

1 A proxy for atmospheric daytime gaseous sulfuric acid 2 concentration in urban Beijing

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23
24 **Abstract.** Gaseous sulfuric acid is known as one of the key precursors for atmospheric
25 new particle formation processes, but its measurement remains a difficulty. A proxy
26 method that is able to derive gaseous sulfuric acid concentrations from parameters that
27 can be measured relatively easily and accurately is therefore highly desirable for the
28 atmospheric chemistry community. Although such methods are available for clean
29 atmospheric environments, a proxy that works well in a polluted atmosphere, such as
30 those in Chinese megacities, is yet to be developed. In this study, the gaseous sulfuric
31 acid concentration was measured in February-March, 2018, in urban Beijing by a nitrate
32 based - Long Time-of-Flight Chemical Ionization Mass Spectrometer (LToF-CIMS). A
33 number of atmospheric parameters were recorded concurrently including the ultraviolet
34 radiation B (UVB) intensity, concentrations of O₃, NO_x (sum of NO and NO₂), SO₂ and
35 HONO, and aerosol particle number size distributions. A proxy for atmospheric
36 daytime gaseous sulfuric acid concentration was derived through a statistical analysis

37 method by using the UVB intensity, [SO₂], condensation sink (CS), [O₃], and [HONO]
38 (or [NO_x]) as the predictor variables. In this proxy method, we considered the formation
39 of gaseous sulfuric acid from reactions of SO₂ and OH radicals during the daytime, and
40 loss of gaseous sulfuric acid due to its condensation onto the pre-existing particles. In
41 addition, we explored formation of OH radicals from the conventional gas-phase
42 photochemistry using ozone as a proxy and from the photolysis of HONO using HONO
43 (and subsequently NO_x) as a proxy. Our results showed that the UVB intensity and [SO₂]
44 are dominant factors for the production of gaseous sulfuric acid, and that the simplest
45 proxy could be constructed with the UVB intensity and [SO₂] alone. When the OH
46 radical production from both homogeneously- and heterogeneously-formed precursors
47 were considered, the relative errors were reduced up to 20 %.

48 **1 Introduction**

49 Gaseous sulfuric acid (H_2SO_4) is a key precursor for atmospheric new particle
50 formation (NPF) processes (Kerminen, 2018; Kirkby et al., 2011; Kuang et al., 2008;
51 Kulmala and Kerminen, 2008; Sipilä et al., 2010). A number of atmospheric nucleation
52 mechanisms including H_2SO_4 - H_2O binary nucleation (Benson et al., 2008; Duplissy et
53 al., 2016; Kirkby et al., 2011), H_2SO_4 - NH_3 - H_2O ternary nucleation (Kirkby et al., 2011;
54 Korhonen et al., 1999; Kürten et al., 2015), and H_2SO_4 -DMA- H_2O ternary nucleation
55 (Almeida et al., 2013; Jen et al., 2014; Kürten et al., 2014; Petäjä et al., 2011; Yao et
56 al., 2018) involve the participation of gaseous sulfuric acid molecules. In addition, the
57 condensation of gaseous sulfuric acid onto newly-formed particles contributes to their
58 initial growth (Kuang et al., 2012; Kulmala et al., 2013). Quantitative assessments of
59 the contribution of gaseous sulfuric acid to both the new particle formation rates and
60 the particle growth rates require real-time measurements of gaseous sulfuric acid
61 concentrations prior to and during the NPF events (Nieminen et al., 2010; Paasonen et
62 al., 2010).

63 Measurements of gaseous sulfuric acid in the lower troposphere are challenging
64 because its ambient concentration is typically quite low (10^6 - 10^7 molecule cm^{-3})
65 (Kerminen et al., 2010; Mikkonen et al., 2011). Reported real-time measurements of
66 gaseous sulfuric acid are currently based on Chemical Ionization Mass Spectrometry
67 with NO_3^- and its ligands as reagent ions (nitrate CIMS) because nitrate CIMS with an
68 atmospheric pressure interface (API) has a low detection limit for the atmospheric
69 concentration range of gaseous sulfuric acid (Jokinen et al., 2012), and a constant
70 fraction of sulfuric acid present in the air sample will be ionized by excessive nitrate
71 ions in CIMS under constant instrumental conditions (Kürten et al., 2012; Zheng et al.,
72 2010), which makes the quantification of gaseous sulfuric acid feasible.

73 Arnold and Fabian (1980) measured the negative ions in the stratosphere and
74 derived the concentration of stratospheric gaseous sulfuric acid from the fractional
75 abundances of a series of stratospheric negative ions as well as the rate constants. Later,
76 real-time measurement of sulfuric acid in the lower troposphere was performed using
77 nitrate CIMS (Eisele and Tanner, 1993), with laboratory calibrations by production of
78 known concentrations of OH radicals that are titrated into gaseous sulfuric acid.
79 Thereafter, measurements of sulfuric acid using CIMS have been performed around the
80 world (e.g., Berresheim et al., 2000; Bianchi et al., 2016; Chen et al., 2012; Jokinen et
81 al., 2012; Kuang et al., 2008; Kürten et al., 2014; Kurtén et al., 2011; Petäjä et al., 2009;

82 Weber et al., 1997; Zheng et al., 2011), and CIMS has been proven to be a robust tool
83 for gaseous sulfuric acid detection. However, sulfuric acid measurements are still rather
84 sparse because of the high cost of the CIMS instrument and the extensive demand of
85 specialized expertise on the instrument calibration, maintenance, and data processing,
86 etc. Therefore, a proxy for gaseous sulfuric acid concentration is highly desirable.

87 Proxies for the estimation of atmospheric gaseous sulfuric acid concentrations
88 were previously developed to approximate measurement results of sulfuric acid
89 in Hyytiälä, Southern Finland (Petäjä et al., 2009), assuming that gaseous sulfuric acid
90 is formed from reactions between SO₂ and OH radicals, and lost due to its condensation
91 onto pre-existing particles. The derived simplest proxy can be written as Eq. (1) below,
92 and the authors recognized that the proxies might be site-specific and should be verified
93 prior to utilization in other environments.

