Author' Response to Referees' Comments

Anonymous Referee #1

Yang et al. present VOC observational results at a rural site in the Beijing metropolitan area before, during, and after a strict emission control implementation for the APEC summit event. They mainly discussed on changes in the VOC composition during three different periods. Then, they move ahead to discuss the potential sources from the PFM analysis and the impacts of VOCs in the context of ozone and aerosol formation rates. Overall the manuscript is very clearly written, and the goals seem to be well achieved.

[1] However, I have a couple of concerns regarding whether this manuscript is well fit in this particular journal. Although the significance in understanding of Chinese pollution problems cannot be overstated considering the large population in China, this manuscript is too narrowly focused on Chinese local air pollution problems without discussion on the bigger context such as regional or global impacts.

Reply: Thanks for the insightful comments. Maybe we have not interpretated our findings properly and fully in the manuscript. We think the findings from our study are implicative at least in two aspects:

1) While exposure to indoor air pollution from burning solid fuels (biomass and coal) affects nearly half of the world's population, and household air pollution has been considered as a major environmental cause of death (Martin et al., 2011; Lim et al., 2012; Subramanian, 2014), a previous study (Liu et al., 2016) revealed that in the Beijing-Tianjian-Hebei (BTH) region, residential use of solid fuels might be a major and underappreciated ambient pollution source for PM_{2.5} (particularly BC and OC) during winter heating period based on the Multiresolution Emission Inventory of China (MEIC; www.meicmodel.org) for January and February 2010. Here we demonstrated that, based on our field monitoring, burning solid fuels (mainly coal) to heat homes would also be a major source of volatile organic compounds (VOCs) during winter in the region. As VOCs are important precursors of secondary organic aerosols and ozone, our study also suggests that cleaner residential energy use for cooking and heating not only benefit lowering indoor air pollution with great health benefits for the residents, but also benefit lowering the exposure to ambient air pollution for a wider range of people.

2) Due to high secondary aerosol contribution to particulate pollution during haze events (Huang et al., 2014), reducing emissions of VOCs as precursors of secondary organic aerosols is important for combating fine particle air pollution and heavy hazes. Enhancing the controls over emissions from vehicles and industry sector would be effective for reducing ambient VOCs, as demonstrated in our study during the APEC before the start of winter heating period. During the winter heating period, since residential coal/biomass burning was found to be a major source for ambient VOCs even in the Beijing metropolitan area, solely enhancing the emission control in the traffic and industry sectors would be not so effective as did in the non-heating period. This is a important message for regions, particularly less developed regions, to control emissions of VOCs to combat air pollution due to ozone and PM_{2.5}.

Therefore, although we conducted our study at a rural site in Beijing, the findings are not just locally significant but also have important implications for other regions. We have incorporated these aspects into our revised manuscript. We have rewrite our introduction, and modified our conclusions and abstract as well.

[2] Moreover, the scientifically relevant analyses such as ozone forming potential and secondary aerosol forming potential are just adapted from previous publications without detailed discussion whether the method is relevant to this particular photochemical environment.

Reply: Thanks. This comment is scientifically very important. It reminds us of remembering that the potentials are related to particular photochemical environments. As for the ozone formation potentials, the Maximum Incremental Reactivity (MIR) scale, originally developed by Carter (1994), has been widely used as a simplified approach to evaluate the relative ground-level ozone impacts of volatile organic compounds. The MIR scale in its nature represents conditions where ambient ozone is most sensitive to changes in VOC emissions, therefore the potentials based on the MIR scale are maximums that can hardly achieved under real atmospheric conditions, particularly depending on the relative availability of NOx (Dondge, 1984; Carter and Atkinson, 1989). However, for the convenience of regulating VOCs based on calculations of their relative ground-level ozone impacts, in the present study we adopted the metrics widely used for calculating OFPs in many previous studies, and therefore we remain the calculation in its present state but indicate in the revised manuscript that it is only a simplified approach.

Organic aerosol formation potentials are comparatively much more complicated. They are largely affected by factors such as the reactivity of the parent compound and volatility of the product species (Odum et al., 1997). The reactivity of the parent species can be directly measured by their reaction rate constants with oxidants. The oxidation products, however, are both numerous and difficult to quantify analytically. Therefore, the SOA yield (Y), defined as mass of SOA formed divided by mass of VOCs reacted, has been used as an indirect measure for a specific VOC species to indicate its ability to form SOA (Odum et al., 1997). This way the secondary organic aerosol formation potentials (SOAFPs) by a mixture of VOCs can be estimated as $\Sigma_i X_i \times Y_i$, where X_i is the mass concentration ($\mu g m^{-3}$) and Y_i (%) is the SOA yield of precursor i. SOA yield data have been obtained in controlled smog chamber studies. In this study, the SOA yields are taken from studies by Ng et al (2007), Lim and Ziemann (2009) and Loza et al (2014). As SOA formation depends on nitrogen oxides (NOx) (Ng et al., 2007), SOAFPs are typically calculated under low-NOx and high-NOx conditions, approximating the higher and lower limits, respectively. Although widely used in a lot of literatures, this kind of calculation is also a simplified approach to indicate SOA potentially formed if the observed VOCs are completely oxidized in the atmosphere.

In the present study, we put our focus mainly on how the control measures or human activities would impact the VOCs occurring in the ambient air, so we just followed the widely adopted approaches to indicate their ozone and SOA formation potentials, although they are simplified and even scientifically not solid enough.

[3] I recommend expanding discussion at least to compare VOC speciation from other locations in the region and other metropolitan areas in the world.

Reply: Thanks for the suggestion. In the revised manuscript, we have added the comparison as beolw:

"Table S1 shows a comparison of VOCs from our study with those observed at other metropolitan areas in the world. Mixing ratios of VOCs from this study at a rural site in Beijing during period I (23.41 ppb) and period III (21.71 ppbv) were comparable to that in urban Shanghai from January 2007 to March 2010 (Cai et al., 2010), but lower than those in Beijing during June 2008 (Wang et al., 2010), Guangzhou from June 2011 to May 2012 (Zou et al., 2015), Lille, French from May 1997 to April 1999 (Borbon et al., 2002) and Houston in August-September 2006 (Leuchner and

Rappengluck, 2010). Average mixing ratios of VOCs during period II (11.25 ppbv) with enhanced emission control in the present study were significantly lower than those reported in other metropolitan areas. As for the most abundant VOC species including ethane, propane, ethylene, benzene, toluene and ethyne, the mixing ratios of ethane and ethylene at UCAS were similar to that at Beijing during June 2008 (Wang et al., 2010) and urban Guangzhou from June 2011 to May 2012 (Zou et al., 2015), but significantly lower than that in urban Beijing during 2014 APEC (Li et al., 2015). Mixing ratios of propane in present study are comparable with that in Hong Kong from September 2002 to August 2003 (Guo et al., 2007) and Lille, French from May 1997 to April 1999 (Borbon et al., 2002), but factors of 2-3 lower than that reported in urban Shanghai from January 2007 to March 2010 (Cai et al., 2010) and Guangzhou from June 2011 to May 2012 (Zou et al., 2015). Mixing ratios of benzene and toluene in Lille, French from May 1997 to April 1999 (Borbon et al., 2002) were over 2 times higher than that in present study. Mixing ratios of ethylene, benzene and toluene in present study were comparable to those observed in Houston during August-September 2006 (Leuchner and Rappengluck, 2010), while ethyne, a tracer of incomplete combustion, had mixing ratios 3-4 times higher than that in Houston."

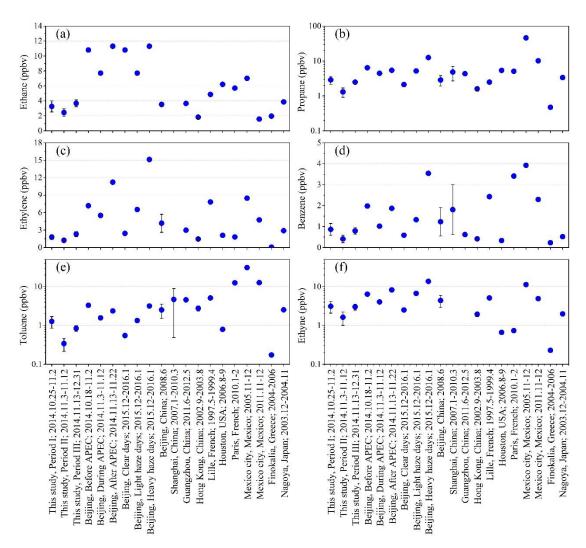


Figure S1. Comparison of (a) ethane, (b) propane, (c) ethylene, (d) benzene, (e) toluene and (f) ethyne observed at UCAS with those from other studies at metropolitan areas in the world.

[4] In addition, I would recommend more thorough descriptions on ozone formation potential and SOA formation potential particularly in the context of whether the metrics are relevant in this photochemical environment.

Reply: As responding to comment [2] above, we fully agree that we should consider if the metrics are relevant in a specific photochemical environment. However, in the present study since we put our focus mainly on how the control measures or human activities would impact the VOCs occurring in the ambient air, so we just followed the widely adopted approaches to indicate their ozone and SOA formation potentials, although they are over simplified and even scientifically not solid enough.

References:

Baudic, A., Gros, V., Sauvage, S., Locoge, N., Sanchez, O., Sarda-Estève, R., Kalogridis, C., Petit,

- J. E., Bonnaire, N., and Baisnée, D.: Seasonal variability and source apportionment of volatile organic compounds (VOCs) in the Paris megacity (France), Atmos. Chem. Phys, 16, 11961-11988, https://doi.org/10.5194/acp-16-11961-2016, 2016.
- Borbon, A., Locoge, N., Veillerot, M., Galloo, J. C., and Guillermo, R.: Characterisation of NMHCs in a French urban atmosphere: overview of the main sources, Sci. Total Environ., 292, 177-191, https://doi.org/10.1016/S0048-9697(01)01106-8, 2002.
- Cai, C. J., Geng, F. H., Tie, X. X., Yu, Q., and An, J. L.: Characteristics and source apportionment of VOCs measured in Shanghai, China, Atmos. Environ., 44, 5005-5014, http://dx.doi.org/10.1016/j.atmosenv.2010.07.059, 2010.
- Carter, W. P. L., and Atkinson, R.: Computer modeling study of incremental hydrocarbon reactivity, Environ. Sci. Technol., 23, 864-880, http://doi.org/10.1021/es00065a017, 1989.
- Carter, W. P. L.: Development of Ozone Reactivity Scales for Volatile Organic Compounds, Air & Waste, 44, 881-899, http://doi.org/10.1080/1073161x.1994.10467290, 1994.
- Dodge, M. C.: Combined effects of organic reactivity and NMHC/NOx ratio on photochemical oxidant formation-a modeling study, Atmos. Environ., 18, 1657-1665, http://doi.org/10.1016/0004-6981(84)90388-3, 1984.
- Garzón, J. P., Huertas, J. I., Magaña, M., Huertas, M. E., Cárdenas, B., Watanabe, T., Maeda, T., Wakamatsu, S., and Blanco, S.: Volatile organic compounds in the atmosphere of Mexico City, Atmos. Environ., 119, 415-429, https://doi.org/10.1016/j.atmosenv.2015.08.014, 2015.
- Guo, H., So, K. L., Simpson, I. J., Barletta, B., Meinardi, S., and Blake, D. R.: C1–C8 volatile organic compounds in the atmosphere of Hong Kong: Overview of atmospheric processing and source apportionment, Atmos. Environ., 41, 1456-1472, http://dx.doi.org/10.1016/j.atmosenv.2006.10.011, 2007.
- Huang, R. J., Zhang, Y. L., Bozzetti, C., Ho, K. F., Cao, J. J., Han, Y. M., Daellenbach, K. R., Slowik,
 J. G., Platt, S. M., Canonaco, F., Zotter, P., Wolf, R., Pieber, S. M., Bruns, E. A., Crippa, M.,
 Ciarelli, G., Piazzalunga, A., Schwikowski, M., Abbaszade, G., Schnelle-Kreis, J.,
 Zimmermann, R., An, Z. S., Szidat, S., Baltensperger, U., El Haddad, I., and Prevot, A. S. H.:
 High secondary aerosol contribution to particulate pollution during haze events in China,
 Nature., 514, 218-222, http://doi.org/10.1038/nature13774, 2014.
- Leuchner, M., and Rappenglück, B.: VOC source-receptor relationships in Houston during

- TexAQS-II, Atmos. Environ., 44, 4056-4067, https://doi.org/10.1016/j.atmosenv.2009.02.029, 2010.
- Li, J., Xie, S. D., Zeng, L. M., Li, L. Y., Li, Y. Q., and Wu, R. R.: Characterization of ambient volatile organic compounds and their sources in Beijing, before, during, and after Asia-Pacific Economic Cooperation China 2014, Atmos. Chem. Phys., 15, 7945-7959, http://dx.doi.org/10.5194/acp-15-7945-2015, 2015.
- Liakakou, E., Bonsang, B., Williams, J., Kalivitis, N., Kanakidou, M., and Mihalopoulos, N.: C2–C8 NMHCs over the Eastern Mediterranean: Seasonal variation and impact on regional oxidation chemistry, Atmos. Environ., 43, 5611-5621, https://doi.org/10.1016/j.atmosenv.2009.07.067, 2009.
- Lim, S. S et al.: A comparative risk assessment of burden of disease and injury attributable to 67 risk factors and risk factor clusters in 21 regions, 1990-2010: a systematic analysis for the Global Burden of Disease Study 2010, Lancet, 380, 2224-2260, http://dx.doi.org/10.1016/s0140-6736(12)61766-8, 2012.
- Lim, Y. B., and Ziemann, P. J.: Effects of molecular structure on aerosol yields from OH radical-initiated reactions of linear, branched, and cyclic alkanes in the presence of NOx, Environ. Sci. Technol., 43, 2328-2334, http://dx.doi.org/10.1021/es803389s, 2009.
- Liu, C. T., Ma, Z. B., Mu, Y. J., Liu, J.F., Zhang, C.L., Zhang, Y. Y., Liu, P. F., and Zhang, H. X.: The levels, variation characteristics, and sources of atmospheric non-methane hydrocarbon compounds during wintertime in Beijing, China, Atmos. Chem. Phys, 17, 10633-10649, https://doi.org/10.5194/acp-17-10633-2017, 2017.
- Liu, J., Mauzerall, D. L., Chen, Q., Zhang, Q., Song, Y., Peng, W., Klimont, Z., Qiu, X. H., Zhang, S. Q., Hu, M., Lin, W. L., Smith, K. R., and Zhu, T.: Air pollutant emissions from Chinese households: A major and underappreciated ambient pollution source, P. Natl. Acad. Sci. USA., 113, 7756-7761, http://dx.doi.org/10.1073/pnas.1604537113, 2016.
- Loza, C.L., Craven, J.S., Yee, L.D., Coggon, M.M., Schwantes, R.H., Shiraiwa, M., Zhang, X., Schilling, K.A., Ng, N.L., Canagaratna, M.R., Ziemann, P.J., Flagan, R.C., Seinfeld, J.H.: Secondary organic aerosol yields of 12-carbon alkanes, Atmos. Chem. Phys., 14, 1423-1439, http://doi.org/10.5194/acp-14-1423-2014, 2014.
- Martin, W. J., Glass, R. I., Balbus, J. M., and Collins, F. S.: A Major Environmental Cause of Death,

- Science, 334, 180-181, http://dx.doi.org/10.1126/science.1213088, 2011.
- Ng, N. L., Kroll, J. H., Chan, A. W. H., Chhabra, P. S., Flagan, R. C., and Seinfeld, J. H.: Secondary organic aerosol formation from m-xylene, toluene, and benzene, Atmos. Chem. Phys., 7, 3909-3922, http://doi.org/10.5194/acp-7-3909-2007, 2007.
- Odum, J. R., Jungkamp, T. P. W., Griffin, R. J., Flagan, R. C., and Seinfeld, J. H.: The atmospheric aerosol-forming potential of whole gasoline vapor, Science., 276, 96-99, http://doi.org./10.1126/science.276.5309.96, 1997.
- Saito, S., Nagao, I., and Kanzawa, H.: Characteristics of ambient C2–C11 non-methane hydrocarbons in metropolitan Nagoya, Japan, Atmos. Environ., 43, 4384-4395, https://doi.org/10.1016/j.atmosenv.2009.04.031, 2009.
- Subramanian, M.: Deadly dinners, Nature, 509, 548-551, http://doi.org/10.1038/509548a, 2014.
- Wang, B., Shao, M., Lu, S. H., Yuan, B., Zhao, Y., Wang, M., Zhang, S. Q., and Wu, D.: Variation of ambient non-methane hydrocarbons in Beijing city in summer 2008, Atmos. Chem. Phys., 10, 5911-5923, http://doi.org/10.5194/acp-10-5911-2010, 2010.
- Zou, Y., Deng, X. J., Zhu, D., Gong, D. C., Wang, H., Li, F., Tan, H. B., Deng, T., Mai, B. R., Liu, X. T., and Wang, B. G.: Characteristics of 1 year of observational data of VOCs, NOx and O₃ at a suburban site in Guangzhou, China, Atmos. Chem. Phys., 15, 6625-6636, 10.5194/acp-15-6625-2015, 2015.

Author' Response to Referees' Comments

Anonymous Referee #3

Household air pollution from burning biomass and coal for cooking foods and heating rooms has long been a major environmental problem. Previous studies suggested that the uncontrolled and inefficient combustion of solid fuels for heating in winter also contributed substantially to outdoor PM_{2.5}, BC, OC, SO₂ and NOx in many regions, such as in the North China Plain (Liu et al., 2016, PNAS). The field observations at the rural site of Beijing in this study further demonstrated that the combustion of solid fuels for heating in winter made remarkable contribution to ambient volatile organic compounds (VOCs). The authors also took advantage of the temporary intervention measures for emission control during the APEC to evaluate the actual effect of the control measures on the ambient VOCs levels through reduction of the source contributions. The comparison between the heating and non-heating periods offered robust results indicating the influential emission from winter heating. Overall, the data quality of this manuscript is quite good, and the interpretation of the results is appropriate and convincing. Therefore, I strongly recommend publication of this manuscript.

Reply: Thanks for the comments. We have revised our manuscript with your constructive comments and suggestions as below.

Majors:

[1] As stated above, an important finding of this study is that residential coal burning, especially during winter heating period, could be a major contributor to the ambient VOCs. Apart from the emission factors available for residential coal burning, is it possible for the authors to have the amounts of coals consumed in the residential sector especially during winter and give an in-depth explanation why this source could contribute substantially?

Reply: Thanks for the suggestions. The information about coal consumptions and an in-depth explanation to the substantial contribution from residential coal burning have been added into the revised manuscript (section 3.3.2, line 519-534):

"Coal is consumed in residential, industrial and power sectors in Beijing. As showed in Fig. S5a, while annual total coal consumptions dropped rapidly during 2006-2015, the annual residential coal

consumptions remained almost unchanged with their percentages in total coal consumptions rising from 8.7% in 2006 to 23.4% in 2015 (Beijing Municipal Bureau of Statistics, 2016; Yu et al., 2018). As a matter of fact, over 60% of the residential coal consumption occurred in rural areas of Beijing (Fig. S5b), and residential coal is mainly burned in the cool winter season for house heating (Xue et al., 2016). While emission factors of VOCs from residential coal burning have been found to be a factor of 20 greater than those from coal-fired power plants (Liu et al., 2017), the differences in coal quality between the urban and rural areas augment emissions in rural areas: coal used in urban area was entirely anthracite with comparatively much lower emissions of volatiles than other types of coal (Xu et al., 2017); instead only 5-15% of coal used in rural area was anthracite (Xue et al., 2016). Consequently, residential coal combustion could have been a major contributor to the ambient VOCs in rural areas of Beijing during winter."

