Vertical and horizontal distribution of regional new particle formation events in Madrid

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Abstract. The vertical profile of new particle formation (NPF) events was studied by comparing the aerosol size number distributions measured aloft and at surface level in a suburban environment in Madrid, Spain using airborne instruments. The horizontal distribution and regional impact of the NPF events was investigated with data from three urban, urban background and suburban stations in the Madrid metropolitan area. Intensive regional NPF episodes followed by particle growth were simultaneously recorded at three stations in and around Madrid, in a field campaign in July 2016. The urban stations presented larger formation rates compared to the suburban station. Condensation and coagulation sinks followed a similar evolution at all stations, with higher values at urban stations. However, total number concentration of particles larger than 2.5 nm was lower at the urban station and peaked around noon, when BC levels are minimum. The vertical soundings demonstrated that ultrafine particles (UFP) are formed exclusively inside the mixed layer. As convection becomes more effective and the mixed layer grows, UFP particles are detected at higher levels. The morning soundings revealed the presence of a residual layer in the upper levels in which aged particles (nucleated and grown on previous days) prevail. The particles in this layer also grow in size, with growth rates significantly smaller than those inside the mixed layer. Under conditions with strong enough convection, the soundings revealed homogeneous number size distributions and growth rates at all altitudes, which follow the same evolution in the other stations considered in this study. This indicates that NPF occurs quasi-homogenously in an area spanning at least 17 km horizontally. The NPF events extend over the full vertical extension of the mixed layer, which can reach as high as 3000 m in the area, according to previous studies. On some days a marked decline in particle size (shrinkage) was observed in the afternoon, associated with a change in air masses. Additionally, a few nocturnal nucleation mode bursts were observed in the urban stations, for which further research is needed to elucidate their origin.

1   Introduction

In urban areas, traffic emissions are a major source of UFP (Kumar et al., 2014; Ma and Birmili, 2015; Pey et al., 2008; Pey et al., 2009; Dall'Osto et al., 2012; Salma et al., 2014; Paasonen et al., 2016). These emissions include primary UFP exhaust emissions (Shi and Harrison, 1999; Shi et al., 2000; Charron and Harrison, 2003; Uhrner et al., 2012); cooling of engine exhaust emissions and condensation of semi-volatile phases vapor species that creates new UFP during dilution (Charron and Harrison, 2003; Kittelson et al., 2006; Robinson et al., 2007; Rönkkö et al., 2017). These are also considered primary particles, since they are formed near the source. Other relevant UFP sources include industrial emissions (Keuken et al., 2015; El Haddad et al, 2013), city waste incineration (Buonanno and Morawska, 2015), shipping (Kecorius et al, 2016; Johnson et al., 2014), airports (Cheung et al., 2011; Hudda et al., 2014; Keuken et al., 2015) and construction works (Kumar and Morawska, 2014).

New particle formation (NPF) from gaseous precursors has been shown to cause high ultrafine particle (UFP) episodes in relatively clean atmospheres due to low condensation sinks (CS) originating from low pre-existing particle concentration (e.g. Kulmala et al., 2000; Boy and Kulmala, 2002; Wiedensohler et al.,2002; Kulmala et al., 2004; Wehner et al., 2007; O'Dowd et al., 2010; Sellegri et al., 2010; Vakkari et al., 2011; Cusack et al., 2013a; Cusack et al., 2013b; Tröstl et al., 2016; Kontkanen et al., 2017). However, at mountain sites, precursors' availability seems to be the most influential parameter in NPF events, with higher values of CS during NPF events than during non NPF events (Boy et al., 2008; Boulon et al., 2010; García et al., 2014, Nie et al., 2014, among others). Tröstl et al. (2016) reported experimental results on nucleation driven by oxidation of Volatile organic compounds (VOCs), and Kirkby et al. (2016) reported pure biogenic nucleation.

NPF events contribute also significantly to ambient UFP concentrations in urban environments (Costabile et al., 2009; Wegner et al., 2012; von Bismarck-Osten et al., 2013;2014; Ma and Birmili, 2015; Hofman et al., 2016, Kontkanen et al., 2017). Common features enhancing urban NPF are high insolation, low relative humidity, availability of SO_2 and organic condensable vapors, and low condensation and coagulation sinks (Kulmala et al., 2004; Kulmala and Kerminen, 2008; Sipilä, et al., 2010, Salma et al., 2016). Urban NPF episodes can be either regionally or locally driven and may or may not impact regional background areas (Dall'Osto et al., 2013; Brines et al., 2015; Salma et al., 2016). Cheung et al. (2011) and Brines et al. (2015) reported that in urban areas nucleation bursts without growth of particles are common; whereas the frequently occurring 'banana like' nucleation bursts at regional background sites are scarcely detected at urban sites, probably because the high CS at traffic rush hours limits the duration of the particle growth. These processes seem to prevail in summer and spring in Southern European urban areas (Dall'Osto et al., 2013; Brines et al, 2014; Brines et al., 2015). Brines et al. (2015) also reported that in urban environments the highest O_3 levels occur simultaneously with NPF events, as well as the highest SO_2 concentrations and insolation, and the lowest relative humidity and NO and NO_2 levels. This close association between O_3 and UFP may be due to ambient conditions that favor two different but simultaneous processes, or to the fact that they are two products of photochemical reactions in the same overall process.

Reche et al. (2011) evaluated the prevalence of primary versus newly formed UFP in several European cities and found a different daily pattern for the southern European cities, where the newly formed particles contributed substantially to the annual average concentrations, probably because of high insolation and possible site-specific chemical precursors. Brines et al. (2015) determined that NPF events lasting for 2 h or more occurred on 55 % of the days and those extending 4 h on 28 % of the days, being NPF the main contributor on 14-19 % of the time in Mediterranean and Sub-tropical climates (Barcelona, Madrid, Roma, Los Angeles and Brisbane). The latter percentages reached 2 % and 24-28 % in Helsinki and Budapest (Wegner et al., 2012 and Salma et al., 2016, respectively). Furthermore, Brines et al. (2015) calculated that 22 % of the annual average UFP number concentration recorded at an urban background site of Barcelona originated from NPF. Ma and Birmili (2015) reported that the annual contribution of traffic on UFP number concentration was 7, 14 and 30 % at roadside, urban background and rural sites respectively, in and around Leipzig, Germany. On the other hand, traffic emissions

contributed 44-69 % to UFP concentrations in Barcelona (Pey et al., 2009; Dall'Osto et al., 2012, Brines et al., 2015), 65 % in London (Harrison et al., 2011; Beddows et al., 2015) and 69 % in Helsinki (Wegner et al., 2012).

Minguillón et al. (2015) and Querol et al. (2017) demonstrated that intensive NPF episodes take place inside the planetary boundary layer (PBL) in Barcelona, occurring around midday at surface levels when insolation and dilution of pollution are at their maxima. Earlier in the morning NPF can only take place at upper atmospheric levels, at an altitude where pollutants are diluted, since at surface levels a high CS prevents particle formation.

While many studies have investigated NPF around the world, only a few have focused on the vertical distribution of these events (Stratmann et al., 2003; Wehner et al., 2010). In view of this, we devised a campaign with the aim to study photochemical episodes, including high O_3 levels and NPF in Madrid metropolitan area. In a twin article (Querol et al., 2018) the study of the temporal and spatial variability of O_3 is presented. In this work we will focus exclusively on the phenomenology of the NPF events, comparing the aerosol size distribution at surface level at urban, urban background and suburban stations in Madrid and the outskirts of a residential village 17 km from Madrid. We also study the vertical distribution of the events using airborne instrumentation carried by tethered balloons.

2 Methodology

2.1 The study area

The Madrid Metropolitan Area (MMA) lies in the center of the Iberian Peninsula at an elevation of 667 m above sea level (a.s.l.). It is surrounded by mountain ranges and river basins that channel the winds in a NE-SW direction. Having an inland Mediterranean climate, winters are cool and summers are hot, and precipitation occurs mainly in autumn and spring. Road traffic and residential heating in winter are the main sources of air pollutants, with small contributions of industrial and aircraft emissions (Salvador et al., 2015).

In summer, the area is characterized by strong convection, which results in PBL heights as high as 3000 m above ground level (a.g.l.), and mesoscale recirculation caused by anabatic and katabatic winds in the surrounding mountain ranges (Plaza et al., 1997, Crespí et al., 1995), which can lead to accumulation of pollutants if the recirculation persists for several days.

Cold and warm advection of air masses associated with the passage of upper level troughs and ridges over the area give rise to a sequence of accumulation and venting periods, respectively. During accumulation periods, pollutants accumulate in the area and concentrations increase for 2-6 days, until a trough aloft brings a cold advection and a venting period starts. For a detailed description of the meteorological context during the campaign see Querol et al. (2018).

A few studies have focused on NPF events in the area. For instance, Gómez Moreno et al. (2011) reported NPF episodes in Madrid to be 'not a frequent phenomenon', since only 63 events per year were detected, 17 % of the total days, occurring mostly in spring and summer. However, Brines et al. (2015) reported both intensive summer and winter NPF episodes at the same station that accounted for 58 % of the time as an annual average, considering the prevalence of nucleation bursts for 2 hours or more. Alonso-Blanco et al. (2017) described the phenomenology of particle shrinking events, i.e. a decline in particle size caused by particle-to-gas conversion, at an urban background station in Madrid (CIEMAT), stating that they occur mainly between May and August in the afternoon, due to either a change in wind direction or the reduction of photochemical processes. Particle shrinkage following their growth is not a common phenomenon but has been observed in a few areas around the world. Yao et al. (2010), Cusack et al. (2013a and 2013b), Young et al. (2013), Skrabalova et al. (2015) and Alonso-Blanco et al. (2017) and references therein, reported shrinkage rates ranging from –1.0 to –11.1 nm h^{-1}.

2.2   Instrumentation

The data used in this study was collected during a summer campaign in and around Madrid in July 2016. Three air quality supersites were used, an urban station, an urban background station and a suburban station, in addition to a setting in a suburban environment with two tethered balloons that allowed study of the vertical distribution of aerosol and air pollutants. All stations are located within a range of 17 km. A map displaying all locations is shown in Fig. S1.

The CSIC urban station, operative from 9 to 20 July, was located in the Institute of Agricultural Sciences (40°26'25" N, 03°41'17" W, 713 m a.s.l.) in central Madrid. The instrumentation in this station was installed in the sixth floor of the building. NO_x and equivalent black carbon (BC) concentrations were measured with a chemiluminescence based analyzer (Teledyne API, 200EU) and an Aethalometer (AE33, Magee Scientific), respectively. The aerosol number size distribution was measured with a Scanning Mobility Particle Spectrometer (SMPS) (TSI, 3082) equipped with a nano-Differential Mobility Analyzer (DMA), for the size range 8-120 nm, and a Particle Size Magnifier (PSM) (AirModus) in scanning mode for the size range 1.2-2.5 nm. PSM data were post-processed using tailored software provided by Airmodus.

The CIEMAT urban background station, operative from 4 to 20 July was located in the outskirts of Madrid, 4 km from the CSIC station (40°27'23" N, 03°43'32" W, 669 m a.s.l.). NO_x, O_3 and BC concentrations were measured with a chemiluminescence based analyzer (THERMO 17i), an ultraviolet photometry analyzer (THERMO 49i) and an Aethalometer (AE33), respectively. The aerosol number size distribution was measured with an SMPS (TSI 3080) for the size range 15-660 nm and a 1 nm SMPS (TSI 3938E77) for the size range 1-30 nm. In the overlapping range (15-30 nm), the nanoSMPS yielded slightly higher concentration values. In order to correct that and to obtain a continuous size distribution, the daily nanoSMPS values were corrected to adapt those of the SMPS. We compared the resulting merged particle size distribution with CPC measurements, to check that there was a good agreement in the total particle concentration. Temperature (4 m a.g.l.), relative humidity (4 m a.g.l.), solar radiation (35 m a.g.l.) and wind speed and wind direction (55 m a.g.l.) were measured at a meteorological tower at the station.

