

1 **Local and regional contributions to fine particulate matter in the 18 cities of**  
2 **Sichuan Basin, southwestern China**

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4 Xue Qiao<sup>1,2</sup>, Hao Guo<sup>2</sup>, Ya Tang<sup>3</sup>, Pengfei Wang<sup>2</sup>, Wenye Deng<sup>4</sup>, Xing Zhao<sup>5</sup>, Jianlin Hu<sup>6</sup>, Qi  
5 Ying<sup>7,\*</sup>, Hongliang Zhang<sup>2,\*</sup>

6  
7 <sup>1</sup>Institute of New Energy and Low-carbon Technology & Healthy Food Evaluation Research  
8 Center, Sichuan University, No. 24, South Section One, First Ring Road, Chengdu, Sichuan  
9 610065, China

10 <sup>2</sup>Department of Civil and Environmental Engineering, Louisiana State University, Baton Rouge  
11 LA 70803, USA

12 <sup>3</sup>College of Architecture and Environment & Healthy Food Evaluation Research Center, Sichuan  
13 University, Chengdu 610065, China

14 <sup>4</sup>Xinjiang Academy of Environmental Protection Science, Urumqi 830011, China

15 <sup>5</sup>Department of Epidemiology and Biostatistics, West China School of Public Health, Sichuan  
16 University, No. 17, Section 3, South Renmin Road, Chengdu, Sichuan 610041, China

17 <sup>6</sup>Jiangsu Key Laboratory of Atmospheric Environment Monitoring and Pollution Control,  
18 Jiangsu Engineering Technology Research Center of Environmental Cleaning Materials,  
19 Nanjing University of Information Science & Technology, Nanjing 210044, China

20 <sup>7</sup>Zachry Department of Civil Engineering, Texas A&M University, College Station, TX 77843,  
21 USA

22 \*Corresponding authors: Qi Ying, [qying@civil.tamu.edu](mailto:qying@civil.tamu.edu); Hongliang Zhang, [hlzhang@lsu.edu](mailto:hlzhang@lsu.edu).

23 **Abstract**

24 The Sichuan Basin (SCB) is one of the regions suffering from severe air pollution in China, but  
25 fewer studies have been conducted for this region than for the more developed regions in East and  
26 North China. In this study, a source-oriented version of the Community Multi-scale Air Quality  
27 (CMAQ) model was used to quantify contributions from nine regions to PM<sub>2.5</sub> (i.e., particulate  
28 matter (PM) with an aerodynamic diameter less than 2.5 μm) and its components in the 18 cities  
29 within the SCB in the winter (December 2014 to February 2015) and summer (June to August,  
30 2015). In the winter, citywide average PM<sub>2.5</sub> concentrations are 45~126 μg m<sup>-3</sup>, with 21~51% and  
31 39~66% due to local and non-local emissions, respectively. In the summer, 15~45% and 25~52%  
32 of citywide average PM<sub>2.5</sub> (14~31 μg m<sup>-3</sup>) are due to local and non-local emissions, respectively.  
33 Compared to primary PM (PPM), the inter-region transport of secondary inorganic aerosols (SIA,  
34 including ammonia, nitrate, and sulfate ions (NH<sub>4</sub><sup>+</sup>, NO<sub>3</sub><sup>-</sup>, and SO<sub>4</sub><sup>2-</sup>, respectively)) and their gas-  
35 phase precursors are greater. The region to the east of SCB (R7, including the central and eastern  
36 China and others) is the largest contributor outside the SCB, and it can contribute approximately  
37 80% of PM<sub>2.5</sub> in the eastern, northeastern, and southeastern rims of the SCB, but only 10% in other  
38 SCB regions in both seasons. Under favorable transport conditions, regional transport of air  
39 pollutants from R7 could account for up to 35~100 μg m<sup>-3</sup> of PM<sub>2.5</sub> in each of the SCB cities in  
40 the winter. This study demonstrates that it is important to have joint emission control efforts among  
41 cities within the SCB and regions to the east in order to reduce PM<sub>2.5</sub> concentrations and prevent  
42 high PM<sub>2.5</sub> days for the entire basin.

43 **Keywords:** Sichuan Basin, local emission, regional transport, PM<sub>2.5</sub>, source apportionment

## 44 1. Introduction

45 Particulate matter (PM) is one of the major air pollutants in China, including primary and  
46 secondary components. Primary PM (PPM) is directly released from emission sources, while  
47 secondary PM is formed from their precursors, such as sulfur dioxides (SO<sub>2</sub>), nitrogen oxides  
48 (NO<sub>x</sub>), and ammonia (NH<sub>3</sub>). All of them are released from local sources or transported for a long  
49 distance (Ying et al., 2014; Zhao et al., 2018). The relative contributions of secondary components  
50 to total PM<sub>2.5</sub> (PM with an aerodynamic diameter less than 2.5 μm) usually increases as PM<sub>2.5</sub>  
51 concentration elevates in megacities (Huang et al., 2014; Qiao et al., 2018).

52 Air pollution in major economic centers in China, including the North China Plain (NCP),  
53 Yangtze River Delta (YRD), and Pearl River Delta (PRD), has been extensively studied in recent  
54 years. Regional transport of air pollutants has been identified an important source of PM in the  
55 three regions, particularly the precursors of secondary PM (Zhang et al., 2013; Zhao et al., 2013a;  
56 Ying et al., 2014; Jiang et al., 2015; Li et al., 2015; Wang et al., 2015; Zheng et al., 2015; Tang et  
57 al., 2016; Yang et al., 2018). Several urbanized areas in western China also have been suffering  
58 from air pollution due to rapid industrial and urban development, but fewer studies have been  
59 conducted compared with the NCP, YRD, and PRD. One such area is the Sichuan Basin (SCB),  
60 which covers an area about 0.22 million km<sup>2</sup> and is home to more than 100 million residents in 18  
61 cities, among which Chengdu and Chongqing are the largest two cities in western China (National  
62 Bureau of Statistics of China (NBSC), 2015; Table S1). The SCB is topographically isolated, with  
63 mountains or plateaus on all sides. They are the Qinghai-Tibetan Plateau (QTP), Yunnan-Guizhou  
64 Plateau (YGP), Wushan Mountains (WUM), and Dabashan Mountains (DBM) to the west, south,  
65 east, and north of the SCB, respectively. As a result of the basin topography, emissions released  
66 from the SCB tend to accumulate in the basin, causing severe air pollution (Ning et al., 2018a;  
67 Zhao et al., 2018). In addition, east and central China, which are to the east of the SCB, have  
68 considerable contributions to PM<sub>2.5</sub> for the SCB. For example, Ying (2014) predicted that central  
69 and east China had a combined contribution of 29.6% to the total mass of NO<sub>3</sub><sup>-</sup> and SO<sub>4</sub><sup>2-</sup> for  
70 Chongqing in January 2009. Due to high emissions within the basin and deep basin landform,  
71 annual average concentrations of PM<sub>2.5</sub> in the SCB were similar to that of NCP and Central China  
72 (Figure S1). Annual PM<sub>2.5</sub> measured in Chengdu and Chongqing in 2015 were 64 and 57 μg m<sup>-3</sup>,  
73 respectively, about six times the World Health Organization (WHO) guideline (10 μg m<sup>-3</sup>) (WHO,  
74 2006; NBSC, 2015).

75 To design effective PM<sub>2.5</sub> control strategies for the SCB, it is necessary to quantify the source  
76 contributions and inter-/intra-region transport of PM<sub>2.5</sub> and its precursors within the SCB and its  
77 surrounding regions. There are many types of source apportionment methods, such as receptor-  
78 based models, air parcel trajectory models, remote sensing, and chemical transport models (CTMs).  
79 Receptor-based models, such as the Positive Matrix Factorization (PMF) (Paatero and Tapper,  
80 1994; Qiu et al., 2019), the Chemical Mass Balance (CMB) (Watson et al., 1990) and a local  
81 contribution model proposed by Zhao et. al. (2019), are semi-quantitative and cannot quantitatively  
82 determine the source contributions from an exact emission sector or a specific location. They also

83 require a large number of monitoring data and can only resolve source contributions at the  
84 monitoring sites (Hopke, 2016). Remote sensing and air parcel trajectory models, such as the  
85 Potential Source Contribution Function (PSCF) and the Hybrid Single Particle Lagrangian  
86 Integrated Trajectory Model (HYSPLIT) can just reflect the atmospheric dynamics so they are not  
87 quantitative for source apportionment of secondary species (Begum et al., 2005; Uno et al., 2009;  
88 Stein et al., 2015; Liu et al., 2018; Wu et al., 2018). Compared to above methods, CTMs are more  
89 quantitative, as they can be track the source contributions to both primary and secondary air  
90 pollutants from a specific region or sectorial source for studies at the local, regional, or global  
91 scales (Bove et al., 2014; Kim et al., 2015; Lelieveld et al., 2015; Itahashi et al., 2017; Shi et al.,  
92 2017). In CTMs, source contributions are quantified through two methods, namely, sensitivity  
93 analysis and tagged-tracer methods (Burr and Zhang, 2011). Sensitivity analysis such as the brute-  
94 force method is more suitable to estimate air quality changes due to emission perturbations, as  
95 emissions from certain sources would be eliminated or reduced in each simulation of sensitivity  
96 analysis (Burr and Zhang, 2011; Han et al., 2018; Huang et al., 2018). In the tagged-tracer method,  
97 source-tagged-species are used to track air pollutants from specific emission regions and sectors  
98 and they would go through all the non-linear chemical and physical processes in the model, thus  
99 this method is considered to provide more realistic evaluations of the contributions of different  
100 source sectors or source regions to the current level of air pollution under the current emission  
101 intensities (Wang et al., 2009; Burr and Zhang, 2011; Chen et al., 2017).