The coal consumptions in Beijing from 2006 to 2015 were also added to the supplement information as shown below (Fig. S5).

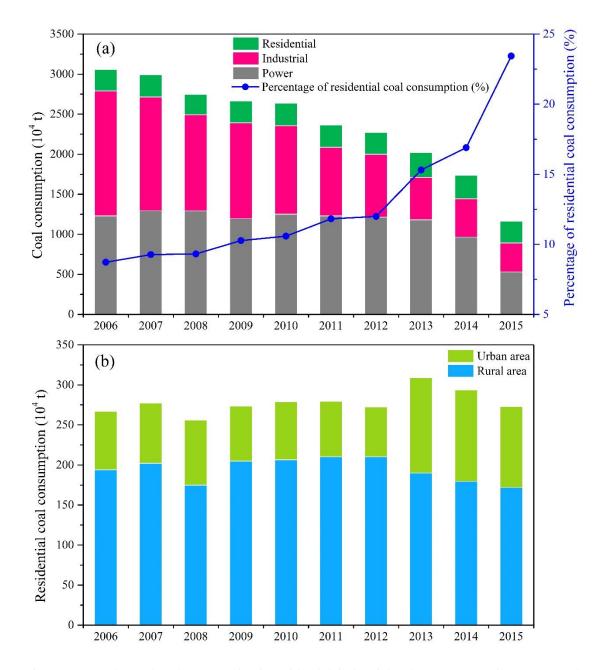


Figure S5. (a) The total coal consumption in residential, industrial and power generation sectors and the percentage of residential coal consumption in total coal consumption in Beijing during 2006-2015; (b) Residential coal consumption in urban and rural areas of Beijing during 2006-2015.

[2] The study suggests clean energy use in residential sector is vital for reducing VOCs in the heavily polluted winter. This aspect should be incorporated into the discussion or the conclusion part of the manuscript.

Reply: Thanks. In the revised manuscript we have re-write the introduction part to stress that residential burning of solid fuels for cooking and heating is not only a problem of indoor air

pollution, but also an important source of outdoor air pollutants. In the conclusion part, based on the results that residential burning of solid fuels contributed nearly halves of VOCs in ambient air

during wintertime heating, we added:

"However, as observed in this study, even in megacities like Beijing, burning raw coal or biomass for household heating in winter could contribute near half of VOCs in ambient air. If the emission control over residential burning of solid fuels is underappreciated, the intervention measures targeted on traffic and industry sectors would be not so effective in the wintertime heating period as did in non-heating periods either to lower PM_{2.5} as indicated by Liu et al. (2016) or to lower VOCs in ambient air as indicated by this study. If fact, a study by Yu et al. (2018) during the same field campaign of this study demonstrated that, without emission control over residential burning of solid fuels, ambient PM_{2.5}-bound toxic polycyclic aromatic hydrocarbons in rural Beijing during the 2014 APEC summit remained unchanged despite of the temporary intervention control measures, and they were largely aggravated after the start of wintertime heating. Therefore, cleaner energy use instead of poor-technology burning of solid fuels for household heating would have tremendous health benefits in lowering both indoor and outdoor air pollution particularly in heavily polluted winter. It worth noting that this study was conducted in a rural area of the megacity Beijing. Emission from residential burning of solid fuels would be a source of greater importance and thus

Minors:

Although I am not a native English speaker, I would say that there is still room for improving English writing of the manuscript. The authors better find a native English speaker to check the English.

Reply: Thanks for your careful check and your great patience in listing the errors/mistakes. For the revised manuscript, we have also asked a native English speaker to re-check the English writing.

Line 33-34: Change "during wintertime severe haze events" to "during severe wintertime haze events":

Reply: Revised as suggested.

Line 35: Change "comparatively much less" to "not well";

deserves more concern in less developed regions."

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Reply: Revised as suggested.
Line 37: Change "inside" to "on";
Reply: Revised as suggested.
Line 37: add "the" before "University";
Reply: Revised as suggested.
Line 38: Change "northeast" to "northeastern";
Reply: Revised as suggested.
Line 39: Remove "that";
Reply: Revised as suggested.
Line 39: Change "during" to "on";
Reply: Revised as suggested.
Line 40: Add "the" before "air quality";
Reply: Revised as suggested.
Line 41: Change "in" to "on";
Reply: Revised as suggested.
Line 41: Remove "that";
Reply: Revised as suggested.
Line 41: Change "since" to "on";
Reply: Revised as suggested.
Line 42: Change "it is" to "this sample collection period provided";
Reply: Revised as suggested.
Line 43: Remove "the" before "temporary" and before "wintertime";
Reply: Revised as suggested.
Line 44, 45: Add "the" before "temporary";
Reply: Revised as suggested.
Line 46: Change "about" to "approximately";
Reply: Revised as suggested.
Line 47: Change "that" to "the values";
Reply: Revised as suggested.
Line 46-48: Change "that of 23.41 ppb before the APEC (25 October-2 November; Period I) or
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21.71 ppb after the APEC (13 November-31 December; Period III)" to "the values of 23.41 ppb in
Period I (25 October-2 November) before the APEC and 21.71 ppb in period III (13 November-31
December) after the APEC";
Reply: Revised as suggested.
Line 48: Change "Their" to "The";
Reply: Revised as suggested.
Line 49: Change "drop" to "decrease";
Reply: Revised as suggested.
Line 50: Change "of" to "over";
Reply: Revised as suggested.
Line 51: Remove "the" before "southerly";
Reply: Revised as suggested.
Line 52: Change "that in the northerly ones during period I, II and III" to "those in northerly air
masses during periods I, II and III";
Reply: Revised as suggested.
Line 53: Remove "and";
Reply: Revised as suggested.
Line 54: Change "south" to "southern";
Reply: Revised as suggested.
Line 56: Change "changed" to "the altered";
Reply: Revised as suggested.
Line 57, 58: Remove "the" before "Period";
Reply: Revised as suggested.
Line 58: Change "emission" to "emissions";
Reply: Revised as suggested.
Line 59: Add "the" before "ambient";
Reply: Revised as suggested.
Line 60: Change "emission" to "emissions";
Reply: Revised as suggested.
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Line 60: Remove "they";

Reply: Revised as suggested.

Line 61-62: Change "VOCs during the period I, II and III" to "the VOCs during periods I, II and III":

Reply: Revised as suggested.

Line 62-63: Change "became the dominant source which accounted for 45.1% of the VOCs" to "became the largest source, accounting for 45.1% of the VOCs";

Reply: Revised as suggested.

Line 64-65: Change "with a remarkably lower average contribution percentage (38.2%) in the southerly air masses than that of 48.8% in the northerly air masses" to "with a specifically lower average contribution percentage in southerly air masses (38.2%) than in northerly air masses (48.8%)";

Reply: Revised as suggested.

Line 72: Change "emission" to "emissions";

Reply: Revised as suggested.

Line 73: Change "benefit improving" to "improve the";

Reply: Revised as suggested.

Line 74: Change "an extensive concern" to "a widespread concern";

Reply: Revised as suggested.

Line 75, 79: Change "summertime" to "summer";

Reply: Revised as suggested.

Line 75: Change "wintertime" to "winter";

Reply: Revised as suggested.

Line 80: Remove "as well";

Reply: Revised as suggested.

Line 82: Change "comparatively the role of VOCs in the wintertime with PM_{2.5}" to "comparatively, the effect of VOCs on wintertime PM_{2.5}";

Reply: Revised as suggested.

Line 88: Change "in the control of air pollution by PM_{2.5} in wintertime" to "in the control of PM_{2.5} air pollution in winter";

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Line 90: Change "under" to "on";
Reply: Revised as suggested.
Line 92: Change "vehicle exhausts are" to "vehicle exhaust is an";
Reply: Revised as suggested.
Line 96: Change "particularly in north China in wintertime" to "particularly in northern China in
winter";
Reply: Revised as suggested.
Line 99: Add "the" before "VOCs";
Reply: Revised as suggested.
Line 100: Change "is widely occurring in the rural areas" to "widely occur in rural areas";
Reply: Revised as suggested.
Line 101-102: Change "how the enhanced emission" to "how enhanced emissions";
Reply: Revised as suggested.
Line 109: Change "north" to "northern";
Reply: Revised as suggested.
Line 110: Add "the" before "air quality";
Reply: Revised as suggested.
Line 111: Remove "so";
Reply: Revised as suggested.
Line 111: Remove "when";
Reply: Revised as suggested.
Line 112: Change "especially in wintertime with" to "especially in winter, with";
Reply: Revised as suggested.
Line 116: Change "kind" to "type";
Reply: Revised as suggested.
Line 118: Change "As for" to "For";
Reply: Revised as suggested.
Line 121: Change "drops of" to "decreases in";
Reply: Revised as suggested.
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Line 123: Change "inside" to "on";

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Reply: Revised as suggested.
Line 124: Add "the" before "total";
Reply: Revised as suggested.
Line 124: Add "were" before "reduced";
Reply: Revised as suggested.
Line 125: Add "those in" before "the period";
Reply: Revised as suggested.
Line 126: Change "about" to "of";
Reply: Revised as suggested.
Line 127, 134: Remove "the" before "urban";
Reply: Revised as suggested.
Line 127: Remove "entirely";
Reply: Revised as suggested.
Line 128: Add "on" before "a regional scale";
Reply: Revised as suggested.
Line 132-133: Change "The objectives of present study are" to "The objectives of the present study
are as follows";
Reply: Revised as suggested.
Line 135, 136: change "crucial" to "the major";
Reply: Revised as suggested.
Line 136: Change "wintertime" to "winter";
Reply: Revised as suggested.
Line 137: Add "the" before "APEC";
Reply: Revised as suggested.
Line 141: Change "_" to "_";
Reply: Revised as suggested.
Line 142: Change "inside the campus of" to "on the campus of the";
Reply: Revised as suggested.
Line 142: Add "the" before "Huairou";
Reply: Revised as suggested.
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Line 143-144: Change "The UCAS is located about 60 km northeast of central Beijing and about 150 km northwest of the Tianjin city" to "UCAS is located approximately 60 km northeast of the center of Beijing and approximately 150 km northwest of the city of Tianjin";

Reply: Revised as suggested.

Reply: Revised as suggested.

Line 145-147: Change "16 meters above ground on the top of a four-story building, about 100 m west of a national road and only 1.5 km far away from the APEC main conference hall" to "16 meters above the ground on the top of a four-story building, approximately 100 m west of a national road and only 1.5 km away from the main APEC conference hall";

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Reply: Revised as suggested.
Line 152: Remove "of";
Reply: Revised as suggested.
Line 153: Add "and" before "one";
Reply: Revised as suggested.
Line 154: Add "was" before "less than";
Reply: Revised as suggested.
Line 154: Add "a" before "relative";
Reply: Revised as suggested.
Line 154: Add "of" before "less";
Reply: Revised as suggested.
Line 158: Remove "the time span";
Reply: Revised as suggested.
Line 163-164: Change "and average temperature was 11.4, 7.0, and 0.6°C during periods I, II and
III" to "and the average temperature was 11.4, 7.0 and 0.6°C during periods I, II and III";
Reply: Revised as suggested.
Line 224: Remove "the" before "period";
Reply: Revised as suggested.
Line 226: Change "Total" to "The total";
Reply: Revised as suggested.
Line 226: Change "inside" to "at";
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Line 227: Change "in" to "on";
Reply: Revised as suggested.
Line 229: Change "halves of 57.45, 36.17, and 56.56 ppb" to "half the values (57.45, 36.17, and
56.56 ppb)";
Reply: Revised as suggested.
Line 231: Remove "both";
Reply: Revised as suggested.
Line 234: Add "a" before "more than";
Reply: Revised as suggested.
Line 234: Remove "the";
Reply: Revised as suggested.
Line 234: Change "about" to "an approximately";
Reply: Revised as suggested.
Line 237: Change "or" to "and the";
Reply: Revised as suggested.
Line 238: Change "densities" to "density";
Reply: Revised as suggested.
Line 240: Change "shared by" to "of";
Reply: Revised as suggested.
Line 240: Remove "quite";
Reply: Revised as suggested.
Line 243-244: Change "percentages shared by aromatics became lower during period II (12%) when
compared to that in period I (21%) or period III (15%)" to "the percentage of aromatics was lower
during period II (12%) than during period I (21%) and period III (15%)";
Reply: Revised as suggested.
Line 246-247: Change "decreased by 49.0, 32.5, 72.8, and 48.1%, respectively, when compared to
those during period I" to "were 49.0, 32.5, 72.8 and 48.1% lower than those during period I,
respectively";
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Line 247-248: Change "Aromatics evidently had a more substantial drop" to "Aromatics evidently

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underwent a larger decrease";
Reply: Revised as suggested.
Line 249: Remove "as";
Reply: Revised as suggested.
Line 252: Change "ozone formation potentials (OFPs)" to "ozone formation potential (OFP)";
Reply: Revised as suggested.
Line 253: Change "in average during periods I, II and III were" to "on average during periods I, II
and III was";
Reply: Revised as suggested.
Line 255: Change "Their" to "The";
Reply: Revised as suggested.
Line 255: Change "potentials (SOAFPs)" to "potential (SOAFP)";
Reply: Revised as suggested.
Line 257, 259: Add "the" to "total";
Reply: Revised as suggested.
Line 261: Change "VOCs" to "VOC";
Reply: Revised as suggested.
Line 262: Change "drop" to "decreases";
Reply: Revised as suggested.
Line 264: Change "changed contribution by aromatic" to "the altered contribution of aromatics";
Reply: Revised as suggested.
Line 266: Change "condition" to "conditions and from";
Reply: Revised as suggested.
Line 267: Change "condition" to "conditions";
Reply: Revised as suggested.
Line 270: Change "showed" to "shown";
Reply: Revised as suggested.
Line 271: Change "with the increase in" to "with an increase in the";
Reply: Revised as suggested.
Line 272: Change "like that" to "as those";
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Reply: Revised as suggested.
Line 276: Add "the" before "wind";
Reply: Revised as suggested.
Line 280: Change "showed" to "shows";
Reply: Revised as suggested.
Line 281: Change "the" to "a";
Reply: Revised as suggested.
Lines 282-285: "It clearly demonstrated that the mixing ratios of VOCs increased rapidly, and the
back trajectories indicated that air masses changed from northerly to southerly and then declined
sharply while the air masses changed back from southerly to northerly", rewrite the sentences.
Reply: Revised as suggested.
Lines 285-286: Change "The southern areas of UCAS are the central Beijing with stronger
emissions" to "The southern areas of UCAS are in central Beijing where emissions are stronger";
Line 286: Add "that";
Reply: Revised as suggested.
Line 287: Change "increase of" to "increase in the";
Reply: Revised as suggested.
Line 288: Change "of source regions" to "in the source region";
Reply: Revised as suggested.
Line 289: Change "showed" to "shown";
Reply: Revised as suggested.
Line 291: Change "changed" to "altered";
Reply: Revised as suggested.
Line 295: Add "and" before "2)";
Reply: Revised as suggested.
Line 296: should be "::: Mongolia and quickly:::";
Reply: Revised as suggested.
Line 297: Change "emission" to "emissions";
Reply: Revised as suggested.
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Line 302: Change "period" to "periods";