94

$$95 \quad [H_2SO_4] = k \cdot \frac{[SO_2] \cdot (UVB \text{ or } Global \text{ radiation})}{CS} \quad (1)$$

96

97 Mikkonen et al. (2011) later developed a couple of statistical proxies based on
98 measurements of sulfuric acid in six European and North American sites, including
99 urban, rural and forest areas. Their results showed that the radiation intensity and [SO₂]
100 are the most important factors to determine the concentration of sulfuric acid, and that
101 the impact of condensation sink (CS) for gaseous sulfuric acid, is generally negligible.
102 In several proxies developed by Mikkonen et al. (2011), the correlation between the
103 gaseous sulfuric acid concentration and CS is positive, which is against what one would
104 expect because a larger CS normally leads to a faster loss for gaseous sulfuric acid. In
105 addition, the performance of a proxy equation is site-specific because of varying
106 atmospheric conditions from one site to another, which implies that the proxy suggested
107 by Mikkonen et al. (2011) might not work well in locations with atmospheric
108 environments different from those in the six sites of that study.

109 Beijing is a location with typical values of CS (e.g., 0.01-0.24 s⁻¹ in the 5-95%
110 percentiles in this study) being 10-100 times higher (Herrmann et al., 2014; Wu et al.,
111 2007; Xiao et al., 2015; Yue et al., 2009; Zhang et al., 2011) and typical SO₂
112 concentrations being 1-10 times higher (Wang et al., 2011a; Wu et al., 2017) than those
113 in Europe and North America (Dunn et al., 2004; Mikkonen et al., 2011), yet measured
114 gaseous sulfuric acid concentrations are relatively similar in these environments (Chen

115 et al., 2012; Smith et al., 2008; Wang et al., 2011b; Zheng et al., 2011). Whether
116 previous proxies developed for European and North American sites work in Beijing
117 remains to be tested. Furthermore, in addition to the gas phase reaction between O(¹D)
118 and water molecules (Crutzen and Zimmermann, 1991; Logan et al., 1981), photolysis
119 of HONO could be a potentially important source of OH radicals in the atmosphere in
120 the early morning (Alicke et al., 2002, 2003; Elshorbany et al., 2009; Li et al., 2012)
121 and during the daytime (Acker et al., 2005; Aumont et al., 2003; Kleffmann, 2007). An
122 experimental study measuring HONO near the surface layer estimated that HONO was
123 a main contributor to OH production in Beijing, with HONO's contribution being larger
124 than 70 % at around 12:00-13:00, except for summer when the contribution of O₃
125 dominated (Hendrick et al., 2014). Given the distinct characteristics of these two OH
126 radical formation pathways, they both should be included and evaluated separately
127 when a proxy for atmospheric gaseous sulfuric acid concentration is being built. The
128 reactions between SO₂ and Criegee intermediates formed from the ozonolysis of
129 atmospheric alkenes could be a potential source of sulfuric acid only in the absence of
130 solar radiation (Boy et al., 2013; Mauldin et al., 2012), so these reactions are expected
131 to provide a minor contribution to the formation of gaseous sulfuric acid during the
132 daytime in urban Beijing.

133 In this study, gaseous sulfuric acid concentration was measured by a Long Time-
134 of-Flight Chemical Ionization Mass Spectrometer (LToF-CIMS) in February - March,
135 2018, in urban Beijing. A number of atmospheric parameters were recorded
136 concurrently, including the ultraviolet radiation B (UVB) intensity, concentrations of
137 O₃, NO_x, SO₂ and HONO, and particle number size distributions. The objective of this
138 study is to develop a robust daytime gaseous sulfuric acid concentration proxy for
139 Beijing, a representative Chinese megacity with urban atmospheric environments.

140

141 **2 Ambient measurements**

142 An intensive campaign was carried out from 9 February to 14 March, 2018 on the
143 fifth floor of a teaching building in the west campus of Beijing University of Chemical
144 Technology (39° 94' N, 116° 30' E). This monitoring site is 2 km to the west of the West
145 3rd Ring Road and surrounded by commercial properties and residential dwellings.
146 Hence, this station can be regarded as a representative urban site.

147 The sulfuric acid concentration was measured by a LToF-CIMS (Aerodyne
148 Research, Inc.) equipped with a nitrate chemical ionization source. Ambient air was

149 drawn into the ionization source through a stainless-steel tube with a length of 1.6 m
 150 and a diameter of 3/4 inch. A mixture of a 3 standard cubic centimeter per minute (sccm)
 151 ultrahigh purity nitrogen flow containing nitric acid and a 20 standard liter per minute
 152 (slpm) pure air flow supplied by a zero-air generator (Aadco 737, USA), together as a
 153 sheath flow, was introduced into a PhotoIonizer (Model L9491, Hamamatsu, Japan) to
 154 produce nitrate reagent ions. This sheath flow was then introduced into a co-axial
 155 laminar flow reactor concentric to the sample flow. Nitrate ions were pushed to the
 156 middle of the sample flow under an electric field and subsequently charged sample
 157 molecules. For example, the atmospheric H₂SO₄ molecules would be charged by nitrate
 158 reagent ion NO₃⁻(HNO₃)₀₋₂ and mainly produce HSO₄⁻ ions (m/z = 96.9601 Th),
 159 HSO₄⁻·HNO₃ ions (m/z = 159.9557 Th), and HSO₄⁻·(HNO₃)₂ ions (m/z = 222.9514 Th).
 160 In addition, HSO₄⁻·H₂SO₄ ions (m/z = 194.9275 Th) were formed from ion-induced
 161 clustering of neutral sulfuric acid and bisulfate ions within the LToF-CIMS ion reaction
 162 zone, and also from the evaporation of dimethylamine (DMA) and the replacement of
 163 one molecule of H₂SO₄ with one bisulfate ion, HSO₄⁻, during the NO₃⁻ reagent ion
 164 charging of a stabilized neutral sulfuric acid dimer in the real atmosphere in presence
 165 of DMA or a molecule that works in the same way as DMA. During the campaign, the
 166 sample flow rate was kept at 8.8 slpm, since mass flow controllers fixed the sheath flow
 167 rate and the excess flow rate, and the flow into the mass spectrometer (around 0.8 slpm)
 168 was fixed by the size of a pinhole between the ionization source and the mass
 169 spectrometer. The concentration of gaseous sulfuric acid was then determined by Eq.
 170 (2).

$$[H_2SO_4] = \frac{HSO_4^- \cdot (HNO_3)_{0-2} + HSO_4^- \cdot H_2SO_4}{NO_3^-(HNO_3)_{0-2}} \cdot C \quad (2)$$

171
 172
 173 where C is the calibration coefficient, and NO₃⁻(HNO₃)₀₋₂, HSO₄⁻·(HNO₃)₀₋₂ and
 174 HSO₄⁻·H₂SO₄ represent the signals of corresponding ions and are in units of counts per
 175 second (cps). The unit of resulting [H₂SO₄] is molecule cm⁻³. The CIMS was calibrated
 176 during the campaign with a home-made calibration box that can produce adjustable
 177 concentrations of gaseous sulfuric acid from SO₂ and OH radicals following the
 178 protocols in previous literatures (Kürten et al., 2012; Zheng et al., 2015). We obtain a
 179 calibration coefficient of 3.79 × 10⁹ molecule cm⁻³ for our instrument and use
 180 1.1 × 10¹⁰ molecule cm⁻³ as the effective calibration coefficient, after taking into
 181 account the diffusion losses in the stainless-steel tube and the nitrate chemical
 182

183 ionization source. The obtained mass spectra were analyzed with a tofTools package
184 based on the MATLAB software (Junninen et al., 2010).

