



1     **Local and regional contributions to fine particulate matter in the 18 cities of**  
2                                    **Sichuan Basin, southwestern China**

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## 23 Abstract

24 The Sichuan Basin (SCB) is one of the regions suffering from severe air pollution in China, but  
25 fewer studies have been conducted for this region than the more developed regions in North and  
26 East China. In this study, a source-oriented version of the Community Multi-scale Air Quality  
27 (CMAQ) model was used to quantify contributions from nine regions to  $PM_{2.5}$  (i.e., particulate  
28 matter (PM) with an aerodynamic diameter less than  $2.5 \mu m$ ) and its components in the 18 cities  
29 within the SCB in the winter (December 2014 to February 2015) and summer (June to August,  
30 2015). In the winter, citywide average  $PM_{2.5}$  concentrations are  $45\sim 126 \mu g m^{-3}$ , with 21~51%  
31 and 39~66% due to local and non-local emissions, respectively. In the summer, 15~45% and  
32 25~52% of citywide average  $PM_{2.5}$  ( $14\sim 31 \mu g m^{-3}$ ) are due to local and non-local emissions,  
33 respectively. Compared to primary PM (PPM), the inter-region transport of secondary inorganic  
34 aerosols (SIA, including ammonia ( $NH_4^+$ ), nitrate ( $NO_3^-$ ), and sulfate ( $SO_4^{2-}$ )) is greater. The  
35 region to the east of SCB (R7) is the largest contributor outside the SCB, and it can contribute  
36 approximately 80% in the northeast, east, and southeast rims of the SCB, but only 10% in the  
37 other regions in both seasons. Under favorable transport conditions, regional transport of air  
38 pollutants from R7 could account for up to  $35\sim 100 \mu g m^{-3}$  of  $PM_{2.5}$  in each of the SCB cities in  
39 the winter. This study demonstrates that it is important to have joint emission control efforts  
40 among cities within the SCB and neighbor regions to the east in order to reduce  $PM_{2.5}$   
41 concentrations and prevent high  $PM_{2.5}$  days for the entire basin.

42 **Keywords:** Sichuan Basin, local emission, regional transport,  $PM_{2.5}$ , source apportionment



## 43 1. Introduction

44 Air pollution in major economic centers in China, including the North China Plain (NCP),  
45 Yangtze River Delta (YRD), and Pearl River Delta (PRD), has been extensively studied in recent  
46 years. Regional transport has been identified as an important process of particulate matter (PM)  
47 (Jiang et al., 2015; Li et al., 2015; Tang et al., 2016; Wang et al., 2015; Yang et al., 2018; Ying et al.,  
48 2014; Zhang et al., 2013; Zhao et al., 2013b; Zheng et al., 2015). Several urbanized areas in  
49 western China also have been suffering from severe air pollution due to rapid industrial and  
50 urban development, but fewer studies have been conducted. One of such areas is the Sichuan  
51 Basin (SCB), which covers an area of  $\sim 0.22$  million  $\text{km}^2$  and is home for more than 100 million  
52 residents (NBSC, 2015). The SCB is topographically isolated, with mountains on all sides.  
53 Emissions from two megacities (Chongqing and Chengdu) and 16 other cities in SCB tend to  
54 accumulate in the basin, causing severe air pollution (Zhao et al., 2018; Ning et al., 2018). In  
55 2015, annual average concentrations of  $\text{PM}_{2.5}$  (i.e., PM with an aerodynamic diameter less than  
56  $2.5 \mu\text{m}$ ) measured in Chengdu and Chongqing were  $64$  and  $57 \mu\text{g m}^{-3}$ , respectively, about six  
57 times of the World Health Organization (WHO) guideline ( $10 \mu\text{g m}^{-3}$ ) (NBSC, 2015; WHO,  
58 2006).

59 To design effective  $\text{PM}_{2.5}$  control strategies for the SCB, it is necessary to quantify the  
60 source contributions and inter-/intra-region transport of  $\text{PM}_{2.5}$  within the SCB and its  
61 surrounding regions. Different methods have been used to quantify source and regional  
62 contributions of pollutants, including receptor-based models, chemical transport models (CTMs),  
63 air parcel trajectory models, and remote sensing. Receptor-based models, such as the Positive  
64 Matrix Factorization (PMF) (Paatero and Tapper, 1994) and the Chemical Mass Balance (CMB)  
65 (Watson et al., 1990), are semi-quantitative and cannot determine the source contributions from  
66 an exact emission sector or a specific location. They also require a large number of monitoring  
67 data and can only resolve source contributions at the monitoring sites (Hopke, 2016). In addition,  
68 receptor-based models. Air parcel trajectory models, such as the Potential Source Contribution  
69 Function (PSCF) and the Hybrid Single Particle Lagrangian Integrated Trajectory Model  
70 (HYSPPLIT) (Liu et al., 2018; Begum et al., 2005; Stein et al., 2015) are used to determine the  
71 sources of air pollutants, but they are not directly quantitative either. Remote sensing has been  
72 used to show the transport of dust storms (Uno et al., 2009) and wildfire plumes (Wu et al.,  
73 2018), but it cannot quantitatively determine source-region contributions. CTMs have also been  
74 used to quantify contributions from different regions and sectors to air pollutants for areas at  
75 local to global scales (Lelieveld et al., 2015; Shi et al., 2017; Kim et al., 2015; Bove et al.,  
76 2014; Itahashi et al., 2017). Sensitivity analysis and tagged-tracer methods are the main types of  
77 methods used in CTMs to quantify source contributions (Burr and Zhang, 2011). Sensitivity  
78 analysis such as the brute-force method is more suitable to estimate air quality changes due to  
79 emission perturbations, as emissions from certain sources would be eliminated or reduced in  
80 each simulation of sensitivity analysis (Burr and Zhang, 2011; Huang et al., 2018; Han et al.,  
81 2018). In the tagged-tracer method, source-tagged-species are used to track air pollutants from



82 specific emission regions and sectors and would go through all the non-linear chemical and  
83 physical processes in the model, thus this type of source apportionment method is considered to  
84 provide more realistic evaluations of the contributions of different sources or source regions to  
85 the current level of air pollution under the current emission intensities (Burr and Zhang,  
86 2011; Chen et al., 2017; Wang et al., 2009).

87 Transport of PM<sub>2.5</sub> and its precursors has been studied for Chengdu, Chongqing, and the  
88 entire SCB region (Zhu et al., 2018). Based on in-situ observations and the HYSPLIT model,  
89 studies have found that dust storms from northwestern China and biomass burning activities  
90 would cause high PM days, particularly in spring (Zhao et al., 2010; Tao et al., 2013; Chen and  
91 Xie, 2014; Chen et al., 2015). Using the PSCF model, a study reported that the main potential  
92 sources of PM<sub>2.5</sub> for Chengdu were from southeastern cities and western margin of the SCB in  
93 addition to local emissions from December 2013 to February 2014 (Liao et al., 2017). Based on  
94 the HYSPLIT and PSCF analyses, air pollution was determined mainly from the south during  
95 persistent extreme haze days in Chengdu from 6<sup>th</sup> to 16<sup>th</sup> January 2015 (Li et al., 2017). Using a  
96 source-oriented version of the Community Multi-scale Air Quality (CMAQ) model, Ying et al.  
97 (2014) predicted that non-local emissions account for approximately 30% and 50% of total of  
98 nitrate (NO<sub>3</sub><sup>-</sup>) and sulfate (SO<sub>4</sub><sup>2-</sup>) ions in the SCB in summer and winter, respectively. However,  
99 the aforementioned studies based on the HYSPLIT and PSCF models are not fully quantitative in  
100 terms of emission contributions. Their accuracy is also limited by the meteorological inputs to  
101 drive these models, which are often in very coarse resolutions (0.5 to 1.0 degree) that are not  
102 enough to accurately predict air pollutant transport within the SCB. Also, previous country-level  
103 modeling study did not have sufficient spatial resolution to properly quantify the transport  
104 among different regions within the SCB.