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Line 303: Change "that" to "those";
Reply: Revised as suggested.
Line 304: Change "OFPs in the southerly air masses were" to "the OFP in southerly air masses was";
Reply: Revised as suggested.
Line 305: Change "and SOAFPs in the" to "and the SOAFP in";
Reply: Revised as suggested.
Line 305: Change "were" to "was";
Reply: Revised as suggested.
Line 308: Change "This indicates that the north and south" to "These results indicate that the
northern and southern";
Reply: Revised as suggested.
Line 313-314: Change "OFPs and SOAPFs" to "OFP and SOAFP";
Reply: Revised as suggested.
Line 314: Add "that" before "the changes";
Reply: Revised as suggested.
Line 315: Remove "to";
Reply: Revised as suggested.
Line 316: Change "when compared to that during period I average" to "compared to that during
period I, the average";
Reply: Revised as suggested.
Line 319: Change "OFPs decreased by 48.1% and SOAFPs" to "the OFP decreased by 48.1% and
the SOAFP";
Reply: Revised as suggested.
Line 320: Change "when compared to that" to "compared to those";
Reply: Revised as suggested.
Line 321: Add "the" before "average";
Reply: Revised as suggested.
Line 323: Remove "when";
Reply: Revised as suggested.
```

Line 323-324: Change "OFPs decreased by 48.9% and SOAFPs decreased by over 70% during period II relative to period I" to "the OFP decreased by 48.9%, and the SOAFP decreased by over 70% during period II relative to those in period I";

Reply: Revised as suggested.

Line 325: Add "a" before "more";

Reply: Revised as suggested.

Line 326: Change "emission control in" to "control over emissions from";

Reply: Revised as suggested.

Line 327: Change "less changes in mixing ratios" to "decreased changes in the mixing ratios";

Reply: Revised as suggested.

Line 328: Change "of" to "over";

Reply: Revised as suggested.

Line 331: Change "that" to "those";

Reply: Revised as suggested.

Line 331: Change "This difference in the increase rates" to "These different increases";

Reply: Revised as suggested.

Line 332: Change "are" to "were";

Reply: Revised as suggested.

Line 333: Change "heating supply was only available since" to "heat sources were only available after";

Reply: Revised as suggested.

Line 334: Change "already" to "have";

Reply: Revised as suggested.

Line 336: Change "apportioning" to "apportionment";

Reply: Revised as suggested.

Line 338: Add "the" before "campaign";

Reply: Revised as suggested.

Lines 338-339: Change "might be resulted from changed contribution by emission sources, such like" to "might have resulted from the altered contributions from emission sources, such as";

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Line 341: add "the" before "characteristic";
Reply: Revised as suggested.
Line 344: Change "in average during period" to "on average during periods";
Reply: Revised as suggested.
Line 345: Change "were approaching" to "approached";
Reply: Revised as suggested.
Line 347, 351: Remove "the" before "period";
Reply: Revised as suggested.
Line 348: Remove ", which is characteristic of vehicular exhaust";
Reply: Revised as suggested.
Line 349: Add "the" before "incomplete";
Reply: Revised as suggested.
Line 354: Change "in" to "on";
Reply: Revised as suggested.
Line 355: Change "that coal burning contributed more" to "an increased contribution of coal
burning";
Reply: Revised as suggested.
Line 357: Change "emission" to "emissions";
Reply: Revised as suggested.
Line 359: Change "are mainly from" to "mainly originate from";
Reply: Revised as suggested.
Line 361: Change "were" to "are";
Reply: Revised as suggested.
Line 364: Change "when compared to that of" to "than the values of";
Reply: Revised as suggested.
Line 365: Change "or" to "and of";
Reply: Revised as suggested.
Line 366: Change "This drop in aromatics/CO ratios" to "This decrease in the ratios of aromatics to
CO";
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Line 367: Change "of" to "over";
Reply: Revised as suggested.
Line 370, 371: Add "those in" before "period I";
Reply: Revised as suggested.
Line 371-372: Change "Apparently larger decrease in TEX/CO ratios in the northerly air masses
reflected" to "A larger decrease in the TEX/CO ratios in northerly air masses reflects the fact that";
Reply: Revised as suggested.
Line 374: Change "3.3.2 Source Apportioning by PMF" to "3.3.2 Source Apportionment by PMF";
Reply: Revised as suggested.
Line 375: Change "Thirty-five" to "The 35";
Reply: Revised as suggested.
Line 375: Add "and" before "ethyne";
Reply: Revised as suggested.
Line 376: Change "sources" to "source";
Reply: Revised as suggested.
Line 377: Add "use with" before "the PMF";
Reply: Revised as suggested.
Line 385: Change "the gasoline vehicle emission" to "gasoline vehicle emissions";
Reply: Revised as suggested.
Line 386: should be "trichloroethylene and tetrachloroethylene";
Reply: Revised as suggested.
Line 387: Change "of" to "by";
Reply: Revised as suggested.
Line 388: Change "manufacturing industrials" to "industrial manufacturing";
Reply: Revised as suggested.
Line 389-390: Change "by industries for make" to "in industry to prepare";
Reply: Revised as suggested.
Line 390: Add "during the" before "production";
Reply: Revised as suggested.
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Line 392: Change "emission" to "emissions";

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Reply: Revised as suggested.
Line 393: Change "a larger percentage of the" to "larger percentages of";
Reply: Revised as suggested.
Line 394: Chang "It is known that TEX are" to "TEX is known to be";
Reply: Revised as suggested.
Line 396: Change "They are also" to "These compounds are also the";
Reply: Revised as suggested.
Line 396-397: Change "auto factory painting and building coating" to "automobile factories, paint
and building coatings";
Reply: Revised as suggested.
Line 398: Change "as" to "to be";
Reply: Revised as suggested.
Line 403-404: Change "top 3 species of" to "the top 3 species emitted during";
Reply: Revised as suggested.
Line 405: Change "were" to "are";
Reply: Revised as suggested.
Line 405-406: Change "and aromatics like benzene" to "as well as aromatics such as benzene";
Reply: Revised as suggested.
Line 406-407: Change "So factor 5 is related to the coal/biomass burning" to "Thus, factor 5 is
related to coal/biomass burning";
Reply: Revised as suggested.
Line 408: Change "period" to "periods";
Reply: Revised as suggested.
Line 409: Add "the" before "VOCs";
Reply: Revised as suggested.
Line 411: Add the "the" before "temporary" and remove "the" before "period";
Reply: Revised as suggested.
Line 412: Change "by" to "of";
Reply: Revised as suggested.
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Line 413: Change "drop" to "decrease";

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Reply: Revised as suggested.
Line 414: Change "by" to "from";
Reply: Revised as suggested.
Line 415: Change "Quite similar" to "Similar";
Reply: Revised as suggested.
Line 417: Remove "the";
Reply: Revised as suggested.
Line 418: Change "from" to "on";
Reply: Revised as suggested.
Line 419: Change "emission" to "emissions";
Reply: Revised as suggested.
Line 420: Add "the" before "VOCs";
Reply: Revised as suggested.
Line 421: Change "were showed" to "are shown";
Reply: Revised as suggested.
Line 422: Change "by" to "of";
Reply: Revised as suggested.
Line 422: Change "with the" to "with the increase in";
Reply: Revised as suggested.
Line 423: Remove "instead";
Reply: Revised as suggested.
Line 427: Change "consumptions" to "consumption";
Reply: Revised as suggested.
Line 429: Change "The residential coal combustion is prevailing" to "Residential coal combustion
is primarily carried out";
Reply: Revised as suggested.
Line 429: Change "by using" to "with";
Reply: Revised as suggested.
Line 430: Add "the" before "rural";
Reply: Revised as suggested.
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Line 431: Change "wintertime" to "winter";
Reply: Revised as suggested.
Line 431: Remove "the";
Reply: Revised as suggested.
Line 431: Change "accounts" to "accounted";
Reply: Revised as suggested.
Line 433: Change "contribute predominately to" to "have been the predominant contributor to the";
Reply: Revised as suggested.
Line 437, 438, 444, 446, 449: Remove "the" before "period";
Reply: Revised as suggested.
Line 437, 440, 441: Change "by" to "from";
Reply: Revised as suggested.
Line 441: Change "reduced" to "decreased";
Reply: Revised as suggested.
Line 443: Change "showed" to "made";
Reply: Revised as suggested.
Line 443: Change "period" to "periods";
Line 447: Change "account" to "accounted";
Reply: Revised as suggested.
Line 451: Change "period" to "periods";
Reply: Revised as suggested.
Line 451: Remove "respectively";
Reply: Revised as suggested.
Line 452: Change "emission (gasoline and diesel vehicles) was" to "emissions (gasoline + diesel)
were";
Reply: Revised as suggested.
Line 453: Change "VOCs during the period" to "the VOCs during periods";
Reply: Revised as suggested.
Line 454: Remove "instead";
Reply: Revised as suggested.
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Line 456: Change "period" to "periods";
Reply: Revised as suggested.
Line 459: Change "Contributions" to "The contributions";
Reply: Revised as suggested.
Line 460: Change "were" to "are";
Reply: Revised as suggested.
Line 460-461: Change "was mainly coming" to "mainly originated";
Reply: Revised as suggested.
Line 467, 468, 470: Change "of" to "in";
Reply: Revised as suggested.
Line 472: Change "emission" to "emissions";
Reply: Revised as suggested.
Line 474: Change "3.3.3 Source contributions to SOAFPs" to "3.3.3 Source contributions to the
SOAFP";
Reply: Revised as suggested.
Line 475: Change "apportioning" to "apportionment";
Reply: Revised as suggested.
Line 475: Change "SOAFPs by" to "the SOAFP of";
Reply: Revised as suggested.
Line 476: Change "showed" to "shown";
Reply: Revised as suggested.
Line 477: Change "condition" to "conditions";
Reply: Revised as suggested.
Line 477: Change "SOAFPs by solvent use were much higher than that by" to "SOAFP of solvent
use was much higher than that of";
Reply: Revised as suggested.
Line 478: Change "were" to "was";
Reply: Revised as suggested.
Line 479, 481: Change "period" to "periods";
Reply: Revised as suggested.
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Line 482: Change "of SOAFPs" to "in SOAFP";
Reply: Revised as suggested.
Line 483: Change "by" to "from";
Reply: Revised as suggested.
Line 485: Change "condition," to "conditions, the";
Reply: Revised as suggested.
Line 485: Change "of" to "in";
Reply: Revised as suggested.
Line 486: Change "by" to "from";
Reply: Revised as suggested.
Line 490: Change "or" to "and";
Reply: Revised as suggested.
Line 490: Change "from" to "during";
Reply: Revised as suggested.
Line 494: Change "SOAFPs" to "the SOAFP";
Reply: Revised as suggested.
Line 494: Remove "only";
Reply: Revised as suggested.
Line 496: Change "is a large of" to "are large";
Reply: Revised as suggested.
Line 500-501: Change "SOA often shared higher factions in" to "SOAs often composed higher
fractions of";
Reply: Revised as suggested.
Line 502: Change "are much less" to "not well";
Reply: Revised as suggested.
Line 504: Change "inside" to "on";
Reply: Revised as suggested.
Line 505: Change "during" to "on";
Reply: Revised as suggested.
Line 505, 506: Remove "and, in fact";
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Reply: Revised as suggested.
Line 507: Change "could take" to "took";
Reply: Revised as suggested.
Line 509: Change "with" to "from";
Reply: Revised as suggested.
Line 510: Change "could also compare" to "also compared";
Reply: Revised as suggested.
Line 511: Add "use";
Reply: Revised as suggested.
Line 512: Change "investigate" to "investigated";
Reply: Revised as suggested.
Line 513: Change "the" to "a";
Reply: Revised as suggested.
Line 514-515: Change "period II (3-12 November)" to "period (period II; 3-12 November), the";
Reply: Revised as suggested.
Line 515: Change "when compared to that" to "compared to those";
Reply: Revised as suggested.
Line 516: Change "And their" to "In addition, the";
Reply: Revised as suggested.
Line 516: Change "potentials" to "potential";
Reply: Revised as suggested.
Line 518: Change "drop" to "decrease";
Reply: Revised as suggested.
Line 519: Change "of" to "over";
Reply: Revised as suggested.
Line 520: Change "apportioning" to "apportionment";
Reply: Revised as suggested.
Line 523: Chang "about" to "of";
Reply: Revised as suggested.
Line 524: Change "With" to "Through";
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Reply: Revised as suggested.
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Line 525: Change "of wind directions" to "in wind direction";

Reply: Revised as suggested.

Line 526: Change "Total" to "The total";

Reply: Revised as suggested.

Line 527: Change "that" to "those";

Reply: Revised as suggested.

Line 531: Change "were" to "was";

Reply: Revised as suggested.

Line 533: Change "by" to "of";

Reply: Revised as suggested.

Line 534: Change "in average due to drops in the percentages by" to "on average due to decreases in the percentages of";

Reply: Revised as suggested.

Line 535-536: Change "that accounted" to "accounting";

Reply: Revised as suggested.

Line 537: Change "south" to "southern";

Reply: Revised as suggested.

Line 538: Change "north" to "northern".

Reply: Revised as suggested.

References:

Beijing Municipal Bureau of Statistics (BMBS).: Bejing Statistical Yearbook 2016. China Statistics Press, Beijing, 2016.

Liu, C. T., Zhang, C. L., Mu, Y. J., Liu, J. F., and Zhang, Y. Y.: Emission of volatile organic compounds from domestic coal stove with the actual alternation of flaming and smoldering Pollut., combustion processes, Environ. 221, 385-391, http://dx.doi.org/10.1016/j.envpol.2016.11.089, 2017.

Liu, J., Mauzerall, D. L., Chen, Q., Zhang, Q., Song, Y., Peng, W., Klimont, Z., Qiu, X. H., Zhang, S. Q., Hu, M., Lin, W. L., Smith, K. R., and Zhu, T.: Air pollutant emissions from Chinese

- households: A major and underappreciated ambient pollution source, P. Natl. Acad. Sci. USA., 113, 7756-7761, http://dx.doi.org/10.1073/pnas.1604537113, 2016.
- Xu, J. Y., Zhuo, J. K., Zhu, Y. N., Pan, Y., and Yao, Q.: Analysis of volatile organic pyrolysis products of bituminous and anthracite coals with single-photon ionization time-of-flight mass spectrometry and gas chromatography/mass spectrometry, Energ Fuel, 31, 730-737, https://doi.org/10.1021/acs.energyfuels.6b02335, 2017.
- Xue, Y. F., Zhou, Z., Nie, T., Wang, K., Nie, L., Pan, T., Wu, X. Q., Tian, H. Z., Zhong, L. H., Li, J., Liu, H. J., Liu, S. H., and Shao, P. Y.: Trends of multiple air pollutants emissions from residential coal combustion in Beijing and its implication on improving air quality for control measures, Atmospheric Environment, 142, 303-312, 10.1016/j.atmosenv.2016.08.004, 2016.
- Yu, Q. Q., Yang, W. Q., Zhu, M., Gao, B., Li, S., Li, G. H., Fang, H., Zhou, H. S., Zhang, H. N., Wu, Z. F., Song, W., Tan, J. H., Zhang, Y. L., Bi, X. H., Chen, L. G., and Wang, X. M.: Ambient PM_{2.5}-bound polycyclic aromatic hydrocarbons (PAHs) in rural Beijing: Unabated with enhanced temporary emission control during the 2014 APEC summit and largely aggravated after the start of wintertime heating, Environ. Pollut., 238, 532-542, http://dx.doi.org/10.1016/j.envpol.2018.03.079, 2018.

- 1 Volatile organic compounds at a rural site in Beijing:
- 2 Influence of temporary emission control and wintertime
- 3 **heating**
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Abstract

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While residential coal/biomass burning might be a major and underappreciated 33 34 emission source for PM_{2.5} especially during winter season, it is not well constrained whether burning solid fuels contributes substantially to ambient volatile organic 35 compounds (VOCs), which are precursors to secondary organic aerosols (SOA) that 36 typically have a higher contribution to particulate matters during winter haze 37 eventsSecondary organic aerosols (SOA) contribute substantially to PM2.5 during 38 wintertime severe haze events in north China, yet ambient volatile organic compounds 39 (VOCs) as SOA precursors are comparatively much less characterized in winter 40 especially in rural areas. In this study, ambient air samples were collected in 2014 from 41 25 October to 31 December at a rural site inside on the campus of the University of 42 Chinese Academy of Sciences (UCAS) in northeastern Beijing for the analysis of VOCs. 43 44 Since that temporary intervention measures were implemented during on 3-12 November to improve the air quality for the Asian-Pacific Economic Cooperation 45 (APEC) summit held in on 5-11 November in Beijing, and that wintertime central 46 47 heating started since on 15 November in Beijing after the APEC summit, it is this sample <u>collection period provided</u> a good opportunity to study the influence of the temporary 48 control measures and the wintertime heating on the ambient VOCs. As a result of the 49 temporary intervention measures implemented during 3-12 November (period II), the 50 total mixing ratios of non-methane hydrocarbons averaged 11.25 ppb, about 51 approximately 50% lower than that the values of 23.41 ppb beforein period I the APEC 52 53 (25 October-2 November; Period I) orand 21.71 ppb after their period III APEC (13 November-31 December; Period III). Their ozone and SOA formation potentials 54