185 Ambient particle number size distributions down to about 1 nm were measured
186 using a combination of a scanning mobility particle sizer spectrometer (SMPS)
187 equipped with a diethylene glycol-based condensation particle counter (DEG-CPC, ~1-
188 10 nm) and a conventional particle size distribution system (PSD, ~3 nm - 10 μ m)
189 consisting of a pair of aerosol mobility spectrometers developed by Tsinghua
190 University (Cai et al., 2017; Jiang et al., 2011; Liu et al., 2016). The values of CS were
191 calculated following Eq. (3) (Dal Maso et al., 2002):

192

$$193 \quad CS = 2\pi D \int_0^{\infty} D_p \beta_m(D_p) n(D_p) dD_p = 2\pi D \sum_i \beta_i D_{pi} N_i \quad (3)$$

194

195 where D_{pi} is the geometric mean diameter of particles in the size bin i and N_i is the
196 particle number concentration in the corresponding size bin. D is the diffusion
197 coefficient of gaseous sulfuric acid, and β_m represents a transition-regime correction
198 factor dependent on the Knudsen number (Fuchs and Sutugin, 1971; Gopalakrishnan
199 and Hogan Jr., 2011).

200 SO₂, O₃ and NO_x concentrations were measured using a SO₂ analyzer (Model 43i,
201 Thermo, USA), a O₃ analyzer (Model 49i, Thermo, USA) and a NO_x analyzer (Model
202 42i, Thermo, USA) with the detection limits of 0.1 ppbv, 0.5 ppbv and 0.4 ppbv,
203 respectively. The above instruments were pre-calibrated before the campaign. The UVB
204 (280 - 315 nm) intensity (UV-S-B-T, KIPP&ZONEN, The Netherlands) was measured
205 on the rooftop of the building. Atmospheric HONO concentrations were measured by
206 a home-made HONO analyzer with a detection limit of 0.01 ppbv (Tong et al., 2016).

207 Particle number size distributions and concentrations of gaseous sulfuric acid, SO₂,
208 O₃, NO_x and HONO were recorded with a time resolution of 5 min, and the UVB
209 intensity with time resolution of 1 min. A linear interpolation method was used for
210 deriving the variables with the same time intervals, *i.e.*, 5 min. Only data between local
211 sunrise and sunset were used in the subsequent analysis.

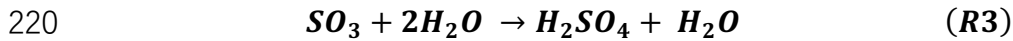
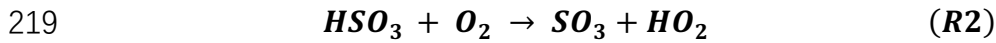
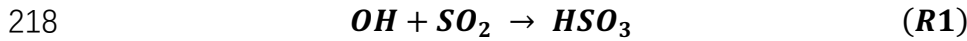
212

213 **3 Development of a proxy for atmospheric gaseous sulfuric acid**

214 We derived the gaseous sulfuric acid concentration proxy on the basis of currently
215 accepted formation pathways of sulfuric acid in the atmosphere (R1-R3) (Finlayson-

216 Pitts and Pitts, 2000; Stockwell and Calvert, 1983):

217



221

222 The reaction (R1) is the rate-limiting step of this formation pathway (Finlayson-Pitts
223 and Pitts, 2000), so our proxy will consider the two major processes that determine the
224 abundance of gaseous sulfuric acid: the formation of gaseous sulfuric acid from
225 reactions between SO₂ and OH radicals, and the loss of gaseous sulfuric acid due to its
226 condensation onto pre-existing particles (Dal Maso et al., 2002; Kulmala et al., 2012;
227 Pirjola et al., 1999).

228 The rate of change of sulfuric acid concentration can be written as Eq. (4)
229 (Mikkonen et al., 2011):

230

$$231 \quad d[\text{H}_2\text{SO}_4]/dt = k \cdot [\text{OH}] \cdot [\text{SO}_2] - [\text{H}_2\text{SO}_4] \cdot CS \quad (4)$$

232

233 where k is a temperature-dependent reaction constant given by Eq. (5) (DeMore et al.,
234 1997; Mikkonen et al., 2011).

235

$$236 \quad k = \frac{A \cdot k_3}{(A + k_3)} \cdot \exp \left\{ k_5 \cdot \left[1 + \log_{10} \left(\frac{A}{k_3} \right)^2 \right]^{-1} \right\} \text{ cm}^3(\text{molecule} \cdot \text{s})^{-1} \quad (5)$$

237

238 where $A = k_1 \cdot [M] \cdot \left(\frac{300}{T}\right)^{k_2}$, $[M]$ represents the density of the air in molecule cm⁻³
239 as calculated by $0.101 \cdot (1.381 \cdot 10^{-23} \cdot T)^{-1}$, $k_1 = 4 \cdot 10^{-31}$, $k_2 = 3.3$, $k_3 = 2 \cdot$
240 10^{-12} and $k_5 = -0.8$.

241 To simplify the calculation, the production and loss of sulfuric acid can be assumed
242 to be at pseudo steady-state (Mikkonen et al., 2011; Petäjä et al., 2009). Then the
243 sulfuric acid concentration can be written as Eq. (6).