105 Since both inter-regional transport within the cities in the basin and from outside the region  
106 can greatly affect PM<sub>2.5</sub> concentrations, systematically quantifying contributions from different  
107 regions to PM<sub>2.5</sub> for all the 18 cities in the SCB is urgently needed as emission controls are  
108 further tightened to improve air quality in this region. In this study, an improved source-oriented  
109 CMAQ model was used to quantify the contributions from nine regions (five within the SCB and  
110 four outside) to PM<sub>2.5</sub> and its components for the 18 SCB cities. The assumption is that the  
111 transport of air pollutants is evident among the SCB cities and some cities in the rims of the SCB  
112 are greatly affected by emissions from outside SCB. Therefore, the objectives of this study are to  
113 quantitatively determine (1) the inter-region transport of air pollutants emitted in the SCB and its  
114 contributions to PM<sub>2.5</sub> within the basin and (2) the contributions of emissions outside the basin to  
115 PM<sub>2.5</sub> in the SCB cities. This is a companion paper to our previous paper, which reports the  
116 model performance and analyzes the characteristics of PM<sub>2.5</sub> and ozone (O<sub>3</sub>) pollution in the  
117 basin (Qiao et al., 2018).

## 118 **2. Methods and materials**

### 119 **2.1. Model description**

120 The source-oriented CMAQ model used in this study is based on the CMAQ model version  
121 5.0.1. The gas phase and aerosol mechanisms are extended from the standard SAPRC-99



122 photochemical mechanism and aerosol module version 6 (AERO6). This version of the source-  
123 oriented CMAQ is capable of simultaneously tracking both primary particulate matter (PPM) and  
124 secondary inorganic aerosol (SIA, including ammonia ( $\text{NH}_4^+$ ),  $\text{NO}_3^-$ , and  $\text{SO}_4^{2-}$ ) from multiple  
125 source sectors and regions. It unifies the two previous models individually developed for PPM  
126 (Hu et al., 2015) and SIA (Ying et al., 2014; Shi et al., 2017) into a single consistent model  
127 framework. For SIA, multiple source-tagged reactive species are introduced in both gas and  
128 particle phases to represent the same species originated from different source sectors or regions.  
129 The corresponding photochemical mechanisms, aerosol and cloud modules are expanded so that  
130 SIA and their precursors from different regions can be tracked separately throughout the model  
131 calculations. For PPM, source-tagged non-reactive tracers are added to track the total amount of  
132 PPM emitted from different source sectors and regions. Source contributions to SOA are not  
133 resolved in the current model.

## 134 2.2. Model application

135 The source-oriented CMAQ model was applied to quantify nine source-region contributions  
136 to  $\text{PM}_{2.5}$  and its components (PPM and SIA) for the 18 cities in the winter (from December 2014  
137 to February 2015) and summer (June to August 2015) using nested domains settings. The  
138 locations of domains, nine source-regions, and the 18 cities of the SCB are shown in Figure 1.  
139 The horizontal resolutions of the parent and nested domains are 36-km and 12-km, respectively.  
140 There are 18 vertical layers with an overall height of 20 km and the layer closest to the land  
141 surface is up to 35 m. As Chengdu and Chongqing are the two largest cities in western China, we  
142 classified western Chongqing, eastern Chongqing, and Chengdu into three individual regions (R1,  
143 R5, and R4, respectively). The five cities in the northeastern SCB (Bazhong, Dazhou,  
144 Guangyuan, Guang'an, and Nanchong) are grouped into R2 and they have relatively lower  
145 anthropogenic emission densities compared to most of the other cities in the basin (Qiao et al.,  
146 2018). As shown in Figures 2a and 3a, R1, R2, and R5 are generally the upwind region within  
147 the SCB. The rest SCB cities are grouped into R3. Region 8 is the area of Sichuan Province  
148 excluding those cities in the SCB. R6 includes three provinces to the south of the SCB and R7  
149 has the Chinese provinces to the east and northeast of the SCB. R9 includes Xinjiang, Qinghai,  
150 Gansu, and Tibet in China and other countries.

151 Meteorological inputs were generated using the Weather Research and Forecasting (WRF)  
152 model version 3.9 based on the 6-hourly FNL (Final) Operational Global Analysis data from the  
153 National Center for Atmospheric Research (NCAR) with a spatial resolution of  $1.0 \times 1.0^\circ$   
154 (<http://dss.ucar.edu/datasets/ds083.2/>). The anthropogenic emission inventory used was the  
155 Emission Database for Global Atmospheric Research (EDGAR) version 4.3.1 for the year of  
156 2012 (Crippa et al., 2018). The inventory was directly used for the model year of 2014-2015 as  
157 no reliable sources for emission changes in the SCB are available. The monthly EDGAR  
158 inventories have a spatial resolution of  $0.1 \times 0.1^\circ$  ( $\sim 10 \text{ km} \times 10 \text{ km}$ ) and were re-projected to the  
159 model domains using the Spatial Allocator (<https://www.cmascenter.org/sa-tools/>). Temporal  
160 profiles specific to sources were used to allocate the monthly emission rates to hourly values for  
161 CMAQ modeling (Olivier et al., 2003; Streets et al., 2003; Wang et al., 2010). The EDGAR  
162 inventory includes carbon monoxide (CO), nitrogen oxides ( $\text{NO}_x$ ), sulfur dioxide ( $\text{SO}_2$ ),  
163 ammonia ( $\text{NH}_3$ ), non-methane volatile organic compounds (NMVOCs),  $\text{PM}_{2.5}$ ,  $\text{PM}_{10}$  (PM with  
164 an aerodynamic diameter less than  $10 \mu\text{m}$ ), elemental carbon (EC), and organic carbon (OC)



165 from various sources. Emission sources are grouped into six categories: energy, industries,  
166 residential activities, on-road transportation, off-road transportation, and agriculture. In addition  
167 to these six anthropogenic sources, contributions of biogenic sources were also determined using  
168 emissions generated by the Emissions of Gases and Aerosols from Nature (MEGAN) model  
169 version 2.1 (Guenther et al., 2012). The contributions from open burning were estimated based  
170 on the fire emissions generated from the Fire Inventory from the National Center for  
171 Atmospheric Research (NCAR FINN) (Wiedinmyer et al., 2010). Contributions of dust and sea  
172 salt emissions were determined based on in-line generated emissions during CMAQ simulations.  
173 The initial and boundary conditions (ICs and BCs, respectively) for the 36-km domain were  
174 based on the default CMAQ profiles, and those for the 12-km domain were generated using the  
175 CMAQ outputs from the 36-km simulations. More details about the setup and configurations of  
176 the WRF/CMAQ modeling system can be found in a previous publication for China (Kang et al.,  
177 2016).