The ISCIII suburban station was located in the Institute of Health Carlos III, in Majadahonda, 15 km from the CSIC station (40°27'27" N, 03°51'54" W, 739 m a.s.l.) and was operative from 4 to 20 July. An SMPS (TSI 3080) measured the aerosol number size distribution in the size range 9-360 nm, and a PSM (AirModus) in scanning mode in the size range 1.2-4.0 nm. PTR-ToF-MS (Ionicon Analytik, PTR-TOF 8000) operating in $H3O+$ mode was used to measure VOCs. Detailed description of the instrument can be found in Graus et al. (2010). Operation procedure of the PTR-ToF-MS is fully described in Querol et al. (2018). Results regarding these measurements are briefly presented in Sect. S1.

TSI instruments at all stations were collocated next to windows or walls where holes were available for inlets, and equipped with individual ¼ inch, 20 cm long conductive silicone tubing inlets for PSM. SMPS and CPC also had individual 30 cm conductive silicone tubing inlets. Being the inlets individual, each instrument had its own flow rate. TSI instrument data were corrected for diffusion losses and multiple charge losses using the instruments' own software.

Regarding the vertical measurements, two tethered balloons carrying miniaturized instrumentation were based at Majadahonda (MJDH) rugby field (40°28'29.9" N 3°52'54.6" W, 728 m a.s.l.), 17 km from CSIC. 28 flights up to 1200 m a.g.l. were carried out from 11 to 14 July. A miniaturized SMPS (Hy-SMPS) measured the particle size distribution in the range 8-245 nm with a time resolution of 45 s and flow of 0.125 L/min (Lee et al., 2015). However, only particles larger than 10 nm could be detected due to a lower efficiency for finer particles. The instrument was intercompared with a TSI-SMPS (Standard DMA with 3776 CPC) for 50-nm monodisperse NaCl particles and polydisperse aerosol (Fig. S2). Particle concentration in the range 3-1000 nm was measured with a miniaturized butanol-based CPC (Hy-CPC). The time resolution was 1 s, and sample flow was 0.125 L/min (Lee et al., 2014). Temperature, relative humidity, pressure, wind speed and wind

direction were also measured. The instrumentation was also equipped with a Global Positioning System (GPS). An additional set of the miniaturized instrumentation was placed at surface level for comparison.

2.3   Data analysis techniques

Identification of NPF events was made by the method proposed by Dal Maso et al. (2005). After examination of the daily particle size distribution, if the day was classified as an event day we proceeded to calculate growth rates (GR), shrinking rates (SR), condensation and coagulation sinks (CS and CoagS) and formation rates (J_{Dp}).

The algorithm proposed by Hussein et al. (2005) was used to fit log-normal modes to the particle size distribution, from which GR were calculated following Eq. (1):

$$GR = \frac{dD_p}{dt}, \tag{1}$$

where D_p are the selected geometric mean diameters corresponding to growing particle modes. Unless stated otherwise, in this work growth rates are calculated for particles growing from 9 to 25 nm. When calculating growth rates with PSM data, the range was selected accordingly to the measuring range of each instrument (see Sect. 2.2). SR were calculated analogously when a decrease in the diameter of the fitted modes was observed.

CS, a measure of the removal rate of condensable vapor molecules due to their condensation onto the pre-existing particles (Kulmala et al., 2012), is calculated using Eq. (2):

$$CS = 2\pi D \sum_i \beta_i D_{p,i} N_i , \tag{2}$$

where D is the diffusion coefficient of the condensing vapor (here we use H_2SO_4), $D_{p,i}$ and N_i are the particle diameter and particle concentration for the size class i. β_i is the transitional correction factor:

$$\beta_i = \frac{1+K_i}{1+\left(\frac{4}{3\alpha}+0.337\right)K_i+\frac{4}{3\alpha}K_i^2}, \tag{3}$$

being K the Knudsen number $K_i = \frac{2\lambda}{D_{p,i}}$, where λ is the mean free path of the condensing vapor in air.

The formation rates of particles were calculated as 30-minutes averages, following Eq. (4):

$$J_{Dp} = \frac{dN_{Dp}}{dt} + CoagS_{Dp} \cdot N_{Dp} + \frac{GR}{\Delta D_p} N_{Dp}, \tag{4}$$

where we use the PSM measuring range for N_{Dp}, and CoagS is a quantification of the ability of the preexisting aerosol to scavenge newly formed particles. For its calculation we take the geometric mean diameter of the size ranges 1-25 nm, using a merged PSM and SMPS particle size distribution. CoagS can be calculated using Eq. (5):

$$CoagS_{Dp} = \sum_{D_{p'}} K(D_p, D_p')N_{Dp'} , \tag{5}$$

where K is the coagulation coefficient. For a detailed description of the parameters and their derivation, see Kulmala et al. (2012).

A rough estimation of the mixed layer height was determined using Hy-CPC measurements. The top of the mixed layer was considered at an altitude in which particle concentration decreases an order of magnitude quasi-instantaneously and remains constant above. All UFP profiles are included in Querol et al. (2018).

Additionally, bivariate polar plots of concentration have been used to relate wind speed and direction with total particle concentration using PSM data by means of the R package *openair* (Carslaw and Ropkins, 2012).

3 Results and discussion

3.1 Meteorological context

Figure S3 shows the evolution of temperature, relative humidity, wind speed and wind direction measured at CIEMAT from 5 to 20 July 2016. The evolution of temperatures during this period evidences a succession of accumulation and venting episodes. Rain gauges collected significant precipitation only on 6 July at midnight (not shown).

The balloons field campaign, held from 11 to 14 July, coincided with the start of a venting period, coinciding with the passage of an upper level trough, and the transition to an accumulation period, when the trough has moved to the east of the Iberian Peninsula and a ridge passes over the area of study (see Fig. S4). Maxima and minima temperatures drop, while strong westerly winds predominate until they veer to NE on 12 July 18:00 UTC. High nocturnal wind speed peaks are recorded in this period, often accompanied by a change in wind direction. For detailed information on meteorological parameters during this campaign see Querol et al. (2018).

3.2 Comparison of NPF events at urban and suburban surface stations

In the following discussions, we group CSIC (urban) and CIEMAT (urban background) as urban stations and compare them to ISCIII (suburban). This is because of the availability of data during the period of interest. However, it has to be noted that CSIC is more influenced by traffic than CIEMAT, therefore it is more representative of urban environments, and for this reason CSIC data is chosen when possible. 18 NPF episodes have been identified on a total of 7 days throughout the campaign. In Table 1 a summary of these events is presented. Out of these, a total of 14 events on 6 days had simultaneous data available for at least one of the urban stations (CSIC, CIEMAT) and the suburban station (ISCIII). These episodes, marked with a star in Table 1, are selected for further analysis in this section. Figure 1 represents the aerosol particle size distribution of the selected episodes (12-18 July 2016).

3.2.1 Episode characteristics

In the selected episodes, intensive daytime nucleation and subsequent condensational growth processes took place simultaneously at urban and suburban stations, located 17 km apart, and accordingly we classify these as regional NPF episodes. Being all stations differently influenced by traffic (influenced, slightly and not influenced by traffic), we can affirm that these episodes are regional events and not representative of primary emissions. Otherwise, we would observe significant differences at our stations. Additional arguments are the fact that number concentrations of sub-25 nm particles peak at noon, when BC levels are at their minimum, as well as concentration of particles measured by PSM being higher at the suburban station, compared to the urban station, implying that the particles are not originated from traffic sources.

At urban stations particles of the order of 10 nm are detected throughout the day, even at night time. Conversely, at the suburban station such small particles are only detected during daytime. Additionally, during some days a very intense short nucleation burst is registered around midnight local time at urban stations that are not detected at the suburban station. This phenomenon is analyzed in section 3.4.2.

Despite the detection of sub-10 nm particles as early as 04:00 UTC (06:00 local time) at the urban stations, only after around 09:00 UTC the growth of the particles is observed, occurring roughly at the same time in both urban and suburban stations.

Newly formed particles grow until they have reached sizes of up to 50 nm, usually around 13:00 UTC (15:00 local time). After this, shrinkage is observed on 10 days, corresponding to 71 % of the days with available data. Consequently, the evolution of the particle size distribution is arc-shaped in these cases.

It should be noted that nucleation episodes coincide in time with the early increases in O_3 concentrations in the morning, whereas the occurrence of maximum O_3 concentration (120 to 150 μg m^{-3} hourly daily maxima between 14:00 and 16:00 UTC; see Fig. S5) takes place during the UFP growth stage, since oxidation of VOCs and inorganic gases is also accelerated with photochemistry and the presence of O_3 and OH radicals, among others (Coleman et al., 2008; Wang et al., 2017; Saiz-Lopez et al., 2017).

3.2.2   Comparison of GR, J_1, CS and CoagS

For the observed daily regional NPF events, GR for the nucleation mode, J_1, CS and CoagS have been determined using PSM and SMPS aerosol size distribution measurements. Here, the growth rate is calculated from the time of detection of the smallest mode until either the particle reaches 25 nm or it stops growing before reaching that size. We considered only the events that are observed simultaneously at the suburban station and at least at one urban station (highlighted in Table 1). Figure 2 shows the growth rates presented in Table 1 according to urban and suburban surface stations. GR regarding the vertical measurements are discussed in the following section due to differing sampling periods. Growth rates ranged from 2.9 to 7.6 nm h^{-1} at the suburban site, with a mean value of 4.5±2.1 nm h^{-1}, and from 1.4 to 4.0 nm h^{-1} at the urban stations with a mean value of 2.8±1.0 nm h^{-1}. We cannot affirm that the mean value of the suburban station is higher than that of the urban stations because the mean value of the GR at urban stations is included in the confidence interval of the GR at the suburban station. It also has to be considered that only a few days of measurements are available for this calculations. The GR calculated are consistent with those observed GR by Alonso-Blanco et al. (2017), ranging 1.4-10.6 nm h^{-1} at CIEMAT.

The GR calculated in this study are also consistent with those observed in other urban and suburban areas. Kulmala et al. (2004) concluded that typical growth rates are 1-20 nm h^{-1} in mid-latitudes. In particular, Stolzenburg et al. (2005) observed GR ranging 2.4-8.5 nm h^{-1} in regional events in an urban environment in Atlanta, US. Qian et al. (2007) reported regional events with median GR of 5.1 nm h^{-1} in an urban environment in St. Louis, US. Ahlm et al. (2012) reported average GR of 7.3 nm h^{-1} at Bakersfield, US. Manninen et al. (2010) characterized NPF events in 12 European sites. Cabauw (The Netherlands) and San Pietro Capiofume (Italy), are stations located in environments comparable to that in our suburban station, ISCIII. For these stations the observed median growth rates were 7-8 nm h^{-1}, corresponding well with our calculated GR in the suburban station.

