# A proxy for atmospheric daytime gaseous sulfuric acid concentration in urban Beijing

3 Yiqun Lu<sup>1</sup>, Chao Yan<sup>2</sup>, Yueyun Fu<sup>3</sup>, Yan Chen<sup>4</sup>, Yiliang Liu<sup>1</sup>, Gan Yang<sup>1</sup>, Yuwei Wang<sup>1</sup>,

4 Federico Bianchi<sup>2</sup>, Biwu Chu<sup>2</sup>, Ying Zhou<sup>5</sup>, Rujing Yin<sup>3</sup>, Rima Baalbaki<sup>2</sup>, Olga Garmash<sup>2</sup>,

- 5 Chenjuan Deng<sup>3</sup>, Weigang Wang<sup>4</sup>, Yongchun Liu<sup>5</sup>, Tuukka Pet äj ä<sup>2,5,6</sup>, Veli-Matti Kerminen<sup>2</sup>,
- 6 Jingkun Jiang<sup>3</sup>, Markku Kulmala<sup>2,5</sup>, Lin Wang<sup>1,7,8\*</sup>
- 7<sup>1</sup> Shanghai Key Laboratory of Atmospheric Particle Pollution and Prevention (LAP<sup>3</sup>),
- 8 Department of Environmental Science & Engineering, Jiangwan Campus, Fudan University,
  9 Shanghai 200438, China
- 10 <sup>2</sup> Institute for Atmospheric and Earth System Research / Physics, Faculty of Science, University
- 11 of Helsinki, 00014 Helsinki, Finland
- 12<sup>3</sup> State Key Joint Laboratory of Environment Simulation and Pollution Control, School of
- 13 Environment, Tsinghua University, Beijing 100084, China
- <sup>4</sup> Institute of Chemistry, Chinese Academy of Sciences, Beijing 100190, China
- <sup>5</sup> Aerosol and Haze Laboratory, Advanced Innovation Center for Soft Matter Science and
   Engineering, Beijing University of Chemical Technology, Beijing 100029, China
- 17 <sup>6</sup> Joint International Research Laboratory of Atmospheric and Earth System Sciences
- 18 (JirLATEST), School of Atmospheric Sciences, Nanjing University, Nanjing 210023, China
- <sup>7</sup> Institute of Atmospheric Sciences, Jiangwan Campus, Fudan University, Shanghai 200438,
   China
- <sup>8</sup> Shanghai Institute of Pollution Control and Ecological Security, Shanghai 200092, China
- 22 \* Corresponding Author: L.W., email, lin wang@fudan.edu.cn; phone, +86-21-31243568.
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24 Abstract. Gaseous sulfuric acid is known as one of the key precursors for atmospheric 25 new particle formation processes, but its measurement remains a difficulty. A proxy 26 method that is able to derive gaseous sulfuric acid concentrations from parameters that 27 can be measured relatively easily and accurately is therefore highly desirable for the 28 atmospheric chemistry community. Although such methods are available for clean 29 atmospheric environments, a proxy that works well in a polluted atmosphere, such as 30 those in Chinese megacities, is yet to be developed. In this study, the gaseous sulfuric 31 acid concentration was measured in February-March, 2018, in urban Beijing by a nitrate 32 based - Long Time-of-Flight Chemical Ionization Mass Spectrometer (LToF-CIMS). A 33 number of atmospheric parameters were recorded concurrently including the ultraviolet 34 radiation B (UVB) intensity, concentrations of O<sub>3</sub>, NO<sub>x</sub> (sum of NO and NO<sub>2</sub>), SO<sub>2</sub> and HONO, and aerosol particle number size distributions. A proxy for atmospheric 35 daytime gaseous sulfuric acid concentration was derived through a statistical analysis 36

37 method by using the UVB intensity, [SO<sub>2</sub>], condensation sink (CS), [O<sub>3</sub>], and [HONO] 38 (or [NO<sub>x</sub>]) as the predictor variables. In this proxy method, we considered the formation of gaseous sulfuric acid from reactions of SO<sub>2</sub> and OH radicals during the daytime, and 39 40 loss of gaseous sulfuric acid due to its condensation onto the pre-existing particles. In 41 addition, we explored formation of OH radicals from the conventional gas-phase 42 photochemistry using ozone as a proxy and from the photolysis of HONO using HONO (and subsequently  $NO_x$ ) as a proxy. Our results showed that the UVB intensity and [SO<sub>2</sub>] 43 are dominant factors for the production of gaseous sulfuric acid, and that the simplest 44 45 proxy could be constructed with the UVB intensity and [SO<sub>2</sub>] alone. When the OH 46 radical production from both homogenously- and heterogeneously-formed precursors 47 were considered, the relative errors were reduced up to 20 %.

#### 48 1 Introduction

49 Gaseous sulfuric acid (H<sub>2</sub>SO<sub>4</sub>) is a key precursor for atmospheric new particle formation (NPF) processes (Kerminen, 2018; Kirkby et al., 2011; Kuang et al., 2008; 50 51 Kulmala and Kerminen, 2008; Sipilä et al., 2010). A number of atmospheric nucleation 52 mechanisms including H<sub>2</sub>SO<sub>4</sub>-H<sub>2</sub>O binary nucleation (Benson et al., 2008; Duplissy et 53 al., 2016; Kirkby et al., 2011), H<sub>2</sub>SO<sub>4</sub>-NH<sub>3</sub>-H<sub>2</sub>O ternary nucleation (Kirkby et al., 2011; 54 Korhonen et al., 1999; Kürten et al., 2015), and H<sub>2</sub>SO<sub>4</sub>-DMA-H<sub>2</sub>O ternary nucleation 55 (Almeida et al., 2013; Jen et al., 2014; Kürten et al., 2014; Petäjä et al., 2011; Yao et al., 56 2018) involve the participation of gaseous sulfuric acid molecules. In addition, the 57 condensation of gaseous sulfuric acid onto newly-formed particles contributes to their 58 initial growth (Kuang et al., 2012; Kulmala et al., 2013). Quantitative assessments of 59 the contribution of gaseous sulfuric acid to both the new particle formation rates and 60 the particle growth rates require real-time measurements of gaseous sulfuric acid concentrations prior to and during the NPF events (Nieminen et al., 2010; Paasonen et 61 62 al., 2010).

