggests that biomass burning may also generate the precursors of 2-methyltetrols
366 followed by photochemical reactions (Xie et al., 2014). Besides, 2-methyltetrols correlated
367 with C₂₉ *n*-alkane (the dominant species of terrestrial higher plant waxes) as well ($p < 0.001$, r
368 = 0.74, $N = 51$), suggesting these organic tracers originate from higher plants or from similar
369 source regions as well. The correlations above illustrate that terrestrial emissions (e.g.
370 biomass burning and higher plants) contributed significantly to the levels of 2-methyltetrols in
371 the marine atmosphere.



372 C₅-Alkene triols are formed from photooxidation of isoprene under low-NO_x conditions
373 (Surratt et al., 2006; Lin et al., 2013). They were detected in all the samples ranging from
374 (0.03–14.6 ng m⁻³, 2.2 ng m⁻³), which are higher than those reported in Arctic aerosols (0.01–
375 0.15 ng m⁻³) (Fu et al., 2009) and other marine aerosols (0.002–4.6 ng m⁻³, 0.65 ng m⁻³) (Fu
376 et al., 2011); but lower than the atmospheric levels of mountain aerosols (Fu et al., 2010; Fu
377 et al., 2014) and subtropical urban aerosols from Hong Kong (Hu et al., 2008). Such a
378 difference illustrates the outflow of continental aerosols, which can be confirmed by the
379 significant positive correlations between C₅-alkene triols and tracers of terrestrial emissions,
380 such as C₂₉ *n*-alkanes ($p < 0.001$, $r = 0.78$, $N = 51$) and levoglucosan ($p < 0.001$, $r = 0.87$, $N =$
381 51). It was found that 2-methyltetrols correlated well with C₅-alkene triols in marine aerosols
382 over ECS ($p < 0.001$, $r = 0.86$, $N = 51$), suggesting a similar formation mechanism or
383 common sources. However, the mass concentration ratios of C₅-alkene triols to 2-
384 methyltetrols showed significant variation in different sampling sites (Fig. 6a), indicating
385 different formation pathways, consistent with a previous report (Fu et al., 2010).

386 Concentrations of 2-methylglyceric acid (2-MGA), formed by photooxidation of isoprene
387 under high-NO_x (NO_x = NO + NO₂) conditions (Surratt et al., 2006), were in the range of
388 0.09–8.3 ng m⁻³ (1.4 ng m⁻³) being greatly lower than those in mountain aerosols (Cahill et al.,
389 2006; Fu et al., 2010; Fu et al., 2014), implying much stronger influence of NO_x on isoprene
390 SOA formation in continental aerosols. 2-MGA is a possible further oxidation product of
391 methacrolein and methacrylic acid, which are two major gas-phase oxidation products of
392 isoprene (Claeys et al., 2004b; Edney et al., 2005; Fu et al., 2009). 2-MGA was related to C₂₉
393 *n*-alkanes ($p < 0.001$, $r = 0.84$, $N = 51$) and levoglucosan ($p < 0.001$, $r = 0.81$, $N = 51$) as well,
394 again suggesting a terrestrial input. These isoprene-SOA tracers in marine aerosols over the
395 ECS may stem from terrestrial higher plants and biomass burning, and then were oxidized
396 during the transport to the oceanic atmosphere. Similar correlations between isoprene-derived
397 SOA tracers and levoglucosan were mentioned in previous study, which reported that biomass
398 burning enhanced the isoprene-SOA formation (Li et al., 2018). Additionally, 2-methyltetrols
399 are higher generation products than 2-MGA (Fu et al., 2014), but the ratios of 2-MGA to 2-
400 methyltetrols did not vary significantly with cruise track. However, a couple of high ratios
401 were observed in the ocean-air influenced aerosols, suggesting the importance of marine
402 source for fresh isoprene-derived SOAs in the atmosphere (Fig. 6b).

403 NO₃⁻ was found to be related to 2-MGA, 2-methyltetrols and C₅-alkene triols ($p < 0.001$, $r =$
404 0.60–0.71, $N = 51$). The ratio of 2-MGA to 2-methyltetrols was found to be negatively



405 correlated with particulate NO_3^- ($p < 0.05$, $r = -0.31$, $N = 51$). These relationships reveal that
406 there may be a close connection between formation of isoprene SOA and NO_3^- in the marine
407 boundary layer.

408 3.2.2 Monoterpene SOA tracers

409 The detected α/β -pinene oxidation products in the study consist of pinonic, pinic acids, 3-
410 hydroxyglutaric acid (3-HGA) and 3-methyl-1,2,3-butanetricarboxylic acid (MBTCA). They
411 are derived from the photooxidation of α/β -pinene with O_3 and OH radicals (Hoffmann et al.,
412 1997; Yu et al., 1999; Glasius et al., 2000; Iinuma et al., 2004). Since monoterpenes account
413 for around 35% of the global biogenic VOCs' emissions, these compounds have been utilized
414 to estimate the role of monoterpene oxidation in the SOA formation (Griffin et al., 1999).
415 Monoterpenes were chiefly emitted from needle leaf trees. In this study, monoterpene-SOA
416 tracers were found to be positively correlated to levoglucosan with $p < 0.001$ (r ranges from
417 0.68–0.82), indicating terrestrial biomass burning made substantial contributions to the
418 formation of monoterpenes, then being transported to the oceanic atmosphere. Generally, total
419 monoterpene-SOA tracers in our study showed a major peak in spring (Fig. S1g), in
420 agreement with a previous study (Zhu et al., 2016). Pinonic and pinic acids, well-known
421 tracers for α/β -pinene, ranged from (0.02–1.6 ng m^{-3} , 0.47 ng m^{-3}) and (0.16–14.9 ng m^{-3} , 3.4
422 ng m^{-3}), respectively (Table 1). Their concentrations were higher than those in high Arctic
423 aerosols (average 69 pg m^{-3} and 514 pg m^{-3} , respectively) (Fu et al., 2009). In addition,
424 concentrations of pinic acid were 7 times higher than pinonic acid, similar to an earlier report
425 (Fu et al., 2009). The vapour pressure of pinic acid is about 2 orders of magnitude lower than
426 pinonic acid (Bhat and Fraser, 2007). Thus, pinic acid can saturate and readily nucleate, being
427 expected to have a higher fraction in the aerosol phase than pinonic acid.

428 Two novel monoterpene SOA tracers, 3-hydroxyglutaric acid (3-HGA) and 3-methyl-1,2,3-
429 butanetricarboxylic acid (MBTCA), were also detected in these marine aerosols. Both of them
430 are reported to be higher-generation products of α -pinene photooxidation (Szmigielski et al.,
431 2007; Kourtchev et al., 2009). The concentration ranges of 3-HGA in marine aerosols over
432 ECS were lower than those of aerosols in Mt. Tai, central east China (Fu et al., 2010), but
433 comparable to other studies about marine aerosols (Fu et al., 2011). The average abundance of
434 atmospheric MBTCA for all samples were 5.6 ng m^{-3} with mean concentrations of 6.8 ng m^{-3}
435 and 4.5 ng m^{-3} during the day and night, respectively, more abundant than 3-HGA (2.2 ng m^{-3}
436 3). Interestingly, these mean values of MBTCA were comparable to those of mountain



437 aerosols (Fu et al., 2010), but still higher than the marine aerosols collected on the remote
438 seas (Fu et al., 2011). Since the ECS is adjacent to Mainland China, the atmospheric boundary
439 layer in these regions would inevitably be affected by continent-derived air masses, as
440 supported by the spatial pattern of individual SOA tracers (Fig. 4 and Fig. S1), in particular of
441 the peaks occurring in the coastal areas. Furthermore, in contrast to isoprene SOA tracers,
442 monoterpene SOA tracers commonly exhibit higher average daytime concentrations (Table 1
443 and Fig. 5), but T-test showed no significant difference ($p > 0.05$) between daytime and
444 nighttime concentrations for both total and individual monoterpene SOA tracers. The
445 concentration ratio of 3-HGA plus MBTCA to pinic acid ((3HGA+MBTCA)/pinic) showed
446 higher values in the terrestrially influenced aerosols (Fig. 6c). Besides, this mass
447 concentration ratios were basically higher during the daytime (mean 2.0) than nighttime (1.9),
448 indicating that aged aerosols are more abundant in the daytime.