Figure 3 shows the average daily cycles of particle concentration in the size ranges 9-25 nm (N_{9-25}) and 1.2-2.5 nm ($N_{1.2-2.5}$), total particle concentration measured by the PSM ($N_{>2.5}$), CS and CoagS during the regional NPF events at urban and suburban stations. Average N_{9-25} daily mean values are 3.7 x 10^3 cm^{-3} and 2.2 x 10^3 cm^{-3} at urban and suburban stations, respectively.. $N_{>2.5}$ have average daily mean values of 1.6 x 10^4 cm^{-3} and 2.1 x 10^4 cm^{-3} at urban and suburban stations, respectively. It has to be highlighted that $N_{1.2-2.5}$ is considerably larger at the suburban station throughout the day, with mean values of 2.5 x 10^3 cm^{-3}, compared to 0.5 x 10^3 cm^{-3} at the urban stations. CS and CoagS have average daily mean values of 3.4 x 10^{-3} s^{-1} (2.5 x 10^{-3} s^{-1} at the suburban station) and 2.4 x 10^{-5} s^{-1}, respectively. After dawn, anthropogenic activities start, and N_{9-25}, $N_{1.2-2.5}$, $N_{>2.5}$, CS and CoagS start to increase at the same time, both in urban and suburban environments. Around 07:00 UTC, once the morning traffic rush diminishes, N_{9-25}, $N_{>2.5}$ and the sinks increase more slowly; moreover, total particle concentration decreases in the suburban station, indicating that in this environment the impact of the traffic emissions in total particle concentration is smaller than near the city center, as expected. Shortly after, at 09:00 UTC the photochemical processes are strong enough to start NPF, as suggested by the increase in particle concentrations, while the sinks get to a

relative minimum. N_{9-25} reaches its maximum at midday (9×10^3 cm^{-3} and 5×10^3 cm^{-3} at the urban and suburban stations, respectively), and then decrease because the particles start growing to diameters greater than 25 nm, adding to the sinks, which increase gradually until the evening. NPF leads to maximal UFP concentrations around midday in all stations, as suggested by the peak in $N_{>2.5}$, which is recorded in coincidence with very low BC levels (see Figs. S5 and S6). Around 19:00 UTC the effect of the afternoon traffic rush is evident, the variables evolving equivalently to that in the morning. Finally, at 23:00 UTC a sharp and short increase in N_{9-25} is observed, associated with the aircraft emissions discussed in Sect. 3.4.2.

Growth rates (GR$_{PSM}$) and total formation rates of 1.2-4.0 nm particles (J_1) were calculated from PSM data at CSIC and ISCIII stations. GR$_{PSM}$ were calculated from 11 to 18 July 2016, averaging 4.3 nm h^{-1} at the urban station and 3.7 nm h^{-1} at the suburban station. J_1 were calculated only for the days in which NPF is identified. The results for these days are included in Table 1. Average J_1 values are higher at the urban station (8.9 cm^{-3} s^{-1}) compared to the suburban station (5.3 cm^{-3} s^{-1}). Concentrations of 1.2-4.0 nm particles are lower at the urban station (Figure S6), which could lead to lower formation rates. However, the coagulation sink is greater at the urban station, as discussed before, which contributes to the second factor in Eq. (4). It has to be noted that only 3 days of overlapping between PSM and SMPS data were available for NPF events at the urban station. A longer dataset could lead to different results.

The average values of the formation rates agree with those reported at similar stations around the world. For instance, Woo et al. (2001) reported J_3 ranging 10-15 cm^{-3} s^{-1} in Atlanta, US. Wehner and Wiedensohler (2003) reported average J_3 of 13 cm^{-3} s^{-1} in Leipzig, Germany. Hussein et al. (2008) reported nucleation rates (D_p<25 nm) ranging 2.1-3.0 cm^{-3} s^{-1} in summer in Helsinki.

3.3 Vertical distribution of NPF events

3.3.1 UFP concentrations

Querol et al. (2018) studied the vertical profiles of UFP and O3 concentrations measured during the campaign using the balloon soundings at Majadahonda. UFP concentrations are homogeneous throughout the mixing layer and present a sharp decrease at the top. As the day progresses, the convection is more effective and high UFP levels reach higher altitudes, as the mixing layer heightens. Moreover, the concentrations tend to increase until midday. Afterwards, they remain constant or slightly decrease, always showing homogeneous levels from surface levels to the top of the PBL. Concentrations of UFP markedly increased from 11 to 14 July, both at surface and at upper levels. This is consistent with the observed decrease of the convective activity in that period, evidenced by a decrease in temperatures, but also with an increase in the formation rates, calculated in this study. Therefore, the increase in particle concentration is probably the result of both a decline in PBL height and more intense nucleation episodes.

3.3.2 Particle size distribution and NPF episodes

The NPF events described in Sect. 3.2 that took place between 12 and 14 July were not only detected at surface level but also in upper layers with the balloons soundings in Majadahonda. However, the measurements were not continuous, since the balloons could not be operated safely if the wind speed was above 8 m s^{-1} at any vertical level.

Figure 4 shows the fitted modes to the particle size distribution measured in the soundings on 12 July. The fact that sub-40 nm particles are not detected at the higher levels of the first flights suggests that convection is not very effective yet, and the sounding goes through different atmospheric layers, most likely the mixed layer and the residual layer. In the residual layer

Aitken-mode particles formed on previous days prevail (Stull, 1988). The interphase between the mixed layer and the residual layer, i.e. the mixed layer height, has been derived using the UFP vertical profiles (see Querol et al., 2018). From 10:00 UTC onwards, once the convection has fully developed, the mixed layer covers all the sounding and we see a homogeneous distribution at all levels, which is also comparable to those recorded with the instrumentation measuring at the surface. This agrees with the fact that UFP are homogeneously distributed in the mixed layer and are detected at higher altitudes as the mixed layer rises.

In the early morning the size distribution is dominated by a 60 nm mode at all altitudes, which grows to 100 nm at 11:00 UTC. Even though it is detected at all levels, the mode slightly decreases its size when the sounding ascends above the mixed layer limit, more clearly visible on the second flight, around 9:00 UTC. This result suggests that there are lower vapor concentrations in the residual layer, which inhibits particle growth, whereas the mixed layer is more polluted, thus the particles can grow faster. The growth rates calculated for this mode were 1.8 nm h^{-1} in the residual layer, and 7.3 nm h^{-1} in the mixed layer. The concentration and size of the Aitken mode decrease after midday, which might be related to an increase of wind speed that entailed dilution and evaporation, leading to shrinking of the particles. Because of the increase in wind speed the balloons could not be safely operated, and no additional flights were made on that day.

Moreover, during the morning we observed particles growing inside the mixing layer from 10 nm at 7:00 UTC, to 30 nm at midday. This mode is observed simultaneously at ISCIII and therefore we consider it for calculation. The growth rate obtained is 3.5 nm h^{-1}. The fact that the growth rate is the same throughout the mixing layer even though we expect VOCs to be higher near the surface upholds the assumption that the convection is very efficient, and the entire layer is well-mixed. After 13:00 UTC, because of the increase in wind speed particles start to shrink. While concentrations were not as high as other episodes, the evolution is remarkably similar to the NPF event measured at the same time at ISCIII, which had a growth rate of 3.0 nm h^{-1}.

The size distribution and the corresponding fitted modes for the soundings made on 13 July are presented in Fig. 5. Although the balloons could not fly until 10:30 UTC for safety reasons, at least 2 modes are detected from early morning at the sounding location. A mode starting roughly at 40 nm at 07:00 UTC grows to 100 nm at 15:00 UTC. With a growth rate of 8.5 nm h^{-1}, this mode was detected at all altitudes once the soundings started, indicating that the convection was already effective by 10:30 UTC and all the measured altitudes were completely mixed, leading to a homogeneous particle distribution throughout the soundings. This mode is the prolongation of the Aitken mode detected the day before, which shrank from midday until the following morning. It is also detected at ISCIII and CSIC, with growth rates of 7.5 nm h^{-1} and 6.9 nm h^{-1}. A nucleation mode grows from the detection limit of the instrument, around 10 nm at 08:30 UTC to 40 nm at 15:00 UTC. Comparing with other stations, we considered this mode only after 9:00 UTC, and calculated the growth rates from that time. We consider this a regional NPF event, since the start of the particle growth is registered simultaneously at all the stations. The growth rates at the sounding location, ISCIII and CSIC are 5.3 nm h^{-1}, 4.6 nm h^{-1} and 2.0 nm h^{-1}, respectively.

Finally, Fig. 6 shows the particle size distribution and fitted modes for the soundings made on 14 July. Correspondingly, in Fig. 7 the vertical distribution of particles for some of the soundings is presented. The earliest soundings revealed the existence of a residual layer aloft. In order to verify this result two constant altitude flights were made during the morning. The extension of the wire was not modified during these flights. However, changing wind conditions varied slightly the altitude of the instruments. The altitude was chosen so that the instruments remained initially outside the mixing layer, i.e. inside the residual layer. As the insolation increased, so did the altitude of the mixing layer, until it reached the altitude at which the balloons were positioned. As the mixing layer reached the balloons, total particle concentration sharply increased from 4×10^3 to 2×10^4 cm^{-3}, demonstrating that newly-formed particles remain inside the mixing layer.

According to the abrupt decline in particle concentration, the boundary between the mixing and residual layers was located at 1000 m at 09:00 UTC, 1200 m at 10:00 UTC, 1350 m at 11:00 UTC and beyond 1800 m after 12:00 UTC. This can be taken as an indicator of the effectiveness of convection, meaning that after 12:00 UTC all the measured particle population was well mixed throughout the sounding range. Inside the residual layer particles had a slower growth rate (0.5 nm h^{-1} compared to 8.45 nm h^{-1} for the 40 nm mode – note that due to the use of a log-scale this might be unnoticeable visually), and no particles smaller than 20 nm were observed.

Nucleation mode particles were detected exclusively inside the mixing layer from 08:00 UTC to 12:00 UTC, whereas growth was only observed from 09:00 to 11:00 and from 12:00 onwards. The time spacing between both growing periods coincides with a marked decrease in wind speed. During the first period growth rates at the sounding station, ISCIII and CSIC were 6.2, 5.4 and 1.4 nm h^{-1}, respectively. However, during the second stage particles grew faster at the urban station (8.6 nm h^{-1}) than at the sounding location (4.5 nm h^{-1}). As the latter is a suburban environment, this contrasts with the results obtained in Sect. 3.2.2. This fact could be explained by the veer of NE winds to weaker southerly winds in Madrid, which is not observed in Majadahonda.

Overall, the soundings revealed that there is simultaneous growth and shrinking of nucleation and Aitken modes, and that both of them grow and shrink at different rates. This was also observed in the surface measurements when comparing urban and suburban stations (see Sect. 3.2.2).

3.4   Other observations

3.4.1   Prevalence of particles and shrinkage

A further interesting feature is the presence of the Aitken mode on most days. Usually in the size range 50-100 nm, reaching 110 nm in some cases, this mode doesn't correspond to newly formed particles, but it follows a parallel evolution (condensational growth and potential shrinkage). When looking at the evolution of aerosol size distributions on consecutive days, it is possible to see a connection between this 50-100 nm mode and the distribution of the previous days. The nucleated and grown mode of one day is still present the following day and it continues to grow until it eventually fades away or grows beyond the detection limits of the instruments. In some occasions the Aitken mode can be tracked for two or more consecutive days, alternating stages of growth and shrinkage.

The start of the shrinking phase coincides with a marked increase in wind speed (Fig. S7), therefore it is associated with dilution, which favors the evaporation of semi-volatile vapors, resulting in a decline in particle diameter and concentrations, as observed in most cases. Figure 8 shows the shrinkage rates according to the starting diameter of the shrinking particles and the stations. Data used for this figure including start and end times and diameters is included in Table S1. The calculated shrinkage rates for particles with a starting diameter below 40 nm range from -1.1 to -8.0 nm h^{-1}. For particles in the Aitken mode above 40 nm the values fall between -4.9 and -20.5 nm h^{-1}. The results confirm that shrinkage is a regional phenomenon in the Madrid area, as already suggested by Alonso-Blanco et al. (2017), the processes being faster in the suburban station compared to the urban station. It is also observed that particles shrink faster the larger the starting diameter is.