63 Measurements of gaseous sulfuric acid in the lower troposphere are challenging because its ambient concentration is typically quite low  $(10^6 - 10^7 \text{ molecule cm}^{-3})$ 64 (Kerminen et al., 2010; Mikkonen et al., 2011). Reported real-time measurements of 65 gaseous sulfuric acid are currently based on Chemical Ionization Mass Spectrometry 66 67 with NO3<sup>-</sup> and its ligands as reagent ions (nitrate CIMS) because nitrate CIMS with an atmospheric pressure interface (API) has a low detection limit for the atmospheric 68 69 concentration range of gaseous sulfuric acid (Jokinen et al., 2012), and a constant 70 fraction of sulfuric acid present in the air sample will be ionized by excessive nitrate 71 ions in CIMS under constant instrumental conditions (Kürten et al., 2012; Zheng et al., 72 2010), which makes the quantification of gaseous sulfuric acid feasible.

73 Arnold and Fabian (1980) measured the negative ions in the stratosphere and 74 derived the concentration of stratospheric gaseous sulfuric acid from the fractional 75 abundances of a series of stratospheric negative ions as well as the rate constants. Later, 76 real-time measurement of sulfuric acid in the lower troposphere was performed using 77 nitrate CIMS (Eisele and Tanner, 1993), with laboratory calibrations by production of 78 known concentrations of OH radicals that are titrated into gaseous sulfuric acid. 79 Thereafter, measurements of sulfuric acid using CIMS have been performed around the 80 world (e.g., Berresheim et al., 2000; Bianchi et al., 2016; Chen et al., 2012; Jokinen et 81 al., 2012; Kuang et al., 2008; Kürten et al., 2014; Kurtén et al., 2011; Petäjä et al., 2009;

Weber et al., 1997; Zheng et al., 2011), and CIMS has been proven to be a robust tool for gaseous sulfuric acid detection. However, sulfuric acid measurements are still rather sparse because of the high cost of the CIMS instrument and the extensive demand of specialized expertise on the instrument calibration, maintenance, and data processing, etc. Therefore, a proxy for gaseous sulfuric acid concentration is highly desirable.

Proxies for the estimation of atmospheric gaseous sulfuric acid concentrations were previously developed to approximate measurement results of sulfuric acid in Hyyti ä ä Southern Finland (Petäjä et al., 2009), assuming that gaseous sulfuric acid is formed from reactions between  $SO_2$  and OH radicals, and lost due to its condensation onto pre-existing particles. The derived simplest proxy can be written as Eq. (1) below, and the authors recognized that the proxies might be site-specific and should be verified prior to utilization in other environments.

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- $[H_2SO_4] = k \cdot \frac{[SO_2] \cdot (UVB \text{ or Global radiation})}{CS}$ (1)
- Mikkonen et al. (2011) later developed a couple of statistical proxies based on 97 98 measurements of sulfuric acid in six European and North American sites, including urban, rural and forest areas. Their results showed that the radiation intensity and [SO<sub>2</sub>] 99 100 are the most important factors to determine the concentration of sulfuric acid, and that the impact of condensation sink (CS) for gaseous sulfuric acid, is generally negligible. 101 102 In several proxies developed by Mikkonen et al. (2011), the correlation between the 103 gaseous sulfuric acid concentration and CS is positive, which is against what one would 104 expect because a larger CS normally leads to a faster loss for gaseous sulfuric acid. In 105 addition, the performance of a proxy equation is site-specific because of varying 106 atmospheric conditions from one site to another, which implies that the proxy suggested 107 by Mikkonen et al. (2011) might not work well in locations with atmospheric 108 environments different from those in the six sites of that study.

Beijing is a location with typical values of CS (*e.g.*,  $0.01-0.24 \text{ s}^{-1}$  in the 5-95% percentiles in this study) being 10-100 times higher (Herrmann et al., 2014; Wu et al., 2007; Xiao et al., 2015; Yue et al., 2009; Zhang et al., 2011) and typical SO<sub>2</sub> concentrations being 1-10 times higher (Wang et al., 2011a; Wu et al., 2017) than those in Europe and North America (Dunn et al., 2004; Mikkonen et al., 2011), yet measured gaseous sulfuric acid concentrations are relatively similar in these environments (Chen

115 et al., 2012; Smith et al., 2008; Wang et al., 2011b; Zheng et al., 2011). Whether 116 previous proxies developed for European and North American sites work in Beijing remains to be tested. Furthermore, in addition to the gas phase reaction between  $O(^{1}D)$ 117 118 and water molecules (Crutzen and Zimmermann, 1991; Logan et al., 1981), photolysis 119 of HONO could be a potentially important source of OH radicals in the atmosphere in 120 the early morning (Alicke et al., 2002, 2003; Elshorbany et al., 2009; Li et al., 2012) 121 and during the daytime (Acker et al., 2005; Aumont et al., 2003; Kleffmann, 2007). An 122 experimental study measuring HONO near the surface layer estimated that HONO was 123 a main contributor to OH production in Beijing, with HONO's contribution being larger 124 than 70 % at around 12:00-13:00, except for summer when the contribution of O<sub>3</sub> 125 dominated (Hendrick et al., 2014). Given the distinct characteristics of these two OH 126 radical formation pathways, they both should be included and evaluated separately 127 when a proxy for atmospheric gaseous sulfuric acid concentration is being built. The 128 reactions between SO<sub>2</sub> and Criegee intermediates formed from the ozonolysis of 129 atmospheric alkenes could be a potential source of sulfuric acid only in the absence of 130 solar radiation (Boy et al., 2013; Mauldin et al., 2012), so these reactions are expected 131 to provide a minor contribution to the formation of gaseous sulfuric acid during the 132 daytime in urban Beijing.

In this study, gaseous sulfuric acid concentration was measured by a Long Timeof-Flight Chemical Ionization Mass Spectrometer (LToF-CIMS) in February - March, 2018, in urban Beijing. A number of atmospheric parameters were recorded concurrently, including the ultraviolet radiation B (UVB) intensity, concentrations of  $O_3$ , NO<sub>x</sub>, SO<sub>2</sub> and HONO, and particle number size distributions. The objective of this study is to develop a robust daytime gaseous sulfuric acid concentration proxy for Beijing, a representative Chinese megacity with urban atmospheric environments.