55 decreased by ~50% and ~70%, respectively, during period II relative to period I, with the larger dropdecrease in SOA formation potentials attributed to more effective control 56 57 of over aromatic hydrocarbons mainly from solvent use. Back trajectory analysis revealed that the average mixing ratios of VOCs in the southerly air masses were 2.3. 58 59 2.3 and 2.9 times thatthose in the northerly onesair masses during periods I, II and III, respectively; and all VOC episodes occurred under the influence of southerly winds, 60 suggesting much stronger emissions in the southern urbanized regions than in the 61 62 northern rural areas. Based on thea positive matrix factorization (PMF) receptor model, 63 the alteredchanged contributions from traffic emissions and solvent use could explain 47.9% and 37.6% of the reduction in ambient VOCs, respectively, during the Period 64 65 period II relative to the Period period I, indicating that the temporary control measures 66 on vehicle emissions and solvent use were effective inat lowering the ambient levels of VOCs. Coal/biomass burning, gasoline exhaust, and industrial emissions were among 67 the vital major sources, and they altogether contributed 60.3%, 78.6% and 78.7% of the 68 69 VOCs during the periods I, II and III, respectively. Coal/biomass burning, mostly residential coal burning, became the dominant source, which accountinged for 45.1% 70 71 of the VOCs during the wintertime heating period, with a specifically remarkably lower average contribution percentage (38.2%) in the southerly air masses (38.2%) than that 72 of 48.8% in the northerly air masses (48.8%). The results suggests that emission control 73 in the industry and traffic sectors is more effective in lowering ambient reactive VOCs 74 in non-heating season; however, during the winter heating season reducing emissions 75 from residential burning of solid fuels would be of greater importance and would have 76

health co-benefits from lowing both indoor and outdoor air pollution.

1. Introduction

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Volatile organic compounds (VOCs) are precursors of tropospheric ozone and secondary organic aerosols (SOA) (Forstner et al., 1997; Odum et al., 1997; Atkinson, 2000; O'Dowd et al., 2002; Sato et al., 2010). As ozone formation in urban areas is largely VOC-limited (Shao et al., 2009; Tang et al., 2010) and SOA are important components of fine particles or PM_{2.5} (particulate matter with an aerodynamic diameter less than 2.5 µm) (Cabada et al., 2004; Lonati et al., 2005; Huang et al., 2014), reducing emissions of VOCs would be very important for benefit improving the air quality in megacities, such as China's capital city Beijing, where air pollution has become an extensive widespread concern, with increasing surface ozone levels during summer time and severe PM_{2.5} pollution during wintertime (Streets et al., 2007; Ji et al., 2012; Wang et al., 2014). While exposure to indoor air pollution from burning solid fuels affects nearly half of the world's population and household air pollution has long been a major environmental cause of death (Martin et al., 2011; Lim et al., 2012; Subramanian, 2014), emissions from residential energy use such as heating and cooking, prevalent in India and China, also have the largest impact on global premature mortality from outdoor air pollution (Lelieveld et al., 2015). In fact, Juda-Rezler et al. (2011) showed that coal combustion in residential boiler during winter was a major source of PM₁₀ in four cities of central eastern European. A recent study revealed that even in Beijing and its surrounding regions, residential use of solid fuels might be a major and underappreciated ambient pollution source for PM_{2.5} (particularly BC and OC) during

winter heating period (Liu et al., 2016). Since substantial amounts of VOCs are released from poor-technology burning of coal and biomass/biofuels (Yokelson et al., 2008; Shrivastava et al., 2015; Fang et al., 2017; Liu et al., 2017; Cheng et al., 2018), it is of wide concern how residential use of solid fuels, particularly for wintertime household heating, would influence ambient levels and compositions of VOCs. In residential areas of Izmir, Turkey, for example, household burning of coal on uncontrolled burners for domestic heating during winter was found to be a larger source of VOCs than the local traffic (Sari and Bayram, 2014). Due to a wide variety of emission sources of VOCs and large uncertainties of the emission inventories of VOCs, to formulate emission control measures on reducing ambient VOCs is a highly challengeable task. More field measurements are therefore needed to characterize VOCs in ambient air and in emission sources for better source attribution. In Beijing, for example, since higher levels of ozone mostly occur during summer and hot seasons ozone formation in urban areas is largely VOC-limited (Shao et al., 2009; Tang et al., 2010), many field measurements of VOCs in Beijing were conducted in summertime with a focus on their sources (Song et al., 2007; Lu et al., 2007; Yuan et al., 2009; Wang et al., 2010a) and their mixing ratios (Liu et al., 2009; An et al., 2012; Zhang et al., 2012a; Liu et al., 2013) as well, particularly during ozone episodes (Liu et al., 2009; An et al., 2012; Zhang et al., 2012a; Liu et al., 2013). However, comparatively the role of VOCs in the wintertime with PM_{2.5} pollution is much less understood. On the other hand, during extremely severe and persistent haze events in Beijing in China, organic matter (OM) could contribute 30-70% of the total

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PM _{2.5} with higher fractions of SOA in OM (Guo et al., 2014; Huang et al., 2014; Zhang
et al., 2014a). Therefore However, the effect of VOCs on wintertime PM _{2.5} pollution is
much less understood, although the control of VOCs, as SOA precursors, is also of great
importance in the control of <u>PM_{2.5}</u> air pollution by PM_{2.5} time . A previous study already
demonstrated that levels of aromatic hydrocarbons and carbonyls increased
significantly <u>underon</u> haze days in urban Beijing from 2008 to 2010 (Zhang et al.,
2014b), yet few reports are available about wintertime precursor VOCs, especially
about the potential contribution by burning solid fuels for household heating.
In urban areas, vehicle exhaust is an important sources of SOA precursors (McDonald
et al., 2015; Liu et al., 2015a; Ortega et al., 2016; Deng et al., 2017; Gentner et al.,
2017). However, biomass/biofuel burning and coal burning may also contribute
substantially to SOA precursors (Yokelson et al., 2008; Shrivastava et al., 2015; Fang
et al., 2017), particularly in northern China in winter when raw coal and biofuels are
widely used for household heating (Liu et al., 2016; Zhang et al., 2016a; Liu et al.,
2017). In fact, a study by Wang et al. (2013) in 2011-2012 revealed that even at an urban
site in Beijing, coal combustion could account for 28-39% of the VOCs observed in
ambient air. As raw coal and/or biofuel burning widely occur in rural areas in winter
(Liu et al., 2016), it is necessary to investigate how enhanced emissions due to
wintertime household heating would influence the levels and compositions of VOCs in
rural areas, as forming SOA or ozone is an issue of regional scale.
Apart from health benefits from lowering indoor air pollution, controlling
emissions in the residential sector would be of greater importance to further improve

outdoor air quality worldwide in the future (Liu et al., 2016), although previously efforts have been overwhelmingly targeted on reducing emissions from industrial and traffic sectors in many regions. In northern China, for example, Due to a wide variety of emission sources of VOCs and large uncertainties of the emission inventories of VOCs. to assess the effect of emission control measures on reducing ambient VOCs is a highly challengeable task. The Chinese government has implemented long term pollution control actions and air quality has been greatly improved in north China in recent years due to the implementation of long-term pollution control actions (Hao and Wang, 2005; Wang et al., 2009; Zhang et al., 2012b; Liu et al., 2015a; Kelly and Zhu, 2016). However, the air quality there in Beijing is not so satisfactory when compared to that in cities in the United States and Europe, especially in winter, time with frequent haze events and high PM_{2.5} levels. Consequently, during critical international events such as the 2008 Olympic Games (Wang et al., 2010b; Huang et al., 2010) and the 2014 Asia-Pacific Economic Cooperation (APEC) summit, temporary intervention measures were adopted to guarantee better air quality. This kindtype of temporary intervention provided a good opportunity to assess the relatively importance of different sources and to study the effectiveness of various control measures on the reduction of ambient air pollutants including VOCs (Yao et al., 2013; Huang et al., 2017). As for For the 21th Asia-Pacific Economic Cooperation (APEC) summit held in Beijing on 5-11 November 2014, temporary control measures in Beijing and its surrounding regions resulted in significant dropsdecreased in of air pollutants, including PM_{2.5} and NOx (Huang et al., 2015; Liu et al., 2015b; Wang et al., 2015; Xu et al., 2015; Zhang et al., 2016a). For the

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VOCs in ambient air, as observed by Li et al. (2015) at an urban site inside on the campus of Peking University, the total mixing ratios of VOCs were reduced by 44% during the APEC summit control period when compared to those in the period before. As the 2008 Olympic Games or the 2014 APEC were all held in the non-heating periods, it is not certain in what extent the temporary intervention measures mostly targeted on industrial and traffic emissions would take effect in the winter heating period. Since most observation based evaluations about the effectiveness of temporary emission control measures were made with monitoring data in the urban areas, it is entirely necessary to further investigate the influence in rural areas or a regional scale. In urban areas, While vehicle exhaust is an important source of SOA precursors VOCs in urban areas (McDonald et al., 2015; Liu et al., 2015c; Ortega et al., 2016; Deng et al., 2017; Gentner et al., 2017). However, coal burning and biomass/biofuel burning and coal burning may also contribute substantially to SOA precursors (Yokelson et al., 2008; Shrivastava et al., 2015; Fang et al., 2017), VOCs particularly in northern China in winter when raw coal and biofuels are widely used for household heating in regions like the northern China (Liu et al., 2016; Zhang et al., 2016b; Liu et al., 2017). In fact, a study by Wang et al. (2013) in 2011-2012 revealed that even at an urban site in Beijing, coal combustion could account for 28-39% of the VOCs observed in ambient air. As raw coal and/or biofuel burning widely occur in rural areas in winter (Liu et al., 2016), it is necessary to investigate extensively on how enhanced emissions due to wintertime household heating would influence the levels and compositions of ambient VOCs particularly in rural areas, as forming SOA or ozone

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is an issue of regional scale.

In this study, ambient air samples were collected at a rural site in the north of Beijing from 25 October to 31 December 2014, covering the period with the enhanced temporary emission control (3-12 November) for the APEC summit and the wintertime heating period starting from 15 November. The objectives of the present study are: (1) to study changes in the mixing ratios and compositions of VOCs at a rural site in Beijing in response to the emission control during the APEC summit and the wintertime heating; (2) to identify the crucial sources of VOCs in Beijing and their changes during the PM-polluted wintertime; (3) to evaluate the impact of control measures implemented during the APEC summit on the reduction of VOCs in ambient air in rural areas; and (4) to assess the contribution of residential use of solid fuels for household heating to ambient VOCs during winter.

2. Methodology

2.1 Sampling Site and Field Sampling

The ambient air samples were collected at a site (40.41° N, 116.68° E; Fig. 1) inside on the campus of the University of Chinese Academy of Science (UCAS) in the Huairou district of Beijing. The UCAS is located about approximately 60 km northeast of the central of Beijing and about approximately 150 km northwest of the city of Tianjin city. It is surrounded by several small villages and farmlands. The samples were collected about 16 meters above the ground on the top of a four-story building, about approximately 100 m west of a national road and only 1.5 km-far away from the main APEC-main conference hall.

Ambient air samples were collected from 25 October-31 December 2014 using cleaned and evacuated 2 L silica-lined stainless steel canisters. During field sampling, a model 910 canister sampler (Xonteck Inc., California, USA) with a constant flow rate of 66.7 ml min⁻¹ was adopted to allow each canister to be filled in 60 min. Samples were collected at approximately 10:00 and 15:00 of local time (LT) on sunny days, and one or two more samples were collected at 12:00 and/or 18:00 LT on haze days when the visibility was less than 10 km at a relative humidity of less than 90% (Fu et al., 2016). A total of 153 samples were collected during sampling. According to the air pollution control measures, the field campaign was divided into periods I (25 October-2 November), II (3-12 November) and III (13 November-31 December). Period II was the time span when temporary control (Table 1: measures http://www.zhb.gov.cn/gkml/hbb/qt/201411/t20141115_291482.htm) implemented for better air quality. Wintertime heating started on 15 November just after the cease of temporary control measures on 13 November. During the sampling periods, the prevailing winds were mostly from north to northwest (315-360°); the average wind speeds were 3.5, 3.9, and 4.1 m s⁻¹; and the average temperature was 11.4, 7.0, and 0.6° C during periods I, II and III, respectively.

2.2 Laboratory Analysis of VOCs and Carbon Monoxide

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All ambient air samples were analyzed with a Model 7100 pre-concentrator (Entech Instruments Inc., California, USA) coupled with an Agilent 5973N gas chromatography-mass selective detector/flame ionization detector (GC-MSD/FID, Agilent Technologies, USA). Detailed cryogenically concentration steps are described

elsewhere (Zhang et al., 2012c). Briefly, 500 ml ambient air samples in the canister were-was first draw through a primary liquid-nitrogen cryogenic trappumped into the primarily trap with glass beads and then concentrated with liquid nitrogen cryogenic trap at -180160°C to get VOCs trapped. Following tThe primary trap was then heated to 10°C, and all target compounds were transferred by pure helium to a secondary trap at -50°C with Tenax-TA as adsorbents. Majority of H₂O and CO₂ were removed through these two traps. Then the secondary trap then was heated to get VOCs transferred by helium to a third cryo-focus trap at -170°C. After the focusing step, the third trap was rapidly heated and the VOCs were transferred to the GC-MSD/FID system. The mixture were first separated by a DB-1 capillary column (60 m×0.32 mm×1.0 μm, Agilent Technologies, USA) with helium as carrier gas, and then split into two ways, one is a PLOT-Q column (30 m×0.32 mm×20.0 μm, Agilent Technologies, USA) followed by FID detector, another is to a 0.35 m×0.10 mm I.D. stainless steel line followed by MSD detection. The GC oven temperature was programmed to be initially at 10°C, holding for 3 min; next-increased to 120°C at 5°C min⁻¹, and then at 10°C min⁻¹ to 250°C with a final holding time of 7 min. The MSD was run in selected ion monitoring (SIM) mode and the ionization method was electron impacting. Carbon monoxide (CO) in the ambient air samples were also analyzed with an Agilent model 6890 gas chromatography equipped with a FID and a packed column (5Å Molecular Sieve 60/80 mesh, 3 m×1/8 inch). CO was first separated by packed column, then converted to CH₄ by Ni-based catalyst and finally detected by FID (Zhang et al., 2016b).

2.3 Quality Control and Quality Assurance

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Before sampling, all canisters were flushed at least five times by repeatedly filling and evacuating humidified zero air. In order to check if there was any contamination in the canisters, all canisters were evacuated after the cleaning procedures, re-filled with pure nitrogen, stored in the laboratory for at least 24 h, and then analyzed the same way as field samples to make sure that all the target VOCs compounds were not present.

Target compounds were identified based on their retention times and mass spectra, and quantified by external calibration methods. The calibration standards were prepared by dynamically diluting the Photochemical Assessment Monitoring Stations (PAMS) standard mixture and TO-14 standard mixture (100 ppbv, Spectra Gases Inc., New Jersey, USA) to 0.5, 1, 5, 15 and 30 ppb. The calibration curves were obtained by running the five diluted standards plus humidified zero air the same way as the field samples. The humidified zero air was initially analyzed every day to ensure the cleanness of system and then the analytical system was challenged daily with a one-point (typically 1 ppb) calibration before running air samples. If the responses wereas beyond ±+/-10% of the initial calibration curve, recalibration was performed. The method detection limits (MDL) for each VOCs species were presented in Table 2.

2.4 Positive Matrix Factorization (PMF)