3.4.2   Nocturnal UFP peaks

Although out of the major focus of this study (photochemical nucleation), other interesting events were detected taking place during night time. From 6 to 11 July and 17 to 19 July, high concentrations of 1.2-4 nm particles are registered shortly after

sunset for several hours, simultaneously at urban and suburban stations (see Fig. 1). BC, NO and NO_2 concentrations also increase during that time (see Fig. S5). Therefore, these processes are probably related to local traffic emissions and the decrease of the mixing layer after sunset. On the other hand, from 12 to 14 July high concentrations of sub-25 nm particles are also detected, but only registered at the urban stations around 23:00 UTC. These are sudden, shorter and more intense,

5   with concentrations greater than 10^5 cm^{-3}. They appear as intense bursts that last one hour or less, with no subsequent growth. These are not accompanied by simultaneous high BC or NO concentrations, thus they are not linked to traffic emissions, although NO_2 levels are significant. Furthermore, these episodes occur outside local traffic rush hours, and are registered together with strong NE winds, which suggest that they might be transported from a stationary source and not formed locally. To better support this hypothesis, Fig. S8 shows PSM data together with wind direction and wind speed,

10   showing that the episodes coincide with strong NE winds.

In order to determine the origin of these sub-25 nm particles, bivariate polar plots of concentration have been used to relate wind speed and direction measured at CIEMAT with total particle concentration of 1.2-2.5 nm particles, BC, NO_2 and NO measured at CSIC, separately analyzing daylight and night time periods (Fig. 9). These plots must be carefully interpreted, since the color scale only represents the average value for a given wind speed and direction. The results are consistent with

15   what we previously stated: the highest nocturnal 1.2-2.5 nm particle concentrations are linked with strong winds from NE direction. Air masses transported from this direction have the lowest BC levels, and moderate NO_2 concentrations. NO concentrations are insignificant at nighttime considering any direction, probably because of titration due to the high concentrations of O_3 observed during daytime.

In the discussion paper we pointed out the airport Adolfo Suárez Madrid-Barajas, located NE of the city, as a possible source

20   of these high UFP concentrations. However, the UFP peaks lasted for about one hour on all days, whereas strong NE winds prevailed a few hours. Moreover, the airport has flights during all night, therefore a longer period with high UFP should be observed. Although other studies have linked aircraft emissions with nucleation bursts without growth (Cheung et al., 2011, Masiol et al., 2017), with this study we cannot affirm that the airport is the origin of these bursts. As mentioned before, these episodes were unexpected and were not the main focus of this study. To elucidate the origin of these UFP bursts further

25   research will be required.

4   Conclusions

We investigated the phenomenology of regional and secondary New Particle Formation (NPF) episodes in central Spain. To this end we set up 3 supersites (an urban, an urban background and a suburban) 17 km away in and around Madrid. We were able to characterize 6 NPF events, and in all cases the evolution of the particle size distribution (PSD) was very similar at all

30   stations: around sunrise nucleation mode particles appear and start growing and in the afternoon a decline in particle sizes, i.e. shrinkage, is observed. The regional origin of the NPF is supported by the simultaneous variation in PSD in the nucleation mode and particle number concentrations, growth and shrinkage rates. Furthermore, temporal evolutions of condensation and coagulation sinks (CS and CoagS) were similar at all stations, having minimum values shortly before sunrise and increasing after dawn towards the maximum value after midday in the early afternoon. In spite of the 17 km

35   scale simultaneous processes affecting particle number concentrations, the following relevant differences between urban and suburban stations were observed: i) the urban stations presented larger formation rates as compared to the suburban stations; ii) in general, the sinks were higher at the urban stations.

Regarding the vertical soundings of the NPF events, we observed that in the early morning the vertical distribution of newly formed particles is differentiated in two layers. The lower layer (mixed layer, ML) in which convection is effective, is well-

40   mixed and has a homogeneous PSD. This ML heightens throughout the day, as insolation is more pronounced, extending

beyond the sounding limits around midday. NPF occurs throughout this ML, and growth rates and concentrations are homogeneous. The upper layer is a stable residual one (RL) in which particles formed or transported the previous days prevail. In the RL growth is inhibited or even completely restrained, compared with the same particles in the ML. Overall, the soundings demonstrate that particles are formed inside the ML, but they can prevail and be displaced and stored at upper levels and continue to evolve on following days.

In this campaign we could not measure in the earliest stages of NPF due to safety requirements of the balloon flights early in the morning. We think it is important for future work to carry out soundings during the nucleation phase of the episodes. However, miniaturized instruments able to measure smaller particles would be needed, which are not available at the present time. This would allow us to determine whether secondary NPF takes place throughout the ML or occurs at the surface and is transported upwards by convection afterwards. If the former were true, then locations with high ML could produce more secondary particles than we have considered, and they could affect a larger population, or influence climate to a greater extent.

Additionally, a few nocturnal bursts of nucleation mode particles were observed in the urban stations, for which further research is needed to elucidate their origin.

We cannot determine whether the NPF episodes were triggered by the pollution generated in the city that extended to the region, or the events are caused by a broader phenomenon. In either way, it can be concluded that in summer the particle number concentrations are dominated by NPF in a wide area. The impact of traffic emissions on concentrations of UFP is much smaller than those of NPF, even near the city center where the pollution load is at the highest. This result is in line with other studies performed in cities from high insolation regions (e.g. Kulmala et al., 2016). Given the extent of the episodes, the health effects of NPF can affect a vast number of people, considering that the Madrid metropolitan area with more than 6 million inhabitants is the most populated area in Spain, and one of the most populated in Europe (UN, 2008). For this reason, we believe that the study of health effects related to newly-formed particle inhalation is crucial.

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References

Ahlm, L., Liu, S., Day, D.A., Russell, L.M., Weber, R., Gentner, D.R., Goldstein, A.H., Digangi, J.P., Henry, S.B., Keutsch, F.N., Vandenboer, T.C., Markovic, M.Z., Murphy, J.G., Ren, X., Scheller, S.: Formation and growth of ultrafine particles from secondary sources in Bakersfield, California. Journal of Geophysical Research Atmospheres 117. doi:10.1029/2011JD017144, 2012.

5   Alonso-Blanco, E., Gómez-Moreno, F. J., Núñez, L., Pujadas, M., Cusack, M., Artíñano, B.: Aerosol particle shrinkage event phenomenology in a South European suburban area during 2009–2015, Atmos. Environ. 160, 154-164, 2017.

Beddows, D.C.S., Harrison, R.M., Green, D.C., Fuller, G.W.: Receptor modelling of both particle composition and size distribution from a background site in London, UK. Atmos. Chem. Phys. Discuss. 15, 10123-10162, 2015.

Boulon, J., Sellegri, K., Venzac, H., Picard, D., Weingartner, E., Wehrle, G., Collaud Coen, M., Bütikofer, R., Flückiger, E.,
10   Baltensperger, U., Laj, P.: New particle formation and ultrafine charged aerosol climatology at a high-altitude site in the Alps (Jungfraujoch, 3580 m a.s.l., Switzerland). Atmos. Chem. Phys. 10, 9333-9349, 2010.

Boy, M. and Kulmala, M.: Nucleation events in the continental boundary layer: Influence of physical and meteorological parameters. Atmos. Chem. Phys. 2, 1-16, 2002.

Boy, M., Karl, T., Turnipseed, A., Mauldin, R. L., Kosciuch, E., Greenberg, J., Rathbone, J., Smith, J., Held, A., Barsanti,
15   K., Wehner, B., Bauer, S., Wiedensohle, A., Bonn, B., Kulmala, M., Guenther, A.: New particle formation in the Front Range of the Colorado Rocky Mountains, Atmos. Chem. Phys. 8, 1577- 1590, 2008.

Brines, M., Dall'Osto, M., Beddows, D.C.S., Harrison, R.M., Querol, X.: Simplifying aerosol size distributions modes simultaneously detected at four monitoring sites during SAPUSS. Atmos. Chem. Phys. 14, 2973-2986, 2014.

Brines, M., Dall'Osto, M., Beddows, D.C.S., Harrison, R.M., Gómez-Moreno, F., Núñez, L., Artíñano, B., Costabile, F.,
20   Gobbi, G.P., Salimi, F., Morawska, L., Sioutas, C., Querol, X.: Traffic and nucleation events as main sources of ultrafine particles in high-insolation developed world cities. Atmos. Chem. Phys. 15, 5929-5945, 2015.

Buonanno, G. and Morawska, L.: Ultrafine particle emission of waste incinerators and comparison to the exposure of urban citizens. Waste Manage. 37, 75-81, 2005.

Carslaw, D. C. and Ropkins, K.: openair - an R package for air quality data analysis. Environmental Modelling & Software.
25   Volume 27-28, 52-61, 2012.

Charron, A. and Harrison, R.M.: Primary particle formation from vehicle emissions during exhaust dilution in the roadside atmosphere, Atmos. Environ. 37, 4109-4119, 2003.

Cheung, H.C., Morawska L., Ristovski Z.D.: Observation of new particle formation in subtropical urban environment. Atmos. Chem. Phys. 11, 3823–3833, 2011.

30   Coleman, B.K., Lunden, M.M., Destaillats, H., Nazaroff, W.W.: Secondary organic aerosol from ozone-initiated reactions with terpene-rich household products Atmospheric Environment 42, 8234-8245, 2008.

Costabile, F., Birmili, W., Klose, S., Tuch, T., Wehner, B., Wiedensohler, A., Franck, U., König, K., and Sonntag, A.: Spatiotemporal variability and principal components of the particle number size distribution in an urban atmosphere. Atmos. Chem. Phys. 9, 3163-3195, 2009.

35   Crespí, S.N., Artíñano, B., Cabal, H.: Synoptic classification of the mixed-layer height evolution. Journal of Applied Meteorology, 34, 1668-1677, 1995.

Cusack, M., Pérez, N., Pey, J., Alastuey, A., Querol, X.: Variability of submicrometer particle number size distributions in the western Mediterranean regional background. Tellus B. 65, 19243, doi:10.3402/tellusb.v65i0.19243, 2013a.

Cusack, M., Alastuey, A., Querol, X.: Case studies of new particle formation and evaporation processes in the western Mediterranean regional background. Atmospheric Environment 81, 651-659, 2013b.

Dall'Osto, M., Beddows, D.C.S., Pey, J., Rodriguez, S., Alastuey, A., Harrison, R.M., Querol, X.: Urban aerosol size distributions over the Mediterranean city of Barcelona, NE Spain. Atmos. Chem. Phys. 12, 10693-10707, 2012.

Dall'Osto, M., Querol, X., Alastuey, A., O'Dowd, C., Harrison, R.M., Wenger, J., Gómez-Moreno, F.J.: On the spatial distribution and evolution of ultrafine particles in Barcelona. Atmos. Chem. Phys. 13, 741-759, 2013.

Dal Maso, M., Kulmala, M., Riipinen, I., Wagner, R., Hussein, T., Aalto, P. P., Lehtinen, K. E. J.: Formation and growth of fresh atmospheric aerosols: eight years of aerosol size distribution data from SMEAR II, Hyytiälä, Finland. Boreal Env. Res. 10, 323-336, 2005.

El Haddad, I., D'Anna, B., Temime-Roussel, B., Nicolas, M., Boreave, A., Favez, O., Voisin, D., Sciare, J., George, C., Jaffrezo, J.-L., Wortham, H., and Marchand, N.: Towards a better understanding of the origins, chemical composition and aging of oxygenated organic aerosols: case study of a Mediterranean industrialized environment, Marseille, Atmos. Chem. Phys., 13, 7875-7894, doi:10.5194/acp-13-7875-2013, 2013.

García, M.I., Rodríguezs, S., González, Y., García, R.D.: Climatology of new particle formation at Izaña mountain GAW observatory in the subtropical North Atlantic, Atmos. Chem. Phys. 14, 8, 3865-3881, 2014.

Gómez-Moreno, F.J., Pujadas, M., Plaza, J., Rodríguez-Maroto, J.J., Martínez-Lozano, P., Artíñano, B.: Influence of seasonal factors on the atmospheric particle number concentration and size distribution in Madrid, Atmos. Environ. 45, 3199-3180, 2011.