449 3.2.3 Sesquiterpene SOA tracers

450 The analysis of sesquiterpenes is a great challenge due to their high reactivity and relatively
451 low vapour pressure. However, the aging of β -caryophyllene derived SOA has proved its
452 significant effects on all CNN-relevant properties (Asa-Awuku et al., 2009). β -caryophyllinic
453 acid is one of the most abundant species among sesquiterpenes originating from plants (Duhl
454 et al., 2008). It is an ozonolysis or photooxidation product of β -caryophyllene (Jaoui et al.,
455 2007). Moreover, sesquiterpenes accumulated in leaves and woods can be emitted during
456 biomass combustion (Ciccioli et al., 2014).

457 The atmospheric levels of β -caryophyllinic acid were $0.16\text{--}17.2 \text{ ng m}^{-3}$ (mean 2.9 ng m^{-3}),
458 with $0.17\text{--}17.2 \text{ ng m}^{-3}$ (3.5 ng m^{-3}) during daytime and $0.16\text{--}9.6 \text{ ng m}^{-3}$ (2.3 ng m^{-3}) during
459 nighttime, respectively (Table 1). As expected, β -caryophyllinic acid correlated well with
460 levoglucosan ($p < 0.001$, $r = 0.61$, $N = 51$) in agreement with earlier report (Ding et al., 2016),
461 indicating substantial contribution of terrestrial biomass burning to sesquiterpenes' loading in
462 the marine atmosphere. On the other hand, biomass-burning processes (e.g. crop straw
463 combustion and forest fires) can raise ambient and/or leaf temperature to some extent,
464 consequently probably enhancing the emission of sesquiterpenes from trees and plants.

465 Many researches have proved that sesquiterpene emissions increase significantly with
466 increasing ambient temperatures (Tarvainen et al., 2005; Jaoui et al., 2007). Ambient
467 temperature seems to be the dominant factor controlling temporal variation in sesquiterpene
468 emission although other factors contribute (Duhl et al., 2008). An early study reported that



469 emission rates of sesquiterpene were 1.2–3 times higher in the daytime as well (Duhl et al.,
470 2008). All the reports described above perfectly interpret much higher concentrations of β -
471 caryophyllinic acid during daytime in our study. The abundance of β -caryophyllinic acid in
472 this study are higher than those reported in the remote marine aerosols during a round-the-
473 world cruise, but comparable to the maximum concentration of 2.5 ng m^{-3} observed at
474 California coast (Fu et al., 2011); however, the concentrations of β -caryophyllinic acid were
475 much lower than those of Mt. Tai aerosols, central east China (average 12 ng m^{-3} for both
476 daytime and nighttime aerosols) (Fu et al., 2010). The spatial distributions of β -caryophyllinic
477 acid also presented higher levels in coastal regions than other sampling sites (Fig. 4h), again
478 proving contribution of terrestrial aerosols.

479 3.3 Contributions of biogenic primary and secondary sources

480 To assess the relative abundances of organic aerosols from primary and secondary emission
481 sources, tracer-based methods are employed to evaluate their contributions to the marine
482 ambient OC. For example, mannitol and arabitol determined in marine aerosols were used to
483 calculate the contributions of fungal spores to OC (Bauer et al., 2008); the average mass
484 percent ratios of levoglucosan to OC (8.14%) are used to investigate the biomass burning
485 derived OC (Fu et al., 2014). Besides, biogenic SOA tracers detected in the present study are
486 utilized to evaluate the SOC formation resulting from the oxidation of isoprene, α -pinene and
487 β -caryophyllene through a tracer-based method reported by Kleindienst et al. (2007). This
488 method used the laboratory-derived mass fractions (f_{soc}) of marker species generated from
489 known precursors into SOC concentrations. Specifically, the f_{soc} values used in our study is
490 0.155 ± 0.039 for isoprene, 0.231 ± 0.111 for α -pinene and 0.023 ± 0.005 for β -caryophyllene,
491 respectively (Kleindienst et al., 2007). Through dividing the sum of tracer compounds
492 measured in these marine samples by f_{soc} , an estimate of the contribution of each SOA to the
493 total marine SOC concentration was determined and the results were presented in Table 2.

494 Biomass burning derived OC in the whole samples is in the range of $1.1\text{--}790 \text{ ngC m}^{-3}$ (89.6
495 ngC m^{-3}), with higher levels in the nighttime ($1.3\text{--}790 \text{ ngC m}^{-3}$, 103 ngC m^{-3}) compared to
496 those ($1.1\text{--}506 \text{ ngC m}^{-3}$, 75.5 ngC m^{-3}) in the daytime. Higher concentrations of biomass-
497 burning OC near the Asian continent than those over the remote oceans (Fig. 7) suggest that
498 continental biomass-burning tracers were possibly removed by dry and/or wet deposition of
499 airborne particles, photodegradation by free radicals in the atmosphere or other atmospheric
500 dilution mechanism during long-range transport to the western North Pacific. Fungal-spore-



501 derived OC for all samples ranged from 1.2–1840 ngC m⁻³ (180 ngC m⁻³), accounting for
502 0.03–19.8% (3.1%) of OC, higher than those of biomass-burning-derived OC (Table 2). The
503 nighttime fungal-spore-derived OC (1.2–1840 ngC m⁻³, 203 ngC m⁻³) is higher than those of
504 daytime ones (1.3–911 ngC m⁻³, 157 ngC m⁻³), possibly associated with intensified activities
505 of yeasts and fungi during nighttime (Fu et al., 2012). As we all known, evenings tend to have
506 higher moist content, which exerts a stronger influence on microbial activity than temperature
507 (Liang et al., 2003). Similar higher levels of fungal-spore-derived OC were also observed in
508 the aerosols collected in coastline waters, probably due to more intensified microbial
509 activities in the coastal regions as discussed previously. β -caryophyllene SOC (6.9–747 ngC
510 m⁻³, 126 ngC m⁻³, with 7.4–747 ngC m⁻³, mean 153 ngC m⁻³ during the day and 6.9–416 ngC
511 m⁻³, mean 100 ngC m⁻³ at night) is found to be dominant contributor among the measured
512 biogenic SOCs over the ECS (the average concentrations were 39.8 ngC m⁻³ and 50.2 ngC m⁻³
513 for isoprene SOC and α/β -pinene SOC, respectively). Moreover, β -caryophyllene SOC
514 account for 0.36–5.3% (2.4%) of OC, about 2–3 times as high as those of isoprene (0.13–
515 3.8%, 0.83%) and monoterpene SOC (0.08–3.5%, 0.98%). Therefore, an emission inventory
516 for sesquiterpenes in marine aerosols over the ECS will be helpful for further understanding
517 formation of biogenic SOA in this region. Zhu et al. (2016) also reported that the
518 sesquiterpene-derived SOC was more abundant than isoprene- and monoterpene-derived SOC
519 for the aerosols collected in Okinawa, Japan.