### 178 3. Results and discussion

#### 179 3.1. Model performance

180 The model performance on meteorological parameters and 24-hr PM<sub>2.5</sub> in the 12-km domain  
181 for the two seasons has been evaluated in a companion paper (Qiao et al., 2018) and is briefly  
182 summarized here. As the meteorology predictions are important in modeling air pollutant  
183 transport (Zhao et al., 2009), the WRF model performance on wind speed (WS), wind direction  
184 (WD), temperature (T), and relative humidity (RH) were evaluated using hourly observation data  
185 were downloaded from the National Climate Data Center (NCDC;  
186 <ftp://ftp.ncdc.noaa.gov/pub/data/noaa/>, last accessed on June 20, 2018). The mean biases (MBs)  
187 of predicted RH (-10.8% to -1.1%) and T (-0.9 to -0.1°C) in each month are comparable to other  
188 studies in China (Wang et al., 2010; Zhao et al., 2013a; Wang et al., 2013). The MB of WD in  
189 each month (-5° to 6°) meet the benchmark of  $< \pm 10^\circ$  suggested by Emery et al. (2001). Although  
190 the MB of WS in each month (0.5 to 1.1 m s<sup>-1</sup>) does not meet the benchmark of  $< \pm 0.5$  m s<sup>-1</sup>, the  
191 gross errors (GEs: 1.4-1.9 m s<sup>-1</sup>) are within the benchmark of  $< 2.0$ . For 24-hr PM<sub>2.5</sub>  
192 concentrations, the statistical measures of model performance are generally within the criteria  
193 recommended by Emery et al. (2017) for regulatory applications, with only a few cities  
194 exceeding the normalized mean bias (NMB) metric of  $< \pm 30\%$  in the winter (Chongqing 42%;  
195 Guangyuan 41%; Mianyang 37%; Meishan 31%; Ziyang 48%) and in the summer (Dazhou -39%)  
196 (Figure S1). The 24-hr PM<sub>2.5</sub> predictions meet the goals of normalized mean error (NME  $< \pm 35\%$ ),  
197 fractional bias (FB  $< \pm 30\%$ ), and fractional error (FE  $< \pm 50\%$ ) in all the cities in both seasons,  
198 except for the NME of Ziyang (58%) in the winter. The predictions of major PM<sub>2.5</sub> components  
199 (including OC, EC, NH<sub>4</sub><sup>+</sup>, NO<sub>3</sub><sup>-</sup>, and SO<sub>4</sub><sup>2-</sup>) in Chengdu and Chongqing are comparable with  
200 observations. Both predictions and observations suggest that OC and SIA are the largest  
201 contributors to PM<sub>2.5</sub> in summer and winter, with combined contributions of ~70% (Qiao et al.,  
202 2018).

#### 203 3.2. Seasonal average contributions

##### 204 3.2.1. Contributions at the city centers

205 The predicted PM<sub>2.5</sub> concentrations and source-region contributions at the 18 SCB city  
206 centers (the grid cell which is located in the center of all the national air quality stations within



207 the urban areas of a city) are presented in Table 1 for winter and Table S1 for summer. In all the  
208 city centers, the predicted PM<sub>2.5</sub> concentrations are much higher in the winter (60~191 μg m<sup>-3</sup>)  
209 than in the summer (14~64 μg m<sup>-3</sup>). The city centers are considerably affected by both local and  
210 regional emissions in both seasons. Emissions within the SCB are the major contributor to PM<sub>2.5</sub>  
211 in Chengdu and Chongqing in both seasons (~80%) and emissions outside the SCB contribute  
212 approximately 7~15%. Among the regions within the SCB, local emissions (i.e., emissions from  
213 the region where the city center is in) are the largest contributor to PM<sub>2.5</sub> in Chongqing and  
214 Chengdu in both seasons (about 70% and 58%, respectively). However, emissions from R3 (i.e.,  
215 the 11 cities in the northwestern, western, and southwestern SCB) also have considerable  
216 contributions in Chengdu (~20% and 14% in the winter and summer, respectively). For the R3  
217 cities, the contributions of emissions within the SCB (64-83%) are also larger than that from  
218 outside the SCB (8~26%) in both seasons. Local emissions are the largest contributor for R3  
219 cities (40~60%), except that Suining has only ~13% due to its local region. For the five cities in  
220 the northern SCB (R2), emissions within the SCB account for 40~70% in both seasons, including  
221 37~57% from local emissions. Emissions outside the SCB also have large contributions to the  
222 R2 cities (21~36% and 17~28% in the winter and summer, respectively), as R2 is located in one  
223 of the regions where winds from R7 intrude the basin (Figure 2a). In the winter, contributions  
224 from SOA and others (including IC, BC, windblown dust, and sea salt) are less than 8% each. In  
225 the summer, SOA and others each contribute 9~28% and less than 10%, respectively, but the  
226 SOA contributions larger than 15% are found only in the city centers, where summer PM<sub>2.5</sub>  
227 concentrations are less than 30 μg m<sup>-3</sup>. In summary, local emissions are the largest contributor  
228 for all the city centers in both seasons, except for Suining. In summary, the non-local  
229 contributions for the city centers are in the ranges of 25~52% in the winter (except for 75% in  
230 Suining) and of 14~40% in the summer (except for 61% in Suining), and emissions outside the  
231 SCB account for 7~36% in the seasons.

232

### 233 3.2.2. Spatial variations and citywide area-weighted averages

234 The spatial variations of source-region contributions to PM<sub>2.5</sub> in the winter and summer are  
235 presented in Figures 2 and 3, respectively. In both seasons, local emissions are generally the  
236 largest contributor in each city, except that R7 has contributions similar to or larger than that of  
237 local emissions for most regions in eastern Chongqing (R5) and R2. Specifically, the  
238 contributions from R7 to PM<sub>2.5</sub> in R2 and R5 are approximately 20~80% in the winter and 20~60%  
239 in the summer. R7 also has contributions larger than 20% for few areas in R3. The regions of R6,  
240 R8, and R9 outside the SCB each has contributions of <5% across the basin, except for some  
241 very limited areas in the western and southern rims of the SCB. The contributions from R6, R8,  
242 and R9 are low because these areas are less urbanized and industrialized. In addition, the  
243 mountains to the west and south of the SCB also prevent the transport of air pollutants from  
244 these regions into the SCB (Figures 1c, 2a, and 3a). In summary, R7 is the sole non-SCB region  
245 that can have >20% contributions to PM<sub>2.5</sub> in the SCB, and its impact decreases from the  
246 northeast, east, and southeast to others in the basin.