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### 141 **2** Ambient measurements

142 An intensive campaign was carried out from 9 February to 14 March, 2018 on the 143 fifth floor of a teaching building in the west campus of Beijing University of Chemical 144 Technology  $(39^{\circ}94' N, 116^{\circ}30' E)$ . This monitoring site is 2 km to the west of the 145 West 3<sup>rd</sup> Ring Road and surrounded by commercial properties and residential dwellings. 146 Hence, this station can be regarded as a representative urban site.

147 The sulfuric acid concentration was measured by a LToF-CIMS (Aerodyne 148 Research, Inc.) equipped with a nitrate chemical ionization source. Ambient air was

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149 drawn into the ionization source through a stainless-steel tube with a length of 1.6 m 150 and a diameter of 3/4 inch. A mixture of a 3 standard cubic centimeter per minute (sccm) 151 ultrahigh purity nitrogen flow containing nitric acid and a 20 standard liter per minute 152 (slpm) pure air flow supplied by a zero-air generator (Aadco 737, USA), together as a 153 sheath flow, was introduced into a PhotoIonizer (Model L9491, Hamamatsu, Japan) to produce nitrate reagent ions. This sheath flow was then introduced into a co-axial 154 155 laminar flow reactor concentric to the sample flow. Nitrate ions were pushed to the 156 middle of the sample flow under an electric field and subsequently charged sample 157 molecules. For example, the atmospheric H<sub>2</sub>SO<sub>4</sub> molecules would be charged by nitrate 158 reagent ion NO<sub>3</sub><sup>-</sup>(HNO<sub>3</sub>)<sub>0-2</sub> and mainly produce HSO<sub>4</sub><sup>-</sup> ions (m/z = 96.9601 Th),  $HSO_4$  ·  $HNO_3$  ions (m/z = 159.9557 Th), and  $HSO_4$  ·  $(HNO_3)_2$  ions (m/z = 222.9514 Th). 159 160 In addition,  $HSO_4 \cdot H_2SO_4$  ions (m/z = 194.9275 Th) were formed from ion-induced 161 clustering of neutral sulfuric acid and bisulfate ions within the LToF-CIMS ion reaction 162 zone, and also from the evaporation of dimethylamine (DMA) and the replacement of 163 one molecule of H<sub>2</sub>SO<sub>4</sub> with one bisulfate ion, HSO<sub>4</sub>, during the NO<sub>3</sub> reagent ion 164 charging of a stabilized neutral sulfuric acid dimer in the real atmosphere in presence 165 of DMA or a molecule that works in the same way as DMA. During the campaign, the 166 sample flow rate was kept at 8.8 slpm, since mass flow controllers fixed the sheath flow 167 rate and the excess flow rate, and the flow into the mass spectrometer (around 0.8 slpm) 168 was fixed by the size of a pinhole between the ionization source and the mass 169 spectrometer. The concentration of gaseous sulfuric acid was then determined by Eq. 170 (2).

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174 where C is the calibration coefficient, and NO3<sup>-</sup>(HNO3)0-2, HSO4<sup>-</sup>(HNO3)0-2 and 175 HSO<sub>4</sub>-H<sub>2</sub>SO<sub>4</sub> represent the signals of corresponding ions and are in units of counts per second (cps). The unit of resulting [H<sub>2</sub>SO<sub>4</sub>] is molecule cm<sup>-3</sup>. The CIMS was calibrated 176 during the campaign with a home-made calibration box that can produce adjustable 177 concentrations of gaseous sulfuric acid from SO2 and OH radicals following the 178 179 protocols in previous literatures (Kürten et al., 2012; Zheng et al., 2015). We obtain a calibration coefficient of  $3.79 \times 10^9$  molecule cm<sup>-3</sup> for our instrument and use  $1.1 \times$ 180  $10^{10}$  molecule cm<sup>-3</sup> as the effective calibration coefficient, after taking into account 181 the diffusion losses in the stainless-steel tube and the nitrate chemical ionization source. 182

 $[H_2SO_4] = \frac{HSO_4^- \cdot (HNO_3)_{0-2} + HSO_4^- \cdot H_2SO_4}{NO_3^- (HNO_3)_{0-2}} \cdot C$ 

(2)

183 The obtained mass spectra were analyzed with a tofTools package based on the184 MATLAB software (Junninen et al., 2010).

Ambient particle number size distributions down to about 1 nm were measured using a combination of a scanning mobility particle sizer spectrometer (SMPS) equipped with a diethylene glycol-based condensation particle counter (DEG-CPC, ~1-10 nm) and a conventional particle size distribution system (PSD, ~3 nm - 10  $\mu$ m) consisting of a pair of aerosol mobility spectrometers developed by Tsinghua University (Cai et al., 2017; Jiang et al., 2011; Liu et al., 2016). The values of CS were calculated following Eq. (3) (Dal Maso et al., 2002):

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$$CS = 2\pi D \int_0^\infty D_p \beta_m(D_p) n(D_p) dD_p = 2\pi D \sum_i \beta_i D_{pi} N_i \qquad (3)$$

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195 where  $D_{pi}$  is the geometric mean diameter of particles in the size bin *i* and  $N_i$  is the 196 particle number concentration in the corresponding size bin. *D* is the diffusion 197 coefficient of gaseous sulfuric acid, and  $\beta_m$  represents a transition-regime correction 198 factor dependent on the Knudsen number (Fuchs and Sutugin, 1971; Gopalakrishnan 199 and Hogan Jr., 2011).

SO<sub>2</sub>, O<sub>3</sub> and NO<sub>x</sub> concentrations were measured using a SO<sub>2</sub> analyzer (Model 43i, Thermo, USA), a O<sub>3</sub> analyzer (Model 49i, Thermo, USA) and a NO<sub>x</sub> analyzer (Model 42i, Thermo, USA) with the detection limits of 0.1 ppbv, 0.5 ppbv and 0.4 ppbv, respectively. The above instruments were pre-calibrated before the campaign. The UVB (280 - 315 nm) intensity (UV-S-B-T, KIPP&ZONEN, The Netherlands) was measured on the rooftop of the building. Atmospheric HONO concentrations were measured by a home-made HONO analyzer with a detection limit of 0.01 ppbv (Tong et al., 2016).