Graus, M., Muller, M., Hansel, A.: High Resolution PTR-TOF: Quantification and Formula Confirmation of VOC in Real Time, Journal of the American Society For Mass Spectrometry, 21, 1037-1044, 2010.

Hofman, J., Staelensa, J., Cordell, R., Stroobants, C., Zikova, N., Hama, S.M.L., Wyche, K.P., Kosf, G.P.A., Van Der Zeeg, S., Smallbone, K.L., Weijers, E.P., Monks, P.S.: Ultrafine particles in four European urban environments: Results from a new continuous long-term monitoring network. Atmospheric Environment 136, 68-81, 2016.

Hudda, N., Gould, T., Hartin, K., Larson, T.V., Fruin, S.A.: Emissions from an International Airport Increase Particle Number Concentrations 4-fold at 10 km Downwind. Environ. Sci. Technol., 48 (12), 6628-6635, 2014.

Hussein, T., Dal Maso, M., Petäjä, T., Koponen, I., Paatero, P., Aalto, P., Hämeri, K., Kulmala, M.: Evaluation of an automatic algorithm for fitting the particle number size distributions. Boreal Env. Res., 10, 337-355, 2005.

Hussein, T., Martikainen, J., Junninen, H., Sogacheva, L., Wagner, R., Dal Maso, M., Riipinen, I., Aalto, P. P., Kulmala, M.: Observation of regional new particle formation in the urban atmosphere. Tellus B, 60: 509-521, 2008. doi:10.1111/j.1600-0889.2008.00365.x

Johnson, G.R., Juwono, A.M., Friend, A.J., Cheung, H.-C., Stelcer, E., Cohen, D., Ayoko, G.A., Morawska, L.: Relating urban airborne particle concentrations to shipping using carbon based elemental emission ratios. Atmospheric Environment. 95, 525-536, 2014.

Kecorius, S., Kivekäs, N., Kristenssond, A., Tuch, T., Covert, D.S., Birmili, W. Lihavainen, H., Hyvärinen, A.-P., Martinsson, J., Sporre, M.K., Swietlicki, E., Wiedensohler, A., Ulevicius, V.: Significant increase of aerosol number concentrations in air masses crossing a densely trafficked sea area. Oceanologia 58, 1, 1-12, 2016.

Keuken, M.P., Moerman, M., Zandveld, P., Henzing, J.S., Hoek, G.: Total and size-resolved particle number and black carbon concentrations in urban areas near Schiphol airport (the Netherlands). Atmospheric Environment 104, 132-142, 2015.

Kirkby, J., Duplissy, J., Sengupta, K., Frege, C., Gordon, H., Williamson, C., Heinritzi, M., Simon, M., Yan, C., Almeida, J., Trostl, J., Nieminen, T., Ortega, I.K., Wagner, R., Adamov, A., Amorim, A., Bernhammer, A.K., Bianchi, F., Breitenlechner, M., Brilke, S., Chen, X., Craven, J., Dias, A., Ehrhart, S., Flagan, R.C., Franchin, A., Fuchs, C., Guida, R., Hakala, J., Hoyle, C.R., Jokinen, T., Junninen, H., Kangasluoma, J., Kim, J., Krapf, M., Kurten, A., Laaksonen, A., Lehtipalo, K., Makhmutov, V., Mathot, S., Molteni, U., Onnela, A., Perakyla, O., Piel, F., Petaja, T., Praplan, A.P., Pringle, K., Rap, A., Richards, N.A.D., Riipinen, I., Rissanen, M.P., Rondo, L., Sarnela, N., Schobesberger, S., Scott, C.E., Seinfeld, J.H., Sipila, M., Steiner, G., Stozhkov, Y., Stratmann, F., Tomé, A., Virtanen, A., Vogel, A.L., Wagner, A.C., Wagner, P.E., Weingartner, E., Wimmer, D., Winkler, P.M., Ye, P., Zhang, X., Hansel, A., Dommen, J., Donahue, N.M., Worsnop, D.R., Baltensperger, U., Kulmala, M., Carslaw, K.S., Curtius, J.: Ion-induced nucleation of pure biogenic particles. Nature 533, 521–526, doi:10.1038/nature17953, 2016.

Kittelson, D.B., Watts, W.F., Johnson, J.P.: On-road and laboratory evaluation of combustion aerosols – Part1: Summary of diesel engine results. J. Aerosol Sci. 37, 913-930, 2006.

Kontkanen, J., Lehtipalo, K., Ahonen, L., Kangasluoma, J., Manninen, H.E., Hakala, J., Rose, C., Sellegri, K., Xiao, S., Wang, L., Qi, X., Nie, W., Ding, A., Yu, H., Lee, S., Kerminen, V.-M., Petäjä, T., Kulmala, M.: Measurements of sub-3 nm particles using a particle size magnifier in different environments: from clean mountain top to polluted megacities. Atmos. Chem. Phys., 17, 2163-2187, 2017.

Kulmala, M., Pirjola, L., Mäkelä, J.M.: Stable Sulphate Clusters as a Source of New Atmospheric Particles, Nature, 404, 66-69, 2000.

Kulmala, M., Vehkamehk, H., Pet, P.T., Dal Maso, M., Lauri, A., Kerminen, V.-M., Birmili, W., McMurry, P.: Formation and growth rates of ultrafine atmospheric particles: a review of observations. J. Aerosol Sci. 35, 143-176, 2004.

Kulmala, M. and Kerminen, V.-M.: On the formation and growth of atmospheric nanoparticles, Atmos. Research 90, 132-150, 2008.

Kulmala, M., Petäjä, T., Nieminen, T., Sipilä, M., Manninen, H. E., Lehtipalo, K., Dal Maso, M., Aalto, P. P., Junninen, H., Paasonen, P., Riipinen, I., Lehtinen, K. E. J., Laaksonen, A., Kerminen, V-M.: Measurement of the nucleation of atmospheric aerosol particles. Nature Protocols, 7, 1651-1667, 2012.

Kulmala, M., Luoma, K., Virkkula, A., Petäjä, T., Paasonen, P., Kerminen, V.M., Nie, W., Qi, X., Shen, Y., Chi, X., Ding, A.: On the mode-segregated aerosol particle number concentration load: Contributions of primary and secondary particles in Hyytiälä and Nanjing. Boreal Environ. Res. 21, 319–331, 2016.

Kumar, P. and Morawska, L.: Recycling Concrete: An Undiscovered Source of Ultrafine Particles. Atmos. Environ 90, 51-58, 2014.

Kumar, P., Morawska, L., Birmili, W., Paasonen, P., Hu, M., Kulmala, M., Harrison, R.M., Norford, L., Britter, R.: Ultrafine particles in cities, Environ. Int. 66, 1-10, 2014.

Lee, H.-K., Hwang, I.-K., Ahn, K.-H.: Development and Evaluation of Hy-CPC. Particle and Aerosol Research 10, 93-97, 2014.

Lee, H.-K., Eun, H.-R., Lee, G.-H., Ahn, K.-H.: Development and evaluation of Hy-SMPS, Particle and Aerosol Research 11, 57-61, 2015.

Ma, N. and Birmili, W.: Estimating the contribution of photochemical particle formation to ultrafine particle number averages in an urban atmosphere. Science of the Total Environment 512–513, 154-166, 2015.

Manninen, H. E., Nieminen, T., Asmi, E., Gagné, S., Häkkinen, S., Lehtipalo, K., Aalto, P., Vana, M., Mirme, A., Mirme, S., Hõrrak, U., Plass-Dülmer, C., Stange, G., Kiss, G., Hoffer, A., Töro, N., Moerman, M., Henzing, B., De Leeuw, G., Brinkenberg, M., Kouvarakis, G. N., Bougiatioti, A., Mihalopoulos, N., O'Dowd, C., Ceburnis, D., Arneth, A., Svenningsson, B., Swietlicki, E., Tarozzi, L., Decesari, S., Facchini, M. C., Birmili, W., Sonntag, A., Wiedensohler, A., Boulon, J., Sellegri, K., Laj, P., Gysel, M., Bukowiecki, N., Weingartner, E., Wehrle, G., Laaksonen, A., Hamed, A., Joutsensaari, J., Petäjä, T., Kerminen, V. M. and Kulmala, M.: EUCAARI ion spectrometer measurements at 12 European sites-analysis of new particle formation events, Atmos. Chem. Phys., 10(16), 7907–7927, doi:10.5194/acp-10-7907-2010, 2010.

Masiol, M., Harrison, R. M., Vu, T. V., Beddows, D. C. S.: Sources of sub-micrometre particles near a major international airport, Atmos. Chem. Phys., 17, 12379-12403, doi:10.5194/acp-17-12379-2017, 2017.

Minguillón, M.C., Brines, M., Pérez, N., Reche, C., Pandolfi, M., Fonseca, A.S., Amato, F., Alastuey, A., Lyasota, A., Codina, B., Lee, H.-K., Eun, H.-R., Ahn, K.-H., Querol, X.: New particle formation at ground level and in the vertical column over the Barcelona area. Atmospheric Research 164–165, 118–130, 2015.

Nie, W., Ding, A., Wang, T., Kerminen, V.M., George, C., Xue, L., Wang, W., Zhang, Q., Petäjä, T., Qi, X., Gao, X., Wang, X., Yang, X., Fu, C., Kulmala, M.: Erratum: Polluted dust promotes new particle formation and growth, Sci. Rep., doi:10.1038/srep08949, 2015.

O'Dowd, C., Monahan, C., Dall'Osto, M.: On the occurrence of open ocean particle production and growth events, Geophys. Res. Lett., 37, L19805, doi:10.1029/2010GL044679, 2010.

Paasonen, P., Kupiainen, K., Klimont, Z., Visschedijk, A., Denier van der Gon, H. A. C., and Amann, M.: Continental anthropogenic primary particle number emissions, Atmos. Chem. Phys., 16, 6823-6840, doi:10.5194/acp-16-6823-2016, 2016.

Park, J.H., Goldstein, A.H., Timkovsky, J., Fares, S., Weber, R., Karlik, J., Holzinger, R.: Active atmosphere–ecosystem exchange of the vast majority of detected volatile organic compounds. Science, 341, 643–647, 2013.

Pey, J., Rodríguez, S., Querol, X., Alastuey, A., Moreno, T., Putaud, J.P., Van Dingenen, R.: Variations of urban aerosols in the western Mediterranean. Atmos. Environ. 42, 9052-9062, 2008.

Pey, J., Querol, X., Alastuey, A., Rodríguez, S., Putaud, J. P., Van Dingenen, R.: Source Apportionment of urban fine and ultrafine particle number concentration in a Western Mediterranean city. Atmos. Environ. 43, 4407-4415, 2009.

Plaza, J., Pujadas, M., Artíñano, B.: Formation and Transport of the Madrid Ozone Plume. J. Air & Waste Manage. Assoc. 47, 766-774, 1997.

Qian, S., Sakurai, H., McMurry, P.H.: Characteristics of regional nucleation events in urban East St. Louis. Atmospheric Environment, 41, 4119-4127, 2007.

Querol, X., Gangoiti, G., Mantilla, E., Alastuey, A., Minguillón, M. C., Amato, F., Reche, C., Viana, M., Moreno, T., Karanasiou, A., Rivas, I., Pérez, N., Ripoll, A., Brines, M., Ealo, M., Pandolfi, M., Lee, H.-K., Eun, H.-R., Park, Y.-H., Escudero, M., Beddows, D., Harrison, R.M., Bertrand, A., Marchand, N., Lyasota, A., Codina, B., Olid, M., Udina, M., Jiménez-Esteve, B., Soler, M.R., Alonso, L., Millán, M., Ahn, K.-H.: Phenomenology of high-ozone episodes in NE Spain. Atmos. Chem. Phys. 17, 2817-2838, 2017.

