

1 Is water vapor a key player of the wintertime haze in North China Plain?

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3 Jiarui Wu^{1,4,7}, Naifang Bei², Bo Hu³, Suixin Liu^{1,4}, Meng Zhou⁵, Qiyuan Wang^{1,4}, Xia Li^{1,4,7}, Lang Liu^{1,4,7}, Tian
4 Feng¹, Zirui Liu³, Yichen Wang¹, Junji Cao^{1,4}, Xuexi Tie^{1,4}, Jun Wang⁵, Luisa T. Molina⁶, and Guohui Li^{1,4*}

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6 ¹Key Lab of Aerosol Chemistry and Physics, SKLLQG, Institute of Earth Environment, Chinese Academy of
7 Sciences, Xi'an, Shaanxi, China

8 ²School of Human Settlements and Civil Engineering, Xi'an Jiaotong University, Xi'an, Shaanxi, China

9 ³State Key Laboratory of Atmospheric Boundary Layer Physics and Atmospheric Chemistry, Institute of
10 Atmospheric Physics, Chinese Academy of Sciences, Beijing, 100029, China

11 ⁴CAS Center for Excellence in Quaternary Science and Global Change, Xi'an, China

12 ⁵Department of Chemical and Biochemical Engineering & Interdisciplinary Graduate Program in
13 Geo-Informatics, University of Iowa, Iowa City, Iowa, USA

14 ⁶Molina Center for Energy and the Environment, La Jolla, California, USA

15 ⁷University of Chinese Academy of Science, Beijing, China

16 *Correspondence to: Guohui Li (ligh@ieecas.cn)

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19 **Abstract.** Water vapor has been proposed to amplify the severe haze pollution in China by
20 enhancing the aerosol-radiation feedback (ARF). Observations have revealed that the
21 near-surface PM_{2.5} concentrations ([PM_{2.5}]) generally exhibits an increasing trend with
22 relative humidity (RH) in North China Plain (NCP) during 2015 wintertime, indicating that
23 the aerosol liquid water (ALW) caused by hygroscopic growth could play an important role
24 in the PM_{2.5} formation and accumulation. Simulations during a persistent and heavy haze
25 pollution episode from 05 December 2015 to 04 January 2016 in NCP were conducted using
26 the WRF-Chem model to comprehensively quantify contributions of the ALW effect to
27 near-surface [PM_{2.5}]. The WRF-Chem model generally performs reasonably well in
28 simulating the temporal variations of RH against measurements in NCP. The factor
29 separation approach (FSA) was used to evaluate the contribution of the ALW effect on the
30 ARF, photochemistry, and heterogeneous reactions to [PM_{2.5}]. The ALW not only augments
31 particle sizes to enhance aerosol backward scattering, but also increases the effective radius
32 to favor aerosol forward scattering. The contribution of the ALW effect on the ARF and
33 photochemistry to near-surface [PM_{2.5}] is not significant, generally within 1.0 μg m⁻³ on
34 average in NCP during the episode. Serving as an excellent substrate for heterogeneous
35 reactions, the ALW substantially enhances the secondary aerosol (SA) formation, with an
36 average contribution of 71%, 10%, 26%, and 48% to near-surface sulfate, nitrate, ammonium,
37 and secondary organic aerosol concentrations. Nevertheless, the SA enhancement due to the
38 ALW decreases the aerosol optical depth and increases the effective radius to weaken the
39 ARF, reducing near-surface primary aerosols. The contribution of the ALW total effect to
40 near-surface [PM_{2.5}] is 17.5% on average, which is overwhelmingly dominated by enhanced
41 SA. Model sensitivities also show that when the RH is less than 80%, the ALW progressively

42 increases near-surface $[PM_{2.5}]$, but commences to decrease when the RH exceeding 80% due
43 to the high occurrence frequencies of precipitation.

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49 1 Introduction

50 Atmospheric aerosols or fine particle matters ($PM_{2.5}$) influence the climate directly by
51 scattering and absorbing the solar radiation, and indirectly by serving as cloud condensation
52 nuclei and ice nuclei (Ackerman, 1977; Ackerman and Baker, 1977; Jacobson, 1998, 2002;
53 Penner et al., 2001). Moreover, high levels of $PM_{2.5}$ in the atmosphere also cause severe haze
54 pollution, impairing visibility and exerting deleterious effect on ecological system and human
55 health (Chan and Yao, 2008; Zhang et al., 2013; Kurokawa et al., 2013; Weinhold, 2008;
56 Parrish and Zhu, 2009). In addition to anthropogenic emissions, the poor air quality is
57 generally influenced by stagnant meteorological situations with weak winds and high relative
58 humidity (RH) (Flocas et al., 2009; Quan et al., 2013; Zhang et al., 2014; Bei et al., 2016; Wu
59 et al., 2017; Ding et al., 2017). RH, as an important meteorological factor in the atmosphere,
60 considerably affects the formation, chemical composition, and physical properties of
61 atmospheric aerosols (Seinfeld et al., 2001; Hallquist et al., 2009; Poulain, 2010; Nguyen et
62 al., 2011).

63 As the main constituent in the atmosphere, water vapor directly participates in the
64 atmospheric physical and chemical processes. Since many components of atmospheric
65 aerosols are hygroscopic, they take up water as RH increases (Covert et al., 1972; Pilinis et
66 al., 1989), thereby influencing the aerosol size distribution, chemical composition, mass
67 concentration, and corresponding optical properties as well as radiative effects (Im et al.,
68 2001; Carrico et al., 2003; Randles et al. 2004; Cheng et al., 2008). Wang et al. (2016) have
69 indicated that the ratio of SO_4^{2-} to SO_2 exhibits an exponential increase with RH. Tie et al.
70 (2017) have shown that the sulfate, nitrate, and ammonium concentrations increase from 16
71 to $25 \mu g m^{-3}$, 15 to $23 \mu g m^{-3}$, and 11 to $17 \mu g m^{-3}$, respectively, when RH increases from 60%
72 to 80%. Field measurements in Beijing have demonstrated that the inorganic aerosol fraction
73 increases with increasing RH (Wu et al., 2018). In addition, water vapor also serves as an

74 important medium in the formation of secondary aerosols (SA) through liquid-phase
75 reactions and heterogeneous reactions (Seinfeld and Pandis, 1986; Pilinis et al., 1989). For
76 example, Li et al. (2017) have indicated that the aerosol liquid water (ALW) induced by the
77 wet growth could play a significant role in the sulfate formation and emphasized the
78 importance of bulk aqueous-phase oxidation of SO₂ in ALW and heterogeneous reaction of
79 SO₂ on aerosol surfaces involving ALW. ALW also plays an important role in secondary
80 organic aerosol (SOA) formation (Hastings et al., 2005; Healy et al., 2009; Kamens et al.,
81 2011; Koehler et al., 2004). Numerous studies have investigated the effect of RH on SOA
82 formed from different aromatics during their photochemical oxidation processes (Blando and
83 Turpin, 2000; Cocker et al., 2001; Seinfeld et al., 2001; Zhou et al., 2011; Jia and Xu., 2014).
84 Furthermore, Zhang et al. (2015) have revealed that, as the RH increases from 40% to 85% in
85 the Yangtze River Delta of China, the aerosol scattering and backscattering coefficients
86 increase by 58% and 25%, respectively, and the calculated aerosol direct radiative forcing
87 caused by hygroscopic growth is increased by 47%.

88 In recent years, China has experienced persistent haze pollution with unprecedentedly
89 high PM_{2.5} concentrations during wintertime, particularly in North China Plain (NCP) (Chan
90 and Yao, 2008; He et al., 2001; Kan et al., 2012; Guo et al., 2014; Wang et al., 2014; Fu et al.,
91 2014). A conceptual model based on the aerosol radiation feedback (ARF) has been
92 established to interpret the wintertime heavy haze formation, in which water vapor is
93 considered to play a key role in the progressive accumulation and formation of PM_{2.5}. In
94 winter, when the atmospheric condition is stagnant, air pollutants commence to accumulate in
95 the planetary boundary layer (PBL), favorable for the PM_{2.5} formation. Increasing PM_{2.5}
96 scatters or absorbs the incoming solar radiation to lower the surface temperature and cause
97 anomalous temperature inversion, subsequently suppressing the vertical turbulent diffusion
98 and decreasing the planetary boundary layer height (PBLH) to further trap more air pollutants

99 and water vapor to increase the RH in the PBL. Increasing RH enhances aerosol hygroscopic
100 growth and multiphase reactions and augments the particle size and mass, causing further
101 dimming and decrease of the surface temperature and PBL height (Quan et al., 2013; Tie et
102 al., 2017; Ding et al., 2017). However, few studies have been performed in China to
103 comprehensively quantify the effect of water vapor in the atmospheric physical and chemical
104 process on the PM_{2.5} pollution to further verify the haze formation.

105 The purpose of the present study is to quantitatively evaluate the contribution of aerosol
106 water induced by the aerosol wet growth to PM_{2.5} concentrations in NCP using the Weather
107 Research and Forecast model with Chemistry (WRF-Chem). The model configuration and
108 methodology are described in Section 2. Results and discussions are presented in Section 3,
109 and conclusions and summaries are given in Section 4.

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111 **2 Model and methodology**

112 **2.1 WRF-Chem Model and Configuration**

113 A persistent air pollution episode with high levels of PM_{2.5} from 05 December 2015 to
114 04 January 2016 in NCP was simulated using the WRF-Chem model with modifications by
115 Li et al. (2010, 2011a, b, 2012) from the Molina Center for Energy and the Environment.
116 Figure 1 shows the WRF-Chem model simulation domain and Table 1 provides the model
117 configurations. Detailed model description can be found in Wu et al. (2018a).

118 **2.2 Data and Methodology**

119 The model performance of RH was validated using the hourly measurements in
120 Luancheng, Yucheng, and Jiaozhouwan observed from the Chinese Ecosystem Research
121 Network (CERN). Furthermore, the NCEP reanalysis data was used to compare to the
122 simulated RH distribution. The detailed information of other data used for validation can be
123 found in Wu et al. (2018a).

124 The mean bias (*MB*), root mean square error (*RMSE*) and the index of agreement (*IOA*)
125 were utilized to evaluate the performance of the WRF-Chem model simulations against
126 measurements. To assess the contributions of ALW to the near-surface concentrations of air
127 pollutants in NCP, the factor separation approach (FSA) was used in this study (Stein and
128 Alpert, 1993; Gabusi et al., 2008; Li et al., 2014). Generally, the formation of the secondary
129 atmospheric pollutants, such as O₃, secondary organic aerosol, and nitrate, is a complicated
130 nonlinear process in which its precursors from various emissions sources and transport react
131 chemically or reach equilibrium thermodynamically. Nevertheless, it is not straightforward to
132 evaluate the contributions from different factors in a nonlinear process (Wu et al., 2017). The
133 factor separation approach (FSA) proposed by Stein and Alpert (1993) can be used to isolate
134 the effect of one single factor from a nonlinear process and has been widely used to evaluate
135 source effects. The total effect of one factor in the presence of others can be decomposed into
136 contributions from the factor and that from the interactions of all those factors. Considering
137 that there are two factors *X* and *Y* that influence the formation of secondary pollutants in the
138 atmosphere and also interact with each other. Denoting f_{XY} , f_X , f_Y , and f_0 as the
139 simulations including both of two factors, factor *X* only, factor *Y* only, and none of the two
140 factors, respectively. The contributions of factor *X* and *Y* can be defined as $f_{XY} - f_Y$ and
141 $f_{XY} - f_X$, respectively. Detailed description of the methodology can be found in Wu et al.
142 (2017).

143

144 3 Results and discussions

145 3.1 Relationship between RH and near-surface PM_{2.5} concentrations

146 High RH has been suggested to be an important factor facilitating the SA formation in
147 the atmosphere and aggravating the haze pollution (Sun et al., 2013; Cheng et al., 2015).
148 Figure 2 presents the scatter plot of the near-surface PM_{2.5} concentrations ([PM_{2.5}]) and RH in

149 the winter of 2015 at six typical polluted cities in NCP, including Beijing, Tianjin,
150 Shijiazhuang, Tangshan, Baoding, and Chengde. The observed near-surface [PM_{2.5}] at those
151 six cities display a growing trend with increasing RH, suggesting that the ALW induced by
152 the hygroscopic growth under high RH conditions has potentials to accelerate the PM_{2.5}
153 formation and accumulation. Increasing RH facilitates the aerosol hygroscopic growth and
154 further enhances the ALW, which serves as an efficient medium for promoting the
155 liquid-phase and heterogeneous reactions and accelerating the transformation of reactive
156 gaseous pollutants into the particle phase. Increased ALW also augments the particle size,
157 enhancing the ARF to increase the near-surface [PM_{2.5}]. However, the attenuation of
158 incoming solar radiation caused by the ALW also decreases the photolysis rates, unfavorable
159 for photochemical activities and lowering the atmospheric oxidation capability (AOC). Field
160 measurements show that large fraction of SA in PM_{2.5} has been observed in NCP during
161 wintertime (Sun et al., 2013; Guo et al., 2014; Xu et al., 2015). Therefore, decreased AOC
162 generally does not facilitate the SA formation, particularly with regards to SOA and nitrate,
163 to partially counteract the PM_{2.5} enhancement caused by the ALW. It is also worth noting that
164 since high RH frequently corresponds to atmospheric stagnation, near-surface [PM_{2.5}] also
165 build up under high RH conditions. For example, the humid air mass is subject to being
166 transported from south to NCP under the stagnant weather with weak south winds and
167 meanwhile the PM_{2.5} also accumulates due to the unfavorable dispersion condition.
168 Additionally, when the RH is very high, there also exist the low near-surface [PM_{2.5}] shown
169 in Figure 2, demonstrating that other factors, such as emissions, horizontal transport, vertical
170 exchange, and precipitation, also substantially influence near-surface [PM_{2.5}]. Generally, high
171 occurrence frequency of precipitation coincides with high RH, thus the precipitation washout
172 might constitute one of the most possible reasons for the low near-surface [PM_{2.5}] under high
173 RH situations. Therefore, it is still imperative to verify quantitatively the contribution of the

174 ALW to near-surface [PM_{2.5}].

175 3.2 Model validation

176 The WRF-Chem model simulation of the haze pollution episode in NCP has been
177 comprehensively validated using available measurements in Wu et al. (2018a). In general, the
178 model simulates well the spatial distribution and temporal variation of PM_{2.5}, O₃, NO₂, SO₂
179 and CO mass concentrations compared to observations in NCP. The predicted aerosol species
180 are also in good agreement with the measurement in Beijing. Moreover, the model performs
181 reasonably well in simulating the aerosol optical depth and single scattering albedo, PBL
182 height and downward shortwave flux against measurements.

183 In order to verify the effect of the ALW on near-surface [PM_{2.5}] during the haze
184 pollution episode, the simulated temporal variation of RH was first compared to
185 measurements at Luancheng, Yucheng, and Jiaozhouwan in NCP from 05 December 2015 to
186 04 January 2016 (Figure 3). The WRF-Chem model generally performs well in simulating
187 the hourly variation of RH in these three cities, with *IOAs* of 0.73, 0.83, and 0.69,
188 respectively. RH is a key meteorological component, sensitive to the atmospheric
189 thermodynamic (e.g., temperature) and dynamic (e.g., winds) conditions. Even when the
190 simulated water vapor content is the same as the observation, the overestimation or
191 underestimation of temperature still causes underestimation or overestimation of RH. Biases
192 of wind speeds and directions considerably influence the origination of air mass at the
193 observation site. In general, the northerly wind carries dry air, and is opposite for the
194 southerly wind during wintertime in NCP. Therefore, the uncertainties from meteorological
195 field simulations might constitute one of the most possible reasons for the RH bias (Bei et al.,
196 2017). Figure 4 presents the pattern comparison of the average simulated RH and the NCEP
197 reanalysis during the episode. The simulated RH distribution is generally consistent with that
198 from the reanalysis, e.g., dry air in West China and fairly humid air in South China. However,

199 the air over NCP in the simulation is more humid than the analysis, and the average simulated
200 RH is 70.6%, about 20% higher than the reanalyzed RH, which might be caused by the
201 temperature decrease due to the ARF considered in the simulation.

202 3.3 Sensitivity studies

203 The ALW not only enlarges the particle size to increase the aerosol optical depth (AOD),
204 likely enhancing the ARF to facilitate the PM_{2.5} accumulation or to alter photolysis rates to
205 affect the AOC, but also influences the SA formation serving as a medium for multiphase
206 reactions. Therefore, sensitivity studies are used to quantitatively evaluate the effect of the
207 ALW on the PM_{2.5} concentration during the haze pollution episode.