207 Particle number size distributions and concentrations of gaseous sulfuric acid,  $SO_2$ , 208  $O_3$ ,  $NO_x$  and HONO were recorded with a time resolution of 5 min, and the UVB 209 intensity with time resolution of 1 min. A linear interpolation method was used for 210 deriving the variables with the same time intervals, *i.e.*, 5 min. Only data between local 211 sunrise and sunset were used in the subsequent analysis.

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## 213 **3 Development of a proxy for atmospheric gaseous sulfuric acid**

We derived the gaseous sulfuric acid concentration proxy on the basis of currently accepted formation pathways of sulfuric acid in the atmosphere (R1-R3) (Finlayson216 Pitts and Pitts, 2000; Stockwell and Calvert, 1983):

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- 218  $OH + SO_2 \rightarrow HSO_3$  (R1)
- 219  $HSO_3 + O_2 \rightarrow SO_3 + HO_2 \qquad (R2)$

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$$SO_3 + 2H_2O \rightarrow H_2SO_4 + H_2O$$
 (R3)

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The reaction (R1) is the rate-limiting step of this formation pathway (Finlayson-Pitts and Pitts, 2000), so our proxy will consider the two major processes that determine the abundance of gaseous sulfuric acid: the formation of gaseous sulfuric acid from reactions between SO<sub>2</sub> and OH radicals, and the loss of gaseous sulfuric acid due to its condensation onto pre-existing particles (Dal Maso et al., 2002; Kulmala et al., 2012; Pirjola et al., 1999).

The rate of change of sulfuric acid concentration can be written as Eq. (4) (Mikkonen et al., 2011):

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$$d[H_2SO_4]/dt = k \cdot [OH] \cdot [SO_2] - [H_2SO_4] \cdot CS \quad (4)$$

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where *k* is a temperature-dependent reaction constant given by Eq. (5) (DeMore et al.,
1997; Mikkonen et al., 2011).

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$$k = \frac{A \cdot k_3}{(A+k_3)} \cdot exp\left\{k_5 \cdot \left[1 + \log_{10}\left(\frac{A}{k_3}\right)^2\right]^{-1}\right\} \quad cm^3(molecule \cdot s)^{-1} \quad (5)$$

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238 where  $A = k_1 \cdot [M] \cdot (\frac{300}{T})^{k_2}$ , [M] represents the density of the air in molecule cm<sup>-3</sup> 239 as calculated by  $0.101 \cdot (1.381 \cdot 10^{-23} \cdot T)^{-1}$ ,  $k_1 = 4 \cdot 10^{-31}$ ,  $k_2 = 3.3$ ,  $k_3 = 2 \cdot 10^{-12}$  and  $k_5 = -0.8$ .

To simplify the calculation, the production and loss of sulfuric acid can be assumed to be at pseudo steady-state (Mikkonen et al., 2011; Petäjä et al., 2009). Then the sulfuric acid concentration can be written as Eq. (6).

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$$[H_2SO_4] = k \cdot [OH] \cdot [SO_2] \cdot CS^{-1} \quad (6)$$

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Atmospheric OH radical measurements represent a major challenge as well. Since

248 previous studies suggest that the OH radical concentration is strongly correlated with the intensity of UVB, [OH] could be replaced with UVB intensity in the proxy equation 249 (Petäjä et al., 2009; Rohrer and Berresheim, 2006). Although photolysis of O3 250 251  $(\lambda < 320 \text{ nm})$  and subsequent reactions with H<sub>2</sub>O are considered to be the dominant 252 source of OH radicals in the atmosphere (Logan et al., 1981), recent studies argue that photolysis of HONO ( $\lambda < 400$  nm) is a potentially important OH radical formation 253 254 pathway (Hendrick et al., 2014; Kleffmann, 2007; Su et al., 2011; Villena et al., 2011). 255 Thus, we attempt to introduce both O<sub>3</sub> and HONO into the proxy equation and evaluate 256 their effects on the concentration of OH radicals.

257 In practice, the exponents for variables in nonlinear fitting procedures are rarely equal to 1 (Mikkonen et al., 2011), so we replaced the factors  $x_i$  with  $x_i^{w_i}$  in the proxy, 258 where  $x_i$  can be an atmospheric variable such as UVB and [SO<sub>2</sub>], and  $w_i$  defines  $x_i$ ' 259 260 weight in the proxy. Since k is a temperature-dependent reaction constant and varies 261 within a 10 % range in the atmosphere temperature range of 267.6 - 292.6 K, i.e., the 262 actual atmospheric temperature variation in this study, we approximately regard k as a 263 constant and use a new scaling factor  $k_0$ . This methodology has been used previously 264 in the proxies of gaseous sulfuric acid in Hyyti äl ä, Southern Finland (Petäjä et al., 2009). 265 As a result, the general proxy equation can be written as Eq. (7), with the UVB intensity, [SO<sub>2</sub>], condensation sink (CS), [O<sub>3</sub>], and [HONO] (or [NO<sub>x</sub>]) as predictor variables: 266

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$$[H_2SO_4] = f(k_0, x_i^{\omega_i}), \qquad x_i = UVB, [SO_2], CS, [O_3], [HONO] \dots$$
(7)

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The nonlinear curve-fitting procedures using iterative least square estimation for the proxies of gaseous sulfuric acid concentration based on Eq. (7) were performed by a custom-made MATLAB software. In addition to the correlation coefficient (R), relative error (RE) is used to evaluate the performance of proxies in the statistical analysis and can be written as Eq. (8).