520 The higher levels of sesquiterpene-SOC than monoterpene- and isoprene-SOC may be due to
521 the differences in the gas/particle partitioning of oxidation products from different VOCs,
522 given that longer chain sesquiterpenes (C₁₅H₂₄) have more carbon atoms than monoterpenes
523 (C₁₀H₁₆) and isoprene (C₅H₈), decreasing the vapour pressures of their oxidation products (Fu
524 et al., 2016). The levels of SOCs stemmed from isoprene, monoterpenes and sesquiterpene in
525 the marine aerosols over the ECS were much lower than those of PM_{2.5} samples in Hong
526 Kong, China during summer (Hu et al., 2008) and those observed in Mt. Tai aerosols (Fu et
527 al., 2010). The sum of biogenic SOCs over the ECS is 11.3–1060 ngC m⁻³ (216 ngC m⁻³),
528 much lower than that of Mt. Fuji aerosols (227–1120 ngC m⁻³, 542 ngC m⁻³) (Fu et al., 2014)
529 and those in the Mt. Tai aerosols (420–3100 ngC m⁻³) (Fu et al., 2010), but higher than those
530 observed in marine aerosols collected during a round-the-world cruise covering most remote
531 oceans (Fu et al., 2011) and those reported in the Arctic aerosols (average 14.6 ngC m⁻³) (Fu
532 et al., 2009). Such difference between terrestrial aerosols and marine aerosols highlights the
533 outflow of continental aerosols again. In terms of spatial distributions, biogenic SOCs



534 calculated from these compounds showed higher loadings in the locations close to the
535 coastline or significantly influenced by terrestrial air (Fig. 8a-b), also validating strong
536 influence of continental origin. Generally, the total concentrations of biogenic SOC (216 ngC
537 m^{-3}) are higher than those of biomass-burning-derived OC (89.6 ngC m^{-3}) and fungal-spore-
538 derived OC (180 ngC m^{-3}) (Table 2), presenting greater contribution from biogenic SOA to
539 the marine aerosols, in agreement with the report by Fu et al. (2014).

540 In order to compare the relative contribution of marine and continental sources to total OC in
541 the oceanic atmosphere, the whole sampling area was divided into five regions from south to
542 north according to the spatial distribution of aerosol samples, i.e. northern waters of ECS,
543 nearby waters in the ECS, seas adjacent to Fujian and Zhejiang provinces, eastern waters of
544 Taiwan. Five-day HYSPLIT back trajectory analysis showed that the atmosphere over waters
545 north of ECS were mainly controlled by marine air masses, while aerosol samples achieved
546 on June 12 showed great influence from South Korea and North China as well (Fig. S2).
547 Aerosol samples in waters of ECS were basically under the control of marine air masses,
548 except for the samples collected on May 20, which were also affected by terrestrial air from
549 the Asian mainland (Fig. S3).

550 In general, aerosols collected off the eastern coast of Taiwan Island were affected by air
551 masses from the remote sea, but some aerosols collected on May 22–23 were also influenced
552 by air masses coming from Southeast Asia (Fig. S4). In contrast, air masses from Asian
553 mainland had substantial impacts on the samples collected near Fujian and Zhejiang province
554 (Fig. S5-S6). On the whole, the aerosols strongly affected by terrestrial sources (e.g. Asian
555 mainland, Southeast Asia and South Korea) tend to own higher levels of sugars and SOA
556 tracers (Fig. 2 and Fig. 4). The contributions of biomass-burning-derived OC, fungal-spore-
557 derived OC, and biogenic SOCs to OC (%) in these five sampling regions were presented in
558 Fig. 9. Generally, the average contributions of biogenic SOCs, biomass-burning OC and
559 fungal-spore OC to OC near Zhejiang and Fujian waters were higher than the other sampling
560 areas, especially for the fungal-spores-OC (7.5 ± 6.8 for Zhejiang waters, 4.0 ± 2.8 for Fujian
561 waters, 2.6 ± 3.2 for the northern waters of ECS, 1.9 ± 1.7 for eastern waters of Taiwan, and
562 1.5 ± 1.4 for ECS, respectively) and biomass-burning-OC (2.8 ± 2.6 for Zhejiang waters,
563 2.7 ± 1.8 for Fujian waters, 1.3 ± 0.92 for the northern waters of ECS, 0.52 ± 0.36 for eastern
564 waters of Taiwan, and 0.96 ± 1.7 for ECS, respectively). Such spatial variations were closely
565 associated with different contributions of land and marine sources to the oceanic atmosphere.



566 Figures S2-S6 display a strong influence from land air masses in waters around Zhejiang and
567 Fujian provinces, while the atmosphere over northern waters of ECS, ECS and eastern waters
568 of Taiwan Island basically came under the influence of relatively clean marine air.
569 Interestingly, the average percentage of isoprene SOC in the eastern waters of Taiwan was
570 slightly larger compared to other regions. In light of back trajectory analysis, aerosols in this
571 region were mainly affected by terrestrial photosynthetic vegetation (e.g. trees and plants) in
572 Southeast Asia and/or marine biota (e.g. phytoplankton, seaweed and bacteria) (Fig. S4).
573 Sesquiterpene-derived SOC was found to be the most abundant SOC species in all five areas
574 (2.9 ± 0.87 for Zhejiang waters, 3.1 ± 1.7 for Fujian waters, 2.1 ± 0.80 for the northern waters of
575 ECS, 2.2 ± 1.2 for eastern waters of Taiwan, and 2.2 ± 1.3 for ECS, respectively) in comparison
576 with other SOCs derived from isoprene and monoterpene. On the other hand, the nighttime
577 contributions of biomass-burning OC, fungal-spores OC and biogenic SOCs to OC were
578 commonly greater than the daytime ones in all five regions (Fig. S9). For instance, the
579 contributions of biomass-burning-OC and fungal-spores-OC to OC during nighttime in the
580 seas near Zhejiang (3.9 ± 3.5 and 9.3 ± 8.0 , respectively) are significantly greater than those in
581 the daytime (1.9 ± 1.2 and 6.0 ± 6.3 , respectively). Such enhanced contributions in the evening
582 were likely to be in connection with intensified emissions, decreased height of PBL and land-
583 sea breeze circulations. The downward movement of PBL does not facilitate dispersion of
584 pollutants and lead to increases in aerosol concentration in the lower PBL (Li et al., 2017).
585 The prevailing land breeze in the nighttime in coastal areas could bring plentiful terrestrial
586 particles to the clean marine atmosphere. Such difference between daytime and nighttime
587 contributions illustrates that land-sea breeze circulation and PBL can be another important
588 factor influencing organics in marine aerosols.

589 **3.4 Source apportionment by PMF**

590 **3.4.1 Analysis of source profiles**

591 After testing runs with different number of factors (5–9), eight factors were chosen on basis of
592 the minimum value of Q (goodness of fit parameters) and probable source profile expected
593 from the study region.

594 Figure 10 illustrates the first source had high loadings of Na^+ (90.8%), suggesting a
595 contribution from sea salt. This profile also contained a significant amount of SO_4^{2-} , which



596 can react with sea-salt particles and release HCl gas, leaving lower Cl^-/Na^+ ratio (0.1) than
597 that of sea water (1.8) (Boreddy et al., 2014).