PMF is a multivariate factor analysis tool that decomposes a matrix of sample data into two matrices: factor contributions (G) and factor profiles (F). The method is reviewed briefly here and described in greater detail elsewhere (Paatero and Tapper, 1994; Paatero, 1997). PMF uses both concentration and user-provided uncertainty associated with the data to weight individual points. Data values below the MDL were

substituted with MDL/2; missing data values were substituted with median concentrations. If the concentration is less than or equal to the MDL provided, the uncertainty is calculated using the equation of $Unc = 5/6 \times MDL$; if the concentration is greater than the MDL provided, the uncertainty is calculated as $Unc = [(Error faction \times mixing ratio)^2 + (MDL)^2]^{1/2}$. The number of factors in PMF was initially chosen based on the result of PCA/APCS model (Zhang et al., 2012c).

3. Results and discussion

3.1 Changing mixing ratios and compositions

As mentioned above, during the period II (3-12 November), temporary emission control measures were implemented to improve air quality during the 2014 APEC summit. The Total mixing ratios of VOCs observed at the rural site insideat UCAS during the period II was 11.25±3.22 ppb inon average, significantly lower than that the value of 23.41±5.76 ppb during period I and 21.71±2.97 ppb during period III (Fig. 2). These levels were less than halves of half the values (57.45, 36.17, and 56.56 ppb) observed by Li et al. (2015) at an urban site in Beijing before, during and after the APEC summit, respectively. However, both our measurements at a rural site in this study and the measurements at an urban site by Li et al. (2015) consistently demonstrated that the temporary emission control resulted in a large decrease in ambient VOCs during the APEC summit, with a more than 30% reduction in the urban areas (Li et al., 2015) and about an approximately 50% reduction in rural areas, as observed in this study. This reduced ambient mixing ratios of VOCs during the period II was also in line with the decreased PM2.5 concentrations observed in Beijing during

the APEC summit (Liu et al., 2015b), orand the reduced NO₂ vertical column densities density (VCD) and aerosol optical depth (AOD) in Beijing during the APEC summit based on remote sensing (Huang et al., 2015).

The percentages shared byof alkanes, alkenes, and ethyne in total VOCs were quite similar: alkanes accounted for 54, 57 and 54% of VOCs; alkenes accounted for 12, 16 and 17%; and ethyne accounted for 13, 14 and 14% of VOCs during periods I, II and III, respectively. Instead, the percentages shared byof aromatics becamewas lower during period II (12%) when compared to that inthan during period I (21%) or and period III (15%).

The mean mixing ratios of alkanes, alkenes, aromatics and ethyne during period II were 6.47, 1.83, 1.33, and 1.62 ppb (Fig. 2), and they decreased bywere 49.0, 32.5, 72.8, and 48.1%_, respectively, when compared to lower than those during period I, respectively. Aromatics evidently hadunderwent a more substantial dropdecrease. Benzene, toluene, ethylbenzene, and m,p-xylene, which are the most abundant aromatics and usually collectively termed as-BTEX, were 52.8, 73.1, 78.8, and 80.5% lower during period II than during period I, respectively.

Table S1 shows a comparison of VOCs from our study with those observed at other metropolitan areas in the world. Mixing ratios of VOCs from this study at a rural site in Beijing during period I (23.41 ppb) and period III (21.71 ppbv) were comparable to that in urban Shanghai from January 2007 to March 2010 (Cai et al., 2010), but lower than those in Beijing during June 2008 (Wang et al., 2010a), Guangzhou from June 2011 to May 2012 (Zou et al., 2015), Lille, French from May 1997 to April 1999

(Borbon et al., 2002) and Houston in August-September 2006 (Leuchner and
Rappengluck, 2010). Average mixing ratios of VOCs during period II (11.25 ppbv) with
enhanced emission control in the present study were significantly lower than those
reported in other metropolitan areas. As for the most abundant VOC species including
ethane, propane, ethylene, benzene, toluene and ethyne, the mixing ratios of ethane and
ethylene at UCAS were similar to that at Beijing during June 2008 (Wang et al., 2010a)
and urban Guangzhou from June 2011 to May 2012 (Zou et al., 2015), but significantly
lower than that in urban Beijing during 2014 APEC (Li et al., 2015). Mixing ratio of
propane in present study are comparable to that in Hong Kong from September 2002 to
August 2003 (Guo et al., 2007) and Lille, French from May 1997 to April 1999 (Borbon
et al., 2002), but factors of 2-3 lower than that reported in urban Shanghai from January
2007 to March 2010 (Cai et al., 2010) and Guangzhou from June 2011 to May 2012
(Zou et al., 2015). Mixing ratios of benzene and toluene in Lille, French from May 1997
to April 1999 (Borbon et al., 2002) were over 2 times higher than that in present study.
Mixing ratios of ethylene, benzene and toluene in present study were comparable to
those observed in Houston during August-September 2006 (Leuchner and Rappengluck,
2010), meanwhile ethyne, a tracer of incomplete combustion, had mixing ratios 3-4
times higher than that in Houston.
The total ozone formation potentials (OFPs), based on the simplified approach of
MIR (maximum incremental reactivity) scale (Carter, 2009), inon average werewas
60.64, 28.51, and 61.47 ppb (Table \$\frac{\text{S1}}{\text{S2}}\) during periods I, II and III, respectively, with
a 53.0% reduction during period II relative to the period I (Fig. 2). Theeir secondary

organic aerosol formation potentials (SOAFPs) under high-NOx and low-NOx conditions (Ng et al., 2007; Lim and Ziemann, 2009) were also calculated (Table \$2\$3). As showedshown in Fig. 2, the total SOAFPs under low-NOx conditions decreased by 71.0% from 8.77 µg m⁻³ during the period I to 2.54 µg m⁻³ during period II, and the total SOAFPs under high-NOx conditions decreased by 64.4% from 4.02 µg m⁻³ during period I to 1.43 µg m⁻³ during period II. This significant decrease in OFPs and SOAFPs during period II is related to lowered VOCs mixing ratios, especially larger dropdecreases in reactive alkenes and aromatics: alkenes and aromatics explain 26% and 52% of the reduction in total OFPs, respectively, while the decrease in total SOAFPs is mostly due to -changedthe altered contribution byof aromatics (Table \$2\$S3), whose SOAFPs decreased from 7.30 μg m⁻³ during period I to 1.93 μg m⁻³ during period II under low-NOx conditions and from, 2.39 µg m⁻³ during period I to 0.75 µg m⁻³ during period II under high-NOx conditions. The results suggest that enhancing the emission control of reactive alkenes and aromatics would be more is especially effective for reducing OFPs and SOAFPs of ambient VOCs reduction.

3.2 Pollution episodes and influence of source regions

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As showedshown in Fig. 3d and 3e, a number of episodes with mixing ratios of VOCs over 30 ppb were recorded along with thean increase in the CO and SO₂ concentrations (Fig. 3d) during the campaign, such likeas thatthose on 4-5 November, 15-16 November, 18-21 November, 28-30 November, 17 December, and 26-28 December. During the episode on 3-5 November, for example, the total mixing ratio of VOCs was 14.30 ppb on 3 November, reached 31.96 ppb on 4 November, and then

decreased again to 13.83 ppb on 5 November. As shown in Fig. 3a, the wind speeds were all below 2 m s⁻¹ during 3-5 November, and the planetary boundary layer (PBL) height on 4 November (477 m) was approximately 83% of that on 3 November (578 m) (Fig. 3c). This lower PBL height on 4 November could only partly explain the higher levels of VOCs. Figure S1a S2a showedshows the 72-h back trajectories (HYSPLIT, ver. 4.0; http://www.arl.noaa. gov/ready/hysplit4.html) of air masses from 3-5 November at thea height of 100 m in 12-h intervals and the corresponding mixing ratios of VOCs. It demonstrated that the mixing ratios of VOCs increased rapidly while air masses changed from the northerly to the southerly, and then declined sharply while the air masses turned back from the southerly to the northerly again. The southern areas of UCAS are thein central Beijing withwhere stronger emissions are stronger; consequently, air masses that passed through these areas would carry higher levels of pollutants to the sampling site, leading to the quick increase of mixing ratios of VOCs. This rapid change of source regions could reasonably explain more than the PBL height during the pollution episode of VOCs. As showedshown in Fig. S1bS2b, 1e2c, and 1d2d, back trajectories also suggested that the episodes on 18-21 November, 28-30 November and 26-28 December are related to the changedaltered source regions. According to the 72-h back trajectories, air masses arriving at the sampling site could be categorized into two types (Fig. 4): 1) southerly (S) air masses, which passed through Hebei, Shandong, Tianjin, and central Beijing with high-density emissions before reaching UCAS, and; 2) northerly (N) air masses, which originated from Mongolia and, quickly passed through areas with less anthropogenic activity and low-

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density emissions before reaching UCAS. The pollution episodes with higher mixing ratios of VOCs and CO, including the cases on 26-30 October, 4-5 November, 15-16 November, 18-20 November, 25-26 November and 26-28 December (Fig. 3d and 3e), all occurred under the influence of southerly air masses, also suggesting the impacts of emissions in the south.

During periods I, II and III, the average mixing ratios of VOCs for southerly air masses were 2.3, 2.3 and 2.9 times thatthose for northerly air masses (Fig. 4), respectively; the OFPs in the southerly air masses werewas 2.0, 2.0 and 3.3 times that in the northerly air masses, respectively; and the SOAFPs in the air masses from the south werewas 1.7, 3.3, and 3.7 times that in the air masses from the north under low-NOx conditions, and 1.9, 2.7, and 3.5 times that in the air masses from the north under high-NOx conditions, respectively. This These results indicates that the northern and southern regions are completely different in their source strengths. Developing strict Stricter control measures in the southern region would be is an effective way for abating VOCs pollution in Beijing.

As mentioned above, the mixing ratios of VOCs, as well as their OFPs and SOAPFs, decreased greatly during period II. We can further see tThe changes in the southerly and northerly air masses to indicate the changes in emission from different source regions. In the southerly air masses, when compared to that during period I, the average mixing ratios of alkanes, alkenes, aromatics, and ethyne during period II were 8.32, 2.16, 1.93, and 2.23 ppb, with reduction rates of 46.0, 33.3, 64.3, and 44.7%, respectively; accordingly, the OFPs decreased by 48.1% and the SOAFPs decreased by

63.5 % (low-NOx conditions) and 57.6% (high-NOx conditions) during period II when compared to that those during period I (Fig. 4). In the northerly air masses, the average mixing ratios of alkanes, alkenes, aromatics, and ethyne decreased 37.7, 4.8, 87.0, and 18.4% during period II when-compared to that during period I, respectively; the OFPs decreased by 48.9% and SOAFPs decreased by over 70% during period II relative to those in period I (Fig. 4). As discussed below, a more drastic decrease in aromatics in both the northerly and southerly air masses implied more effective emission control over emissions infrom industrial solvent use during the APEC summit, and the much less changes in the mixing ratios of alkenes in the northerly air masses were related to the less effective control of over domestic coal/biomass burning in the northern regions. The mixing ratios of VOCs in the southerly and northerly air masses during period III were 36.1% and 7.2% higher than that those during period I, respectively. This difference in the Those different increases rates might be explained by the fact that the urban areas in the south arewere largely central heating areas where heating supply was sources were only available since after 15 November, and the northern areas were largely rural areas where individual household heating might already have started during period I.

- 426 *3.3 Source attribution and apportioning apportionment*
- 427 *3.3.1 Indication from tracers*

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The great changes in the mixing ratios of VOCs during the campaign might behave resulted from changed the altered contributions by from emission sources, such like as enhanced emission control during the APEC summit or intensified emissions due to

wintertime heating. These changes could be indicated by <u>the</u> characteristic fingerprints of different sources (Guo et al., 2007).

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The toluene/benzene (T/B) ratio, a widely used indicator for sources of aromatics, was 1.09, 0.67 and 0.70 inon average during periods I, II and III, respectively (Fig. S2aS3a). While the T/B ratios during periods II and III were approaching approached 0.6, which is characteristic of coal/biomass burning (Liu et al., 2008; Liu et al., 2015d), the ratios during the period I fell between that of coal/biomass burning (0.6) and vehicle exhaust (1.6), which is characteristic of vehicular exhaust (Wang et al., 2002; Liu et al., 2009; Zhang et al., 2013a). Carbon monoxide (CO), a typical tracer of the incomplete combustion of biomass or fossil fuels (Parrish et al., 2009; Zhang et al., 2015a), showed highly significant correlations with benzene during the period II ($r^2=0.96$, Fig. S2bS3b) and the period III ($r^2=0.88$, Fig. S2bS3b). SO₂, a good indicator of coal burning (Li et al., 2017), had similar concentrations during period II and period I, but its concentrations increased 56.5% inon average during period III compared to that during period I (Fig. 3d), suggesting that an increased contribution of coal burning contributed more after the start of central heating. Methyl tert-butyl ether (MTBE), a specific indicator of gasoline related traffic emissions (Song et al., 2007; Cai et al., 2010), showed better correlation with benzene during period I (r²=0.88, Fig. S2eS3c) than during periods II and III.

As toluene, ethylbenzene and xylene (TEX), are mainly originate from solvent use in painting, decorations and coatings (Guo et al., 2007; Zhang et al., 2012c), the ratios of TEX to CO were are widely used to examine the impact of solvent use relative to

combustion emissions (Zhang et al., 2013a). The ratios of T/CO, E/CO and X/CO were 0.61±0.09, 0.23±0.06 and 0.35±0.07 (ppb/ppm) during period II, obviously lower when compared to that ofthan the values of 1.16±0.49, 0.59±0.24 and 0.99±0.41 during period I, or and of 1.34±0.27, 0.40±0.06 and 0.83±0.09 during period III (Fig. 5B), respectively. This dropdecrease in the ratios of aromatics to /CO-ratios during period II also reflected more effective control of over solvent use during the APEC summit.

If further categorized according to the air masses trajectories, the ratios of T/CO, E/CO and X/CO decreased 29.5, 45.7 and 45.7% in the southerly air masses during period II relative to those in period I₂₇ and decreased 68.0, 80.3 and 83.0% in the northerly air masses during period II relative to those in period I, respectively (Fig. 5A). Apparently A larger decrease in the TEX/CO ratios in the northerly air masses reflected reflects the fact that the control of solvent use was more effective in northern regions.

3.3.2 Source Apportioning Apportionment by PMF

Thirty five The 35 most abundant VOCs, including alkanes, alkenes, aromatics, and ethyne, and sources tracers, such as chloromethane, trichloroethylene, tetrachloroethylene and MTBE, plus SO₂ and CO, were selected for use with the PMF receptor model. Figure 6 shows the 5 sources retrieved by the model.

Factor 1 has high values of MTBE and C₅-C₆ alkanes. MTBE is a common gasoline additive in China, and 2,2-dimethylbutane is used to enhance the octane levels of gasoline (Chang et al., 2004; Song et al., 2007; Cai et al., 2010), Ethyne can be formed during fuel combustion (Blake and Rowland, 1995; Song et al., 2007;

- Suthawaree et al., 2010).; C₅-C₆ alkanes are associated with unburned vehicular
- emissions (Guo et al., 2004; Cai et al., 2010; Zhang et al, 2013b). Consequently factor
- 1 is related to the gasoline vehicle emissions.
- Factor 2 is distinguished by a strong presence of trichloroethylene, and
- 479 tetrachloroethylene and moderate contributions of py propene and butenes.
- Trichloroethylene and tetrachloroethylene are species from <u>industrial</u> manufacturing
- 481 industrials (Yuan et al., 2013; Zhang et al., 2015b); propene and butenes are gases
- widely used by industries in industry forto makeprepare organic chemicals (Guo et al.,
- 483 2007), such as <u>during the</u> production of synthetic rubber in the petrochemical industry
- (Lau et al., 2010). Thus, factor 2 was identified as industrial emissions.
- Factor 3 accounts for a larger percentages of the toluene, ethylbenzene, m/p-xylene
- and o-xylene. <u>ItTEX</u> is known that <u>TEX</u> are to be the primary constituents of solvent
- 487 (Guo et al., 2004; Yuan et al., 2009; Zheng et al., 2013; Zhang et al., 2014c; Ou et al.,
- 488 2015). They These compounds are also the main components in emissions from
- automobile factories, y painting and building coatings (Liu et al., 2008; Yuan et al.,
- 490 2010). Therefore, this source is considered as to be solvent use related to painting and
- 491 architecture.