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$$RE = \frac{1}{n} \cdot \sum_{i=1}^{n} \frac{|[H_2SO_4]_{proxy,i} - [H_2SO_4]_{meas,i}|}{[H_2SO_4]_{meas,i}}$$
(8)

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278 4 Results and discussion

# 279 4.1 General Characteristics of daytime sulfuric acid and atmospheric parameters

Table 1 summarizes the mean, median and 5-95 % percentiles of gaseous sulfuric

281 acid concentrations and other variables measured during the daytime of the campaign. 282 The 5-95 % percentile ranges of the UVB intensity, [SO<sub>2</sub>], [NO<sub>x</sub>] and [O<sub>3</sub>] were 0-0.45 W m<sup>-2</sup>, 0.9-11.4 ppbv, 3.3-61.4 ppbv and 3.5-23.3 ppbv, respectively. Compared with 283 284 the sites in the study by Mikkonen et al. (2011), Beijing was characterized with a factor of 1.4-13.1 higher mean [SO<sub>2</sub>] but a factor of 3.4-5.4 lower mean [O<sub>3</sub>]. The 5-95 % 285 percentile range of CS in Beijing was 0.01-0.24 s<sup>-1</sup>, which is about 10-100 times higher 286 than corresponding value ranges in Europe and North America. The concentration of 287 gaseous sulfuric acid during this campaign was  $(2.2 - 10.0) \times 10^6$  molecule cm<sup>-3</sup> 288 289 in the 5-95 % percentile range, relatively similar to observed elsewhere around the world. A diurnal mean concentration of 0.74 ppbv for HONO was observed in this 290 291 campaign, consistent with previous long-term HONO measurements of about 0.48-1.8 292 ppbv (averaged values) in winter in Beijing (Hendrick et al., 2014; Spataro et al., 2013; 293 Wang et al., 2017), which is a factor of 4-10 higher than HONO concentrations measured in Europe (Alicke et al., 2002, 2003). In addition, Beijing is dry in winter 294 295 with a mean ambient relative humidity of 28 % during the campaign.

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#### 4.2 Correlations between [H2SO4] and atmospheric variables

298 Table 2 summarizes the correlation coefficients between [H<sub>2</sub>SO<sub>4</sub>] and atmospheric variables using a Spearman-type correlation analysis. Clearly, the UVB intensity is an 299 300 isolated variable that is independent of all the other variables but that imposes a positive 301 influence on O<sub>3</sub> because of photochemical formation of ozone, and a negative influence 302 on HONO because of HONO's photochemical degradation. The sulfuric acid 303 concentration shows positive correlations with all the other variables. The correlation 304 coefficients between [H<sub>2</sub>SO<sub>4</sub>] and [SO<sub>2</sub>] and between [H<sub>2</sub>SO<sub>4</sub>] and UVB intensity are 305 0.74 and 0.46, respectively, which indicate that [SO<sub>2</sub>] and UVB have important 306 influences on the formation of atmospheric gaseous sulfuric acid. Accordingly, [O<sub>3</sub>] and 307 [HONO] show positive correlations with [H<sub>2</sub>SO<sub>4</sub>] because both O<sub>3</sub> and HONO could 308 be precursors of OH radicals. Surprisingly, a high positive correlation coefficient (0.6) 309 was found between [H<sub>2</sub>SO<sub>4</sub>] and CS, which is in contrast to the conventional thought 310 that CS describes the loss of gaseous sulfuric acid molecules onto pre-existing particles 311 and thus should show a negative correlation. CS correlates well with  $[SO_2]$  (r = 0.83) and  $[NO_x]$  (r = 0.77): a high CS value, as an indicator of atmospheric particle pollution, 312 313 is thus usually accompanied with a high concentration of both SO<sub>2</sub> and NO<sub>x</sub> in urban 314 China, indicating co-emissions. A strong correlation between [HONO] and  $[NO_x]$  (r =

0.88 in our measurement is supported by the fact that HONO can be either heterogeneously formed by reactions of NO<sub>2</sub> on various surfaces (Calvert et al., 1994) or homogeneously formed by the gas phase NO + OH reaction, between which the former likely dominate for the daytime HONO production in urban Beijing (Liu et al., 2014).

Since the UVB intensity and [SO<sub>2</sub>] have been reported as the dominating factors 320 321 for the formation of sulfuric acid (Mikkonen et al., 2011; Petäjä et al., 2009), we further 322 explored the relationship of the measured sulfuric acid concentrations with the UVB 323 intensity and [SO<sub>2</sub>] using the nonlinear curve-fitting method with a single variable. 324 Figure 1a presents a scatter plot of [H<sub>2</sub>SO<sub>4</sub>] against the UVB intensity, color-coded by 325 [SO<sub>2</sub>]. A good correlation layering with [SO<sub>2</sub>] is evident, indicating that the UVB 326 intensity and [SO<sub>2</sub>] together play an important role in the formation of sulfuric acid. A 327 similar scatter plot (Figure 1b) of [H<sub>2</sub>SO<sub>4</sub>] against [SO<sub>2</sub>], color-coded by the UVB 328 intensity, leads to a similar conclusion.

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# 330 4.3 Proxy construction

Similar to the non-linear proxies suggested by Mikkonen et al. (2011), we tested a number of proxies for gaseous sulfuric acid, listed in Table 3 with their respective fitting parameters and performance summarized in Table 4. The scatter plots of observed  $[H_2SO_4]$  *versus* predicted values given by proxies are presented in Fig. S1. In these proxies, the concentration of a gaseous species is in the unit of molecule cm<sup>-3</sup>, the unit of the UVB intensity is W m<sup>-2</sup>, the unit of CS is s<sup>-1</sup>, and  $k_0$  is a scaling factor.

337 The proxy N1 was built by using the UVB intensity and [SO<sub>2</sub>] as the source terms 338 and CS as the sink term, which follows the conventional idea of the H<sub>2</sub>SO<sub>4</sub> formation 339 and loss in the atmosphere. CS was then removed from this proxy to examine the 340 performance of the proxy N2 that has the UVB intensity and [SO<sub>2</sub>] as the only predictor 341 variables. Since the formation of OH radicals in the atmosphere depends on precursors 342 in addition to UVB, we further attempted to introduce the OH precursor term into the 343 H<sub>2</sub>SO<sub>4</sub> proxy. The proxies N3 and N4 were built by introducing O<sub>3</sub> as the only OH 344 precursor to evaluate its influence on the formation of sulfuric acid. Furthermore, we 345 added HONO as another potential precursor for OH radicals, resulting in the proxies 346 N5 and N6. Lastly, the proxy N7 was built by replacing [HONO] with [NO<sub>x</sub>] because 347 firstly, HONO is not regularly measured, and secondly, a good linear correlation 348 between [HONO] and [NO<sub>x</sub>] was generally observed in the daytime during this

campaign, although higher  $[HONO]/[NO_x]$  ratios were observed in the morning due to the accumulation of HONO during the night (Figure 2). RH was not considered in the current study because a test by introducing RH into the proxies do not result in a significantly better performance, which is consistent with those conclusions in the Mikkonen et al. study (2011).