247 As shown in Figures 2 and 3, PM<sub>2.5</sub> concentrations and its source contributions from a given  
248 region may vary greatly within a city in both seasons. For example, about 20~80% of PM<sub>2.5</sub>  
249 across Chengdu (R4) and western Chongqing (R1) are due to local emissions in each season, and  
250 higher PM<sub>2.5</sub> concentrations are generally related to higher local contributions. For the downwind



251 regions of Chengdu and western Chongqing, they receive considerable contributions from the  
252 two mega-cities. For example, over half areas of Meishan and Ya'an, which are downwind of  
253 Chengdu, have 20~40% and 20~60% of PM<sub>2.5</sub> concentrations due to Chengdu in the winter,  
254 respectively. In the two seasons, western Chongqing contributes to about 10~40% of PM<sub>2.5</sub>  
255 concentrations in its neighboring cities, except that most area of eastern Chongqing (R5) is not  
256 affected by emissions from western Chongqing, as R5 is upwind of western Chongqing (Figure  
257 2a). Because of the large spatial variations of PM<sub>2.5</sub> and its source contributions in the basin, we  
258 further calculated their citywide area-weighted averages (Table 2). In the winter, the citywide  
259 average PM<sub>2.5</sub> concentrations in Chengdu and urban Chongqing are 99 and 110 μg m<sup>-3</sup>, with only  
260 38% and 47% due to local emissions, respectively. Non-local emissions also have high  
261 contributions in other SCB cities, with citywide averages of 39~66% and 25~52% in the winter  
262 and summer, respectively. The above suggest the importance of regional emission control to  
263 reduce PM<sub>2.5</sub> concentrations for the entire basin.

### 264 3.2.3. Differences in PPM and SIA

265 The transport distances of PPM, NH<sub>4</sub><sup>+</sup>, NO<sub>3</sub><sup>-</sup>, and SO<sub>4</sub><sup>2-</sup> might be different, because of the  
266 differences in chemical and physical processes that affect their concentrations in the atmosphere  
267 (Hu et al., 2015; Ying et al., 2014). This leads to significant differences in their regional  
268 distributions and thus requires different control strategies. From the source-region contributions  
269 to PPM and SIA for each city center shown in Figure 4, it is obvious that the regional transport  
270 of SIA is more significant than that of PPM. In the city centers of Chengdu and Chongqing,  
271 55~65% of PPM and 25~45% of SIA are due to local emissions in the two seasons. In the city  
272 centers of R2, PPM is also more from local emissions (65~80%) than SIA (25~45%) in both  
273 seasons. Similarly, local emissions have larger contributions to PPM (50~85%) than to SIA  
274 (34~50%) in all the city centers of R3 except for Suining, which is not significantly affected by  
275 local emissions. The spatial distributions of source-region contributions to PPM and SIA also  
276 indicate more significant transport of SIA (Figure S1-4) than PPM. For example, R3 contributes  
277 to >20% of SIA across entire Chengdu, but only half areas of Chengdu are similarly affected  
278 (>20%) by R3 for PPM. From the north to south in R2, the contributions from R7 to PPM  
279 decrease from ~55% to ~10%, while the contributions of R7 to SIA decrease from ~75% to  
280 ~20%. The contributions to NH<sub>4</sub><sup>+</sup>, NO<sub>3</sub><sup>-</sup>, and SO<sub>4</sub><sup>2-</sup> in each city center from local emissions and  
281 emissions within and outside SCB are further analyzed (Tables S3 and S4). In each city center,  
282 concentrations of SO<sub>4</sub><sup>2-</sup> (3.8~12.6 and 12~41 μg m<sup>-3</sup>) are much higher than that of NH<sub>4</sub><sup>+</sup> (1.4~4.0  
283 and 6.0~17.0 μg m<sup>-3</sup>) and NO<sub>3</sub><sup>-</sup> (0.3~2.4 and 6~20 μg m<sup>-3</sup>) in the summer and winter,  
284 respectively. Also, the transport of SO<sub>4</sub><sup>2-</sup> is greater than the other two ions, as the percentage  
285 contributions from emissions outside the SCB to SO<sub>4</sub><sup>2-</sup> is higher than that to NH<sub>4</sub><sup>+</sup> and NO<sub>3</sub><sup>-</sup> in  
286 each city center. In both seasons, emissions outside SCB contribute <25% of NH<sub>4</sub><sup>+</sup> in the city  
287 centers, except for Chongqing (26%) in the summer and Bazhong (36%) and Guangyuan (33%)  
288 in the winter. As for NO<sub>3</sub><sup>-</sup>, emissions outside SCB also contribute <25% in the city centers in  
289 both seasons, except for Bazhong (49%), Dazhou (34%), and Guangyuan (25%) in the summer  
290 and all the cities of R2 (27~57%) in the winter. In the two seasons, emissions outside SCB  
291 account for 22~33% of SO<sub>4</sub><sup>2-</sup> in Chengdu and Chongqing, while they contribute 52~70% of  
292 SO<sub>4</sub><sup>2-</sup> for the R2 cities. For the R3 cities, emissions outside SCB account for 25~53% of SO<sub>4</sub><sup>2-</sup> in  
293 the city centers in the seasons, except for Meishan (21%) in the winter. All the above suggest



294 that it would be more efficient to control the SIA (particularly  $\text{SO}_4^{2-}$ ) and its precursors than  
295 PPM in order to reduce the transport of air pollutants within and into the basin.

### 296 **3.3. Maximum daily contributions from a given region**

297 The maximum daily contribution from a given region to  $\text{PM}_{2.5}$  (MDC,  $\mu\text{g m}^{-3}$ ) in each city  
298 center is shown in Table 3 for winter and Table S4 for summer. The largest MDC for each city  
299 center (79~291 and 13~147  $\mu\text{g m}^{-3}$  in the winter and summer, respectively) are found associated  
300 with local emissions, except for Guangyuan and Suining. In Guangyuan and Suining, the largest  
301 MDCs in the winter are from R7 (62  $\mu\text{g m}^{-3}$ ) and R2 (110  $\mu\text{g m}^{-3}$ ), both are slightly higher than  
302 that from the local region of 60 and 105  $\mu\text{g m}^{-3}$ , respectively. Table 3 also shows that the inter-  
303 basin transport of air pollutants can have large contributions ( $>50 \mu\text{g m}^{-3}$ ) on high  $\text{PM}_{2.5}$  days  
304 ( $>150 \mu\text{g m}^{-3}$ ). For example, R7 contributes 99  $\mu\text{g m}^{-3}$  to total  $\text{PM}_{2.5}$  (200  $\mu\text{g m}^{-3}$ ) in Chongqing  
305 on a winter day. In Nanchong, the MDC due to western Chongqing (R1) is 58  $\mu\text{g m}^{-3}$ , when  
306 daily  $\text{PM}_{2.5}$  is 180  $\mu\text{g m}^{-3}$  on that day. In Chengdu, R3 and R7 can contribute up to 86 and 63  $\mu\text{g}$   
307  $\text{m}^{-3}$  on the days with daily  $\text{PM}_{2.5}$  of 267 and 151  $\mu\text{g m}^{-3}$ , respectively. In Deyang and Meishan,  
308 the MDCs from Chengdu are 147 and 138  $\mu\text{g m}^{-3}$  on the days having daily  $\text{PM}_{2.5}$  of 288 and 235  
309  $\mu\text{g m}^{-3}$ , respectively. Table S4 shows that air pollutant regional transport is also significant on  
310 certain days in the summer. For example, the highest summer MDC from R7 among the 18  
311 central cities is found for Bazhong (36  $\mu\text{g m}^{-3}$ ), when daily  $\text{PM}_{2.5}$  is 63  $\mu\text{g m}^{-3}$ . Chengdu  
312 contributes about 44, 16, 55, 13, 7, and 21  $\mu\text{g m}^{-3}$  to Deyang, Leshan, Meishan, Ya'an, and  
313 Mianyang on the summer days, when daily  $\text{PM}_{2.5}$  are 89, 56, 100, 85, 22, and 54  $\mu\text{g m}^{-3}$ ,  
314 respectively. All the above suggest that joint effects should be made by neighbor cities and  
315 provinces to reduce  $\text{PM}_{2.5}$  pollution for the entire SCB.