598 The second source shows high loadings of levoglucosan (66.3%), arabitol (55.0%), mannitol
599 (51.1%) and trehalose (50.4%), representing mixed sources of biomass burning and fungal
600 spores (Simoneit et al., 1999; Medeiros et al., 2006a; Bauer et al., 2008). Yang et al. (2012)
601 found an enhanced abundance of fungal tracers on account of biomass burning activities,
602 during which large numbers of fungi could be dispersed into the surrounding atmosphere or
603 be carried upward with the warm plume to other fields via long-distance atmospheric
604 transport.

605 The third source can be interpreted as crustal dust because of high loadings of Mg^{2+} (66.1%)
606 and Ca^{2+} (64.2%), characteristic elements of soil/crustal dust (Xu et al., 2016). This source
607 may include airborne road dust, construction dust and windblown soil particles, which are
608 derived primarily from terrestrial source. This source also contains large amount of sulfate,
609 which can react with crustal aerosols forming metal sulphates such as CaSO_4 and MgSO_4 .

610 The fourth factor is characterized by dominance of NH_4^+ (58.4%) and SO_4^{2-} (49.7%), which
611 can be classified as secondary ammonium sulfate. The molar ratio of NH_4^+ and SO_4^{2-} was 2.9
612 in this profile, suggesting that $(\text{NH}_4)_2\text{SO}_4$ was the dominated sulfate form in the marine
613 atmosphere over the ECS. Due to its long lifetime in the atmosphere, terrestrial SO_4^{2-} could
614 be transported long distance to coastal areas and even to the remote sea (Itahashi et al., 2017),
615 affecting chemical composition of oceanic atmosphere.

616 The fifth factor exhibits high loadings of di-isobutyl (DiBP) and di-n-butyl (DnBP), dominant
617 species of phthalates in the marine aerosols over the ECS, which is assigned to plastic
618 materials' emission likely from industry, agriculture and domestic application in coastal
619 regions.

620 The sixth factor has high loadings from sucrose (85.3%) followed by fructose (50.9%),
621 implies a significant emission from airborne pollen grains to the marine atmosphere over the
622 ECS during late spring to early summer (Fu et al., 2012).

623 The seventh profile presents high loading of Cl, likely to be associated with coal combustion,
624 which provides significant releases of chlorine (McCulloch et al., 1999; Sun et al., 2013).
625 Another confirmation is high concentration of SO_4^{2-} from this profile, since coal consumption
626 can produce mass sulfate as well.



627 The eighth factor illustrates high loadings of NO_3^- (67.5%) and biogenic SOA tracers,
628 especially the monoterpene SOA tracers, i.e. PA (67.5%) and MBTCA (65.2%). This factor
629 could be attributed to the photochemical oxidation products stemmed from emission of
630 vehicle and biogenic VOCs. Formation of secondary nitrate depends on NO_x , which is mainly
631 produced from power plants and mobile sources (Kim et al., 2006; Heo et al., 2009).
632 Anthropogenic NO_x could also enhance biogenic SOA formation via nitrate radical oxidation
633 of monoterpenes (Xu et al., 2015). The positive correlations between NO_x and isoprene-
634 derived SOA tracers as discussed before in our study also suggest NO_3^- and BSOA may share
635 common formation pathways. Previous study once pointed out that nitrogen-containing
636 species act a pivotal part in the formation and fate of SOA through varying radical and
637 oxidant regimes and particle properties, such as volatility and hygroscopicity (Chen et al.,
638 2017), agreeing well with our results.

639 Overall, the eight sources based on PMF were sea salt, biomass burning and fungal spores,
640 crustal dust, secondary sulfate ammonia, plastic emission, pollen grains, coal combustion, and
641 secondary nitrate and BSOA, which contributed to the TSP over the ECS of 16.9%, 3.5%,
642 8.0%, 28.5%, 6.3%, 2.0%, 14.3% and 20.4% in average, respectively (Fig. 11). The results of
643 PMF present that secondary origin and marine natural emissions could be the main sources
644 for the aerosols over the ECS.

645 **3.4.2 Temporal and spatial variation in sources**

646 Figure 12 shows the temporal and spatial variation in each profile during the day and the night.
647 The contribution of each source changes over time and varies with distance from the continent.
648 In general, higher levels contributed by biomass burning, fungal spores, crustal dust, pollen
649 grains, coal combustion, secondary nitrate and BSOA, were basically observed in coastal
650 aerosols and/or terrestrially influenced aerosols, suggesting strong influence of continental air
651 from East Asia and Southeast Asia in light of the back trajectories and wind directions during
652 the sampling periods (Fig. S2-S6 and S10). However, the contributions from sea salt,
653 indicative of oceanic emission, tend to be higher in the aerosols mainly affected by marine air
654 masses. Our study demonstrates that primary and secondary OM of terrestrial origin play an
655 important role in the marine aerosol chemistry over the western North Pacific through long-
656 range atmospheric transport in addition to natural emission of ocean.

657



658 **4 Conclusions**

659 In summary, atmospheric concentration, spatial distribution and source appointment of sugars
660 and biogenic SOA tracers were studied for the coastal and remote marine aerosols. Higher
661 concentrations of sugars and BSOA tracers were observed in the atmosphere around coastal
662 waters and/or in the terrestrially influenced regions in comparison with the remote oceans,
663 suggesting that continent origin contributed a lot to the abundance of sugars and BSOA
664 tracers in the marine atmosphere. Glucose was the dominant sugar species, followed by
665 mannitol among the total identified sugar compounds. Biogenic SOC were characterized by a
666 predominance of β -caryophyllene oxidation products in comparison with isoprene and α/β -
667 pinene tracers. The contributions of biomass-burning-derived OC, fungal-spore-derived OC,
668 and biogenic SOC to OC (%) were greater in the marine aerosols affected by land air masses.
669 The results of PMF present that secondary origin and marine natural emissions may be the
670 main sources for the aerosols over the ECS. Our study demonstrates that both primary and
671 secondary organic aerosols of terrestrial origin have great influences on the marine aerosol
672 chemistry over the western North Pacific through long-range atmospheric transport.

673

674 **Data availability.** The data for this paper are available upon request from the corresponding
675 author (fupingqing@tju.edu.cn).

676

677 **Competing interests.** The authors declare that they have no conflict of interest.

678

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683 **References**

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1072 **Table 1.** Concentrations (ng m^{-3}) of saccharides and biogenic SOA tracers measured in the
1073 marine aerosols collected over the East China Sea during May to June 2014.