354 As shown in Table 4, the correlation coefficients are in the range of 0.83-0.86 and 355 REs are in the range of 19.1-20.0 %. The exponents for the UVB intensity range from 356 0.13 to 0.16, and those for [SO<sub>2</sub>] generally range from 0.38 to 0.41, except in case of 357 the proxy N6 (b=0.33). The obtained exponent b for [SO<sub>2</sub>] is significantly smaller than 358 1 unlike the assumption in Eq. (6), mainly because [SO<sub>2</sub>] is also an indicator of air 359 pollution that usually influences the sinks of both OH radicals and sulfuric acid. The 360 exponent for [SO<sub>2</sub>] ranged from 0.5 to 1.04 in the previous proxy study for European 361 and North American sites (Mikkonen et al., 2011), including values from 0.48 to 0.69 362 in Atlanta, GA, USA, which was probably quite a polluted site because the 363 measurements were conducted only 9 km away from a coal-fired power plant. The 364 obtained value range of the exponent b for [SO<sub>2</sub>] in our study is probably related to 365 the urban nature of Beijing. The value of exponent c for CS in the proxy N1 is as low 366 as 0.03, which either might be due to the covariance of CS and certain H<sub>2</sub>SO<sub>4</sub> sources 367 that cancels the dependence on CS, or it might indicate that CS is actually insufficient 368 in regulating the H<sub>2</sub>SO<sub>4</sub> concentration, as recently suggested by Kulmala et al. (2017). 369 By comparing the proxies N1 and N2, we can see that CS plays a minor role because 370 the exponents of  $[SO_2]$  and UVB, the overall correlation coefficient and the REs are 371 almost identical with and without CS. We can see the negligible role of CS also when comparing the results of the proxies N3 and N4 where O3 is considered. However, the 372 373 role of CS becomes evident between the proxies N5 and N6 when HONO is introduced: 374 the exponents of [SO<sub>2</sub>], [O<sub>3</sub>], and [HONO] significantly increased when taking into 375 account the CS, suggesting that the covariance between HONO and CS can explain, at 376 least partially, the close-to-zero exponent of CS in the proxies N1-N4. In addition, when 377 [O<sub>3</sub>] is introduced as the only precursor for OH radicals, minor improvements in the 378 correlation coefficient and RE were obtained, as suggested by comparing the proxies 379 N3 and N1. When both [O<sub>3</sub>] and [HONO] were introduced as OH precursors in the proxies N5-N7, REs have noticeable improvements, and correlation coefficients 380 381 improved slightly. Altogether, these observations suggest that it is crucial to introduce 382 HONO into the proxy, both in our study and also likely for the previous work where

383 the exponent of CS is close-to-zero (Mikkonen et al., 2011).

384 Although so far the proxy N5 had the best fitting quality, it is impractical to explicitly include [HONO] because HONO measurements are very challenging. As 385 shown in Fig. 2, [HONO] and [NO<sub>x</sub>] tended to correlate linearly with each other in the 386 387 daytime during this campaign, with a linearly fitted [HONO]/[NO<sub>x</sub>] ratio of around 0.03 388 and a relative error of 0.42. Occasionally higher [HONO]/[NO<sub>x</sub>] ratios could be seen in 389 the morning, which might be due to the fact that HONO concentration could have an 390 accumulation process during the nighttime and lead to a deviation from the steady state. 391 Therefore, due to the good correlation, the proxy N7 replaces [HONO] by [NO<sub>x</sub>], a 392 more easily measured variable, and performs equally well with the proxy N5.

393 Clearly, the proxy N2 provides the simplest parameterization, but the proxies N5 394 and N7 result in the best fitting quality because of the introduction of [HONO]. Figure 395 3 presents the RE values for the proxies N2 and N7, respectively, as a function of linear bins of measured sulfuric acid concentrations. The performance of the proxy N7 is 396 397 considerably better than that of the proxy N2 in the sulfuric acid concentration range of  $(2.2 - 10) \times 10^6$  molecule cm<sup>-3</sup>, which covers the 5-95% percentiles of sulfuric acid 398 concentration in this study. In the worst scenario, RE of proxy N2 is 1.2 times as high 399 as that of proxy N7, e.g., REs are 16.75 % and 13.99 %, respectively, in the sulfuric 400 acid concentration bin of  $(4-5) \times 10^6$  molecule cm<sup>-3</sup>, and 16.71 % and 14.42 %, 401 respectively, in the bin of  $(7 - 8) \times 10^6$  molecule cm<sup>-3</sup>. 402

403

# 404 **4.4 Comparison of measured and predicted [H<sub>2</sub>SO<sub>4</sub>]**

405 A comparison between measured and predicted [H<sub>2</sub>SO<sub>4</sub>] was performed. Figure 4 406 includes calculated results from the proxies N2 and N7 as well as from a proxy constructed according to measurement in a boreal forest site, Finland, *i.e.*, Eq (1) (Petäjä 407 et al., 2009). The measured daytime [H<sub>2</sub>SO<sub>4</sub>] on 10 March, 2018, was above 4  $\times 10^6$ 408 molecules  $cm^{-3}$  when averaged to a time resolution of 5 min. The predicted [H<sub>2</sub>SO<sub>4</sub>] 409 410 using the proxies N2 and N7 both track the measured [H<sub>2</sub>SO<sub>4</sub>] pretty well, even when 411 an unexpected dip in the sulfuric acid concentration was observed at around 10:00-11:00. The performance of the proxy N7 is better than that of proxy N2 during the entire 412 413 day, consistent with our results in Fig. 3. The proxy by Petäjä et al. (2009) 414 underestimated the concentrations of sulfuric acid by a factor of 20 or so, which can be 415 attributed to the very different values of CS between Beijing and the boreal forest. The fact that [H<sub>2</sub>SO<sub>4</sub>]<sub>Petäjä et al.</sub> does not track the measured [H<sub>2</sub>SO<sub>4</sub>] even after including 416