208 The FSA method was used to evaluate the contribution of the ALW to near-surface
209 [PM_{2.5}] by differentiating two model simulations with and without the ALW effect. Besides
210 the base case with all the ALW effect (hereafter referred to as f_{base}), additional four
211 sensitivity simulations have been performed, in which the ALW effect on the ARF,
212 photolysis, multiphase reactions, and the total is excluded, respectively (hereafter referred to
213 as $f_{\text{alw_rad0}}$, $f_{\text{alw_j0}}$, $f_{\text{alw_het0}}$, and $f_{\text{alw_tot0}}$, respectively). It is worth noting that in all the
214 sensitivity simulations, the aerosol hygroscopic growth is not turned off. In the sensitivity
215 simulation $f_{\text{alw_rad0}}$, only the ALW contribution to the AOD is not included in the ARF. In
216 the $f_{\text{alw_j0}}$, only the ALW contribution to the AOD is not considered in the photolysis
217 calculation. In the $f_{\text{alw_het0}}$, only the heterogeneous formation of secondary aerosols (SA)
218 involving the ALW is turned off, including the SO₂ heterogeneous oxidation by O₂ catalyzed
219 by Fe³⁺, N₂O₅ heterogeneous hydrolysis, and the heterogeneous reaction of glyoxal and
220 methylglyoxal. For the $f_{\text{alw_tot0}}$, the ALW contribution to the AOD is not considered in the
221 ARF and photolysis calculation, and the SA heterogeneous formation involving the ALW is
222 excluded. Detailed description about the sensitivity simulations can be found in Table 2. The
223 difference between f_{base} and $f_{\text{alw_rad0}}$ represents the ALW effect on the ARF during the

224 study episode, and so does for the ALW effect on photochemistry, multiphase reactions, and
225 the total ALW effect.

226

227 3.3.1 ALW effect on the ARF

228 The ALW, caused by the aerosol hygroscopic growth, augments the particle size to
229 increase AOD, potentially enhancing the ARF and aggravating the haze pollution. Figure 5
230 shows the distribution of the average AOD contribution due to the ALW during the haze
231 episode, evaluated by differentiating f_{base} and f_{alw_rad0} . Apparently, the ALW
232 substantially increases the AOD in NCP, with the contribution ranging from 30% to more
233 than 50%, indicating that ALW is an important contributor of the AOD. Substantial increase
234 of the AOD due to the ALW is anticipated to attenuate the incoming solar radiation,
235 decreasing the surface temperature and suppressing the PBL development, therefore,
236 deteriorating the haze pollution, as proposed by recent studies (Tie et al., 2017; Liu et al.,
237 2018).

238 Figure 6 presents the distribution of the average near-surface $PM_{2.5}$ contribution of the
239 ALW effect on the ARF (hereafter referred as to ALW-ARF) during the haze episode.
240 Interestingly, the ALW-ARF does not increase the near-surface $[PM_{2.5}]$ consistently in NCP,
241 as expected. The ALW-ARF enhances the near-surface $[PM_{2.5}]$ most strikingly in the south of
242 Hebei, with a contribution of more than $15 \mu g m^{-3}$ (less than 7%). However, in some areas,
243 such as the east of Shandong, the near-surface $PM_{2.5}$ contribution of the ALW-ARF is
244 negative, or the ALW-ARF decreases $[PM_{2.5}]$. On average, the ALW-ARF increases
245 near-surface $[PM_{2.5}]$ in NCP during the episode by about $1.1 \mu g m^{-3}$, so cannot constitute an
246 important factor for the heavy haze formation.

247 It is worth noting that enlarged particles due to the ALW not only increase the AOD to
248 enhance aerosol backward scattering, but also cause the aerosol spectrum to successively

249 shift toward larger sizes. Based on the Mie scattering theory, when the particle size is similar
250 to the wavelength of incoming solar radiation, the radiation is favored to be scattered in the
251 forward directions. In order to further verify the ALW effect on solar radiation and
252 near-surface $[PM_{2.5}]$, an ensemble method is employed similar to that reported in Wu et al.
253 (2018a). The daytime near-surface $[PM_{2.5}]$ in NCP during the episode in f_{base} are first
254 subdivided into 30 bins with the interval of $20 \mu g m^{-3}$. The AOD at 550 nm, aerosol effective
255 radius ($Reff$), downward shortwave radiation at the surface (SWDOWN), surface temperature
256 (TSFC), PBLH, and near-surface $[PM_{2.5}]$ in f_{base} and f_{alw_rad0} in the same grid cell are
257 assembled as the bin $[PM_{2.5}]$, respectively, and an average of these variables in each bin are
258 calculated. Figure 7 shows the variation of AOD and $Reff$ in f_{base} and f_{alw_rad0} as a
259 function of bin $[PM_{2.5}]$. The ALW not only significantly enhances the AOD, with an average
260 contribution of 46% in NCP during the episode, but also increases the $Reff$ considerably
261 (Figures 7a and 7b). The $Reff$ enhancement due to the ALW is the most striking with
262 near-surface $[PM_{2.5}]$ between 40 and $160 \mu g m^{-3}$, exceeding 60%. On average, the ALW
263 increases the $Reff$ from $0.31 \mu m$ to $0.48 \mu m$, close to the peak band of solar radiation.
264 Therefore, the ALW increases the AOD to scatter more incoming solar radiation, decreasing
265 the SWDOWN, but augments particle sizes to favor the forward scattering, increasing the
266 SWDOWN. Generally, the decrease of the SWDOWN caused by the ALW is not significant,
267 less than 1% with near-surface $[PM_{2.5}]$ less than $200 \mu g m^{-3}$ and ranging from 2% to 3% with
268 $[PM_{2.5}]$ exceeding $240 \mu g m^{-3}$ (Figure 8a). Correspondingly, the ALW-ARF effect on the
269 daytime TSFC is also marginal and the TSFC is decreased by less than $0.2^{\circ}C$ (Figure 8b).
270 Furthermore, the ALW-ARF generally increases the daytime PBLH with near-surface $[PM_{2.5}]$
271 less than $140 \mu g m^{-3}$, but is opposite with $[PM_{2.5}]$ exceeding $140 \mu g m^{-3}$ (Figure 8c). Hence,
272 the contribution of the ALW-ARF to near-surface $[PM_{2.5}]$ is highly uncertain (Figure 8d),

273 depending on the relative importance of the ALW induced enhancement of aerosol backward
274 and forward scattering.

275 3.3.2 ALW effect on the photochemistry

276 In addition to the ARF, the ALW also exerts an impact on the photochemistry by
277 altering the aerosol backward and forward scattering to affect the photolysis, further
278 influencing the ozone (O₃) and SA formation. Previous studies have shown that the ALW
279 modifies the vertical profile of photolysis rate of NO₂ (J_{NO_2}), inhibiting it at the ground level
280 and accelerating it in the upper PBL (Tao et al., 2014; Dickerson et al., 1997). The
281 combination reaction between the ground state oxygen atom (O³P), produced from NO₂
282 photolysis, and oxygen molecules (O₂) forms O₃, representing the only important source of
283 O₃ in the troposphere:



286 Thus, the variation of J_{NO_2} considerably affects the O₃ formation in the troposphere,
287 changing the AOC and further the SA formation.

288 Figures 9a and 9b present the distribution of the percentage variation of average daytime
289 J_{NO_2} due to the ALW during the haze episode at the 1st and 5th model layer, respectively. At
290 the 1st model layer, except in the north of Jiangsu, the ALW generally decreases J_{NO_2}
291 slightly in NCP. However, at 5th model layer, the region with enhanced J_{NO_2} caused by the
292 ALW is obviously increased compared to that at 1st model layer. Apparently, the ALW
293 induced enlargement of particle sizes increases the AOD and enhances the aerosol backward
294 scattering to reduce the solar radiation reaching the ground level, decreasing the photolysis
295 rate, but the enhanced forward scattering still potentially accelerates the photolysis, such as in
296 the north of Jiangsu. In addition, the enhanced aerosol backward scattering also increases the
297 photolysis rate in the upper and above PBL, which is consistent with previous studies (Tao et

298 al., 2014; Dickerson et al., 1997). The variation of average daytime O₃ concentrations is not
299 consistent with that of J_{NO_2} at the 1st model layer (Figure 9c). For example, although the
300 J_{NO_2} is decreased in Hebei and Shandong, the O₃ concentration is still enhanced by the ALW
301 in some areas of the two provinces. One of the possible reasons is the vertical transport of O₃
302 from the upper layers where the O₃ is plausibly enhanced due to the increased photolysis rate
303 (Figure 9b). At the 5th model layer, the area with enhanced O₃ concentrations is much larger
304 than that at 1st model layer, which is in agreement with the variation of the J_{NO_2} (Figure 9d).
305 On average, the ALW decreases near-surface (1st model layer) daytime O₃ concentrations by
306 about 0.2 μg m⁻³ (or 0.45%), playing a minor role in O₃ formation.

307 Figure 10 shows the distribution of the average near-surface PM_{2.5} contribution of the
308 ALW effect on the photolysis frequencies (hereafter referred as to ALW-J) during the haze
309 episode. Except in some areas in Shandong and Anhui, the ALW-J slightly decreases
310 near-surface [PM_{2.5}] in NCP, with an average reduction of about 0.87 μg m⁻³ (or 0.64%).
311 Therefore, the ALW-J does not play an important role in mitigating the haze pollution.

312 3.3.3 ALW effect on heterogeneous reactions

313 The ALW provides an excellent substrate for heterogeneous reactions in the atmosphere,
314 which has been proposed to play a key role in the SA formation during haze days (Li et al.,
315 2017; Xing et al., 2018).

316 Figure 11 presents the distribution of contributions of the ALW on heterogeneous
317 reactions (hereafter referred to ALW-HET) to near-surface sulfate, nitrate, and ammonium
318 concentrations averaged during the haze episode. A parameterization of sulfate
319 heterogeneous formation involving ALW has been developed and implemented into the
320 WRF-Chem model, which has successfully reproduced the observed rapid sulfate formation
321 during haze days (Li et al., 2017). The sulfate heterogeneous formation from SO₂ is
322 parameterized as a first order irreversible uptake by ALW surfaces, with a reactive uptake

323 coefficient of 0.5×10^{-4} assuming that there is enough alkalinity to maintain the high
324 iron-catalyzed reaction rate. The contribution of the ALW-HET to the sulfate formation is
325 substantial in NCP, exceeding $5 \mu\text{g m}^{-3}$ (or 50%) in NCP (Figures 11a and 11b). The
326 ALW-HET contributes about 71.3% of the sulfate in NCP during the episode, indicating that
327 the heterogeneous formation involving the ALW is the dominant sulfate source during haze
328 days.

329 The heterogeneous hydrolysis of N_2O_5 on the surface of deliquescent aerosols leads to
330 the HNO_3 formation, which is the key contributor to the nitrate aerosol loading (Riemer et al.,
331 2003; Pathak et al., 2011; Chang et al., 2016). In this study, the parameterization of the
332 heterogeneous hydrolysis of N_2O_5 , proposed by Riemer et al. (2003), has been included in the
333 WRF-Chem model for considering the effect of ALW-HET on the nitrate formation. The
334 N_2O_5 uptake coefficient on wet aerosol surfaces ranges from 0.002 to 0.02, depending on the
335 sulfate and nitrate aerosol mass. The ALW-HET generally increases the near-surface nitrate
336 mass concentration by 2 to $8 \mu\text{g m}^{-3}$ in NCP, with an average contribution of $2.8 \mu\text{g m}^{-3}$ (or
337 10%) (Figures 11c and 11d). It is worth noting that the ALW-HET does not consistently
338 increase the nitrate formation and the nitrate concentration is considerably decreased in some
339 areas in east China. One of the possible reasons is that the ALW-HET substantially enhances
340 the sulfate formation and the increased sulfate competes with nitrate for ammonia (NH_3),
341 suppressing the nitrate formation. Figure 12 shows the distribution of the NH_3 emission rate
342 in December. High NH_3 emissions are concentrated in NCP, Central China, Sichuan basin,
343 and Northeast China, where the nitrate concentration is generally increased by the
344 ALW-HET.

345 Ammonium serves as the main alkali in the atmosphere to neutralize acidic aerosols
346 such as sulfate and nitrate, so its formation is not only dependent on its precursor (NH_3), but
347 is also influenced by acid aerosols. The ALW-HET enhances the near-surface ammonium

348 mass concentration most strikingly in NCP with high NH₃ emissions and increased sulfate
349 and nitrate aerosols. The contribution of the ALW-HET to the ammonium concentration
350 varies from 2 to more than 10 μg m⁻³, with an average of 4.2 μg m⁻³ (or 25.6%), showing that
351 the heterogeneous formation constitutes an important ammonium source (Figures 11e and
352 11f).

353 Heterogeneous reactions are also an important pathway for the SOA formation (Fu et al.,
354 2008; Li et al., 2013). Laboratory and field studies have indicated that glyoxal and
355 methylglyoxal play an important role in SOA formation via aerosol uptake or cloud
356 processing (Liggio et al., 2005; Volkamer et al., 2007; Li et al., 2011). In this study, the
357 heterogeneous reaction of SOA formation from glyoxal and methylglyoxal is parameterized
358 as a first-order irreversible uptake by aerosol particles, with a reaction uptake coefficient of
359 3.7×10^{-3} (Liggio et al., 2005; Zhao et al., 2006; Volkamer et al., 2007; Li et al., 2011).
360 During the haze episode, the average near-surface SOA contribution of the ALW-HET is 7.4
361 μg m⁻³ (or 48%) in NCP, ranging from 3 to over 15 μg m⁻³ (or 30 to over 50%) in NCP,
362 showing that heterogeneous reactions of glyoxal and methylglyoxal on wet aerosol surfaces
363 play a critical role in SOA formation (Figure 13). Xing et al. (2018) have demonstrated that,
364 in Beijing-Tianjin-Hebei, the near-surface SOA contribution from glyoxal and methylglyoxal
365 is around 30% during a haze episode in January 2014, much less than that (48%) in the study.
366 The possible reason is that the RH in the episode is higher than that in Xing et al. (2018),
367 causing more SOA formation from glyoxal and methylglyoxal.

368 High SA contribution to the haze pollution has been observed in China (e.g., Huang et
369 al., 2014), so the ALW-HET substantially enhances the SA formation, constituting an
370 important factor for heavy haze formation. Figure 14 shows the contribution of the
371 ALW-HET to average near-surface [PM_{2.5}] during the study episode. The ALW-HET
372 enhances near-surface [PM_{2.5}] by more than 40 μg m⁻³ in the central part of NCP, and on

373 average, the $PM_{2.5}$ contribution of the ALW-HET is $21.7 \mu\text{g m}^{-3}$ (or 15.9%), less than the
374 total enhancement of $25.1 \mu\text{g m}^{-3}$ (or 18.4%) from sulfate, nitrate, ammonium, and SOA. The
375 inconsistency indicates that the ALW-HET induced SA enhancement causes a decrease in
376 primary aerosols. Figure 15a and 15b show the percentage decrease of black carbon (BC) and
377 primary organic aerosols (POA) as a function of bin $[PM_{2.5}]$, respectively. Interestingly, the
378 average BC and POA concentrations are decreased by 5.1% and 5.2% due to the ALW-HET,
379 respectively. The ALW-HET induced SA enhancement augments the particle size and should
380 increase the AOD. However, on the contrary, the ALW-HET decreases the AOD, although
381 considerably increases the $Reff$ (Figures 16a and 16b). The reason is that the enhanced SA
382 enlarges particle sizes, facilitating the coagulation to decrease the aerosol number
383 concentration and surface area, as shown in Figures 16c and 16d. Hence, the decreased AOD
384 and increased $Reff$ due to the ALW-HET enhance the SWDOWN and TSFC, further
385 increasing the PBLH and decreasing near-surface primary aerosols, as shown in Figure 17.

386 It is worth noting that the ALW content (ALWC) and aerosol composition are mutually
387 influenced, i.e., the ALWC depends on the existence of hygroscopic aerosols (mainly
388 inorganic components) and the RH in the atmosphere, and the SA, particularly inorganic
389 components, formed via heterogeneous and aqueous reactions involving the ALW further
390 increases the ALWC. Hence the initial hygroscopic aerosols might play a seeding role in the
391 contribution of the ALW-HET, likely constituting the most important factor in determining
392 the ALWC and additional SA formed via heterogeneous or aqueous reactions. However, in
393 model simulations, the initial hygroscopic aerosols or the seed particles, are insignificant to
394 the ALWC, since even without consideration of multiphase formation, the other processes
395 still dominate the hygroscopic aerosol concentration in the atmosphere, including direct
396 emissions, nucleation, condensation et al., particularly after model spin-up. Therefore, the
397 effect of initial hygroscopic aerosols or seed particles on the ALW and additional

398 heterogeneous SA is generally negligible.