Compounds	Whole period (<i>n</i> = 51)		Daytime (<i>n</i> = 25)		Nighttime (<i>n</i> = 26)	
	Mean	Range	Mean	Range	Mean	Range
<i>Anhydrosugars</i>						
Galactosan	0.59	0.02–3.9	0.56	0.04–3.6	0.62	0.02–3.9
Mannosan	1.4	0.06–10.2	1.2	0.08–5.2	1.6	0.06–10.2
Levogluosan	7.3	0.09–64.3	6.1	0.09–41.2	8.4	0.10–64.3
<i>Sugar Alcohols</i>						
Erythritol	1.3	n.d. ^a –16.9	0.87	0.02–4.60	1.6	n.d.–16.9
Arabitol	5.1	0.02–51.0	4.5	0.06–26.7	5.7	0.02–51.0
Mannitol	16.3	0.03–169	14.1	0.08–81.3	18.4	0.03–169
Inositol	0.57	0.01–7.7	0.47	0.01–2.8	0.67	0.01–7.7
<i>Sugars</i>						
Fructose	9.1	0.09–106	8.1	0.09–56.5	10.2	0.32–106
Glucose	18.8	0.31–209	17.8	0.31–106	19.7	0.47–209
Sucrose	11.7	0.09–216	9.7	0.14–136	13.5	0.09–216
Trehalose	9.4	0.08–96.0	9.1	0.08–60.7	9.7	0.10–96.0
Subtotal	81.5	1.8–950	72.6	1.0–525	90.1	1.8–950
<i>Isoprene SOA tracers</i>						
2-methylthreitol	1.5	0.01–12.7	1.2	0.04–5.2	1.9	0.01–12.7
2-methylerythritol	3.2	0.02–29.1	2.5	0.08–11.8	3.9	0.02–29.1
Sum of 2-methyltetrols	4.8	0.03–41.9	3.8	0.11–17.0	5.7	0.03–41.9
2-Methylglyceric acid (2-MGA)	1.4	0.09–8.3	1.2	0.20–3.6	1.6	0.09–8.33
<i>C</i> ₅ -Alkene triols ^b	2.2	0.03–14.6	2.1	0.06–14.6	2.3	0.03–13.8
Subtotal	8.4	0.15–64.0	7.1	0.42–35.2	9.6	0.15–64.0
<i>α/β-pinene (monoterpene) SOA tracers</i>						
3-Hydroxyglutaric acid (3-HGA)	2.2	0.03–14.1	2.3	0.08–14.1	2.1	0.03–7.7
Pinonic acid (PNA)	0.47	0.02–1.6	0.49	0.06–1.6	0.46	0.02–1.6
Pinic acid (PA)	3.4	0.16–14.9	3.4	0.28–14.9	3.3	0.16–13.6
MBTCA	5.6	n.d.–56.9	6.8	n.d.–56.9	4.5	n.d.–24.5
Subtotal	11.6	0.26–87.2	13.0	0.45–87.2	10.3	0.26–43.7
<i>β-caryophyllene (sesquiterpene) SOA tracers</i>						
β -Caryophyllinic acid	2.9	0.16–17.2	3.5	0.17–17.2	2.3	0.16–9.6
Total measured tracers	22.9	1.1–135	23.6	1.4–135	22.2	1.1–115

^a n.d. denotes not detected. ^b *C*₅-alkene triols: cis-2-methyl-1,3,4-trihydroxy-1-butene, 3-methyl-2,3,4-trihydroxy-1-butene and trans-2-methyl-1,3,4-trihydroxy-1-butene.

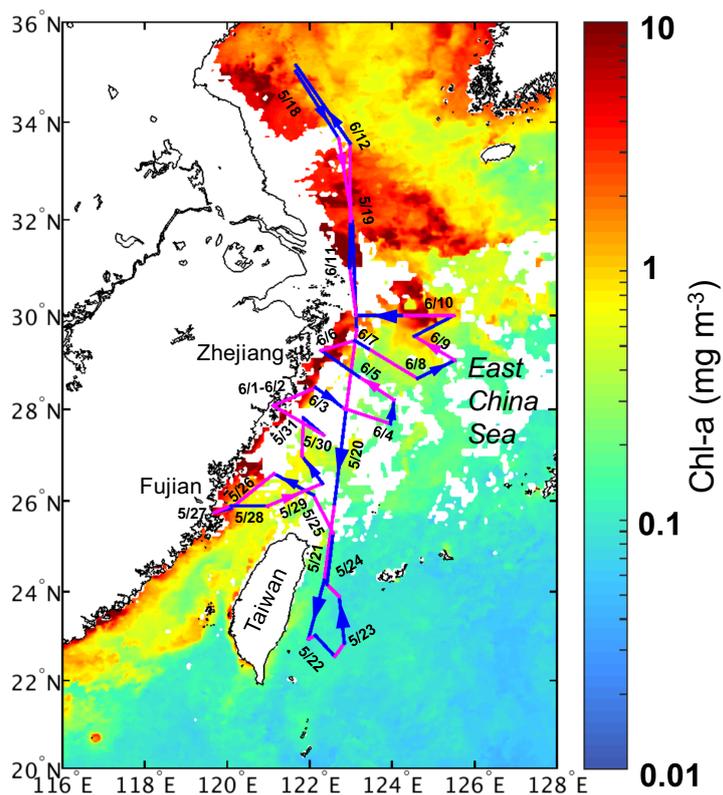
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1075 **Table 2.** Concentrations of organic carbon (OC) (ngC m^{-3}) from biogenic primary emission
 1076 (biomass-burning OC and fungal-spore OC) and biogenic SOC and their contributions in
 1077 aerosol OC (%) in marine aerosols over the East China Sea.

Component	Total			Daytime			Nighttime		
	Range	Mean	std	Range	Mean	std	Range	Mean	std
Concentration (ngC m^{-3})									
Aerosol OC	424–14100	4260	3480	1030–14100	4940	3800	424–9830	3600	3080
Biomass burning OC	1.1–790	89.6	156	1.1–506	75.5	117	1.3–790	103	187
Fungal spore OC	1.2–1840	180	334	1.3–911	157	249	1.2–1840	203	403
Isoprene SOC ^a	0.78–324	39.8	55.4	2.3–133	32.1	37.6	0.78–324	47.2	68.3
Monoterpene SOC	1.1–377	50.2	72.0	1.9–377	56.1	84.3	1.1–189	44.6	58.9
Sesquiterpene SOC	6.9–747	126	153	7.4–747	153	183	6.9–416	100	114
Sum of biogenic SOC	11.3–1060	216	250	13.7–1060	241	279	11.3–824	192	222
Subtotal	16.1–3460	486	688	29.8–2480	473	600	16.1–3460	498	776
Percentage in aerosol OC (%)									
Biomass burning OC	0.05–8.5	1.5	1.8	0.05–4.4	1.1	1.2	0.13–8.5	1.8	2.3
Fungal spore OC	0.03–19.8	3.1	4.0	0.03–16.6	2.6	3.7	0.18–19.7	3.5	4.2
Isoprene SOC	0.13–3.8	0.83	0.87	0.13–2.0	0.60	0.54	0.14–3.8	1.0	1.1
Monoterpene SOC	0.08–3.5	0.98	0.90	0.10–3.3	1.00	0.95	0.08–3.5	0.96	0.87
Sesquiterpene SOC	0.36–5.3	2.4	1.2	0.36–5.3	2.5	1.4	0.87–5.1	2.4	1.1
Sum of biogenic SOC	0.67–9.3	4.2	2.2	0.67–9.3	4.1	2.2	1.8–8.8	4.4	2.2
Subtotal	1.5–37.0	8.7	6.9	1.5–27.4	7.8	6.3	2.8–37.0	9.7	7.4

^a The total mass concentrations of SOC produced by isoprene (2-methylglyceric acid and 2-methyltetrols were used), α / β -pinene, and β -caryophyllene were estimated using the tracer-based method by Kleindienst et al. (2007).