Querol, X., Alastuey, A., Gangoiti, G., Perez, N., Lee, H. K., Eun, H. R., Park, Y., Mantilla, E., Escudero, M., Titos, G., Alonso, L., Temime-Roussel, B., Marchand, N., Moreta, J. R., Revuelta, M. A., Salvador, P., Artíñano, B., García dos Santos, S., Anguas, M., Notario, A., Saiz-Lopez, A., Harrison, R. M., Millán, M., and Ahn, K.-H.: Phenomenology of

summer ozone episodes over the Madrid Metropolitan Area, central Spain, Atmos. Chem. Phys., 18, 6511-6533, https://doi.org/10.5194/acp-18-6511-2018, 2018.

Reche, C., Querol, X., Alastuey, A., Viana, M., Pey, J., Moreno, T., Rodríguez, S., González, Y., Fernández-Camacho, R., de la Rosa, J., Dall'Osto, M., Prévôt, A.S.H., Hueglin, C., Harrison, R.M., Quincey, P.: New considerations for PM, Black Carbon and particle number concentration for air quality monitoring across different European cities. Atmos. Chem. Phys. 11, 6207-6227, 2011.

Roberts, J. M.: Measurement of the Henry's law coefficient and first order loss rate of PAN in n-octanol, Geophys. Res. Lett., 32, L08803, doi:10.1029/2004gl022327, 2005.

Robinson, A.L., Donahue, N.M., Shrivastava, M.K., Weitkamp, E.A., Sage, A.M., Grieshop, A.P., Lane, T.E., Pierce, J.R., Pandis, S.N.: Rethinking organic aerosols: semivolatile emissions and photochemical aging. Science 31, 1259-1262, 2007.

Rönkkö, T., Kuuluvainen, H., Karjalainen, P., Keskinen, J., Hillamo, R., Niemi, J. V, Pirjola, L., Timonen, H.J., Saarikoski, S., Saukko, E., Järvinen, A., Silvennoinen, H., Rostedt, A., Olin, M., Yli-Ojanperä, J., Nousiainen, P., Kousa, A., Dal Maso, M.: Traffic is a major source of atmospheric nanocluster aerosol. Proc. Natl. Acad. Sci. U. S. A. 114, 7549–7554, doi:10.1073/pnas.1700830114, 2017.

Saiz-Lopez, A., Borge, R., Notario, A., Adame, J.A., De la Paz, D., Querol, X., Artíñano, B., Gomez-Moreno, F.J., Cuevas, C.A.: Unexpected increase in the oxidation capacity of the urban atmosphere of Madrid, Spain, Sci. Rep., 7, 45956, doi:10.1038/srep45956, 2017.

Salma, I., Borsós, T., Weidinger, T., Aalto, P., Hussein, T., Dal Maso, M., Kulmala, M.: Production, growth and properties of ultrafine atmospheric aerosol particles in an urban environment. Atmos. Chem. Phys. 11, 1339-1353, 2011.

Salma, I., Borsos T., Nemeth Z., Weidiger T., Aalto P., Kulmala M.: Comparative study of ultrafine atmospheric aerosol within a city. Atmos. Environ., 92, 154-161, 2014.

Salma, I., Németh, Z., Kerminen, V-M., Aalto, P., Nieminen, T., Weidinger, T., Molnár, Á., Imre, K., Kulmala, M.: Regional effect on urban atmospheric nucleation, Atmospheric Chemistry and Physics 16, 8715-8728, 2016.

Salvador, P., Artíñano, B., Viana, M., Alastuey, A., Querol, X.: Multicriteria approach to interpret the variability of the levels of particulate matter and gaseous pollutants in the Madrid metropolitan area, during the 1999-2012 period. Atmospheric Environment 109, 205-216, 2015.

Sellegri, K., Laj, P., Venzac, H., Boulon, J., Picard, D., Villani, P., Bonasoni, P., Marinoni, A., Cristofanelli, P., Vuillermoz, E.: Seasonal variations of aerosol size distributions based on longterm measurements at the high altitude Himalayan site of Nepal Climate Observatory-Pyramid (5079 m), Nepal. Atmos. Chem. Phys. 10, 10679-10690, 2010.

Sipila, M., Berndt, T., Petaja, T., Brus, D., Vanhanen, J., Stratmann, F., Patokoski, J., Mauldin, R. L., Hyvärinen, A. P., Lihavainen, H. and Kulmala, M.: The role of sulfuric acid in atmospheric nucleation, Science, 327(5970), 1243–1246, doi:10.1126/science.1180315, 2010.

Shi, J.P. and Harrison, R.M.: Investigation of ultrafine particle formation during diesel exhaust dilution. Environ. Sci. Technol. 33, 3730-3736, 1999.

Shi, J.P., Mark, D., Harrison, R.M.: Characterization of particles from a current technology heavy-duty diesel engine, Environ. Sci. Technol., 34, 748-755, 2000.

Skrabalova, L., Zikova, N., Zdimal, V.: Shrinkage of newly formed particles in an urban environment. Aerosol Air Qual. Res. 15, 1313-1324, 2015.

Stolzenburg, M.R., McMurry, P.H., Sakurai, H., Smith, J.N., Lee, M.R., Eisele, F.L., Clement, C.F.: Growth rates of freshly nucleated atmospheric particles in Atlanta. Journal of Geophysical Research, 110, D22S05, 2005.

Stratmann, F., Siebert, H., Spindler, G., Wehner, B., Althausen, D., Heintzenberg, J., Hellmuth, O., Rinke, R., Schmieder, U., Seidel, C., Tuch, T., Uhrner, U., Wiedensohler, A., Wandinger, U., Wendisch, M., Schell, D., and Stohl, A.: New-particle formation events in a continental boundary layer: first results from the SATURN experiment, Atmos. Chem. Phys., 3, 1445-1459, doi:10.5194/acp-3-1445-2003, 2003.

Tröstl, J., Chuang, W. K., Gordon, H., Heinritzi, M., Yan, C., Molteni, U., Ahlm, L., Frege, C., Bianchi, F., Wagner, R., Simon, M., Lehtipalo, K., Williamson, C., Craven, J. S., Duplissy, J., Adamov, A., Almeida, J., Bernhammer, A. K., Breitenlechner, M., Brilke, S., Dias, A., Ehrhart, S., Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Gysel, M., Hansel, A., Hoyle, C. R., Jokinen, T., Junninen, H., Kangasluoma, J., Keskinen, H., Kim, J., Krapf, M., Kürten, A., Laaksonen, A., Lawler, M., Leiminger, M., Mathot, S., Möhler, O., Nieminen, T., Onnela, A., Petäjä, T., Piel, F. M., Miettinen, P., Rissanen, M. P., Rondo, L., Sarnela, N., Schobesberger, S., Sengupta, K., Sipilä, M., Smith, J. N., Steiner, G., Tomè, A., Virtanen, A., Wagner, A. C., Weingartner, E., Wimmer, D., Winkler, P. M., Ye, P., Carslaw, K. S., Curtius, J., Dommen, J., Kirkby, J., Kulmala, M., Riipinen, I., Worsnop, D. R., Donahue, N. M. and Baltensperger, U.: The role of low-volatility organic compounds in initial particle growth in the atmosphere, Nature, 533(7604), 527–531, doi:10.1038/nature18271, 2016.

Stull, R.B.: An introduction to boundary layer meteorology. Kluwer Academic Publishers, Dordrecht, Boston and London, 1988.

United Nations, Department of Economic and Social Affairs: World Urbanization Prospects (2007 revision). https://www.un.org/esa/population/publications/wup2007/2007WUP_Highlights_web.pdf, 2008.

Uhrner, U., von Lowis, S., Vehkamaki, H., Wehner, B., Brasel, S., Hermann, M., Stratmann, F., Kulmala, M., Wiedensohler, A.: Dilution and aerosol dynamics within a diesel car exhaust plume – CFD simulations of on-road conditions. Atmos. Environ. 41, 7440-7461, 2007.

Vakkari, V., Laakso, H., Kulmala, M., Laaksonen, A., Mabaso, D., Molefe, M., Kgabi, N., Laakso, L.: New particle formation events in semi-clean South African Savannah. Atmos. Chem. Phys. 11, 3333-3346, 2011.

von Bismarck-Osten, C., Birmili, W., Ketzel, M., Massling, A., Petäjä, T., Weber, S.: Characterization of parameters influencing the spatio-temporal variability of urban particle number size distributions in four European cities. Atmos. Environ. 77, 415-429, 2013.

Wang, N., Sun, X., Chen, J., Li, X.: Sci. Heterogeneous Nucleation of Trichloroethylene Ozonation Products in the Formation of New Fine Particles. Rep., 7: 42600, doi:10.1038/srep42600, 2017.

Wegner, T., Hussein, T., Hämeri, K, Vesala, T., Kulmala, M., Weber, S.: Properties of aerosol signature size distributions in the urban environment as derived by cluster analysis. Atmos. Environ. 61, 350-360, 2012.

Wehner, B. and Wiedensohler, A.: Long term measurements of submicrometer urban aerosols: statistical analysis for correlations with meteorological conditions and trace gases, Atmos. Chem. Phys., 3, 867-879, https://doi.org/10.5194/acp-3-867-2003, 2003.

Wehner, B., Siebert, H., Stratmann, F., Tuch, T., Wiedensohler, A., Petäjä, T., Dal Maso, M., Kulmala, M.: Horizontal homogeneity and vertical extent of new particle formation events. Tellus 59B, 362-371, 2007.

Wehner, B., Siebert, H., Ansmann, A., Ditas, F., Seifert, P., Stratmann, F., Wiedensohler, A., Apituley, A., Shaw, R. A., Manninen, H. E., and Kulmala, M.: Observations of turbulence-induced new particle formation in the residual layer, Atmos. Chem. Phys., 10, 4319-4330, doi:10.5194/acp-10-4319-2010, 2010.

Wiedensohler, A., Wehner, B., Birmili, W.: Aerosol number concentrations and size distributions at mountain-rural, urbaninfluenced rural, and urban-background sites in Germany. J. Aerosol Med. 15, 237-243, 2002.

Woo, K.S., Chen, D.R., Pui, D.Y.H., McMurry, P.H.: Measurement of Atlanta aerosol size distributions: observations of ultrafine particle events. Aerosol Sci. Tech. 34, 75-87, 2001.

5 Yao, X., Choi, M.Y., Lau, N.T., Lau, A.P.S., Chan, C.K., Fang, M.: Growth and shrinkage of new particles in the atmosphere in Hong Kong. Aerosol Sci. Tech. 44, 639-650, 2010.

Young, L.-H., Lee, S.-H., Kanawade, V.P., Hsiao, T.-C., Lee, Y.L., Hwang, B.-F., Liou, Y.-J., Hsu, H.-T., Tsai, P.-J.: New particle growth and shrinkage observed in subtropical environments, Atmos. Chem. Phys. 13, 547-564, 2013.

Table 1: Summary of new particle formation events recorded during the campaign showing the starting time, considered as the moment of first detection of the nucleation mode, the final time, considered when the mode reaches 25 nm, the growth rate calculated in that period using SMPS and PSM data, and formation rates at starting time. A star marks the events that are detected simultaneously at all stations and were chosen for further analysis in this work.