399 3.3.4 ALW total effect

400 Above discussions have shown that the ALW influences near-surface $[PM_{2.5}]$ through
401 complicated physical and chemical processes, which interact with each other. Figure 18
402 shows the near-surface $PM_{2.5}$ contribution of the total ALW effect (hereafter referred as to
403 ALW-TOT) during the haze episode, evaluated by differentiating f_{base} and f_{alw_tot0} . The
404 ALW-TOT plays an important role in the $PM_{2.5}$ formation during the wintertime haze
405 pollution in NCP, with an average contribution of $23.8 \mu g m^{-3}$ (or 17.5%), ranging from 5 to
406 over $40 \mu g m^{-3}$. About 78% of the enhanced near-surface $[PM_{2.5}]$ due to the ALW-TOT is
407 contributed by secondary inorganic aerosols, in which the contributions from sulfate, nitrate,
408 and ammonium are 45%, 14% and 19%, respectively, and around 32% is contributed by SOA.
409 Therefore, about 10% of near-surface $[PM_{2.5}]$ enhancement from SA is counterbalanced by
410 the decrease in the primary aerosols. Therefore, the ALW induced enhancement of
411 near-surface $[PM_{2.5}]$ is overwhelmingly determined by the ALW-HET, and the ALW-ARF
412 and ALW-J are subject to decreasing $[PM_{2.5}]$.

413 Figure 19 shows the variation of near-surface $[PM_{2.5}]$ caused by the ALW-TOT,
414 ALW-HET, ALW-ARF, and ALW-J, respectively, as a function of bin near-surface RH in
415 NCP during the haze episode to further assess the ALW effect. The hourly near-surface RH
416 in f_{base} is first divided into 28 bins, ranging from 30% to 100%, with interval of 2.5%. The
417 $PM_{2.5}$ contribution from the four sensitivity simulations is assembled in the same grid cell as
418 the bin RH, and an average of $PM_{2.5}$ contribution in each bin is calculated. The ALW does
419 not continuously increase near-surface $[PM_{2.5}]$ with the RH. When the RH is less than 80%,
420 the near-surface $PM_{2.5}$ contribution of the ALW-TOT generally increases rapidly with the RH.
421 However, when the RH exceeds 80%, the contribution commences to decrease and fluctuates
422 between 20 and $30 \mu g m^{-3}$, showing the effect of high occurrence frequencies of precipitation.

423 In addition, the ALW-HET dominates the PM_{2.5} contribution, particularly with RH less than
424 50%. The ALW-RAD generally decreases [PM_{2.5}] slightly with the RH less than 52.5% and
425 vice versa with the RH more than 52.5%. The PM_{2.5} contribution of the ALW-J is negative
426 and less than 1.5 μg m⁻³, except when the RH is between 92.5% and 97.5%.

427

428 4 Conclusions and summaries

429 The good relationships between near-surface [PM_{2.5}] concentration and RH during the
430 wintertime of 2015 in NCP indicate the possibility that high RH plays an important role in
431 the PM_{2.5} formation during the haze pollution. A severe haze pollution episode from 05
432 December 2015 to 04 January 2016 is simulated using the WRF-Chem model to investigate
433 the impact of ALW caused by the accumulated moisture on near surface [PM_{2.5}] in NCP. The
434 air over NCP during the haze episode is humid, with an average simulated RH of about 71%.
435 In general, the WRF-Chem model reproduces reasonably well the temporal variations of RH
436 compared to observations at three sites in NCP, although the model biases still exist due to
437 the uncertainties in simulated meteorological fields.

438 The FSA method is used to evaluate the contribution of the ALW effect on ARF,
439 photochemistry and heterogeneous reactions to the wintertime PM_{2.5} concentration in NCP.
440 Model results show that the ALW substantially increases the daytime AOD with an average
441 contribution of 46% in NCP during the episode, enhancing the aerosol backward scattering,
442 and also augments the Reff from 0.31 μm to 0.48 μm, approaching the peak band of solar
443 radiation, favoring the aerosol forward scattering. The ALW does not significantly attenuate
444 the incoming solar radiation at the ground surface to enhance the ARF, and the average
445 near-surface PM_{2.5} contribution of the ALW-ARF is about 1.1 μg m⁻³ in NCP during the
446 episode. Therefore, the ALW-ARF is not an important factor for heavy haze formation and
447 its contribution relies on the relative importance of the ALW induced enhancement of aerosol

448 backward and forward scattering. The ALW generally decreases the photolysis rate at the
449 surface level due to enhanced backward aerosol scattering, but the favored forward scattering
450 still possibly accelerates the photolysis. Additionally, the ALW increases the photolysis rate
451 in the upper and above PBL. On average, the ALW decreases near-surface daytime O₃
452 concentrations by 0.20 μg m⁻³ and [PM_{2.5}] by 0.87 μg m⁻³, playing a minor role in mitigating
453 the haze pollution.

454 The ALW substantially enhances the SA formation by serving as an important medium
455 for liquid-phase and heterogeneous reactions. The ALW contribution to near-surface sulfate,
456 nitrate, ammonium, and SOA concentrations is 71%, 10%, 26%, and 48% in NCP during the
457 episode, respectively. However, the enhanced SA due to the ALW-HET enlarges particle
458 sizes to facilitate the coagulation, decreasing the aerosol number concentration and surface
459 area and further AOD. Therefore, the decreased AOD and increased Reff enhance the
460 incoming solar radiation reaching the ground surface, further increasing the surface
461 temperature and PBLH and decreasing near-surface primary aerosols. The ALW-HET
462 contributes 15.9% of near-surface [PM_{2.5}], less than the total SA enhancement of 18.4%, and
463 the rest is counterbalanced by the decrease in primary aerosols.

464 The near-surface PM_{2.5} contribution of the ALW total effect is 17.5% in NCP, indicating
465 that ALW plays an important role in the PM_{2.5} formation during the wintertime haze
466 pollution. Moreover, the ALW-HET overwhelmingly dominates the PM_{2.5} enhancement due
467 to the ALW. The ALW does not consistently enhance near-surface [PM_{2.5}] with increasing
468 RH. When the RH exceeds 80%, the contribution of the ALW commences to decrease caused
469 by the high occurrence frequencies of precipitation.

470 Although the model performs reasonably well in simulating air pollutants, aerosol
471 species and optical properties, and RH during the episode in NCP, the uncertainties from
472 meteorological fields and emission inventory still exist, leading to model biases. In addition,

473 simulation period of one month might not be sufficient to provide a comprehensive view of
474 the ALW effect on the PM_{2.5} formation. More wintertime case studies will need to be
475 performed in the future to further investigate the ALW effect.

476

477

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480 conducted a research, designed the experiments, carried the methodology out, performed the
481 simulation, processed the data, prepared the data visualization, and prepared the manuscript
482 with contributions from all authors. Naifang Bei provided the treatment of meteorological
483 data, analyzed the study data, validated the model performance, and reviewed the manuscript.
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485 reviewed the paper. Suixin Liu, Meng Zhou, Qiyuan Wang, Zirui Liu, and Yichen Wang
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499 Reference

- 500 Ackerman, T., and Baker, M. B.: Shortwave radiative effects of unactivated aerosol-particles
501 in clouds, *J. Appl. Meteorol.*, 16, 63-69,
502 10.1175/1520-0450(1977)016<0063:sreoua>2.0.co;2, 1977.
- 503 Ackerman, T. P.: Model of effect of aerosols on urban climates with particular applications to
504 Los-angeles basin, *J. Atmos. Sci.*, 34, 531-547,
505 10.1175/1520-0469(1977)034<0531:amoteo>2.0.co;2, 1977.
- 506 Bei, N. F., Li, G. H., Huang, R. J., Cao, J. J., Meng, N., Feng, T., Liu, S. X., Zhang, T.,
507 Zhang, Q., and Molina, L. T.: Typical synoptic situations and their impacts on the
508 wintertime air pollution in the Guanzhong basin, China, *Atmos. Chem. Phys.*, 16,
509 7373-7387, 10.5194/acp-16-7373-2016, 2016a.
- 510 Bei, N. F., Xiao, B., Meng, N., and Feng, T.: Critical role of meteorological conditions in a
511 persistent haze episode in the Guanzhong basin, China, *Sci. Total Environ.*, 550,
512 273-284, 10.1016/j.scitotenv.2015.12.159, 2016b.
- 513 Bei, N. F., Wu, J. R., Elser, M., Feng, T., Cao, J. J., El-Haddad, I., Li, X., Huang, R. J., Li, Z.
514 Q., Long, X., Xing, L., Zhao, S. Y., Tie, X. X., Prevot, A. S. H., and Li, G. H.: Impacts
515 of meteorological uncertainties on the haze formation in Beijing-Tianjin-Hebei (BTH)
516 during wintertime: a case study, *Atmos. Chem. Phys.*, 17, 14579-14591,
517 10.5194/acp-17-14579-2017, 2017.
- 518 Blando, J. D., and Turpin, B. J.: Secondary organic aerosol formation in cloud and fog
519 droplets: a literature evaluation of plausibility, *Atmos. Environ.*, 34, 1623-1632,
520 10.1016/s1352-2310(99)00392-1, 2000.
- 521 Carrico, C. M., Kus, P., Rood, M. J., Quinn, P. K., and Bates, T. S.: Mixtures of pollution,
522 dust, sea salt, and volcanic aerosol during ACE-Asia: Radiative properties as a function
523 of relative humidity, *J. Geophys. Res.-Atmos.*, 108, 18, 10.1029/2003jd003405, 2003.
- 524 Chan, C. K. and Yao, X.: Air pollution in mega cities in China, *Atmos. Environ.*, 42, 1-42,
525 2008.
- 526 Chang, W. L., Brown, S. S., Stutz, J., Middlebrook, A. M., Bahreini, R., Wagner, N. L., Dube,
527 W. P., Pollack, I. B., Ryerson, T. B., and Riener, N.: Evaluating N₂O₅ heterogeneous
528 hydrolysis parameterizations for CalNex 2010, *J. Geophys. Res.-Atmos.*, 121,
529 5051-5070, 10.1002/2015jd024737, 2016.
- 530 Chen, F. and Dudhia, J.: Coupling an advanced land surface-hydrology model with the Penn
531 State-NCAR MM5 modeling system. Part I: Model implementation and sensitivity, *Mon.
532 Weather Rev.*, 129(4), 569-585, 2001.
- 533 Cheng, Y., He, K. B., Du, Z. Y., Zheng, M., Duan, F. K., and Ma, Y. L.: Humidity plays an
534 important role in the PM_{2.5} pollution in Beijing, *Environ. Pollut.*, 197, 68-75,
535 10.1016/j.envpol.2014.11.028, 2015.
- 536 Cheng, Y. F., Wiedensohler, A., Eichler, H., Heintzenberg, J., Tesche, M., Ansmann, A.,
537 Wendisch, M., Su, H., Althausen, D., Herrmann, H., Gnauk, T., Brüggemann, E., Hu,
538 M., and Zhang, Y. H.: Relative humidity dependence of aerosol optical properties and

- 539 direct radiative forcing in the surface boundary layer at Xinken in Pearl River Delta of
540 China: An observation based numerical study, *Atmos. Environ.*, 42, 6373-6397,
541 10.1016/j.atmosenv.2008.04.009, 2008.
- 542 Chou, M.-D. and Suarez, M. J.: A solar radiation parameterization for atmospheric studies,
543 NASA Tech. Rep. NASA/TM-1999- 10460, 15, 38 pp., 1999.
- 544 Chou, M.-D. and Suarez, M. J.: A thermal infrared radiation parameterization for
545 atmospheric studies, NASA/TM-2001-104606, 19, 55 pp., 2001.
- 546 Cocker, D. R., Clegg, S. L., Flagan, R. C., and Seinfeld, J. H.: The effect of water on
547 gas-particle partitioning of secondary organic aerosol. Part I: alpha-pinene/ozone system,
548 *Atmos. Environ.*, 35, 6049-6072, 10.1016/s1352-2310(01)00404-6, 2001a.
- 549 Cocker, D. R., Flagan, R. C., and Seinfeld, J. H.: State-of-the-art chamber facility for
550 studying atmospheric aerosol chemistry, *Environ. Sci. Technol.*, 35, 2594-2601,
551 10.1021/es0019169, 2001b.
- 552 Cocker, D. R., Mader, B. T., Kalberer, M., Flagan, R. C., and Seinfeld, J. H.: The effect of
553 water on gas-particle partitioning of secondary organic aerosol: II. m-xylene and
554 1,3,5-trimethylbenzene photooxidation systems, *Atmos. Environ.*, 35, 6073-6085,
555 10.1016/s1352-2310(01)00405-8, 2001c.
- 556 Covert, D. S., Charlson, R. J., and Ahlquist, N. C.: A Study of the Relationship of Chemical
557 Composition and Humidity to Light Scattering by Aerosols, *J. Appl. Meteorol.*, 11,
558 968-976, 1972.
- 559 Dickerson, R. R., Kondragunta, S., Stenchikov, G., Civerolo, K. L., Doddridge, B. G., and
560 Holben, B. N.: The impact of aerosols on solar ultraviolet radiation and photochemical
561 smog, *Science*, 278, 827-830, 10.1126/science.278.5339.827, 1997.
- 562 Ding, Y. H., Wu, P., Liu, Y. J., and Song, Y. F.: Environmental and Dynamic Conditions for
563 the Occurrence of Persistent Haze Events in North China, *Engineering*, 3, 266-271,
564 10.1016/j.eng.2017.01.009, 2017.
- 565 Flocas, H., Kelessis, A., Helmis, C., Petrakakis, M., Zoumakis, M., and Pappas, K.: Synoptic
566 and local scale atmospheric circulation associated with air pollution episodes in an urban
567 Mediterranean area, *Theo. Appl. Climatol.*, 95, 265-277, 10.1007/s00704-008-0005-9,
568 2009.
- 569 Fu, G. Q., Xu, W. Y., Yang, R. F., Li, J. B., and Zhao, C. S.: The distribution and trends of
570 fog and haze in the North China Plain over the past 30 years, *Atmos. Chem. Phys.*, 14,
571 11949-11958, 10.5194/acp-14-11949-2014, 2014.
- 572 Fu, T. M., Jacob, D. J., Wittrock, F., Burrows, J. P., Vrekoussis, M., and Henze, D. K.:
573 Global budgets of atmospheric glyoxal and methylglyoxal, and implications for
574 formation of secondary organic aerosols, *J. Geophys. Res.-Atmos.*, 113, 17,
575 10.1029/2007jd009505, 2008.
- 576 Gabusi, V., Pisoni, E., and Volta, M.: Factor separation in air quality simulations, *Ecol.*
577 *Model.*, 218, 383-392, 10.1016/j.ecolmodel.2008.07.030, 2008.
- 578 Guenther, A., Karl, T., Harley, P., Wiedinmyer, C., Palmer, P. I., and Geron, C.: Estimates of
579 global terrestrial isoprene emissions using MEGAN (Model of Emissions of Gases and
580 Aerosols from Nature), *Atmos. Chem. Phys.*, 6, 3181-3210,
581 <https://doi.org/10.5194/acp-6-3181-2006>, 2006.

582 Guo, S., Hu, M., Zamora, M. L., Peng, J. F., Shang, D. J., Zheng, J., Du, Z. F., Wu, Z., Shao,
583 M., Zeng, L. M., Molina, M. J., and Zhang, R. Y.: Elucidating severe urban haze
584 formation in China, *P. Natl. Acad. Sci. USA.*, 111, 17373-17378,
585 10.1073/pnas.1419604111, 2014.

586 Hallquist, M., Wenger, J. C., Baltensperger, U., Rudich, Y., Simpson, D., Claeys, M.,
587 Dommen, J., Donahue, N. M., George, C., Goldstein, A. H., Hamilton, J. F., Herrmann,
588 H., Hoffmann, T., Iinuma, Y., Jang, M., Jenkin, M. E., Jimenez, J. L., Kiendler-Scharr,
589 A., Maenhaut, W., McFiggans, G., Mentel, T. F., Monod, A., Prevo, A. S. H., Seinfeld,
590 J. H., Surratt, J. D., Szmigielski, R., and Wildt, J.: The formation, properties and impact
591 of secondary organic aerosol: current and emerging issues, *Atmos. Chem. Phys.*, 9,
592 5155-5236, 10.5194/acp-9-5155-2009, 2009.