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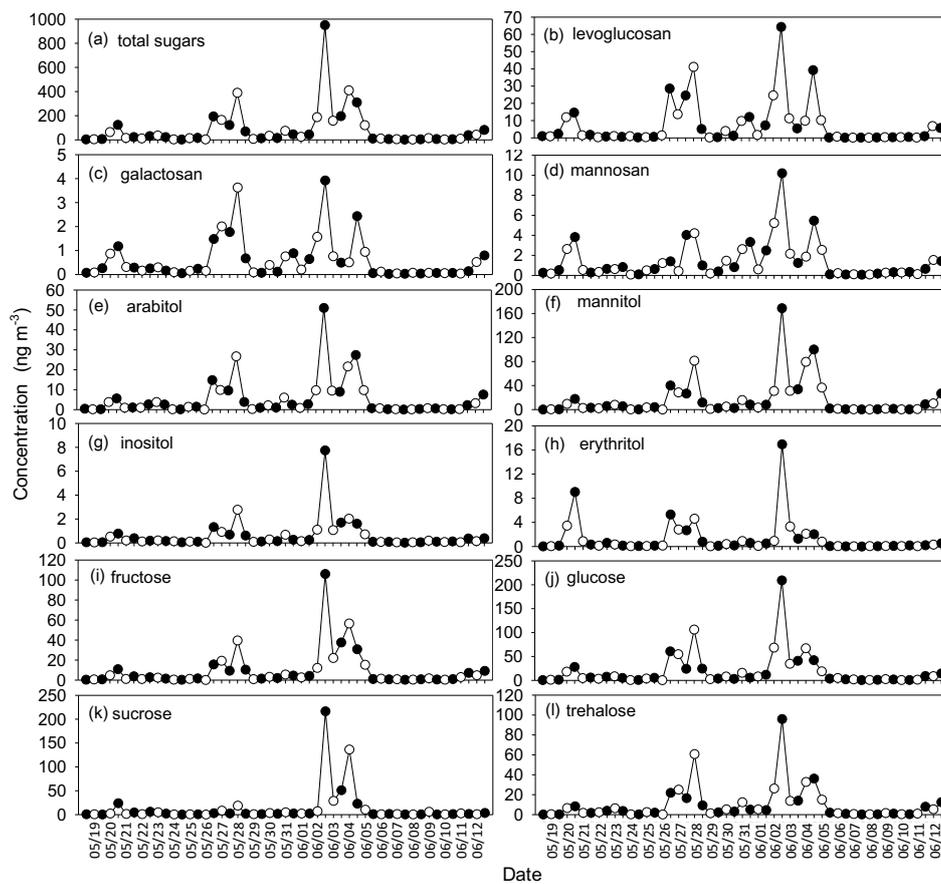
1080 **Figure 1.** Cruise tracks of KEXUE-1 and spatial distribution of satellite-derived chlorophyll-a

1081 concentrations (chl-a , mg m^{-3}) in surface seawater derived from MODIS L3 products during

1082 the sampling period in the East China Sea. The purple and blue lines represent daytime and

1083 nighttime aerosol samples, respectively.

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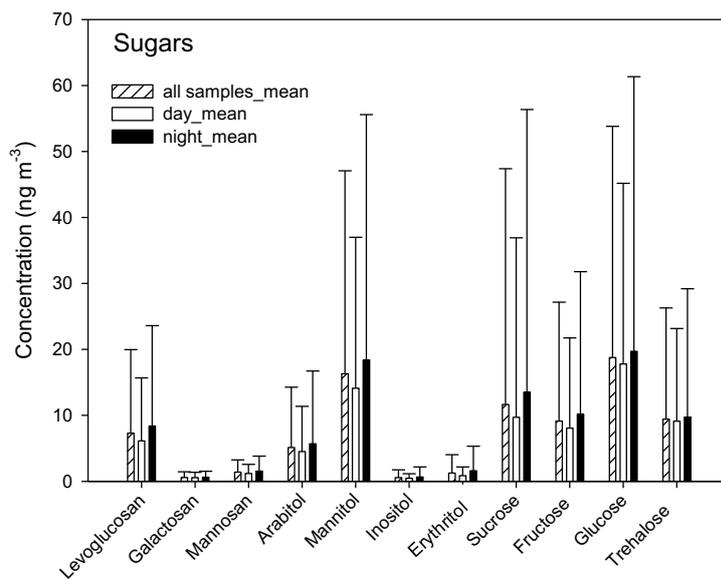
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1087 **Figure 2.** Temporal variations in sugar compounds in marine aerosols collected over the East

1088 China Sea. The open and shaded circles represent daytime and nighttime values, respectively.

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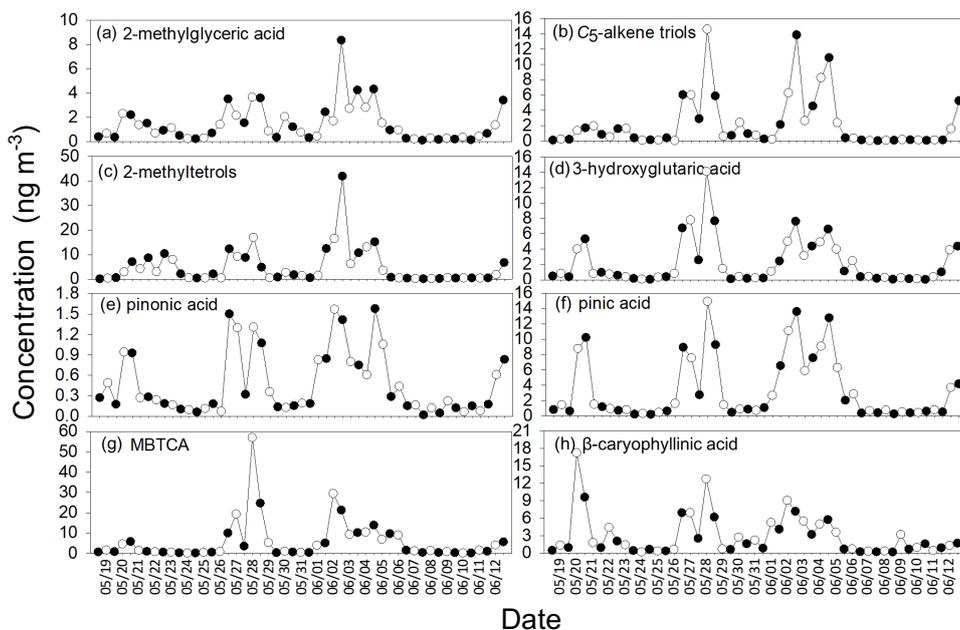
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1092 **Figure 3.** Average concentrations of sugars detected in marine aerosols over the East China

1093 Sea.

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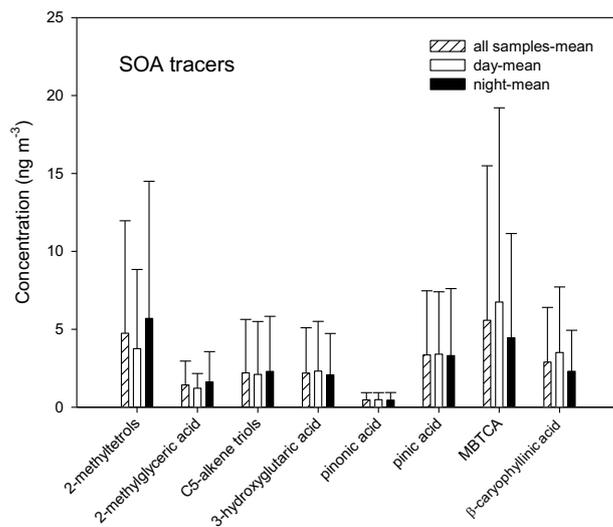
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1097 **Figure 4.** Temporal variations in biogenic SOA tracers detected in marine aerosols over the

1098 East China Sea during May to June 2014. The open and shaded circles represent daytime and

1099 nighttime values, respectively.