### 316 **4. Conclusion**

317 In this study, a source-oriented CMAQ model was applied to quantify contributions of nine  
318 regions to  $\text{PM}_{2.5}$  for the 18 cities in the SCB. The simulations were carried out for winter  
319 (December 2014 to February 2015) and summer (June to August 2015). Predicted citywide area-  
320 weighted average  $\text{PM}_{2.5}$  concentrations are much higher in the winter (60~191  $\mu\text{g m}^{-3}$ ) than in  
321 the summer (14~64  $\mu\text{g m}^{-3}$ ). In the winter, the citywide average  $\text{PM}_{2.5}$  concentrations in  
322 Chengdu and western Chongqing are 99 and 110  $\mu\text{g m}^{-3}$ , with and 44% and 52% due to non-local  
323 emissions, respectively. Non-local emissions also have high contributions in other SCB cities,  
324 with citywide averages of 39~66% and 25~52% in the winter and summer, respectively. Among  
325 the four regions outside the SCB, only the one to the northeast, east, and southeast of the SCB  
326 (R7) has large contributions to  $\text{PM}_{2.5}$  concentrations for the SCB in both seasons (10~80%), and  
327 the contributions decrease from the rims of the northeastern, eastern, and southeastern SCB to  
328 other regions. However, the MDCs from R7 are large (35~99  $\mu\text{g m}^{-3}$ ) for all the city centers in  
329 the winter. On high  $\text{PM}_{2.5}$  days in the winter, emissions outside SCB can contribute up to 99  $\mu\text{g}$   
330  $\text{m}^{-3}$  in a city center, suggesting the importance of regional emission control in not just reducing  
331 averaged  $\text{PM}_{2.5}$  but also preventing severe events. The transport of SIA is greater than that of  
332 PPM, suggested by that local emissions have higher contributions to PPM ( $>55\%$ ) than to SIA  
333 ( $<45\%$ ) in the city centers in both seasons. Among the three ions of SIA, the transport of  $\text{SO}_4^{2-}$  is  
334 the greatest in general, as  $>50\%$  of it in all the city centers is associated with non-local emissions



335 in both seasons, except that the contributions are 37~44% in Chongqing and Chengdu in the  
336 summer and Chongqing in the winter. In conclusion, in order to reduce PM<sub>2.5</sub> concentrations and  
337 prevent high PM<sub>2.5</sub> days for the entire SCB, local emissions and the transport of air pollutants  
338 within and across SCB should be all controlled simultaneously.

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497



498 Table 1. Predicted source-region contributions to PM<sub>2.5</sub> in the 18 SCB city centers in the winter.

Region ID	City	PM <sub>2.5</sub>	Contributions from each region, SOA, and others* (%)													
		(μg m <sup>-3</sup> )	R1	R2	R3	R4	R5	Within SCB	R6	R7	R8	R9	Outside SCB	SOA	Others	Non-local <sup>#</sup>
<b>R1</b>	Chongqing <sup>§</sup>	191	<b>67.7</b>	4.5	3.7	0.4	1.8	<b>78.1</b>	2.7	11.1	0.1	0.9	<b>14.8</b>	3.9	3.3	<b>25.2</b>
<b>R2</b>	Bazhong	65	2.0	<b>42.6</b>	2.8	1.2	2.9	<b>51.5</b>	1.7	31.9	0.1	2.7	<b>36.4</b>	6.9	5.1	<b>45.3</b>
	Dazhou	89	2.4	<b>47.3</b>	1.7	0.5	7.5	<b>59.4</b>	1.5	27	0.1	1.6	<b>30.2</b>	6.0	4.3	<b>42.3</b>
	Guangan	109	8.9	<b>50.4</b>	2.0	0.5	5.4	<b>67.2</b>	2.3	19.6	0.1	1.4	<b>23.4</b>	5.8	3.5	<b>40.2</b>
	Guangyuan	60	2.3	<b>45.5</b>	5.2	1.8	1.7	<b>56.5</b>	1.3	25.5	0.1	4.2	<b>31.1</b>	6.5	5.8	<b>42.1</b>
	Nanchong	120	6.1	<b>56.5</b>	3.2	0.6	3.1	<b>69.5</b>	1.8	18.1	0.1	1.3	<b>21.3</b>	5.7	3.4	<b>34.3</b>
<b>R3</b>	Deyang	143	2.7	4.6	<b>58.0</b>	14.5	0.8	<b>80.6</b>	0.8	9.4	0.1	1.6	<b>11.9</b>	4.2	3.4	<b>34.5</b>
	Leshan	125	2.9	3.1	<b>58.7</b>	15.1	0.6	<b>80.4</b>	1.2	8.2	0.1	1.2	<b>10.7</b>	5.7	3.0	<b>32.4</b>
	Luzhou	149	13.9	4.8	<b>53.9</b>	1.8	1.2	<b>75.6</b>	3.8	11.2	0.1	1.0	<b>16.1</b>	5.4	2.9	<b>37.8</b>
	Meishan	153	2.5	3.1	<b>40.2</b>	36.3	0.6	<b>82.7</b>	0.9	7.7	0.1	1.1	<b>9.8</b>	4.6	2.9	<b>52.3</b>
	Mianyang	114	2.7	6.8	<b>60.3</b>	4.8	1.1	<b>75.7</b>	0.9	12.4	0.1	2.2	<b>15.6</b>	4.9	3.8	<b>31.0</b>
	Neijiang	140	11.3	7.6	<b>51.9</b>	2.8	1.3	<b>74.9</b>	2.4	12.7	0.1	1.1	<b>16.3</b>	5.6	3.1	<b>39.3</b>
	Suining	100	11.7	33.5	<b>14.4</b>	1.1	3.2	<b>63.9</b>	2.6	21.4	0.1	1.7	<b>25.8</b>	6.8	3.6	<b>75.3</b>
	Ya'an	79	3.2	3.6	<b>45.9</b>	20.1	0.7	<b>73.5</b>	1.4	11.0	0.3	2.3	<b>15.0</b>	7.6	3.8	<b>42.6</b>
	Yibin	134	6.8	4.1	<b>60.1</b>	5.5	0.9	<b>77.4</b>	2.8	9.7	0.1	1.1	<b>13.7</b>	6.0	3.0	<b>31.0</b>
	Zigong	145	8.9	6.0	<b>57.1</b>	3.2	1.1	<b>76.3</b>	2.4	11.4	0.1	1.1	<b>15.0</b>	5.5	3.3	<b>34.2</b>
Ziyang	131	5.8	7.3	<b>54.0</b>	7.1	1.2	<b>75.4</b>	1.7	12.6	0.1	1.4	<b>15.8</b>	5.6	3.2	<b>37.2</b>	
<b>R4</b>	Chengdu	144	2.2	3.5	20.4	<b>55.2</b>	0.6	<b>81.9</b>	0.8	8.2	0.1	1.4	<b>10.5</b>	4.2	3.5	<b>37.2</b>

499 \* Others include initial and boundary conditions, windblown dust, and sea salt.

500 <sup>#</sup> Non-local=Within SCB + Outside SCB – Local.