593 Hastings, W. P., Koehler, C. A., Bailey, E. L., and De Haan, D. O.: Secondary organic
594 aerosol formation by glyoxal hydration and oligomer formation: Humidity effects and
595 equilibrium shifts during analysis, *Environ. Sci. Technol.*, 39, 8728-8735,
596 10.1021/es050446l, 2005.

597 He, K. B., Yang, F. M., Ma, Y. L., Zhang, Q., Yao, X. H., Chan, C. K., Cadle, S., Chan, T.,
598 and Mulawa, P.: The characteristics of PM_{2.5} in Beijing, China, *Atmos. Environ.*, 35,
599 4959-4970, 10.1016/s1352-2310(01)00301-6, 2001.

600 Healy, R. M., Temime, B., Kuprovskite, K., and Wenger, J. C.: Effect of Relative Humidity
601 on Gas/Particle Partitioning and Aerosol Mass Yield in the Photooxidation of p-Xylene,
602 *Environ. Sci. Technol.*, 43, 1884-1889, 10.1021/es802404z, 2009.

603 Hong, S.-Y. and Lim, J.-O. J.: The WRF Single-Moment 6-Class Microphysics Scheme
604 (WSM6), *Asia-Pacific J. Atmos. Sci.*, 42, 129-151, 2006.

605 Horowitz, L. W., Walters, S., Mauzerall, D. L., Emmons, L. K., Rasch, P. J., Granier, C., Tie,
606 X. X., Lamarque, J. F., Schultz, M. G., Tyndall, G. S., Orlando, J. J., and Brasseur, G. P.:
607 A global simulation of tropospheric ozone and related tracers: Description and
608 evaluation of MOZART, version 2, *J. Geophys. Res.-Atmos.*, 108, 4784,
609 <https://doi.org/10.1029/2002jd002853>, 2003.

610 Huang, R.-J., Zhang, Y., Bozzetti, C., Ho, K.-F., Cao, J.-J., Han, Y., Daellenbach, K. R.,
611 Slowik, J. G., Platt, S. M., Canonaco, F., Zotter, P., Wolf, R., Pieber, S. M., Bruns, E. A.,
612 Crippa, M., Ciarelli, G., Piazzalunga, A., Schwikowski, M., Abbaszade, G.,
613 Schnelle-Kreis, J., Zimmermann, R., An, Z., Szidat, S., Baltensperger, U., El Haddad, I.,
614 and Prévôt, A. S. H.: High secondary aerosol contribution to particulate pollution during
615 haze events in China, *Nature*, 514, 218-222, doi: 10.1038/nature13774, 2014.

616 Im, J. S., Saxena, V. K., and Wenny, B. N.: An assessment of hygroscopic growth factors for
617 aerosols in the surface boundary layer for computing direct radiative forcing, *J. Geophys.*
618 *Res.-Atmos.*, 106, 20213-20224, 10.1029/2000jd000152, 2001.

619 Jacobson, M. Z.: Studying the effects of aerosols on vertical photolysis rate coefficient and
620 temperature profiles over an urban airshed, *J. Geophys. Res.-Atmos.*, 103, 10593-10604,
621 10.1029/98jd00287, 1998.

622 Jacobson, M. Z.: Analysis of aerosol interactions with numerical techniques for solving
623 coagulation, nucleation, condensation, dissolution, and reversible chemistry among
624 multiple size distributions, *J. Geophys. Res.-Atmos.*, 107, 23, 10.1029/2001jd002044,
625 2002.

626 Janjic', Z. I.: Nonsingular Implementation of the Mellor -Yamada Level 2.5 Scheme in the

- 627 NCEP Meso Model, Ncep Office Note, 436, 2002.
- 628 Jia, L., and Xu, Y. F.: Effects of Relative Humidity on Ozone and Secondary Organic
629 Aerosol Formation from the Photooxidation of Benzene and Ethylbenzene, *Aerosol Sci.*
630 *Technol.*, 48, 1-12, 10.1080/02786826.2013.847269, 2014.
- 631 Kamens, R. M., Zhang, H. F., Chen, E. H., Zhou, Y., Parikh, H. M., Wilson, R. L., Galloway,
632 K. E., and Rosen, E. P.: Secondary organic aerosol formation from toluene in an
633 atmospheric hydrocarbon mixture: Water and particle seed effects, *Atmos. Environ.*, 45,
634 2324-2334, 10.1016/j.atmosenv.2010.11.007, 2011.
- 635 Kan, H. D., Chen, R. J., and Tong, S. L.: Ambient air pollution, climate change, and
636 population health in China, *Environ. Int.*, 42, 10-19, 10.1016/j.envint.2011.03.003,
637 2012.
- 638 Koehler, C. A., Fillo, J. D., Ries, K. A., Sanchez, J. T., and De Haan, D. O.: Formation of
639 secondary organic aerosol by reactive condensation of furandiones, aldehydes, and
640 water vapor onto inorganic aerosol seed particles, *Environ. Sci. Technol.*, 38, 5064-5072,
641 10.1021/es034672b, 2004.
- 642 Kurokawa, J., Ohara, T., Morikawa, T., Hanayama, S., Janssens- Maenhout, G., Fukui, T.,
643 Kawashima, K., and Akimoto, H.: Emissions of air pollutants and greenhouse gases over
644 Asian regions during 2000–2008: Regional Emission inventory in Asia (REAS) version
645 2, *Atmos. Chem. Phys.*, 13, 11019–11058, <https://doi.org/10.5194/acp-13-11019-2013>,
646 2013.
- 647 Li, G., Lei, W., Zavala, M., Volkamer, R., Dusanter, S., Stevens, P., and Molina, L. T.:
648 Impacts of HONO sources on the photochemistry in Mexico City during the
649 MCMA-2006/MILAGO Campaign, *Atmos. Chem. Phys.*, 10, 6551-6567,
650 10.5194/acp-10-6551-2010, 2010.
- 651 Li, G., Bei, N., Tie, X., and Molina, L. T.: Aerosol effects on the photochemistry in Mexico
652 City during MCMA-2006/MILAGRO campaign, *Atmos. Chem. Phys.*, 11, 5169-5182,
653 10.5194/acp-11-5169-2011, 2011a.
- 654 Li, G., Zavala, M., Lei, W., Tsimpidi, A. P., Karydis, V. A., Pandis, S. N., agaratna, M. R.,
655 and Molina, L. T.: Simulations of organic aerosol concentrations in Mexico City using
656 the WRF-CHEM model during the MCMA-2006/MILAGRO campaign, *Atmos. Chem.*
657 *Phys.*, 11, 3789-3809, 10.5194/acp-11-3789-2011, 2011b.
- 658 Li, G., Lei, W., Bei, N., and Molina, L. T.: Contribution of garbage burning to chloride and
659 PM_{2.5} in Mexico City, *Atmos. Chem. Phys.*, 12, 8751-8761, 10.5194/acp-12-8751-2012,
660 2012.
- 661 Li, G. H., Bei, N. F., Zavala, M., and Molina, L. T.: Ozone formation along the California
662 Mexican border region during Cal-Mex 2010 field campaign, *Atmos. Environ.*, 88,
663 370-389, 10.1016/j.atmosenv.2013.11.067, 2014.
- 664 Li, G., Bei, N., Cao, J., Huang, R., Wu, J., Feng, T., Wang, Y., Liu, S., Zhang, Q., Tie, X.,
665 and Molina, L. T.: A possible pathway for rapid growth of sulfate during haze days in
666 China, *Atmos. Chem. Phys.*, 17, 3301–3316, <https://doi.org/10.5194/acp-17-3301-2017>,
667 2017a.
- 668 Li, N., Fu, T. M., Cao, J. J., Lee, S. C., Huang, X. F., He, L. Y., Ho, K. F., Fu, J. S., and Lam,
669 Y. F.: Sources of secondary organic aerosols in the Pearl River Delta region in fall:

- 670 Contributions from the aqueous reactive uptake of dicarbonyls, *Atmos. Environ.*, 76,
671 200-207, 10.1016/j.atmosenv.2012.12.005, 2013.
- 672 Liggio, J., Li, S. M., and McLaren, R.: Reactive uptake of glyoxal by particulate matter, *J.*
673 *Geophys. Res.-Atmos.*, 110, doi: 10.1029/2004jd005113, 2005.
- 674 Liu, Q., Jia, X. C., Quan, J. N., Li, J. Y., Li, X., Wu, Y. X., Chen, D., Wang, Z. F., and Liu, Y.
675 G.: New positive feedback mechanism between boundary layer meteorology and
676 secondary aerosol formation during severe haze events, *Sci. Rep.*, 8, 8,
677 10.1038/s41598-018-24366-3, 2018.
- 678 Nguyen, T. B., Roach, P. J., Laskin, J., Laskin, A., and Nizkorodov, S. A.: Effect of humidity
679 on the composition of isoprene photooxidation secondary organic aerosol, *Atmos. Chem.*
680 *Phys.*, 11, 6931-6944, 10.5194/acp-11-6931-2011, 2011.
- 681 Parrish, D. D. and Zhu, T.: Clean Air for Megacities, *Science*, 326, 674–675,
682 <https://doi.org/10.1126/science.1176064>, 2009.
- 683 Pathak, R. K., Wang, T., and Wu, W. S.: Nighttime enhancement of PM_{2.5} nitrate in
684 ammonia-poor atmospheric conditions in Beijing and Shanghai: Plausible contributions
685 of heterogeneous hydrolysis of N₂O₅ and HNO₃ partitioning, *Atmos. Environ.*, 45,
686 1183-1191, 10.1016/j.atmosenv.2010.09.003, 2011.
- 687 Penner, J. E., Hegg, D., and Leitch, R.: Unraveling the role of aerosols in climate change,
688 *Environ. Sci. Technol.*, 35, 332A-340A, 10.1021/es0124414, 2001.
- 689 Poulain, L., Wu, Z., Petters, M. D., Wex, H., Hallbauer, E., Wehner, B., Massling, A.,
690 Kreiden- weis, S. M., and Stratmann, F.: Towards closing the gap between hygroscopic
691 growth and CCN activation for secondary organic aerosols - Part 3: Influence of the
692 chemical compo- sition on the hygroscopic properties and volatile fractions of aerosols,
693 *Atmos. Chem. Phys.*, 10, 3775–3785, doi:10.5194/acp-10-3775-2010, 2010.
- 694 Pilinis, C., Seinfeld, J. H., and Grosjean, D.: Water content of atmospheric aerosols, *Atmos.*
695 *Environ.*, 23, 1601–1606, 1989.
- 696 Quan, J. N., Gao, Y., Zhang, Q., Tie, X. X., Cao, J. J., Han, S. Q., Meng, J. W., Chen, P. F.,
697 and Zhao, D. L.: Evolution of planetary boundary layer under different weather
698 conditions, and its impact on aerosol concentrations, *Particuology*, 11, 34-40,
699 10.1016/j.partic.2012.04.005, 2013.
- 700 Randles, C. A., Russell, L. M., and Ramaswamy, V.: Hygroscopic and optical properties of
701 organic sea salt aerosol and consequences for climate forcing, *Geophys. Res. Lett.*, 31, 4,
702 10.1029/2004gl020628, 2004.
- 703 Riemer, N., Vogel, H., Vogel, B., Schell, B., Ackermann, I., Kessler, C., and Hass, H.:
704 Impact of the heterogeneous hydrolysis of N₂O₅ on chemistry and nitrate aerosol
705 formation in the lower troposphere under photo-smog conditions, *J. Geophys.*
706 *Res.-Atmos.*, 108, 21, 10.1029/2002jd002436, 2003.
- 707 Seinfeld, J. H., Erdakos, G. B., Asher, W. E., and Pankow, J. F.: Modeling the formation of
708 secondary organic aerosol (SOA). 2. The predicted effects of relative humidity on
709 aerosol formation in the alpha-pinene-, beta-pinene-, sabinene-, Delta(3)-Carene-, and
710 cyclohexene-ozone systems, *Environ. Sci. Technol.*, 35, 1806-1817, 10.1021/es001765+,
711 2001.

- 712 Seinfeld, J. H. and Pandis, S. N.: Atmospheric Chemistry and Physics: From Air Pollution to
713 Climate Change, John Wiley & Sons, USA, 1986.
- 714 Sun, Y., Wang, Z., Fu, P., Jiang, Q., Yang, T., Li, J., and Ge, X.: The impact of relative
715 humidity on aerosol composition and evolution processes during wintertime in Beijing,
716 China, *Atmos. Environ.*, *77*, 927-934, 10.1016/j.atmosenv.2013.06.019, 2013.
- 717 Stein, U., and Alpert, P.: Factor separation in numerical simulations, *Journal of the*
718 *Atmospheric Science*, *50*, 2107-2115, 10.1175/1520-0469(1993)
719 050<2107:fsins>2.0.co;2, 1993.
- 720 Tao, J. C., Zhao, C. S., Ma, N., and Liu, P. F.: The impact of aerosol hygroscopic growth on
721 the single-scattering albedo and its application on the NO₂ photolysis rate coefficient,
722 *Atmos. Chem. Phys.*, *14*, 12055-12067, 10.5194/acp-14-12055-2014, 2014.
- 723 Tie, X. X., Huang, R. J., Cao, J. J., Zhang, Q., Cheng, Y. F., Su, H., Chang, D., Poschl, U.,
724 Hoffmann, T., Dusek, U., Li, G. H., Worsnop, D. R., and O'Dowd, C. D.: Severe
725 Pollution in China Amplified by Atmospheric Moisture, *Sci. Rep.*, *7*, 8,
726 10.1038/s41598-017-15909-1, 2017.
- 727 Volkamer, R., Martini, F. S., Molina, L. T., Salcedo, D., Jimenez, J. L., and Molina, M. J.: A
728 missing sink for gas-phase glyoxal in Mexico City: Formation of secondary organic
729 aerosol, *Geophys. Res. Lett.*, *34*, 5, 10.1029/2007gl030752, 2007.
- 730 Wang, G., Zhang, R., Gomez, M. E., Yang, L., Zamora, M. L., Hu, M., Lin, Y., Peng, J., Guo,
731 S., and Meng, J.: Persistent sulfate formation from London Fog to Chinese haze, *P. Natl.*
732 *Acad. Sci. USA.*, *113*, 13630–13635, 2016.
- 733 Wang, Y. S., Yao, L., Wang, L. L., Liu, Z. R., Ji, D. S., Tang, G. Q., Zhang, J. K., Sun, Y.,
734 Hu, B., and Xin, J. Y.: Mechanism for the formation of the January 2013 heavy haze
735 pollution episode over central and eastern China, *Sci. China-Earth Sci.*, *57*, 14-25,
736 10.1007/s11430-013-4773-4, 2014.
- 737 Weinhold, B.: Ozone nation – EPA standard panned by the people, *Environ. Health. Persp.*,
738 *116*, A302–A305, 2008.
- 739 Wu, J. R., Li, G. H., Cao, J. J., Bei, N. F., Wang, Y. C., Feng, T., Huang, R. J., Liu, S. X.,
740 Zhang, Q., and Tie, X. X.: Contributions of trans-boundary transport to summertime air
741 quality in Beijing, China, *Atmos. Chem. Phys.*, *17*, 2035-2051,
742 10.5194/acp-17-2035-2017, 2017.
- 743 Wu, P., Ding, Y. H., and Liu, Y. J.: Atmospheric circulation and dynamic mechanism for
744 persistent haze events in the Beijing-Tianjin-Hebei region, *Adv. Atmos. Sci.*, *34*,
745 429-440, 10.1007/s00376-016-6158-z, 2017.
- 746 Wu, Z. J., Wang, Y., Tan, T. Y., Zhu, Y. S., Li, M. R., Shang, D. J., Wang, H. C., Lu, K. D.,
747 Guo, S., Zeng, L. M., and Zhang, Y. H.: Aerosol Liquid Water Driven by
748 Anthropogenic Inorganic Salts: Implying Its Key Role in Haze Formation over the
749 North China Plain, *Environ. Sci. Technol. Lett.*, *5*, 160-166, 10.1021/acs.estlett.8b00021,
750 2018.
- 751 Xing, L., Wu, J., Elser, M., Tong, S., Liu, S., Li, X., Liu, L., Cao, J., Zhou, J., El-Haddad, I.,
752 Huang, R., Ge, M., Tie, X., Prévôt, A. S. H., and Li, G.: Wintertime secondary organic
753 aerosol formation in Beijing-Tianjin-Hebei (BTH): Contributions of HONO sources and
754 heterogeneous reactions, *Atmos. Chem. Phys. Discuss.*, 2018, 1-25,