	Date	Starting Time (UTC)	Final Time (UTC)	GR (nm h^{-1})	GR$_{PSM}$ (nm h^{-1})	J$_1$ (cm^{-3} s^{-1})
	12/07/2016 (*)	6:20	10:39	3.9	1.9	2.4
CSIC	13/07/2016 (*)	8:30	12:49	2.0	1.1	8.5
	14/07/2016 (*)	8:45	11:53	1.4	6.75	15.7
	12/07/2016 (*)	5:30	9:44	3.0	0.7	1.9
	13/07/2016 (*)	8:50	11:54	4.6	4.3	8.1
ISCIII	14/07/2016 (*)	9:20	10:39	7.6	6.8	6.5
	16/07/2016 (*)	-	-	-	4.3	-
	17/07/2016 (*)	-	-	-	4.4	3.2
	18/07/2016 (*)	10:44	12:20	2.9	1.38	6.8
	13/07/2016 (*)	8:15	13:45	2.5	-	-
CIEMAT	14/07/2016 (*)	9:00	13:10	4.1	-	-
	15/07/2016	8:34	13:08	4.0	-	-
	18/07/2016 (*)	9:09	11:49	2.6	-	-
	12/07/2016	7:27	8:08	3.5	-	-
MJDH -Sounding	13/07/2016	8:39	9:56	5.3	-	-
	14/07/2016	9:00	10:34	6.2	-	-

Figure captions

Figure 1: Particle size distribution at CSIC, CIEMAT and ISCIII from 4 to 20 July 2016. Total particle concentration of particles with diameter greater than 2.5 nm is also shown.

Figure 2: Boxplot of growth rates (GR) determined for the nucleation mode (< 25 nm) during regional new particle formation events at urban (CSIC or CIEMAT) and suburban (ISCIII) stations. The red line represents the median, the upper and lower limits of the boxes represent the 75th and 25th percentiles and the whiskers include 99.3% of the data. Outliers are represented with a red cross.

Figure 3: Averaged daily cycles of (a) total particle concentration in the size range 9-25 nm, (b) total particle concentration in the size range 1.2-2.5 nm measured with PSM at CSIC and ISCIII (c) total concentration of particles >2.5 nm measured with PSM at CSIC and ISCIII (d) Condensation Sink (CS) and (e) Coagulation Sink (CoagS) during regional new particle formation events at urban (CSIC and CIEMAT, solid line) and suburban (ISCIII, dashed line) stations. The hour of the day is UTC. Local time is UTC+2.

Figure 4: Particle size distribution with fitted log-normal modes (black dots) measured during the balloons soundings at Majadahonda on 12 July 2016. An estimation of the mixing layer height is represented with red dots. The altitude of the instrumentation is represented with a white line. Surface level is 630 m above sea level. Time is UTC. Local time is UTC+2.

Figure 5: Particle size distribution with fitted log-normal modes (black dots) measured during the balloons soundings at Majadahonda on 13 July 2016. The altitude of the instrumentation is represented with a white line. Surface level is 630 m above sea level. Time is UTC. Local time is UTC+2.

Figure 6: Particle size distribution with fitted log-normal modes (black dots) measured during the balloons soundings at Majadahonda on 14 July 2016. The altitude of the instrumentation is represented with a white line. An estimation of the mixing layer height is represented with red dots. Surface level is 630 m above sea level. Time is UTC. Local time is UTC+2.

Figure 7: Vertical particle size distribution measured on 14 July during selected soundings.

Figure 8: Boxplot of shrinkage rate (SR) determined during regional NPF events at urban (CSIC and CIEMAT) and suburban (ISCIII) stations according to the starting diameter of the shrinking particles. The considered categories are particles with a starting diameter below 40 nm and particles with a starting diameter between 40 and 100 nm. The red line represents the median, the upper and lower limits of the boxes represent the 75th and 25th percentiles and the whiskers include 99.3% of the data. Outliers are represented with a red cross.

Figure 9: Bipolar plot of (a) total particle concentration in the size range 1.2-4 nm measured with the PSM, (b) Black Carbon, (c) NO₂ and (d) NO concentrations at CSIC urban station, using the wind data registered at CIEMAT. Daylight and nighttime hours are separated according to sunrise (5 UTC) and sunset (20 UTC) hours. The data correspond to the period 11-15 July 2016.
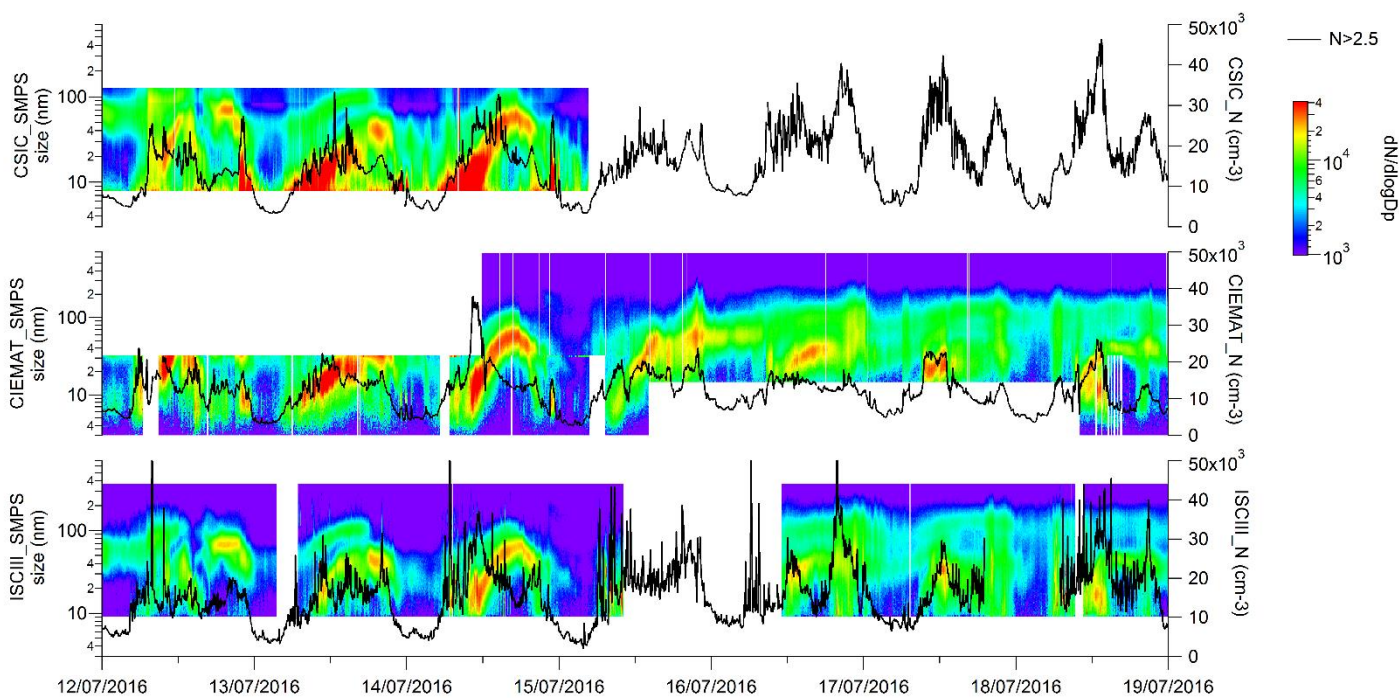
Figure 1
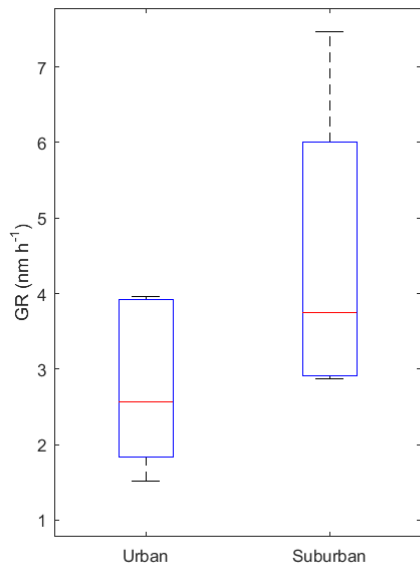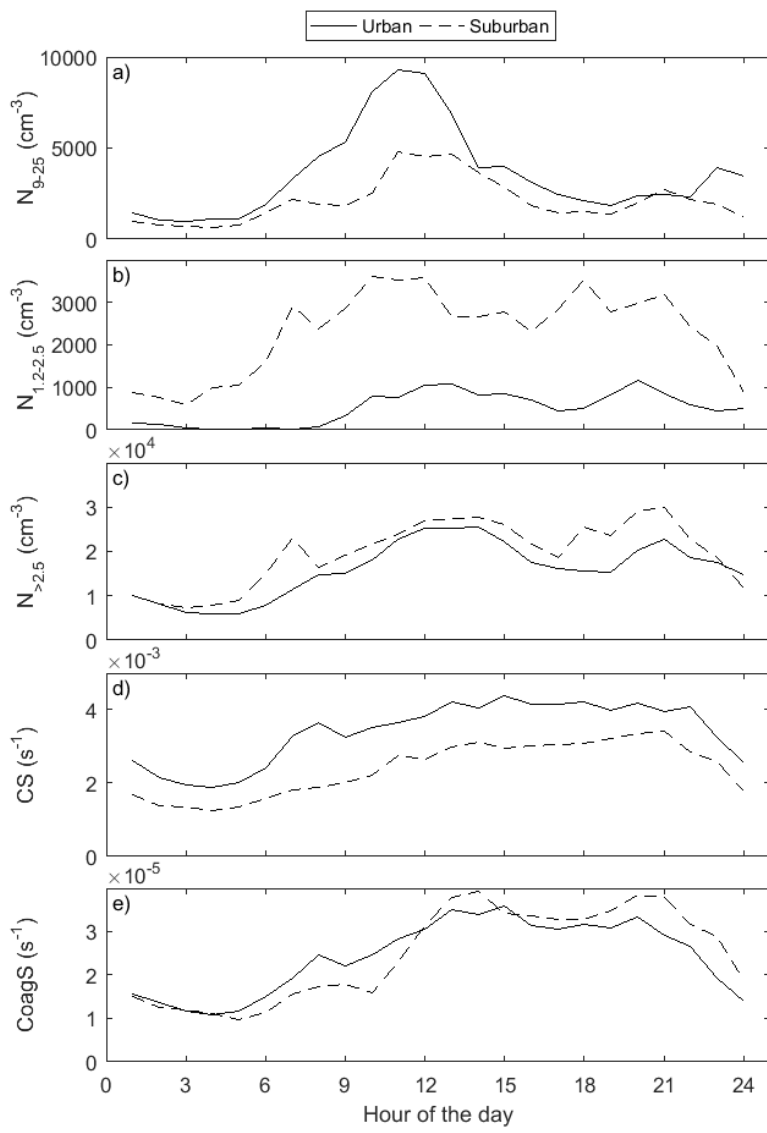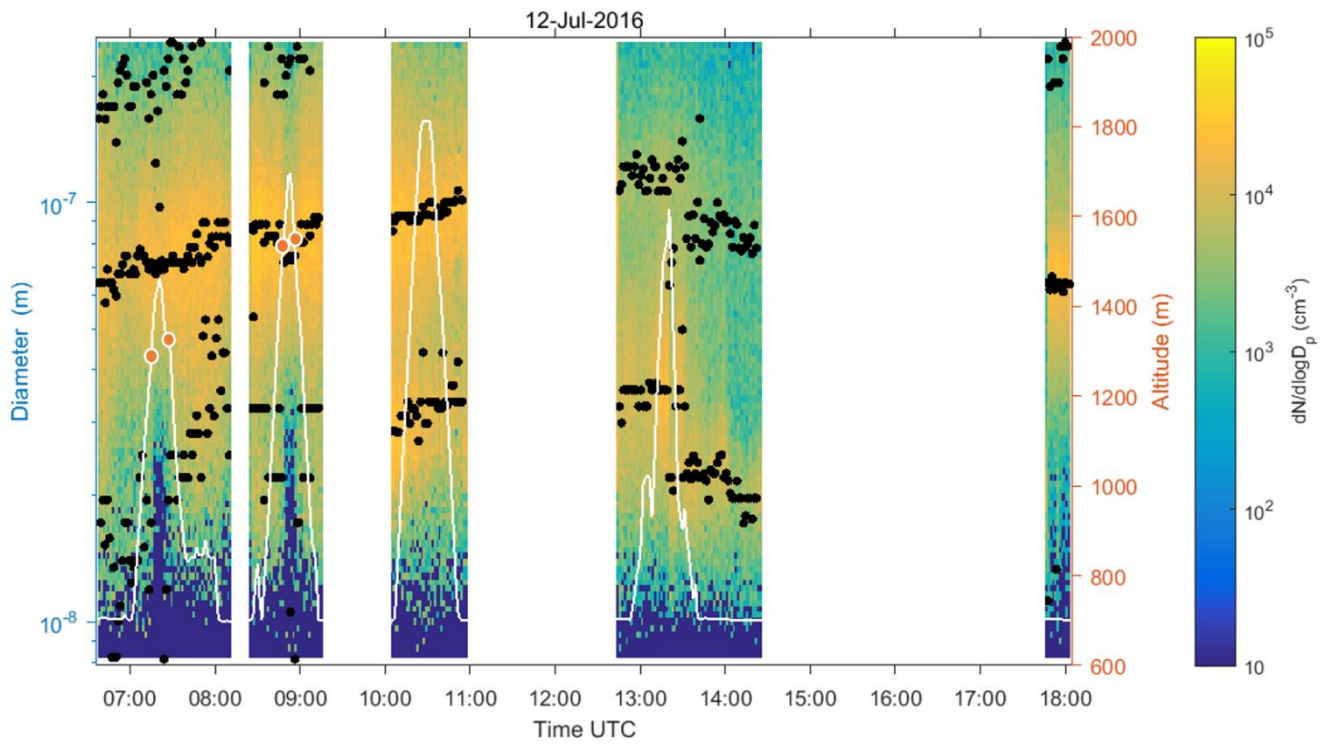
Figure 2