102 Transport of PM<sub>2.5</sub> and its precursors has been studied for Chengdu, Chongqing, and the entire  
103 SCB region (Zhu et al., 2018). Based on in-situ observations and the HYSPLIT model, studies  
104 have found that dust storms from northwestern China and biomass burning activities would cause  
105 high PM days in the two cities, particularly in spring (Zhao et al., 2010; Tao et al., 2013; Chen and  
106 Xie, 2014; Chen et al., 2015). Using the PSCF model, a study reported that the main potential  
107 sources of PM<sub>2.5</sub> for Chengdu were from southeastern cities and the western margin of the SCB in  
108 addition to local emissions from December 2013 to February 2014 (Liao et al., 2017). Based on  
109 the HYSPLIT and PSCF analyses, air pollution was determined mainly from the south during  
110 persistent extreme haze days in Chengdu from 6<sup>th</sup> to 16<sup>th</sup> January 2015 (Li et al., 2017). However,  
111 the aforementioned studies based on the HYSPLIT and PSCF models are not quantitative in terms  
112 of emission contributions. Their accuracy is also limited by the meteorological inputs to drive these  
113 models, which are often in very coarse resolutions (0.5 to 1.0 degree) that are not enough to  
114 accurately predict air pollutant transport within the SCB, as meteorological conditions and  
115 pollutant concentrations may vary greatly within short distance (Shi et al., 2017). Also, previous  
116 country-level modeling study did not have sufficient spatial resolution to properly quantify the  
117 transport among cities within the SCB.

118 Since both inter-regional transport within the cities in the basin and from outside the region  
119 can greatly affect PM<sub>2.5</sub> concentrations in the 18 SCB cities, systematically quantifying  
120 contributions from different regions to PM<sub>2.5</sub> for all the 18 cities in the SCB is urgently needed as  
121 emission controls are further tightened to improve air quality in this region. In this study, an

122 improved source-oriented community multi-scale air quality (CMAQ) model was used to quantify  
123 the contributions from nine regions (five within the SCB and four outside) to PM<sub>2.5</sub> and its  
124 components for the 18 SCB cities. The assumption is that the transport of air pollutants is evident  
125 among the SCB cities and some cities in the rims of the SCB are greatly affected by emissions  
126 outside SCB. Therefore, the objectives of this study are to quantitatively determine (1) the inter-  
127 region transport of air pollutants emitted in the SCB and its contributions to PM<sub>2.5</sub> in the 18 SCB  
128 cities and (2) the contributions of emissions outside the basin to PM<sub>2.5</sub> in the SCB. In this study,  
129 the percentage contributions and maximum mass contributions from each region to PM<sub>2.5</sub> in each  
130 city are both presented to better understand the extent of air pollutant transport. We modeled PM<sub>2.5</sub>  
131 and its source contributions only for two seasons, as PM<sub>2.5</sub> concentrations in the SCB are highest  
132 in winter and summer, respectively (Ning et al., 2018a).

## 133 **2. Methods and materials**

### 134 **2.1. Model description**

135 The source-oriented CMAQ model used in this study is based on the CMAQ model version  
136 5.0.1. The gas phase and aerosol mechanisms are extended from the standard SAPRC-99  
137 photochemical mechanism and aerosol module version 6 (AERO6). This version of the source-  
138 oriented CMAQ is capable of simultaneously tracking both primary particulate matter (PPM) and  
139 secondary inorganic aerosols (SIA, including NH<sub>4</sub><sup>+</sup>, NO<sub>3</sub><sup>-</sup>, and SO<sub>4</sub><sup>2-</sup>) from multiple source sectors  
140 and regions. It unifies the two previous models individually developed for PPM (Hu et al., 2015)  
141 and SIA (Ying et al., 2014; Shi et al., 2017) into a single consistent model framework. For SIA,  
142 multiple source-tagged reactive species are introduced in both gas and particle phases to represent  
143 the same species originated from different source sectors or regions. The corresponding  
144 photochemical mechanisms, aerosol and cloud modules are expanded so that SIA and their  
145 precursors from different regions can be tracked separately throughout the model calculations. For  
146 example, NO<sub>2</sub>\_S1 and NH<sub>3</sub>\_S2 can be used to represent NO<sub>2</sub> from region 1 and NH<sub>3</sub> from region  
147 2, respectively. After the photochemical mechanism is expanded, the source-tagged species are  
148 allowed to go through all processed to form (NH<sub>4</sub>\_S2)(NO<sub>3</sub>\_S1) based on additional reactions of  
149 NO<sub>2</sub> + OH → HNO<sub>3</sub> and NH<sub>3</sub>+ HNO<sub>3</sub> →NH<sub>4</sub>NO<sub>3</sub>. Thus, the contributions of region 1 to NO<sub>3</sub><sup>-</sup>  
150 and region 2 to NH<sub>4</sub><sup>+</sup> are quantified. For PPM, source-tagged non-reactive tracers are added to  
151 track the total amount of PPM emitted from different source sectors and regions. SOA is included  
152 in the current model but its source contributions are not resolved.

### 153 **2.2. Model application**

154 The source-oriented CMAQ model was applied to quantify nine source-region contributions  
155 to PM<sub>2.5</sub> and its components (PPM and SIA) for the 18 cities in the winter (from December 2014  
156 to February 2015) and summer (June to August 2015) using nested domain settings. The locations  
157 of domains, nine source-regions, and the 18 cities of the SCB are shown in Figure 1. The horizontal  
158 resolutions of the parent and nested domains are 36-km and 12-km, respectively. There are 18

159 vertical layers with an overall height of 20 km and the layer closest to the land surface is up to 35  
160 m. The geographical regions of emissions are classified into nine source-regions. As Chengdu and  
161 Chongqing are the two largest cities in western China and within the SCB, we classified Chengdu,  
162 eastern Chongqing, and western Chongqing into three individual regions (R4, R5, and R1,  
163 respectively). Western Chongqing is well urbanized and eastern Chongqing is mostly rural areas.  
164 The five cities in the northeastern SCB (Bazhong, Dazhou, Guangyuan, Guang'an, and Nanchong)  
165 are grouped into R2, as they have relatively lower anthropogenic emission densities compared to  
166 most of the other SCB cities and they are located in the upwind areas within the SCB (Qiao et al.,  
167 2019). The rest SCB cities are grouped into R3. Sichuan Province excluding those cities within  
168 the SCB is R8, most of which remote rural areas. R6 includes three provinces to the south of the  
169 SCB and R7 has the Chinese provinces to the east and northeast of the SCB. R9 includes the other  
170 jurisdictions to the west of the SCB, including Xinjiang, Qinghai, Gansu, Tibet, and other countries.