- 755 10.5194/acp-2018-770, 2018.
- 756 Xu, J. W., Martin, R. V., van Donkelaar, A., Kim, J., Choi, M., Zhang, Q., Geng, G., Liu, Y.,
757 Ma, Z., Huang, L., Wang, Y., Chen, H., Che, H., Lin, P., and Lin, N.: Estimating
758 ground-level PM_{2.5} in eastern China using aerosol optical depth determined from the
759 GOCI satellite instrument, *Atmos. Chem. Phys.*, 15, 13133-13144,
760 10.5194/acp-15-13133-2015, 2015.
- 761 Zhang, L., Sun, J. Y., Shen, X. J., Zhang, Y. M., Che, H., Ma, Q. L., Zhang, Y. W., Zhang, X.
762 Y., and Ogren, J. A.: Observations of relative humidity effects on aerosol light scattering
763 in the Yangtze River Delta of China, *Atmos. Chem. Phys.*, 15, 8439-8454,
764 10.5194/acp-15-8439-2015, 2015.
- 765 Zhang, Q., Streets, D. G., Carmichael, G. R., He, K. B., Huo, H., Kannari, A., Klimont, Z.,
766 Park, I. S., Reddy, S., Fu, J. S., Chen, D., Duan, L., Lei, Y., Wang, L. T., and Yao, Z. L.:
767 Asian emissions in 2006 for the NASA INTEX-B mission, *Atmos. Chem. Phys.*, 9,
768 5131-5153, <https://doi.org/10.5194/acp-9-5131-2009>, 2009.
- 769 Zhang, R. H., Li, Q., and Zhang, R. N.: Meteorological conditions for the persistent severe
770 fog and haze event over eastern China in January 2013, *Sci. China-Earth Sci.*, 57, 26-35,
771 10.1007/s11430-013-4774-3, 2014.
- 772 Zhang, R., Jing, J., Tao, J., Hsu, S.-C., Wang, G., Cao, J., Lee, C. S. L., Zhu, L., Chen, Z.,
773 Zhao, Y., and Shen, Z.: Chemical characterization and source apportionment of PM_{2.5} in
774 Beijing: seasonal perspective, *Atmos. Chem. Phys.*, 13, 7053–7074,
775 <https://doi.org/10.5194/acp-13-7053-2013>, 2013.
- 776 Zhao, J., Levitt, N. P., Zhang, R. Y., and Chen, J. M.: Heterogeneous reactions of
777 methylglyoxal in acidic media: implications for secondary organic aerosol formation,
778 *Environ. Sci. Technol.*, 40, 7682–7687, 2006.
- 779 Zhou, Y., Zhang, H. F., Parikh, H. M., Chen, E. H., Rattanavaraha, W., Rosen, E. P., Wang,
780 W. X., and Kamens, R. M.: Secondary organic aerosol formation from xylenes and
781 mixtures of toluene and xylenes in an atmospheric urban hydrocarbon mixture: Water
782 and particle seed effects (II), *Atmos. Environ.*, 45, 3882-3890,
783 10.1016/j.atmosenv.2010.12.048, 2011.
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785 Table 1 WRF-Chem model configurations
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| Regions | East Asia |
| Simulation period | December 05, 2015 - January 04, 2016 |
| Domain size | 400 × 400 |
| Domain center | 35°N, 114°E |
| Horizontal resolution | 12km × 12km |
| Vertical resolution | 35 vertical levels with a stretched vertical grid with spacing ranging from 30 m near the surface, to 500 m at 2.5 km and 1 km above 14 km |
| Microphysics scheme | WSM 6-class graupel scheme (Hong and Lim, 2006) |
| Boundary layer scheme | MYJ TKE scheme (Janjić, 2002) |
| Surface layer scheme | MYJ surface scheme (Janjić, 2002) |
| Land-surface scheme | Unified Noah land-surface model (Chen and Dudhia, 2001) |
| Longwave radiation scheme | Goddard longwave scheme (Chou and Suarez, 2001) |
| Shortwave radiation scheme | Goddard shortwave scheme (Chou and Suarez, 1999) |
| Meteorological boundary and initial conditions | NCEP 1°×1° reanalysis data |
| Chemical initial and boundary conditions | MOZART 6-hour output (Horowitz et al., 2003) |
| Anthropogenic emission inventory | Developed by Zhang et al. (2009) and Li et al. (2017), 2012 base year, and SAPRC-99 chemical mechanism |
| Biogenic emission inventory | MEGAN model developed by Guenther et al. (2006) |
| Model spin-up time | 28 hours |

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792 Table 2 Description of the sensitivity simulations.

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| Case | Aerosol hygroscopic growth | ALW contribution to AOD | | Multiphase reactions involving the ALW |
|-----------------|-------------------------------|-------------------------|-------------------|---|
| | | in the ARF | in the photolysis | |
| f_{base} | On | On | On | On |
| f_{alw_rad0} | On | Off | On | On |
| f_{alw_j0} | On | On | Off | On |
| f_{alw_het0} | On | On | On | Off |
| f_{alw_tot0} | On | Off | Off | Off |

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Figure Captions

800 Figure 1 (a) WRF-Chem simulation domain with topography and (b) North China Plain. In
801 (a), the blue circles represent centers of cities with ambient monitoring sites and the
802 size of circles denotes the number of ambient monitoring sites of cities. In (b), the red
803 capitals denote six typical polluted cities in NCP. A: Beijing; B: Tianjin; C:
804 Shijiazhuang; D: Baoding; E: Tangshan; F: Chengde. The blue numbers denote the
805 CERN sites with the hourly RH measurement. 1: Jiaozhouwan; 2: Yucheng; 3:
806 Luancheng.

807 Figure 2 Scatter plots of near-surface $[PM_{2.5}]$ and RH at six typical polluted cities in NCP
808 during the 2015 wintertime. The red diamond shows the bin average of near-surface
809 $[PM_{2.5}]$, and the red line denotes the variation of the bin average of near-surface
810 $[PM_{2.5}]$ with RH.

811 Figure 3 Comparison of measured (black dots) and predicted (red line) diurnal profiles of the
812 RH in (a) Luancheng, (b) Yucheng, and (c) Jiaozhouwan from 05 December 2015 to
813 04 January 2016.

814 Figure 4 Spatial distribution of (a) NCEP reanalyzed and (b) simulated RH averaged from 05
815 December 2015 to 04 January 2016.

816 Figure 5 Average (a) absolute and (b) relative AOD contribution caused by the ALW from 05
817 December 2015 to 04 January 2016.

818 Figure 6 Average near-surface (a) absolute and (b) relative $PM_{2.5}$ contribution caused by the
819 ALW-ARF from 05 December 2015 to 04 January 2016.

820 Figure 7 Average variations of (a) AOD and (b) $Reff$ in f_{base} (red line) and f_{alw_rad0} (blue
821 line) as a function of bin $[PM_{2.5}]$ in NCP during daytime from 05 December 2015 to
822 04 January 2016.

823 Figure 8 Average (a) percentage decrease of SWDOWN at the ground surface, (b) decrease
824 of TSFC, (c) percentage decrease of PBLH, and (d) percentage contribution of
825 near-surface $[PM_{2.5}]$ caused by the ALW-ARF, as a function of the near-surface
826 $[PM_{2.5}]$ in NCP during daytime from 05 December 2015 to 04 January 2016.

827 Figure 9 Average variations of daytime (a)/(b) NO_2 photolysis and (c)/(d) O_3 concentration at
828 1st/5th model layer (around 18 m and 420 m above the ground surface, respectively)
829 caused by the ALW from 05 December 2015 to 04 January 2016 in NCP.

830 Figure 10 Average near-surface (a) absolute and (b) relative $PM_{2.5}$ contribution caused by the
831 ALW-J from 05 December 2015 to 04 January 2016.

832 Figure 11 Average near-surface (a)/(c)/(e) absolute and (b)/(d)/(f) relative contribution to
833 sulfate/nitrate/ammonium concentrations, caused by the ALW-HET from 05
834 December 2015 to 04 January 2016.

835 Figure 12 Spatial distribution of NH_3 emission rate in December.

836 Figure 13 Average near-surface (a) absolute and (b) relative SOA contribution caused by the
837 ALW-HET from 05 December 2015 to 04 January 2016.

838 Figure 14 Average near-surface (a) absolute and (b) relative $PM_{2.5}$ contribution caused by the
839 ALW-HET from 05 December 2015 to 04 January 2016.

840 Figure 15 Average percentage decrease of (a) BC and (b) POA concentrations caused by the
841 ALW-HET, as a function of the near-surface $[PM_{2.5}]$ in NCP from 05 December 2015
842 to 04 January 2016.

843 Figure 16 Average variations of (a) AOD and (b) Reff in f_{base} (red line) and f_{alw_het0}
844 (blue line), respectively, and average percentage decrease of near-surface (c) aerosol
845 number concentration and (d) surface area caused by the ALW-HET, as a function of
846 bin $[PM_{2.5}]$ in NCP from 05 December 2015 to 04 January 2016.

847 Figure 17 Average (a) percentage decrease of SWDOWN at the ground surface, (b) decrease
848 of TSFC, and (c) percentage decrease of PBLH caused by the ALW-HET, as a
849 function of the near-surface $[PM_{2.5}]$ in NCP during daytime from 05 December 2015
850 to 04 January 2016.

851 Figure 18 Average near-surface (a) absolute and (b) relative $PM_{2.5}$ contribution caused by the
852 ALW-TOT from 05 December 2015 to 04 January 2016.

853 Figure 19 Average contributions to near-surface $[PM_{2.5}]$ caused by the ALW-TOT (black
854 line), ALW-HET (red line), ALW-RAD (green line), and ALW-J (blue line),
855 respectively, as a function of the RH in NCP from 05 December 2015 to 04 January
856 2016.

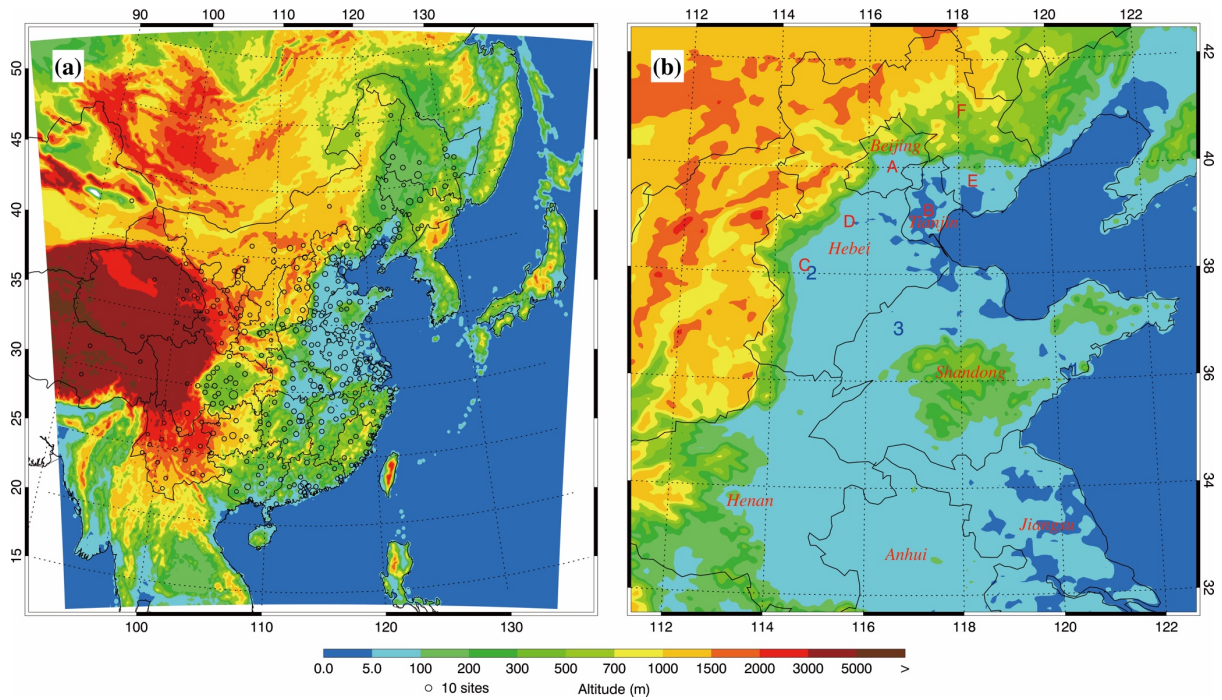
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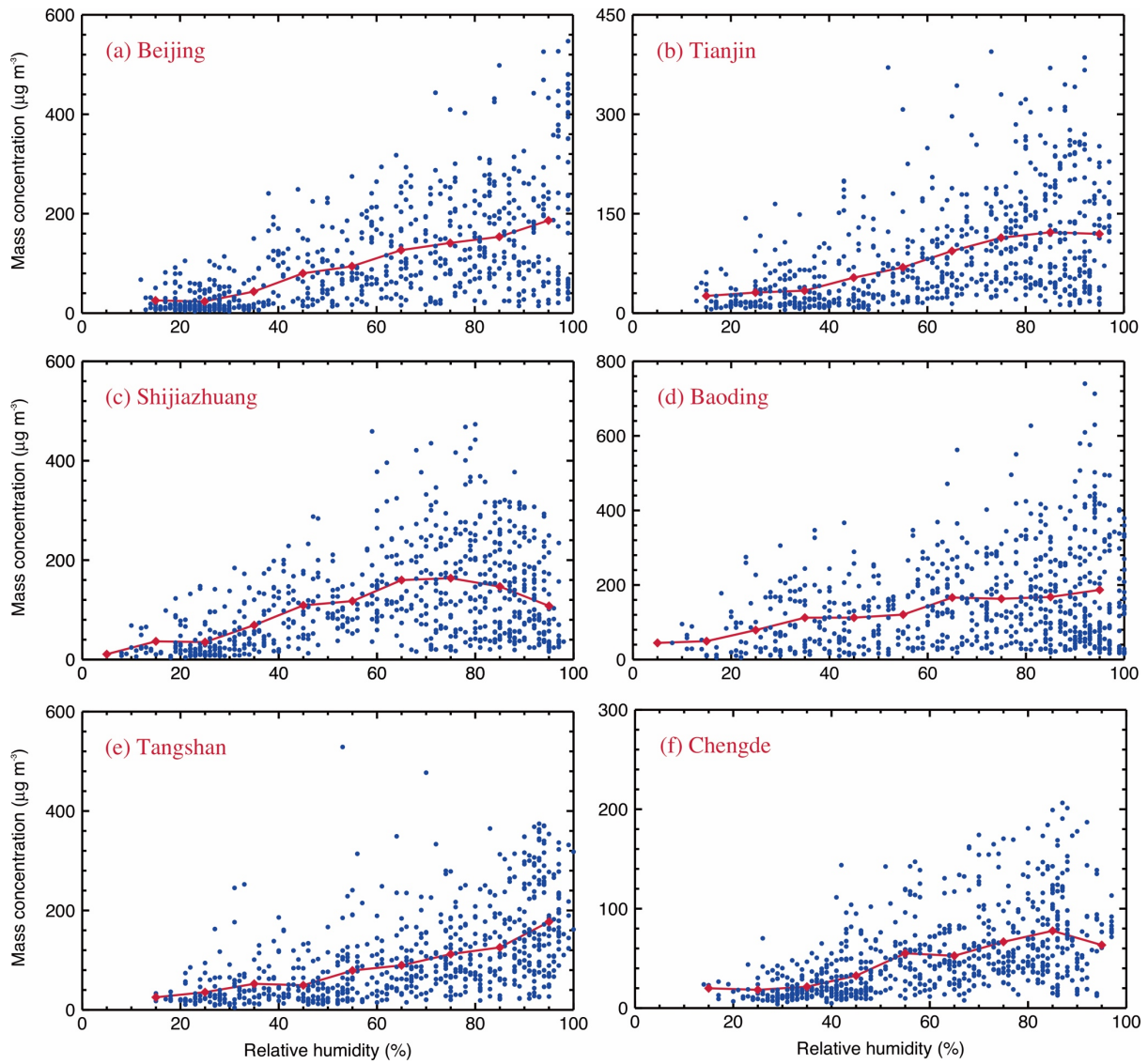
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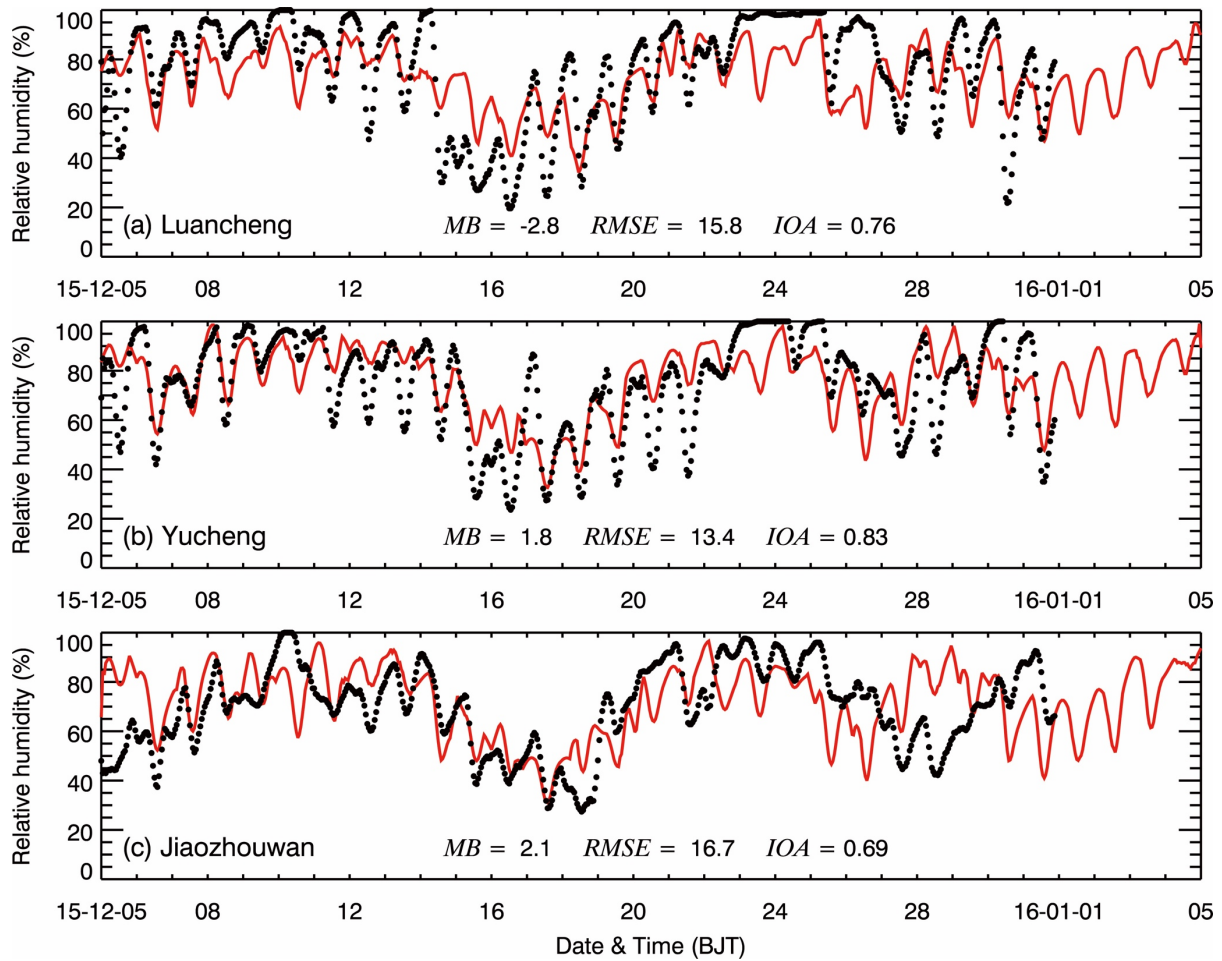
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Figure 1 (a) WRF-Chem simulation domain with topography and (b) North China Plain. In (a), the blue circles represent centers of cities with ambient monitoring sites and the size of blue circles denotes the number of ambient monitoring sites of cities. In (b), the red capitals denote six typical polluted cities in NCP. A: Beijing; B: Tianjin; C: Shijiazhuang; D: Baoding; E: Tangshan; F: Chengde. The blue numbers denote the CERN sites with the RH measurement. 1: Jiaozhouwan; 2: Yucheng; 3: Luancheng.