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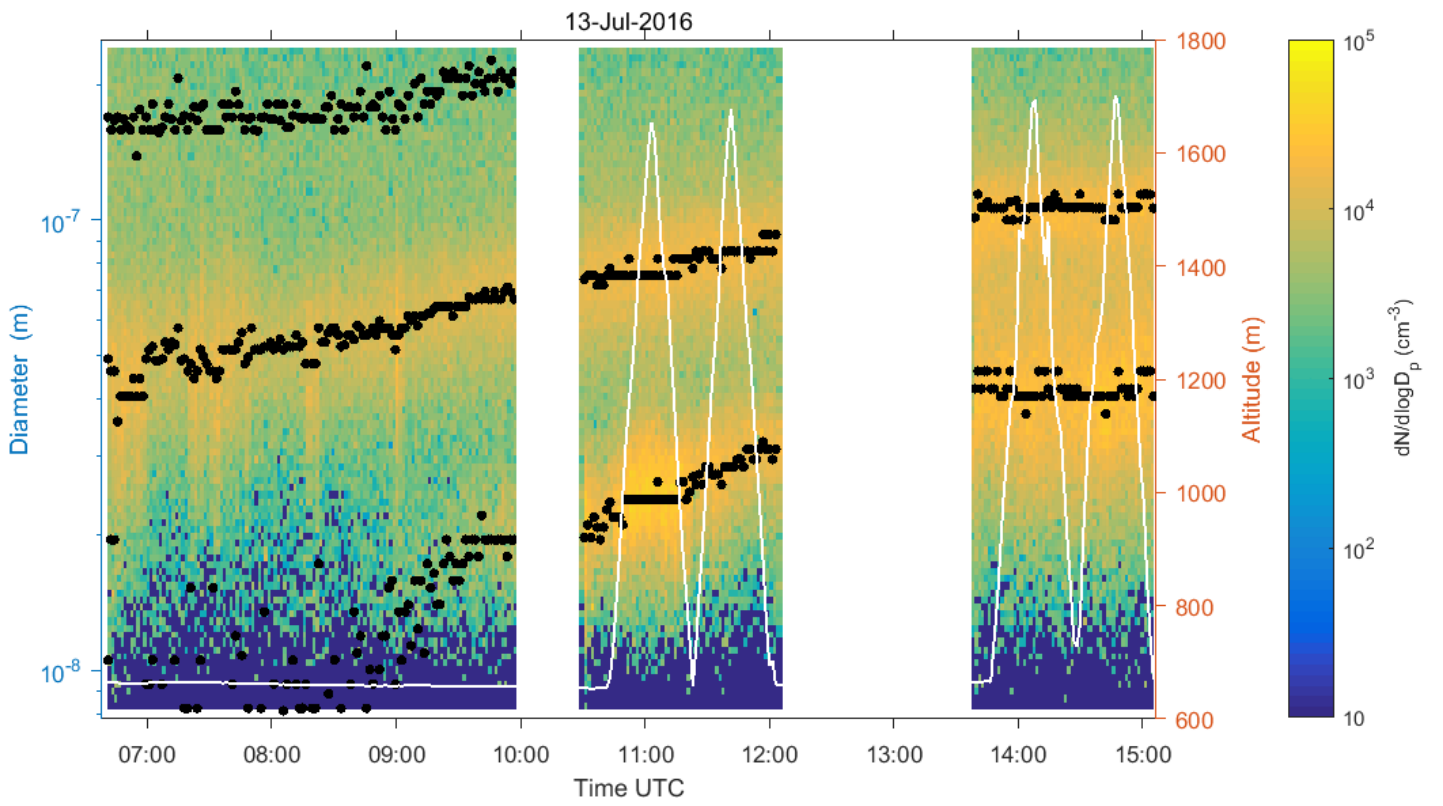
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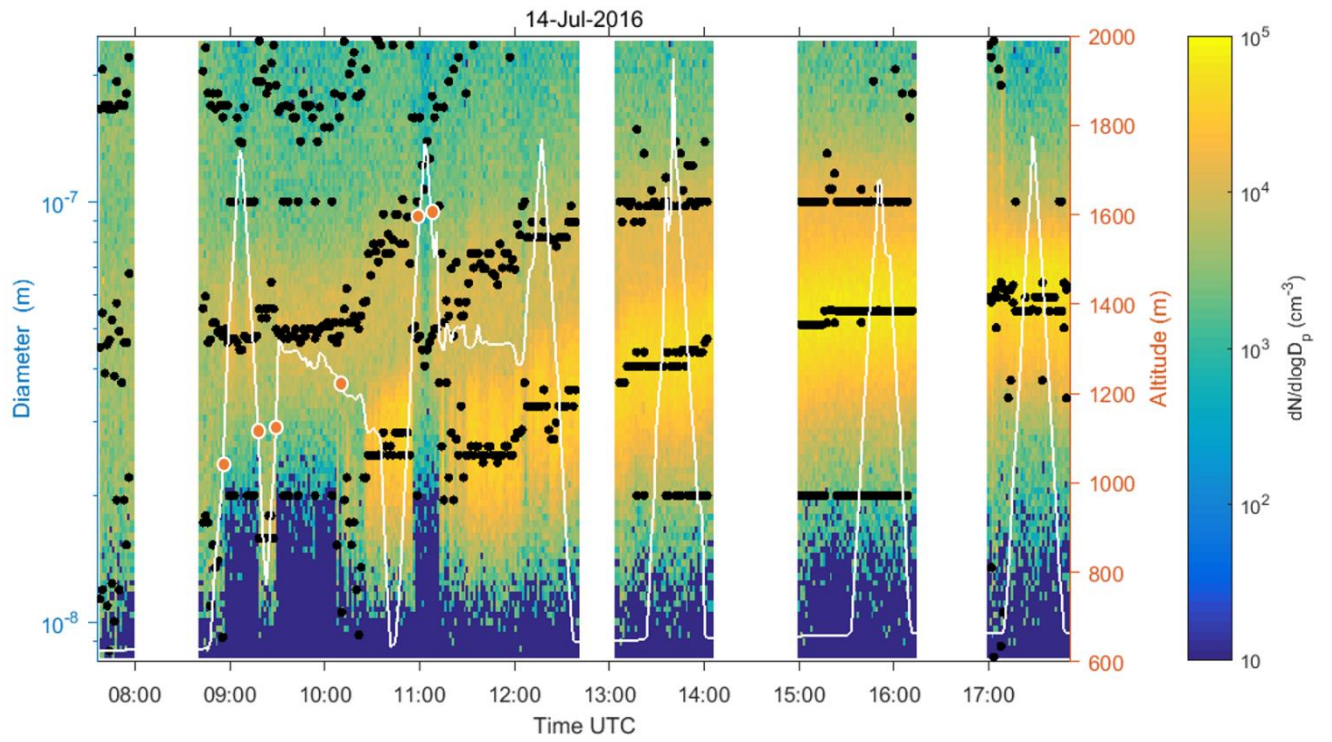
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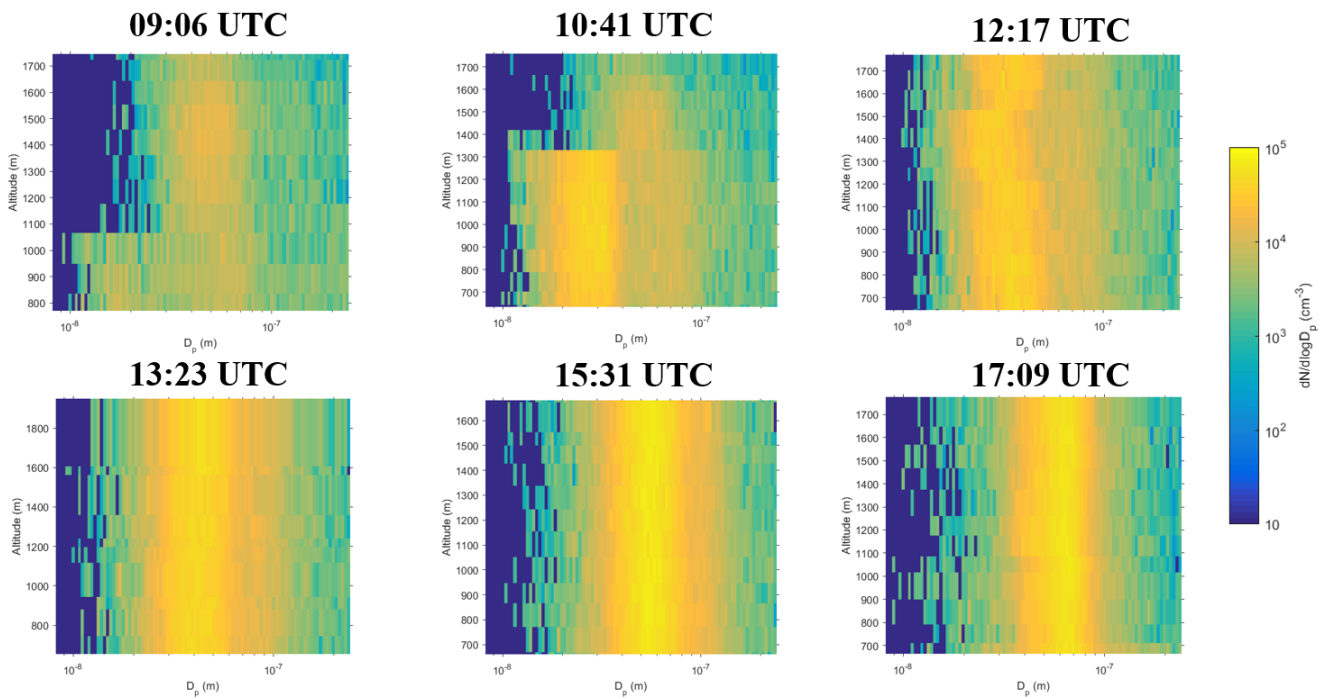
5   Figure 7

Figure 8

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