244

$$245 \quad [\text{H}_2\text{SO}_4] = k \cdot [\text{OH}] \cdot [\text{SO}_2] \cdot CS^{-1} \quad (6)$$

246

247 Atmospheric OH radical measurements represent a major challenge as well. Since

248 previous studies suggest that the OH radical concentration is strongly correlated with
 249 the intensity of UVB, [OH] could be replaced with UVB intensity in the proxy equation
 250 (Petäjä et al., 2009; Rohrer and Berresheim, 2006). Although photolysis of O₃
 251 ($\lambda < 320 \text{ nm}$) and subsequent reactions with H₂O are considered to be the dominant
 252 source of OH radicals in the atmosphere (Logan et al., 1981), recent studies argue that
 253 photolysis of HONO ($\lambda < 400 \text{ nm}$) is a potentially important OH radical formation
 254 pathway (Hendrick et al., 2014; Kleffmann, 2007; Su et al., 2011; Villena et al., 2011).
 255 Thus, we attempt to introduce both O₃ and HONO into the proxy equation and evaluate
 256 their effects on the concentration of OH radicals.

257 In practice, the exponents for variables in nonlinear fitting procedures are rarely
 258 equal to 1 (Mikkonen et al., 2011), so we replaced the factors x_i with $x_i^{w_i}$ in the proxy,
 259 where x_i can be an atmospheric variable such as UVB and [SO₂], and w_i defines x_i '
 260 weight in the proxy. Since k is a temperature-dependent reaction constant and varies
 261 within a 10 % range in the atmosphere temperature range of 267.6 - 292.6 K, *i.e.*, the
 262 actual atmospheric temperature variation in this study, we approximately regard k as a
 263 constant and use a new scaling factor k_0 . This methodology has been used previously
 264 in the proxies of gaseous sulfuric acid in Hyytiälä, Southern Finland (Petäjä et al., 2009).
 265 As a result, the general proxy equation can be written as Eq. (7), with the UVB intensity,
 266 [SO₂], condensation sink (CS), [O₃], and [HONO] (or [NO_x]) as predictor variables:

$$267$$

$$268 \quad [H_2SO_4] = f(k_0, x_i^{w_i}), \quad x_i = UVB, [SO_2], CS, [O_3], [HONO] \dots \quad (7)$$

$$269$$

270 The nonlinear curve-fitting procedures using iterative least square estimation for
 271 the proxies of gaseous sulfuric acid concentration based on Eq. (7) were performed by
 272 a custom-made MATLAB software. In addition to the correlation coefficient (R),
 273 relative error (RE) is used to evaluate the performance of proxies in the statistical
 274 analysis and can be written as Eq. (8).

$$275$$

$$276 \quad RE = \frac{1}{n} \cdot \sum_{i=1}^n \frac{|[H_2SO_4]_{proxy,i} - [H_2SO_4]_{meas.,i}|}{[H_2SO_4]_{meas.,i}} \quad (8)$$

$$277$$

278 **4 Results and discussion**

279 **4.1 General Characteristics of daytime sulfuric acid and atmospheric parameters**

280 Table 1 summarizes the mean, median and 5-95 % percentiles of gaseous sulfuric

281 acid concentrations and other variables measured during the daytime of the campaign.
282 The 5-95 % percentile ranges of the UVB intensity, [SO₂], [NO_x] and [O₃] were 0-0.45
283 W m⁻², 0.9-11.4 ppbv, 3.3-61.4 ppbv and 3.5-23.3 ppbv, respectively. Compared with
284 the sites in the study by Mikkonen et al. (2011), Beijing was characterized with a factor
285 of 1.4-13.1 higher mean [SO₂] but a factor of 3.4-5.4 lower mean [O₃]. The 5-95 %
286 percentile range of CS in Beijing was 0.01-0.24 s⁻¹, which is about 10-100 times higher
287 than corresponding value ranges in Europe and North America. The concentration of
288 gaseous sulfuric acid during this campaign was (2.2 – 10.0) × 10⁶ molecule cm⁻³
289 in the 5-95 % percentile range, relatively similar to observed elsewhere around the
290 world. A diurnal mean concentration of 0.74 ppbv for HONO was observed in this
291 campaign, consistent with previous long-term HONO measurements of about 0.48-1.8
292 ppbv (averaged values) in winter in Beijing (Hendrick et al., 2014; Spataro et al., 2013;
293 Wang et al., 2017), which is a factor of 4-10 higher than HONO concentrations
294 measured in Europe (Alicke et al., 2002, 2003). In addition, Beijing is dry in winter
295 with a mean ambient relative humidity of 28 % during the campaign.

296

297 **4.2 Correlations between [H₂SO₄] and atmospheric variables**

298 Table 2 summarizes the correlation coefficients between [H₂SO₄] and atmospheric
299 variables using a Spearman-type correlation analysis. Clearly, the UVB intensity is an
300 isolated variable that is independent of all the other variables but that imposes a positive
301 influence on O₃ because of photochemical formation of ozone, and a negative influence
302 on HONO because of HONO's photochemical degradation. The sulfuric acid
303 concentration shows positive correlations with all the other variables. The correlation
304 coefficients between [H₂SO₄] and [SO₂] and between [H₂SO₄] and UVB intensity are
305 0.74 and 0.46, respectively, which indicate that [SO₂] and UVB have important
306 influences on the formation of atmospheric gaseous sulfuric acid. Accordingly, [O₃]
307 and [HONO] show positive correlations with [H₂SO₄] because both O₃ and HONO
308 could be precursors of OH radicals. Surprisingly, a high positive correlation coefficient
309 (0.6) was found between [H₂SO₄] and CS, which is in contrast to the conventional
310 thought that CS describes the loss of gaseous sulfuric acid molecules onto pre-existing
311 particles and thus should show a negative correlation. CS correlates well with [SO₂] (*r*
312 = 0.83) and [NO_x] (*r* = 0.77): a high CS value, as an indicator of atmospheric particle
313 pollution, is thus usually accompanied with a high concentration of both SO₂ and NO_x
314 in urban China, indicating co-emissions. A strong correlation between [HONO] and

315 [NO_x] ($r = 0.88$) in our measurement is supported by the fact that HONO can be either
316 heterogeneously formed by reactions of NO₂ on various surfaces (Calvert et al., 1994)
317 or homogeneously formed by the gas phase NO + OH reaction, between which the
318 former likely dominate for the daytime HONO production in urban Beijing (Liu et al.,
319 2014).