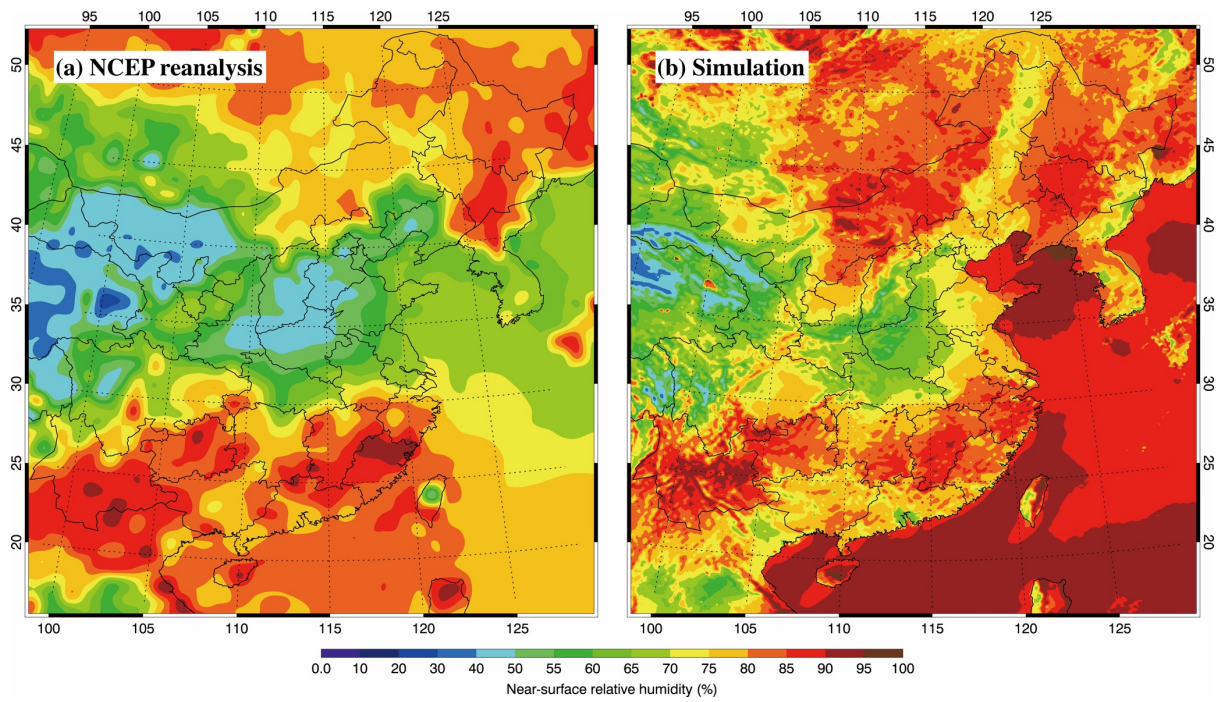
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Figure 2 Scatter plots of near-surface $[PM_{2.5}]$ and RH at six typical polluted cities in NCP during the 2015 wintertime. The red diamond shows the bin average of near-surface $[PM_{2.5}]$, and the red line denotes the variation of the bin average of near-surface $[PM_{2.5}]$ with RH.



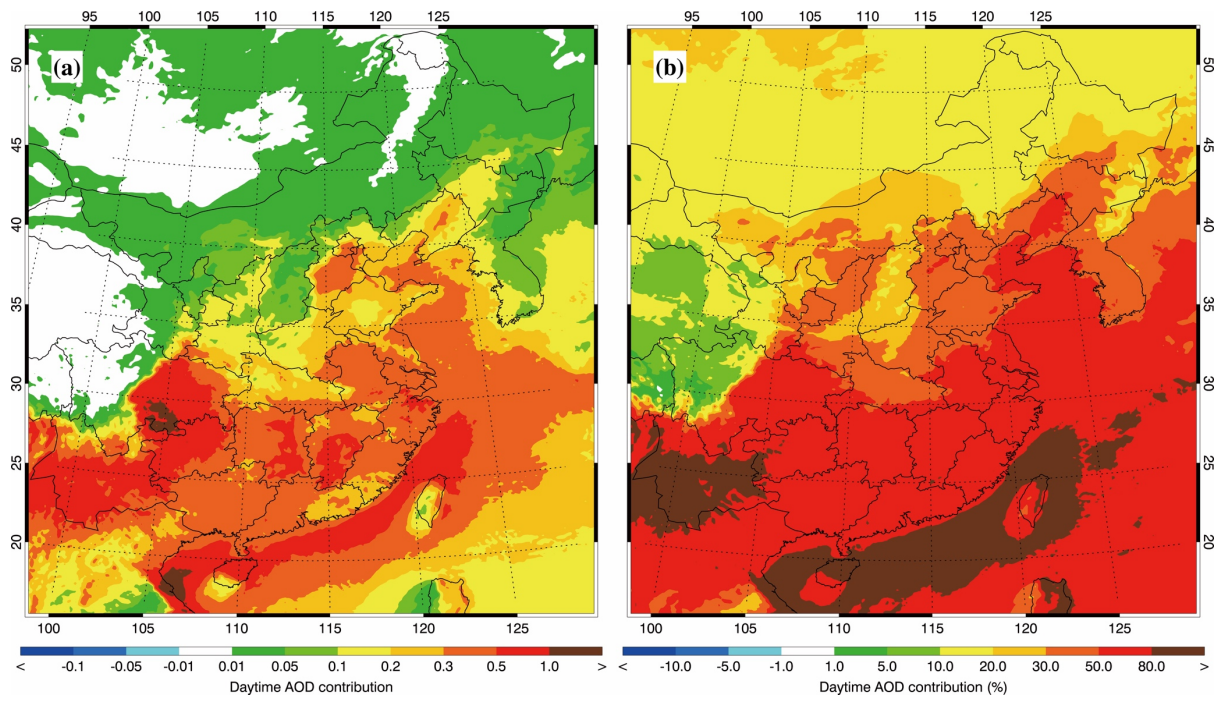
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Figure 3 Comparison of measured (black dots) and predicted (red line) diurnal profiles of the RH in (a) Luancheng, (b) Yucheng, and (c) Jiaozhouwan from 05 December 2015 to 04 January 2016.



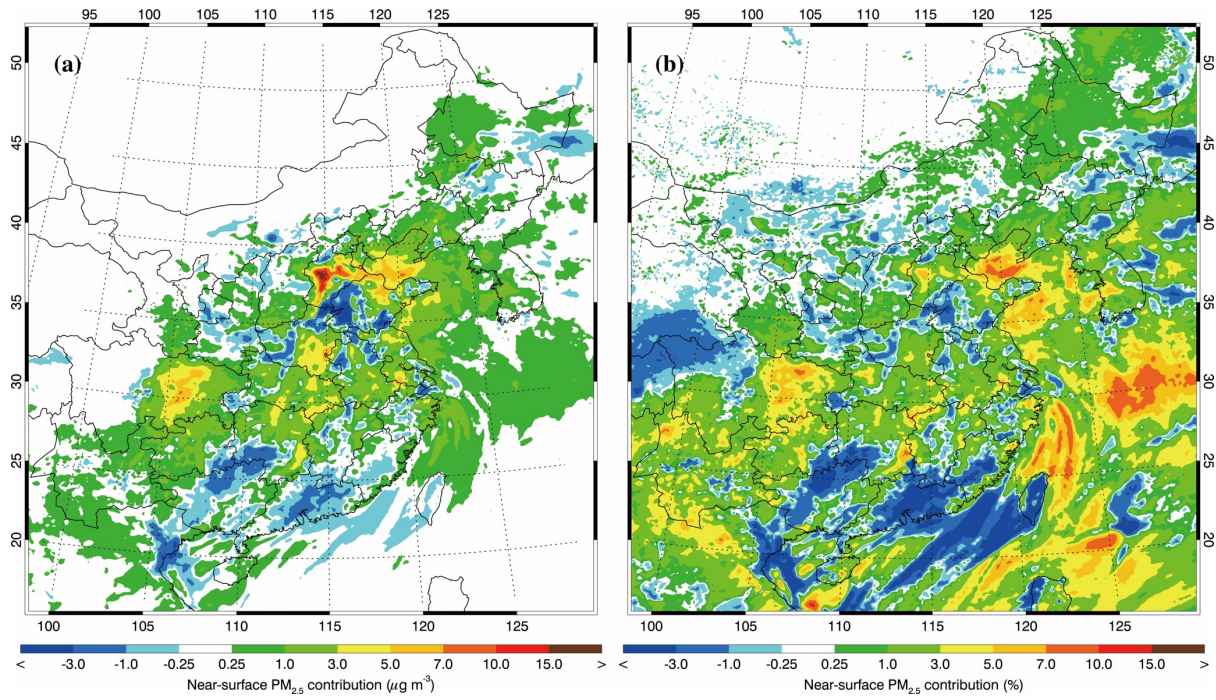
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Figure 4 Spatial distribution of (a) NCEP reanalyzed and (b) simulated RH averaged from 05 December 2015 to 04 January 2016.



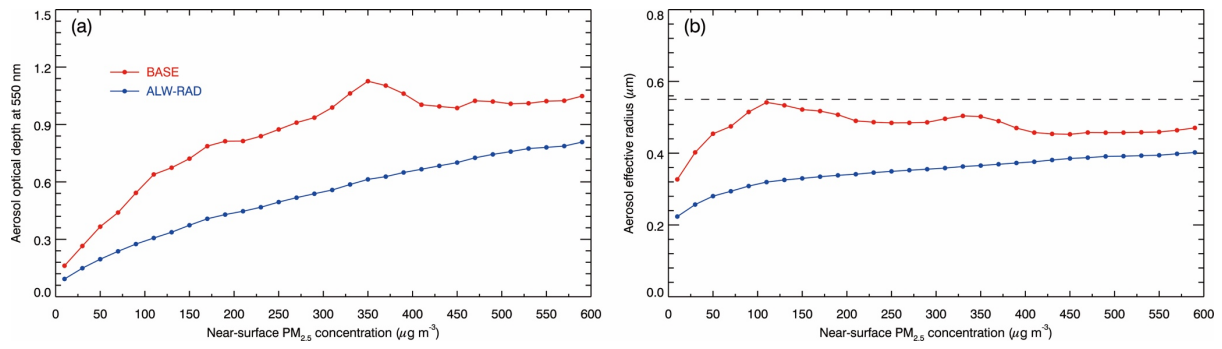
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Figure 5 Average (a) absolute and (b) relative AOD contribution caused by the ALW from 05 December 2015 to 04 January 2016.



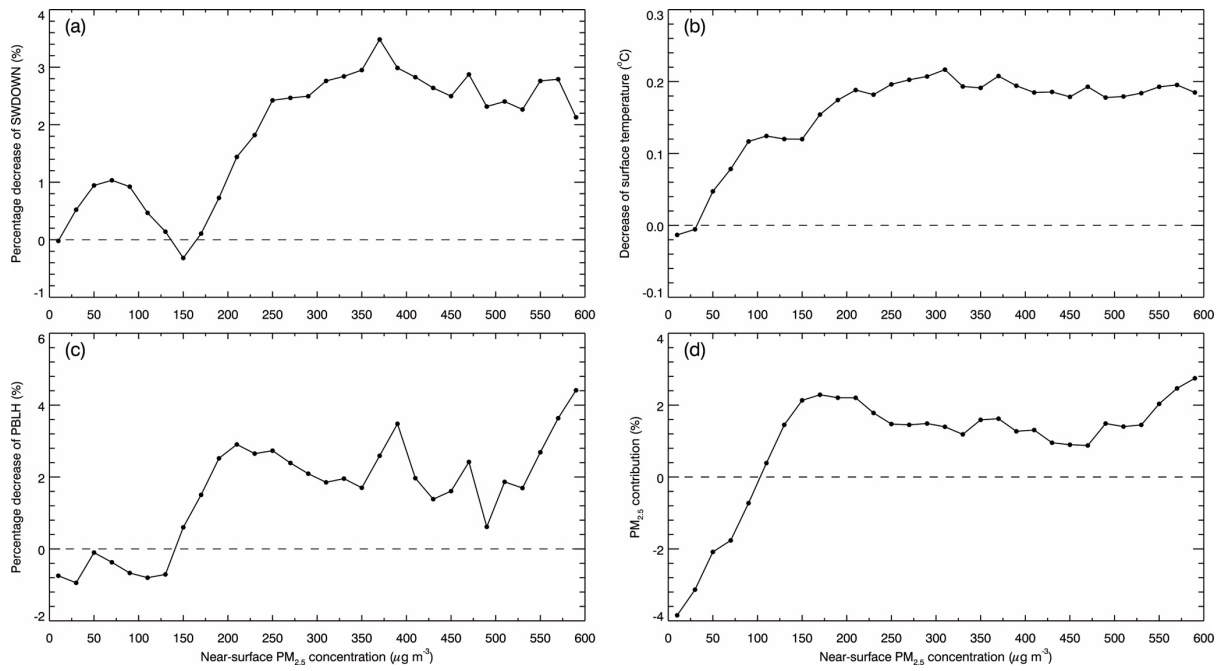
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Figure 6 Average near-surface (a) absolute and (b) relative PM_{2.5} contribution caused by the ALW-ARF from 05 December 2015 to 04 January 2016.