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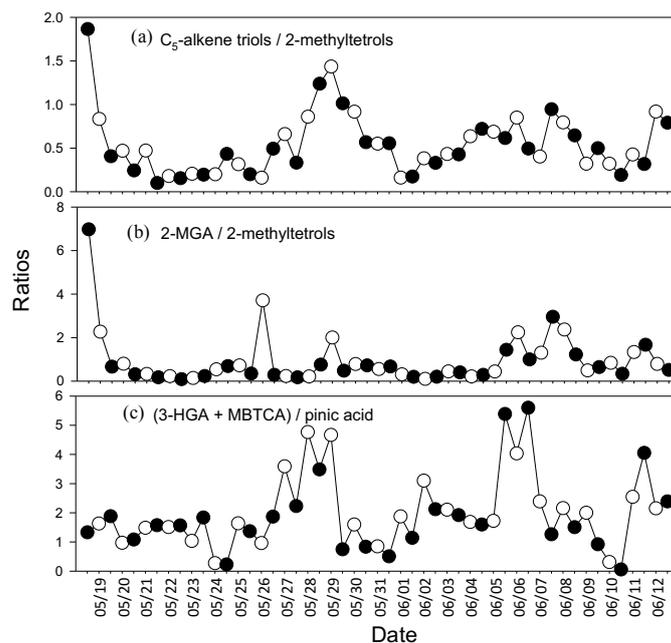
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1103 **Figure 5.** Average concentrations of SOA tracers detected in marine aerosols over the East

1104 China Sea.

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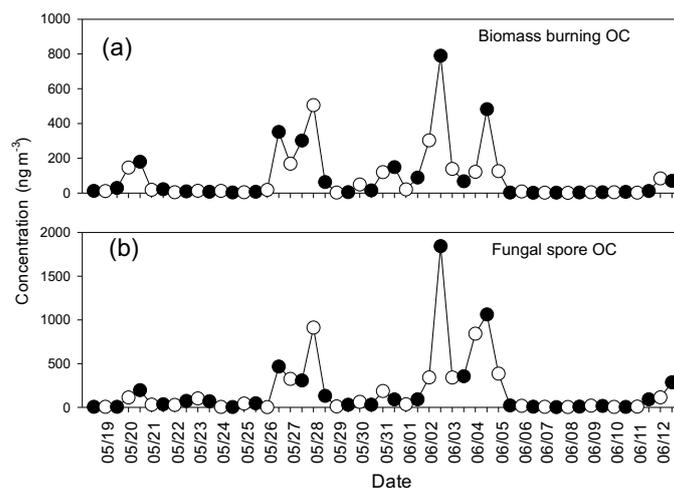
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1108 **Figure 6.** Temporal variations in the concentration ratios of isoprene and α/β -pinene

1109 oxidation products in the marine aerosols over the East China Sea. The open and shaded

1110 circles represent daytime and nighttime samples, respectively.

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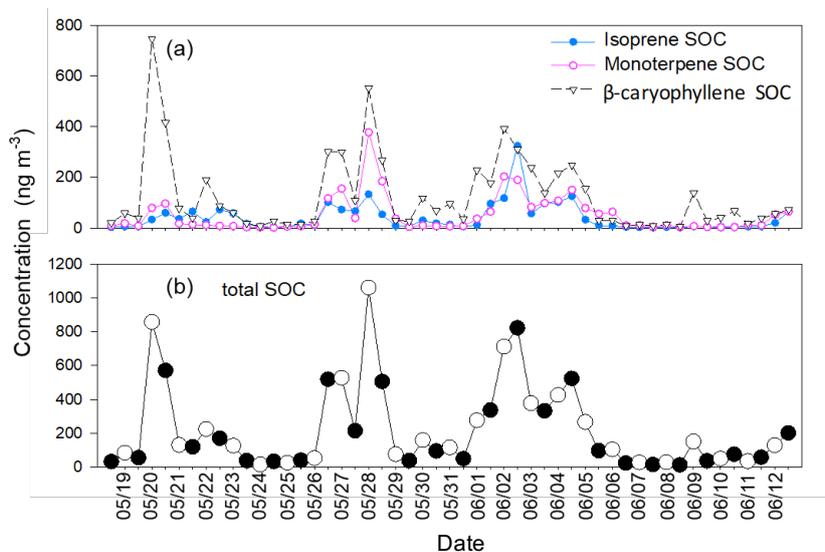


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1114 **Figure 7.** Temporal variations in (a) biomass-burning-derived OC, and (b) fungal-spore-
1115 derived OC in the marine aerosols over the East China Sea. The open and shaded circles
1116 represent daytime and nighttime samples, respectively.

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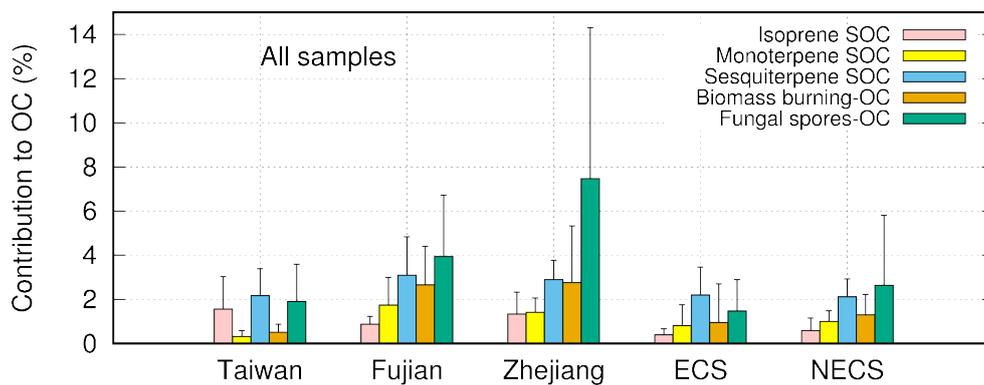
1120 **Figure 8.** Temporal variations in (a) secondary organic carbon (SOC) derived from isoprene,

1121 α/β -pinene and sesquiterpene and (b) the total SOC levels in the marine aerosols over the East

1122 China Sea during May to June 2014. The open and shaded circles represent the daytime and

1123 nighttime samples, respectively.