501 <sup>§</sup> the city center of Chongqing.



502 Table 2. Predicted citywide area-weighted average PM<sub>2.5</sub> concentrations and source-region contributions in the 18 SCB cities in the  
 503 winter and summer.

Region ID	Number of grid cells	Total area (km <sup>2</sup> )	City	Number of grid cells	Total area (km <sup>2</sup> )	Winter			Summer		
						PM <sub>2.5</sub> (µg m <sup>-3</sup> )	Contributions (%)		PM <sub>2.5</sub> (µg m <sup>-3</sup> )	Contributions (%)	
							Local	Non-local <sup>#</sup>		Local	Non-local <sup>#</sup>
<b>R1</b>	248	28011	Western Chongqing	248	28011	<b>99</b>	37.8	52.1	<b>27</b>	36.8	36.0
<b>R2</b>	543	56265	Bazhong	106	10734	<b>51</b>	27.9	58.2	<b>14</b>	19.3	43.6
			Dazhou	139	14689	<b>64</b>	29.1	58.6	<b>16</b>	21.6	45.7
			Guangan	60	5618	<b>100</b>	40.6	49.5	<b>23</b>	32.3	40.3
			Guangyuan	133	14182	<b>50</b>	28.3	57.1	<b>13</b>	23.4	40.5
			Nanchong	105	11042	<b>89</b>	47.9	41.4	<b>20</b>	36.3	34.1
<b>R3</b>	998	98185	Deyang	59	5346	<b>101</b>	49.9	40.2	<b>25</b>	44.5	32.8
			Leshan	107	1159	<b>81</b>	42.5	44.4	<b>14</b>	35.0	29.7
			Luzhou	112	1075	<b>92</b>	35.3	53.8	<b>20</b>	25.6	43.7
			Meishan	74	6570	<b>120</b>	42.6	47.7	<b>26</b>	36.7	35.8
			Mianyang	171	1801	<b>59</b>	36.8	49.2	<b>15</b>	32.9	36.7
			Neijiang	58	4859	<b>122</b>	45.8	44.8	<b>29</b>	40.8	34.7
			Suining	50	4743	<b>104</b>	30.3	59.7	<b>23</b>	25.4	48.4
			Ya'an	131	1349	<b>45</b>	34.5	49.2	<b>6</b>	27.9	35.1
			Yibin	117	1182	<b>101</b>	50.1	39.1	<b>21</b>	43.3	24.9
Zigong	45	3935	<b>126</b>	51.3	39.4	<b>30</b>	47.3	26.9			
Ziyang	74	7042	<b>115</b>	40.3	50.2	<b>25</b>	32.6	41.9			
<b>R4</b>	105	1075	Chengdu	105	1075	<b>110</b>	46.5	44.1	<b>31</b>	44.6	33.0
<b>R5</b>	390	44371	Eastern Chongqing	390	44371	<b>54</b>	21.0	66.0	<b>15</b>	14.7	52.1

504 <sup>#</sup> Non-local = 100% - Local - SOA - Others; Others include initial and boundary conditions, windblown dust, and sea salt.

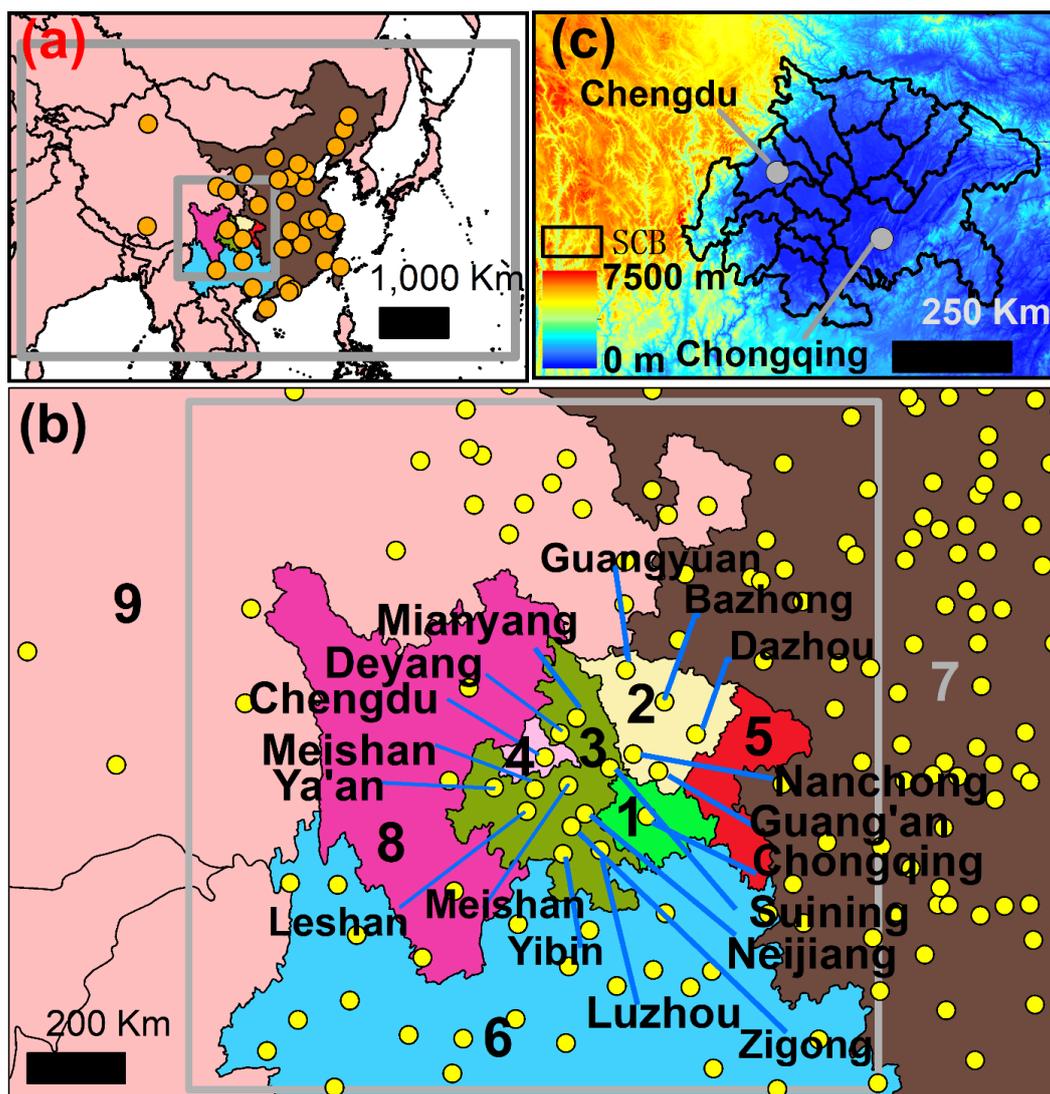


505 Table 3. Predicted maximum daily contribution from a given region (MDCs) in the SCB city center and the corresponding PM<sub>2.5</sub>  
 506 concentrations in the city center on the same day. Only winter data are included in this table. The units are  $\mu\text{g m}^{-3}$ . The numbers in  
 507 the bold present the contributions due to local emissions or that from R7.