417 a scaling factor indicates that proxies are site-specific and do not necessarily work well
418 in locations other than where they have originally been developed for. In addition, the
419 direct performance comparison between the proxy N2 and the proxy by Petäjä et al.
420 (2009) indicates the importance of assigning exponential weights to variables in the

- 421 nonlinear fitting procedures, which is consistent with results by Mikkonen et al. (2011).
- 422

# 423 **5 Summary and conclusions**

Sulfuric acid is a key precursor for atmospheric new particle formation. In this study, we constructed a number of proxies for gaseous sulfuric acid concentration according to our measurements in urban Beijing during the winter. According to the obtained proxies and their performance, the UVB intensity and [SO<sub>2</sub>] were the dominant influencing factors. Hence, the simplest proxy (Proxy N2) only involves UVB intensity and [SO<sub>2</sub>] as shown by Eq. (9). The units of [H<sub>2</sub>SO<sub>4</sub>] and [SO<sub>2</sub>] are molecule cm<sup>-3</sup>, and the unit of UVB is W m<sup>-2</sup>.

432

431

$$[H_2SO_4] = 280.05 \cdot UVB^{0.14} \cdot [SO_2]^{0.40}$$
(9)

433

For a comprehensive consideration of the formation pathways of OH radicals,  $[O_3]$ and [HONO] as well as CS should be included (Proxy N5), as shown by Eq. (10). The units of  $[H_2SO_4]$ ,  $[SO_2]$ ,  $[O_3]$  and [HONO] are molecule cm<sup>-3</sup>, the unit of UVB is W m<sup>-</sup> <sup>2</sup>, and the unit of CS is s<sup>-1</sup>.

- 438
- 439
- 440 441

 $[H_2SO_4] = 0.0072 \cdot UVB^{0.15} \cdot [SO_2]^{0.41} \cdot CS^{-0.17} \cdot ([O_3]^{0.36} + [HONO]^{0.38})$ (10)

- 442 Since HONO measurements are not a regular practice, we can further replace [HONO] 443 with  $[NO_x]$ , shown in Eq. (11), which can be justified by the strong linear correlation 444 between [HONO] and  $[NO_x]$  observed in this study. The units of  $[H_2SO_4]$ ,  $[SO_2]$ ,  $[O_3]$ 445 and  $[NO_x]$  are molecule cm<sup>-3</sup>, the unit of UVB is W m<sup>-2</sup>, and the unit of CS is s<sup>-1</sup>.
- 446

447 
$$[H_2SO_4] = 0.0013 \cdot UVB^{0.13} \cdot [SO_2]^{0.40} \cdot CS^{-0.17} \cdot ([O_3]^{0.44}$$

448 449

450 We consider this last proxy more reasonable than the others due to the following reasons:

+  $[NO_r]^{0.41}$  (11)

451 first, it makes the equation physically meaningful as the CS starts to be involved as a 452 sink term, and second, the RE was reduced considerably compared with the other 453 proxies. Overall, this suggests that the photolysis of  $O_3$  and HONO are both important 454 OH sources in urban Beijing.

455 As a summary, we recommend using the simplest proxy (proxy N2 as shown in 456 Eq. (9)) and a more accurate proxy (Proxy N7 as shown in Eq. (11)) for calculating 457 daytime gaseous sulfuric acid concentrations in the urban Beijing atmosphere. It is clear 458 that the current proxies are based on only a month-long campaign of sulfuric acid 459 measurements in urban Beijing during winter. Given the dramatic reduction in the 460 concentration of SO<sub>2</sub> in recent years (Wang et al., 2018) and the strong dependence of calculated [H<sub>2</sub>SO<sub>4</sub>] on [SO<sub>2</sub>], the performance of the proxies in the past and future years 461 462 remain to be evaluated. Furthermore, the proxies might be site-specific and season-463 specific. Since the proxies were derived with atmospheric parameters in winter, in 464 urban Beijing, the exponents for atmospheric variables in the proxy could have different 465 values for other cities or other seasons. Thus, the proxies in this study should be further 466 tested before their application to other Chinese megacities or other seasons.

467

#### 468 Author contributions

LW designed this study. YL (Yiqun Lu), CY, YF, YC, YL (Yiliang Liu), GY, YW, YZ, RY, RB
and CD conducted the field campaign. YL (Yiqun Lu) analyzed data with contributions from
LW and all the other co-authors. YL (Yiqun Lu) and LW wrote the manuscript with
contributions from all the other co-authors.

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#### 474 Acknowledgement

This study was financially supported by the National Key R&D Program of China
(2017YFC0209505), and the National Natural Science Foundation of China (41575113,
91644213).

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	UVB (W $m^{-2}$ )	[SO <sub>2</sub> ] (ppbv)	$CS (s^{-1})$	[0 <sub>3</sub> ] (ppbv)	[HONO] (ppbv)	$[NO_x]$ (ppbv)	$[H_2SO_4]$ (× 10 <sup>6</sup> molecule cm <sup>-3</sup> )	RH (%)
mean	0.17	4.6	0.11	10.5	0.74	25.3	5.4	28
median	0.14	3.7	0.11	9.0	0.51	23.0	4.9	26
5-95% percentiles	0.00-0.45	0.9-11.4	0.01-0.24	3.5-23.3	0.09-2.65	3.3-61.4	2.2-10.0	9-59

Table 1 Mean, median, 5-95 % percentiles of key atmospheric variables and [H<sub>2</sub>SO<sub>4</sub>] in the daytime.

	UVB	$[SO_2]$	CS	[0 <sub>3</sub> ]	[HONO]	$[NO_2]$	$[H_2SO_4]$
UVB	1	0.01	-0.02	0.14	-0.23	-0.04	0.46
[SO <sub>2</sub> ]		1	0.83	0.25	0.64	0.70	0.74
CS			1	0.36	0.75	0.77	0.60
[0 <sub>3</sub> ]				1	-0.02	-0.04	0.29
[HONO]					1	0.88	0.39
$[NO_2]$						1	0.53
$[H_2SO_4]$							1

Table 2 Correlation coefficients (Spearman type) between  $[H_2SO_4]$  and atmospheric variables in the daytime.