- Factor 4 is diesel exhaust, which is characterized by a significant amount of n-
- undecane and n-dodecane (Song et al., 2007; Zhang et al., 2012c).
- 494 Factor 5 is characterized by the presence of ethane, ethylene, CO, SO₂ and
- chloromethane. Chloromethane is the typical tracer of biomass burning (Liu et al., 2008;
- Cai et al., 2010; Zhang et al., 2014c). Ethylene, ethane and propene are the top 3 species

ofemitted during rice straw burning (Zhang et al., 2013c; Fang et al., 2017). The VOC species from coal burning were are mainly ethyne, C₂-C₃ alkenes and alkanes, and as well as aromatics likesuch as benzene (Liu et al., 2008). SO₂ is mainly from coal burning (Li et al., 2017). So Thus, factor 5 is related to the coal/biomass burning.

Figure 7 shows the source contributions during periods I, II and III. During period I, gasoline exhaust was the largest source and accounted for 24.0% of the VOCs, while during period II, coal/biomass burning became the largest source. The most significant changes due to the temporary emission control during the period II were in the contribution percentages byof coal/biomass burning (22.3% in period I and 42.4% in period II) and by solvent use (21.9% in period I and 5.8% in period II). The large dropdecrease in the contribution byfrom solvent use was consistent with the above discussion about the TEX/CO ratios. Quite similar contributions were observed for industrial emission and diesel exhaust.

In the period III (13 November-31 December), with the central heating starting from 15 November, coal/biomass burning became the largest source (45.1%), and industrial emission, solvent use, diesel exhaust and gasoline exhaust accounted for 25.2, 12.8, 8.7 and 8.2% of the VOCs, respectively. The time series of source contributions during the campaign were showed are shown in Fig. S3S4; the contribution percentages by of coal/biomass burning increased gradually with the increase in the wintertime heating, while that of gasoline exhaust instead decreased.

Coal/biomass burning was an important source of VOCs during winter in Beijing, especially during period III with the start of central heating. In Beijing, coal

consumption was greater than that of residential biomass (Liu et al., 2016). Coal is consumed in residential, industrial and power sectors in Beijing. As showed in Fig. S5a, while annual total coal consumptions dropped rapidly during 2006-2015, the annual residential coal consumptions remained almost unchanged with their percentages in total coal consumptions rising from 8.7% in 2006 to 23.4% in 2015 (Beijing Municipal Bureau of Statistics, 2016; Yu et al., 2018). As a matter of fact, over 60% of the residential coal consumption occurred in rural areas of Beijing (Fig. S5b), and residential coal is mainly burned in the cool winter season for house heating (Xue et al., 2016). While emission factors of VOCs from residential coal burning have been found to be a factor of 20 greater than those from coal-fired power plants (Liu et al., 2017), the differences in coal quality between the urban and rural areas augment emissions in rural areas: coal used in urban area was entirely anthracite with comparatively much lower emissions of volatiles than other types of coal (Xu et al., 2017); instead only 5-15% of coal used in rural area was anthracite (Xue et al., 2016). Consequently, residential coal combustion could have been a major contributor to the ambient VOCs in rural areas of Beijing during winter. During 2008-2014 in Beijing the annual residential coal consumptions increased gradually while the total coal consumption decreased (Beijing Municipal Bureau of Statistics, 2015). The residential coal combustion is prevailing for heating and cooking by using domestic coal stoves in rural areas around urban Beijing particularly during wintertime. In 2014, although the annual residential coal consumption accounts for 17% (2.93×10⁹ kg a⁻¹) of the total coal consumption in Beijing (Beijing Municipal Bureau of Statistics, 2015), residential coal

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burning could contribute predominately to ambient VOCs from coal burning since the emission factors of VOCs from residential coal burning have been found to be a factor of 20 greater than those from coal fired power plants (Liu et al., 2017).

Compared with that in the period I (Fig. S4S6), the contribution byfrom solvent use during the period II was reduced to a greater extent than other sources; it became 4.29 ppb lower and could explain 37.6% of the reduction in ambient VOCs (Table S3S4). The contribution byfrom gasoline vehicles was 3.18 ppb lower and accounted for 27.9% of the total reductions. The contribution byfrom diesel exhaust and industrial emissions reduceddecreased 2.28 ppb and 1.35 ppb; and explained 20.0 and 11.8% of the total reduction, respectively. Coal/biomass burning showedmade similar contributions during periods I and II, with an elevated contribution percentage in the period II due to the reduction in other sources. This is consistent with the fact that during the APEC summit residential coal/biomass burning was not restricted in the rural areas. Traffic-related sources (gasoline and diesel vehicles) and solvent use accounted for 47.9 and 37.6% of the total reduction in ambient VOCs, indicating that control measures (Table 1) related to the control of traffic and solvent use were among the most effective ways to reduce the ambient VOCs.

Figure 8 shows the source contributions in the southerly and northerly air masses during periods I, II and III, respectively. In the southerly air masses, traffic related emission (gasoline and diesel vehicles) waswere the largest source, contributing 44.1 and 41.5% of the VOCs during the periods I and II, respectively; while coal/biomass burning instead was the largest source during period III, contributing 38.2% of the

VOCs. In the northerly air masses, coal/biomass burning contributed 28.8, 51.6 and
48.8% of the VOCs during periods I, II and III, respectively. Overall, gasoline vehicle
exhaust contributed more VOCs in the southern regions (mostly densely populated
urban areas) and coal/biomass burning and diesel exhaust accounted for more emissions
of VOCs in northern regions (mostly rural areas). <u>The Cc</u> ontributions of different
sources to most reactive alkenes and aromatics based on PMF werewas presented in
Fig. 9. Alkenes was mainly comingoriginated from coal/biomass burning, with shares
of 31.2-68.0%, and gasoline exhaust ranked second, with contributions of 3.0-26.5%.
Unlike alkenes, <u>aromatics were mainly from</u> solvent use was the major contributors of
aromatics, accounting for 77.5% during period I and 29.0% during period II in the
northerly air masses; gasoline exhaust contributed 8.2-43.6% of the aromatics during
the campaign. In the southerly air masses, the reductions in solvent use, gasoline
exhaust, and diesel exhaust during the period II relative to the period I could explain
38.1, 31.1, and 15.8% of the total reduction of VOCs, respectively. In the northerly
air masses, reductions ofin solvent use, diesel exhaust, and gasoline exhaust during the
period II relative to the period I could explain 46.5, 35.8 and 11.9% of the total
reduction $\frac{\text{ofin}}{\text{of}}$ VOCs, respectively. Consequently, control measures related to solvent
use and gasoline exhaust were more effective in the southern regions, while the control
of solvent use and diesel exhaust emission \underline{s} were more effective in the northern region.
3.3.3 Source contributions to the SOAFPs
With the PMF source apportioning apportionment results, the contributions of the
SOAFPs byof different sources were further estimated. As showedshown in Fig. 10,

under low-NOx conditions, the SOAFPs by of solvent use werewas much higher than that byof other sources, which werewas 4.88, 0.68 and 2.89 µg m⁻³, accounting for 56.9, 27.2 and 54.7% of the total SOAFPs during periods I, II and III, respectively. Gasoline exhaust contributed 19.2, 29.5 and 10.9%, and diesel exhaust contributed 16.5, 26.8 and 11.3% of the SOAFPs during periods I, II and III, respectively. During the period II, with temporary intervention measures, the reduction of SOAFPs was mainly due to reduced contributions by from solvent use, gasoline exhaust and diesel exhaust, which could explain 69.1, 14.9 and 12.2% of the reduction in SOAFPs, respectively. Under high-NOx conditions, the calculated reduction of SOAFPs during the period II relative to the period I could largely be explained by the reduced contributions by the solvent use, diesel exhaust and gasoline exhaust, which accounted for 54.0, 25.8 and 16.8% of the reduction in SOAFPs, respectively. It is worth noting that recent chamber studies revealed that aromatic hydrocarbons orand traditional VOCs could not fully explain the SOA formed from during atmospheric aging of source emissions (Zhao et al., 2014; 2015; Liu et al., 2015c; Deng et al., 2017; Fang et al., 2017), particularly for emissions from diesel vehicles or

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orand traditional VOCs could not fully explain the SOA formed fromduring atmospheric aging of source emissions (Zhao et al., 2014; 2015; Liu et al., 2015c; Deng et al., 2017; Fang et al., 2017), particularly for emissions from diesel vehicles or biomass burning (Zhao et al., 2015; Deng et al., 2017; Fang et al., 2017). Therefore the discussion on the SOAFPs in this study is only limited to traditional anthropogenic SOA precursor species (mainly aromatic hydrocarbons), and intermediate-volatility organic compounds (IVOCs), which is agre important SOA precursors large of secondary organic aerosol (Zhao et al., 2014), should be further considered in order to fully understand the influence of control measures on the ambient SOA.

4. Conclusions

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During severe wintertime haze events in recent years in Beijing, SOA often sharedcomposed higher fractions in of organic aerosols, yet their precursor VOCs in ambient air during winter are much less not well understood, especially in the rural areas. In this study, we collected ambient air samples from 25 October to 31 December in 2014 at a rural site insideon the campus of UCAS in north Beijing. As tThe APEC summit was held in Beijing duringon 5-11 November 2014, and temporary control measures were adopted to improve the air quality, and in fact the so-called "APEC Blue" was achieved due to the enhanced emission control. Therefore, we could taketook advantage of this opportunity to see how the control measures influenced the ambient VOCs in the rural areas. On the other hand, wintertime heating withfrom coal burning has been regarded as a major contributor to wintertime PM pollution and haze events; thus, we could also compared the ambient VOC levels and compositions before and after the start of central heating—on 15 November, and investigated the influence of wintertime heating on ambient VOCs based on our observation at thea rural site. We observed that during the enhanced emission control period (period II; (3-12) November), the average mixing ratios of VOCs decreased ~50% when compared to that those before or after that period. And In addition, their ozone and SOA formation potentials accordingly decreased by ~50% and ~70%, respectively, as a result of the temporary intervention measures implemented during period II. The larger dropdecrease in SOA formation potentials was attributable to more effective control of over aromatic hydrocarbons mainly from solvent use. Based on PMF source

apportioningapportionment, the control of traffic-related emissions (gasoline and diesel exhaust) and solvent use could explain 47.9 and 37.6% of the reduction in ambient VOCs. This result thus offered an observation-based evaluation aboutof the temporary emission control measures.

With Through back trajectory analysis, we could compare ambient VOCs with the change ofin wind directions and thus further investigate the source emission strengths in different regions. The Ttotal mixing ratios of VOCs in the southerly air masses were 2.3, 2.3 and 2.9 times that those in the northerly air masses before, during and after the period with temporary emission control for the APEC summit. VOC episodes during the campaign all occurred under southerly winds. This confirms that emission control in the southern urbanized regions is crucial for reducing the ambient VOCs.

As residential coal/biomass burning werewas not controlled during the APEC summit, its contribution to the ambient mixing ratios of VOCs was similar between period I and period II, although the contribution percentages byof coal/biomass burning became the largest inon average due to dropsdecreases in the percentages byof other sources. During period III, with emissions from burning solid fuels for household heating, coal/biomass burning became the largest source, that accountinged for 45.1% of the VOCs. Specifically, during period III, coal/biomass combustion contributed 38.2% of the VOCs in the southerly air masses (or in the southern regions), and 48.8% of the VOCs in the northerly air masses (or in the northern regions).

The finding of this study will provide useful information <u>for the emission on the</u>

<u>direction of control strategyies</u> of VOCs <u>for abating both ozone and PM_{2.5} pollution</u>. <u>A</u>

comparison of VOCs between period I (without intervention measures) and period II (with intervention measures) revealed that the temporary intervention measures mostly targeted on the control in traffic and industry (industrial processes and solvent use) sectors are very effective to reduce reactive alkenes and aromatics and thereby to The reduction reduce in totalthe OFPs and SOAFPs during the APEC is largely due to the drop of reactive alkenes and aromatics, so adopting reactivity based emission control would be the effective and economical way to lower the ozone and SOA formation potentials of VOCs in ambient air. As control measures related to solvent use and vehicle exhausts explained most of the reduction in both ambient VOCs and their ozone/SOA formation potentials, Therefore, enhancing emission control for solvent use (especially solvents with aromatic hydrocarbons) and vehicle exhaust would benefit reducing the VOC-related air pollutionimproving air quality in the future. Moreover However, as observed in this study, even in megacities like Beijing, burning raw coal or biomass for household heating in winter could contribute near half of VOCs in ambient air. If the emission control over residential burning of solid fuels is underappreciated, the intervention measures targeted on traffic and industry sectors would be not so effective in the wintertime heating period as did in non-heating periods either to lower PM_{2.5} as indicated by Liu et al. (2016) or to lower VOCs in ambient air as indicted by this study. If fact, a study by Yu et al. (2018) during the same field campaign of this study demonstrated that, without emission control over residential burning of solid fuels, ambient PM_{2.5}-bound toxic polycyclic aromatic hydrocarbons in rural Beijing during the 2014 APEC summit remained unchanged despite of the

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temporary intervention control measures, and they were largely aggravated after the start of wintertime heating. Therefore, cleaner energy use instead of poor-technology burning of solid fuels for household heating would have tremendous health benefits in lowering both indoor and outdoor air pollution particularly in heavily polluted winter. It worth noting that this study was conducted in a rural area of the megacity Beijing. Emission from residential burning of solid fuels would be a source of greater importance and thus deserves more concern in less developed regions. a cleaner way of wintertime household heating would help to lower both primary emission and secondary formation of air pollutants.

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- 690 References
- An, J. L., Wang, Y. S., Wu, F. K., and Zhu, B.: Characterizations of volatile organic
- compounds during high ozone episodes in Beijing, China, Environ. Monit. Assess.,
- 693 184, 1879-1889, http://dx.doi.org/10.1007/s10661-011-2086-7, 2012.
- Atkinson, R.: Atmospheric chemistry of VOCs and NOx, Atmos. Environ., 34, 2063-
- 695 2101, http://dx.doi.org/10.1016/S1352-2310(99)00460-4, 2000.
- Baudic, A., Gros, V., Sauvage, S., Locoge, N., Sanchez, O., Sarda-Estève, R.,
- Kalogridis, C., Petit, J. E., Bonnaire, N., and Baisnée, D.: Seasonal variability and
- source apportionment of volatile organic compounds (VOCs) in the Paris megacity
- (France), Atmos. Chem. Phys., 16, 11961-11988, https://doi.org/10.5194/acp-16-
- 700 11961-2016, 2016.
- Beijing Municipal Bureau of Statistics (BMBS).: Bejing Statistical Yearbook 20165.
- 702 China Statistics Press, Beijing, 201<u>6</u>5.
- 703 Blake, D. R., and Rowland, F. S.: Urban leakage of liquefied petroleum gas and its
- 704 impact on Mexico-city air-quality, Science., 269, 953-956,
- 705 http://dx.doi.org10.1126/science.269.5226.953, 1995.
- Borbon, A., Locoge, N., Veillerot, M., Galloo, J. C., and Guillermo, R.: Characterisation
- of NMHCs in a French urban atmosphere: overview of the main sources, Sci. Total
- Environ., 292, 177-191, https://doi.org/10.1016/S0048-9697(01)01106-8, 2002.
- Cabada, J. C., Pandis, S. N., Subramanian, R., Robinson, A. L., Polidori, A., and Turpin,
- B.: Estimating the secondary organic aerosol contribution to PM_{2.5} using the EC
- 711 tracer method, Aerosol Sci. Technol., 38, 140-155,

- 712 http://dx.doi.org/10.1080/02786820390229084, 2004.