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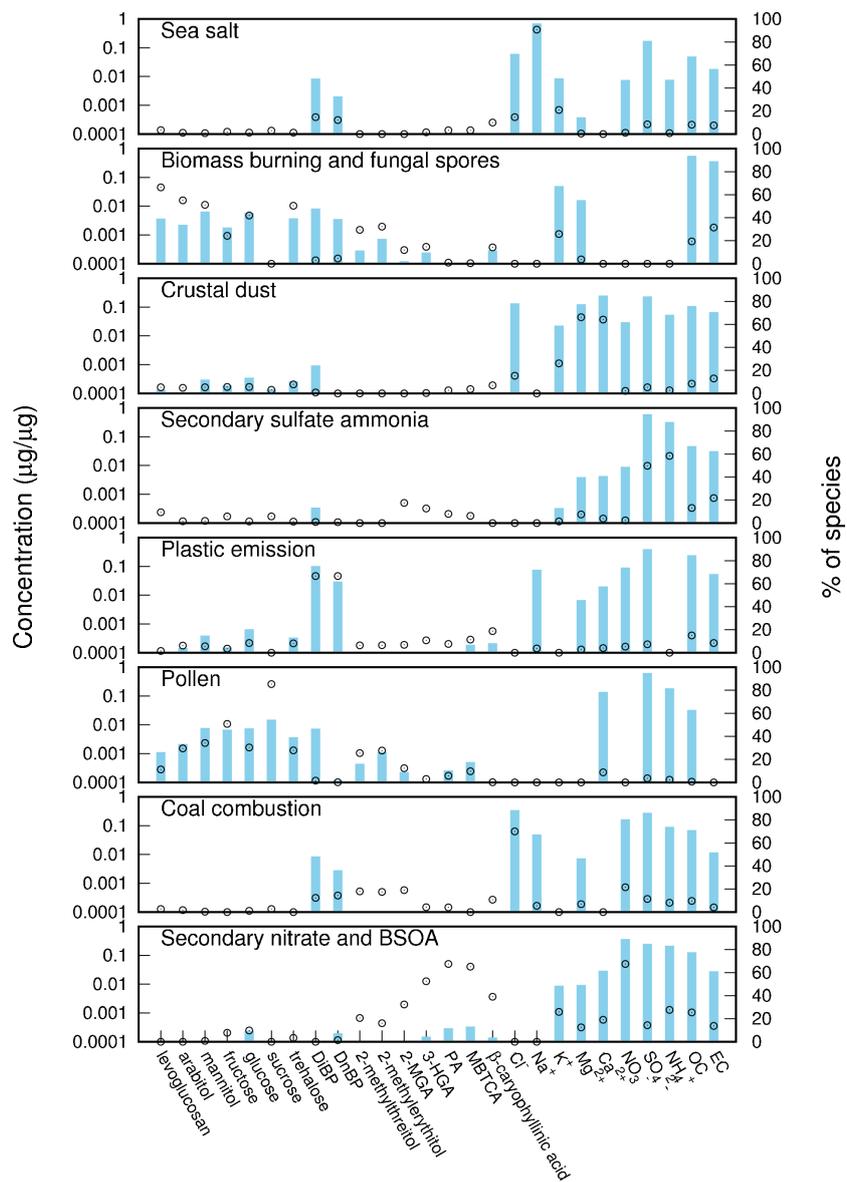
1127 **Figure 9.** Contributions of OC (ngC m^{-3}) from biogenic primary emission (biomass-burning

1128 OC and fungal-spore OC) and biogenic SOC to OC (%) in different sampling regions. Taiwan,

1129 Fujian and Zhejiang refer to the waters around Taiwan, Fujian and Zhejiang; ECS represents

1130 East China Sea waters; NECS represents northern waters of ECS.

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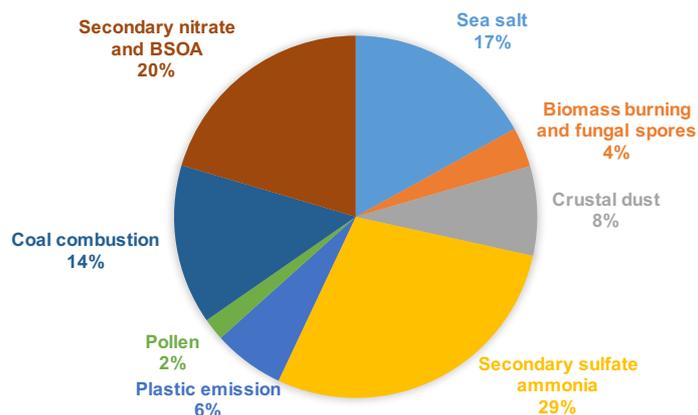


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1134 **Figure 10.** Source profiles identified by PMF. Blue bars represent the mass contribution with
 1135 y-axis on the left, while black dots stand for the percentage of species to the sum with y-axis
 1136 on the right.

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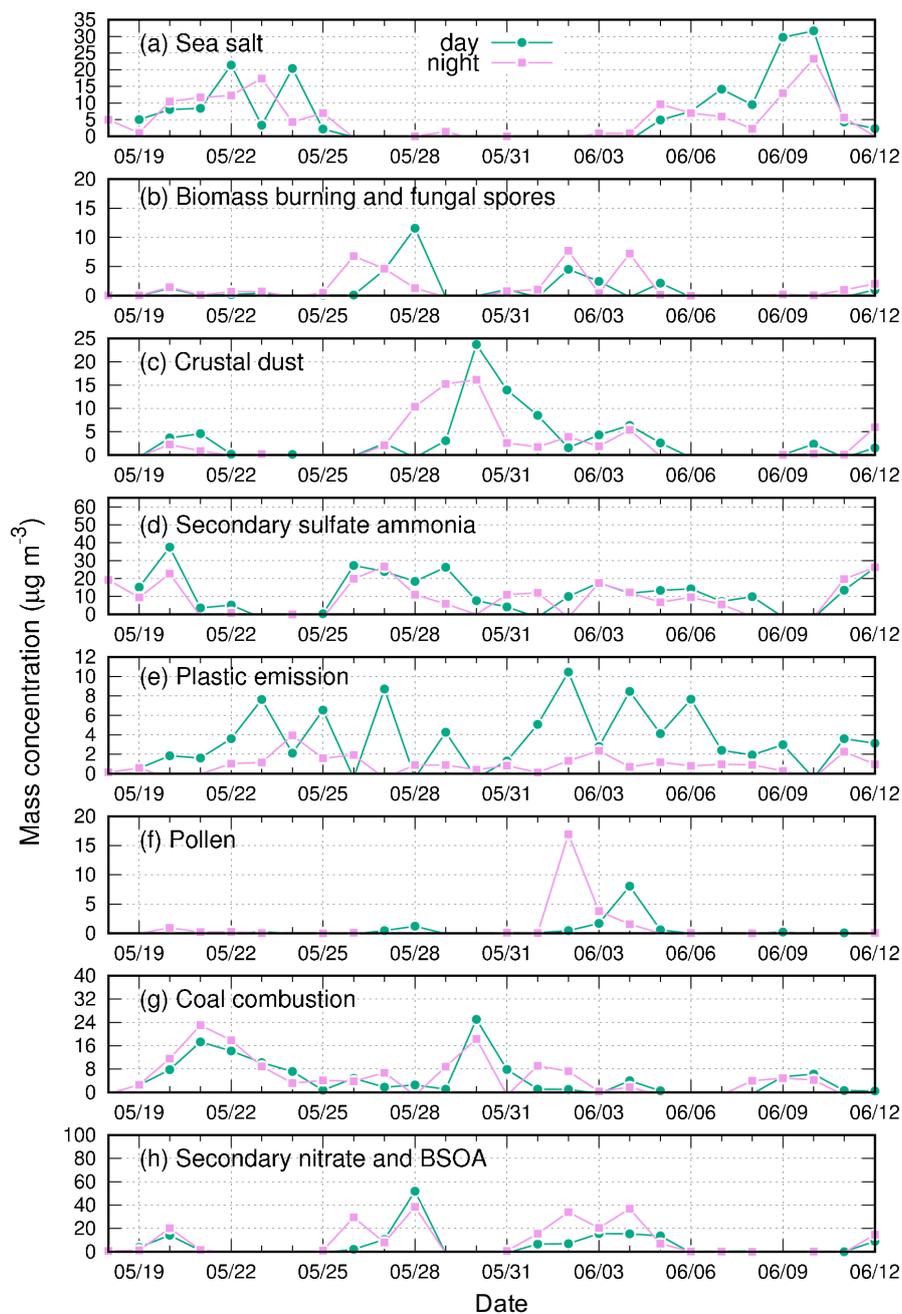
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1140 **Figure 11.** Contributions of different sources to marine organic aerosols over the East China

1141 Sea.

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1145 **Figure 12.** Temporal variation in sources contributed to marine aerosols ($\mu\text{g m}^{-3}$).