171 Meteorological inputs were generated using the Weather Research and Forecasting (WRF)  
172 model version 3.9 based on the 6-hourly FNL (Final) Operational Global Analysis data from the  
173 National Center for Atmospheric Research (NCAR) with a spatial resolution of  $1.0 \times 1.0^\circ$   
174 (<http://dss.ucar.edu/datasets/ds083.2/>). The anthropogenic emission inventory used was the  
175 Emission Database for Global Atmospheric Research (EDGAR) version 4.3.1 for the year of 2012  
176 (Crippa et al., 2018). The inventory was directly used for the model year of 2014-2015 as no  
177 reliable sources for emission changes in the SCB are available. The monthly EDGAR inventories  
178 have a spatial resolution of  $0.1^\circ \times 0.1^\circ$  ( $\sim 10 \text{ km} \times 10 \text{ km}$ ) and were re-projected to the model domains  
179 using the Spatial Allocator (<https://www.emascenter.org/sa-tools/>). Temporal profiles specific to  
180 sources were used to allocate the monthly emission rates to hourly values for CMAQ modeling  
181 (Olivier et al., 2003; Streets et al., 2003; Wang et al., 2010). The EDGAR inventory includes  
182 carbon monoxide (CO),  $\text{NO}_x$ ,  $\text{SO}_2$ ,  $\text{NH}_3$ , non-methane volatile organic compounds (NMVOCs),  
183  $\text{PM}_{2.5}$ ,  $\text{PM}_{10}$  (PM with an aerodynamic diameter less than  $10 \mu\text{m}$ ), elemental carbon (EC), and  
184 organic carbon (OC) from various sources. Emission sources are grouped into six categories:  
185 energy, industries, residential activities, on-road transportation, off-road transportation, and  
186 agriculture. In addition to these six anthropogenic sources, contributions of biogenic sources were  
187 also determined using emissions generated by the Emissions of Gases and Aerosols from Nature  
188 (MEGAN) model version 2.1 (Guenther et al., 2012). The emissions from open burning were  
189 estimated based on the Fire Inventory from the National Center for Atmospheric Research (NCAR  
190 FINN) (Wiedinmyer et al., 2010). Contributions of windblown dust and sea salt emissions were  
191 determined based on in-line generated emissions during CMAQ simulations. It should be noted  
192 that the uncertainties in emission inventories potentially lead to uncertainties in the contributions.

193 The initial and boundary conditions (ICs and BCs, respectively) for the 36-km domain were  
194 based on CMAQ default profiles, and those for the 12-km domain were generated using the CMAQ  
195 outputs from the 36-km simulations. More details about the setup and configurations of the  
196 WRF/CMAQ modeling system can be found in a previous publication for China (Kang et al.,  
197 2016).

### 198 **3. Results and discussion**

### 199 3.1. Model performance

200 The model performance on meteorological parameters and 24-hr PM<sub>2.5</sub> in the 12-km domain  
201 for the two seasons has been evaluated in a companion paper (Qiao et al., 2019) and is briefly  
202 summarized here (Figure S2). As the predictions on wind speed (WS) and wind direction (WD)  
203 are important in modeling air pollutant transport (Zhao et al., 2009), the WRF model performance  
204 on WS, WD, ambient air temperature (T), and relative humidity (RH) were evaluated by using  
205 hourly observations at China's national meteorological stations and the observation data that were  
206 downloaded from the National Climate Data Center (NCDC;  
207 ftp://ftp.ncdc.noaa.gov/pub/data/noaa/, last accessed on June 20, 2018). The mean biases (MBs)  
208 of predicted RH (-10.8% to -1.1%) and T (-0.9 to -0.1°C) in each month are comparable to other  
209 studies in China (Wang et al., 2010; Zhao et al., 2013b; Wang et al., 2013). The MB of WD in  
210 each month (-5° to 6°) meet the benchmark of  $<\pm 10^\circ$  suggested by Emery et al. (2001). Although  
211 the MB of WS in each month (0.5 to 1.1 m s<sup>-1</sup>) does not meet the benchmark of  $<\pm 0.5$  m s<sup>-1</sup>, the  
212 gross errors (GEs: 1.4-1.9 m s<sup>-1</sup>) are within the benchmark of 2.0 m s<sup>-1</sup>. For 24-hr PM<sub>2.5</sub>  
213 concentrations, the statistical metrics of model performance are generally within the criteria  
214 recommended by Emery et al. (2017) for regulatory applications, with only a few cities exceeding  
215 the normalized mean bias (NMB) criteria of  $<\pm 30\%$  in the winter (Chongqing 42%; Guangyuan  
216 41%; Mianyang 37%; Meishan 31%; Ziyang 48%) and in the summer (Dazhou -39%) (Figure S2).  
217 The 24-hr PM<sub>2.5</sub> predictions meet the goals of normalized mean error (NME $<\pm 35\%$ ), fractional  
218 bias (FB $<\pm 30\%$ ), and fractional error (FE $<\pm 50\%$ ) in all the cities in both seasons, except for the  
219 NME of Ziyang (58%) in the winter. The predictions of major PM<sub>2.5</sub> components (including OC,  
220 EC, NH<sub>4</sub><sup>+</sup>, NO<sub>3</sub><sup>-</sup>, and SO<sub>4</sub><sup>2-</sup>) in Chengdu and Chongqing are comparable with observations, and  
221 both predictions and observations suggest that OC and SIA are the largest contributors to PM<sub>2.5</sub> in  
222 summer and winter, with combined contributions about 70% (Qiao et al., 2019).

223

### 224 3.2. Seasonal average contributions

#### 225 3.2.1. Source contributions at the city centers

226 In each city, there are 4 to 17 national air quality stations (NAQs), and almost all the NAQs  
227 are located in the urban areas, where population densities are higher. Thus, coordinates of the  
228 NAQs in the urban area of a given city were averaged to define the city center in order to  
229 understand PM<sub>2.5</sub> concentrations and its sources for the most-populated region of each city. The  
230 predicted PM<sub>2.5</sub> concentrations and source-region contributions at the 18 SCB city centers are  
231 presented in Table 1 for winter and Table S2 for summer. In all the city centers, the predicted  
232 PM<sub>2.5</sub> concentrations are much higher in the winter (60~191 µg m<sup>-3</sup>) than in the summer (14~64  
233 µg m<sup>-3</sup>). The city centers are considerably affected by both local and regional emissions in both  
234 seasons. Emissions within the SCB are the major contributor to PM<sub>2.5</sub> in Chengdu and Chongqing  
235 in both seasons (~80%) and emissions outside the SCB contribute approximately 7~15%. Among

236 the regions within the SCB, local emissions (i.e., emissions from the region where the city center  
237 is located) are the largest contributor to PM<sub>2.5</sub> in Chongqing and Chengdu in both seasons (about  
238 70% and 58%, respectively). However, emissions from R3 (i.e., the 11 cities in the northwestern,  
239 western, and southwestern SCB) also have considerable contributions in Chengdu (~20% and 14%  
240 in the winter and summer, respectively). For the R3 cities, the contributions of emissions within  
241 the SCB (64~83%) are also larger than that from outside the SCB (8~26%) in both seasons. Local  
242 emissions are the largest contributor for R3 cities (40~60%), except that Suining has only ~13%  
243 due to its local region. The low local contribution in Suining might be because it is less  
244 economically developed compared to other cities, except for Bazhong, Guangyuan, and Ya'an, as  
245 suggested by the 2015 gross domestic production (GDP; Table S1). For the five cities in the  
246 northern SCB (R2), emissions within the SCB account for 40~70% in both seasons, including  
247 37~57% from local emissions. Emissions outside the SCB also have large contributions to the R2  
248 cities (21~36% and 17~28% in the winter and summer, respectively), as R2 is located in one of  
249 the regions where winds from R7 intrude the basin (Figure 2a). In the winter, contributions from  
250 SOA and others (including IC, BC, windblown dust, and sea salt) are less than 8% each. In the  
251 summer, SOA and others each contribute 9~28% and less than 10%, respectively, but the SOA  
252 contributions larger than 15% are found only in the city centers where summer PM<sub>2.5</sub>  
253 concentrations are less than 30 µg m<sup>-3</sup>. In summary, local emissions are the largest contributor for  
254 all the city centers in both seasons, except for Suining. The non-local contributions for the city  
255 centers are in the ranges of 25~52% in the winter (except for 75% in Suining) and of 14~40% in  
256 the summer (except for 61% in Suining), and emissions outside the SCB account for 7~36% in the  
257 seasons.

258

### 259 **3.2.2. Spatial variations and citywide area-weighted averages**

260 The spatial variations of source-region contributions to PM<sub>2.5</sub> in the winter and summer are  
261 presented in Figures 2 and 3, respectively. In both seasons, local emissions are generally the largest  
262 contributor in each city, except that R7 has contributions similar to or larger than that of local  
263 emissions for most regions in eastern Chongqing (R5) and R2. Specifically, the contributions from  
264 R7 to PM<sub>2.5</sub> in R2 and R5 are approximately 20~80% in the winter and 20~60% in the summer.  
265 R7 also has contributions larger than 20% for a few areas in R3. The regions of R6, R8, and R9  
266 outside the SCB each has contributions of <5% across the basin, except for some very limited  
267 areas in the western and southern rims of the SCB. The contributions from R6, R8, and R9 are low  
268 because these areas are less urbanized and industrialized. In addition, the mountains to the west  
269 and south of the SCB also prevent the transport of air pollutants from these regions into the SCB  
270 (Figures 1c, 2a, and 3a). In summary, R7 is the sole non-SCB region that can have >20%  
271 contributions to PM<sub>2.5</sub> in the SCB, and its impact decreases from the northeast, east, and southeast  
272 to others in the basin.