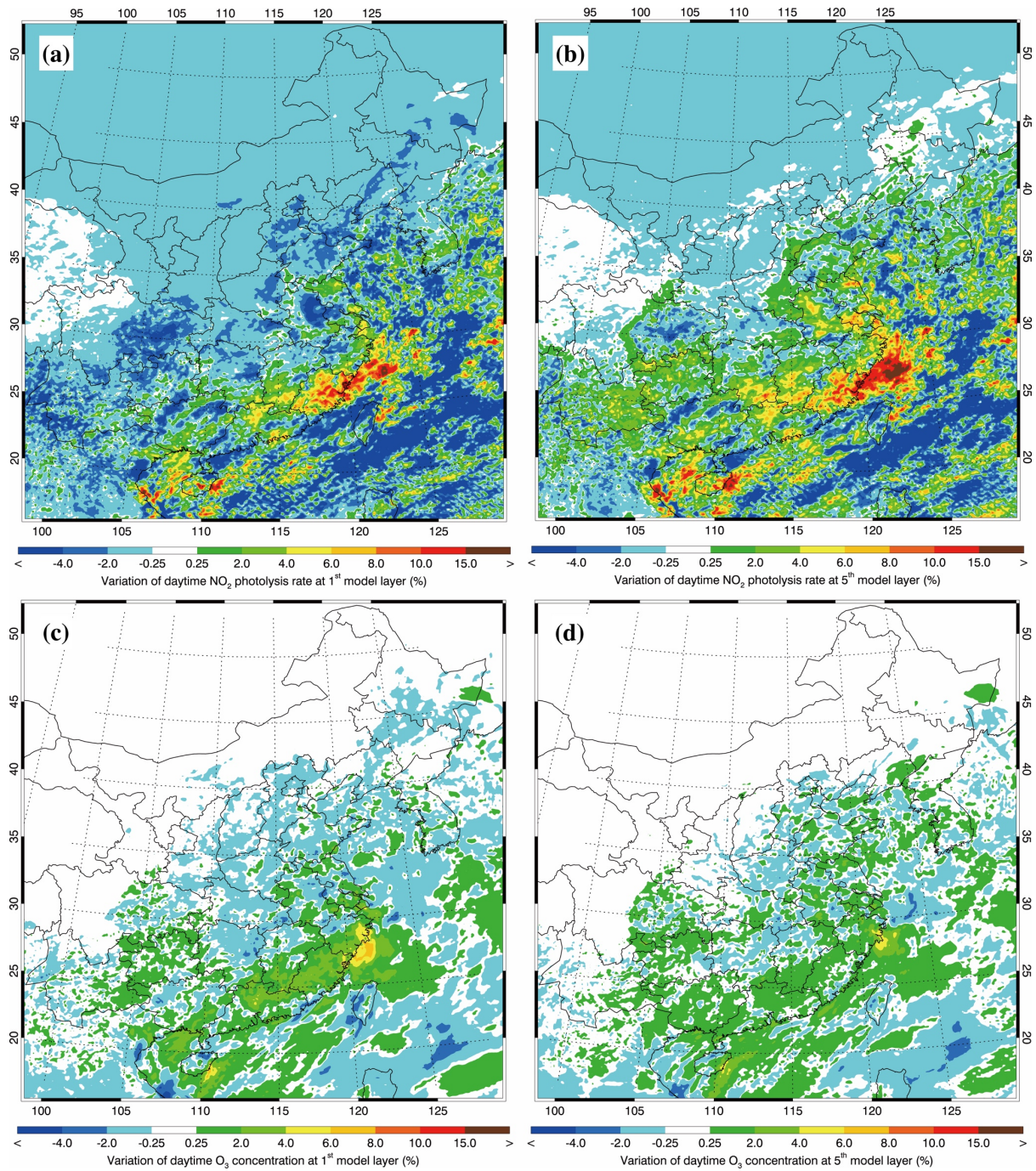
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Figure 7 Average variations of (a) AOD and (b) Reff in f_{base} (red line) and f_{alw_rad0} (blue line) as a function of bin $[PM_{2.5}]$ in NCP during daytime from 05 December 2015 to 04 January 2016.



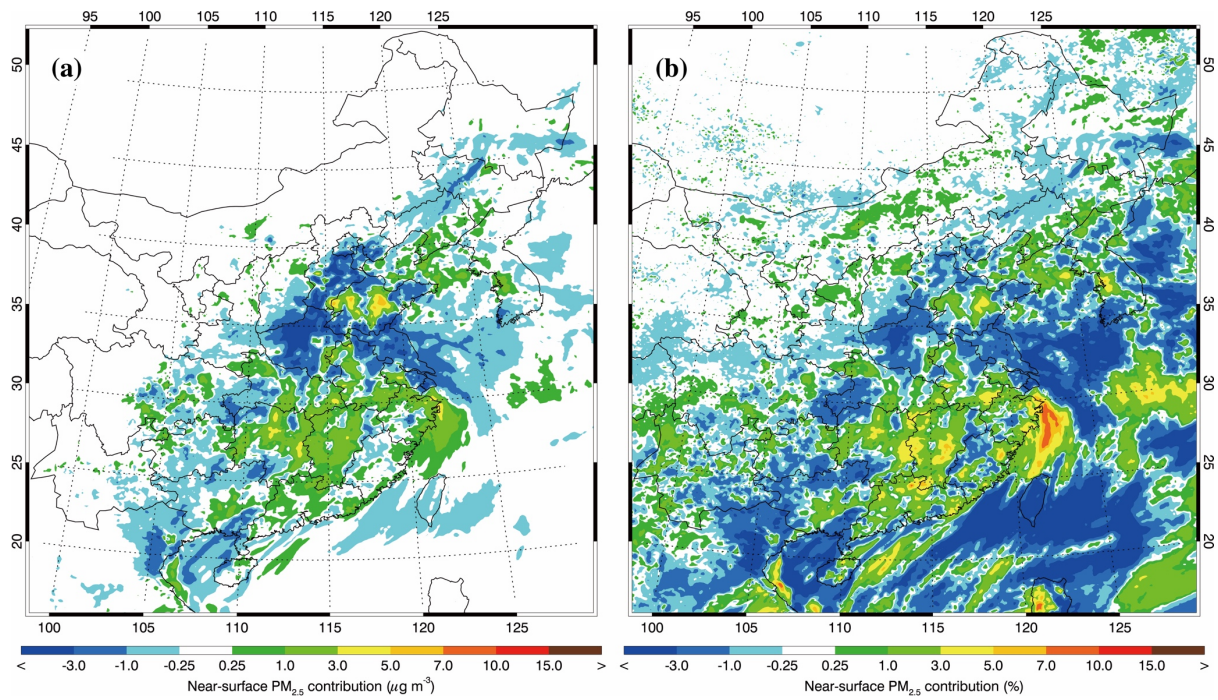
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Figure 8 Average (a) percentage decrease of SWDOWN at the ground surface, (b) decrease of TSFC, (c) percentage decrease of PBLH, and (d) percentage contribution of near-surface [PM_{2.5}] caused by the ALW-ARF, as a function of the near-surface [PM_{2.5}] in NCP during daytime from 05 December 2015 to 04 January 2016.



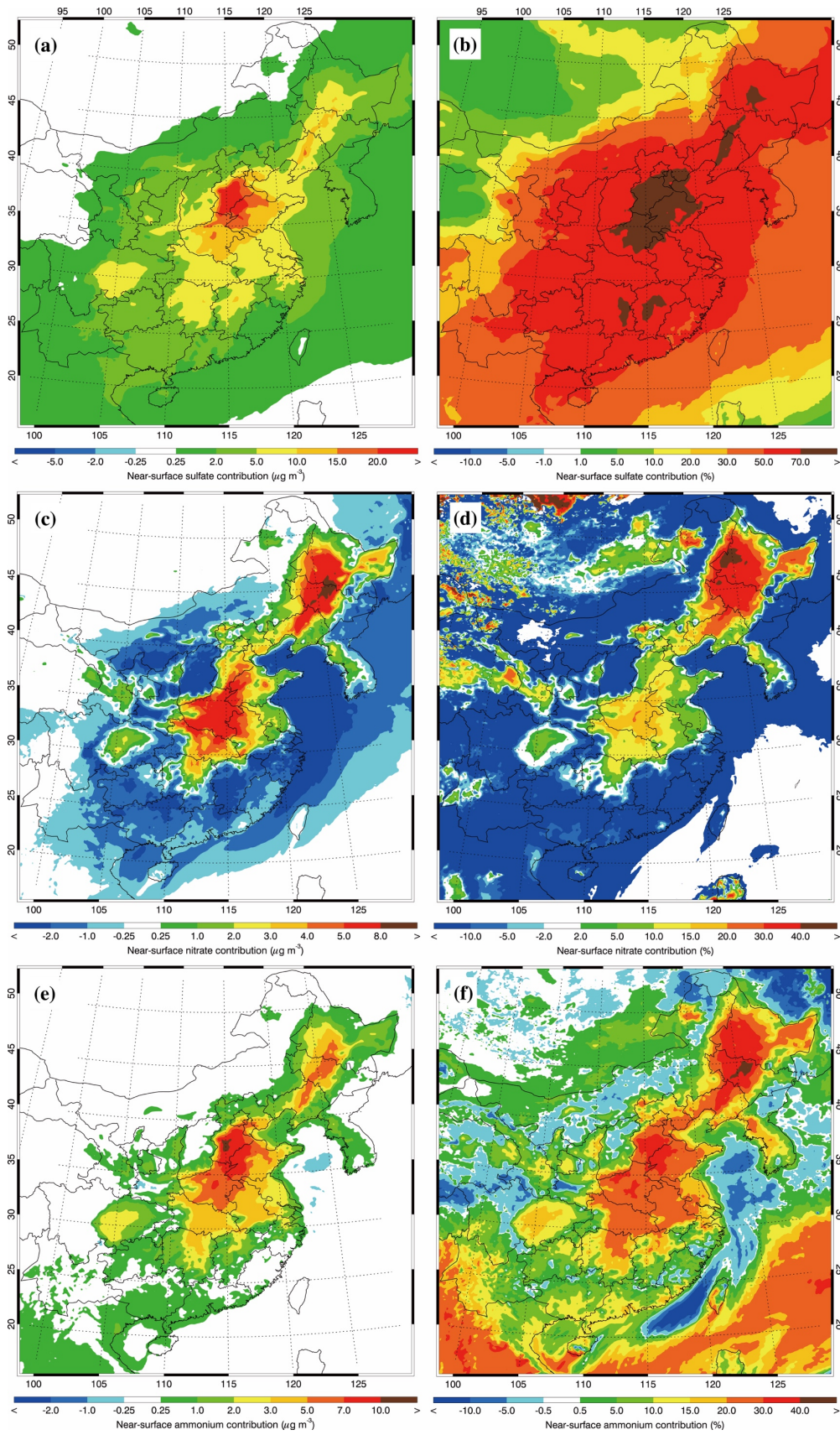
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Figure 9 Average variations of daytime (a)/(b) NO₂ photolysis and (c)/(d) O₃ concentration at 1st/5th model layer (around 18 m and 420 m above the ground surface, respectively) caused by the ALW from 05 December 2015 to 04 January 2016 in NCP.



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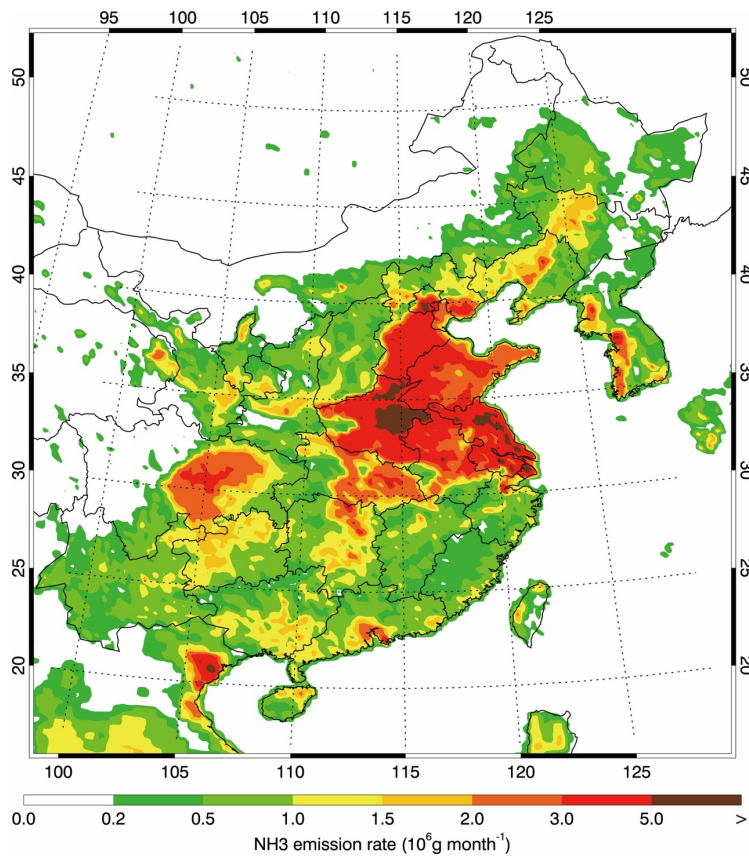
Figure 10 Average near-surface (a) absolute and (b) relative PM_{2.5} contribution caused by the ALW-J from 05 December 2015 to 04 January 2016.



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Figure 11 Average near-surface (a)/(c)/(e) absolute and (b)/(d)/(f) relative contribution to sulfate/nitrate/ammonium concentrations, caused by the ALW-HET from 05 December 2015 to 04 January 2016.

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970 Figure 12 Spatial distribution of NH₃ emission rate in December.