320 Since the UVB intensity and [SO₂] have been reported as the dominating factors
321 for the formation of sulfuric acid (Mikkonen et al., 2011; Petäjä et al., 2009), we further
322 explored the relationship of the measured sulfuric acid concentrations with the UVB
323 intensity and [SO₂] using the nonlinear curve-fitting method with a single variable.
324 Figure 1a presents a scatter plot of [H₂SO₄] against the UVB intensity, color-coded by
325 [SO₂]. A good correlation layering with [SO₂] is evident, indicating that the UVB
326 intensity and [SO₂] together play an important role in the formation of sulfuric acid. A
327 similar scatter plot (Figure 1b) of [H₂SO₄] against [SO₂], color-coded by the UVB
328 intensity, leads to a similar conclusion.

329

330 **4.3 Proxy construction**

331 Similar to the non-linear proxies suggested by Mikkonen et al. (2011), we tested a
332 number of proxies for gaseous sulfuric acid, listed in Table 3 with their respective fitting
333 parameters and performance summarized in Table 4. The scatter plots of observed
334 [H₂SO₄] *versus* predicted values given by proxies are presented in Fig. S1. In these
335 proxies, the concentration of a gaseous species is in the unit of molecule cm⁻³, the unit
336 of the UVB intensity is W m⁻², the unit of CS is s⁻¹, and k_0 is a scaling factor.

337 The proxy N1 was built by using the UVB intensity and [SO₂] as the source terms
338 and CS as the sink term, which follows the conventional idea of the H₂SO₄ formation
339 and loss in the atmosphere. CS was then removed from this proxy to examine the
340 performance of the proxy N2 that has the UVB intensity and [SO₂] as the only predictor
341 variables. Since the formation of OH radicals in the atmosphere depends on precursors
342 in addition to UVB, we further attempted to introduce the OH precursor term into the
343 H₂SO₄ proxy. The proxies N3 and N4 were built by introducing O₃ as the only OH
344 precursor to evaluate its influence on the formation of sulfuric acid. Furthermore, we
345 added HONO as another potential precursor for OH radicals, resulting in the proxies
346 N5 and N6. Lastly, the proxy N7 was built by replacing [HONO] with [NO_x] because
347 firstly, HONO is not regularly measured, and secondly, a good linear correlation
348 between [HONO] and [NO_x] was generally observed in the daytime during this

349 campaign, although higher [HONO]/[NO_x] ratios were observed in the morning due to
350 the accumulation of HONO during the night (Figure 2). RH was not considered in the
351 current study because a test by introducing RH into the proxies do not result in a
352 significantly better performance, which is consistent with those conclusions in the
353 Mikkonen et al. study (2011).

354 As shown in Table 4, the correlation coefficients are in the range of 0.83-0.86 and
355 REs are in the range of 19.1-20.0 %. The exponents for the UVB intensity range from
356 0.13 to 0.15, and those for [SO₂] generally range from 0.38 to 0.41, except in case of
357 the proxy N6 ($b=0.33$). The obtained exponent b for [SO₂] is significantly smaller than
358 1 unlike the assumption in Eq. (6), mainly because [SO₂] is also an indicator of air
359 pollution that usually influences the sinks of both OH radicals and sulfuric acid. The
360 exponent for [SO₂] ranged from 0.5 to 1.04 in the previous proxy study for European
361 and North American sites (Mikkonen et al., 2011), including values from 0.48 to 0.69
362 in Atlanta, GA, USA, which was probably quite a polluted site because the
363 measurements were conducted only 9 km away from a coal-fired power plant. The
364 obtained value range of the exponent b for [SO₂] in our study is probably related to
365 the urban nature of Beijing. The value of exponent c for CS in the proxy N1 is as low
366 as 0.03, which either might be due to the covariance of CS and certain H₂SO₄ sources
367 that cancels the dependence on CS, or it might indicate that CS is actually insufficient
368 in regulating the H₂SO₄ concentration, as recently suggested by Kulmala et al. (2017).
369 By comparing the proxies N1 and N2, we can see that CS plays a minor role because
370 the exponents of [SO₂] and UVB, the overall correlation coefficient and the REs are
371 almost identical with and without CS. We can see the negligible role of CS also when
372 comparing the results of the proxies N3 and N4 where O₃ is considered. However, the
373 role of CS becomes evident between the proxies N5 and N6 when HONO is introduced:
374 the exponents of [SO₂], [O₃], and [HONO] significantly increased when taking into
375 account the CS, suggesting that the covariance between HONO and CS can explain, at
376 least partially, the close-to-zero exponent of CS in the proxies N1-N4. In addition, when
377 [O₃] is introduced as the only precursor for OH radicals, minor improvements in the
378 correlation coefficient and RE were obtained, as suggested by comparing the proxies
379 N3 and N1. When both [O₃] and [HONO] were introduced as OH precursors in the
380 proxies N5-N7, REs have noticeable improvements, and correlation coefficients
381 improved slightly. Altogether, these observations suggest that it is crucial to introduce
382 HONO into the proxy, both in our study and also likely for the previous work where

383 the exponent of CS is close-to-zero (Mikkonen et al., 2011).

384 Although so far the proxy N5 had the best fitting quality, it is impractical to
385 explicitly include [HONO] because HONO measurements are very challenging. As
386 shown in Fig. 2, [HONO] and [NO_x] tended to correlate linearly with each other in the
387 daytime during this campaign, with a linearly fitted [HONO]/[NO_x] ratio of around 0.03
388 and a relative error of 0.42. Occasionally higher [HONO]/[NO_x] ratios could be seen in
389 the morning, which might be due to the fact that HONO concentration could have an
390 accumulation process during the nighttime and lead to a deviation from the steady state.
391 Therefore, due to the good correlation, the proxy N7 replaces [HONO] by [NO_x], a
392 more easily measured variable, and performs equally well with the proxy N5.

393 Clearly, the proxy N2 provides the simplest parameterization, but the proxies N5
394 and N7 result in the best fitting quality because of the introduction of [HONO]. Figure
395 3 presents the RE values for the proxies N2 and N7, respectively, as a function of linear
396 bins of measured sulfuric acid concentrations. The performance of the proxy N7 is
397 considerably better than that of the proxy N2 in the sulfuric acid concentration range of
398 $(2.2 - 10) \times 10^6$ molecule cm⁻³, which covers the 5-95% percentiles of sulfuric acid
399 concentration in this study. In the worst scenario, RE of proxy N2 is 1.2 times as high
400 as that of proxy N7, *e.g.*, REs are 16.75 % and 13.99 %, respectively, in the sulfuric
401 acid concentration bin of $(4 - 5) \times 10^6$ molecule cm⁻³, and 16.71 % and 14.42 %,
402 respectively, in the bin of $(7 - 8) \times 10^6$ molecule cm⁻³.