273 As shown in Figures 2 and 3, PM<sub>2.5</sub> concentrations and its source contributions from a given  
274 region may vary greatly within a city in both seasons. For example, about 20~80% of PM<sub>2.5</sub> across



275 Chengdu (R4) and western Chongqing (R1) are due to local emissions in each season, and higher  
276 PM<sub>2.5</sub> concentrations are generally related to higher local contributions. For the downwind regions  
277 of Chengdu and western Chongqing, they receive considerable contributions from the two mega-  
278 cities. For example, over half areas of Meishan and Ya'an, which are downwind of Chengdu, have  
279 20~40% and 20~60% of PM<sub>2.5</sub> concentrations due to Chengdu in the winter, respectively. In the  
280 two seasons, western Chongqing contributes to about 10~40% of PM<sub>2.5</sub> concentrations in its  
281 neighboring cities, except that most area of eastern Chongqing (R5) is not affected by emissions  
282 from western Chongqing, as R5 is upwind of western Chongqing (Figure 2a). Because of the large  
283 spatial variations of PM<sub>2.5</sub> and its source contributions in the basin, we further calculated their  
284 citywide area-weighted averages (Table 2). In the winter, the citywide average PM<sub>2.5</sub>  
285 concentrations in Chengdu and urban Chongqing are 99 and 110  $\mu\text{g m}^{-3}$ , with only 38% and 47%  
286 due to local emissions, respectively. Non-local emissions also have high contributions in other  
287 SCB cities, with citywide averages of 39~66% and 25~52% in the winter and summer, respectively.  
288 The above suggests the importance of regional emission control to reduce PM<sub>2.5</sub> concentrations  
289 for the entire basin.

290

### 291 **3.2.3. Differences in PPM and SIA**

292 The transport distances of PPM, NH<sub>4</sub><sup>+</sup>, NO<sub>3</sub><sup>-</sup>, and SO<sub>4</sub><sup>2-</sup> might be different, because of the  
293 differences in chemical and physical processes that affect their concentrations in the atmosphere  
294 (Ying et al., 2014; Hu et al., 2015). This leads to significant differences in their regional  
295 distributions and thus requires different control strategies. From the source-region contributions to  
296 PPM and SIA for each city center shown in Figure 4, it is obvious that the regional transport of  
297 SIA is more significant than that of PPM. In the city centers of Chengdu and Chongqing, 55~65%  
298 of PPM and 25~45% of SIA are due to local emissions in the two seasons. In the city centers of  
299 R2, PPM is also more from local emissions (65~80%) than SIA is (25~45%) in both seasons.  
300 Similarly, local emissions have larger contributions to PPM (50~85%) than to SIA (34~50%) in  
301 all the city centers of R3 except for Suining, which is not significantly affected by local emissions.  
302 The spatial distributions of source-region contributions to PPM and SIA also indicate more  
303 significant transport of SIA (Figures S3-6) than PPM. For example, R3 contributes to >20% of  
304 SIA across entire Chengdu, but only half areas of Chengdu are about equally affected (>20%) by  
305 R3 for PPM. From the north to south in R2, the contributions from R7 to PPM decrease from ~55%  
306 to ~10%, while the contributions of R7 to SIA decrease from ~75% to ~20%. The contributions to  
307 NH<sub>4</sub><sup>+</sup>, NO<sub>3</sub><sup>-</sup>, and SO<sub>4</sub><sup>2-</sup> in each city center from local emissions and emissions within and outside  
308 SCB are further analyzed (Tables S3 and S4). In each city center, concentrations of SO<sub>4</sub><sup>2-</sup>  
309 (3.8~12.6 and 12~41  $\mu\text{g m}^{-3}$ ) are much higher than that of NH<sub>4</sub><sup>+</sup> (1.4~4.0 and 6.0~17.0  $\mu\text{g m}^{-3}$ ) and  
310 NO<sub>3</sub><sup>-</sup> (0.3~2.4 and 6~20  $\mu\text{g m}^{-3}$ ) in the summer and winter, respectively. Also, the transport of  
311 SO<sub>4</sub><sup>2-</sup> and its precursor is greater than the other two ions, as the percentage contributions from  
312 emissions outside the SCB to SO<sub>4</sub><sup>2-</sup> is higher than that to NH<sub>4</sub><sup>+</sup> and NO<sub>3</sub><sup>-</sup> in each city center. In  
313 both seasons, emissions outside SCB contribute <25% of NH<sub>4</sub><sup>+</sup> in the city centers, except for

314 Chongqing (26%) in the summer and Bazhong (36%) and Guangyuan (33%) in the winter. As for  
315  $\text{NO}_3^-$ , emissions outside SCB also contribute <25% in the city centers in both seasons, except for  
316 Bazhong (49%), Dazhou (34%), and Guangyuan (25%) in the summer and all the cities of R2  
317 (27~57%) in the winter. In the two seasons, emissions outside SCB account for 22~33% of  $\text{SO}_4^{2-}$   
318 in Chengdu and Chongqing, while they contribute 52~70% of  $\text{SO}_4^{2-}$  for the R2 cities. For the R3  
319 cities, emissions outside SCB account for 25~53% of  $\text{SO}_4^{2-}$  in the city centers in the seasons,  
320 except for Meishan (21%) in the winter. All the above suggest that it would be more efficient to  
321 control the SIA (particularly  $\text{SO}_4^{2-}$ ) and its precursors than PPM in order to reduce the transport  
322 of air pollutants within and into the basin.

323

### 324 **3.3. Maximum daily contributions from a given region**

325 The maximum daily contribution from a given region to  $\text{PM}_{2.5}$  (MDC,  $\mu\text{g m}^{-3}$ ) in each city  
326 center is shown in Table 3 for winter and Table S5 for summer. The largest MDC for each city  
327 center (79~291 and 13~147  $\mu\text{g m}^{-3}$  in the winter and summer, respectively) are found associated  
328 with local emissions, except for Guangyuan and Suining. In Guangyuan and Suining, the largest  
329 MDCs in the winter are from R7 (62  $\mu\text{g m}^{-3}$ ) and R2 (110  $\mu\text{g m}^{-3}$ ), both are slightly higher than  
330 that from the local region of 60 and 105  $\mu\text{g m}^{-3}$ , respectively. Table 3 also shows that the inter-  
331 basin transport of air pollutants can have large contributions (>50  $\mu\text{g m}^{-3}$ ) on high  $\text{PM}_{2.5}$  days  
332 (>150  $\mu\text{g m}^{-3}$ ). For example, R7 contributes 99  $\mu\text{g m}^{-3}$  to total  $\text{PM}_{2.5}$  (200  $\mu\text{g m}^{-3}$ ) in Chongqing  
333 on a winter day. In Nanchong, the MDC due to western Chongqing (R1) is 58  $\mu\text{g m}^{-3}$ , when daily  
334  $\text{PM}_{2.5}$  is 180  $\mu\text{g m}^{-3}$  on that day. In Chengdu, R3 and R7 can contribute up to 86 and 63  $\mu\text{g m}^{-3}$  on  
335 the days with daily  $\text{PM}_{2.5}$  of 267 and 151  $\mu\text{g m}^{-3}$ , respectively. In Deyang and Meishan, the MDCs  
336 from Chengdu are 147 and 138  $\mu\text{g m}^{-3}$  on the days having daily  $\text{PM}_{2.5}$  of 288 and 235  $\mu\text{g m}^{-3}$ ,  
337 respectively. Table S4 shows that air pollutant regional transport is also significant on certain days  
338 in the summer. For example, the highest summer MDC from R7 among the 18 central cities is  
339 found for Bazhong (36  $\mu\text{g m}^{-3}$ ), when daily  $\text{PM}_{2.5}$  is 63  $\mu\text{g m}^{-3}$ . Chengdu contributes about 44, 16,  
340 55, 13, 7, and 21  $\mu\text{g m}^{-3}$  to Deyang, Leshan, Meishan, Ya'an, and Mianyang on the summer days,  
341 when daily  $\text{PM}_{2.5}$  are 89, 56, 100, 85, 22, and 54  $\mu\text{g m}^{-3}$ , respectively. All the above suggest that  
342 joint effects should be made by neighboring cities and the provinces to the east of the SCB in order  
343 to prevent high  $\text{PM}_{2.5}$  episodes for the SCB.