- Cai, C. J., Geng, F. H., Tie, X. X., Yu, Q., and An, J. L.: Characteristics and source
- apportionment of VOCs measured in Shanghai, China, Atmos. Environ., 44, 5005-
- 715 5014, http://dx.doi.org/10.1016/j.atmosenv.2010.07.059, 2010.
- Carter, W. P. L.: Update maximum incremental reactivity scale and hydrocarbon bin
- reactivities for regulatory application, California Air Resources Board Contract
- 718 07-339., 2009.
- Chang, C. C., Chen, T. Y., Chou, C., and Liu, S. C.: Assessment of traffic contribution
- to hydrocarbons using 2,2-dimethylbutane as a vehicular indicator, Terr. Atmos.
- Ocean. Sci., 15, 697-711, https://doi.org/10.3319/TAO.2004.15.4.697(A), 2004.
- Cheng, J., Zhang, Y. S., Wang, T., Xu, H., Norris, P., and Pan, W.-P.: Emission of
- volatile organic compounds (VOCs) during coal combustion at different heating
- rates, Fuel, 225, 554-562, https://doi.org/10.1016/j.fuel.2018.03.185, 2018.
- 725 Deng, W., Hu, Q. H., Liu, T. Y., Wang, X. M., Zhang, Y. L., Song, W., Sun, Y. L., Bi,
- 726 X. H., Yu, J. Z., Yang, W. Q., Huang, X. Y., Zhang, Z., Huang, Z. H., He, Q. F.,
- Mellouki, A., and George, C.: Primary particulate emissions and secondary
- organic aerosol (SOA) formation from idling diesel vehicle exhaust in China, Sci.
- 729 Total Environ., 593–594, 462-469, https://doi.org/10.1016/j.scitotenv.2017.03.088,
- 730 2017.
- 731 Fang, Z., Deng, W., Zhang Y. L., Ding, X., Tang, M. J., Liu, T. Y., Hu, Q. H., Zhu, M.,
- 732 Wang, Z. Y., Yang, W. Q., Huang, Z. H., Song, W., Bi, X. H, Chen, J. M., Sun, Y.
- L., George, C., and Wang, X. M.: Open burning of rice, corn and wheat straws:

- primary emissions, photochemical aging, and secondary organic aerosol formation,
- 735 Atmos. Chem. Phys., 17, 14821-14839, http://doi.org/10.5194/acp-17-14821-
- 736 2017, 2017.
- Forstner, H. J. L., Flagan, R. C., and Seinfeld, J. H.: Secondary organic aerosol from
- the photooxidation of aromatic hydrocarbons: Molecular composition, Environ.
- 739 Sci. Technol., 31, 1345-1358, http://doi.org/10.1021/es9605376, 1997.
- 740 Fu, X. X., Wang, X. M., Hu, Q. H., Li, G. H., Ding, X., Zhang, Y. L., He, Q. F., Liu, T.
- Y., Zhang, Z., Yu, Q. Q., Shen, R. Q., and Bi, X. H.: Changes in visibility with
- PM_{2.5} composition and relative humidity at a background site in the Pearl River
- Delta region, J. Environ. Sci., 40, 10-19, http://doi.org/10.1016/j.jes.2015.12.001,
- 744 2016.
- Garzón, J. P., Huertas, J. I., Magaña, M., Huertas, M. E., Cárdenas, B., Watanabe, T.,
- Maeda, T., Wakamatsu, S., and Blanco, S.: Volatile organic compounds in the
- 747 atmosphere of Mexico City, Atmos. Environ., 119, 415-429,
- 748 <u>https://doi.org/10.1016/j.atmo</u>senv.2015.08.014, 2015.
- Gentner, D. R., Jathar, S. H., Gordon, T. D., Bahreini, R., Day, D. A., El Haddad, I.,
- Hayes, P. L., Pieber, S. M., Platt, S. M., de Gouw, J., Goldstein, A. H., Harley, R.
- A., Jimenez, J. L., Prevot, A. S. H., and Robinson, A. L.: Review of urban
- secondary organic aerosol formation from gasoline and diesel motor vehicle
- 753 emissions, Environ. Sci. Technol., 51, 1074-1093,
- 754 http://dx.doi.org/10.1021/acs.est.6b04509, 2017.
- Guo, H., Wang, T., and Louie, P. K. K.: Source apportionment of ambient non-methane

- 756 hydrocarbons in Hong Kong: Application of a principal component
- analysis/absolute principal component scores (PCA/APCS) receptor model,
- 758 Environ. Pollut., 129, 489-498, http://dx.doi.org/10.1016/j.envpol.2003.11.006,
- 759 2004.
- Guo, H., So, K. L., Simpson, I. J., Barletta, B., Meinardi, S., and Blake, D. R.: C1–C8
- volatile organic compounds in the atmosphere of Hong Kong: Overview of
- atmospheric processing and source apportionment, Atmos. Environ., 41, 1456-
- 763 1472, http://dx.doi.org/10.1016/j.atmosenv.2006.10.011, 2007.
- 764 Guo, S., Hu, M., Zamora, M. L., Peng, J. F., Shang, D. J., Zheng, J., Du, Z. F., Wu, Z.
- J., Shao, M., Zeng, L. M., Molina, M. J., and Zhang, R. Y.: Elucidating severe
- urban haze formation in China, P. Natl. Acad. Sci. USA., 111, 17373-17378,
- 767 http://doi.org/10.1073/pnas.1419604111, 2014.
- Hao, J. M., and Wang, L. T.: Improving urban air quality in China: Beijing case study,
- 769 J. Air Waste Manage. Assoc., 55, 1298-1305,
- 770 http://dx.doi.org/10.1080/10473289.2005.10464726, 2005.
- Huang, K., Zhang, X. Y., and Lin, Y. F.: The "APEC Blue" phenomenon: Regional
- emission control effects observed from space, Atmos. Res., 164-165, 65-75,
- 773 http://dx.doi.org/10.1016/j.atmosres.2015.04.018, 2015.
- Huang, R. J., Zhang, Y. L., Bozzetti, C., Ho, K. F., Cao, J. J., Han, Y. M., Daellenbach,
- K. R., Slowik, J. G., Platt, S. M., Canonaco, F., Zotter, P., Wolf, R., Pieber, S. M.,
- Bruns, E. A., Crippa, M., Ciarelli, G., Piazzalunga, A., Schwikowski, M.,
- Abbaszade, G., Schnelle-Kreis, J., Zimmermann, R., An, Z. S., Szidat, S.,

- Baltensperger, U., El Haddad, I., and Prevot, A. S. H.: High secondary aerosol
- contribution to particulate pollution during haze events in China, Nature., 514,
- 780 218-222, http://doi.org/10.1038/nature13774, 2014.
- Huang, X. F., He, L. Y., Hu, M., Canagaratna, M.R., Sun, Y., Zhang, Q., Zhu, T., Xue,
- L., Zeng, L. W., Liu, X. G.: Highly time-resolved chemical characterization of
- atmospheric submicron particles during 2008 Beijing Olympic Games using an
- Aerodyne High-Resolution Aerosol Mass Spectrometer, Atmos. Chem. Phys., 10,
- 785 8933-8945, http://dx.doi.org/10.5194/acp-10-8933-2010, 2010.
- 786 Huang, X. Y., Zhang, Y. L., Yang, W. Q., Huang, Z. Z., Wang, Y. J., Zhang, Z., He, Q.
- F., Lü, S. J., Huang, Z. H., Bi, X. H., Wang, X. M.: Effect of traffic restriction on
- reducing ambient volatile organic compounds (VOCs): observation-based
- evaluation during a traffic restriction drill in Guangzhou, China, Atmos. Environ.,
- 790 161, 61-70, http://dx.doi.org/10.1016/j.atmosenv.2017.04.035, 2017.
- 791 Ji, D. S., Wang, Y. S., Wang, L. L., Chen, L. F., Hu, B., Tang, G. Q., Xin, J. Y., Song,
- T., Wen, T. X., Sun, Y., Pan, Y. P., and Liu, Z. R.: Analysis of heavy pollution
- episodes in selected cities of northern China, Atmos. Environ., 50, 338-348,
- 794 http://dx.doi.org/10.1016/j.atmosenv.2011.11.053, 2012.
- Juda-Rezler, K., Reizer, M., and Oudinet, J. P.: Determination and analysis of PM₁₀
- source apportionment during episodes of air pollution in Central Eastern European
- urban areas: The case of wintertime 2006, Atmos. Environ., 45, 6557-6566,
- 798 <u>https://doi.org/10.1016/j.atmosenv.2011.08.020, 2011.</u>
- Kelly, F. J., and Zhu, T.: Transport solutions for cleaner air, Science., 352, 934-936,

- 800 http://doi.org/10.1126/science.aaf3420, 2016.
- Lau, A. K. H., Yuan, Z. B., Yu, J. Z., and Louie, P. K. K.: Source apportionment of
- ambient volatile organic compounds in Hong Kong, Sci. Total Environ., 408,
- 4138-4149, http://dx.doi.org/10.1016/j.scitotenv.2010.05.025, 2010.
- Lelieveld, J., Evans, J. S., Fnais, M., Giannadaki, D., and Pozzer, A.: The contribution
- of outdoor air pollution sources to premature mortality on a global scale, Nature,
- 525, 367-371, https://doi.org/10.1038/nature15371, 2015.
- Leuchner, M., and Rappenglück, B.: VOC source–receptor relationships in Houston
- 808 during TexAQS-II, Atmos. Environ., 44, 4056-4067,
- https://doi.org/10.1016/j.atmosenv.2009.02.029, 2010.
- Li, H. Y., Zhang, Q., Zhang, Q., Chen, C. R., Wang, L. T., Wei, Z., Zhou, S., Parworth,
- 811 C., Zheng, B., Canonaco, F., Prevot, A. S. H., Chen, P., Zhang, H. L., Wallington,
- T. J., and He, K. B.: Wintertime aerosol chemistry and haze evolution in an
- extremely polluted city of the North China Plain: Significant contribution from
- coal and biomass combustion, Atmos. Chem. Phys., 17, 4751-4768,
- http://dx.doi.org/10.5194/acp-17-4751-2017, 2017.
- Li, J., Xie, S. D., Zeng, L. M., Li, L. Y., Li, Y. Q., and Wu, R. R.: Characterization of
- ambient volatile organic compounds and their sources in Beijing, before, during,
- and after Asia-Pacific Economic Cooperation China 2014, Atmos. Chem. Phys.,
- 15, 7945-7959, http://dx.doi.org/10.5194/acp-15-7945-2015, 2015.
- 820 Liakakou, E., Bonsang, B., Williams, J., Kalivitis, N., Kanakidou, M., and
- Mihalopoulos, N.: C2–C8 NMHCs over the Eastern Mediterranean: Seasonal

- variation and impact on regional oxidation chemistry, Atmos. Environ., 43, 5611-
- 823 <u>5621</u>, https://doi.org/10.1016/j.atmosenv.2009.07.067, 2009.
- 824 Lim, S. S et al.: A comparative risk assessment of burden of disease and injury
- attributable to 67 risk factors and risk factor clusters in 21 regions, 1990-2010: a
- systematic analysis for the Global Burden of Disease Study 2010, Lancet, 380,
- 827 2224-2260, http://dx.doi.org/10.1016/s0140-6736(12)61766-8, 2012.
- Lim, Y. B., and Ziemann, P. J.: Effects of molecular structure on aerosol yields from
- OH radical-initiated reactions of linear, branched, and cyclic alkanes in the
- presence of NOx, Environ. Sci. Technol., 43, 2328-2334,
- http://dx.doi.org/10.1021/es803389s, 2009.
- Liu, C. T., Zhang, C. L., Mu, Y. J., Liu, J. F., and Zhang, Y. Y.: Emission of volatile
- organic compounds from domestic coal stove with the actual alternation of flaming
- and smoldering combustion processes, Environ. Pollut., 221, 385-391,
- http://dx.doi.org/10.1016/j.envpol.2016.11.089, 2017.
- 836 Liu, J. F., Mu, Y. J., Zhang, Y. J., Zhang, Z. M., Wang, X. K., Liu, Y. J., and Sun, Z. Q.:
- Atmospheric levels of BTEX compounds during the 2008 Olympic Games in the
- urban area of Beijing, Sci. Total Environ., 408, 109-116,
- http://dx.doi.org/10.1016/j.scitotenv.2009.09.026, 2009.
- Liu, J., Mauzerall, D. L., Chen, Q., Zhang, Q., Song, Y., Peng, W., Klimont, Z., Qiu, X.
- H., Zhang, S. Q., Hu, M., Lin, W. L., Smith, K. R., and Zhu, T.: Air pollutant
- emissions from Chinese households: A major and underappreciated ambient
- 843 pollution source, P. Natl. Acad. Sci. USA., 113, 7756-7761,

- http://dx.doi.org/10.1073/pnas.1604537113, 2016.
- Liu, J. G., Xie, P. H., Wang, Y. S., Wang, Z. F., He, H., and Liu, W. Q.: Haze observation
- and control measure evaluation in Jing-Jin-Ji (Beijing, Tianjin, Hebei) area during
- the period of the Asia-Pacific Economic Cooperation (APEC) Meeting, Bulletin
- 848 of Chinese Academy of Sciences., 30, 368-377,
- http://dx.doi.org/10.16418/j.issn.1000-3045.2015.03.011, 2015b., (in Chinese).
- 850 Liu, K. K., Quan, J. N., Mu, Y. J., Zhang, Q., Liu, J. F., Gao, Y., Chen, P. F., Zhao, D.
- L., and Tian, H. J.: Aircraft measurements of BTEX compounds around Beijing
- 852 city, Atmos. Environ., 73, 11-15,
- http://dx.doi.org/10.1016/j.atmosenv.2013.02.050, 2013.
- Liu, K. K., Zhang, C. L., Cheng, Y., Liu, C. T., Zhang, H. X., Zhang, G., Sun, X., and
- Mu, Y. J.: Serious BTEX pollution in rural area of the North China Plain during
- 856 winter season, J. Environ. Sci., 30, 186-190,
- http://dx.doi.org/10.1016/j.jes.2014.05.056, 2015d.
- 858 Liu, T., Wang, X., Deng, W., Hu, Q., Ding, X., Zhang, Y., He, Q., Zhang, Z., Lü, S., Bi,
- X., Chen, J., and Yu, J.: Secondary organic aerosol formation from photochemical
- aging of light-duty gasoline vehicle exhausts in a smog chamber, Atmos. Chem.
- Phys., 15, 9049-9062, http://dx.doi.org/10.5194/acp-15-9049-2015, 2015c.
- Liu, Y., Shao, M., Fu, L. L., Lu, S. H., Zeng, L. M., and Tang, D. G.: Source profiles of
- volatile organic compounds (VOCs) measured in China: Part I, Atmos. Environ.,
- 42, 6247-6260, http://dx.doi.org/10.1016/j.atmosenv.2008.01.070, 2008.
- Liu, Z. R., Hu, B., Wang, L. L., Wu, F. K., Gao, W. K., and Wang, Y. S.: Seasonal and

- diurnal variation in particulate matter (PM₁₀ and PM_{2.5}) at an urban site of Beijing:
- Analyses from a 9-year study, Environ. Sci. Pollut. R., 22, 627-642,
- 868 http://dx.doi.org/10.1007/s11356-014-3347-0, 2015a.
- Lonati, G., Giugliano, M., Butelli, P., Romele, L., and Tardivo, R.: Major chemical
- components of PM_{2.5} in Milan (Italy), Atmos. Environ., 39, 1925-1934,
- http://dx.doi.org/10.1016/j.atmosenv.2004.12.012, 2005.
- Lu, S. H., Liu, Y., Shao, M., and Huang, S.: Chemical speciation and anthropogenic
- sources of ambient volatile organic compounds (VOCs) during summer in Beijing,
- 874 2004, Front. Environ. Sci. Engin. China., 1, 147-152,
- http://dx.doi.org/10.1007/s11783-007-0026-0, 2007.
- Martin, W. J., Glass, R. I., Balbus, J. M., and Collins, F. S.: A major environmental
- cause of death, Science, 334, 180-181, http://dx.doi.org/10.1126/science.1213088,
- 878 <u>2</u>011.
- McDonald, B. C., Goldstein, A. H., and Harley, R. A.: Long-term trends in California
- mobile source emissions and ambient concentrations of black carbon and organic
- 881 aerosol, Environ. Sci. Technol., 49, 5178-5188,
- http://dx.doi.org/10.1021/es505912b, 2015.
- Ng, N. L., Kroll, J. H., Chan, A. W. H., Chhabra, P. S., Flagan, R. C., and Seinfeld, J.
- H.: Secondary organic aerosol formation from m-xylene, toluene, and benzene,
- 885 Atmos. Chem. Phys., 7, 3909-3922, http://doi.org/10.5194/acp-7-3909-2007,
- 886 2007.
- 887 O'Dowd, C. D., Aalto, P., Hameri, K., Kulmala, M., and Hoffmann, T.: Aerosol

- formation Atmospheric particles from organic vapours, Nature., 416, 497-498,
- http://doi.org./10.1038/416497a, 2002.
- Odum, J. R., Jungkamp, T. P. W., Griffin, R. J., Flagan, R. C., and Seinfeld, J. H.: The
- atmospheric aerosol-forming potential of whole gasoline vapor, Science., 276, 96-
- 892 99, http://doi.org./10.1126/science.276.5309.96, 1997.
- 893 Ortega, A. M., Hayes, P. L., Peng, Z., Palm, B. B., Hu, W. W., Day, D. A., Li, R.,
- Cubison, M. J., Brune, W. H., Graus, M., Warneke, C., Gilman, J. B., Kuster, W.
- 895 C., de Gouw, J., Gutierrez-Montes, C., and Jimenez, J. L.: Real-time
- measurements of secondary organic aerosol formation and aging from ambient air
- in an oxidation flow reactor in the Los Angeles area, Atmos. Chem. Phys., 16,
- 898 7411-7433, http://dx.doi.org/10.5194/acp-16-7411-2016, 2016.
- 899 Ou, J. M., Guo, H., Zheng, J. Y., Cheung, K., Louie, P. K. K., Ling, Z. H., and Wang,
- D. W.: Concentrations and sources of non-methane hydrocarbons (NMHCs) from
- 2005 to 2013 in Hong Kong: A multi-year real-time data analysis, Atmos. Environ.,
- 902 103, 196-206, http://dx.doi.org/10.1016/j.atmosenv.2014.12.048, 2015.
- Paatero, P., and Tapper, U.: Positive matrix factorization: A non-negative factor model
- with optimal utilization of error estimates of data values, Environmetrics., 5, 111-
- 905 126, http://dx.doi.org/10.1002/env.3170050203, 1994.
- 906 Paatero, P.: Least squares formulation of robust non-negative factor analysis,
- 907 Chemometr. Intell. Lab., 37, 23-35, http://dx.doi.org/10.1016/S0169-
- 908 7439(96)00044-5, 1997.
- Parrish, D. D., Kuster, W. C., Shao, M., Yokouchi, Y., Kondo, Y., Goldan, P. D., de

- Gouw, J. A., Koike, M., and Shirai, T.: Comparison of air pollutant emissions
- 911 among mega-cities, Atmos. Environ., 43, 6435-6441,
- 912 http://dx.doi.org/10.1016/j.atmosenv.2009.06.024, 2009.
- 913 Saito, S., Nagao, I., and Kanzawa, H.: Characteristics of ambient C2–C11 non-methane
- hydrocarbons in metropolitan Nagoya, Japan, Atmos. Environ., 43, 4384-4395,
- 915 <u>https://doi.org/10.1016/j.atmosenv.2009.04.031, 2009.</u>
- 916 Sari, D., and Bayram, A.: Quantification of emissions from domestic heating in
- residential areas of İzmir, Turkey and assessment of the impact on local/regional
- 918 air-quality, Sci. Total Environ., 488-489, 429-436,
- 919 <u>https://doi.org/10.1016/j.scitotenv.2013.11.033, 2014.</u>
- 920 Sato, K., Takami, A., Isozaki, T., Hikida, T., Shimono, A., and Imamura, T.: Mass
- spectrometric study of secondary organic aerosol formed from the photo-oxidation
- of aromatic hydrocarbons, Atmos. Environ., 44, 1080-1087,
- 923 http://dx.doi.org/10.1016/j.atmosenv.2009.12.013, 2010.
- Shao, M., Lu, S. H., Liu, Y., Xie, X., Chang, C. C., Huang, S., and Chen, Z. M.: Volatile
- organic compounds measured in summer in Beijing and their role in ground-level
- ozone formation, J. Geophys. Res.-Atmos., 114, 13,
- 927 http://dx.doi.org/10.1029/2008jd010863, 2009.
- 928 Shrivastava, M., Easter, R. C., Liu, X. H., Zelenyuk, A., Singh, B., Zhang, K., Ma, P.-
- L., Chand, D., Ghan, S., Jimenez, J. L., Zhang, Q., Fast, J., Rasch, P. J., and Tiitta,
- 930 P.: Global transformation and fate of SOA: Implications of low-volatility SOA and
- gas-phase fragmentation reactions, J. Geophys. Res.-Atmos., 120, 4169-4195,

- 932 http://dx.doi.org/10.1002/2014jd022563, 2015.
- Song, Y., Shao, M., Liu, Y., Lu, S. H., Kuster, W., Goldan, P., and Xie, S. D.: Source
- apportionment of ambient volatile organic compounds in Beijing, Environ. Sci.
- 935 Technol., 41, 4348-4353, http://dx.doi.org/10.1021/es0625982, 2007.
- 936 Streets, D. G., Fu, J. S., Jang, C. J., Hao, J. M., He, K. B., Tang, X. Y., Zhang, Y. H.,
- Wang, Z. F., Li, Z. P., Zhang, Q., Wang, L. T., Wang, B. Y., and Yu, C.: Air quality
- during the 2008 Beijing Olympic Games, Atmos. Environ., 41, 480-492,
- 939 https://doi.org/10.1016/j.atmosenv.2006.08.046, 2007.
- 940 Subramanian, M.: Deadly dinners, Nature, 509, 548-551,
- 941 http://doi.org/10.1038/509548a, 2014.
- 942 Suthawaree, J., Kato, S., Okuzawa, K., Kanaya, Y., Pochanart, P., Akimoto, H., Wang,
- 2., and Kajii, Y.: Measurements of volatile organic compounds in the middle of
- Central East China during Mount Tai Experiment 2006 (MTX2006): Observation
- of regional background and impact of biomass burning, Atmos. Chem. Phys., 10,
- 946 1269-1285, http://doi.org/10.5194/acp-10-1269-2010, 2010.
- Tang, X., Wang, Z. F., Zhu, J., Gbaguidi, A. E., Wu, Q. Z, Li, J., and Zhu, T.: Sensitivity
- of ozone to precursor emissions in urban Beijing with a Monte Carlo scheme,
- 949 Atmos. Environ., 44, 3833-3842, http://dx.doi.org/10.1016/j.atmosenv.
- 950 2010.06.026, 2010.
- 951 Wang, B., Shao, M., Lu, S. H., Yuan, B., Zhao, Y., Wang, M., Zhang, S. Q., and Wu, D.:
- Variation of ambient non-methane hydrocarbons in Beijing city in summer 2008,
- 953 Atmos. Chem. Phys., 10, 5911-5923, http://dx.doi.org/10.5194/acp-10-5911-2010,

- 954 2010a.
- 955 Wang, M., Zhu, T., Zheng, J., Zhang, R. Y., Zhang, S. Q., Xie, X. X., Han, Y. Q., and
- Li, Y.: Use of a mobile laboratory to evaluate changes in on-road air pollutants
- during the Beijing 2008 Summer Olympics, Atmos. Chem. Phys., 9, 8247-8263,
- 958 http://dx.doi.org/10.5194/acp-9-8247-2009, 2009.
- 959 Wang, M., Shao, M., Lu, S. H., Yang, Y. D., and Chen, W. T.: Evidence of coal
- combustion contribution to ambient VOCs during winter in Beijing, Chin. Chem.
- 961 Lett., 24, 829-832, http://dx.doi.org/10.1016/j.cclet.2013.05.029, 2013.
- 962 Wang, S. W., Zhao, M., Xing, J., Wu, Y., Zhou, Y., Lei, Y., He, K. B., Fu, L. X., Hao, J.
- M.: Quantifying the Air Pollution Emission Reduction during the 2008 Olymic
- Games in Beijing, Environ. Sci. Technol., 44, 2490-2496, doi:10.1021/es9028167,
- 965 2010b.
- 966 Wang, X. M., Sheng, G. Y., Fu, J. M., Chan, C. Y., Lee, S. G., Chan, L. Y., and Wang,
- 2. S.: Urban roadside aromatic hydrocarbons in three cities of the Pearl River Delta,
- People's Republic of China, Atmos. Environ., 36, 5141-5148,
- 969 http://dx.doi.org/10.1016/s1352-2310(02)00640-4, 2002.
- 970 Wang, Y. S., Yao, L., Wang, L. L., Liu, Z. R., Ji, D. S., Tang, G. Q., Zhang, J. K., Sun,
- Y., Hu, B., and Xin, J. Y.: Mechanism for the formation of the January 2013 heavy
- haze pollution episode over central and eastern China, Sci. China Earth Sci., 57,
- 973 14-25, http://dx.doi.org/10.1007/s11430-013-4773-4, 2014.
- 974 Wang, Z. S., Li, Y. T., Chen, T., Li, L. J., Liu, B. X., Zhang, D. W., Sun, F., Wei, Q.,
- Jiang, L., and Pan, L. B.: Changes in atmospheric composition during the 2014

- APEC conference in Beijing, J. Geophys. Res.-Atmos., 120, 12695-12707,
- 977 http://dx.doi.org/10.1002/2015jd023652, 2015.
- 978 Xu, J. Y., Zhuo, J. K., Zhu, Y. N., Pan, Y., and Yao, Q.: Analysis of volatile organic
- 979 pyrolysis products of bituminous and anthracite coals with single-photon
- 980 <u>ionization time-of-flight mass spectrometry and gas chromatography/mass</u>
- 981 spectrometry, Energ Fuel, 31, 730-737,
- 982 <u>https://doi.org/10.1021/acs.energyfuels.6b02335, 2017.</u>
- 983 Xu, W. Q., Sun, Y. L., Chen, C., Du, W., Han, T. T., Wang, Q. Q., Fu, P. Q., Wang, Z.
- F., Zhao, X. J., Zhou, L. B., Ji, D. S., Wang, P. C., and Worsnop, D. R.: Aerosol
- composition, oxidation properties, and sources in Beijing: Results from the 2014
- Asia-Pacific Economic Cooperation summit study, Atmos. Chem. Phys., 15,
- 987 13681-13698, http://dx.doi.org/10.5194/acp-15-13681-2015, 2015.
- 988 <u>Xue, Y. F., Zhou, Z., Nie, T., Wang, K., Nie, L., Pan, T., Wu, X. Q., Tian, H. Z., Zhong,</u>
- 289 <u>L. H., Li, J., Liu, H. J., Liu, S. H., and Shao, P. Y.: Trends of multiple air pollutants</u>
- emissions from residential coal combustion in Beijing and its implication on
- 991 <u>improving air quality for control measures, Atmospheric Environment, 142, 303-</u>
- 992 312, 10.1016/j.atmosenv.2016.08.004, 2016.
- 993 Yao, Z. L., Zhang, Y. Z., Shen, X. B., Wang, X. T., Wu, Y., He, K. B.: Impacts of
- 994 temporary traffic control measures on vehicular emissions during the Asian Games
- 995 in Guangzhou, China, J. Air Waste Manag. Assoc., 63, 11-19,
- 996 https://doi.org/10.1080/10962247.2012.724041, 2013.
- 997 Yokelson, R. J., Christian, T. J., Karl, T. G., and Guenther, A.: The tropical forest and
- 998 fire emissions experiment: Laboratory fire measurements and synthesis of

- 999 campaign data, Atmos. Chem. Phys., 8, 3509-3527, http://dx.doi.org/10.5194/acp-
- 1000 8-3509-2008, 2008.
- 1001 Yu, Q. Q., Yang, W. Q., Zhu, M., Gao, B., Li, S., Li, G. H., Fang, H., Zhou, H. S., Zhang,
- 1002 H. N., Wu, Z. F., Song, W., Tan, J. H., Zhang, Y. L., Bi, X. H., Chen, L. G., and
- Wang, X. M.: Ambient PM_{2.5}-bound polycyclic aromatic hydrocarbons (PAHs) in
- rural Beijing: Unabated with enhanced temporary emission control during the
- 2014 APEC summit and largely aggravated after the start of wintertime heating,
- Environ. Pollut., 238, 532-542, http://dx.doi.org/10.1016/j.envpol.2018.03.079,
- 1007 **2018**.
- Yuan, B., Shao, M., Lu, S. H., and Wang, B.: Source profiles of volatile organic
- compounds associated with solvent use in Beijing, China, Atmos. Environ., 44,
- 1010 1919-1926, http://dx.doi.org/10.1016/j.atmosenv.2010.02.014, 2010.
- 1011 Yuan, Z. B., Lau, A. K. H., Shao, M., Louie, P. K. K., Liu, S. C., and Zhu, T.: Source
- analysis of volatile organic compounds by positive matrix factorization in urban
- and rural environments in Beijing, J. Geophys. Res., 114, 14,
- 1014 http://dx.doi.org/10.1029/2008jd011190, 2009.
- Yuan, Z. B., Zhong, L. J., Lau, A. K. H., Yu, J. Z., and Louie, P. K. K.: Volatile organic
- 1016 compounds in the Pearl River Delta: Identification of source regions and
- recommendations for emission-oriented monitoring strategies, Atmos. Environ.,
- 76, 162-172, http://dx.doi.org/10.1016/j.atmosenv.2012.11.034, 2013.
- Zhang, J. K., Sun, Y., Wu, F. K., Sun, J., and Wang, Y. S.: The characteristics, seasonal
- variation and source apportionment of VOCs at Gongga Mountain, China, Atmos.

- Environ., 88, 297-305, http://dx.doi.org/10.1016/j.atmosenv.2013.03.036, 2014c.
- 1022 Zhang, J. K., Sun, Y., Liu, Z. R., Ji, D. S., Hu, B., Liu, Q., and Wang, Y. S.:
- 1023 Characterization of submicron aerosols during a month of serious pollution in
- Beijing, 2013, Atmos. Chem. Phys., 14, 2887-2903, http://dx.doi.org/10.5194/acp-
- 1025 14-2887-2014, 2014a.
- Zhang, J. K., Wang, L. L., Wang, Y. H., and Wang, Y. S.: Submicron aerosols during
- the Beijing Asia-Pacific Economic Cooperation conference in 2014, Atmos.
- Environ., 124, 224-231, http://dx.doi.org/10.1016/j.atmosenv.2015.06.049, 2016b.
- Zhang, Q., He, K. B., and Huo, H.: Policy: Cleaning China's air, Nature., 484, 161-162,
- 1030 http://dx.doi.org/10.1038/484161a, 2012b.
- Zhang, Y. J., Mu, Y. J., Liu, J. F., and Mellouki, A.: Levels, sources and health risks of
- carbonyls and BTEX in the ambient air of Beijing, China, J. Environ. Sci., 24,
- 1033 124-130, http://dx.doi.org/10.1016/S1001-0742(11)60735-3, 2012a.
- 1034 Zhang, Y. J., Mu, Y. J., Meng, F., Li, H., Wang, X. Z., Zhang, W. Q., Mellouki, A., Gao,
- J., Zhang, X.M., Wang, S.L., and Chai, F.H.: The pollution levels of BTEX and
- carbonyls under haze and non-haze days in Beijing, China, Sci. Total Environ.,
- 490, 391-396, http://dx.doi.org/10.1016/j.scitotenv.2014.05.025, 2014b.
- 1038 Zhang, Y. S., Shao, M., Lin, Y., Luan, S. J., Mao, N., Chen, W. T., and Wang, M.:
- Emission inventory of carbonaceous pollutants from biomass burning in the Pearl
- River Delta Region, China, Atmos. Environ., 76, 189-199,
- 1041 http://dx.doi.org/10.1016/j.atmosenv.2012.05.055, 2013c.
- Thang, Y. L., Wang, X. M., Barletta, B., Simpson, I. J., Blake, D. R., Fu, X. X., Zhang,

- Z., He, Q. F., Liu, T. Y., Zhao, X. Y., and Ding, X.: Source attributions of hazardous
- aromatic hydrocarbons in urban, suburban and rural areas in the Pearl River Delta
- 1045 (PRD) region, J. Hazard. Mater., 250–251, 403-411,
- 1046 http://dx.doi.org/10.1016/j.jhazmat.2013.02.023, 2013a.
- 1047 Zhang, Y. L., Wang, X. M., Zhang, Z., Lü, S. J., Shao, M., Lee, F. S. C., and Yu, J. Z.:
- Species profiles and normalized reactivity of volatile organic compounds from
- gasoline evaporation in China, Atmos. Environ., 79, 110-118,
- 1050 http://dx.doi.org/10.1016/j.atmosenv.2013.06.029, 2013b.
- Zhang, Y. L., Wang, X. M., Zhang, Z., Lü, S. J, Huang, Z. H., and Li, L. F.: Sources of
- 1052 C2–C4 alkenes, the most important ozone nonmethane hydrocarbon precursors in
- the Pearl River Delta region, Sci. Total Environ., 502, 236-245,
- 1054 http://dx.doi.org/10.1016/j.scitotenv.2014.09.024, 2015b.
- 1055 Zhang, Y. L., Wang, X. M., Blake, D. R., Li, L. F., Zhang, Z., Wang, S. Y., Guo, H.,
- Lee, F. S. C., Gao, B., Chan, L. Y., Wu, D., and Rowland, F. S.: Aromatic
- hydrocarbons as ozone precursors before and after outbreak of the 2008 financial
- crisis in the Pearl River Delta region, south China, J. Geophys. Res.-Atmos., 117,
- 1059 16, http://dx.doi.org/10.1029/2011jd017356, 2012c.
- Zhang, Z., Wang, X. M., Zhang, Y. L., Lü, S. J., Huang, Z. H., Huang, X. Y., and Wang,
- 1061 Y. S.: Ambient air benzene at background sites in China's most developed coastal
- regions: Exposure levels, source implications and health risks, Sci. Total Environ.,
- 511, 792-800, http://dx.doi.org/10.1016/j.scitotenv.2015.01.003, 2015a.
- 1064 Zhang, Z., Zhang, Y. L., Wang, X. M., Lü, S. J., Huang, Z. H., Huang, X. Y., Yang, W.

- Q., Wang, Y. S., and Zhang, Q.: Spatiotemporal patterns and source implications
- of aromatic hydrocarbons at six rural sites across China's developed coastal
- 1067 regions, J. Geophys. Res.-Atmos., 121, 6669-6687,
- 1068 http://dx.doi.org/10.1002/2016jd025115, 2016a.
- Zhao, Y. L., Hennigan, C. J., May, A. A., Tkacik, D. S., de Gouw, J. A., Gilman, J. B.,
- 1070 Kuster, W. C., Borbon, A., and Robinson, A. L.: Intermediate-volatility organic
- 1071 compounds: a large source of secondary organic aerosol, Environ. Sci. Technol.,
- 48, 13743-13750, doi:10.1021/es5035188, 2014.
- Zhao, Y. L., Nguyen, N. T., Presto, A. A., Hennigan, C. J., May, A. A., and Robinson,
- 1074 A. L.: Intermediate volatility organic compound emissions from on-road diesel
- vehicles: Chemical composition, emission factors, and estimated secondary
- organic aerosol production, Environ. Sci. Technol., 49, 11516-11526,
- doi:10.1021/acs.est.5b02841, 2015.
- 1078 Zheng, J. Y., Yu, Y. F., Mo, Z. W., Zhang, Z., Wang, X. M., Yin, S. S., Peng, K., Yang,
- 1079 Y., Feng, X. Q., and Cai, H. H.: Industrial sector-based volatile organic compound
- 1080 (VOC) source profiles measured in manufacturing facilities in the Pearl River
- 1081 Delta, China, Sci. Total Environ., 456–457, 127-136,
- http://dx.doi.org/10.1016/j.scitotenv.2013.03.055, 2013.
- 083 Zou, Y., Deng, X. J., Zhu, D., Gong, D. C., Wang, H., Li, F., Tan, H. B., Deng, T., Mai,
- B. R., Liu, X. T., and Wang, B. G.: Characteristics of 1 year of observational data
- of VOCs, NOx and O₃ at a suburban site in Guangzhou, China, Atmos. Chem.
- Phys., 15, 6625-6636, 10.5194/acp-15-6625-2015, 2015.

Table 1. Enhanced temporary air pollution control measures during the 2014 Asian-Pacific Economic Cooperation (APEC) summit.

Control types	Details	Control areas
Traffic	 1.—Yellow label vehicles were banned to run inside the sixth ring of Beijing and the Huairou urban area; 2.—The number of private vehicles in operation reduced by 50% through an odd/even-number-plate rule throughout Beijing; 3.—Trucks were limited to drive inside the sixth ring of Beijing between 6 am and 24 pm; 4.—70% of government cars also ordered off the 	Inside the sixth ring of Beijing an the Huairou urban area
I., d.,	roads in Beijing.	D-:::
Industrial	 1) 1.—9289 enterprises were suspended, 3900 enterprises were ordered to limit production; 2) 2.—More than 40000 construction sites were shut down. 	Beijing, some areas of Tianjin, Hebei, Shanxi, Shandong and Inner Mongolia
Other fields	 1.—Open fire was completely controlled at North China Plain; 2.—Increasing road cleaning and water spraying in Beijing; 3.—Other relate control measures carried out in surrounding areas. 	North China Plain Tianjin, Hebei, Shanxi, Shandong and Inner Mongolia

Table 2. The mixing ratios, ranges and 95% confidence intervals (95% C.I.) of VOCs during period I, II and III at the rural site inside UCAS (in parts per trillion by volume, pptv).

Species	MDLa		Period I		Period II		Period III
		Range	Mean (95% C.I.)	Range	Mean (95% C.I.)	Range	Mean (95% C.I.)
Ethane	39	1172-7855	` /		2442(491)	1082-12714	3674(465)
Propane	31	427-6145	2880(720)		1296(384)	598-7604	2479(329)
i-Butane	14	53-2755	1121(312)	59-1400	474(187)	106-2741	754(129)
n-Butane	21	158-2947	1283(302)	83-1735	562(196)	174-3047	841(136)
i-Pentane	14	94-3729	1425(354)	39-1388	561(167)	72-12590	1018(279)
n-Pentane	8	47-1697	615(182)	30-910	247(106)	26-4808	456(112)
2,2-Dimethylbutane	14	15-68	30(6)	17-32	24(3)	BDL ^b -75	26(2)
Cyclopentane	12	13-135	64(15)	BDL-64	35(7)	13-274	50(8)
2,3-Dimethylbutane	12	13-140	45(15)	22-51	32(4)	13-235	38(6)
2-Methylpentane	8	13-679	171(68)	12-257	77(31)	9-1077	124(27)
3-Methylpentane	7	12-548	150(54)	14-220	68(26)	16-792	104(20)
n-Hexane	6	115-1033	505(97)	102-921	324(89)	108-7393	1400(257)
Methylcyclopentane	9	10-283	100(30)	13-195	59(23)	BDL-535	88(17)
2,4-Dimethylpentane	4	BDL-43	15(5)	BDL-15	10(2)	BDL-90	16(2)
Cyclohexane	6	10-458	167(51)	10-107	43(14)	7-646	76(17)
2-Methylhexane	6	10-304	68(27)	7-56	22(6)	7-318	51(10)
2,3-Dimethylpentane	9	BDL-139	31(12)	BDL-24	15(2)	BDL-102	28(3)
3-Methylhexane	6	12-436	93(38)	8-97	41(11)	9-367	70(12)
2,2,4-Trimethylpentane	9	12-126	44(12)	BDL-41	25(4)	BDL-127	38(5)
n-Heptane	10	12-358	89(33)	12-71	30(8)	13-441	82(14)
Methylcyclohexane	5	BDL-162	51(17)	BDL-66	21(7)	BDL-162	44(8)
2,3,4-Trimethylpentane	6	BDL-38	14(4)	BDL-12	9(1)	BDL-59	16(2)
2-Methylheptane	4	8-175	31(16)	BDL-31	13(3)	BDL-91	22(3)
3-Methylheptane	5	BDL-231	26(20)	BDL-15	8(1)	BDL-74	17(2)
n-Octane	6	8-104	42(11)	BDL-31	18(3)	BDL-160	40(6)
n-Nonane	6	9-99	40(11)	BDL-37	18(4)	BDL-171	38(6)
n-Decane	6	14-777	129(74)	8-110	36(14)	BDL-600	73(17)
n-Undecane	7	47-317	151(35)	27-206	66(20)	11-374	94(12)
n-Dodecane	7	9-646	129(57)	25-313	75(30)	8-316	63(9)
Ethylene	41	367-3495	1788(391)	553-3572	1254(352)	319-13911	2313(428)
Propene	31	117-1264	430(118)	170-766	371(67)	176-3222	820(128)
1-Butene	17	19-161	107(18)	BDL-100	55(12)	19-581	137(22)
1,3-Butadiene	20	21-403	154(44)	23-234	79(27)	BDL-2140	252(74)
trans-2-Butene	5	BDL-41	18(4)	BDL-35	12(4)	BDL-425	39(10)
cis-2-Butene	7	9-50	23(4)	BDL-38	14(5)	BDL-276	37(7)
1-Pentene	20	BDL-47	33(3)	21-25	23(1)	BDL-127	52(6)
Isoprene	13	BDL-623	163(56)	16-143	62(15)	17-765	200(24)
trans-2-Pentene	10	BDL-37	17(4)	BDL-19	15(3)	BDL-65	23(3)
cis-2-Pentene	6	BDL-24	11(3)	BDL-9	8(0)	BDL-46	15(2)
2-Methyl-2-butene	12	13-50	21(4)	17-20	18(1)	BDL-61	24(2)
Benzene	14	75-2695	868(279)	43-1465	410(179)	72-2916	795(151)
Toluene	9	120-3585	1273(419)	47-1186	343(126)	62-3425	840(146)
Ethylbenzene	6	25-2210	684(240)	12-611	145(67)	23-2450	317(75)
m/p-Xylene	9	39-2106	765(248)	16-620	149(67)	25-3285	422(91)
Styrene	8	15-578	167(71)	BDL-99	32(11)	10-1267	151(38)
o-Xylene	4	11-965	334(104)	9-284	71(31)	15-1224	178(36)
Isopropylbenzene	4	5-66	24(7)	BDL-21	11(2)	BDL-77	18(3)
n-Propylbenzene	4	6-231	71(27)	BDL-55	20(7)	5-239	38(8)
m-Ethyltoluene	3	13-593	136(67)	4-91	28(11)	4-854	85(23)
p-Ethyltoluene	3	6-295	61(29)	4-59	17(6)	4-245	41(9)
1,3,5-Trimethylbenzene	3	7-217	48(21)	BDL-35	12(4)	4-179	38(6)
o-Ethyltoluene	3	5-246	64(26)	4-58	17(6)	5-230	40(8)
1,2,4-Trimethylbenzene	6	22-984	220(93)	13-219	58(22)	8-803	127(26)
1,2,3-Trimethylbenzene	5	12-442	82(37)	BDL-92	24(9)	6-300	56(11)
1,3-Diethylbenzene	4	11-135	35(12)	BDL-26	15(3)	BDL-126	26(4)
1,4-Diethylbenzene	4	14-461	80(40)	5-69	23(7)	BDL-292	51(10)
1,2-Diethylbenzene	4	BDL-30	15(4)	BDL-8	6(1)	BDL-76	15(2)
Ethyne	57	406-10539	3128(1043)	290-6260	1625(615)	584-10378	3008(509)

^a MDL, method detection limits, pptv; ^b BDL, bellowed detection limit.

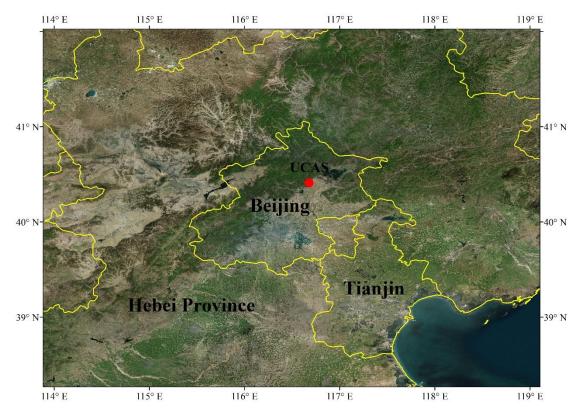


Figure 1. Location of the rural sampling site at a rural inside the campus of University of Chinese Academy of Science (UCAS) in the north Beijing.

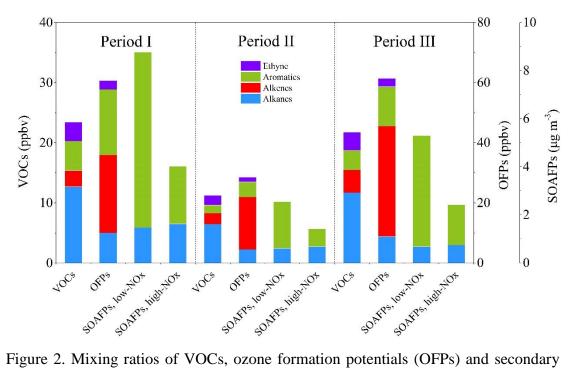


Figure 2. Mixing ratios of VOCs, ozone formation potentials (OFPs) and secondary organic aerosol formation potentials (SOAFPs) of ambient VOCs during period I, II and III at UCAS, respectively.

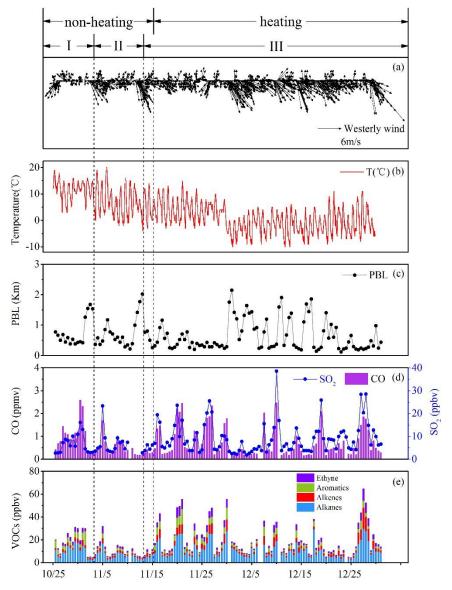


Figure 3. Time series of (a) wind speed and wind direction, (b) temperature, (c) planetary boundary layer height, (d) mixing ratios of CO and SO₂, (e) mixing ratios of VOCs, at the sampling site inside UCAS. The heating periods started on 15 November. Period I: 25 October-2 November; Period II: 3-12 November; Period III: 13 November-31 December.

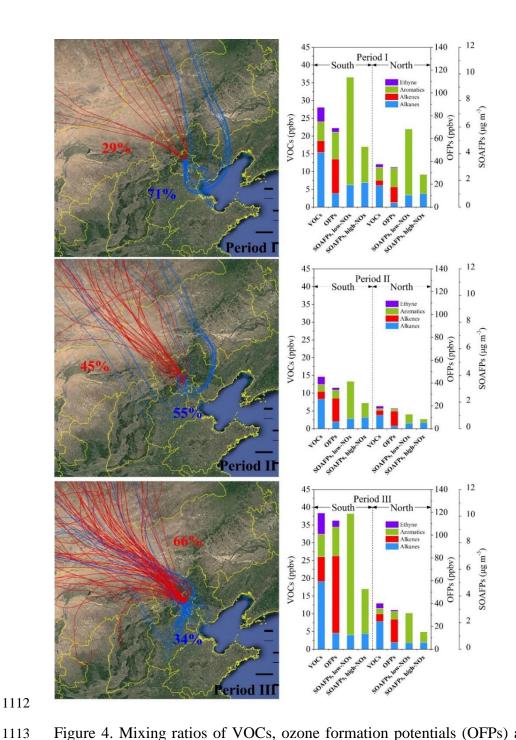


Figure 4. Mixing ratios of VOCs, ozone formation potentials (OFPs) and secondary organic aerosol formation potentials (SOAFPs) of VOCs in the air masses from the south and north regions (right) and corresponding back trajectories at 100 meters above the ground level during period I, II and III, respectively (Left).

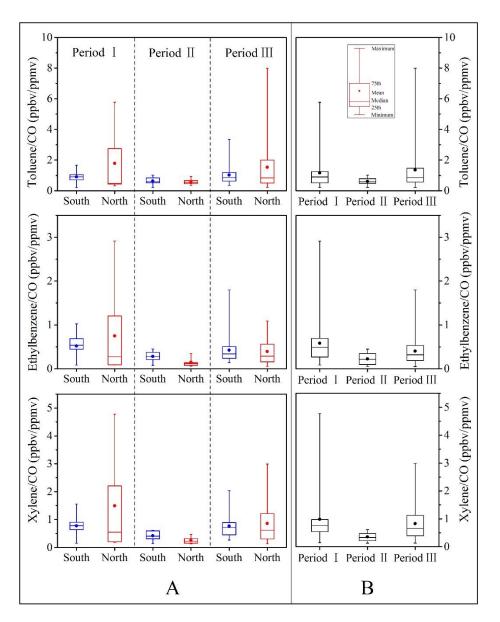


Figure 5. Ratios of aromatic hydrocarbons to carbon monoxide (CO) (A) in the air masses from the south and north regions and (B) in all samples during period I, II and III. (The lower and upper boundaries of the box represent the 25th and 75th percentiles, respectively; the whiskers below and above the box indicate the minimum and maximum, respectively; the line within the box marks the median; the dot represent the mean).

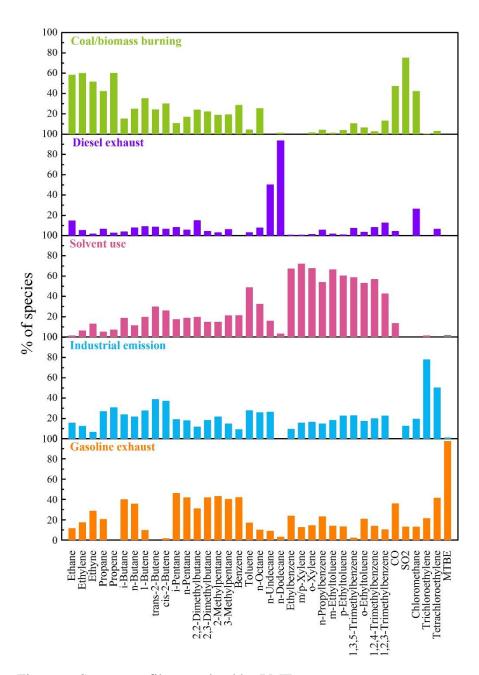


Figure 6. Source profiles revolved by PMF.

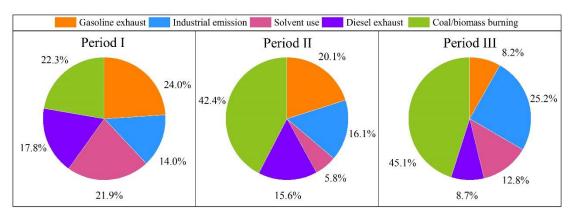


Figure 7. Contributions to VOCs in percentages (%) by different sources during period I, II and III.

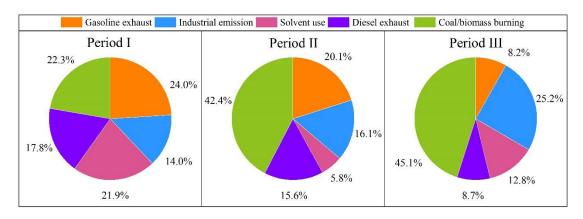


Figure 8. Sources contributions (%) to VOCs in the air masses from the south and north regions during period I, II and III.

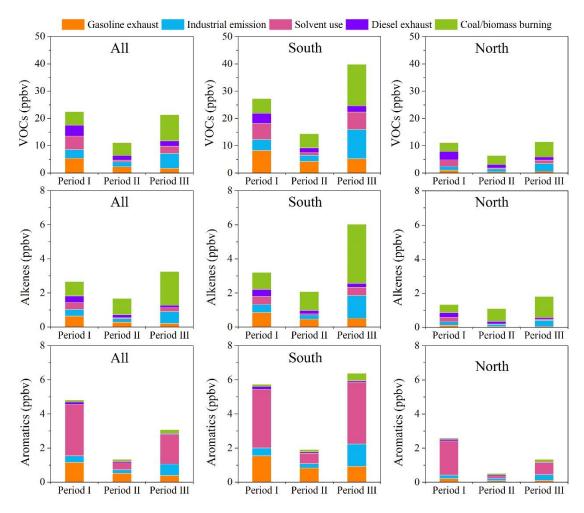


Figure 9. Sources contributions of VOCs and reactive alkenes<u>and</u>/aromatics<u>at UCAS</u>, in all samples and in air masses from the south and north regions during period I, II and III.

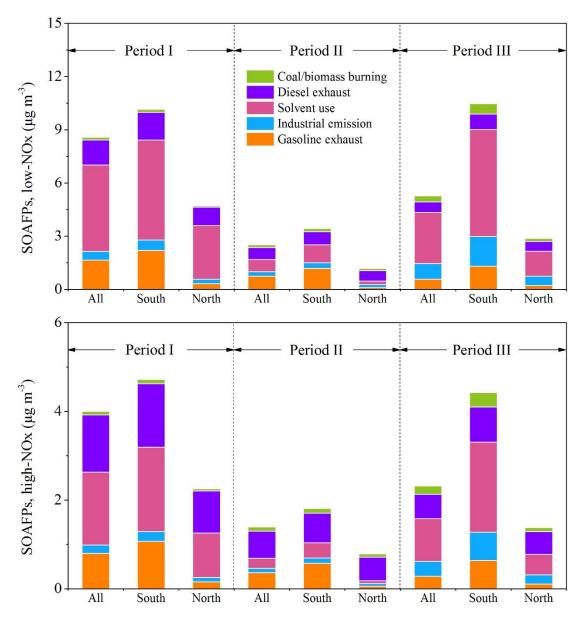


Figure 10. Contributions to SOAFPs by different sources in the air masses from the south and north regions during period I, II and III.