403

404 **4.4 Comparison of measured and predicted [H₂SO₄]**

405 A comparison between measured and predicted [H₂SO₄] was performed. Figure 4
406 includes calculated results from the proxies N2 and N7 as well as from a proxy
407 constructed according to measurement in a boreal forest site, Finland, *i.e.*, Eq (1) (Petäjä
408 et al., 2009). The measured daytime [H₂SO₄] on 10 March, 2018, was above 4×10^6
409 molecules cm⁻³ when averaged to a time resolution of 5 min. The predicted [H₂SO₄]
410 using the proxies N2 and N7 both track the measured [H₂SO₄] pretty well, even when
411 an unexpected dip in the sulfuric acid concentration was observed at around 10:00-
412 11:00. The performance of the proxy N7 is better than that of proxy N2 during the entire
413 day, consistent with our results in Fig. 3. The proxy by Petäjä et al. (2009)
414 underestimated the concentrations of sulfuric acid by a factor of 20 or so, which can be
415 attributed to the very different values of CS between Beijing and the boreal forest. The
416 fact that $[H_2SO_4]_{\text{Petäjä et al.}}$ does not track the measured [H₂SO₄] even after including

417 a scaling factor indicates that proxies are site-specific and do not necessarily work well
 418 in locations other than where they have originally been developed for. In addition, the
 419 direct performance comparison between the proxy N2 and the proxy by Petäjä et al.
 420 (2009) indicates the importance of assigning exponential weights to variables in the
 421 nonlinear fitting procedures, which is consistent with results by Mikkonen et al. (2011).

422

423 **5 Summary and conclusions**

424 Sulfuric acid is a key precursor for atmospheric new particle formation. In this
 425 study, we constructed a number of proxies for gaseous sulfuric acid concentration
 426 according to our measurements in urban Beijing during the winter. According to the
 427 obtained proxies and their performance, the UVB intensity and [SO₂] were the
 428 dominant influencing factors. Hence, the simplest proxy (Proxy N2) only involves
 429 UVB intensity and [SO₂] as shown by Eq. (9). The units of [H₂SO₄] and [SO₂] are
 430 molecule cm⁻³, and the unit of UVB is W m⁻².

432

$$431 \quad [H_2SO_4] = 280.05 \cdot UVB^{0.14} \cdot [SO_2]^{0.40} \quad (9)$$

433

434 For a comprehensive consideration of the formation pathways of OH radicals, [O₃]
 435 and [HONO] as well as CS should be included (Proxy N5), as shown by Eq. (10). The
 436 units of [H₂SO₄], [SO₂], [O₃] and [HONO] are molecule cm⁻³, the unit of UVB is W m⁻²,
 437 and the unit of CS is s⁻¹.

438

$$439 \quad [H_2SO_4] = 0.0072 \cdot UVB^{0.15} \cdot [SO_2]^{0.41} \cdot CS^{-0.17} \cdot ([O_3]^{0.36} \\ 440 \quad + [HONO]^{0.38}) \quad (10)$$

441

442 Since HONO measurements are not a regular practice, we can further replace [HONO]
 443 with [NO_x], shown in Eq. (11), which can be justified by the strong linear correlation
 444 between [HONO] and [NO_x] observed in this study. The units of [H₂SO₄], [SO₂], [O₃]
 445 and [NO_x] are molecule cm⁻³, the unit of UVB is W m⁻², and the unit of CS is s⁻¹.

446

$$447 \quad [H_2SO_4] = 0.0013 \cdot UVB^{0.13} \cdot [SO_2]^{0.40} \cdot CS^{-0.17} \cdot ([O_3]^{0.44} \\ 448 \quad + [NO_x]^{0.41}) \quad (11)$$

449

450 We consider this last proxy more reasonable than the others due to the following reasons:

451 first, it makes the equation physically meaningful as the CS starts to be involved as a
452 sink term, and second, the RE was reduced considerably compared with the other
453 proxies. Overall, this suggests that the photolysis of O₃ and HONO are both important
454 OH sources in urban Beijing.

455 As a summary, we recommend using the simplest proxy (proxy N2 as shown in
456 Eq. (9)) and a more accurate proxy (Proxy N7 as shown in Eq. (11)) for calculating
457 daytime gaseous sulfuric acid concentrations in the urban Beijing atmosphere. It is clear
458 that the current proxies are based on only a month-long campaign of sulfuric acid
459 measurements in urban Beijing during winter. Given the dramatic reduction in the
460 concentration of SO₂ in recent years (Wang et al., 2018) and the strong dependence of
461 calculated [H₂SO₄] on [SO₂], the performance of the proxies in the past and future years
462 remain to be evaluated. Furthermore, the proxies might be site-specific and season-
463 specific. Since the proxies were derived with atmospheric parameters in winter, in
464 urban Beijing, the exponents for atmospheric variables in the proxy could have different
465 values for other cities or other seasons. Thus, the proxies in this study should be further
466 tested before their application to other Chinese megacities or other seasons.

467

468 **Author contributions**

469 LW designed this study. YL (Yiqun Lu), CY, YF, YC, YL (Yiliang Liu), GY, YW, YZ, RY, RB
470 and CD conducted the field campaign. YL (Yiqun Lu) analyzed data with contributions from
471 LW and all the other co-authors. YL (Yiqun Lu) and LW wrote the manuscript with
472 contributions from all the other co-authors.

473

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Table 1 Mean, median, 5-95 % percentiles of key atmospheric variables and [H₂SO₄] in the daytime.

	UVB (W m ⁻²)	[SO ₂] (ppbv)	CS (s ⁻¹)	[O ₃] (ppbv)	[HONO] (ppbv)	[NO _x] (ppbv)	[H ₂ SO ₄] (× 10 ⁶ molecule cm ⁻³)	RH (%)
mean	0.17	4.6	0.11	10.5	0.74	25.3	5.4	28
median	0.14	3.7	0.11	9.0	0.51	23.0	4.9	26
5-95% percentiles	0.00-0.45	0.9-11.4	0.01-0.24	3.5-23.3	0.09-2.65	3.3-61.4	2.2-10.0	9-59

Table 2 Correlation coefficients (Spearman type) between [H₂SO₄] and atmospheric variables in the daytime.