344

### 345 **3.4. Impacts of topography on $\text{PM}_{2.5}$ concentrations**

346 While air pollutant emissions are the root of air pollution, topography and meteorological  
347 conditions play a very important role on determining the fate of pollutants including dispersion,  
348 accumulation, and transformation (Arya, 1999; Zhang et al., 2015; He et al., 2017). It has been  
349 widely noticed that heavy air pollution often occurs in well urbanized and/or industrialized cities  
350 associated with mountains and basins, such as Beijing, Chengdu, Xi'an, and Lanzhou in China

351 (Chambers et al., 2015; Bei et al., 2017; Bei et al., 2018; Ning et al. 2018), Mexico City, Salt Lake  
352 City, and Los Angeles in the North America (Langford et al., 2010; Witeman et al., 2014;  
353 Calderón-Garcidueñas et al., 2015), and megacities in the Mediterranean Basin of the Europe  
354 (Kanakidou et al., 2011). The SCB is surrounded by the QTP to the west, YGP to the south, WUM  
355 to the east, and DBM to the northeast. Mainly affected by the high elevations of the QTP and YGP,  
356 near-surface winds mainly intrude the basin from the north, east, and southeast in the summer and  
357 winter, as shown in Figures 2(a) and 3(a). Consequently, R7 is the largest contributor outside the  
358 basin, contributing 20~60% of  $PM_{2.5}$  in the eastern, northeastern, and southeastern parts of the  
359 SCB (Figures 2(h) and 3(h)), where  $PM_{2.5}$  concentrations are relatively lower in the SCB ( $<75$  and  
360  $25 \mu g m^{-3}$  in the winter and summer, respectively). The contributions from R6 (including YGP)  
361 and R8 (including QTP) are  $<10\%$  along the western and southern rims of the SCB. Within the  
362 basin, near-surface winds travel anti-clockwise wind and form a cyclone near Yibin, Zigong,  
363 Neijiang, and Luzhou in the south (Figures 1(b), 2(a), and 3(a)) (Lin, 2015), causing air pollutants  
364 transported to and accumulated at the cyclone.  $PM_{2.5}$  concentrations in the cyclone-affected region  
365 (mostly 100-150 and 30~50  $\mu g m^{-3}$  in the winter and summer, respectively) are generally lower  
366 than that of Chengdu and Chongqing but are higher than that of most of other regions. In Yibin,  
367 Zigong, Neijiang, and Luzhou, at least 39~53% and 25~44% of citywide average  $PM_{2.5}$   
368 concentrations are not due to their own emissions in the winter and summer, respectively (Tables  
369 2 and S2). R7 only contributes about 10% to  $PM_{2.5}$  in the cyclone-affected region. In order to  
370 reduce seasonal and annual concentrations of  $PM_{2.5}$  within the SCB, the emissions and inter-city  
371 transport of air pollutants within the basin should receive the priorities to be controlled.

372

#### 373 4. Conclusion

374 In this study, a source-oriented CMAQ model was applied to quantify contributions of nine  
375 regions to  $PM_{2.5}$  for the 18 cities in the SCB. The simulations were carried out for winter  
376 (December 2014 to February 2015) and summer (June to August 2015). Predicted citywide area-  
377 weighted average  $PM_{2.5}$  concentrations are much higher in the winter ( $60\sim 191 \mu g m^{-3}$ ) than in the  
378 summer ( $14\sim 64 \mu g m^{-3}$ ). In the winter, the citywide average  $PM_{2.5}$  concentrations in Chengdu and  
379 western Chongqing are 99 and 110  $\mu g m^{-3}$ , with 44% and 52% due to non-local emissions,  
380 respectively. Non-local emissions also have high contributions in other SCB cities, with citywide  
381 averages of 39~66% and 25~52% in the winter and summer, respectively. Among the four regions  
382 outside the SCB, only the one to the northeast, east, and southeast of the SCB (R7) has large  
383 contributions to  $PM_{2.5}$  concentrations for the SCB in both seasons (10~80%), and the contributions  
384 decrease from the rims of the northeastern, eastern, and southeastern SCB to other regions.  
385 However, the MDCs from R7 are large ( $35\sim 99 \mu g m^{-3}$ ) for all the city centers in the winter. On  
386 high  $PM_{2.5}$  days in the winter, emissions outside SCB can contribute up to 99  $\mu g m^{-3}$  in a city  
387 center, suggesting the importance of regional emission control in not just reducing averaged  $PM_{2.5}$   
388 but also preventing severe PM pollution events. The transport of SIA is greater than that of PPM,  
389 suggested by that local emissions have higher contributions to PPM ( $>55\%$ ) than to SIA ( $<45\%$ )

390 in the city centers in both seasons. Among the three ions of SIA, the transport of  $\text{SO}_4^{2-}$  and its gas-  
391 phase precursor ( $\text{SO}_2$ ) is the greatest in general, as >50% of it in all the city centers is associated  
392 with non-local emissions in both seasons, except that the contributions are 37~44% in Chongqing  
393 and Chengdu in the summer and Chongqing in the winter. In conclusion, in order to reduce  $\text{PM}_{2.5}$   
394 concentrations and prevent high  $\text{PM}_{2.5}$  days for the entire SCB, local emissions and the transport  
395 of air pollutants within and across SCB should be controlled simultaneously.

396

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593 particulate matter in China: Insights from source apportionment studies published in 1987–2017, *Environ.*  
594 *Int.*, 115, 343-357, 2018.

595 Table 1. Predicted source-region contributions to PM<sub>2.5</sub> in the 18 SCB city centers in the winter.

Region ID	City	PM <sub>2.5</sub> ( $\mu\text{g m}^{-3}$ )	Contributions from each region, SOA, and others* (%)													
			R1	R2	R3	R4	R5	Within SCB	R6	R7	R8	R9	Outside SCB	SOA	Others	Non-local <sup>#</sup>
<b>R1</b>	Chongqing <sup>\$</sup>	191	<b>67.7</b>	4.5	3.7	0.4	1.8	<b>78.1</b>	2.7	11.1	0.1	0.9	<b>14.8</b>	3.9	3.3	<b>25.2</b>
<b>R2</b>	Bazhong	65	2.0	<b>42.6</b>	2.8	1.2	2.9	<b>51.5</b>	1.7	31.9	0.1	2.7	<b>36.4</b>	6.9	5.1	<b>45.3</b>
	Dazhou	89	2.4	<b>47.3</b>	1.7	0.5	7.5	<b>59.4</b>	1.5	27	0.1	1.6	<b>30.2</b>	6.0	4.3	<b>42.3</b>
	Guangan	109	8.9	<b>50.4</b>	2.0	0.5	5.4	<b>67.2</b>	2.3	19.6	0.1	1.4	<b>23.4</b>	5.8	3.5	<b>40.2</b>
	Guangyuan	60	2.3	<b>45.5</b>	5.2	1.8	1.7	<b>56.5</b>	1.3	25.5	0.1	4.2	<b>31.1</b>	6.5	5.8	<b>42.1</b>
	Nanchong	120	6.1	<b>56.5</b>	3.2	0.6	3.1	<b>69.5</b>	1.8	18.1	0.1	1.3	<b>21.3</b>	5.7	3.4	<b>34.3</b>
<b>R3</b>	Deyang	143	2.7	4.6	<b>58.0</b>	14.5	0.8	<b>80.6</b>	0.8	9.4	0.1	1.6	<b>11.9</b>	4.2	3.4	<b>34.5</b>
	Leshan	125	2.9	3.1	<b>58.7</b>	15.1	0.6	<b>80.4</b>	1.2	8.2	0.1	1.2	<b>10.7</b>	5.7	3.0	<b>32.4</b>
	Luzhou	149	13.9	4.8	<b>53.9</b>	1.8	1.2	<b>75.6</b>	3.8	11.2	0.1	1.0	<b>16.1</b>	5.4	2.9	<b>37.8</b>
	Meishan	153	2.5	3.1	<b>40.2</b>	36.3	0.6	<b>82.7</b>	0.9	7.7	0.1	1.1	<b>9.8</b>	4.6	2.9	<b>52.3</b>
	Mianyang	114	2.7	6.8	<b>60.3</b>	4.8	1.1	<b>75.7</b>	0.9	12.4	0.1	2.2	<b>15.6</b>	4.9	3.8	<b>31.0</b>
	Neijiang	140	11.3	7.6	<b>51.9</b>	2.8	1.3	<b>74.9</b>	2.4	12.7	0.1	1.1	<b>16.3</b>	5.6	3.1	<b>39.3</b>
	Suining	100	11.7	33.5	<b>14.4</b>	1.1	3.2	<b>63.9</b>	2.6	21.4	0.1	1.7	<b>25.8</b>	6.8	3.6	<b>75.3</b>
	Ya'an	79	3.2	3.6	<b>45.9</b>	20.1	0.7	<b>73.5</b>	1.4	11.0	0.3	2.3	<b>15.0</b>	7.6	3.8	<b>42.6</b>
	Yibin	134	6.8	4.1	<b>60.1</b>	5.5	0.9	<b>77.4</b>	2.8	9.7	0.1	1.1	<b>13.7</b>	6.0	3.0	<b>31.0</b>
	Zigong	145	8.9	6.0	<b>57.1</b>	3.2	1.1	<b>76.3</b>	2.4	11.4	0.1	1.1	<b>15.0</b>	5.5	3.3	<b>34.2</b>
	Ziyang	131	5.8	7.3	<b>54.0</b>	7.1	1.2	<b>75.4</b>	1.7	12.6	0.1	1.4	<b>15.8</b>	5.6	3.2	<b>37.2</b>
<b>R4</b>	Chengdu	144	2.2	3.5	20.4	<b>55.2</b>	0.6	<b>81.9</b>	0.8	8.2	0.1	1.4	<b>10.5</b>	4.2	3.5	<b>37.2</b>