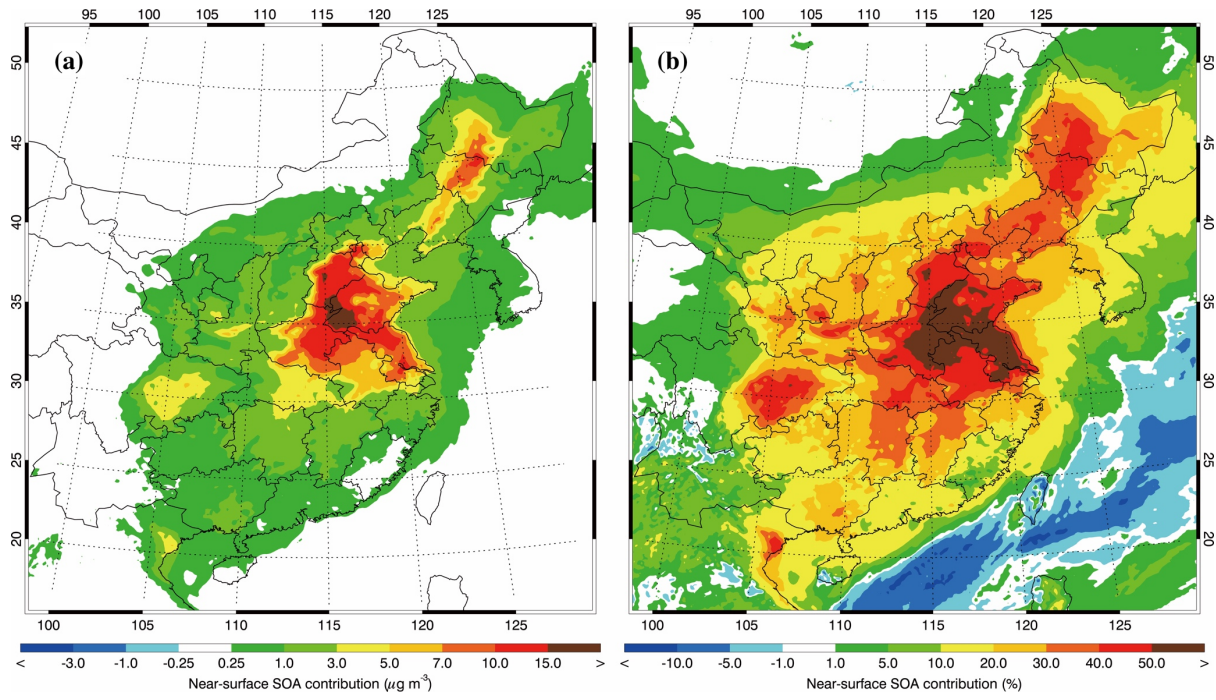
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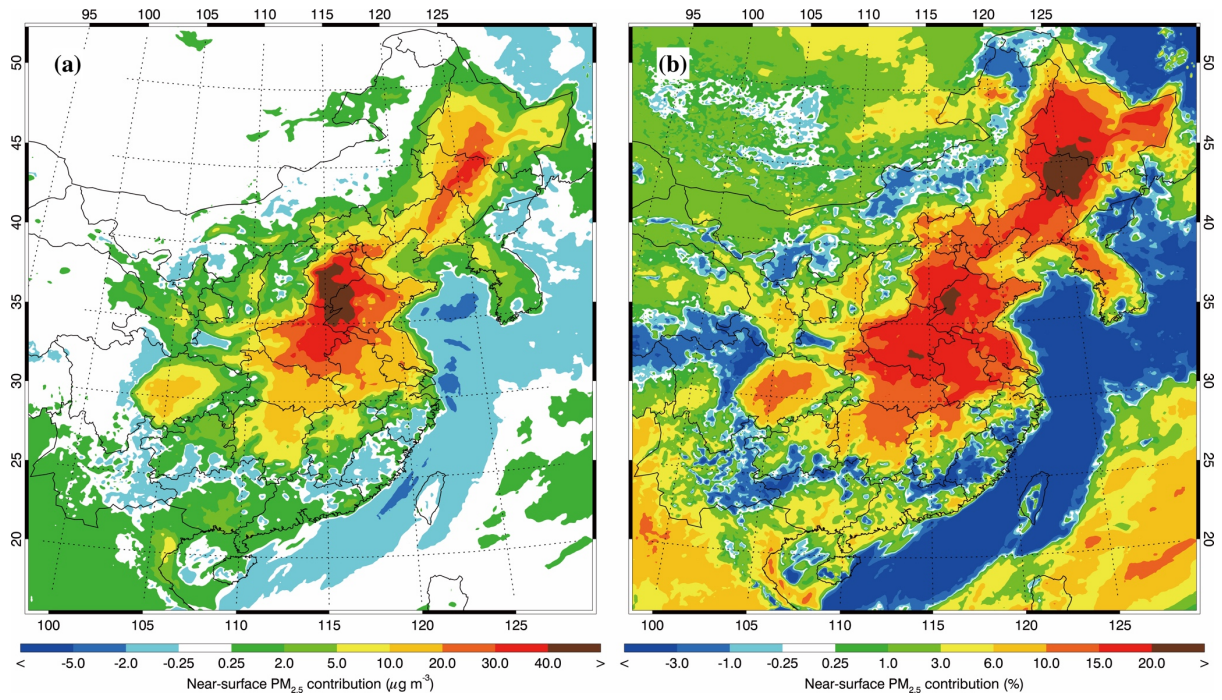
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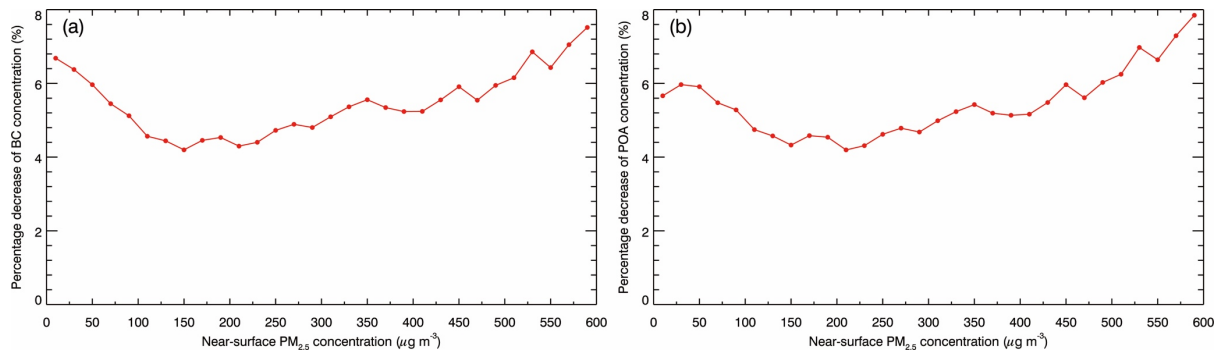
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Figure 13 Average near-surface (a) absolute and (b) relative SOA contribution caused by the ALW-HET from 05 December 2015 to 04 January 2016.



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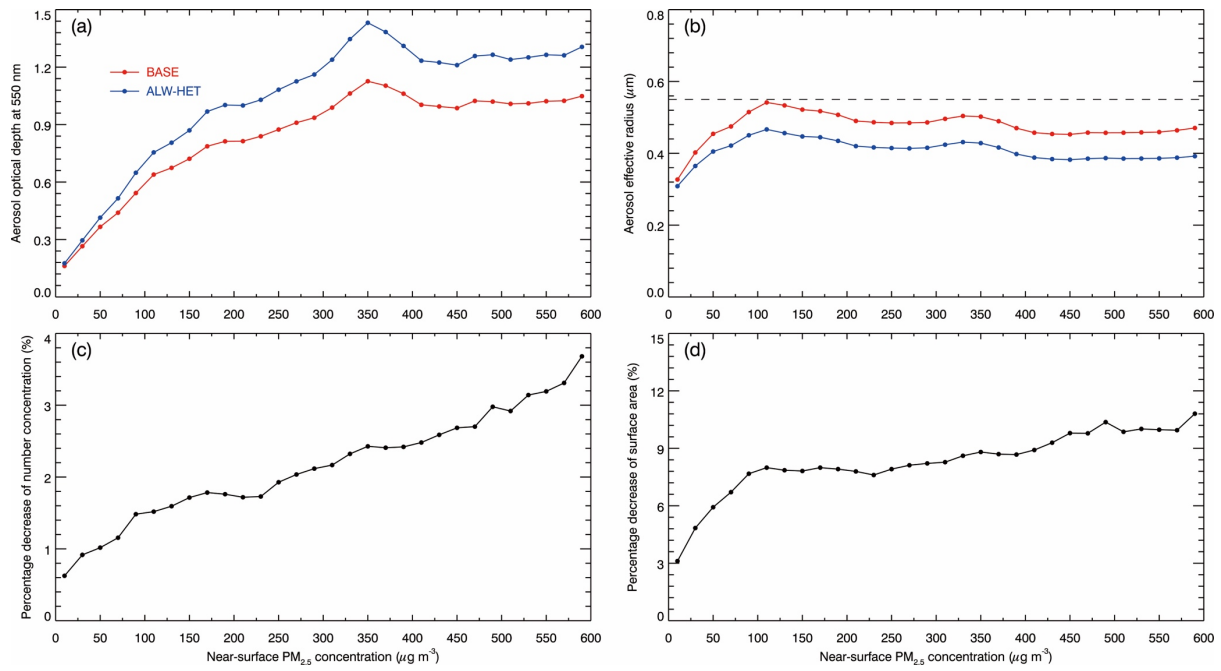
Figure 14 Average near-surface (a) absolute and (b) relative PM_{2.5} contribution caused by the ALW-HET from 05 December 2015 to 04 January 2016.



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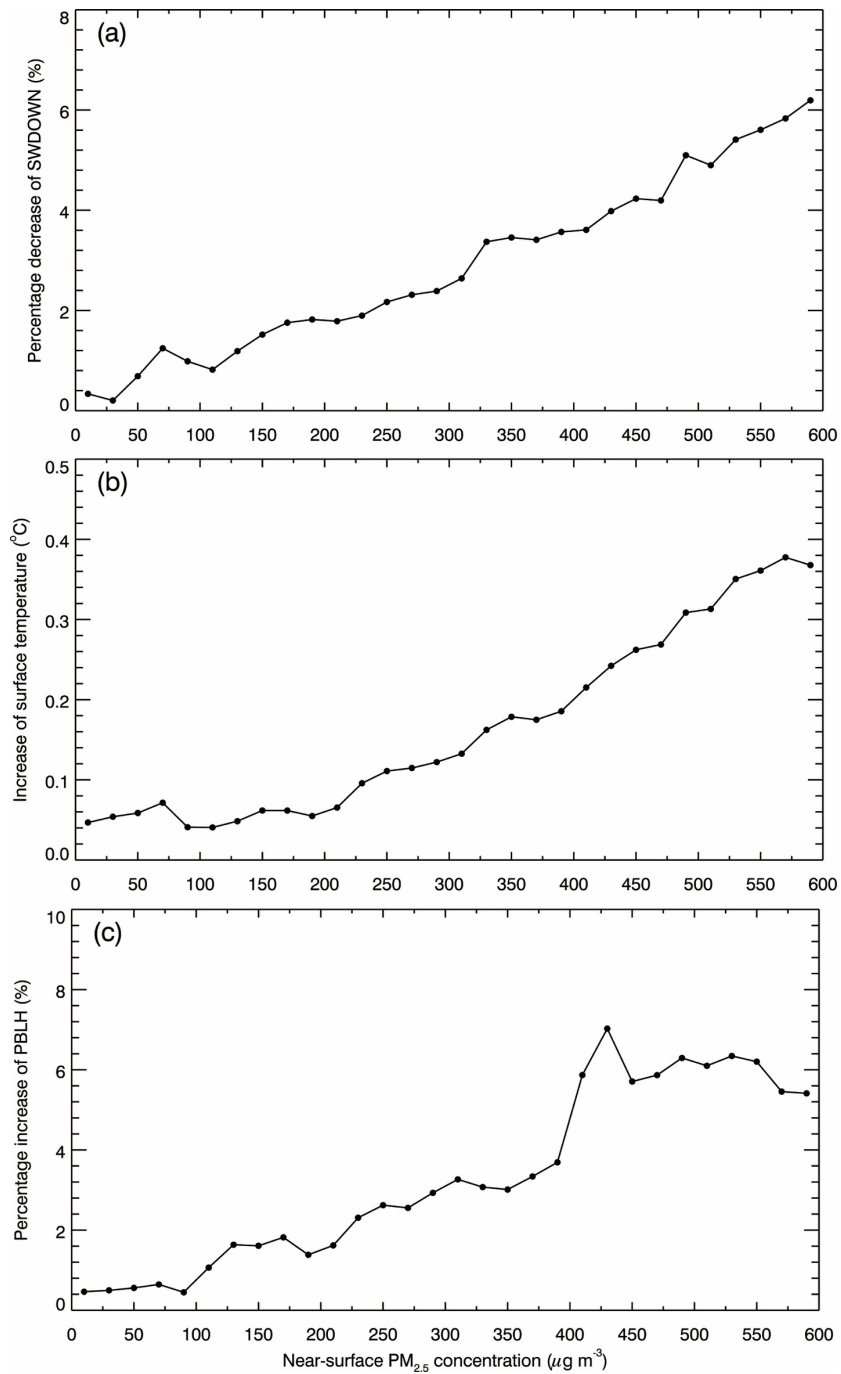
996 Figure 15 Average percentage decrease of (a) BC and (b) POA concentrations caused by the
997 ALW-HET, as a function of the near-surface [PM_{2.5}] in NCP from 05 December 2015 to 04
998 January 2016.

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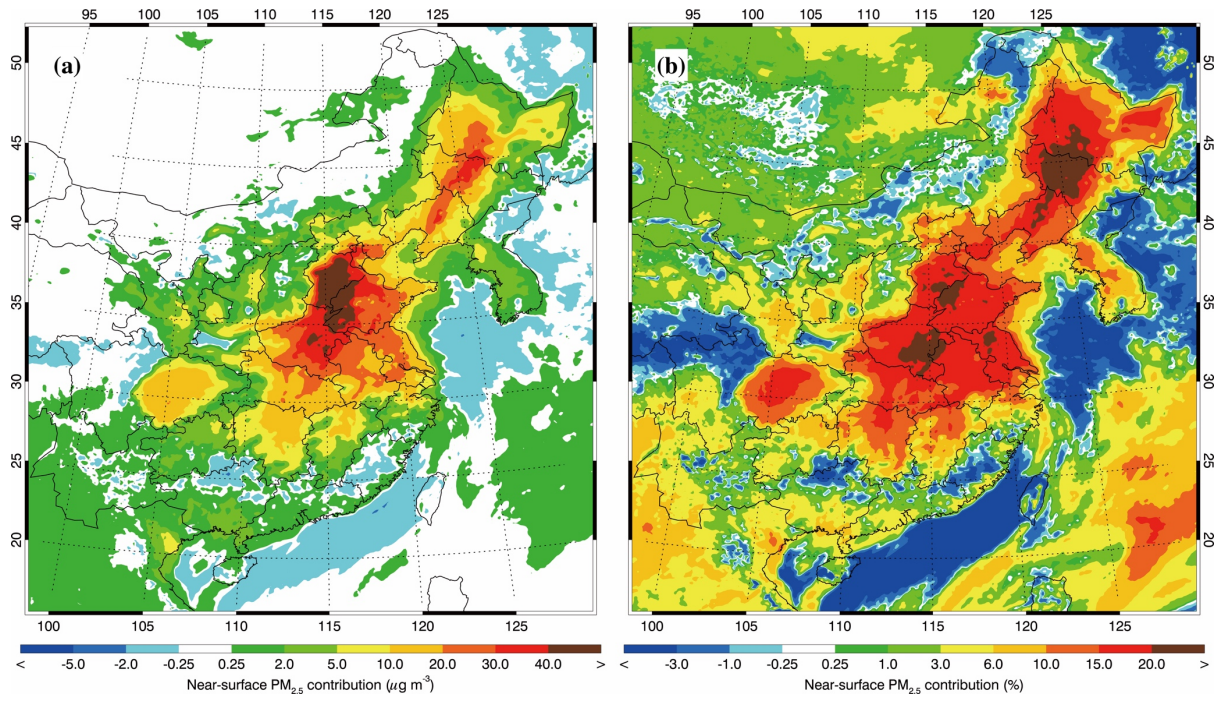
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Figure 16 Average variations of (a) AOD and (b) Reff in f_{base} (red line) and f_{alw_het0} (blue line), respectively, and average percentage decrease of near-surface (c) aerosol number concentration and (d) surface area caused by the ALW-HET, as a function of bin [PM_{2.5}] in NCP from 05 December 2015 to 04 January 2016.



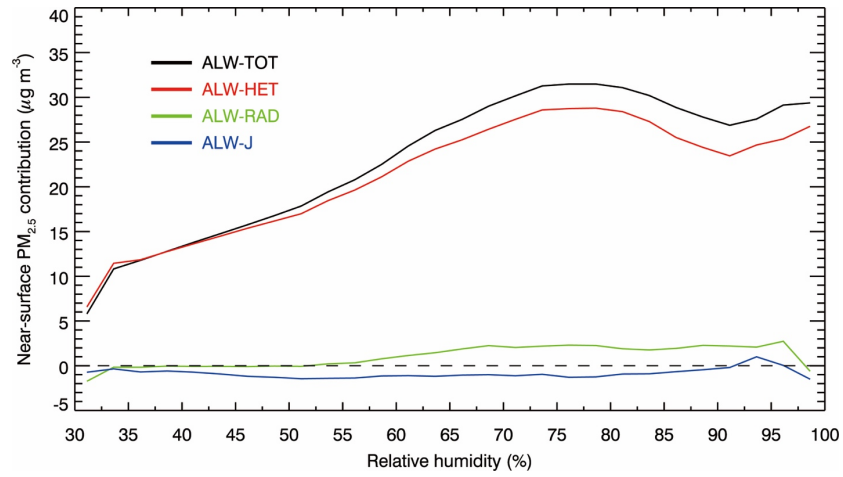
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Figure 17 Average (a) percentage decrease of SWDOWN at the ground surface, (b) decrease of TSFC, and (c) percentage decrease of PBLH caused by the ALW-HET, as a function of the near-surface [PM_{2.5}] in NCP during daytime from 05 December 2015 to 04 January 2016.



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Figure 18 Average near-surface (a) absolute and (b) relative PM_{2.5} contribution caused by the ALW-TOT from 05 December 2015 to 04 January 2016.



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Figure 19 Average contributions to the near-surface [PM_{2.5}] caused by the ALW-TOT (black line), ALW-HET (red line), ALW-RAD (green line), and ALW-J (blue line), respectively, as a function of the RH in NCP from 05 December 2015 to 04 January 2016.