Region ID	Cities	MDCs (total PM <sub>2.5</sub> concentrations)										
		Within SCB					Outside SCB					SOA
R1	R2	R3	R4	R5	R6	R7	R8	R9				
<b>R1</b>	Chongqing <sup>§</sup>	<b>291 (353)</b>	54 (414)	48 (294)	16 (302)	15 (143)	22 (290)	<b>99 (200)</b>	1 (182)	8 (212)	12 (160)	12 (294)
	Bazhong	8 (160)	<b>83 (139)</b>	23 (160)	18 (160)	7 (64)	14 (68)	<b>73 (109)</b>	0 (140)	6 (90)	14 (327)	11 (34)
	Dazhou	46 (219)	<b>123 (216)</b>	16 (171)	6 (171)	34 (219)	12 (77)	<b>79 (190)</b>	0 (173)	5 (54)	18 (367)	8 (54)
<b>R2</b>	Guangan	72 (159)	<b>129 (205)</b>	31 (180)	10 (180)	22 (225)	14 (102)	<b>76 (165)</b>	0 (234)	5 (129)	16 (219)	7 (75)
	Guangyuan	11 (122)	<b>60 (107)</b>	35 (122)	15 (122)	4 (56)	10 (86)	<b>62 (100)</b>	0 (91)	11 (46)	16 (226)	10 (25)
	Nanchong	58 (180)	<b>152 (242)</b>	50 (200)	12 (243)	18 (238)	13 (141)	<b>78 (166)</b>	0 (177)	5 (136)	17 (178)	10 (71)
	Deyang	32 (226)	30 (257)	<b>170 (296)</b>	147 (288)	4 (257)	14 (215)	<b>70 (132)</b>	0 (215)	9 (137)	13 (91)	10 (67)
	Leshan	15 (270)	15 (166)	<b>163 (270)</b>	57 (222)	4 (166)	10 (222)	<b>47 (122)</b>	1 (108)	5 (118)	15 (222)	9 (105)
	Luzhou	115 (245)	33 (207)	<b>211 (261)</b>	21 (240)	10 (207)	18 (259)	<b>71 (272)</b>	1 (192)	6 (90)	17 (307)	10 (182)
	Meishan	17 (189)	22 (379)	<b>155 (379)</b>	138 (235)	4 (176)	11 (263)	<b>54 (263)</b>	1 (379)	5 (80)	15 (164)	9 (100)
	Mianyang	24 (209)	30 (170)	<b>219 (294)</b>	67 (305)	6 (168)	11 (144)	<b>70 (112)</b>	0 (337)	9 (86)	16 (294)	11 (46)
<b>R3</b>	Neijiang	93 (214)	39 (168)	<b>161 (205)</b>	49 (309)	6 (168)	17 (165)	<b>79 (198)</b>	1 (192)	5 (198)	19 (181)	9 (164)
	Suining	75 (183)	110 (242)	<b>105 (186)</b>	25 (225)	14 (210)	17 (151)	<b>83 (167)</b>	0 (154)	5 (79)	17 (198)	10 (41)
	Ya'an	14 (126)	17 (205)	<b>79 (170)</b>	88 (234)	3 (205)	8 (170)	<b>35 (205)</b>	1 (123)	6 (40)	15 (189)	9 (43)
	Yibin	74 (283)	21 (162)	<b>160 (217)</b>	45 (164)	5 (236)	16 (188)	<b>58 (166)</b>	1 (122)	4 (147)	14 (170)	9 (127)
	Zigong	64 (281)	34 (171)	<b>162 (214)</b>	51 (291)	6 (171)	17 (192)	<b>71 (195)</b>	1 (195)	5 (199)	16 (283)	10 (162)
	Ziyang	39 (192)	33 (137)	<b>164 (298)</b>	63 (176)	7 (120)	19 (181)	<b>73 (155)</b>	1 (298)	6 (176)	18 (203)	9 (126)
<b>R4</b>	Chengdu	16 (232)	23 (225)	86 (267)	<b>250 (327)</b>	4 (139)	11 (267)	<b>63 (151)</b>	1 (166)	7 (132)	16 (228)	9 (101)

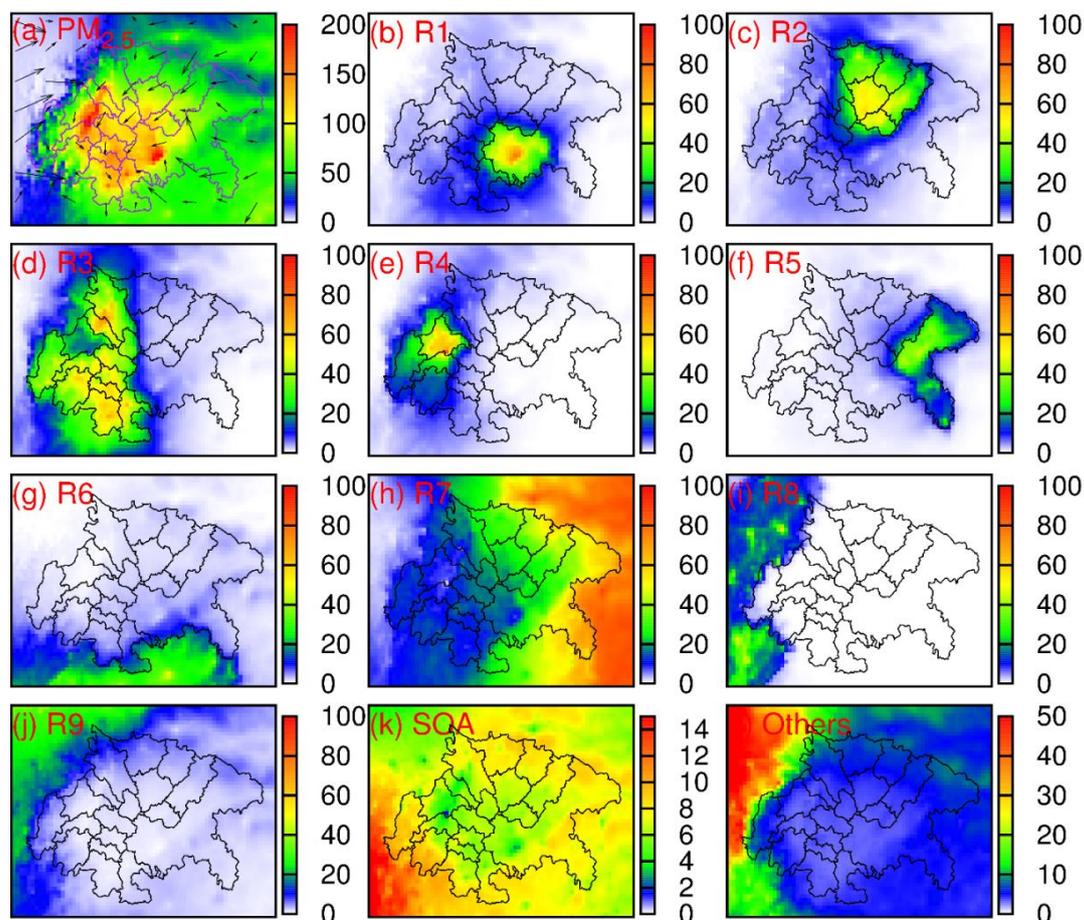
508 <sup>#</sup> Others include initial and boundary conditions, windblown dust, and sea salt.

509 <sup>§</sup> includes the city center of Chongqing.