596 \* Others include initial and boundary conditions, windblown dust, and sea salt.

597 # Non-local=Within SCB + Outside SCB – Local.

598 \$ the city center of Chongqing.

599 Table 2. Predicted citywide area-weighted average PM<sub>2.5</sub> concentrations and source-region contributions in the 18 SCB cities in the  
 600 winter and summer.

Region ID	No. of grid cells	Total area (km <sup>2</sup> )	City	No. of grid cells	Total area (km <sup>2</sup> )	Winter			Summer		
						PM <sub>2.5</sub> (μg m <sup>-3</sup> )	Contributions (%)		PM <sub>2.5</sub> (μg m <sup>-3</sup> )	Contributions (%)	
							Local	Non-local <sup>#</sup>		Local	Non-local <sup>#</sup>
<b>R1</b>	248	28011	Western Chongqing	248	28011	<b>99</b>	37.8	52.1	<b>27</b>	36.8	36.0
<b>R2</b>	543	56265	Bazhong	106	10734	<b>51</b>	27.9	58.2	<b>14</b>	19.3	43.6
			Dazhou	139	14689	<b>64</b>	29.1	58.6	<b>16</b>	21.6	45.7
			Guangan	60	5618	<b>100</b>	40.6	49.5	<b>23</b>	32.3	40.3
			Guangyuan	133	14182	<b>50</b>	28.3	57.1	<b>13</b>	23.4	40.5
			Nanchong	105	11042	<b>89</b>	47.9	41.4	<b>20</b>	36.3	34.1
<b>R3</b>	998	98185	Deyang	59	5346	<b>101</b>	49.9	40.2	<b>25</b>	44.5	32.8
			Leshan	107	11599	<b>81</b>	42.5	44.4	<b>14</b>	35.0	29.7
			Luzhou	112	10758	<b>92</b>	35.3	53.8	<b>20</b>	25.6	43.7
			Meishan	74	6570	<b>120</b>	42.6	47.7	<b>26</b>	36.7	35.8
			Mianyang	171	18012	<b>59</b>	36.8	49.2	<b>15</b>	32.9	36.7
			Neijiang	58	4859	<b>122</b>	45.8	44.8	<b>29</b>	40.8	34.7
			Suining	50	4743	<b>104</b>	30.3	59.7	<b>23</b>	25.4	48.4
			Ya'an	131	13493	<b>45</b>	34.5	49.2	<b>6</b>	27.9	35.1
			Yibin	117	11827	<b>101</b>	50.1	39.1	<b>21</b>	43.3	24.9
			Zigong	45	3935	<b>126</b>	51.3	39.4	<b>30</b>	47.3	26.9
Ziyang	74	7042	<b>115</b>	40.3	50.2	<b>25</b>	32.6	41.9			
<b>R4</b>	105	10753	Chengdu	105	10753	<b>110</b>	46.5	44.1	<b>31</b>	44.6	33.0
<b>R5</b>	390	44371	Eastern Chongqing	390	44371	<b>54</b>	21.0	66.0	<b>15</b>	14.7	52.1

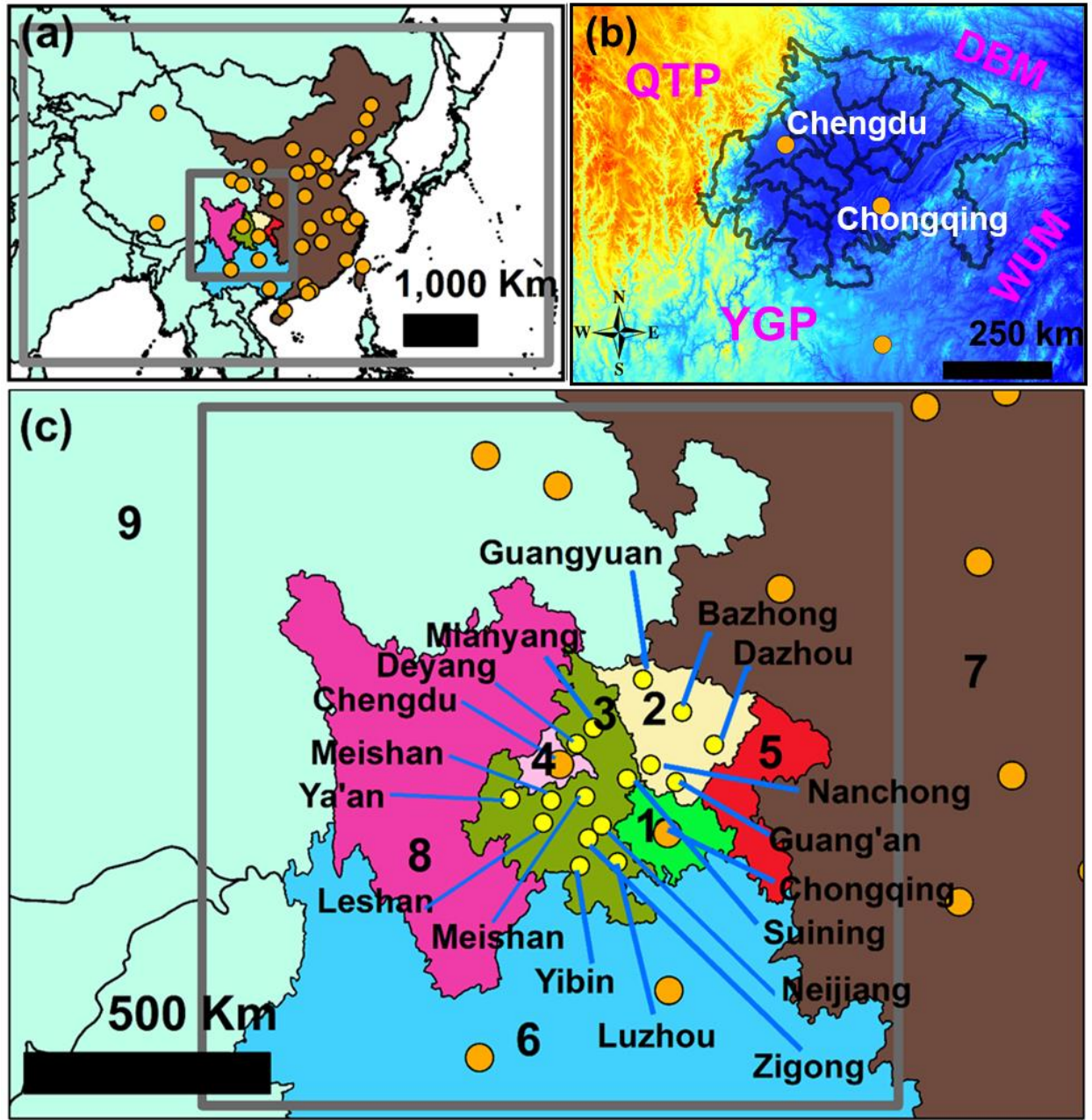
601 <sup>#</sup> Non-local =100%-Local-SOA-Others; Others include initial and boundary conditions, windblown dust, and sea salt.

602 Table 3. Predicted maximum daily contribution from a given region (MDCs) in the SCB city center and the corresponding PM<sub>2.5</sub>  
603 concentrations in the city center on the same day. Only winter data are included in this table. The units are  $\mu\text{g m}^{-3}$ . The numbers in the  
604 bold present the contributions due to local emissions or that from R7.