	UVB	[SO ₂]	CS	[O ₃]	[HONO]	[NO _x]	[H ₂ SO ₄]
UVB	1	0.01	-0.02	0.14	-0.23	-0.04	0.46
[SO ₂]		1	0.83	0.25	0.64	0.70	0.74
CS			1	0.36	0.75	0.77	0.60
[O ₃]				1	-0.02	-0.04	0.29
[HONO]					1	0.88	0.39
[NO _x]						1	0.53
[H ₂ SO ₄]							1

Table 3 Proxy functions for the nonlinear fitting procedure.

Proxy	Function [#]
N1	$k_0 \cdot UVB^a \cdot [SO_2]^b \cdot CS^c$
N2	$k_0 \cdot UVB^a \cdot [SO_2]^b$
N3	$k_0 \cdot UVB^a \cdot [SO_2]^b \cdot CS^c \cdot [O_3]^d$
N4	$k_0 \cdot UVB^a \cdot [SO_2]^b \cdot [O_3]^d$
N5	$k_0 \cdot UVB^a \cdot [SO_2]^b \cdot CS^c \cdot ([O_3]^d + [HONO]^e)$
N6	$k_0 \cdot UVB^a \cdot [SO_2]^b \cdot ([O_3]^d + [HONO]^e)$
N7	$k_0 \cdot UVB^a \cdot [SO_2]^b \cdot CS^c \cdot ([O_3]^d + [NO_x]^f)$

[#]UVB is the intensity of ultraviolet radiation in $W\ m^{-2}$; $[SO_2]$ is the concentration of sulfur dioxide in molecule cm^{-3} ; CS is the condensation sink in s^{-1} ; $[O_3]$ is the concentration of ozone in molecule cm^{-3} ; $[HONO]$ is the concentration of nitrous acid in molecule cm^{-3} ; $[NO_x]$ is the concentration of nitrogen dioxide in molecule cm^{-3} ; k_0 is a scaling factor.

Table 4 Results of the nonlinear fitting procedure for different proxy functions, together with correlation coefficient (R , Pearson type) and relative error (RE).

Proxy	k_0	a	b	c	d	e	f	R	RE (%)
N1	515.74	0.14	0.38	0.03				0.83	20.04
N2	280.05	0.14	0.40					0.83	20.00
N3	9.95	0.13	0.39	-0.01	0.14			0.85	19.95
N4	14.38	0.13	0.38		0.14			0.85	19.95
N5	0.0072	0.15	0.41	-0.17	0.36	0.38		0.86	19.11
N6	2.38	0.14	0.33		0.24	0.24		0.85	19.66
N7	0.0013	0.13	0.40	-0.17	0.44		0.41	0.86	19.34

Figure Captions

Figure 1. Correlations (a) between $[\text{H}_2\text{SO}_4]$ and UVB intensity, and (b) between $[\text{H}_2\text{SO}_4]$ and $[\text{SO}_2]$ during the campaign from 9 February to 14 March, 2018.. k_m is a constant term.

Figure 2. Correlation between $[\text{HONO}]$ and $[\text{NO}_x]$ during the campaign from 9 February to 14 March, 2018. The black line represents a linear fitting with a zero intercept.

Figure 3. Performance assessments of proxy N2 and proxy N7. The REs are used to evaluate the performances of proxy N2 and N7, respectively as a function of linear bins of measured sulfuric acid concentrations.

Figure 4. Comparison of measured $[\text{H}_2\text{SO}_4]$, $[\text{H}_2\text{SO}_4]_{\text{N2}}$, $[\text{H}_2\text{SO}_4]_{\text{N7}}$ and $[\text{H}_2\text{SO}_4]_{\text{Petäjä et al.}}$ on 10 March, 2018 with a time resolution of 5 min.