Proxy	Function <sup>#</sup>
N1	$k_0 \cdot UVB^a \cdot [SO_2]^b \cdot CS^c$
N2	$k_0 \cdot UVB^a \cdot [SO_2]^b$
N3	$k_0 \cdot UVB^a \cdot [SO_2]^b \cdot CS^c \cdot [O_3]^d$
N4	$k_0 \cdot UVB^a \cdot [SO_2]^b \cdot [O_3]^d$
N5	$k_0 \cdot UVB^a \cdot [SO_2]^b \cdot CS^c \cdot ([O_3]^d + [HONO]^e)$
N6	$k_0 \cdot UVB^a \cdot [SO_2]^b \cdot ([O_3]^d + [HONO]^e)$
N7	$k_0 \cdot UVB^a \cdot [SO_2]^b \cdot CS^c \cdot ([O_3]^d + [NO_x]^f)$

Table 3 Proxy functions for the nonlinear fitting procedure.

<sup>#</sup>UVB is the intensity of ultraviolet radiation b in W cm<sup>-3</sup>; [SO<sub>2</sub>] is the concentration of sulfur dioxide in molecule cm<sup>-3</sup>; CS is the condensation sink in s<sup>-1</sup>; [O<sub>3</sub>] is the concentration of ozone in molecule cm<sup>-3</sup>; [HONO] is the concentration of nitrous acid in molecule cm<sup>-3</sup>; [NO<sub>x</sub>] is the concentration of nitrogen dioxide in molecule cm<sup>-3</sup>;  $k_0$  is a scaling factor.

Proxy	$k_0$	а	b	С	d	е	f	R	RE (%)
N1	515.74	0.14	0.38	0.03				0.83	20.04
N2	280.05	0.14	0.40					0.83	20.00
N3	9.95	0.13	0.39	-0.01	0.14			0.85	19.95
N4	14.38	0.13	0.38		0.14			0.85	19.95
N5	0.0072	0.15	0.41	-0.17	0.36	0.38		0.86	19.11
N6	2.38	0.14	0.33		0.24	0.24		0.85	19.66
N7	0.0013	0.13	0.40	-0.17	0.44		0.41	0.86	19.34

Table 4 Results of the nonlinear fitting procedure for different proxy functions, together with correlation coefficient (R, Pearson type) and relative error (RE).

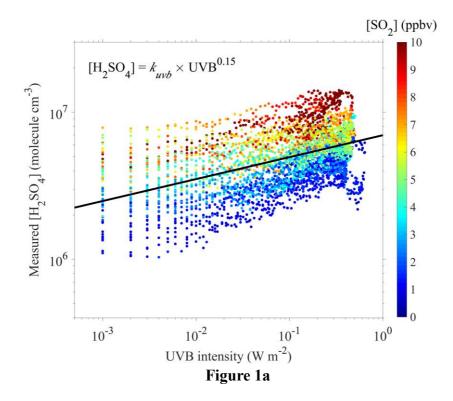
# **Figure Captions**

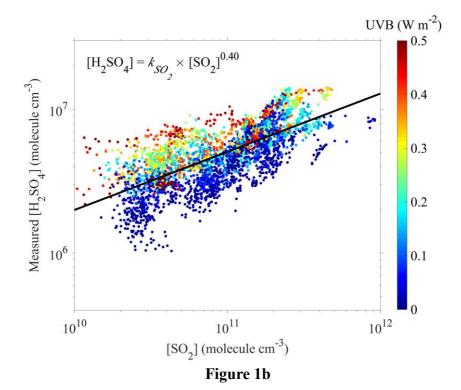
Figure 1. Correlations (a) between  $[H_2SO_4]$  and UVB intensity, and (b) between  $[H_2SO_4]$  and  $[SO_2]$  during the campaign from 9 February to 14 March, 2018..  $k_m$  is a constant term.

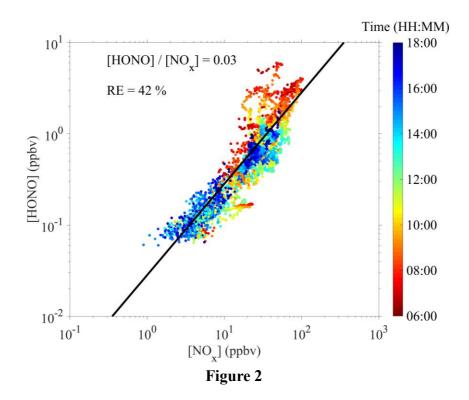
Figure 2. Correlation between [HONO] and  $[NO_x]$  during the campaign from 9 February to 14 March, 2018. The black line represents a linear fitting with a zero intercept.

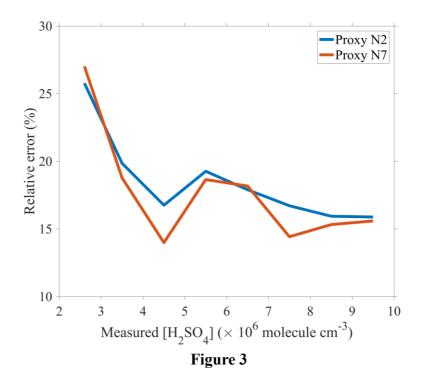
**Figure 3.** Performance assessments of proxy N2 and proxy N7. The REs are used to evaluate the performances of proxy N2 and N7, respectively as a function of linear bins of measured sulfuric acid concentrations.

**Figure 4.** Comparison of measured  $[H_2SO_4]$ ,  $[H_2SO_4]_{N2}$ ,  $[H_2SO_4]_{N7}$  and  $[H_2SO_4]_{Petäjä \ et \ al.}$  on 10 March, 2018 with a time resolution of 5 min.









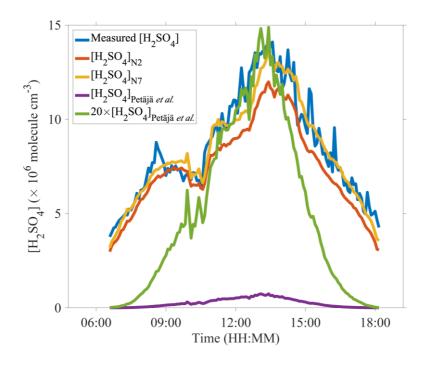


Figure 4