Region ID	Cities	MDCs (total PM <sub>2.5</sub> concentrations)										
		Within SCB					Outside SCB					SOA
		R1	R2	R3	R4	R5	R6	R7	R8	R9		
<b>R1</b>	Chongqing <sup>\$</sup>	<b>291 (353)</b>	54 (414)	48 (294)	16 (302)	15 (143)	22 (290)	<b>99 (200)</b>	1 (182)	8 (212)	12 (160)	12 (294)
	Bazhong	8 (160)	<b>83 (139)</b>	23 (160)	18 (160)	7 (64)	14 (68)	<b>73 (109)</b>	0 (140)	6 (90)	14 (327)	11 (34)
	Dazhou	46 (219)	<b>123 (216)</b>	16 (171)	6 (171)	34 (219)	12 (77)	<b>79 (190)</b>	0 (173)	5 (54)	18 (367)	8 (54)
<b>R2</b>	Guangan	72 (159)	<b>129 (205)</b>	31 (180)	10 (180)	22 (225)	14 (102)	<b>76 (165)</b>	0 (234)	5 (129)	16 (219)	7 (75)
	Guangyuan	11 (122)	<b>60 (107)</b>	35 (122)	15 (122)	4 (56)	10 (86)	<b>62 (100)</b>	0 (91)	11 (46)	16 (226)	10 (25)
	Nanchong	58 (180)	<b>152 (242)</b>	50 (200)	12 (243)	18 (238)	13 (141)	<b>78 (166)</b>	0 (177)	5 (136)	17 (178)	10 (71)
	Deyang	32 (226)	30 (257)	<b>170 (296)</b>	147 (288)	4 (257)	14 (215)	<b>70 (132)</b>	0 (215)	9 (137)	13 (91)	10 (67)
	Leshan	15 (270)	15 (166)	<b>163 (270)</b>	57 (222)	4 (166)	10 (222)	<b>47 (122)</b>	1 (108)	5 (118)	15 (222)	9 (105)
	Luzhou	115 (245)	33 (207)	<b>211 (261)</b>	21 (240)	10 (207)	18 (259)	<b>71 (272)</b>	1 (192)	6 (90)	17 (307)	10 (182)
	Meishan	17 (189)	22 (379)	<b>155 (379)</b>	138 (235)	4 (176)	11 (263)	<b>54 (263)</b>	1 (379)	5 (80)	15 (164)	9 (100)
	Mianyang	24 (209)	30 (170)	<b>219 (294)</b>	67 (305)	6 (168)	11 (144)	<b>70 (112)</b>	0 (337)	9 (86)	16 (294)	11 (46)
<b>R3</b>	Neijiang	93 (214)	39 (168)	<b>161 (205)</b>	49 (309)	6 (168)	17 (165)	<b>79 (198)</b>	1 (192)	5 (198)	19 (181)	9 (164)
	Suining	75 (183)	110 (242)	<b>105 (186)</b>	25 (225)	14 (210)	17 (151)	<b>83 (167)</b>	0 (154)	5 (79)	17 (198)	10 (41)
	Ya'an	14 (126)	17 (205)	<b>79 (170)</b>	88 (234)	3 (205)	8 (170)	<b>35 (205)</b>	1 (123)	6 (40)	15 (189)	9 (43)
	Yibin	74 (283)	21 (162)	<b>160 (217)</b>	45 (164)	5 (236)	16 (188)	<b>58 (166)</b>	1 (122)	4 (147)	14 (170)	9 (127)
	Zigong	64 (281)	34 (171)	<b>162 (214)</b>	51 (291)	6 (171)	17 (192)	<b>71 (195)</b>	1 (195)	5 (199)	16 (283)	10 (162)
	Ziyang	39 (192)	33 (137)	<b>164 (298)</b>	63 (176)	7 (120)	19 (181)	<b>73 (155)</b>	1 (298)	6 (176)	18 (203)	9 (126)
<b>R4</b>	Chengdu	16 (232)	23 (225)	86 (267)	<b>250 (327)</b>	4 (139)	11 (267)	<b>63 (151)</b>	1 (166)	7 (132)	16 (228)	9 (101)

605 <sup>#</sup> Others include initial and boundary conditions, windblown dust, and sea salt.

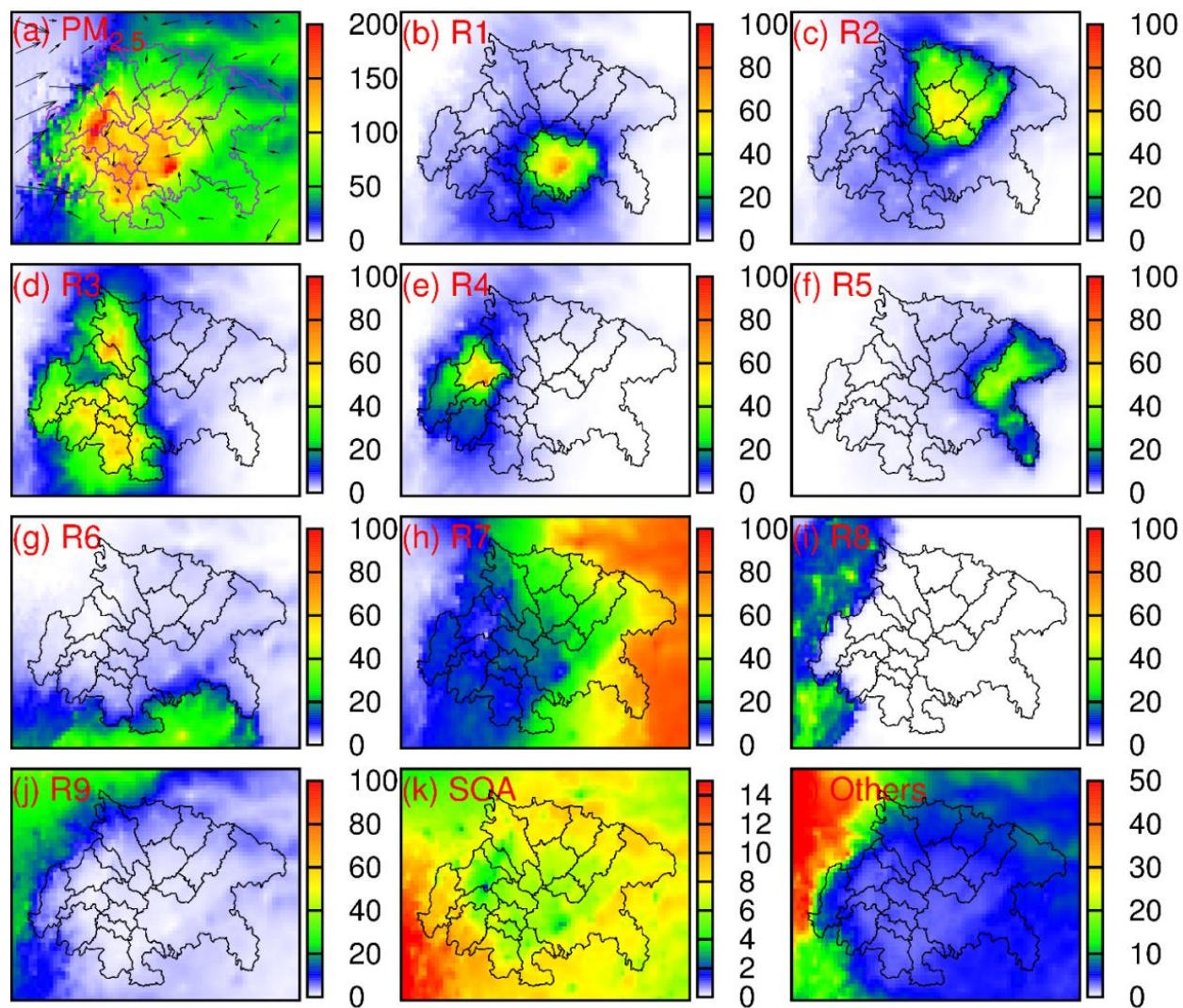
606 <sup>\$</sup> includes the city center of Chongqing.