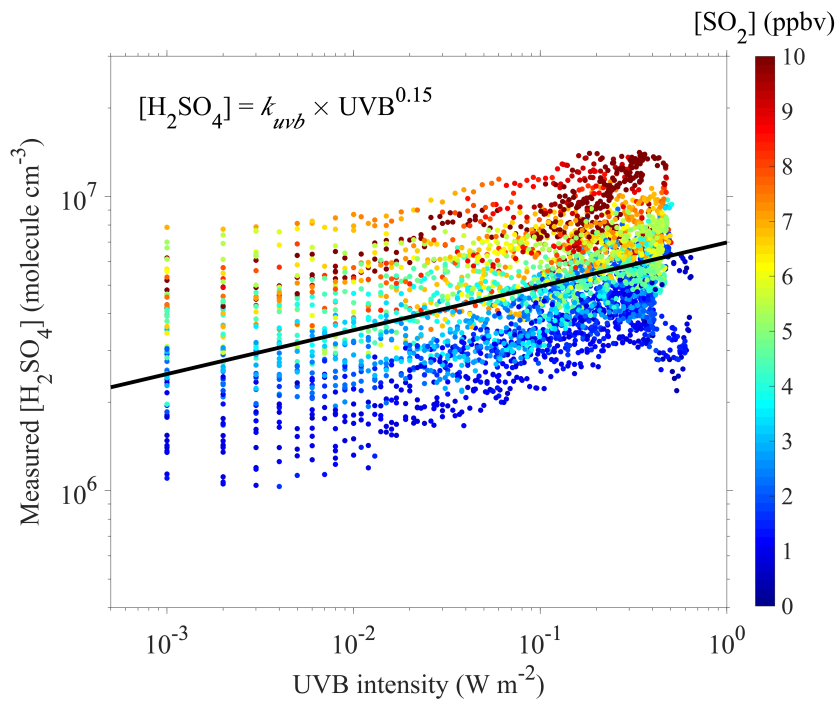


Figure 1a

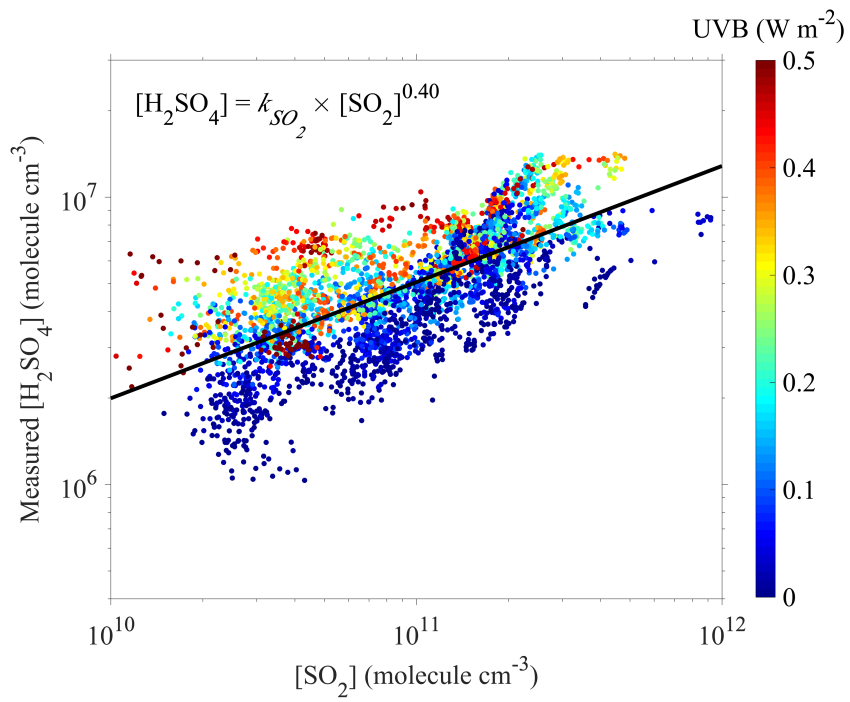


Figure 1b

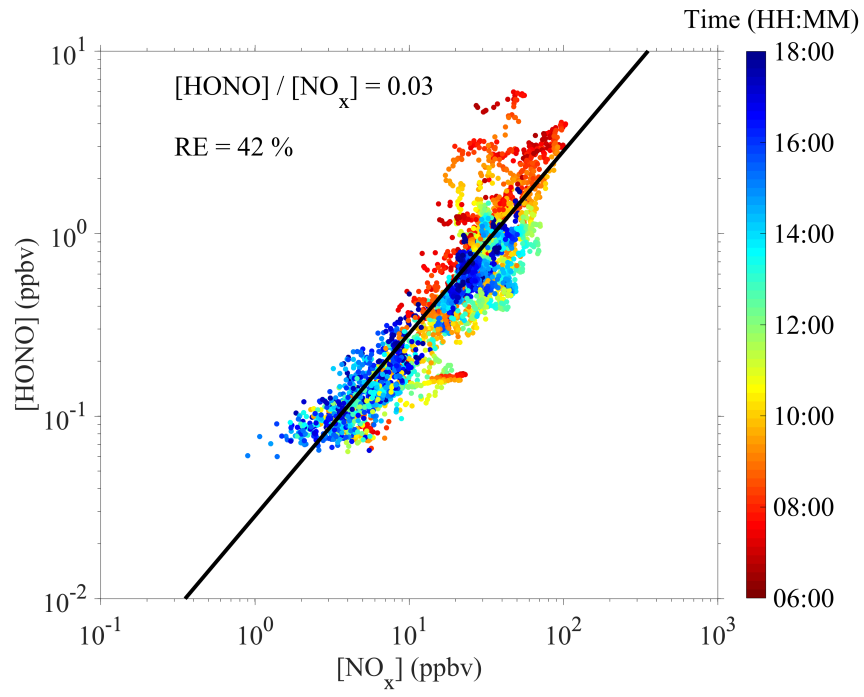


Figure 2

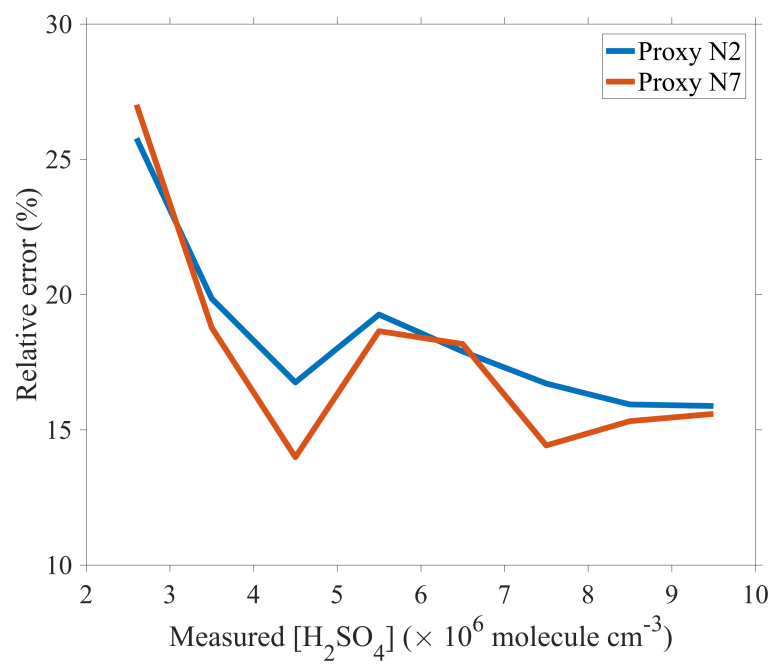


Figure 3

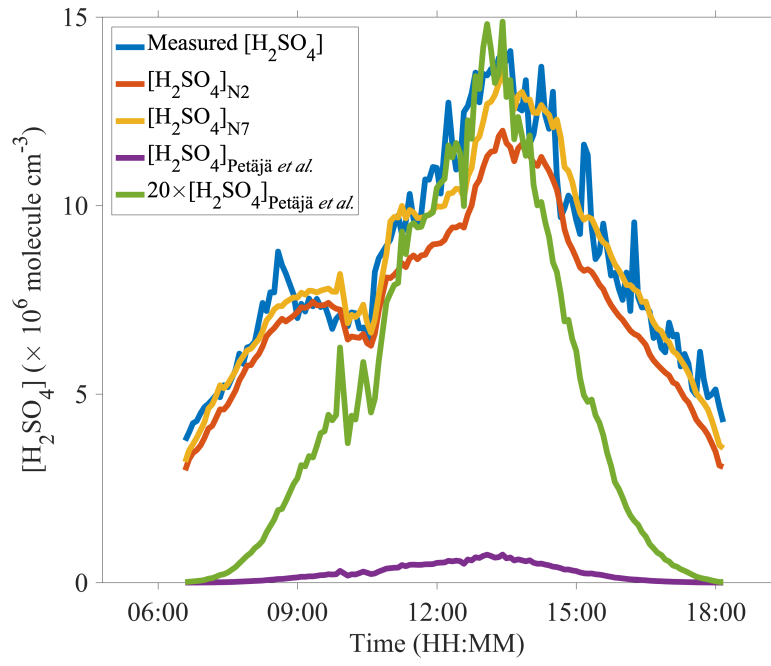


Figure 4