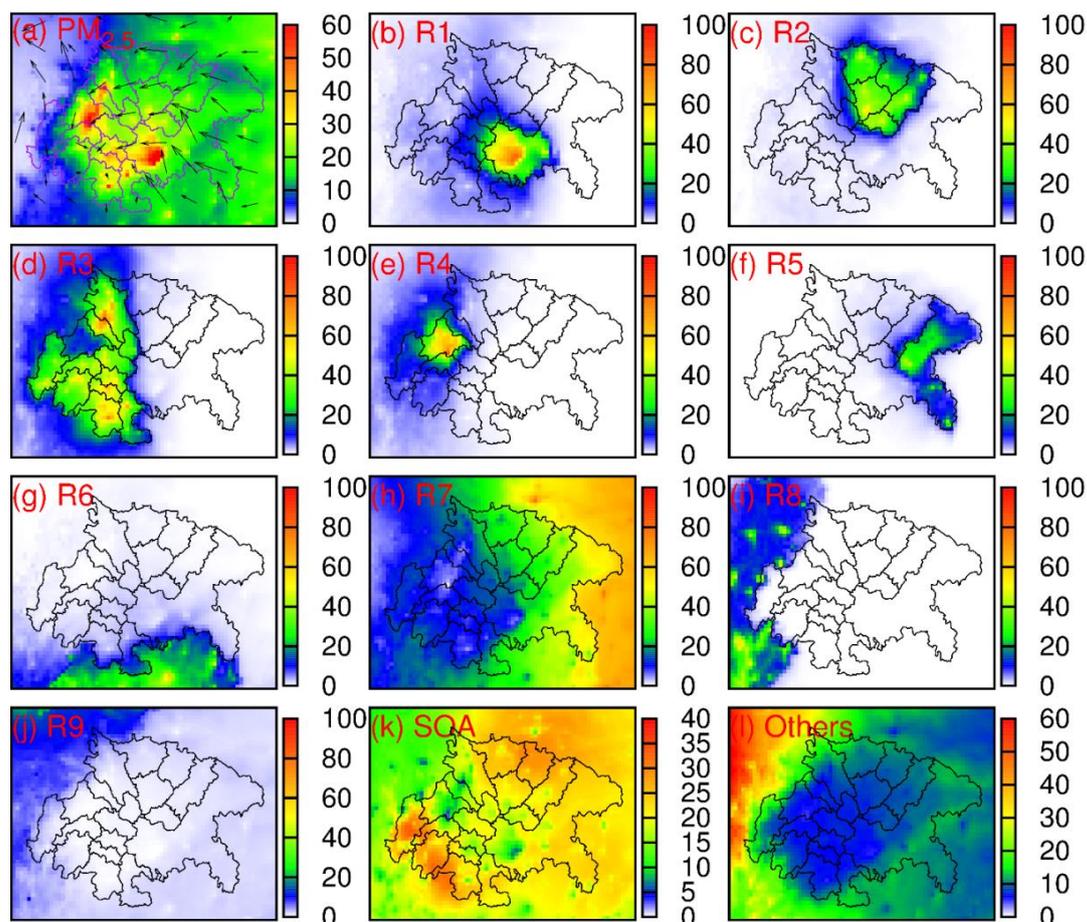


510

511 Figure 1. (a) Locations of the 12-km and 36-km domains (grey rectangles) and the locations of  
 512 provincial capitals and municipalities (orange circles), (b) locations of region categories 1~9 and  
 513 the prefecture-level cities (yellow circles), and (c) terrain within and surrounding the 18 cities of  
 514 the SCB (black line). Regions 1~5 are the cities within the SCB. Regions 1, 4, and 5 are western  
 515 Chongqing, Chengdu, and eastern Chongqing, respectively. The city center of Chongqing is  
 516 located in western Chongqing. Region 8 is the area of Sichuan Province excluding those cities in  
 517 the SCB.



518  
519 Figure 2. (a) Spatial distributions of predicted PM<sub>2.5</sub> concentrations ( $\mu\text{g m}^{-3}$ ) and (b-l) source-  
520 region contributions to PM<sub>2.5</sub> (%) in the winter. Others include IC, BC, windblown dust, and sea  
521 salt. Black arrows in (a) are wind vectors.

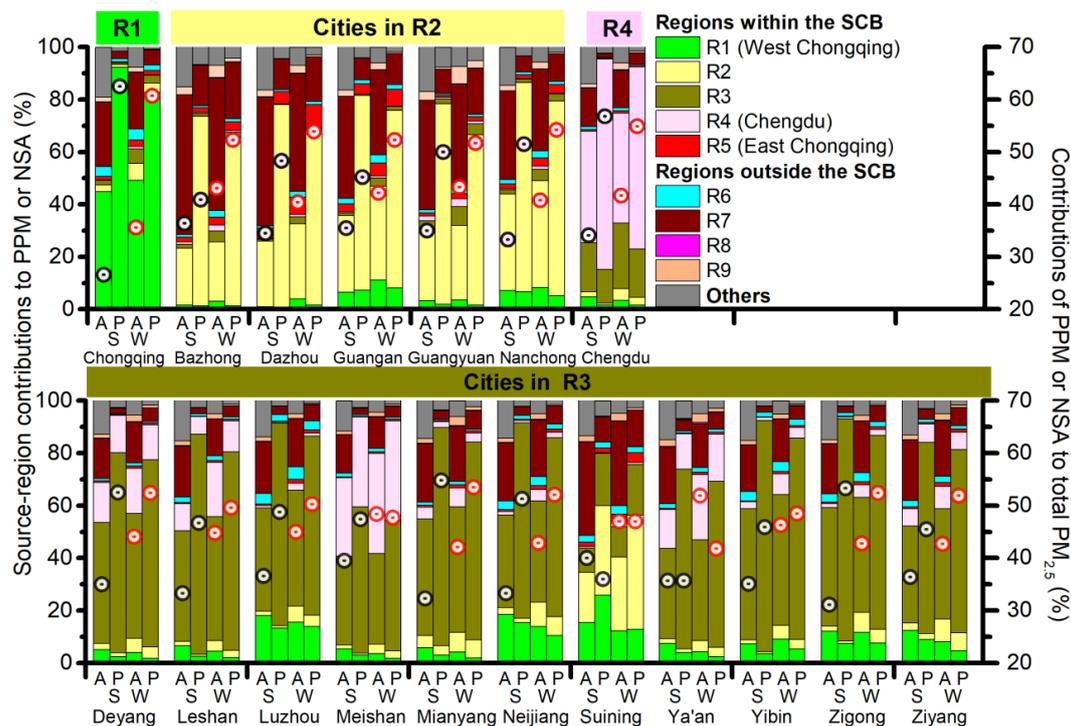


522

523 Figure 3. (a) Spatial distributions of predicted  $PM_{2.5}$  concentrations ( $\mu g m^{-3}$ ) and (b-l) the  
524 source-region contributions to  $PM_{2.5}$  (%) in the summer. Others include IC, BC, windblown dust,  
525 and sea salt.



526



527

528 Figure 4. Predicted source-region contributions to SIA (A) and PPM (P) (bars, left y-axis) and  
 529 the predicted proportions of SIA and PPM in  $PM_{2.5}$  (circles, right y-axis) for the 18 city centers  
 530 of the SCB in the summer (S) and winter (W). Others include IC, BC, windblown dust, and sea  
 531 salt.