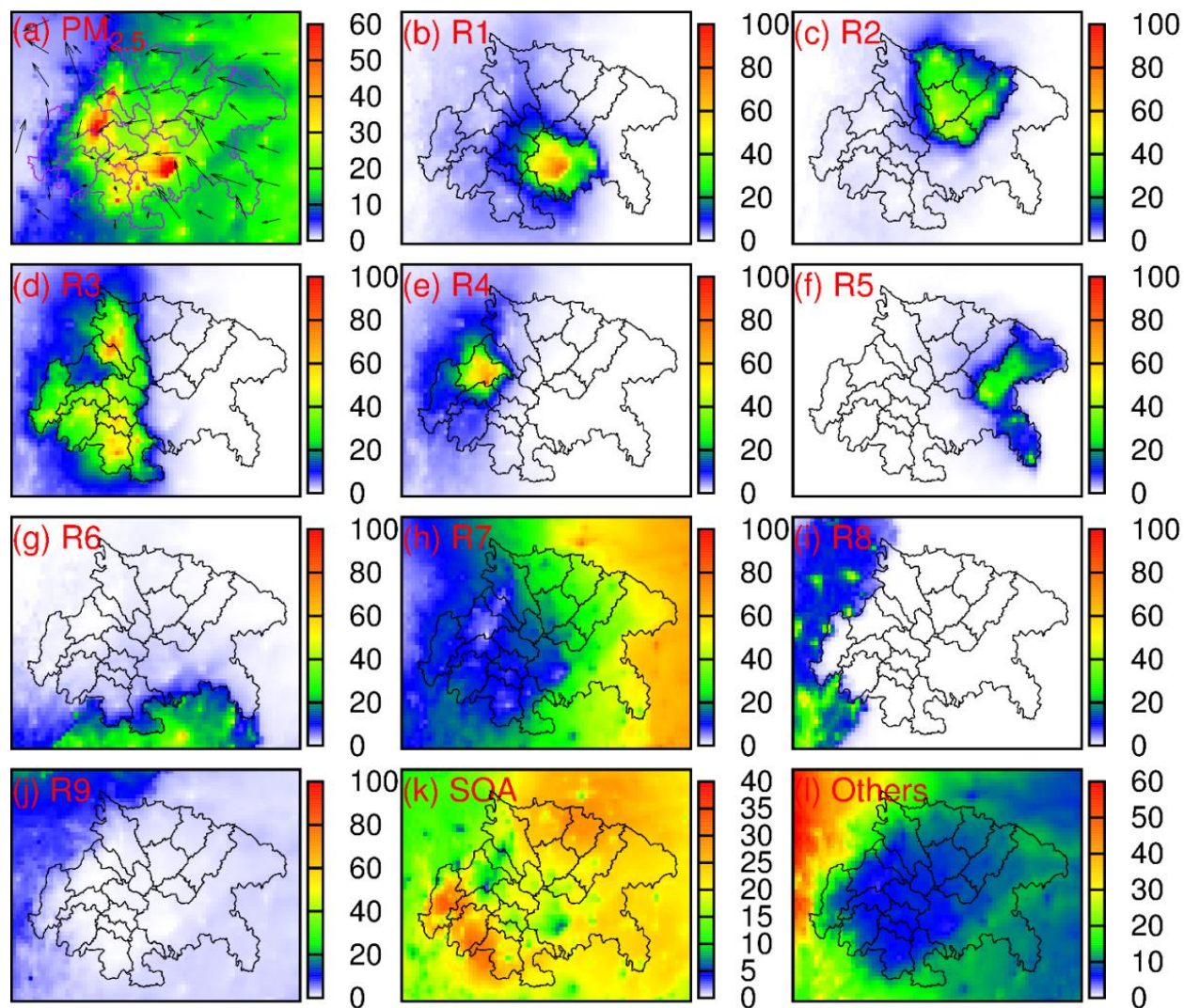
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609 Figure 1. (a) Locations of the 12-km and 36-km domains (grey rectangles) and the locations of  
 610 provincial capitals and municipalities (orange circles), (b) terrain within and surrounding the 18  
 611 cities of the SCB (black line), and (c) locations of region categories 1~9 and the prefecture-level  
 612 cities (yellow circles). Regions 1~5 are the cities within the SCB. Regions 1, 4, and 5 are western  
 613 Chongqing, Chengdu, and eastern Chongqing, respectively. The city center of Chongqing is  
 614 located in western Chongqing. Region 8 is the area of Sichuan Province excluding those cities in  
 615 the SCB. QTP, Qinghai-Tibetan Plateau; YGP, Yunnan-Guizhou Plateau; DBM, Dabashan  
 616 Mountains; WUM, Wushan Mountains.



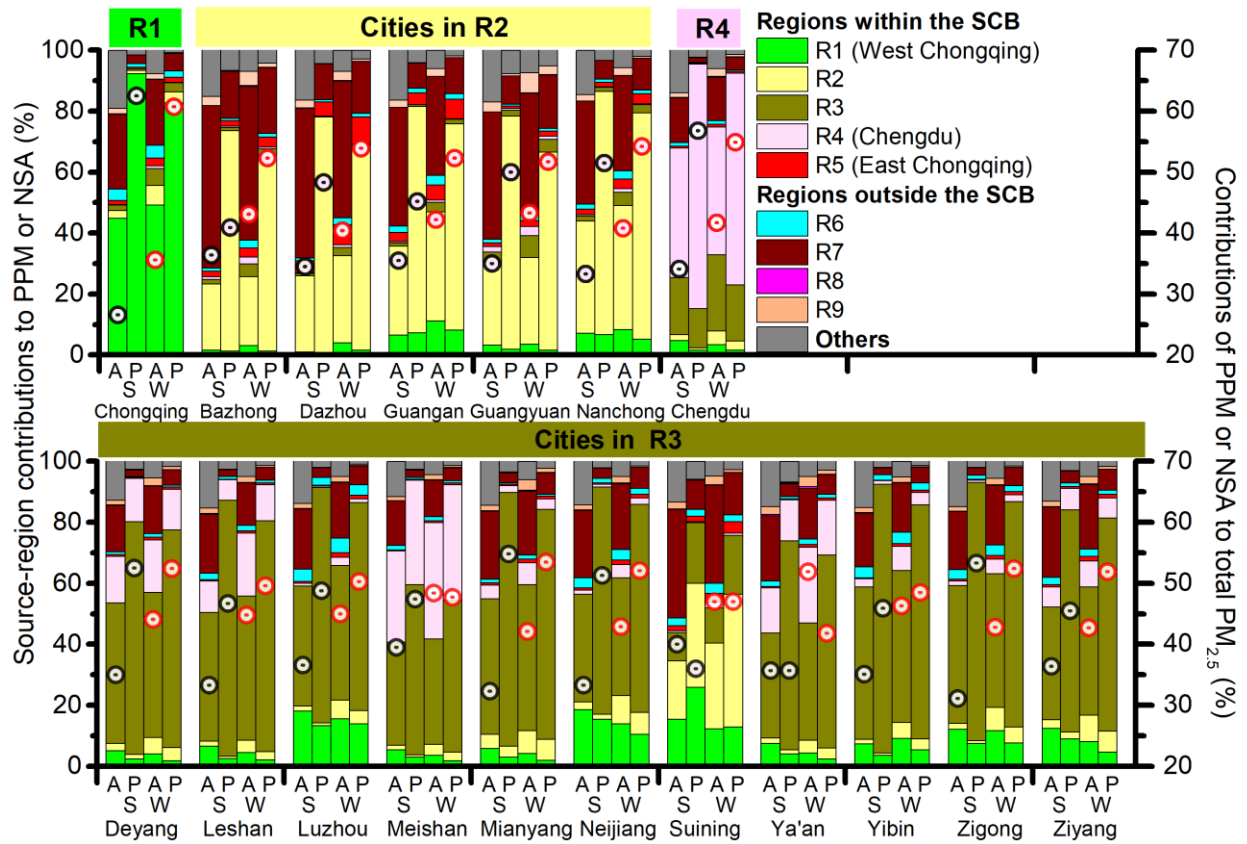


617  
 618 Figure 2. (a) Spatial distributions of predicted PM<sub>2.5</sub> concentrations ( $\mu\text{g m}^{-3}$ ) and (b-l) source-  
 619 region contributions to PM<sub>2.5</sub> (%) in the winter. Others include IC, BC, windblown dust, and sea  
 620 salt. Black arrows in (a) are wind vectors.



621  
 622 Figure 3. (a) Spatial distributions of predicted PM<sub>2.5</sub> concentrations ( $\mu\text{g m}^{-3}$ ) and (b-l) the source-  
 623 region contributions to PM<sub>2.5</sub> (%) in the summer. Others include IC, BC, windblown dust, and sea  
 624 salt.

625



626

627 Figure 4. Predicted source-region contributions to SIA (A) and PPM (P) (bars, left y-axis) and the  
 628 predicted proportions of SIA and PPM in PM<sub>2.5</sub> (circles, right y-axis) for the 18 city centers of the  
 629 SCB in the summer (S) and winter (W). Others include IC, BC, windblown dust, and sea salt.

630